

- 1 **Long-Term Variability of PM<sub>2.5</sub> and PM<sub>10</sub> in Seoul:**
- 2 **Impacts of Asian Dust Events and Synoptic Circulation**

3 **Abstract**

4           This study utilized PM<sub>2.5</sub> data from two stations in Seoul, South Korea, providing a long record from  
5 2000 to the present, during winter (DJF) and spring (MAM) seasons. Both PM<sub>2.5</sub> and PM<sub>10</sub> exhibited a decreasing  
6 trend in both seasons, with a slower decline in the last decade (2011-2020) compared to the earlier decade (2000-  
7 2010). Concerning the correlation between PM<sub>2.5</sub> and PM<sub>10</sub>, the winter season exhibited a complex behavior with  
8 the last decade showing a stronger correspondence. By contrast, the spring season showed a robust interannual  
9 correlation throughout the whole period. It appears that Asian dust events reduced the correlation between PM<sub>2.5</sub>  
10 and PM<sub>10</sub> concentrations. Regarding daily variation, PM<sub>2.5</sub> and PM<sub>10</sub> generally fluctuated together during winter  
11 corresponding to migratory synoptic systems originating from northwest China. Such systems can remain  
12 stationary over Korea for multiple days, alternating PM behaviors. These features persisted into the spring season,  
13 with a spatial pattern similar to winter but with weaker intensity. The results offer insights into PM<sub>2.5</sub> variability  
14 across different time scales and its correlation with PM<sub>10</sub> over an extended period, with potential implications for  
15 PM<sub>2.5</sub> mitigating planning.

16 **Plain language summary**

17 This study looked at air pollution in Seoul, South Korea, focusing on smaller particles (PM<sub>2.5</sub>) and larger  
18 particles (PM<sub>10</sub>) during winter and spring from 2000 to now. We found that both PM<sub>2.5</sub> and PM<sub>10</sub> levels have been  
19 going down over time, but the decrease slowed in the last 10 years (2011-2020) compared to the previous 10 years  
20 (2000-2010). In winter, the connection between PM<sub>2.5</sub> and PM<sub>10</sub> was more complex, with a stronger link in the  
21 recent decade. In spring, this connection stayed strong throughout the whole period. We also found that natural  
22 dust storms from China and Mongolia (Asian dust events) weakened the link between PM<sub>2.5</sub> and PM<sub>10</sub> levels.  
23 Daily changes in PM<sub>2.5</sub> and PM<sub>10</sub> levels were usually similar during winter, influenced by synoptic scale weather  
24 systems from northwest China that can stay over Korea for several days and affect pollution levels. This pattern  
25 continued into spring but was less strong. These findings help us understand how PM<sub>2.5</sub> levels change over time  
26 and how they relate to PM<sub>10</sub> levels based on unique long-term records. This information is useful for planning  
27 ways to reduce PM<sub>2.5</sub> pollution and improve air quality in the future.

28

29 **Key Points**

30 1. Both PM<sub>2.5</sub> and PM<sub>10</sub> exhibit a decreasing trend from 2000 in Seoul, South Korea. The  
31 decline is more pronounced in the earlier decade compared to the last decade, indicating changes in  
32 air quality over time and likely reflecting enhanced air quality regulation.

33 2. In general, PM<sub>2.5</sub> and PM<sub>10</sub> display comparable interannual variability. Winter exhibits  
34 intricate behavior, with the correlation being significantly weaker in the earlier decade compared to  
35 the last decade. Nevertheless, a robust correlation between the two PM types can be observed in  
36 spring, regardless of the period.

37 3. The daily variations of PM<sub>2.5</sub> and PM<sub>10</sub> during winter are closely linked to a migratory  
38 synoptic anticyclone circulation system originating from northwest China. This system also  
39 influences PM concentrations in spring, albeit with reduced intensity.

40 4. The impact of Asian dust events on the correlation between PM<sub>2.5</sub> and PM<sub>10</sub> has been  
41 demonstrated to be detrimental, affecting both interannual variability and daily fluctuations.

## 42 **1. Introduction**

43 Particulate matter (PM) pollution significantly impacts human health and the environment. Concerns  
44 about air pollution, particularly the high concentrations of PM during winter and spring seasons, have  
45 intensified in South Korea due to rapid economic growth in recent decades. Public concern about PM has surged,  
46 driven by the daily discomfort it causes. Since regular measurements began in 1995, the concentration of PM<sub>10</sub>  
47 has gradually declined, thanks to government-led policies focused on reducing anthropogenic emissions.  
48 However, PM concentrations in South Korea are also influenced by Asian dust, a natural phenomenon (Chun  
49 et al., 2001). Asian dust transports coarse particles to South Korea, consequently increasing the proportion of  
50 PM<sub>10</sub> (Kim & Kim, 2003). Moreover, frequent Asian dust events significantly affect the annual average  
51 concentration of PM<sub>10</sub> (Ghim et al., 2015). To enhance prediction accuracy, linking PM levels to meteorological  
52 conditions is crucial. In response, the Ministry of Environment initiated nationwide forecasts for PM<sub>10</sub> and  
53 PM<sub>2.5</sub> in 2014 and 2015, respectively.

54 Meteorological conditions near South Korea manifest synoptic scale patterns, significantly  
55 influencing both pollutant transport and local air quality (Kim et al., 2016). These patterns affect PM<sub>10</sub>  
56 concentrations by either facilitating transport from external sources or causing stagnation from internal sources  
57 (Lee et al., 2022; Lee et al., 2011). The number of high PM<sub>10</sub> events in Seoul corresponds to the seasonality of  
58 Asian dust, and a characteristic synoptic scale condition favorable for high PM<sub>10</sub> events has been identified  
59 (Jung et al., 2019). Long-range transport of emissions from east-central China heavily influences South Korea's  
60 air quality, with domestic emissions also playing a role (Lee et al., 2019). Atmospheric blocking (Oh et al.,  
61 2015; Yun & Yoo, 2019) and large-scale circulation patterns resembling wave trains in the upper troposphere  
62 (Ku et al., 2021) also affect PM<sub>10</sub> concentrations. Based on these associations, there have been attempts to  
63 predict PM<sub>10</sub> levels using related atmospheric factors.

64 Previous studies have identified that temperature, wind, and pressure could be potential predictors of  
65 PM<sub>10</sub> concentrations (Callahan et al., 2019; Lee et al., 2022; Lee et al., 2018; Lee et al., 2011). They have  
66 examined PM<sub>10</sub> concentrations and meteorological conditions at different time scales. Lee et al. (2018)  
67 developed an index representing atmospheric conditions during high PM<sub>10</sub> cases by performing regression  
68 analysis of daily PM<sub>10</sub> concentrations and weather conditions. Jeong et al. (2022) also used statistical and  
69 dynamical models to develop a prediction model for the winter mean PM<sub>10</sub> concentration. The model  
70 demonstrated significant forecasting abilities, validating the feasibility of seasonal air quality forecasting based  
71 on climate conditions. Climate conditions are underscored by their physical linkage with PM<sub>10</sub> and their reliable

72 predictability through climate models.

73           This study is timely because we now have the extended PM<sub>2.5</sub> data from two long-term, reliable  
74 observation stations in Seoul. Although PM<sub>2.5</sub> has historically received less attention than PM<sub>10</sub> due to its shorter  
75 observation period, we can overcome this limitation by using the longer dataset available. We comprehensively  
76 analyze PM<sub>2.5</sub> concentration characteristics during winter and spring and their associated meteorological  
77 conditions on a multi-time scale. We also analyze the long-term trend of PM<sub>2.5</sub> and establish the relationship  
78 between PM<sub>2.5</sub> and PM<sub>10</sub> concentrations from daily to interannual variability. Additionally, we characterize the  
79 meteorological conditions associated with their daily relationship. Understanding the intricate dynamics  
80 dictating Seoul's air quality is crucial for future research directions and policy strategies aimed at maintaining  
81 and improving air quality standards.

## 82 2. Data and Methods

### 83 2.1. PM<sub>2.5</sub> and PM<sub>10</sub> concentration data sources and processing

84 We obtained daily mean PM<sub>10</sub> concentrations from 2000 to 2021 by averaging data from 25  
85 observation stations in Seoul, provided by the National Institute of Environmental Research (NIER) (available  
86 at <https://www.airkorea.or.kr/>). Due to the relatively short period of available PM<sub>2.5</sub> data from NIER, starting  
87 only after 2015, we extended the data period for PM<sub>2.5</sub> back to December 2000 using hourly PM<sub>2.5</sub> observations  
88 from the Seoul Research Institute of Public Health and Environment (SRIPHE) (available at  
89 <https://data.seoul.go.kr/dataList/OA-15526/S/1/datasetView.do#>), which offers the longest PM<sub>2.5</sub> observation  
90 record in South Korea (Chang et al., 2021; Jeong et al., 2023).

91 Among the 25 monitoring stations in Seoul, two stations (Seongdong-gu and Gwangjin-gu) have data  
92 from December 2000 (Supplementary Figure 1). However, the dataset lacked calibration and contained outliers  
93 due to various factors such as calibration errors, power disconnections, maintenance activities, and data  
94 abnormalities. Additionally, in December 2018, the Seongdong-gu Observatory was relocated from Seoul  
95 Forest to the Seongsu 1-ga 1-dong Residents' Center. To obtain reliable daily data, we implemented the  
96 following steps:

- 97 1. Outlier removal for hourly data: We removed outliers from the hourly data collected at each  
98 station to exclude abnormal values.
- 99 2. Daily mean calculation: After eliminating outliers, we computed the daily mean by  
100 aggregating the remaining hourly data points to smooth out irregularities.
- 101 3. Outlier removal for daily mean: We excluded any day with more than 25% outliers to  
102 maintain data integrity, following NIER's data screening approach.
- 103 4. Data combination from two stations: To maximize data coverage, we combined daily data  
104 from two stations. If one station lacked data for a specific day, we imputed it using the data from the  
105 other station, extending the dataset's temporal coverage.

106 These steps improved data quality and reliability by systematically addressing calibration issues and  
107 outliers. The PM<sub>2.5</sub> daily data resulting from our study showed a total of 8 missing values during the winter  
108 season (DJF) from 2000 to 2020 over 1895 days and 14 missing values during the spring season (MAM) from  
109 2001 to 2020 over 1840 days. Most missing data instances occurred in the earlier decade (2000-2010) for both  
110 seasons. These dates are listed in Supplementary Table 1.

111 This data is suitable for use in the study as a representation of PM<sub>2.5</sub> because it closely matches NIER

112 PM<sub>2.5</sub> data, as shown in Supplementary Figure 2. During the common winter (spring) period of 2015-2020, we  
113 compared the average NIER PM<sub>2.5</sub> data from the same two stations with the SRIPHE PM<sub>2.5</sub> data produced in  
114 this study and found a high correlation of 0.99 (0.99), indicating a close match. Although we used the same  
115 stations, discrepancies in certain values likely result from NIER's additional data screening processes. However,  
116 the data's reliability is supported by the high correlation for both seasons and the lack of any cases where PM<sub>2.5</sub>  
117 concentrations exceeded PM<sub>10</sub>. Furthermore, PM<sub>10</sub> concentrations, widely used as a proxy for PM<sub>2.5</sub> in air  
118 quality studies, also showed a high correlation of 0.84 (0.83) over a common period during winter (spring) for  
119 the same two stations (Supplementary Figure 3). Thus, we used the daily PM<sub>2.5</sub> data for the winter months of  
120 2000 to 2020 and the spring months of 2001 to 2020, generated as described above.

121

## 122 **2.2. Definition of Asian dust events**

123 Since Asian dust primarily affects larger particulate matter, we accounted for its more pronounced  
124 impact on PM<sub>10</sub> levels than PM<sub>2.5</sub> (Kim & Kim, 2003). We excluded Asian dust events from the analysis to  
125 discern their influence on the variability and concentration levels of both PM<sub>10</sub> and PM<sub>2.5</sub>. We identified days  
126 designated as Asian dust events based on records reported by the Korea Meteorological Administration  
127 (available at <https://www.weather.go.kr>). Throughout the winters (springs) of 2000-2020 (2001-2020), a total  
128 of 33 (138) Asian dust events were recorded.

129 Asian dust is a natural occurrence that results in high PM episodes, which also occur during non-  
130 Asian dust conditions. To compare the meteorological background fields, we conducted a composite analysis.  
131 During winter (spring), we defined 251 (216) high PM<sub>2.5</sub> cases as days with concentrations exceeding 1 standard  
132 deviation of the winter (spring) mean for 2000-2020, 57.98  $\mu\text{g}/\text{m}^3$  (2001-2020, 52.78  $\mu\text{g}/\text{m}^3$ ).

133

## 134 **2.3. Meteorological data**

135 To examine the relationship between fluctuations in PM levels and weather patterns, we utilized the  
136 6-hourly Japanese 55-year Reanalysis (JRA-55) as the meteorological dataset. JRA-55 has a horizontal  
137 resolution of 1.25° x 1.25° with 37 vertical levels (Kobayashi et al., 2015). Previous studies (Jung et al., 2019;  
138 Ku et al., 2021; Lee et al., 2022; Lee et al., 2018; Lee et al., 2011; Oh et al., 2015) have found the relationship  
139 between PM<sub>10</sub> levels and meteorological conditions in Korea, specifically pressure, temperature, and wind. We  
140 used the winter and spring daily mean geopotential height (Z), air temperature (T), and horizontal winds (U and  
141 V) from 2000 to 2020, considering the local time of Seoul. We calculated daily anomalies by subtracting the

142 corresponding daily climatologies.

143

#### 144 **2.4. Methods for analyzing trends, variability, and correlations**

145 To understand the relationship between the interannual variability of  $PM_{2.5}$  and  $PM_{10}$ , it is necessary  
146 to remove trends. The winter mean of  $PM_{2.5}$  and  $PM_{10}$  between 2000 and 2020 displayed marked trends  
147 attributed to government policies aimed at emission reduction (Heo et al., 2017). However, the linear trend  
148 derived through the least-squares method was insufficient to capture the nuanced trends observed in winter  
149 mean PM levels across the study period (Supplementary Table 2). Distinct differences in trends were apparent  
150 between the initial and final decades. Therefore, we omitted the second-order polynomial regression curve for  
151 the winter mean trends to depict the trend spanning the entire timeframe accurately. On the other hand, we  
152 successfully detrended the trends in spring mean PM using the least-squares method.

153 We evaluated the interannual and daily relationship between  $PM_{2.5}$  and  $PM_{10}$  using Pearson  
154 correlation coefficients with a two-tailed Student's t-test to assess statistical significance. To examine the  
155 association between weather conditions and  $PM_{2.5}$  concentrations in space and time, we performed lead-lag  
156 correlation analysis with weather conditions using a time lag of 7 days before and after.

### 157 3. Results and Discussions

#### 158 3.1. Decreasing long-term trends of PM<sub>2.5</sub> and PM<sub>10</sub> with recent decadal weakening

159 The long-term variation in winter mean PM<sub>2.5</sub> and PM<sub>10</sub> shows a notable downward trend from 2000  
160 to 2020 (Figure 1(a) and Table 1). Both PM<sub>2.5</sub> and PM<sub>10</sub> exhibit significant decreasing trends, reflecting the  
161 efficacy of Korean government policies in reducing PM levels. The spring season shows similar characteristics.  
162 Both PM<sub>2.5</sub> and PM<sub>10</sub> display steadily decreasing long-term trends in spring as well (Figure 1(b) and Table 2).  
163 However, among the similar trends, there is a marked disparity between the trends in the earlier decade (2000-  
164 2010, P1) and the recent decade (2011-2020, P2). During P2, the trends for both PM<sub>2.5</sub> and PM<sub>10</sub> weaken  
165 compared to P1. Specifically, PM<sub>2.5</sub> shows an almost flattened trend (-2.476 µg/m<sup>3</sup>/decade) compared to PM<sub>10</sub> (-  
166 10.33 µg/m<sup>3</sup>/decade). In spring, the trends for P1 are steeper compared to P2 and both PM<sub>2.5</sub> and PM<sub>10</sub> exhibit  
167 similar decreasing trends (Table 1 and Table 2). Yeo and Kim (2019) observed fluctuations of the decreasing  
168 trend in the annual mean concentrations from 2001, and this pattern is even more pronounced in the winter and  
169 spring means observed in this study. Excluding Asian dust events does not result in a significant difference in  
170 trend compared to the original trends. This demonstrates that Asian dust does not impact the long-term decrease  
171 driven by pollution-reduction policies.

172 The observed long-term decreasing trends in PM<sub>2.5</sub> and PM<sub>10</sub> from 2000 to 2020 indicate the success  
173 of South Korea's air quality management policies. These policies appear particularly effective in the earlier  
174 decade, as evidenced by the steeper decline in PM concentrations during this period compared to the recent  
175 decade. This suggests that initial policy measures had a strong impact, but the effectiveness may have diminished  
176 over time, possibly due to changes in emission sources or saturation of policy impact.

177 The more pronounced weakening in the PM<sub>2.5</sub> trend compared to PM<sub>10</sub> in the recent decade raises  
178 concerns. PM<sub>2.5</sub>, being finer and more harmful to human health, requires continued and possibly more strict  
179 control measures. Furthermore, the seasonal analysis highlights that the policy impacts vary by season, with  
180 winter and spring showing different degrees of trend reductions. This seasonal variability suggests that policies  
181 may need to be tailored to address specific seasonal emission sources and meteorological conditions.

182 **Table 1. DJF mean trends of PM<sub>10</sub> and PM<sub>2.5</sub> concentrations (brackets: excluding Asian dust events).**

DJF mean trend ( $\mu\text{g}/\text{m}^3/\text{decade}$ )	2000-2020	P1 (2000-2010)	P2 (2011-2020)
PM <sub>10</sub>	-14.21* (-14.28*)	-13.27 (-16.21*)	-10.33 (-9.014)
PM <sub>2.5</sub>	-14.07* (-14.01*)	-30.90* (-31.16*)	-2.476 (-2.719)

183 \*: significant at a 95% significance level.

184

185 **Table 2. MAM mean trends of PM<sub>10</sub> and PM<sub>2.5</sub> concentrations (brackets: excluding Asian dust**  
 186 **events).**

MAM mean trend ( $\mu\text{g}/\text{m}^3/\text{decade}$ )	2001-2020	P1 (2001-2010)	P2 (2011-2020)
PM <sub>10</sub>	-24.32* (-15.88*)	-46.93* (-28.49*)	-12.0 (-6.513)
PM <sub>2.5</sub>	-15.38* (-12.42*)	-42.26* (-33.22*)	3.352 (4.096)

187 \*: significant at a 95% significance level.

188

### 189 3.2. Seasonal differences in the interannual relationship between PM<sub>2.5</sub> and PM<sub>10</sub>

190 Table 3 shows that Asian dust disrupts the interannual relationship between PM<sub>2.5</sub> and PM<sub>10</sub>. Without  
 191 Asian dust events, PM<sub>2.5</sub> and PM<sub>10</sub> show similar long-term interannual variabilities, with a correlation coefficient  
 192 of 0.50 over the 2000-2020 period during winter. However, disparities emerge between the earlier decade (2000-  
 193 2010, P1) and the recent decade (2011-2020, P2): P1 has a lower coefficient of 0.45 compared to a stronger 0.66  
 194 correlation in P2. Here, the separation of these periods is solely determined by the mid-point. The statistically  
 195 significant correlations in P2 emphasize the yet similar complex nature of interannual variability between PM<sub>2.5</sub>  
 196 and PM<sub>10</sub>. In spring, PM<sub>2.5</sub> and PM<sub>10</sub> exhibit highly similar interannual variability (Table 4). During P1, spring  
 197 maintains a high correlation throughout both periods, whereas the winter correlation is significantly lower. By  
 198 removing Asian dust events, the correlation between PM<sub>2.5</sub> and PM<sub>10</sub> concentrations increases, particularly in  
 199 winter. By contrast, the spring season maintains a strong correlation throughout the study period, with or without  
 200 Asian dust events. These findings highlight the modulation of Asian dust events on the year-to-year variability of  
 201 PM<sub>2.5</sub> and PM<sub>10</sub> in winter when cold-air outbreaks are frequent and more widespread.

202 The interannual variability analysis reveals that Asian dust events significantly disrupt the correlation  
 203 between PM<sub>2.5</sub> and PM<sub>10</sub>, particularly in winter. The higher correlation in the recent decade (2011-2020) suggests  
 204 that, despite overall decreasing trends, episodic events like Asian dust storms still play a critical role in PM  
 205 variability. By comparison, the weaker interannual correlation in 2000-2010 (P1) may be influenced by heightened  
 206 variability in PM levels due to the pronounced effects of emission reduction policies. The GEOS-Chem (Goddard  
 207 Earth Observing System-Chemistry) experiment with fixed emissions shows no significant trends. However, it  
 208 reveals consistent correlations in both P1 and P2. Specifically, monthly PM<sub>2.5</sub> and PM<sub>10</sub> data from this model  
 209 (Jeong et al., 2018) indicate nearly flat trends over the same period. They also show a high interannual correlation  
 210 of 0.91 (Supplementary Figure 4).

211 Furthermore, the declining PM<sub>2.5</sub>/PM<sub>10</sub> fraction in P1 supports this result (Supplementary Figure 5). A  
 212 stable fraction indicates a strong correlation, while P2, with a higher interannual correlation, exhibits a less  
 213 pronounced fraction variation than P1. The reduction in policy influence suggests that the proportion of the impact  
 214 of other factors may increase in the relationship between the two PM types. These findings indicate that underlying  
 215 meteorological conditions continue to drive PM levels even as emissions decrease. This points to the importance  
 216 of integrating meteorological considerations into air quality management strategies.

217

218 **Table 3. Correlation coefficients between winter mean PM<sub>2.5</sub> and PM<sub>10</sub> concentrations.**

DJF mean correlation	With Asian dust	Without Asian dust
<b>2000-2020</b>	0.34	0.50*
<b>2000-2010 (P1)</b>	0.38	0.45
<b>2011-2020 (P2)</b>	0.32	0.66*

219 \*: significant at a 95% significance level.

220

221 **Table 4. Correlation coefficients between spring mean PM<sub>2.5</sub> and PM<sub>10</sub> concentrations.**

MAM mean correlation	With Asian dust	Without Asian dust
<b>2000-2020</b>	0.77*	0.80*
<b>2001-2010 (P1)</b>	0.77*	0.75*
<b>2011-2020 (P2)</b>	0.79*	0.87*

222 \*: significant at a 95% significance level.

223

### 224 3.3. Daily variation of PM<sub>2.5</sub> and PM<sub>10</sub> under the synoptic-scale circulation system

225 PM<sub>2.5</sub> and PM<sub>10</sub> show a robust correlation in their daily variations, with an expected influence from  
226 Asian dust events. During winter from 2000 to 2020, PM<sub>2.5</sub> and PM<sub>10</sub> exhibit similar daily variabilities, with a  
227 correlation coefficient of 0.80 when including Asian dust events, and 0.91 without. Figure 2(a) confirms this  
228 influence, showing weaker correlations during winters with Asian dust events included and stronger correlations  
229 when dust days are excluded. In spring, PM<sub>2.5</sub> and PM<sub>10</sub> display highly similar daily variations (Figure 2(b)),  
230 especially when excluding Asian dust events, resulting in a correlation coefficient of 0.89. Including these events  
231 reduces the correlation coefficient to 0.85. This further confirms that Asian dust events disrupt the relationship  
232 between PM<sub>2.5</sub> and PM<sub>10</sub> in both daily and interannual variability.

233 The impact of emissions-driven long-term trend differences in winter mean PM levels, which previously  
234 caused interannual correlation disparities, appears minimal on a daily scale. This is evident in the significant and  
235 relatively consistent daily correlation observed in P1 and P2. Comparing these periods reveals consistent  
236 correlations, albeit higher when excluding Asian dust events (winter P1: 0.84 and P2: 0.71; spring P1: 0.87 and  
237 P2: 0.72 when including Asian dust events; winter P1: 0.90 and P2: 0.88; spring P1: 0.89 and P2: 0.87 when  
238 excluding Asian dust events). Hence, the daily variability appears less affected by fluctuations in emission trends.

239 We analyzed the covarying patterns between PM<sub>2.5</sub> and meteorological conditions to understand the role  
240 of synoptic circulation patterns in winter and spring. Correlation maps between PM<sub>2.5</sub> and Z500, T850, U500, and  
241 V850 illustrate the synoptic-scale circulation system over Korea (Figure 3(A)(a), (c), (e), and (g)). This system is  
242 characterized by several features. First, mid-tropospheric positive pressure anomalies (Z500) are centered over  
243 Korea. Second, low-tropospheric northerly winds (V850) are weakened. Third, positive temperature anomalies  
244 (T850) occur due to reduced cold air inflow. Mid-tropospheric zonal wind patterns (U500) exhibit north-south  
245 symmetry, further reducing northerlies and impeding the removal of PM<sub>2.5</sub>. Considering the reduced role of  
246 government policies in the recent decade as discussed in section 3.2, the influence of synoptic patterns may have  
247 become more pronounced compared to the previous decade. Their relationship remains significant throughout the  
248 period, although some years exhibit lower correlations (Figure 3(A)(b), (d), (f), and (h)). The spring patterns are  
249 similar to winter patterns but differ in intensity and location, and the relationship shows more year-to-year  
250 variability than the winter (Figure 3(B)).

251 The synoptic patterns align with the study on PM<sub>2.5</sub> in Beijing, approximately 1000 km from Seoul (Cai  
252 et al., 2017). Furthermore, the similarity to synoptic patterns associated with pollutant transport underscores the  
253 relationship with the meteorological conditions (Kim et al., 2016; Lee et al., 2011). This indicates that similar

254 synoptic patterns affect PM levels across East Asia. Therefore, understanding these patterns is crucial for  
255 predicting high PM events and developing responsive air quality management measures.

256 The temporal evolution of the synoptic circulation system is shown in Figure 4 through time lead-lag  
257 correlation maps. During winter, a migratory anticyclone originating from northwest China shows significant  
258 positive correlation patterns with PM<sub>2.5</sub>, beginning two days prior and persisting over the Korean peninsula for  
259 two additional days due to blocking by low pressure over the Northern Pacific. This system brings warm  
260 temperature anomalies and weakened northerlies in the low troposphere over Korea, leading to high PM<sub>2.5</sub> levels.  
261 The migratory system might be associated with the propagating wave train pattern observed in high PM<sub>10</sub> cases  
262 (Ku et al., 2021), and the stagnation could be linked to atmospheric blocking, which impacts PM<sub>10</sub> concentrations  
263 (Yun & Yoo, 2019). The significant relationship with PM<sub>2.5</sub> concentrations persists for approximately five days  
264 (Figure 4(A)(i)). While spring spatial patterns resemble those of winter, their intensity is weaker. Consistent  
265 durations of covariability are observed in pressure and other meteorological factors such as U500, T850, and V850  
266 (not shown). The result suggests that forecasting these patterns could provide a few days' lead time for  
267 implementing mitigative actions. This highlights the potential for using synoptic-scale forecasts in operational air  
268 quality management.

269

### 270 **3.4. Contrasting meteorological conditions during Asian dust events**

271 We conducted a composite analysis to identify the background conditions when Asian dust events occur.  
272 In Figure 5, high PM<sub>2.5</sub> cases without Asian dust events are associated with mid-tropospheric positive pressure  
273 anomalies centered over Korea, weakened low-tropospheric northerly winds, and positive temperature anomalies  
274 for both seasons. The patterns are similar to the correlation maps in Figure 3 showing the stagnant high-pressure  
275 system. This emphasizes the critical role of synoptic circulation systems in the variation of PM<sub>2.5</sub> concentrations.

276 In contrast, the composite maps of Asian dust events reveal contrasting features. Negative Z500 and  
277 intensified V850 are shown, and these conditions are known to facilitate the transportation of Asian dust from its  
278 source region in arid northern China (Jung et al., 2019). The low T850 extends from continental Asia to the Korean  
279 Peninsula, diverging from the positive T850 patterns observed in high PM<sub>2.5</sub> cases. These temperature anomaly  
280 patterns resemble cold surge patterns over Korea, aligning with negative pressure patterns associated with the cold  
281 surge phenomenon. The negative Z500 pattern extending toward the Korean Peninsula resembles the migratory  
282 anticyclone patterns from the Siberian High to the Korean Peninsula (Jeong et al., 2006; Ryoo et al., 2005). Thus,  
283 the contrasting conditions highlight how Asian dust events disrupt the relationship between PM<sub>2.5</sub> and PM<sub>10</sub>

284 variability. They also emphasize the complexity of managing PM levels, requiring different strategies for different  
285 types of high PM events. For example, reducing local emissions might be more effective for non-dust high PM  
286 events, while regional cooperation is essential for addressing transboundary dust transport.

287 **4. Conclusion**

288 This study reports an extended analysis of the long-term trends and interannual variability of PM<sub>2.5</sub> and  
289 PM<sub>10</sub> in Seoul, considering the impact of Asian dust events and synoptic circulation patterns. Both PM<sub>2.5</sub> and PM<sub>10</sub>  
290 exhibit significant declining trends from 2000 to 2020, reflecting the effectiveness of emission reduction policies.  
291 Excluding Asian dust events shows that while they significantly affect PM<sub>10</sub> levels, they do not alter the overall  
292 declining trend of PM<sub>2.5</sub> and PM<sub>10</sub>. However, the correlation between PM<sub>2.5</sub> and PM<sub>10</sub> is influenced by Asian dust  
293 events, which disrupt their relationships on both daily and interannual scales.

294 The study also reveals that the interannual correlation during winter between PM<sub>2.5</sub> and PM<sub>10</sub> is stronger  
295 in the recent decade (2011-2020) compared to the earlier decade (2000-2010), likely due to the variability in PM  
296 levels caused by emission reduction policies. During spring, the relationship remains very stable throughout the  
297 study period. On a daily scale, PM<sub>2.5</sub> and PM<sub>10</sub> show robust correlations, but these are weakened by the presence  
298 of Asian dust events. The findings emphasize the role of synoptic-scale circulation patterns in the daily and  
299 seasonal variability of PM concentrations.

300 The results highlight the importance of considering Asian dust events in air quality management  
301 strategies. Although policies aimed at reducing emissions have effectively decreased PM levels over the past two  
302 decades, the presence of Asian dust can obscure these improvements. Transport of dust also implies PM from  
303 remote sources. Understanding the synoptic meteorological conditions that influence PM variability can enhance  
304 the accuracy of PM forecasts and aid in developing more effective mitigation strategies. This study underscores  
305 the need for integrated air quality management that addresses both local emissions and transboundary pollution  
306 events to protect public health.

307 Future research should focus on the detailed mechanisms through which meteorological factors  
308 influence PM variability. There is a need to explore the seasonal differences in the impact of synoptic circulation  
309 patterns on PM levels. Additionally, further studies should investigate the potential impacts of future policy  
310 interventions on PM trends, considering both local and regional sources of pollution. Enhanced monitoring and  
311 modeling efforts are necessary to improve the accuracy of PM forecasts and to develop effective mitigation  
312 strategies.

313

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