

Vertical distributions of blooming cyanobacteria populations in a freshwater lake from LIDAR observations

Timothy S. Moore^a, James H. Churnside^b, James M. Sullivan^c, Michael S. Twardowski^c, Aditya R. Nayak^{c,d}, Malcolm N. McFarland^c, Nicole D. Stockley^c, Richard W. Gould^g, Thomas H. Johengen^e, Steven A. Ruberg^f

^a*University of New Hampshire, Durham, NH*

^b*NOAA Earth System Research Laboratory, Boulder, CO*

^c*Harbor Branch Oceanographic Institute, Florida Atlantic University, Fort Pierce, FL*

^d*Department of Ocean and Mechanical Engineering, Florida Atlantic University, Boca Raton, FL*

^e*Cooperative Institute for Limnology and Ecosystems Research, University of Michigan, Ann Arbor, MI USA*

^f*NOAA Great Lakes Environmental Research Laboratory, Ann Arbor, MI*

^g*Bio-Optical/Physical Processes and Remote Sensing Section, Naval Research Laboratory, Code 7331, Stennis Space Center, MS*

Abstract

The vertical distributions of freshwater cyanobacteria populations are important to plankton community structure, ecology and for influencing water column optical properties relevant to remote sensing. In August of 2014, we examined the vertical structure of a cyanobacteria bloom across the western basin of Lake Erie with new technologies, including LIDAR and a digital holographic system. In addition, vertical profiles of environmental and optical properties were made. The active LIDAR penetrated the water column, and provided a detailed picture of the particle distribution for the whole water column. The holographic system provided digital images processed for particle

size, count and identification of *Microcystis* and *Planktothrix* - the two main cyanobacteria genera that were present. The correlations between the LIDAR backscatter intensity and the cyanobacteria cell counts from holography averaged to 0.53 and ranged from -0.13 to 0.96 based on nearest matchups. The vertical structure of the overall cyanobacteria population was influenced by wind speed, and to a lesser degree the solar heating of surface waters. On a more detailed level, *Microcystis* populations were consistently nearer to the surface relative to *Planktothrix*. Pigments from surface samples revealed a higher degree of photoprotection for *Planktothrix*-dominated communities. The vertical distributions of the cyanobacteria genera were related to light intensity in the water column and known tolerances and/or preferences for each genus. Vertical profiles of optical properties supported the patterns seen in the LIDAR and holographic data, and had direct implications on the exiting light field. These combined data provide a unique view into the natural variations in spatial (vertical and horizontal) distribution patterns of cyanobacteria and resulting impacts on remote sensing detection and associated interpretations, and demonstrate the potential for these technologies to observe cyanobacteria in lake environments.

Keywords: remote sensing, harmful algal blooms, LIDAR, cyanobacteria, species distributions, holography

1. Introduction

Cyanobacteria inhabit lake systems worldwide (Harke et al., 2016), and have a long evolutionary history resulting in a diversity of ecological strategies and traits from long-term adaptations (Uyeda et al. 2016; Blank 2013; Sánchez-Baracaldo et al. 2005). One trait shared across a large number of cyanobacteria genera is the ability to regulate buoyancy with the aid of gas vesicles (Carey et al. 2012; Walsby 1994). The role of buoyancy in cyanobacteria ecology has long been a research topic and viewed as an important factor for ecological success (Harke et al. 2016; Walsby et al. 1997; Ibelings et al. 1994; Humphries and Lyne 1988). This capability gives an organism an advantage for accessing light and nutrient resources for optimizing photosynthesis and growth.

Buoyancy regulation in gas-vacuolate cyanobacteria is a function of overall cell density, largely derived from the balance between intracellular gas vesicle volume and the amount of other cellular components, especially carbohydrates (Visser et al. 1997; Oliver 1994). Gas vesicles can be synthesized or collapsed by cells to increase or decrease buoyancy, respectively (Oliver 1994). Cell ballast change through carbohydrate content is another mechanism to regulate buoyancy. Carbohydrate content can fluctuate rapidly in cells, usually on shorter time scales than gas vesicle changes (Oliver, 1994). By adjusting these compartments, cells and larger colonial aggregates can alter buoyancy to migrate upwards or downwards in the water column with rates of several meters per hour (Visser et al., 1997).

24 There are multiple environmental factors that influence buoyancy regula-
25 tion in cyanobacteria, led by light level exposure and external nutrient levels
26 (Walsby et al. 2004; Oliver 1994; Konopka 1989). Buoyancy responses to
27 changing light conditions have been reported in laboratory experiments and
28 in natural settings. In laboratory experiments of *Planktothrix* populations, a
29 genus with filamentous morphologies, dark-adapted cells lost buoyancy when
30 exposed to increased light levels, and increased buoyancy in lower light inten-
31 sities (e.g., Visser et al. 1997; Oliver and Walsby 1984; Walsby and Booker
32 1980). Across different natural settings, *Planktothrix* populations have been
33 observed to rise and/or sink in response to light exposure (Walsby et al. 2004;
34 Davis et al. 2003; Kromkamp and Walsby 1990). In these cases, the cyanobac-
35 teria populations maintained neutral buoyancy during stratified periods; that
36 is, vertical position was maintained at depth without escaping to the surface
37 or to the bottom outside the preferred zone. Based on these behavioral pat-
38 terns to light, *Planktothrix* are considered shade-adapted (Walsby et al. 2004;
39 Davis et al. 2003; Halvstedt et al. 2007).

40 The degree to which cells respond to light in terms of buoyancy regulation
41 varies between genera, species and even within species depending on physio-
42 logical state (Oliver, 1994). As a result, differing light tolerances/preferences
43 can create niche separation along vertical gradients allowing for co-existence.
44 This has been observed with different *Planktothrix* species (Kokocinski et al.
45 2010; Davis et al. 2003), and marine cyanobacteria (Stomp et al., 2007).
46 *Microcystis* is another commonly found cyanobacteria genus and has been

47 studied for decades, with its ecological success linked to buoyancy regulation
48 (Paerl and Otten 2013; Davis et al. 2003; Paerl et al. 1985). *Microcys-*
49 *tis* is well known for colony formation from aggregating cells and forming
50 surface blooms during calm wind periods. Colonial aggregates are believed
51 to enhance vertical migration, and can promote excessive or 'overbuoyancy'
52 (Oliver 1994; Paerl et al. 1983) leading to surface scum formation. Studies
53 by Paerl et al. (1983) and Paerl et al. (1985) elaborated on the ability of
54 *Microcystis* surface blooms to withstand the high light environment. The
55 studies determined that not only did photoprotective pigments (e.g., zeax-
56 anthin) shield cells from otherwise dangerous excessive light, the photosyn-
57 thetic efficiencies often increased in surface blooms of *Microcystis*. Given
58 their widespread distribution around the globe and long geologic history, it
59 is an aspect of their evolutionary success that enables *Microcystis* to not only
60 withstand but in certain cases thrive at high light, giving them a competitive
61 advantage under those conditions and potentially shading out competitors
62 (Paerl et al., 1985).

63 The vertical distributions of cyanobacteria have implications for remote
64 sensing applications. Gas-vacuolate cyanobacteria have high scatter and
65 backscatter efficiency (Matthews and Bernard 2013; Moore et al. 2017) which
66 elevates light backscattered out of the water the nearer they are to the surface.
67 Kutser (2004) describes three states of cyanobacteria vertical distributions
68 in the context of remote sensing. The states include vertically uniform dis-
69 tributions, near-surface distributions, and floating scums - each of which has

70 a different impact on the spectral remote sensing reflectance ($R_{rs}(\lambda)$), the
 71 quantity that is detected by radiometers including those on satellites. The
 72 $R_{rs}(\lambda)$ is proportional to the light scattered back out of the water from a
 73 layer extending from the surface to a variable depth, which is a function of
 74 the attenuation coefficient of light and wavelength (Kirk, 1994). This depth
 75 can be 1 meter or less in dense, near-surface cyanobacteria blooms (Kutser,
 76 2004). In these cases, $R_{rs}(\lambda)$ is enhanced to the extreme point of resembling
 77 land foliage. Conversely, if a cyanobacteria population is concentrated in a
 78 deeper sub-surface layer, they may be below the detection depth of satellite
 79 sensors, and may contribute weakly or not at all to $R_{rs}(\lambda)$. This type of
 80 distribution does not fit into one of the three states, although it has been
 81 observed in nature (e.g., Davis et al. 2003; Walsby et al. 2004). Knowing
 82 the state or vertical distribution of cyanobacteria populations in natural en-
 83 vironments improves the understanding of cyanobacteria ecology and the
 84 interpretations of the remote sensing observations. These two aspects are
 85 important to assessing and determining lake water quality attributes.

86 Vertical profile measurements of optical properties in freshwater lakes for
 87 remote sensing studies are scarce (see Xue et al. 2017 and references therein).
 88 Most remote sensing studies of cyanobacteria blooms focus on surface con-
 89 ditions. However, recent modeling studies have investigated the impacts
 90 of gas-vacuolate cyanobacteria on $R_{rs}(\lambda)$ (Xue et al. 2017; Matthews and
 91 Bernard 2013; Kutser et al. 2008; Metsamaa et al. 2006). These studies re-
 92 vealed the dependencies of the $R_{rs}(\lambda)$ on the vertical structure of biomass

93 and their associated inherent optical properties (IOPs) such as the absorp-
94 tion and backscatter coefficients. A general conclusion from these studies
95 is that non-uniform vertical structure is a complication for remote sensing
96 algorithms used for quantifying cyanobacteria biomass.

97 We previously reported on the horizontal distributions of surface optical
98 properties in western Lake Erie during a cyanobacteria bloom (Moore et al.,
99 2017), and identified two genera of cyanobacteria dominating the microbial
100 community - *Microcystis* and *Planktothrix*. During our field sampling in Au-
101 gust 2014, another experiment led by the Naval Research Laboratory (NRL)
102 was sampling the same area with similar optical packages, and measuring
103 for a suite of surface pigments. In addition, a Light Detection and Ranging
104 (LIDAR) instrument, developed in-house at the National Oceanic and Atmo-
105 spheric Administration (NOAA), was flown on-board a Twin Otter aircraft
106 taking measurements over the western and central basins of Lake Erie. The
107 LIDAR instrument is able to detect vertical profiles of optical backscattering
108 from particles (e.g., cyanobacteria cells and colonies, suspended sediments)
109 in surface waters, and is the only remote sensing technique that can profile
110 the upper water column from above the surface (Churnside, 2014). Compar-
111 isons of LIDAR returns with in water measurements of optical backscattering
112 have shown good agreement (Lee et al. 2013; Churnside et al. 2017). The
113 LIDAR can measure particle distributions with a vertical resolution of less
114 than 1 *m* and a horizontal resolution of 5 to 15 *m* (Churnside and Donaghay
115 2009; Churnside 2015). In clear oceanic water, these profiles can reach 50 *m*

116 in depth. During the study period, the penetration was less in Lake Erie be-
117 cause of higher density of particles, but the full water column was measured
118 in many places of the western basin.

119 *1.1. Objectives*

120 The field and aircraft measurements collected during August of 2014 from
121 this location provide a unique data set to assess the vertical distributions of
122 cyanobacteria populations in natural settings and their impacts on remote
123 sensing. The goals of this study are to 1) examine the associations between
124 the vertical distributions of particle fields from LIDAR data and vertical
125 distributions of optical properties and cyanobacteria counts measured in Lake
126 Erie; and 2) to assess these findings in the context of cyanobacteria ecology
127 and impacts on bio-optical algorithms used in remote sensing applications.

128 **2. Methodology**

129 *2.1. Study Area*

130 Lake Erie comprises three connected basins, distinguished by bathymetry
131 and other natural features - the western, central and eastern basins. Obser-
132 vations from aircraft and direct sampling from this study were made in the
133 western basin, and the western edge of the central basin. Mean flow is from
134 west to east, and thus water flows from the western basin to the central
135 basin. The mean residence time of water in the western basin is 50 days
136 (Millie et al., 2009). The two largest and most important rivers flowing into

137 the western basin are the Maumee River, entering the basin in the southwest
138 corner (Maumee Bay) through the city of Toledo, and the Detroit River en-
139 tering the basin in the northwest corner. The Maumee River watershed is
140 dominated by agricultural land (Joosse and Baker, 2011), and supplies much
141 of the nutrient load to the western basin (Stumpf et al. 2012; Michalak et al.
142 2013; IJC 2014), despite the Detroit River delivering over 80% of the annual
143 basin-wide discharge by volume. The Detroit River waters are poorer in nu-
144 trient content, and are optically different from waters to the south in Maumee
145 Bay (Moore et al., 2017). The western basin transitions to the deeper cen-
146 tral basin in an area populated by islands known as the Lake Erie Islands,
147 with the two largest comprising Pelee Island (to the north) and Kellys Island
148 (to the south). West of these islands are three smaller islands called North,
149 Middle and South Bass Islands. Other smaller islands are also within this
150 island vicinity. Sandusky Bay is a shallow, enclosed water body receiving
151 water from the Sandusky River, and discharges into the southwestern central
152 basin and also plays a role as a source of cyanobacteria to Lake Erie (Davis
153 et al. 2015; Kane et al. 2014).

154 To aid in understanding patterns in the data, we have further identified
155 nine sub-regions within the overall study area corresponding to geographic
156 and hydrographic features (Figure 1). Not all of these regions were directly
157 sampled in the field, but all were observed with the LIDAR. These areas are:
158 1 - southeastern western basin; 2 - Detroit River plume front (a transition re-
159 gion between the Detroit River and Maumee Bay water); 3 - Detroit River (a

quasi-permanent hydrographic feature in the northwest corner); 4 - Maumee Bay (the shallowest part of the basin that directly receives Maumee River discharge); 5 - Islands West (an area to the west of the islands in the western basin); 6 - Islands Central (a transition region encompassing the islands); 7 - Islands East (an area to the east of the islands extending into the deeper central basin); 8 - Islands Southeast (an area in the central basin outside the entrance of Sandusky Bay); 9 - Sandusky Bay. Maumee and Sandusky Bays were observed with LIDAR only. Echograms from these two areas (4 and 9) are contained in a supplement at the end of this manuscript.

2.2. Data Sets

Field and aircraft data were collected between August 17, 2014 and August 28, 2014 (Table 1). The field data were derived from two separately conducted but simultaneous field surveys (Table 2). The first data set comprised 20 stations led by the University of New Hampshire (UNH), and included surface water discrete samples, vertical profiles of inherent optical properties (IOPs), above-water $R_{rs}(\lambda)$, and digital holographic profiles. The second data set was generated from a simultaneous field survey led by the Naval Research Laboratory (NRL) and comprised vertical IOP profiles (N=11), $R_{rs}(\lambda)$ and surface water discrete samples for pigments. Not all stations had the same suite of measurements. These are indicated on the map in Figure 1.

181 2.2.1. *Discrete measurements*

182 Discrete measurements of surface water quality parameters for the UNH
183 data set included suspended particulate matter (SPM), chlorophyll-a con-
184 centration (*Chl-a*), phycocyanin concentration (PC). The details of the pro-
185 cessing are contained in Moore et al. (2017). We further determined volatile
186 (organic) and non-volatile SPM by combusting filters for 4 hours at 450°C,
187 cooling, and reweighing (APHA 1998).

188 The NRL discrete data comprised surface water samples processed for
189 high pressure liquid chromatography (HPLC). Some of these did not coin-
190 cide with any IOP profiles. Thus, in addition to the 11 NRL stations with
191 profiles, another 9 stations contained HPLC samples only (see Figure 1).
192 The HPLC data were processed with an Agilent RR1200 system, and extrac-
193 tion and pigment concentration followed the protocol detailed in (Heukelem
194 and Thomas, 2001). A set of pigments were quantified, and included but
195 not limited to total chlorophyll-a and zeaxanthin. In our analysis, we only
196 used these pigments, with zeaxanthin being a major photoprotective pig-
197 ment found in cyanobacteria (Jeffrey et al., 1997), including *Microcystis* and
198 *Planktothrix*. We used these two pigments to quantify the degree of internal
199 photoprotection in the surface algal populations across the study region.

200 2.2.2. *IOP Vertical profiles*

201 Vertical profiles of optical and hydrographic properties from the UNH
202 data set were collected at 14 stations. The vertical profiling system included

203 a WET Labs (Philomath, OR) ac-9 measuring absorption and attenuation at
 204 9 wavelengths: 412, 440, 488, 510, 532, 555, 650, 676 and 715 *nm*, backscat-
 205 tering meters (WET Labs ECO-VSF, ECO-BB3, and ECO-BB9 sensors)
 206 and a SeaBird (Bellevue, WA) SBE49 CTD. The ac-9 was calibrated with
 207 Milli-Q ultrapure water, and absorption $a(\lambda)$ was corrected for scattering
 208 effects using the proportional method of Zaneveld et al. (1994). Data were
 209 corrected for temperature and salinity effects using the coefficients of Twar-
 210 dowski et al. (1999) using the CTD data. Two profiles were made at each
 211 station. The first profile was taken without any filters to derive total ab-
 212 sorption (a_t). A 0.2 μM filter was fitted on the ac-9 for the second profile
 213 to derive dissolved absorption (a_g). From the two profiles, particulate ab-
 214 sorption (a_p) was derived by subtracting water (a_w) and dissolved absorption
 215 from total absorption. Of these, we report only on the particulate absorption
 216 in the Results. Particulate scattering $b_p(\lambda)$ was derived from the ac-9 as the
 217 difference between attenuation $c(\lambda)$ and absorption $a(\lambda)$. We further derived
 218 the particle backscatter ratio - $\widetilde{b_{bp}}$ - as the particle backscatter coefficient
 219 ($b_{bp}(\lambda)$) divided by $b_p(\lambda)$. This parameter provides insight into the nature of
 220 the particle composition, especially in detecting the presence of gas-vacuolate
 221 cyanobacteria Moore et al. (2017). All data were averaged into 0.5 *m* depth
 222 bins. Further details of the package and data processing are contained in
 223 Moore et al. (2017).

224 For the NRL data set, 11 stations were sampled with a vertical profiling
 225 package that included dual WET Labs ac-9 systems - one equipped with

226 a $0.2 \mu M$ filter and one for non-filtered water - measuring absorption and
227 attenuation at the same wavelengths as above. Processing of the ac-9 data
228 followed the same processing protocol applied to the above ac-9 data. Further
229 details are available on the NASA SeaBASS website Gould (2014).

230 2.2.3. LIDAR data

231 Between August 17 through August 28, 2014, over 50 LIDAR tracks were
232 flown over Lake Erie. The LIDAR, developed in-house at NOAA, used lin-
233 early polarized light at a wavelength of $532 nm$. The laser transmitter pro-
234 duced $100 mJ$ in $10 ns$ pulses at a rate of $30 Hz$. The beam was expanded
235 to $5 mrad$, so the illuminated spot diameter depended on flight altitude. For
236 the data reported here, the spot diameter was generally between 5 and $8.5 m$,
237 except for track $T23$, where it was $15 m$. Two receiver telescopes collected
238 the returns that were co-polarized and cross-polarized with respect to the
239 transmitted light. Each channel used a photomultiplier tube as a detector,
240 followed by a logarithmic amplifier and an eight bit digitizer with a $1 GHz$
241 sample rate. For this study, the cross-polarized channel was used, because
242 it provides better sensitivity to large, irregularly shaped particles such as
243 *Microcystis* and *Planktothrix* colonial aggregates. The reason for this sensi-
244 tivity is that a co-polarized or unpolarized receiver is sensitive to light that
245 is specularly reflected from the surface, scattered by spherical particles like
246 bubbles in the water, and scattered by lake water. None of these components
247 depolarize, so their contribution can mask the co-polarized scattering from

248 the algal cells of interest. The most common effect would be an enhanced
249 signal near the surface due to the specular reflection and bubbles near the
250 surface. This enhanced signal might be mistaken for a surface algal layer
251 unless the cross-polarized return is used. The system was not calibrated to
252 provide cell counts, but all data are presented on the same relative scale.

253 Processing of the LIDAR data involved several steps. First, segments
254 of the data were identified where the aircraft was flying straight and level.
255 Then, the raw digitization levels for these segments were converted to photo-
256 cathode current values using the measured system response. The surface was
257 identified in each LIDAR return, and the depth of each subsurface sample
258 calculated from the time difference between the surface and that sample. The
259 exponential attenuation of the signal in water was estimated for each return
260 using a linear regression to the logarithm of the return over the depth range
261 of 2-4 *m*, and the data were corrected to remove the effects of attenuation.
262 Finally, the data were multiplied by the square of aircraft altitude, so data
263 taken at different altitudes can be compared directly.

264 Background level and system noise level were calculated as the mean and
265 standard deviation of the last 100 samples of each shot. Background level was
266 subtracted from each sample. Penetration depth was defined as the depth
267 at which the signal first dropped to less than 3 times the system noise level.
268 For all flights, the median penetration depth was 14 m, and the penetration
269 depth was over 6 m for 90% of the data. This is consistent with penetration
270 depths found in other turbid waters, such as Chesapeake Bay (Churnside

et al. 2011). Data from below the calculated penetration depth were not used in the analysis to eliminate possible artifacts.

From the processing, cross-polarized attenuation, cross-polarized penetration depth and cross-polarized echograms were generated. We will be primarily showing the echograms from a number of tracks, which show along-track vertical distributions of particles. These have been smoothed in the vertical dimension by the finite laser pulse length, which corresponds to a sliding window of about 1 *m*. Some tracks were initiated with north-to-south or east-to-west orientation. In our subsequent figures, we mirror-reverse some echograms to match the orientation of the lake. All echograms were displayed to the same absolute color scale of 0-7.8 Am^2 , which corresponds to a LIDAR signal current of 0-7.8 μA at a flight altitude of 1000 *m*. The intensity of the LIDAR signal current is directly proportional to the particle concentration, and thus the color is a reflection of particle concentration. Each x-axis is also on a dimensionless relative scale. Absolute distance and other track information (time, date, location) are contained in Table 3. In all cases, the data were inspected visually to ensure that the echograms were not influenced by reflections from the lake bottom.

2.3. Diffuse attenuation coefficient and optical depth

The diffuse attenuation coefficient at 490 *nm*, K_d490 , was estimated from the IOP data based on Lee et al. (2005):

$$K_d(490) = (1 + .005 * solz) * a + 4.18 * (1 - 0.52 * exp(-10.8 * a)) * b_b \quad (1)$$

292

293 where *solz* is the solar zenith angle, *a* is the total absorption at 490 nm, and
 294 *b_b* is the total backscatter at 490 nm. The optical depth at which 10% of the
 295 light remains is calculated as:

$$Z_{10\%} = \frac{2.3}{K_d(490)} \quad (2)$$

296

297

298 2.3.1. HOLOCAM Particle count and cell identification

299 Phytoplankton populations were identified and counted using the HOLO-
 300 CAM, an *in situ* holographic imaging system (Twardowski et al. 2016; Za-
 301 mankhan et al. 2016; Nayak et al. 2018). Briefly, digital holography involves
 302 illuminating a region of interest (sample volume) with a coherent beam of
 303 light (e.g., laser beam). The diffraction patterns that are a result of the in-
 304 terference between light scattered by particles in the volume, and the undis-
 305 turbed portion of the laser beam, are recorded on an imaging device. Numer-
 306 ical schemes are then used to reconstruct the hologram in 2-D cross-sections
 307 within the sampling volume, which results in recording of all in-focus particles

308 within that particular plane. Thus, repeating it over multiple cross-sections
309 across the entire sampling volume, enables the detection and segregation of
310 discrete particles within this entire 3-D space. Further details on the holo-
311 graphic imaging technique and its applications can be found elsewhere (e.g.,
312 Katz and Sheng 2010; Talapatra et al. 2013).

313 The HOLOCAM consists of a 660 nm laser which acts as the coherent il-
314 lumination source, collimating optics, and a camera recording the holograms
315 at 15 Hz. The entire system is designed to be lowered and raised through
316 the water column, while continuously recording holograms which encapsu-
317 late information about the particle fields within the sampling volume. The
318 HOLOCAM was deployed from the boat by hand, at a slow rate to minimize
319 disturbance while traversing through the water column vertically. This en-
320 abled the characterization of particle fields within a size range of $\sim 1 \mu\text{m}$ to
321 $\sim 10 \text{ mm}$ in their true state, i.e., without inducing any particle breakage, at
322 least during the downward profiles. It is to be noted that during the upcast,
323 while the system is being retrieved, the sample volume lies in the wake of the
324 system and thus sees well-mixed, turbulent flow which can lead to fragmen-
325 tation of particles. In fact, prior comparisons of particle size distributions
326 between downcasts and upcasts, have shown a 10-15 % decrease in large
327 particle counts ($>150 \mu\text{m}$) during upcasts, and a corresponding increase in
328 smaller particle populations. Thus, while data was recorded during upcasts,
329 to avoid adding this uncertainty to the data, only the downcasts have been
330 processed for this analysis. At several stations, the system profiled the wa-

331 ter column twice in succession, resulting in two downcasts for the relevant
332 station. In such cases, data has been averaged over both downcasts before
333 being presented.

334 In general, holographic post-processing scheme used here involves three
335 steps: background subtraction, hologram reconstruction, and composite im-
336 age formation. First, for a given profile, the average image obtained from all
337 the holograms is generated and subsequently subtracted from each hologram.
338 This helps minimize background intensity variations, as well as facilitates re-
339 moval of static particles, e.g., dust on imaging windows. Second, hologram
340 reconstruction was carried out in 500 μm incremental depth steps over the
341 entire 4 cm sampling volume. Finally, in-focus particles in each reconstructed
342 plane were then consolidated into one composite image. Once the particle list
343 is generated, parameters including area, aspect ratio, major axis length, etc.,
344 are used to further isolate both *Microcystis* and *Planktothrix* colonies. Fig-
345 ure 2 illustrates the entire post-processing methodology as applied to a single
346 hologram containing *Planktothrix* colonies. A more thorough overview of the
347 image processing routines/methodology in creating the particle list from each
348 raw hologram is provided in Nayak et al. (2018). Repeating this procedure
349 for each hologram in a depth profile, provides a vertical distribution of colony
350 number for each species. For *Microcystis*, the empirical relationship of Joung
351 et al. (2006) was used to derive the cell count in each colony based on the
352 surface area. For *Planktothrix* on the other hand, each cell is assumed to
353 be 3.5 μm in length (Churro et al., 2017). Based on this, and knowing the

length of the filament of each colony, the number of *Planktothrix* cells were
estimated. Cell counts were then binned at 0.5 m depths to generate vertical
profiles of cell counts for the two species.

2.3.2. Winds

Wind data (speed, direction) were obtained from four different sites with
anemometers (Figure 1) through on-line resources managed and maintained
by the National Oceanic and Atmospheric Administration (NOAA). These
were Toledo Light 2 (TOL2) - a coast guard tower in western Lake Erie and a
part of the Great Lakes Real-time Coastal Observation Network (ReCON), a
Coastal-Marine Automated Network (C-MAN) station on South Bass Island
(SBIO), a C-MAN station at Marblehead (MRHO) on land near Sandusky
Bay, and a C-MAN station near Toledo Harbor (THRO) on land near the
mouth of the Maumee River. Raw data points were downloaded and syn-
chronized to the field data collected during the study time period.

2.3.3. Remote sensing reflectance measurements

Above-surface $R_{rs}(\lambda)$ measurements were made with a Field Spec Pro
VNIR-NIR1 portable spectrometer system from Analytical Spectral Devices
(Boulder, Colorado) for both UNH and NRL data sets. A sequence of radi-
ance measurements of a gray plaque ($L_g(\lambda)$), water surface ($L_t(\lambda)$) and sky
($L_{sky}(\lambda)$) were made and used to derive $R_{rs}(\lambda)$. Briefly, the L_t and L_{sky} mea-
surements were used to derive an estimate of spectral water-leaving radiance
 L_w :

376

$$L_w(\lambda) = L_t(\lambda) - \rho L_{sky}(\lambda) \quad (3)$$

377

378

379 The reflectance, ρ , represents the proportion of incident light, which is
 380 reflected by a flat water surface at the angle of observation, as determined
 381 by Fresnel's Equation (Kirk 1994). The Fresnel reflectance used was 0.028
 382 (Austin 1972). The downwelling irradiance $E_d(\lambda)$ was calculated from $L_g(\lambda)$
 383 assuming that the gray plaque is a Lambertian diffuser as:

384

$$E_d(\lambda) = \frac{\pi L_g}{R_g} \quad (4)$$

385

386

387 The R_g derivation from L_g was based on the spectral reflectivity of the
 388 gray plaque (approximately 10% reflection). Above surface $R_{rs}(\lambda)$ was cal-
 389 culated as the ratio of $L_w(\lambda)$ to $E_d(\lambda)$.

390 For the NRL measurements, $R_{rs}(\lambda)$ was computed following the same
 391 basic protocol, with some differences in the how surface and sky reflectance
 392 were computed. A "white" normalization algorithm was applied over a range
 393 of 700 *nm* to 825 *nm* rather than the 750 *nm* specified in (Carder and

394 Steward, 1985). Further details are provided in the NRL SeaBASS files
395 Gould (2014), and in the Ocean Optics Protocols, Vol III, Chapter 3 Method
396 2 Mueller et al. (2003).

397 **3. Results**

398 During the study period, a cyanobacteria bloom was occurring throughout
399 the southern portion of the western basin extending from Maumee Bay to
400 beyond the islands into the central basin. From a macro point of view,
401 this is considered one bloom, but in fact there was a *Microcystis* bloom
402 occurring in the southern half of the western basin and a *Planktothrix* bloom
403 occurring in the southwest region of the central basin. The northern extent
404 of the Maumee Bay bloom was bounded by the transition zone of the Detroit
405 River plume. The northeastern boundary near the island region was more
406 complex (see section 3.2). The location of the Detroit River plume front
407 was dynamic, and changed with meteorological conditions. Strong sustained
408 winds were observed from August 12 through August 14 on several of the
409 wind stations (Figure 3). This altered the hydrography of the western basin,
410 and transported the Detroit River water to the south and encroached into
411 Maumee Bay. Winds decreased from August 17 through August 23, the
412 main window when LIDAR measurements, IOP and holographic profiles were
413 collected. Winds decreased to near or at zero m/s on August 21 across the
414 entire basin. After August 23, wind speeds increased across the basin.

415 3.1. Horizontal distributions of particles

416 The surface SPM varied across the region (Figure 4), with highest values
417 in Maumee Bay (Area 4) with a median of 19.4 g/L , and lowest measured
418 values in the Detroit River (Area 3) with a median of 3.3 g/L . At sites in
419 the island region (Area 6), the SPM values were intermediate with a median
420 value of 5.1 g/L . The particle fields comprised organic and inorganic types,
421 with the lowest organic content in the Detroit River (Area 3) with a median
422 organic/Total ratio of 0.27, and highest in the southeastern western basin
423 (Area 1) with a median of 0.62. The highest ratio was found in this area as
424 well, with a value of 0.79. Inorganic particles were continuously present even
425 in bloom areas, a result of the shallowness of the basin.

426 The algal populations were primarily composed of cyanobacteria during
427 the sampling period, although other groups likely co-existed but were not
428 abundant and not recorded. Based on the holographic image data (Figure
429 4), we observed *Microcystis* and *Planktothrix* genera present in the western
430 and central basins (Moore et al., 2017). The western basin regions (Areas
431 1 through 5) were dominated by *Microcystis*, and the central basin regions
432 (Areas 7 through 9) were dominated by *Planktothrix*. The transition region
433 (Areas 6) contained mixtures of the two genera.

434 3.2. Vertical distributions of particles

435 The vertical particle fields of the different sub-basin areas varied. The
436 driving factors governing the variations in the vertical particle distributions

were the wind speed, presence/absence of cyanobacteria and their taxonomic type. The vertical particle patterns were also the driving factor governing the in-water optical properties and associated remote sensing reflectance. These distributions are explored in the following sections in more detail. We present the data organized by sub-region, and examine the coincident measurements where possible.

3.2.1. Area 1: The southeastern region of the western basin

The highest concentrations of algal particles and *Chl-a* were found in this region during the study period. This area was sampled directly on August 20 and 21, 2014, and was observed by the aircraft LIDAR on three consecutive days from August 17 through August 19, 2014 (Figure 5). In all tracks, particles were concentrated towards the surface in a layer one to two meters thick - a ubiquitous feature from this area. The surface features also followed bathymetric features in some but not all echograms. These similarities exist in other transect data from other areas (e.g., Area 2). We believe the surface features result from particles and not artifacts of the processing. These are interesting features nonetheless, but we do not explore this subject further.

The surface features were prominent over a number of days when winds were 5 *m/s* or less. Particle fields in tracks *T11* and *T2* also showed surface layers losing form and diminishing in northward directions on the edges of the cyanobacteria bloom to non-bloom waters. Particles in these transitions appear to dilute and spread vertically from the surface layer towards the

459 bottom, and a similar pattern was present in the eastern end of track *T10*
460 (near the transition into the central basin).

461 In this area, holographic image analysis showed a large portion of the
462 particle field was dominated by cyanobacteria, specifically *Microcystis*. We
463 compared the LIDAR vertical patterns to overall cyanobacteria cell counts
464 from holographic station profiles from matchups. The LIDAR tracks and
465 the station profiles were not exactly coincident in space and time. Although
466 stations were generally within several *km* to the nearest track location (Table
467 5), the temporal difference was several days. This is not ideal, as these
468 differences introduce mismatch errors into the comparisons. However, the
469 surface layer feature in this area was a persistent feature over the course of
470 days. The correlations between cyanobacteria cell counts and the LIDAR
471 return signal strength were high for the stations this area, ranging from 0.85
472 to 0.96 (Figure 5F-H). The LIDAR patterns were a good descriptor for the
473 cyanobacteria distributions here.

474 Examining the distributions of the two main genera, *Microcystis* cell con-
475 centrations were highest near the surface at the three stations (*S18*, *S19* and
476 *S20*) with levels exceeding 1×10^6 cells mL^{-1} (Figure 6A). *Planktothrix* cells
477 were also observed, but were lower in number (less than 2×10^4 cells mL^{-1})
478 and followed different vertical profile structures (Figure 6B). None of these
479 three stations showed a surface maxima for *Planktothrix*. Stations *S19* and
480 *S20* showed sub-surface maxima at 2 *m* and 4 *m* deep, respectively, below
481 the maxima of the *Microcystis*, while *S18* showed a more uniform distribu-

tion. Vertically-normalized cell count ratios (by integrated column sum) of *Microcystis* to *Planktothrix* showed a decrease from the surface to a depth of about 4 m, then increased towards the bottom at all stations (Figure 6C). There was an uneven vertical distribution of these two populations, with *Planktothrix* increasing over *Microcystis* through depth. The increase in *Microcystis* cells at the bottom was likely from *Microcystis* cells/colonies having sunk.

Water temperature profiles indicate some thermal variation with warmest waters near the surface (Figure 6D). This enhanced stratification would favor *Microcystis* surface accumulation/retention. Surface b_{bp} at 443 nm (in Area 1) ranged over a factor of two ($0.2\ m^{-1}$ to $0.4\ m^{-1}$), and station *LE5* was oversaturated, as evident in the b_{bp} profile (Figure 6E). Omitting this station, this area had the highest b_{bp} values recorded during the study period. There were also variations in the profile structure, with *S20* showing a strong vertical gradient. High values were found at the surface and towards the bottom. High $\widetilde{b_{bp}}$ (> 0.03) and particle absorption coefficients indicate the dominance of gas-vacuolate cyanobacteria on the backscattering efficiency (Figure 6F,G).

The R_{rs} spectra from this area contain features in the red/NIR (Figure 6H), also consistent with high biomass - a trough at 675 nm and high peaks at 555 nm and 709 nm. The high features in R_{rs} beyond 700nm from *S20* optically resemble land vegetation (Hu et al. 2010; Kutser 2004). Although the particle fields in Area 1 are part of the same broader population, the R_{rs}

505 NIR signal from $S18$ and $S19$ are well below that of $S20$, and highlights a
506 difference in R_{rs} between populations just below the water interface and at
507 the surface Kutser (2004).

508 *3.2.2. Area 2, 3 and 5: northern transition edges in the western basin*

509 The Detroit River plume (Area 3) was sampled in the field and by the
510 aircraft on the same day on August 19, 2014. There were few particles
511 (and virtually no cyanobacteria) in the water relative to other areas, and
512 contained few noteworthy features in the vertical structure (not shown). The
513 more interesting particle features were the transitions from cyanobacteria to
514 non-cyanobacteria waters, located between the southern and northern half of
515 the western basin. Echograms from tracks in Areas 2 and 5 highlight these
516 transitions (Figure 7). These tracks were flown on different days (Table 3),
517 and likely reflect some changes in water structure from winds. Particles were
518 concentrated near the surface and sharply discontinuing at transitions into
519 waters associated with the Detroit River plume in the western half of tracks
520 $T41$ and $T44$, and northern part of track $T42$. Track $T44$ was flown a few
521 days after track $T41$, during a period when winds were decreasing to minimal
522 levels on August 21 when track $T44$ was flown.

523 A station was sampled towards the eastern segments of these tracks (sta-
524 tion $S11$) with the HOLOCAM. The cell counts and LIDAR strength of the
525 nearest track point on $T44$ have a low correlation ($R=0.30$) (Figure 7G).
526 Winds were strong enough (5 to 10 knots) from August 19 to August 20

527 when track *T41* and station *S11* were observed to prevent the surface layers
 528 of cyanobacteria forming in this area, which were low in overall concentration
 529 relative to Area 1 stations. Track *T41*, parallel to track *T44*, was flown on
 530 August 19, a day before the station profile. The nearest LIDAR profile from
 531 this track to station *S11* shows a lower particle maximum than track *T44*,
 532 and it appears the whole profile is offset by a meter or so. The features in
 533 this profile are also slightly offset from features in the holographic profile,
 534 leading to a negative correlation ($R=-0.13$). From these three profiles, taken
 535 on consecutive days, it appears that the particle field was moving upward
 536 as winds were decreasing, resulting in a weak near-surface layer to form on
 537 August 21.

538 Tracks *T21* and *T23* were flown days later after wind increases across the
 539 basin between August 23 and August 25 (Figure 3). Particles were more
 540 dispersed throughout the water column and a more gradual northward tran-
 541 sition occurred on track *T21* compared to track *T42*. Particle distributions
 542 from along track *T23* (August 28) showed a transition between north and
 543 south, with a high number of particles distributed through the water column
 544 in the southern half. The southern portions of these tracks are connected to
 545 the particle fields from Area 1 and belong to the same cyanobacteria bloom.
 546 Station *S17* was sampled near track *T23*. The correlation between holo-
 547 graphic cell counts and the nearest LIDAR track point was poor ($R=0.06$)
 548 (Figure 7H). In this case, the time difference was over 7 days, and not ex-
 549 pected to be highly correlated, although both LIDAR and cell counts were

low relative to other stations.

Four profiles of IOPs (two paired with HOLOCAM profiles) were made in this area. Cyanobacteria counts for both *Microcystis* and *Planktothrix* were low relative to Area 1 (where the bloom was most intense), particularly Station *S17* which was between the bloom area and the Detroit River plume front (Figure 8A-C). The vertical profiles of the cyanobacteria counts showed less structure and were more uniform compared to the profiles from Area 1, but normalized cell ratios decreased for *Microcystis* relative to *Planktothrix* at station *S17*.

Profiles of water temperature showed a mostly uniform structure at all stations, and the IOP profiles also were uniform vertically (Figure 8D). The b_{bp} values ranged from $0.02\ m^{-1}$ to $0.15\ m^{-1}$ throughout the water column, much lower than those from Area 1 (Figure 8E). In comparison, station *S11* contained higher cyanobacteria counts, $\widetilde{b_{bp}}$ and R_{rs} . Optically and ecologically, this station was still in the bloom region. Station *LE7* (no HOLOCAM data), near Station *S11*, also had high $\widetilde{b_{bp}}$ in the IOP profile (Figure 8F) and similar R_{rs} shape but lower magnitude than *S11* (Figure 8H). For *S17* and *LE3*, R_{rs} were much lower with no spectral features in the red/NIR, and are considered as outside the bloom.

3.2.3. Area 6: transition zones in the island region

In contrast to Area 1, the particle distributions were more dispersed from about 2 *m* depth down to 8 *m* approaching the bottom in the island region

(Figure 9). There were no distinct surface layers present in the echograms. There was horizontal variability along some tracks, notable track *T15* and *T16* with particles appearing more concentrated in the northern ends. Tracks *T17* and *T18* were taken in the same region but a few days after tracks *T15* and *T16*, and show weaker particle concentration yet the same dispersed pattern from the surface down to 8 *m*. The echogram for track *T27*, an east-west track linking the island region to the central basin, shows moderately high particle concentrations distributed throughout the water column, with a slight sub-surface maximum forming at the western edge of the track closer to the islands. The hologram profile near this track showed low cell counts overall, and good agreement with the LIDAR data ($R=0.78$) (Figure 9H). Further in the south below Kellys Island, a diffuse particle distribution was detected with higher concentrations away from the surface. A nearby station with holographic data also showed a weak, diffuse cell count profile (station *S13*) and the correlation is weaker ($R=0.45$), but both LIDAR strength and cell count totals were low. The time difference (three days) may explain the low correlation for this pairing.

We found mixtures of *Microcystis* and *Planktothrix* at stations within and around the islands (Figure 10A-C). Relative to stations from other areas, there were high amounts of *Planktothrix* at station *S12* (3×10^4 cells mL^{-1}) in the middle of the island formation, with a mostly uniform vertical profile and a weak increase towards the surface. Very few *Microcystis* colonies were observed here (less than 2×10^5 cells mL^{-1}). We note that *Mi-*

595 *crocostis* cells/colonies were difficult to visually identify from other particles
 596 in the image processing of the HOLOCAM data. These particles were low in
 597 concentration throughout the water column, but were counted as *Microcystis*
 598 and may be an overestimate. The other two stations with holographic data
 599 (*S13* and *S15*) - both at the southern end of the island channel connecting
 600 the central and western basins - had similar vertical profiles for *Planktothrix*.
 601 Both stations contained high cell counts (greater than 3×10^4 cells mL^{-1})
 602 with non-uniform vertical distributions concentrated at depths of about 3 *m*.
 603 These sub-surface maxima were about twice as high as the surface concen-
 604 trations, and remained high from this maxima layer to the bottom at 7 *m*.
 605 Station *S12* had a uniform temperature structure, but stations *S13* and *S15*
 606 showed a surface warming with potential gradients setting up. Sub-surface
 607 *Planktothrix* maxima are below this, while in Area 1 *Microcystis* maxima
 608 were in the warming surface layer (Figure 10D).

609 Optical properties increased weakly over depth for b_{bp} , $\widetilde{b_{bp}}$ and a_p (Figure
 610 10E-G), consistent with holographic cell counts. The R_{rs} spectra for all three
 611 stations exhibit a broad peak at 550 *nm* with weaker but identifiable red/NIR
 612 features, indicative of moderate particle concentrations in the surface (Figure
 613 10H). The "U" shape between 670 *nm* and 709 *nm* however is evident but
 614 the depression at 620 *nm* (from phycocyanin absorption) is not pronounced.

615 *3.2.4. Area 8: Southeast of Islands*

616 This area contained the most discrete stations and is well represented by
617 IOP profiles and LIDAR tracks. However, only one station (*S14*) contained a
618 HOLOCAM profile. The echograms from this area contain varying patterns,
619 but all show dispersed particles evenly distributed through the water column
620 and no surface accumulation. Particle concentrations were low in the water
621 column in tracks *T1*, *T19* and *T26* (Figure 11). These tracks are all outside
622 Sandusky Bay to the northeast of the mouth. Conversely, moderate particle
623 concentrations were disbursed evenly, with a slight indication of sub-surface
624 maxima several meters below the surface (2 to 3 *m* depth) in tracks *T8*, *T27*
625 and *T28* (in the middle of the islands in Area 6). Particle concentrations were
626 elevated throughout the water column on the northern end of track *T30* near
627 Area 1, and were lowest overall in the eastern segment of track *T24*. An
628 elevated sub-surface particle field extending from 3 *m* to 10 *m* with even
629 distribution is seen towards the western end of the segment. This pattern
630 continues for the remainder of track *T24* with an abrupt increase in particle
631 concentration near the western edge. This track continued into track *T27*
632 (see section 3.2.3).

633 The lone holographic profile (station *S14*) was matched to LIDAR track
634 *T26* (Figure 11H). The cell counts for this station were lower relative to
635 stations from other areas, although not the lowest. The vertical structures
636 of cell counts and LIDAR signal strength follow the same pattern - lower at
637 the surface with a deeper, relatively constant level, and showed a moderate

638 correlation ($R=0.65$), with a time difference of less than a day. The vertical
639 patterns for both *Microcystis* and *Planktothrix* were similar (Figure 12A-
640 C). There was a cell maximum was below the surface at around 3 *m* depth
641 accompanied with weak, featureless changes in normalized cell ratios. Water
642 temperature profiles had more vertical variation than other areas, and most
643 stations exhibited warmer temperatures at the surface (Figure 12D). Station
644 *LE12*, located at the northern end of track *T26*, showed the strongest thermal
645 gradient with stratification setting in in the top 2-3 *m*. The other stations
646 displayed more gradual changes.

647 Surface b_{bp} were low overall relative to Area 1, and showed weak verti-
648 cal structure except station *LE12* where a shallow thermocline developed
649 (Figure 12E). Some profiles also contained bottom increase, which could be
650 re-suspended sediments or colonies that have sunk. The $\widetilde{b_{bp}}$ varied over a
651 wide range at the surface (Figure 12F), with highest value at station *LE12*
652 (along with stations *LE9* and *LE14*), which is an indication of dominance
653 by gas-vacuolate cyanobacteria. Stations with lower values (*S14*, *LE10* and
654 *LE15*) may have contained cyanobacteria, but the particle fields may not
655 have been dominated by them. Vertical structures of $\widetilde{b_{bp}}$ at stations *LE11*
656 and *LE12* showed near-surface maxima, while the profiles at stations *LE10*
657 and *LE15* showed increasing values with depth. Values are above 0.03 start-
658 ing at depths below 6 *m*, indicating a deeper cyanobacteria population. The
659 a_p values were high at the surface in some stations, and also at the bottom
660 (*LE11* and *LE12*) which also indicate a bottom population (Figure 12G).

661 The R_{rs} spectra all contained a peak at 550 nm (Figure 12H), as well as
 662 red/NIR features associated with cyanobacteria - shoulder peaks at 650 and
 663 709 nm forming the "U" shape. These features are weaker and less promi-
 664 nent compared to R_{rs} from other areas. Notable among this group, station
 665 LE12 containing the highest R_{rs} magnitude and deep spectral features in
 666 the red/NIR. This was an unusual station that stands out from the other
 667 stations in this area, and the pronounced surface thermal layer that may
 668 have acted to maintain cyanobacteria cells near the surface, consistent with
 669 this station's R_{rs} spectra and IOP profiles.

670 3.3. Light levels and cell distributions

671 The spatial distribution of light attenuation at 490 nm (K_d490) was ex-
 672 amined for connections with vertical population structure (Figure 13A). The
 673 derived K_d490 varied across the region geographically, and the mean value
 674 for the southern half of the western basin ($=2.17\ m^{-1}$) was double that of the
 675 central basin (mean of $1.02\ m^{-1}$). The algal communities themselves modified
 676 K_d490 through cellular absorption and scattering processes, and was more
 677 pronounced in the surface waters in the western basin (Area 1). The K_d490
 678 patterns explain differences in sub-surface maxima of the *Planktothrix* com-
 679 munities between areas. In the western basin (Area 1), *Planktothrix* maxima
 680 were closer to the surface where light was attenuated more rapidly compared
 681 to the areas in the island region and central basin.

682 Zeaxanthin, a photoprotective pigment found in both *Microcystis* and

683 *Planktothrix* (Descy et al. 2009; Schagerl and Müller 2006), and the ra-
684 tio of zeaxanthin to total *Chl-a* ($zea/Chl - a$) in surface waters were also
685 mapped (Figure 13B). The highest concentrations of zeaxanthin occurred in
686 the western basin, but the higher $zea/Chl - a$ occurred in the island and
687 central basin region (mean of 0.080), and lower $zea/Chl - a$ (mean of 0.054)
688 were measured in the western basin. *Planktothrix* has been reported to have
689 a higher $zea/Chl - a$ than *Microcystis* in laboratory culture at similar light
690 levels Schlüter et al. (2006), suggesting that *Planktothrix* cells required more
691 photoprotection from light.

692 From K_d490 , the optical depth (Z_{10}) was derived (equation 2), yielding
693 the depth where 10% of the light remains. The depths of the *Planktothrix*
694 cell maxima were derived from the HOLOCAM profiles (Figure 13C). Deeper
695 *Planktothrix* maxima were observed in clear waters (higher Z_{10}). From this,
696 we believe that low winds, light intensity and water clarity were prime de-
697 termining factors in influencing the vertical *Planktothrix* distributions across
698 the region.

699 The surface concentrations of all cyanobacteria cells were an order of
700 magnitude less than the water column total when integrated over the full
701 depth range from the holographic dataset. There was a consistent log-linear
702 relationship between surface cell concentration and total column counts for
703 both genera, combining to form a continuum (Figure 13D). The lower cell
704 concentration range ($< 10^5$) is occupied by *Planktothrix*, and the higher range
705 is occupied by *Microcystis*. The highest cell counts for *Planktothrix* were

below the lowest cell counts for *Microcystis*. The mean ratio of *Microcystis* surface cells to column integrated cells was 0.22, whereas the mean ratio for *Planktothrix* was 0.13. On a relative basis, *Microcystis* cells were nearly two times the concentration of *Planktothrix* cells, a majority of which were dispersed throughout the water column.

Penetration depths of the LIDAR were greater than the optical depths reported above. For the Detroit River plume (not shown), LIDAR penetration depths ranged from 12.4 *m* to 13.2 *m*. Around the islands, we saw values from 10.7 *m* (track *T1*) to 12.3 *m* (track *T26*). Penetration depths in the western basin ranged from 5.5 *m* (track *T3*) to 8.2 *m* (track *T2*), with even lower values in Maumee Bay (4.4 *m*, track *T34*). These depths are deep enough to reach the bottom in most places in the western basin. Penetration depths were highest in the central basin (17.4 *m*, track *T5*), but often did not reach the bottom. The effect of limited penetration depth is that cyanobacteria below that depth will not be measured by the LIDAR. For example, *Microcystis* cell counts increased at depths below about 6 *m* at station *S18*. This increase was not seen in the nearby LIDAR profiles, which had penetration depths of about 5.5 *m*.

The LIDAR penetration depth and the optical depth are related, but both describe different aspects of light transmission/attenuation in the water column. The optical depth was derived from surface values for the IOPs using a marine model and predicts the depth where 10% of the light remains assuming homogenous water and does not account for variations within the

729 water column, whereas the LIDAR penetration depth was based on the de-
730 tection of return of photons throughout the water column. Despite these and
731 time/space differences already mentioned, there is a positive correlation be-
732 tween the LIDAR penetration depth and Z_{10} based on the discrete stations
733 and the LIDAR tracks ($R=0.93$) (Figure 14). A non-linear relationship is
734 evident between the two variables, but this is expected, as the light decay is
735 exponential and the IOP vertical structures are heterogeneous in many loca-
736 tions. Based on this comparison, we believe the LIDAR resolves the particle
737 structure at greater depths than expected when just considering Z_{10} depths.

738 4. Discussion

739 4.1. Vertical distributions of particles using LIDAR and Holography

740 A main objective of the study was to use a combination of LIDAR and
741 cell counts from holography to describe the three-dimensional distribution of
742 a cyanobacteria bloom in Lake Erie. The LIDAR observations and the profile
743 structure of cell counts largely agree, despite the differences in time and space
744 between the nearest stations and LIDAR track. These measurements were
745 not planned to be coincident, as they were from individual projects that were
746 not coordinated. Nonetheless, there was good opportunity to compare the
747 two data sets, as the number of LIDAR tracks was extensive over the study
748 region. The correlation between cyanobacteria cell counts and the LIDAR
749 signal strength ranged from -0.13 to 0.96 with an average of 0.53. These
750 were surprisingly high given the differences in time and space. We also note

751 that the LIDAR signal is a function of all particles in the water, and we only
752 quantified the cyanobacteria cell counts for two genera which were dominant.

753 We believe that most of the LIDAR return signal was governed by the
754 cyanobacteria populations for the prime reason that the LIDAR signal is more
755 sensitive to larger, irregularly shaped particles. During our field surveys,
756 the largest particles in the waters were dominated by cyanobacteria cells,
757 strands and colonies. Because of the use of the cross-polarized return, water
758 molecules, very small particles, and spherical particles did not contribute
759 to the LIDAR signal. Thus, the abrupt transitions seen in echograms from
760 north-south transects in Areas 2 and 5 make sense when accounting for this
761 view, and is really the only plausible way to interpret the LIDAR data.

762 Concerning cell identification in the holographic imagery, assumptions
763 had to be made for cell dimensions for the cyanobacteria. *Planktothrix*
764 strands were readily identifiable, but we are not certain of the mean cell
765 length during this bloom event. Our dimensions were based on literature
766 values, but the ultimate impact of this on the error budget for the cell counts
767 is unknown. The same assumptions hold for the *Microcystis* colonies. While
768 readily identifiable, the number of cells per colony was based on reported sizes
769 from the literature. As this was the first field deployment of the HOLOCAM
770 in a freshwater cyanobacteria bloom, the uncertainties of cell counts are not
771 quantified yet for this instrument in this environment.

772 With these considerations, the correlations provide some metric of con-
773 firmation that the LIDAR was detecting the vertical cyanobacteria distri-

774 bution. The merging of these observations and the IOP profiles, the re-
775 flectance measurements and the pigment distributions fit together and form
776 a three-dimensional picture of a multi-species cyanobacteria bloom at peak
777 development.

778 4.2. Plankton distributions

779 The two main cyanobacteria genera present in Lake Erie in August of
780 2014 - *Planktothrix* and *Microcystis* - exhibited different vertical distribution
781 patterns across the lake basins. The *Planktothrix* populations were lower in
782 cell count relative to *Microcystis*, and were found in the southwestern central
783 basin near Sandusky Bay, and the southeastern edge of the western basin.
784 Sandusky Bay directly flows into the southwestern central basin, and is a
785 source of cyanobacteria to Lake Erie. The prominent cyanobacteria species
786 in Sandusky Bay is *Planktothrix agardhii* (Davis et al. 2015; Chaffin and
787 Bridgeman 2014; Rinta-Kanto and Wilhelm 2006), which is commonly found
788 in metalimnion layers (Halvstedt et al., 2007) and shallow turbid freshwaters
789 (Scheffer et al., 1997). Sandusky Bay fits this latter description and is an ideal
790 habitat for *Planktothrix agardhii*. In contrast, the optically clearer surface
791 waters of central basin are not ideal for *Planktothrix agardhii*. At the time
792 of our study, it is plausible that as Sandusky Bay waters entered and mixed
793 with the clearer, less turbid waters of the central basin, the *Planktothrix* cells
794 and colonies were exposed to higher light, and maintained a photoprotective
795 strategy through pigment enhancement and descent through the water col-

796 umn by buoyancy regulation until preferred light levels were reached.

797 The concentrations of *Planktothrix* populations increased with depth in
798 both western and central basins. In contrast, *Microcystis* cells and colonies
799 were found in abundance near the surface, usually within a few meters. The
800 surface layer exhibited remarkable stability and was a consistent and promi-
801 nent feature that stretched from Maumee Bay to the islands in an east-west
802 direction and roughly half way across the western basin in a north-south di-
803 rection. These distributions were observed during a low-wind period, and a
804 near-surface layer roughly 1 meter to 2 meters thick was present over suc-
805 cessive days throughout the week of observation. In this layer, *Microcystis*
806 cells/colonies formed the primary organic component of the surface particle
807 field and were numerically an order of magnitude greater than *Planktothrix*
808 cell counts.

809 The instances where the two genera co-occurred revealed vertical dif-
810 ferentiation by light preferences/tolerances, allowing for co-existence. *Plank-*
811 *tothrix* colonies were consistently deeper in the water column than the surface
812 *Microcystis* layer, but depths of the cell maxima varied in relation to light
813 availability. Dense concentrations of near-surface *Microcystis* populations
814 increased light attenuation within the water column, which would explain
815 the shallower position of the *Planktothrix* maxima in these waters. In this
816 context, *Planktothrix* migrated both up and down to depths where light was
817 optimal for their photosynthesis. *Microcystis* was largely absent in the cen-
818 tral basin, which not only removes resource competition for *Planktothrix*,

819 but also the light shielding function that *Microcystis* provided in the western
820 basin. Our previous analysis of surface IOPs indicated that the central basin
821 waters had higher relative absorption than the western basin, mostly due to
822 colored dissolved organic matter (Moore et al., 2017). This would provide
823 some additional light shielding for the central basin populations, but not to
824 the degree of dense *Microcystis* surface layers as the attenuation coefficients
825 were higher in western basin compared to the central basin.

826 A succession of different cyanobacteria species has been previously ob-
827 served during summer/fall periods in western Lake Erie. Changes in nutrient
828 concentrations (particularly nitrogen) from replete to deplete conditions has
829 been linked to the collapse of *Microcystis* and rise of *Anabaena* communities
830 (Chaffin and Bridgeman 2014; Michalak et al. 2013). Although *Microcystis*
831 has been typically found as the dominant species during summer (Stumpf
832 et al., 2012), co-occurrence of multiple cyanobacteria species is not uncom-
833 mon in Lake Erie (Kutovaya et al. 2012; Millie et al. 2009) and elsewhere
834 (Gitelson 2017; Gagala et al. 2010; Davis et al. 2003). Morphological and
835 genetic differences may account for resource partitioning. Among these, light
836 is a resource that elicits differential responses by cyanobacteria to its vertical
837 variation in spectral quality and intensity.

838 Vertical regulation through buoyancy control gives planktonic organisms
839 an advantage on maintaining a position to their light tolerances. The high
840 light tolerance of *Microcystis* is well known (Walsby, 1991) allowing for sur-
841 vival in surface waters, and gives a competitive advantage in certain con-

842 ditions. The shade-tolerance of *Planktothrix* is also well known (Konopka,
843 1982), typically resulting in deeper populations to depths depending on the
844 clarity of the overlying waters. In a comparison between depths where *Plank-*
845 *tothrix rubescens* populations reached their maxima, the depth occurred
846 deeper in the clearer Lake Zürich (Walsby et al., 2004) than the English
847 lake Blehnam Tarn and was attributed to light intensity (Davis et al., 2003).
848 The interactions between coexisting *Microcystis* and *Planktothrix* are less
849 known, especially in regards to vertical distributions. The patterns of verti-
850 cal maxima for *Planktothrix* in this study resemble the patterns in these other
851 systems; the depth of the cell maxima varied according to light intensity with
852 deeper depths associated with more transparent overlying waters.

853 4.3. On the detection of cyanobacteria blooms from remote sensing

854 One objective of satellite monitoring of cyanobacteria is to determine
855 water column concentration for water quality assessments (e.g., Wynne and
856 Stumpf 2015). Currently, available algorithms do not differentiate between
857 species or genera (e.g., Hunter et al. 2010; Wynne et al. 2010; Simis et al.
858 2005). Although lower in cell number by an order of magnitude, the *Plank-*
859 *tothrix* population observed in this study was nonetheless an important
860 part of the cyanobacteria community, and was the dominant algal popu-
861 lation in the central basin. The lower cell number and more diffuse and
862 deeper distributions challenge the capabilities of detection through passive
863 remote sensing, which center on spectral features that are expressed from 620

864 nm (PC absorption) through the red/NIR region (high biomass backscat-
865 ter/absorption).

866 The R_{rs} peaks in the red/NIR region have been observed for decades
867 across many freshwater systems (see Schalles et al. 1998 and references therein).
868 In earlier studies, the source of the peaks was not well understood (Gitelson,
869 1992), but it was noted early on that spectral variations in the red/NIR, as
870 well as at 620 nm , could be exploited for algorithms to detect cyanobacteria
871 blooms. Since then, a variety of algorithms have been developed to quan-
872 tify total algal biomass (Gitelson et al., 2008), cyanobacteria bloom inten-
873 sity (Wynne et al., 2010) and floating vegetation (Hu et al. 2010; Matthews
874 et al. 2012) using reflectance from a combination of red/NIR bands. Early re-
875 mote sensing studies relied on aircraft imagery (Dierberg and Carrlker 1994).
876 Kutser (2004) showed the capability of a satellite in observing these spectral
877 features with the Hyperion sensor, a prototype hyperspectral radiometer that
878 extended bands into the red and NIR. However, it was not until the MERIS
879 sensor that a global orbiting satellite was equipped with a channel centered
880 at 709 nm . Gower et al. (2005) was one of the first studies to publish data
881 on the use of the MERIS 709 nm channel for detecting special blooms of
882 phytoplankton. Relations of the peak height (between 680 nm and 750 nm)
883 and *Chl-a* showed tight relationships in various water bodies (Gower et al.,
884 2005), and has become an important channel for the detection of freshwater
885 cyanobacteria and eutrophic conditions.

886 Gitelson (1992) and Schalles et al. (1998) attributed the R_{rs} peak between

887 700 *nm* and 710 *nm* to algal biomass. This peak is governed by the compet-
 888 ing processes on the fate of photons between water absorption and particle
 889 backscatter (Kutser et al., 2008). Near-surface particles can overcome water
 890 absorption of photons. The overall strength of the R_{rs} peak is a function of
 891 the vertical position of particles, their number and backscatter efficiency. In
 892 the case of gas-vacuolate cyanobacteria, the backscatter efficiency is greatly
 893 enhanced due to the intracellular gas vacuoles (Matthews and Bernard 2013
 894 and references therein). The resulting positive buoyancy can position and
 895 maintain these cells near the surface, as is often the case with *Microcystis*.
 896 Buoyant cyanobacteria near the surface will enhance the red/NIR R_{rs} fea-
 897 tures relative to other algal groups because of these traits. Binding et al.
 898 (2011) estimated a 3-4 fold increase in R_{rs} from vertically mixed to a surface
 899 concentration of the same population of gas-vacuolate cyanobacteria, also
 900 shown in a modeling study by Kutser et al. (2008). However, a sufficient
 901 number of cyanobacteria cells in surface waters is needed to impact red/NIR
 902 R_{rs} . While the threshold for number of cells is theoretically lower than other
 903 non-vacuolate species because of the higher backscatter, low levels of surface
 904 populations of cyanobacteria leave a weak or undistinguished trace on R_{rs} ,
 905 as was the case for the central basin open water *Planktothrix* bloom. The de-
 906 gree to which remote sensing can specifically detect cyanobacteria blooms is
 907 influenced by the ecology and light tolerances/preferences of species. There
 908 are limitations to what remote sensing can provide, as the near-surface com-
 909 munities may be different from deeper populations and column integrated

910 biomass may be far different than what is being detected at the surface.

911 These red/NIR R_{rs} characteristics of *Microcystis* blooms are not exclu-
912 sive. Floating sargassum have been detected in marine environments (Gower
913 et al., 2006) based on red/NIR R_{rs} features that are similar to surface scums
914 of cyanobacteria blooms. Other cyanobacteria in brackish and freshwater
915 also can produce R_{rs} features similar to *Microcystis* when they are blooming
916 near the surface. Walsby et al. (1997) reported on surface scum formation in
917 the brackish Baltic Sea by the gas-vacuolate cyanobacteria *Aphanizommon*
918 *flos-aquae*, and Shaw et al. (1999) reported on thick, brown surface scums of
919 *Aphanizommon ovalisporum* in Australian lakes. Binding et al. (2011) asso-
920 ciated blooms of *Aphanizommon flos-aquae* with red/NIR features in Lake
921 of the Woods, Minnesota. Schalles et al. (1998) reported similar R_{rs} features
922 in the red/NIR for blooming *Synedra* sp. - a diatom - and *Anabaena* sp. -
923 a cyanobacteria - at different times of year in a eutrophic, freshwater lake.
924 In contrast, low-light adapted *Planktothrix* often dominate in shallow turbid
925 lakes (Scheffer et al., 1997), but are usually found deeper in the water column
926 in clearer lakes away from where remote sensing can detect their presence
927 (Davis et al. 2003; Walsby et al. 2004).

928 The impact of dense, near-surface particles with enhanced backscatter
929 properties (e.g., vacuolate cyanobacteria) on $R_{rs}(\lambda)$ results in a broad eleva-
930 tion of spectral magnitude, accentuated at certain wavelengths dictated by
931 the interplay of light absorption and scattering. Kutser (2004) examined the
932 depth range of light extinction in the context of near-surface cyanobacteria

933 blooms, and found the optical depth of light ($Z_{10\%}$) decreased from a few
 934 meters to zero when cyanobacteria biomass ranged from 1 mg/m^3 to levels
 935 of surface scum in the Baltic Sea. Calculations for $Z_{10\%}$ from this study were
 936 similar, although we note that these are an estimate based on the surface
 937 layer. In the Detroit River plume, the optical depths were deepest averaging
 938 to 4.67 m . At stations around the islands and in the central basin, optical
 939 depths ranged from 1.2 m to 4.2 m averaging to 2.80 m . Stations in the
 940 western basin with *Microcystis* were shallowest ranging from 0.84 m to 2.10
 941 m and an average of 1.4 m . In the case of *Microcystis* with near-surface pop-
 942 ulations occupying the top 2 meters, 90% of photons reaching the satellites
 943 leaving from the water originate from the about the first meter or less. As
 944 a consequence, biomass estimates based on passive remote sensing data can
 945 miss a major fraction of the microbial community, as the penetration depth
 946 is a function of the particle density and associated optical clarity of the wa-
 947 ter column. The LIDAR showed a deeper penetration of light and ability to
 948 resolve a greater vertical range for determining particle distributions. The
 949 LIDAR penetrations depths were highly correlated with the optical depth.
 950 These results seem to favor the use LIDAR technology for assessing the par-
 951 ticle distribution and concentrations over passive remote sensing, even in a
 952 highly turbid environment such as Lake Erie.

953 5. Conclusions

954 The combination of active and passive remote sensing measurements
955 with *in situ* profiles of optical properties and cell counts generated a three-
956 dimensional view of the particle distributions highlighting horizontal and
957 vertical distributions of gas-vacuolate cyanobacteria. An advantage of LI-
958 DAR measurements is the deeper and finer resolution of the vertical particle
959 distribution relative to passive 'ocean color' sensors. Despite time and space
960 differences in matchup quality, the LIDAR-derived vertical particle struc-
961 ture was explained by cyanobacteria cell counts determined with a profiling
962 holographic digital system. Based on these results, additional studies that
963 specifically coordinate LIDAR with *in situ* optical and holographic measure-
964 ments would improve the quantitative uses of LIDAR data. In this study, we
965 used the LIDAR data in a qualitative way to describe overall particles distri-
966 butions across a large lake area. We believe the cyanobacteria populations
967 that comprised the particle fields were well described by these observations,
968 which could be applied to other systems. However, deeper lakes with sub-
969 surface populations may be more problematic as there are limitations to the
970 penetration of a LIDAR system. The use of holography was also important
971 to understanding these vertical distributions in terms of particle composi-
972 tion. The holographic system used in this study was capable of observing
973 the particles undisturbed, as there was no pumping of water into camera
974 fields which minimized cell/colony disruption. The system detects particles
975 over a large size range, from several microns to several millimeters. Given

976 the large colony and aggregate sizes that cyanobacteria can form, this system
977 is well suited for enumerating and identifying cyanobacteria genera in these
978 types of waters and conditions.

979 From a macroscopic point of view, the cyanobacteria formed one large
980 bloom, but details of the data set reveal two co-occurring populations - *Microcystis*
981 *microcystis* and *Planktothrix*. Dense *Microcystis* populations were concentrated
982 near the surface and increased light attenuation, while *Planktothrix* popula-
983 tions were more diffuse, and showed variable vertical maxima depending on
984 the degree of surface light attenuation. These preferences may allow for a
985 degree of niche partitioning and co-existence. *Microcystis* cell counts were an
986 order of magnitude higher than counts for *Planktothrix*, although the latter
987 was dominant in the southwestern central basin.

988 We have shown that the cyanobacteria bloom in western Lake Erie com-
989 prised two different genera and have linked their distributions to light tol-
990 erances/preferences. However, we know less about other ecological aspects,
991 such as cell sources, nutrient sources and life stages. We do not know what
992 part of either populations were ascending or descending, or to what degree
993 nutrient competition was playing a role in these distributions, if at all. We
994 do not have a lot of detail on the vertical distributions of toxins. However,
995 the information we have collated presents a unique picture for understanding
996 how these two cyanobacteria genera co-exist along a light gradient.

997 These ecological life attributes have impacts on the light field relevant
998 to passive remote sensing. Detecting and quantifying cyanobacteria from re-

999 mote sensing and bio-optical algorithms that exploit red/NIR features have
1000 limitations that are dependent on the vertical distribution and concentra-
1001 tions of the cells. When high amounts of cells are concentrated near or at
1002 the surface, the signal is more pronounced and detection is more reliable than
1003 diffusely distributed populations. This would favor more reliable detection
1004 of *Microcystis* compared to *Planktothrix*. We can now add a fourth state to
1005 the three defined by Kutser (2004) - a sub-surface vertically varying popu-
1006 lation with non-homogenous structure. This state could encompass a range
1007 of vertical shapes. These states could switch from one to another rapidly,
1008 within the span of hours, depending on external conditions (e.g., wind, so-
1009 lar irradiation), without changing the column population and overall column
1010 community. There is a continuum of optical outcomes from many permu-
1011 tations of cell density, vertical position, species composition, and pigment
1012 content - all which determine light reflectance. There are likely constrained
1013 ranges of optical outcomes though within each state apart from the whole,
1014 and defining these ranges and linking them to these states would be benefi-
1015 cial towards connecting remote sensing data to cyanobacteria distributions
1016 and associated water quality indices.

1017 **Conflict of Interest Statement**

1018 The authors declare no conflict of interest. Copyright permission has
1019 been obtained for re-production of Figure 3 from the copyright holders.

Author Contributions

TM, JS, MT and SR designed the main field study for the UNH data set. TM, JS, MT, and NS acquired the field data. JC collected and processed the LIDAR data. SR provided the logistical use of NOAA lab space for instrument assembly and water sample processing, use of the NOAA R/V for the study and field support. TJ, JS, MT, AN and MM processed and analyzed the discrete water samples. AN processed the HOLOCAM imagery for cell and particles counts and concentrations. TM and JS made the radiometric measurements. JS, MT and NS processed and analyzed the IOP data. All authors contributed to, commented on and edited the paper.

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1048 **References**

- 1049 APHA (1998). *Standard methods for the examination of water and waste*
1050 *water*. American Public Health Association, Washington, D.C.
- 1051 Austin, R. (1972). The remote sensing of spectral radiance from below the
1052 ocean surface. In Jerlov, N. and Nielsen, S., editors, *Optical Aspects of*
1053 *Oceanography*, pages 201–213. Academic, New York.
- 1054 Binding, C., Greenberg, T., Berome, J., Bukata, R., and Letourneau, G.
1055 (2011). An assessment of MERIS algal products during an intense bloom
1056 in Lake of the Woods. *Journal of Plankton Research*, 33:793–806.
- 1057 Blank, C. (2013). Origin and early evolution of photosynthetic eukaryotes
1058 in freshwater environments: reinterpreting proterozoic paleobiology and
1059 biogeochemical processes in light of trait evolution. *Journal of Phycology*,
1060 130:243–253.
- 1061 Bridgeman, T., Chaffin, J., and Filbrun, J. (2013). A novel method for
1062 tracking western Lake Erie *Microcystis* blooms, 2002–2011. *Journal of Great*
1063 *Lakes Research*, 39:83–89.
- 1064 Carder, K. and Steward, R. (1985). A remote-sensing reflectance model of
1065 a red tide dinoflagellate off West Florida. *Limnology and Oceanography*,
1066 30:286–298.
- 1067 Carey, C., Ibelings, B., Hoffman, E., Hamilton, D., and Brookes, J. (2012).

1068 Eco-physiological adaptations that favour freshwater cyanobacteria in a
1069 changing climate. *Water Research*, 46:1394–1407.

1070 Chaffin, J. and Bridgeman, T. (2014). Organic and inorganic nitrogen utiliza-
1071 tion by nitrogen-stressed cyanobacteria during bloom conditions. *Journal*
1072 *of Applied Phycology*, 26:299–309.

1073 Churnside, J. (2014). Review of profiling oceanographic lidar. *Optical Engi-*
1074 *neering*, 53:051405.

1075 Churnside, J. (2015). Subsurface plankton layers in the Arctic Ocean. *Geo-*
1076 *physical Research Letters*, 42:4896–4902.

1077 Churnside, J. and Donaghay, P. (2009). Thin scattering layers observed by
1078 airborne lidar. *ICES Journal of Marine Science*, 66:778–789.

1079 Churnside, J., Marchbanks, R., Lembke, C., and Beckler, J. (2017). Optical
1080 backscattering measured by airborne lidar and underwater glider. *Remote*
1081 *Sensing*, 9:379.

1082 Churnside, J., Sharov, A., and Richter, R. (2011). Aerial surveys of fish
1083 in estuaries: a case study in chesapeake bay. *ICES Journal of Marine*
1084 *Science*, 68:239–244.

1085 Churro, C., Azevedo, J., Vasconcelos, V., and Silva, A. (2017). Detection of a
1086 *Planktothrix agardhii* bloom in Portuguese marine coastal waters. *Toxins*,
1087 9:391.

- 1088 Davis, P., Dent, M., Parker, J., Reynolds, C., and Walsby, A. (2003). The
1089 annual cycle of growth rate and biomass change in *Planktothrix* spp. in
1090 Blelham Tarn, English Lake District. *Freshwater Biology*, 48:852–867.
- 1091 Davis, T., Bullerjahn, G., Tuttle, T., McKay, R., and Watson, S. (2015).
1092 Effects of increasing nitrogen and phosphorus concentrations on phyto-
1093 plankton community growth and toxicity during *Planktothrix* blooms in
1094 Sandusky Bay, Lake Erie. *Environmental Science and Technology*, 49:7197–
1095 7207.
- 1096 Descy, J.-P., Sarmiento, H., and Higgins, H. (2009). Variability of phyto-
1097 plankton pigment ratios across aquatic environments. *European Journal*
1098 *of Phycology*, 44:319–330.
- 1099 Dierberg, F. and Carrlker, N. (1994). Field testing two instruments for re-
1100 motely sensing water quality in the Tennessee valley. *Environmental Sci-*
1101 *ence and Technology*, 28:16–25.
- 1102 Gagala, I., Izydorczyk, K., Skowron, A., Kamecka-Plaskota, D., Stefaniak,
1103 K., Kokocinski, M., and Mankiewicz-Boczek, J. (2010). Appearance of tox-
1104 igenous cyanobacteria in two Polish lakes dominated by *Microcystis aerugi-*
1105 *nosa* and *Planktothrix agardhii* and environmental factors influence. *Eco-*
1106 *hydrology and Hydrobiology*, 10:25–34.
- 1107 Gitelson, A. (1992). The peak near 700nm on radiance spectra of algae

1108 and water: relationships of its magnitude and position with chlorophyll
1109 concentration. *International Journal of Remote Sensing*, 13:3367–3373.

1110 Gitelson, A. (2017). Unusual cohabitation and competition between *Plank-*
1111 *tothrix rubescens* and *Microcystis* sp. (cyanobacteria) in a subtropical
1112 reservoir (Hammam Debagh) located in Algeria. *PLOS ONE*, 13:3367–
1113 3373.

1114 Gitelson, A., Dall’Olmo, G., Moses, W., Rundquist, D., Barrow, T., Fisher,
1115 T., Gurlin, D., and Holz, J. (2008). A simple semi-analytical model for
1116 remote estimation of chlorophyll-a in turbid waters: Validation. *Remote*
1117 *Sensing of Environment*, 112:3582–3593.

1118 Gould, R. (2014). [https://seabass.gsfc.nasa.gov/archive/NRL/Lake_](https://seabass.gsfc.nasa.gov/archive/NRL/Lake_Eerie_Aug_2014/rv_muksie/documents)
1119 [Eerie_Aug_2014/rv_muksie/documents](https://seabass.gsfc.nasa.gov/archive/NRL/Lake_Eerie_Aug_2014/rv_muksie/documents). NASA SeaWiFS Bio-optical
1120 Storage System (SeaBASS).

1121 Gower, J., Hu, C., Borstad, G., and King, S. (2006). Ocean color satellites
1122 show extensive lines of floating sargassum in the gulf of mexico. *IEEE*
1123 *Trans. Geosci. Remote Sens.*, 44:3619–3625.

1124 Gower, J., King, S., Borstad, G., and Brown, L. (2005). Detection of intense
1125 plankton blooms using the 709 nm band of the meris imaging spectrometer.
1126 *International Journal of Remote Sensing*, 26:2005–2012.

1127 Halvstedt, C., Rohrlack, T., Andersen, T., Skulberg, O., and Evardsen, B.
1128 (2007). Seasonal dynamics and depth distribution of *Planktothrix* spp. in

1129 Lake Steinsfjorden (Norway) related to environmental factors. *Journal of*
1130 *Plankton Research*, 29:471–482.

1131 Harke, M., Steffen, M., Gobler, C., Otten, T., Wilhelm, S., Wood, S., and
1132 Paerl, H. (2016). A review of the global ecology, genomics, and biogeogra-
1133 phy of the toxic cyanobacterium, *Microcystis* spp. *Harmful Algae*, 54:4–20.

1134 Heukelem, L. V. and Thomas, C. (2001). Computer-assisted high-
1135 performance liquid chromatography method development with applica-
1136 tions to the isolation and analysis of phytoplankton pigments. *Journal*
1137 *of Chromatography. A.*, 910:31–49.

1138 Hu, C., Lee, Z., Ma, R., Yu, K., Li, D., and Shang, S. (2010). Moderate Reso-
1139 lution Imaging Spectroradiometer (MODIS) observations of cyanobacteria
1140 blooms in Taihu Lake, China. *Journal of Geophysical Research*, 115.

1141 Humphries, S. and Lyne, V. (1988). Cyanophyte blooms: The role of cell
1142 buoyancy. *Limnology and Oceanography*, 33:79–91.

1143 Hunter, P., Tyler, A., Carvalho, L., Codd, G., and Maberly, S. (2010). Hy-
1144 perspectral remote sensing of cyanobacterial pigments as indicators for cell
1145 populations and toxins in eutrophic lakes. *Remote Sensing of Environment*,
1146 114:2705–2718.

1147 Ibelings, B., Kroon, B., and Mur, C. (1994). Acclimation of photosystem II
1148 in a cyanobacterium and a eukaryotic green alga to high and fluctuating

1149 photosynthetic photon flux densities, simulating light regimes induced by
1150 mixing in lakes. *New Phytologist*, 128:407–424.

1151 IJC (2014). A balanced diet for Lake Erie: Reducing phosphorus loadings
1152 and harmful algal blooms. ISBN: 978-1-927336-07-6.

1153 Jeffrey, S., Mantoura, R., 78, S. W. G., Wright, S., Staff, U., of Scien-
1154 tific Unions. Scientific Committee on Oceanic Research, I. C., and Unesco
1155 (1997). *Phytoplankton Pigments in Oceanography: Guidelines to Modern*
1156 *Methods*. Monographs on oceanographic methodology. UNESCO Publish-
1157 ing.

1158 Joosse, P. and Baker, D. (2011). Context for re-evaluating agricultural source
1159 phosphorus loadings to the Great Lakes. *Canadian Journal of Soil Science*,
1160 91:317327.

1161 Joung, S.-H., Kim, C.-J., Ahn, C.-Y., Jang, K.-Y., Boo, S., and Oh, H.-M.
1162 (2006). Simple method for a cell count of the colonial cyanobacterium,
1163 *Microcystis* sp. *The Journal of Microbiology*, 44:562565.

1164 Kane, D., Conroy, J., Richards, R., Baker, D., and Culver, D. (2014). Re-
1165 eutrophication of Lake Erie: Correlations between tributary nutrient loads
1166 and phytoplankton biomass. *Journal of Great Lakes Research*, 40:496–501.

1167 Katz, J. and Sheng, J. (2010). Applications of holography in fluid mechanics
1168 and particle dynamics. *Annu. Rev. Fluid Mech.*, 42:531–555.

- 1169 Kirk, J. (1994). *Light and photosynthesis in aquatic ecosystems*. Cambridge
1170 University Press, Cambridge, England.
- 1171 Kokocinski, M., K.Stefaniak, Mankiewicz-Boczek, J., Izydorzyc, K., and
1172 Soininen, J. (2010). The ecology of the invasive cyanobacterium *Cylin-*
1173 *drospermopsis raciborskii* (nostocales, cyanophyta) in two hypereutrophic
1174 lakes dominated by *Planktothrix agardhii* (oscillatoriales, cyanophyta). *Eu-*
1175 *ropean Journal of Phycology*, 45:365–374.
- 1176 Konopka, A. (1982). Buoyancy regulation and vertical migration by *Oscil-*
1177 *latoria rubescens* in Crooked Lake, Indiana. *British Phycological Journal*,
1178 17:427–442.
- 1179 Konopka, A. (1989). Metalimnetic cyanobacteria in hard-water lakes: Buoy-
1180 ancy regulation and physiological state. *Limnology and Oceanography*,
1181 34:1174–1184.
- 1182 Kromkamp, J. and Walsby, A. (1990). A computer model of buoyancy and
1183 vertical migration in cyanobacteria. *Journal of Plankton Research*, 12:161–
1184 183.
- 1185 Kutovaya, O., McKay, R., Beall, B., Wilhelm, S., Kane, D., Chaffin, J.,
1186 Bridgeman, T., and Bullerjahn, G. (2012). Evidence against fluvial seeding
1187 of recurrent toxic blooms of *Microcystis* spp. in Lake Eries western basin.
1188 *Harmful Algae*, 15:71–77.

- 1189 Kutser, T. (2004). Quantitative detection of chlorophyll in cyanobacterial
1190 blooms by satellite remote sensing. *Limnology and Oceanography*, 49:2179–
1191 2189.
- 1192 Kutser, T., Metsamaa, L., and Dekker, A. (2008). Influence of the vertical
1193 distribution of cyanobacteria in the water column on the remote sensing
1194 signal. *Estuarine, Coastal and Shelf Science*, 78:649–654.
- 1195 Lee, J., Churnside, J., Marchbanks, R., Donaghay, P., and Sullivan, J. (2013).
1196 Oceanographic lidar profiles compared with estimates from in situ optical
1197 measurements. *Appl. Opt.*, 52:786–794.
- 1198 Lee, Z.-P., Du, K.-P., and Arnone, R. (2005). A model for the diffuse at-
1199 tenuation coefficient of downwelling irradiance. *Journal of Geophysical*
1200 *Research*, 110.
- 1201 Matthews, M. and Bernard, S. (2013). Using a two-layered sphere model to
1202 investigate the impact of gas vacuoles on the inherent optical properties of
1203 *Microcystis aeruginosa*. *Biogeosciences*, 10:8139–8157.
- 1204 Matthews, M., Bernard, S., and Robertson, L. (2012). An algorithm for
1205 detecting trophic status (chlorophyll-a), cyanobacterial-dominance, surface
1206 scums and floating vegetation in inland and coastal waters. *Remote Sensing*
1207 *of Environment*, 124:637–652.
- 1208 Metsamaa, L., Kutser, T., and Strombeck, N. (2006). Recognizing cyanobac-

1209 teria blooms based on their optical signature: a modelling study. *Boreal*
1210 *Environment Research*, 11:493–506.

1211 Michalak, A., Anderson, E., Beletsky, D., Boland, S., Bosch, N., Bridge-
1212 man, T., et al. (2013). Record-setting algal bloom in lake erie caused by
1213 agricultural and meteorological trends consistent with future conditions.
1214 *Proceedings of the National Academy of Sciences*.

1215 Millie, D., Fahnenstiel, G., Dyble-Bressie, J., Pigg, R., Rediske, R., Klarer,
1216 D., Tester, P., and Whitaker, R. (2009). Late-summer phytoplankton in
1217 western Lake Erie (Laurentian Great Lakes): bloom distributions, toxicity,
1218 and environmental influences. *Aquatic Ecology*, 43:915–934.

1219 Moore, T., Mouw, C., Sullivan, J., Twardowski, M., Burtner, A., Ciochetto,
1220 A., McFarland, M., Nayak, A., Paladino, D., Stockley, N., Johengen, T.,
1221 Yu, A., Ruberg, S., and Weidemann, A. (2017). Bio-optical properties of
1222 cyanobacteria blooms in western Lake Erie. *Frontiers in Marine Science*,
1223 4:300.

1224 Mueller, J., Fargion, G., and McClain, C. (2003). Ocean optics protocols for
1225 satellite ocean color sensor validation, revision 4, volume IV. Technical
1226 report, National Aeronautical and Space Administration, Goddard Space
1227 Flight Space Center, Greenbelt, Maryland.

1228 Nayak, A., McFarland, M., Sullivan, J., and Twardowski, M. (2018). Ev-

1229 idence of ubiquitous preferential particle orientation in representative
1230 oceanic shear flows. *Limnology and Oceanography*, 63:122–143.

1231 Oliver, R. (1994). Floating and sinking in gas-vacuolate cyanobacteria. *Jour-*
1232 *nal of Phycology*, 30:161–173.

1233 Oliver, R. and Walsby, A. (1984). Direct evidence for the role of light-
1234 mediated gas vesicle collapse in the buoyancy regulation of *Anabaena flos-*
1235 *aquae* (cyanobacteria). *Limnology and Oceanography*, 29:879–886.

1236 Paerl, H., Bland, P., Bowles, D., and Haibach, M. (1985). Adaptation to
1237 high-intensity, low-wavelength light among surface blooms of the cyanobac-
1238 terium *Microcystis aeruginosa*. *Applied and Environmental Microbiology*,
1239 49:1046–1052.

1240 Paerl, H. and Otten, T. (2013). Blooms bite the hand that feeds them.
1241 *Science*, 342:433–434.

1242 Paerl, H., Tucker, J., and Bland, P. (1983). Carotenoid enhancement and
1243 its role in maintaining blue-green algal (*Microcystis aeruginosa*) surface
1244 blooms. *Limnology and Oceanography*, 28:847–857.

1245 Rinta-Kanto, J. and Wilhelm, S. (2006). Diversity of microcystin-producing
1246 cyanobacteria in spatially isolated regions of lake erie. *Applied and Envi-*
1247 *ronmental Microbiology*, 72:50835085.

1248 Sánchez-Baracaldo, P., Hayes, P., and Blank, C. (2005). Morphological and

1249 habitat evolution in the cyanobacteria using a compartmentalization ap-
1250 proach. *Geobiology*, 3:145–165.

1251 Schagerl, M. and Müller, B. (2006). Acclimation of chlorophyll a and
1252 carotenoid levels to different irradiances in four freshwater cyanobacteria.
1253 *Journal of Plant Physiology*, 163:709–716.

1254 Schalles, J., Gitelson, A., Yacobi, Y., and Kroenke, A. (1998). Estimation
1255 of chlorophyll a from time series measurements of high spectral resolution
1256 reflectance in an eutrophic lake. *Journal of Phycology*, 34:383–390.

1257 Scheffer, M., Rinaldi, S., Gragnani, A., Mur, L., and Nes, E. V. (1997).
1258 On the dominance of filamentous cyanobacteria in shallow, turbid lakes.
1259 *Ecology*, 78:272–282.

1260 Schlüter, L., Lauridsen, T., Krogh, G., and Jørgensen, T. (2006). Iden-
1261 tification and quantification of phytoplankton groups in lakes using new
1262 pigment ratios a comparison between pigment analysis by hplc and mi-
1263 croscopy. *Freshwater Biology*, 51:1474–1485.

1264 Shaw, G., Sukenik, A., Livne, A., Chiswell, R., Smith, M., Seawright, A.,
1265 Norris, R., Eaglesham, G., and Moore, M. (1999). Blooms of the cylin-
1266 dropermopsin containing cyanobacterium, *Aphanizomenon ovalisporum*
1267 (Forti), in newly constructed lakes, Queensland, Australia. *Environmental*
1268 *Toxicology*, 14:167–177.

- 1269 Simis, S., Peeters, S., and Gons, H. (2005). Remote sensing of the cyanobacte-
 1270 rial pigment phycocyanin in turbid inland water. *Limnology and Oceanog-*
 1271 *raphy*, 50:237245.
- 1272 Stomp, M., Huisman, J., Vordos, L., Pick, F., Laamanen, M., Haverkamp, T.,
 1273 and Stal, L. (2007). Colourful coexistence of red and green picocyanobac-
 1274 teria in lakes and seas. *Ecology Letters*, 10:290–298.
- 1275 Stumpf, R., Wynne, T., Baker, D., and Fahnenstiel, G. (2012). Interannual
 1276 variability of cyanobacterial blooms in Lake Erie. *PLoS ONE*, 7.
- 1277 Talapatra, S., Hong, J., McFarland, M., Nayak, A., Zhang, C., Katz, J.,
 1278 Sullivan, J., Twardowski, M., Rines, J., and Donaghay, P. (2013). Char-
 1279 acterization of biophysical interactions in the water column using in situ
 1280 digital holography. *Marine Ecology Progress Series*, 473:29–51.
- 1281 Twardowski, M., Sullivan, J., and Dalgeish, F. (2016). Novel technologies to
 1282 study undisturbed particle fields in the ocean. *Sea Technology*, 57:15–19.
- 1283 Twardowski, M., Sullivan, J., Donaghay, P., and Zaneveld, J. (1999). Mi-
 1284 croscale quantification of the absorption by dissolved and particulate mate-
 1285 rial in coastal waters with an ac-9. *J. Atmos. Ocean. Technol.*, 16:691–707.
- 1286 Uyeda, J., Harmon, L., and Blank, C. (2016). A comprehensive study
 1287 of cyanobacterial morphological and ecological evolutionary dynamics
 1288 through deep geologic time. *PLoS One*, 11.

1289 Visser, P., Passarge, J., and Mur, L. (1997). Modelling vertical migration of
1290 the cyanobacterium *Microcystis*. *Hydrobiologia*, 349:99–109.

1291 Walsby, A. (1991). The mechanical properties of the *Microcystis* gas vesicle.
1292 *Journal of General Microbiology*, 137:24012408.

1293 Walsby, A. (1994). Gas vesicles. *Microbiological Reviews*, 58:94–144.

1294 Walsby, A. and Booker, M. (1980). Changes in buoyancy of a planktonic
1295 blue-green alga in response to light intensity. *British Phycological Journal*,
1296 15:311319.

1297 Walsby, A., Hayes, P., Boje, R., and Stal, L. (1997). The selective advantage
1298 of buoyancy provided by gas vesicles for planktonic cyanobacteria in the
1299 Baltic Sea. *New Phytologist*, 136:407–417.

1300 Walsby, A., Ng, G., Dunn, C., and Davis, P. (2004). Comparison of the
1301 depth where *Planktothrix rubescens* stratifies and the depth where the daily
1302 insolation supports its neutral buoyancy. *New Phytologist*, 162:133145.

1303 Wynne, T. and Stumpf, R. (2015). Spatial and temporal patterns in the
1304 seasonal distribution of toxic cyanobacteria in western Lake Erie from
1305 20022014. *Toxins*, 7:1649–1663.

1306 Wynne, T., Stumpf, R., Tomlinson, M., and Dyble, J. (2010). Characterizing
1307 a cyanobacterial bloom in western Lake Erie using satellite imagery and
1308 meteorological data. *Limnology and Oceanography*, 55:2025–2036.

- 1309 Xue, K., Zhang, Y., Ma, R., and Duan, H. (2017). An approach to correct the
1310 effects of phytoplankton vertical nonuniform distribution on remote sensing
1311 reflectance of cyanobacterial bloom waters. *Limnology and Oceanography:*
1312 *Methods*, 15:302–319.
- 1313 Zamankhan, H., Westrick, J., Anscombe, F., Stumpf, R., Wynne, T., Sulli-
1314 van, J., Twardowski, M., Moore, T., and Choi, H. (2016). *Sustainable Wa-*
1315 *ter Management and Technologies*, chapter Chapter 3: Sustainable moni-
1316 toring of algal blooms. Taylor and Francis Group, Boca Raton, FL.
- 1317 Zaneveld, J., Kitchen, J., and Moore, C. (1994). The scattering error correc-
1318 tion of reflecting tube absorption meters. *Ocean Optics XII, Proc. SPIE*,
1319 2258:44–55.

Table 1: Matrix of *in situ* measurements and *LIDAR* track availability by area (see Figure 1). Y = potential measurement availability.

Area	Type	8/17	8/18	8/19	8/20	8/21	8/22	8/23	8/24	8/28
1 SE Western Basin	<i>LIDAR</i> <i>IOP</i> <i>HOLOCAM</i>	Y	Y	Y Y	Y Y	Y Y			Y	
2 Detroit River front	<i>LIDAR</i> <i>IOP</i> <i>HOLOCAM</i>	Y							Y	
3 Detroit River	<i>LIDAR</i> <i>IOP</i> <i>HOLOCAM</i>			Y Y Y						
4 Maumee Bay	<i>LIDAR</i> <i>IOP</i> <i>HOLOCAM</i>		Y Y	Y						
5 Islands West	<i>LIDAR</i> <i>IOP</i> <i>HOLOCAM</i>	Y		Y	Y Y	Y				
6 Islands Central	<i>LIDAR</i> <i>IOP</i> <i>HOLOCAM</i>		Y		Y Y					
7 Islands East	<i>LIDAR</i> <i>IOP</i> <i>HOLOCAM</i>				Y Y	Y Y	Y	Y Y		
8 Islands Southeast	<i>LIDAR</i> <i>IOP</i> <i>HOLOCAM</i>		Y	Y	Y Y	Y				
9 Sandusky Bay	<i>LIDAR</i> <i>IOP</i> <i>HOLOCAM</i>									Y

Table 2: Data sets.

Source	Chl-a	SPM	PC	HOLOCAM	HPLC	Vertical IOP	AOP	Dates
UNH	20	20	20	13	N/A	14	20	8/19/14-8/21/14
NRL	20	N/A	N/A	N/A	20	11	11	8/18/14-8/28/14

Table 3: Details of LIDAR tracks.

Track	Area	Date (mm/dd/yy)	Length (km)	Time (UTC)	Starting Lat (degrees)	Starting Lon (degrees)	Ending Lat (degrees)	Ending Lon (degrees)
1	8	8/23/14	8.5	15:55	41.58984	-82.62874	41.51921	-82.66784
2	1	8/19/14	12.1	14:16	41.65756	-82.89373	41.68256	-83.03602
3	1	8/18/14	9.9	14:04	41.69559	-83.21689	41.67456	-83.10139
8	8	8/28/14	7.6	18:28	41.47615	-82.57850	41.54406	-82.58718
10	1	8/17/14	17.2	20:02	41.65555	-83.09680	41.61375	-82.89759
11	1	8/23/14	12.0	14:14	41.75874	-83.08436	41.65276	-83.11452
15	6	8/19/14	4.4	17:49	41.64078	-82.80970	41.60128	-82.80908
16	6	8/19/14	2.5	17:34	41.59858	-82.83004	41.62069	-82.83145
17	6	8/21/14	6.5	15:09	41.60887	-82.74957	41.63797	-82.81722
18	6	8/21/14	0.8	15:21	41.61580	-82.80749	41.61181	82.79889
19	8	8/23/14	5.8	15:49	41.53201	-82.61277	41.56912	-82.66157
21	2	8/24/14	4.5	18:25	41.77124	-83.11100	41.73364	-83.12988
23	2	8/28/14	11.0	20:29	41.79158	-83.22393	41.87998	-83.16398
24	7	8/18/14	19.6	13:39	41.64355	-82.4261	41.64171	-82.66182
26	8	8/21/14	8.3	14:38	41.54451	-82.5388	41.6194	-82.53375
27	7	8/18/14	7.8	13:44	41.64124	-82.72043	41.64541	-82.81355
28	6	8/19/14	5.0	14:15	41.63664	-82.72060	41.64354	-82.77937
30	6	8/19/14	5.2	17:47	41.59639	-82.81330	41.64295	-82.81724
32	4	8/18/14	10.8	20:53	41.80326	-83.31335	41.76710	-83.43442
34	4	8/18/14	11.9	20:21	41.73262	-83.27136	41.69067	-83.40374
35	4	8/18/14	8.8	15:11	41.72162	-83.32837	41.71043	-83.22338
41	5	8/19/14	27.1	17:28	41.83845	-83.05896	41.64809	-82.85441
42	5	8/19/14	11.3	17:22	41.68220	-83.22277	41.76506	-83.14341
44	5	8/21/14	19.3	14:50	41.69493	-82.83242	41.75650	-83.05016
48	9	8/28/14	3.8	19:16	41.47721	-82.73857	41.47657	-82.69286
49	9	8/28/14	2.8	19:11	41.47299	-82.76120	41.47056	-82.79522
50	9	8/28/14	4.0	19:20	41.48353	-82.72931	41.46095	-82.76728

Table 4: Station biogeochemistry - surface.

Station Units	Date	Area	Chl-a <i>ug/L</i>	SPM <i>g/L</i>	PC <i>ug/L</i>	apg490 <i>m⁻¹</i>	bbp 490 <i>m⁻¹</i>	bbp:bp 490 N/A	Zeaxanthin <i>ug/L</i>	Zea:Chl-a N/A	Kd490 <i>m⁻¹</i>
S11	8/20/14	5	22.0	7.4	6.6	0.458	0.132	0.035	N/A	N/A	1.17
S12	8/20/14	6	17.5	3.8	3.3	0.279	0.057	0.044	N/A	N/A	0.59
S13	8/20/14	6	22.7	5.1	3.9	0.371	0.073	0.040	N/A	N/A	0.76
S14	8/20/14	8	7.8	2.7	4.4	0.303	0.043	0.026	N/A	N/A	0.55
S15	8/20/14	6	19.9	6.6	6.8	0.473	0.118	0.029	N/A	N/A	1.09
S17	8/21/14	2	5.8	1.8	3.1	0.201	0.028	0.026	N/A	N/A	0.39
S18	8/21/14	1	41.1	12.6	74.8	0.493	0.223	0.045	N/A	N/A	1.55
S19	8/21/14	1	60.5	18.9	156.9	0.621	0.276	0.039	N/A	N/A	1.91
S20	8/21/14	1	106.3	22.7	213.3	0.500	0.195	0.040	N/A	N/A	1.44
LE3	8/18/14	2	6.1	N/A	N/A	0.251	0.250	0.028	0.301	0.049	0.51
LE4	8/19/14	1	124.3	N/A	N/A	1.024	0.347	0.035	3.248	0.026	2.72
LE5	8/19/14	1	133.8	N/A	N/A	1.301	0.350	0.017	6.203	0.046	3.14
LE7	8/20/14	5	23.9	N/A	N/A	0.490	0.136	0.036	1.223	0.051	1.16
LE8	8/20/14	1	52.9	N/A	N/A	0.909	0.332	0.038	3.135	0.059	2.74
LE9	8/21/14	8	12.1	N/A	N/A	0.413	0.067	0.030	0.922	0.076	0.80
LE10	8/21/14	8	22.7	N/A	N/A	1.349	0.088	0.023	1.550	0.068	2.17
LE11	8/22/14	8	38.5	N/A	N/A	0.708	0.060	0.038	2.904	0.076	1.11
LE12	8/22/14	8	16.0	N/A	N/A	0.941	0.097	0.053	1.434	0.090	1.83
LE14	8/23/14	8	10.2	N/A	N/A	0.422	0.049	0.029	0.776	0.076	0.82
LE15	8/23/14	8	13.7	N/A	N/A	0.528	0.054	0.021	1.086	0.080	1.01

Table 5: LIDAR - Discrete Station Matchups

Station Units	Track ID	Nearest Track Lat (degrees)	Nearest Track Lon (degrees)	Distance (km)	Time difference (days)	R
S11	T41	41.694	-82.904	3.6	1.3	-0.13
S11	T44	41.707	-82.881	1.4	1.0	0.30
S12	T27	41.641	-82.748	3.1	2.0	0.78
S13	T19	41.577	-82.669	5.0	3.0	0.45
S14	T26	41.547	-82.537	1.7	0.9	0.64
S17	T23	41.827	-83.197	0.1	7.3	0.06
S18	T2	41.673	-83.003	2.0	2.0	0.85
S19	T10	41.644	-83.039	0.1	3.8	0.96
S20	T3	41.675	-83.114	2.0	3.1	0.89
Average				1.86	2.7	0.53

Figures

Figure captions

Figure 1. Top: Map of LIDAR flight tracks in August, 2014, color coded by LIDAR attenuation coefficient in m^{-1} , superimposed on the MODIS true-color image from August 25. Sub-region identification scheme for LIDAR track data: 1 - southeastern western basin; 2 - Detroit River plume front; 3 - Detroit River; 4 - Maumee Bay; 5 - islands west; 6 - islands central; 7 - islands east; 8 - islands southeast; 9 - Sandusky Bay. Bottom: locations of discrete stations. Station legend: diamonds - UNH data; circles - NRL data; brown - Holocam, IOPs, SPM, R_{rs} ; red - IOP, SPM blue - R_{rs} , IOP, HPLC; green - HPLC only; yellow squares: wind stations- Toledo Light 2 (TOL2), NOS Toledo (THRO), South Bass Island (SBIO) and Marblehead (MRHO).

Figure 2. HOLOCAM images from Station *S14*. A: sample raw hologram with *Planktothrix*; B: Background-subtracted hologram; C: Post -recon image with all particles in-focus; and D: Isolated *Planktothrix* chains after segmentation and thresholding.

Figure 3. Wind speed measurements from 4 NOAA meteorological stations positioned around the study area (see Fig 1). A: Toledo Light 2 (TOL2); B: South Bass Island (SBIO); C: Marblehead (MRHO); D: NOS Toledo (THRO). Grey area is period of LIDAR observations; grey vertical dashed lines indicate days with IOP profiles.

Figure 4. Discrete surface SPM (top left) and *Chl-a* (bottom left) across the study area. Size of circles in top left proportional to organic fraction of

SPM; Size of circles in bottom left proportional to ratio of particle backscatter to total particle scatter at 443 nm. Right panels show HOLOCAM frame shots from three different areas - A: Area 3 (Detroit River plume; B: Area 1; C: Area 8.

Figure 5. Echograms from LIDAR tracks and matched holographic profiles in Area 1. A: RGB image from Landsat-8 showing selected LIDAR transect lines and stations *S18*, *S19* and *S20* (sampled on August 21); B: Track *T11* from August 23, 2014; C: Track *T10* from August 17, 2014; D: Track *T3* from August 18, 2014; E: Track *T2* from August 19, 2014; Note: y-axis for all echograms extends from 8 m depth to 2 m above the surface; x-axis for all echograms are relative along-track distances scaled to actual distances contained in Table 3. From left to right, track orientations are south to north (Track *T11*) or west to east (Tracks *T10*, *T3* and *T2*). Color scale has a range of 0-7.8 Am^2 proportional to particle concentration; F, G, H: cell count totals of *Planktothrix* and *Microcystis* versus depth (red), along with the depth profile of LIDAR return signal strength (blue) from nearest track and location for stations *S18*, *S19* and *S20*. Note: horizontal scales for cell counts (top of plot) and LIDAR (bottom of plot) are on absolute scales for comparison. Further track details contained in Table 3.

Figure 6. Holographic and IOP vertical profile data are highlighted for stations *S18*, *S19* and *S20* in Area 1 sampled on Aug. 21, 2014, and IOP-only profiles for stations *LE4*, *LE5* (sampled on Aug. 19, 2014) and *LE8* (sampled on Aug. 20, 2014). A: *Microcystis* counts; B: *Planktothrix* counts;

1367 C: Cell count ratio normalized to column integrated sum for *Planktothrix*
1368 and *Microcystis*; D: water temperature; E: particle backscatter at 443 nm;
1369 F: particle backscatter ratio at 443 nm; F: particle absorption at 443 nm; H:
1370 above-water R_{rs} . NOTE: grey lines in plots A through H indicate profiles of
1371 all other stations not in the area for comparative visualization.

1372 Figure 7. Echograms from Areas 2 and 5 flown across particle heavy
1373 southern areas to particle-free northern areas of western basin. A: RGB im-
1374 age from MODIS-Aqua showing selected LIDAR transect lines and station
1375 locations; B, C and D: Tracks T_{21} , T_{23} and T_{42} in a south to north orienta-
1376 tion (left to right); E and F: Tracks T_{41} and T_{44} in a west to east orientation
1377 (left to right); G and H: cell count totals of *Planktothrix* and *Microcystis* ver-
1378 sus depth (red), along with the depth profile of LIDAR return signal strength
1379 (blue) from nearest track and location for stations S_{11} and S_{17} , respectively.
1380 Station S_{11} was also matched to a second track - T_{41} (yellow line). Note:
1381 horizontal scales for cell counts (top of plot) and LIDAR (bottom of plot)
1382 are on absolute scales for comparison; echogram color scale same as Figure
1383 5. Further track details contained in Table 3.

1384 Figure 8. Holographic and IOP vertical profile data for stations S_{11} and
1385 S_{17} are highlighted. IOP-only profiles for $LE3$ and $LE7$ are included. A:
1386 *Microcystis* counts; B: *Planktothrix* counts; C: Cell count ratio normalized
1387 to column integrated sum for *Planktothrix* and *Microcystis*; D: water tem-
1388 perature; E: particle backscatter at 443 nm; F: particle backscatter ratio at
1389 443 nm; F: particle absorption at 443 nm; H: above-water R_{rs} . NOTE: gray

1390 lines in plots A through H indicate profiles of all other stations not in the
1391 area for comparative visualization.

1392 Figure 9. Echograms from tracks in Area 6. A: RGB image from MODIS-
1393 Aqua with selected LIDAR transect lines; B, C and D: Tracks *T15*, *T16* and
1394 *T19* in a south to north orientation (left to right); E, F and G: Tracks *T17*,
1395 *T18* and *T27* in a west to east orientation (left to right); H and I: cell count
1396 totals of *Planktothrix* and *Microcystis* versus depth (red), along with the
1397 depth profile of LIDAR return signal strength (blue) from nearest track and
1398 location for stations *S12* and *S13*. Note: horizontal scales for cell counts (top
1399 of plot) and LIDAR (bottom of plot) are on absolute scales for comparison;
1400 echogram color scale same as Figure 5; dashed lines indicate tracks continuing
1401 off map. Further track details contained in Table 3.

1402 Figure 10. Holographic and IOP vertical profile data are highlighted for
1403 stations *S12*, *S13* and *S15* in the island area (Area 6) sampled on August
1404 20, 2014. A: *Microcystis* counts; B: *Planktothrix* counts; C: Cell count ratio
1405 normalized to column integrated sum for *Planktothrix* and *Microcystis*; D:
1406 water temperature; E: particle backscatter at 443 nm; F: particle backscatter
1407 ratio at 443 nm; G: particle absorption at 443 nm; H: above-water R_{rs} .
1408 NOTE: gray lines in plots A through H indicate profiles of all other stations
1409 not in the area for comparative visualization.

1410 Figure 11. Echograms from tracks in Area 8. A: RGB image from
1411 Landsat-8 with selected LIDAR transect lines; B and C: Tracks *T24* and
1412 *T28* in a west to east orientation (left to right); D, E, F and G: Tracks *T1*,

1413 *T26*, *T8* and *T30* in a south to north orientation (left to right); H: cell count
1414 totals of *Planktothrix* and *Microcystis* versus depth (red), along with the
1415 depth profile of LIDAR return signal strength (blue) from nearest track and
1416 location for station *S14*. Note: horizontal scales for cell counts (top of plot)
1417 and LIDAR (bottom of plot) are on absolute scales for comparison; echogram
1418 color scale same as Figure 5; yellow: echograms for these tracks are contained
1419 in Figure 9; dashed lines indicate tracks continuing off map. Further track
1420 details contained in Table 3;

1421 Figure 12. Holographic and IOP vertical profile data are highlighted for
1422 stations *S14* taken on August 19, 2014 in Area 8. IOP-only profiles for *LE9*
1423 and *LE10* (August 21, 2014), *LE11* and *LE12* (August 22, 2014), and *LE14*
1424 and *LE15* (August 22, 2014). A: *Microcystis* counts; B: *Planktothrix* counts;
1425 C: Cell count ratio normalized to column integrated sum for *Planktothrix*
1426 and *Microcystis*; D: water temperature; E: particle backscatter at 443 nm;
1427 F: particle backscatter ratio at 443 nm; F: particle absorption at 443 nm; H:
1428 above-water R_{rs} . NOTE: gray lines in plots A through H indicate profiles of
1429 all other stations not in the area for comparative visualization.

1430 Figure 13. A: Map of *Chl-a* (circle size) color-coded by the attenuation co-
1431 efficient at 490 nm (K_d490); B: Map of zeaxanthin (circle size) color-coded by
1432 the ratio of zeaxanthin to *Chl-a*; C: depth of *Planktothrix* (blue) and *Micro-*
1433 *cystis* (red) maximum cell count versus optical depth, line fit to *Planktothrix*
1434 only; D: column-integrated cell counts for *Microcystis* and *Planktothrix* ver-
1435 sus surface cell count. Stations *S11*, *S12* and *S14* showed weak *Microcystis*

1436 vertical structure, and positions on graph may not be highly accurate.

1437 Figure 14. Optical depth at the 10% light level (Z_{10}) versus LIDAR
1438 penetration depth (LPD) from stations and nearest LIDAR track point shown
1439 in Table 5. A quadratic curve was fitted to the data points (red).

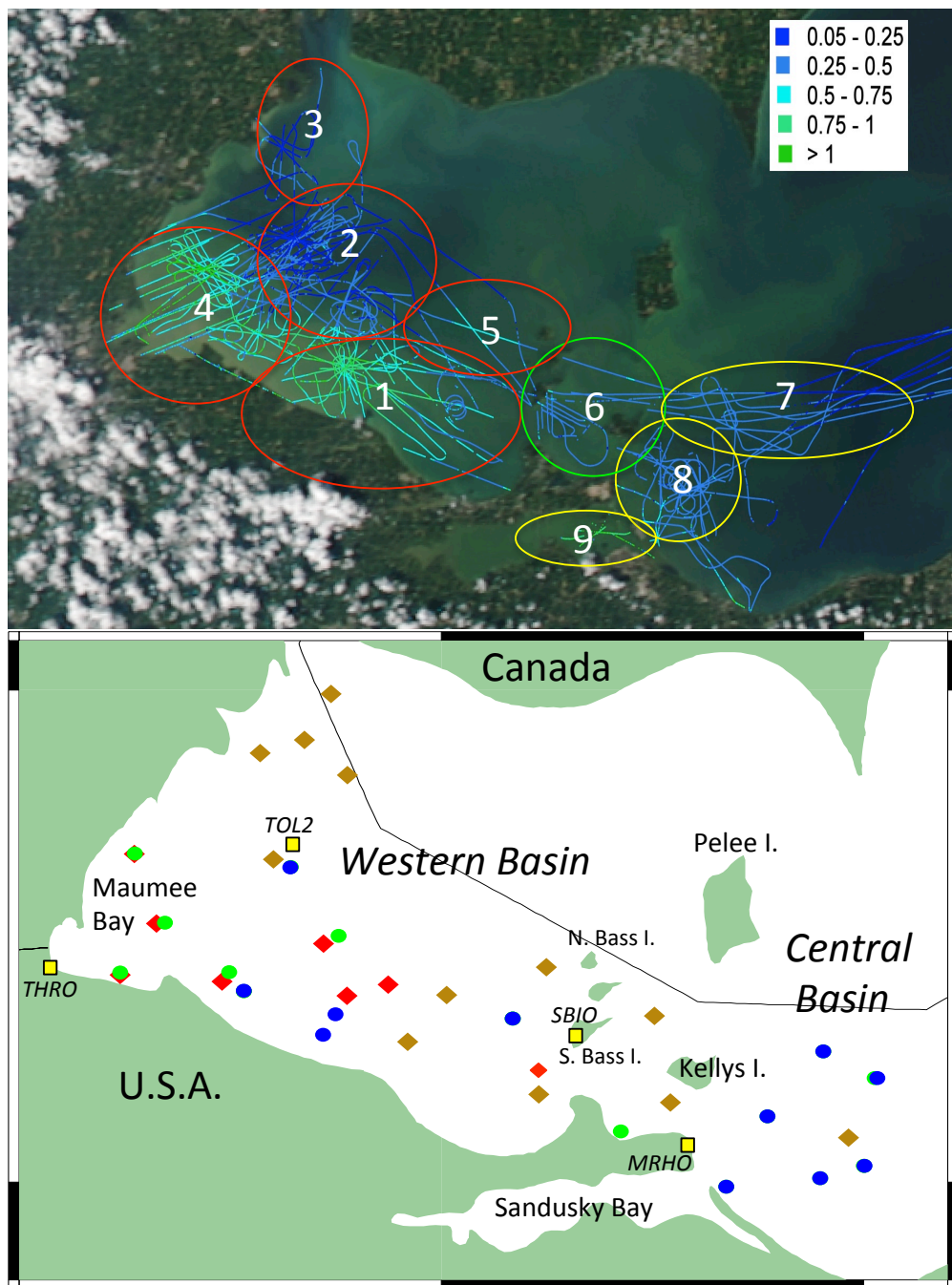


Figure 1

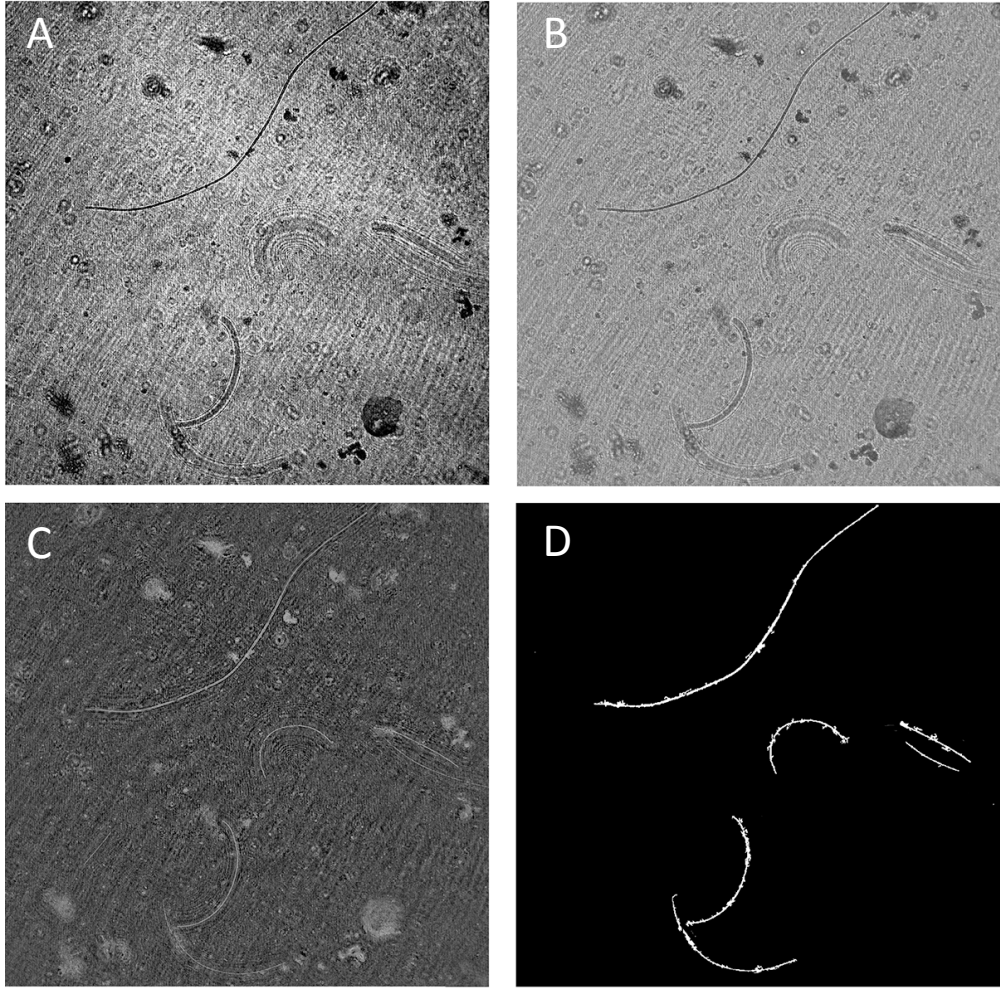


Figure 2

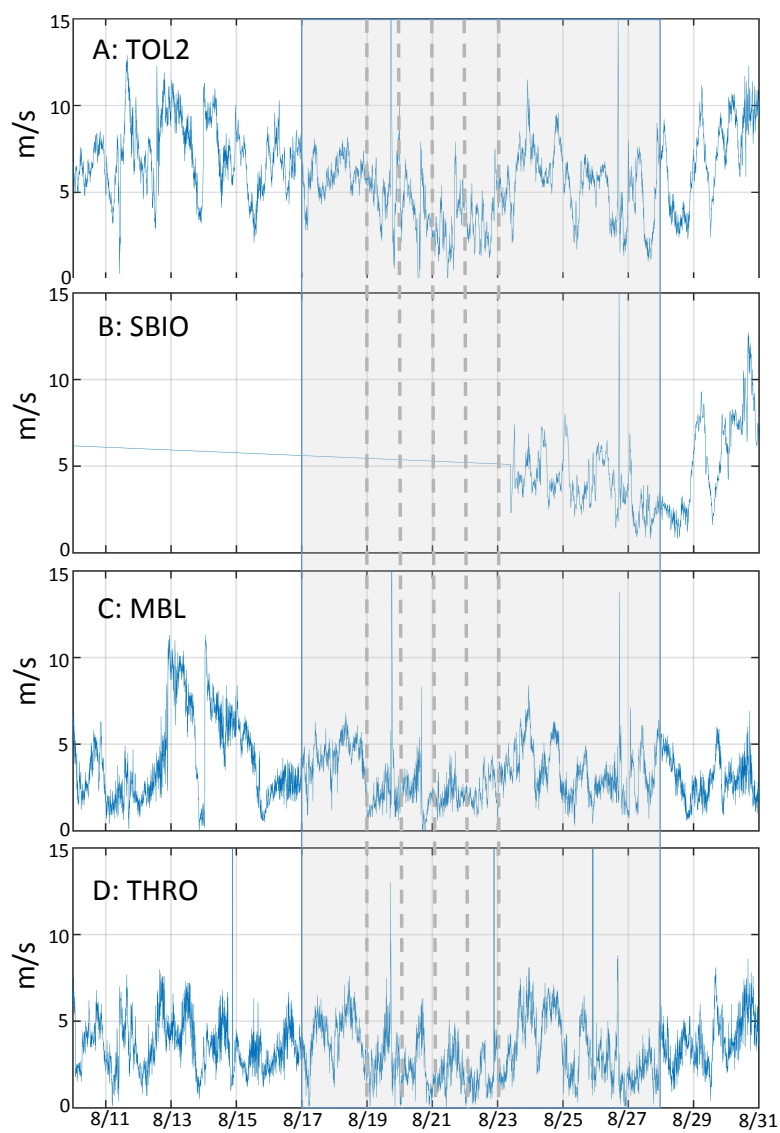


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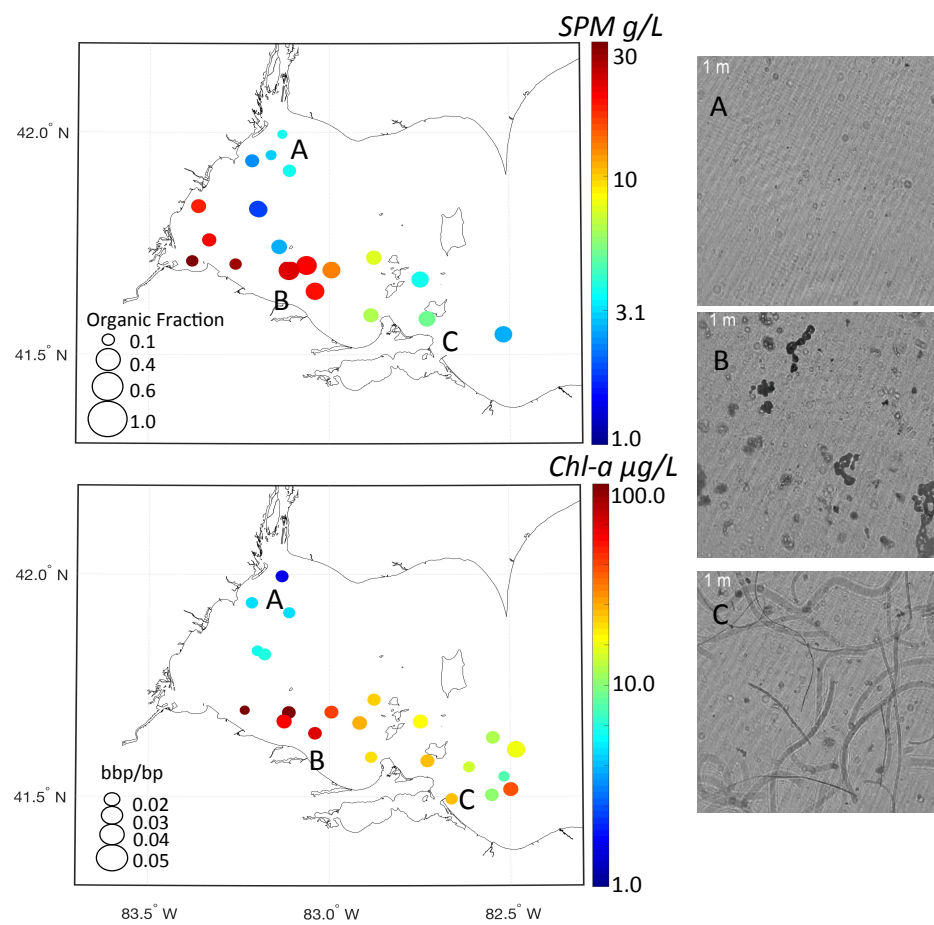


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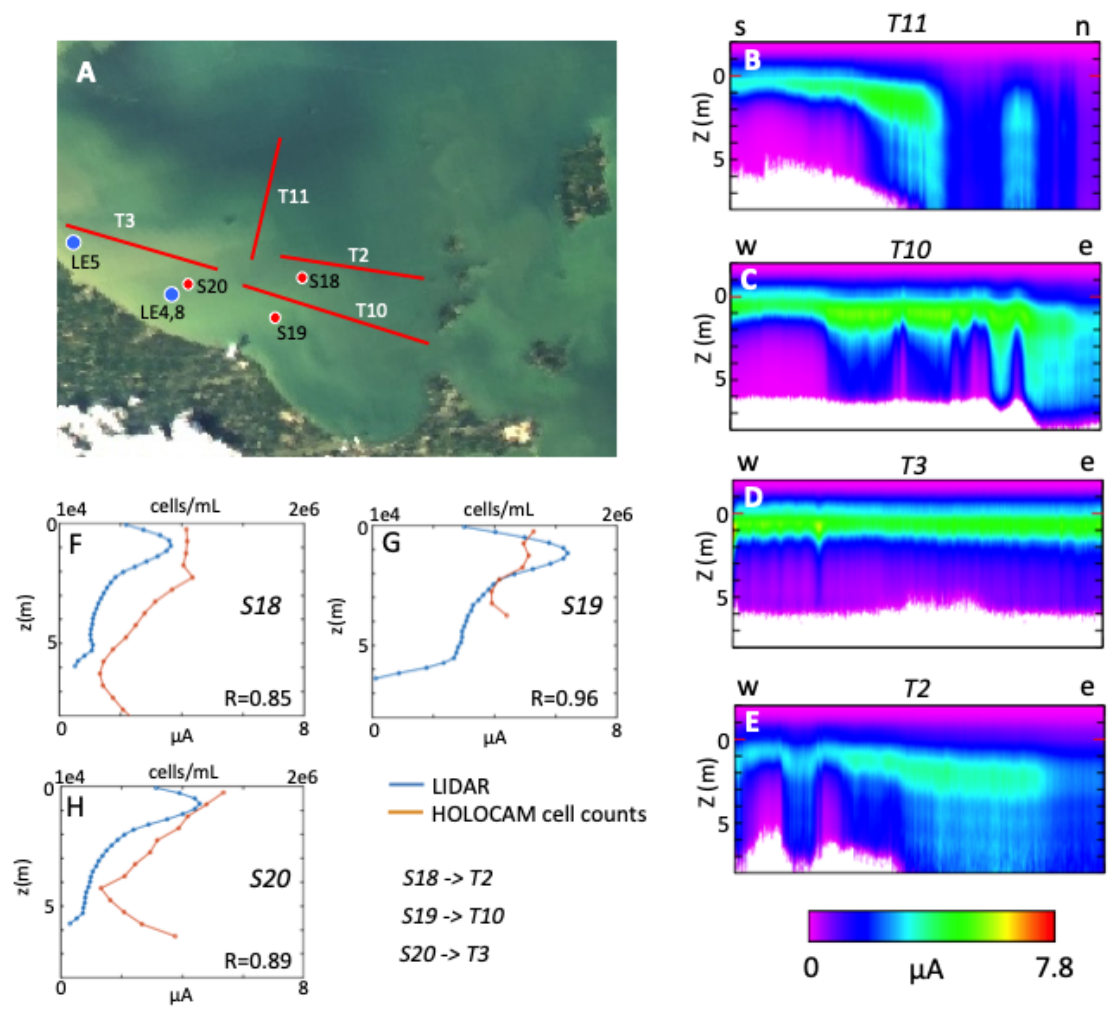


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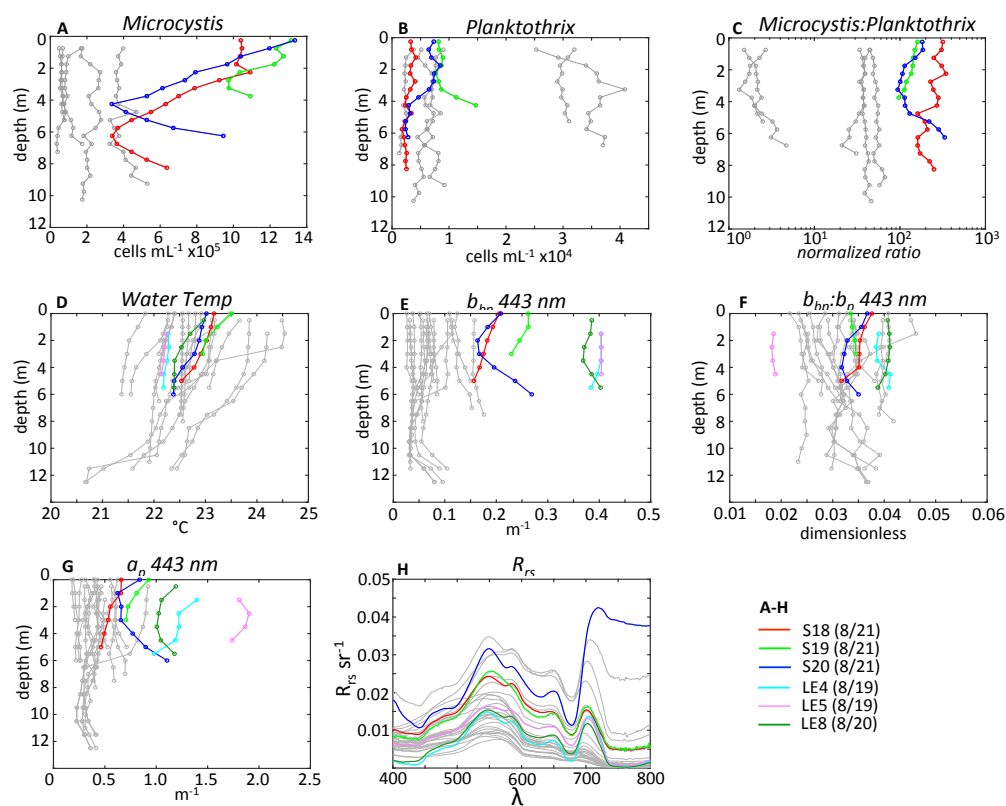


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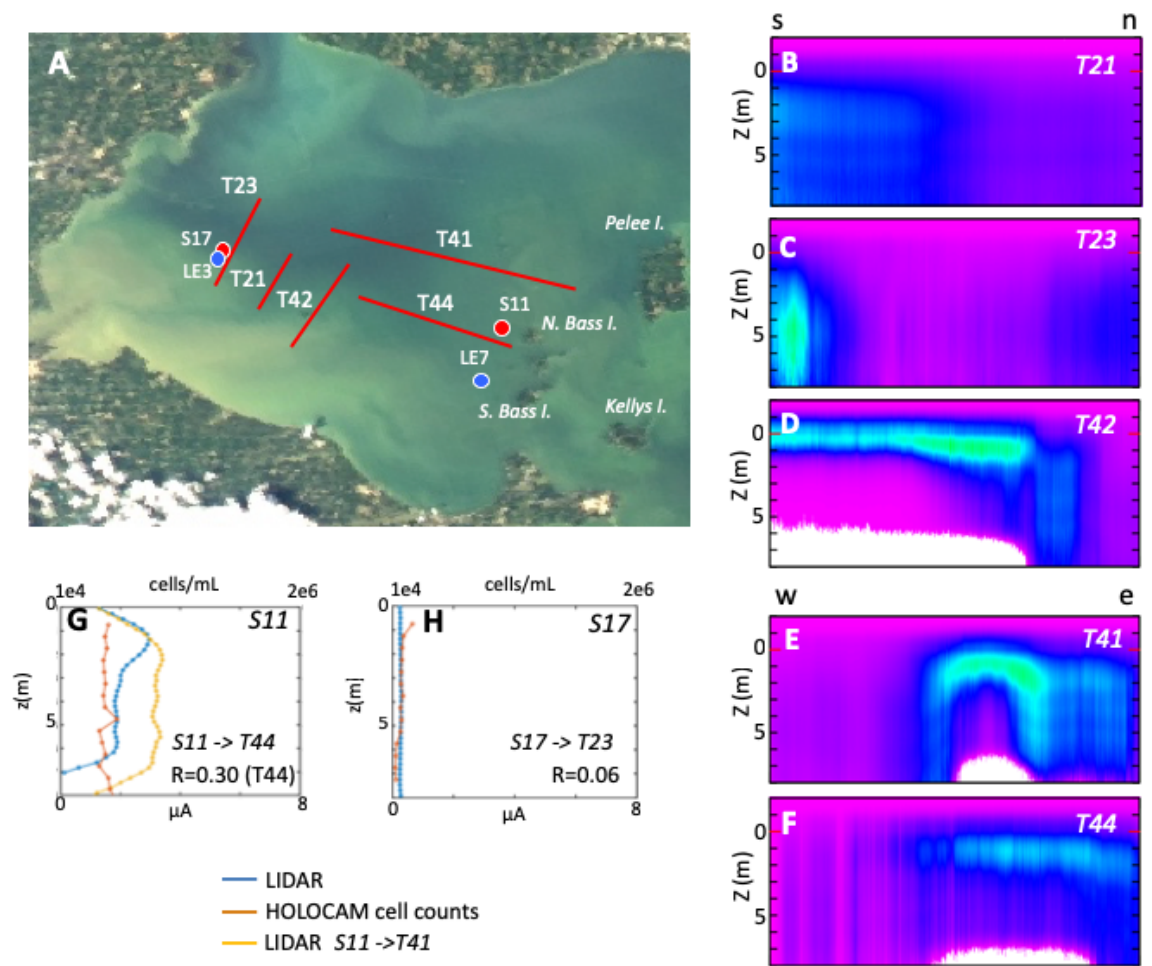


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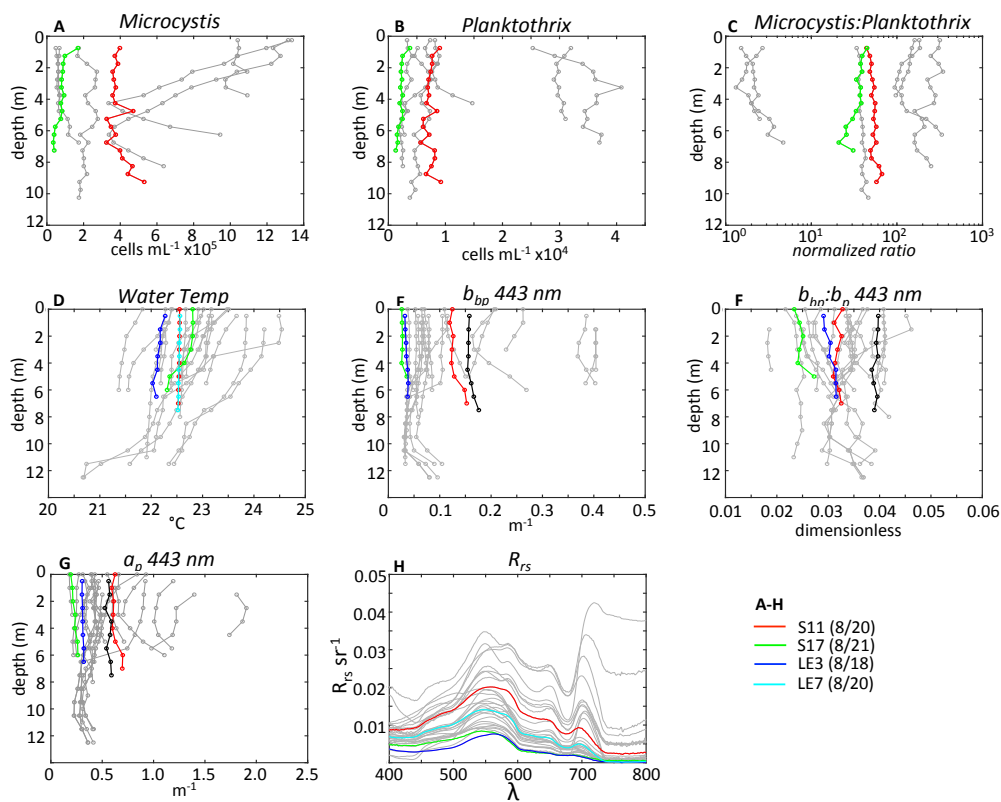


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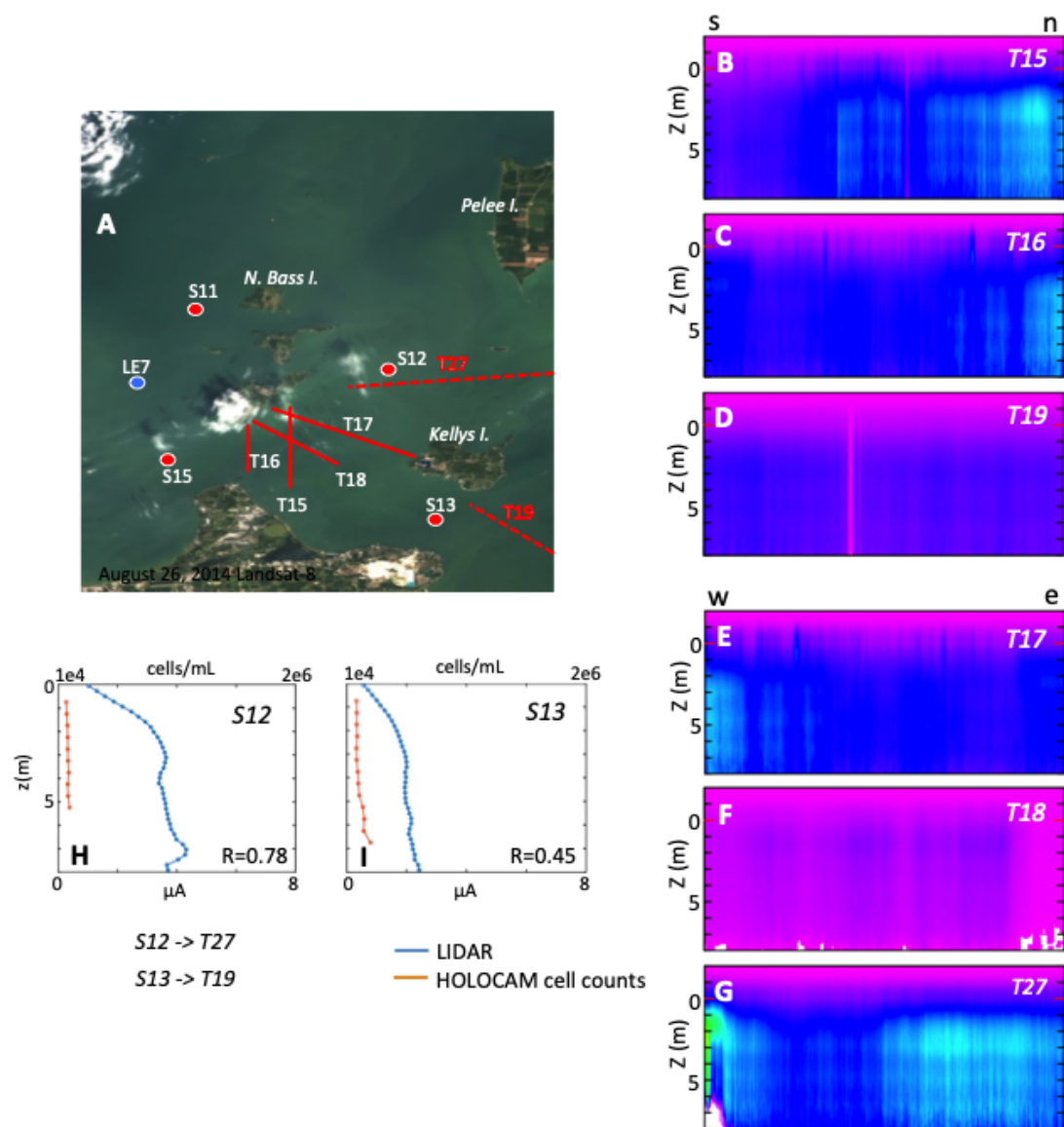


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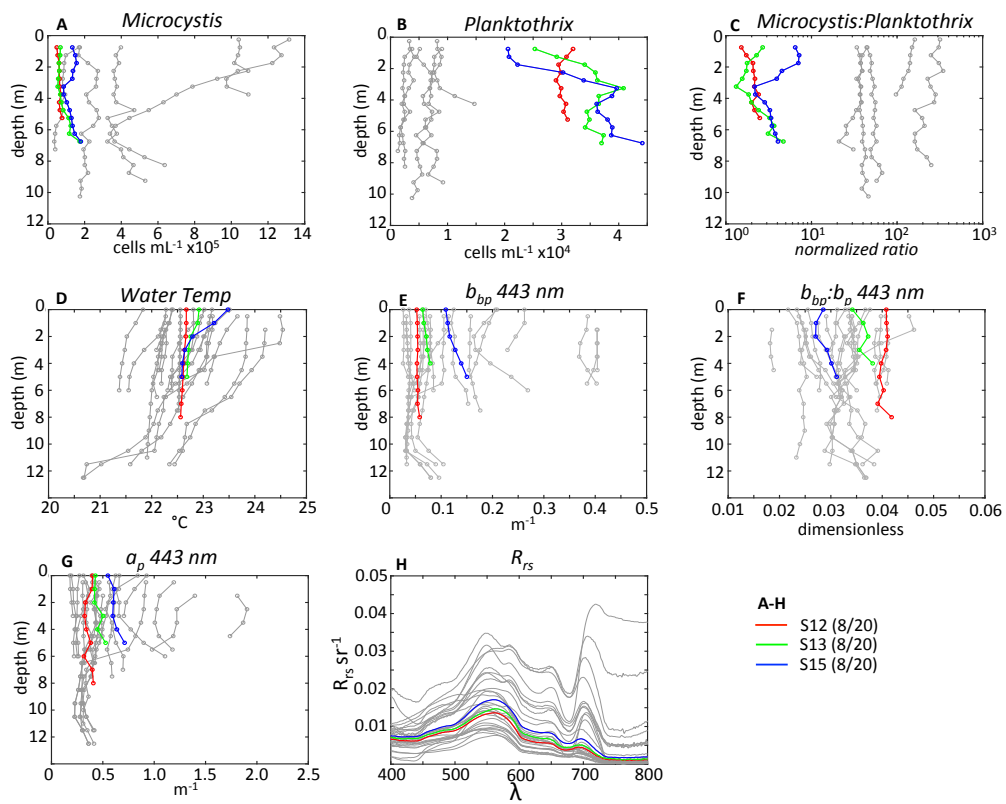


Figure 10

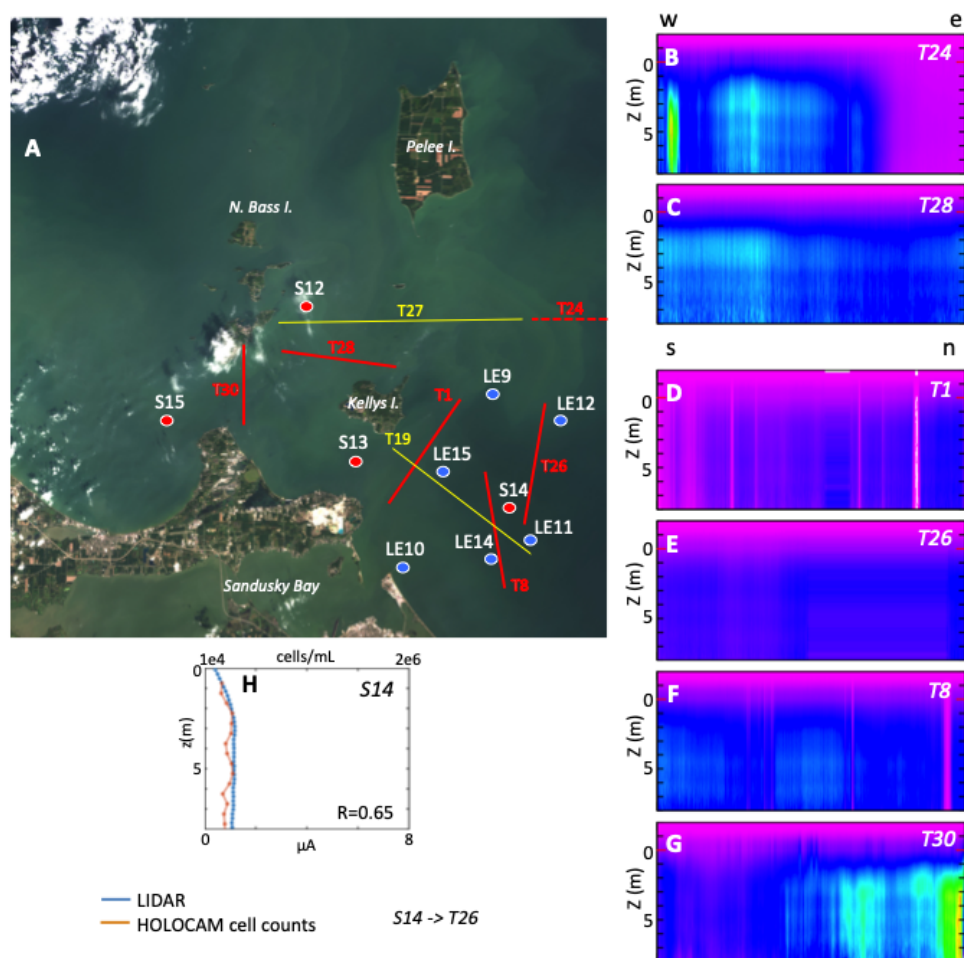


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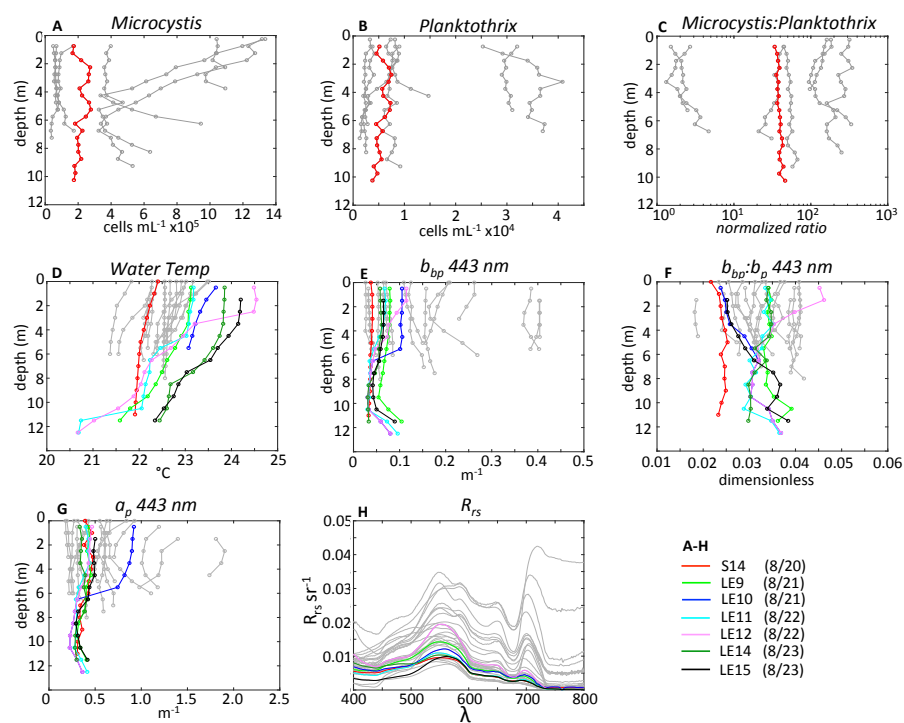


Figure 12

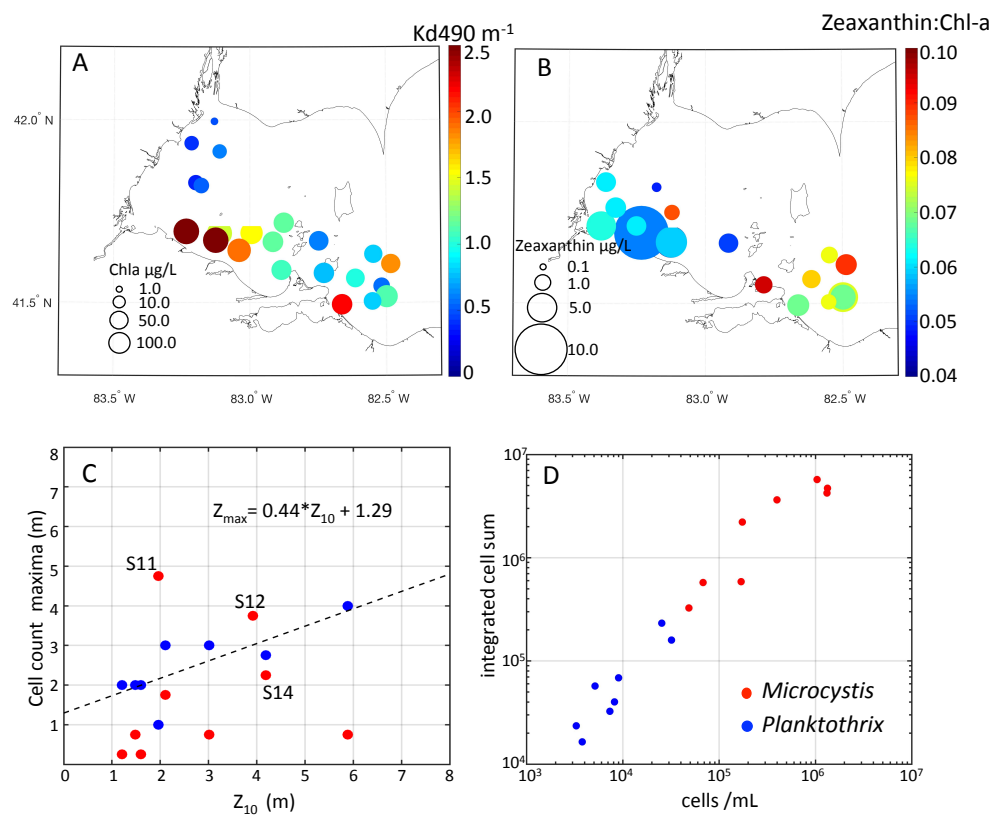


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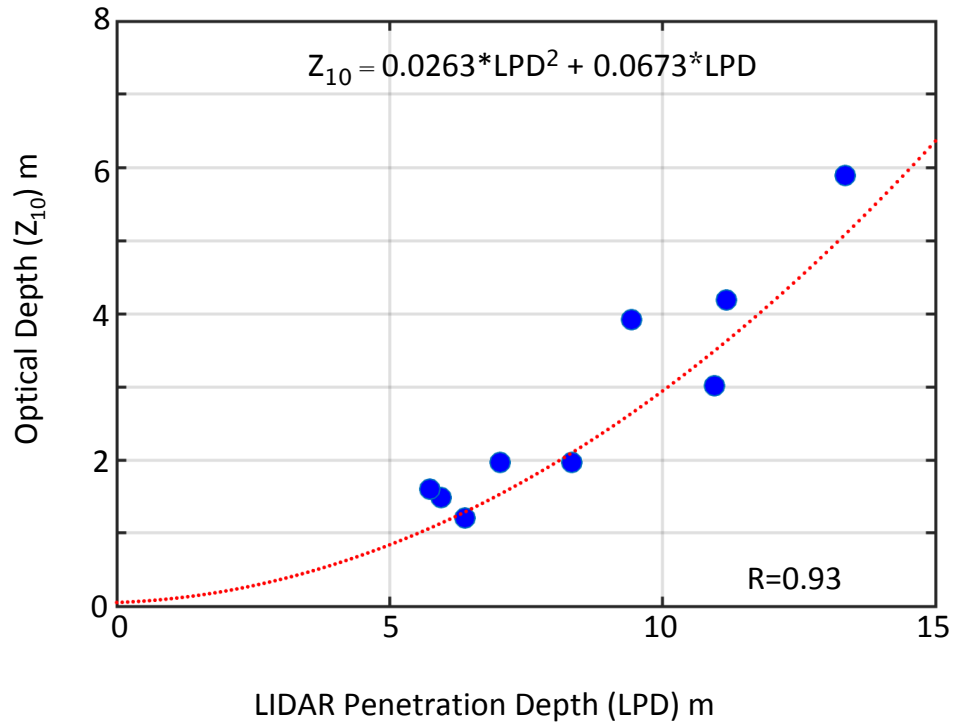


Figure 14

1440 Supplement

1441 5.1. Area 4: Maumee Bay

1442 Maumee Bay, located in the southwestern corner of Lake Erie, has histor-
 1443 ically been viewed as a HAB source region for western Lake Erie (Bridgeman
 1444 et al., 2013). It typically contains high amounts of cyanobacteria biomass
 1445 in summertime and, due to its shallowness, a high amount of resuspended
 1446 particles. Unfortunately, we did not sample this region with the HOLOCAM

1447 or the MASCOT in 2014 (see Moore et al. (2017) for measurements from
 1448 2013). However, LIDAR tracks and surface water samples were taken on
 1449 the same day on August 18, 2014. Discrete measurements of *Chl-a* indi-
 1450 cated high biomass in this area (Figure 4). Echograms from several track
 1451 lines showed near-surface concentrations of particles in all tracks (Figure
 1452 15). The two transects perpendicular to the shoreline (tracks *T32* and *T34*)
 1453 show a decrease in surface particles towards the offshore (eastern) end of the
 1454 tracks. This agrees with the features in the satellite image from the same day,
 1455 where the track ends extended past the turbid zone into the clearer waters.
 1456 Particle distributions from the alongshore track (*T35*) were concentrated in
 1457 a continuous near-surface particle layer. The particles in this region were
 1458 likely dominated by *Microcystis* colonies/cells mixed with suspended inor-
 1459 ganic particles.

1460 5.2. Area 9: Sandusky Bay

1461 The LIDAR transects in Sandusky Bay (Area 9) were flown on August
 1462 28, 2014 and several days after the *in situ* field sampling. The transects
 1463 were flown over the eastern half of the bay towards the mouth (Figure 16).
 1464 Particles were concentrated near the surface in thin layers above the shallow
 1465 bottom. The echogram from track *T48*, near the mouth of the bay/open
 1466 water transition, shows a gradual expansion of the surface layer towards the
 1467 bottom traversing west to east. It is unclear if this particle distribution is the
 1468 result of a mixing of particles from the surface layer, descending cells from

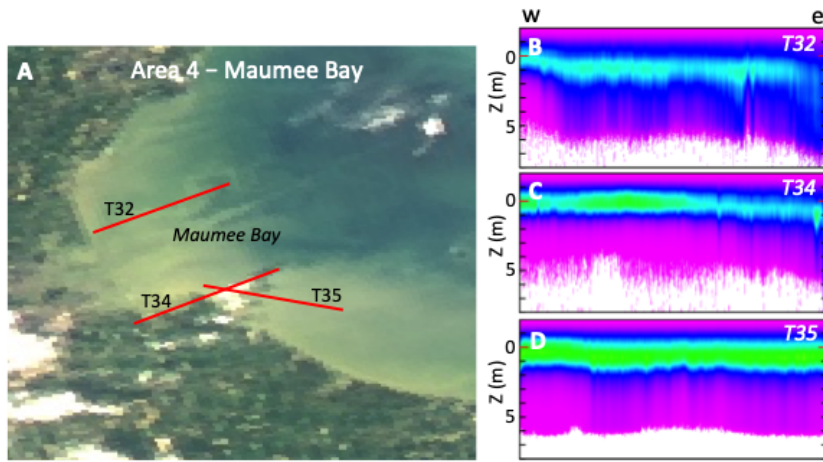


Figure 15: Echograms from 3 tracks in Maumee Bay (Area 4) on August 18, 2014. A: RGB image from Modis-Aqua showing selected LIDAR transect lines; B: Track *T32*; C: Track *T34*; D: Track *T35*. From left to right, all tracks are west to east orientation. Echogram color scale same as Figure 5. Track details contained in Table 3.

vertical migration, from bottom re-suspension or some combination. Overall, this vertical particle structure is more similar to Maumee Bay (Area 4) than to the adjacent, deeper waters of the central basin (Area 8).

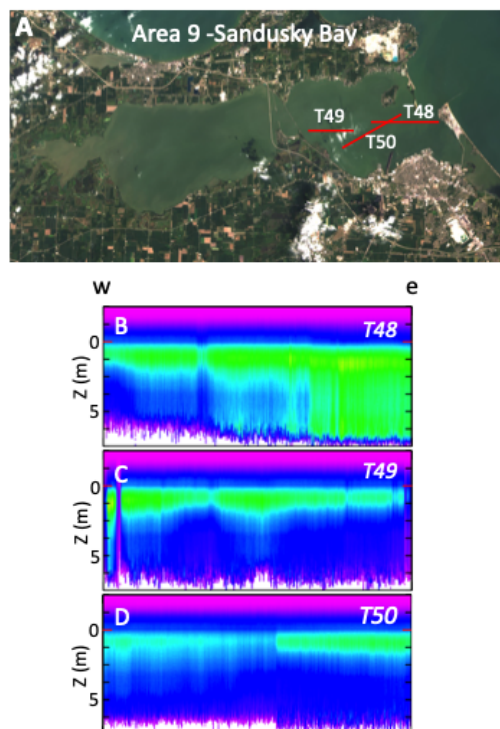


Figure 16: Echograms from LIDAR tracks in Area 9 in Sandusky Bay on August 28, 2014. A: RGB image from Modis-Aqua showing selected LIDAR transect lines; B: Track *T48*; C: Track *T49*; D: Track *T50*. From left to right, track orientations are west to east. Echogram color scale same as Figure 5. Track details contained in Table 3.