

1 **Good practices, trade-offs, and precautions for model diagnostics in integrated stock**  
2 **assessments**

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10

11 **Abstract**

12 Carvalho et al. (2021) provided a “cookbook” for implementing contemporary model  
13 diagnostics, which included convergence checks, examinations of fits to data, retrospective and  
14 hindcasting analyses, likelihood profiling, and model-free validation. However, it remains  
15 unclear whether these widely-used diagnostics exhibit consistent behavior in the presence of  
16 model misspecification, and whether there are trade-offs in diagnostic performance that the  
17 assessment community should consider. This illustrative study uses a statistical catch-at-age  
18 simulation framework to compare diagnostic performance across a spectrum of correctly  
19 specified and mis-specified assessment models that incorporate compositional, survey, and catch  
20 data. Results are used to contextualize how reliably common diagnostic tests perform given the  
21 degree and nature of known model issues, including parameter and model process  
22 misspecification, and combinations thereof, and trade-offs among model fits, prediction skill,  
23 and retrospective bias that analysts must consider as they evaluate diagnostic performance. A  
24 surprising number of mis-specified models were able to pass certain diagnostic tests, although  
25 there was a trend of more frequent failure with increased mis-specification for most diagnostic  
26 tests. Nearly all models that failed multiple tests were mis-specified, indicating the value of  
27 examining multiple diagnostics during model evaluation. Diagnostic performance was best (most  
28 sensitive) when recruitment variability was low and historical exploitation rates were high, likely  
29 due to the induction of better contrast in the data, particularly indices of abundance, under this  
30 scenario. These results suggest caution when using standalone diagnostic results as the basis for  
31 selecting a “best” assessment model, a set of models to include within an ensemble, or to inform  
32 model weighting. The discussion advises stock assessors to consider the interplay across multiple  
33 dynamics. Future work should evaluate how the resolution of the production function, quality  
34 and quantity of data time series, and exploitation history can influence diagnostic performance.

35 **1. Introduction**

36 Sustainable exploitation of renewable natural resources requires quantitative scientific guidance.  
37 Integrated population dynamics models (e.g., Fournier and Archibald (1982)) have flourished as  
38 the tool of choice to evaluate the status and possible future outcomes for exploited, threatened or  
39 managed populations (Maunder and Punt, 2014; Tempel et al., 2014). Integrated population  
40 models use mathematical relationships (processes) to specify how changes in population  
41 abundance occur and to link model predictions to data (observations). The processes are  
42 themselves governed by parameters that can either be estimated during the model fitting process  
43 or pre-specified based on independent studies. With the increase of computing power and the  
44 popularization of integrated stock assessment modeling (Maunder et al., 2009), the complexity of  
45 modern stock assessment modeling for fisheries management has increased. Analysts have  
46 multiple ways to model the observed data and underlying population processes when conducting  
47 a stock assessment, yet the tradeoffs among modeling choices are not always obvious.  
48 Developing and presenting multiple candidate models has become commonplace for many  
49 national and international fishery management agencies (Karp et al., 2022). Sometimes, analysts  
50 and/or review bodies must choose among candidate models, or systematically filter and combine  
51 models in an ensemble (Jardim et al., 2021).

52  
53 The complexity of data types and model structures available to fishery stock assessment, and the  
54 desire for objective means of evaluating multiple candidate models, have led to a growing  
55 interest in diagnostic tests. Diagnostic tests can serve an important role in model validation, a  
56 crucial step in the assessment process that establishes the credibility and robustness of the advice  
57 that proceeds from an assessment model. Model validation communicates confidence in model  
58 outputs to stakeholders not directly involved in model construction, though validation of key  
59 derived quantities (current biomass and fishing mortality, for example) is not possible, as these  
60 values cannot be directly observed. Therefore, diagnostic tests used for model validation  
61 typically focus on evaluating how well the model fits to the observed data, whether the model  
62 meets its statistical assumptions and seems ecologically plausible, and the robustness of the  
63 model to new or removed data. These tests remain an integral component to the evaluation and  
64 provision of management advice, and are often required in assessment reports.

65  
66 A large number of studies over the last decade have specifically investigated the reliability of  
67 such tests, particularly in the context of model mis-specification. Previous work in model  
68 diagnostics has evaluated only a single source of mis-specification in relatively simple models  
69 (e.g., Carvalho et al. (2017); Piner et al. (2011)). Model specification decisions are related to the  
70 functional form of the process, the variables they depend on (e.g., age or length, Lee et al.  
71 (2019)) and potential temporal variability in those processes. Inappropriate specification of a  
72 population dynamics model can occur in the observation (data) and/or population processes  
73 (Maunder and Piner, 2017). **These can include spatial variability, local depletion, movement  
74 dynamics, and the precision and accuracy of data inputs, among others.** In contrast to  
75 population processes, the parameters of the observation model are nearly always estimated  
76 because they address sampling uncertainties that are largely unknowable without an estimate of  
77 the population dynamics. Mis-specification occurs when a process is assumed to be governed by  
78 the wrong functional form, a parameter is set to the wrong value, or a process is modelled such  
79 that temporal variability is not correctly accounted for (or even ignored). Mis-specification of the  
80 population processes can lead to biased estimates of the parameters and hence quantities of

81 management interest while mis-specification of observation processes can lead to the data not  
82 providing the correct information about the estimated parameters (e.g., Langseth et al., 2016;  
83 Maunder et al., 2023). Moreover, mis-specification in one process can lead to poor fits to data  
84 directly linked to that process and to data indirectly linked via the population dynamics because  
85 all data and model processes are linked through the population dynamics equations (Lee et al.,  
86 2019; Taylor et al., 2013). The linkage of all data via the population dynamics equations is the  
87 strength of integrated modelling, but this strength also makes locating mis-specifications  
88 challenging. Almost certainly, diagnosing and correcting model mis-specification becomes more  
89 difficult when multiple processes are simultaneously mis-specified.

90  
91 Carvalho et al. (2017) presented a simulation-based exploration into how popular diagnostic tests  
92 respond to standalone misspecifications for a singular stock. The chief finding of that study was  
93 that the examined diagnostic tests are not equally reliable at detecting model mis-specification  
94 (Carvalho et al., 2017). Residual analyses appeared best at detecting mis-specification of the  
95 observation model, while only the age-structured production model (ASPM, Maunder and Piner,  
96 2015) could detect a mis-specification of the system dynamics (Carvalho et al., 2017). No single  
97 diagnostic could reliably identify the process of a given misspecification for complex models  
98 (such as those with many fisheries and/or data types). The key limitations of that work were that  
99 the simulations 1) did not consider varied life history strategies, particularly those that result in  
100 highly stochastic population trajectories, 2) did not consider varied levels of fishing mortality,  
101 which can impact the degree of contrast in simulated or real datasets and therefore affect  
102 parameter estimability (Magnusson et al 2007), and 3) model misspecifications were explored  
103 individually, so synergistic effects of misspecifying multiple parameters and/or processes on  
104 diagnostic performance remain unexplored. Finally, the results were not contextualized alongside  
105 the relative error in management quantities, so it was impossible to compare the presence and  
106 degree of bias in the models that failed diagnostic tests, versus those that did not. These caveats  
107 are especially important given that diagnostic tests are employed across a diverse range of  
108 stocks, particularly those managed by international organizations (Karp et al., 2022), whose life  
109 and exploitation histories may vary considerably from the initial study, and the fact that most  
110 stock assessors manipulate several model components at once (Maunder and Punt 2014),  
111 presenting many opportunities for misspecification to be introduced or eliminated. All of these  
112 limitations are revisited in the present study.

113  
114 Carvalho et al. (2021) provided a "cookbook" for implementing modern model diagnostics and  
115 model performance evaluation. That work suggested that a model would be considered adequate  
116 for management advice if the model a) optimizes successfully, b) fits to the data (e.g., passes a  
117 residual analysis), c) provides reliable estimates of population trends and scale, d) produces  
118 consistent results when provided new data, or if data are removed (e.g., retrospective analysis),  
119 and e) can make adequate future predictions (e.g., hindcasting). An ideal situation would allow  
120 for the suite of diagnostic tests presented in Carvalho et al. (2021) to be used to select among  
121 candidate models, or to evaluate or weight a set of candidate models for inclusion in an ensemble  
122 (aka Jardim et al., 2021). Such a one-size-fits-all approach is not yet realized, particularly since  
123 real-world applications of the cookbook have encountered tradeoffs between passing all or most  
124 diagnostics. Implementing the cookbook has also become complicated because some diagnostics  
125 do not have clear thresholds for passing, or if they do, the applicability of such thresholds to a  
126 diversity of stocks has not been rigorously tested. For example, since the publication of

127 threshold-like values for  $\rho$  (Hurtado-Ferro et al., 2012), many management agencies have  
128 made a practice of selecting among management models based on whether they fall among the  
129 cutoffs presented in that paper (Carvalho et al., 2021; Merino et al. 2022). Recent work has  
130 shown that these cutoffs should not be considered universal (e.g., Breivik et al. 2023), and  
131 proposed alternative approaches to model selection (e.g., the “Rose” approach, Legault, 2020).  
132 Much uncertainty remains about the appropriateness of strict thresholds for many diagnostic  
133 criteria, the order in which they should be applied, and how to consider models that perform well  
134 on some, but not all diagnostics.

135  
136 This paper synthesizes the lessons learned from previous simulation work on diagnostic  
137 performance (Carvalho et al., 2017; Carvalho et al., 2021) and a series of workshops held with  
138 stock assessors (Karp et. al., 2022, Maunder et al., 2022) to explore and propose “good  
139 practices” for the application of diagnostics to integrated stock assessment models used for  
140 fisheries management. To contribute to the simulation-based literature on this topic, this paper  
141 presents an illustrative (but not exhaustive) study using a statistical catch-at-age simulation  
142 framework to compare diagnostic performance across a spectrum of correctly specified and mis-  
143 specified assessment models. The results are used to contextualize how reliably various  
144 diagnostic tests perform given the degree and nature of known mis-specifications in parameters  
145 and processes and trade-offs, among model fits, prediction skill, and retrospective bias that  
146 analysts must consider as they evaluate diagnostic performance.

147  
148 The field of model diagnostics for fisheries assessment is emerging; the development of state-  
149 space modeling applications also warrant new diagnostic approaches as they become more  
150 commonly used in assessments (Li et al., 2024). The discussion includes good practices and an  
151 evaluation of the tradeoffs in diagnostic performance that analysts must consider when  
152 developing and selecting models used for fisheries management.

## 154 **2. Methods**

### 155 **2.1. Overview**

156 We use a combinatorial simulation approach (Figure S3) to introduce a variety of  
157 misspecifications into the estimation method, and to evaluate the estimation performance  
158 diagnostic tests. The simulation procedure first involves the specification of a model of true  
159 population dynamics, the operating model (OM), using the R package *ss3sim* (Johnson et al.,  
160 2019). The OM is used to generate typical data for a fish population (time-series of catches in  
161 weights from a fishery fleet, an index of abundance from a survey, and the proportions-at-length  
162 and -age for both the fishery and surveys). All three types of data are generated from the OM for  
163 one hundred years, with a period of early recruitment deviations extending for 26 years prior to  
164 the start of the model. These generated data are used in a set of estimation methods (EMs). The  
165 EMs fit to the data and estimate the quantities of management interest. These estimates are then  
166 compared to the true values from the OM.

### 167 **2.2. Operating model (OM)**

168 The OM is an age-structured population dynamics model implemented in the Stock Synthesis  
169 software (SS version 3.30.16, Methot and Wetzel, 2013). Key systems and observation processes  
170 and their parameter values are listed in Table 1, and some biological assumptions (e.g., stock-

171 recruitment steepness, natural mortality, and growth) were originally estimated for Pacific cod  
172 (*Gadus macrocephalus*, Anderson et al., 2014. A general description of the OM is as follows: it  
173 is a one-area, single-sex model, with time-invariant length-weight, length-at-age, and maturity-  
174 at-age relationships, and natural mortality ( $M$ ). Recruitment is assumed to follow a time-invariant  
175 Beverton and Holt (1957) relationship with steepness (expected recruitment at 20% of the  
176 expected pre-fishery biomass,  $h$ ) set to 0.65 and randomly-generated stock-recruitment deviation.  
177 The observation process involves a single fishing fleet and survey.

178 The relative probability of capture at length (selectivity) for the fishery fleet and survey is time-  
179 invariant; the length at 50% selectivity is 52 cm and 51 cm for the fishery and survey,  
180 respectively. All ages are available to the survey and fishery fleets. The initial conditions were  
181 specified so that there was no impact of fishing prior to the first year.

### 182 **2.3. OM Scenarios**

183 Six OM scenarios were designed using combinations of fishing mortality ( $F$ ) vectors and various  
184 levels of recruitment variability: 0.1, 0.4 or 1.0. Simulations were designed to produce unbiased  
185 estimates of spawning biomass in the absence of misspecification (Figure 1a). For each of the six  
186 OM scenarios, 16 OM replicates were generated by resampling the data given a) process error,  
187 sampling a vector of recruitment deviates (Figures 1b and S1) from a normal distribution with  
188 mean zero and the applicable variance for that scenario, and b) observation error for the catch,  
189 survey, and compositional datasets (Figure 1c, d, and e), described below. This number of  
190 replicates was chosen to compromise between simulation run times and balance among the  
191 randomized experimental design. Two vectors of annual fishing mortality rates were generated:  
192 1) a “high” fishing mortality scenario increases to a maximum of twice the true  $F_{MSY}$  and then  
193 decreases to  $0.9F_{MSY}$ , and 2) a “low” fishing mortality scenario is defined by the overall  
194 maximum fishing mortality rate equal to one-fourth of natural mortality (Figure S1). Process  
195 error in each OM replicate arises from variability in annual recruitment deviations (Figure 1c)  
196 see Reproduction section below) and fishing mortality time series. Sixteen replicates of the OM  
197 were generated for each of the six possible fishing mortality and recruitment variation  
198 combinations (e.g., low fishing mortality and recruitment variance of 0.4, Figure S1), for a total  
199 of 96 unique OMs.

200

### 201 **2.4. Data Generation**

202 Data used in the EMs are the time-series of catches in weight from the fishing fleet, a time series  
203 of relative abundance from the survey, and length- and age-composition data that provide a  
204 measure of the size and age structure of the survey and the fishery (Figure S2). The catch  
205 observations were assumed to be known without error (coefficient of variation = 0.01). Each  
206 abundance observation was assumed to be proportional to the available absolute abundance,  
207 called “catchability” in fisheries applications, and was generated from a log-normal distribution  
208 with a coefficient of variation of 0.2 (Figure 1b). Each length- and age-composition observation  
209 was generated from a multinomial distribution with variability described by an effective sample  
210 size of 50 (Figure 1d,e). No additional data weighting was applied to any component. Below, we  
211 describe the model components that were manipulated in our simulation experiments and how  
212 the misspecifications were implemented. The order of the corrections varied with each  
213 simulation, such that all possible unique combinations of corrections were explored.

214

## 215 **2.5. Mis-specified processes**

### 216 *Growth*

217 The growth curve in the OM is modeled using the von Bertalanffy (1957) growth function, a  
218 common relationship used in fisheries assessment to model the length (cm) of an average fish  
219 with respect to its age (years). The model is parameterized using asymptotic length (the inferred  
220 length at infinite age) and the growth rate (the rate at which the average fish reaches asymptotic  
221 length). Researchers may obtain inaccurate input values of this parameter via unrepresentative or  
222 imprecise sampling, which fails to capture or correctly measure individuals at large lengths  
223 and/or older ages (Shelton & Mangel 2012). In the correctly-specified estimation method all  
224 main growth parameters ( $L_I$ ,  $L_\infty$ ,  $K$ , and CVs of length at ages) were freely estimated. For  
225 estimation methods exploring a mis-specification in growth,  $L_I$  and  $K$  were set to correct values  
226 and  $L_\infty$  was set to the mis-specified value, while CVs of length at ages remained estimable.

227 Estimation methods with  $L_\infty$  mis-specified are denoted by the letter L.

228

### 229 *Natural mortality*

230 Generally, it is difficult to obtain empirical estimates of natural mortality for any fish species  
231 (e.g., Hamel, 2014; Punt et al., 2021; Maunder et al., 2023). In fisheries, several methods infer  
232 this value from the maximum age or length (Then et al., 2015 and Hamel and Cope, 2022) or via  
233 a meta-analysis of similar species within a genus (Thorson et al., 2017). In the OM, natural  
234 mortality,  $M$ , is time- and age-invariant and set at  $0.2 \text{ yr}^{-1}$  (Table 1). Estimation methods with  
235 natural mortality mis-specified at incorrect values are denoted by the letter M.

236

### 237 *Reproduction*

238 Steepness of the stock-recruitment relationship, the common measure of stock resilience, is a  
239 highly uncertain yet critical quantity in fishery stock assessment and management. Estimating  
240 steepness inside a stock assessment model is difficult and estimates have lower precision and  
241 higher bias (Lee et al., 2012). Because of the difficulty of estimating steepness, this parameter is  
242 typically not estimated within assessment models. Annual reproduction  $R$  in the OM is calculated  
243 based on a Beverton-Holt function (Equation 1) of the system-wide reproductive biomass in a  
244 given year ( $SB$ ), expected unfished recruitment  $R_0$  and biomass  $SB_0$  and  $h$ , i.e.:

245

$$246 \quad R_y = \frac{4hR_0SB_y}{SB_0(1-h) + SB_y(5h-1)} e^{-0.5\sigma_R^2 + \tilde{R}_y}; \tilde{R}_y \sim N(0, \sigma_R^2) \quad \text{Eq. 1}$$

246

247 Annual recruitment deviates, governed by a recruitment variability error term ( $\sigma_R^2$ ), measure the  
248 log-distance from the deterministic curve given in Equation 1 and is a source of process error in  
249 the OM. The variance in recruitment deviates was set to either 0.1, 0.4 or 1.0. Recruitment  
250 deviates are randomly generated once for each OM replicate; steepness and  $R_0$  are not estimated.  
251 In the estimation methods, the recruitment deviates and  $R_0$  are estimated with steepness set to  
252 either the correct or a mis-specified value (Table S1);  $\sigma_R^2$  is set to the correct value from the  
253 applicable OM. Estimation methods with steepness mis-specified are denoted by the letter H.

254

## 255 *Fishery selectivity*

256 In the OM, the fishery and survey have a length-based double normal selectivity pattern with the  
257 initial selectivity at first bin and final selectivity at last bin parameters set to low numbers to  
258 avoid numerical estimation. This creates an asymptotic selectivity curve, meaning that all  
259 individuals above a certain size have a close to equal probability of being captured. When the  
260 selectivity is correct, estimation methods estimate the selectivity parameters under the  
261 assumption that selectivity is an asymptotic function of length for the fishery. Estimation  
262 methods with selectivity mis-specified are denoted by the letter X, indicating that the model sets  
263 the ascending limb (e.g., the length at 50% selectivity) to a mis-specified value (Table S1).

## 265 *Determining misspecification thresholds*

266 Instead of arbitrarily choosing mis-specified parameter values, the two values nearest to those  
267 used in the OM that led to a 10% change in the final-year depletion (the ratio between final-year  
268 biomass and expected unfished biomass) were solved for. The threshold detection was performed  
269 by fitting a series of estimation methods to the same dataset generated by the OM across a broad  
270 range of fixed values for each parameter in turn: for example, 19 estimation methods with  
271 steepness  $h$  set to 0.05, 0.10, 0.15, ..., 0.95 and other parameter values estimated. This was done  
272 for each unique combination of recruitment deviation variance and fishing mortality. The relative  
273 error in final-year depletion was calculated between the estimates from each estimation method  
274 and the values in the OM and used to find the parameter values nearest to the OM values that  
275 corresponded to relative errors of -10% and 10% (Table S1). Preliminary investigations included  
276 OM values that resulted in relative errors of as much as 20%, but these often involved most  
277 parameters hitting their bounds;  $\pm 10\%$  was selected to keep most parameters within their  
278 plausible ranges (and to avoid having to discard models where estimates were on bounds). This  
279 step ensured that the mis-specifications implemented in the experimental design are known to  
280 impact estimated outputs to the same extent. This led to two mis-specified parameter values, one  
281 above and one below the values used in the OM, for all parameters except for steepness. In cases  
282 where no values above the OM value met the mis-specification threshold criteria (e.g. steepness),  
283 the value above the OM value that corresponded to the greatest relative error was selected.

## 285 **2.6. Estimation methods and experimental design**

287 The EMs were implemented in Stock Synthesis version 3.30 (Methot and Wetzel, 2013). The  
288 experimental design followed a systematic procedure (Figure 2), which enabled the  
289 determination of how well model diagnostics could detect the nature and extent of model  
290 misspecification. Calculation of model diagnostics across 1,536 EMs was facilitated by using the  
291 OpenScienceGrid HTCondor high-throughput computing network (Pordes et al., 2007; Sfiligoi et  
292 al., 2009) and the *ssgrid* package in R (Ducharme-Barth, 2022). The experimental workflow was  
293 as follows:

- 295 1. Generate an operating model “replicate” with process errors (recruitment deviations and  
296 fishing mortalities) and observation errors (generation of survey abundance indices and  
297 compositional data).
- 298 2. Sample a vector of four values for each replicate, each with an even probability of being  
299 either a 0 or 1. This vector determines how each mis-specification, H, M, X, or L, is  
300 implemented. A value of 0 indicates the mis-specification is below the true value whereas

a value of 1 indicates the mis-specification is above the true value. For example, the first OM replicate may have the draw [0, 0, 0, 0] in which all four parameters would be specified below the true value for all EMs fit to those OM data. The next OM replicate may have a different vector draw, ensuring that variation caused by differences in process and observation errors are balanced against the directionality of mis-specifications.

3. Fit EMs for each of the 16 functionally unique combinations corresponding to the mis-specified categories (Table S1) to each replicate. All unique combinations of mis-specifications were evaluated. For example the combination “HMXL” denotes a model with all four mis-specifications, while “MX” denotes a model with only natural mortality (M) and selectivity (X) mis-specified. Note that “MX” is functionally equivalent to “XM” so only the former is investigated. EMs using all components correctly specified and using the correctly-stratified data from the corresponding OM replicate are labeled as “correct”.
4. Repeat steps for each of sixteen resampled OM replicates. This protocol ensures the effect of the mis-specifications was not influenced by the high/low nature of the random vector assigned to each combination. In total, the study design fit 1,536 EMs (16 unique estimation methods fitted to 96 OM replicates).

## 2.7. Performance metrics

### *Relative Error*

The results were summarized by the deviation between the estimates of the management quantities and the corresponding OM values. In lieu of fisheries-specific management quantities (e.g., the ratio of current biomass to the biomass that corresponds to maximum sustainable yield), we examined values common across the EMs, namely the time series in reproductive biomass (here, spawning stock biomass, *SSB*) and reproductive output (here, recruitment). In addition to the general trend in these estimated values, we also evaluated results based on the mean *SSB* over the last ten years. Together, these statistics aim to capture temporal variation in estimation performance as well as model performance during the recent period, which is typically of more interest to managers. The deviations between EM and OM values by year *y*, replicate *i*, combination *j* and scenario *k* were summarized using relative (equation 2) or absolute (equation 3) relative errors and then averaged across replicates.

$$MRE_{SSB_y} = \sum_i \frac{\hat{SSB}_y^{EM_{ijk}} - SSB_y^{OM_i}}{SSB_y^{OM_i}} / i \quad \text{Eq. 2}$$

Both measures indicate the magnitude of difference between estimated quantities and the OM values. Relative error (positive or negative) enables us to investigate whether there are systematic and/or directional biases induced by the various mis-specifications. Using absolute relative error disregards the direction of the difference, and is useful for highlighting the scale of the effects of various mis-specifications. The mean absolute relative (MARE) errors for the terminal ten years of *SSB* are calculated via:



$$MARE_{SSB_y} = \sum_i \frac{\left| \sum_{y=91}^{100} \frac{\widehat{SSB}_y^{EM_{i,j,k}}}{10} - \sum_{y=91}^{100} \frac{SSB_y^{OM_i}}{10} \right|}{\sum_{y=91}^{100} \frac{SSB_y^{OM_i}}{10}} / i \quad \text{Eq. 3}$$

341  
342

## 343 2.8. Model Diagnostics: Review and Application to Simulations

344 We applied model diagnostics following the recommendations of the cookbook using the  
345 associated R package *ss3diags* (Carvalho et al., 2021). The following section provides a brief  
346 summary of the logic and method behind each diagnostic, and how it was applied to our  
347 simulations.

348

### 349 *Convergence*

350 Models were assumed to have converged if no parameters were estimated at a bound, the  
351 gradient was relatively small (less than 1e-4) and the Hessian matrix was invertible, as  
352 recommended in Carvalho et al. (2021). The results shown here are comprised of converged  
353 models only.

354

### 355 *Residual analysis*

356 We explored non-random variation in residual patterns using a non-parametric runs test, wherein  
357 the 2-sided p-value is calculated for the distribution of residuals about a model estimate  
358 (typically estimated indices of survey abundance). If this p value is greater than or equal to 0.05,  
359 there is no evidence to reject the hypothesis that the residuals are randomly residuals and the  
360 model is determined to pass the runs test. We also calculated the root mean square error (RMSE)  
361 for the survey and compositional time series data as a measure of the standard deviation of the  
362 residuals from the model estimates. A small RMSE ( $\leq 0.3$ ) indicates a reasonably precise model  
363 fit to relative abundance indices (Winker et al., 2018).

364

### 365 *R<sub>0</sub> likelihood profile*

366 We constructed likelihood profiles on unfished recruitment ( $R_0$ ) using the `profile()` function  
367 from R package *r4ss* (Taylor et al., 2011). This approach sequentially fixes unfished recruitment  
368 at a pre-specified value and re-runs the estimation method with whatever other parameter  
369 settings were specified in the original experiment. This was repeated for all of the unique OM  
370 replicate-estimation method combinations described previously. The range of  $R_0$  values used  
371 were chosen for each OM replicate, to encompass one unit of  $R_0$  (in log space) both above and  
372 below the MLE for the correct estimation method associated with that replicate, in increments of  
373 **0.2**.

374 This profile enables evaluation of the stability of the parameter estimate, which is influential in  
375 terms of model scale, and the relative influence of individual data sources upon the parameter. A  
376 poorly-estimated parameter is revealed by a profile that is flat (**delta** likelihood values below  
377  $\sim 1.96$  across a large parameter range), and/or may be characterized by data conflicts (where one  
378 or more data sources achieves a minimum likelihood at a much higher or lower parameter value  
379 than the others, or than the total likelihood. Wang et al. (2017) proposed the “psi” statistic, which

380 quantifies whether the maximum likelihood estimate of  $R_0$  for a specific data component falls  
381 within the 95% confidence interval for the total likelihood. This method has the potential to  
382 measure of the information content of a given likelihood component (lower values indicate less  
383 information, and are a rough measure of the degree of mismatch between the total likelihood for  
384 a given EM and the profile obtained for that data component). We did not implement the psi  
385 statistic in this study, as it is not widely used and the comparison of psi statistics from models  
386 with dissimilar parameterizations was not clear. Indications of poor parameter estimation or data  
387 conflict suggest that either model assumptions or data inputs need be re-evaluated.

388

### 389 *Retrospective analysis*

390 A retrospective analysis is a useful approach for addressing the consistency of terminal-year  
391 estimates. The analysis sequentially removes a year of data (a peel) at a time and reruns the  
392 model. The typical interpretation of this analysis is that serial over- or under-estimation of  
393 quantities such as SSB or fishing mortality are indicative of unidentified process error, and  
394 require a revisitation of model assumptions. The severity of over- or under-estimation is  
395 normally evaluated by eye and by the calculation of rho (equation 4) which is then compared to  
396 pre-determined thresholds (Hurtado-Ferro et al., 2015). We conducted retrospective analyses  
397 using the retro function from *r4ss* and mean rho over five, one-year peels was calculated as:

398

$$rho = \frac{1}{h} \sum_{t=1}^h \left( \frac{X_{T-t} - \widehat{X}_{T-t}}{\widehat{X}_{T-t}} \right) \quad \text{Eq. 4}$$

399 where  $X$  is the SSB or fishing mortality,  $\widehat{X}$  is the corresponding estimate from the reference  
400 model (model fitted to the full dataset),  $T$  is the terminal year of the model, and  $h$  is the number  
401 of peels (Hurtado-Ferro et al., 2015). Models with rho values less than -0.15 or greater than 0.20  
402 would fail the retrospective diagnostic based on the rule of thumb as proposed by Hurtado-Ferro  
403 et al. (2015).

404

### 405 *Age-structured production model*

406 Maunder and Piner (2015) proposed an age-structured production model (ASPM) as a model  
407 diagnostic for complex age-structured integrated assessments. Briefly, this approach fixes  
408 selectivity, assumes average recruitment, and disregards compositional data. The tool can be  
409 used to determine whether the stock dynamics are readily explained by the production function  
410 and catches alone, which would suggest that the survey time series provides information  
411 regarding absolute abundance (Minte-Vera et al., 2017). A discrepancy between the ASPM and  
412 age-structured stock trajectory might indicate mis-specification of the components which make  
413 up the production function.

414

415 The ASPM performs best in situations characterized by high and low periods of fishing effort  
416 (also known as “contrast”) and where observations (i.e., catch, life history, and index) are  
417 reasonable representations of the actual states. It has been shown **to be sensitive to**  
418 misspecification of key systems-modeled processes that control the shape of the production  
419 function (Carvalho et al., 2017). However, failure of the ASPM is not necessarily indicative of  
420 model mis-specification and could be due to several factors. The stock could be recruitment  
421 driven (e.g., short-lived fishes with high recruitment variability) and/or lightly exploited such  
422 that the fishing signal is not strong enough to drive change in the stock.

423  
 424 A deterministic recruitment model is a similar means to diagnose a model’s ability to capture the  
 425 production function. Deterministic recruitment model is a simpler alternative to the ASPM as it  
 426 only requires recruitment to be constrained to what would be predicted by the stock-recruit  
 427 relationship without deviation (Merino et al. 2022). For both the ASPM and deterministic  
 428 recruitment model, we calculated the relative difference in model estimates of  $R_0$ ,  $MSY$ , and the  
 429 mean absolute difference (MARE) in predicted SSB between the full model and the  
 430 ASPM/deterministic recruitment model; these metrics are taken to measure how well-defined  
 431 and influential the production function is upon stock dynamics, given the mis-specifications  
 432 investigated in our study.

433  
 434 *Hindcast cross-validation (MASE)*

435 The accuracy and precision of a model’s prediction skill can be measured with hindcast cross-  
 436 validation, which involves comparing observations to predicted future values (Kell et al., 2022)..  
 437 It is similar to retrospective analysis in that it involves peeling one year of data away at a time  
 438 and re-fitting the model but involves an extra step of predicting the removed observation. The  
 439 predicted values are cross-validated by comparing the model’s one-step-ahead forecast, or  
 440 expected value, of the observation at time  $t$  ( $y_t$ ) versus a “naïve” forecasted value equal to the  
 441 last observation ( $y_{t-1}$ ) for a given number of hindcasting time steps ( $h$ ). The prediction skill can  
 442 be calculated using the mean absolute scaled error (MASE) between models, where values less  
 443 than one indicate that the model did better than the naïve approach:  
 444

$$MASE = \frac{\frac{1}{h} \sum_{t=T-h+1}^T |y_t - \hat{y}_t|}{\frac{1}{h} \sum_{t=T-h+1}^T |y_t - y_{t-1}|} \quad \text{Eq. 5}$$

445  
 446 MASE was calculated for relative abundance indices and composition data using ten hindcast  
 447 steps.

448  
 449 *Recruitment trend*

450 The principle of the goodness-of-fit tests (runs and RMSE) described above is that residual  
 451 patterns in model estimates can be indicative of model mis-specification and un-modeled  
 452 process. The estimation of recruitment deviates is a principle way that process error is  
 453 incorporated into stock assessments, and residual trends therein may similarly indicate a mis-  
 454 specification (uncaptured process error). Following Merino et al. (2022) the existence of a  
 455 significant linear trend in the recruitment deviates was quantified, and monotonic trends, and  
 456 non-monotonic (any) trends in recruitment deviates were tested. Additionally, it was calculated if  
 457 first-order autocorrelation in the deviates was different from 0 and runs tests (using a threshold  
 458 of  $p \geq 0.05$  to pass) were applied to test for non-randomness.

459  
 460 **3. Results**

461 **3.1. Convergence**

462 Overall, 82% (1264/1536) of the models converged; results are only presented for converged models (

463 Table 2). All correctly-specified EMs converged, and convergence frequency across all  
464 scenarios declined as the number of mis-specifications increased to a minimum of 79% for four  
465 misspecifications. Proportionally fewer mis-specified models converged when fishing mortality  
466 was high, regardless of recruitment variability. These were typically disqualified due to gradients  
467 above the threshold. Models with low recruitment variability and low fishing mortality  
468 converged the most frequently overall, though convergence rates declined with increasing mis-  
469 specifications.

470

### 471 **3.2. Relative error**

472 The magnitude of error in estimated SSB and depletion varied among OM replicates, with  
473 systematic changes in error given by the fishing mortality vector and level of recruitment  
474 variability. The MRE of terminal SSB was highest with greater model-misspecification and  
475 greater variation in recruitment for both exploitation ( $F$ ) scenarios, though the absolute value of  
476 error was greater under the high  $F$  scenario. This same pattern was present for MRE of estimated  
477 depletion, though the overall scale of error was smaller (ranging from -50 to 50%, Figure 1a).

478

### 479 **3.3. Residual analysis**

480 The residual analyses examined fits to the survey abundance time-series, as well as the  
481 calculation of the root-mean-square error (RMSE) for the survey abundance time-series, length,  
482 and age composition data. RMSE for fishery and survey compositional data increased  
483 systematically with an increasing number of mis-specifications, while the average RMSE for the  
484 survey index of abundance did not dramatically increase even in the presence of 3 or 4 mis-  
485 specifications (Figure 2a). Scenarios with high fishing mortality and/or low recruitment  
486 variability exhibited the largest increases in RMSE for composition data as the number of mis-  
487 specifications increased. Importantly, no models resulted in RMSE values above the 30%  
488 threshold indicated in Winker et al. (2018).

489

490 The majority (97%) of correct models passed the runs test; while pass rates declined with  
491 increasing numbers of misspecification, the overall failure rate only ranged from 4%-10%  
492 (Figure 3 and Table 2). Visual inspection of models that failed the runs test showed slightly  
493 worse fits to the data for the highly mis-specified models compared to the correct model (Figure  
494 3). There were similarities between the performance of the runs test and RMSE. Firstly, most  
495 diagnostic responsiveness (e.g., increased failure rates with increased degree of misspecification)  
496 emerged for the compositional data while p-values for the survey index of abundance were less  
497 responsive (Figure 3). There also appeared to be greater sensitivity (more failures) to increased  
498 mis-specification when fishing mortality was high and/or recruitment variability was low, as was  
499 seen for RMSE.

500

### 501 **3.4. $R_0$ likelihood profile**

502 Of the 254,064 unique models run as part of the profiling exercise, 99% converged and were  
503 included in this analysis. Figure 4 presents the likelihood profiles for the total objective function,  
504 survey data and length and age composition data, scaled so that the x-axis represents the  
505 difference between the fixed  $R_0$  for the model at hand and the value for  $R_0$  from the OM. The  
506 MLE for  $R_0$ , indicated by the total likelihood, was well-defined for correct EMs. For EMs with  
507 zero or one misspecifications, the total likelihood agreed with the survey and length-composition  
508 data, while the age composition data indicated  $R_0$  values slightly lower than the other data

509 sources. The likelihood profiles differed systematically from those obtained using the correct EM  
510 upon the introduction of two or more mis-specifications. The qualitative and relative behavior of  
511 the profiles was strikingly consistent within EMs: the survey data were always the broadest, and  
512 the length and age composition profiles were consistently narrower than the survey and shifted  
513 slightly below the total MLEs. Profiles for the age composition data had lower specificity overall  
514 (many statistically indistinguishable models above and below the MLE). Conflict between the  
515 best  $R_0$  values of the survey and length composition data versus the age composition data was  
516 present in all EMs.

517

### 518 **3.5. Retrospective Analysis**

519 The thresholds proposed by Hurtado-Ferro et al. (2015) had little ability to detect model mis-  
520 specification in our framework. Overall, 5% of the EMs had rho values for the spawning biomass  
521 and fishing mortality time series outside of the thresholds [-0.15, 0.2] (Figure 5). For correct  
522 EMs at all levels of fishing mortality and recruitment variability, rho values for both *SSB* and  
523 fishing mortality were centered around 0 with very few models falling outside the thresholds,  
524 though higher rho values occurred with the highest level of recruitment variability explored. For  
525 the low *F* scenarios, rho values for *SSB* and fishing mortality did diverge from zero to a greater  
526 degree than the high *F* scenarios. The change in rho was most pronounced for scenarios with low  
527 *F* and high recruitment variability. EMs with high fishing mortality did not show clear trends in  
528 the magnitude or direction of rho values with increasing levels of misspecification nor across  
529 recruitment variability levels.

530

### 531 **3.6. Age-structured production model**

532 The performance of the ASPM varied by scenario. Generally, the ASPM estimated *SSB*  
533 trajectories that were higher in scale and smoother through time when fishing mortality was low  
534 (Figure 6). The ASPM was better able to capture the scale and dynamics of the *SSB* trajectory  
535 under the high *F* scenario, with minimal difference from the base model under high *F* and low  
536 recruitment variability (Figure 6). The ASPM and model with deterministic recruitment  
537 consistently resulted in lower *MSY*.

538

539 Both the ASPM and deterministic recruitment results showed virtually identical patterns in terms  
540 of relative error in  $R_0$  from the full model, and the MARE of *SSB* (Figure 7). Both the ASPM and  
541 deterministic recruitment were able to estimate  $R_0$  well. However, the MARE of *SSB* was  
542 consistently over estimated. There were differences between the trends of MARE of *SSB* across  
543 the number of mis-specifications between the ASPM and deterministic recruitment model. For  
544 the ASPM models, MARE of *SSB* showed a general increase as the number of mis-specifications  
545 increased across all levels of fishing mortality and recruitment variability. For the deterministic  
546 recruitment models at all recruitment variability levels, MARE of *SSB* decreased as the number  
547 of mis-specifications increased for models with low fishing mortality but increased as the  
548 number of mis-specifications increased for models with high fishing mortality Both ASPM and  
549 deterministic recruitment models with high recruitment variability had the smallest difference in  
550 *MSY* from the full model and models with low recruitment variability had the greatest difference  
551 in *MSY* from the full model.

552

553 **3.7. Hindcast Cross Validation (MASE)**

554 The MASE statistic indicated that models had better predictive performance than the null (e.g.  
555  $MASE < 1$ ) for all levels of misspecification, with increasing predictive performance with fewer  
556 misspecifications (Figure 8). However, the proportion of models that passed this diagnostic only  
557 ranged from 54% (fully misspecified) to 66% (correct model). There were not strong patterns in  
558 MASE statistics across data types nor  $F$  levels, though it seemed that the lowest passing rates  
559 occurred under lower levels of recruitment variability (54% at the lowest, to 68% under the  
560 highest values of  $\sigma_R$ ). Of the models that had worse predictive power than a null model  
561 ( $MASE > 1$ ), the failed statistic most commonly occurred for age-composition data (Figure 8).  
562 Overall, the hindcast statistic had the highest failure rates of all diagnostics regardless of  
563 exploitation level or recruitment variability (Table 2).

564

565 **3.8. Recruitment trend**

566 Significant linear trends in recruitment deviates usually indicated the presence of at least one  
567 model mis-specification at least for low fishing mortality scenarios, and increasing the number of  
568 mis-specifications tended to increase the proportion of model runs that showed significant linear  
569 trends in recruitment deviates (Figure 9). However, a substantial proportion of mis-specified  
570 models did not indicate significant trends in recruitment deviates (false negatives), and some  
571 correctly specified models showed significant linear trends in the recruitment deviates (false  
572 positives). Additionally, rates of false positives and false negatives were not consistent across  
573 OM replicates. Testing for the presence of monotonic trends or any (non-monotonic) trend  
574 showed similar results as the test for linear trends in recruitment deviates. Runs tests of the  
575 recruitment deviates and testing for any non-zero first order autocorrelation indicated a poor  
576 ability to discriminate between correctly specified models and mis-specified models under the  
577 low fishing mortality scenario.

578 **4. Discussion**

579

580 **4.1. Limitations**

581 *Data Richness*

582 Several characteristics of our study design limit the interpretation of our results and form the  
583 basis for future research regarding the utility and robustness of diagnostic tools. The simulations  
584 explored here are centered on a data-rich, age-structured assessment model, with a longer time  
585 series of data (particularly compositional data) than is likely available even for the most heavily-  
586 monitored stocks (Maunder et al., 2014; Ono et al., 2015). The experiment presented here was  
587 designed to eliminate data concerns so that the performance of diagnostics tests could be  
588 evaluated in a “best-case scenario”; we did not wish to construct candidate EMs that were so  
589 mis-specified that they would be dismissed out of hand by any competent analyst (e.g., an extinct  
590 population, survey estimates completely out of range). A potential risk of our simulation design  
591 is that these data are *so* informative and abundant that models are able to approximate the correct  
592 solution (i.e., fit the survey time series to a satisfactory degree) even when parameters are mis-  
593 specified. This could explain the apparent lack of power that the diagnostics appear to have for  
594 discerning between correctly specified and mis-specified models, particularly for the survey  
595 time-series. Model diagnostics that did not perform well in our study are unlikely to perform well  
596 for similar stocks with fewer or worse data. An additional research avenue related to this topic is  
597 the diagnostic use of changes to the effective sample size for compositional data under a  
598 Dirichlet-multinomial (D-M) distribution (Thorson et al., 2023). However, a minority of global  
599 models have reliable compositional data to begin with, and a minority of those use the D-M  
600 distribution in estimation routines. This highlights the primacy of developing and testing  
601 diagnostic tools that are applicable to a range of model types.

602

603 A simulation that explores how diagnostic performance varies with a reduction in time series  
604 length of frequency, smaller compositional sample sizes, or larger observational errors would test  
605 this hypothesis. Relatedly, we did not introduce temporal variability into our simulation  
606 framework, which may have dampened our ability to detect a retrospective trend. Model-specific  
607 confidence intervals can be calculated for rho (Miller & Legault, 2017), though this approach has  
608 not been adapted widely. We suggest further research into the topic of retrospective thresholds;  
609 recent work has indicated that retrospective performance indeed varies across model complexity  
610 and the amount of data provided to the model (Breivik et al., 2023) or the breadth of model  
611 configurations considered in an ensemble (Brooks & Brodziak, 2024). Our results suggest that  
612 even in the absence of temporal variability, the combination of low exploitation levels, high  
613 recruitment variability and/or high levels of model misspecification can produce patterning in  
614 rho values, so it is not inconceivable that thresholds specific to recruitment and exploitation  
615 histories could be developed.

616

617 *Data Quality*

618 Most assessments, particularly those that rely on fisheries-dependent data sources, will utilize  
619 data that are biased to an unknown degree. This study assumes that all data used in the EMs are  
620 representative and unbiased relative to the dynamics of the OM, again a deliberate decision to  
621 represent a ‘best-case scenario’. In addition to the data availability issues discussed above, the  
622 impact of data *quality* on diagnostic performance remains an open question for future research  
623 (Punt 2023, this issue; Liljestr and et al., 2024). Processes such as hyperstability, effort creep, or

624 the under-reporting of catch can result in non-proportional indices of relative abundance. The  
625 presence of these dynamics could manifest through the residual runs test, poor ASPM, hindcast  
626 cross-validation, or by inducing a trend in the recruitment deviates. It is possible that some  
627 diagnostics are more useful for identifying *data* mis-specifications rather than parameter or  
628 model mis-specifications. An urgent area of future research is to investigate the performance of  
629 diagnostic tests in models with well-specified processes and parameters but poorly-representative  
630 data.

631

### 632 *Recruitment Driven Dynamics and Model Parsimony*

633 This study is also limited because the operating model appears to be recruitment-driven, meaning  
634 that the biomass dynamics suggested by the age-structured model are distinct from what the  
635 underlying production function would suggest, so the recruitment time series (and deviations  
636 thereof) explain the stock's trajectory. (In contrast, a "production driven" stock would be one  
637 where the time series of biomass is well-explained by the mean stock-recruitment relationship  
638 and historical fishery removals). This is likely because of the "data-richness" of the simulation,  
639 in that the composition data (from which recruitment estimates are derived) was abundant and  
640 continuous throughout the time series. The varied performance of the ASPM diagnostic across  
641 scenarios is consistent with findings that such tools perform best when applied to production, not  
642 recruitment-driven stocks (Minte-Vera et al, 2017). This relates to the above discussion of model  
643 and data complexity, and is important considering that many global stocks do not consider age  
644 structure at all for management purposes. We emphasize that several tools, particularly  
645 goodness-of-fit tests and explorations of model convergence are applicable across a range of  
646 model types.

647

648 All operating models tested here used one of two vectors for fishing mortality. Some diagnostics,  
649 like MASE, might simply echo the stock's responsiveness to fishing pressure, which in this case  
650 will be more pronounced in trajectories that have lower *SSB* because of reduced recruitment. As  
651 stated in Punt et al. (2023, this issue), process error can occur in multiple model processes,  
652 including selectivity. This study does not investigate the impacts of time-varying selectivity  
653 curves, or allowing the estimation of the descending limb of the double normal curve, which  
654 could enable the model to compensate for additional mis-specified processes. However, given  
655 that many mis-specified models were able to pass various diagnostic tests, we anticipate that  
656 introducing further flexibility into the model structures would reinforce the ability for mis-  
657 specified models to satisfy diagnostic criteria. An investigation into the relative performance of  
658 these production-related diagnostics for stocks with and without well-informed production  
659 functions would be informative.

660

## 661 **4.2. How do individual diagnostics perform?**

### 662 *Fits to the Data and Parameter Estimation*

663 Model convergence was the strongest indicator of the number of misspecifications, which is  
664 consistent with our recommendation (and that of the Cookbook) that it be the first diagnostic test  
665 performed, and alternative structures explored if the test is failed. Out of all diagnostics other  
666 than model convergence, the RMSE test most reliably returned higher (poorer) values with an  
667 increasing amount of misspecifications. This is reassuring evidence that goodness-of-fit tests can  
668 be a useful first step in evaluating a model. In contrast, the runs test using the traditional cutoff of  
669 0.05 was one of the least reliable diagnostics. A majority of highly mis-specified models passed



670 the runs test at this threshold, corresponding to the fact that all models seemed to visually fit the  
671 survey index. Given this result, it is possible that the statistical cutoff for passing the test is not  
672 appropriate. Our results suggest that correct models have, on average, p-values of 0.5 or higher –  
673 though the range was uninformatively large ( $<0.05$  to 0.95). It is also illustrative that diagnostic  
674 performance was most robust (i.e., failure rates higher with increased mis-specification) in  
675 scenarios with high fishing mortality and/or low recruitment variability. This corresponds to  
676 previous studies that have indicated that model contrast is often required to inform stock  
677 dynamics (Magnusson and Hilborn, 2007), and suggests caution for analysts applying goodness-  
678 of-fit tests to lightly exploited stocks. Similarly, both residual diagnostics exhibited greater  
679 variation when applied to compositional data, while survey time-series scores remained flat. This  
680 result is likely related to the production question described above, and underscores the  
681 importance of running diagnostics on multiple data sources (when available). Analysts should  
682 consider whether visually satisfactory fits to survey abundance time series are sufficient for  
683 model acceptance, and be warned that there are circumstances (high observation errors,  
684 abundance of other data sources) that can lead a mis-specified model to fit the survey time series  
685 well.

686  
687 Likelihood profiles appear to remain internally consistent, with the relative degree of conflict  
688 stable across mis-specified models (e.g. survey abundance data were always less informative  
689 than fishery and survey length composition data, with broader profiles more distinct from the  
690 total likelihood). This indicates that likelihood profiles can remain a useful tool for determining  
691 data conflicts and information content regardless of the degree of misspecification in an  
692 assessment model, but would not alert the analyst to the presence of misspecification.

### 693 *Model Consistency*

695 The present study complements recent research to offer new insights into the process of  
696 identifying and addressing retrospective patterns in fish stock assessments. Legault (2020)  
697 compared the rho-adjustment to the “Rose” approach, a time-intensive process whereby a  
698 retrospective pattern is eliminated across an ensemble of models, allowing the analyst to change  
699 multiple processes or data inputs. That evaluation determined that both approaches are viable for  
700 removing retrospective patterns, though neither identifies the cause(s), and the choice between  
701 approaches depends on the time and expertise available. rho calculated from retrospective  
702 analyses in our study was a surprisingly poor correlate to model misspecification (given the  
703 traditional cutoff range of -0.15 to 0.2 for SSB, Hurtado-Ferro et al. 2015), though retrospective  
704 performance did degrade with increasing mis-specifications. This does not indicate that the  
705 retrospective diagnostic is a poor tool, rather that the presence of a retrospective pattern (and  
706 associated failure of the rho cutoff) is not a guaranteed outcome when the parameters we  
707 examined are mis-specified. This finding is similar to those of Breivik et al. (2023) who  
708 indicated that the acceptable range for the traditional rho diagnostic varies with the amount of  
709 data and type of model used. The authors proposed an alternative “post-sample rho significance  
710 test” with the aim to reduce subjectivity in decisions about significant retrospective patterns in  
711 state-space assessment models. The new statistic, which conditions the distribution of rho values  
712 on the data prior to the retrospective period, enables the analyst to evaluate whether the  
713 retrospective pattern is truly anomalous or reasonable given the model used. This presents a  
714 promising avenue for future research, though readers are reminded that retrospective patterns can

715 be reduced while reference points remain biased (Szuwalski et al. 2018). We do not propose  
716 alternative, universal ranges for the original rho statistic.

717

### 718 *Prediction skill*

719 For most data components, MASE scores were related to the number of model misspecifications,  
720 suggesting that the hindcast diagnostic can evaluate model performance and can potentially  
721 detect model misspecification. However, the MASE criterion used here to quantify prediction  
722 skill was not sensitive enough to detect mis-specification across all models; this is consistent  
723 with earlier work indicating that good MASE performance for hindcasting is likely to occur if  
724 the stock is production driven and the production function is estimable from the data (Minte-  
725 Vera et al., 2021), which is not the case in our example. It is notable that the MASE statistic was  
726 the most commonly failed across all levels of F and recruitment variability (Table 2), and that the  
727 high F – low recruitment scenario did not exhibit improved performance of this diagnostic as it  
728 did for other tests. Given that MASE tests explicitly for prediction skill, it's understandable that  
729 scores less than 1 are harder to obtain when the precision of the data is high and the biomass  
730 trend is relatively flat; this is the reason why users may elect to set a precision threshold for the  
731 naïve prediction error below which the statistic is no longer penalized, an area that requires  
732 further research.

733

734 The ASPM and deterministic recruitment model diagnostics appear versatile and promising.  
735 Carvalho et al., (2017) showed via simulation analysis that ASPM was the only diagnostic  
736 capable of detecting mis-specification of the key systems-modeled processes that control the  
737 shape of the production function. Here, the ASPM and the deterministic recruitment model were  
738 not able to provide evidence for a production function, essentially confirming this example as a  
739 recruitment-driven model. The deterministic recruitment model returned virtually the same  
740 results as the ASPM, so either diagnostic could be used as an alternative to measure the effects of  
741 fishing. Our findings suggest that the ASPM performs best under scenarios with low-to-medium  
742 recruitment variability and when fishing mortality is high, such that contrast is induced in the  
743 time series – an emergent theme among both estimation and diagnostic performance. These  
744 findings, as well as those for the MASE diagnostic, underscore the importance of data contrast  
745 and, relatedly, the presence of a production function in determining diagnostic performance,  
746 which was

747

748 Estimated recruitments are one of the primary ways process error is modeled in stock  
749 assessments, so examining the recruitment deviates for trend and non-randomness makes  
750 intuitive sense as a potential model diagnostic (Merino et al. 2022). Merino et al. (2022) explored  
751 using a test for statistical significance of a linear trend in the recruitment residuals as a potential  
752 diagnostic for identifying model mis-specification within an ensemble of models. Our study is  
753 the first time (to our knowledge) that this diagnostic has been formally evaluated within a  
754 simulation framework. As currently formulated, this diagnostic may have some discriminatory  
755 power in identifying mis-specified models from correctly specified models. While it was more  
756 likely that models with significant linear trend in the residuals were mis-specified, there was still  
757 a chance (~7%) that the model was correctly specified (false-positive). This is close to the  
758 assumed false positive rate of the statistical test ( $p \leq 0.05$ ). However, there remains a large false  
759 negative rate for mis-specified models. Further simulation testing is needed to refine either the

760 statistical thresholds used to identify significant residual trend to see if that improves  
761 discriminatory power or the types of mis-specifications this test may be used to identify.

762

### 763 **4.3. Good practices in applying model diagnostics**

#### 764 *Updating the “Cookbook”*

765 The original cookbook (Carvalho et al., 2021) proposed a linear workflow of diagnostic tools,  
766 whereby a model is required to “pass” a set of diagnostics in a given order. That workflow  
767 inherently prioritized certain tests such as residual diagnostics and the runs test before likelihood  
768 profiling or retrospective analyses. The spirit of that approach – that a model should converge  
769 and reasonably fit the data to be considered a candidate – is unchanged, and we point readers to  
770 both the original workflow as well as the ordered list provided in Table 2 when applying  
771 diagnostic tests. Our findings further contextualize how modelers should use the outcomes of  
772 diagnostics: firstly, it is evident that the degree of recruitment variability and exploitation history  
773 together modulate diagnostic performance (likely through the induction of contrast in the data),  
774 so quantitative thresholds for most diagnostics, if desired, would need to be developed with those  
775 factors in mind. Secondly, diagnostics of prediction skill (namely the MASE statistic) appear less  
776 insensitive to model misspecification overall, though with more promising performance for age-  
777 composition data than for survey biomass. Further research into the best way to test for and  
778 improve prediction skill, particularly the use of thresholds or minima for such diagnostics, is  
779 warranted. Finally, our results show that it is possible to develop plausible, realistic stock  
780 assessments that fit data well and still perform poorly on some model diagnostics. We suggest  
781 that the community should strive for a balance among the considerations of model realism and  
782 diagnostic performance.

783

#### 784 *Tradeoffs in Model Development*

785 The primary challenge in developing diagnostic workflows arises because stock assessors must  
786 evaluate a small subset of total possible models representing a population, far fewer than the  
787 hundreds of thousands of models run for this simulation analysis. In our study, the RMSE,  
788 ASPM and likelihood profile diagnostics were the most internally consistent and responsive to  
789 the presence of misspecification. Yet an assessment scientist would only see results for, at most,  
790 a dozen models, and have no knowledge of how divergent the selected model is from reality.  
791 Furthermore, the information gleaned from diagnostics such as the RMSE is not much more  
792 useful than a simple visual inspection of the model fits; it is likely that models with poor RMSE  
793 scores would have been discarded in the first place based on their poor fits to the survey data.  
794 This means that the scientist’s holistic evaluation of the model’s ecological plausibility remains  
795 necessary.

796

797 The tentative “good practices” and associated precautions presented in Table 3 warrant a  
798 comment about the general push towards automation of assessment procedures. We assert that  
799 stock assessment modeling requires the experience and the subjective evaluation of competing  
800 priorities, which are not replaceable by a set of diagnostic algorithms – particularly when the true  
801 recruitment trend might be unknown, as discussed above. Tools such as machine learning  
802 (particularly for image classification), boosted regression trees, and artificial intelligence present  
803 a promising avenue that may improve data collection (Zhang et al., 2022) and detect patterns in  
804 population dynamics (Mendoza et al, 2012; Memarzadeh et al., 2019). Furthermore, the nature of  
805 assessment science requires analysts to place value on sometimes competing priorities, whether

806 in a formal framework such as a management strategy evaluation (Punt et al., 2016), or in the  
807 process of data or model weighting (Francis, 2017). These subjective tasks invite the  
808 consideration of socio-economic topics and the participation of fishery stakeholders, which could  
809 lead to model configurations being selected despite poor performance on one or more diagnostic  
810 criteria. For this reason, as well as the growing body of evidence that standardized cutoffs for  
811 diagnostic performance are not ideal for the selection of management models, we discourage the  
812 use of automatic pass/fail criteria for most diagnostic tests. Instead, analysts are encouraged to  
813 couple the results of diagnostic tests with their expert evaluation of the model's plausibility,  
814 given the biological and historical context of the stock, and in consultation with managers. We  
815 caution assessors against reverting to simplified assessment types (e.g., data-limited methods,  
816 Legault et al. 2023) in order to pass diagnostic tests as they carry the risk of poor management  
817 performance.

#### 818 819 **4.1. Conclusion**

820 This study substantially expands the simulation framework developed by Carvalho et al. (2017)  
821 and updates the framework for applying diagnostics to integrated fisheries assessments presented  
822 by Carvalho et al. (2021). There remains several outstanding research avenues as the community  
823 continues to refine (or discard) quantitative diagnostic criteria. Further investigation of  
824 diagnostic performance should evaluate 1) the impact of changes to data quality or availability  
825 (particularly for the case of data-limited stocks (e.g., no age or length structure); 2) if there are  
826 correlations between the operating model characteristics (e.g. general stock trajectory or  
827 production/recruitment driven dynamics, or trends in process error) and diagnostic performance,  
828 and 3) whether the introduction of time-varying components (such as recruitment regime shifts,  
829 or time blocks in selectivity) impact the performance of diagnostic tests, particularly those  
830 associated with prediction skill. The scientific assessment community should continue to  
831 investigation via simulation without neglecting the subjective expertise and decision-making  
832 skill required to produce analyses for scientific management.

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837

838 **6. References**

- 839 Anderson, S.C., Monnahan, C.C., Johnson, K.F., Ono, K., Valero, J.L., 2014. ss3sim: An R  
840 Package for Fisheries Stock Assessment Simulation with Stock Synthesis. PLOS ONE 9,  
841 e92725. <https://doi.org/10.1371/journal.pone.0092725>  
842
- 843 Beverton, R. J. H. and S. J. Holt. On the Dynamics of Exploited Fish Populations. Fisheries  
844 Investment Series 2. 19. U. K. Min. of Agr. and Fish. Chapman and Hall: London (1957).  
845
- 846 Breivik, O.N., Aldrin, M., Fuglebakk, E., Nielsen, A., 2023. Detecting significant retrospective  
847 patterns in state space fish stock assessment. Can. J. Fish. Aquat. Sci. 80, 1509–1518.  
848 <https://doi.org/10.1139/cjfas-2022-0250>  
849
- 850 Brooks, E.N., Brodziak, J.K.T., 2024. Simulation testing performance of ensemble models when  
851 catch data are underreported. ICES Journal of Marine Science fsae067.  
852 <https://doi.org/10.1093/icesjms/fsae067>  
853
- 854 Carvalho, F., Punt, A.E., Chang, Y.-J., Maunder, M.N., Piner, K.R., 2017. Can diagnostic tests  
855 help identify model misspecification in integrated stock assessments? Fisheries Research. 192,  
856 28–40. <https://doi.org/10.1016/j.fishres.2016.09.018>.  
857
- 858 Carvalho, F., Winker H., Courtney D., Kapur M., Kell L., Cardinale M., Schirripa M., Kitakado  
859 T., Yemane D., Piner K. R., Maunder M. N., Taylor I. Wetzal C. R., Doering K., Johnson K. F.,  
860 and Methot R. D. 2021. A cookbook for using model diagnostics in integrated stock assessments.  
861 Fisheries Research. <https://doi.org/10.1016/j.fishres.2021.105959>  
862
- 863 Ducharme-Barth, N. 2022. ssgrid: ssgrid: Stock Sythesis - OpenScienceGrid - utilities.  
864 <https://github.com/N-DucharmeBarth-NOAA/ssgrid>, [https://n-ducharmebarth-](https://n-ducharmebarth-noaa.github.io/ssgrid/)  
865 [noaa.github.io/ssgrid/](https://n-ducharmebarth-noaa.github.io/ssgrid/).  
866
- 867 Fisch, N., Shertzer, K., Camp, E., Maunder, M., Ahrens, R., 2023. Process and sampling  
868 variance within fisheries stock assessment models: estimability, likelihood choice, and the  
869 consequences of incorrect specification. ICES Journal of Marine Science fsad138.  
870 <https://doi.org/10.1093/icesjms/fsad138>  
871
- 872 Francis, R.I.C.C., 2017. Revisiting data weighting in fisheries stock assessment models.  
873 Fisheries Research 192, 5–15. <https://doi.org/10.1016/j.fishres.2016.06.006>  
874
- 875 Fournier, D., Archibald, C.P., 1982. A General Theory for Analyzing Catch at Age Data.  
876 Canadian Journal of Fisheries and Aquatic Sciences 39, 1195–1207. [https://doi.org/10.1139/f82-](https://doi.org/10.1139/f82-157)  
877 [157](https://doi.org/10.1139/f82-157)  
878

879 Hamel, O.S., 2015. A method for calculating a meta-analytical prior for the natural mortality rate  
880 using multiple life history correlates. *ICES Journal of Marine Science* 72, 62–69.  
881 <https://doi.org/10.1093/icesjms/fsu131>  
882

883 Hamel, O.S., Cope, J.M., 2022. Development and considerations for application of a longevity-  
884 based prior for the natural mortality rate. *Fisheries Research* 256, 106477.  
885 <https://doi.org/10.1016/j.fishres.2022.106477>  
886

887 Hurtado-Ferro, F., Szuwalski, C.S., Valero, J.L., Anderson, S.C., Cunningham, C.J., Johnson,  
888 K.F., Licandeo, R., McGilliard, C.R., Monnahan, C.C., Muradian, M.L., Ono, K., Vert-Pre,  
889 K.A., Whitten, A.R., Punt, A.E., 2015. Looking in the rear-view mirror: bias and retrospective  
890 patterns in integrated, age-structured stock assessment models. *ICES Journal of Marine Science*  
891 72, 99–110. <https://doi.org/10.1093/icesjms/fsu198>  
892

893 Ian G. Taylor, Kathryn L. Doering, Kelli F. Johnson, Chantel R. Wetzel, Ian J. Stewart, 2021.  
894 Beyond visualizing catch-at-age models: Lessons learned from the r4ss package about software  
895 to support stock assessments, *Fisheries Research*, 239:105924.  
896 <https://doi.org/10.1016/j.fishres.2021.105924>.  
897

898 Jardim, E., Azevedo, M., Brodziak, J., Brooks, E.N., Johnson, K.F., Klibansky, N., Millar, C.P.,  
899 Minto, C., Mosqueira, I., Nash, R.D.M., Vasilakopoulos, P., Wells, B.K., 2021. Operationalizing  
900 ensemble models for scientific advice to fisheries management. *ICES Journal of Marine Science*  
901 78, 1209–1216. <https://doi.org/10.1093/icesjms/fsab010>  
902

903 Johnson KF, Anderson SC, Doering K, Monnahan CC, Stawitz CC, Taylor IG (2019). ss3sim:  
904 Fisheries Stock Assessment Simulation Testing with Stock Synthesis. R package version 1.0.3.  
905

906 Karp, M.A., Kuriyama, P., Blackhart, K., Brodziak, J., Carvalho, F., Curti, K., Dick, E.J.,  
907 Hanselman, D., Ianelli, J., Sagarese, S., Shertzer, K., Taylor, I., 2022. Common model  
908 diagnostics for fish stock assessments in the United States (No. NMFS-F/SPO240A). National  
909 Marine Fisheries Service.  
910

911 Kell, L.T., Sharma, R., Kitakado, T., Winker, H., Mosqueira, I., Cardinale, M., Fu, D., 2021.  
912 Validation of stock assessment methods: is it me or my model talking? *ICES Journal of Marine*  
913 *Science* 78, 2244–2255. <https://doi.org/10.1093/icesjms/fsab104>  
914

915 Langseth, B.J., Schueller, A.M., Shertzer, K.W., Craig, J.K., Smith, J.W., 2016. Management  
916 implications of temporally and spatially varying catchability for the Gulf of Mexico menhaden  
917 fishery. *Fisheries Research* 181, 186–197. <https://doi.org/10.1016/j.fishres.2016.04.013>  
918

919 Li, C., Deroba, J.J., Miller, T.J., Legault, C.M., Perretti, C.T., 2024. An evaluation of common  
920 stock assessment diagnostic tools for choosing among state-space models with multiple random  
921 effects processes. *Fisheries Research* 273, 106968. <https://doi.org/10.1016/j.fishres.2024.106968>  
922

923 Legault, C.M., Wiedenmann, J., Deroba, J.J., Fay, G., Miller, T.J., Brooks, E.N., Bell, R.J.,  
Langan, J.A., Cournane, J.M., Jones, A.W., Muffley, B., 2023. Data-rich but model-resistant: an

924 evaluation of data-limited methods to manage fisheries with failed age-based stock assessments.  
925 Can. J. Fish. Aquat. Sci. 80, 27–42. <https://doi.org/10.1139/cjfas-2022-0045>  
926

927 Legault, C.M., 2020. Rose vs. rho: a comparison of two approaches to address retrospective  
928 patterns in stock assessments. ICES Journal of Marine Science 77, 3016–3030.  
929 <https://doi.org/10.1093/icesjms/fsaa184>  
930

931 Lee, H.-H., Maunder, M.N., Piner, K.R., Methot, R.D., 2012. Can steepness of the stock–  
932 recruitment relationship be estimated in fishery stock assessment models? Fisheries Research  
933 125–126, 254–261. <https://doi.org/10.1016/j.fishres.2012.03.001>  
934

935 Lee, H., Piner, K.R., Taylor, I.G., Kitakado, T., 2019. On the use of conditional age at length  
936 data as a likelihood component in integrated population dynamics models. Fisheries Research  
937 216, 204–211. <https://doi.org/10.1016/j.fishres.2019.04.007>  
938

939 Legault, C.M., Wiedenmann, J., Deroba, J.J., Fay, G., Miller, T.J., Brooks, E.N., Bell, R.J.,  
940 Langan, J.A., Cournane, J.M., Jones, A.W., Muffley, B., 2023. Data-rich but model-resistant: an  
941 evaluation of data-limited methods to manage fisheries with failed age-based stock assessments.  
942 Can. J. Fish. Aquat. Sci. 80, 27–42. <https://doi.org/10.1139/cjfas-2022-0045>  
943

944 Liljestrand, E.M., Bence, J.R., Deroba, J.J., 2024. The effect of process variability and data  
945 quality on performance of a state-space stock assessment model. Fisheries Research 275,  
946 107023. <https://doi.org/10.1016/j.fishres.2024.107023>  
947

948 Magnusson, A. and Hilborn, R. (2007), What makes fisheries data informative?. Fish and  
949 Fisheries, 8: 337-358. <https://doi.org/10.1111/j.1467-2979.2007.00258.x>  
950

951 Maunder, M.N., Schnute, J.T., Ianelli, J. 2009. Computers in fisheries population dynamics. In:  
952 Megrey, B.A., Moksness, E. (Eds.), Computers in Fisheries Research. Springer, pp. 337–372.  
953

954 Maunder, M.N., Piner, K.R., 2017. Dealing with data conflicts in statistical inference of  
955 population assessment models that integrate information from multiple diverse data sets.  
956 Fisheries Research 192, 16–27. <https://doi.org/10.1016/j.fishres.2016.04.022>  
957

958 Maunder, M.N., Punt, A.A., 2014. A review of integrated analysis in fisheries stock assessment.  
959 Fisheries Research 142, 61–74. <https://doi.org/10.1016/j.fishres.2012.07.025>  
960

961 Maunder, M.N. and Piner, K.R., 2015. Contemporary fisheries stock assessment: many issues  
962 still remain. ICES Journal of Marine Science, 72(1), pp.7-18.  
963

964 Maunder, M., Punt, A., Carvalho, F., Winker, H., Valero, J., Minte-Vera, C., 2022. 1st  
965 Workshop On Improving The Risk Analysis For Tropical Tunas In The Eastern Pacific Ocean:  
966 Model Diagnostics In Integrated Stock Assessments 1st Workshop On Improving The Risk  
967 Analysis For Tropical Tunas In The Eastern Pacific Ocean: Model Diagnostics In Integrated  
968 Stock Assessments (No. WSRSK-01).  
969

970 Maunder, M.N., Hamel, O.S., Lee, H.-H., Piner, K.R., Cope, J.M., Punt, A.E., Ianelli, J.N.,  
971 Castillo-Jordán, C., Kapur, M.S., Methot, R.D., 2023. A review of estimation methods for  
972 natural mortality and their performance in the context of fishery stock assessment. *Fisheries*  
973 *Research* 257, 106489. <https://doi.org/10.1016/j.fishres.2022.106489>  
974

975 Memarzadeh, M., Britten, G.L., Worm, B., Boettiger, C., 2019. Rebuilding global fisheries under  
976 uncertainty. *Proceedings of the National Academy of Sciences* 116, 15985–15990.  
977 <https://doi.org/10.1073/pnas.1902657116>  
978

979 Mendoza, M., Pennino, M.G., Bellido, J.M., 2012. Tree-based machine learning analysis for  
980 fisheries research. *Fishery Management* 61–75.  
981

982 Merino, G., Urtizberea, A., Fu, D., Winker, H., Cardinale, M., Lauretta, M., Murua, H.,  
983 Kitakado, T., Arrizabalaga, H., Scott, H., Pilling, G., Minte-Vera, C., Xu, H., Laborda, A.,  
984 Erauskin-Extramiana, M., Santiago, J. 2022. Investigating trends in process error as a diagnostic  
985 for integrated fisheries stock assessments. *Fisheries Research*.  
986 <https://doi.org/10.1016/j.fishres.2022.106478>.  
987

988 Methot, R.D., Wetzel, C.R. 2013. Stock synthesis: a biological and statistical framework for fish  
989 stock assessment and fishery management. *Fisheries Research*. 142, 86–99.  
990 <https://doi.org/10.1016/j.fishres.2012.10.012>.  
991

992 Timothy J. Miller, Christopher M. Legault, Statistical behavior of retrospective patterns and their  
993 effects on estimation of stock and harvest status, *Fisheries Research*, Volume 186, Part 1,  
994 2017, Pages 109-120, ISSN 0165-7836, <https://doi.org/10.1016/j.fishres.2016.08.002>.  
995

996 Minte-Vera, C.V., Maunder, M.N., Aires-da-Silva, A.M., Satoh, K., Uosaki, K., 2017. Get the  
997 biology right, or use size-composition data at your own risk. *Fisheries Research*. 192, 114–125.  
998 <https://doi.org/10.1016/j.fishres.2017.01.014>.  
999

1000 Minte-Vera, C.V., Maunder, M.N., Aires-da-Silva, A.M., 2021. Auxiliary diagnostic analyses  
1001 used to detect model misspecification and highlight potential solutions in stock assessments:  
1002 application to yellowfin tuna in the eastern Pacific Ocean. *ICES Journal of Marine Science* 78,  
1003 3521–3537. <https://doi.org/10.1093/icesjms/fsab213>  
1004

1005 Ono, K., Licandeo, R., Muradian, M.L., Cunningham, C.J., Anderson, S.C., Hurtado-Ferro, F.,  
1006 Johnson, K.F., McGilliard, C.R., Monnahan, C.C., Szuwalski, C.S., Valero, J.L., Vert-Pre, K.A.,  
1007 Whitten, A.R., Punt, A.E., 2015. The importance of length and age composition data in statistical  
1008 age-structured models for marine species. *ICES Journal of Marine Science* 72, 31–43.  
1009 <https://doi.org/10.1093/icesjms/fsu007>  
1010

1011 Piner, K.R., Lee, H.-H., Maunder, M.N., Methot, R.D., 2011. A Simulation-Based Method to  
1012 Determine Model Misspecification: Examples Using Natural Mortality and Population Dynamics  
1013 Models. *Marine and Coastal Fisheries* 3, 336–343.  
1014 <https://doi.org/10.1080/19425120.2011.611005>  
1015



1016 Pordes, R., Petravick, D., Kramer, B., Olson, D., Livny, M., Roy A., Avery, P., Blackburn, K.,  
1017 Wenaus, T., Würthwein, F., Foster, I., Gardner, R., Wilde, M., Blatecky, A., McGee, J., Quick,  
1018 R. 2007. The open science grid. *Journal of Physics: Conference Series*. 78:12057. doi:  
1019 10.1088/1742-6596/78/1/012057  
1020  
1021 Punt, A.E., Butterworth, D.S., de Moor, C.L., De Oliveira, J.A.A., Haddon, M., 2016.  
1022 Management strategy evaluation: Best practices. *Fish and Fisheries* 17, 303–334.  
1023 <https://doi.org/10.1111/faf.12104>  
1024  
1025 Punt, A.E., Castillo-Jordán, C., Hamel, O.S., Cope, J.M., Maunder, M.N. and Ianelli, J.N., 2021.  
1026 Consequences of error in natural mortality and its estimation in stock assessment models.  
1027 *Fisheries Research*, 233, p.105759  
1028  
1029 Punt, A.E., 2023. Those who fail to learn from history are condemned to repeat it: A perspective  
1030 on current stock assessment good practices and the consequences of not following them.  
1031 *Fisheries Research* 261, 106642. <https://doi.org/10.1016/j.fishres.2023.106642>  
1032  
1033 Sfiligoi, I., Bradley, D. C., Holzman, B., Mhashilkar, P., Padhi, S., Wurthwein, F. 2009. The  
1034 Pilot Way to Grid Resources Using glideinWMS. 2009. WRI World Congress on Computer  
1035 Science and Information Engineering, Vol. 2, pp. 428–432. doi:10.1109/CSIE.2009.950.  
1036  
1037 Szuwalski, C.S., Thorson, J.T., 2017. Global fishery dynamics are poorly predicted by classical  
1038 models. *Fish and Fisheries* 18, 1085–1095.  
1039  
1040 Shelton AO, Mangel M. Estimating von Bertalanffy parameters with individual and  
1041 environmental variations in growth. *J Biol Dyn*. 2012;6 Suppl 2:3-30. doi:  
1042 10.1080/17513758.2012.697195. Epub 2012 Jun 28. PMID: 22882022.  
1043  
1044 Szuwalski, C.S., Ianelli, J.N., Punt, A.E., 2018. Reducing retrospective patterns in stock  
1045 assessment and impacts on management performance. *ICES Journal of Marine Science* 75, 596–  
1046 609. <https://doi.org/10.1093/icesjms/fsx159>  
1047  
1048 Szuwalski, C., 2022. Estimating time-variation in confounded processes in population dynamics  
1049 modeling: A case study for snow crab in the eastern Bering Sea. *Fisheries Research* 251, 106298.  
1050 <https://doi.org/10.1016/j.fishres.2022.106298>  
1051  
1052  
1053 Taylor, I.G., Methot, R.D., 2013. Hiding or dead? A computationally efficient model of selective  
1054 fisheries mortality. *Fisheries Research* 142, 75–85. <https://doi.org/10.1016/j.fishres.2012.08.021>  
1055  
1056 Tempel, D.J., Peery, M.Z., Gutiérrez, R.J., 2014. Using Integrated Population Models to Improve  
1057 Conservation Monitoring: California Spotted Owls as a Case Study. *Ecological Modelling* 289,  
1058 86–95. <https://doi.org/10.1016/j.ecolmodel.2014.07.005>  
1059  
1060 Then, A.Y., Hoenig, J.M., Hall, N.G., Hewitt, D.A., Handling editor: Ernesto Jardim, 2015.  
1061 Evaluating the predictive performance of empirical estimators of natural mortality rate using

1062 information on over 200 fish species. *ICES Journal of Marine Science* 72, 82–92.  
1063 <https://doi.org/10.1093/icesjms/fsu136>  
1064  
1065 Thorson, J.T., Munch, S.B., Cope, J.M., Gao, J., 2017. Predicting life history parameters for all  
1066 fishes worldwide. *Ecological Applications* 27, 2262–2276. <https://doi.org/10.1002/eap.1606>  
1067  
1068 **Thorson, J.T., Monnahan, C.C., Hulson, P.-J.F., 2023. Data weighting: An iterative process**  
1069 **linking surveys, data synthesis, and population models to evaluate mis-specification.**  
1070 ***Fisheries Research* 266, 106762. <https://doi.org/10.1016/j.fishres.2023.106762>**  
1071  
1072 **Trijoulet, V., Albertsen, C.M., Kristensen, K., Legault, C.M., Miller, T.J., Nielsen, A., 2023.**  
1073 **Model validation for compositional data in stock assessment models: Calculating residuals**  
1074 **with correct properties. *Fisheries Research* 257, 106487.**  
1075 **<https://doi.org/10.1016/j.fishres.2022.106487>**  
1076  
1077 Von Bertalanffy, L. (1957) Quantitative Laws in Metabolism and Growth. *Quarterly Review of*  
1078 *Biology*, 3, 218.  
1079  
1080 Wang, S.P., Maunder, M.N., 2017. Is down-weighting composition data adequate for dealing with  
1081 model misspecification, or do we need to fix the model? *Fisheries Research* 192, 41–51.  
1082 <https://doi.org/10.1016/j.fishres.2016.12.005>  
1083  
1084 Winker, H., Carvalho, F., Kapur, M., 2018. JABBA: Just Another Bayesian Biomass  
1085 Assessment. *Fisheries Research* 204, 275–288. <https://doi.org/10.1016/j.fishres.2018.03.010>  
1086  
1087 Winker, H., Carvalho, F., Thorson, J.T., Kell, L.T., Parker, D., Kapur, M., Sharma, R., Booth,  
1088 A.J., Kerwath, S.E., 2020. JABBA-Select: Incorporating life history and fisheries’ selectivity  
1089 into surplus production models. *Fisheries Research*.  
1090 <https://doi.org/10.1016/j.fishres.2019.105355>  
1091  
1092  
1093 Zhang, D., O’Conner, N.E., Simpson, A.J., Cao, C., Little, S., Wu, B., 2022. Coastal fisheries  
1094 resource monitoring through A deep learning-based underwater video analysis. *Estuarine,*  
1095 *Coastal and Shelf Science* 269, 107815. <https://doi.org/10.1016/j.ecss.2022.107815>  
1096  
1097  
1098  
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1104 **7. Tables**1105 *Table 1. Key systems and observation processes and parameter values.*

<b>Parameter</b>	<b>Value</b>
Natural mortality, $M$ ( $\text{yr}^{-1}$ )	0.2
Reference age, $A_{min}$ (yr)	0
Maximum age, $A_{max}$ (yr)	25
Length at $A_{min}$ , $L_{Amin}$ (cm)	20.5
Length at $A_{max}$ , $L_{Amax}$ (cm)	135.3
Growth rate, $k$ ( $\text{yr}^{-1}$ )	0.19
CV of length $< L_{Amin}$	0.10
CV of length $< L_{Amax}$	0.08
Length-weight coefficient	6.8e-6
Length-weight exponent	3.101
Length at 50% maturity, $L_{mat50}$ (cm)	38.18
Slope of maturity ogive	-0.276
Unfished recruitment ( $\text{Log } R_0$ )	19.0
Spawner-recruitment steepness ( $h$ )	0.65
Catchability ( $\text{Log } q$ )	0.045
Length selectivity for fishery*	50.8, -3, 5.08, 6.99, -999, 999
Length selectivity for survey*	41.8, -4, 4.97, 6.49, -99, 99

1106 \*Values for parameterization of double-normal selectivity curve; see Figure S2. For details, see Methot and Wetzel  
1107 (2013).

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Table 2. Percentage of runs (values) that passed each diagnostic across the F (row groups) and recruitment scenarios (column groups) for various numbers of misspecifications (columns). The values in the cells and the cell color denote the percentage of replicates that passed the tests, ranging from 0% (dark red) to 100% (dark green). Not all diagnostics have published or widely used quantitative thresholds, so pass rates were not obtainable for every diagnostic examined.

Diagnostic	sigma R = 0.1					sigma R = 0.4					sigma R = 1.0						
	0	1	2	3	4	0	1	2	3	4	0	1	2	3	4		
% Converged	100	94	96	98	100	100	97	96	97	100	100	98	97	98	100	Low F	
Mohn's Rho (SSB)	100	97	97	95	94	94	95	95	97	94	75	76	82	86	81		
Mohn's Rho (F)	94	97	95	94	94	94	95	97	95	100	75	76	81	86	81		
Hindcast Fishery Age	56	55	54	51	50	69	68	66	63	62	75	73	74	78	81		
Hindcast Survey Age	56	53	45	30	12	62	61	51	40	38	62	59	58	54	50		
Hindcast Survey Bio	75	58	57	51	50	50	58	49	52	44	75	73	68	70	75		
Hindcast Fishery Len	88	77	76	78	81	62	68	72	74	75	62	65	66	67	69		
Hindcast Survey Len	62	53	42	38	38	62	56	51	50	50	75	70	66	62	62		
Runs Test Fishery Age	100	93	95	95	94	94	95	93	94	94	94	92	89	89	81		
Runs Test Fishery Len	100	100	99	98	100	100	98	97	100	100	100	98	96	94	94		
Runs Test Survey Age	81	80	79	78	69	100	100	96	89	75	100	97	90	87	81		
Runs Test Survey Bio	100	98	93	89	94	100	100	100	100	94	94	94	94	94	94		
Runs Test Survey Len	94	95	96	95	94	100	95	95	95	94	100	95	91	89	75		
% Converged	100	77	68	55	56	100	81	66	64	69	100	64	65	53	50		High F
Mohn's Rho (SSB)	100	98	98	97	100	100	100	100	100	100	100	98	98	100	100		
Mohn's Rho (F)	100	98	98	97	100	100	100	100	100	100	94	95	98	100	100		
Hindcast Fishery Age	62	59	63	51	44	62	62	65	61	64	81	78	74	76	88		
Hindcast Survey Age	56	55	42	37	22	56	56	41	41	27	62	56	47	47	50		
Hindcast Survey Bio	69	67	68	74	67	62	60	63	51	45	50	56	55	50	38		
Hindcast Fishery Len	69	65	60	63	67	56	60	57	54	64	69	66	60	62	62		
Hindcast Survey Len	50	47	45	43	44	69	62	65	61	55	75	73	69	50	25		
Runs Test Fishery Age	100	98	94	94	100	94	98	98	100	91	88	93	95	97	100		
Runs Test Fishery Len	94	94	95	91	89	100	98	98	100	100	100	100	98	97	100		
Runs Test Survey Age	94	92	85	86	89	100	92	81	73	55	100	90	84	76	62		
Runs Test Survey Bio	100	98	100	100	100	100	100	100	100	100	100	100	98	97	100		
Runs Test Survey Len	94	96	94	91	78	100	96	94	95	100	100	93	90	79	62		

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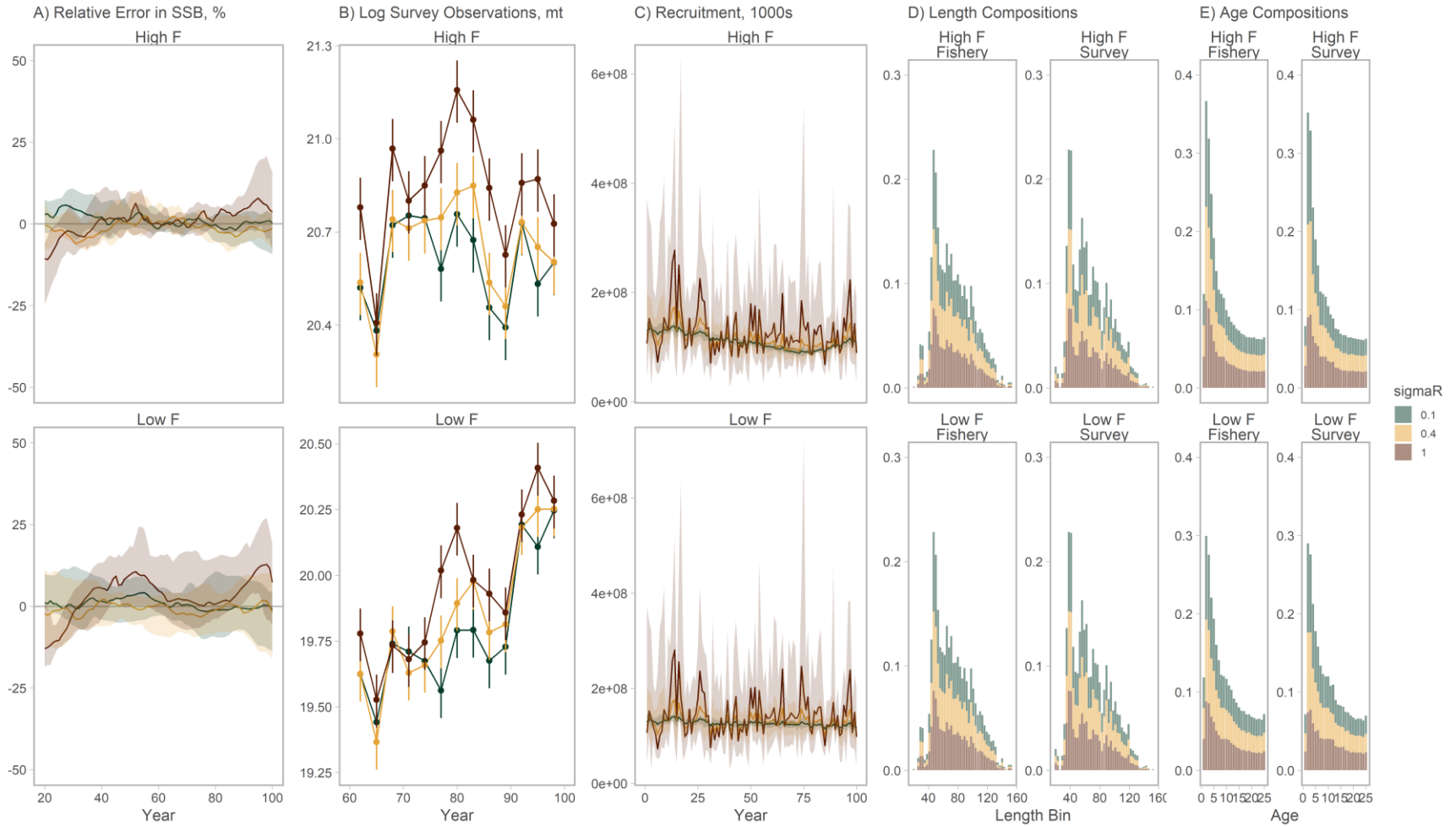
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Table 3. Tentative good practices and precautions for applying diagnostic tests to assessments. *This table is meant to be taken as an ordered guide; should a model fail the diagnostic “good practice” for a given row, we suggest exploration of alternative model structure(s).*

Diagnostic	Good Practice	Precaution
Plausibility	Contextualize model in ecological, life-history and fishery dynamics (realism)	Risk of model over-complication (e.g., too many or improper time-varying processes, Szuwalski 2017; Fisch 2023,) Particularly difficult when working with multi-species and multi-area assessments
Convergence and Check for Global Solution	Final gradient below pre-specified minimum (e.g., 1E-4); Hessian matrix is invertible; No parameters are on bound Consider Bayesian approaches when applicable	Avoid massaging data or exhaustive chain lengths (jitter, MCMC) to force convergence; set terms of analyses ahead-of-time
<b>Goodness of fit</b>		
Residual Diagnostics	Visual inspection of residuals and model fits	The p-value of 0.05 (runs test) might be too low; RMSE cutoff of 30% might be too high Tests appear more sensitive when applied to compositional data than indices Beware small time series Consider One-Step-Ahead residuals for compositional data (Trijoulet et. al. 2022)
<b>Model Consistency</b>		
R0 likelihood profile	Profile over key model parameters ( $R_0$ , M and steepness if applicable); Check for minima outside of 95% CI of base model; Evaluate data conflicts and likelihood surface	Consider how prior likelihoods are included; Model specification and data weighting can impact behavior (Wang et al., 2017)
Age structured production model	Explore when there are multiple data sources, especially for compositions	Recruitment-driven models (e.g. short-lived species, low recruitment variability, and/or low exploitation history) might have poorly defined production functions; Biomass scale might be poorly informed when fishing mortality is low (Mintev Vera et al. 2022)
Retrospective analysis	Visually inspect retrospective	Rho within fixed threshold cannot rule

	<p>patterns;  Rose approach, resource permitting (Legault, 2020);  Consider post-sample rho (Breivik et al. 2023);  Consider model-specific confidence intervals for rho (Miller and Legault, 2017)</p>	<p>out parameter misspecification;  Consider time-varying processes and data weighting;  Reference points can remain biased even when retrospective patterns disappear (Szuwalski et al. 2018)</p>
Prediction Skill	<p>Consider leave-one-out cross validation, especially when time series are sparse or few</p>	<p>MASE criterion within threshold cannot rule out parameter misspecification; more research needed</p>
Recruitment trend	<p>Visually inspect recruitment deviates/calculate quantitative metrics for trend and non-randomness in the deviates (Merino et al., 2022)</p>	<p>Models with significant linear trend in the recruitment deviates are likely to be mis-specified; the <i>absence</i> of linear trend is not evidence that the model is correctly specified.</p>

1118 **8. Figures**  
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Figure 1. A) Relative error (%) in depletion for estimation models with no misspecifications. B) Survey observations from the operating model (points and lines) with 95% simulation intervals (values are summarized across all OM replicates). C) Annual recruitment in 1000s of individuals. D) Observed length compositions, aggregated across time, for the fishery and survey fleets. E) Observed age compositions, aggregated across time, for the fishery and survey fleets. In all plots, colors correspond to the recruitment variability scenarios. The shaded ribbons in A) and C) correspond to the 95% simulation interval

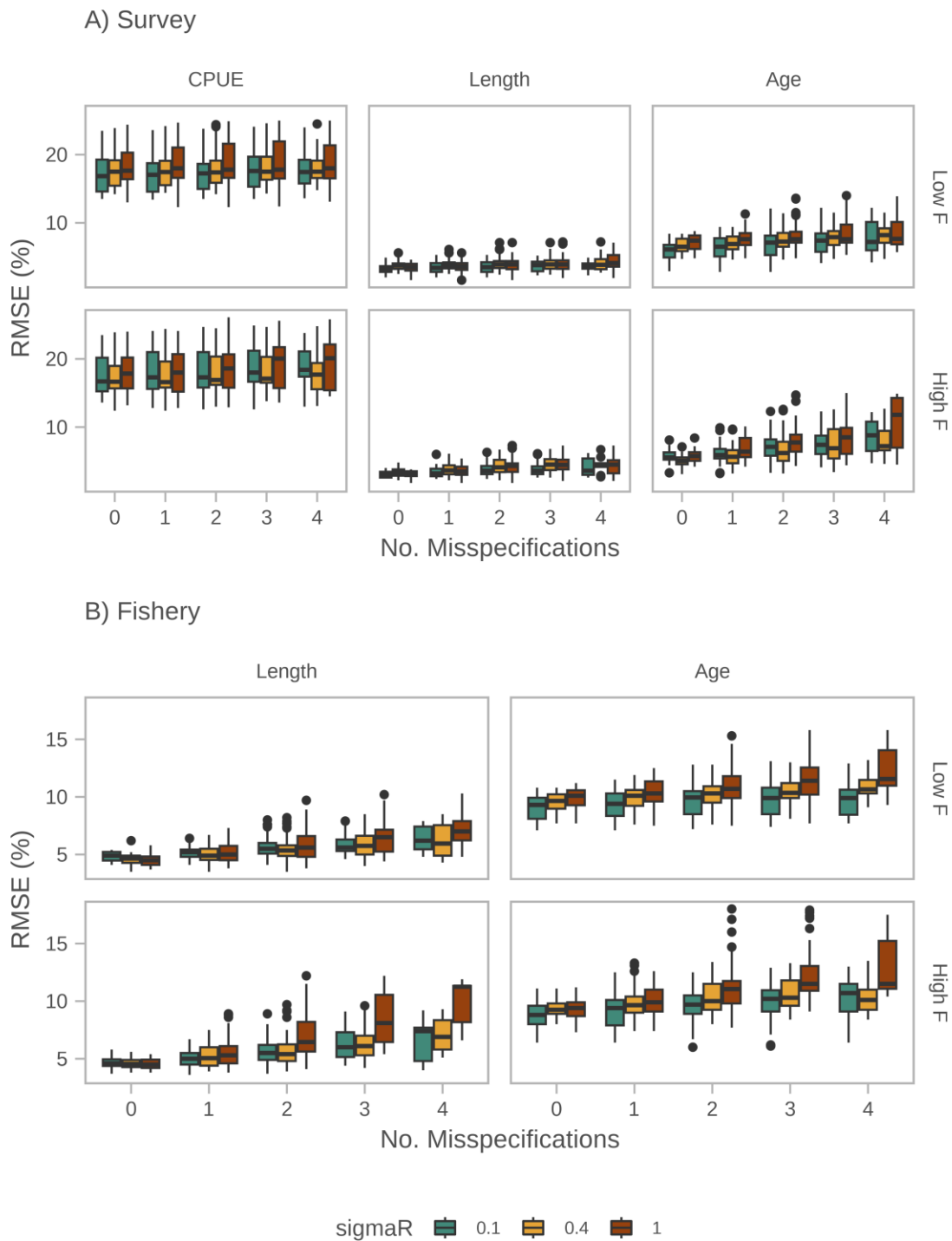


Figure 2. Boxplots of RMSE of A) survey indices of abundance, length and age composition data and B) fishery length and age composition data (bottom) for two levels of fishing mortality (rows). The x-axis represents the number of misspecifications present in the estimation method (0 mis-specifications corresponds to the correct estimator). Colors correspond to the value of recruitment variability used in the OM.



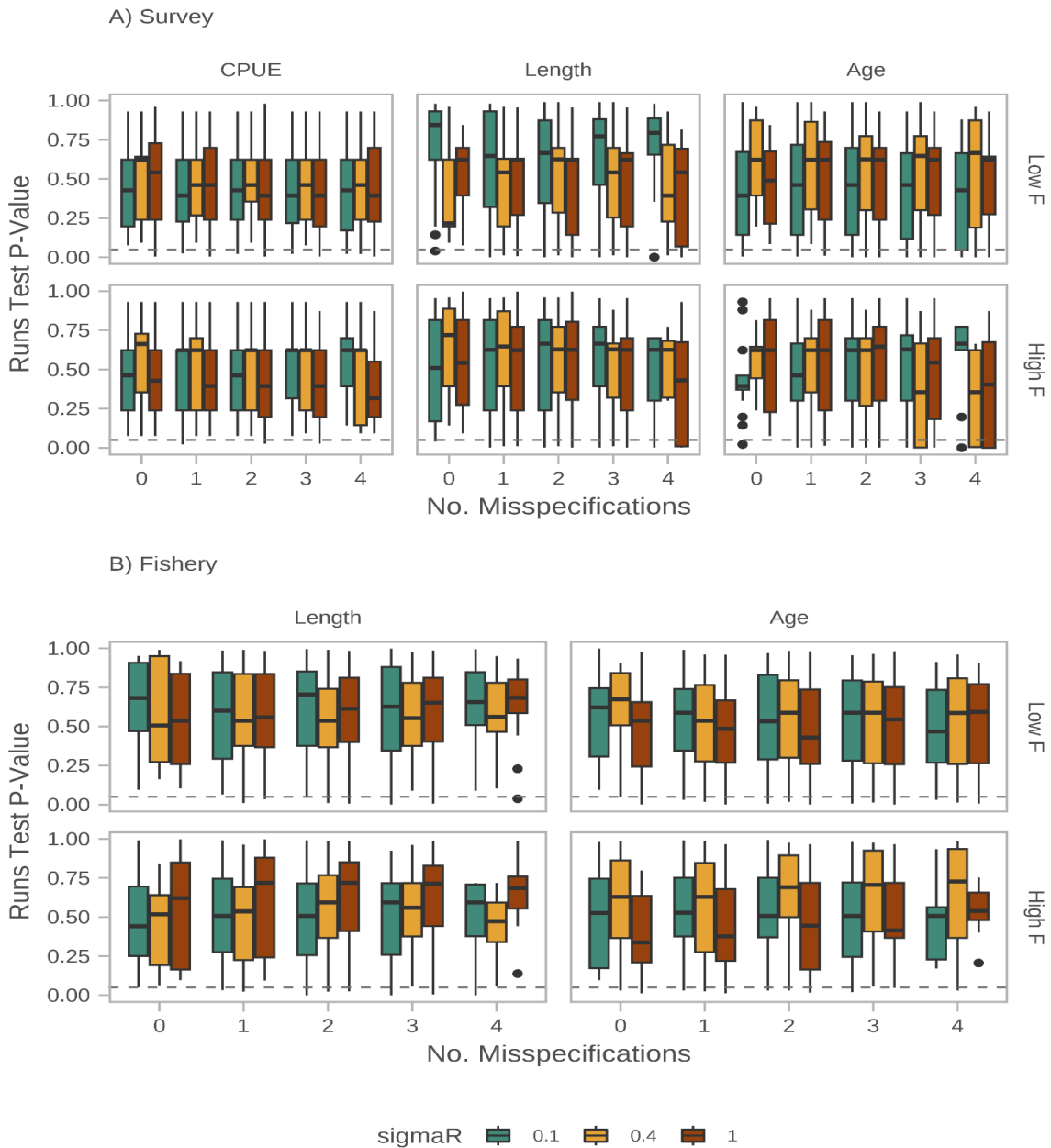


Figure 3. Box plots of  $p$ -values from Runs test for A) survey indices of abundance, length and age composition data and B) fishery length and age composition data (bottom) for two levels of fishing mortality (rows). The traditional interpretation of this test is that  $p$ -values greater than 0.05 (dashed line) indicate no evidence to reject the null hypothesis that residuals are normally distributed (thus values above the line “pass” the test). The  $x$ -axis represents the number of misspecifications present in the estimation method (0 mis-specifications corresponds to the correct estimator). Colors correspond to the value of recruitment variability used in the OM.

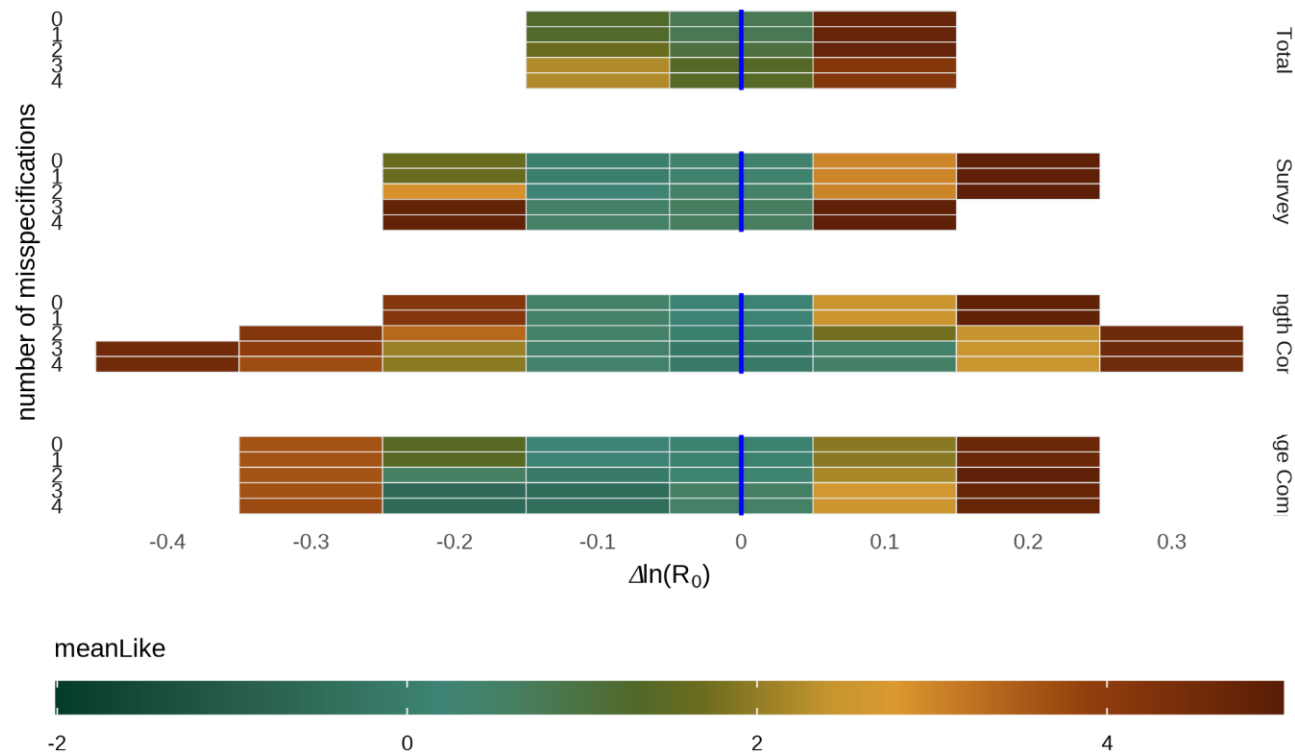


Figure 4. Likelihood profiles for  $\log(R_0)$  shown for a subset of estimation methods with zero through four misspecifications. Each panel corresponds to either the total likelihood (top), or survey, length or age composition components (bottom three panels). The x-axis has been re-centered to the corresponding MLE from the correct (not mis-specified) estimation method (vertical blue line); profiles have been filtered to only display model runs with changes in the scaled negative log-likelihood less than 10 units. Green tiles indicate models closer to the minimum negative log-likelihood; red values are higher.

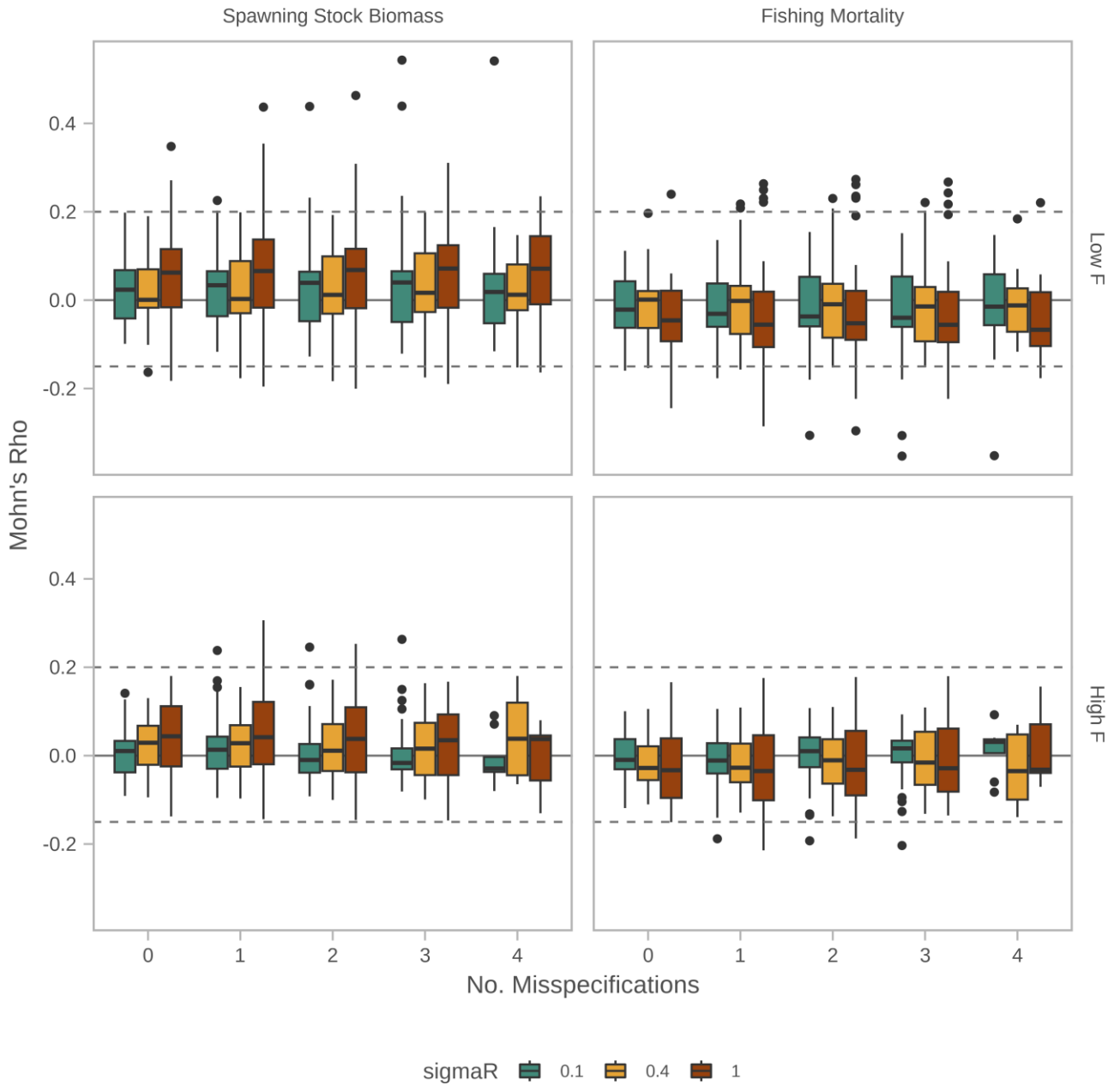


Figure 5. Boxplots of  $\rho$  from 5-year retrospectives in SSB (left) and fishing mortality (right) for two levels of fishing mortality (rows). The x-axis represents the number of mis-specified parameters in the estimation method (0 misspecifications corresponds to the correct estimator). Colors correspond to the value of recruitment variability used in the OM. Dashed lines indicate the thresholds suggested by Hurtado-Ferro et al. (2015) for poor  $\rho$  values.

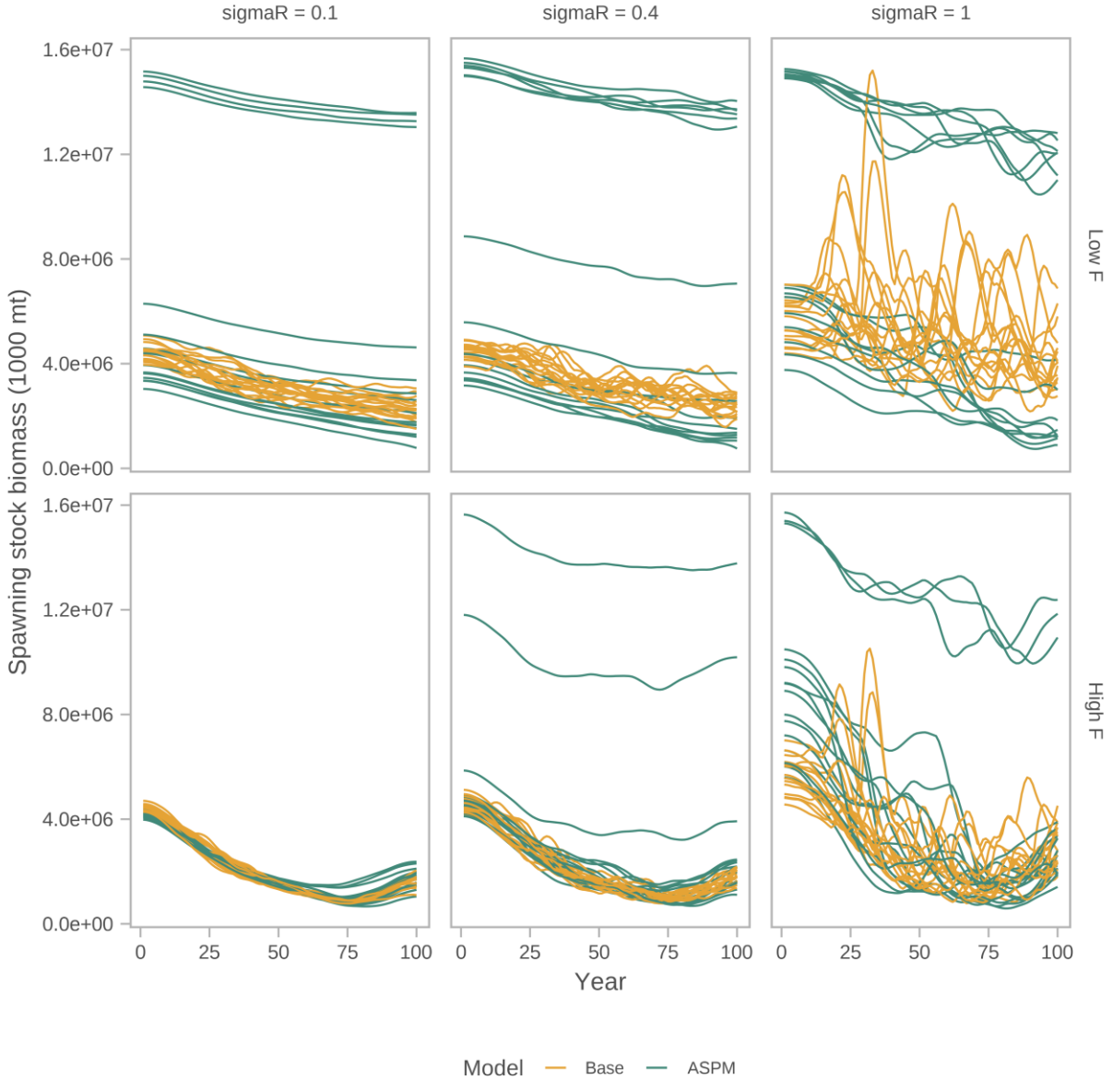


Figure 6. Estimated spawning stock biomass trajectories for correctly specified ASPM EMs (green lines) and correctly specified age-structured integrated model (yellow lines).

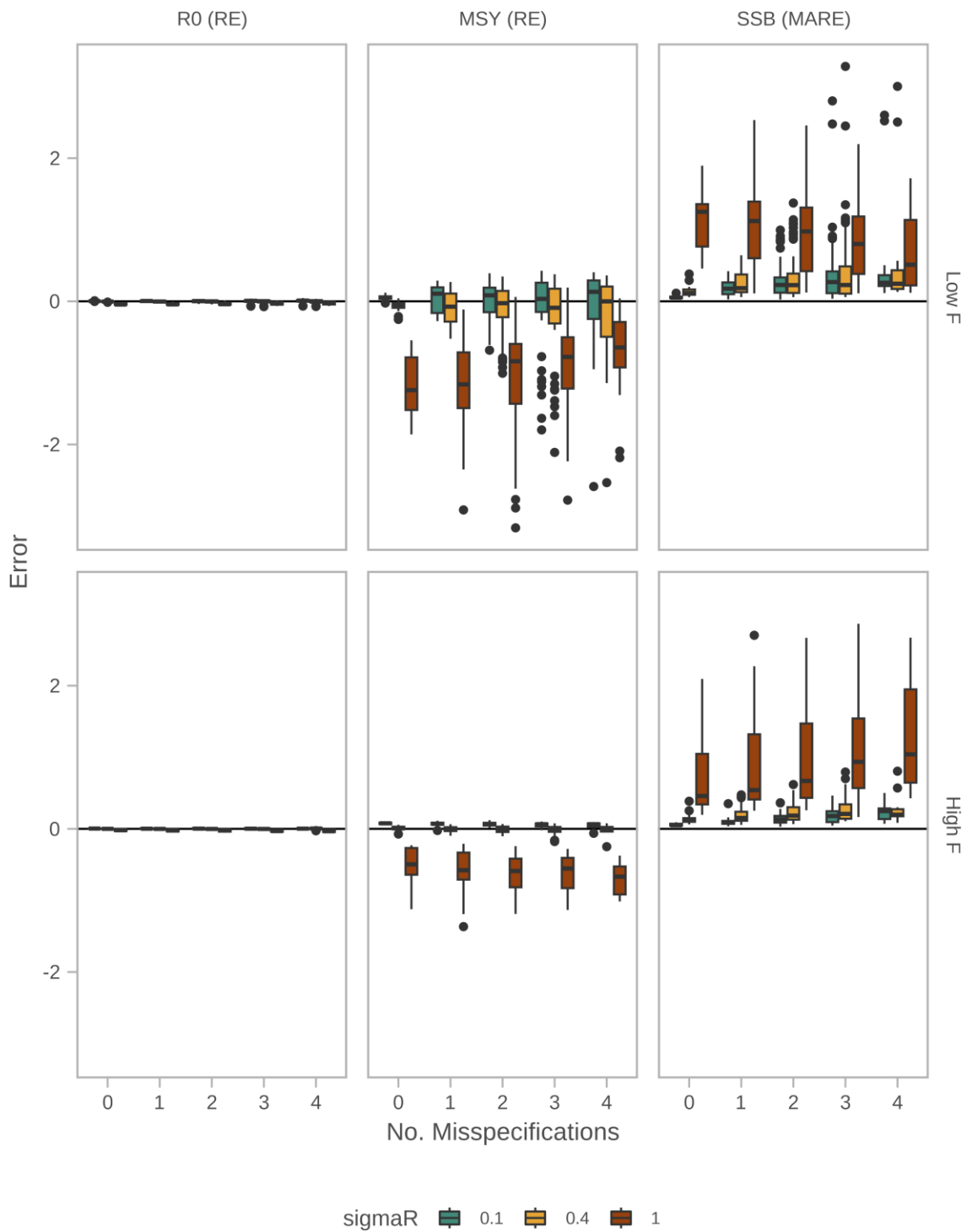


Figure 7. Boxplots of the relative difference in a deterministic recruitment model and full model of  $R_0$ ,  $MSY$ , and  $MARE$  of  $SSB$  for two levels of fishing mortality (rows). The x-axis represents the number of mis-specifications present in the estimation method (0 misspecifications corresponds to the correct estimator). Colors correspond to the value of recruitment variability used in the OM.

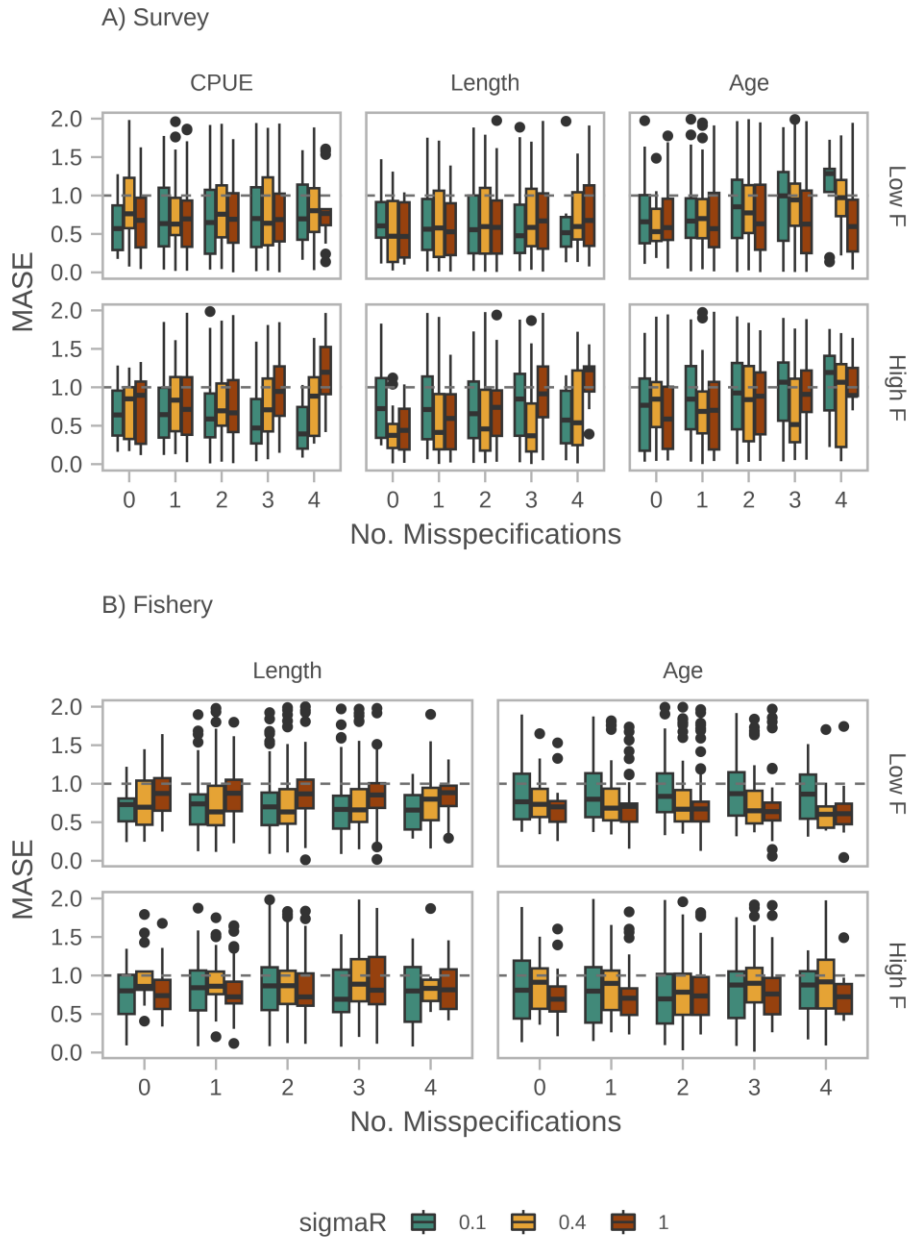


Figure 8. Boxplots of hindcast cross-validation MASE for A) survey indices of abundance, length and age composition data and B) fishery length and age composition data (bottom) for two levels of fishing mortality (rows). The x-axis represents the number of mis-specifications present in the estimation method (0 misspecifications corresponds to the correct estimator). Colors correspond to the value of recruitment variability used in the OM. MASE scores below 1 (dashed line) have greater predictive power than a null model.

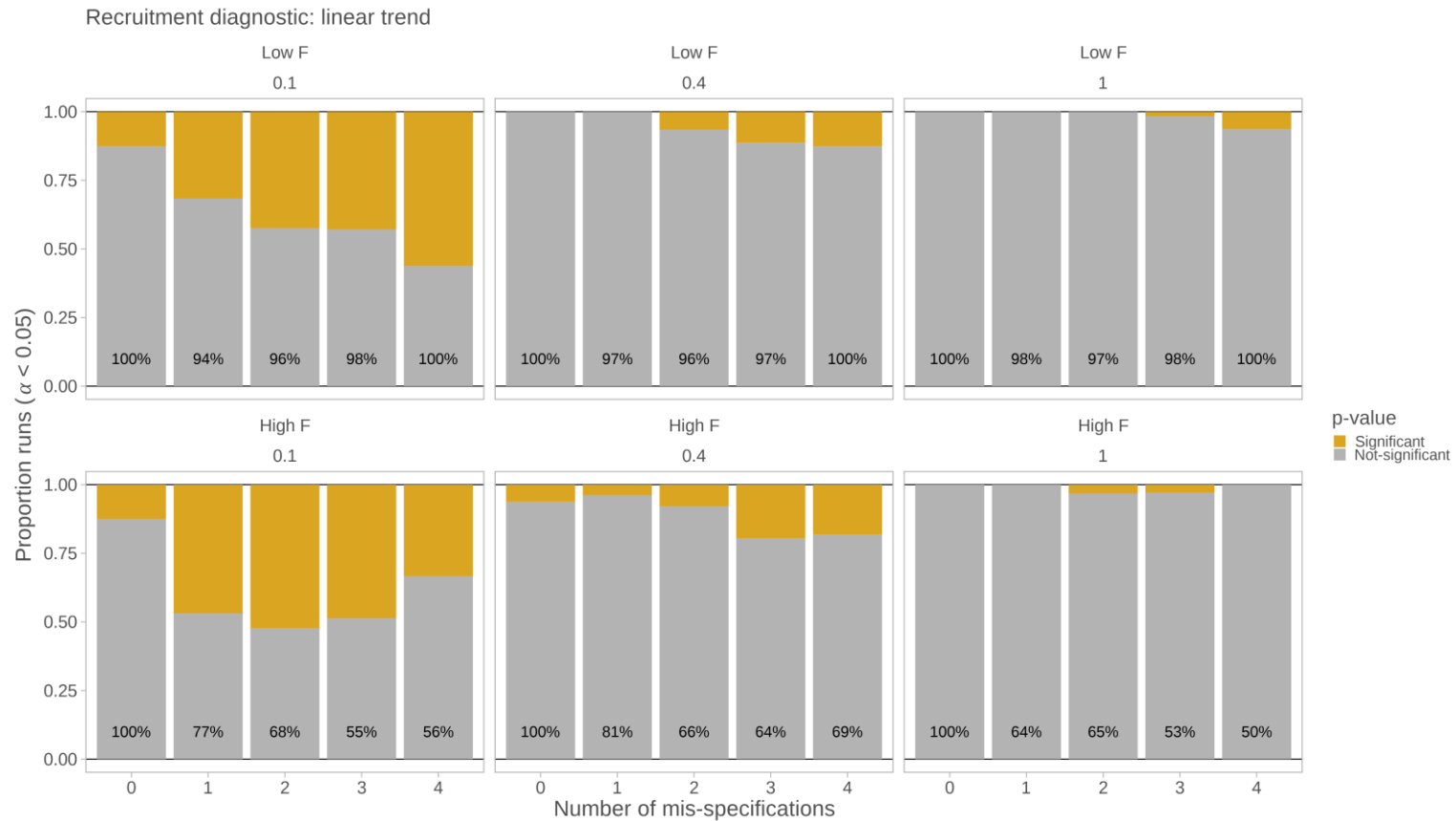


Figure 9. Proportion of converged models with significant results for tests of linear trend in the recruitment deviates by number of mis-specifications, levels of fishing mortality, and assumed recruitment variability. The percentage shown at the bottom of each bar is the percentage of converged models relative to the total number of models run.

## 9. Supplementary material

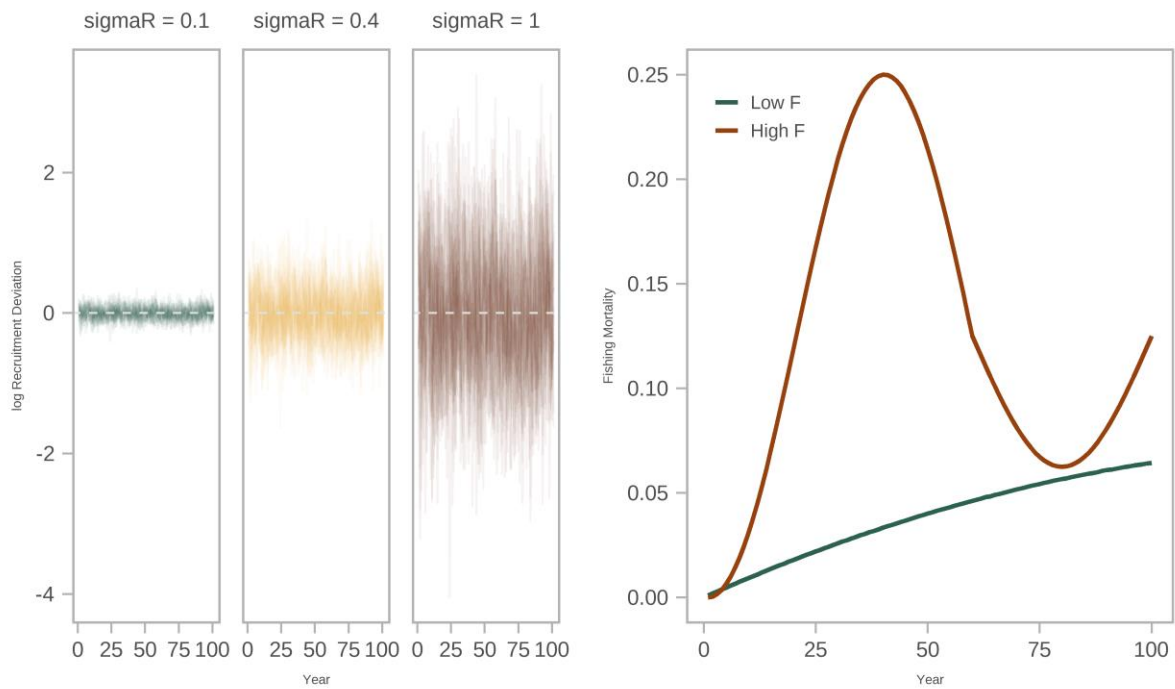


Figure S1. (Left) time series of simulated recruitment deviations at three different levels of variation. (Right) Input fishing mortality vectors for the “low” (green) and “high” fishing mortality scenario.



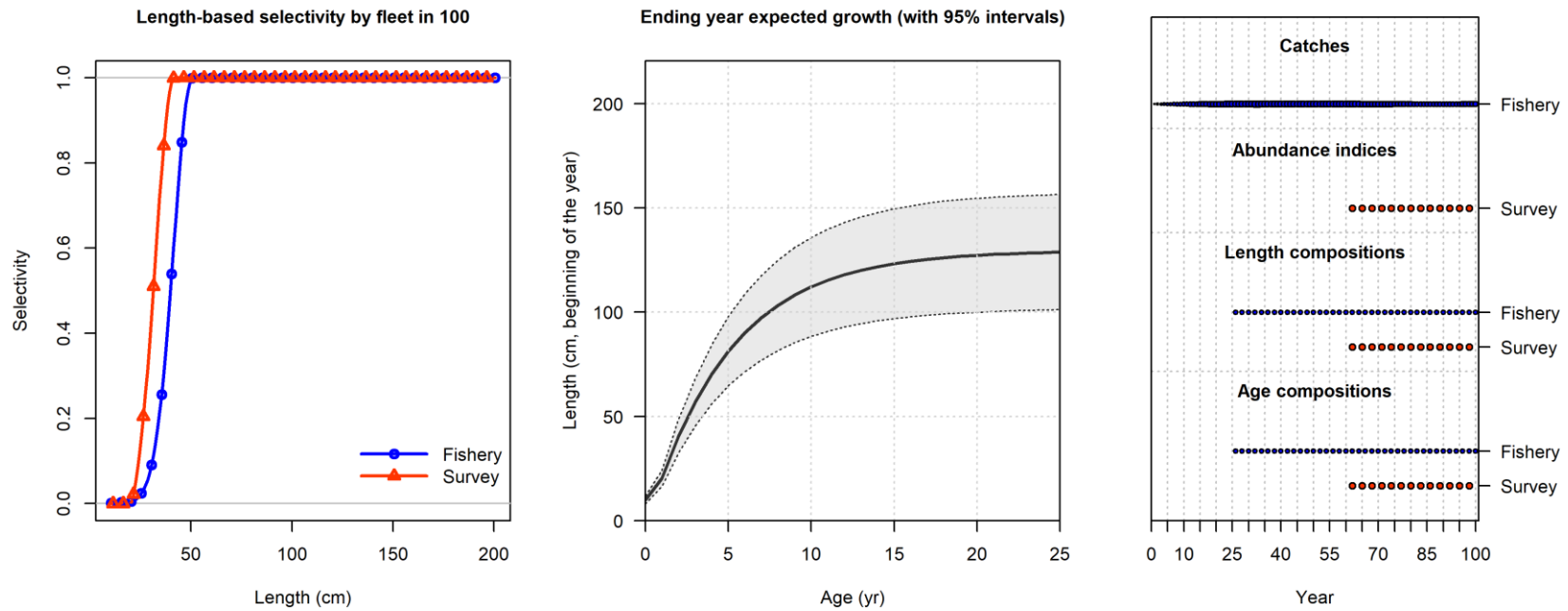


Figure S2. Operating model values for selectivity (left) and growth (center) and data availability by year for the fishery fleet and survey (right). All OM replicates and associated EMs have the same years of data available.

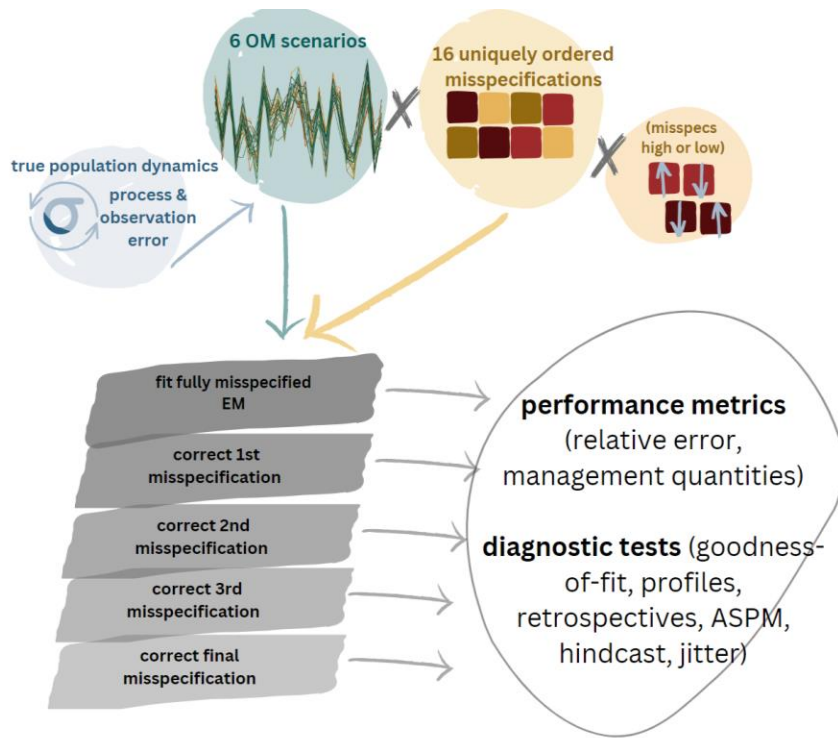


Figure S3. Schematic of the main experimental design. Operating model replicates are made by simulating datasets given process and observation error (catches are assumed known). All uniquely ordered sets of misspecifications (including parameter identity and direction of misspecification) are fit to each replicate and corrected sequentially. For all unique estimation methods, we run a suite of diagnostic tests as described in Carvalho et al. (2021).

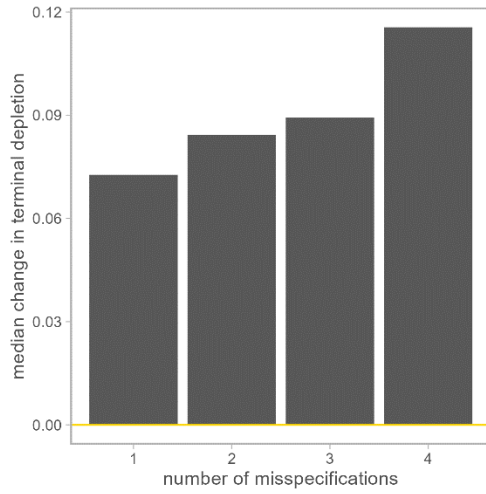


Figure S4. Median change in terminal depletion versus number of misspecifications, for converged results only. This illustrates that the error in terminal depletion values cumulatively increases with increasing misspecifications.

Table S1. Summary of mis-specified values for natural mortality, steepness, selectivity, and asymptotic length identified during the thresholding experiments for unique combinations of fishing mortality and variation in recruitment deviations. The true values are shown in OM Value, and the mis-specified values above and below the true value are shown in the last two columns.

OM Scenario	Parameter	OM Values [bounds]	Value that results in ~10% decrease in terminal SSB	Value that results in ~10% increase in terminal SSB
F=low sigmaR=0.1	Natural Mortality	0.2 [0.01,1.8]	0.188	0.212
F=low sigmaR=0.4			0.188	0.212
F=low sigmaR=1			0.188	0.214
F=high sigmaR=0.1			0.188	0.216
F=high sigmaR=0.4			0.19	0.214
F=high sigmaR=1			0.186	0.214
F=low sigmaR=0.1	Steepness	0.65 [0.2,1]	0.4615	0.988
F=low sigmaR=0.4			0.4615	0.988
F=low sigmaR=1			0.4615	0.988
F=high sigmaR=0.1			0.5915	0.715
F=high sigmaR=0.4			0.585	0.7215
F=high sigmaR=1			0.5915	0.7345
F=low sigmaR=0.1	Length at 50% selectivity	50.8 [5.08,101.6]	44.196	58.42
F=low sigmaR=0.4			43.688	58.42
F=low sigmaR=1			42.672	58.928
F=high sigmaR=0.1			43.688	59.436
F=high sigmaR=0.4			44.196	57.912
F=high sigmaR=1			40.64	62.484
F=low sigmaR=0.1	Asymptotic length	129.5 [6.6,660]	126.72	138.6
F=low sigmaR=0.4			126.72	138.6
F=low sigmaR=1			126.72	138.6
F=high sigmaR=0.1			125.4	139.92
F=high sigmaR=0.4			120.12	138.6
F=high sigmaR=1			122.76	141.24

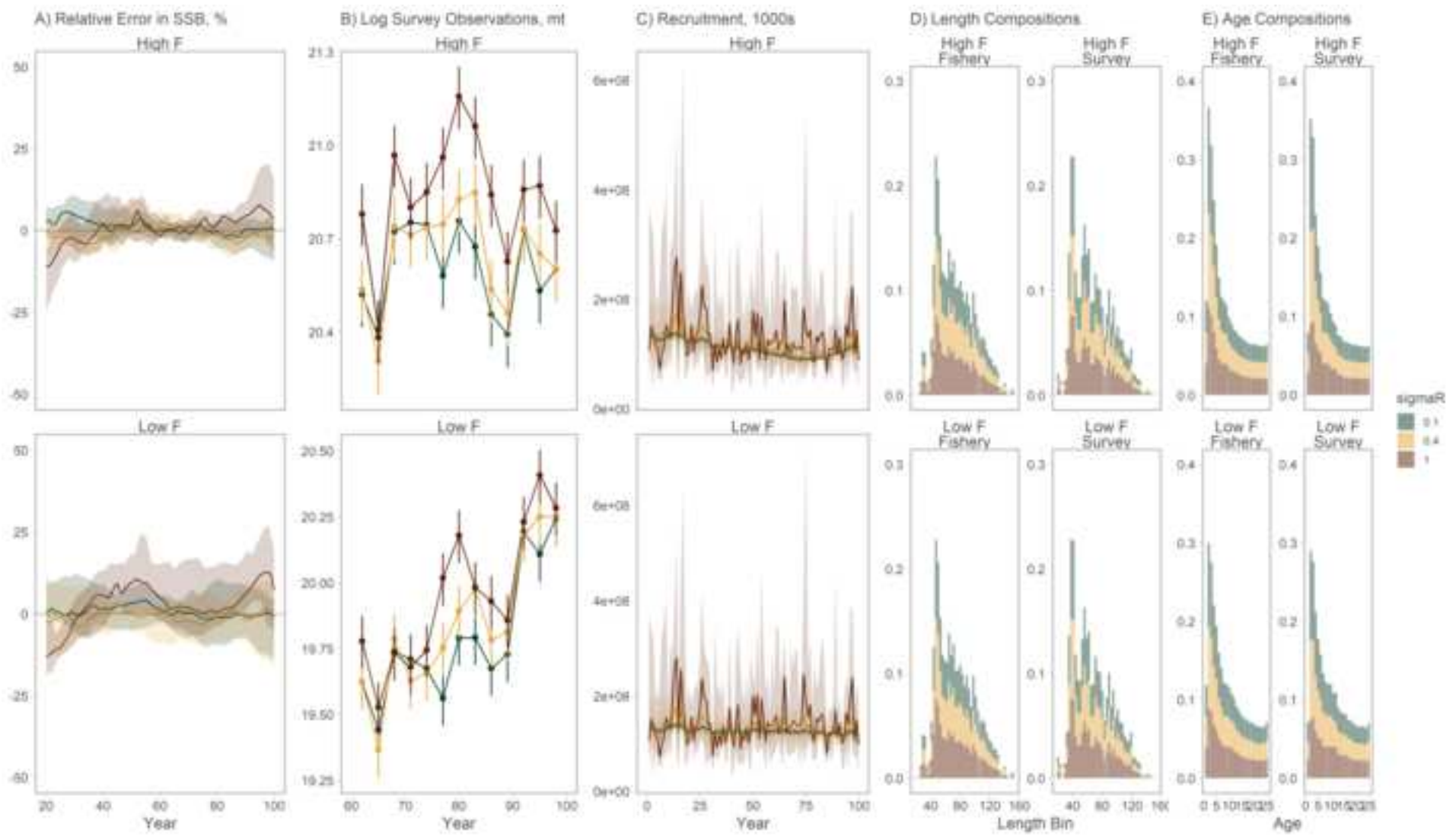


Diagnostic	sigma R = 0.1					sigma R = 0.4					sigma R =		
	0	1	2	3	4	0	1	2	3	4	0	1	2
% Converged	100	94	96	98	100	100	97	96	97	100	100	98	97
Mohn's Rho (SSB)	100	97	97	95	94	94	95	95	97	94	75	76	82
Mohn's Rho (F)	94	97	95	94	94	94	95	97	95	100	75	76	81
Hindcast Fishery Age	56	55	54	51	50	69	68	66	63	62	75	73	74
Hindcast Survey Age	56	53	45	30	12	62	61	51	40	38	62	59	58
Hindcast Survey Bio	75	58	57	51	50	50	58	49	52	44	75	73	68
Hindcast Fishery Len	88	77	76	78	81	62	68	72	74	75	62	65	66
Hindcast Survey Len	62	53	42	38	38	62	56	51	50	50	75	70	66
Runs Test Fishery Age	100	93	95	95	94	94	95	93	94	94	94	92	89
Runs Test Fishery Len	100	100	99	98	100	100	98	97	100	100	100	98	96
Runs Test Survey Age	81	80	79	78	69	100	100	96	89	75	100	97	90
Runs Test Survey Bio	100	98	93	89	94	100	100	100	100	94	94	94	94
Runs Test Survey Len	94	95	96	95	94	100	95	95	95	94	100	95	91
% Converged	100	77	68	55	56	100	81	66	64	69	100	64	65
Mohn's Rho (SSB)	100	98	98	97	100	100	100	100	100	100	100	98	98
Mohn's Rho (F)	100	98	98	97	100	100	100	100	100	100	94	95	98
Hindcast Fishery Age	62	59	63	51	44	62	62	65	61	64	81	78	74
Hindcast Survey Age	56	55	42	37	22	56	56	41	41	27	62	56	47
Hindcast Survey Bio	69	67	68	74	67	62	60	63	51	45	50	56	55
Hindcast Fishery Len	69	65	60	63	67	56	60	57	54	64	69	66	60
Hindcast Survey Len	50	47	45	43	44	69	62	65	61	55	75	73	69
Runs Test Fishery Age	100	98	94	94	100	94	98	98	100	91	88	93	95
Runs Test Fishery Len	94	94	95	91	89	100	98	98	100	100	100	100	98
Runs Test Survey Age	94	92	85	86	89	100	92	81	73	55	100	90	84
Runs Test Survey Bio	100	98	100	100	100	100	100	100	100	100	100	100	98
Runs Test Survey Len	94	96	94	91	78	100	96	94	95	100	100	93	90

1.0		
3	4	
98	100	Low F
86	81	
86	81	
78	81	
54	50	
70	75	
67	69	
62	62	
89	81	
94	94	
87	81	
94	94	
89	75	
53	50	
100	100	
100	100	
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50	25	
97	100	
97	100	
76	62	
97	100	
79	62	

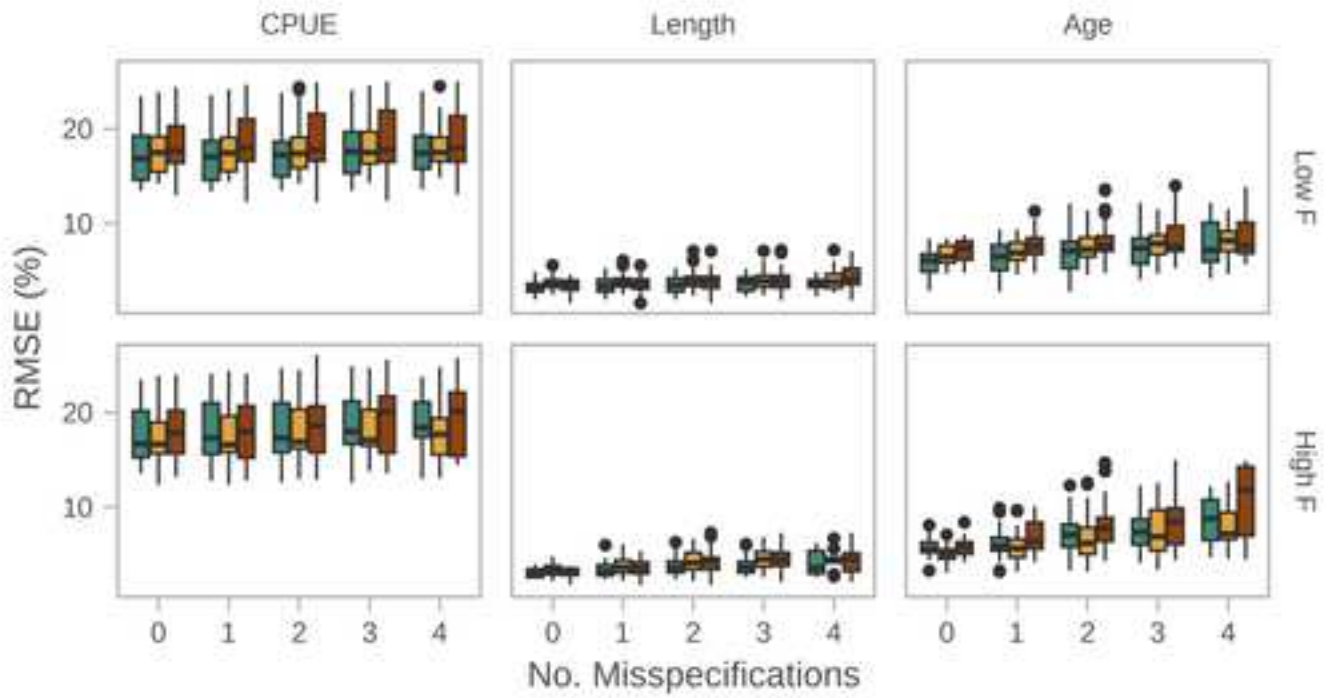
Figure 1

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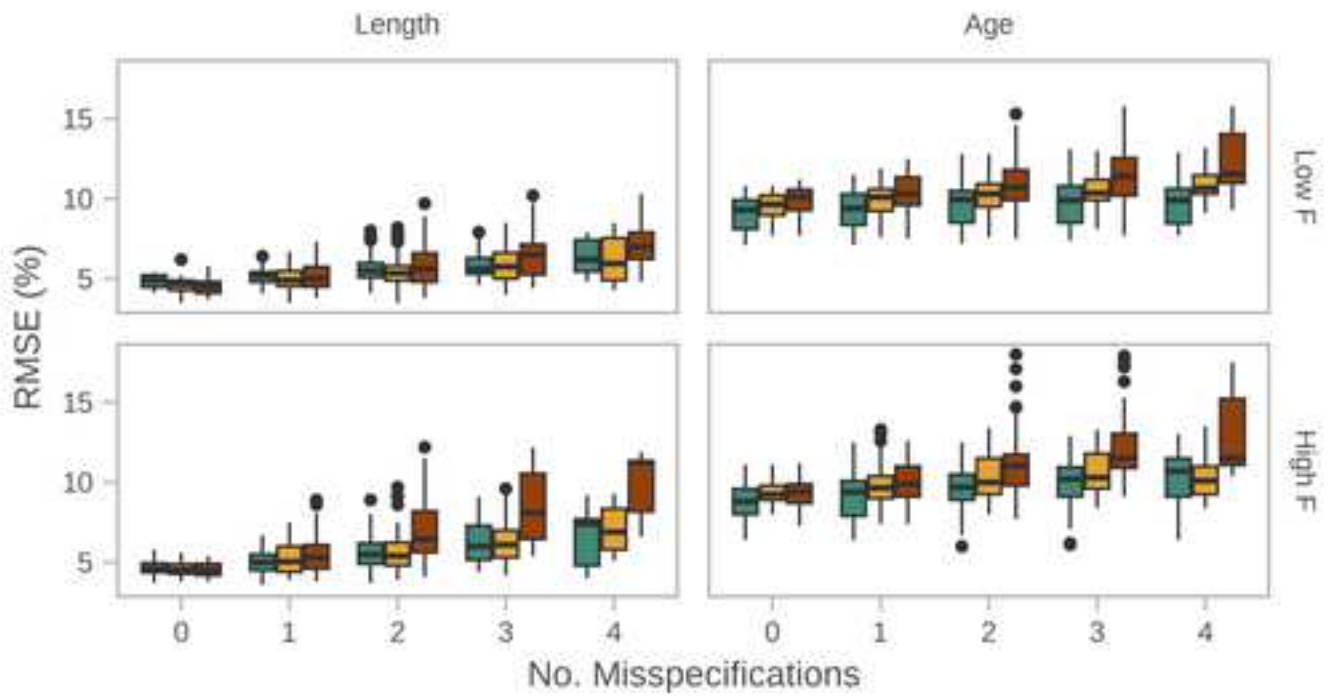




### A) Survey

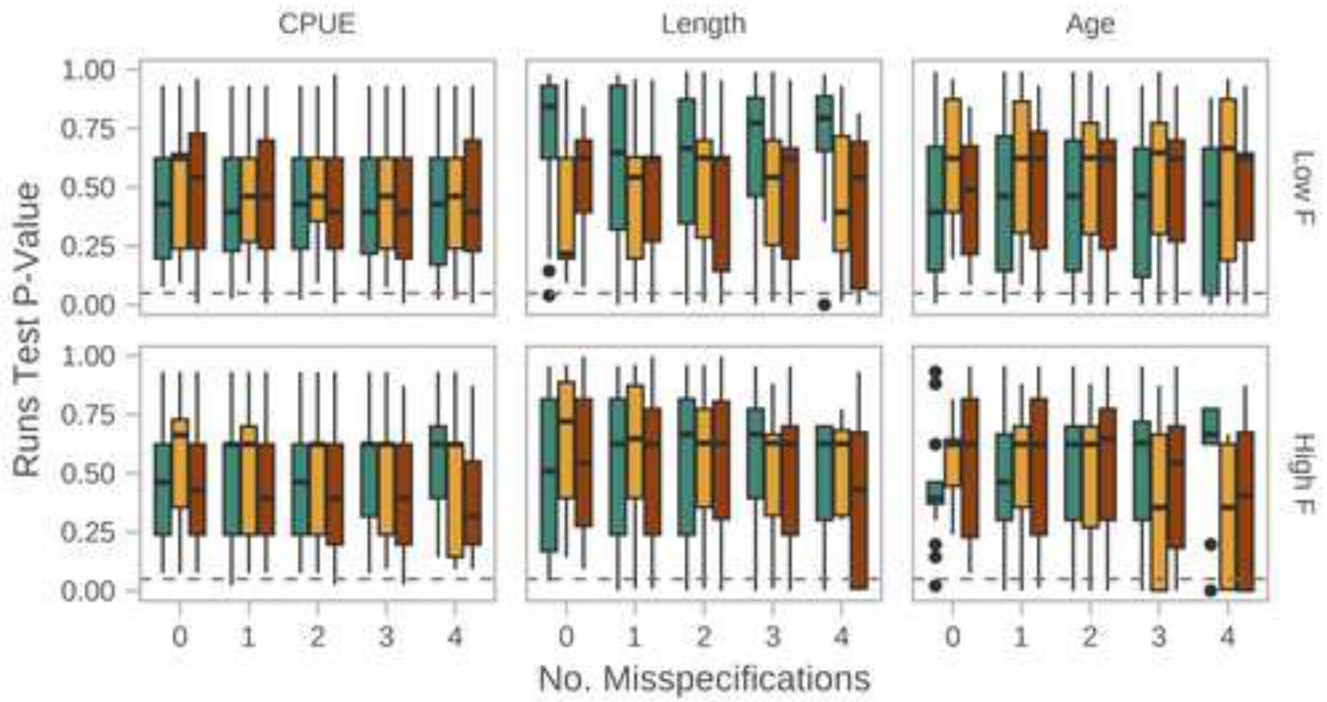


### B) Fishery

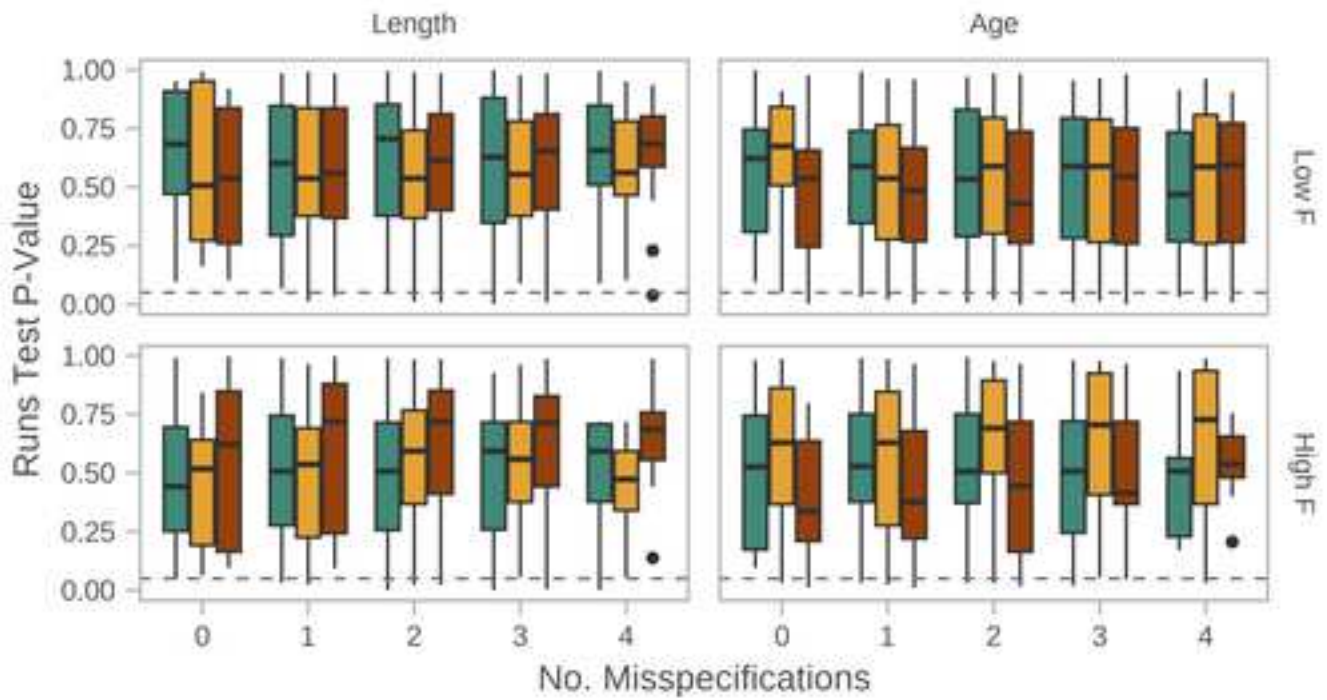


sigmaR 0.1 0.4 1

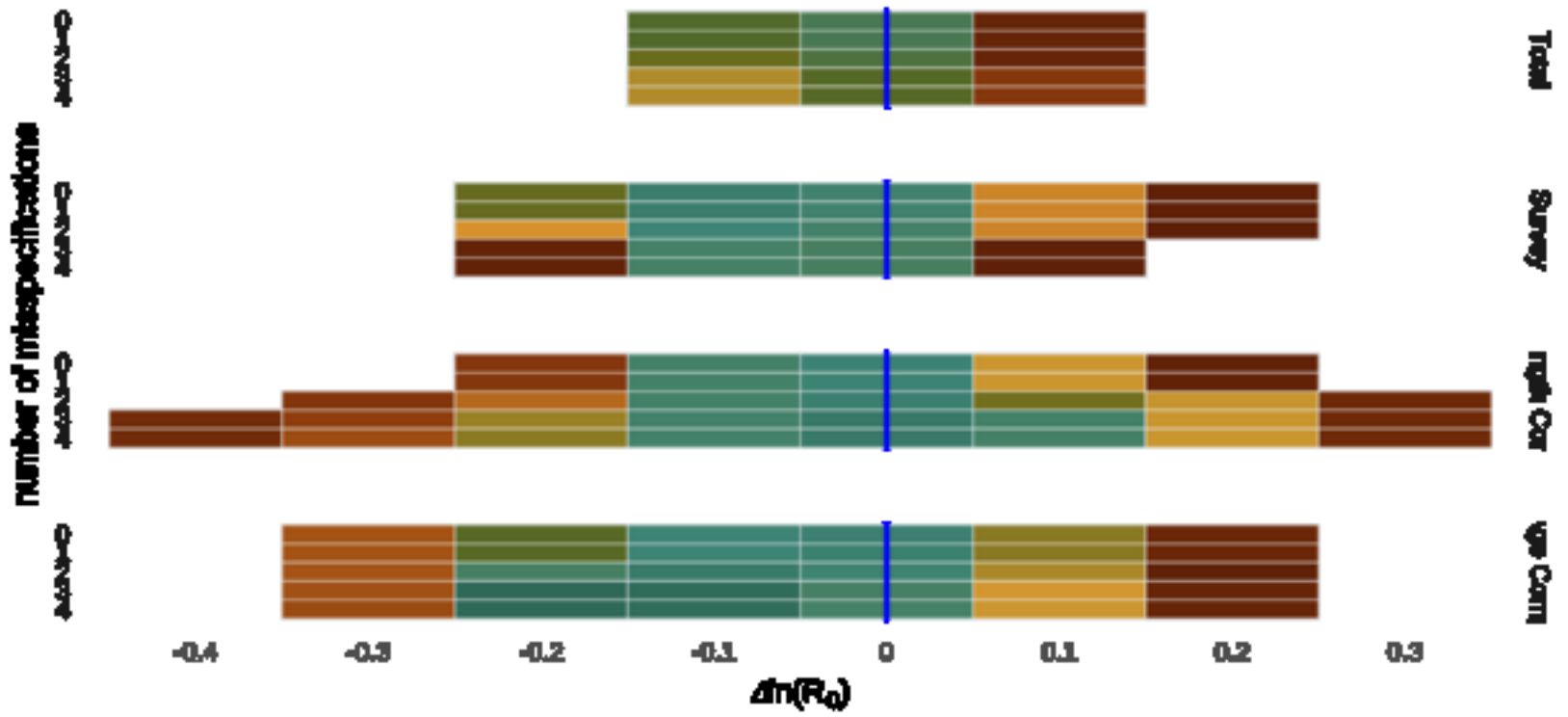
A) Survey



B) Fishery

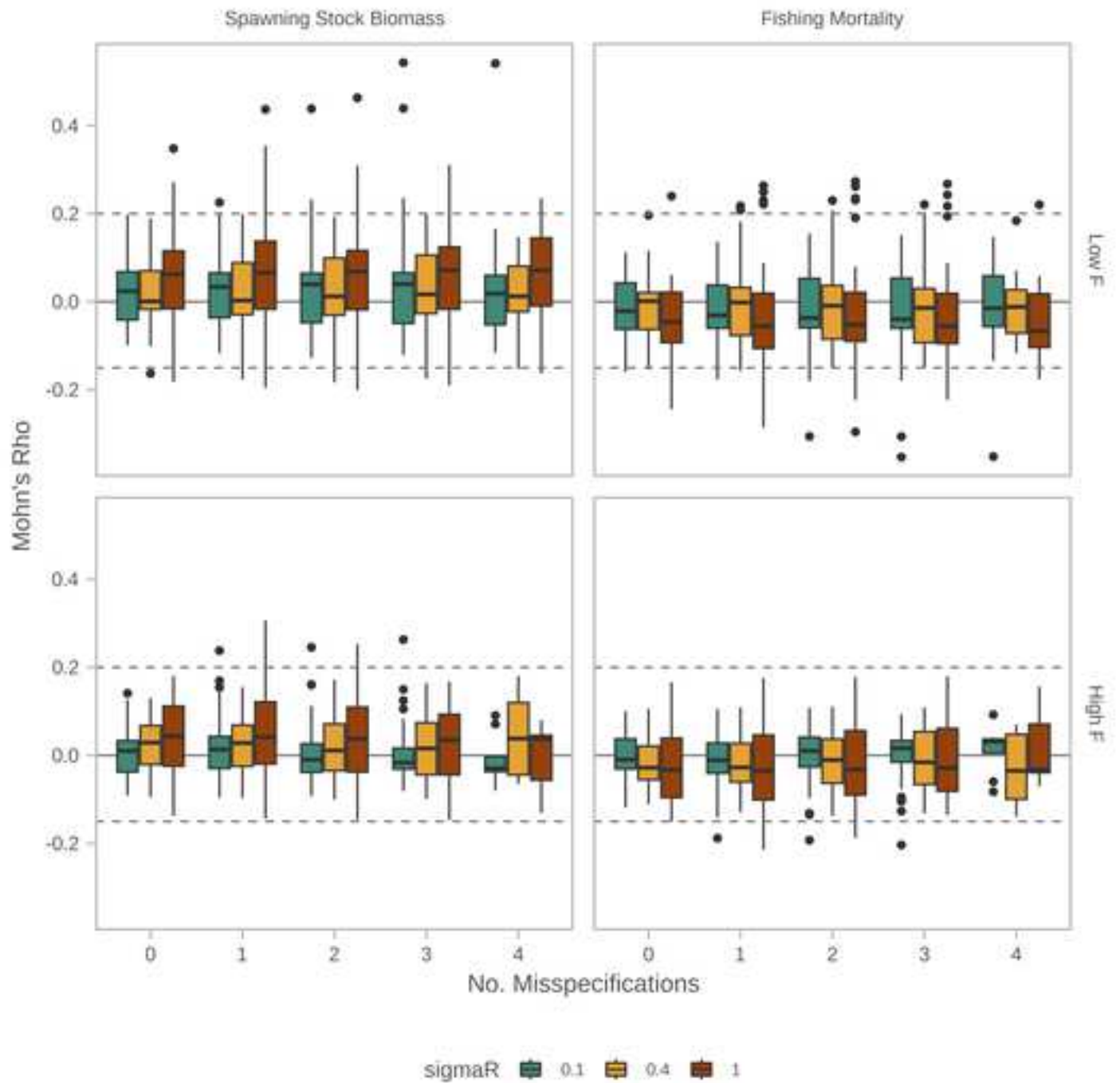


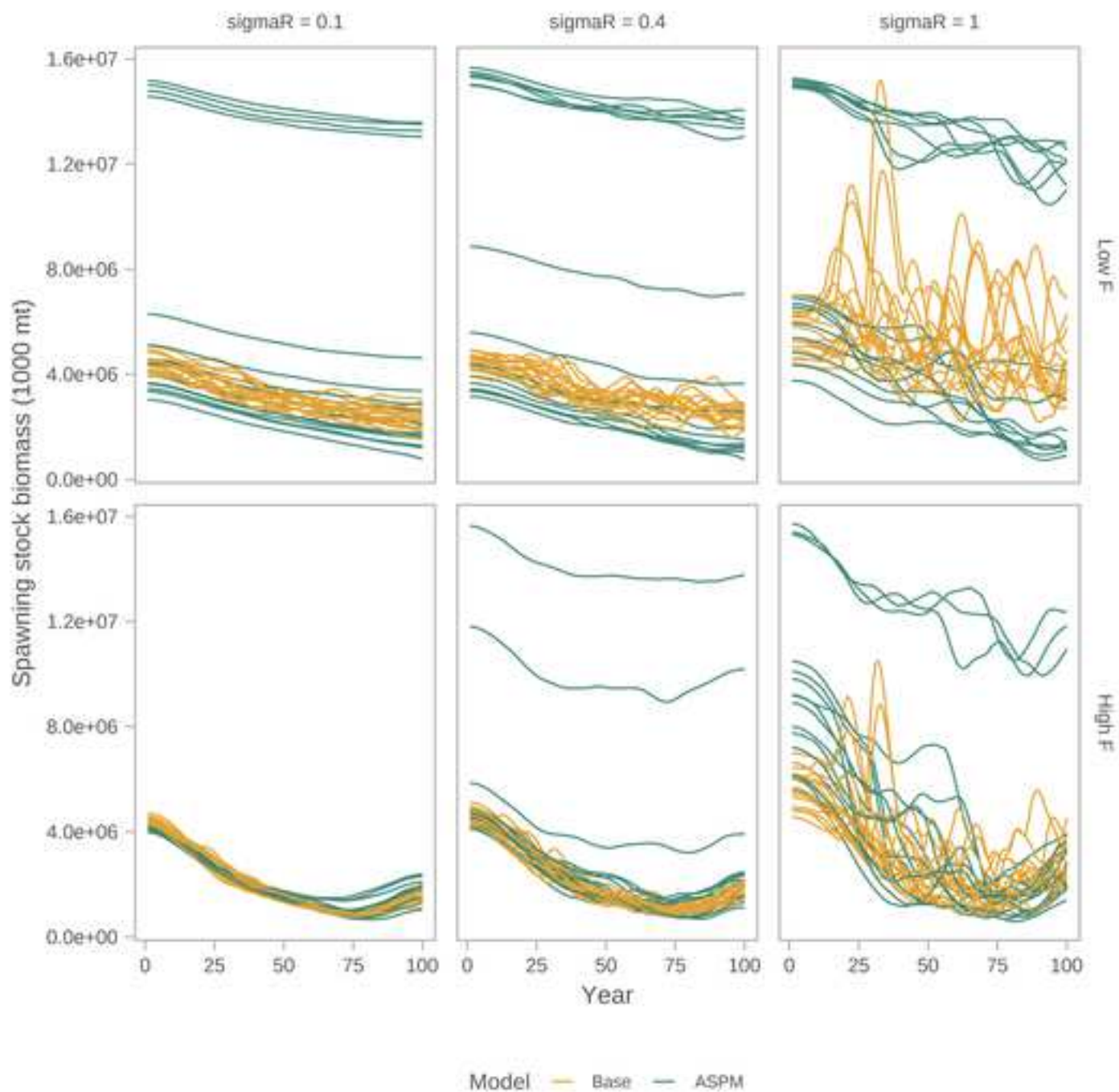
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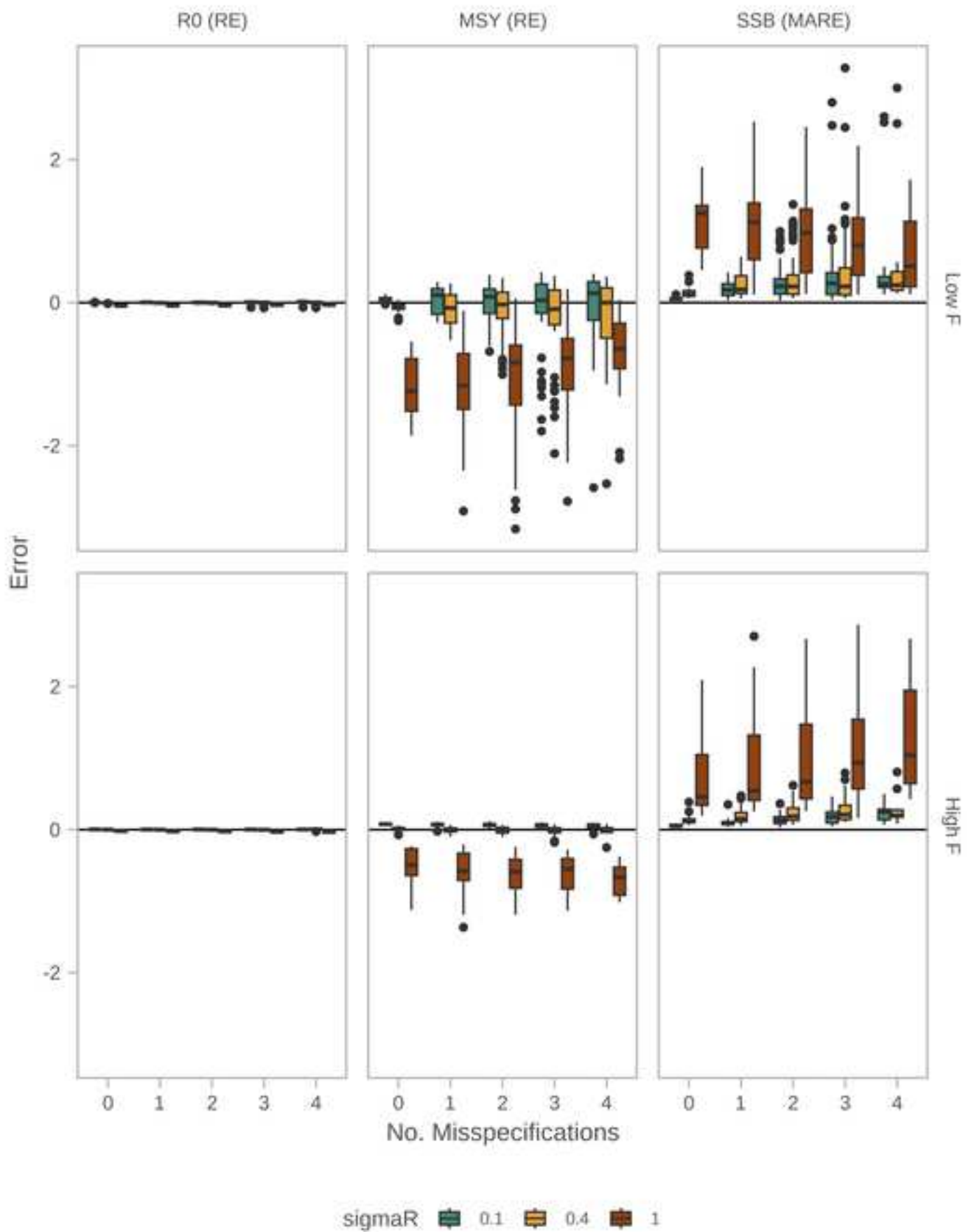


meanLike

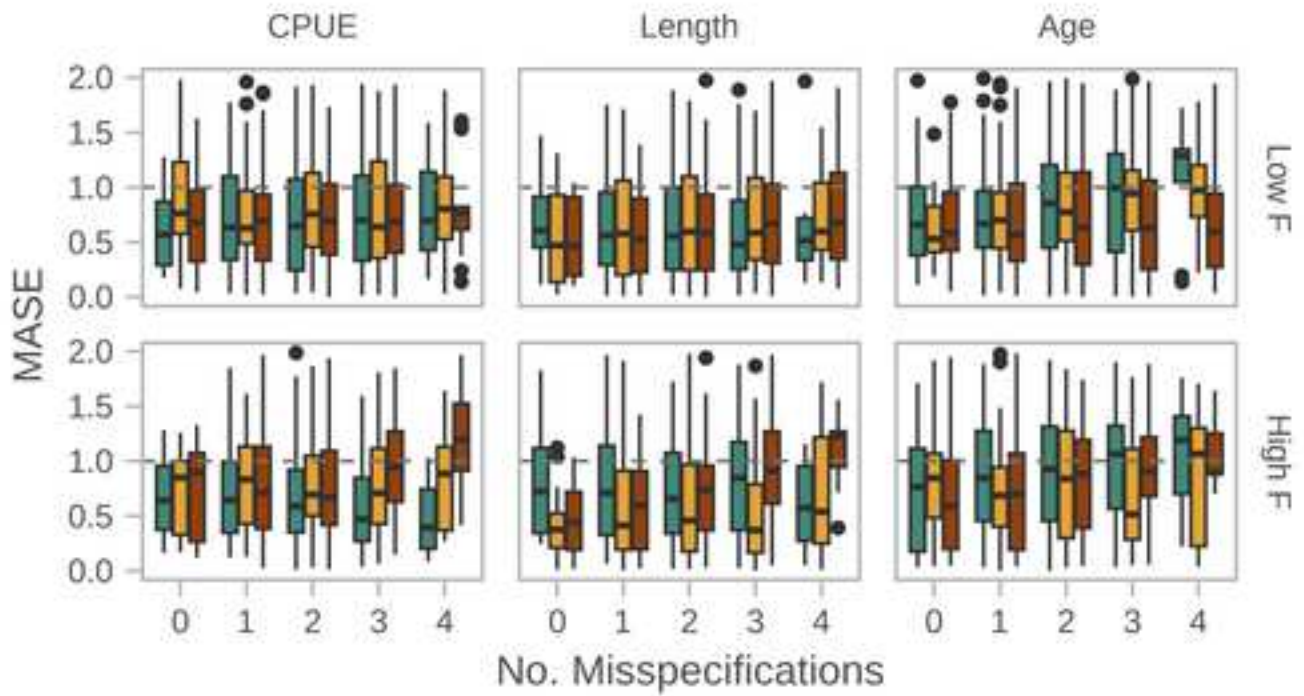




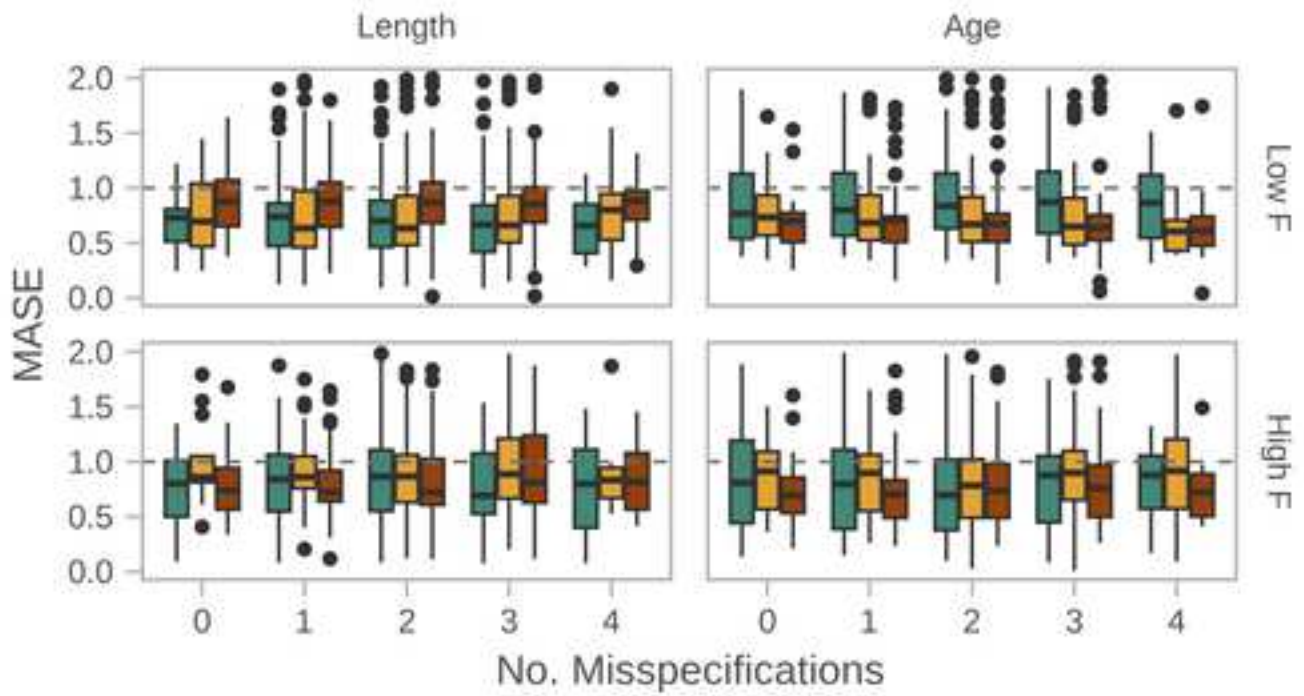




### A) Survey



### B) Fishery



sigmaR 0.1 0.4 1

