

1 Title: Evaluating the influence of press and pulse disturbances on community dynamics of the
2 Northeast US Large Marine Ecosystem
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16 Running Head: The nature of pulse and press disturbances

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18 Abstract: As climate change intensifies, there is a pressing concern as to how ecological
19 communities respond to disturbances occurring at different intensities and time scales. We
20 explored how the type of disturbance influences the dynamics of a marine community. A pulse
21 disturbance is an abrupt, high magnitude shift in conditions that can cause immediate and
22 significant impacts to an ecological community. Alternatively, press disturbances are long term,
23 multi-generational pressures acting on communities over time. The Northeast US Continental

24 Shelf Large Marine Ecosystem (NES LME) is one of the fastest warming regions in the world
25 and has experienced historic overfishing. Assemblage shifts in the NES LME have previously
26 been characterized, however, these were prior to an unprecedented pulse disturbance marine
27 heatwave event (MHW) in 2012 followed by punctuated MHWs over the last decade. We
28 quantified community change across the NES LME using a community trajectory analysis, a
29 multivariate tool that utilizes geometric analyses and comparisons of community trajectories, to
30 quantify shifts in dynamic beta diversity. We hypothesized the pulse MHWs would strongly
31 influence ecosystem structure, however, no significant impact was detected. Our analysis
32 indicates that the NES LME continues to tropicalize. However, it was not the pulse marine
33 heatwave events that seemed to drive change, but rather ecosystem overfishing and rising
34 temperatures. Here, we quantified beta diversity over time in marine communities undergoing
35 abrupt environmental changes and press disturbances. When expanded globally, this analysis can
36 compare how variable disturbance pressures may result in different manifestations of beta
37 diversity change within marine assemblages.

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40 Key words: large marine ecosystem, community trajectory analysis, ecosystem overfishing,
41 pulse disturbance, press disturbance

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47 1. INTRODUCTION

48 One of ecology's central aims is to understand how the nature of disturbances affect
49 community assemblages. Regular small disturbances can maintain diversity (sensu Hutchinson
50 1961, Connell 1978) while larger and more sudden disturbances can completely reorganize food
51 webs (Filbee-Dexter & Scheibling 2014, Gårdmark et al. 2015). Long-term, multi-generational
52 pressures on communities, such as rising ocean temperatures and fishing pressures, are defined
53 as press disturbances (Bender et al. 1984, Harris 2014). Alternatively, short-term or pulse
54 disturbances, such as marine heat waves (MHWs), disease outbreaks, and storms can have a
55 significant influence on the dynamics and interactions within communities as well (Jentsch &
56 White 2019, Raymond et al. 2022). Pulse events are abrupt, as defined by their magnitude in
57 relation to their duration and have demonstrated significant immediate ramifications for marine
58 communities and negative effects on commercially important species (Jones et al. 2018, Smale et
59 al. 2019, Von Biela et al. 2019, Diamond et al. 2020, Hillebrand & Kunze 2020). However,
60 MHWs do not always disrupt fish biomass and accelerate tropicalization, and these disturbance
61 pressures require additional research to understand (Fredston et al. 2023). Furthermore, an
62 increase in both the spatial extent and potency of press and pulse disturbances in marine
63 ecosystems has been observed across the globe as a result of increasing anthropogenic stressors
64 and climate change (Herring et al. 2015).

65 These disturbance events have pronounced impacts on the social and economic well-
66 being of fisheries-dependent people around the world (Hamilton & Haedrich 1999, He &
67 Silliman 2019, Mendenhall et al. 2020). Ecological communities provide valuable ecosystem
68 services, support global economic markets, ensure food security across the globe, and capture
69 stunning levels of biodiversity (Costanza et al. 2014, Costello et al. 2020, Trew & Maclean

70 2021). The loss of such services and the stability of a system to provide them is a legitimate
71 concern due to increasing, unprecedented rates of climate change (Amaya et al. 2023). For
72 example, in 2012 a MHW caused by a combination of atmospheric warming and oceanographic
73 change in the Gulf of Maine initiated earlier onshore migration of lobster and an increase in
74 longfin squid (Mills et al. 2013, Pershing et al. 2018). This shift proved detrimental to the lobster
75 fishery due to the market surplus and subsequent plummet in prices within the US and Canada,
76 however, some fishers took advantage of the expanded squid fishery (Mills et al. 2013). This
77 pulse event caused major disruptions to the socioeconomic systems in the Gulf of Maine, but
78 how it altered the species assemblage at both short- and long-term scales is unclear.

79 Disturbances and deviations from ideal ecological conditions are often approached
80 through a single or multi-species lens, however assemblage level assessments may capture the
81 dynamics of the whole system where higher scales of biological organization may behave
82 differently than individual components (Levin 1992, Schindler et al. 2015). Biological
83 assemblages are valuable response indicators of changes in the environment as they integrate
84 multiple drivers across trophic levels such that temporal changes in community assemblages
85 suggest system-wide ecosystem change (Legendre & Gauthier 2014, Sturbois et al. 2021). Beta
86 diversity metrics quantify compositional and abundance changes in community members
87 (Dornelas et al. 2013), inform spatial and temporal patterns of community change (Legendre &
88 De Cáceres 2013), and can be used to compare different communities to one another or the same
89 community over time (Sturbois et al. 2021). Here, we quantify changes in beta diversity,
90 community composition over time, in the marine nekton assemblages along the eastern coast of
91 the United States in the Northeast United States Continental Shelf Large Marine Ecosystem
92 (NES LME) using multivariate methods (De Cáceres et al. 2019, Sturbois et al. 2021).

93 The NES LME is an ocean warming hotspot (Pershing et al. 2015) boasting extensive
94 oceanographic and fisheries data coverage, productive fisheries impacted by intense fishing
95 (Fogarty & Murawski 1998), and unprecedented warming (Pershing et al. 2015, Le Bris et al.
96 2018). Recently there has been an increase in pulse MHW events (Mills et al. 2013, Chen et al.
97 2015, Gawarkiewicz et al. 2019) providing a unique opportunity to explore how the region has
98 responded to disturbance pressures of different time scales.

99 Within the NES LME, the exploitation of commercially and economically relevant
100 species by humans can be traced back centuries (Rosenberg et al. 2005, Grasso 2008, Boudreau
101 et al. 2017). Ecosystem overfishing occurs when the total harvest exceeds what the ecosystem
102 can support and is indicated by a decline in the total ecosystem catch, a decline in the total catch
103 per unit effort of the entire ecosystem, and/or the ratio of total landings to primary production in
104 the system surpassing established thresholds (Link & Watson 2019). Forty to fifty percent of
105 temperate and tropical large marine ecosystems exceed ecosystem overfishing thresholds (Link
106 & Watson 2019). Even with sustainable harvest, ecosystem overfishing can occur if there are
107 declines in primary productivity due to disturbances. Combining these variables captures the
108 important interaction between long term fishing pressures and shifts in primary productivity that
109 may contribute to shifts in community composition.

110 Multi-decadal warming and shifts in oceanographic conditions play an important role
111 within the NES LME where many of the ecological changes observed in the last 60 years can be
112 explained by climate variability (Lucey & Nye 2010, Nye et al. 2011, Nye et al. 2014, Pershing
113 et al. 2015). Long-term warming has been expressed in species range shifts in response to
114 shifting temperature conditions (Nye et al. 2009, Pinsky et al. 2013, Mills et al. 2024). For
115 example, species rarely seen in the Gulf of Maine such as longfin squid, black sea bass, and blue

116 crab were commonly sighted in 2012 (Johnson 2015). This region is greatly affected by basin-
117 scale climate variability due to the convergence of the Gulf Stream and Labrador Current
118 systems (Greene et al. 2013, Lotze et al. 2022). Recent oceanographic alterations, such as the
119 increased occurrence of warm core rings since 2000 (Gangopadhyay et al. 2019) and shifts in the
120 Gulf Stream in 2008 (Gonçalves Neto et al. 2021), raise concerns about their impact on species
121 assemblages. Increasing water column stratification and changes in nutrient concentrations (e.g.
122 silicate) (Pershing et al. 2005, Ouellet-Bernier & de Vernal 2018) have also facilitated shifts in
123 phytoplankton and zooplankton communities of the NES LME (Morse et al. 2017, Pershing et al.
124 2021, Lotze et al. 2022, Pershing & Kemberling 2024). It is unclear how these recent changes in
125 oceanographic conditions have impacted the species assemblage; however, implications for the
126 recruitment of commercially important species (Perretti et al. 2017) and range compression with
127 changes in prey quality and availability have already been documented (e.g. the North Atlantic
128 right whale, Record et al. 2019, Meyer-Gutbrod et al. 2021). These alterations in zooplankton
129 populations may disrupt food web pathways (Pershing et al. 2021), with their size ratio
130 potentially signaling broader community-level responses under disturbance pressures.

131 As anthropogenic pressures increase, it is crucial to determine whether pulse MHW
132 events have a significant effect on the ecosystem or if gradual warming or other slow, but
133 persistent drivers are more important. The NES LME, spanning from the Gulf of Maine to Cape
134 Hatteras, is categorized into four ecological production units (EPUs) based on variations in
135 physiographic, oceanographic, and biotic conditions (Lucey & Fogarty 2013). With persistent
136 high mean temperatures prevalent in the NES LME, we anticipate the gradual tropicalization of
137 the region's EPUs demonstrated in previous studies will continue (Lucey & Nye 2010). Herein,
138 tropicalization refers to the current assemblage becoming more similar to the historic species

139 assemblage of the adjacent southern EPU (Lucey & Nye 2010). However, our main goal is to
140 understand drivers of change and magnitude of the community response. First, we predict the
141 pulse MHW events will greatly disrupt community-level dynamics, as has been illustrated for
142 individual species and populations (Mills et al. 2013, Pearce & Feng 2013, Frölicher &
143 Laufkötter 2018, Amaya et al. 2023). Second, we expect that the extreme 2012 MHW and
144 subsequent events in 2016 to 2017 caused a distinct impact on the marine community of the NES
145 LME due to prolonged exposure to anomalous temperatures. Third, we hypothesize that the slow
146 press of long-term warming and fishing pressure in the NES LME have significantly influenced
147 community composition. An increase in fishing effort relative to the amount of productivity,
148 captured by the Fogarty Ecosystem overfishing index, can result in adjusted energy flow patterns
149 within the system because of declines in cumulative production, biomass of apex predators, and
150 mean trophic level which can signal community level change (Pauly et al. 1998, Libralato et al.
151 2008, Link et al. 2015). Specifically, we anticipate the Fogarty index, as an indicator of both top-
152 down and bottom-up pressures, may explain variation in community structure in the NES LME.
153 Our approach considers community change within the context of short term disturbance events
154 and long term ecosystem change and demonstrates quantitative approaches that can provide
155 novel insights into drivers of ecological change.

156 2. MATERIALS & METHODS

157 This study investigates shifts in dynamic beta diversity, changes in community
158 composition over time, within the NES LME regional EPUs over the last 50 years. To test the
159 regional variation in community change during the study period we used a community trajectory
160 analysis (CTA), a multivariate tool that extends the analysis of community dissimilarity matrices
161 commonly used by community ecologist. The CTA calculates distances between assemblages

162 (trajectory segment lengths) and the direction of change (trajectory angles) using geometric
163 analyses of community composition over time (De Cáceres et al. 2019). Next, a redundancy
164 analysis was computed to identify the variables associated with community composition in the
165 NES LME through the study period. Lastly, a cross-correlation analysis was used to explore the
166 community-level implications of the intense pulse MHWs.

167 2.1 STUDY AREA

168 The NES LME, defined as the area from the Gulf of Maine to Cape Hatteras, is divided
169 into four EPU's classified by patterns in physiographic, oceanographic, and biotic conditions
170 (Lucey & Fogarty 2013). Starting in the north, these include the Scotian Shelf, Gulf of Maine
171 (GOM), Georges Bank (GB), and the Mid-Atlantic Bight (MAB)(Figure 1A). We analyzed data
172 from the GOM, GB, and MAB over the years 1969-2019; because of gaps in spatial and
173 temporal coverage of surveys in the Scotian Shelf, this region was not included in our analyses.

174 2.2 SURVEY DATA

175 In this analysis, the EPU's of the NES LME represent distinct ecological communities.
176 We calculated the mean stratified biomass for species in each EPU from the Northeast Fisheries
177 Science Center's fall bottom trawl survey (Grosslein 1969, Azarovitz 1981). We used only
178 offshore strata within the distinct EPU's that were sampled annually, for a total of 50 years
179 included in the analysis. The year 2017 was dropped from the MAB due to ship issues leading to
180 low coverage. The survey uses a stratified random design to sample fish and macroinvertebrates
181 (Politis et al. 2014). We used only species that composed at least 0.01% of the total biomass in at
182 least one EPU from 1969 to 2019; e.g., a species that met the threshold in the MAB but not in the
183 GOM was included for all ecoregions. This provided a total of 76 finfish and invertebrate species
184 (Table S1).

185 We used mean stratified biomass (kg per tow) rather than abundance (numbers per tow)
186 because it is a better measure of energy within the system and considers not only changes in
187 numbers of individuals, but the weight of those individuals. All analyses were performed using R
188 Statistical Software (v4.1.2; R Core Team 2021). Mean stratified biomass observations for each
189 species within the separate EPU were extracted via the “survdat” R package (v1.0, Lucey &
190 Beet 2015).

191 2.3 COMMUNITY TRAJECTORY ANALYSIS

192 We conducted a CTA on the mean stratified biomass of the finfish and
193 macroinvertebrate assemblage comprised of 76 selected species over a period of 50 years to
194 explore variability in community trajectories (De Cáceres et al. 2019) (Table S1). A dissimilarity
195 matrix calculates the dissimilarity between two assemblages in space or time using biomass of all
196 species at each site. Here we quantified this dissimilarity with the Hellinger distance coefficient,
197 which computes the Euclidean distance on square root transformed species relative biomass
198 (Legendre & Gallagher 2001). A Hellinger transformation on these data places little to no
199 emphasis on zeros and rare species, is the recommended transformation for evaluating
200 compositional resemblances over time, and fulfills metric and Euclidean properties (Legendre &
201 Gallagher 2001, Legendre and De Cáceres 2013, Sturbois et al. 2021, Toumi et al. 2023).

202 Two key geometric properties of the CTA that can be used for comparison are the
203 segment length and direction of the trajectories between community states over time and space.
204 The trajectory segment length is representative of the distance between the two observation
205 endpoints in the multidimensional space of the analysis; it illustrates how dissimilar the two
206 consecutive state observations are (De Cáceres et al. 2019). The traditional trajectory segment
207 length measures interannual differences in year-to-year community state observations (see De

208 Caceres et al. 2019 and Sturbois et al. 2021 for detailed formulas). We also computed segment
209 lengths relative to the starting assemblage, which extracts the distance calculated when
210 comparing the current year's community state to the initial community state (Sturbois et al.
211 2021). Segment lengths relative to the starting assemblage visualize how far away from the
212 reference condition the community becomes over the time series. Here, the reference condition is
213 defined by our community observations in 1969. Comparatively, the traditional trajectory
214 segment lengths quantify the magnitude of year-to-year changes in community state.

215 The CTA also captures change in community state by its directionality. The value of the
216 angle within the CTA falls in a range of 0 to 180 degrees, which is the measured angle of the
217 vector created between the first and second observation and the second and third. In a
218 hypothetical community of three species, if 0, the angle represents that the community is
219 changing in the same way over time, e.g. in both time step 1 and time step 2 species A increases,
220 species B decreases, species 3 stays the same. If 180, then the orientation of the vectors is the
221 same, but is opposite "in sense", e.g. now in time step 2 we would see species A decrease,
222 species B increase, and species C decrease (see De Cáceres et al. 2019 and Sturbois et al. 2021
223 for detailed formulas). When there is a high angle between two consecutive segments, that means
224 an abrupt change in direction has been detected.

225 To compute the CTA and extract trajectory segment lengths, segment lengths relative to
226 the starting assemblage, trajectory angles, and ordination diagrams we used the "ecotraj"
227 package (v.0.0.1, methods per De Cáceres et al. 2019, Sturbois et al. 2021). The trajectory
228 segment lengths and segment lengths relative to the starting assemblage for each individual EPU
229 were extracted for the entirety of the sampling period to determine the magnitude of change in
230 community state from year to year and relative to the reference state in 1969. Similarly, the

231 trajectory angles for each individual EPU were extracted for the entirety of the sampling period
232 to provide the directional change in community state.

233 2.4 COVARIATES

234 We compared a series of covariates to the changes the NES LME finfish and
235 macroinvertebrate assemblage experienced. The selection of covariates set out to provide proxies
236 for pulse disturbances and long-term pressures within the system in three categories: biological,
237 physical, and socio-economic variables. A description of all covariates is displayed in Table 1.
238 These covariates included annual observations of: zooplankton size ratio, the mean Gulf Stream
239 index position, bottom temperature anomalies, MHW cumulative intensity, and the Fogarty
240 ecosystem overfishing index. Ecosystem overfishing represented a major socio-economic driver
241 based upon carrying capacity limits to production (Link 2021). When overfished, the system's
242 catch level results in a decline of catch per unit effort which can result in the systemic depletion
243 of the ecosystem's resources. Oppositely, continued ecosystem underfishing can suggest excess
244 productivity or the possible underutilization of the fishery that can be further explored (Link &
245 Watson 2019). Annual variable observations were extracted via the "Documentation of
246 Ecosystem Indicator Reporting" package "ecodata"(v.0.1.0, Bastille & Hardison 2018) except for
247 the zooplankton size ratio data. Due to constraints in MHW data availability, covariate
248 observations from 1982 to 2019 were used for this study (Table 1).

249 2.5 CROSS CORRELATION ANALYSIS

250 Cross correlations were then conducted to detect if pulse marine heatwaves had lagged
251 effects on the magnitude and directionality of change in the NES LME communities. Cross
252 correlations determine the relationship between two time series at time t and $t+k$ ($k = \text{time lag}$)
253 and estimate a correlation coefficient for each time lag k considered within the analysis. Here,

254 lags in the correlation up to seven years were considered. The threshold of seven years was set to
255 avoid further exploring potentially spurious environmental relationships. A cross correlation is
256 significant when the absolute value of the correlation coefficient exceeds:

257
$$2 \div \sqrt{(n - |k|)} \text{ (Yoo et al. 2023).}$$

258 Where n is the number of observations and k is the time lag. Here, we used a time series of the
259 difference in MHW cumulative intensity as our possible predictor and the trajectory metrics,
260 both segment lengths and angles, as our response variable. A significant cross correlation
261 coefficient here may suggest MHW changes “lead” significant changes in trajectory segment
262 lengths and angles. This analysis was also completed using the “vegan” R package (v.2.5-7,
263 Oksanen et al. 2020) with data from 1982 to 2019 due to MHW data availability.

264 2.6 REDUNDANCY ANALYSIS

265 To test the relative importance of pulse and press disturbances on changes to the
266 community assemblage over time we used a redundancy analysis on each EPU to summarize the
267 variation due to the set of explanatory covariates being tested. We modeled relationships
268 between community composition and each of the covariates from 1982 to 2017. A complete time
269 series was required to compute the redundancy analysis so data was limited given the MHW time
270 series began in 1982 and MAB data coverage was disrupted in 2017. The analysis was completed
271 using the “vegan” R package (v.2.5-7, Oksanen et al. 2020). All covariate data were
272 standardized, scaled to zero mean and unit variance, for comparison and inclusion in the
273 redundancy analysis. The species data included in the redundancy analysis were Hellinger
274 transformed as they were in the CTA. Forward stepwise model selection was used to determine
275 the best model to explain community composition through the condition that the AIC of the new
276 model including the additional variables had a lower AIC than the simpler model (Table S2). We

277 defined the best model as the one that had the lowest AIC and if the difference in AIC values of
278 two models was less than two, we choose the most parsimonious model (Burnham & Anderson
279 2002) (Table S2). We used variable partitioning to account for the possible correlation between
280 covariates and to extract the marginal and conditional effects of each variable included in the
281 final model. An analysis of variance (ANOVA) was conducted on each global, marginal, and
282 conditional model to determine the significance ($p < 0.05$) of variance between models (Table
283 S3).

284 3. RESULTS

285 The principal coordinate analysis generated by the CTA displays the temporal shifts in
286 community state observations within the NES LME marine assemblage over the past fifty years
287 (Fig 1B). The axes represent the principal coordinates, linear combinations of the original
288 variables that explain the maximum variance in the data. Axis 1 and Axis 2 represent the two
289 most significant principal coordinates, collectively explaining 67% of the total variation in
290 community state observations over the fifty-year period (Fig. 1B) Samples that cluster closely
291 together indicate similar community compositions, while those farther apart suggest significant
292 shifts in assemblage structure. The trajectory segment lengths, represented by arrows connecting
293 each consecutive point, quantify the magnitude of change the community is experiencing from
294 one year to the next (Fig. 1B). The variance in the MAB was better described by PCoA 1 (x-axis,
295 Fig 1B). The variance in the GOM and GB EPU's was orthogonal to the MAB and better
296 explained by PCoA 2 (y-axis in Fig 1B). Over time, the MAB became more distinct from the
297 other two EPU's and showed a higher degree of dissimilarity from its historic observations earlier
298 in the time series (Fig.2C and Fig. 3).

299 The ordination of ecoregions over the sampling period revealed a long-term
300 tropicalization across EPU's (Figure 2A-C). In the GOM, the later observations in the time series
301 resemble the historic assemblage of the ecoregion to its south, GB (Fig. 2A). Similarly, within
302 GB the current assemblage more closely resembles the historic MAB assemblage (Fig. 2B). In
303 recent years the more northward ecoregions demonstrate community compositions that are more
304 like the compositions of the southward regions at the beginning of the time series. The MAB did
305 not have a southern neighboring region explored within this analysis; however, we did observe
306 the MAB shifting uniquely over time (Fig. 2C). Over the entire time series, the MAB shifted the
307 furthest from its 1969 reference state (Fig. 3). Segment lengths relative to the starting assemblage
308 are highest in the MAB and lowest in GB (Fig. 3C and 3D). Observations of segment lengths
309 relative to the starting assemblage increase monotonically in the first 20 years of the time series
310 in all three EPU's (Fig. 3A). In the 1990's these relative segment lengths peak and then are
311 relatively smooth with a slight average uptick across all three EPU's after 2010.

312 A redundancy analysis was conducted on each EPU individually to investigate the
313 variables associated with the variation in beta diversity. The covariates within this study follow
314 similar trends within all three EPU's, apart from the GOM which experienced higher levels of
315 ecosystem overfishing prior to 2000 (Fig. 4). Each of the EPU's is influenced by different
316 covariates (Fig. 5), however, the community-level influence of ecosystem overfishing manifests
317 in all three EPU's. Forward stepwise model selection selected ecosystem overfishing as a
318 significant variable in all three EPU's (Table S2). MHWs in the MAB and the Gulf Stream index
319 in GB were also including in the best fitting models for these regions (Fig. 5B and 5C, Table S2).

320 Variance in community composition explained by the model selected covariates ranged
321 between 13% in the MAB to 18.6% in GB (Table 2). Within the GOM, the Fogarty ecosystem

322 overfishing index is a significant covariate. 13.43% of the variation in community composition
323 over the sampling period can be attributed to this term. Within GB, the Gulf Stream Index and
324 the Fogarty ecosystem overfishing index explained 18.58% of the variation in community
325 composition over the sampling period. After variable partitioning, the conditional effect of the
326 Gulf Stream Index is approximately 5% and the conditional effect of the Fogarty ecosystem
327 overfishing index is 3%. The shared variance between the two is 5.8%. The effect of both
328 variables is significant. The MAB model selection determined the Fogarty ecosystem overfishing
329 index and MHW cumulative intensity to be significant with 13% of variation over the sampling
330 period explained by these two variables. After variable partitioning, the conditional effect of both
331 the Fogarty index and MHW cumulative intensity is 3.4%, with a shared variation between the
332 two of 1%. Significant global, marginal, and conditional models in our EPU also suggests that
333 the included variables are significant and distinct thus supporting our conclusions here (Table
334 S3).

335 No significant influence of MHW pulse disturbances on the speed and directionality of
336 community change was detected in this study (Fig. 6, Table S4) (see *Supplemental Methods - 1.1*
337 *Cross Correlation Interpretation* for details).

338 4. DISCUSSION

339 4.1 TROPICALIZATION IN THE NES LME

340
341 The NES LME finfish and macroinvertebrate assemblage continues to tropicalize (Lucey
342 & Nye 2010), but the recent MHW pulse events did not result in an abrupt shift in the
343 assemblage and did not appear to accelerate this change. Rather ecosystem overfishing was the
344 one variable that most described changes in all three NES LME communities. While MHWs in
345 the MAB and Gulf Stream index in GB also added to the explanatory power of our models, their

346 impact was small. However, the fact that individual indicators of enhanced warming in the NES
347 LME did not add explanatory power to our models does not mean that warming is not important.
348 The Fogarty ecosystem overfishing index incorporates both fishing and productivity into one
349 metric. Temperature and stratification can greatly impact productivity (Behrenfeld et al. 2006,
350 Polovina et al. 2008, Van De Poll et al. 2013), such that the combined effect of heavy fishing and
351 a decline in productivity could have a much larger effect on the species assemblage than
352 warming alone. As such, understanding how both fishing and climate impact species
353 assemblages and ecosystem resilience are critical to effectively manage fisheries (Lotze et al.
354 2019).

355 We were surprised pulse MHW events did not cause sudden and large changes in species
356 assemblages, but an important distinction to make is that we are investigating these disturbances
357 at the community level. Pulse disturbances can cause significant impacts within marine
358 communities on short time scales (Johnston & Keough 2002, Walker & Schlacher 2011, Wijaya
359 et al. 2022) and to specific taxa (Von Biela et al. 2019, Diamond et al. 2020, Raymond et al.
360 2022), however, how these impacts could extend to the community and persist at longer time
361 scales is poorly understood. Within the NES LME, we have observed pronounced but smooth
362 tropicalization in the marine assemblage over time compared to the abrupt shifts we anticipated.
363 Climate driven shifts in populous cold and warm water species synergized by fishing have been
364 observed in North Atlantic waters and perhaps a similar dynamic may be at play here (Gushing
365 & Dickson 1977, Sundby & Nakken 2008, Drinkwater & Kristiansen 2018). Compositional
366 changes in the NES LME could be the result of long-term overfishing of commercially important
367 benthic cold water species (e.g. cod, haddock, yellowtail flounder) that allowed warm water
368 species to increase as has been observed in temperate marine ecosystems (Wernberg et al. 2016).

369 As pulse MHWs become more frequent and water temperatures increase, the incremental shifts
370 in composition may accumulate and lead to tremendous change (Jacquet & Altermatt 2020), but
371 not in the abrupt fashion we expected, as observed here. Both pulse and press events can affect
372 individual species more than the community assemblage on short time scales, and their
373 interactions may not ripple through the entire community immediately but could lead to lower
374 resilience in the future.

375 4.2 PULSE DISTURBANCES – MARINE HEATWAVES

376 When explicitly testing metrics from the CTA, we do not see any significant influence of
377 pulse MHWs on the magnitude and directionality of community change in the entire NES LME.
378 This result is contrary to our initial hypothesis, as we anticipated the drastic nature of MHWs
379 would drive a rapid and directional change in the aftermath of the disturbance across all EPU.
380 While these pulse disturbances create striking environmental contrast for community members to
381 adapt and adjust to, marine systems can demonstrate higher levels of resilience to these pulse
382 disturbances as a function of their high connectivity and functional redundancy where complex
383 food webs provide a wide range of ways for the community to recover successfully (Hillebrand
384 & Kunze 2020). Another factor that may contribute to these results is that species in this study
385 tend to be longer-lived, as opposed to annual species that experience less ability to buffer from
386 intense events. The spatial and temporal scale at which we analyzed the community can mask the
387 effects of pulse events (Heim et al. 2021). Nekton could be changing their distribution within our
388 EPU to compensate for change without changing the overall community assemblage. This
389 suggests that the NES LME is a highly resilient ecosystem, even though it has changed and
390 tropicalized over the last 50 years (O’Leary et al. 2017).

391 The NES LME, like many western boundary current systems, is warming and is predicted
392 to warm more quickly with anthropogenic climate change because of the poleward displacement
393 of the western boundary currents (Wu et al. 2012, Saba et al. 2015, Yang et al. 2016). MHWs
394 have been prominent in frequency and intensity in boundary current waters, but the effects have
395 not been reported in western boundary current literature as much as eastern boundary current
396 areas (Holbrook et al. 2019). Perhaps this is because the MHWs tend to be shorter in duration in
397 these systems and/or because the western boundary current communities are adapted to high
398 seasonality and interannual variability (Holbrook et al. 2019). Even so, the expectation of more
399 intense and frequent MHWs begets the question of how these regional dynamics may respond in
400 the future if pulse dynamics become a more consistent environmental condition.

401 We did find evidence in our redundancy analysis that MHWs impacted the MAB EPU,
402 and that the MAB EPU is changing differently than the GOM and GB. Within the NES LME,
403 there is a significant influence of subregions on community structure (Kleisner et al. 2016,
404 Roberts et al. 2022). Cape Hatteras, North Carolina serves as the boundary between the MAB
405 and the South Atlantic Bight. In this region, the Hatteras Front, and the convergence of cooler,
406 low salinity waters in the MAB meeting the warmer, high salinity waters of the South Atlantic
407 Bight creates a strong physical barrier and highly dynamic region (Savidge 2002). Even mobile
408 species in the area can be restricted in moving beyond this barrier and/or be genetically distinct
409 from northern stocks (e.g., short-finned pilot whales, Thorne et al. 2017 and Atlantic croaker,
410 Schaffler et al. 2009). Species in the MAB can also be more vulnerable at the edge of their
411 optimal thermal and geographic ranges (e.g., Atlantic surfclam, Timbs et al. 2019). Together,
412 these dynamics may drive the different pattern of change that we saw in the MAB to anomalous
413 temperatures and may be the mechanism associating the stronger influence of pulse disturbances

414 like MHWs. Within the GOM and GB EPU, we did not see a strong impact of MHWs; although
415 individual species may be affected, in a resilient system the community as a whole can remain
416 relatively stable.

417 4.3 PRESS DISTURBANCES – ECOSYSTEM OVERFISHING

418 In contrast to pulse disturbances, across all three EPUs, ecosystem overfishing explained
419 a fraction of the observed variation in community composition, as indicated by the significance
420 of the Fogarty ecosystem overfishing index in the redundancy analyses. The Fogarty index is
421 unique in that it in essence captures both environmental change and fishing by relating primary
422 productivity dynamics to the total landings in the system (Link & Watson 2019). The same
423 magnitude of landings in two time periods could result in ecosystem overfishing when
424 productivity declines presumably because of some environmental change (Behrenfeld et al.
425 2006, Polovina et al. 2008, Van De Poll et al. 2013). Historically we have also observed
426 exhaustive, intense fishing pressure within the NES LME; predatory fish biomass reconstructions
427 suggest biomass declines of greater than 70 to 80% from 1900 to 2000 in the region (Christensen
428 et al. 2003). Using multi-faceted covariates, like the Fogarty index, may prove to be a more
429 auspicious approach to capturing dynamics at the community level by accounting for multiple
430 important factors influencing the system.

431 For comparison to our findings, a recent ecological assessment of eastern spotted skunk
432 den use patterns with a redundancy analysis observed 19.6% of variation in den type selection
433 attributed to the constrained covariates considered (Thorne & Ford 2022). In another study, a
434 redundancy analysis was deployed to explore the relationship between soil heavy metal
435 concentrations and environmental variables and found a total of 43.3% of variation in heavy
436 metal concentrations, 36.3% and 16.6% respectively, attributed to soil properties and

437 anthropogenic variables (Song et al. 2016). Their assessment considered a single sampling event
438 of 99 sites in Wenling, China. While low, we consider an observed range of 13-18% of variation
439 attributed to our constrained variables within an analysis of 76 species over 35 years to be
440 noteworthy. Assemblage level dynamics are complex, thus finding comparable explained
441 variance to single species studies and research with a narrower temporal scope within a marine
442 assemblage is promising.

443 Prolonged pressures from fishing and shifts in productivity regimes within the NES LME
444 act over the course of multiple generations and cause shifts in the reproductive success, species-
445 level interactions, and health of marine community members through incremental, but persistent
446 change (Carozza et al. 2018). The Fogarty index was consistently highest in the GOM, and this
447 system experienced the most directional change. We observed a shared variance between the
448 Fogarty index and mean Gulf Stream Index supporting the strong link between oceanographic
449 processes and productivity dynamics in this ecoregion (Lee et al. 1991, Boyce et al. 2017).
450 Ocean circulation and transport play critical roles in the nutrient transport and water column
451 conditions necessary for increased primary productivity (Dai et al. 2023). Altered nutrient ratios
452 and oceanographic regime shifts within the GOM are well-documented (Smith et al. 2012). In
453 particular, the Fogarty index peaked in 1990-2000, and this leads to the period when we see
454 marked change in the assemblage of zooplankton and GOM nekton. In all three EPU's the
455 Fogarty index peaks prior to 2005, and the monotonic increase in segment lengths relative to the
456 starting assemblage across all regions in this period may result in part from the press of
457 ecosystem overfishing on assemblage conditions. The changes in composition we observed here
458 could be indicative of the community repartitioning interactions to maintain system-level

459 functionality even as individual species have varied over time (Hautier et al. 2014, Hoover et al.
460 2014), especially due to the large-scale tropicalization of the region.

461 5. CONCLUSION

462 Climate and fishing not only drive shifts in interactions and species success, but the two
463 pressures interact with one another to further degrade historic community conditions (Planque et
464 al. 2010, Free et al. 2019). Discerning the difference between these pressures and the influence
465 of humans and climate is difficult. The Fogarty ecosystem overfishing index encompasses both
466 social-ecological interactions within the system as well as biological changes. No single variable
467 could explain the change in community assemblage. We suggest more nuanced variables that
468 address both anthropogenic and biological interactions may demonstrate a higher impact than
469 any one environmental variable, even those with strong signals such as MHWs. One mechanism
470 for further improving this work is to explore long-term pressures in the broader temporal context
471 of marine ecosystems. Exploration of the paleo oceanographic and archaeological records to
472 establish premodern baselines would provide conditions for a more comprehensive
473 understanding of both climate and human impacts, well modeled recently in the GOM and
474 Scotian Shelf (Lotze et al. 2022).

475 The amount of variability explained in the region-specific redundancy analyses was very
476 low. Since the redundancy analysis is a constrained multivariate analysis, this suggests that
477 some environmental variability or variables that we did not include in the redundancy analysis
478 might be able to explain the changes in community assemblage that we observed in the
479 unconstrained CTA analysis. We chose environmental variables that represented the dominant
480 physical, biological, and socio-economic drivers in the region. We did not explicitly consider
481 trophic interactions even though the nature of trophic interactions is critical in the resilience of

482 ecosystems (Berlow 1999, McCann 2000). Trophic interactions likely moderate community
483 dynamics within the NES LME and we suggest they be evaluated in future endeavors. While
484 these mechanisms and pressures are more difficult to quantify over multiple generations, they
485 represent important forces in how marine communities operate now and in the future.

486 Even though we did not see the impact of MHW dynamics we anticipated, this analysis
487 was a successful collaboration of multiple techniques to explore disturbance dynamics. We
488 believe the NES LME to be a resilient system, and this analysis is the beginning of further
489 research to quantify beta diversity metrics through time and explore their utility as ecological
490 indicators. The methods applied here should be applied to other assemblages to compare the
491 apparent resiliency of different ecosystems; mechanisms described in this study may drive the
492 NES LME's lack of dynamic response to severe, short-term events and serve as a launching
493 point for supplemental research endeavors. Notably, methods to detrend MHW events to account
494 for prolonged heat stress and water column location are emerging and represent another facet of
495 pulse perturbation stress that we recommend be applied to this system in future study (Wang et
496 al. 2022).

497 As climate change continues to alter marine ecosystems and impact fisheries, directing
498 management actions to buffer the impacts of climate change or enhance the ability of people to
499 respond to disturbance events is critical. Reducing fishing pressure, ameliorating eutrophication,
500 and restoring habitats may yield more immediate benefits for ecosystem management than
501 attempting to manage or predict nature's stochasticity (Scheffer 2009). As we determine
502 appropriate management actions for stressed ecological systems, being attentive to how changes
503 in fishing pressure can be compounded by abiotic factors and climate cycles will help align
504 harvest levels and fishing strategies with stock productivity and biomass (Rogers & Dougherty

505 2018, Cline et al. 2019, Beaugrand et al. 2022). Our application of this emerging method in a
506 large marine ecosystem creates an opportunity to explore how these metrics can serve as
507 ecological indicators in response to disturbance pressures. While recent advancements in
508 modeling and predictive research within marine fisheries have made significant strides, there
509 remains a notable gap in community-level assessments encompassing both abiotic and biotic
510 environmental conditions and interactions, which this study aims to advance.

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516 7. DATA AVAILABILITY

517 The data and the code used for data analysis, visualization, and modeling is openly accessible on
518 GitHub at [https://github.com/IleanaF/MEPS_pulse_press]. Researchers are encouraged to refer
519 to the provided references for further exploration and validation.

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 866 9. TABLES

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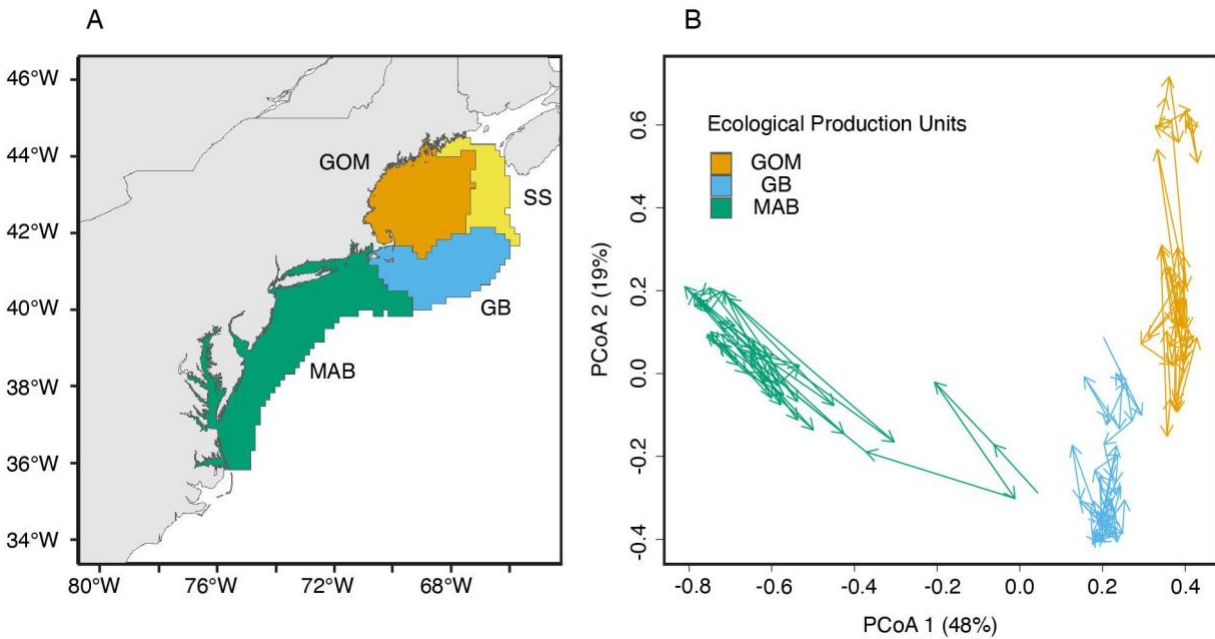
Variable	Variable Category	Disturbance Type	Description	Location and Time Series Length
Small to large zooplankton ratio	Biological	press	The ratio of abundance of small zooplankton (<i>Pseudocalanus spp</i> , <i>Centropages typicus</i> , <i>Centropages hamatus</i> , and <i>Temora longicornis</i>) to large zooplankton (<i>Calanus finmarchicus</i>). Abundance is recorded as number of individuals per 10 m ² , extracted from the Northeast Fisheries Science Center Ecosystem Monitoring program observations (Bucklin et al. 2019).	Measured within each individual EPU 1977-2019
Gulf Stream index position	Physical	press	Annual time series of the relative position of the north wall of the Gulf Stream index (Joyce et al. 2019). The Gulf Stream is a warm ocean current flowing northward along the eastern United States that has a long-term increasing trend with decadal variability in its position. A positive index indicates that the Gulf Stream is in a more poleward position which is positively correlated with temperature on the NES shelf (Nye et al. 2011).	One value for all regions within the NES LME. 1954 - 2021
Bottom temperature anomalies	Physical	press	The anomaly time series depicts the variance between regional mean temperatures and predicted reference temperatures derived from historical regional averages (1981-2010) for the same time of year using a multiple linear regression model (Mountain 1991).	Measured within each individual EPU 1977 - 2021
Ecosystem overfishing - the Fogarty index	Socio-Economic	press	The ratio of total catches to total primary productivity in an ecosystem measured in parts per thousand (Link & Watson 2019).	Measured within each individual EPU 1960 - 2019
Marine heatwave cumulative intensity	Physical	pulse	A measure of the severity of marine heatwaves calculated by multiplying marine heatwave intensity (temperature anomaly in Celsius) and duration (cumulative degree days) using satellite sea surface temperature measurements. Marine heatwaves are defined as five or more days when sea surface temperatures are warmer than the 90th percentile based on a 30-year historical baseline period (Hobday et al. 2016).	Measured within each individual EPU 1982 - 2019

868 Table 1. Key variables examined for influence on community structure within the Northeast US
 869 Continental Shelf Large Marine Ecosystem (NES LME). Annual observations of each variable
 870 from 1982 to 2017 were considered within the redundancy analysis. Here we pulled
 871 observations from the three ecological production units (EPUs) within the NES LME used in this

872 study.
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Ecological Production Unit	Partition	Inertia	Proportion	Overall Explained Variance	Adjusted R ²
Gulf of Maine	Total	0.12379	1		
Gulf of Maine	Constrained	0.01662	0.1343	13.43%*	0.10882*
Gulf of Maine	Unconstrained	0.10717	0.8657		
Georges Bank	Total	0.09926	1		
Georges Bank	Constrained	0.01845	0.1858	18.58%*	0.13651*
Georges Bank	Unconstrained	0.08081	0.8142		
Mid Atlantic Bight	Total	0.11563	1		
Mid Atlantic Bight	Constrained	0.01505	0.1301	13.01%*	0.07740*
Mid Atlantic Bight	Unconstrained	0.10059	0.8699		

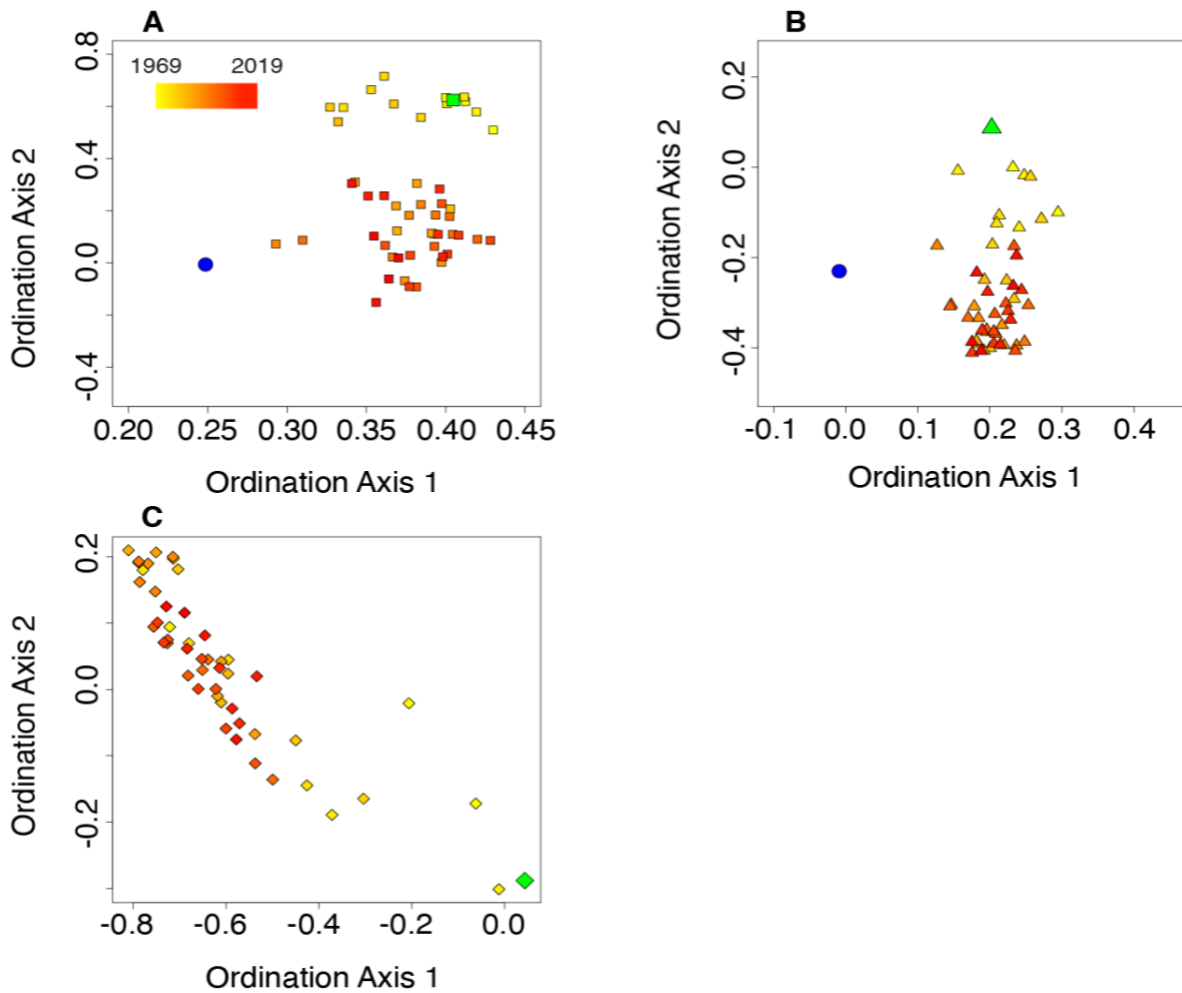
875
 876 Table 2. Variance partitioning results for three Ecological Production Units (EPUs): Gulf of
 877 Maine, Georges Bank, and Mid-Atlantic Bight. The table presents the total inertia, constrained
 878 inertia, and unconstrained inertia for each EPU, along with the proportion of variance explained,
 879 adjusted R squared, and whether the ANOVA determined the results to be significant ($p < 0.05$,
 880 here all values are $p < 0.001$), designated by a *. Model selection and ANOVA reporting are
 881 found in Tables S2 and S3, respectively.
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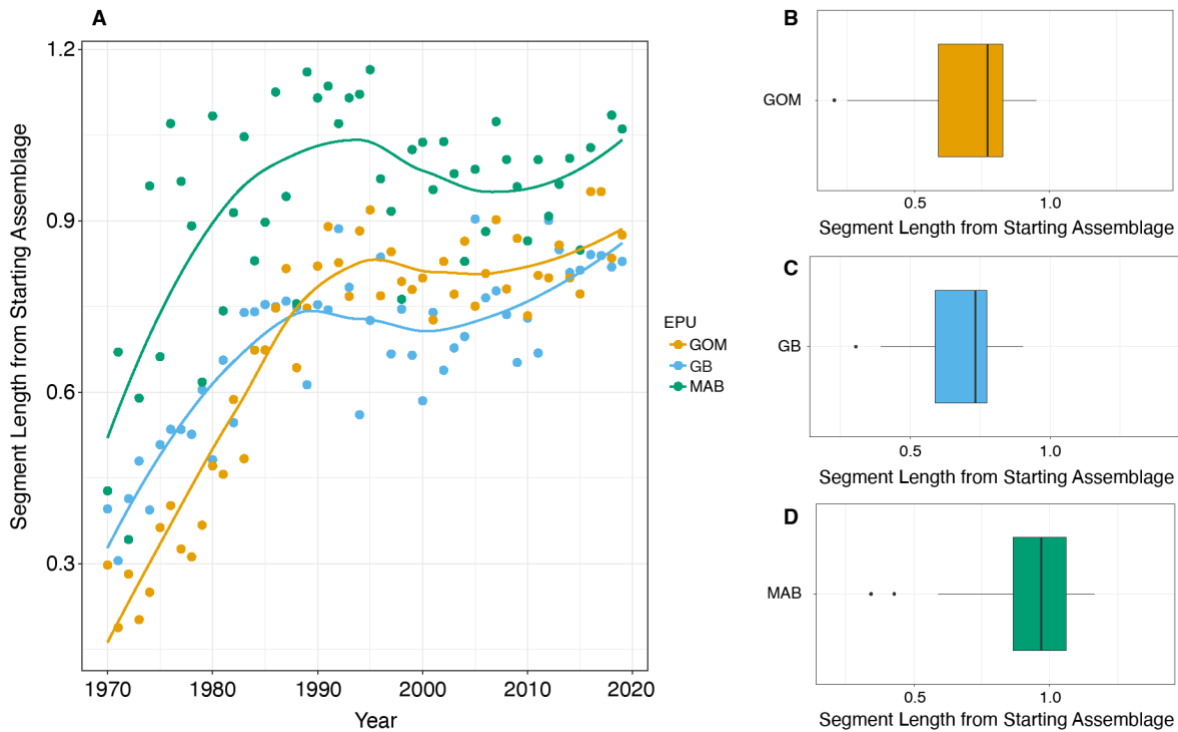
885 Figure 1. (A) The Northeast US Continental Shelf Large Marine Ecosystem (NES LME) is
886 broken into four ecological production units: the Scotian Shelf (SS), Gulf of Maine (GOM),
887 Georges Bank (GB), and the Mid Atlantic Bight (MAB). (B) The community trajectory analysis for
888 the NES LME (excluding SS due to data limitations) where the arrows between points visualize
889 the different states of the community from year to year. PCoA Axis 1 captures 48% of variation
890 and PCoA Axis 2 captures 19% for a total of 67%.

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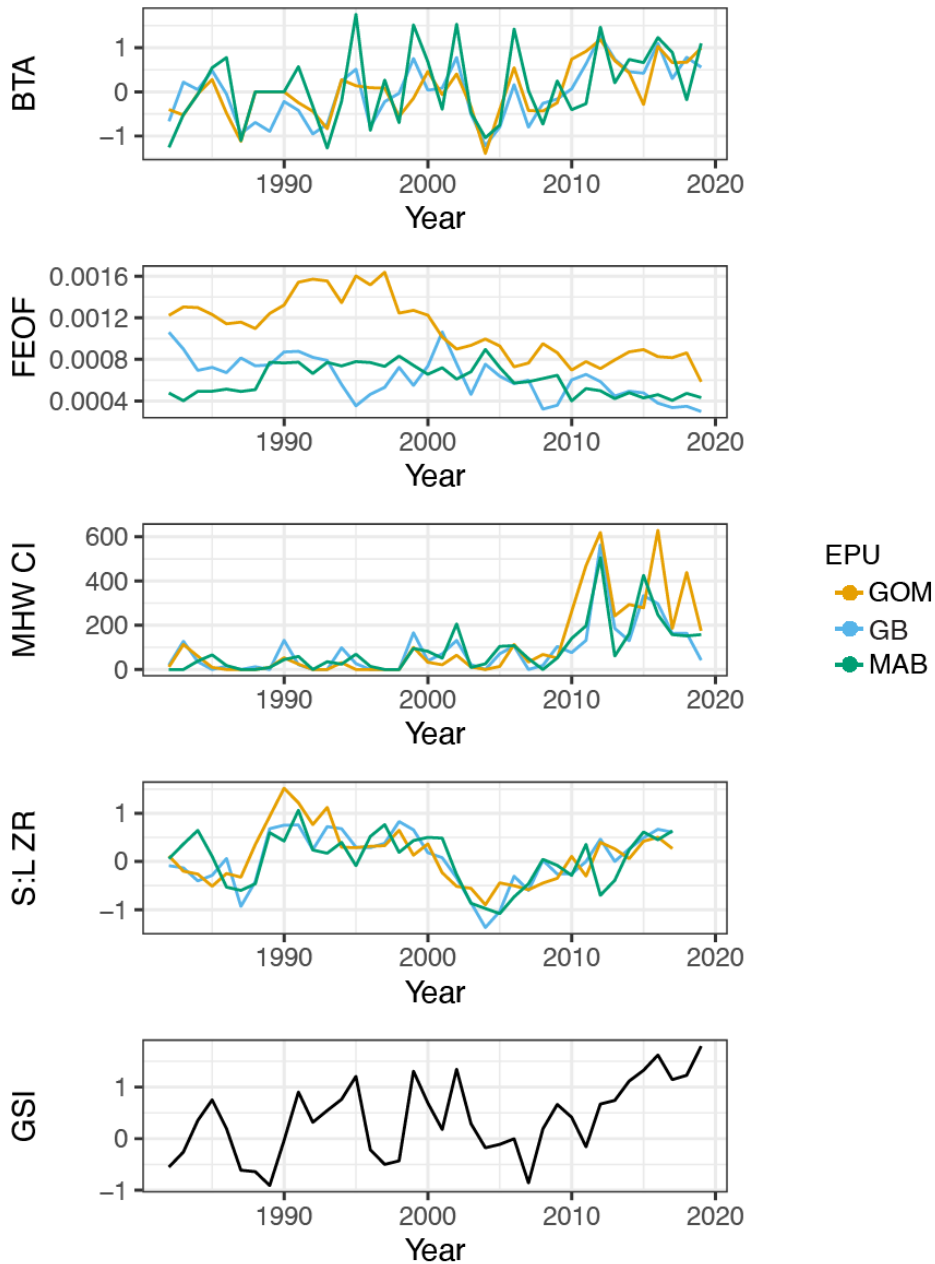


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 894 Figure 2. The ordination diagram for each of the three ecological production units examined; (A)
 895 Gulf of Maine, (B) Georges Bank, and (C) Mid Atlantic Bight. There are two reference points
 896 within each ordination. The blue circular point represents the average location of community
 897 observations from the 1960s of the ecoregion to its south, such that the blue point in (A) the Gulf
 898 of Maine references the average Georges Bank assemblage of the 1960s, (B) Georges Bank
 899 references the average Mid Atlantic Bight of the 1960s and (C) The Mid Atlantic Bight does not
 900 have a southern ecoregion included in this analysis so is excluded. The green diamond denotes
 901 the first community state observation used in the trajectory analysis in 1969. Segment lengths
 902 relative to the starting assemblage are calculated from this point.

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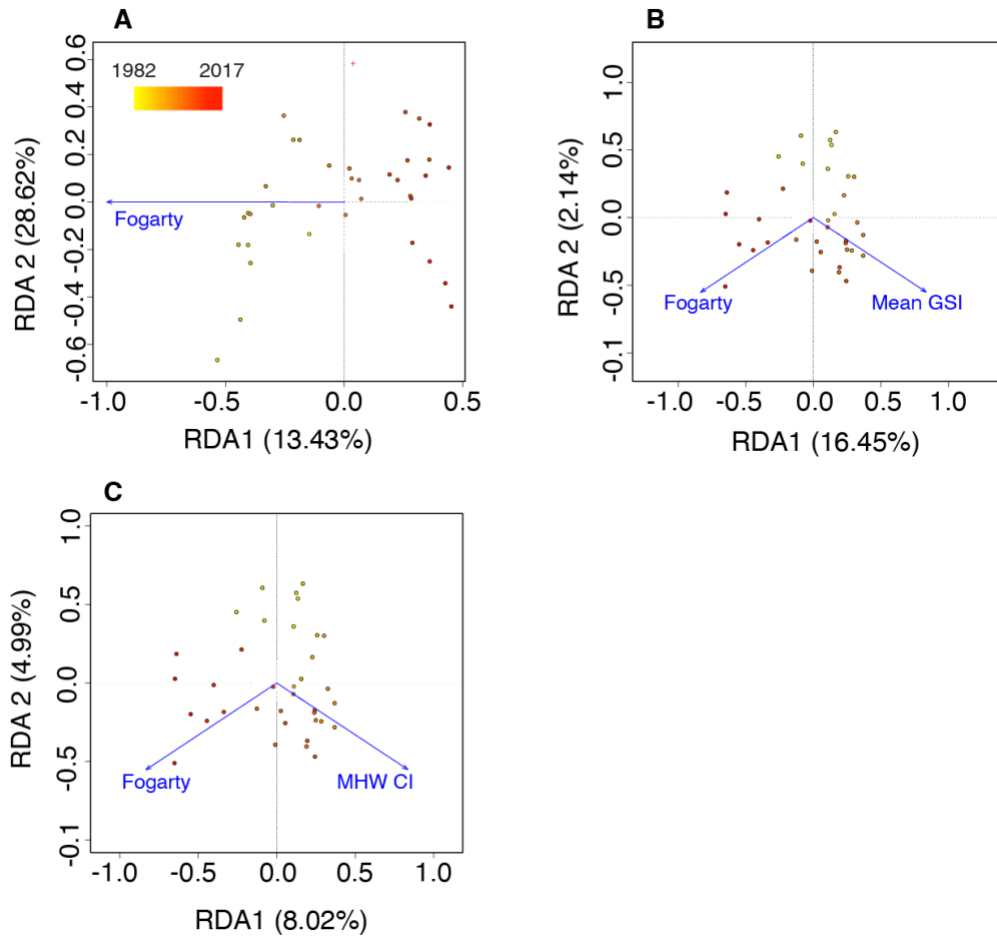


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 906 Figure 3. (A) Scatter plot displaying distribution of segment lengths relative to the starting
 907 assemblage in 1969 within each ecological production unit (EPU). The larger the value the
 908 further from the reference condition the community state in the given year is. A locally weighted
 909 scatterplot smoothing (LOESS) curve is overlaid. (B) Boxplot displaying the distribution of
 910 segment lengths from the starting assemblage in the Gulf of Maine (GOM), (C) Georges Bank
 911 (GB), and (D) the Mid Atlantic Bight (MAB). Each box represents the interquartile range within
 912 the respective category, with the median indicated by the horizontal line within the box. The
 913 whiskers extend to the minimum and maximum values within 1.5 times the interquartile range,
 914 while outliers are plotted individually.
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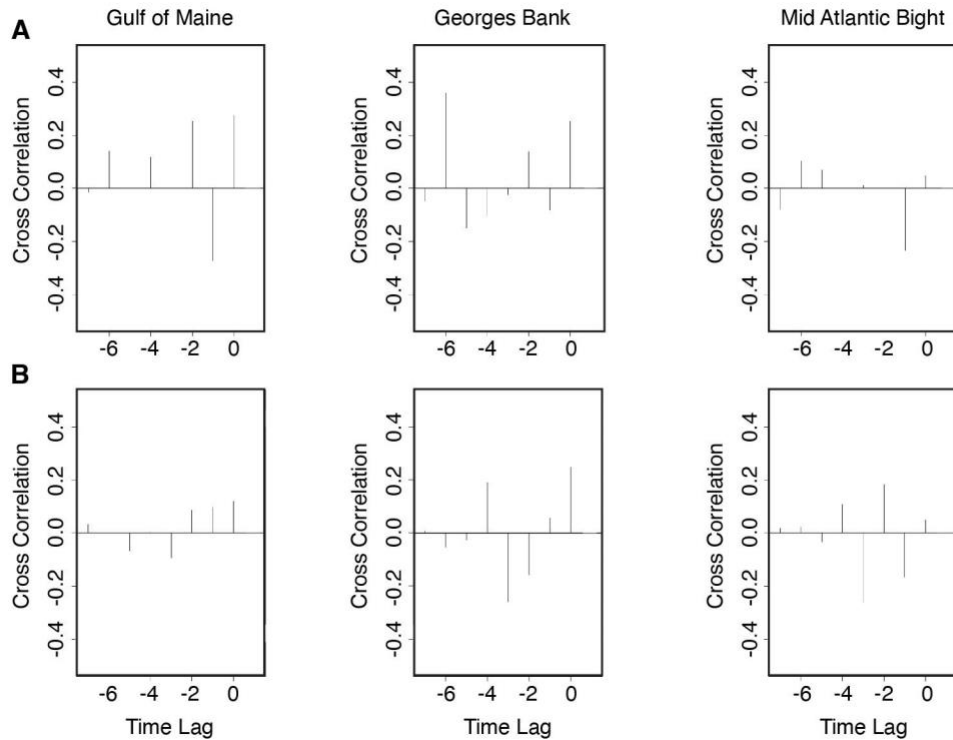


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 917 Figure 4. The time series of annual observations for bottom temperature anomalies (BTA), the
 918 Fogarty ecosystem overfishing index (FEOF), marine heatwave cumulative intensity (MHW CI),
 919 the small to large zooplankton ratio (S:L ZR), and mean Gulf Stream index position (GSI) limited
 920 by years with full data availability (starting in 1982 for MHW CI). A full description of these
 921 variables is in Table 1.

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 925 Figure 5. Redundancy analysis (RDA) of covariate influence on community structure and
 926 composition over the study period for all three ecological production units (A) the Gulf of Maine,
 927 (B) Georges Bank, and (C) the Mid Atlantic Bight with data from 1982 to 2017. Circles are the
 928 RDA scores for each sampling year with time represented in the gradient of colors from yellow
 929 (earliest years) to red (latest). MHW CI is the annual marine heatwave cumulative intensity and
 930 Mean GSI is the mean Gulf Stream index position (Table 1).
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Figure 6. Cross correlation relationships between marine heatwave cumulative intensity and (A) the magnitude of community change (community trajectory analysis segment lengths), and (B) the directionality of change (community trajectory analysis angles) on a lag of 7 years. Within all three ecological production units no cross correlation exceeds calculated thresholds for significance.