

1 **Source Attribution of Black Carbon Affecting Regional Air Quality,**
2 **Premature Mortality and Glacial Deposition**

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17 **Abstract**

18 Black carbon (BC) mitigation can reduce adverse environmental impacts on climate, air
19 quality, human health, and water resource availability. To facilitate identification of
20 mitigation priorities, we use a state-of-the-science global coupled chemistry-climate model
21 (AM3), with tagged regional (East Asia, South Asia, Europe and North America) and
22 sectoral (land transport, residential, industry) anthropogenic BC emissions to identify
23 which sources have the largest impacts on air quality, human health and glacial deposition.
24 We find that within each tagged region, domestic emissions dominate BC surface
25 concentrations and associated premature mortality (generally over 90%), as well as BC
26 deposition on glaciers (~40-95% across glaciers). BC emissions occurring domestically
27 within each tagged source region contribute roughly 1-2 orders of magnitude more to
28 domestic BC concentrations, premature mortality, and BC deposition on regional glaciers
29 than the same quantity of BC emitted from foreign sources. At the sectoral level, the South
30 Asian residential sector contributes ~60% of BC associated premature mortality in South
31 Asia and ~40-60% of total BC deposited on southern Tibetan glaciers. Our findings imply
32 that BC mitigation within a source region, particularly from East and South Asian
33 residential sectors, will bring the largest reductions in BC associated air pollution,
34 premature mortality, and glacial deposition.

35

36 **Keywords:**

37 BC, Tagging method, Source-receptor matrix, Attribution efficiency, Human health.

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39

40 **1. Introduction**

41 Climate change is a complex, long-term, and large-scale issue which is challenging to
42 address in the short term due in part to the high political and economic costs of CO₂
43 mitigation ([Keohane and Victor, 2016](#)). However, mitigating black carbon (BC), which has
44 a lifetime of only several days in the atmosphere and has a strong positive direct regional
45 radiative forcing, is recognized as an effective strategy to address near-term global
46 warming by both the scientific and diplomatic communities ([CCAC, 2012](#); [EPA, 2012](#);
47 [Jacobson, 2002](#); [Kopp and Mauzerall, 2010](#); [Ramanathan and Carmichael, 2008](#)). Also,
48 BC mitigation has immediate benefits for air quality, human health and water resources
49 (via impacts on glaciers), providing strong motivation for its near-term mitigation ([Bond](#)
50 [et al., 2013](#); [Hansen and Nazarenko, 2004](#); [Peng et al., 2016](#); [Shindell et al., 2012](#);
51 [UNEP/WMO, 2011](#)). To help prioritize BC mitigation strategies, we quantify the regional
52 and sectoral contributions of BC emissions on air quality, premature mortality and glacier
53 deposition.

54 Epidemiological studies have linked long-term exposure to fine particulate matter in
55 ambient air, of which BC is a key component, with an elevated risk of premature mortality
56 ([Hoek et al., 2013](#); [Janssen, 2012](#); [REVIHAAP, 2013](#)). [UNEP/WMO \(2011\)](#) and [Shindell](#)
57 [et al. \(2012\)](#) identified potential BC mitigation measures and estimated that by 2030 a full
58 global implementation of these measures could avoid 2.3 million premature deaths
59 annually due to reduced outdoor air pollution. Both studies quantified the health benefits
60 of BC abatement under future scenarios using PM_{2.5} concentration-response functions from
61 the American Cancer Society (ACS) cohort studies ([Krewski, 2009](#)). However, the relative
62 risk of PM_{2.5} exposure found in the ACS study, which is based on ambient concentrations

63 typical of the U.S., does not accurately represent the risk associated with high levels of
64 PM_{2.5} exposure in Asia, which is better represented in the Global Burden of Disease (GBD)
65 analysis ([Burnett et al., 2014](#)). Here we conduct additional analyses of source attribution
66 of regional and sectoral BC emissions to BC associated health effects, based on relative
67 risks at both high and low levels of PM_{2.5} exposures, in order to prioritize mitigation efforts.

68 In addition to air quality and health considerations, BC deposition on glaciers is also of
69 immediate concern because of its effects on snow albedo, atmospheric radiative forcing,
70 and the melting rates of snow and ice that may modify the seasonality and trends of melt-
71 water supply ([Kopacz et al., 2011](#); [Ramanathan et al., 2005](#)). Using the GEOS-Chem
72 adjoint model, Kopacz et al. (2011) identified India and China as the primary emission
73 sources of BC deposited on five glaciers in the Himalayas and Tibetan Plateau (HTP). Here
74 we further characterize the regional and sectoral anthropogenic sources depositing BC on
75 glaciers globally to prioritize BC mitigation.

76 This study goes beyond previous work and conducts an integrated assessment of the
77 source attribution of regional and sectoral present-day (2000) BC emissions on air quality,
78 premature mortality and deposition of BC on northern hemisphere glaciers. We apply a
79 source-tagging approach within a global 3-D fully coupled chemistry-climate atmospheric
80 model, Atmospheric Model version 3 (AM3). This model allows us to evaluate inter-
81 regional and inter-continental transport of BC from specific sources and regions and
82 analyze their impacts during average conditions of climate and air quality around the year
83 2000 on air quality, health and glacial deposition. This, to our knowledge, is among the
84 first trials, to conduct such an integrated assessment to evaluate the impact of BC globally.

85 The tagging technical has been applied in many literature ([Brandt et al., 2012](#); [Liu et](#)

86 [al., 2005, 2009a; Wang et al., 2014](#)) and the model (AM3/GFDL) we applied also has been
87 applied in many recent atmospheric physics/chemistry, air quality and health studies ([Fang](#)
88 [et al., 2013; Lamarque et al., 2013a; Lamarque et al., 2013b; Lee et al., 2013; Naik et al.,](#)
89 [2013; Parrish et al., 2014; Schnell et al., 2015; Shen et al., 2017; Shindell et al., 2013; Silva](#)
90 [et al., 2013; Stevenson et al., 2013; Voulgarakis et al., 2013; Young et al., 2013](#)). Using
91 AM3, a climate-chemistry coupled model, we aim to reflect the average condition of
92 climate and air quality around the year 2000. The simulation we conduct for this study is
93 part of an extensive assessment of the impact of BC mitigation, not only on air quality,
94 health and glacier deposition (as shown in this paper), but also on climate ([Lamarque et al.,](#)
95 [2013b; Lee et al., 2013](#)). As a result, it uses a present-day (2000) emission inventory
96 developed for the Intergovernmental Panel on Climate Change Fifth Assessment Report
97 (IPCC AR5) by Lamarque et al. (2010). The emission inventory used in our study is a
98 standard one that the climate-chemistry community made and used in their work to
99 simulate the 2000s climatology and chemistry ([Fang et al., 2013; Lamarque et al., 2013a;](#)
100 [Lamarque et al., 2013b; Lee et al., 2013; Naik et al., 2013; Parrish et al., 2014; Schnell et](#)
101 [al., 2015; Shen et al., 2017; Shindell et al., 2013; Silva et al., 2013; Stevenson et al., 2013;](#)
102 [Voulgarakis et al., 2013; Young et al., 2013](#)). Our simulation follows the standard protocol
103 of the Atmospheric Chemistry and Climate Model Inter-Comparison Project (ACCMIP)
104 and uses standard emission inventories for the year 2000, so that people can inter-compare
105 and evaluate their simulated air quality and climate impacts. In addition to running
106 simulations for global total BC emissions from all regions and sources, we tag land
107 anthropogenic BC emitted from four heavily-populated industrial regions (East Asia, South
108 Asia, North America and Europe) and three sectors (land transportation, residential, and

109 industry) in East Asia and South Asia and identify their effects on receptors including the
110 source regions listed above and the rest of the world (Fig. 1). Our analyses aim to identify
111 regions and sectors where mitigation of anthropogenic BC emissions yields the greatest
112 benefits.

113

114 **2. Method**

115 **2.1 Model description**

116 The global AM3 model is developed by the Geophysical Fluid Dynamics Laboratory
117 (GFDL)/NOAA as the atmospheric component of their coupled general circulation model
118 (CM3). AM3 is designed to address key emerging issues in climate science, including
119 aerosol-cloud interactions and chemistry-climate feedbacks, and has been widely applied
120 in many recent studies ([Fang et al., 2013](#); [Lamarque et al., 2013a](#); [Lamarque et al., 2013b](#);
121 [Lee et al., 2013](#); [Naik et al., 2013](#); [Parrish et al., 2014](#); [Schnell et al., 2015](#); [Shen et al.,](#)
122 [2017](#); [Shindell et al., 2013](#); [Silva et al., 2013](#); [Stevenson et al., 2013](#); [Voulgarakis et al.,](#)
123 [2013](#); [Young et al., 2013](#)). A finite-volume dynamical core with a cubed-sphere horizontal
124 grid is used (2° latitude \times 2.5° longitude), with horizontal resolution varying from 163 km
125 (at the corners of each face) to 231 km (near the center of each face). Vertically, AM3
126 extends from the surface up to 0.01 hpa (~86 km) with 48 vertical hybrid sigma pressure
127 levels. Vertical resolution ranges from 70 m near the earth's surface to 1 to 1.5 km near the
128 tropopause and 3 to 4 km in much of the stratosphere ([Donner et al., 2011](#)). The model
129 includes interactive tropospheric and stratospheric chemistry (85 species, including BC)
130 and uses emissions (both natural and anthropogenic) to drive its chemistry and aerosol
131 simulations.

132 To improve agreement between the simulated lifetime and concentrations of BC with
133 recent observations, a BC aging scheme with updated parameters for estimating BC dry
134 and wet deposition is implemented ([Liu et al., 2011](#); [Zhang et al., 2015a](#)). This improved
135 BC aging scheme considers the conversion of hydrophobic BC to hydrophilic BC, as well
136 as the mixing of BC with other aerosols (i.e. sulfate) via characterizing the lifetime of BC
137 with being proportional to OH radical concentrations parameterized according to
138 observations ([Liu et al., 2011](#); [Shen et al., 2017](#)). In terms of wet deposition, AM3 includes
139 in-cloud and below-cloud scavenging by large-scale and convective clouds. The wet
140 deposition flux is directly proportional to the local concentration. In-cloud scavenging of
141 aerosols is calculated following the work of [Giorgi and Chameides \(1985\)](#)). In the case of
142 convective clouds, wet deposition is only computed within the updraft plume. Below-cloud
143 washout is only considered for large-scale precipitation and is parameterized as by [Li et al.](#)
144 [\(2008\)](#) for aerosols. Such simplified wet deposition scheme generally provides a
145 reasonable simulation of the global mean and regional patterns of aerosol optical depth
146 (AOD) ([Donner et al., 2011](#)) and of the spatial patterns in surface PM_{2.5} ([Fang et al., 2013](#)).
147 However, AM3 assumes that both snow and rain have the same scavenging factor and also
148 implements “limiters” to enforce monotonicity and positive definiteness in the tracer
149 mixing ratio profiles, which was later found to cause less effective convective removal than
150 realistic ([Paulot et al., 2016](#); [Zhao et al., 2018](#)). More details of parameterizations of wet
151 deposition, particularly for aerosols, in cloud and below cloud by snow and rain are
152 described in [Fang et al. \(2011\)](#)). However, AM3 assumes the same scavenging factor by
153 snow and rain, which leads to less effective convective removal. This is because aerosols’
154 in-cloud scavenging does not depend on the type of precipitation (rain or snow) in AM3,

155 thus the same collection efficiency (0.001, dimensionless) with which aerosols are
156 collected by raindrops and snow is used. More details of parameterizations of wet
157 deposition, particularly for aerosols, in cloud and below cloud by snow and rain are
158 described in [Fang et al. \(2011\)](#).

159 Earlier studies have extensively evaluated the physical (i.e., optical characteristics of
160 aerosols) and chemical (O₃ and PM_{2.5} concentrations) parameters in AM3 ([Donner et al.,
161 2011](#); [Fang et al., 2013](#); [Naik et al., 2013](#)). Also, AM3 has been used in BC study and
162 shown good comparison with other similar models ([Liu et al., 2011](#); [Shen et al., 2017](#); [Lee
163 et al., 2013](#)). Liu et al. (2011) found that AM3 has better performance in simulating BC
164 concentrations in the Arctic than all HTAP models considered by [Shindell et al. \(2008\)](#).
165 They also found that AM3 would be among the best models evaluated by [Koch et al.
166 \(2009\)](#). Shen et al. (2017) found that AM3 simulation results for BC in the Arctic compares
167 well to the models in the Arctic Monitoring and Assessment Programme (AMAP), which
168 causes an average underestimation of a factor of 2 for BC at Alert and Barrow. For this
169 paper, we chose AM3, a chemistry-climate model because it allows us to conduct a
170 complete and integrated assessment of the BC impacts, not only on air quality, health and
171 glacial deposition as discussed in this paper, but also on the climate (in a parallel project).
172 Furthermore, it allows us to evaluate inter-regional and inter-continental transport of BC
173 and its impacts during average conditions of climate and air quality around the year 2000
174 (e.g., 1996-2005, a specific time slice on which the Atmospheric Chemistry and Climate
175 Model Inter-Comparison Project -ACCMIP project focused), and thus are comparable to
176 the ACCMIP results.

177 **2.3 Tagging BC from various regional and sectoral sources**

178 In addition to the 85 atmospheric chemical species included in the standard AM3 model,
179 we add “tagged” BC species to the model to identify the original source of BC affecting
180 air quality, premature mortality and glacier deposition. Tagged BC species undergo all the
181 dynamical, physical and chemical processes as the untagged BC, but only originate from
182 specific sources and do not affect the simulated climate. Using this method, [Liu et al.](#)
183 [\(2009b\)](#) found that the differences between the sum of non-reactive tagged tracers and the
184 concentrations of the same untagged species are less than 1% at most locations thus
185 justifying the use of tagging method for source attribution analyses.

186 Thus, total BC emissions in this study include BC emissions from all anthropogenic
187 (land anthropogenic, aviation, and shipping) and natural sources. On top of that, we also
188 tag BC from land anthropogenic sources because they can be most easily mitigated and
189 collectively represent 66% of global land anthropogenic and biomass burning BC
190 emissions. To balance tracking specific emissions with computational capacity, we add
191 tagged BC tracers only from four highly-populated source regions (East Asia, South Asia,
192 Europe, and North America (see Fig.1), which contribute 18%, 7%, 6%, and 5% to global
193 land anthropogenic and biomass burning BC emissions, respectively) and sector-specific
194 tracers from the three largest emission sectors (land transportation, residential, and industry)
195 from East Asia and South Asia.

196 Our tagged BC species include BC_EA_Trans, BC_EA_Res, BC_EA_Ind,
197 BC_EA_Anthro, BC_SA_Trans, BC_SA_Res, BC_SA_Ind, BC_SA_Anthro,
198 BC_EU_Anthro, and BC_NA_Anthro. Here, ‘EA’, ‘SA’, ‘EU’ and ‘NA’, refer to BC
199 originating from East Asia, South Asia, Europe and North America, respectively. ‘Trans’,
200 ‘Res’, and ‘Ind’ refer to land transportation, residential, and industrial sectors, respectively.

201 Our tagged transportation BC emissions only include BC from land transport. BC
202 emissions from maritime transport (international shipping, domestic shipping and fishing)
203 and aviation are included in our global total BC emissions ([Lamarque et al., 2010](#)).
204 Residential sector includes emissions from both fossil fuel and biofuel combustion. Also,
205 industrial sector includes emissions from both combustion (including the power sector) and
206 non-combustion industrial processes ([Lamarque et al., 2010](#)). “Anthro” is the sum of land
207 anthropogenic emissions from these three sectors (including both fossil fuel and biofuel
208 combustion) and from agricultural waste burning (not including forest or grassland fires)
209 and waste management. Implementing these tagged BC species allows us to quantify the
210 contributions of specific anthropogenic sources (both regional and sectoral) to surface BC
211 concentrations, the resulting premature mortality, and BC depositions on glaciers.

212 **2.4 Estimating Premature Mortality**

213 Some studies suggest that combustion-related particulate matter (including BC) is more
214 toxic than other PM_{2.5} components and that BC itself may act as a “universal carrier” of
215 toxic particles ([Cooke et al., 2007](#); [Grahame and Schlesinger, 2007](#); [Hoek et al., 2013](#);
216 [Morton et al., 2013](#); [REVIHAAP, 2013](#)). However, while efforts are growing to isolate the
217 specific health effects of different components of fine particles (PM_{2.5}), evidence for
218 differential toxicity is not fully conclusive ([EPA., 2009](#); [Janssen, 2012](#); [REVIHAAP, 2013](#)).
219 Therefore, we treat BC as a fractional component of PM_{2.5} to estimate BC associated
220 premature mortality in the main analysis. Specifically, we use the PM_{2.5} relative risk (RR)
221 functions for adults (≥ 25 years old) from Burnett et al. (2014) for ischemic heart disease
222 (IHD), cerebrovascular disease (stroke), chronic obstructive pulmonary disease (COPD),
223 and lung cancer (LC), as their study covers the global range of PM_{2.5} exposures, using the

224 equation below. Burnett et al. (2014) provides concave-shaped disease-specific integrated
225 exposure–response (IER) functions to capture the variations in premature deaths due to
226 changes in PM_{2.5} exposures. Details for the disease-specific IHD functions can be referred
227 to in Fig.1 in Burnett et al. (2014) and the Global Burden of Disease Study 2010 (GBD
228 2010) - Ambient Air Pollution Risk Model 1990 – 2010 database
229 ([http://ghdx.healthdata.org/record/global-burden-disease-study-2010-gbd-2010-ambient-](http://ghdx.healthdata.org/record/global-burden-disease-study-2010-gbd-2010-ambient-air-pollution-risk-model-1990-2010)
230 [air-pollution-risk-model-1990-2010](http://ghdx.healthdata.org/record/global-burden-disease-study-2010-gbd-2010-ambient-air-pollution-risk-model-1990-2010)).

$$231 \quad \Delta Mort_i = POP_{25} * Mortbase_i * [1 - RR_i(C_{PM2.5-BC_s})/RR_i(C_{PM2.5})] \quad (1)$$

232 In each AM3 model grid cell, we estimate BC_s associated premature mortality from disease
233 i ($\Delta Mort_i$) based on the adult population (POP_{25}), the baseline mortality rate in adults for
234 disease i ($Mortbase_i$), the relative risk for disease i (RR_i) at the simulated total surface
235 PM_{2.5} concentrations ($C_{PM2.5}$), and the relative risk for disease i at the PM_{2.5} concentrations
236 without the contribution of the specific BC source ($C_{PM2.5-BC_s}$). BC_s refers to individual
237 BC species, either the total untagged BC or each of our tagged regional or sectoral BC
238 tracers. We obtain the 2000 global population data at 1km² grid level from [CIESIN \(2005\)](http://www.ciesin.com).
239 Country-level age distribution and baseline mortality rates for IHD, stroke, COPD, and LC
240 for the year 2000 are from the GBD project (<http://www.healthdata.org/gbd>) and are
241 assumed uniform within each country.

242 In our main analysis, we assume equal toxicity between undifferentiated PM_{2.5} and BC.
243 In order to test the sensitivity of our estimated BC associated health impacts to this
244 assumption, we also estimate health impacts using BC-specific exposure-response
245 functions for all-cause mortality from a meta-analysis by Hoek et al. (2013). Unlike the
246 concave integrated exposure-response functions of Burnett et al. (2014), the BC-specific

247 functions are (log-) linear. Therefore, for comparison, we also include a second sensitivity
248 study using a recently-updated (log-) linear exposure-response function for all-cause
249 mortality for undifferentiated PM_{2.5} ([Forestiére et al., 2014](#)). Details of the sensitivity
250 studies are reported in the SI.

251 **2.5 Calculating Attribution Efficiency**

252 To identify regions and sectors that impose the largest negative environmental impacts
253 from each unit of BC emission, we estimate the BC attribution efficiency for air quality,
254 premature mortality, and glacial deposition. The “attribution efficiency” for BC surface
255 concentrations, BC associated premature mortality, or BC deposited to glaciers is defined
256 as the magnitude of BC concentrations ($\mu\text{g}/\text{m}^3$), the number of premature deaths (cases),
257 or the amount of BC deposition ($\mu\text{g}/\text{m}^2/\text{day}$) in receptor regions resulting from each Gg of
258 BC emitted from a specific source region or sector ($AE_{r-s,k} = \frac{I_{r-s,k}}{E_s}$), respectively.
259 $AE_{r-s,k}$ is the attribution efficiency for a specific type of environmental impact k (i.e., BC
260 associated premature mortality) in receptor region r due to tagged BC tracer from source
261 s , $I_{r-s,k}$ is the environmental impact k in receptor region r due to tagged BC tracer from
262 source s , E_s is the emission of source s . A high attribution efficiency for a specific BC
263 tracer indicates that a unit reduction from that source will likely result in a relatively large
264 reduction in the negative environmental impact at the receptor region.

265

266 **3. Results**

267 **3.1 Regional and Sectoral Attribution of Annual Average Simulated BC Surface** 268 **Concentrations**

269 We evaluate the effect of total global BC emissions and emissions originating from each
270 of our tagged regions and sectors on simulated annual average surface BC concentrations.
271 Simulated global annual average BC surface concentrations exhibit hotspots in Asia,
272 particularly in eastern China and the Indo-Gangetic plain of India (Fig. 2), with the highest
273 concentrations found in Shandong province in China ($\sim 3.5 \mu\text{g}/\text{m}^3$). BC surface
274 concentrations in Europe and the United States are much lower - roughly $0.05\text{-}1 \mu\text{g}/\text{m}^3$ and
275 $0.05\text{-}0.5 \mu\text{g}/\text{m}^3$, respectively. BC emissions from each of our four tagged regions are
276 primarily concentrated within their respective source regions (Fig. 2).

277 We also present the simulated annual average population-weighted (P-W) BC surface
278 concentrations ($C_{PW}(R) = \frac{\sum_{i=1}^n \text{Pop}_i(R) * C_i(R)}{\sum_{i=1}^n \text{Pop}_i(R)}$) in the four tagged regions in Fig. 3. $C_{PW}(R)$
279 refers to P-W annual average BC surface concentrations in each tagged region R , Pop_i
280 refers to population in grid i , and $C_i(R)$ refers to model simulated BC concentrations in
281 grid i in region R . As shown in Fig. 3, East Asia ($1.7 \mu\text{g}/\text{m}^3$) and South Asia ($1.1 \mu\text{g}/\text{m}^3$),
282 the two regions with the largest population, also have the highest P-W concentrations,
283 followed by Europe ($0.5 \mu\text{g}/\text{m}^3$) and North America ($0.4 \mu\text{g}/\text{m}^3$). We find that domestic
284 anthropogenic BC emissions dominate that region's P-W BC surface concentrations,
285 accounting for 96%, 93%, 89% and 93% of surface BC concentrations over East Asia,
286 South Asia, Europe and North America, respectively. Any individual foreign contributions
287 to P-W BC concentrations are generally below 1%. Within each region, the sectors that
288 contribute most to domestic P-W BC concentrations are the industrial sector in East Asia
289 ($\sim 50\%$) and the residential sector in South Asia ($\sim 60\%$). We find that in East Asia and
290 South Asia, sectoral contributions to P-W BC surface concentrations are roughly
291 proportional to their domestic emission shares.

292 **3.2 Regional and Sectoral Attribution of BC Associated Premature Mortality**

293 We estimate global total BC associated premature mortality to be 106,000 annually (95%
294 confidence interval, CI, 45,000-143,000, see Table 1 and Supplementary Fig. S1), with
295 ranges reflecting the low and high estimates of relative risks for PM_{2.5} exposure reported
296 in Burnett et al. (2014). As shown in Fig. 4 and Fig. S2, high BC associated premature
297 mortality primarily occurs in East Asia (over 10 BC associated premature deaths/1000 km²
298 in eastern China) and South Asia (1-10 BC associated premature deaths/1000 km² across
299 India), both regions with high BC concentrations and dense population. Accordingly,
300 approximately half of the global total BC associated premature deaths occur in East Asia,
301 followed by 17% in South Asia. In contrast, BC associated premature mortality is relatively
302 low in Europe (0.1-10 premature deaths/ 1000 km²) and North America (0.1-1 premature
303 deaths/ 1000 km² in eastern U.S.), contributing approximately 12% and 4% of the global
304 total BC associated premature deaths, respectively.

305 We further quantify the source attribution of regional and sectoral BC emissions on BC
306 associated premature mortality (Table 1). In each region, emissions from that region
307 (domestic emissions) generally account for over 90% of total BC associated premature
308 mortality while the contribution from any given foreign region is typically below 1%. From
309 the sectoral perspective, domestic BC emissions from the industrial (~50%) and residential
310 (~35%) sectors make up ~85% of BC associated premature deaths occurring in East Asia.
311 The residential sector dominates (~60%) BC associated premature deaths in South Asia.
312 Similar to BC surface concentrations, in both East Asia and South Asia, sectoral
313 contributions to BC associated premature mortality are roughly proportional to their
314 respective emission shares.

315 Our sensitivity analyses (Fig. S1) demonstrate that the health impacts from BC exposure
316 would be much higher if BC toxicity is greater than that of typical PM_{2.5}. We find that
317 using BC-specific relative risks leads to a mortality burden that is 14 times higher than
318 what we obtain with the cause-specific, concave functions for PM_{2.5} reported in Burnett et
319 al. (2014). Only a relatively small part of the difference is attributable to assumptions about
320 the shape of the exposure-response relationship (log-linear versus concave) and the number
321 of causes (all causes versus four causes) (Fig. S1). The primary difference results from the
322 difference in toxicity of the PM_{2.5} components. This indicates that our main estimates of
323 the mortality burden associated with BC exposure, which are based on changes to total
324 PM_{2.5} concentrations, are conservative should evidence of enhanced toxicity of BC be
325 confirmed.

326 **3.3 Regional and Sectoral Attribution of BC Deposition on Northern Hemispheric** 327 **Glaciers**

328 We analyze BC deposition at the specific locations of individual glaciers in each region.
329 Detailed geographic information of glaciers evaluated are shown in Fig. S3 and Table S1.
330 Due to notable spatial heterogeneities and seasonal variations for BC deposition, we
331 quantify the seasonal contributions of regional and sectoral anthropogenic BC emissions
332 to BC deposition on glaciers in Asia, Europe, North America and the Arctic, respectively
333 (Fig. S4).

334 **3.3.1 Asia Glaciers**

335 In Asia, we focus on the Himalayan Tibetan Plateau (HTP) glaciers. For nearly all these
336 glaciers, seasonal contributions of land anthropogenic BC from South Asia and East Asia
337 represent between ~40%-95% of total BC deposition. Southern Tibetan glaciers receive a
338 relatively large total BC deposition (Fig. 5a). On these glaciers, South Asian land
339 anthropogenic BC emissions alone account for ~60-90% of total BC deposition across
340 seasons, of which ~40-60% is attributed to the residential sector. This is consistent with
341 earlier findings that South Asia is the major source of BC in southern HTP ([Zhang et al.,](#)
342 [2015b](#)). Besides, due to the Asian monsoon that carries emissions from South Asia
343 northwards, these glaciers usually have the largest total deposition in summer, when South
344 Asian emissions account for 80-90% of total BC deposition (Fig. 5a). In contrast, the
345 largest East Asian contribution to BC deposition on these glaciers occurs in October (i.e.,
346 ~ 30% in Zuoqiupu), in part because of the returning winter monsoonal flow. In other
347 seasons, East Asia contributes less than 10% of total BC deposition. Regional contributions
348 to BC deposition on central and southern HTP glaciers are similar, though total deposition

349 on central HTP glaciers is much lower than on southern HTP glaciers (Fig. 5a). In the
350 northern Tibetan Plateau, we generally find a dominant East Asian contribution in summer
351 among our tagged tracers, which has also been identified in previous studies ([Li et al., 2016](#);
352 [Zhang et al., 2015b](#)). However, our study finds that south Asian land anthropogenic BC
353 emissions mostly dominate BC deposition in the northern Tibetan Plateau in other seasons,
354 though East Asian emissions are also important (Fig. 5b). Additionally, European land
355 anthropogenic BC emissions deposit 1-10% of total BC on northern Tibetan glaciers,
356 particularly on Haxilegen River and Mt. Muztagh glaciers, with the largest contributions
357 in summer. Kopacz et al. (2011) identify a strong snow-albedo driven seasonal cycle
358 (lowest in winter and highest in summer) in radiative forcing on northern Tibetan glaciers,
359 indicating an enhancement of potential glacial melting due to glacial deposition of BC in
360 summer. Thus, East Asian land anthropogenic BC emissions, which contribute most to
361 northern HTP glacial deposition in summer, are likely to enhance northern HTP glacial
362 melting most.

363 **3.3.2 European and North American Glaciers**

364 As shown in Fig. 5c, anthropogenic BC emissions within each source region dominate BC
365 deposition over both European and North American glaciers in all seasons. For European
366 glaciers, domestic land anthropogenic BC emissions contribute ~50-90% of total BC
367 deposition, depending on the season. In comparison, contributions from East Asia, South
368 Asia and North America are each generally below 5%, with the remaining originating from
369 either wildfires or BC emission sources outside the tagged regions.

370 Similarly, North American land anthropogenic BC emissions contribute ~40-80% of
371 BC deposition on North American glaciers. Notably, East Asian land anthropogenic BC
372 emissions contribute ~20% to BC deposition on western US glaciers (i.e., Washington state
373 and the Sierra Nevada mountains) during non-summer seasons, with roughly 9% and 7%
374 attributable to the East Asian industrial and residential sectors, respectively.

375 **3.3.3 Arctic Glaciers**

376 BC deposition on Arctic glaciers is of particular concern due to their critical role in
377 affecting the climate via positive feedbacks on melting sea ice ([McConnell et al., 2007](#);
378 [Ramanathan and Carmichael, 2008](#)). We find that European (~25%), East Asian (~13%),
379 North American (~7%), and South Asian (~3%) land anthropogenic emissions together
380 contribute about half of annual average total BC deposition to the Arctic (> 67° N), though
381 large variations exist depending on the geographic location of glaciers and the season (Fig.
382 5d and Fig. S4). The other half of BC deposition over the Arctic may come from emissions
383 from biomass burning (wildfires) and other untagged sources. For example, BC from
384 biomass burning has been found to make up 40% of BC in Arctic air ([Winiger et al., 2016](#)),

385 and to account for over 90% of BC in Arctic snow when counted along with biofuel ([Hegg](#)
386 [et al., 2009](#)).

387 Fig. 5d shows the seasonal contributions of our tagged anthropogenic BC emissions to
388 BC deposition on Arctic glaciers in Alaska, Greenland, and Russia. Our tagged
389 anthropogenic BC emissions from four tagged regions together explain roughly 60-80% of
390 the total BC deposition on Alaskan and Greenland glaciers and about 40%-60% on Russian
391 glaciers in non-summer seasons. The lower contribution of tagged emissions on Russian
392 glaciers suggests that untagged Russian emissions (particularly domestic biomass burning
393 may contribute roughly half of total Russian BC emissions) are an important source of BC
394 deposition ([Lamarque et al., 2010](#)). Contributions from untagged wildfire emissions are
395 particularly notable in summer, when our tagged anthropogenic BC emissions together
396 only explain ~20%, ~45%, and ~25% of total BC deposition over Alaska, Greenland, and
397 Russian glaciers. This is most likely due to large contributions to Arctic BC from European
398 and Siberian boreal fires in summer ([Bond et al., 2013](#); [Hegg et al., 2009](#); [McConnell et](#)
399 [al., 2007](#)).

400 In our simulations, Alaskan glaciers, depending on the season, receive ~10-35% of their
401 total BC deposition from East Asian land anthropogenic BC emissions, significantly more
402 than from Europe (~3-20%), North America (~2-10%) or South Asia (~1-5%). For
403 southern Greenland glaciers (i.e., Dye2), North American land anthropogenic BC
404 emissions, depending on the season, contribute ~20-30% of BC deposition, followed by
405 Europe with ~10-20%. In comparison, among our tagged emissions, European land
406 anthropogenic BC emissions dominate BC deposition onto northern Greenland glaciers
407 (~20-50% depending on the season). We also find that of our tagged emissions, European

408 and East Asian land anthropogenic BC emissions dominate BC deposition over western
409 and eastern Russian glaciers, respectively.

410 **3.4 Attribution Efficiency of BC Impacts on Air Quality, Premature Mortality, and** 411 **Deposition on Glaciers**

412 We estimate the attribution efficiency for annual average P-W BC concentrations, BC
413 associated premature mortality, and BC deposition on glaciers (Supplementary Tables S2-
414 S4).

415 As shown in Supplementary Table S2, across regions, each unit of BC emission in South
416 Asia contributes the most (ie. has the largest attribution efficiency) to BC surface
417 concentrations within its own source region (approximately 60%, 80%, and 95% higher
418 than attribution efficiency in East Asia, Europe, and North America, respectively). This is
419 primarily because the annual average lifetime of BC from South Asia is longer than from
420 other regions mainly due to less efficient wet deposition ([Shindell et al., 2008](#)). In
421 comparison, East Asia has the largest attribution efficiency in BC-associated premature
422 deaths within its own source region, roughly 35 (11-47) premature deaths per Gg BC. In
423 comparison, South Asia, Europe, and North America have attribution efficiencies of 28
424 (13-45), 25 (15-31) and 12 (7-12) premature deaths per Gg BC of their own emissions,
425 respectively (Supplementary Table S3). Ranges of premature mortality attribution
426 efficiency are due to uncertainties in the PM_{2.5} relative risks. Higher attribution efficiency
427 for BC associated premature mortality in East Asia primarily occurs because the adult
428 population in East Asia was ~1.5, ~2.2, and ~4.2 times higher than in South Asia, Europe,
429 and North America in the year 2000, respectively. Nevertheless, the number of adults in
430 India has grown rapidly since 2000

431 (<http://esa.un.org/unpd/wpp/Download/Standard/Population/>), and is estimated to surpass
432 China around 2040 (Fig. S5). Thus, the attribution efficiency for BC associated premature
433 mortality will likely increase in South Asia.

434 We find that South Asian emissions have the largest BC deposition attribution
435 efficiency over Asian glaciers (Table S4). Specifically, in southern and central (northern)
436 Tibetan glaciers, each unit of South Asian emissions results in 10-100 times (1-4 times)
437 more BC deposition than the same quantity of East Asian emissions. Similarly, for
438 European and North American glaciers, domestic BC emissions have 1-2 orders of
439 magnitude larger attribution efficiencies for BC deposition than BC emissions from any
440 foreign region. Notably, for most Arctic glaciers, anthropogenic BC emissions from
441 Europe generally have the largest BC deposition attribution efficiency among our tagged
442 tracers. However, the largest attribution efficiency for southern Greenland glaciers is from
443 North America.

444 Sectoral differences affecting the attribution efficiency of BC concentrations,
445 premature mortality, and deposition on glaciers are generally small. However, our
446 estimated premature mortality attributed to emissions from the residential sector does not
447 account for indoor air pollution, which resulted in over 3 million premature deaths
448 worldwide in 2000 (<http://www.healthdata.org/gbd>). Thus, including the effects of indoor
449 air pollution would substantially increase the attribution efficiency of total (indoor plus
450 outdoor) BC associated premature mortality associated with the residential sector.

451

452 **4. Discussion**

453 To evaluate the potential benefits of various BC mitigation strategies, this study
454 systematically quantifies the source attribution of regional and sectoral BC emissions on
455 air quality, premature mortality and deposition on glaciers by tagging land anthropogenic
456 BC tracers from four regions and three sectors within a global coupled chemistry-climate
457 model.

458 We find that intra-regional emissions usually contribute over 90% of BC surface
459 concentrations and associated premature mortality, as well as ~40-95% of BC deposition
460 on glaciers within each region. Additionally, domestic BC emissions generally contribute
461 approximately 1-2 orders of magnitude more to BC concentrations, associated premature
462 deaths, and BC deposition on glaciers than the same quantity of foreign emissions. This
463 indicates that reducing domestic BC emissions can bring both the largest total reduction
464 potential and the highest reduction efficiency for the BC associated environmental impacts
465 we evaluate in this work.

466 Across regions, people in East Asia and South Asia are exposed to the highest BC
467 surface concentrations and suffer from the largest resulting premature mortality. Among
468 the 106,000 (95% CI: 45,000-143,000) cases of BC associated premature deaths globally,
469 ~70% occur in Asia. Within either East Asia or South Asia, sectoral contributions to BC
470 surface concentrations and associated premature mortality, and to BC deposition on
471 glaciers are roughly proportional to their emission shares. For instance, the South Asian
472 residential sector, which accounts for approximately 60% of total regional land
473 anthropogenic BC emissions, contributes ~60% of domestic population-weighted (P-W)
474 BC concentrations and associated premature mortality, as well as ~40-60% of total BC
475 deposition on southern Tibetan glaciers.

476 Regarding attribution efficiency, the largest attribution efficiency for health impacts
477 occurs in East Asia, where it is approximately 20%, 40%, and 200% larger than in South
478 Asia, Europe, and North America, respectively, primarily due to the co-location of dense
479 population and high BC emissions in East Asia. Importantly, the health impacts from each
480 unit of BC emission have increased significantly in South Asia over the past decade due to
481 rapid population growth. Sectoral differences in attribution efficiency for BC associated
482 premature mortality from outdoor air pollution are usually small. However, BC emission
483 from the residential sector may be particularly important given the adverse impacts of
484 indoor air pollution and associated health burden ([Smith and Mehta, 2003](#)). Overall, this
485 implies that BC mitigation may be particularly beneficial in Asia, especially from the
486 residential sector.

487 Although domestic BC dominates regional impacts, foreign contributions to premature
488 mortality and BC deposition on glaciers do occur, particularly for regions with little
489 domestic BC emission sources. For instance, East Asian anthropogenic BC emissions are
490 a large contributor to BC deposition over the Arctic, especially on Alaskan glaciers (~10-
491 35% across seasons). Thus, even though the Arctic countries can work together through
492 the Arctic Council to mitigate emissions of their own short-lived climate forcers ([Sand et
493 al., 2016](#)), our study suggests it would be beneficial for them to facilitate East Asian BC
494 emission reductions, particularly for Alaskan glaciers. Over the past decade, international
495 efforts to facilitate BC mitigation in developing countries by institutions such as the
496 Climate and Clean Air Coalition ([CCAC, 2012](#)) and Global Alliance for Clean Cookstoves
497 ([GACC, 2010](#)) have made some progress at reducing emissions. Continuing international
498 collaboration for existing projects as well as local capacity building, particularly in South

499 Asia and East Asia, will be beneficial to ensure sustainable and cost-effective BC
500 mitigation in these regions, especially in the residential sector ([Baron et al., 2015](#); [UNEP,](#)
501 [2011](#)).

502 We recognize that results in this study are subject to various uncertainties from emission
503 inventories, the representation of BC in our model, and the relative toxicity of BC in
504 comparison to other PM_{2.5} components. BC emission inventory studies report large
505 uncertainties (i.e., a factor of 2-4 in Asia) due to challenges in quantifying particle
506 emissions from incomplete combustion and in determining BC emitting activity levels
507 ([Bond et al., 2013](#); [Cohen and Wang, 2014](#); [Qin and Xie, 2011a, b, 2012](#); [Zhao et al., 2011](#)).

508 Also, although this version of the AM3 model uses an updated representation of BC aging
509 ([Liu et al., 2011](#); [Zhang et al., 2015a](#)), it still does not perfectly capture the chemical-
510 physical transformation of BC. Thus, better quantification of the BC emission inventory
511 and model development advances will be valuable in improving BC source attribution
512 estimates. In addition, our study identifies substantial uncertainties (roughly a factor of 15)
513 in BC associated health impacts primarily due to differences in estimates of the toxicity of
514 BC compared with PM_{2.5}. Thus increasing efforts to differentiate the toxicity of various
515 PM_{2.5} components are needed to better identify the health impacts of BC. As our simulated
516 BC concentrations from the global coarse-resolution chemistry-climate model tend to
517 underestimate BC concentrations especially in urban areas, and we assume the same
518 toxicity for BC with other PM_{2.5} components, our estimated BC associated premature
519 deaths are conservative. However, we believe that such underestimates are unlikely to
520 change our qualitative conclusions in terms of the dominant local benefits of BC
521 mitigation.

522 Our study focuses on year 2000 to be comparable to ACCMIP results. However, BC
523 emissions have evolved over the past two decades ([Bond et al., 2013](#); [Mao et al., 2016](#); [Qin
524 and Xie, 2012](#)). [Granier et al. \(2011\)](#) have made a comprehensive comparison of various
525 bottom-up anthropogenic and biomass burning emission inventories between 1980-2010.
526 Their comparison indicates that globally, all emission inventories agree that anthropogenic
527 BC emissions increased from 2000-2010 although the rate of increase decreased around
528 2005. [Granier et al. \(2011\)](#) also found that all emission inventories suggested a strong
529 increase in anthropogenic BC emissions from China and India during this period, though
530 different emission inventories identified either small increases or decreases in
531 anthropogenic BC emissions in Western and Central Europe as well as in the United States.
532 The increasing trend in China and India is consistent with a few later Asian-focused studies.
533 For example, Bond et al. (2013) found that Asian BC emissions increased from ~7.5 Tg/yr
534 in 2000 to 9.8 Tg/yr in 2005. Likewise, Qin and Xie (2012) found that BC emissions in
535 China, the world's largest BC emitter, increased from ~1.2 Tg/yr in 2000 to ~1.9 Tg/yr in
536 2009. Similar global and regional trends are also captured in the Representative
537 Concentration Pathways (RCP) scenarios. The RCP8.5 has been especially considered to
538 be a reasonable extension of the ACCMIP simulations beyond 2000 ([Granier et al., 2011](#)).
539 All RCP scenarios also show increases in BC emissions from 2000 to 2010, though they
540 also suggest that total BC emissions peaked in 2010 and will reach a similar emission level
541 to 2000 in 2020 (Fig. S6a). Nevertheless, at the regional level, BC emissions show a
542 consistent increasing trend in Asia from 2000 to 2020, though it has been decreasing in the
543 OECD countries (including Europe and North America) during the same period (Fig. S6b).
544 Meanwhile, global population has increased from ~6 billion in 2000 to ~7 billion in 2015,

545 with particularly large growth in South and East Asia. These region-specific trends in both
546 BC emissions and population growth suggest that the source attribution of south and east
547 Asian BC emissions on air quality, health and glacier deposition is likely to be even more
548 important in more recent years than what we identified in this study.

549

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553

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556

557

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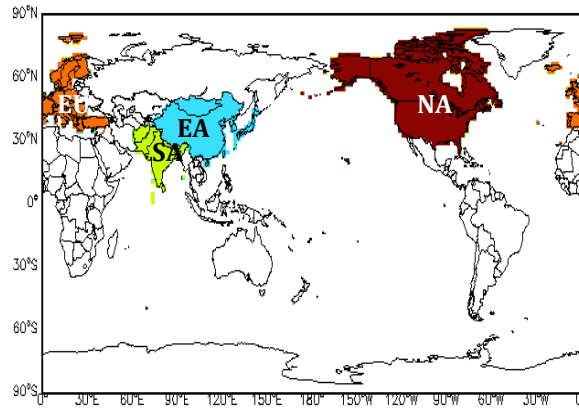
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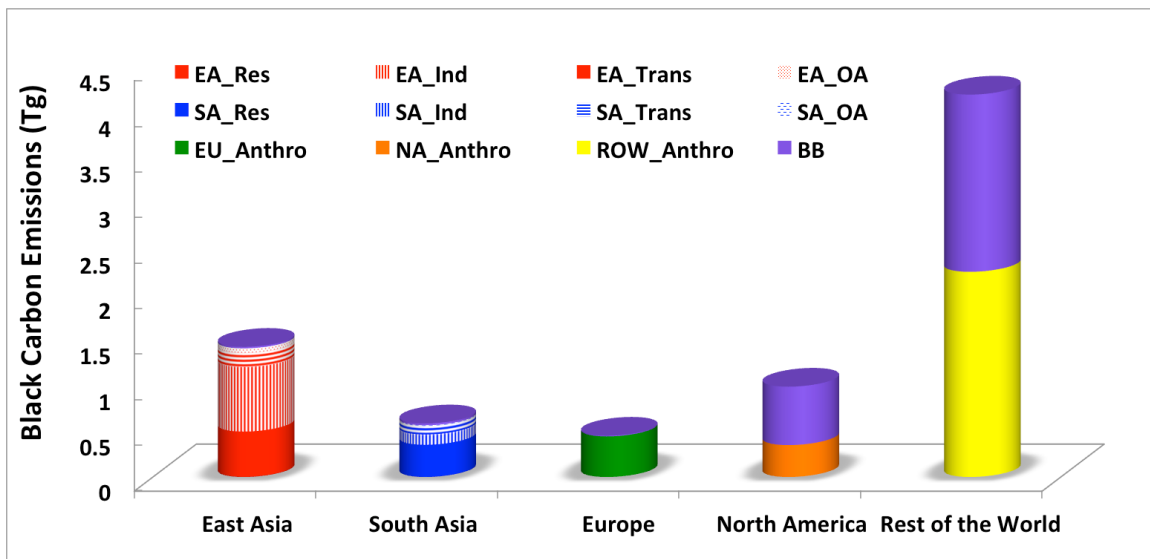
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(a)



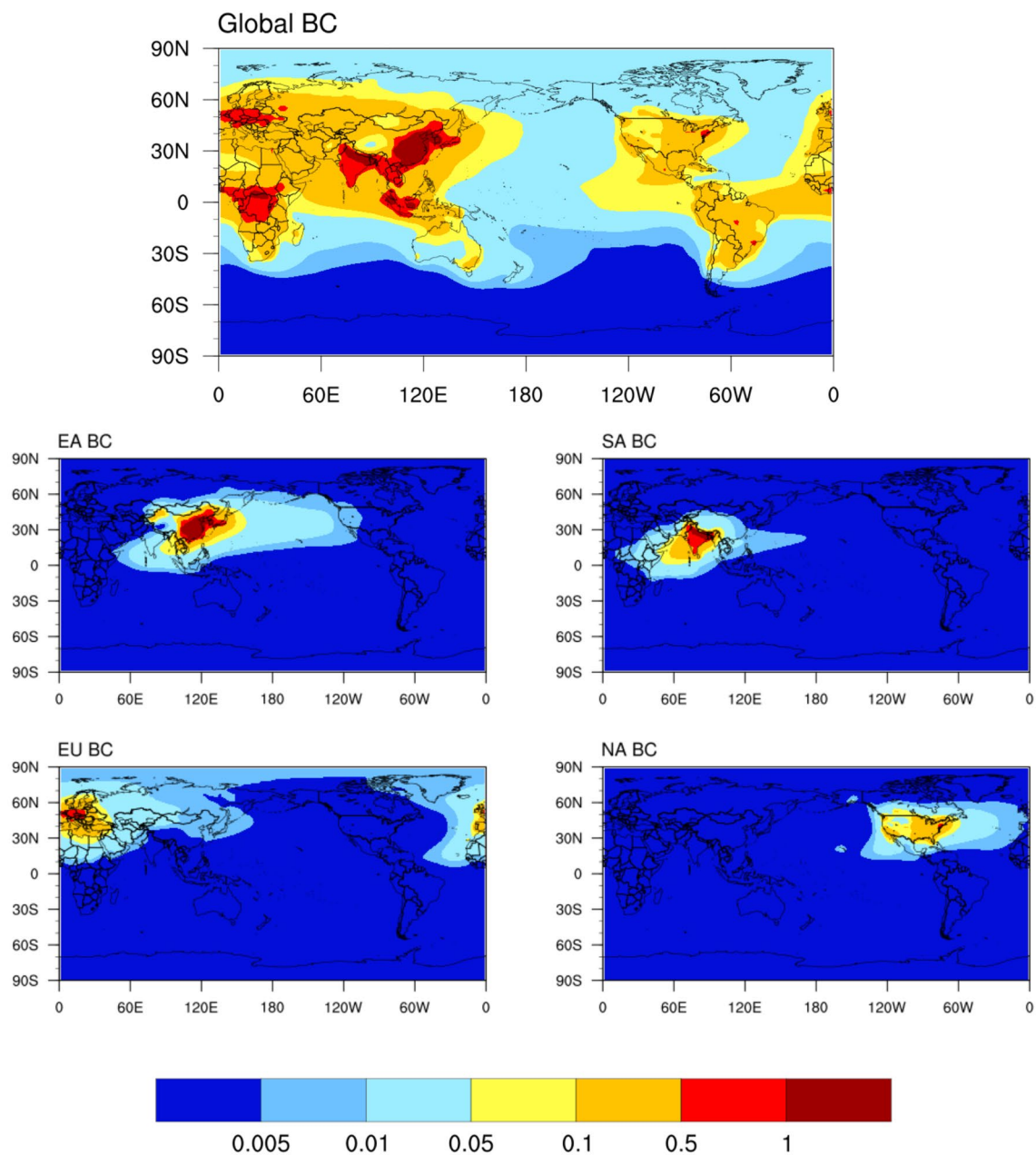
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(b)

769 **Fig. 1 (a) Location of the four-tagged regions: East Asia (EA), South Asia (SA),**
 770 **Europe (EU), and North America (NA). (b) Global annual mean land anthropogenic**
 771 **and biomass burning black carbon (BC) emissions (Tg) from our tagged regions and**
 772 **different sectors. Land anthropogenic (Anthro) sources include fossil fuel and biofuel**
 773 **from the residential (Res), industry (Ind), land transportation (Trans), and other**
 774 **(agricultural waste burning and waste management) (OA) sectors. BB (Biomass**
 775 **burning) emissions include forest and grassland fires. BB emissions are not tagged**
 776 **but are included in total BC emissions.**

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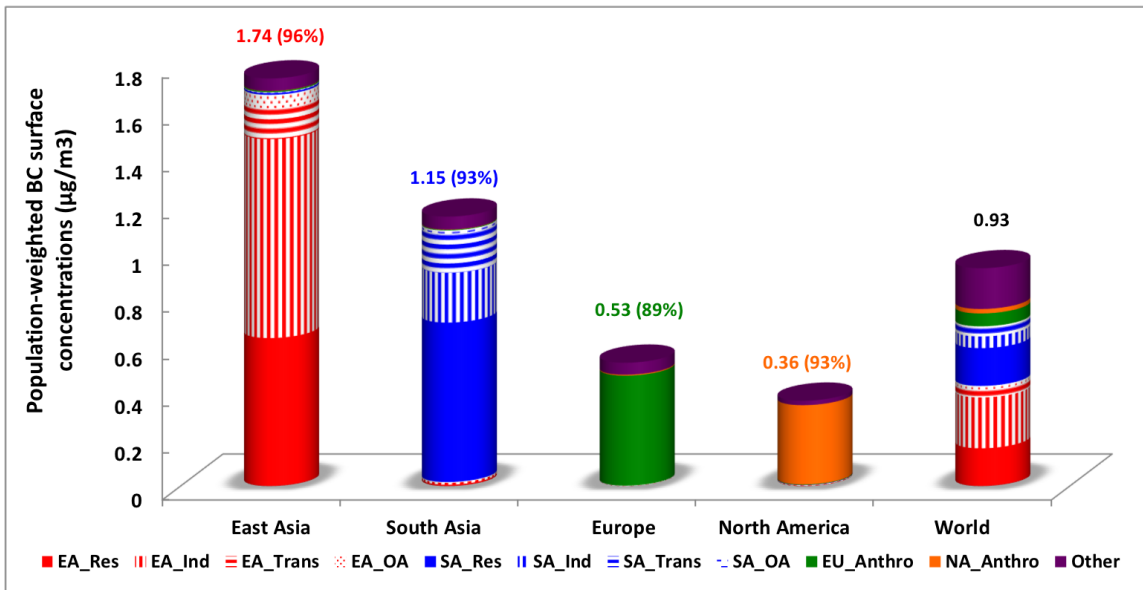
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779 **Fig. 2 Annual average BC surface concentrations ($\mu\text{g}/\text{m}^3$) resulting from global total**

780 **BC emissions, and land anthropogenic BC emissions from each tagged region in 2000:**

781 **East Asia (EA), South Asia (SA), Europe (EU), and North America (NA).**

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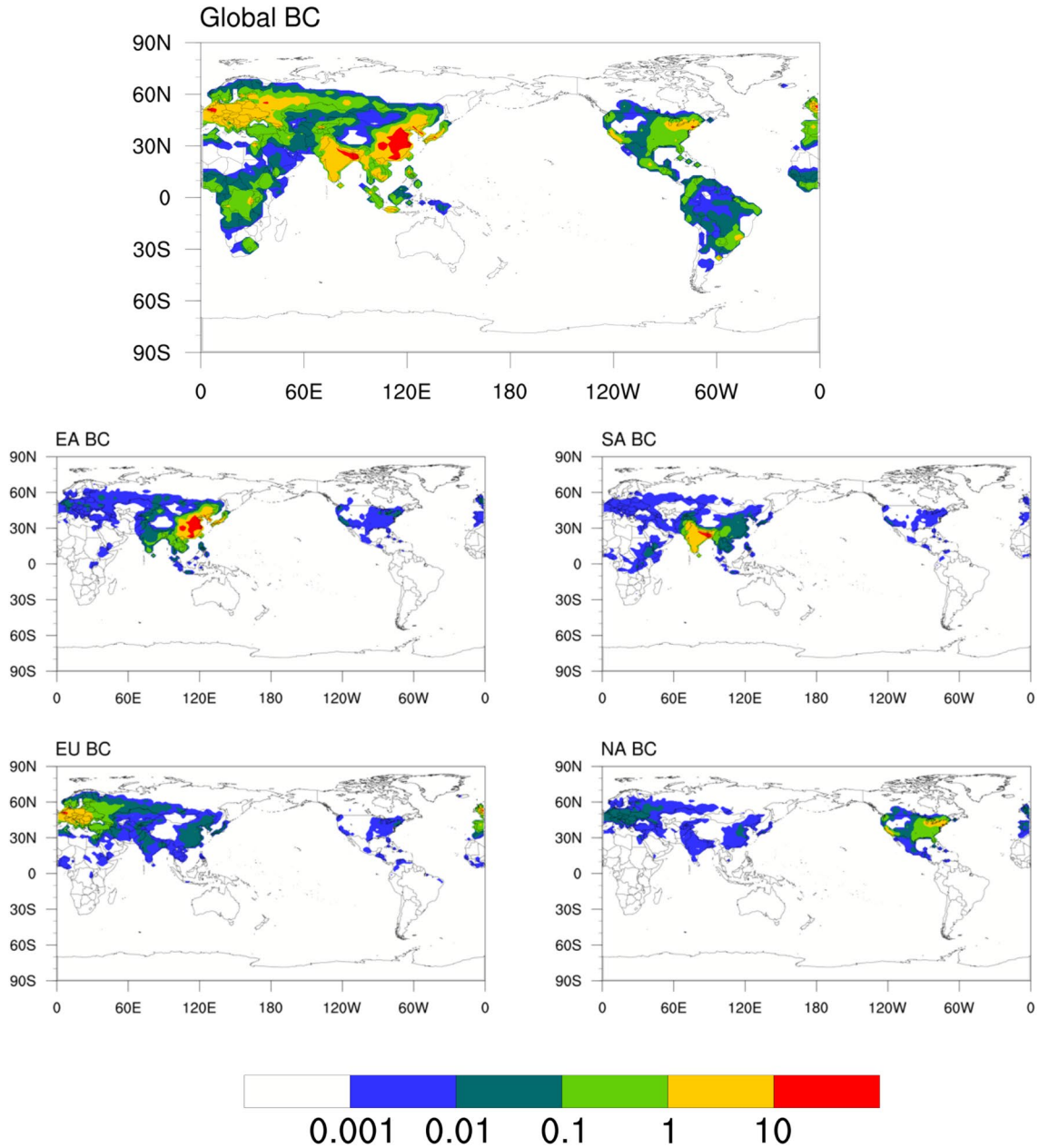


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785 **Fig. 3 Annual average population-weighted BC surface concentrations in receptor**
 786 **regions (µg/m³). Percentages in parentheses represent the intra-regional contribution**
 787 **to annual average BC concentrations in the surface layer over each receptor region.**
 788 **EA, SA, EU, and NA refer to regions defined in Fig. 1a. Res, Ind, Trans, OA, and**
 789 **Anthro refer to the residential, industrial, and land transportation sectors plus other**
 790 **land anthropogenic sources (e.g., agricultural waste burning and waste management),**
 791 **and total land anthropogenic BC emissions, respectively. Other emissions include**
 792 **untagged anthropogenic emissions and biomass burning emissions.**

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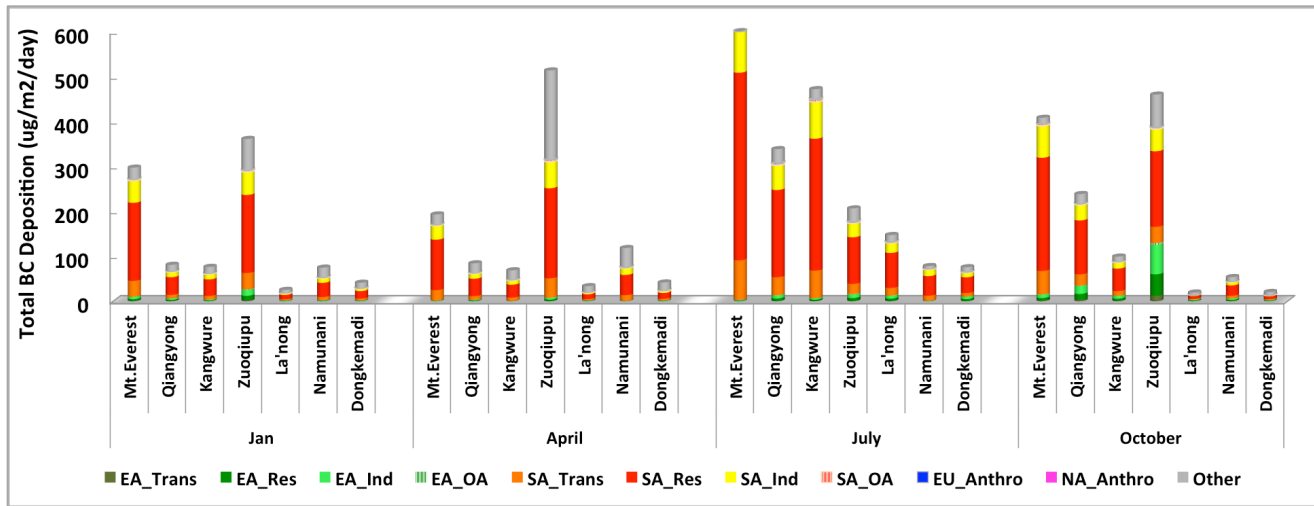
795 **Fig. 4 Premature mortality (deaths per 1000 km²) associated with global total BC**

796 **emissions, and land anthropogenic BC emissions from each tagged region in 2000:**

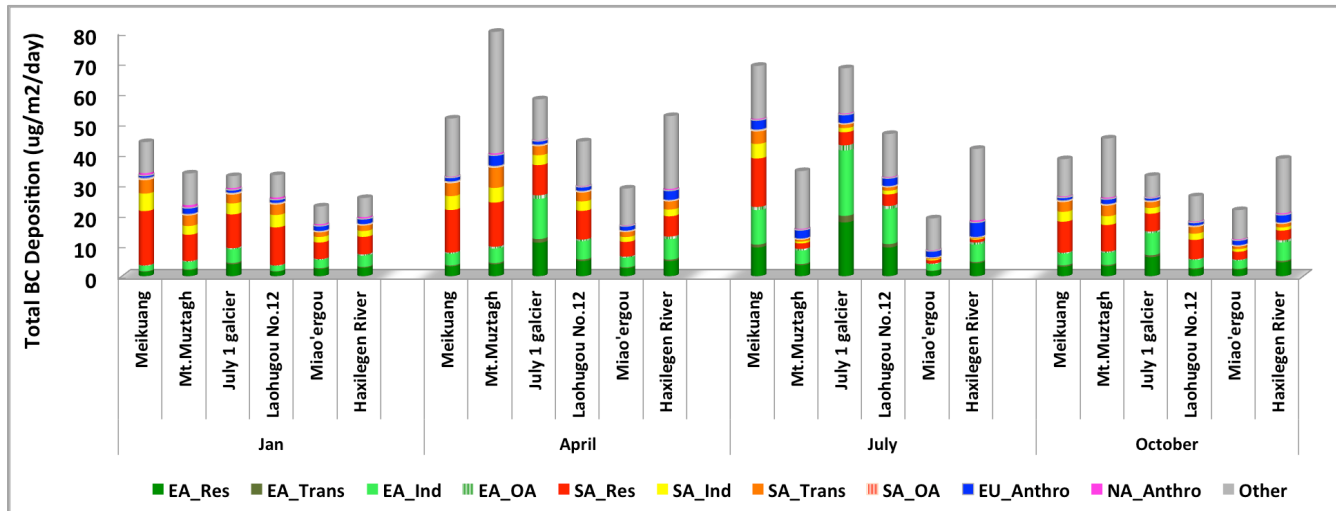
797 **East Asia (EA), South Asia (SA), Europe (EU), and North America (NA).**

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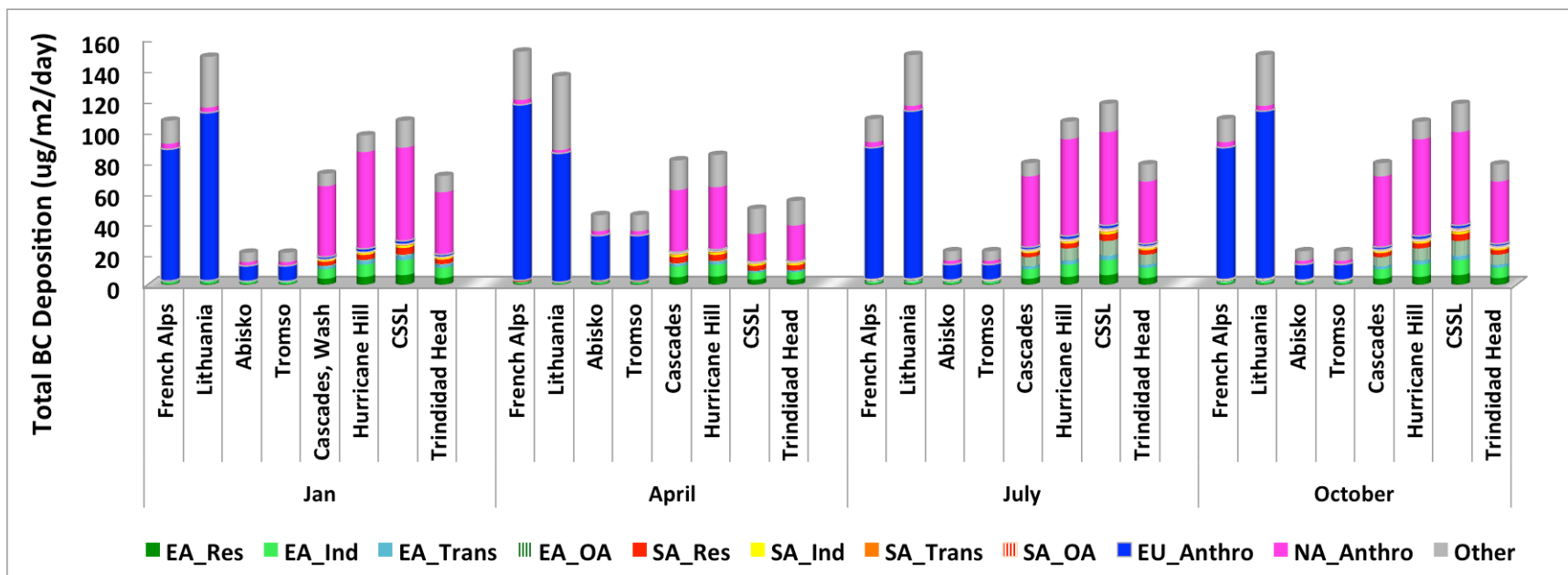
(a)



(b)

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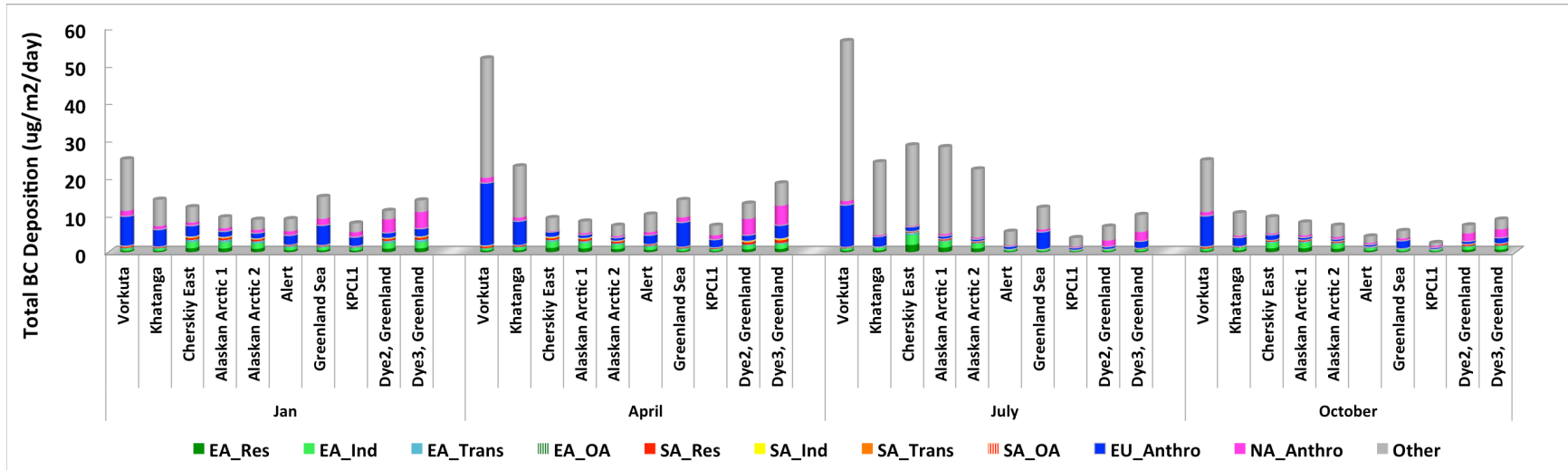
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(c)



(d)

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812 **Fig. 5 Total BC deposition on glaciers ($\mu\text{g}/\text{m}^2/\text{day}$) in (a) Southern and Central Himalayas and Tibetan Plateau (HTP), (b)**
 813 **northern HTP, (c) Europe and North America, and (d) Arctic. Arctic glaciers located in northwestern Russia (Vorkuta), north**
 814 **central Russia (Khatanga), northeastern Russia (Cherskiy East), Alaska (Alaskan Arctic 1 and 2), northern Greenland (Alert,**
 815 **Greenland Sea, KPL1), and southern Greenland (Dye2 and Dye3). See Fig. S3 and Table S1 for glacier locations.**

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818 **Table 1. Annual average premature mortality associated with BC exposure in each receptor region resulting from specific**
819 **regional and sectoral sources (unit: deaths).**

Sources	Receptors				
	EA	SA	EU	NA	World
Total BC	51,500 (15,600-69,400)	17,600 (8,000-27,900)	12,300 (7,300-15,500)	4,400 (2,800-4,700)	106,100 (44,900-143,400)
EA_Anthro	48,900 (95%) (14,800-65,700)	310 (2%) (96-506)	32 (0.3%) (20-42)	46 (1%) (25-65)	50,100 (47%) (15,300-67,500)
EA_Trans	3,800 (7%) (1,500-5,000)	30 (0.2%) (9-48)	4 (0.0%) (3-5)	6 (0.1%) (3-8)	3,900 (4%) (1,500-5,300)
EA_Res	18,100 (35%) (5,300-24,300)	120 (0.7%) (36-190)	11 (0.1%) (7-14)	16 (0.4%) (8-22)	18,500 (17%) (5,500-25,000)
EA_Ind	24,200 (47%) (7,100-33,000)	160 (0.9%) (49-255)	16 (0.1%) (10-21)	23 (0.5%) (12-32)	24,700 (23%) (7,400-33,800)
SA_Anthro	380 (0.7%) (129-616)	16,100 (92%) (7,200-25,800)	11 (0.1%) (7-16)	14 (0.3%) (8-19)	16,800 (16%) (7,500-26,800)
SA_Trans	65 (0.1%) (21-97)	2,600 (15%) (800-3,900)	2 (0.0%) (1-3)	2 (0.1%) (1-3)	2,700 (3%) (840-4,100)
SA_Res	240 (0.5%) (80-390)	10,400 (59%) (4,100-16,300)	7 (0.1%) (4-10)	9 (0.2%) (5-12)	10,800 (10%) (4,300-17,000)
SA_Ind	75 (0.1%) (25-115)	3,200 (18%) (1,000-5,000)	2 (0.0%) (1-3)	3 (0.1%) (2-4)	3,400 (3%) (1,100-5,300)
EU_Anthro	170 (0.3%) (62-215)	55 (0.3%) (19-81)	11,000 (89%) (6,500-13,800)	18 (0.4%) (12-22)	13,100 (12%) (7,700-16,000)
NA_Anthro	31 (0.1%) (11-40)	14 (0.1%) (5-21)	81 (0.7%) (50-103)	4,100 (93%) (2,600-4,300)	4,300 (4%) (2,700-4,600)

820 Total BC include emissions from all anthropogenic (land anthropogenic, aviation, and shipping) and natural sources. EA, SA, EU and
821 NA refer to tagged BC emissions from East Asia, South Asia, Europe and North America, respectively. Trans, Res and Ind refer to
822 tagged BC emissions from land transportation, residential and industrial sectors, respectively. Anthro refer to the sum of land
823 anthropogenic BC emissions in each tagged region.

824 Percentages shown in this table refer to the contributions of BC emissions from each source region and sector to total premature
825 deaths in the receptor region (i.e., EA total land anthropogenic BC emissions are responsible for 95% of total premature mortality in
826 East Asia associated with global total BC emissions).

827 Number ranges shown in this table refer to the lower and upper bound estimates in BC associated premature deaths due to
828 uncertainties in PM_{2.5} relative risks.

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