1 The PM_{2.5}-climate penalty: quantifying sensitivities in a

2 global climate model

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19 Abstract

20 Climate change can influence fine particulate matter concentrations (PM_{2.5}) through changes 21 in air pollution meteorology. The extent to which climate change can exacerbate or alleviate 22 air pollution in the future is an important aspect of robust climate and air pollution policy 23 decision-making. To examine the influence of climate on PM2.5, we use the Geophysical Fluid 24 Dynamics Laboratory Coupled Model version 3 (GFDL CM3), a fully-coupled chemistry-25 climate model, combined with future emissions and concentrations provided by the four 26 Representative Concentration Pathways (RCPs). For each of the RCPs, we conduct future 27 simulations in which emissions of aerosols and their precursors are held at present-day levels 28 while other climate forcing agents evolve in time, such that only climate (and thus 29 meteorology) can influence PM_{2.5} surface concentrations. We find a small increase in global

PM_{2.5} of about 0.21 µg m⁻³ (5%) for RCP8.5, a scenario with maximum warming. Changes in 1 2 global PM_{2.5} are at a maximum in the fall and are mainly controlled by sulfate followed by organic aerosol with minimal influence of black carbon. RCP2.6 is the only scenario that 3 projects a decrease in global PM_{2.5} with future climate changes, albeit only by -0.06 µg m⁻³ 4 (1.5%) by the end of the 21st century. Regional and local changes in PM_{2.5} are larger, reaching 5 upwards of 2 µg m⁻³ for polluted (eastern China) and dusty (western Africa) locations on an 6 7 annually averaged basis in RCP8.5. Using multiple linear regression, we find that future PM_{2.5} 8 concentrations are most sensitive to local temperature, surface wind, and precipitation, in that 9 order. Fine particulate matter concentrations are robustly positively associated with temperature, while negatively related with precipitation and wind speed. Present-day (2006-10 11 2015) modeled sensitivities of PM2.5 to meteorological variables are evaluated against 12 observations and found to agree reasonably well with observed sensitivities (within 10-50% 13 over the eastern United States for several variables), although the modeled PM_{2.5} is less 14 sensitive to precipitation than in the observations due to weaker convective scavenging. We 15 conclude that the hypothesized "climate penalty" of future increases in fine particulate matter is relatively minor on a global scale compared to the influence of emissions on PM2.5 16 17 concentrations.

18 **1** Introduction

19 In 2012, poor air quality resulted in up to 7 million premature deaths worldwide (WHO, 20 2014). Particles less than 2.5 µm in diameter, known as PM_{2.5}, are the major contributor to 21 poor air quality and have been linked numerous times to increases in mortality (Dockery et al., 22 1993; Dominici et al., 2006; Pope, 1999; Pope et al., 2004, 2006). The chemical composition 23 of PM_{2.5} is dependent on many factors, but in general, inorganic species such as sulfate, 24 nitrate, and ammonium along with a plethora of organic species comprise the majority of the 25 particulate mass (Jimenez et al., 2009a). Emissions of particulate matter and their precursors, from sources such as energy use and biomass burning, are the dominant contributors to PM_{2.5} 26 concentrations in the atmosphere (Chan and Yao, 2008; West et al., 2013). However, PM_{2.5} 27 28 levels depend on meteorological factors such as temperature, precipitation, and wind speed, 29 suggesting that PM_{2.5} can be influenced by climate change. For example, wet deposition is the main sink for atmospheric particulate matter, so increases in precipitation (particularly 30 31 frequency of precipitation) will lead to decreases in particle concentrations (Bernard et al., 2001; Dawson et al., 2007; Jacob and Winner, 2009; Jimenez-Guerrero et al., 2011; Tai et al., 32

1 2010). PM_{2.5} is also thought to be negatively sensitive to wind speed, as decreases in 2 ventilation will cause build-up of PM_{2.5} levels (Jacob and Winner, 2009; Porter et al., 2015). 3 PM_{2.5} may be positively associated with relative humidity (RH), as RH increases result in 4 increased aqueous uptake of semi-volatile materials (such as nitrate and organic aerosol); 5 however, an anti-correlation has been reported between sulfate and cloud cover, which calls the positive RH-PM_{2.5} relationship into question (Koch et al., 2003; Wise and Comrie, 2005). 6 7 Warmer temperatures can increase reaction rates for oxidation of sulfur dioxide (SO₂) to 8 sulfate (SO₄) aerosol by the hydroxyl radical (OH), but warmer temperatures are also less 9 favorable for condensation of semi-volatile materials to the aerosol phase. With increased 10 warming, variable changes in precipitation, and other significant climate changes projected 11 over the remainder of the 21st century, it is important to understand the effect that these 12 changes might have on PM_{2.5} and air quality.

13 Previous modeling studies on the effects of climate change on future air quality have yielded 14 mixed results. The majority of these studies have employed offline meteorological fields that 15 are then fed into a chemical transport or general circulation model. A review by Jacob and 16 Winner (2009) reported mostly an increase in $PM_{2.5}$ from climate change, ranging from $\pm 0.1 -$ 1 µg m⁻³, depending on the time horizon, region, and species (Heald et al., 2008; Jacobson, 17 2008; Liao et al., 2006; Racherla and Adams, 2006; Spracklen et al., 2009; Tagaris et al., 18 19 2007; Unger et al., 2006). However, two papers featured in the review, Pye et al. (2009) and 20 Avise et al. (2009), reported a change of -0.3 to +0.3 μ g m⁻³ (2050 vs 2000, US, sulfate, annual mean) and -1 µg m⁻³ (2050 vs 2000, US, PM_{2.5}, July mean), respectively. More 21 recently, Jiang et al. (2013) reported an annually averaged change in 2050 of -1.5 to +0.8 µg 22 m⁻³ for aerosol concentrations in eastern China due to climate change alone. Megaritis et al. 23 24 (2013) reported strictly decreases in aerosol concentrations in Central Europe in 2050 due to climate change because of dominating decreases in ammonium nitrate and organic aerosol 25 26 levels. Reflecting the uncertainty and discrepancies in these various estimates, the 27 Intergovernmental Panel on Climate Change (IPCC) assigns "no confidence level" to the 28 overall impact of climate change on PM_{2.5}, and only "low confidence" to the expectation that 29 increases in precipitation lead to decreases in PM2.5 (Kirtman et al. 2013) because of 30 confounding effects of the timing and location of changes in precipitation and PM2.5 (Fang et al., 2011). Although there is a significant amount of variability of the magnitude and even the 31 32 sign of the effect of climate on future air quality, virtually all studies agree that the magnitude of climate-driven changes in surface PM2.5 are dwarfed by emissions-driven changes (Fiore et 33

al., 2015; Gonzalez-Abraham et al., 2015; Hedegaard et al., 2013; Jiang et al., 2013; Kelly et

2 al., 2012; Trail et al., 2014; Unger et al., 2006; Val Martin et al., 2015; West et al., 2013).

Here, we build upon previous studies in several ways. First, we utilize a fully coupled 3 4 chemistry-climate model (GFDL CM3, see Sect. 2) with online meteorology, whereas almost all previous studies have relied on offline (input) meteorology. The offline meteorology 5 approach allows for shorter simulations (i.e., beginning and end simulations for the time 6 7 domain). A shortcoming of this method is that the short simulations make it more difficult to 8 separate climate change-induced signal versus climate variability in surface PM_{2.5} 9 concentrations. Our work employs a three-member ensemble of 95-year simulations, allowing 10 us to better estimate the climate-forced signal in surface PM_{2.5} from the noise. Second, we 11 conduct these full 21st century simulations using all four of the Representative Concentration 12 Pathways (RCPs) for future climate and emissions (modified such that aerosol and precursor emissions are held fixed at 2005 levels). To our knowledge this is the first study to analyze the 13 14 effect of future climate on PM2.5 using all four RCPs in fully transient simulations in a chemistry-climate model. Third, we focus on both global analyses and specific regions, which 15 16 exhibit larger PM_{2.5} changes and also contain large population centers. Fourth, we use a multiple linear regression (MLR) model to estimate sensitivities of surface PM_{2.5} to several 17 18 meteorological parameters, allowing for analysis of the magnitude and sign of the PM_{2.5}-19 meteorology association, following the method of Tai et al. (2010). We also compare the 20 present-day modeled MLR sensitivities to observed sensitivities based on PM2.5 and 21 meteorological data taken over the United States from 1998-2008 (Tai et al., 2010). Finally, 22 we include both annual and seasonal analyses as well as speciated results for individual PM_{2.5} 23 components. Although previous studies may have partially covered some of our goals, we aim 24 to provide a comprehensive, robust analysis as an additional piece of evidence to a growing 25 understanding of the effect of climate change on surface PM_{2.5}.

26 2 Models and analysis

27 2.1 GFDL CM3

We employ the Geophysical Fluid Dynamics Laboratory Coupled Model version 3 (GFDL CM3) in this work (Donner et al., 2011; Naik et al., 2013). CM3 is a fully coupled chemistryclimate model, featured in the latest Coupled Model Intercomparison Project (CMIP) as well as the Intergovernmental Panel on Climate Change Assessment Report 5 (IPCC AR5). A brief description of the model is provided below. CM3 has been extensively evaluated against
 observations including many of the model variables considered in this work. Further model
 description and evaluation details can be found in Donner et al. (2011) as well as Naik et al.
 (2013) and references therein.

The model consists of a cubed sphere finite-volume dynamical core with a horizontal grid 5 6 consisting of 6 faces with roughly a 200-km by 200-km spatial resolution. Vertical resolution consists of 48 levels extending from the surface up to about 80 km (0.01 hPa). Anthropogenic 7 8 emissions of reactive gases and aerosols and their precursors are based on estimates from 9 Lamarque et al. (2010) for the historical period (1860-2000) and from van Vuuren et al. 10 (2012) and Lamarque et al. (2011) for the RCP projections (2006-2100). Tropospheric chemistry is interactive with emissions and radiation modules and is based on Horowitz et al. 11 (2003) with updates from Horowitz (2006). Aerosol species include sulfate, black carbon, 12 primary organic aerosol (POA), secondary organic aerosol (SOA), sea salt, and mineral dust. 13 14 Sulfate is formed via oxidation of sulfur dioxide and dimethyl sulfide (DMS) by the hydroxyl radical (OH), ozone (O₃), and hydrogen peroxide (H₂O₂). SOA formation is parameterized by 15 16 apportioning fixed yields of terpene emissions and butane oxidation as biogenic and anthropogenic SOA, respectively. The model's SOA source of 40 Tg C yr⁻¹ globally (natural 17 plus anthropogenic) is likely to be an underestimate (Heald et al., 2008). Additionally, CM3 18 19 lacks inclusion of sophisticated treatment of SOA chemistry and includes few SOA 20 precursors. Advanced treatment of SOA chemistry and precursor species is undergoing 21 development at GFDL. Hydrophobic organic carbon aerosols (OA) and black carbon (BC) 22 aerosols are converted to hydrophilic aerosol with an e-folding time of 1.44 days. Sea salt and 23 mineral dust aerosol are treated with a five-section size distribution ranging from 0.1 to 10 µm 24 dry radius, and have emissions that depend on model meteorology. Changes in climate do not feedback on biogenic VOC or wildfire emissions, or partitioning of semi-volatile organics 25 26 from the gas phase into SOA, which are not temperature-dependent in this configuration of 27 CM3. The model also does not include any changes in the dust source availability in response 28 to climate-induced changes in soil moisture or vegetation. These omissions may lead to errors 29 in our estimate of the total influence of climate change on future PM2.5 levels (Heald et al., 30 2008; Spracklen et al., 2009). Particularly, increases in future wildfires and biogenic VOC 31 emissions due to temperature changes will likely result in a larger estimate PM2.5 climate 32 penalty.

1 Species included in our surface PM_{2.5} definition are limited to sulfate, BC, OA, and dust. We do not include sea salt particles in our analysis as our focus is mostly on anthropogenic 2 aerosols which impact human health more directly and comprise the vast majority of PM2.5 in 3 major world population centers (Jimenez et al., 2009b; Vallius et al., 2005; Yang et al., 2011). 4 5 Dust and sea salt are also typically coarser in size such that a large portion of their mass would fall outside of the 2.5 µm upper size cutoff (Pérez et al., 2008). Dust and sea salt levels 6 7 have been found to comprise no more than 10% of total PM2.5 mass concentration in several 8 Chinese megacities (Yang et al., 2011), but in five Mediterranean European cities, dust 9 contributed anywhere between 5-32% of total PM2.5 (Salameh et al., 2015). Thus, our inclusion of dust is important for estimating climate driven PM2.5, especially considering dust 10 11 emission changes are driven by climate (wind speed) alone in CM3. We do not currently 12 consider nitrate aerosol in PM2.5 changes, due to the relatively small current contribution of 13 nitrate to total aerosol mass in CM3 and the simplified manner in which nitrate chemistry and 14 thermodynamics is treated in the model (Paulot et al., 2016). However, in the future nitrate may play a larger role in PM_{2.5} due to decreasing sulfate (Bellouin et al., 2011), and a more 15 advanced treatment of nitrate including its interaction with radiation and clouds is undergoing 16 17 development at GFDL (Paulot et al., 2016).

18 **2.2** Simulations with 21st century climate scenarios

19 We use GFDL CM3 to evaluate the effect of future climate changes on PM_{2.5} concentrations. Since these simulations were previously carried out by Westervelt et al. (2015), we follow the 20 21 same simulation naming convention as used in that work. Simulations made up of three 22 ensemble members for each RCP were run from 2006-2100. These simulations are denoted RCPx.x 2005AER, where x.x = 2.6, 4.5, 6.0, or 8.5 for each of the four RCPs. In each 23 24 simulation, anthropogenic and biomass burning emissions of aerosols and their precursors are 25 held fixed at 2005 levels throughout the 95 years of the simulation, while other climate forcing agents including greenhouse gases follow the RCP scenario, thus allowing only 26 27 changes in climate and meteorology to perturb aerosol levels. Tracer concentrations and climate variables for each of the three ensemble members for each RCP were also initialized 28 29 from the ending conditions of 1860-2005 transient historical runs in CM3. We use these historical simulations from 1980-2005 in which aerosol emissions are allowed to vary in time 30 for comparison with climate-only changes in the 21st century. We present all results as 31 32 ensemble means.

1 The Representative Concentration Pathways (RCPs) contain emissions projections for all 2 long- and short-lived climate forcers, including the aerosol and aerosol precursor species SO₂, OA, and BC. The four pathways include a strong mitigation scenario (RCP2.6), two 3 stabilization scenarios in which radiative forcing stabilizes shortly after 2100 (RCP4.5 and 4 5 RCP6), and one high emissions/no mitigation scenario (RCP8.5). Details of the implementation of the RCPs into GFDL CM3 can be found in Westervelt et al. (2015). Since 6 7 recent GHG emissions are tracking at or above RCP8.5 (Peters et al., 2012; Sanford et al., 8 2014) we will focus mainly on RCP8.5 2005AER, and RCP8.5 2005AER should be viewed 9 as the most realistic scenario as of now, leading to the largest warming and other climate signals. We also present RCP2.6 2005AER as a lower bound. The other RCPs are included 10 11 mostly in the Supplementary Information.

12 2.3 Multiple linear regression

In order to determine the magnitude and sign of the effect of various climate parameters on surface PM_{2.5} concentrations, we use a multiple linear regression (MLR) approach (Tai et al., 2010). Unlike a correlation coefficient from a single linear regression, MLR allows for analysis of multiple independent variables (meteorological variables) affecting the dependent variable (PM_{2.5} concentrations) at the same time. The form of the model is:

18 $y = \beta_0 + \sum_{i=1}^6 \beta_i x_i + interaction terms$

(Equation 1)

19 In Eq. 1, the dependent variable y represents PM_{2.5} concentrations (annual or seasonal mean) 20 and is a function of the coefficient β_0 (y-intercept), the slope coefficients β_i , and the 21 independent variables x_i . Each of the independent variables x_i represents a climate or 22 meteorological variable that determines the values of $v_{1,0}$ or the PM_{2.5} concentration. We chose 23 six meteorological variables for inclusion in the MLR model: surface temperature (K), precipitation (mm d⁻¹), total cloud amount (in all vertical layers) (%), 10 m wind speed 24 magnitude (m s⁻¹), sea level pressure (hPa) and relative humidity (%). Both the modeled PM_{2.5} 25 26 and meteorological variables are monthly averages beginning in January 2006 and ending in 27 December 2100. The slope coefficients β_i represent the amount of change in PM_{2.5} concentrations for a unit change in the meteorological variables x_i if all other climate variables 28 x are held constant. These values can be interpreted as "sensitivities" of $PM_{2.5}$ to a change in 29 climate, and have units of µg m⁻³ D⁻¹, where D refers to the units for the particular 30 31 meteorological variable (e.g., temperature in K). Positive coefficients refer to an increase in 1 PM_{2.5} concentrations for an increase in the meteorological variable (positive correlation), 2 whereas negative values refer to decreases in PM_{2.5} (anti-correlation). Interaction terms, which 3 refer to the product of two or more independent variables and represent the higher-order linear 4 effects when meteorological variables are interacting with each other in affecting PM_{2.5}, are 5 found to be small and thus excluded from the analysis.

6 We also present MLR slope coefficients for observations over the United States domain for comparison against model results. Meteorological monthly mean data (derived from daily 7 8 data) from 1998 to 2008 from the National Center for Environmental Prediction/National 9 Center for Atmospheric Research (NCEP/NCAR) Reanalysis 1 is used for the MLR 10 estimatation (http://www.esrl.noaa.gov/psd/data/gridded/data.ncep.reanalysis.html) (Kalnay et al., 1996). The Reanalysis data is gridded on a 2.5° by 2.5° latitude-longitude grid. For PM_{2.5}, 11 we calculated monthly means of daily mean surface concentrations of total PM2.5 from the 12 EPA Air Quality System (EPA-AQS, http://www.epa.gov/aqs), which operates a network of 13 about 1000 stations in the United States. The modeled and observed PM2.5 and meteorological 14 values were detrended and deseasonalized by subtracting the three-month running means from 15 each original month. Details on both the meteorological and PM data can be found in Tai et 16 al. (2010), which applied the same methodology but on the daily deviation of PM_{2.5} from a 17 18 30-day moving mean to focus on the synoptic scale.

19

20 3 PM_{2.5} changes

We first present our PM_{2.5} changes over time with our 2005 fixed aerosol emissions simulations (RCPx.x_2005AER) in Sect 3.1, Fig. 2, and Table 1. Then, spatial distributions of the change of PM_{2.5} from the beginning of the 21^{st} century to the end (Δ PM_{2.5}, Eq. 2) are presented in Figs. 3 and 4. Finally, we present the regional results in Sect. 3.2 and Fig. 3 and Tables 2 and 3. Region definitions are provided in Fig. 1.



- 1 Figure 1: Region definitions. North America (NM), South and Central America (SM), Europe
- 2 (EU), Russia (RU), East Asia (EA), South Asia (SA), Australia (AU), Africa (AF), Middle
- 3 East (ME)
- We report values of population-weighted surface $\Delta PM_{2.5}$ as the change in concentration from a ten-year ensemble mean of 2006-2015 to a ten-year ensemble mean of 2091-2100, i.e. the change from the beginning to the end of the 21st century. All results are for surface-level PM_{2.5}. In equation form:

8
$$\Delta PM_{2.5} = PM_{2.5}(2091 - 2100) - PM_{2.5}(2006 - 2015)$$
 (Equation 2)

9 Population weighted concentrations were calculated by summing the product of population10 and concentration for a specific grid cell divided by the total population:

11
$$C_{w} = \frac{\sum_{j} C_{j} P_{j}}{\sum_{j} P_{j}}$$
 (Equation 3)

12 where C_j indicates concentration at each grid cell, P_j is population at each grid cell, and C_w

13 is the weighted concentration. Population data was obtained from the CIESEN 2005 Gridded

14 Population product (http://sedacciesincolumbiaedu/plue/gpw).

15 **3.1 Global mean** Δ**PM**_{2.5} from 1980-2100

16 Figure 2 shows the time evolution of globally and annually averaged speciated and total surface PM_{2.5} from 1980-2100. "Total" PM_{2.5} refers here to the sum of sulfate (and its 17 18 associated ammonium), black carbon, organic aerosol, and dust less than 2.5 µm in diameter. Results from the historical runs are shown in purple from 1980-2005, followed by results 19 20 from the four RCPx.x 2005AER simulations for the rest of the 21st century. For the historical 21 time series in PM2.5, aerosol and precursor emissions were not fixed; they instead were 22 allowed to vary according to the historical emissions inventory of Lamarque et al. (2010). Starting in 2005, all anthropogenic and biomass burning aerosol emissions are held fixed at 23 24 2005 levels for the remainder of the simulation of the 21st century, allowing us to isolate only the effect of meteorology. The results plotted for the 25-year period 1980-2005, during which 25 26 emissions vary, has about the same range or an even larger range of PM_{2.5} concentrations $(\Delta PM_{2.5} \text{ of } \sim 0.15 \text{ } \mu \text{g m}^{-3})$ as the effect of meteorology over the entire 21st century for any of 27 the RCPs (0.21 μ g m⁻³ for RCP8.5 2005AER). This is particularly noticeable in the black 28

1 carbon and organic aerosol time series (Fig. 2, b and c) in which the variable-emissions trend from 1980-2005 increases dramatically, followed by little to no increase for the next 95 years 2 3 due to meteorology only. For sulfate (Fig. 2a) and the sum of sulfate, BC, OA, and dust, the 4 $\Delta PM_{2.5}$ from 2006 to 2100 is of comparable magnitude to the emissions-driven changes of 5 ΔPM_{2.5} from 1980-2005, at least in RCP8.5 and RCP2.6. Thus, since a century of climatedriven changes in PM_{2.5} are less than or equal to emissions-driven changes over only 25 years, 6 7 we conclude that changes in emissions are the dominant driver for future PM2.5 concentrations, with fairly minor meteorological changes, at least on the global, annual scale. 8 9 In particular, black carbon aerosols are very weakly affected by precipitation perhaps owing 10 to their lower water affinity than sulfate aerosols, which are most strongly affected. 11 Additionally, production of secondary aerosols such as sulfate is climate-dependent, whereas 12 emissions of primary aerosols such as BC are not as affected by climate.

Westervelt et al. (2015) found similar results for aerosol optical depth (AOD) which was far
more sensitive to RCP emissions reductions than climate changes with emissions held fixed
(see Fig. S3 of Westervelt et al. (2015)).





18

17 Figure 2: Global, annual mean trend in $PM_{2.5}$ concentration for sulfate (a), black carbon (b),

organic aerosol (c), and total (including fine dust) (d) for each of the RCPx.x 2005AER

19 simulations. Emissions vary in time only for the 1980-2005 pre-RCP time period.

1 As shown in Fig. 2, PM_{2.5} concentrations are weakly affected by meteorology on a global 2 scale. The high-warming scenario, RCP8.5, projects a consistent but slight increase in PM_{2.5} as the 21st century progresses. This increase in PM_{2.5} can be considered a "climate penalty" 3 (Wu et al., 2008), due to increases in PM_{2.5} caused solely by climate change (Fang et al., 4 5 2013). Table 1 presents the $\Delta PM_{2.5}$ concentrations differences for ten-year averages from the beginning to the end of the century. Table 1 and Fig. 2 clearly show that the meteorological 6 impact on PM2.5 is strongly dependent on the RCP scenario. For example, total $\Delta PM_{2.5}$ for 7 RCP2.6 2005AER is -55.4 ng m⁻³ and +210 ng m⁻³ for RCP8.5 2005AER, indicating that 8 9 even the sign of the change (increase or decrease in PM2.5) is dependent on the particular 10 climate change scenario. Exceptions to this are OA and BC, which have a positive $\Delta PM_{2.5}$ for 11 all RCPs. The order of the four trends over the 21st century, especially visible in the sulfate 12 and total PM_{2.5} concentrations (Fig. 2, a and d), mirrors the order of warming in each of the 13 simulations. In general, RCP8.5 2005AER results in the largest $\Delta PM_{2.5}$, followed by 14 RCP6 2005AER, RCP4.5 2005AER, and RCP2.6 2005AER. This order comes expected, as 15 RCP8.5 2005AER results in the largest amount of warming and other meteorological change (e.g., precipitation), while RCP2.6 2005AER features stringent climate policies and thus 16 relatively little warming (Table 1). 17

Our globally averaged $\Delta PM_{2.5}$ results are comparable with Fang et al. (2013), although their estimate (0.28 µg m⁻³) using a prototypical version of GFDL AM3 (atmospheric component only) and an older set of emissions scenarios (IPCC SRES) is slightly above the RCP8.5_2005AER global average of 0.21 µg m⁻³. This small positive enhancement in PM_{2.5} due to climate change also falls in the range of PM sensitivity to climate (±0.1 – 1 µg m⁻³) as defined by the Jacob and Winner (2009) review paper and reiterated by Dawson et al. (2014).

24 3.2 Spatial distribution of seasonal and annual $\Delta PM_{2.5}$

25 Spatial distributions of seasonally and annually averaged total $\Delta PM_{2.5}$ (end of 21st century 26 minus present day) are shown for RCP8.5_2005AER in Fig. 3.





Figure 3: Seasonal (a-d) and annual (e) average ΔPM_{2.5} for a difference of 2091-2100 average
and a 2006-2015 average in RCP8.5 2005AER

4 In each season, the impact of meteorology on PM2.5 concentrations is largely an increase in most regions, with some exceptions. While annual globally averaged $\Delta PM_{2.5}$ were less than 5 0.3 µg m⁻³, over certain regions, annual concentration changes can reach 1-2 µg m⁻³. Increases 6 are typically found over populated continental regions such as Europe, China, and North 7 8 America. Changes (mainly increases) over Western Africa reflect increases in dust PM2.5 due 9 to increases in wind speed. PM2.5 increases over the tropical Pacific Ocean are caused by 10 decreases in wet deposition in CM3 associated with a warming climate (Westervelt et al., 11 2015). In certain seasons, PM_{2.5} concentration changes are even larger, resulting in a larger "climate penalty". For example, in the autumn months (SON, Fig. 3d) total $\Delta PM_{2.5}$ 12 differences are upwards of +2 μ g m⁻³, with a global average increase of 0.31 μ g m⁻³, the 13 14 largest of any season. Pye et al. (2009) also found peak aerosol concentrations in fall when looking solely at meteorological driven changes. In SON, sulfate PM_{2.5} increases more than 15 16 any other season with organic aerosol increases being at second highest values out of the four 17 seasons, which explains the large increases in PM2.5. Total PM2.5 increases are at a minimum global mean value of 0.16 µg m⁻³ in the northern hemisphere summer months (JJA) due to a 18

1 corresponding minimum in the sulfate aerosol increases (dominant species by mass), despite a 2 summertime peak in the increases for organic aerosol. Details of speciated seasonal results for 3 sulfate, OA, and BC are shown in the Supplement Fig. S1. The seasonal averages reveal 4 pockets of negative $\Delta PM_{2.5}$ (aerosol decreases) that are largely masked in the annual spatial 5 distribution. We find that not only the magnitude but also the sign of $\Delta PM_{2.5}$ is highly dependent on the season in some areas. For example, South America exhibits positive $\Delta PM_{2.5}$ 6 7 in DJF and SON but negative $\Delta PM_{2.5}$ in the other months. The speciated results (Fig. S1) 8 reveal that these decreases are driven entirely by decreases in organic aerosol, as sulfate and 9 black carbon either increase or do not change. This finding implies a strong meteorological 10 influence on organic aerosol aerosol in the northern hemisphere winter (DJF) and fall (SON). 11 Fang et al. (2013) also found decreases in PM2.5 in South America and attributed the change 12 to changing precipitation patterns. We confirm that PM_{2.5} is negatively associated with 13 precipitation in South America and globally, which is discussed in more detail in Sect. 4.

- 14 The spatial distribution of $\Delta PM_{2.5}$ in RCP2.6_2005AER differs from RCP8.5_2005AER (Fig. 15 4 vs. Fig. 3), with large decreases over most of the continents balanced with pockets of
- 16 increases in RCP2.6 2005AER.



17

18 Figure 4: Same as Figure 3, but for RCP2.6 2005AER

1 Annually, this results in a slight overall decrease in global mean PM_{2.5} due to meteorological 2 changes (-0.03 µg m⁻³), suggesting a slight "climate benefit" overall. Decreases are occurring largest over the summer months (JJA, Fig. 4c), in which concentration decreases are found 3 over nearly all of Europe and US, for a global average of -0.08 µg m⁻³. As shown in 4 5 Supplement Fig. S2, the large summer decreases across Europe and the eastern US come largely from sulfate decreases, with secondary influence of organic aerosol. Globally averaged 6 JJA sulfate $\Delta PM_{2.5}$ is about -0.08 µg m⁻³, whereas BC and OA show essentially zero change 7 8 globally, with decreases in some regions offset by increases. The large decreases in the 9 eastern US and Europe are consistent across all seasons, with the slight exception of European 10 $\Delta PM_{2.5}$ in the winter months (DJF). However, as found for RCP8.5 2005AER, $\Delta PM_{2.5}$ can be 11 very different for the same location in different seasons. As is the case with 12 RCP8.5 2005AER, the largest changes in magnitude (positive or negative) in global mean 13 PM_{2.5} occur in the summer and fall, with relative minima in the spring and winter. Annual and 14 seasonal spatial distributions of $\Delta PM_{2.5}$ for RCP4.5 2005AER and RCP6.0 2005AER are 15 shown in Supplemental Figs. S3 and S4, and fall between the results of RCP2.6 2005AER 16 and RCP8.5 2005AER.

17 **3.3 Regional average** Δ **PM**_{2.5}

18 Figure 5 presents the total $PM_{2.5}$ (sulfate + OA + BC + fine dust) annual mean time series 19 from 1980-2100 for four key regions (North America, East Asia, South Asia, and Russia) as 20 defined by Fig. 1. Total $\Delta PM_{2.5}$ is presented in Table 2 for RCP8.5 2005AER and Table 3 for 21 RCP2.6 2005AER for all of the regions in Fig. 1. Four regions were chosen for further 22 analysis because they represent some of the largest PM2.5 changes driven by meteorology and 23 include major population centers. As with Fig. 2, the historical model emissions are varying in 24 time (not fixed) from 1980-2005. These historical trends are guite different for each of the 25 regions. For example, North America had largely begun to enact and enforce air quality regulations by the latter part of the 20th century, which is reflected by a decreasing trend in the 26 PM_{2.5} concentrations. East and South Asia (Fig 5, b and c) PM_{2.5} increases by up to 3 to 4 µg 27 m⁻³, reflecting rapid industrialization in these parts of the world. Concentrations of PM_{2.5} in 28 29 Russia (Fig. 5, panel d) first increase until about 1990 and then decrease strongly for the next decade, indicative of a slowing economy after the collapse of the Soviet Union. By the 30 beginning of the 21st century, PM_{2.5} are again trending upwards due to economic recovery in 31 32 Russia.



Figure 5: Annually averaged timeseries of total PM_{2.5} concentration for each of the
RCPx.x_2005AER simulations for the North American region (a), East Asia (b), South Asia
(c), and Russia (d) (see Fig. 1 for region definitions).

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Compared to the global time series in Fig. 2 and Table 1, 21st century meteorologically-driven 6 7 changes in regional PM2.5 are much larger. For example, for RCP8.5 2005AER East Asia (average of China, Japan, and Korea) has an annual average $\Delta PM_{2.5}$ of about 0.4 µg m⁻³, and 8 the value for Russia is 0.44 µg m⁻³ (see Table 2). The largest regional increase is in Africa at 9 $0.64 \ \mu g \ m^{-3}$, however this is one of the few regions where dust PM_{2.5} concentrations increase 10 11 due to climate change (wind speed increases), whereas they decrease significantly over East Asia, South Asia, Europe, Middle East, and Russia. Grid boxes over certain populous cities in 12 these regions exhibit $\Delta PM_{2.5}$ of up to 1 to 2 µg m⁻³, representing a more significant climate 13 14 penalty. Though subject to uncertainty and variability in specific studies, epidemiologists 15 have surmised up to a 6% and 11% increase in all-cause and cardiovascular mortality for every incremental 10 µg m⁻³ increase in daily PM_{2.5} concentrations (Hoek et al., 2013), 16 17 suggesting that our projected climate-driven concentration increases may exacerbate local air 18 pollution related mortality. However, these meteorologically-driven PM2.5 changes represent 19 less than a 10% change in total PM_{2.5} over highly polluted areas such as eastern China. 20 Additionally, the linear relationship between increased mortality and PM_{2.5} cited above does not hold in highly polluted environments with high base concentrations. Still, in less or 21

1 moderately polluted areas, such as the Pacific Northwest U.S., the relative increase in $PM_{2.5}$ 2 can be as high as 40%.

As was the case with the global results, the total $\Delta PM_{2.5}$ from the beginning of the 21st century 3 4 to the end is much smaller than the change from 1980-2005, highlighting the dominance of emissions over meteorology in driving PM2.5 changes at the regional level. The order of the 5 6 PM_{2.5} concentrations in 2100 also reflects differences inherent to the RCPs, with RCP8.5 2005AER having the largest changes in meteorological values such as temperature, 7 8 precipitation, etc., and therefore the largest impact of changing climate on PM_{2.5}. Regional 9 increases in temperature and precipitation for each of the RCPs can be found in Westervelt et 10 al. (2015). This order is most prevalent in the North America and Russia regions. PM2.5 11 concentrations in South Asia increase the most in RCP6 2005AER, and East Asian PM_{2.5} 12 trend splits into two regimes in which RCP8.5 2005AER and RCP6.0 2005AER increase throughout the century nearly identically, while RCP4.5 2005AER and RCP2.6 2005AER 13 14 decrease nearly identically by the same amount. Indeed the $\Delta PM_{2.5}$ for East Asia in RCP2.6 2005AER and RCP4.5 2005AER is -0.51 and -0.55 µg m⁻³, and +0.40 and +0.43 µg 15 m⁻³ for RCP8.5 2005AER and RCP6.0 2005AER (see Tables 2-3 and Tables S1-S2). 16

17 Speciated results shown in Tables 2-3 provide some insight into which components of PM_{2.5} 18 are driving the 21st century trends. For RCP8.5 2005AER, the increases are nearly half driven 19 by sulfate and half driven by organic aerosol for many regions, with a few exceptions. South America has a trivial organic aerosol increase (0.014 μ g m⁻³) such that 87% of the Δ PM_{2.5} 20 increase is due to sulfate, and Europe has a smaller sulfate increase such that only 35% of the 21 22 increase is caused by sulfate increases. BC plays a small role in total PM2.5 changes due to a 23 lack of atmospheric abundance by mass, contributing at most 6-9% in East and South Asia but 24 significantly less elsewhere. For RCP2.6 2005AER, overall $\Delta PM_{2.5}$ concentrations are negative in all regions except Africa. This is entirely caused by strong sulfate decreases, 25 which occur in all regions. On the other hand, positive values of organic aerosol $\Delta PM_{2.5}$ 26 27 (increases) are still predicted for all but one region (South America). We therefore conclude 28 that organic aerosol concentrations primarily increase due to climate change and that this 29 increase is robust across virtually all RCPs and all regions. BC concentration changes, although minor, are also consistently positive. Sulfate $\Delta PM_{2.5}$ changes may be either positive 30 31 or negative depending on the RCP (and thus the climate scenario). Sulfate $\Delta PM_{2.5}$ tends to be 32 positive (increases) globally and across most regions for RCP8.5 2005AER and 1 RCP6.0_2005AER, the two scenarios with the most extreme climate changes. Sulfate $\Delta PM_{2.5}$ 2 is largely negative (decreases) for RCP4.5 2005AER and RCP2.6 2005AER.

3 4 PM_{2.5} dependence on meteorology

4 4.1 Modeled PM_{2.5}

5 In order to determine the relationship between changing meteorology and PM_{2.5} changes, we use the multiple linear regression (MLR) model described in Sect. 2.3. Figures 6 and 7 show 6 7 plots of the annual average coefficients β_i , or slope coefficients, for RCP8.5 2005AER and RCP2.6 2005AER for the 21st century. These coefficients are essentially "sensitivities" of 8 9 PM_{2.5} to unit changes in each of 6 meteorological variables assuming all other meteorological variables are held fixed. For a seasonal analysis of the PM_{2.5} dependence on meteorological 10 11 parameters, Figs. S5 and S6 show DJF and JJA averages for RCP8.5 2005AER. Figure 6 12 provides a broad look at the sign of the relationship between each meteorological variable and PM_{2.5}, with red values representing a positive PM-meteorological correlation, and blue values 13 representing a negative correlation. The MLR model explains 50% to 80% of the variation in 14 PM_{2.5} for RCP8.5 2005AER, with the largest percentages found over the continents. "The 15 16 correlation matrix shown in the supplemental information shows that each of the independent 17 variables (the meteorological variables) are not strongly correlated with one another. Thus, we 18 consider multicollinearity to be minimal."

For RCP8.5 2005AER, the PM2.5-temperature relationship is robustly positive across the 19 20 continents and oceans, with minor exceptions in locales of East and South Asia. This is 21 consistent with studies compiled by Jacob and Winner (2009). In GFDL CM3, this positive 22 correlation likely comes temperature-dependent reaction rates and changes in oxidant 23 abundances, which can enhance oxidation of aerosol precursors (such as SO₂) with higher 24 temperature, leading to more PM_{2.5}. There are other possible feedbacks of temperature on PM, 25 such as the relationship between organic aerosol volatility and temperature, but these 26 mechanisms are not included in this version of the model. PM2.5 sensitivity to temperature can 27 also manifest in changes in stagnation and circulation. Taking the average slope coefficient in the eastern US as 0.15 µg m⁻³ K⁻¹ and projected RCP8.5 temperature increases of 4 K leads to 28 a PM_{2.5} increase of 0.6 µg m⁻³, which would represent most if not all of the observed PM_{2.5} 29 30 increase due to all meteorological variables in that region.

1 Similar scaling arguments can be made for the other meteorological parameters. PM_{2.5} is also 2 found to be sensitive to precipitation, which is consistently anti-correlated due to the large scavenging sink of wet deposition. Precipitation changes are typically more positive than 3 negative from the beginning to the end of the 21st century (increases) although there is 4 5 significant variation across regions. As expected, the relationship between PM2.5 and 6 precipitation is largely negative (anti-correlated), indicating that areas with precipitation 7 decreases yield PM_{2.5} increases and vice versa. The largest negative sensitivities (Fig. 6b) are 8 found over the Middle East, parts of Africa, and Europe. The strong sensitivity of PM_{2.5} to 9 precipitation over most of Africa and the Middle East, some of the driest regions in the world, 10 suggests that precipitation is an important PM2.5 sink even in areas where it infrequently happens. Using an average value of the slope coefficients of roughly -1.0 µg m⁻³ mm⁻¹ d and a 11 precipitation increase in Europe of about 0.1 mm d⁻¹ results in a -0.1 µg m⁻³ change in PM_{2.5} 12 13 due to precipitation only, suggesting that the precipitation effect is moderately strong but likely not larger than the influence of temperature (~0.5 µg m⁻³) in CM3. The moderate 14 sensitivity of PM_{2.5} to precipitation is in part caused by a weak convective scavenging sink in 15 this version of CM3 (Paulot et al., 2016; Fang et al. 2011). 16

We also find in the model that PM_{2.5} is consistently anti-correlated with wind speed (Fig. 6d), 17 as negative values of the slope coefficients appear over most continental regions. Lower wind 18 19 speeds lead to less ventilation, causing an increase in PM2.5, as is discussed in Jacob and Winner (2009). Similarly, higher wind speeds cause more mixing and dilution, which may 20 21 decrease PM_{2.5} levels. An exception is in Saharan Africa, the Middle East, and Australia, 22 which have a large positive correlation between wind speed and PM_{2.5}. This is likely due to 23 the strong wind speed dependence of dust emissions. Sea salt is not included in PM2.5 in this 24 study, but would result in large positive correlations with wind speed over the Southern 25 Ocean. Typical wind speed changes from the beginning to the end of the century are much larger over the oceans and are typically no greater than ± 0.4 m s⁻¹ over the continents. With 26 wind speeds in CM3 slightly decreasing through the 21st century over the continents (due to 27 decreases in temperature gradients), sensitivities of -0.5 µg m⁻³ m⁻¹ s could lead to PM_{2.5} 28 29 increases of up to 0.2 µg m⁻³, which is a substantial portion of projected regional and global $\Delta PM_{2.5}$. Since relatively humidity, sea level pressure, and cloud cover sensitivities are not 30 substantial (discussed in detail below), we conclude that after temperature, wind speed 31 decreases may be the second-most important contributor to PM_{2.5} increases. 32

Fig. 6e shows the slope coefficients β_i for the relationship between PM_{2.5} and sea level 1 pressure. We find that the slopes are positive nearly everywhere, with somewhat modest 2 sensitivities (~0.1-0.2 µg m⁻³ hPa⁻¹) except for in the tropical Pacific near Malaysia, 3 Indonesia, and the Philippines. Despite being a typically low-pressure region at present, future 4 5 shifting of the Hadley circulation and the ITCZ due to greenhouse gas forcing could lead to an increase in sea level pressure which in turn causes the strong sensitivity of PM_{2.5} over the 6 7 remainder of the 21st century seen here. Generally, high-pressure systems can lead to stable 8 atmospheric conditions and a shallow mixing layer, resulting in poor ventilation and therefore 9 an increase in PM_{2.5}.

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Figure 6: Multiple linear regression coefficients (β_i, "slope coefficients") for the dependence
of PM_{2.5} on six meteorological parameters in RCP8.5_2005AER

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The effect of relative humidity on PM_{2.5} (Fig. 6f) is small and mostly limited to tropical regions in South Asia. In these regions RH is found to be mostly anti-correlated with dry PM_{2.5}, with minor exceptions most notably in India and southwestern North America. Since RH-dependent inorganic aerosol formation is not currently included in this version of the

1 model, changes in moisture and cloud cover likely explain this negative relationship. Indeed, 2 the relationship between total cloud fraction and PM_{2.5} is found to be negative in this region 3 (Fig. 6c), as well as large-scale anti-correlation across most of Eurasia. In regions with this 4 anti-correlation, clouds can partially inhibit photochemical oxidation of aerosol precursors 5 (such as sulfur dioxide and VOCs) from occurring. Regions where cloudiness and PM2.5 are 6 positively correlated, such as India and the southeastern United States, are likely indicative of 7 enhancements in in-cloud production of sulfate aerosols. However, with relatively small slope 8 coefficients in both the PM-fcld and PM-RH regressions, the impact of relative humidity on 9 PM_{2.5} in CM3 is small.

10 Figure 7 shows the same annually averaged slope coefficients β_i except for RCP2.6 2005AER. There are not many major qualitative differences from RCP8.5 2005AER 11 (Fig. 6) in the RCP2.6 2005AER coefficients across each of the six parameters. Some 12 exceptions include temperature in Fig. 7a, which for RCP2.6 2005AER includes many 13 regions of negative correlation with PM2.5. There are also some significant differences in the 14 PM-wind regression (Fig. 7d), especially over North America where sensitivity values are 15 much less negative in RCP2.6 compared to RCP8.5 2005AER. There are several major 16 17 differences in the trends in meteorological variables in RCP2.6 2005AER and 18 RCP8.5 2005AER; for example, global temperatures in RCP2.6 2005AER are only projected 19 to increase by 2 K at most, compared to 4-5 K in RCP8.5 2005AER. These smaller changes 20 in temperature in RCP2.6 2005AER result in smaller increases in PM2.5, which combined 21 with large increases in precipitation in RCP2.6 2005AER, result in an overall decrease in 22 RCP2.6 2005AER PM_{2.5} (Fig. 2). The regression plots for both RCPs are largely consistent. 23 Intuitively, the magnitude or sign of the meteorological perturbation should not fundamentally 24 change the nature of PM2.5 sensitivity to meteorology. In other words, whether or not 25 precipitation increases or decreases by a certain amount, PM2.5 should respond in a way such that the sensitivities (slope coefficients) are consistent with each other since the underlying 26 27 mechanisms driving the relationship do not change, i.e. PM2.5 concentration decreases for 28 precipitation increases and vice versa.

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2 Figure 7: Same as Figure 6 but for RCP2.6_2005AER

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4 4.2 Comparison of modeled PM_{2.5} sensitivities to observations

5 Though several modeling studies have performed similar regression analyses for PM2.5 and have compared results to each other with mixed agreement, none have evaluated their model 6 7 against regression coefficients derived from observations. Here, we conduct such an analysis. 8 Our goal is to compare monthly PM_{2.5} sensitivities in the model and observations for similar 9 time periods with longer-term trends removed (e.g. from emissions) in order to isolate the 10 impact on meteorological variability on PM2.5 concentrations. We do not aim to compare PM_{2.5} sensitivities over time periods influenced by emissions trends. Figure 8 shows the 11 12 modeled and observed MLR slope coefficients over the United States. Monthly mean modeled slopes from 2006-2015 in the RCP8.5 2005AER simulations are compared to 13 monthly mean slopes from 1998-2008 in the observations. Both observed and modeled 14 15 monthly values are detrended and deseasonalized by subtracting the 3-month moving average 16 from each original monthly value. These monthly model slopes over the US therefore different from those described in Sect 4.2 in that they represent the short term variability in 17

early 21st century PM_{2.5} sensitivity rather than longer term climate-driven changes. This 1 2 conversion is necessary to facilitate consistent comparison with the observations. As can be seen in Figs. 8a and 8b, GFDL CM3 captures the PM2.5-T relationship well in the eastern and 3 4 midwestern US, with large positive slopes across the entire domain. Modeled slope 5 coefficients for temperature are uniformly positive across the Western US, whereas the 6 observations suggest slopes near zero or at times negative, with the exception of southern 7 California and parts of the mountain west. The domain average modeled and observed slopes are 0.09 and 0.07 µg m⁻³ K⁻¹, respectively. For precipitation (Figs. 8c and 8d), GFDL CM3 is 8 9 underestimating the magnitude of the slope coefficients, although capturing the sign (anti-10 correlation) correctly in the mountain west and northeast in particular. PM_{2.5} in GFDL CM3 is 11 much less sensitive to changes in precipitation than seen in the observations. CM3 shows some spatial agreement in the sign of the slopes in certain locales such as the upper midwest 12 13 and parts of the eastern US, but overall the model fails to capture the magnitude of the slopes. This is evidenced by the domain mean model slope of $-0.06 \ \mu g \ m^{-3} \ mm^{-1} \ d \ versus -0.41 \ \mu g \ m^{-3}$ 14 ³ mm⁻¹ d in the observations. This discrepancy is likely due to weak convective removal of 15 aerosols in this version of CM3, which has recently been investigated by Paulot et al. (2016). 16



Figure 8: Detrended and deseasonalized monthly multiple linear regression coefficients over
the United States for CM3 (left column) and observations from EPA-AQS and NCEP
reanalysis (right column).

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5 PM_{2.5}-wind sensitivities in both the model and the observations agree well spatially in sign 6 and magnitude primarily in the Eastern US. Model predicted slopes are strongly negative 7 across most of the US, with the largest values occurring in the eastern and midwestern US (-1 to -2 µg m⁻³ m⁻¹ s). Observed values of PM-wind slopes are large over California, whereas the 8 model predicts modest negative slopes (-1 µg m⁻³ m⁻¹ s) over this region. The domain 9 averages in the model and the observations are in excellent agreement at -0.294 and $-0.286 \mu g$ 10 m⁻³ m⁻¹ s. The PM_{2.5} versus RH in both the model and the observations agree in that most of 11 the PM_{2.5} domain is largely insensitive to RH. However, observed PM-RH regression slopes 12 in the model and observations correlate poorly (domain mean of -0.02 and 0.01 μ g m⁻³ %⁻¹ in 13 14 the model and observations, respectively). Finally, total cloud amount is positively correlated

1 with PM_{2.5} in most of the eastern US in CM3, but the observations suggest positive slopes 2 only sporadically in the eastern US. Domain mean modeled and observed slopes for PM2.5-feld are both small in magnitude but disagree in sign (0.005 and -0.006 µg m⁻³ %⁻¹). However, 3 CM3 is able to capture an observed negative relationship over the Pacific northwestern US. 4 5 Overall, the model is able to capture the relationship between PM_{2.5} and temperature and wind speed over the United States with skill, suggesting that the model is useful for projections of 6 7 the impact of climate change on future air quality at large regional scales. One caveat to this 8 finding is the slightly differing time periods over which the model was run and the 9 observations were reported.

10 **5** Conclusions

11 Using the GFDL CM3 global chemistry-climate model, we have assessed the impact of climate change on PM_{2.5} concentrations from present-day until the end of the 21st century. We 12 conducted simulations of 21st century climate in which anthropogenic and biomass burning 13 emissions were held fixed at 2005 levels (denoted RCPx.x 2005AER), in order to isolate the 14 15 effects of changing climate on PM_{2.5} concentrations. We use multiple linear regression 16 modeling to establish associations between changes in meteorological variables and PM_{2.5}. Our analysis focuses on multiple spatial domains (global and regional averages) as well as 17 18 multiple temporal domains (annual and seasonal averages).

19 We find that climate impacts on global PM_{2.5} are relatively small, ranging from an annual PM_{2.5} decrease of 0.055 µg m⁻³ to an increase of 0.21 µg m⁻³ from present-day to the end of 20 the 21st century ($\Delta PM_{2.5}$), depending on the choice of RCP future climate scenario. 21 Simulations using three of the four RCPx.x 2005AER scenarios project a "climate penalty" 22 for PM_{2.5}, meaning future climate will raise PM_{2.5} levels, albeit by small amounts (up to 5%). 23 24 RCP8.5 2005AER, which projects a future with the most extreme climate impacts, has the largest positive $\Delta PM_{2.5}$ both globally and in most regions, while the stringent climate 25 26 mitigation scenario of RCP2.6 2005AER results in modest PM2.5 decreases due to increases 27 in precipitation and decreases in windspeed that dominate modest temperature changes. In 28 general, the trends in climate-driven PM_{2.5} closely mirror the inherent projected climate of 29 each of the RCPs, as one would expect. Global $\Delta PM_{2.5}$ is found to be largest in the fall months 30 (SON) and at a minimum in the summer (JJA), largely because of sulfate aerosol, which is found to be the most susceptible to climate change. Sulfate is also found to both increase and 31 32 decrease in concentration depending on the climate scenario, whereas OA and BC aerosol

almost always increase over the course of the 21st century regardless of the scenario due to a 1 2 lower sensitivity to precipitation than sulfate. Regional averages of $\Delta PM_{2.5}$ are much larger, and specific population centers may see annual average increases of up to 2 µg m⁻³ due to 3 4 climate changes only. Although seemingly small, epidemiology studies have found up to a 10% increase in mortality for every 10 µg m⁻³ increase in PM_{2.5}, suggesting that these regional 5 6 changes are not trivial (Hoek et al., 2013). Additionally, current PM2.5 standards are expected 7 to become more stringent in the future, amplifying the relative contribution of small changes 8 like those that are due to climate if the sensitivities to climate are independent of emissions. 9 Still, global climate-driven changes in PM_{2.5} are found to be minor compared to emissionsdriven changes, as changes in PM from 1980-2005 (while emissions were changing) are equal 10 11 to or sometimes larger than changes from 2005-2100 (while emissions were held fixed), 12 despite occurring over a much shorter time period.

PM_{2.5} is found to be positively correlated with and sensitive to changes in temperature. 13 14 Temperature increases are likely the major factor explaining PM_{2.5} increases in RCP8.5 2005AER. As a caveat, the GFDL CM3 model lacks certain aerosol-temperature 15 16 feedbacks that may be negative, such as the effect of increasing temperature on semi-volatile species such as nitrate aerosol and on source of biogenic SOA. PM2.5 is inversely related to 17 18 precipitation, with increasing precipitation rates leading to decreases in PM_{2.5} concentrations. PM2.5 is also anti-correlated with wind speed, although projected wind speed decreases over 19 20 the century will result in less mixing and more stagnation, ultimately yielding higher PM2.5 21 levels. PM2.5 is found to be mostly anti-correlated with relative humidity, although this is 22 found to play a minor role in influencing PM_{2.5} concentrations. We find that PM2.5 over the 23 21st century is most sensitive to temperature, wind speed, and precipitation in that order, 24 although precipitation estimates may be biased low due to weak convective scavenging. We 25 present a model-observation comparison of PM2.5 sensitivity to climate variables over the 26 United States where observations of PM_{2.5} and meteorology are available. We find overall 27 some agreement between the model and observations, especially for the PM-temperature and PM-wind sensitivities, which in addition to being dominant on the global scale, also represent 28 29 the largest sensitivities in the US.

30 Our model results are consistent with recent findings that climate impacts on PM_{2.5} are

31 relatively minor compared to emissions impacts (West et al., 2013). We are also in agreement

32 with the majority of studies that have found the impact of climate on $PM_{2.5}$ more likely to be a

"penalty" (increase in concentrations) rather than a "benefit" (decrease) (Jacob and Winner, 1 2 2009; Fiore et al. 2015). Since we find PM_{2.5} to mostly increase in the future due to climate, air pollution control and climate policies will need to be strong in order to avoid exacerbation 3 of air-pollution related mortality. This is especially true for locations where we find the 4 "climate penalty" to be more severe (1-2 μ g m⁻³). To our knowledge this is the first study to 5 6 utilize a fully coupled chemistry-climate model, all four RCPs, and full century transient 7 simulations to determine the PM_{2.5} concentrations driven by meteorological changes in a 8 future climate. This work has focused on relatively broad temporal and spatial domains, with 9 the finest resolution being at the regional and seasonal domains. We emphasize that local 10 meteorology may impact local PM2.5 much more strongly, as has been shown for winter 2013 11 in Beijing, China (Wang et al., 2014). Additionally, although these findings are consistent with other models, these results are of a single GCM. Similar experiments with fixed aerosol 12 13 emissions in order to isolate the "climate penalty" could be conducted perhaps as part of a 14 Model Intercomparison Project (MIP) in order to fully determine the robustness of these 15 results. Future work could focus on incorporating temperature-dependent aerosol volatility, feedbacks on biogenic VOC and other natural emissions, and better treatment of ammonium 16 nitrate thermodynamics and chemistry. However, this is unlikely to change the overall results 17 18 of this work, such as the finding that PM_{2.5} is much more strongly influenced by emissions 19 than it is climate.

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21 Acknowledgements

D. Westervelt was supported by a fellowship from the Science, Technology and Environmental Policy (STEP) program at the Woodrow Wilson School of Public and International Affairs at Princeton University. This article was made possible by EPA-STAR Grant 83520601. Its contents are solely the responsibility of the grantee and do not necessarily represent the official view of the EPA.

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15 Table 1: Global average change in population-weighted fine partialate matter concentration 16 $(\Delta PM_{2.5}, \text{ ng m}^{-3})$ and temperature (K) between end of century (2091-2100 mean) and

17	beginning of the centur	y (2006-2015) for e	ach RCPx.x_2	2005AER simulation
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	Sulfate	OA	BC	Dust	Total	ΔT
RCP2.6_2005AER	-57.8	12.6	1.40	-11.6	-55.4	0.8
RCP4.5_2005AER	-19.8	15.6	0.61	6.41	2.88	1.8
RCP6.0_2005AER	17.7	43.1	3.20	19.7	83.9	2.5
RCP8.5_2005AER	104	82.0	4.47	10.7	210	3.9

Table 2: Average change in fine particulate matter ($\Delta PM_{2.5}$, ng m⁻³) between the end of the

century (2091-2100 mean) and present day (2006-2015 mean) for each region (see Fig. 1 for

	Sulfate	OA	BC	Dust	Total
North	148	191	4.61	4.34	387
America					
South	128	13.8	4.97	116	262
America					
East Asia	245	276	37.8	-159	401
South Asia	132	183	20.5	-37.2	299
Europe	67.7	186	7.34	-68.1	193
Russia	234	282	6.86	-81.0	441
Australia	132	41.5	2.23	-43.5	132
Africa	281	258	21.3	80.1	640
Middle East	159	142	12.1	-295	17.7

region definitions) for RCP8.5_2005AER

	Sulfate	OA	BC	Dust	Total
North America	-177	28.3	0.595	1.05	-147
South America	-18.6	-9.79	0.006	10.9	-17.5
East Asia	-548	57.4	7.42	-27.6	-510
South Asia	-234	68.5	9.61	21.7	-134
Europe	-328	89.5	9.74	8.0	-221
Russia	-82.2	59.7	1.10	-3.3	-24.7
Australia	-44.7	18.7	0.657	107	81.1
Africa	-26.5	164	12.5	31.9	182
Middle East	-163	18.4	0.744	16.9	-126