

Contribution to the Theme Section 'Jellyfish bloom research: advances and challenges'



# Jellyfish and forage fish spatial overlap on the eastern Bering Sea shelf during periods of high and low jellyfish biomass

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**ABSTRACT:** Forage fishes and scyphozoan jellyfish are both voracious planktivores within the productive eastern Bering Sea (EBS) ecosystem. To determine the potential competition between the dominant jellyfish *Chrysaora melanaster* and forage fishes, we compared the spatial distributions of *C. melanaster* and 4 forage fish species in the EBS as observed in annual surveys of the upper 30 m. We calculated spatial metrics (centers of gravity, inertia and global index of collocation) of *C. melanaster* and each fish species and examined the degree of jellyfish–forage fish spatial overlap using several geostatistical methods during 2004 to 2012, a period that included high and low jellyfish biomass. Overall, EBS jellyfish occupied large areas where they overlapped with dominant forage fishes; however, the degree of overlap varied inter-annually with fluctuations in jellyfish and forage fish biomass and with climate conditions on the shelf. The spatial overlap between jellyfish and age-0 walleye pollock *Gadus chalcogrammus* was consistent in both low jellyfish biomass (2004 to 2007) and high jellyfish biomass (2009 to 2012) periods, whereas degree of jellyfish overlap with Pacific herring *Clupea pallasi*, capelin *Mallotus villosus* and age-0 Pacific cod *Gadus macrocephalus* varied with climate regimes. Competition between these 2 mid-trophic level groups is important because, while forage fishes are a critical link between plankton and higher trophic levels, jellyfish support few predator groups. Also, jellyfish are potential predators of the early life stages of fish. In locations where overlap is high, jellyfish predation on plankton, fish eggs and larvae may be important in driving dynamics of commercially important fish species.

**KEY WORDS:** Scyphomedusa · Planktivores · Spatial overlap · Inter-annual variability · Pelagic fishes

## INTRODUCTION

Blooms of gelatinous plankton predators (i.e. jellyfish) are frequent occurrences in many coastal ecosystems, with some evidence of increasing populations (Brotz et al. 2012, Condon et al. 2013). Jellyfish are important consumers of zooplankton and can sig-

nificantly restructure food webs when their abundance is high (Deason & Smayda 1982, Suchman et al. 2008, Condon et al. 2011). Field studies in other ecosystems indicate that jellyfish can negatively impact fisheries because they compete with zooplanktivorous fish and feed on early life stages of fish (Purcell & Grover 1990, Purcell & Sturdevant 2001,

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Brodeur et al. 2008b, 2011, Shoji et al. 2009). Likewise, modeling studies have shown that jellyfish indirectly compete with other groups within the food web by diverting plankton production away from upper trophic levels (Ruzicka et al. 2012, Robinson et al. 2014, 2015, Schnedler-Meyer et al. 2016).

The eastern Bering Sea (EBS) is a biologically productive and economically valuable ecosystem, supporting rich populations of zooplankton, forage fish, groundfish, crabs, seabirds and marine mammals. The EBS ecosystem also supports large jellyfish populations. Jellyfish biomass, consisting primarily of the northern sea nettle *Chrysaora melanaster*, has fluctuated greatly over the past 3 decades (Brodeur et al. 2008a, Decker et al. 2014). A steep increase was documented over the EBS shelf during the 1990s (Brodeur et al. 2002). Biomass peaked in summer 2000 and then declined precipitously, stabilizing at a moderately low level between 2001 and 2008. The onset of the biomass increase during the 1990s and the biomass decline in 2000 coincided with transitions between climatic regimes. Brodeur et al. (2008a) examined a 27 yr time series of EBS jellyfish biomass relative to physical conditions, current patterns, zooplankton biomass and associated fish biomass. These analyses indicated that jellyfish biomass fluctuations during the 1982 to 2004 period were influenced regionally by interacting variables such as sea ice cover, sea surface temperature, currents, wind mixing and prey availability.

The EBS shelf is also a highly dynamic ecosystem, forced by a variable climate and oceanographic conditions (Bond & Adams 2002, Aydin & Mueter 2007, Stabeno et al. 2016). The EBS ecosystem has recently experienced anomalously warm (2001 to 2005) and cold (2006 to 2012) conditions, as indicated by late summer and fall water temperatures and changes in winter sea ice conditions (Stabeno et al. 2012, Eisner et al. 2014, Sigler et al. 2016). Changes in ice conditions and timing of the spring bloom have been hypothesized to affect EBS zooplankton biomass (Hunt et al. 2011, Sigler et al. 2016). Recent ecosystem studies indicate that with the cooling that occurred between 2006 and 2010, as the system transitioned from warm to cold, populations of large zooplankton (i.e. *Calanus marshallae* and *Thysanoessa* spp.) increased (Coyle et al. 2011, Hunt et al. 2011, Eisner et al. 2014, Sigler et al. 2016). Likewise, trawl surveys between 2009 and 2012 indicate that jellyfish biomass has increased once again to late-1990 levels (Decker et al. 2014). Peaks in zooplankton biomass during the time series preceded increases in jellyfish biomass, suggesting that food availability is a key

factor contributing to fluctuations in Bering Sea jellyfish populations.

Due to their high abundance and spatial overlap (e.g. Brodeur et al. 1999) with some forage fish species in the Bering Sea, there is a potential for jellyfish, and specifically *C. melanaster*, to negatively impact fish through direct predation upon fish eggs and larvae or through competition for resources. In other systems, research has shown jellyfish and the early life stages of commercial fish and forage fish to overlap significantly in space and time (Brodeur et al. 2008b, 2014, Eriksen et al. 2012, Eriksen 2016). In the North Sea, Lynam et al. (2005) found inverse relationships between jellyfish abundance and Atlantic herring *Clupea harengus* recruitment, indicating that jellyfish may have negative effects on commercial fisheries. Inverse relationships between jellyfish and forage fish biomass have also been observed in the EBS, as well as in other coastal ecosystems (Robinson et al. 2014), suggesting that jellyfish are negatively impacting age-0 fish via predation and competition. Thus, in locations where jellyfish–forage fish overlap is high, predation by jellyfish on fish and their planktonic prey may be an important factor affecting the dynamics of commercially important fishes. However, we currently lack an understanding of how EBS jellyfish affect trophic energy flow through the ecosystem and how different forage fish species are impacted.

Here, we examined the extent of spatial overlap and co-occurrence of *C. melanaster* and 4 dominant pelagic fish species: mixed age class (includes age-0 to adult) Pacific herring *Clupea pallasii*, mixed age class capelin *Mallotus villosus*, age-0 walleye pollock *Gadus chalcogrammus* and age-0 Pacific cod *Gadus macrocephalus* during summer surveys in 2004 to 2007 and 2009 to 2012. We examined differences in spatial overlap during periods of low (2004 to 2007) and high (2009 to 2012) jellyfish biomass, as well as inter-annual changes.

## MATERIALS AND METHODS

Jellyfish and forage fish were collected on the National Marine Fisheries Service Alaska Fisheries Science Center (NMFS AFSC) Bering–Arctic Subarctic Integrated Survey (BASIS) between 2004 and 2012. These surface trawl surveys collect forage fish and large medusae from mid-August to early October. The sampling grid covered the shelf off western Alaska, from 160 to 172°W longitude and 55 to 64°N latitude (Fig. 1). All stations were approxi-

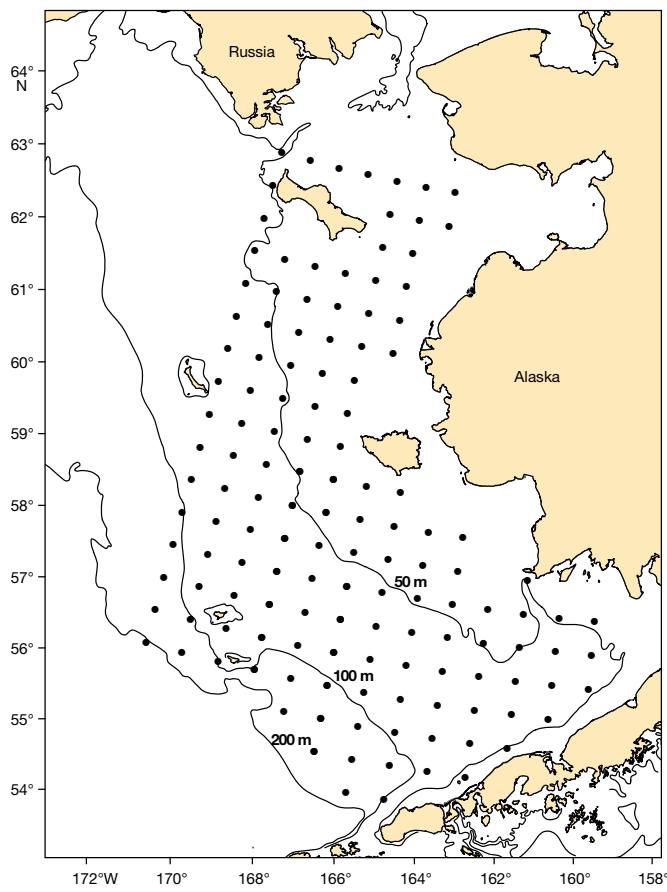


Fig. 1. Eastern Bering Sea Bering–Arctic Subarctic Integrated Survey (BASIS) grid for stations sampled between 2004 and 2012. The annual survey occurs during August and September and focuses on fisheries. Forage fish, jellyfish and oceanography are sampled at each station (black dots). The 50, 100 and 200 m bottom depths are indicated by isobath lines on the map, which also delineate the inner, middle and outer domains, respectively, in the Bering Sea

mately 30 nautical miles (55.6 km) apart. All tows were standardized for gear (50 m wide by 18 m deep) and duration of tow (30 min at 3.5 to 5 knots, covering 2.8 to 4.6 km). Fish and jellyfish were collected using a Cantrawl model 400/601 (Cantrawl Pacific) midwater rope trawl, at or near the surface, with typical spreads of 66.4 m horizontally and 14.6 m vertically. Sea-surface temperatures were collected with a CTD (conductivity, temperature, depth) sensor between 0 and 10 m. All net sampling was performed during daylight hours. We omitted the 2008 survey data from our spatial analyses, as the station coverage in 2008 differed significantly from that in the other years of our study.

Pelagic fish and jellyfish distributions are highly patchy. Pelagic trawl survey data are skewed and highly variable—most individuals may be collected

within a small number of relatively large hauls while a large number of hauls may collect no individuals of a particular species. To accurately estimate annual mean biomass densities of *Chrysaora melanaster* and each forage fish species within the eastern Bering Sea from the BASIS pelagic trawl survey data (Fig. 1), we adopted the  $\Delta$ -distribution method of Pennington (1996). For each species, this method calculates the mean biomass density from a log-normal model of the non-zero hauls, weighted by the proportion of non-zero hauls among all hauls. The mean estimated using the  $\Delta$ -distribution method is more precise than the sample average and for large sample sizes ( $n > 20$ ) is reasonably robust against deviations from the log-normal model (Pennington 1996).

We calculated spatial metrics (centers of gravity and inertia) to describe the distribution of the main forage fish species and *C. melanaster* yearly in the Bering Sea. Center of gravity (CG) represents the mean spatial location of the sampled population and was calculated for each species in each year. Inertia ( $I$ ) is the variance of the locations of positive catches (or the mean squared distance between positive catch locations) (Woillez et al. 2009). Using CG and  $I$ , we calculated the global index of collocation (GIC) between a given forage fish species and *C. melanaster* in a given year. GIC was calculated by taking the distance between their CGs and respective  $I$  values (Bez & Rivoirard 2000, Woillez et al. 2009), ranging between 0 (no individuals of either jellyfish or forage fish species co-occurred at any survey station) and 1 (where the CG of both species were at the same location). For equations describing the calculations for CG,  $I$  and GIC, we refer the reader to Woillez et al. (2007, 2009).

Station-level overlap was measured using the Cramér-von Mises randomization test (Syrjala 1996). This is a non-parametric test of the null hypothesis that *C. melanaster* and the forage fishes have different distribution patterns for all the years of sampling. The test statistic ( $\Psi$ ) is calculated and compared to 9999 random permutations of both populations redistributed among the survey stations as implemented in the 'ecespa' package (De la Cruz 2008) in R (R Core Team 2013). This test measures differences in the way the populations are distributed in the study area but is insensitive to abundance differences between populations.

We determined the degree of spatial overlap of each forage fish species with *C. melanaster* using geostatistical methods following procedures described by Brodeur et al. (2008b, 2014). This was done annually and for all 4 forage fish species for all BASIS sta-

tions. Briefly, outliers in jellyfish and fish data sets were removed by finding the maximum biomass value and replacing it with the next largest value in the data set. Biomass data were then  $\log_{10}(x + 1)$  transformed to improve normality. Transformed data were projected in ArcMap using the North American Datum 1983 Alaska Albers coordinate system, and tested for overall significant spatial autocorrelation using Moran's  $I$  ( $\alpha = 0.05$ ). Geostatistical models were then fitted to each transformed jellyfish and fish species data set using the Geostatistical Analyst package (ArcGIS v.10.2; ESRI). An inverse distance weighting model was used if no overall or directional (i.e. anisotropy) spatial autocorrelation was detected; ordinary kriging models were used if spatial autocorrelation was observed. Three predictive models were generated for each data set if kriging was used. Model fit was evaluated using 4 metrics: mean error, root-mean-square-error, mean standard error, and root-mean-square standardized error, following Johnston et al. (2001). The best fitting model was used to estimate biomass in the annual survey area. The spatial extent ( $\text{km}^2$ ) was then calculated for 3 biomass levels: 0,  $>0$ , and  $\geq 75^{\text{th}}$  quartile value. Overlap between jellyfish and each fish species was estimated as the area where co-occurring biomass for each taxon was greater than zero. Exceptionally high overlap was determined as areas where jellyfish and fish biomasses were both  $\geq 75^{\text{th}}$  quartile values.

## RESULTS

### Fluctuations in jellyfish and forage fish biomass

*Chrysaora melanaster* biomass has varied in the AFSC surface trawl survey since their monitoring began in 2004 (Fig. 2a). Similar to what has been observed on the AFSC bottom trawl surveys (Decker et al. 2014), jellyfish biomass sampled by the surface trawl was low in the mid-2000s, but then increased after 2008.

Both jellyfish and forage fish biomass in the upper 30 m changed dramatically during the 2 periods we examined (2004 to 2007 and 2009 to 2012). Jellyfish biomass was uniformly low during 2004 to 2007 (Fig. 2a), which was also a time when the surface layer was relatively warm in the survey area (Fig. 3). By contrast, jellyfish biomass was variable but higher between 2009 and 2012 (Fig. 2a), a period when cool surface waters were present on the EBS shelf. Similarly, capelin *Mallotus villosus* biomass was higher in the later period (Fig. 2b), while biomass of age-0

walleye pollock *Gadus chalcogrammus* was higher during the earlier period (Fig. 2e). Age-0 Pacific cod *Gadus macrocephalus* and adult Pacific herring *Clupea pallasii* biomass was variable during both periods, but appeared to be higher during 2004 to 2007 (Fig. 2c,d).

### Centers of jellyfish and forage fish distributions

Distribution of the CG and  $I$  for *C. melanaster* and the 4 forage fishes during the 2004 to 2012 surface trawl surveys showed different spatial and temporal patterns (Fig. 4). Overall, *C. melanaster* was centered in the middle of the survey area, occupying the middle and inner shelf areas. During the cold year of 2012, the center of the *C. melanaster* distribution shifted to the southeastern extent of the survey area. The distribution of Pacific herring was centered inshore in the northern part of the survey, while capelin were centered in the northern middle shelf. Similar to *C. melanaster*, the center capelin distribution shifted to the south when conditions were cold in 2012. By contrast, walleye pollock and Pacific cod biomass was centered in the southern middle shelf.

Capelin and walleye pollock had the highest degree of overlap in most years, while herring and Pacific cod showed more moderate degrees of collocation with jellyfish (Fig. 5). Overall, the degree of overlap between jellyfish and forage fish differed in low jellyfish (2004 to 2007) and high jellyfish (2009 to 2012) periods; i.e. collocation was greater when *C. melanaster* biomass was low (Fig. 5).

Based upon the Cramér-von Mises randomization test (Syrjala 1996) results, the forage fish in most years (22 out of 32 combinations) had significantly overlapping distributions with *C. melanaster*, but there were notable exceptions (Table 1). All years except 2004 had at least one non-significant overlap; but in no year were more than half of the comparisons non-significant. Capelin and walleye pollock showed the least overlap with *C. melanaster* with half of the years showing no significant differences (Table 1).

### Spatial overlap of jellyfish and forage fish distributions

Geospatial mapping revealed different patterns of overlap among *C. melanaster* and the 4 forage fish species (see Figs. 6–10). Spatial overlap between *C. melanaster* and herring was consistent in all years

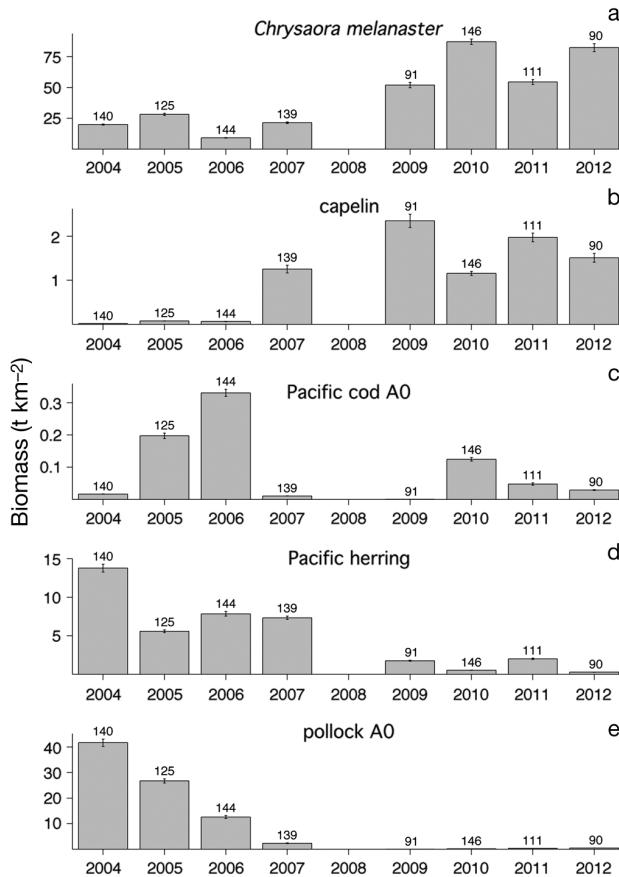


Fig. 2. Time series of mean biomass density of (a) *Chrysaora melanaster*, (b) capelin *Mallotus villosus*, (c) age-0 Pacific cod *Gadus macrocephalus*, (d) Pacific herring *Clupea pallasi* and (e) age-0 walleye pollock *Gadus chalcogrammus* in the eastern Bering Sea as estimated from the Bering–Arctic Subarctic Integrated Survey (BASIS) pelagic trawl data. Values above error bars (1 SE) indicate number of trawls sampled each year

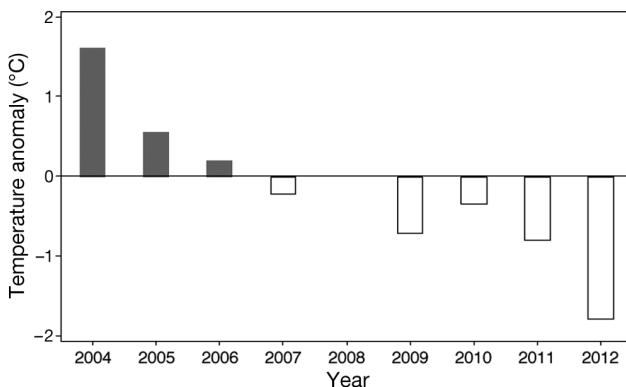


Fig. 3. Annual fall sea-surface temperature anomalies from the Bering–Arctic Subarctic Integrated Survey (BASIS) stations in the eastern Bering Sea from 2004 to 2012. Temperatures were averaged over the top 5 m from CTD casts and the reference value was based on August and September data from the eastern Bering Sea during 2003 to 2016.

Black bars = positive, white bars = negative

and there was no apparent difference in overlap during the 2 periods of 2004 to 2007 and 2009 to 2012 (Figs. 6 & 7). By contrast, spatial overlap of *C. melanaster* and capelin changed over time. In the warm period when jellyfish biomass was low, there was a reduced area where overlap occurred (Fig. 8). The increase in jellyfish biomass that occurred between 2009 and 2012 (Fig. 2a) resulted in greater overlap of capelin and *C. melanaster* throughout the middle shelf (Fig. 8).

For both age-0 walleye pollock and age-0 Pacific cod, there were larger regions of high overlapping biomass, where both forage fish and jellyfish were in the upper 75th percentile for biomass (Figs. 9 & 10). Spatial overlap maps also revealed that walleye pollock and *C. melanaster* co-occurred in similar habitats, that is, primarily on the middle shelf. During the early period (2004 to 2007), Pacific cod and *C. melanaster* co-occurrence reached into the northern portion of the survey, whereas during the later period (2009 to 2012), the overlap between Pacific cod and *C. melanaster* was restricted to the southern half of the survey (Fig. 10).

## DISCUSSION

The goal of this study was to estimate the scale of spatial overlap and potential for competition between forage fish and jellyfish in the EBS by determining their spatial overlap across several years. Our study on *Chrysaora melanaster* builds on work done in the northern California Current (NCC) on the congener *C. fuscescens* (Brodeur et al. 2014). Similar to Brodeur et al. (2014), we used fisheries-independent data collected over several years and varying oceanographic conditions to examine inter-annual changes in spatial overlap between a dominant jellyfish and forage fishes. Brodeur et al. (2014) found that higher interspecific spatial overlap in the NCC occurred during cool and highly productive years compared to warm years when distributions were more dissimilar. In the NCC study and in the present EBS study, the abundance of *Chrysaora* spp. was higher during the cooler period (2009 to 2012), allowing for a wider spatial distribution which likely increased spatial overlap with all pelagic species.

Scyphomedusa biomass on the Bering Sea shelf has fluctuated widely over the last few decades as indicated by surveys conducted for fisheries assessments (Brodeur et al. 2008a, Decker et al. 2014). Climate variability is known to be an important driver of change in the Bering Sea ecosystem. Variable envi-

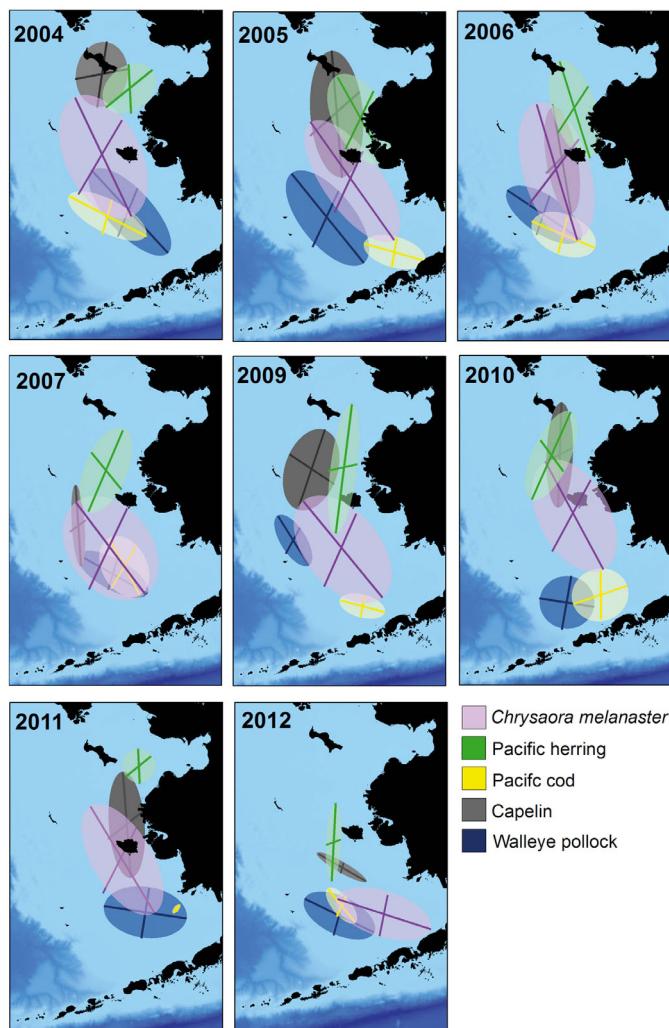


Fig. 4. Distribution of the center of gravity and axes of inertia for *Chrysaora melanaster* and the 4 forage fishes collected during the Bering–Arctic Subarctic Integrated Survey (BASIS) survey. Color of lines and ellipses represent each species as indicated in the key

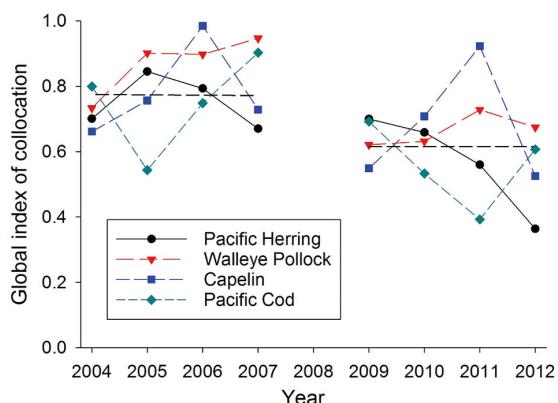


Fig. 5. Inter-annual variability of the global index of collocation (GIC) for *Chrysaora melanaster* and the 4 forage fishes. Dashed horizontal black lines indicate mean GIC in low jellyfish bio-mass (left) and high jellyfish biomass (right) periods

ronmental conditions have been shown to affect multiple trophic levels ranging from primary producers to zooplankton including jellyfish, and up to the top trophic levels (Hunt et al. 2002, 2011, Coyle et al. 2011, Sigler et al. 2016). A simple ecosystem model indicates that high productivity and trophic control, along with fishing pressure and reduced water clarity, results in increased jellyfish biomass and a decrease in forage fish (Schnedler-Meyer et al. 2016).

When pelagic fish occur in areas where jellyfish biomass is high, they can be impacted in several ways. The most direct is through jellyfish predation on the early life stages (Möller 1984, Purcell & Grover 1990, Purcell & Arai 2001, Tilves et al. 2016). Although there are limited data on the diet of *C. melanaster* in the Bering Sea, the data that do exist show that they do consume early stages of walleye pollock, and thus have the potential to impact pollock recruitment due to their high densities in this region (Brodeur et al. 2002). Evidence for this is not limited to the Bering Sea; in the North Sea (Lynam et al. 2005) and the Black Sea (Oguz et al. 2008), jellyfish predation upon fish larvae and competition between fish and jellyfish for zooplankton food may have contributed to diminished fish recruitment. Jellyfish also have the potential to compete for limited prey resources with older age classes of fish when diets of jellyfish and fish overlap (Purcell & Sturdevant 2001, Brodeur et al. 2008b, Shoji et al. 2009). Although such competition has not been demonstrated for the Bering Sea, avoidance of competition among fish species has been observed, in that early life stages of walleye pollock and

Table 1. Cramér-von Mises test p-values for the difference between the spatial distributions of *Chrysaora melanaster* and the 4 forage fishes. Most pairwise comparisons were significant, indicating overlap in the distributions of the jellyfish and forage fish species. Non-significant differences at  $p > 0.01$  and  $p > 0.05$  are highlighted in **bold** with a single and double asterisk, respectively

Year	Capelin	Pollock	Cod	Herring
2004	0.002	<0.001	<0.001	<0.001
2005	<b>0.026*</b>	0.004	0.001	<b>0.031*</b>
2006	<b>0.613**</b>	<b>0.014*</b>	<0.001	0.003
2007	0.001	<b>0.117**</b>	<0.001	<0.001
2009	<0.001	<b>0.064**</b>	0.003	0.010
2010	<b>0.015*</b>	<0.001	<0.001	0.002
2011	<b>0.117**</b>	<b>0.012*</b>	<0.001	<0.001
2012	<0.001	0.006	<b>0.101**</b>	0.002

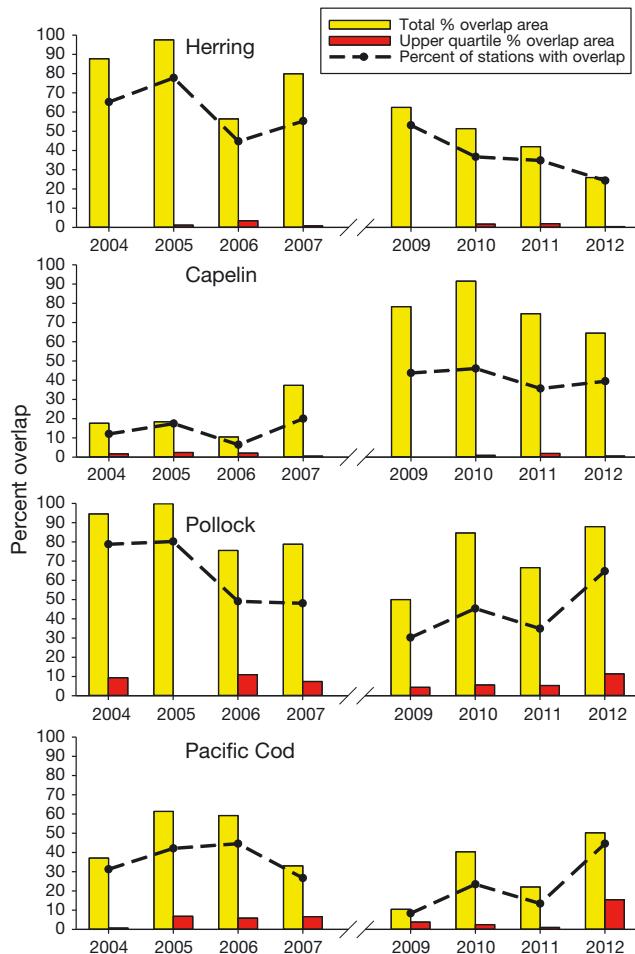


Fig. 6. Inter-annual variability of spatial overlap of *Chrysaora melanaster* and 4 forage fish species, based on GIS-interpolated maps. Shown are the total overlap areas as a percent of the total area sampled (yellow bars), the area of overlap of the pixels containing >75% of the biomass of both *Chrysaora melanaster* and forage fishes (red bars), and the total percentage of stations where overlap occurred (points and dashed lines) on each survey

Pacific cod partitioned prey resources during cold, high productivity conditions in the southeastern Bering Sea, rather than competing for them as they might during low productivity conditions (Strasburger et al. 2014). Finally, jellyfish blooms have been shown to affect commercial fisheries directly through net fouling, spoilage of the catch and limiting fishing opportunities in areas of high jellyfish abundance (Purcell et al. 2007, Kim et al. 2012, Quiñones et al. 2013, Graham et al. 2014).

Changes in overlap of jellyfish and forage fish may be driven by changes in the distributions of both groups in response to changing environmental conditions. During the cold period (2006 to 2011), capelin were distributed into the southeastern Bering Sea

and were present in a large portion of the Inner and Middle Domains, whereas during the warm period (2002 to 2005), capelin were restricted to the northeastern Bering Sea (Andrews et al. 2016). In contrast, mixed-age Pacific herring *Clupea pallasii* catches showed less variation in the EBS during warm and cold periods (Andrews et al. 2016). Similarly, the distribution of EBS age-0 Pacific cod *Gadus macrocephalus* tends to be relatively stable across thermal regimes (Hurst et al. 2012, Parker-Stetter et al. 2016). Age-0 walleye pollock *Gadus chalcogrammus* generally occur in stratified waters of the middle domain (50 to 100 m depth), and they exhibit a broader geographical distribution in the surface waters during warm compared to cool periods (Hollowed et al. 2012).

During periods of high spatial overlap, forage fish may actively avoid regions of high jellyfish biomass. We saw evidence of this between 2009 and 2012, a time of increased jellyfish abundance in the survey area (Fig. 2a). That is, the mean GIC of all forage fish species was lower in 2009 to 2012 compared to the earlier period (Fig. 5). This may be an indication that the fish moved to other areas after the jellyfish became numerous. In the North Sea, an inverse correlation between jellyfish abundance and subsequent herring production may be indicative of active avoidance of herring from regions of high jellyfish abundance and movement into less favorable foraging environments (Lynam et al. 2005).

Shifts in forage fish distribution influence the degree of overlap between jellyfish and fish. We observed greater overlap between *C. melanaster* and capelin in cold years than in warm years (Figs. 6 & 8). In warm years, capelin shifted to the north, away from the center of the survey area where jellyfish biomass was most concentrated. However, because Pacific herring distributions varied little in the EBS during warm and cold periods (Andrews et al. 2016), the decrease in the total overlap area of jellyfish and forage fish during 2009 to 2012 (Figs. 6 & 7) may have been driven by herring avoidance of high jellyfish biomass regions, or indicative of low herring abundance. Likewise, while distributions of age-0 walleye pollock and age-0 Pacific cod remained similar during the 2 periods we examined (Parker-Stetter et al. 2016), the percent overlap of these 2 species with jellyfish declined during the latter period when jellyfish biomass was high (Figs. 6, 9 & 10).

The horizontal distributions of EBS jellyfish are also influenced by environmental conditions (Decker et al. 2013), especially between warm and cold years. However, fish might be more likely to respond to

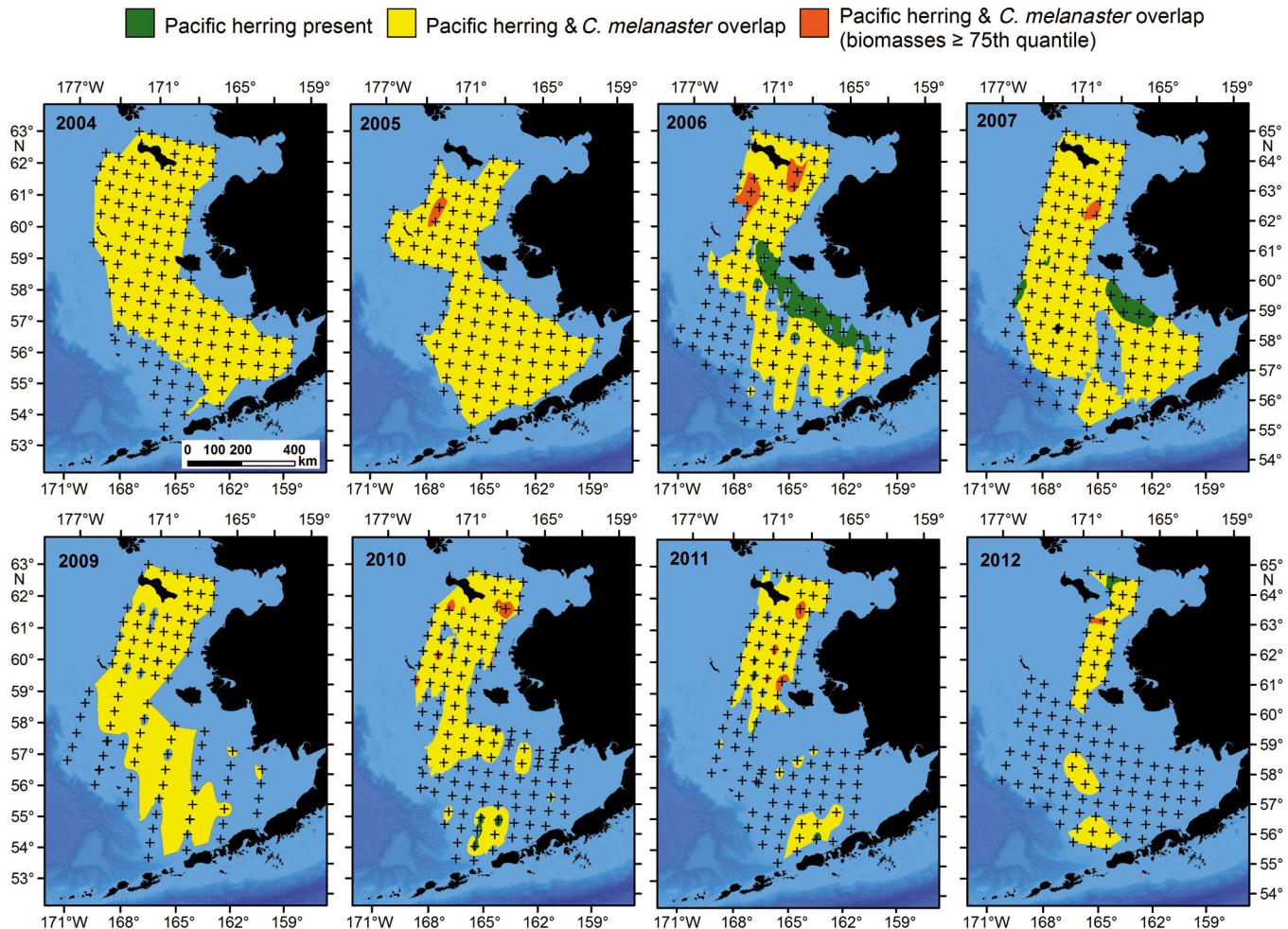


Fig. 7. Areas of no, moderate and high overlap between *Chrysaora melanaster* and Pacific herring *Clupea pallasi*. Shown are sampling stations (crosses), occurrence of herring (green area), overlap between herring and jellyfish (yellow area) and region of maximum overlap (>75th percentile of biomass of both herring and jellyfish; red area)

changing environmental conditions by moving horizontally than jellyfish, which may be constrained by the locations where their polyps occur (Chen et al. 2014) and by their more limited swimming abilities. Ocean currents in the EBS vary between warm and cold years (Stabeno et al. 2012), which may influence the degree of displacement of medusae from their benthic polyp sources, i.e. rocky coastlines in the EBS (Chen et al. 2014). Changes in winter sea ice and general ocean warming may result in poleward shifts of forage populations and their prey resources (Mueter & Litzow 2008), but it is presently uncertain how changing conditions will affect jellyfish production or distribution in the Bering Sea (Brodeur et al. 2008a). Further studies are needed to examine how projected future changes in the physical environment could drive changes in spatial overlap between jellyfish and forage fish.

It is also possible that the change in mean overlap, as indicated by GIC (Fig. 5), was affected by variation in the vertical distribution of fish (or jellyfish), which would, in turn, affect their catchability by pelagic surface trawls. The vertical distributions of forage fish and *C. melanaster* may be affected by environmental conditions on both a seasonal and inter-annual basis. The collections in this study were made in the upper 20 m of the water column during the day in the boreal late summer and early fall, which would capture a substantial proportion of the *C. melanaster* population but miss some of the diel vertical migrants (Brodeur 1998, Decker et al. 2014). For the forage fishes, Parker-Stetter et al. (2013) examined the vertical distribution of capelin, age-0 pollock and Pacific cod from acoustic data and trawl collections from the BASIS surveys used in the present study and found that a sizeable proportion of the

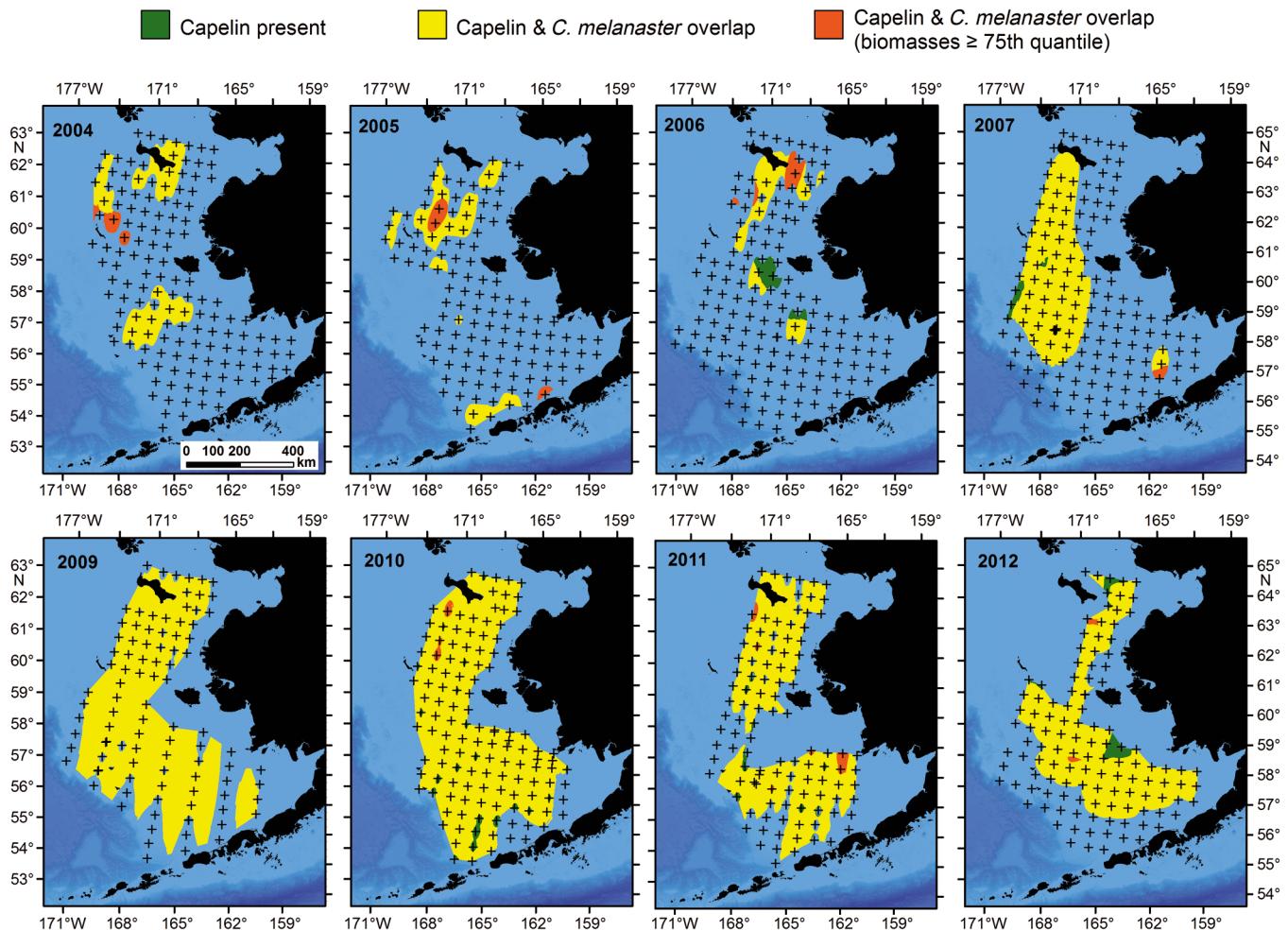


Fig. 8. Areas of no, medium and high overlap between *Chrysaora melanaster* and capelin *Mallotus villosus*. Shown are sampling stations (crosses), occurrence of capelin (green area), overlap between capelin and jellyfish (yellow area) and region of maximum overlap ( $>75$ th percentile of biomass of both capelin and jellyfish; red area)

population of all 3 species inhabited depths below those that were trawled, especially capelin. They also found substantial inter-annual variability depending on oceanographic changes in water column structure and mixed layer depth. The vertical distribution of age-0 walleye pollock and presumably Pacific cod may change seasonally due to the autumn breakdown of the thermocline, when pollock undergo an ontogenetic movement to deeper layers (Parker-Stettler et al. 2015). Given this information and also the differential catchability of jellyfish and forage species with the trawl used (e.g. De Robertis et al. 2017), our overlap measurements based on the abundance of the 2 groups should only be considered as approximate.

We observed that fish densities themselves changed between these 2 periods (Fig. 2b–e), and it is possible that during times of high jellyfish abundance, *C.*

*melanaster* had a substantial predatory impact on the early life stages of these forage fish. However, to date we have no evidence that *C. melanaster* have a large predatory impact on fish larvae in the Bering Sea. Several of the diet studies on this species (Brodeur et al. 2002, Zavolokin et al. 2008, Decker et al. 2014) have shown evidence of feeding on younger stages of fish including walleye pollock, but these were conducted during the summer period whereas most fish spawn in late winter or spring in the Bering Sea. Thus, any estimates of the predatory effects of *C. melanaster* on Bering Sea fish larvae must await diet studies during the critical spring season and in larval rearing habitats.

Field and modeling studies in other ecosystems indicate that jellyfish can negatively impact fisheries because they have the potential to directly compete with zooplanktivorous fish, prey on fish eggs and

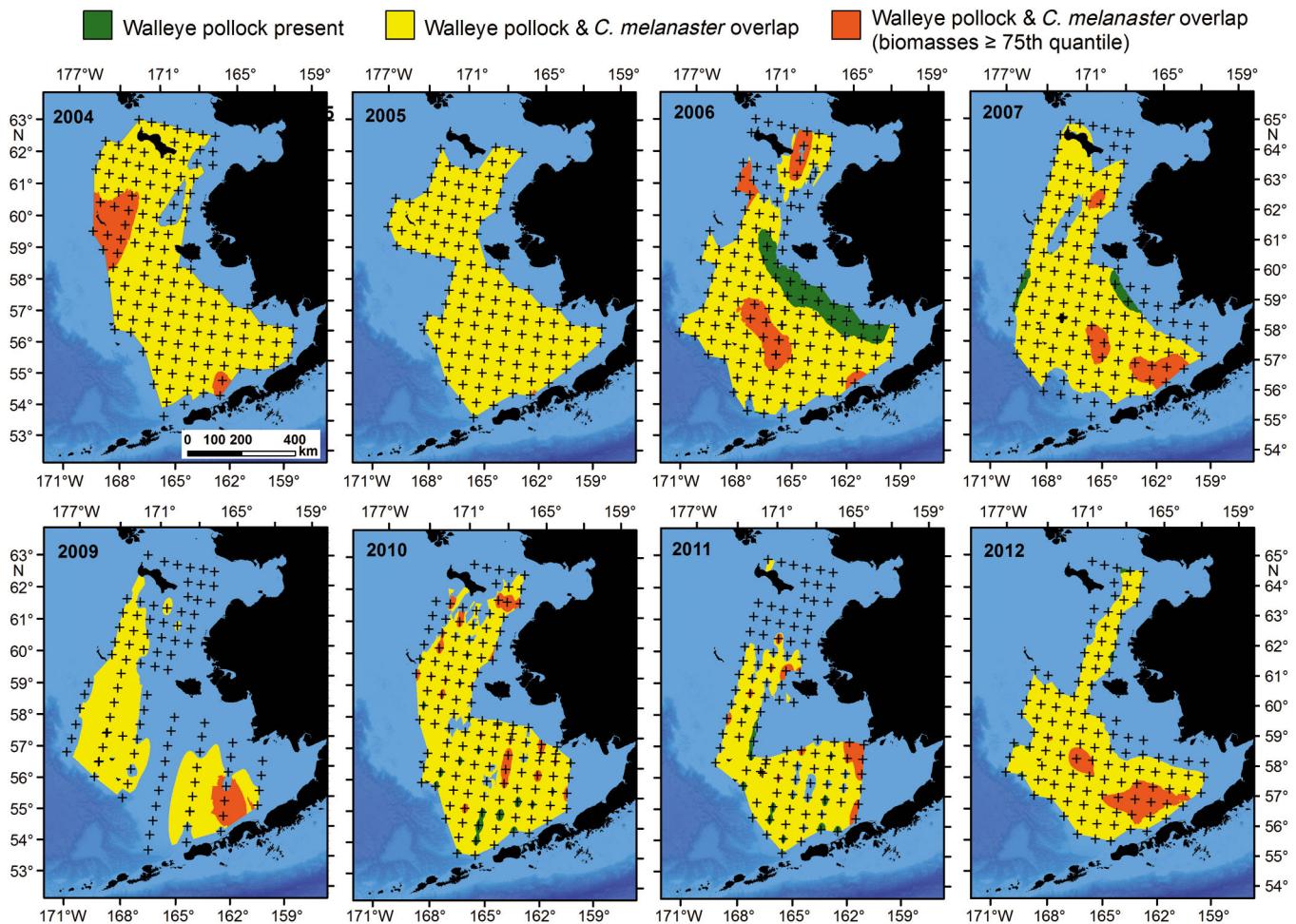


Fig. 9. Areas of no, medium and high overlap between *Chrysaora melanaster* and age-0 walleye pollock *Gadus chalcogrammus*. Shown are sampling stations (crosses), occurrence of pollock (green area), overlap between pollock and jellyfish (yellow area) and region of maximum overlap ( $>75$ th percentile of biomass of both pollock and jellyfish; red area)

larvae, physically harm fish via stinging nematocysts and affect the food web as a whole by diverting plankton production away from upper trophic levels (Brodeur et al. 2008b, 2011, Ruzicka et al. 2012, Robinson et al. 2015). In the North Sea, there are inverse correlations between jellyfish biomass and Atlantic herring *Clupea harengus* recruitment. This may be attributed to predation upon larval herring or to competition with larval herring for zooplankton (Lynam et al. 2005). In the NCC, high local abundances of the jellyfish *Chrysaora fuscescens* are associated with poor feeding conditions for juvenile salmon *Oncorhynchus* spp., as evidenced by less food in salmon stomachs in the presence of jellyfish blooms (Ruzicka et al. 2016). Even smaller juvenile salmon do not themselves prey heavily upon zooplankton within the NCC. They prey upon small juvenile fishes, which in turn prey upon zooplankton (Daly et al. 2009). These examples indicate the im-

portance of monitoring jellyfish populations in systems that support commercially important fish species (Brodeur et al. 2016).

Compared to other regions of the globe, the Bering Sea has been substantially less affected by humans, but we need to consider factors such as fishing, pollution and coastal development as drivers of forage fish and jellyfish population changes and the impacts that these population changes have upon the rest of the ecosystem (e.g. Greene et al. 2015). Because Bering Sea jellyfish populations fluctuate widely annually (Brodeur et al. 2008a, Decker et al. 2014) and because jellyfish blooms may be becoming larger and more frequent in some regions (Brotz et al. 2012, Condon et al. 2012, 2013), an understanding of the trophic roles of gelatinous predators is required for the successful management of not only the EBS ecosystem, but of marine ecosystems in general.

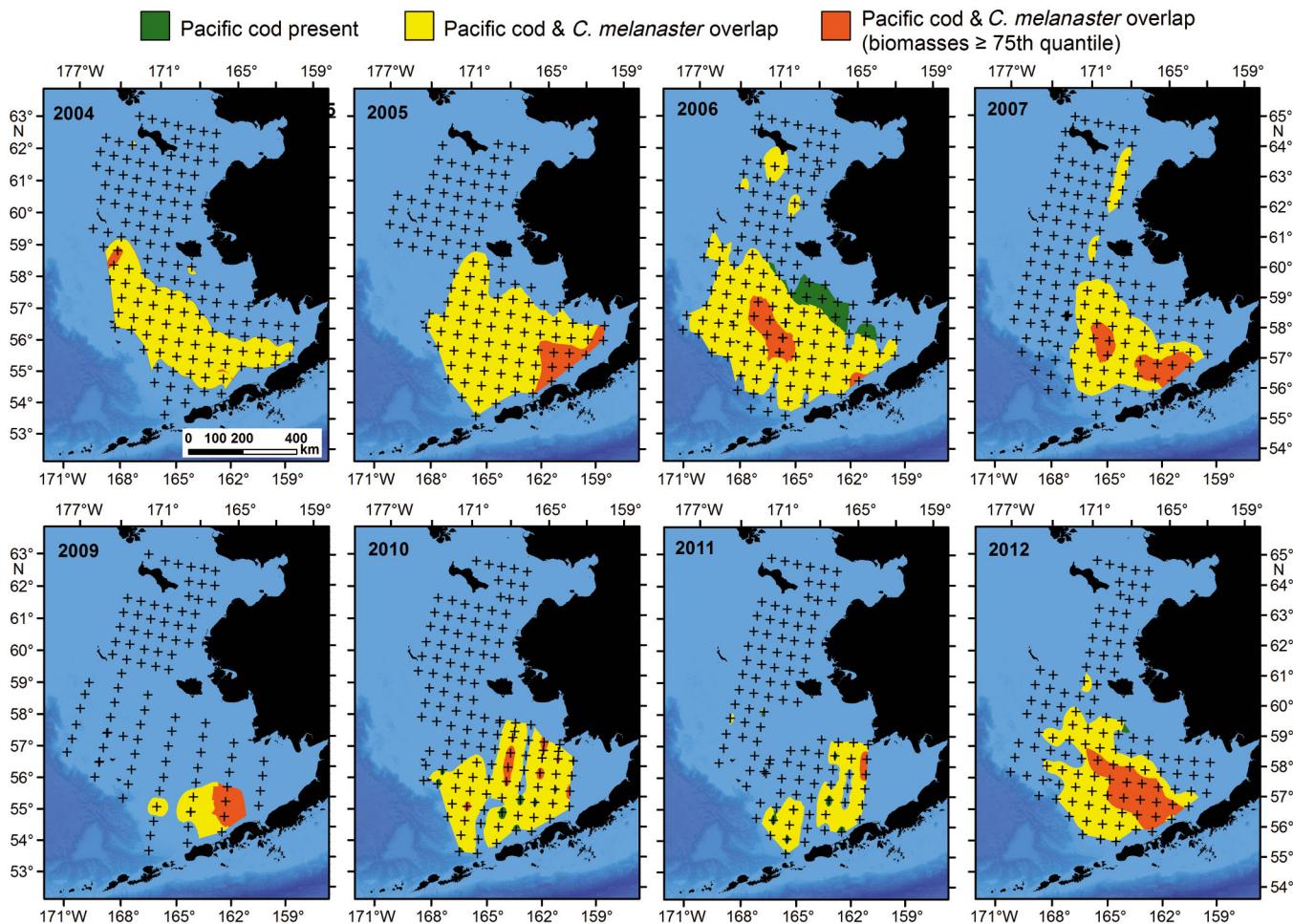


Fig. 10. Areas of no, medium and high overlap between *Chrysaora melanaster* and age-0 Pacific cod *Gadus macrocephalus*. Shown are sampling stations (crosses), occurrence of Pacific cod (green area), overlap between Pacific cod and jellyfish (yellow area) and region of maximum overlap ( $\geq 75$ th percentile of biomass of both Pacific cod and jellyfish; red area)

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