- **Title:** Blubber steroid hormone profiles as indicators of physiological state in free-
- ranging common bottlenose dolphins (Tursiops truncatus)
- **Authors:** Thomas M. Galligan a,b*, Ashley S.P. Boggs c, Brian C. Balmer d, Teri Rowles e,
- 4 Cynthia R. Smith ^d, Forrest Townsend ^f, Randall S. Wells ^g, Nicholas M. Kellar ^h, Eric S.
- 5 Zolman^d, Lori H. Schwacke^d

Affiliations:

Abstract

Blubber has been proposed as a possible alternative to blood in the assessment of endocrine physiology in marine mammals because it can be collected via remote biopsy, which removes some of the confounding variables and logistical constraints associated with blood collection. To date, few studies have directly assessed the relationships between circulating versus blubber steroid hormone profiles in marine mammals, and these studies have been limited to a small subset of steroid hormones, which collectively limit the current utility of blubber steroid hormone measurements. In this study, we used liquid-chromatography tandem-mass spectrometry (LC-MS/MS) to screen for 16 steroid hormones in matched blood and blubber samples from free-ranging common bottlenose dolphins (Tursiops truncatus). Seven steroid hormones were detected and quantified, including two progestogens, two androgens, and three corticosteroids. Using principal components analysis (PCA), we explored relationships between hormones in both matrices and three physiological states: sexual maturity in males, pregnancy, and acute stress response. Plasma and blubber testosterone and its precursors, 17-hydroxyprogesterone and androstenedione, loaded to the first principal component (PC1), and PC1 scores were higher in mature males. Plasma and blubber progesterone loaded to PC2, and pregnant/probable pregnant females had significantly higher PC2 scores. Pregnant females also had higher PC1 scores than other females, suggesting differences in androgen profiles between these groups. There was disagreement between plasma and blubber corticosteroid profiles, as indicated by their loading to different PCs; plasma corticosteroids loaded to PC3 and blubber corticosteroids to PC4. PC3 scores were significantly predicted by elapsed time to blood collection (i.e., time between initiating the capture process and blood collection), while elapsed time to blubber collection significantly predicted PC4 scores, indicating that corticosteroid profiles shift in both tissues during acute stress. Corticosteroid profiles were not related to demographic group, site-month, body mass index, water temperature, or time spent outside of the water on the processing boat. Overall, these results demonstrate that blubber steroid hormone profiles reflect changes in endocrine function that occur over broad temporal scales.

1. Introduction

 Blood is the most common matrix used for endocrine assessments in vertebrates, but collecting blood from free-ranging wildlife typically requires capture and restraint, which is a stressful event and inherently induces shifts in circulating hormone measurements, particularly of stress hormones. As such, it is difficult to measure endocrinological baselines in wildlife, including marine mammals, using blood and current sampling techniques. Furthermore, in free-ranging cetaceans, collection of blood is a labor-intensive and expensive process (Balmer et al., 2014). Using remotely collected sample matrices for marine mammal endocrine assessments could minimize stress to the animals, allow for the measurement of endocrinological baselines, reduce sampling costs for researchers, and increase the number of animals that can be feasibly sampled. Several such matrices have been used, including: feces (Champagne et al., 2018; Wasser et al., 2017), respiratory vapor (Hunt et al., 2014a), baleen (Hunt et al., 2014b), earwax (Trumble et al., 2013), and blubber. Blubber, a form of subcutaneous adipose tissue in marine mammals, can be collected via remote biopsy (Noren and Mocklin, 2012), and contains measurable concentrations of numerous steroid hormones (Boggs et al., 2017; Champagne et al., 2017; Kellar et al., 2015; Kellar et al., 2009; Kellar et al., 2006; Kershaw and Hall, 2016; Kershaw et al., 2017; Mansour et al., 2002; Pallin et al., 2018a; Pallin et al., 2018b; Pérez et al., 2011; Trego et al., 2013; Vu et al., 2015). Thus, blubber could potentially serve as an alternative to blood in marine mammal endocrine assessments involving lipophilic hormones, such as steroids. Blubber steroid hormone measurements vary with physiological states in cetaceans, including stress (Kellar et al., 2015), sexual maturity (Inoue et al., 2018; Kellar et al., 2009), and pregnancy (Kellar et al., 2006; Mansour et al., 2002; Pérez et al., 2011; Trego et al., 2013). In general, blubber steroid hormone profiles qualitatively reflect circulating profiles. In adult male cetaceans, both circulating and blubber testosterone (T) values increase during the breeding season (Harrison and Ridgway, 1971; Kellar et al., 2009; Schroeder and Keller, 1989; Vu et al., 2015). Similarly, 88 progesterone (P_4) is elevated in the blubber and blood of pregnant cetaceans (Kellar et al., 2006; Kirby and Ridgway, 1984; Mansour et al., 2002; Pallin et al., 2018a; Pallin et al., 2018b; Pérez et al., 2011; Sawyer-Steffan et al., 1983; Trego et al., 2013). Stressor-induced activation of the hypothalamo-pituitary-adrenal axis leads to elevated cortisol (F) concentrations in both blood and blubber in cetaceans (Champagne et al., 2018; Houser

et al., 2011; Kellar et al., 2015; Kershaw and Hall, 2016; Schroeder and Keller, 1989; St. Aubin et al., 1996; Thomson and Geraci, 1986). Blubber cortisol levels were also linked to body condition in harbor porpoises (Phocoena phocoena) (Kershaw et al., 2017). However, few studies have explicitly characterized the relationships between circulating and blubber hormone concentrations (Champagne et al., 2017; Champagne et al., 2018; Kellar et al., 2013).

Endocrine glands secrete hormones into blood, which then delivers hormones to peripheral tissues, including hormone target tissues, sites of peripheral hormone metabolism, and blubber. Thus, when changes in central endocrine function occur, circulating hormone values will change rapidly and these changes can be detected nearly instantaneously, while changes in hormone concentrations in peripheral tissue must lag behind changes in circulating concentrations and central endocrine function. 105 Such a relationship was observed in domesticated pigs, in which peak adipose P_4 106 concentrations exhibited a one-to-two day lag behind peak plasma P_4 concentrations and returned to baseline concentrations more gradually than plasma concentrations (Hillbrand and Elsaesser, 1983). Thus, it is likely that blubber integrates circulating hormones over some period of time, and blubber hormone concentrations reflect an average circulating value over that period. Blubber hormone profiles are likely also influenced by in situ metabolism of steroid hormones—as demonstrated by Galligan et al. (2018b)— blubber perfusion rates, and perhaps other factors (e.g., lipid composition, concentration of steroid binding proteins in blood, etc.), which are also dynamic.

We currently have a poor understanding of the temporal relationships between circulating and blubber steroid hormone concentrations. In common bottlenose dolphins under human care, circulating F values were elevated 15 min following exposure to acute stress, remained high during the 120 min of the stress exposure, and then had returned to baseline within one hour post exposure (Champagne et al., 2018). Blubber was sampled at 0 min, 60 min, and 120 min into the stress exposure, and blubber F values had significantly increased at both 60 min and 120 min, though the magnitude of increase was lower compared to blood (Champagne et al., 2018). Notably, blubber was not sampled prior to 60 min, thus it is unclear how soon after the initiation of exposure that a change in blubber F could be detected. Furthermore, Champagne et al. (2018) did not sample blubber after cessation of exposure, meaning we do not know how long blubber F values are elevated after an acute change in circulating F concentration.

Therefore, while an increase in circulating F would indicate stressor exposure within the past several minutes-to-hours, elevated blubber F would indicate stressor exposure at some currently undetermined point(s) in the past, meaning that blubber cannot necessarily be used interchangeably with blood for assessment of acute stress. Conversely, blubber may be interchangeable with blood when measuring pregnancy-131 related shifts in P_4 physiology. In bowhead whales (*Balaena mysticetus*), during 132 pregnancy when P₄ secretion increases and remains high for a prolonged period, 133 circulating and blubber P_4 values are strongly correlated with one another (Kellar et al., 134 2013). This is likely because the persistent increase in P_4 secretion that occurs during 135 pregnancy results in relatively stable circulating P₄ concentrations over a long period 136 allowing sufficient time for blubber P_4 values to equilibrate with blood P_4 . Taken together, this evidence suggests that blubber steroid measurements likely reflect changes in systemic endocrine function over a broader temporal scale, which has implications for how we interpret blubber hormone values in relation to physiological states. Importantly, the blubber steroid hormone literature has primarily focused on F, T, and P4 (Champagne et al., 2017; Kellar et al., 2015; Kellar et al., 2013; Kellar et al., 2009; Kellar et al., 2006; Mansour et al., 2002; Pérez et al., 2011; Trego et al., 2013), and no studies to date have directly studied the relationship between T concentrations in blood and blubber. Furthermore, several additional steroid hormones – 17- hydroxyprogesterone (17OHP4), 11-deoxycorticosterone (DOC), corticosterone (B), 11- deoxycortisol (S), cortisone (E), and androstenedione (AE) – have recently been measured in free-ranging common bottlenose dolphin blubber and blood (Boggs et al., 2019; Boggs et al., 2017; Galligan et al., 2019; Galligan et al., 2018a). Galligan et al. (2018a) and Boggs et al. (2019) explored relationships between these hormones and various physiological states in blood and blubber, respectively. Both noted positive 151 correlations between several of the hormones in the Δ_4 androgen biosynthesis pathway (specifically, 17OHP4, AE, and T) (Figure 1), and increases in AE during pregnancy. Herein, we build from these studies and conduct a comprehensive assessment of the relationships between physiological state and steroid hormone profiles in dolphin blood and blubber. In this study we used individual-matched blood and blubber samples to assess

the relationships between hormones in both matrices in the context of various

physiological states, including sexual maturity, pregnancy, and acute stress response.

We hypothesized that generally hormone profiles would be comparable across the two 160 sample matrices. Furthermore, we predicted that the hormones in the Δ_4 -androgen biosynthesis pathway (Figure 1) would be elevated in adult males because sexual maturity is marked by an increase in T secretion and is detectable in both matrices (Harrison and Ridgway, 1971; Kellar et al., 2009; Schroeder and Keller, 1989). Progestogens should be elevated in pregnant females, as has been observed previously (Kellar et al., 2006; Kirby and Ridgway, 1984; Mansour et al., 2002; Pérez et al., 2011; Sawyer-Steffan et al., 1983; Trego et al., 2013), and in sexually mature females (Inoue et al., 2018). We suspected that androgens would also be elevated in both tissues in pregnant females as reported in bottlenose dolphins and killer whales (Boggs et al., 2019; Galligan et al., 2018a; Robeck et al., 2017; Steinman et al., 2016). Additionally, corticosteroids should be elevated during pregnancy (Valenzuela-Molina et al., 2018). Finally, the hormones in the glucocorticoid pathway should be positively correlated with elapsed time to sample collection because capture and handling stress induces cortisol secretion (Kellar et al., 2015; Kellar et al., 2013; St. Aubin et al., 1996; Thomson and Geraci, 1986) and impacted by body condition (Kershaw et al., 2017). This comprehensive comparison of blood-blubber hormone profiles will improve our ability to use remotely collected blubber biopsies to study endocrine function in free-ranging marine mammals.

2. Materials and Methods

2.1 Animals, Field Data Collection, and Sample Collection

Matched blubber and blood samples were collected from free-ranging common 182 bottlenose dolphins ($n = 77$) from three locations in the southeastern United States during late spring and late summer/early fall (Barataria Bay, LA [June 2013, 2014; n = 34]; Brunswick, GA [September 2015; n = 16]; and Sarasota Bay, FL [May 2013-2016; n $185 = 27$). Methods for the temporary capture, restraint, sampling, and release have been previously described (Schwacke et al., 2014; Smith et al., 2017; Wells et al., 2005). Briefly, a seine net was deployed encircling a dolphin or group of dolphins, and if necessary, the radius of the space enclosed by the net was reduced to force the dolphin to become entangled or enable handlers to safely restrain the dolphin without entanglement. When a dolphin entangled itself in the net, handlers would immediately

restrain the animal and disentangle it. The time at which the capture net was deployed

was recorded, and is considered the start of the capture process. The times at which blood and blubber samples were collected were also recorded, and time elapsed between capture initiation (i.e., net deployment) and sample collection was calculated. Blood and blubber were not collected concurrently; blood was always collected first then blubber was collected second. On average, the interval between these collections was 197 87.9 min (standard deviation = 31.6 min; minimum = 9.00 min, median = 80.0 min, maximum = 185 min). Animals were transferred to a boat to measure body mass, and time of transfer was recorded. The blubber biopsy was generally collected on the boat (five animals were returned to the water prior to biopsy collection). Six individuals did not have body mass measured. Water temperature at time of sampling was also recorded for all but four animals. Sarasota Bay sampling was performed under National Marine Fisheries Service (NMFS) Scientific Research Permit No. 15543 and annually renewed IACUC approvals through Mote Marine Laboratory. Barataria Bay and Brunswick sampling were conducted under NMFS permit no. 932-1905/MA-009526 with protocols reviewed and approved by National Oceanic and Atmospheric Administration IACUC.

 This sample set includes individuals from different demographic groups defined by different physiological states, including subadult and adult males and pregnant, probable pregnant, and non-pregnant females (adult and subadult) (Table 1). Males 210 were sampled from Brunswick ($n = 13$) and Sarasota ($n=13$). Pregnant females were 211 largely sampled from Barataria Bay ($n = 14$), with only one pregnant individual sampled 212 from Brunswick. All probable pregnant females were sampled from Barataria ($n= 4$). 213 Non-pregnant females were sampled from Barataria ($n = 16$), Sarasota ($n = 14$), and 214 Brunswick $(n = 2)$. Age was determined either through lifelong observation (i.e., known birth date) or, when possible, through examination of growth layer patterns in teeth using methods that have been described previously (Hohn et al., 1989; McFee et al., 2010); age was not determined in all individuals. Age classification was dictated by age, if 218 known (individuals \geq 10 years old were classified as adults; n = 47), or total length if age 219 was not known (individuals with total length \geq 240 cm were classified adults; n = 30). Pregnancy status was diagnosed by ultrasound examination of the uterus and ovaries 221 and preliminary assessment of circulating P_4 by immunoassay (Schwacke et al., 2014; Smith et al., 2013; Smith et al., 2017; Wells et al., 2014). Females with a corpus luteum 223 present on either ovary, serum progesterone concentrations greater than 5 ng m L^{-1} , and presence of a fetus and uterine fluid were classified as pregnant, per Smith et al. (2017).

225 Those with a corpus luteum present and serum $P_4 > 5$ ng mL⁻¹ but without a fetus observed were classified as probable pregnant. Body mass index (BMI) was calculated from body length and weight measurements per Hart et al. (2013); no individuals we

228 classified as having low BMI (below the lower $95th$ percentile threshold).

Full-depth blubber samples were collected by surgical or punch biopsy (Schwacke et al., 2014). Blood was collected from the vasculature of the ventral fluke into sodium heparin vacutainers, and plasma was produced by centrifugation at site of 232 capture. After removal of the skin, blubber (average blubber biopsy mass = 0.400 g \pm 0.168) and plasma (in 1 mL to 5 mL aliquots) were immediately frozen in a liquid 234 nitrogen dry shipper at approximately -150 \degree C, and shipped to the National Institute of Standards and Technology (NIST) Environmental Specimen Bank at Hollings Marine 236 Laboratory (Charleston, SC, USA), where they were stored at -80 \degree C until analysis.

2.2 Calibration and Internal Standards

Calibration and isotopically-labeled internal standard manufacturer and purity information are reported in Supplemental Table 1. Calibration (cal) and internal standard (IS) mixture solutions were diluted in methanol, with the concentration of each compound in the final mixture calculated gravimetrically (ng compound q^{-1} mixture). Average mass of each IS compound amended to tubes is reported in Supplemental Table 2.

2.3 Hormone Extraction

Blubber hormone extraction was completed using methods described by Boggs et al. (2017) with a kit (Agilent, Santa Clara, CA, USA) that utilizes a salting-out assisted liquid:liquid extraction (SALLE) to dispersive solid phase extraction (SPE) process (kits: Agilent Bond Elut QuEChERS EN Extraction kit, p/n 5982-5650, and Agilent Bond Elut QuEChERS dispersive-SPE kit for Drug Residues in Meat, 15 mL, p/n 5982-4956) (Boggs et al., 2017; Fu and Zhai, 2010). Plasma hormones were extracted by reverse phase solid phase extraction (SPE) via methods described by Galligan et al. (2018a). For both matrices, process blanks containing only IS were extracted alongside samples and cals.

2.5 Instrumental Methods and Quantitation

Chromatographic separation and quantification of steroids in both blubber and plasma extracts proceeded according to methods described by Galligan et al. (2018a), using an Agilent 1200 Series HPLC system with a binary pump and an autosampler

linked to an AB Sciex (Framingham, MA, USA) API 4000 QTRAP hybrid triple quadrupole/linear ion trap mass spectrometer. This method allows for quantification of the following steroid hormones: pregnenolone, 17-hydroxypregnenolone, P4, 17OHP4, AE, T, dihydrotestosterone, dehydroepiandrosterone, S, F, E, DOC, B, estradiol, estrone, and estriol (Figure 1). Two transitions were monitored per compound, and the transition with the larger signal was used for quantification. We used Sciex Analyst software (version 1.5) to integrate peaks. Steroid concentration was determined by interpolating analyte area ratios (analyte area:IS area) on a standard curve comprised of calibration standards which fully encompassed the range of sample values 267 (Supplemental Tables 3 and 4). Observed reporting limits (RL_{obs}) were defined as the lowest calibration standard used in the calibration curve; calculated reporting limits 269 (RL_{calc}) were defined as three times the standard deviation of the mean of process blank measurements plus the mean of the process blanks (Supplemental Table 5). The larger of the two RL values was used as the censoring threshold in statistical analyses. Censoring determination was based on raw area rather than calculated hormone value, but RLs are reported as hormone values for clarity. See section 2.6 "Statistics" for further details about censoring methods.

2.6 Statistics

Statistical analyses were performed with IBM SPSS Statistics 24 (IBM, North Castle, NY, USA) and R (version 3.6.1) with RStudio (R Core Team, 2018; RStudio 278 Team, 2016). For all hypothesis tests, α = 0.05. A principal components analysis (PCA) was performed to examine relationships between hormones in both matrices. A uniform distribution was assumed between 0 and RL, and hormone values below RL were 281 substituted with a random value within this range. Data were log_{10} transformed, mean centered, and unit scaled. Suitability for PCA was confirmed by ensuring all variables had Kaiser-Meyer-Olkin (KMO) measure of sampling adequacy > 0.5; that the KMO measure of adequacy for the entire data set was > 0.5; and that Bartlett's test of 285 sphericity was significant ($p < 0.05$) (Dziuban and Shirkey, 1974). Factors with eigenvalues > 1.0 were extracted. Varimax rotation was utilized to simplify interpretation. Two samples were lost during extraction (one plasma sample from a non-pregnant female and one blubber sample from a pregnant female), and were thus excluded from analysis (Supplemental Table 6).

We assessed relationships between PC scores and various demographic, morphometric, and sampling variables via stepwise linear modeling in which we 292 removed variables based on p-value until all remaining variables were significant at $α =$ 0.05. For PC1 and PC2, we included demographic group, site-month, and their interaction as potential predictor variables. For PC3, we included elapsed time to blood collection, BMI, demographic group, and site-month as potential predictor variables. Due to a violation of assumptions, PC3 scores were rank transformed prior to analysis. For PC4, we included elapsed time to blubber collection, water temperature, time spent on boat, BMI, site-month, and demographic group as potential predictor variables. Individuals that were brought onto the boat but then returned to the water prior to blubber biopsy collection were excluded from the models including "time spent on boat" as a variable because it would be impossible to interpret the effect of "time spent on boat" in these animals. Pairwise comparisons via t-test, where appropriate, were performed with Benjamini-Hochberg correction. We also assessed relationships between age, length, and weight and PC1 scores (males) and PC2 scores (non-pregnant females) via Pearson correlation or Kendall's tau correlation.

3. Results

Six hormones were detected and quantified in both blubber and plasma: P4, 17OHP4, AE, T, F, and E (Tables 1 and 2). S was detected and quantified in blubber only (Tables 1 and 2). We did not detect pregnenolone, 17-hydroxypregnenolone, DOC, B, dehydroepiandrosterone, dihydrotestosterone, estradiol, estrone, or estriol in any samples. Hormone concentrations for each individual are reported in Supplemental Table 6.

In the PCA, four components with eigenvalues > 1 were extracted explaining 27.2 %, 21.8 %, 12.8 %, and 9.19 % of the variance respectively (71.0 % cumulatively), and simple structure was obtained by Varimax rotation (Table 3; Figure 2A and B). 317 Plasma and blubber T, AE, and 17OHP₄ loaded (i.e., absolute value of loading > 0.4) 318 positively to PC1; plasma and blubber P_4 loaded positively to PC2; plasma F and E loaded positively to PC3; and blubber S, F, and E loaded positively to PC4 (Table 3; Figure 2A and B). 321 PC1 scores were significantly predicted (linear model, $R^2 = 0.799$, $F_{9.65} = 28.8$, p

322 \leq 0.001) by demographic group (F₄ = 55.8, p \lt 0.001), site-month (F₂ = 10.9, p \lt 0.001),

323 and their interaction ($F_3 = 4.66$, $p = 0.005$). Based on the interaction plot (Figure 2C), the site-month factor was important for males, particularly adult males, and not females. Therefore, we considered adult and subadult males collected from different sites-months to be distinct groups when performing pairwise comparisons. We found that adult males, regardless of site-month, exhibited higher scores than all other groups per pairwise t-test with Benjamini-Hochberg correction (Figure 2D). Among males, Sarasota (May) individuals exhibited higher PC1 scores than Brunswick (September) within both age classes (Figure 2D). Among females (inclusive of all sites-months), pregnant individuals had higher PC1 scores than non-pregnant and probable pregnant individuals per pairwise t-test with Benjamini-Hochberg correction (Figure 2D). Relationships between females and subadult males depended on site-month; Sarasota (May) subadult males had PC1 scores comparable to pregnant females and higher than non-pregnant and probable pregnant females, while Brunswick (September) subadult males exhibited PC1 scores comparable to non-pregnant and probable pregnant females and lower than pregnant females per pairwise t-test with Benjamini-Hochberg correction (Figure 2D). For PC2 scores, demographic group was the only significant predictor (linear model, R^2 339 = 0.793, $F_{4,70}$ = 66.9, p < 0.001), with pregnant and probable pregnant females having higher scores than all other groups per pairwise t-test with Benjamini-Hochberg correction (Figure 2E).

Elapsed time to blood collection was the only significant predictor of PC3 scores 343 (rank-transformed linear model, $R^2 = 0.256$, $F_{1,74} = 25.4$, p < 0.001) Similarly, elapsed time to blubber collection was the only significant predictor of PC4 scores (linear model, $R^2 = 0.267$, $F_{1,73} = 26.6$, $p < 0.001$) (Figure 3B).

 346 Since three of four hormones in the Δ_4 androgen pathway (Figure 1)—specifically 17OHP4, AE, and T—loaded to PC1, we assessed PC1 scores by age and age-related morphometric variables (i.e., body length and weight) in males and non-pregnant females to further examine variation in androgen profiles by sexual maturity. Within males, we stratified this analysis by site-month due to the interaction between demographic group and site-month (Figure 2C). In males from both sites-months, PC1 scores were significantly and positively correlated with age (Sarasota [May]: Kendall tau correlation, τ = 0.623, z = 2.94, p = 0.003; Brunswick [September]: Pearson correlation, r 354 = 0.833, $t_5 = 3.37$, $p = 0.020$), body length (Sarasota [May]: Kendall tau correlation, $\tau =$ 355 0.632, $z = 3.00$, $p = 0.003$; Brunswick [September]: Pearson correlation, $r = 0.722$, $t_{11} =$

356 3.46, $p = 0.005$), and body weight (Sarasota [May]: Kendall tau correlation, $\tau = 0.667$, T 357 = 65, p < 0.001; Brunswick [September]: Pearson correlation, $r = 0.873$, $t_6 = 4.39$, p = 0.004) (Figure 5).

There were no significant relationships between PC2 scores and age (Kendall 360 tau, $\tau = 0.0522$, $z = 0.339$, $p = 0.735$), length (Pearson correlation, $r = 0.210$, $t_{28} = 1.14$, p 361 = 0.266), or weight (Pearson correlation, $r = 0.295$, $t_{25} = 1.54$, $p = 0.136$) in non-pregnant females (not shown).

4. Discussion

The goal of this study was to explore the relationships between blood and blubber steroid hormone profiles in common bottlenose dolphins, and thereby, provide information to subsequently improve our ability to use remotely collected blubber biopsies to assess endocrine status in marine mammals. We accomplished this using LC-MS/MS to measure a broad suite of steroid hormones in matched plasma and blubber samples from free-ranging common bottlenose dolphins. We performed a PCA to explore relationships among hormones, examine the relationships between hormone profiles in the two matrices, and study hormone profiles associated with three important physiological states: sexual maturity, pregnancy, and acute stress response.

The PCA allowed us to assess the relationships among hormones both within and across each matrix and thereby, provided information about the agreement between hormone profiles in each tissue. We also used the PCA to assess how hormone profiles (collectively in both tissues) vary by physiological state. In general, the variable loading patterns suggest that androgen and progestogen profiles were similar across matrices, 379 and that most of the hormones in the Δ_4 androgen pathway are associated with one another. Conversely, corticosteroid profiles were poorly correlated across matrices, as evidenced by the fact that plasma and blubber corticosteroids loaded to separate PCs. 382 The range of F values reported here (blubber: 0.0596 ng g^{-1} to 21.0 ng g^{-1} ; 383 plasma: 1.67 ng g^{-1} to 30.2 ng g^{-1} are comparable to previous studies in bottlenose 384 dolphins and other cetaceans (blubber: approximately 1 ng q^{-1} to 70 ng q^{-1} ; plasma: 385 approximately 3.5 ng g⁻¹ to 60 ng g⁻¹, assuming density of plasma is approximately 1.025 386 g mL $^{-1}$) (Champagne et al., 2018; Kellar et al., 2015). Capture and handling induces the secretion of corticosteroids in cetaceans (St. Aubin et al., 1996; Thomson and Geraci, 1986). We found that only elapsed time to blood collection was a significant predictor of

PC3 (plasma corticosteroids) scores. Thus, corticosteroid secretion increased during capture, handling, and sample collection, as would be expected during the acute stress response, while demographic group, BMI, and site-month had no impact on circulating corticosteroid profiles.

Since circulating corticosteroid concentrations were increasing during the capture, handling, and sampling process, we expected elapsed time to blubber collection to predict PC4 scores (blubber corticosteroids) as greater elapsed time would allow more time for corticosteroid secretion and for circulating corticosteroids to become incorporated into blubber. As expected, elapsed time to blubber collection was a significant predictor of PC4 score. It is also important to note that there was temporal mismatch between sample collection times for each tissue. Plasma was collected first (ranging from 6 min to 66 min post capture onset) and blubber was collected second (53 min to 215 min post capture onset). This mismatch likely allowed more time for equilibration between the matrices than if the two samples had been collected simultaneously, but the interval between blood and blubber collection varied between individuals, meaning individuals likely differed in their levels of blood-blubber equilibration at the time of blubber collection, which could have contributed to the lack of association between plasma and blubber corticosteroid profiles. Additionally, we cannot determine the rate at which corticosteroid profiles in blubber changed because we only have single timepoint measurements, as opposed to repeated measures, and our earliest sample occurred at 53 min, which is comparable to when Champagne et al. (2018) began assessing changes in blubber F in common bottlenose dolphins under human care exposed to acute stress. We conclude that corticosteroid concentrations increased in both matrices due to capture stress. Concurrently sampling blood and blubber repeatedly over a prolonged period of time would better clarify these temporal relationships. Developing novel, remote sampling devices that collect both blubber and blood simultaneously would greatly enhance our understanding of differences in corticosteroid profiles between these two matrices (e.g., an animal borne blood sampling device in development for pinnipeds (Takei et al., 2016)). Corticosteroids are important regulators of energy metabolism. Therefore, we hypothesized that body condition (BMI) may be related to corticosteroid profiles in blood and blubber, as seen in harbor porpoises (Kershaw et al., 2017). However, we found no

 such relationship, potentially because no individuals in this study had low BMI (below the 422 lower 95th percentile threshold) (Hart et al., 2013). Future studies with broader ranges of BMI should further assess this relationship.

It is likely that the rate of perfusion also influences corticosteroid delivery to blubber (i.e., more blood flow to blubber would increase hormone delivery). Thus, factors influencing perfusion immediately prior to and during capture, handling, and sampling would also influence blubber corticosteroid profiles and, thus, PC4 scores. Water temperature and time spent on the boat prior to blubber collection could affect blubber perfusion since cetaceans modulate blood flow to the blubber and skin to adjust heat flux with the environment for thermoregulation. We did not observe an effect of water 431 temperature or time spent on the boat on PC4 scores, potentially due to the fact that the dolphins were constantly sponged with water to cool their skin and keep it wet during out-of-water processing. This could potentially indicate that there was minimal influence of sampling procedure/temperature on perfusion and/or that changes in perfusion during sampling do not appreciably impact blubber hormone profiles. To explicitly elucidate these relationships, future studies will need to directly measure changes in perfusion in relation to circulating and blubber hormone profiles.

Some of the mismatch between blood and blubber corticosteroid profiles may arise from the fact that common bottlenose dolphin blubber has the capacity to metabolize corticosteroids (Galligan et al., 2018b); i.e., after corticosteroids are delivered to blubber via blood, they are metabolized by blubber, which could cause blubber corticosteroid profiles to shift away from circulating profiles. While the rates of metabolism are poorly defined, we would expect the influence of in situ metabolism to increase over time as this would allow greater quantities of hormone to be metabolized, leading to wider divergence between blood and blubber profiles in animals with greater interval between blood and blubber collection. Future research should seek to better characterize the metabolism of corticosteroids—and potentially other steroids—in blubber and examine how such metabolism may impact the relationships between blood and blubber steroid hormone profiles.

T secretion is elevated during breeding season in sexually mature male cetaceans; this leads to a seasonal increase in circulating and blubber T (Harrison and Ridgway, 1971; Kellar et al., 2009; Schroeder and Keller, 1989). Therefore, since sampling occurred between late spring and late summer, when breeding is likely occurring in these populations (McFee et al., 2014; Urian et al., 1996), we anticipated

that adult males would exhibit elevated T concentrations in both plasma and blubber. Furthermore, based on Galligan et al. (2018a) and Boggs et al. (2019), we expected that 457 the upstream hormones in the Δ_4 androgen pathway – P_4 , 17OHP₄, and AE – would also be elevated in both matrices because increased production of these hormones is required to support increased T production. As a result, these hormones should be positively correlated in both matrices and exhibit maturity-dependent differences. Our 461 results largely support these hypotheses. Blubber and plasma T, AE, and 17OHP₄ loaded positively to PC1, indicating a positive association between these variables. Furthermore, PC1 scores were significantly higher in adult males, regardless of site-month, compared to all other groups, and PC1 scores were positively correlated with age, length, and weight in males from both sites-months, which indicates that the 466 combined variance in T, AE, and $17OHP₄$ was related to maturity in males. The effect of site-month on PC1 scores within males may indicate that androgen profiles vary among populations or between months within the breeding season. One might also expect to 469 observe elevated P_4 concentrations to support 17OHP₄ production, but P_4 was not detected in any male blubber and was rarely detected in male plasma. This is likely due 471 in part to the abnormally high RL for blubber P_4 in this study, which may have been due 472 to use of a higher IS concentration for P_4 than Boggs et al. (2017) (18.9 ng vs. 5.26 ng), 473 who achieved a lower detection limit for blubber P_4 (0.246 ng) which was comparable to immunoassay techniques (Inoue et al., 2018; Kellar et al., 2006; Mansour et al., 2002; Pallin et al., 2018a; Pallin et al., 2018b; Pérez et al., 2011; Trego et al., 2013). 476 Nonetheless, our high RL for blubber P_4 impedes our ability to assess relationships 477 between blubber P_4 and other hormones, especially in males and non-pregnant females, and thus is a key limitation to this study. P4 secretion increases during pregnancy in cetaceans. This increase can be observed in both plasma and blubber (Kellar et al., 2006; Kirby and Ridgway, 1984; Mansour et al., 2002; Pérez et al., 2011; Sawyer-Steffan et al., 1983; Trego et al., 2013).

482 As expected, plasma and blubber P_4 loaded to the same principal component (PC2) in the PCA, and pregnant/probable pregnant females had significantly higher PC2 scores

- compared to non-pregnant females and other demographic groups. We expected that
- pregnant females would have higher corticosteroid levels, based on previous work in
- humpback whales (Megaptera novaeangliae) (Valenzuela-Molina et al., 2018), but found
- no such difference. Androgens, including T and AE, are also elevated during pregnancy

488 in bottlenose dolphins and killer whales (*Orcinus orca*) (Boggs et al., 2019; Galligan et al., 2018a; Robeck et al., 2017; Steinman et al., 2016) , potentially to support ovarian secretion of P4 (Carrizo et al., 1994; Telleri´a et al., 1995; Waddell et al., 1992). Therefore, it is unsurprising that pregnant females also had higher PC1 scores compared to non-pregnant females.

Importantly, probable pregnant female PC1 scores were significantly lower than pregnant, but not non-pregnant, females. Probable pregnant females are either newly pregnant or in the luteal phase of their estrous cycle (Smith et al., 2017). Thus, this finding suggests that blood and blubber androgen measurements could potentially be used to differentiate pregnant females from early pregnant/luteal phase females without the need for ultrasound tests. This could prove useful in assessing reproductive dynamics and dysfunction in free-ranging cetacean populations. However, with low 500 sample size of probable pregnant females $(n = 4)$, our conclusion here is limited and should be the subject of further investigation. Additionally, it should be noted that all the pregnant females in this study were in the first trimester of pregnancy when sampled. Endocrine profiles likely change throughout pregnancy (Robeck et al., 2017; Robeck et al., 2016; Steinman et al., 2016); therefore, future studies should examine the relationships between circulating and blubber hormone profiles throughout pregnancy.

 P4 has been used to classify sexual maturity in female cetaceans. According to a review of cetacean endocrinology, females in many cetacean species exhibit an 508 increase in circulating P_4 at the onset of sexual maturity, with sexually immature 509 individuals exhibiting concentrations $<$ 1 ng/mL of P_4 (reviewed: Atkinson and Yoshioka, 510 2007). However, this threshold is misleading because P_4 concentrations will vary 511 significantly by pregnancy status and during the estrous cycle. In reality, circulating P_4 concentrations only rise above this threshold when a female is pregnant or in the luteal phase of the estrous cycle (Kirby and Ridgway, 1984; Sawyer-Steffan et al., 1983). As 514 such, a circulating P_4 concentration < 1 ng/mL cannot be considered a marker of sexual immaturity, but simply an indication that the female has not recently ovulated and/or is 516 not pregnant. Nonetheless, blubber P_4 concentration was a useful marker of maturity in 517 minke whales (Balaenoptera acutorostrata) (Inoue et al., 2018). We found no evidence to suggest that PC2 scores are significantly related to age, length, or weight in non-pregnant bottlenose dolphins. Inoue et al. (2018) used enzyme immunoassays to 520 measure P₄ and achieved lower a detection limit (0.2 ng g⁻¹ vs. 5.62 ng g⁻¹), which likely

 improved their ability to use P_4 to differentiate between mature and immature females. 522 Had we achieved a lower limit of detection for blubber P_4 we would likely have been better able to address this question. However, over half of the mature non-pregnant 524 minke whales in Inoue et al. (2018) had blubber P_4 levels within our range of detection; thus, if such a relationship existed in our study, we should have detected it despite our 526 high RL. This suggests that there is either a difference between species and/or a 527 seasonal component to the relationship between blubber P_4 and female maturity (notably, Inoue et al. (2018) only sampled in December, January, and February). Pallin et al. (2018b) found that most of the non-pregnant humpback whales in their study 530 exhibited blubber P_4 concentrations that were indistinguishable from immature females, further supporting this conclusion. Thus, relationships between female maturity and steroid hormone profiles in cetaceans require further investigation.

Taken together, our results support our overall hypothesis that blubber is well-suited for assessing changes in endocrine function over relatively broad temporal scales but cannot currently be used to study instantaneous endocrine status. Our findings demonstrate that blubber hormone profiles can be used to study shifts in steroid hormone profiles associated with important physiological states, including sexual maturity, pregnancy status, and acute stress. In the future, with further assessment of these relationships using these techniques, it is possible that blubber could be used to classify an individual's sex, maturity, and reproductive status without having to perform physical/ultrasound exams or assessing genetic sex or morphometrics. This is particularly important because demographic, morphometric, and health data often cannot currently be collected when using remote sampling techniques. Overall, this study advances our understanding of cetacean endocrinology and improves our ability to assess cetacean reproductive, developmental, and stress physiology with remotely collected samples. Additionally, this study may provide important insights into the use of adipose tissue to assess endocrine physiology in other species and/or the use of other alternative matrices in marine mammals.

5. Acknowledgements

Funding for this work was primarily provided by the NMFS Marine Mammal Health and

- Stranding Program. Additional support was provided by the Medical University of South
- Carolina, the National Institute of Standards and Technology, and the National Marine

Mammal Foundation. This research was also made possible in part by a grant from the Gulf of Mexico Research Initiative; those data are publicly available through the Gulf of Mexico Research Initiative Information & Data Cooperative (GRIIDC) at https://data.- gulfresearchinitiative.org (doi: http://dx.doi.org/10.7266/N7GF0S16). Samples in Barataria Bay were collected as part of the Natural Resource Damage Assessment 559 following the *Deepwater Horizon* oil spill. The authors would like to thank all members of the NIST Biorepository for their assistance with sample collection, archiving, and management. We also thank Brian Quigley (NMMF) for his help with data collection and management, and Kevin Huncick (NIST) for instrumental support, maintenance, and troubleshooting.

6. Conflicts of Interest

The authors declare that they have no conflict of interest in the publication of this

manuscript. Commercial equipment, instruments, or materials are identified to specify

adequately the experimental procedure. Such identification does not imply

recommendation or endorsement by the National Institute of Standards and Technology

nor the National Oceanographic and Atmospheric Administration, nor does it imply that

the materials or equipment identified are necessarily the best available for the purpose.

			Male	Female				
	RL	Adult	Subadult	Non- Pregnant Pregnant		Probable Pregnant		
	(ng)	$n = 11$	$n = 15$	$n = 32$	$n = 15$	$n = 4$		
Blubber Hormone								
17-hydroxyprogesterone (17OHP ₄)	0.535	63.6%	0.00%	0.00%	42.9%	0.00%		
Testosterone (T)	0.260	63.6%	0.00%	0.00%	0.00%	0.00%		
Androstenedione (AE)	0.206	72.7%	6.67%	0.00%	28.6%	0.00%		
Progesterone (P_4)	5.62	0.00%	0.00%	0.00%	100 %	100 %		
Cortisone (E)	0.0838	90.9%	93.3%	90.6%	64.3%	100 %		
Cortisol (F)	0.856	81.8%	66.6%	71.9%	21.4%	100 %		
11-deoxycorticosterone (S)	0.0779	27.3%	13.3%	34.4%	0.00%	50.0%		
Plasma Hormone								
17-hydroxyprogesterone (17OHP ₄)	0.114	90.9%	33.3 %	38.7%	73.3%	50.0%		
Testosterone (T)	0.259	90.9%	33.3%	16.1%	13.3 %	25.0%		
Androstenedione (AE)	0.0545	100 %	73.3%	3.23%	73.3 %	0.00%		
Progesterone (P ₄)	0.459	0.00%	0.00%	6.45%	100 %	100 %		
Cortisone (E)	0.200	90.9%	100 %	96.8%	100 %	100 %		
Cortisol (F)	0.853	90.9%	100 %	96.8%	100 %	100 %		
11-deoxycorticosterone (S)	NQ.	0.00%	0.00%	0.00%	0.00%	0.00%		

Table 1. Sample size and hormone detection frequency by demographic group and sample matrix.

11-deoxycorticosterone (S) NQ 0.00 % 0.00 % 0.00 % 0.00 % 0.00 % RL = reporting limit, NQ = not quantified Non-pregnant female plasma and pregnant female blubber excludes one sample each which were lost during processing (Supplemental Table 6)

		Blubber (ng g^{-1} wet weight blubber)						Plasma (ng g^{-1} wet weight plasma)						
		170HP4		AE	P ₄	Е	F	\mathbf{s}	170HP4	Т	AE	P ₄	Е	F
Adult Male	Min.	1.59	1.08	3.45	ND	0.492	0.0178	0.0383	0.245	2.04	0.315	ND	1.71	6.76
	Med.	4.02	4.24	10.3	ND	0.803	2.96	0.173	1.57	17.7	1.02	ND	3.23	12.8
	Max.	13.7	17.4	63.3	ND	2.31	14.6	0.780	12.7	56.9	2.86	ND	6.64	30.2
Subadult Male	Min.	ND.	ND	3.48	ND	0.349	0.0596	0.159	0.115	0.395	0.0571	ND	1.23	4.62
	Med.	ND	ND.	3.48	ND	0.709	3.13	0.247	0.129	0.571	0.136	ND	1.84	9.53
	Max.	ND	ND	3.48	ND	1.23	8.69	0.334	0.138	0.732	0.449	ND	3.20	17.4
Non- Pregnant Female	Min.	ND	ND	ND	ND	0.184	0.0940	0.0779	0.143	0.191	0.0352	0.402	0.886	2.89
	Med.	ND	ND	ND	ND	1.02	4.54	0.227	0.180	0.342	0.0352	1.71	2.18	10.6
	Max.	ND	ND	ND	ND	2.63	21.0	1.00	0.287	0.526	0.0352	3.02	4.19	17.5
Pregnant Female	Min.	2.05	ND	3.92	23.4	0.295	0.723	ND	0.265	0.203	0.0582	4.25	0.651	1.67
	Med.	2.12	ND	5.49	95.8	0.486	0.873	ND	0.371	0.237	0.223	13.8	1.06	7.63
	Max.	2.91	ND	8.88	174	0.900	6.13	ND	1.19	0.272	0.732	20.6	2.25	13.9
Probable Pregnant Female	Min.	ND	ND	ND	29.4	0.598	0.154	0.0791	0.213	0.213	ND	14.3	0.977	5.74
	Med.	ND	ND	ND	90.2	0.688	1.28	0.116	0.255	0.213	ND	16.8	1.62	9.85
	Max.	ND	ND	ND	146	0.859	5.00	0.154	0.297	0.213	ND	20.6	1.78	12.4

Table 2. Minimum, median, and maximum detected values of hormones by demographic group and sample matrix. Shaded cells indicate hormones that were not detected in specific matrices within groups.

ND = not detected

Shindiatod component mathodol Dolada raidoo mgmight randolo load	PC ₁	PC ₂	PC ₃	PC4
% Variance	27.24	21.75	12.83	9.191
Blubber Hormone				
17-hydroxyprogesterone $(17OHP4)$	0.667	0.304	0.130	0.034
Testosterone (T)	0.699	-0.156	-0.127	0.124
Androstenedione (AE)	0.804	-0.002	-0.122	0.001
Progesterone (P_4)	0.082	0.912	-0.061	-0.079
Cortisone (E)	0.002	-0.352	0.201	0.751
Cortisol (F)	-0.053	-0.140	-0.024	0.772
11-deoxycorticosterone (S)	0.157	0.176	0.183	0.602
Plasma Hormone				
17-hydroxyprogesterone $(17OHP4)$	0.781	0.194	0.017	0.068
Testosterone (T)	0.768	-0.275	0.064	0.106
Androstenedione (AE)	0.825	0.025	0.114	-0.174
Progesterone (P_4)	-0.073	0.900	-0.132	-0.123
Cortisone (E)	0.087	-0.170	0.943	0.170
Cortisol (F)	-0.082	-0.036	0.970	0.121

Table 3. PCA rotated component matrices. Bolded values highlight variable loading with an absolute value greater than 0.4.

Figure 1. Steroidogenesis pathway. Boxed pathway indicates the Δ4 androgen pathway. Parentheticals indicate hormone abbreviation.

Figure 2. (A and B) Principal components analysis score and loading plots inclusive of both plasma and blubber hormone measurements from all bottlenose dolphin blubber and blood samples; markers indicate individuals (color and shape indicate demographic group and site per the key) and arrows indicate magnitude and direction of variable loading (the prefix "b_" indicates blubber hormone variable, while the prefix "p_" indicates plasma hormone variable). (**C**) Interaction plot for PC1 scores (i.e., mean PC1 scores by demographic group and site-month); error bars indicate standard deviation. (**D and E**) Differences in PC1 and PC2 scores, respectively, by demographic group and site-month; horizontal lines indicate demographic groups in which all sites-months were combined into a single group for pairwise comparison; groups with different letter headings are significantly different per pairwise t-test with Benjamini-Hochberg correction (α = 0.05); numbers below boxes indicate sample size. AM = adult male, SM $=$ subadult male, NPF = non-pregnant female, PF = pregnant female, PPF = probable pregnant female.

Figure 3. (A) Relationship between PC3 score (circulating corticosteroid profile) and elapsed time to blood collection; markers indicate individuals (color and shape indicate demographic group and site per the key); **(B)** Relationship between PC4 score (blubber corticosteroid profile) and elapsed time to blubber collection; markers indicate individuals (color and shape indicate demographic group and site-month per the key). Demographic group and site-month are only shown for visualization, neither of these variables are included in the final models for PC3 or PC4 scores. $AM =$ adult male, $SM =$ subadult male, $NPF = non-prequant$ female, $PF = prequant$ female, $PPF = probable prequant$.

Figure 4. Relationship between age (**A**), body length (**B**), and body weight (**C**) and PC1 scores (androgen profile) in male bottlenose dolphins stratified by site-month (nonpregnant females shown with reduced size markers for comparison); markers indicate individuals (color and shape indicate demographic group and site-month per the key). $AM =$ adult male, $SM =$ subadult male, $NPF =$ non-pregnant female.

References:

- Atkinson, S., Yoshioka, M., 2007. Endocrinology of reproduction, in: Reproductive biology and phylogeny of cetacea: Whales, porpoises and dolphins, D.L. Miller, 171-192.
- Balmer, B.C., Wells, R.S., Schwacke, L.H., Schwacke, J.H., Danielson, B., George, R.C., Lane, S.M., Mclellan, W.A., Pabst, D.A., Sparks, K., 2014. Integrating multiple techniques to identify stock boundaries of common bottlenose dolphins (Tursiops truncatus). Aquatic Conservation: Marine and Freshwater Ecosystems 24, 511-521.
- Boggs, A.S., Ragland, J.M., Zolman, E.S., Schock, T.B., Morey, J.S., Galligan, T.M., Dalle Luche, G., Balmer, B.C., Wells, R.S., Kucklick, J.R., Schwacke, L.H., 2019. Remote blubber sampling paired with liquid chromatography tandem mass spectrometry for steroidal endocrinology in free-ranging bottlenose dolphins (Tursiops truncatus). Gen. Comp. Endocrinol. 281, 164-172.
- Boggs, A.S.P., Schock, T.B., Schwacke, L.H., Galligan, T.M., Morey, J.S., McFee, W.E., Kucklick, J.R., 2017. Rapid and reliable steroid hormone profiling in Tursiops truncatus blubber using liquid chromatography tandem mass spectrometry (LC-MS/MS). Anal. Bioanal. Chem. 409, 5019-5029.
- Carrizo, D.G., Rastrilla, A.M., Tellería, C.M., Aguado, L.I., 1994. Androstenedione stimulates progesterone production in corpora lutea of pregnant rats: an effect not mediated by oestrogen. The Journal of steroid biochemistry and molecular biology 51, 191-197.
- Champagne, C.D., Kellar, N.M., Crocker, D.E., Wasser, S.K., Booth, R.K., Trego, M.L., Houser, D.S., 2017. Blubber cortisol qualitatively reflects circulating cortisol concentrations in bottlenose dolphins. Marine Mammal Science 33, 134-153.
- Champagne, C.D., Kellar, N.M., Trego, M.L., Delehanty, B., Boonstra, R., Wasser, S.K., Booth, R.K., Crocker, D.E., Houser, D.S., 2018. Comprehensive endocrine response to acute stress in the bottlenose dolphin from serum, blubber, and feces. Gen. Comp. Endocrinol. 266, 178-193.
- Dziuban, C.D., Shirkey, E.C., 1974. When is a correlation matrix appropriate for factor analysis? Some decision rules. Psychol. Bull. 81, 358-361.
- Fu, R., Zhai, A., 2010. Determination of hormones in shrimp by Agilent 1290 Infinity LC, Poroshell 120 LC column and QuEChERS sample prep. Aglient.
- Galligan, T.M., Balmer, B.C., Schwacke, L.H., Bolton, J.L., Quigley, B.M., Rosel, P.E., Ylitalo, G.M., Boggs, A.S., 2019. Examining the relationships between blubber steroid hormones and persistent organic pollutants in common bottlenose dolphins. Environmental Pollution 249, 982-991.
- Galligan, T.M., Schwacke, L.H., Houser, D.S., Wells, R.S., Rowles, T., Boggs, A.S.P., 2018a. Characterization of circulating steroid hormone profiles in the bottlenose dolphin (Tursiops truncatus) by liquid chromatography–tandem mass spectrometry (LC– MS/MS). Gen. Comp. Endocrinol. 263, 80-91.
- Galligan, T.M., Schwacke, L.H., McFee, W.E., Boggs, A.S.P., 2018b. Evidence for cortisol– cortisone metabolism by marine mammal blubber. Mar. Biol. 165, 114.
- Harrison, R.J., Ridgway, S.H., 1971. Gonadal activity in some bottlenose dolphins (Tursiops truncatus). J. Zool. 165, 355-366.
- Hart, L.B., Wells, R.S., Schwacke, L.H., 2013. Reference ranges for body condition in wild bottlenose dolphins Tursiops truncatus. Aquat. Biol. 18, 63-68.
- Hillbrand, F.W., Elsaesser, F., 1983. Concentrations of progesterone in the backfat of pigs during the oestrous cycle and after ovariectomy. J. Reprod. Fertil. 69, 73-80.
- Hohn, A.A., Scott, M.D., Wells, R.S., Sweeney, J.C., Irvine, A.B., 1989. Growth layers in teeth from known-age, free-ranging bottlenose dolphins. Marine Mammal Science 5, 315-342.
- Houser, D.S., Yeates, L.C., Crocker, D.E., 2011. Cold Stress Induces an Adrenocortical Response in Bottlenose Dolphins (Tursiops truncatus). J. Zoo Wildl. Med. 42, 565-571.
- Hunt, K.E., Rolland, R.M., Kraus, S.D., 2014a. Detection of steroid and thyroid hormones via immunoassay of North Atlantic right whale (*Eubalaena glacialis*) respiratory vapor. Marine Mammal Science 30, 796-809.
- Hunt, K.E., Stimmelmayr, R., George, C., Hanns, C., Suydam, R., Brower, H., Rolland, R.M., 2014b. Baleen hormones: a novel tool for retrospective assessment of stress and reproduction in bowhead whales (Balaena mysticetus). Conservation physiology 2.
- Inoue, S., Yasunaga, G., Pastene, L.A., 2018. Determining sexual maturity in female Antarctic minke whales during the feeding season based on concentrations of progesterone in blubber. SC/67B/SCSP/05. I.W. Commission
- Kellar, N.M., Catelani, K.N., Robbins, M.N., Trego, M.L., Allen, C.D., Danil, K., Chivers, S.J., 2015. Blubber cortisol: a potential tool for assessing stress response in free-ranging dolphins without effects due to sampling. PLoS ONE 10, e0115257.
- Kellar, N.M., Keliher, J., Trego, M.L., Catelani, K.N., Hanns, C., George, J.C., Rosa, C., 2013. Variation of bowhead whale progesterone concentrations across demographic groups and sample matrices. Endangered Species Research 22, 61-72.
- Kellar, N.M., Trego, M.L., Marks, C.I., Chivers, S.J., Danil, K., Archer, F.I., 2009. Blubber testosterone: a potential marker of male reproductive status in short - beaked common dolphins. Marine Mammal Science 25, 507-522.
- Kellar, N.M., Trego, M.L., Marks, C.I., Dizon, A.E., 2006. Determining pregnancy from blubber in three species of delphinids. Marine Mammal Science 22, 1-16.
- Kershaw, J.L., Hall, A.J., 2016. Seasonal variation in harbour seal (Phoca vitulina) blubber cortisol-A novel indicator of physiological state? Sci. Rep. 6, 21889.
- Kershaw, J.L., Sherrill, M., Davison, N.J., Brownlow, A., Hall, A.J., 2017. Evaluating morphometric and metabolic markers of body condition in a small cetacean, the harbor porpoise (Phocoena phocoena). Ecol. Evol. 7, 3494-3506.
- Kirby, V., Ridgway, S., 1984. Hormonal evidence of spontaneous ovulation in captive dolphins, Tursiops truncatus and Delphinus delphis. Rep Int Whal Commn 6, 459-464.
- Mansour, A.A.H., Mkay, D.W., Lien, J., Orr, J.C., Banoub, J.H., Øien, N., Stenson, G., 2002. Determination of pregnancy status from blubber samples in minke whales (Balaenoptera acutorostrata). Marine Mammal Science 18, 112-120.
- McFee, W.E., Schwacke, J.H., Stolen, M.K., Mullin, K.D., Schwacke, L.H., 2010. Investigation of growth phases for bottlenose dolphins using a Bayesian modeling approach. Marine Mammal Science 26, 67-85.
- McFee, W.E., Speakman, T.R., Balthis, L., Adams, J.D., Zolman, E.S., 2014. Reproductive seasonality of a recently designated bottlenose dolphin stock near Charleston, South Carolina, U.S.A. Marine Mammal Science 30, 528-543.
- Noren, D.P., Mocklin, J.A., 2012. Review of cetacean biopsy techniques: Factors contributing to successful sample collection and physiological and behavioral impacts. Marine Mammal Science 28, 154-199.
- Pallin, L., Baker, C.S., Steel, D., Kellar, N.M., Robbins, J., Johnston, D.W., Nowacek, D.P., Read, A.J., Friedlaender, A.S., 2018a. High pregnancy rates in humpback whales (Megaptera novaeangliae) around the Western Antarctic Peninsula, evidence of a rapidly growing population. Royal Society Open Science 5, 180017.
- Pallin, L., Robbins, J., Kellar, N., Bérubé, M., Friedlaender, A., 2018b. Validation of a blubberbased endocrine pregnancy test for humpback whales. Conservation Physiology 6.
- Pérez, S., García-López, Á., De Stephanis, R., Giménez, J., García-Tiscar, S., Verborgh, P., Mancera, J., Martínez-Rodriguez, G., 2011. Use of blubber levels of progesterone to determine pregnancy in free-ranging live cetaceans. Mar. Biol. 158, 1677-1680.

R Core Team, 2018. R: A Language and Environment for Statistical Computing, Vienna, Austria.

Robeck, T.R., Steinman, K.J., O'Brien, J.K., 2017. Characterization and longitudinal monitoring of serum androgens and glucocorticoids during normal pregnancy in the killer whale (Orcinus orca). Gen. Comp. Endocrinol. 247, 116-129.

Robeck, T.R., Steinman, K.J., O'Brien, J.K., 2016. Characterization and longitudinal monitoring of serum progestagens and estrogens during normal pregnancy in the killer whale (Orcinus orca). Gen. Comp. Endocrinol. 236, 83-97.

RStudio Team, 2016. RStudio: Integrated Development for R. RStudio, Inc., Boston, MA.

- Sawyer-Steffan, J.E., Kirby, V.L., Gilmartin, W.G., 1983. Progesterone and estrogens in the pregnant and nonpregnant dolphin, Tursiops truncatus, and the effects of induced ovulation. Biology of Reproduction 28, 897-901.
- Schroeder, J.P., Keller, K.V., 1989. Seasonality of serum testosterone levels and sperm density in Tursiops truncatus. J. Exp. Zool. 249, 316-321.
- Schwacke, L.H., Smith, C.R., Townsend, F.I., Wells, R.S., Hart, L.B., Balmer, B.C., Collier, T.K., De Guise, S., Fry, M.M., Guillette, L.J., Lamb, S.V., Lane, S.M., McFee, W.E., Place, N.J., Tumlin, M.C., Ylitalo, G.M., Zolman, E.S., Rowles, T.K., 2014. Health of common bottlenose dolphins (Tursiops truncatus) in Barataria Bay, Louisiana, following the Deepwater Horizon oil spill. Environ. Sci. Technol. 48, 93-103.
- Smith, C.R., Jensen, E.D., Blankenship, B.A., Greenberg, M., D'Agostini, D.A., Pretorius, D.H., Saenz, N.C., Noll, N., Venn-Watson, S.K., 2013. Fetal omphalocele in a common bottlenose dolphin (Tursiops truncatus). J. Zoo Wildl. Med. 44, 87-92.
- Smith, C.R., Rowles, T.K., Hart, L.B., Townsend, F.I., Wells, R.S., Zolman, E.S., Balmer, B.C., Quigley, B., Ivancˇic΄, M., McKercher, W., Tumlin, M.C., Mullin, K.D., Adams, J.D., Wu, Q., McFee, W., Collier, T.K., Schwacke, L.H., 2017. Slow recovery of Barataria Bay dolphin health following the *Deepwater Horizon* oil spill (2013-2014), with evidence of persistent lung disease and impaired stress response. Endangered Species Research 33, 127-142.
- St. Aubin, D.J., Ridgway, S.H., Wells, R.S., Rhinehart, H., 1996. Dolphin thyroid and adrenal hormones: circulating levels in wild and semidomesticated Tursiops truncatus, and influence of sex, age, and season. Marine Mammal Science 12, 1-13.
- Steinman, K.J., Robeck, T.R., O'Brien, J.K., 2016. Characterization of estrogens, testosterone, and cortisol in normal bottlenose dolphin (Tursiops truncatus) pregnancy. Gen. Comp. Endocrinol. 226, 102-112.
- Takei, Y., Suzuki, I., Wong, M.K., Milne, R., Moss, S., Sato, K., Hall, A., 2016. Development of an animal-borne blood sample collection device and its deployment for the determination of cardiovascular and stress hormones in phocid seals. American Journal of Physiology-Regulatory, Integrative and Comparative Physiology 311, R788-R796.
- Telleri´a, C.M., Stocco, C.O., Stati, A.O., Rastrilla, A.M., Carrizo, D.G., Aguado, L.I., Deis, R.P., 1995. Dual regulation of luteal progesterone production by androstenedione during spontaneous and RU486-induced luteolysis in pregnant rats. The Journal of steroid biochemistry and molecular biology 55, 385-393.
- Thomson, C., Geraci, J., 1986. Cortisol, aldosterone, and leucocytes in the stress response of bottlenose dolphins, Tursiops truncatus. Can. J. Fish. Aquat. Sci. 43, 1010-1016.
- Trego, M.L., Kellar, N.M., Danil, K., 2013. Validation of blubber progesterone concentrations for pregnancy determination in three dolphin species and a porpoise. PloS ONE 8, e69709.
- Trumble, S.J., Robinson, E.M., Berman-Kowalewski, M., Potter, C.W., Usenko, S., 2013. Blue whale earplug reveals lifetime contaminant exposure and hormone profiles. Proceedings of the National Academy of Sciences 110, 16922-16926.
- Urian, K., Duffield, D., Read, A., Wells, R., Shell, E., 1996. Seasonality of reproduction in bottlenose dolphins, Tursiops truncatus. J. Mammal. 77, 394-403.
- Valenzuela-Molina, M., Atkinson, S., Mashburn, K., Gendron, D., Brownell, R.L., 2018. Fecal steroid hormones reveal reproductive state in female blue whales sampled in the Gulf of California, Mexico. General and Comparative Endocrinology 261, 127-135.
- Vu, E.T., Clark, C., Catelani, K., Kellar, N.M., Calambokidis, J., 2015. Seasonal blubber testosterone concentrations of male humpback whales (Megaptera novaeangliae). Marine Mammal Science 31, 1258-1264.
- Waddell, B.J., Albrecht, E.D., Pepe, G.J., 1992. Utilization of maternal and fetal androstenedione for placental estrogen production at mid and late baboon pregnancy. The Journal of steroid biochemistry and molecular biology 41, 171-178.
- Wasser, S.K., Lundin, J.I., Ayres, K., Seely, E., Giles, D., Balcomb, K., Hempelmann, J., Parsons, K., Booth, R., 2017. Population growth is limited by nutritional impacts on pregnancy success in endangered Southern Resident killer whales (Orcinus orca). PLoS ONE 12, e0179824.
- Wells, R.S., Smith, C.R., Sweeney, J.C., Townsend, F.I., Fauquier, D.A., Stone, R., Langan, J., Schwacke, L.H., Rowles, T.K., 2014. Fetal survival of common bottlenose dolphins (Tursiops truncatus) in Sarasota Bay, Florida. Aquat. Mamm. 40, 252.
- Wells, R.S., Tornero, V., Borrell, A., Aguilar, A., Rowles, T.K., Rhinehart, H.L., Hofmann, S., Jarman, W.M., Hohn, A.A., Sweeney, J.C., 2005. Integrating life-history and reproductive success data to examine potential relationships with organochlorine compounds for bottlenose dolphins (Tursiops truncatus) in Sarasota Bay, Florida. Science of the Total Environment 349, 106-119.

60 100 140 180 220 260 300

Body Weight (kg)

