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Ecological Genomics Predicts Climate Vulnerability in an Endangered Southwestern Songbird

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44

45 *Data Accessibility:* The Willow flycatcher genome and annotations are available through DRYAD
46 [DRYAD NUMBER] and population-level RAD-Seq data are available through NCBI's Sequence Read
47 Archive [SRA NUMBER].

48 **Few regions have been more severely impacted by climate change in the United States than the**
49 **Desert Southwest. Here we use ecological genomics to assess the potential for adaptation to rising**
50 **global temperatures in a widespread songbird, the willow flycatcher (*Empidonax traillii*), and find**
51 **the endangered desert southwestern subspecies (*E. t. extimus*) most vulnerable to future climate**
52 **change. Highly significant correlations between present abundance and estimates of genomic**
53 **vulnerability – the mismatch between current and predicted future genotype-environment**
54 **relationships – indicate small, fragmented populations of the Southwestern willow flycatcher will**
55 **have to adapt most to keep pace with climate change. Links between climate-associated genotypes**
56 **and genes important to thermal tolerance in birds provide a potential mechanism for adaptation to**
57 **temperature extremes. Our results demonstrate that the incorporation of genotype-environment**
58 **relationships into landscape-scale models of climate vulnerability facilitates more precise**
59 **predictions of climate impacts that can guide conservation in threatened and endangered groups.**

60

61 **Introduction**

62 The effects of climate change on biodiversity are forecast to be one of the leading causes of extinction
63 over the next century (Dawson *et al.* 2011; Warren *et al.* 2013; Pacifici *et al.* 2015; Urban 2015).

64 Evidence of climate-induced local extinctions are now widespread among plant and animal species
65 (Sinervo *et al.* 2010; Wiens 2016) and the velocity of climate change impacts in desert biomes is
66 predicted to be among the fastest (Loarie *et al.* 2009). Recent climate change has altered community

67 composition by favoring generalist taxa over habitat specialists and rare species (Menéndez *et al.* 2006;
68 Estrada *et al.* 2016), but the ability to measure climate impacts below the species level is often lacking.
69 Fine-scale estimates of vulnerability to climate change require an understanding of both the capacity for
70 populations to shift their ranges to track climate conditions, as well as their capacity to tolerate climate
71 alterations *in situ* via phenotypic plasticity or adaptation. Despite the fact that intraspecific variation in
72 climate tolerances may factor critically in the ability of species to move or adapt to environmental change,
73 most modeling efforts ignore local adaptation. However, genomic tools are facilitating assessments of
74 local adaptation in non-model species with increasing reliability (Savolainen *et al.* 2013) and such
75 information can be used to improve climate vulnerability estimates. Here we combine genome-wide
76 sequencing with environmental data to improve predictions of how genotype-environment relationships
77 may be disrupted by future environmental change in an endangered songbird native to the Desert
78 Southwest of the United States, the Southwestern willow flycatcher.

79
80 Until recently, assessing species vulnerability to climate change focused largely on using current range-
81 climate associations to predict distributions under models of future climate (Parmesan & Yohe 2003;
82 Pacifici *et al.* 2015). However, complex biotic interactions (competition, specialization, coevolution, etc.)
83 and or limits to dispersal imposed by physical barriers may limit range shifts, making it important to
84 understand a species' potential to adapt to climate change *in situ* (Williams *et al.* 2008). Methodologies
85 in the field of ecological genomics have provided tools to help incorporate information on local
86 adaptation into climate vulnerability models by identifying regions where climate-induced selective
87 pressure will be highest (Fitzpatrick & Keller 2015), but such methods have yet to be widely implemented.
88 These approaches calculate the difference between current genotype-environment relationships and those
89 predicted under future climate change to identify the geographic regions of greatest mismatch. More
90 specifically, they can be used to ask, "How much would allele frequencies across the range have to
91 change to keep pace with projected changes in climate?". In the absence of a range shift, populations in
92 regions where the mismatch is greatest will either need to adapt or may suffer population declines, as was
93 recently shown in the North American songbird, the Yellow warbler (*Setophaga petechia*) (Bay *et al.*
94 2018).

95
96 Few regions in North America will be more severely impacted by temperature extremes than the desert
97 Southwest (Diffenbaugh *et al.* 2008; Hsiang *et al.* 2017). While most large-scale analyses of climate
98 impacts in birds have focused on changes in geographic ranges or shifts in migratory phenology to better
99 synchronize arrival times with earlier spring onset (Both & Visser 2001; Both *et al.* 2006; Stephens *et al.*
100 2016), these changes will do little to offset the impact of summer heat waves in desert regions. Recent

101 work suggests that small desert passerines, in particular, will experience higher rates of mortality due to
102 dehydration and hyperthermia as the frequency of extreme temperature events increases (Albright *et al.*
103 2017). In addition, work in poultry has shown that high temperatures can cause heart strain, or in some
104 cases heart failure, as birds attempt to dissipate heat through increased blood circulation. Further, this
105 work has shown that such stress is not just physiological in nature, but is associated with differential
106 expression in a suite of ~300 genes (Zhang *et al.* 2017). Based on these studies, we predict that genes
107 important to thermal cooling will be under strong selection in small desert passerines as the frequency of
108 heat waves increases.

109
110 The endangered Southwestern willow flycatcher provides an example of a desert passerine for which a
111 better understanding of climate vulnerability has important implications for its conservation. This desert
112 subspecies is one of four subspecies within the willow flycatcher whose combined ranges span the
113 continental United States (Fig. 1; Pacific Northwestern form, *E. t. brewsteri*; Western Central form, *E. t.*
114 *adastus*; and Eastern form, *E. t. traillii*). The presence of the Southwestern willow flycatcher in particular
115 is associated with riparian woodlands along streams and waterways (Sedgwick 2000) and such habitats
116 are thought to provide important refuges from temperature extremes (Chen 1999; McLeod *et al.* 2008).
117 At the turn of the century, the Southwestern willow flycatcher was described as common wherever its
118 specialized habitat existed (Grinnell & Miller 1944), but by 1995 when it was listed under the Endangered
119 Species Act, the number of known breeding pairs had been reduced to between 300 and 500 (Unitt 1987;
120 Sogge *et al.* 1997). Population declines have been attributed to loss of riparian habitats in the Southwest
121 following dam-building, water diversions, groundwater pumping, urbanization, agricultural development,
122 and livestock grazing (Service 2002), but the role that climate change may have played in declines is
123 unknown. Some researchers have questioned the subspecies designation of the Southwestern willow
124 flycatcher, suggesting that it is a peripheral population of an otherwise widespread species with no
125 evidence for ecological distinctiveness (Zink 2015), although this suggestion has been questioned
126 (Theimer *et al.* 2016). Here we use ecological genomics to investigate the potential for ecological
127 distinctiveness within the willow flycatcher as well as the potential role of rising global temperatures on
128 its future persistence.

129
130 To investigate potential genomic signals of local adaptation in the willow flycatcher, we tested for
131 significant genotype-environment correlations using 105,000 SNP markers from 219 individuals spanning
132 24 populations across the breeding range (Fig. 1; Table 1). To identify the genomic locations of climate-
133 associated SNPs in relation to genes and gene regions important to adaptation under climate change, we
134 also assembled and annotated the first willow flycatcher genome. Significant genotype-environment

135 correlations for a subset of loci were further validated by genotyping an additional 274 individuals
136 spanning 25 populations. To identify geographical regions where climate-induced selective pressure is
137 predicted to be greatest under future climate change, we used gradient forest modeling and calculated an
138 index of genomic vulnerability - defined as the mismatch between current and future genotype-
139 environment relationships. We compared genomic vulnerability across the four subspecies and examined
140 relationships between abundance and genomic vulnerability in order to understand which geographical
141 regions will be most severely impacted by climate-induced selective pressure.

142

143

144

145

146 **Materials and Methods**

147 *Sample Collection and DNA extraction*

148 We compiled a collection of 493 willow flycatcher blood or tissue samples from 41 locations across the
149 breeding range using a combination of samples from previous studies, museum donations and new field
150 collections (Paxton 2000). 219 individuals from 24 populations were used to test for genome-wide
151 genotype-environment correlations, while 274 individuals spanning 25 populations (8 replicate and 17
152 new populations) were used to validate a subset of significant genotype-environment correlations
153 identified in the genome-wide analysis ($N_{\text{total_indiv}} = 493$; $N_{\text{total_pops}} = 41$; Table 1; Fig. 1). The willow
154 flycatcher range map and associated subspecies boundaries was taken from the most current United States
155 Geological Survey map used for willow flycatcher surveys (Sogge *et al.* 1997). Samples within one
156 degree latitude and longitude and with no more than 10% difference in any environmental variable (as
157 indicated by our environmental analysis, below) were lumped into a single population. DNA was purified
158 using the Qiagen™ DNeasy Blood and Tissue extraction kit and quantified using the Qubit® dsDNA HS
159 Assay kit (Thermo Fisher Scientific).

160

161 *Genome Sequencing, Assembly, and Annotation*

162 The genomic DNA library was created using a single Southwestern willow flycatcher individual from
163 Roosevelt Lake, AZ and the Illumina TruSeq DNA PCR-Free LT kit (Illumina), with adjustments. One
164 μg of DNA was diluted in 100 μl of AE buffer and fragmented to an average insert size of $\sim 400\text{bp}$. The
165 resulting library was sequenced on two lanes of an Illumina HiSeq2500 using 250bp paired-end
166 sequencing at the QB3 Vincent J. Coates Genomics Sequencing Laboratory, UC Berkeley. Two mate-pair
167 libraries were also created, using 4kb and 8kb inserts and sequenced on one-third of a 100bp paired-end
168 Illumina HiSeq 2500 lane at the Huntsman Cancer Center at the University of Utah. The 250bp paired

169 end reads were used to assemble contigs with the Discover DeNovo assembler from the Broad Institute
170 (<http://www.broadinstitute.org>), discarding contigs less than 1000 bp in length. Mate pair reads were
171 trimmed and separated from paired end reads using NxTrim (O'Connell *et al.* 2015) and contigs were
172 scaffolded with SSPACE (overlap requirement k=3) (Boetzer *et al.* 2010) using both paired end and mate
173 pair libraries. We used reapr (Hunt *et al.* 2013) and mapping of the 8kb insert library to break the
174 assembly at likely error regions. SSPACE scaffolding was repeated with k=5 and scaffolds <5kbp were
175 discarded for the final assembly.

176
177 For annotation purposes, repetitive regions were replaced with N's using RepeatMasker (-species birds)
178 (Tarailo-Graovac & Chen 2009). For annotation, we used two different *ab initio* gene predictions within
179 the MAKER pipeline (Cantarel *et al.* 2008): SNAP and AUGUSTUS. SNAP was trained iteratively using
180 Zebra Finch cDNA and protein sequences downloaded from Ensembl and AUGUSTUS was run using the
181 available chicken training dataset. We used Interproscan (Zdobnov & Apweiler 2001) to add Pfam protein
182 annotation and gene ontology (GO) terms and identified 15,489 genes. Scaffolds were aligned to the
183 Zebra Finch genome (version 3.2.4) using the software promer, part of the MUMmer package (Delcher *et al.*
184 *et al.* 2003). After alignment, we retained the longest consistent alignment (-q) for each chromosome while
185 filtering for similarity (-i 50) and alignment length (-l 500). We then determined the location of the
186 longest alignment for each scaffold and ordered scaffolds accordingly for visualization purposes.

187 188 *SNP Discovery and SNP Filtering*

189 Genome scans on 219 individuals were conducted following the BestRAD library preparation protocol
190 with some modifications (Ali *et al.* 2016). 100ng of DNA was digested using the SbfI restriction enzyme
191 (New England Biolabs, NEB) and fragments were ligated with SbfI adapters prepared with biotinylated
192 ends. Adapter-ligated samples were pooled and cleaned using 1X Agencourt® AMPure XP beads
193 (Beckman Coulter). All DNA fragments were sheared to an average length of 400bp and adapter-ligated
194 fragments were bound to M-280 streptavidin magnetic Dynabeads (Life Technologies). Blunt end repair
195 and ligation of NEBNext Adapters was performed using the Illumina NEBNext Ultra DNA Library Prep
196 Kit (New England Biolabs, NEB) and Agencourt® AMPure XP beads (Beckman Coulter) were used to
197 size select an average of 500bp fragments. The final library was cleaned and run on a Bioanalyzer at the
198 UCLA Technology Center for Genomics & Bioinformatics to check for the size distribution and the
199 absence of contaminants. Two libraries, each comprised of 96 individuals, were initially sequenced in
200 two lanes of 100bp paired end reads on an Illumina HiSeq 2500 at the UC Davis Genome Center. In a

201 third lane, 69 individuals with low coverage from the first two libraries were re-sequenced and an
202 additional 27 individuals were sequenced.

203
204 The program Stacks (Catchen *et al.* 2013) was used to demultiplex, filter and trim adapters from the data
205 with the `process_radtags` function and to remove duplicate read pairs using the `clone_filter` function.
206 Reads were mapped to our genome assembly using bowtie2 (Langmead & Salzberg 2012) and the
207 Haplotype Caller in the Genome Analysis Toolkit was used to identify single nucleotide polymorphisms
208 (SNPs), following best practices from the Broad Institute (<http://www.broadinstitute.org>). Finally, we
209 discarded low quality and rare variants (genotype quality<30; depth<8; minor allele frequency<0.01), as
210 well as indels and non-biallelic SNPs using vcftools (Danecek *et al.* 2011). We used the R package
211 `genoscapeRtools` (DOI: 10.5281/zenodo.848279) to visualize the tradeoff between discarding SNPs with
212 low coverage and discarding individuals with missing genotypes in order to determine the final number of
213 SNPs and individuals retained (SI Fig. 1).

214
215 *Environmental data*

216 For each sampling location, we obtained environmental data from publicly available databases. These 25
217 variables included 19 climate variables downloaded from WorldClim (Hijmans *et al.* 2005) which
218 represented average climate between the years 1960-1990, as well as vegetation indices (Carroll *et al.*
219 2004) (NDVI and NDVIstd, average for the year 2003), Tree Cover (Sexton *et al.* 2013) and elevation
220 data from the Global Land Cover Facility (<http://www.landcover.org>) and a measure of surface moisture
221 characteristics from the NASA Scatterometer Climate Record Pathfinder project (QuickSCAT mean and
222 standard deviation, downloaded from scp.byu.edu).

223
224 *Assessing the role of geography and environment*

225 To assess the relative contributions of geography and the environment to genetic divergence in the willow
226 flycatcher, we compared genetic, environmental, and geographic distance matrices and used multiple tests
227 designed to account for spatial autocorrelation. For locations with > 4 individuals (Table 1), we
228 calculated pairwise F_{ST} across all quality-filtered SNPs using the R package `SNPrelate` (Zheng *et al.* 2012)
229 and pairwise geographic distances from longitude and latitude using the R package `geosphere` (Hijmans
230 *et al.* 2012). We then calculated environmental distance between each pair of sites by removing highly
231 correlated climate variables (Pearson's $r > 0.7$; Table 2; SI Table 1), scaling and centering each
232 environmental variable to account for differences in magnitude, and then calculating pairwise Euclidean
233 differences between sites. Mantel, Partial Mantel, and multiple regression of distance matrices were used

234 to test for associations between linearized F_{ST} ($F_{ST}/1-F_{ST}$) and genetic and environmental distance after
235 accounting for geographic distance.

236

237 *Gradient forest prediction of genomic mismatch*

238 We identified the environmental variables that best explained genetic variation using gradient forest
239 analysis with the R package gradientForest (Ellis *et al.* 2012). Because rare alleles are more likely to yield
240 false positives, we only used SNPs with minor allele frequency >10%. The gradient forest analysis
241 (ntree=500, nbin=201, corr.threshold=0.5) provided a ranked list based on the relative predictive power of
242 all environmental variables (Table 2). To ensure that our model was explaining more variation than we
243 would expect by chance, we compared the number of SNPs with positive R^2 and the mean R^2 across these
244 ‘predictive’ loci (those with positive R^2) to 10 runs with randomized environments. Visualization of the
245 gradient forest model across the range of the willow flycatcher (Buschke *et al.* 2016), was done by
246 generating and extracting uncorrelated BIOCLIM values for 100,000 random points. The final gradient
247 forest model was used to predict the genomic composition from uncorrelated environmental variables for
248 each random point (Table 2). Principal components analysis (PCA) was used to summarize values. To
249 visualize the different adaptive environments across the breeding range, colors were assigned based on
250 the top 3 principal components axes, as recommended by the authors (Ellis *et al.* 2012).

251

252 We extended the gradient forest analysis to predict “genomic vulnerability” using the method presented
253 by Fitzpatrick and Keller (2015). Here, “genomic vulnerability” (termed “genetic offset” by Fitzpatrick
254 and Keller) is a measure of the mismatch between genotype and future predicted environment using
255 associations across current gradients as a baseline. We used the baseline gradient forest model calculated
256 using current BIOCLIM values to predict genomes under future environmental conditions (based on RCP
257 2.6 2050 projections) at the same 100,000 random points. The Euclidean distance between these weighted
258 current and predicted values is what we refer to as “genomic vulnerability” (Bay *et al.* 2018).

259

260 *Identification of SNPs as candidates for environmental selection*

261 To identify SNPs (with minor allele frequency >0.1) that were most highly associated with the top
262 environmental variables while accounting for underlying population structure, we used Latent Factor
263 Mixed Models (LFMM) (Frichot *et al.* 2013). For each of the top 8 environmental variables from the
264 gradient forest analysis, we ran five separate MCMC runs with a latent factor of $K=4$, based on the
265 number of reported subspecies and previous morphological and genetic analysis based upon neutral
266 markers (Paxton 2000). P-values from all five runs were combined and adjusted for multiple tests using a
267 false discovery rate (FDR) correction. We annotated each significant SNP with genes within 25kb

268 upstream or downstream which we assume is within the distances before which LD should break down
269 (Backstrom *et al.* 2006).

270

271 *Validation of climate associated SNPs*

272 To validate genotype-environment correlations identified in the LFMM analysis, we genotyped the top
273 ranking 18 SNPs that were significantly associated with the top 8 climate variables and could be
274 converted to SNPtype Assays in an additional 274 breeding individuals from 25 locations. DNA was
275 extracted from feather samples using the KingFisher™ Cell and Tissue DNA Kit and SNP genotyping
276 was performed on the Fluidigm™ 96.96 IFC controller following manufacturer guidelines. Nine
277 individuals with greater than 8% of missing data were removed from downstream analysis and final allele
278 frequencies were calculated for each SNP at each location. Standard linear regression was used to test for
279 significant associations between climate and allele frequency (FDR-corrected p-value <0.05).

280

281 *Association between genomic vulnerability and abundance*

282 To assess the relationship between genomic vulnerability and abundance and determine which subspecies
283 may be most vulnerable to future climate change, we correlated estimates of genomic vulnerability with
284 willow flycatcher relative abundance from the North American Breeding Bird Survey (BBS) for 2011-
285 2015, including all sites where the species was detected at least once during the history of the survey
286 (Pardieck 2017). In order to associate the two datasets, vector-based BBS relative abundance estimates
287 derived from inverse-distance weighting interpolation (2010-15; Sauer *et al.* 2017; [https://www.mbr-
288 pwrc.usgs.gov/bbs/shape_ra15.html](https://www.mbr-pwrc.usgs.gov/bbs/shape_ra15.html)) of route-level mean counts was converted to raster format with grid
289 resolution of approximately 15 × 15 km. We then extracted values of relative abundance and genomic
290 vulnerability for grid cells including BBS routes using bilinear interpolation (Hijmans 2015). For cells
291 with BBS routes where detections had been recorded, but for which model-based estimates of abundance
292 were not available due to low abundance and isolation from other sites with detections, we assigned mean
293 count values (~ 9% of routes; mean count = 0.06). Significant differences in genomic vulnerability
294 between subspecies were assessed using boxplots with 95% confidence intervals around median
295 vulnerability scores (Chambers *et al.* 1983).

296

297 **Results**

298 *Genome assembly, SNP discovery, and SNP/population filtering*

299 The final Southwestern willow flycatcher genome assembly was 1.2 Gb in length and consisted of 7,791
300 scaffolds (contig N50=79,613bp; scaffold N50=895,074bp). In total, we identified 6,355,061 SNPs
301 across the genome. Discarding low quality SNPs and low coverage individuals resulted in a final set of

302 105,000 SNPs and 175 individuals (SI Fig. 1), with less than 7.4% missing genotypes per SNP
303 (mean=2.3%), less than 15.6% missing SNPs per individual (mean=2.3%), and minor allele frequency
304 greater than 1%. Because F_{ST} is robust to low sample size when a large number of SNPs are employed
305 (Nazareno *et al.* 2017), we retained all populations with a minimum of 4 (mean = 8) individuals for
306 analysis based upon F_{ST} (distance matrix comparisons), resulting in a final dataset of 168 individuals
307 from 22 sampling locations. Alternatively, to avoid bias associated with low sample size in analyses
308 requiring estimates of allele frequency (Gradient Forest and LFMM), we used only populations with a
309 minimum of six individuals (average = 10), resulting in a final dataset of 136 individuals from 14
310 sampling locations (Fig. 1; Table 1).

311

312 *Assessing the role of geography and environment in shaping genetic structure*

313 Pairwise F_{ST} across all quality-filtered SNPs ranged from 0 - 0.11 (SI Table 2). Mantel tests revealed
314 highly significant correlations between genetic and geographic distance ($r = 0.70$, $P = 1 \times 10^{-6}$), genetic
315 and environmental distance ($r = 0.56$, $P = 1 \times 10^{-6}$), and geographic and environmental distance ($r = 0.42$,
316 $p = 1.8 \times 10^{-4}$) (SI Fig. 2A). Partial Mantel tests revealed the correlation between genetic and
317 environmental distance remained significant after accounting for the relationship between genetic and
318 geographic distance ($r = 0.42$, $p = 3 \times 10^{-4}$; SI Fig. 2) and both geographic and environmental distances
319 were significant in a multiple regression of distance matrices (MRM: $R^2=0.59$; geography $P = 1 \times 10^{-5}$;
320 environment $P = 3 \times 10^{-5}$).

321

322 *Gradient forest mapping of genotype environment correlations*

323 More genetic variation was explained by our gradient forest than those generated under randomized
324 environments (SI Fig. 3). A total of 9015 SNPs were correlated with environment with mean $R^2=0.18$,
325 compared to a mean R^2 of 0.13-0.15 across 3489-5633 SNPs for randomized data. We used gradient
326 forest models to identify which climate and vegetation variables were most important in structuring
327 genetic variation in the willow flycatcher and visualize climate-associated allelic variation across the
328 breeding range (Fig. 2 A & B). Seven temperature variables and one precipitation variable were most
329 strongly correlated with genetic variation across the breeding range of willow flycatchers (Table 2).
330 Mapping principal components of gradient forest output revealed putative signals of local adaptation
331 across the US Southwest, the East, the Inter-Mountain West, and the Pacific Northwest geographic
332 regions (Fig. 2C).

333

334 *Identification of candidate SNPs for environmental selection*

335 To investigate genomic regions potentially involved in climate adaptation, we identified genomic regions
336 associated with the top 8 climatic variables (which explained 49% of the total variation) using Latent
337 Factor Mixed Models (25) (Table 2, SI Table 3). We found 77, 100, 104, 97, 97, 58, 107 and 70 SNPs
338 significantly associated with BIO11, BIO10, BIO5, BIO1, BIO6, BIO9, BIO4 and BIO17 respectively
339 (FDR-corrected $p < 0.05$), with 1 SNP located on chromosome 16, Climate_20, shared among 7 variables.
340 The SNPs were broadly distributed across the genome and within 25 KB of 202 genes with a variety of
341 functions (SI Table 3). We identified 5 genes (BRACA1, RND2, CIITA, ICOS, and UBE2C) that were
342 among the ~300 genes found to be differentially expressed in an RNA-seq analysis of thermal tolerance in
343 chickens (Zhang *et al.* 2017), two of which were physically linked (BRACA1 and RND2), and an
344 additional 5 genes (Ecel1, SLC23A2, NOX4, PIRT, and GR1N1) with GO terms related to other aspects
345 of thermal tolerance, including respiratory system process, oxidative stress, and response to heat (Rimoldi
346 *et al.* 2015) (SI Table 4). Three of the 5 genes from the poultry thermal stress study as well as 3 of the
347 genes with potentially relevant GO terms were found to be outliers in association with BIO6 (Fig. 3A).
348 Further, targeted genotyping using Fluidigm SNPtype assays for 18 of the top candidate SNPs in an
349 additional 274 birds from 24 locations validated climate associations in 8/18 SNPs (FDR-corrected
350 $p < 0.05$; SI Table 5). In particular, we found a highly significant relationship between the Climate_20
351 SNP and 7 of the 8 top ranked climate variables in both the genome scan and validation results. While no
352 link between Climate_20 and genes linked to thermal tolerance in birds was found, the highly significant
353 relationship between this SNP and climate variables reflective of the intensity of summer heat waves,
354 such as Mean Temperature of the Warmest Quarter (BIO10), suggests a potential role for this region in
355 climate adaptation (Fig. 3 B, C, and D).

356

357 *Prediction of genomic mismatch and association between vulnerability and abundance*

358 Under a model of future climate change, genomic vulnerability was predicted to be highest in the
359 southern part of the willow flycatcher range (Fig. 4A), corresponding to the range of the Southwestern
360 willow flycatcher subspecies range. Overall, highest genomic vulnerability occurred at sites with
361 especially low abundance, resulting in a significant negative correlation between abundance and genetic
362 vulnerability ($r = -0.18$; $P < 0.001$; $df = 1382$; Fig. 4B, C). Abundance of southwestern willow flycatcher
363 was low across sites and correlation between abundance and vulnerability for this subspecies was
364 especially strong ($r = -0.49$; $P = 0.016$; $df = 27$) and weakest for the eastern subspecies region (*traillii*; $r =$
365 -0.11 ; $P < 0.001$; $df = 957$). While there were regions of high and low genomic vulnerability across the
366 range, the southwestern willow flycatcher subspecies had the highest overall median genomic
367 vulnerability score (Fig. 4D).

368

369 **Discussion**

370 Climate envelope models are widely used to predict future species distributions (Parmesan & Yohe 2003;
371 Pacifici *et al.* 2015), but such models do not account for complex biotic interactions (competition,
372 specialization, coevolution, etc.) or barriers to dispersal that may limit range shifts (Williams *et al.* 2008).
373 Here we show evidence for local adaptation in the willow flycatcher, supporting the idea that climate
374 vulnerability estimates based on a single species distribution model across the entire North American
375 breeding range (Zink 2015) could potentially result in misplaced conservation efforts. Here we move
376 beyond species distribution modeling to identify populations that will need to adapt most to keep pace
377 with climate change - a critically important question for the endangered Southwestern willow flycatcher
378 whose dispersal is known to be limited by proximity to water sources (Friggens & Finch 2015). By
379 calculating the difference between current genotype-environment relationships and those predicted under
380 future climate change, we identify regions of highest vulnerability in the southern part of the willow
381 flycatcher range. A comparison of the average genomic vulnerability across all currently recognized
382 subspecies strongly supports the view that allele frequencies in the endangered Southwestern willow
383 flycatcher will have to evolve most to keep pace with climate change. Significant correlations between
384 estimates of genomic vulnerability and abundance from Breeding Bird Survey data confirm that already
385 rare populations in the Southwest and throughout the range have the highest genomic vulnerability,
386 suggesting that climate change may have already had an impact on population declines in regions at the
387 edge of the species niche. Our results demonstrate how the incorporation of genotype-environment
388 relationships into models of climate vulnerability can improve predictions of climate-induced impacts
389 below the species level.

390

391 Assessing the extent of intraspecific variation in climate tolerances is an important first step towards
392 understanding species vulnerability to climate change. Here we investigate the relationship between
393 genetic, geographic and environmental distance in the willow flycatcher and find consistent support for
394 the conclusion both geography and environment are important to genetic divergence in the willow
395 flycatcher (SI Fig. 2). Mapping putatively adaptive genetic variation using gradient forest-transformed
396 climate variables supports the idea that the Pacific Northwest, the Southwest, the East, and the Inter-
397 Mountain West harbor unique genotype-environment correlations. More specifically, our results support
398 the idea that high maximum temperatures during the warmest month (BIO5) are important to genotype-
399 environment correlations in the Southwest, while genotype-environment relationships in the Pacific
400 Northwest are driven by environmental variables such as precipitation during the driest quarter (BIO17)
401 and mean temperatures during the coldest quarter (Figs. 2 & 3). In contrast, genotype-environment

402 correlations in Inter-Mountain West and Eastern populations, center closer zero in the PCA (Fig. 2A),
403 indicating a more moderate impact of climate variables underlying climate adaptation in this area. In sum,
404 our results support the idea that genotype-environment correlations in the willow flycatcher are complex,
405 involving multiple environmental variables and genomic regions and such information can be used to help
406 refine estimates of future climate vulnerability.

407
408 Adaptation to local environments often occurs through natural selection acting on a large number of loci,
409 each with a small effect on phenotype (Orr 2005). Here we identify putative loci important to local
410 adaptation in the willow flycatcher, after accounting for underlying population structure, and find
411 between 58 – 107 SNPs significantly associated with each of the top 8 environmental variables (SI Table
412 3). Independent validation of our top climate-associated SNPs in 274 new individuals from 24
413 populations revealed that 8 of our top 18 loci were likely robust to Type 1 error. While such error is a
414 problem common to all association studies (McCarthy *et al.* 2008), the high number of false positives in
415 our data underscores the idea that genotype-environment associations that cannot be validated should be
416 interpreted with caution. Highly significant associations between Climate_20 and 7 of our 8 top-ranked
417 environmental variables in both the genome scan and validation datasets provides the strongest evidence
418 for local adaptation across the willow flycatcher genome (Fig. 3). While no associations between
419 Climate_20 and genes known to be important to thermal tolerance in birds were identified, the
420 relationship between allele frequency variation in this SNP and Mean Temperature of the Warmest
421 Quarter (BIO5) suggests a potential role for this region in adaptation to temperature extremes. Overall,
422 our results are in keeping with the idea that willow flycatchers exhibit region-specific genotype-climate
423 associations that should be considered when assessing the capacity for endangered populations of the
424 Southwestern willow flycatcher to shift their range in response to rising global temperatures.
425 While genotype-environment correlations have been noted across a variety of plant and animal systems,
426 the mechanisms behind such local adaptation remain less well understood. Recent work on birds supports
427 the idea that exposure to high temperatures can result in dehydration and heat stress related mortality
428 (Albright *et al.* 2017; Zhang *et al.* 2017). As a first step towards understanding the genomic basis of
429 adaptation to temperature in the willow flycatcher, we identify genes within 25KB of our top-ranking
430 climate-associated SNPs (SI Table 4). Our strongest evidence for genes and gene regions that may be
431 important to climate adaptation in this species comes from the overlap between five genes in our panel
432 (BRACA1, RND2, CIITA, ICOS, and UBE2C) and those that were also found to be differentially
433 expressed in a thermal tolerance study in poultry (Zhang *et al.* 2017). More specifically, Zang *et al.* (2016)
434 concluded that expression of these genes was linked to the dissipation of heat through increased heart
435 pumping and blood circulation in smaller breeds of chickens. These results are consistent with the recent

436 work by Albright et al (2017) who found that small passerines in the Desert Southwest were particularly
437 prone to mortality resulting from the failure to maintain body temperatures below lethal limits. While
438 more research is needed, it is possible that physiological pathways responsible for overheating are related
439 to those involved in interspecific adaptation to temperature extremes. Further, while limited gene
440 annotation information for non-model organisms makes us cautious about placing significance on GO
441 term analyses (Stein 2001), we also note the presence of four genes (Ecell1, SLC23A2, NOX4, PIRT, and
442 GRIN1) with GO terms related to heat stress, thermal tolerance, and oxidative stress. Future efforts will
443 focus on validating gene environment correlations at putative heat stress related loci as well as
444 investigating the extent to which the genes identified here may serve as a mechanism for adaptation to
445 temperature extremes in the willow flycatcher.

446
447 Desert ecosystems are home to some of the world's rarest species, many of which are already threatened
448 by climate change (Loarie *et al.* 2009). Methods for assessing climate change impacts that rely on single
449 species distribution models may overlook the importance of local adaptation in the ability of populations
450 to respond to environmental shifts, potentially leading to misplaced conservation efforts. The US Fish
451 and Wildlife Service was considering removing the Southwestern willow flycatcher from the endangered
452 species list, in part because of a single species distribution model that showed no evidence of habitat
453 specialization across the range. Here we annotate the first willow flycatcher genome and use population-
454 level, genome-wide sequencing to show that willow flycatchers are not a single homogenous group, but a
455 composite of locally adapted populations with specific genotype-environment relationships related to
456 differences in temperature extremes. Clear evidence for local adaptation across the range highlights the
457 need for management efforts below the species level if locally adapted populations are to be conserved.
458 Estimates of the mismatch between current genotype-environment correlations and those predicted under
459 future climate indicate that the Southwestern subspecies is at the greatest risk of climate-induced
460 extinction. Our findings support the idea that protection or enhancement of riparian thermal refuges (Chen
461 1999) within regions of lower genomic vulnerability in the desert Southwest may be the most effective
462 strategy for conserving remaining populations of flycatchers by buffering them from temperature
463 extremes.

464
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478 **Literature Cited**

479 Albright, T.P., Mutiibwa, D., Gerson, A.R., Smith, E.K., Talbot, W.A., O'Neill, J.J. *et al.* (2017).
480 Mapping evaporative water loss in desert passerines reveals an expanding threat of lethal
481 dehydration. *Proceedings of the National Academy of Sciences*, 201613625.

482 Ali, O.A., O'Rourke, S.M., Amish, S.J., Meek, M.H., Luikart, G., Jeffres, C. *et al.* (2016). RAD
483 capture (Rapture): flexible and efficient sequence-based genotyping. *Genetics*, 202, 389-400.

484 Backstrom, N., Ovarnstrom, A., Gustafsson, L. & Ellegren, H. (2006). Levels of linkage
485 disequilibrium in a wild bird population. *Biol Letters*, 2, 435-438.

486 Bay, R., Harrigan, R.J., Underwood, V.L., Gibbs, H.L., Smith, T.B. & Ruegg, K. (2018).
487 Genomic signals of selection predict climate-driven population declines. *Science*, 83-89.

488 Boetzer, M., Henkel, C.V., Jansen, H.J., Butler, D. & Pirovano, W. (2010). Scaffolding pre-
489 assembled contigs using SSPACE. *Bioinformatics*, 27, 578-579.

490 Both, C., Bouwhuis, S., Lessells, C. & Visser, M.E. (2006). Climate change and population
491 declines in a long-distance migratory bird. *Nature*, 441, 81-83.

492 Both, C. & Visser, M.E. (2001). Adjustment to climate change is constrained by arrival date in a
493 long-distance migrant bird. *Nature*, 411, 296-298.

494 Buschke, F.T., Brendonck, L. & Vanschoenwinkel, B. (2016). Adding energy gradients and long
495 -distance dispersal to a neutral model improves predictions of Madagascan bird diversity.

496 *Ecology and Evolution*, 6, 6919-6929.

497 Cantarel, B.L., Korf, I., Robb, S.M., Parra, G., Ross, E., Moore, B. *et al.* (2008). MAKER: an
498 easy-to-use annotation pipeline designed for emerging model organism genomes. *Genome Res.*,
499 18, 188-196.

500 Carroll, M., DiMiceli, C., Sohlberg, R. & Townshend, J. (2004). 250m MODIS normalized
501 difference vegetation index. *University of Maryland, College Park, Maryland.*

502 Catchen, J., Hohenlohe, P.A., Bassham, S., Amores, A. & Cresko, W.A. (2013). Stacks: an
503 analysis tool set for population genomics. *Molecular ecology*, 22, 3124-3140.

504 Chambers, J.M., Cleveland, W.S., Kleiner, B. & Tukey, P.A. (1983). *Graphical methods for data*
505 *analysis*. Wadsworth Belmont, CA.

506 Chen, J., Saunders, S.C., Crow, T.R., Naiman, R.J., Brosofske, K.D., Mroz, G.D., Brookshire,
507 B.L. and Franklin, J.F. (1999). Microclimate in forest ecosystem and landscape ecology:
508 variations in local climate can be used to monitor and compare the effects of different
509 management regimes. *Bioscience*, 49, 288-297.

510 Danecek, P., Auton, A., Abecasis, G., Albers, C.A., Banks, E., DePristo, M.A. *et al.* (2011). The
511 variant call format and VCFtools. *Bioinformatics*, 27, 2156-2158.

512 Dawson, T.P., Jackson, S.T., House, J.I., Prentice, I.C. & Mace, G.M. (2011). Beyond
513 predictions: biodiversity conservation in a changing climate. *Science*, 332, 53-58.

514 Delcher, A.L., Salzberg, S.L. & Phillippy, A.M. (2003). Using MUMmer to identify similar
515 regions in large sequence sets. *Current Protocols in Bioinformatics*, 10.13. 11-10.13. 18.

516 Diffenbaugh, N.S., Giorgi, F. & Pal, J.S. (2008). Climate change hotspots in the United States.
517 *Geophysical Research Letters*, 35.

518 Ellis, N., Smith, S.J. & Pitcher, C.R. (2012). Gradient forests: calculating importance gradients
519 on physical predictors. *Ecology*, 93, 156-168.

520 Estrada, A., Morales-Castilla, I., Caplat, P. & Early, R. (2016). Usefulness of species traits in
521 predicting range shifts. *Trends in ecology & evolution*, 31, 190-203.

522 Fitzpatrick, M.C. & Keller, S.R. (2015). Ecological genomics meets community-level modelling
523 of biodiversity: mapping the genomic landscape of current and future environmental adaptation.
524 *Ecology Letters*, 18, 1-16.

525 Frichot, E., Schoville, S.D., Bouchard, G. & François, O. (2013). Testing for associations
526 between loci and environmental gradients using latent factor mixed models. *Molecular biology*
527 *and evolution*, 30, 1687-1699.

528 Friggens, M.M. & Finch, D.M. (2015). Implications of Climate Change for Bird Conservation in
529 the Southwestern US under Three Alternative Futures. *Plos One*, 10.

530 Grinnell, J. & Miller, A.H. (1944). *The distribution of the birds of California*. The Cooper
531 Ornithological Society, Berkeley, CA.

532 Hijmans, R.J. (2015). Geographic Data Analysis and Modeling. R package version 2.5-8.

533 Hijmans, R.J., Cameron, S.E., L., P.J., Jones, P.G. & Jarvis, A. (2005). Very high resolution
534 interpolated global terrestrial climate surfaces. *International Journal of Climatology*, 25, 1965-
535 1978.

536 Hijmans, R.J., Williams, E. & Vennes, C. (2012). geosphere: Spherical Trigonometry. R package
537 version 1.2-28. *CRAN. R-project.org/package=geosphere*.

538 Hsiang, S., Kopp, R., Jina, A., Rising, J., Delgado, M., Mohan, S. *et al.* (2017). Estimating
539 economic damage from climate change in the United States. *Science*, 356, 1362-1369.

540 Hunt, M., Kikuchi, T., Sanders, M., Newbold, C., Berriman, M. & Otto, T.D. (2013). REAPR: a
541 universal tool for genome assembly evaluation. *Genome biology*, 14, R47.

542 Langmead, B. & Salzberg, S.L. (2012). Fast gapped-read alignment with Bowtie 2. *Nat. Meth.*,
543 9, 357-359.

544 Loarie, S.R., Duffy, P.B., Hamilton, H., Asner, G.P., Field, C.B. & Ackerly, D.D. (2009). The
545 velocity of climate change. *Nature*, 462, 1052.

546 McCarthy, M.I., Abecasis, G.R., Cardon, L.R., Goldstein, D.B., Little, J., Ioannidis, J.P. *et al.*
547 (2008). Genome-wide association studies for complex traits: consensus, uncertainty and
548 challenges. *Nat. Rev. Genet.*, 9, 356.

549 McLeod, M.A., Koronkiewicz, T.J., Brown, B.T., Langeberg, W.J. & Carothers, S.W. (2008).
550 Southwestern Willow Flycatcher surveys, demography, and ecology along the lower Colorado
551 River and tributaries, 2003–2007. Five-year summary report submitted to U.S. Bureau of
552 Reclamation, Boulder City, Nevada. SWCA Environmental Consultants, Flagstaff, Arizona.

553 Menéndez, R., Megías, A.G., Hill, J.K., Braschler, B., Willis, S.G., Collingham, Y. *et al.* (2006).
554 Species richness changes lag behind climate change. *Proceedings of the Royal Society of London*
555 *B: Biological Sciences*, 273, 1465-1470.

556 Nazareno, A.G., Bemmels, J.B., Dick, C.W. & Lohmann, L.G. (2017). Minimum sample sizes
557 for population genomics: an empirical study from an Amazonian plant species. *Molecular*
558 *ecology resources*.

559 O’Connell, J., Schulz-Trieglaff, O., Carlson, E., Hims, M.M., Gormley, N.A. & Cox, A.J.
560 (2015). NxTrim: optimized trimming of Illumina mate pair reads. *Bioinformatics*, 31, 2035-2037.

561 Orr, H.A. (2005). The genetic theory of adaptation: a brief history. *Nat. Rev. Genet.*, 6, 119-127.

562 Pacifici, M., Foden, W.B., Visconti, P., Watson, J.E., Butchart, S.H., Kovacs, K.M. *et al.* (2015).
563 Assessing species vulnerability to climate change. *Nature Climate Change*, 5, 215-224.

564 Pardieck, K.L., D.J. Ziolkowski Jr., M. Lutmerding, K. Campbell and M.-A.R. Hudson (2017).
565 North American Breeding Bird Survey

566 Parmesan, C. & Yohe, G. (2003). A globally coherent fingerprint of climate change impacts
567 across natural systems. *Nature*, 421, 37-42.

568 Paxton, E.H. (2000). Molecular genetic structuring and demographic history of the Willow
569 Flycatcher (*Empidonax traillii*). Northern Arizona University Flagstaff, AZ, USA.

570 Rimoldi, S., Lasagna, E., Sarti, F.M., Marelli, S.P., Cozzi, M.C., Bernardini, G. *et al.* (2015).
571 Expression profile of six stress-related genes and productive performances of fast and slow
572 growing broiler strains reared under heat stress conditions. *Meta gene*, 6, 17-25.

573 Savolainen, O., Lascoux, M. & Merilä, J. (2013). Ecological genomics of local adaptation. *Nat.*
574 *Rev. Genet.*, 14, 807-820.

575 Sedgwick, J.A. (2000). Willow flycatcher (*Empidonax traillii*). In: *The Birds of North America*
576 (ed. Rodewald, PG). Cornell Lab of Ornithology Ithaca.

577 Service, U.S.F.a.W. (2002). Final Recovery Plan Southwestern Willow Flycatcher (*Empidonax*
578 *traillii* extimus). (ed. Service, USFaW) Albuquerque, New Mexico, p. 210.

579 Sexton, J.O., Song, X.-P., Feng, M., Noojipady, P., Anand, A., Huang, C. *et al.* (2013). Global,
580 30-m resolution continuous fields of tree cover: Landsat-based rescaling of MODIS vegetation
581 continuous fields with lidar-based estimates of error. *International Journal of Digital Earth*, 6,
582 427-448.

583 Sinervo, B., Mendez-De-La-Cruz, F., Miles, D.B., Heulin, B., Bastiaans, E., Villagrán-Santa
584 Cruz, M. *et al.* (2010). Erosion of lizard diversity by climate change and altered thermal niches.
585 *Science*, 328, 894-899.

586 Sogge, M.K., Marshall, R.M., Sferra, S.J. & Tibbitts, T.J. (1997). A southwestern willow
587 flycatcher natural history summary and survey protocol. Technical Report
588 NPS/NAUCPRS/NRTR-97/12. USGS Colorado Plateau Research Station, Northern Arizona
589 University, Flagstaff, AZ.

590 Stein, L. (2001). Genome annotation: from sequence to biology. *Nat. Rev. Genet.*, 2, 493.

591 Stephens, P.A., Mason, L.R., Green, R.E., Gregory, R.D., Sauer, J.R., Alison, J. *et al.* (2016).
592 Consistent response of bird populations to climate change on two continents. *Science*, 352, 84-
593 87.

594 Tarailo-Graovac, M. & Chen, N. (2009). Using RepeatMasker to identify repetitive elements in
595 genomic sequences. *Current protocols in bioinformatics*, 4.10. 11-14.10. 14.

596 Theimer, T.C., Smith, A.D., Mahoney, S.M. & Ironside, K.E. (2016). Available data support
597 protection of the Southwestern Willow Flycatcher Under the Endangered Species Act. *The*
598 *Condor*, 118, 289-299.

599 Unitt, P. (1987). *Empidonax traillii extimus*: An endangered subspecies. *West. Birds*, 18, 137-
600 162.

601 Urban, M.C. (2015). Accelerating extinction risk from climate change. *Science*, 348, 571-573.

602 Warren, R., VanDerWal, J., Price, J., Welbergen, J., Atkinson, I., Ramirez-Villegas, J. *et al.*
603 (2013). Quantifying the benefit of early climate change mitigation in avoiding biodiversity loss.
604 *Nature Climate Change*, 3, 678-682.

605 Wiens, J.J. (2016). Climate-related local extinctions are already widespread among plant and
606 animal species. *PLoS Biol.*, 14, e2001104.

607 Williams, S.E., Shoo, L.P., Isaac, J.L., Hoffmann, A.A. & Langham, G. (2008). Towards an
608 integrated framework for assessing the vulnerability of species to climate change. *PLoS Biol.*, 6,
609 2621-2626.

610 Zdobnov, E.M. & Apweiler, R. (2001). InterProScan—an integration platform for the signature-
611 recognition methods in InterPro. *Bioinformatics*, 17, 847-848.

612 Zhang, J., Schmidt, C.J. & Lamont, S.J. (2017). Transcriptome analysis reveals potential
613 mechanisms underlying differential heart development in fast-and slow-growing broilers under
614 heat stress. *BMC Genomics*, 18, 295.

615 Zheng, X., Levine, D., Shen, J., Gogarten, S.M., Laurie, C. & Weir, B.S. (2012). A high-
616 performance computing toolset for relatedness and principal component analysis of SNP data.
617 *Bioinformatics*, 28, 3326-3328.

618 Zink, R.M. (2015). Genetics, morphology, and ecological niche modeling do not support the
619 subspecies status of the endangered Willow Flycatcher (*Empidonax traillii extimus*). *The*
620 *Condor*, 117, 78-86.

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Table 1. Sample location information. $N_{RAD_nofilter}$ = number of individuals for genome-wide RAD dataset before filtering for read depth and missing data, N_{RAD_filter} = number of remaining post-filtering, $N_{validation}$ = number of individuals in the SNP validation dataset.

Location	Latitude	Longitude	$N_{RAD_}$	N_{RAD_filter}	$N_{validation}$
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nofilter

East Pima, AZ	32.83	-109.7	-	-	8
San Pedro/Gila River confluence, AZ	32.98	-110.77	18	14	17
West Fort Ditch, NM	33.04	-108.54	-	-	11
San Carlos Reservation, AZ	33.2	-110.44	-	-	30
San Diego, CA	33.28	-117.37	14	4	6
Roosevelt Lake, AZ	33.77	-111.24	20	10	18
White Mountains, AZ	34.00	-109.00	15	13	15
Camp Verde, AZ	34.56	-111.84	-	-	17
Santa Ynez River, CA	34.62	-120.18	-	-	8
Zuni/Nutria Diversion Reservation, NM	35.24	-108.64	-	-	8
South Fork Kern River, CA	35.66	-118.46	20	13	11
Southern Ute Reservation, CO	37.12	-107.59	-	-	6
Pahranagat Lake NWR, NV	37.32	-115.13	-	-	6
Owen's River at Bishop, CA	37.41	-118.48	-	-	12
Alamosa National Wildlife Refuge, CO	37.5	-106	-	-	17
Beaver Creek, CO	37.68	-108.38	-	-	6
Clear Creek, CO	37.79	-108.24	-	-	8
Baltimore Area, MD	39.4	-76.99	-	-	8
Escalante State Wildlife Area, CO	39.47	-106.37	-	-	13
Fish Creek, UT	39.78	-111.20	14	11	-
Rio Blanco Lake, CO	40.09	-108.21	-	-	7
Orefield, PA	40.66	-75.67	21	21	-
White River Confl. to the Green River, UT	40.67	-109.68	7	6	-
Willow Slew, IN	40.98	-87.53	4	4	-
Bigelow Meadows, CA	41.26	-121.88	7	6	-
Agusta, MI	42.3	-85.32	-	-	9
Mink Creek, ID	42.75	-112.39	6	6	-
Malheur NWR, OR	42.83	-118.87	7	6	-
FCTC-SABO, MI	42.84	-85.30	4	4	6
Jones Creek, OR	43.04	-123.97	10	10	-
Little White River Rec. Area, SD	43.17	-101.53	4	4	6
Black Creek, NY	43.38	-73.91	6	4	-
Fall Creek 2, ID	43.43	-111.40	7	7	-
Marion Forks, OR	44.37	-122.02	-	-	14
Finley NWR, OR	44.41	-123.35	3	0	-
Priem Road, OR	44.78	-123.38	7	6	-
Elm Creek, MN	45.13	-93.45	4	0	6
Waubay NWR, SD	45.40	-97.33	4	4	-
Hamon Memorial, MT	45.95	-114.13	5	4	-
Carbondale (Edgwick), WA	47.09	-122.05	8	7	-
Fork clearcut, WA	47.97	-124.40	4	4	-
Total			219	168	273

Table 2. Environmental variables used in the gradient forest analysis, ordered by ranked importance of variables

and the cumulative contribution of each variable. The top eight environmental variables represent 49% of the total.

Variable	Definition	GF Rank	Cumulative Contribution
BIO11 *	Mean Temperature of Coldest Quarter	8.03E-04	7.66
BIO10	Mean Temperature of Warmest Quarter	6.71E-04	14.40
BIO1	Annual Mean Temperature	6.41E-04	21.05
BIO5*	Max Temperature of Warmest Month	6.40E-04	27.47
BIO6	Min Temperature of Coldest Month	5.79E-04	32.90
BIO4*	Temperature Seasonality (standard deviation *100)	5.20E-04	38.30
BIO9	Mean Temperature of Driest Quarter	4.91E-04	43.64
BIO17*	Precipitation of Driest Quarter	4.78E-04	48.76
NDVI_Mean	Vegetation Indices	4.50E-04	53.18
BIO15	Precipitation Seasonality (Coefficient of Variation)	4.28E-04	57.39
BIO7	Temperature Annual Range (BIO5-BIO6)	3.75E-04	61.41
TreeCover	Tree Cover	3.72E-04	65.41
BIO14	Precipitation of Driest Month	3.64E-04	69.36
BIO16	Precipitation of Wettest Quarter	3.09E-04	72.75
BIO19	Precipitation of Coldest Quarter	2.96E-04	76.04
BIO2*	Mean Diurnal Range (Mean of monthly (max temp - min temp))	2.90E-04	79.31
BIO8*	Mean Temperature of Wettest Quarter	2.87E-04	82.49
BIO13	Precipitation of Wettest Month	2.82E-04	85.63
STM	Elevation	2.21E-04	88.36
BIO12	Annual Precipitation	2.12E-04	90.98
BIO3	Isothermality (BIO2/BIO7) (* 100)	2.07E-04	93.54
QuickScat	Surface moisture characteristics	2.02E-04	95.87
BIO18	Precipitation of Warmest Quarter	1.92E-04	98.17
NDVI_StDev	Vegetation Indices	1.81E-04	100.00

* Top ranked, uncorrelated climate variables used for Gradient Forest mapping and distance matrix comparison analyses.

These variables were selected by moving down the list of ranked importance for the full model and discarding

variables highly correlated (Pearson's $r > .7$) with a variable of higher importance.

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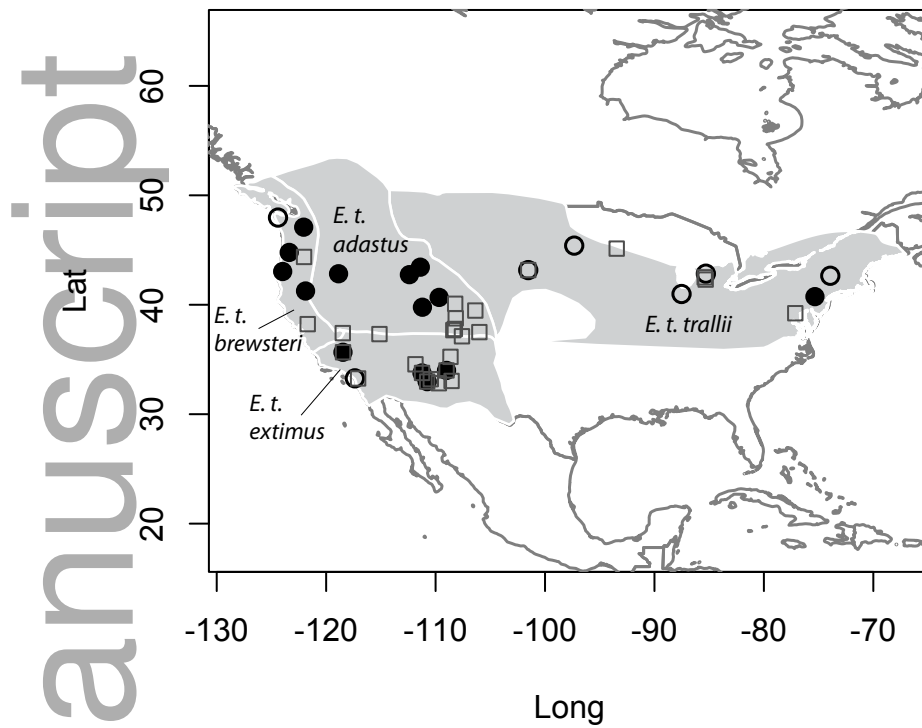


Figure 1. Willow Flycatcher Range Map and Sampling. Open and closed circles represent the data used in distance matrix comparison tests, while only populations represented by closed circles were used in the Gradient Forest analysis. Open gray boxes represent populations used to validate gene-environment correlations. Lines represent currently recognized subspecies boundaries according to Sogge et al (1997). *E. t. brewsteri* = Pacific Coastal, *E. t. adastus* = Interior West, *E. t. trallii* = East, and *E. t. extimus* = Southwest.

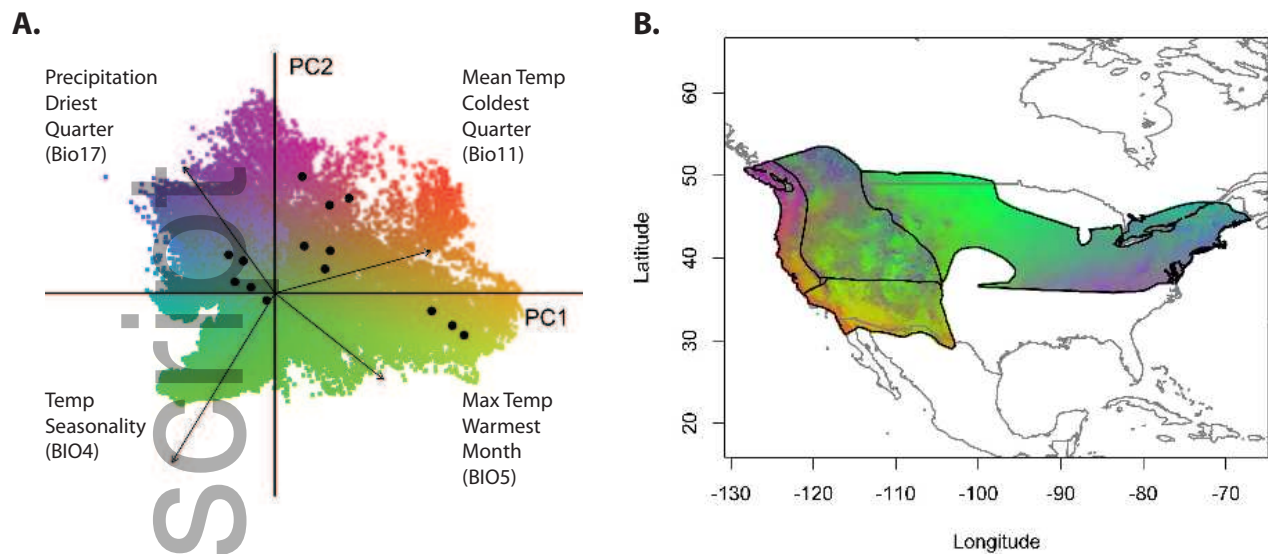


Figure 2. Mapping gene-environment correlations across the willow flycatcher breeding range. A) Principal components analysis of gradient forest-transformed climate variables. Black dots represent the PC scores associated with the sampling locations, while colors are based upon modeled gene-environment correlations from 100,000 random points across the breeding range. Arrows show the loadings of the top ranked uncorrelated environmental variables. B) Gradient forest-transformed climate variables from the PCA mapped to geography support climate adaptation across the breeding range. Black lines designating approximate subspecies locations support the idea that while subspecies are adapted to distinct ecological regions, climate adaptation is complex.

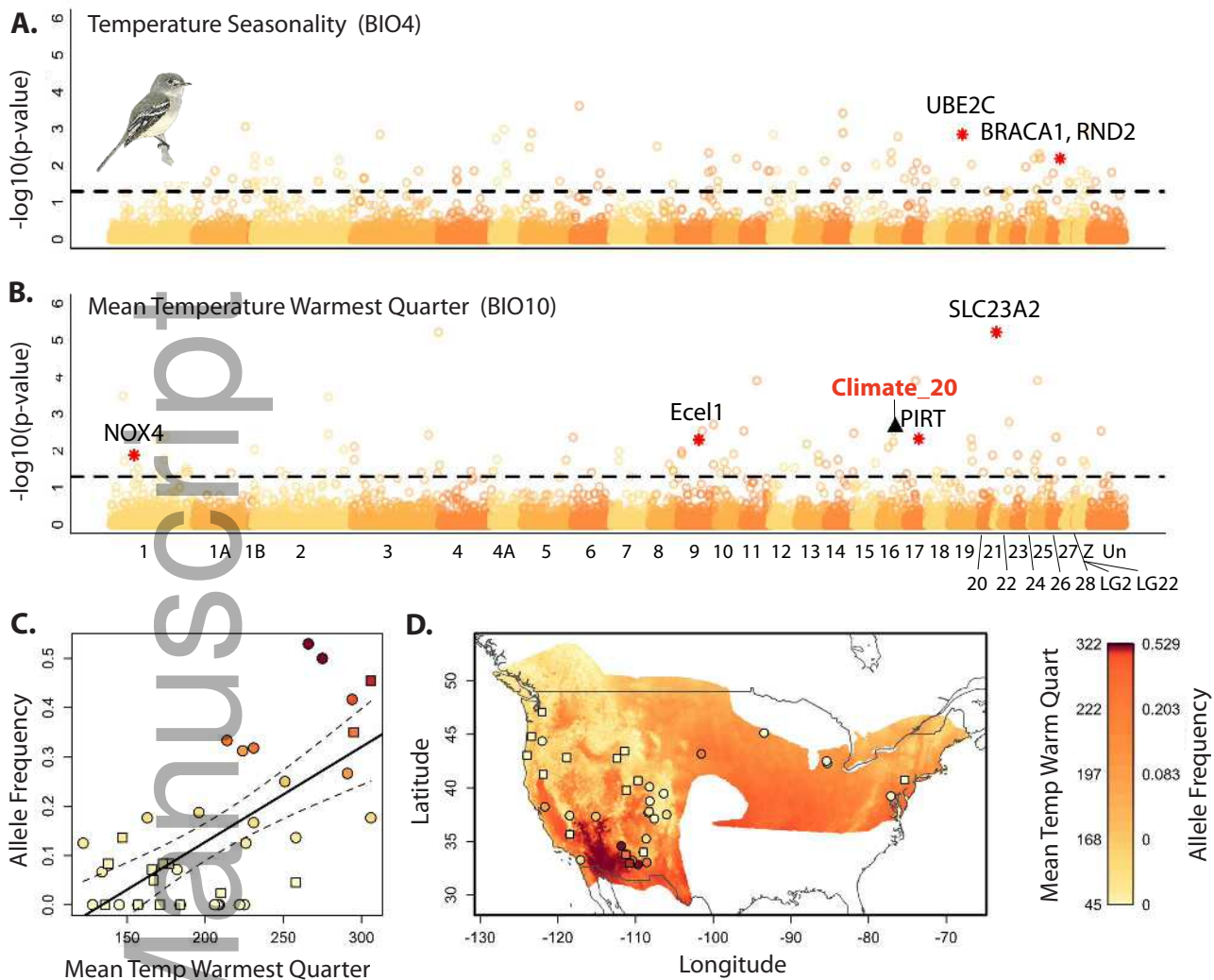


Figure 3. Candidate SNPs linked to temperature in the Willow flycatchers. A) Manhattan plot showing the FDR-corrected significance level for SNPs associated with Temperature Seasonality (BIO4) and B) Mean Temperature of the Warmest Quarter (BIO10). Dashed line represents $p=0.05$. Colors distinguish different chromosomes. Candidate genes linked to thermal tolerance in birds are highlighted by red stars and denoted with gene names, while *Climate_20*, the SNP validated in B and C below, is denoted by a black triangle. No link between *Climate_20* and genes linked to thermal tolerance in birds was found, but the highly significant relationship between this SNP and 7 of the 8 top ranked climate variables (except temperature seasonality shown in A above) in both the genome scan and validation results (SI Table 5) suggest a potential role for this region in climate adaptation. C) Relationship between *Climate_20* and mean temperature of the warmest quarter in genome scan and SNP validation datasets. The allele frequencies from the original genome scan data are denoted by squares, while allele frequencies based upon the validation set are denoted by circles. D) The association between Mean Temperature of the Warmest Quarter (BIO10) and *Climate_20* across geographic space, with population allele frequencies color coded from high frequency (red) to low (yellow).

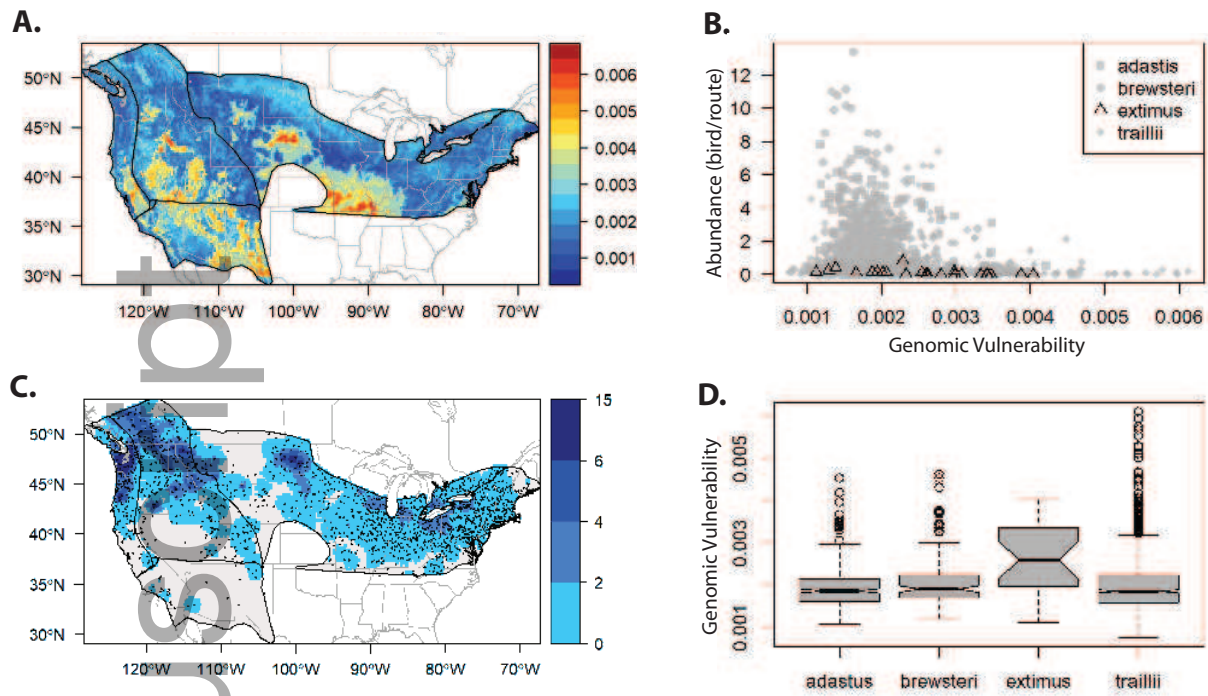


Figure 4. Genomic Vulnerability and abundance in the Willow Flycatcher. A) Map of genomic vulnerability across the Willow Flycatcher breeding range. Red = high genomic vulnerability, blue = low genomic vulnerability, lines indicate subspecies boundaries. B) Genomic Vulnerability versus abundance based upon the estimated mean number of birds/ route in 2011–2015 Breeding Bird Survey. C) Estimates of relative abundance from the BBS based on inverse-distance weighting interpolation. Points indicate the BBS routes where Willow Flycatchers have been recorded. Points in the grey areas fall in regions where abundance was too low or distant from other detection routes to be included in the BBS spatial model. D) Quantile box plots of the median Genomic Vulnerability broken down by subspecies. Open circles represent outliers.