

1 **Desert dust contribution to PM10 loads in Italy: methods and recommendations**
2 **addressing the relevant European Commission Guidelines in support to the Air**
3 **Quality Directive 2008/50**

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18 **Abstract**

19 In 2011 the European Commission (EC) released specific ‘Guidelines’ describing the methods to
20 quantify and subtract the contribution of natural sources from the PM10 values regulated by the
21 European Air Quality Directive (2008/50/EC). This work investigates the applicability to Italy of the
22 EC-Methodology suggested for desert-dust, describes main limitations encountered and proposes
23 specific modifications embedded within a ‘revised-Methodology’ to extend/improve its use. The
24 revised-Methodology capabilities are then evaluated using original, chemically-resolved mineral-
25 dust mass concentration measurements, showing better performances in predicting timing and
26 absolute values of the desert-dust contribution to the daily-PM10 with respect to the current EC-
27 approach. The revised-Methodology is then translated into an automatic (user-independent) tool
28 tailored to the expected final-users. This tool is applied over the Central Italy across a 3-year long
29 period (2012-2014), and over the whole Italian country for a calendar year (2012). The derived
30 results confirm and extend to Italian regions never addressed before some previously observed
31 features of the desert-dust impact over the country, such as a clear latitudinal dependence of the
32 desert-dust impact on the yearly average PM10 (from more than 5 $\mu\text{g}/\text{m}^3$ to less than 0.5 $\mu\text{g}/\text{m}^3$,
33 going from south to north Italy). The modifications introduced within the revised-Methodology also
34 suggest a non-negligible role of desert-dust resuspension in areas characterized by both high traffic
35 levels and soil sealing (urban areas and along the major Italian routes). In the Rome area, such an
36 effect is found to add a contribution of about 2 $\mu\text{g}/\text{m}^3$ (i.e., of 20-30%) to the mean desert-dust load
37 per dust day (about 10 $\mu\text{g}/\text{m}^3$). At the national level, this effect contributes increasing the total
38 number of desert-dust driven exceedances of the PM10 daily limit value even in the northern
39 regions, where the desert-dust impact on the PM10 yearly average is otherwise limited. These
40 results also indicate the direction for possible mitigation strategies to be applied over impacted
41 areas. The successful application of the revised-Methodology over Italy suggests it could represent
42 a valid option for a nationwide standard procedure to quantify the desert-dust contribution to
43 PM10, promoting the homogenisation of the relevant values annually reported to the EC.

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48 **Keywords:** desert-dust, PM10, air quality, Directive 2008/50/EC, resuspension, Italy

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53 **1. Introduction**

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55 A large share of the particulate matter (PM, or aerosols) load into the European atmosphere has
56 an anthropogenic origin (Viana et al., 2012). Still, the contribution of natural sources needs to be
57 quantified to support the EU Member State efforts to meet the PM10-related air quality standards
58 set out by EU legislation (daily limit value, DLV, of 50 $\mu\text{g}/\text{m}^3$, not to be exceeded more than 35 times
59 per year, and yearly average of 40 $\mu\text{g}/\text{m}^3$).

60 In the Euro-Mediterranean regions, the main natural particles contributing to PM10 (i.e., particles
61 with aerodynamic diameter < 10 μm) are desert-dust, sea spray and aerosol from wildfires (e.g.,
62 Viana et al., 2014). In 2011, the European Commission (EC) released specific ‘Guidelines’ (EC, 2011)
63 to provide the Member States with common methods to assess the contribution of particles of
64 natural origin to the PM-related metrics regulated by the Ambient Air Quality Directive 2008/50/EC
65 (EC, 2008). Standardized methods to quantify the contribution of natural particles to the measured

66 PM values are important as, once adopted to prove that limit values are exceeded for natural
67 causes, the EC legislation allows not to consider PM exceedances as such for the purpose of the
68 Directive.

69 In this work we focus on the contribution of desert-dust to PM₁₀ levels, a natural component
70 which has been proved to be significant, especially in southern and western European countries
71 (e.g., Basart et al., 2012a; Pey et al., 2013), and such to cause exceedances of the EU air quality
72 PM₁₀ limit values. In fact, the understanding of the impact of desert-dust on air quality, and the
73 subsequent application of mitigation measures, has been recognized as one of the three major
74 challenges in the field of air pollution chemistry in the recent overview by Gimeno et al. (2013).

75 In this context, the methodology currently suggested by the EC-Guidelines to quantify desert-dust
76 in PM₁₀ (hereafter referred to as the 'EC-Methodology') is certainly a valuable tool to provide
77 European environmental agencies with a common procedure to assess such impact on the regulated
78 PM₁₀ values. The EC-Methodology is however largely based on past, extensive work performed
79 over the Iberian peninsula (Escudero et al., 2005; Escudero et al., 2007; Querol et al., 2009) and,
80 although applied to other regions of Europe (e. g., Matassoni et al., 2009; Aleksandropoulou and
81 Lazaridis, 2013; Pey et al., 2013) there has been little investigation on the actual applicability of this
82 method to other areas, which may have desert-dust regimes different from the western
83 Mediterranean ones in terms of seasonality, driving meteorology and atmospheric processing (e.g.,
84 Israelevich et al., 2012; Gaetani et al., 2012; Salvador et al., 2014; Varga et al., 2014). In some cases,
85 non-negligible differences between the desert-dust contribution derived from the EC-Methodology
86 and that derived by other methods or by chemically-derived desert-dust values have been reported
87 (e.g., Viana et al., 2010; Marconi et al., 2014).

88 A further element needed to promote a widespread exploitation of a shared methodology at the
89 European level is the simplicity of its application by the end-users. This is still a weak aspect of the
90 EC-Methodology, which requires a large number of datasets, images, and modelling resources to be
91 considered in the evaluation process, also leaving large margins of subjectivity in its application.

92 Aim of the present study was testing the applicability of the current EC-Methodology to Italy, and
93 contribute improving it in view of a more widespread use of this or similar methods by
94 regional/national environmental services. The work was carried on in the framework of the
95 DIAPASON Project ('Desert-dust Impact on Air quality through model-Predictions and Advanced
96 Sensors Observations'), funded by the European Union LIFE+ program (www.diapason-life.eu;
97 Gobbi et al., 2017). To reach the study objectives, we first applied the EC-Methodology in Central
98 Italy (chosen as Pilot study area) over a 4-year-long period. Based on this 'pilot application', we then
99 identified some drawbacks in the assumptions the EC-Methodology builds upon, and in its practical
100 implementation in the Italian case (Sections 2.2, 2.3). We then propose specific solutions to
101 overcome the identified drawbacks (Section 2.3), and embed such solutions into a modified,
102 improved version of the EC-Methodology referred to as 'revised-Methodology'. The capabilities of
103 this revised-Methodology in correctly predicting the desert-dust contribution to PM₁₀ values were
104 evaluated over the Pilot study area using original datasets of desert-dust mass concentration
105 derived from chemical measurements on PM₁₀ samples (Section 2.4). The revised-Methodology
106 was then operationally implemented at both the regional and national scale deriving relevant
107 statistics on desert-dust impact on PM₁₀ levels in Italy (Section 3). Besides providing a basis for
108 comparison with the corresponding EC-Methodology outcomes, this analysis allowed to unveil some
109 features of the desert-dust impact over Italy undetectable following the EC-Methodology approach.

110 Due to the large amount of data and evidences contributing to this effort, the more
111 specific/technical aspects of the study have been included in five relevant Appendices to the main
112 text (Appendices A-E).

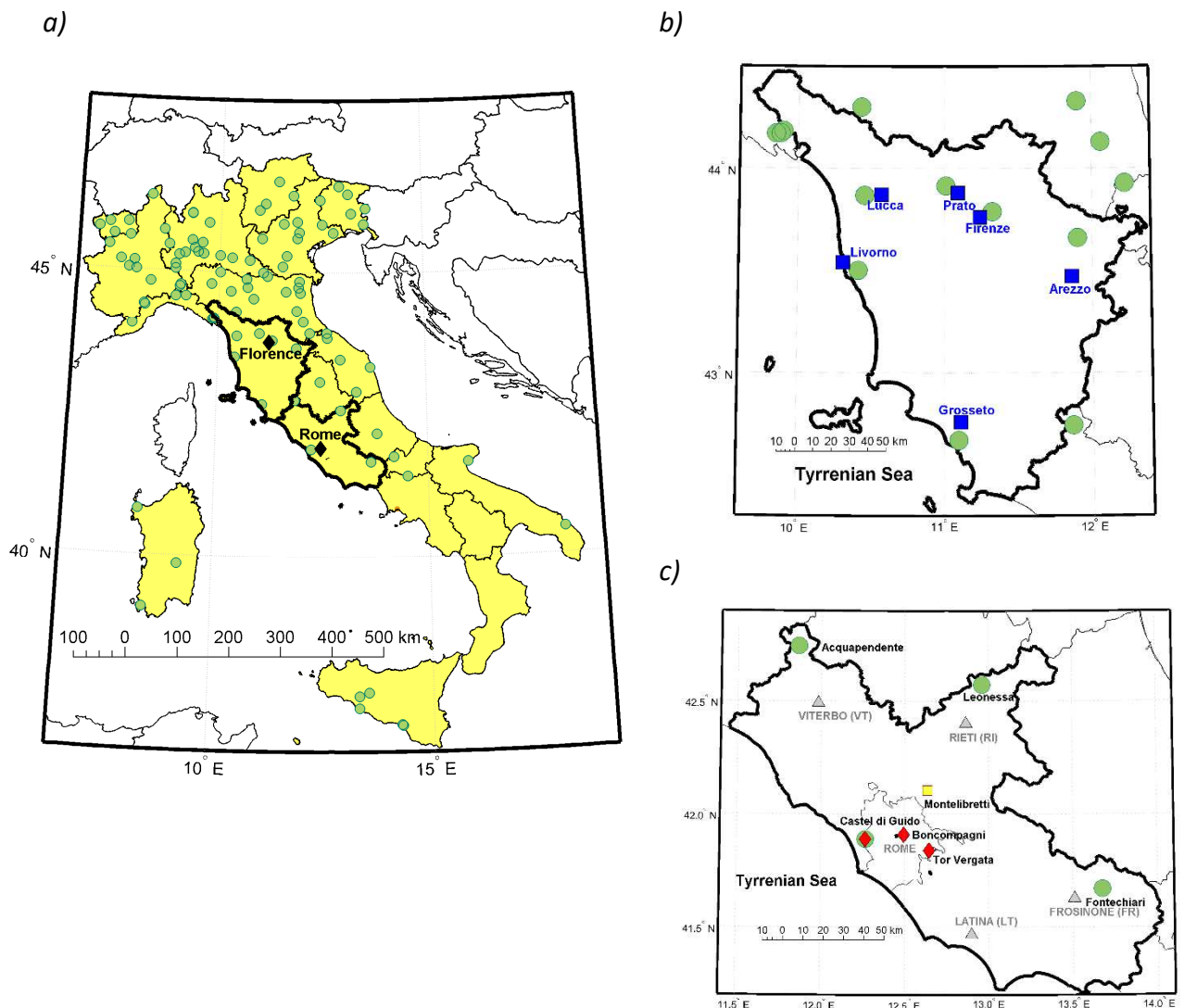
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114 **2. Material and Methods**

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116 **2.1 The study area**

117 Standing in the middle of the Mediterranean Sea, in-between continental Europe and North Africa,
118 Italy is an ideal place where to investigate the transport of Saharan dust to Europe and related
119 impacts (e.g., Gobbi et al., 2004; Gobbi et al., 2006; Matassoni et al., 2009; Mallone et al., 2011;
120 Nava et al., 2012; Marconi et al., 2014; Mona et al., 2014; Stafoggia et al., 2016). Figure 1a shows
121 the country coordinates and location of the official PM10 monitoring sites classified as ‘rural
122 background’ (green bullets) within the Italian air quality network.
123



124 **Figure 1:** a) The Italian territory (yellow), its regional partitioning (thin black lines) and the PM10 monitoring sites
125 classified as ‘rural background’ stations active in the last years (green bullets). The regions of Tuscany and Latium in
126 Central Italy are highlighted (thick black lines), as most of the data presented in this work refer to these two regions.
127 These are zoomed in panels b) and c), respectively. In addition to the relevant ‘rural background’ stations (green bullets),
128 these panels also show location and naming of the sites (coloured symbols) where most of the datasets used in this
129 work were collected. In panel c), boundaries of the Rome Metropolitan area (thin grey line) and location/names of the
130 other 4 main cities (provincial capitals) of the Latium region (grey triangles) are also reported.
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133 A special attention is given in this study to the central Italian regions, and namely to Tuscany
134 (Figure 1b) and Latium (Figure 1c), where most of the original PM10 datasets were collected in the
framework of the PATOS (‘Particolato Atmosferico in Toscana’; Nava et al., 2012; 2015) and

135 DIAPASON (Gobbi et al., 2017) projects, respectively. Figure 1b illustrates location and naming of
136 the PATOS sites (blue squares) and their position with respect to the rural background stations
137 (green bullets) in the area. Figure 1c shows the Latium region, chosen as DIAPASON Pilot Scale and
138 'test area' for both a long-term (4-year) application of the current EC-Methodology (see Sections
139 2.2 and 3.1) and for the test/validation of 'upgrades' to that methodology (Section 2.4). In particular,
140 Figure 1c shows location and naming of the Latium region sites where most of the original
141 observations used in this study were performed: the three DIAPASON monitoring sites in the Rome
142 area (Rome-Tor Vergata, Rome-Boncompagni, and Rome-Castel di Guido, red diamonds) and the
143 EMEP (European Monitoring and Evaluation Programme) station of Montelibretti (yellow square).
144 As in the previous panels, rural background stations are represented by green bullets, among which
145 the DIAPASON site of Rome-Castel di Guido. In the region, routine PM10 measurements are
146 performed by the regional environmental agency (ARPA-Lazio) through a network of 37 air quality
147 monitoring sites (not shown), 20 of which in the Rome metropolitan area and 17 distributed in the
148 areas of the other four main cities of the region (grey triangles in Figure 1c).
149

150 2.2 The EC-Methodology and its operational application

151 We summarize in this section those elements of the EC-Methodology useful for the discussion, its
152 thorough description being available in the specific report released by the European Commission in
153 2011 (EC, 2011). Some more technical details are also provided in Appendix A, in which we describe
154 the specific tool developed within DIAPASON to facilitate its operational implementation over the
155 Pilot Scale (i.e., the Latium region, Figure 1c).

156 The EC-Methodology is based on two steps:

- 157 1) Identification of the dates affected by desert-dust transport;
- 158 2) Quantification of the net desert-dust load on the daily PM10 record.

159
160 1) The identification of the days affected by desert-dust transport requires: a) the daily collection
161 of a series of support information (e.g., model forecasts, back-trajectories, satellite images), mostly
162 available via web, and b) the visual evaluation of these datasets/maps by an operator in order to
163 classify each single date as dust-affected or dust-free. Overall, main weak points of this first
164 step are: the non-negligible efforts required to collect all this material; the high degree of subjectivity in
165 evaluating the presence/absence of desert dust based on it; and the occurrence of cases in which
166 such information provides contradictory indications (e. g., a back-trajectory suggesting a Saharan
167 origin, but no desert-dust predicted by the models). In our operational implementation of the EC-
168 Methodology, this last aspect led to a number of cases classified as 'doubtful-dust' (Appendix A).
169

170 2) The quantification phase of the EC-Methodology requires a continuous record of daily PM10
171 data at a single, selected Regional Background (RB) site within the region under investigation, this
172 acting as 'reference' for the entire region. For those days identified as 'dust-affected' at step 1), the
173 net desert-dust contribution ($PM10_{net_dust}$) to the daily PM10 value recorded at the RB site ($PM10_{RB}$)
174 is estimated as a difference ($\Delta PM10$) with respect to the PM10 values measured at that site within
175 an 'out-of-dust' reference period, i.e.,:

$$176 \quad PM10_{net_dust} = \Delta PM10 = PM10_{RB} - PM10_{outof dust_RB} \quad (1)$$

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180 In the work by Escudero et al. (2007), on which the EC-Methodology is mostly based, the out-of-
181 dust PM10 background ($PM10_{outof dust_RB}$) is computed as the 30th-percentile PM10 value over a 30-
182 days-moving period, ranging from the 15 days before to the 15 days following any identified 'dust

183 day' (and once any other dust-affected date within this time-window is removed). In the same work,
184 the effect of changing this percentile between the 10th and the 50th is investigated showing, as
185 expected from Eq. 1, that the estimated dust load ($PM_{10_{net_dust}}$) decreases as this percentile
186 increases. In the EC-Methodology, the use of the 40th percentile is suggested as more conservative.
187 This has been used for example by Pey et al. (2013) to estimate the dust load over Mediterranean
188 countries in the period 2001-2011.

189 In order to follow a conservative approach, we chose to use the 50th percentile in applying the EC-
190 Methodology, this leading to the lowest estimated desert-dust impact among the tested options
191 (see further details in Appendix A).

192 Overall, the EC-Methodology has the great advantage of being simple and quantitative and,
193 although with the limitations described in this work, both these aspects are worth to be maintained
194 in any revision/upgrade/advancement of this (or similar) tools.

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197 **2.3 EC-Methodology drawbacks and proposed solutions within a 'revised-Methodology'**

198 In addition to the limitations mentioned above, in this section we focus on two further, important
199 drawbacks identified when applying the EC-Methodology to the Pilot area, and mainly related to
200 the 'scientific' assumptions behind it. These are:

201

202 1) The choice of a 30-day-long period over which to compute the out-of-dust 'reference' PM_{10}
203 ($PM_{10_{outofdust_RB}}$ in Eq. 1);

204 2) The need to have (or choose) a Regional Background site for this computation.

205

206 1) In our application to Italy, the time-window length (± 15 days preceding and following the dust
207 date) indicated in the EC-Methodology to evaluate the out-of-dust background PM_{10} was found to
208 be problematic for three main reasons:

209 a) There are important meteorological variations occurring over a 30-days period to consider this
210 as an effective 'background period'.

211 b) The auto-correlation function of PM_{10} series at 'background sites' becomes low (< 0.5) after
212 1-to-3 days of time lag.

213 c) By estimating the desert-dust load as a difference (ΔPM_{10}) between 'in-dust' and 'out-of-dust'
214 PM_{10} conditions, Eq. 1 implicitly assumes this ΔPM_{10} is (or should be) zero in desert-dust-free
215 periods. We tested this 'null hypothesis' and found that it is not. By testing different time-window
216 lengths, we then found that the shorter the time-window, the minimum this difference keeps, with
217 the closest-to-zero values obtained for a time window of ± 3 days.

218 The detailed supporting evidences of each of these points are provided in Appendix B. Overall,
219 analysis of each of the three points listed above indicated that the closer the background period to
220 the desert-dust event, the better it represents the 'most likely' PM_{10} background level over which
221 the desert-dust event builds on. A symmetrical, out-of-dust background period of ± 3 days is found
222 to be particularly suitable for the purpose, having the further advantage of avoiding any possible
223 weekly cycle in the PM_{10} record. This ± 3 days- time-window option is therefore implemented within
224 the proposed revised-Methodology.

225

226 2) An important disadvantage of the current EC-Methodology application to Italy is the need to
227 define/select a Regional Background (RB) site as a 'clean' reference location to quantify the desert-
228 dust contribution. This choice is challenging for a number of reasons:

229 a) A site with the required characteristics does not always exist in many Italian regions (Figure 1).

230 b) In some regions [more than one site could be classified](#) as RB (Figure 1): how to choose the
231 most suitable one among [the possible options](#)?

232 c) Saharan dust plumes may exhibit a not-negligible variability over horizontal scales of the order
233 of 20-30 km (e.g., Schepanski et al, 2015; Yuval et al., 2015). To which extent can a value of Saharan
234 dust load derived [in the selected RB site](#) be representative of the Saharan dust load affecting larger
235 areas? RB sites are usually far from urban areas where PM10 exceedances are more critical, how
236 much is the desert-dust quantification at the RB site representative of the dust loads in a city area?

237 d) Saharan dust plumes typically exhibit a not-negligible variability over the vertical scale (e.g.,
238 Gobbi et al., 2004; Kishcha et al., 2005, Mona et al., 2014; Gkikas et al., 2016). As Regional
239 Background sites are often elevated sites, this may introduce a further element of potential
240 difference between what is observed at the RB site and what is observed at lower altitudes.

241 Examples to substantiate these points based on the DIAPASON and PATOS projects datasets are
242 provided in APPENDIX C.

243 Overall, our results indicate that the need for, selection of, and use of a Regional Background site
244 as required by the EC-Methodology is too constraining, particularly considering that evaluation of
245 the dust load is mostly important at urban sites. We then propose to overcome this limitation by
246 using each single site as a 'background reference' for its own PM10 record ('self-calibration'
247 approach). This means that Eq. 1 is applied at each site on the basis of its own PM10 record, thus
248 deriving site-dependent $PM_{10_{net_dust}}$ values. Besides solving the problems/ambiguities as of point 2
249 above, this 'self calibration' approach has the major advantage of producing a regional dust load
250 field (map) rather than a single dust load value to be considered [valid over](#) a wide region, as in the
251 EC-Methodology.

252 [These suggested modifications \(use of a background period length of \$\pm 3\$ days and of the 'self-](#)
253 [calibration approach'\)](#) represent a main revision of the EC-Methodology embedded within the
254 [proposed revised-Methodology](#).

255 An example of the type of outcome derived applying [the revised-Methodology](#) approach over the
256 Latium region is shown in Figure 2. In particular, Figure 2a and 2b show the estimated desert-dust
257 contribution to the daily-average PM10 as obtained for two days within a Saharan advection in May
258 2014. To insert the results into a more complete picture, in Figure 2 we also include the
259 corresponding view obtained in the same dates from a modelling approach (mean desert-dust-
260 PM10 predicted over Italy by the Chimere model (Menut et al., 2013), Figure 2c, 2d) and from a
261 satellite perspective (AQUA satellite true-colour image for 22 May 2014, Figure 2e). This example
262 depicts the effectiveness of the 'self-calibration' approach at capturing the spatial variability of the
263 dust plume contribution, [which is typically observed over our regions](#). Figure 2a highlights a marked
264 northwest-to-southeast gradient, both predicted by the model and captured by the satellite [image](#),
265 with estimated desert-dust loads above 20-25 $\mu\text{g}/\text{m}^3$ on the north-western part of the Latium region
266 and values lower than 10-15 $\mu\text{g}/\text{m}^3$ on the South-eastern side (a magenta line has been
267 superimposed to [Figure 2a](#) to better highlight this discontinuity). In a similar way, the outcome of
268 the 'self-calibration' procedure on May 24 (Figure 2b) captures the southward retirement of the
269 dust plume predicted by the model (Figure 2d). It is also worth noticing how, in this specific case,
270 the position of the RB Fontechiari site (black symbol) makes it unsuitable to represent the desert-
271 dust conditions over the Rome city (thin [grey line](#)). [This aspect will be further addressed throughout](#)
272 [the text](#).

273