

Winter habitat predictions of a key Southern Ocean predator, the Antarctic fur seal (*Arctocephalus gazella*)

Benjamin Arthur^{* a,b}, Mark Hindell^{a,b}, Marthan Bester^c, P.J. Nico De Bruyn^c, Phil
Trathan^d, Michael Goebel^e and Mary-Anne Lea^a

^a Institute for Marine and Antarctic Studies, University of Tasmania, 20 Castray Esplanade, Battery
Point, Hobart, TAS 7004, Australia.

^b Antarctic Climate and Ecosystems Cooperative Research Centre, University of Tasmania, 20
Castray Esplanade, Battery Point, Hobart, TAS 7004, Australia.

^c Mammal Research Institute, Department of Zoology and Entomology, University of Pretoria,
Private Bag X20, Hatfield, Pretoria, 0028, South Africa.

^d British Antarctic Survey, Cambridge, High Cross, Madingley Road, Cambridge, CB3 0ET, United
Kingdom.

^e Antarctic Ecosystem Research Division, NOAA-NMFS-SWFSC, 8901 La Jolla Shores Drive, La
Jolla, California 92037-1508, USA.

*Corresponding author: Benjamin.Arthur@utas.edu.au, Institute for Marine and Antarctic Studies,
Private Bag 129, Hobart, TAS, 7001, +61 424284321.

Email addresses: Benjamin.Arthur@utas.edu.au (B. Arthur), Mark.Hindell@utas.edu.au (M.
Hindell), mnbest@zoology.up.ac.za (M. Bester), pjndebruyn@zoology.up.ac.za (N. De Bruyn),
pnt@bas.ac.uk (P. Trathan), mike.goebel@noaa.gov (M. Goebel), MaryAnne.Lea@utas.edu.au
(MA. Lea).

25 **Abstract**

26 Quantification of the physical and biological environmental factors that influence the spatial
27 distribution of higher trophic species is central to inform management and develop ecosystem
28 models, particularly in light of ocean changes. We used tracking data from 184 female Antarctic fur
29 seals (*Arctocephalus gazella*) to develop habitat models for three breeding colonies for the poorly
30 studied Southern Ocean winter period. Models were used to identify and predict the broadly
31 important winter foraging habitat and to elucidate the environmental factors influencing these areas.
32 Model predictions closely matched observations and several core areas of foraging habitat were
33 identified for each colony, with notable areas of inter-colony overlap suggesting shared productive
34 foraging grounds. Seals displayed clear choice of foraging habitat, travelling through areas of
35 presumably poorer quality to access habitats that likely offer an energetic advantage in terms of
36 prey intake. The relationships between environmental predictors and foraging habitat varied
37 between colonies, with the principal predictors being wind speed, sea surface temperature,
38 chlorophyll a concentration, bathymetry and distance to the colony. The availability of core
39 foraging areas was not consistent throughout the winter period. The habitat models developed in
40 this study not only reveal the core foraging habitats of Antarctic fur seals from multiple colonies,
41 but can facilitate the hindcasting of historical foraging habitats as well as novel predictions of
42 important habitat for other major colonies currently lacking information of the at-sea distribution of
43 this major Southern Ocean consumer.

44

45 **Key-words:** Foraging behaviour, geographical distribution, habitat model, pinniped, prediction,
46 tracking.

47

48 **1. Introduction**

49 Information on the spatial distribution of marine predators is fundamental to understanding
50 the structure and function of their ecosystems and is ultimately driven by the availability of prey
51 resources that are heterogeneously dispersed in space and time (Russell et al., 1992). The
52 abundance of marine prey is intrinsically linked to physical and biological oceanographic
53 properties, allowing us to relate the distribution and responses (such as breeding success) of higher
54 trophic species with the fundamental bio-physical aspects of their environment (e.g. Friedlaender et
55 al., 2006; Reid and Croxall, 2001). Quantifiable understanding of these factors is necessary to
56 inform and appraise management decisions such as defining marine protected areas (Hyrenbach et
57 al., 2000), fisheries management and by-catch mitigation measures (Burger and Shaffer, 2008), as
58 well as for the development of accurate ecosystem models to assess the effects of future
59 environmental changes.

60 The Antarctic fur seal (*Arctocephalus gazella*, Peters, 1875) is a highly mobile marine
61 predator that inhabits an extremely dynamic environment, the Southern Ocean. Antarctic fur seals
62 are major consumers in the Southern Ocean ecosystem, in particular of krill (Croxall et al., 1985)
63 often competing with other predators such as penguins for this resource (Barlow et al., 2002;
64 Blanchet et al., 2013). Antarctic fur seals breed at 10 major sites in the Southern Ocean across their
65 circumpolar range (Shirihai, 2002), spanning latitudes from the northernmost colony at the Crozet
66 Islands (46°25'S), north of the Polar Front, to the southernmost at the South Shetland Islands
67 (62°27'S), within the zone of winter sea ice. The at-sea habitat use of Antarctic fur seals reflects
68 these geographical differences and the species exhibits a diverse foraging ecology across their
69 range, with highly flexible summer foraging behaviours within and between colonies associated
70 with local environmental conditions (Lea et al., 2006) and differences in prey (Boyd et al., 1994;
71 Lea et al., 2008; Staniland et al., 2010).

72 During the winter, non-breeding animals are free from the constraints of central place
73 foraging (Orians and Pearson, 1979) associated with provisioning their offspring. Consequently,

74 female Antarctic fur seals can make wide-ranging migrations (Boyd et al., 2002) of up to eight
75 months. Moreover, the Southern Ocean in winter is both physically and biologically distinct from
76 the summer, with the growth of sea ice, decline in primary productivity due to decreased irradiance
77 and temperature (Clarke, 1988; Mitchell et al., 1991) and a deeper mixed-layer depth because of
78 strong winds (Sakshaug et al., 1991), being major distinctions. Consequently, foraging animals, and
79 their prey, can be expected to behave differently during this time. Female Antarctic fur seals are
80 also gestating during winter (Boyd, 1996) requiring them to make judicious foraging choices to
81 maximise their energy intake in the pre-breeding period. Recent studies of the winter migrations of
82 female Antarctic fur seals show they utilise a variety of habitats during this time, occupying all
83 inter-frontal zones from pole-ward of the Antarctic Circumpolar Current, including ice associated
84 waters, to north of the sub-Antarctic Front, incorporating both shelf and pelagic habitats (Boyd et
85 al., 2002; Staniland et al., 2012).

86 The variety of foraging habitats utilised by Antarctic fur seals during the winter is reflected
87 in their diet, with differences in the trophic position of their prey between the various inter-frontal
88 zones, both within and between individuals (Walters, 2014). Some female fur seals are also highly
89 faithful to winter foraging grounds returning to the same broad foraging area annually, presumably
90 because of an increased energy acquisition associated with these habitats over the long-term (Arthur
91 et al., 2015). The diversity of habitats used by Antarctic fur seals during the winter suggests that
92 breeding colony location is a key factor in determining the suite of habitat types available for
93 foraging (Mary-Anne Lea, unpublished data). However, despite being a major secondary consumer
94 in the Southern Ocean, the at-sea behaviour of Antarctic fur seals outside the breeding season
95 remains poorly understood. In particular, little is understood about the relationships between marine
96 characteristics and foraging behaviour and how animals from different populations respond to these
97 factors.

Habitat models (or Species Distribution Models) can assist with this process and are often used to describe the environmental drivers of species distribution patterns, providing useful ecological insights (Elith and Leathwick, 2009). Ultimately, they may be used to make predictions of species distributions in un-sampled areas or under changing environmental conditions, and have been employed across a variety of taxa, scales and environments using a range of methodologies (Guisan and Zimmermann, 2000). The fundamental information on the distribution of marine predators that is needed to build such models is often provided by telemetry studies. However, these studies are often restricted to a single site or season. For Antarctic fur seals, Guinet et al. (2001) developed a probabilistic model for the distribution of diving activity of lactating seals at Îles Kerguelen, which predicted where animals should concentrate their foraging based on the oceanographic conditions within that year. The authors note that studies conducted over several years will provide further insights into the effects of oceanographic conditions on the foraging ecology and at-sea distribution of this, and other, marine predator species.

Here, we examine the at-sea distribution and foraging habitats of female Antarctic fur seals from three breeding colonies across multiple inter-breeding periods in the Atlantic and Indian sectors of the Southern Ocean. The study aims to: (1) identify important foraging habitats for Antarctic fur seals during the non-breeding, winter season, (2) describe the environmental factors that characterise these areas and compare these relationships between animals from three major breeding populations and (3) develop predictive models for foraging habitat.

2. Materials and methods

2.1. Study sites and instrumentation

The study was conducted at three Antarctic fur seal breeding colonies: Marion Island (46°54'S, 37°44'E, Prince Edward Islands), Bird Island (54°00'S, 38°03'W, South Georgia) and Cape Shirreff (62°27'S, 60°47'W, South Shetland Islands) (Fig. 1). At Marion Island, the study was

undertaken over five years between 2008 and 2013, at Bird Island for four years between 2008 and 2011 and at Cape Shirreff for three years between 2008 and 2010. Adult females were captured towards the end of lactation between February and April and were instrumented with a global-location sensing (GLS; British Antarctic Survey, Cambridge UK) logger for the duration of their winter migrations. Seals were recaptured and instruments recovered at the start of the following breeding season in November-December when pregnant females return to the colony to pup. Several animals were recaptured in subsequent years. Animal handling, GLS logger architecture, attachment and calibration methods are detailed in Arthur et al. (2015) provided in Supporting Information S1.

132

2.2. Tracking datasets

Locations were produced from the raw light and temperature data from GLS loggers following the Bayesian approach of Sumner et al. (2009) using the R software (R Core Team, 2014) package ‘tripEstimation’ (Sumner and Wotherspoon, 2010). In summary, two location estimates per day (dawn and dusk) were produced from the posterior mean for each twilight period that were summarised from the accepted Markov Chain Monte Carlo (MCMC) samples. Full details of geo-location model design and implementation are presented in Supporting Information S2. Seals undertook between 1-9 foraging trips per winter with the average \pm SD being 2.2 ± 1.5 at Marion Island, 2.5 ± 1.6 at Bird Island, while all animals at Cape Shirreff undertook a single trip. For animals making multiple foraging excursions from their colony, tracks were split into individual trips and analysed independently. Individual trips were identified in the raw light data, with haul-outs typified by distinctly messy light curves resulting from the animal periodically shading the light sensor while on land. Winter foraging trips encompassed the first post-weaning excursion (typified by a marked increase in duration from short trips during lactation) to the animal’s return to the colony the following breeding season. Between 2008-13, 184 GLS tags were recovered from

148 post winter migrations across the three colonies (Table 1). Unprocessed GLS data are publicly
149 available from the Australian Antarctic Data Centre (<http://data.aad.gov.au>) for each site: Marion
150 Island (Lea et al., 2014a), Bird Island (Lea et al., 2014c) and Cape Shirreff (Lea et al., 2014b).

151

152 2.3. *Habitat models*

153 Models were constructed to explain the spatial distribution of Antarctic fur seal habitat use
154 during the winter. The mean time spent in each cell (total time spent divided by the number of seals
155 visiting each cell) of a 60 km x 60 km grid consistent across the spatial extent of locations (Table 1)
156 was calculated for the period of study to quantify habitat use, hereto referred to as *time spent*. A
157 grid of this resolution was chosen to match the error uncertainty surrounding geo-location
158 estimates, which is shown to be 70 ± 35 km for an Antarctic fur seal carrying GLS and Argos tags
159 simultaneously (Mary-Anne Lea, unpublished data). Time spent is a proxy for foraging effort as
160 animals are likely to spend more time in an area which they are actively exploiting than when
161 travelling between foraging areas (Barraquand and Benhamou, 2008; Kareiva and Odell, 1987).
162 Time spent was a continuum from low to high use and can be considered a “usage” approach, being
163 similar to kernel density analysis often applied to tracking data, rather than as a binary presence-
164 absence response contrasting areas where animals did go with areas that they didn’t go. Three
165 winter habitat models were generated: one for each colony with data pooled across all available
166 years. Prior to developing these models, an assessment of the adequacy of the sample size at each
167 colony was undertaken. We assessed the amount of new information (i.e. grid cells) arising from
168 the inclusion of each additional individual seal (averaged over 100 permutations), providing an
169 estimate of the minimum number of individuals needed to adequately represent the spatial
170 distribution patterns of animals from each colony.

171

172 2.3.1. *Environmental parameters*

173 A suite of environmental variables that potentially influenced time spent was included in
174 models to characterise fur seal habitat. Variables were chosen for a priori reasons based on our
175 understanding of the nature of the variables and how they relate to the biology of the seals.
176 Variables included static parameters: bathymetry (BATHY) and distance to colony (d2col) and
177 dynamic parameters: sea surface temperature (SST), sea surface height anomaly (SSHa),
178 chlorophyll a concentration (CHLa), wind speed, surface current magnitude (CURR) and eddy
179 kinetic energy (EKE). Variability of sea surface height anomaly (SSHV) and the gradient of sea
180 surface temperature (SSTG) were also included (Fig. 1). The source, spatial resolution and
181 oceanographic significance of environmental variables are provided in Table A1. Environmental
182 data were extracted for each pixel of the spatial domain at weekly intervals spanning the temporal
183 range of location data at each colony. The grid based approach aggregated tracking data over
184 multiple years, so weekly maps were averaged to produce one mean parameter value per cell for the
185 period of study (in the case of SSHV variance was calculated) to create a temporal climatology
186 (Sumner et al., 2003). These climatologies allow investigation of the influence of environmental
187 factors on seal habitat use across broad spatial and temporal scales. All variables were re-
188 interpolated across a 60 km x 60 km grid to match the time spent response data. All data, including
189 environmental predictors and time-spent response, were then re-projected to Lambert azimuthal
190 equal-area projection. Environmental data were available from the Australian Antarctic Data Centre
191 and extracted using the R package 'raadtools' (Sumner, 2015).

192

193 2.3.2. Model design and predictions

194 Generalized additive models (GAMs) were fitted to the relationship between time spent and
195 environmental predictors. To determine the most appropriate error structure, a comparison was
196 made between Gaussian models with an identity link, log-transformed Gaussian with identity link,
197 and Gamma with a log link models. Log-likelihood and Akaike's Information Criterion (AIC)

198 scores adjusted to account for transformation were used for model comparison and to determine the
199 most appropriate error structure. The distribution of environmental predictors was examined and
200 data were log-transformed where appropriate to meet the assumptions of normality. For numerical
201 stability, predictors were scaled and centred to account for the considerably different scales of
202 measurement. Highly correlated predictor variables (Pearson's $r > 0.9$) were excluded from the
203 models. Model selection was undertaken using the maximum-likelihood approach to minimise the
204 AIC. Models including all combination of variables were compared and ranked by their Akaike
205 weight (w_{AIC}) to represent the relative likelihood of each model.

206 To account for individual variability in the response term, it is possible to include a random
207 term in the GAM framework (Wood, 2006). However, such models are computationally demanding
208 and potentially problematic for smaller relative sample sizes (Raymond et al., 2014), so standard
209 GAMs were utilised. The influence of individual variability was instead reduced by using the
210 average value of time spent across individuals in each cell. A further problem arises with tracking
211 data that are often spatially auto-correlated, which can lead to violations of the assumption of
212 independence of residuals. We therefore included a spatial autocorrelation structure in all models
213 (Dormann et al., 2007).

214 Model performance was evaluated by assessing model fit and predictive performance.
215 Model fit was indicated by the percent deviance explained and by checking model residuals. The
216 predictive performance of models was assessed by calculating the root mean-squared error (RMSE)
217 using a k-fold cross-validation procedure. Grid cells were randomly assigned to one of 10 folds
218 where models were trained on nine folds and tested on the remaining one, with each fold withheld
219 in turn. The RMSE (expressed in the same units as the response) was aggregated across the 10 sets
220 of results. The best model for each population was then fit on the unscaled and uncentred
221 environmental predictor variables with the sole purpose of aiding the interpretability of the
222 smoothed relationships on meaningful scales. Lastly, validated models were used to predict winter

223 habitat use of fur seals by interpolating across the entire spatial domain of the locations observed
224 for each colony. All analyses were conducted in R 3.2.0 (R Development Core Team).

225

226 **3. Results**

227 *3.1. Distribution of time spent in area*

228 Between 2008-13, a total of 320 foraging trips and 83,796 location estimates were observed
229 for 184 female Antarctic fur seals during the austral winter. At Marion Island, 54,051 locations
230 from 227 trips were available for 119 female seals with 12,328 locations from 56 trips and 28 seals
231 at Bird Island, and 17,417 locations across 37 trips collected for 37 animals from Cape Shirreff
232 (Table 1, Fig. A1). For Marion Island, the cumulative information curve showed that the number of
233 newly visited grid cells arising from the inclusion of each additional animal asymptotes at
234 approximately 50 individuals (Fig. 2c), indicating we had an adequate sample to accurately
235 represent the spatial use patterns of the population. The curve for Bird Island closely matched that
236 of Marion Island, however it failed to level out completely, suggesting that additional animals
237 would better represent the distribution patterns of this population. Similarly, at Cape Shirreff, 37
238 individuals were observed and although the curve is beginning to asymptote, it failed to level out
239 entirely (Fig. 2c), indicating a greater number of animals are needed to more fully characterise the
240 variability in habitat use.

241 The distribution of time spent for Marion Island fur seals indicated that animals utilised a
242 diversity of areas during their winter migrations (Fig. 3). Of prominent use were areas located
243 approximately 100-800 km to the north and east of Marion Island associated with the Del Cano
244 Rise, and pelagic waters to the west of the island between 20-30°E. Time spent values were also
245 high in several areas to the south of the colony at approximately 55°S as well as other locations at
246 the extreme east and west of the population's range, notably around Bouvet Island and east of Iles
247 Kerguelen. Seals spent relatively little time in local waters within several hundred kilometres to the

248 west and south of Marion Island, suggesting animals transited through these areas to reach distant
249 foraging grounds.

250 For the Bird Island population, time spent was concentrated mostly in local waters within
251 approximately 300 km of South Georgia, particularly to the northwest of the colony (Fig. 3). There
252 was a further area of high usage to the south of South Georgia. Additional high-use areas were off
253 the Patagonian coast of South America, east of South Georgia towards the Scotia Arc and South
254 Sandwich Islands and on the continental shelf along the western Antarctic Peninsula.

255 At Cape Shirreff, the areas of high usage were along the southern coast of Chile associated
256 with the shelf-break and pelagic waters further west of this region (Fig. 3), and waters to the west of
257 South Georgia. Time spent values were high along the Patagonian shelf-break and several distant
258 pelagic areas at the western extent of the population's range (westward of 110°W longitude)
259 resulting from individual animals concentrating their efforts in these regions for extended periods.

260

261 3.2. Environmental characteristics of high-use areas

262 Across the spatial distribution of all three populations, EKE and CURR oceanographic
263 variables were highly correlated ($r > 0.95$). CURR was subsequently excluded from analyses,
264 leaving nine predictor variables available for model build. For the Marion Island population, the
265 best model explaining mean time spent in a grid cell included all variables but SSTG ($wAIC =$
266 0.691 ; Table 2) fit to a Gamma error structure. Model residuals were normally distributed and the
267 model explained 73.3% of the deviance in the data and had good predictive performance ($r^2 =$
268 0.704 , $RMSE = 14.30$). Omitting the spatial autocorrelation term from the model still produced
269 good model fit ($r^2 = 0.413$, $RMSE = 19.93$), suggesting the broad relationships between time spent
270 and environmental variables were robust. The strongest relationships between foraging effort and
271 the seals' environment were observed with Wind, SST, SSHV and d2col. Antarctic fur seals from
272 Marion Island spent more time, on average, in areas of higher wind speeds ($>12 \text{ m.s}^{-1}$) with greater

SSHV. Seals were found in water temperatures that were either colder ($\sim 0^{\circ}\text{C}$) or warmer (between 6°C and 10°C) than average and areas that were further ($>1500\text{ km}$) from the colony (Fig. 4a).

The best model for the Bird Island population was a reduced Gamma model excluding CHLa and SSHV ($w\text{AIC} = 0.355$; Table 2). Model assessment suggested a good fit to the observed data with the model explaining 85.3% of the deviance in the data and having good predictive performance ($r^2 = 0.828$, $\text{RMSE} = 13.24$). Model performance was good when the spatial autocorrelation term was excluded ($r^2 = 0.591$, $\text{RMSE} = 19.76$). The clearest relationships with time spent were with Wind, BATHY and d2col. These indicated that cells close to the colony ($<500\text{ km}$), with shallow relative water depths ($<2000\text{ m}$) and with lower wind speeds ($<10\text{ m.s}^{-1}$) had high mean time spent values (Fig. 4b).

At Cape Shirreff, the best model explaining time spent was a Gamma model excluding SSTG and BATHY predictor terms ($w\text{AIC} = 0.594$; Table 2). Model residuals and cross validation indicated the model was a good fit to the observations, explaining 72.5% of the deviance in the data ($r^2 = 0.701$, $\text{RMSE} = 15.63$). Model performance was good when the spatial autocorrelation structure was omitted ($r^2 = 0.461$, $\text{RMSE} = 20.55$). The strongest relationships between mean time spent and environmental predictors were for d2col, Wind and SST. The smoothed relationships indicated that seals from Cape Shirreff spent more time in areas of colder (between -1°C and 5°C) or warmer ($>10^{\circ}\text{C}$) than average waters and with high relative wind speeds ($>10\text{ m.s}^{-1}$) when closer to the colony (Fig 4c).

292

293 *3.3. Predicting important winter foraging habitat*

The habitat models were interpolated across the entire spatial domain of observations from each population. The resultant distribution maps of winter foraging habitat are shown in Figure 3. For all three populations these predictions closely matched the time-spent observations recorded in this study, giving us further confidence in the ability of the models to make realistic predictions of

important foraging habitat for Antarctic fur seals in the different oceanic basins. For seals from the Marion Island colony, the model successfully predicted the major areas of observed time spent, being those regions to the east, west and far south of the colony. At Bird Island, likely important foraging habitat was identified in waters local to South Georgia and extending south, as well as on the Patagonian Shelf, north of the Scotia Arc and the West Antarctic Peninsula, closely matching observations. Newly predicated habitat was located north of Tierra del Fuego (southern tip of South America) and at the extreme eastern edge of the range of tracked animals, however, we note this is driven by observations from a single animal only. The Cape Shirreff model predictions also closely matched the observations, with the model identifying the three major focal areas for animals from this population: the southern Chilean coast, the Patagonian Shelf break and around South Georgia. Notably, additional important foraging habitats were predicted for coastal waters on the Patagonian Shelf and oceanic waters around 100°W longitude at the northern extent of the population's range, which were areas with no previous observations.

Predicted important foraging habitats were not wholly distinct between the three populations, with clear overlap of some areas (Fig. 3). Seals from Cape Shirreff and Bird Island in particular, have considerable overlap in their predicted use of habitats around South Georgia, along the Patagonian Shelf and, to a lesser degree, waters of the western Antarctic Peninsula. Important foraging habitat around Bouvet Island in the Southern Atlantic sector is also likely to be shared by the Bird and Marion Island populations.

4. Discussion

Our study considers time spent as a proxy for foraging effort in Antarctic fur seals. Residence time is a suitable proxy as an animal is likely to spend more time in an area that it is actively exploiting (area-restricted search) than when travelling between foraging areas (Barraquand and Benhamou, 2008) and several studies have shown high relative residence times to

323 be associated with increased dive effort and food intake in marine predators (Cotté et al., 2007;
324 Thums et al., 2008), including Antarctic fur seals (Arthur et al., 2016). We therefore refer to
325 foraging habitat and effort henceforth.

326 By using tracking data from multiple sites in the Southern Ocean across several years this
327 study has revealed the broadly important foraging habitats, and the environmental conditions that
328 characterise these, for female Antarctic fur seals during the poorly studied winter period. During
329 that time, when animals are free to range widely, there are several habitats that are important for
330 seals. The performance of habitat models was good, with predictions interpolated across the spatial
331 domain of each population closely matching the observed data. The models also identified several
332 novel areas of importance where no animals had been observed, particularly for the Cape Shirreff
333 population. The cumulative information analysis suggested the minimum number of animals needed
334 to adequately characterise the spatial use patterns of this population was not achieved, unlike at
335 Marion Island where little novel habitat was predicted. We are therefore confident in the ability of
336 the habitat models to make realistic predictions of the foraging habitat for this species.

337 Habitat models can perform well in characterizing the distribution of species within their
338 current range and interpolation is generally reliable providing data and model design are reasonable
339 (Elith and Leathwick, 2009). Habitat modelling has been used to quantify species-environment
340 relationships and predict the distributions of a variety of taxa including terrestrial and aquatic plant
341 species, terrestrial animal species, fish, plant communities, vegetation types and biodiversity (For
342 reviews see Guisan and Thuiller, 2005; Guisan and Zimmermann, 2000). More recently, habitat
343 models have been applied to marine species, including highly mobile top predators in an effort to
344 identify critical oceanic habitats (e.g. Block et al., 2011) including for the Southern Ocean (Hindell
345 et al., 2011; Raymond et al., 2014).

346 For the three Antarctic fur seal populations in this study, distinct foraging areas were
347 identified for the winter period. At Marion Island, core foraging areas were contiguous with the Del

348 Cano Rise, a prominent bathymetric feature associated with the development of eddies and
349 spring/summer phytoplankton blooms (Pollard et al., 2007; Venables et al., 2007) known to be
350 utilised by foraging predators from Marion Island (de Bruyn et al., 2009). Further core habitat was
351 to the west of the colony and south towards the Polar Front, a region of focus for several apex
352 predator species because of the predictable distribution of prey such as mesopelagic fish (Bost et al.,
353 2009). At Bird Island, core habitat was located in the productive waters around South Georgia and
354 downstream, where zooplankton biomass is approximately four to five times higher than in other
355 typical Southern Ocean areas (Atkinson et al., 2001), as well as on the Patagonian Shelf which is an
356 important winter foraging region for other predators from South Georgia such as white-chinned
357 petrels (*Procellaria aequinoctialis*, Linnaeus, 1758) (Phillips et al., 2006). These results broadly
358 match those of the only other studies of the winter movements of Antarctic fur seals from South
359 Georgia (Boyd et al., 2002; Staniland et al., 2012). For Antarctic fur seals from the South Shetland
360 Islands, important habitat was located off the Chilean coast proximate to the high-primary
361 productivity, cold Humboldt Current system and associated upwelling (Daneri et al., 2000) as well
362 as the Patagonian Shelf break and around South Georgia.

363 The use of core foraging areas is ultimately driven by prey availability. Unfortunately, direct
364 measurements of prey distribution have poor spatial and temporal coverage in the Southern Ocean
365 and regional-scale models therefore rely on environmental proxies that characterise ocean processes
366 related to prey distribution (Bost et al., 2009). Nonetheless, investigation of proximate drivers can
367 aid in understanding the bio-physical properties of habitats. The environmental parameters in this
368 study provide indices, effectively summarising the environment across years. Consequently, this
369 limits the inferences that can be made, precluding the investigation of fine-scale spatial or temporal
370 regional relationships. Any relationships that are identified, however, are likely to be broad and
371 generally robust. Although the relationships between foraging effort and environmental parameters

372 differed between populations, the principal predictors in habitat models were wind speed, sea
373 surface temperature, distance to colony, bathymetry and sea surface height variability.

374 Wind speed contributed strongly to all three models. Antarctic fur seals from Marion Island
375 and Cape Shirreff foraged more in windier areas, while seals from Bird Island foraged in areas with
376 low to moderate wind speeds, as was observed for female northern fur seals (*Callorhinus ursinus*,
377 Linnaeus, 1758) (Sterling et al., 2014), a northern hemisphere analogue for Antarctic fur seals.
378 Wind strength and associated winter storms can impact the dispersal routes of predators (Lea et al.,
379 2009) and the vertical distribution of biomass, with prey driven deeper by the increased mixing and
380 turbulence resulting from higher wind stress (Incze et al., 2001). Seals from Bird Island, which feed
381 largely on lower trophic level prey such as krill (Reid and Arnould, 1996), foraged more in areas of
382 reduced wind speeds where prey fields may be higher in the water column and therefore more
383 accessible. Conversely, at Marion Island and Cape Shirreff, seals preferred windier areas. Marion
384 Island animals, which feed mainly on mesopelagic fish and squid in winter (Walters, 2014) also
385 foraged more in areas with higher eddy kinetic energy. Wind is important to the vertical distribution
386 of biomass in mesoscale eddies, with zooplankton distribution typically deeper under high wind
387 events (Mackas et al., 2005). Despite high winds impacting the vertical distribution of biomass the
388 aggregation of prey in eddy features such as those along the South West Indian Ridge around
389 Marion Island (Ansorge and Lutjeharms, 2005) makes them important areas to foraging predators
390 (Nel et al., 2001).

391 Habitat accessibility was also an important determinant of foraging habitat, with animals
392 typically foraging less in areas that were relatively distant from breeding colonies, likely because of
393 the energetic costs associated with travel. The exception was at Marion Island, where animals
394 foraged more in distant areas. Sea surface temperature contributed highly to habitat models for
395 Marion Island and Cape Shirreff, with animals decreasing their foraging effort in areas of average
396 temperatures and preferring relatively cold or warmer waters. Although the use of water masses

397 with certain temperatures by foraging predators can indicate preferences for productive oceanic
398 features such as fronts (e.g. King penguins and the Polar Front; Péron et al., 2012), temperature is
399 inherently coupled with latitude in the Southern Ocean and may simply be a product of how far
400 north or south seals travelled during their wide-ranging migrations. Seals from Marion Island
401 foraged more in areas of elevated average chlorophyll-a concentration, supporting observations for
402 this species during the summer season at Kerguelen Island (Guinet et al., 2001). The habitat model
403 for Bird Island showed a strong positive relationship with bathymetry, indicating seals preferred to
404 forage in the shallower waters (<2000 m) of South Georgia and the Patagonian Shelf (Fig 1) where
405 they feed on neritic prey (Walters, 2014).

406 The availability of important foraging habitats to Antarctic fur seals varies throughout the
407 winter and usage of the major habitats identified here will not be consistent throughout the winter.
408 If the non-breeding winter season is divided into three periods: early winter (March-May), mid-
409 winter (June-August) and late winter (September-December), there are obvious differences in
410 habitat availability. At Marion Island, seals had a reduced longitudinal range in early winter relative
411 to mid or late winter (Fig. 2a) due to their recent departure from the colony. Seals utilised more
412 southerly regions in early and mid-winter, shifting their distribution north in late winter (Fig. 2b),
413 where foraging habitats east and west of the colony are likely to be exploited. At Bird Island, there
414 were no significant variations in latitudinal range across the season (Fig. 2b). Fur seals from Cape
415 Shirreff displayed a notable latitudinal shift in distribution across the winter, utilising areas between
416 60-65°S in early winter and more northerly habitats around 45-55°S as winter progressed (Fig. 2b).

417 The accessibility and use of foraging habitat can vary in response to factors including
418 environmental conditions, prey availability, competition, predation risk, breeding status and age
419 (Field et al., 2005; Heithaus and Dill, 2006; Nakano, 1995; Nordstrom et al., 2013; Weimerskirch et
420 al., 1993). Of particular importance, is the reproductive cycle, with the early post-breeding period a
421 critical time for recovering body condition after the extended lactation of otariid seals, which is

422 energetically costly (Pitcher et al., 1998). Although animals are released from the constraints of
423 parental care and free to travel farther during this time (Lowther et al., 2014), predictable and
424 profitable habitat in close proximity to breeding colonies, such as those to the east of Marion Island,
425 around South Georgia and the Antarctic Peninsula, will be critical for fast energy acquisition. For
426 marine predators lacking specialised adaptations, the growth of winter sea ice can represent a
427 barrier excluding them from an area (Ainley et al., 2003). Although some Antarctic fur seal females
428 are known to utilise sea-ice habitats during winter (Mary-Anne Lea, unpublished data), they are not
429 considered an ice-obligate species and are mostly absent from areas of significant ice cover.
430 Consequently, the availability of southerly habitats to fur seals, particularly around the Western
431 Antarctic Peninsula and south of South Georgia, where krill is a significant dietary component
432 (Walters, 2014), is restricted to the early winter prior to the growth of sea ice (Fig. 2a). As ice cover
433 excludes seals from southern regions in mid to late winter, more northerly habitats are increasingly
434 utilised, especially for animals from Cape Shirreff where areas off the Chilean coast and around
435 South Georgia and the Patagonian Shelf are important. We suggest that habitats close to breeding
436 colonies and those that will be covered by winter sea ice are critical immediately after seals depart
437 the colony in April-May, whereas regions farther north will be increasingly utilised throughout mid
438 to late winter during which time the availability and quality of food resources can strongly effect
439 the risk of abortion (Soto et al., 2004) and reproductive success the following breeding season
440 (Boyd et al., 1995).

441 The important Antarctic fur seal foraging areas identified in this study were not unique to
442 colonies, with some areas used by seals from multiple populations. Animals from Bird Island and
443 Cape Shirreff in particular had considerable overlap of foraging areas, as do the Bird and Marion
444 Island populations although to a lesser degree. Furthermore, there will likely be inter-population
445 overlap with seals from the study populations and other colonies in the Southern Atlantic and Indian
446 Oceans including Crozet, Kerguelen, Heard, Bouvet, the South Sandwich and the South Orkney

447 Islands. Although Antarctic fur seals are generally regarded as philopatric in respect of breeding
448 sites (Lunn and Boyd, 1991), tracking studies demonstrate the species' capacity for widespread
449 dispersal (e.g. Boyd et al., 2002). It is not unexpected, therefore, that animals from multiple
450 breeding sites will migrate to shared productive areas. Some Antarctic fur seals are highly faithful
451 to winter foraging grounds and return to the same broad area annually (Arthur et al., 2015) and the
452 pattern of inter-colony overlap may, in part, be a product of colony memory of major foraging
453 habitats (Bonadonna et al., 2001).

454

455 **5. Conclusions**

456 When considered together, the areas identified in this study constitute the important foraging
457 habitats that are exploited by a key Southern Ocean predator throughout the poorly studied non-
458 breeding winter period. The broad spatial and temporal approach of this study has produced realistic
459 estimates of the foraging habitat of Antarctic fur seals from three populations in the Southern
460 Atlantic and Indian Oceans. Seals display clear choice of foraging habitat, travelling through
461 regions of seemingly poorer quality habitat to access areas with probable elevated prey availability.
462 Such areas can be several thousand kilometres from breeding colonies and consequently the seals
463 balance energy intake with the costs of travel and prey searching (Charnov, 1976).

464 Appreciation of the temporal shifts in availability and use of foraging habitats during this
465 ~9-month period is important not only biologically, but also from a management context. Antarctic
466 fur seals are currently the only pinniped indicator species contributing to the Commission for the
467 Conservation of Antarctic Marine Living Resources' (CCAMLR) Ecosystem Monitoring Program
468 (CEMP), which aims to manage the ecological impacts of commercial harvests in the Southern
469 Ocean. An increased understanding of the habitat use of this species is therefore critical to inform
470 conservation management and will facilitate future investigation of the potential effects of short (i.e.

471 El Niño Southern Oscillation and Southern Annular Mode) and long-term oceanographic changes
472 (i.e. climate change) on the habitat use and foraging efficiency of this species.

473 Recently, the objective of habitat models has shifted towards documenting habitat change
474 and extrapolating model predictions to novel areas (Elith and Leathwick, 2009). The habitat models
475 developed in this study can be used to hindcast foraging habitat, establishing historical distribution
476 ranges that can be compared to current observations to indicate past habitat changes and improve
477 our understanding of future distribution shifts. The development of seasonal habitat models to better
478 elucidate the temporal variation in habitat importance over the winter will also help quantify when
479 key foraging habitats are used by fur seals and to what degree any potential overlap with human
480 activities such as fishing may occur. For marine predators, habitat models are a useful conservation
481 tool to identify critical habitats of understudied populations without the need to undertake time-
482 consuming and expensive tracking programs. Subsequently, habitat models for the three Antarctic
483 fur seal colonies from this study can be used to predict critical foraging habitat for seals from other
484 key Southern Ocean colonies where information on the winter at-sea distribution is currently not
485 available. Among other considerations, information on local habitat availability and preferences
486 will be necessary for accurate extrapolation (Torres et al., 2015) and careful consideration of the
487 environmental and oceanographic similarities of model and prediction populations will be
488 important.

489

490 **Acknowledgements**

491 We are especially grateful to the Marion Island (2008-13), Bird Island (2008-11) and Cape Shirreff
492 (2008-10) field teams for their efforts in deploying and recovering tags. We acknowledge the
493 logistical support provided by the South African National Antarctic Program (SANAP), the British
494 Antarctic Survey (BAS) and the U.S. Antarctic Marine Living Resources (US AMLR) Program.
495 Thank you to Mike Sumner, Ben Raymond and Simon Wotherspoon for analytical assistance and to

Keith Reid for his advice during analysis and manuscript preparation. This work was funded by Sea World Research and Rescue Foundation Inc. Australia (SWR/6/2013), ANZ Trustees Holsworth Wildlife Research Endowment (L0020491), Australian Research Council and Australian Antarctic Science Grant. Travel support to present this work at the 3rd CLIOTOP Symposium was provided by the Antarctic Climate & Ecosystems Cooperative Research Centre and the Australian Wildlife Society. All animal handling and experimentation were approved by the University of Tasmania Animal Ethics Committee (permit A001134), the University of Pretoria Animal Use and Care Committee (permit AUCC 040827–024) and the joint British Antarctic Survey- Cambridge University Animal Ethics Review Committee (does not issue permit numbers). Work at Cape Shirreff was conducted under the USA Marine Mammal Protection Act Permit No. 774-1847-04 granted by the Office of Protected Resources, National Marine Fisheries Service, the Antarctic Conservation Act Permit No. 2008-008, and approved by the NMFS–SWFSC Institutional Animal Care and Use Committee.

509

510 REFERENCES

- 511 Ainley, D.G., Tynan, C.T., Stirling, I., 2003. Sea ice: a critical habitat for polar marine mammals
512 and birds, in: Thomas, D.N., Dieckmann, G., S. (Eds.), *Sea Ice-An Introduction to its*
513 *Physics, Chemistry, Biology and Geology*. Blackwell, Oxford, pp. 240-266.
- 514 Ansorge, I.J., Lutjeharms, J.R., 2005. Direct observations of eddy turbulence at a ridge in the
515 Southern Ocean. *Geophysical Research Letters* 32, L14603.
- 516 Arthur, B., Hindell, M., Bester, M., Trathan, P., Jonsen, I., Staniland, I., Oosthuizen, W.C., Wege,
517 M., Lea, M.-A., 2015. Return Customers: Foraging Site Fidelity and the Effect of
518 Environmental Variability in Wide-Ranging Antarctic Fur Seals. *PLoS ONE* 10, e0120888.
- 519 Arthur, B., Hindell, M., Bester, M.N., Oosthuizen, W.C., Wege, M., Lea, M.-A., 2016. South for
520 the winter? Within-dive foraging effort reveals the trade-offs between divergent foraging
521 strategies in a free-ranging predator. *Functional Ecology*.
- 522 Atkinson, A., Whitehouse, M., Priddle, J., Cripps, G., Ward, P., Brandon, M., 2001. South Georgia,
523 Antarctica: a productive, cold water, pelagic ecosystem. *Marine Ecology Progress Series*
524 216, 279-308.
- 525 Barlow, K.E., Boyd, I.L., Croxall, J.P., Reid, K., Staniland, I.J., Brierley, A.S., 2002. Are penguins
526 and seals in competition for Antarctic krill at South Georgia? *Mar. Biol.* 140, 205-213.
- 527 Barraquand, F., Benhamou, S., 2008. Animal movements in heterogeneous landscapes: identifying
528 profitable places and homogeneous movement bouts. *Ecology* 89, 3336-3348.
- 529 Blanchet, M.-A., Biuw, M., Hofmeyr, G.G., de Bruyn, P.N., Lydersen, C., Kovacs, K.M., 2013. At-
530 sea behaviour of three krill predators breeding at Bouvetøya - Antarctic fur seals, macaroni
531 penguins and chinstrap penguins. *Marine Ecology Progress Series* 477, 285-302.

- Block, B., Jonsen, I., Jorgensen, S., Winship, A., Shaffer, S., Bograd, S., Hazen, E., Foley, D.,
Breed, G., Harrison, A.L., 2011. Tracking apex marine predator movements in a dynamic
ocean. *Nature* 475, 86-90.
- Bonadonna, F., Lea, M.A., Dehorter, O., Guinet, C., 2001. Foraging ground fidelity and route-
choice tactics of a marine predator: the Antarctic fur seal *Arctocephalus gazella*. *Marine
Ecology Progress Series* 223, 287-297.
- Bost, C.A., Cotté, C., Bailleul, F., Cherel, Y., Charrassin, J.B., Guinet, C., Ainley, D.G.,
Weimerskirch, H., 2009. The importance of oceanographic fronts to marine birds and
mammals of the southern oceans. *Journal of Marine Systems* 78, 363-376.
- Boyd, I., 1996. Individual variation in the duration of pregnancy and birth date in Antarctic fur
seals: the role of environment, age, and sex of fetus. *Journal of Mammalogy* 77, 124-133.
- Boyd, I., Arnould, J., Barton, T., Croxall, J., 1994. Foraging behaviour of Antarctic fur seals during
periods of contrasting prey abundance. *Journal of Animal Ecology* 63, 703-713.
- Boyd, I., Croxall, J., Lunn, N., Reid, K., 1995. Population demography of Antarctic fur seals: the
costs of reproduction and implications for life-histories. *Journal of Animal Ecology* 64, 505-
518.
- Boyd, I., Staniland, I., Martin, A., 2002. Distribution of foraging by female Antarctic fur seals.
Marine Ecology Progress Series 242, 285-294.
- Burger, A.E., Shaffer, S.A., 2008. Application of Tracking and Data-Logging Technology in
Research and Conservation of Seabirds. *The Auk* 125, 253-264.
- Charnov, E.L., 1976. Optimal foraging, the marginal value theorem. *Theoretical population biology*
9, 129-136.
- Clarke, A., 1988. Seasonality in the Antarctic marine environment. *Comparative Biochemistry and
Physiology Part B: Comparative Biochemistry* 90, 461-473.
- Cotté, C., Park, Y.-H., Guinet, C., Bost, C.-A., 2007. Movements of foraging king penguins through
marine mesoscale eddies. *Proceedings of the Royal Society B: Biological Sciences* 274,
2385-2391.
- Croxall, J., Prince, P., Ricketts, C., 1985. Relationships between prey life-cycles and the extent,
nature and timing of seal and seabird predation in the Scotia Sea, Antarctic nutrient cycles
and food webs. Springer-Verlag, Berlin, pp. 516-533.
- Daneri, G., Dellarosa, V., Quinones, R., Jacob, B., Montero, P., Ulloa, O., 2000. Primary
production and community respiration in the Humboldt Current System off Chile and
associated oceanic areas. *Marine Ecology Progress Series* 197, 41-49.
- de Bruyn, P.N., Tosh, C.A., Oosthuizen, W.C., Bester, M.N., Arnould, J.P., 2009. Bathymetry and
frontal system interactions influence seasonal foraging movements of lactating subantarctic
fur seals from Marion Island. *Marine Ecology Progress Series* 394, 263-276.
- Dormann, C., McPherson, J., Araújo, M., Bivand, R., Bolliger, J., Carl, G., Davies, R., Hirzel, A.,
Jetz, W., Daniel Kissling, W., Kuhn, I., Ohlemuller, R., Peres-Neto, P., Reineking, B.,
Schroder, B., Schurr, F., Wilson, R.P., 2007. Methods to account for spatial autocorrelation
in the analysis of species distributional data: a review. *Ecography* 30, 609-628.
- Elith, J., Leathwick, J.R., 2009. Species distribution models: Ecological explanation and prediction
across space and time. *Annual Review of Ecology, Evolution, and Systematics* 40, 677-697.
- Field, I.C., Bradshaw, C.J., Burton, H.R., Sumner, M.D., Hindell, M.A., 2005. Resource
partitioning through oceanic segregation of foraging juvenile southern elephant seals
(*Mirounga leonina*). *Oecologia* 142, 127-135.
- Friedlaender, A.S., Halpin, P.N., Qian, S.S., Lawson, G.L., Wiebe, P.H., Thiele, D., Read, A.J.,
2006. Whale distribution in relation to prey abundance and oceanographic processes in shelf
waters of the Western Antarctic Peninsula. *Marine Ecology Progress Series* 317, 297-310.
- Guinet, C., Dubroca, L., Lea, M.A., Goldsworthy, S., Cherel, Y., Duhamel, G., Bonadonna, F.,
Donnay, J.P., 2001. Spatial distribution of foraging in female Antarctic fur seals

582 *Arctocephalus gazella* in relation to oceanographic variables: a scale-dependent approach
583 using geographic information systems. Marine Ecology Progress Series 219, 251-264.

584 Guisan, A., Thuiller, W., 2005. Predicting species distribution: offering more than simple habitat
585 models. Ecology Letters 8, 993-1009.

586 Guisan, A., Zimmermann, N.E., 2000. Predictive habitat distribution models in ecology. Ecological
587 modelling 135, 147-186.

588 Heithaus, M.R., Dill, L.M., 2006. Does tiger shark predation risk influence foraging habitat use by
589 bottlenose dolphins at multiple spatial scales? Oikos 114, 257-264.

590 Hindell, M., Lea, M.-A., Bost, C.A., Charrassin, J.-B., 2011. Foraging habitats of top predators,
591 and Areas of Ecological Significance, on the Kerguelen Plateau, in: Duhamel, G., Welsford,
592 D. (Ed.), The Kerguelen Plateau: Marine Ecosystem and Fisheries, pp. 203-215.

593 Hyrenbach, K.D., Forney, K.A., Dayton, P.K., 2000. Marine protected areas and ocean basin
594 management. Aquatic conservation: marine and freshwater ecosystems 10, 437-458.

595 Incze, L., Hebert, D., Wolff, N., Oakey, N., Dye, D., 2001. Changes in copepod distributions
596 associated with increased turbulence from wind stress. Marine Ecology Progress Series 213,
597 229-240.

598 Kareiva, P., Odell, G., 1987. Swarms of predators exhibit "preytaxis" if individual predators use
599 area-restricted search. American Naturalist 130, 233-270.

600 Lea, M.-A., Guinet, C., Cherel, Y., Duhamel, G., Dubroca, L., Pruvost, P., Hindell, M., 2006.
601 Impacts of climatic anomalies on provisioning strategies of a Southern Ocean predator.
602 Marine Ecology Progress Series 310, 77-94.

603 Lea, M.-A., Guinet, C., Cherel, Y., Hindell, M., Dubroca, L., Thalmann, S., 2008. Colony-based
604 foraging segregation by Antarctic fur seals at the Kerguelen Archipelago. Marine Ecology
605 Progress Series 358, 273-287.

606 Lea, M.-A., Hindell, M., Arthur, B., Bester, M., De Bruyn, P., Oosthuizen, C., 2014a. Marion
607 Island female Antarctic fur seal geolocation tracking data ASAC-2940, Australian Antarctic
608 Division, <http://dx.doi.org/10.4225/15/531FEC9077D6E>

609 Lea, M.-A., Hindell, M., Arthur, B., Goebel, M., 2014b. Cape Shirreff female Antarctic fur seal
610 geolocation tracking data ASAC_2940, Australian Antarctic Data Centre,
611 <http://dx.doi.org/10.4225/15/531FDF2765BAB>

612 Lea, M.-A., Hindell, M., Arthur, B., Trathan, P., Staniland, I., Maloney, D., Edwards, E., 2014c.
613 Bird Island female Antarctic fur seal geolocation tracking data ASAC_2940, Australian
614 Antarctic Data Centre, <http://dx.doi.org/10.4225/15/531FD86AAF564>

615 Lea, M.-A., Johnson, D., Ream, R., Sterling, J., Melin, S., Gelatt, T., 2009. Extreme weather events
616 influence dispersal of naive northern fur seals. Biology letters 5, 252-257.

617 Lowther, A.D., Lydersen, C., Biuw, M., de Bruyn, P.J.N., Hofmeyr, G.J.G., Kovacs, K.M., 2014.
618 Post-breeding at-sea movements of three central-place foragers in relation to submesoscale
619 fronts in the Southern Ocean around Bouvetøya. Antarctic Science 26, 533-544.

620 Lunn, N., Boyd, I., 1991. Pupping-site fidelity of Antarctic fur seals at Bird Island, South Georgia.
621 Journal of Mammalogy 72, 202-206.

622 Mackas, D., Tsurumi, M., Galbraith, M., Yelland, D., 2005. Zooplankton distribution and dynamics
623 in a North Pacific Eddy of coastal origin: II. Mechanisms of eddy colonization by and
624 retention of offshore species. Deep Sea Research Part II: Topical Studies in Oceanography
625 52, 1011-1035.

626 Mitchell, B.G., Brody, E.A., Holm-Hansen, O., McClain, C., Bishop, J., 1991. Light limitation of
627 phytoplankton biomass and macronutrient utilization in the Southern Ocean. Limnology and
628 Oceanography 36, 1662-1677.

629 Nakano, S., 1995. Competitive interactions for foraging microhabitats in a size-structured
630 interspecific dominance hierarchy of two sympatric stream salmonids in a natural habitat.
631 Canadian Journal of Zoology 73, 1845-1854.

Nel, D., Lutjeharms, J., Pakhomov, E., Ansorge, I., Ryan, P., Klages, N., 2001. Exploitation of mesoscale oceanographic features by grey-headed albatross *Thalassarche chrysostoma* in the southern Indian Ocean. *Marine Ecology Progress Series* 217, 15-26.

Nordstrom, C.A., Battaile, B.C., Cotté, C., Trites, A.W., 2013. Foraging habitats of lactating northern fur seals are structured by thermocline depths and submesoscale fronts in the eastern Bering Sea. *Deep Sea Research Part II: Topical Studies in Oceanography* 88–89, 78-96.

Orians, G.H., Pearson, N.E., 1979. On the theory of central place foraging. Ohio State University Press, Columbus.

Péron, C., Weimerskirch, H., Bost, C.A., 2012. Projected poleward shift of king penguins' (*Aptenodytes patagonicus*) foraging range at the Crozet Islands, southern Indian Ocean. *Proceedings of the Royal Society B: Biological Sciences* 279, 2515-2523.

Phillips, R.A., Silk, J.R., Croxall, J.P., Afanasyev, V., 2006. Year-round distribution of white-chinned petrels from South Georgia: relationships with oceanography and fisheries. *Biological Conservation* 129, 336-347.

Pitcher, K.W., Calkins, D.G., Pendleton, G.W., 1998. Reproductive performance of female Steller sea lions: an energetics-based reproductive strategy? *Canadian Journal of Zoology* 76, 2075-2083.

Pollard, R., Venables, H., Read, J., Allen, J., 2007. Large-scale circulation around the Crozet Plateau controls an annual phytoplankton bloom in the Crozet Basin. *Deep Sea Research Part II: Topical Studies in Oceanography* 54, 1915-1929.

R Core Team, 2014. R: A language and environment for statistical computing. R Foundation for Statistical Computing., version 3.1.1 ed.

Raymond, B., Lea, M.A., Patterson, T., Andrews-Goff, V., Sharples, R., Charrassin, J.B., Cottin, M., Emmerson, L., Gales, N., Gales, R., 2014. Important marine habitat off east Antarctica revealed by two decades of multi-species predator tracking. *Ecography* 37, 001-009.

Reid, K., Arnould, J.P., 1996. The diet of Antarctic fur seals *Arctocephalus gazella* during the breeding season at South Georgia. *Polar Biol.* 16, 105-114.

Reid, K., Croxall, J.P., 2001. Environmental response of upper trophic-level predators reveals a system change in an Antarctic marine ecosystem. *Proceedings of the Royal Society of London. Series B: Biological Sciences* 268, 377-384.

Russell, R.W., Hunt Jr, G.L., Coyle, K.O., Cooney, R.T., 1992. Foraging in a fractal environment: spatial patterns in a marine predator-prey system. *Landscape Ecology* 7, 195-209.

Sakshaug, E., Slagstad, D., Holm-Hansen, O., 1991. Factors controlling the development of phytoplankton blooms in the Antarctic Ocean - a mathematical model. *Marine Chemistry* 35, 259-271.

Shrihai, H., 2002. A complete guide to Antarctic wildlife. Princeton University Press, Princeton.

Soto, K.H., Trites, A.W., Arias-Schreiber, M., 2004. The effects of prey availability on pup mortality and the timing of birth of South American sea lions (*Otaria flavescens*) in Peru. *Journal of Zoology* 264, 419-428.

Staniland, I.J., Gales, N., Warren, N.L., Robinson, S.L., Goldsworthy, S.D., Casper, R.M., 2010. Geographical variation in the behaviour of a central place forager: Antarctic fur seals foraging in contrasting environments. *Mar. Biol.* 157, 2383-2396.

Staniland, I.J., Robinson, S., Silk, J.R.D., Warren, N., Trathan, P., 2012. Winter distribution and haul-out behaviour of female Antarctic fur seals from South Georgia. *Mar. Biol.* 159, 291-301.

Sterling, J.T., Springer, A.M., Iverson, S.J., Johnson, S.P., Pelland, N.A., Johnson, D.S., Lea, M.-A., Bond, N.A., 2014. The Sun, Moon, Wind, and Biological Imperative—Shaping Contrasting Wintertime Migration and Foraging Strategies of Adult Male and Female Northern Fur Seals (*Callorhinus ursinus*). *PLoS ONE* 9, e93068.

682 Sumner, M., Wotherspoon, S., 2010. tripEstimation: Metropolis sampler and supporting functions
683 for estimating animal movement from archival tags.
684 Sumner, M.D., 2015. raadttools: Tools for synoptic environmental spatial data.
685 Sumner, M.D., Michael, K.J., Bradshaw, C.J.A., Hindell, M.A., 2003. Remote sensing of Southern
686 Ocean sea surface temperature: implications for marine biophysical models. *Remote*
687 *Sensing of Environment* 84, 161-173.
688 Sumner, M.D., Wotherspoon, S.J., Hindell, M.A., 2009. Bayesian estimation of animal movement
689 from archival and satellite tags. *PLoS ONE* 4, e7324.
690 Thums, M., Bradshaw, C., Hindell, M., 2008. Tracking changes in relative body composition of
691 southern elephant seals using swim speed data. *Marine Ecology Progress Series* 370, 249-
692 261.
693 Torres, L.G., Sutton, P.J.H., Thompson, D.R., Delord, K., Weimerskirch, H., Sagar, P.M., Sommer,
694 E., Dilley, B.J., Ryan, P.G., Phillips, R.A., 2015. Poor Transferability of Species
695 Distribution Models for a Pelagic Predator, the Grey Petrel, Indicates Contrasting Habitat
696 Preferences across Ocean Basins. *PLoS ONE* 10, e0120014.
697 Venables, H.J., Pollard, R.T., Popova, E.E., 2007. Physical conditions controlling the development
698 of a regular phytoplankton bloom north of the Crozet Plateau, Southern Ocean. *Deep Sea*
699 *Research Part II: Topical Studies in Oceanography* 54, 1949-1965.
700 Walters, A., 2014. Quantifying the trophic linkages of Antarctic marine predators. Phd thesis,
701 University of Tasmania, Hobart.
702 Weimerskirch, H., Salamolard, M., Sarrazin, F., Jouventin, P., 1993. Foraging strategy of
703 wandering albatrosses through the breeding season: a study using satellite telemetry. *The*
704 *Auk* 110, 325-342.
705 Wood, S., 2006. Generalized additive models: an introduction with R. CRC press, Boca Raton,
706 Florida, USA.
707
708

709 **Figure legends**

710 **Figure 1.** Maps of eight environmental climatology variables likely to influence fur seal foraging
711 effort during the Southern Ocean winter (April-December). Variables were averaged across all
712 winters in the study period (2008-13), except for SSHa Variance, for which variance was
713 calculated. The locations of the three study colonies are shown on the map in the bottom right
714 panel.

715

716 **Figure 2.** (a) Winter tracks for female Antarctic fur seals (n=184) split into three periods: early
717 (March-May), mid (June-August) and late (September-December) winter. The mean sea-ice extent
718 for each period over the five years (2008-13) is represented by the dashed lines. (b) Density
719 distribution of locations by latitude of fur seals from Marion Island (n=119), Bird Island (n=28) and

720 Cape Shirreff (n=37) during early (red), mid (blue) and late winter (green). Triangles show the
721 latitude of each colony. (c) The average number of new grid cells visited with the inclusion of
722 additional animals for the Marion Island (black), Bird Island (light grey) and Cape Shirreff (dark
723 grey) colonies.

724

725 **Figure 3.** Observed and predicted winter habitat use (time spent in hours per cell of a 60 km x 60
726 km grid) for female Antarctic fur seals from Marion Island, Bird Island and Cape Shirreff. 75%
727 distribution areas are shown by solid black lines. Dashed lines show the average position of the sea-
728 ice edge for early, mid and late winter. Black triangles show the location of each colony.

729

730 **Figure 4.** Predicted time spent per grid cell in relation to key environmental parameters (unscaled
731 and uncentred) for the winter foraging effort of female Antarctic fur seals. Lines show the
732 prediction of a GAM fit to a Gamma error structure for (a) Marion Island, (b) Bird Island and (c)
733 Cape Shirreff. Shading represents the 95% confidence interval for predictions. Black bars show the
734 distribution of observations. SST = sea surface temperature, SSTG = sea surface temperature
735 gradient, SSHa = sea surface height anomaly, SSHV = sea surface height variance, Wind = wind
736 speed, CHLa = chlorophyll a concentration, BATHY = bathymetry, d2col = distance to colony.

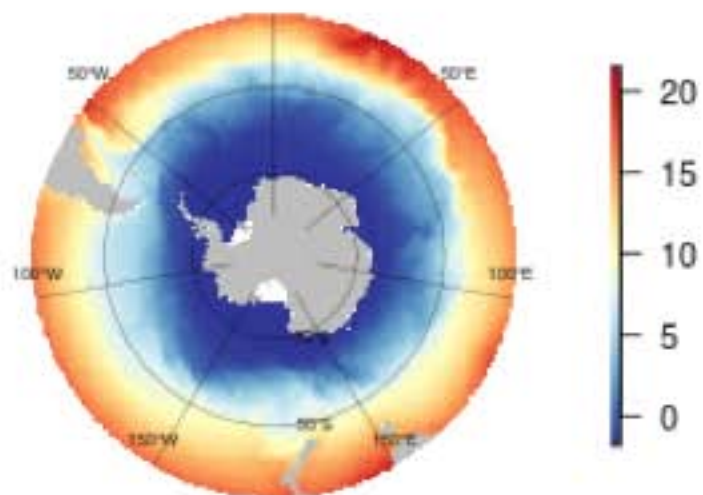
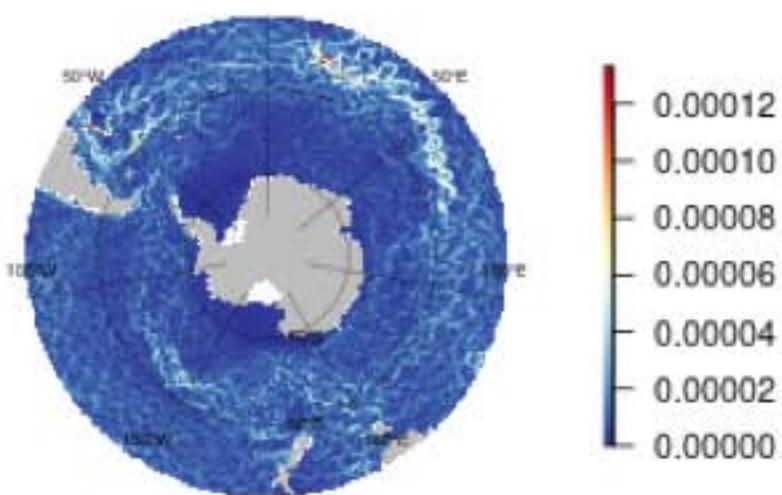
Table 1. Summary of data: Number of tags deployed, recovered, trips recorded and locations estimated by site and year collected for the winter foraging trips of female Antarctic fur seals.

Site	Year	GLS deployed	GLS recovered	N trips	N locations	Latitude
Marion Island	2008	30	20	42	9035	41.4°S
	2009	31	10	27	6509	
	2010	16	8	17	3148	
	2011	42	32	71	13 588	
	2012	30	26	41	16 709	
	2013	30	23	28	5062	
	<i>All years</i>	<i>179</i>	<i>119</i>	<i>227</i>	<i>54 051</i>	
Bird Island	2008	29	3	6	1407	41.7°S
	2009	30	9	18	4665	
	2010	30	10	21	4186	
	2011	30	6	11	2070	
	<i>All years</i>	<i>119</i>	<i>28</i>	<i>56</i>	<i>12 328</i>	
Cape Shirreff	2008	18	14	14	6562	41.3°S
	2009	19	11	11	5309	
	2010	19	12	12	5546	
	<i>All years</i>	<i>56</i>	<i>37</i>	<i>37</i>	<i>17 417</i>	
Total		354	184	320	83 796	

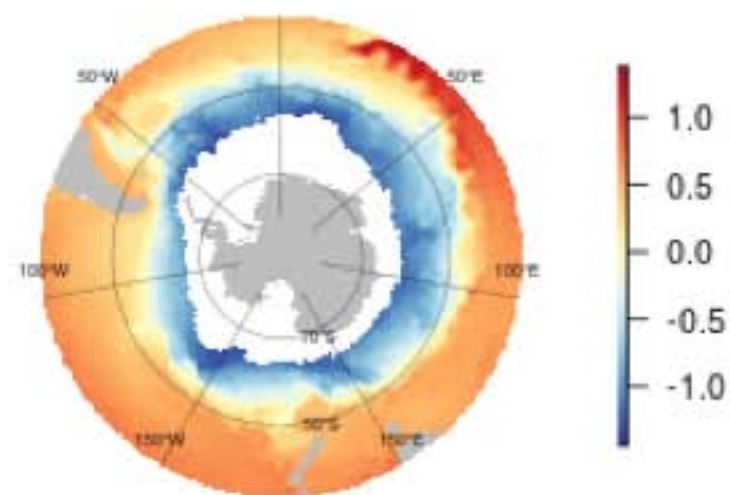
Table 2. Summary of generalised additive model (GAM) comparisons examining the relationship between foraging effort (time spent per grid cell) and environmental variables for: (a) Marion Island, (b) Bird Island and (c) Cape Shirreff colonies. TS = time spent, SST = sea surface temperature, SSTG = sea surface temperature gradient, SSHa = sea surface height anomaly, SSHV = sea surface height variance, Wind = wind speed, CHLa = chlorophyll a concentration, BATHY = bathymetry, d2col = distance to colony, (lon,lat) = spatial autocorrelation term. Only models with a $\Delta\text{AIC} < 10$ are shown and the accepted model is presented in bold.

Candidate models	<i>k</i>	LL	AIC	ΔAIC	<i>w</i> AIC
<i>(a) Marion Island</i>					
1. TS ~ SST + SSHa + SSHV + Wind + CHLa + BATHY + EKE + d2col + (lon,lat)	10	-24833.0	49851.2	0.0	0.691
2. TS ~ SST + SSTG + SSHa + SSHV + Wind + CHLa + BATHY + EKE + d2col + (lon,lat)	11	-24832.5	49852.9	1.7	0.295
3. TS ~ SST + SSHa + SSHV + Wind + CHLa + EKE + d2col + (lon,lat)	9	-24843.5	49859.1	7.9	0.013
<i>(b) Bird Island</i>					
1. TS ~ SST + SSTG + SSHa + Wind + BATHY + EKE + d2col + (lon,lat)	9	-13702.8	27563.4	0.0	0.355
2. TS ~ SST + SSTG + SSHa + SSHV + Wind + CHLa + BATHY + EKE + d2col + (lon,lat)	11	-13695.9	27563.5	0.1	0.334
3. TS ~ SST + SSTG + SSHa + Wind + CHLa + BATHY + EKE + d2col + (lon,lat)	10	-13702.1	27563.8	0.4	0.291
4. TS ~ SST + SSTG + SSHa + Wind + EKE + d2col + (lon,lat)	8	-13706.8	27569.3	5.9	0.017
<i>(c) Cape Shirreff</i>					
1. TS ~ SST + SSHa + SSHV + Wind + CHLa + EKE + d2col + (lon,lat)	9	-27232.9	54637.4	0.0	0.594
2. TS ~ SST + SSHa + SSHV + Wind + CHLa + BATHY + EKE + d2col + (lon,lat)	10	-27232.9	54638.9	1.5	0.287
3. TS ~ SST + SSTG + SSHa + SSHV + Wind + CHLa + BATHY + EKE + d2col + (lon,lat)	11	-27232.8	54640.7	3.3	0.115

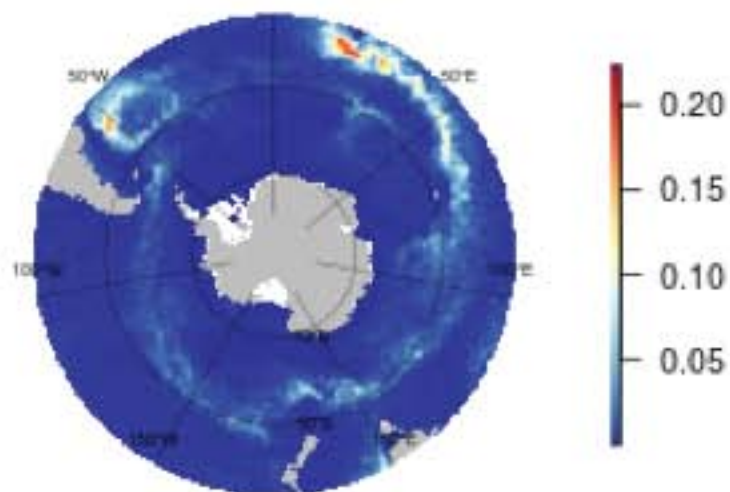
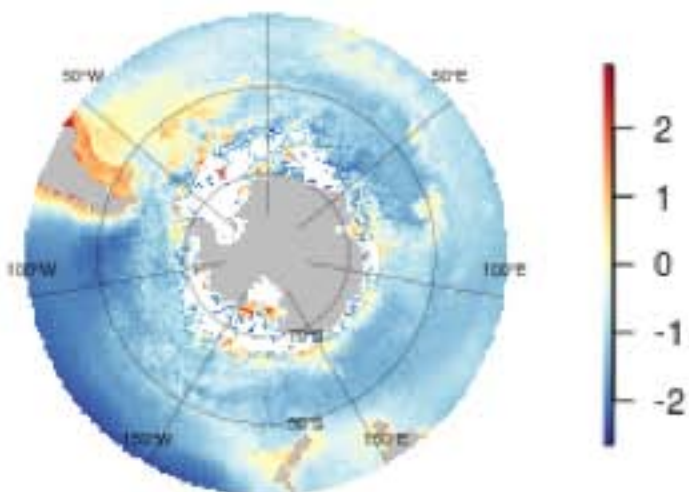
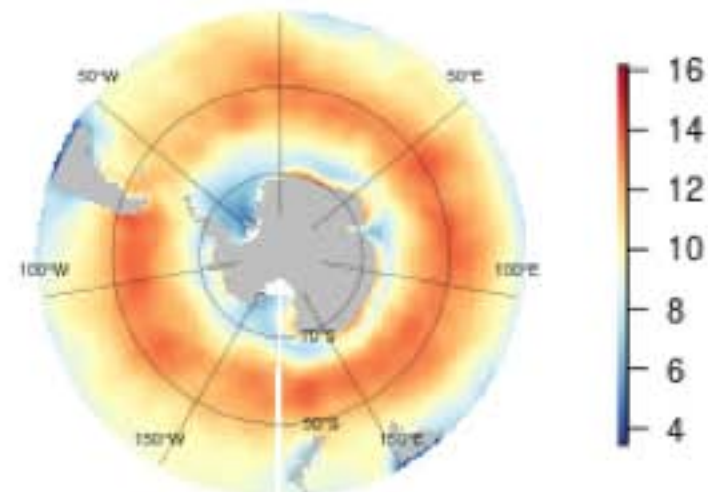
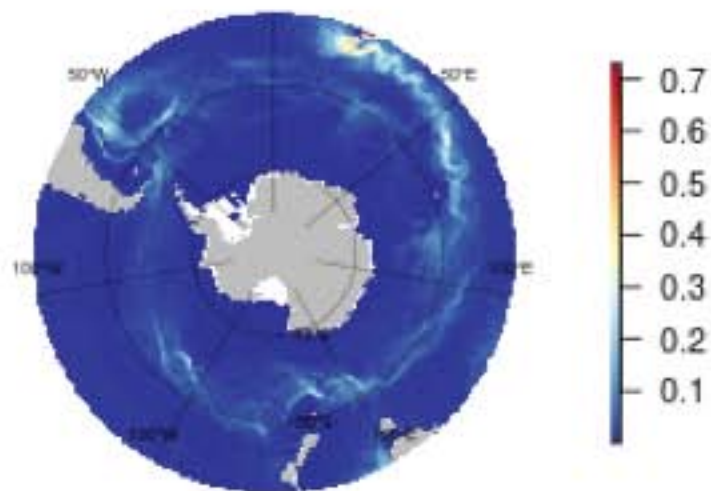
k, number of parameters; LL, log-likelihood; AIC, Akaike's Information Criterion; ΔAIC , difference in AIC from that of the best fitting model; *w*AIC, AIC weight; prop dive vARS, proportion of the dive in vertical area-restricted search behaviour.

SST ($^{\circ}\text{C}$)SST Gradient ($^{\circ}\text{C}$)

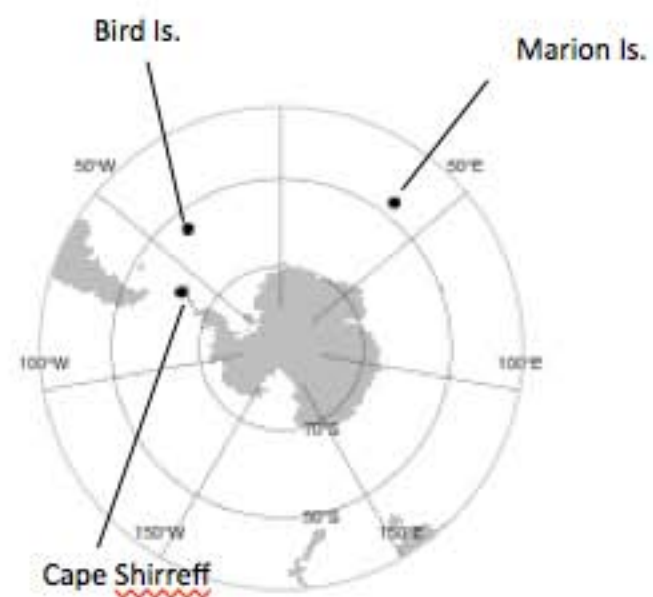
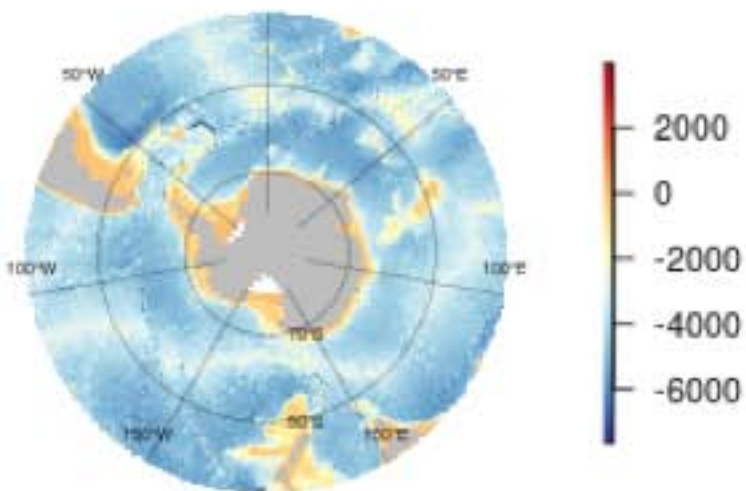
SSHa (m)



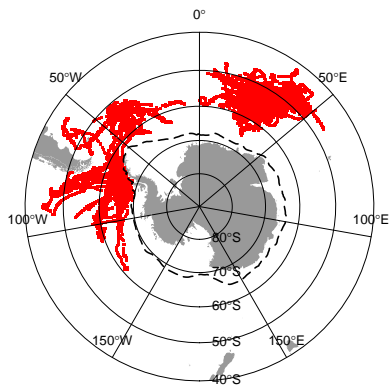
SSHa Variance (m)

CHLa ($\log \text{mg.m}^{-3}$)Wind (m.s^{-1})Eddy Kinetic Energy ($\text{cm}^2 \text{s}^{-2}$)

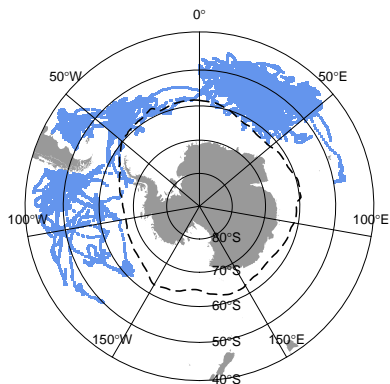
Bathymetry (m)



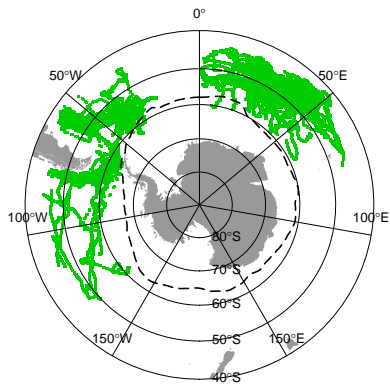
Early winter



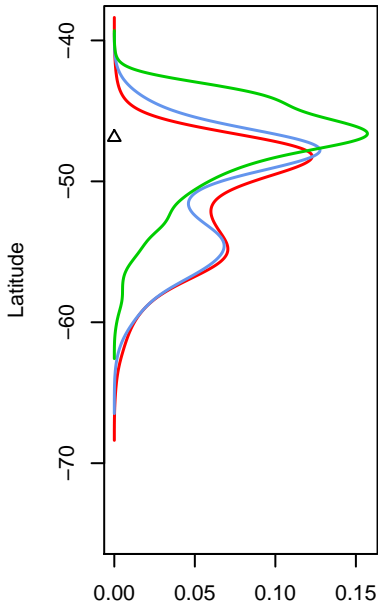
Mid winter



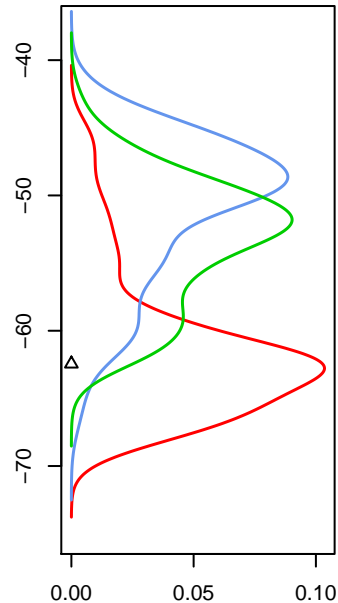
Late winter



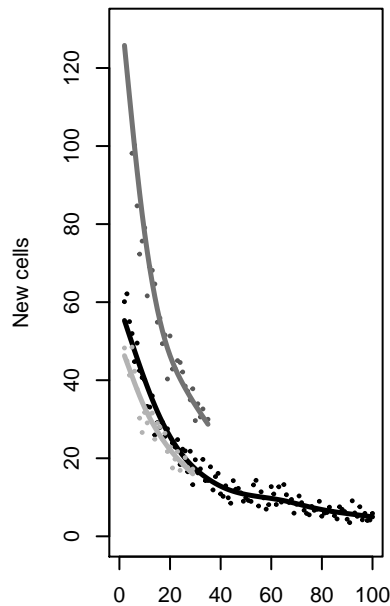
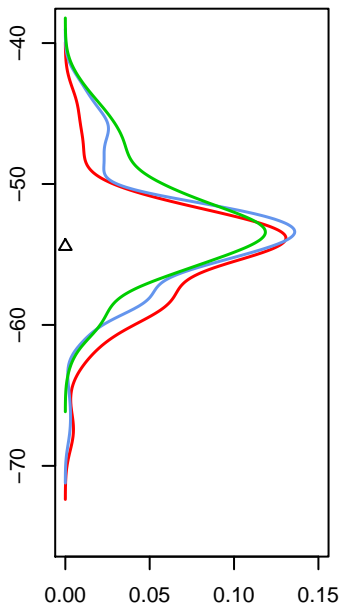
Marion Island (2008–13)



Cape Shirreff (2008–10)



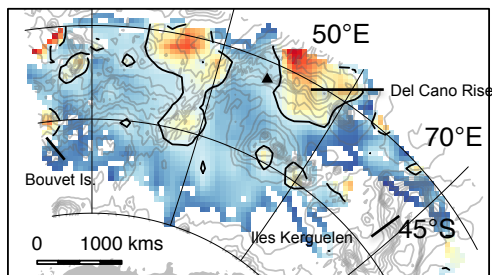
Bird Island (2008–11)



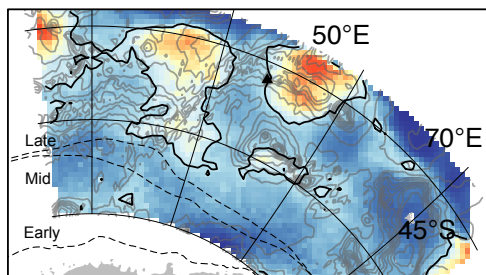
Density

Number of seals

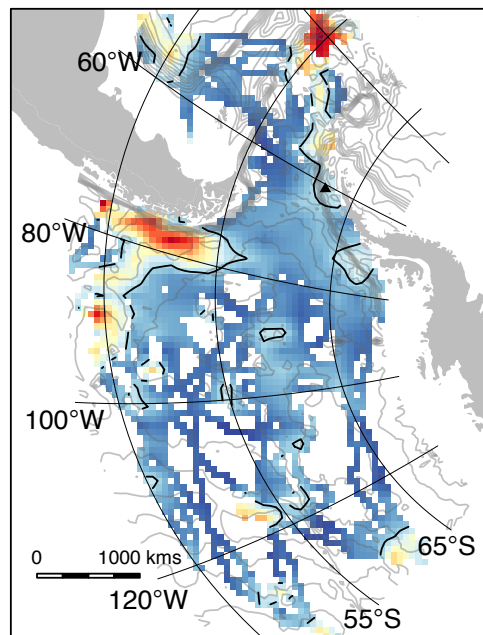
Marion Island – observed



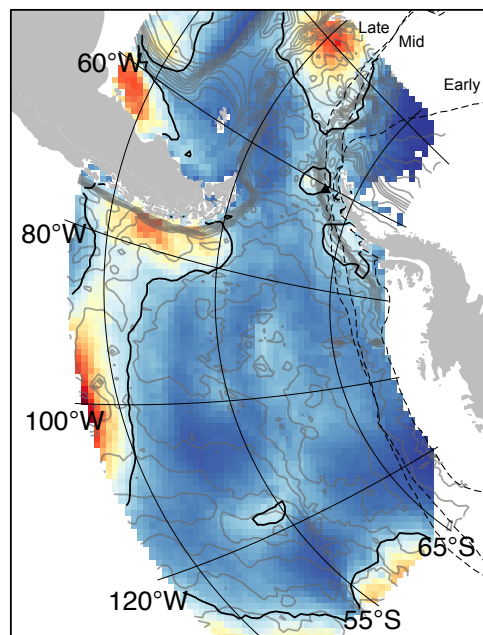
Marion Island – predicted



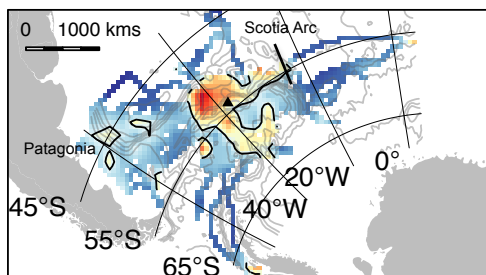
Cape Shirreff – observed



Cape Shirreff – predicted



Bird Island – observed



Bird Island – predicted

