

1 **Philopatry Influences the Genetic Population Structure of the Blacktip Shark (*Carcharhinus***
2 ***limbatus*) at Multiple Spatial Scales**

3 **Genetic Structure of the Blacktip Shark**

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26 Abstract

27 Understanding how interactions among microevolutionary forces generate genetic population
28 structure of exploited species is vital to the implementation of management policies that facilitate
29 persistence. Philopatry displayed by many coastal shark species can impact gene flow and facilitate
30 selection, and has direct implications for the spatial scales of management. Here, genetic structure
31 of the blacktip shark (*Carcharhinus limbatus*) was examined using a mixed-marker approach
32 employing mitochondrial control region sequences and 4,339 SNP-containing loci generated using
33 ddRAD-Seq. Genetic variation was assessed among young-of-the-year sampled in 11 sites in
34 waters of the United States in the western North Atlantic Ocean, including the Gulf of Mexico.
35 Spatial and environmental analyses detected 68 nuclear loci putatively under selection, enabling
36 separate assessments of neutral and adaptive genetic structure. Both mitochondrial and neutral
37 SNP data indicated three genetically distinct units – the Atlantic, eastern Gulf, and western Gulf –
38 that align with regional stocks and suggest regional philopatry by males and females.
39 Heterogeneity at loci putatively under selection, associated with temperature and salinity, was
40 observed among sites within Gulf units, suggesting local adaptation. Furthermore, five pairs of
41 siblings were identified in the same site across timescales corresponding with female reproductive
42 cycles. This indicates that females re-used a site for parturition, which has the potential to facilitate
43 the sorting of adaptive variation among neighboring sites. The results demonstrate differential
44 impacts of microevolutionary forces at varying spatial scales and highlight the importance of
45 conserving essential habitats to maintain sources of adaptive variation that may buffer species
46 against environmental change.

47 Keywords

48 conservation genomics; local adaptation; elasmobranch; parturition site fidelity; male philopatry.

49 **Introduction**

50 Genetic population structure is determined by differences in the distribution of alleles among
51 contemporary populations that result from interactions of microevolutionary forces (Laikre et al.,
52 2005). Because genetic drift and gene flow influence allele frequencies on a genome-wide scale,
53 selectively neutral loci exhibit patterns of variation that can be used to understand historical and
54 contemporary demographic processes (Luikart et al., 2003). By contrast, selection acts upon
55 variation at specific genes and/or genomic regions, and often produces patterns of structure distinct
56 from those observed at neutral loci (Gagnaire et al., 2015; Nielsen, 2001). Disentangling these
57 patterns is especially informative for the management of exploited species. While neutral structure
58 can inform the designation of management units (Waples et al., 2008), loci under selection can be
59 used to infer local adaptation across heterogeneous environments within management units
60 (Nielsen et al., 2009). Understanding levels of gene flow among and within units is also critical
61 because the adaptive potential of populations can facilitate the persistence of species confronted
62 with environmental change (Bowen & Roman, 2005; Garant et al., 2007).

63 Examining the interplay of microevolutionary forces is challenging in marine systems because
64 barriers to gene flow are fewer and often cryptic and they can be more difficult to study than many
65 terrestrial systems (Grummer et al., 2019; Palumbi, 1994). In addition, marine species typically
66 exhibit weak structure that is difficult to detect (Waples, 1998), resulting from the potential for
67 long-distance dispersal (via adults and/or larvae), high fecundity, and large effective population
68 sizes that reduce the magnitude of genetic drift (Poulsen et al., 2006). However, large population
69 sizes and high fecundities provide more opportunities for mutation and increase the efficacy of
70 selection relative to drift (Allendorf et al., 2010; Cormack et al., 1990). Further, many marine
71 species have broad geographic ranges and are distributed across heterogeneous environments,

72 increasing the potential for local adaptation (Bernatchez, 2016). Therefore, selection acting with
73 varying degrees of strength upon a small number of loci can lead to fine-scale adaptive structure
74 while neutral processes produce weaker, genome-wide structure across broader geographic scales
75 (Gagnaire & Gaggiotti, 2016; Hoey & Pinsky, 2018).

76 The life history characteristics of elasmobranchs (i.e., sharks, skates, and rays) have an important
77 role in shaping patterns of genetic structure. In contrast to many bony fishes and marine
78 invertebrates, elasmobranchs mature late, have long life spans, and produce relatively few progeny
79 within and across reproductive efforts (Conrath & Musick, 2012). Frequently, this leads to smaller
80 effective sizes that are more coupled to census sizes (Portnoy et al., 2009). Though elasmobranchs
81 lack a dispersive larval stage, they retain the potential for high levels of gene flow because they
82 can move vast distances during juvenile and adult life stages (Kohler & Turner, 2019). However,
83 many species display fidelity to specific habitats where they mate and give birth or deposit eggs
84 (Chapman et al., 2015; Flowers et al., 2016). Furthermore, this behavior can extend across
85 generations, causing individuals to reproduce in their region of birth (i.e., regional philopatry;
86 Pardini et al., 2001) and even result in females giving birth in the same habitat in which they were
87 born (i.e., natal philopatry; Feldheim et al., 2014).

88 Female philopatry is common among coastal shark species that give birth in bays and estuaries
89 where progeny may remain for extended periods (Heupel et al., 2007; Karl et al., 2011; Keeney et
90 al., 2005). Female regional philopatry has the potential to limit gene flow mediated by females
91 compared with males, and evidence for this has been documented in multiple species based on
92 discrepancies in maternally- and biparentally-inherited DNA (Phillips et al., 2021). Because
93 coastal sharks are heavily exploited around the world (Dulvy et al., 2017), understanding how
94 philopatry influences neutral genetic structure by impacting gene flow is vital for delineating

95 management units that will promote persistence. In addition, parturition sites can be
96 environmentally heterogeneous (Bethea et al., 2015; Matich et al., 2017) and newborn sharks can
97 be subject to higher rates of mortality than other life stages (Heupel & Simpfendorfer, 2002; Lowe,
98 2002; Manire & Gruber, 1993). Therefore, natal philopatry could drive selection for locally
99 adaptive phenotypes and lead to fine-scale adaptive structure (Portnoy et al., 2015; Portnoy &
100 Heist, 2012). This could have further implications for management because parturition sites
101 harboring novel adaptive variants may require individually tailored policies.

102 The blacktip shark (*Carcharhinus limbatus*) is a coastal shark species with a circumglobal
103 distribution in tropical and warm temperate latitudes, that is harvested for meat, fins, and liver oil
104 (Compagno et al., 2005; Rigby et al., 2021). In waters of the United States (hereafter U.S. waters),
105 blacktip sharks are found along the Atlantic coast from Florida to Massachusetts and throughout
106 the Gulf of Mexico, where they are targeted by commercial and recreational fisheries (Castro,
107 1996; SEDAR, 2018, 2020). Commercial fisheries operate year-round and harvest adults in federal
108 and state waters; however, recreational fisheries also operate in state waters, and some may land
109 smaller blacktip sharks closer to shore (SEDAR, 2020). Male and female blacktip sharks mature
110 after four and six years (respectively) and females produce one to eight pups (four on average)
111 every two years (Baremore & Passerotti, 2013; Natanson et al., 2019). Moreover, the species is
112 highly migratory: males and females can move ~1,200 km in fewer than 100 days (Weber et al.,
113 2020) and males have been recorded traveling over 3,400 km per year (Bowers and Kajiura
114 unpublished data). In the spring and early summer, females move into bays and estuaries to give
115 birth (Castro, 1996; Hueter & Tyminski, 2007). Young-of-the-year (YOY) remain in their
116 parturition site until the autumn of their birth year and migrate south and/or offshore when water

117 temperatures decrease (Castro, 1996; Heupel et al., 2004; Heupel, 2007), and many return to the
118 vicinity of their parturition site the following spring (Hueter et al., 2005).

119 Based in part on population genetics studies, the U.S. National Marine Fisheries Service (hereafter
120 NOAA Fisheries) currently manages blacktip sharks as two stocks – one in the Atlantic and one
121 in Gulf – but the Gulf stock is split into two subregions (eastern and western), with the dividing
122 line through Mobile Bay, Alabama (SEDAR, 2018, 2020). An assessment of genetic structure
123 based on YOY sampled in parturition sites from Texas, Florida, and Georgia/South Carolina
124 identified three genetic units using the mitochondrial control region, but did not find significant
125 differences using eight nuclear-encoded microsatellites, suggesting female regional philopatry
126 (Keeney et al., 2005). However, the discordance between nuclear and mitochondrial data could
127 also be due to limited resolution (i.e., too few loci) or insufficient time for differences to accrue
128 (Whitlock & McCauley, 1999). Thus, to inform appropriate management and avoid loss of genetic
129 variation resulting from localized depletion, it is vital to accurately characterize blacktip shark
130 population structure and adaptive potential. An assessment of genetic structure at neutral and
131 putatively adaptive loci is therefore warranted.

132 Here, the genetic structure of blacktip sharks in U.S. waters of the western North Atlantic Ocean
133 was examined using mitochondrial control region and double digest restriction-site associated
134 DNA sequencing (ddRAD-Seq) data. The sampling design targeted YOY within or just outside
135 parturition sites during their spring-autumn residency to ensure that structure reflected differences
136 among reproductive units. By examining thousands of loci spread throughout the genome, a higher
137 resolution assessment of genetic structure at nuclear-encoded loci is possible, and the data can also
138 be used to identify siblings captured in the same habitats across years, a pattern indicative of
139 parturition site fidelity by females. Moreover, by screening for loci putatively under selection, the

140 approach facilitates an assessment of the influence of genetic drift, gene flow, and selection in
141 structuring genomic variation, providing a means to identify habitats harboring adaptive variants
142 that may facilitate the species' persistence.

143 **Materials and Methods**

144 ***Sampling***

145 Tissue samples were collected as fin clips from 503 individual blacktip sharks captured within or
146 near 11 estuaries (sites) off the U.S. Atlantic Coast (hereafter Atlantic) and throughout the northern
147 Gulf of Mexico (hereafter Gulf). The three sites in the Atlantic were along the coast of South
148 Carolina. In the Gulf, there were three sites along the west coast of Florida, one on the coast of
149 Alabama, and four along the coast of Texas. Mobile Bay, the site in Alabama, straddles the 88th
150 meridian that separates the eastern and western blacktip shark Gulf stock subregions (NMFS,
151 2006). Fin clips were immersed in 20% DMSO-0.25M EDTA NaCl-saturated buffer (DMSO,
152 Seutin et al., 1991), or ethanol and then transferred into DMSO, and stored at room temperature
153 until DNA extraction. All sharks were captured between March and November 2012-2019. The
154 location of capture (latitude and longitude) was recorded for each individual, and sex was recorded
155 for all but seven individuals. Body measurements (i.e., at least one of pre-caudal, fork, total, and
156 stretch total lengths) were also recorded. If a fork or total length was not recorded, a customized
157 R script (v3.6.0; R Development Core Team, 2008) was used to assign missing values based on
158 observed relationships among length measurements (Carlson et al., 2006). Of the 503 individuals
159 sampled, 488 were YOY: 227 (~47%) were classified as YOY based on the presence of an
160 umbilical scar (Castro, 1993) and 261 (~53%) were classified as YOY using fork length (< 593
161 mm) if sampled in the Atlantic (Ulrich et al., 2007) or total length (<= 800 mm) if sampled in the
162 Gulf (Parsons & Hoffmayer, 2007). Based on observations that YOY blacktip sharks in the

163 Atlantic and Gulf remain in or near their parturition site into the autumn months of their first year
164 of life (Castro, 1996; Heupel et al., 2004), these 488 individuals were assumed to have been
165 sampled in their parturition site (Table S1).

166 ***ddRAD-Seq Library Preparation and Genotyping***

167 High molecular weight genomic DNA was extracted from fin clips using either Mag-Bind® Blood
168 and Tissue DNA Kits (Omega Bio-Tek) or phenol-chloroform extraction (Sambrook et al., 1989).
169 A modified version of ddRAD-Seq (Peterson et al., 2012) was used to prepare genomic libraries
170 containing the 488 YOY individuals plus 31 technical replicates spread across sites and libraries
171 and sequenced using 11 lanes of an Illumina HiSeq 4000 (paired-end 150 bp; see Supplementary
172 Methods for more information).

173 To map and improve the genotyping efficacy of HiSeq data, a separate library consisting of 27
174 individuals sampled across Atlantic and Gulf locations at multiple life history stages (Table S2)
175 was prepared using the same protocol and sequenced on a single Illumina MiSeq lane (paired-end
176 300 bp). Of these 27 individuals, 12 were included in the HiSeq libraries. All raw HiSeq and MiSeq
177 reads were demultiplexed using *process radtags* (Catchen et al., 2011) and quality-trimmed using
178 default parameters implemented in DDOCENT (Puritz et al., 2014). DDOCENT was also used to
179 assemble MiSeq reads into a reference of contiguous sequence alignments (i.e., contigs)
180 representing putatively single-copy (orthologous) loci. DDOCENT was subsequently used to map
181 HiSeq reads to the MiSeq reference and genotype SNPs.

182 ***ddRAD-Seq Data Filtering***

183 Raw SNPs were filtered using VCFTOOLS (v0.1.14; Danecek et al., 2011) and R functions in a
184 customized workflow, following practices laid out in O’Leary et al. (2018). Filtering initially

185 removed genotypes with < 5 reads and quality < 20 while applying a minor allele count of three.
186 Loci were further filtered based on allele balance, mapping quality, ratio of reference vs. alternate
187 allele, consistency of scoring in forward and reverse directions, proper pairing, depth/quality ratio,
188 and excess heterozygosity to remove potential paralogs and other technical artifacts. Individuals
189 with $> 20\%$ missing data or very negative F_{IS} (< -0.13) indicative of cross-contamination (Petrou
190 et al., 2019) were removed. Retained loci had a mean depth > 20 and were called in at least 90%
191 of individuals, 80% of individuals in each site, and 50% of individuals in each library. Haplotypes
192 were then generated by collapsing SNPs on the same contig to produce a dataset of multi-allelic
193 SNP-containing loci (Willis et al., 2017). In addition, the composite genotypes of technical
194 replicates included within and across libraries were compared to characterize locus-specific
195 genotyping error. Replicates were confirmed by assessing relatedness between each pair of
196 individuals using the dyadic likelihood estimator (Milligan, 2003) executed using the R package
197 *related* (Pew et al., 2015). Loci with systematic genotyping error (i.e., in > 1 replicate pair) and
198 one individual from each pair were removed, along with monomorphic loci. To minimize genotype
199 inconsistencies across libraries (i.e., library effects), individuals were grouped by library and
200 BAYESCAN (Fischer et al., 2011; Foll et al., 2010; Foll & Gaggiotti, 2008) executed to identify and
201 remove loci contributing to differences among libraries.

202 ***Mitochondrial Sequencing and Haplotyping***

203 A 915 bp portion of the mitochondrial control region (1070 bp total length) was amplified for a
204 subset of individuals (323) using a pair of primers within the proline (Pro:
205 GCCCTTGGCTCCCAAAGC) and phenylalanine (Phe: TCATCTTAGCATCTTCAGTGCCA)
206 tRNA genes (Table S3). These primers were designed to amplify the mitochondrial control region
207 of multiple shark species (see Supplementary Methods and Table S4). Amplification was

208 performed using polymerase chain reaction (PCR) in 50 μ l reactions with 1x Green GoTaq buffer
209 (Promega), 2 mM MgCl₂, 200 μ M of each dNTP, 0.5 μ M of each primer, and 1.25 units of GoTaq
210 DNA Polymerase. Amplification consisted of an initial denaturation at 95°C for two minutes,
211 followed by 40 cycles of denaturation at 94°C for 30 seconds, annealing at 54°C for 60 seconds,
212 and extension at 72°C for 90 seconds, with a final extension of 72°C for 10 minutes. PCR products
213 were visualized using gel electrophoresis before being cleaned, quantified, and standardized to 20
214 ng/ μ l. Mitochondrial sequence data was generated by unidirectional Sanger sequencing using the
215 Pro primer and an ABI 3730xl platform.

216 Mitochondrial sequences were aligned using CLUSTAL OMEGA (Sievers et al., 2011) and edited
217 manually in BIOEDIT (Hall, 1999). The R package *haplotypes* was used to identify unique
218 haplotypes. To visualize the distribution of haplotypes among sites, a TCS network (Clement et
219 al., 2000) was produced using POPART (Leigh & Bryant, 2015).

220 **Relatedness**

221 To identify full- and half-siblings, pairwise relatedness was assessed using Wang's estimator
222 corrected for sample size (Wang, 2002) executed using the R package *demerelate* (Kraemer &
223 Gerlach, 2017). Because female blacktip sharks are thought to display regional philopatry (Keeney
224 et al., 2005) and relatedness analysis used to confirm technical replicates already screened for kin
225 sampled between regions, relatedness between individuals was assessed for each region separately
226 (i.e., Atlantic, eastern Gulf, and western Gulf). For each region, 1,000 pairs of simulated full- and
227 half-sibling relationships were generated using empirical allele frequencies. To identify full- and
228 half-siblings, minimum relatedness thresholds were set after trimming the lowest 1% of simulated
229 values to reduce instances of false positives. Mitochondrial haplotypes were then compared for
230 observed sibling pairs to determine if any half-siblings were paternally related (i.e., had distinct

231 haplotypes). Removal of randomly sampled siblings can reduce the precision of population
232 genetics analyses, as can the inclusion of siblings that are non-randomly sampled (Waples &
233 Anderson, 2017). Therefore, full- and half-siblings were considered non-randomly sampled if both
234 individuals were captured in the same site on the same day, in which case one individual from each
235 pair was removed for all downstream analyses.

236 ***F_{ST} Outlier Analysis***

237 Three methods were used to screen for F_{ST} outlier loci putatively under directional selection with
238 individuals grouped by site. The first approach, implemented in *OutFLANK* (Whitlock &
239 Lotterhos, 2015), identifies F_{ST} outliers (q -value < 0.05) based on an inferred distribution of neutral
240 F_{ST} after trimming the lowest and highest 5% of F_{ST} values, thus avoiding implicit assumptions of
241 population structure and demography. The second method generates a null distribution of F_{ST} for
242 neutral loci using a Bayesian approach implemented in *BAYESCAN* (Fischer et al., 2011; Foll et
243 al., 2010; Foll & Gaggiotti, 2008). This method assumes an island model where allele frequencies
244 in each group are correlated through a common ancestral gene pool. *BAYESCAN* was executed with
245 prior odds of 1,000 and a burn-in of 200,000 iterations; 25 pilot runs of 5,000 iterations were used
246 to tune MCMC parameters and following 35,000 sampling iterations with a thinning interval of
247 50, significance was evaluated using a q -value of 0.05. Finally, the *FDIST* method (Beaumont &
248 Nichols, 1996), implemented in *ARLEQUIN* v3.5 (Excoffier & Lischer, 2010), identifies loci with
249 elevated F_{ST} for simulated background heterozygosity under two models: an island model and a
250 hierarchical island model in which sites in the Atlantic and Gulf were grouped. For both models,
251 50,000 simulations were executed, 100 demes were simulated per group, and significance was
252 evaluated using α of 0.05 corrected for multiple comparisons (Benjamini & Hochberg, 1995) by
253 the *p.adjust* function of the R package *stats*.

254 ***Spatial and Environmental Analysis***

255 To examine the effects of spatial and environmental variation on genetic structure, correlations
256 among genomic variation, spatial position, and environmental variables were assessed using
257 redundancy analysis (RDA), as implemented in the R package *vegan* (Oksanen et al., 2018). RDA
258 is a constrained ordination method based on multivariate regression that models how linear
259 combinations of explanatory variables explain variation at a series of response variables, thereby
260 enabling the identification of loci that co-vary with multivariate predictors (Legendre & Legendre,
261 2012). This approach is particularly useful when applied to genomic datasets because it can be
262 performed without grouping individuals by location and does not rely on assumptions of
263 equilibrium between microevolutionary forces, both of which are inherent components of F_{ST} -
264 based analyses. Thus, RDA provides an alternative approach to assess population structure while
265 screening for loci putatively under selection (Forester et al., 2018).

266 The genomic dataset was transformed into a response matrix detailing the allelic composition of
267 each individual across loci (i.e., the number of copies of each allele at each locus for each
268 individual). Two explanatory matrices describing relative spatial positions and environmental
269 measurements for each sampling location were then produced. To ensure that each individual had
270 a unique sampling location, the R package *geoR* (Ribeiro & Diggle, 2001) was used to jitter
271 latitudes and longitudes for individuals caught in the same sampling effort. To generate the spatial
272 matrix, Moran's eigenvector maps (MEMs; Dray et al., 2006) were calculated using the R package
273 *adespatial* (Dray et al., 2019) based on coastal distances estimated between all sample locations
274 using the R package *gdistance* (Van Etten, 2017). The environmental matrix encompassed
275 measurements for coastal locations (Table S5) that were procured from the MARSPEC (35
276 variables; Sbrocco & Barber, 2013) and Bio-ORACLE (447 variables; Assis et al., 2018;

277 Tyberghein et al., 2012) databases using the R package *sdmpredictors* (Bosch & Fernandez, 2021).
278 For each explanatory matrix, forward model selection was used to identify the combination of
279 variables that best explained genomic variation based on adjusted R^2 and significance testing (999
280 permutations; $\alpha < 0.01$; Blanchet et al., 2008). Because collinearity is likely among environmental
281 variables, model selection prohibited the inclusion of variables with variance inflation factors
282 (VIF) > 3 (Zuur et al., 2010).

283 The significance of each axis of the spatial and environmental RDA models was assessed using
284 999 permutation tests with α of 0.05. To visualize the differential effects of space and environment
285 on genetic structure, the approach outlined by Forester et al. (2018) was used to produce individual
286 biplots depicting how spatial and environmental RDA clustered individuals based on the
287 combination of variables that were selected by each analysis. However, because environmental
288 data is almost always spatially autocorrelated (Legendre, 1993), it is vital to disentangle spatial
289 and environmental signals when identifying loci putatively under selection (Hoban et al., 2016).
290 Therefore, partial RDA (pRDA), in which the linear effects of one set of variables are adjusted by
291 accounting for covariates (Capblancq & Forester, 2021), was used to identify alleles most
292 strongly correlated with environmental variables adjusted for spatial position. Allele loadings
293 should form a distribution in which alleles at the center show no relationship with environment,
294 while those with loadings in the tails are strongly associated, and may therefore be considered
295 putatively under selection (Forester et al., 2018). Environmentally-associated loci were defined
296 using a function that sets thresholds three standard deviations from the mean (equivalent to a two-
297 tailed p -value of 0.0027; (Forester et al., 2018)). The significance of the full environmental pRDA
298 model and each axis was assessed using 999 permutation tests with α of 0.05.

299 **Population Structure**

300 Allele frequencies of neutral and adaptive loci are shaped by different sets of interactions among
301 microevolutionary forces and may provide for distinct patterns of genetic structure (Luikart et al.,
302 2003). Therefore, nuclear loci flagged as being putatively under selection by either of the F_{ST}
303 outlier methods or determined to be environmentally associated using pRDA were designated as
304 adaptive. The nuclear data was then divided into adaptive and neutral (i.e., all other loci) datasets.
305 For each of the three datasets (mitochondrial control region, neutral, and adaptive nuclear loci),
306 hierarchical AMOVA (Excoffier et al., 1992) was executed separately using ARLEQUIN. For the
307 mitochondrial data, standard AMOVA was performed. For neutral and adaptive datasets, locus-
308 by-locus AMOVA was performed, with F -statistics calculated as weighted means of locus-specific
309 values to account for uneven levels of missing data among loci (Weir & Cockerham, 1984). Sites
310 were grouped as Atlantic and Gulf, with significance assessed ($\alpha < 0.05$) by permuting individuals
311 among sites 10,000 times and by bootstrapping the nuclear data 20,000 times to create 95%
312 confidence intervals. For each dataset, single-level AMOVA was also executed for Atlantic and
313 Gulf sites separately. Subsequently, *post-hoc* estimates of pairwise Φ_{ST} and F_{ST} between sites were
314 calculated using ARLEQUIN, with 95% confidence intervals produced and significance assessed as
315 above, but corrected for multiple comparisons (Benjamini & Hochberg, 1995). For the nuclear
316 datasets, pairwise F_{ST} was estimated on a locus-by-locus basis. Finally, to test for isolation-by-
317 distance, linear regression was used to determine if pairwise Φ_{ST} , neutral F_{ST} , and adaptive F_{ST}
318 increased with coastal distance between sites.

319 ***Genetic Diversity and Effective Population Size***

320 The diversity of mitochondrial sequence data was assessed for each site based on the number of
321 haplotypes, as well as haplotype (h) and nucleotide sequence (π) diversities (Nei, 1987) calculated
322 in ARLEQUIN. For neutral and adaptive nuclear loci, diversity was assessed separately for each site

323 using Nei's gene diversity (H_e ; Nei, 1978) and rarified allelic richness (A_r ; El Mousadik & Petit,
324 1996) using the R packages *hierfstat* (Goudet, 2005) and *poppr* (Kamvar et al., 2014), respectively.
325 For each nuclear diversity estimate, differences among sites were assessed using Friedman's rank-
326 sum test ($\alpha < 0.05$), and Wilcoxon signed-rank tests were used to assess for *post-hoc* pairwise
327 differences ($\alpha < 0.05$), with p -values corrected for multiple comparisons (Benjamini & Hochberg,
328 1995).

329 Contemporary effective population size (N_e) was estimated for each site using the linkage
330 disequilibrium method (Hill, 1981) implemented in NEESTIMATOR (v2.1; Do et al., 2014). To
331 ensure that the effective sample size was the same for each pair of loci, N_e was estimated using
332 1,823 neutral nuclear loci with no missing data. Singleton alleles were also removed for each site.
333 In addition to point estimates, 95% confidence intervals were estimated using a method that
334 jackknifes over individuals (Jones et al., 2016). To account for downward bias resulting from
335 physical linkage among loci, N_e estimates were adjusted based on the haploid number of
336 chromosomes (43; Asahida et al., 1995) for the blacktip shark, following Waples et al. (2016).

337 All figures were produced in R using the package *ggplot2* (Wickham, 2016).

338 **Results**

339 ***ddRAD-Seq Data Filtering***

340 After demultiplexing and trimming, the mean number of HiSeq and MiSeq reads per sample was
341 3,796,003 and 1,121,052, respectively (standard deviation: 2,240,246 and 322,556). Filtering
342 removed 47 individuals with missing data > 20% and 31 individuals with $F_{IS} < -0.13$. Also, one
343 sample was removed from each of 17 pairs of technical replicates confirmed using the dyadic

344 likelihood estimator. After filtering, 424 individuals genotyped at 4,339 SNP-containing loci (1.54
345 SNPs and 2.39 alleles per locus on average) were retained for subsequent analyses.

346 ***Mitochondrial Sequencing and Haplotyping***

347 Sixteen unique mitochondrial haplotypes were identified among 323 individuals, seven of which
348 were previously identified by Keeney et al. (2003, 2005).

349 ***Relatedness***

350 Minimum values of relatedness used to identify siblings, as determined by simulations, were 0.44-
351 0.45 for full-siblings and 0.19-0.20 for half-siblings (Figure S1). No siblings were identified in the
352 Atlantic. Non-randomly sampled siblings included one full-sibling pair in Terra Ceia Bay (eastern
353 Gulf) and a group of six full- and half-siblings in San Antonio Bay (western Gulf; Table S6).
354 Randomly sampled siblings were detected only in Terra Ceia Bay and included three pairs of full-
355 siblings and 15 pairs of half-siblings (Table S7). Notably, three pairs of half-siblings were sampled
356 two years apart and two pairs were sampled four years apart. All other siblings were sampled
357 within the same year or one year apart. Parent-offspring and avuncular relationships can produce
358 similar relatedness values to full- and half-siblings (respectively). However, blacktip sharks do not
359 mature until after four years, and because all kin were sampled within four years, pairs of kin
360 identified in this study are most likely siblings. Mitochondrial haplotypes were assessed for 12
361 pairs of siblings (67%) and two pairs of half-siblings were found to have distinct haplotypes.

362 After an individual from each non-randomly sampled sibling pair was removed, 418 individuals
363 remained, 77% of which (323) were also haplotyped using the mitochondrial control region
364 (Figure 1).

365 ***F_{ST} Outlier Analysis***

366 Zero F_{ST} outliers were detected by *OutFLANK*, BAYESCAN, or ARLEQUIN.

367 ***Spatial and Environmental Analysis***

368 Ten MEMs describing spatial differences were generated based on coastal distances between
369 sampling locations, and the first two MEMs were chosen by model selection: MEM1 (adjusted R^2
370 = 0.00123; $p < 0.01$; Figure 2C) and MEM2 (adjusted R^2 = 0.00188; $p < 0.01$; Figure 2D). The full
371 spatial RDA model and both axes were significant ($p < 0.001$), and linear combinations of MEMs
372 produced three groups (Figure 2A). While MEM1 clustered individuals into Atlantic and Gulf
373 groups, MEM2 divided Gulf individuals into eastern and western groups. Individuals from Mobile
374 Bay – which straddles the boundary between the eastern and western Gulf stock subunits – grouped
375 predominantly with individuals from Florida. Model selection chose two environmental variables
376 with VIF < 3 (Table S8): minimum annual sea surface temperature (°C; adjusted R^2 = 0.00133; p
377 < 0.01; Figure 2F) and mean sea surface salinity in June (unitless; adjusted R^2 = 0.00193; $p < 0.01$;
378 Figure 2G). The full environmental RDA model and both axes were significant ($p < 0.001$), and
379 linear combinations of environmental variables also produced three groups (Figure 2B). Like
380 MEM1, temperature grouped Atlantic and Gulf individuals separately, and salinity split Gulf
381 individuals into two groups; however, in contrast to MEM2, salinity grouped individuals from
382 Mobile Bay with those from the western Gulf. Furthermore, while 69% of loci (9/13) with high
383 loadings for MEM1 also had high loadings for temperature, an additional 15 loci had high loadings
384 only for temperature (Table S9), and structured Mobile Bay and western Gulf sites by latitude.
385 MEM2 and salinity each had 11 loci with high loadings, including six loci for both variables, and
386 a latitudinal pattern was also observed among Florida sites due to salinity.

387 The full pRDA model (i.e., the effect of temperature and salinity adjusted by MEMs 1 and 2) and
388 each axis were significant ($p < 0.05$). Allele loadings resembled a normal distribution (Figure S2)

389 and 68 environmentally-associated loci (1.6%) were identified and removed to produce putatively
390 adaptive (68 loci) and neutral nuclear datasets (4,271 loci).

391 ***Population Structure***

392 For the mitochondrial dataset, heterogeneity was observed among groups (Atlantic and Gulf; Φ_{CT}
393 = 0.0997; $p < 0.05$) and among sites within groups ($\Phi_{SC} = 0.0795$; $p < 0.0001$; Table 1).
394 Heterogeneity was also observed at neutral nuclear loci among groups ($F_{CT} = 0.0015$; $p < 0.0001$)
395 and among sites within groups ($F_{SC} = 0.0006$; $p < 0.001$; Table 1). By contrast, heterogeneity was
396 observed at adaptive nuclear loci among sites within groups ($F_{SC} = 0.0069$; $p < 0.0001$), but not
397 among groups ($F_{CT} = 0.0002$; $p = 0.3641$; Table 1). Within the Gulf, heterogeneity was observed
398 for the mitochondrial (Φ_{ST} : 0.0826 and $p < 0.0001$), neutral nuclear (F_{ST} : 0.0007 and $p < 0.0001$),
399 and adaptive nuclear (F_{ST} : 0.0085 and $p < 0.0001$) datasets based on single-level AMOVA (Table
400 1). By contrast, homogeneity was found in the Atlantic for all three datasets (Φ_{ST} : 0.0246 and $p =$
401 0.1768; neutral F_{ST} : 0.0004 and $p = 0.1871$; adaptive F_{ST} : 0.0006 and $p = 0.4687$; Table 1).

402 *Post-hoc* estimates of pairwise neutral F_{ST} between sites were statistically significant ($p < 0.05$
403 after corrections) for all but two Atlantic-Gulf comparisons (92%; Table S10). A similar, albeit
404 weaker pattern was observed for the mitochondrial dataset, with differences found for 63% of
405 Atlantic-Gulf comparisons (Table S11). Furthermore, the neutral nuclear dataset indicated
406 differences within the Gulf between Terra Ceia and Waccasassa Bays (both eastern Gulf), as well
407 as between Terra Ceia Bay and each of the four sites in the western Gulf (Table S10). However,
408 after excluding siblings randomly sampled in Terra Ceia Bay, the difference with Waccasassa Bay
409 was no longer significant (Table S12). A similar pattern was observed in the Gulf using the
410 mitochondrial data, but in addition to Terra Ceia Bay being different from all four sites in the
411 western Gulf, Apalachicola Bay was significantly different from San Antonio and Corpus Christi

412 Bays (Table S11). Terra Ceia Bay was also significantly different from Waccasassa Bay, but not
413 after the removal of randomly sampled siblings (Table S13). Consequently, estimates of pairwise
414 Φ_{ST} and F_{ST} calculated after removing randomly sampled siblings from Terra Ceia Bay were used
415 to assess for relationships between pairwise genetic differences and coastal distances between
416 sites. Linear regression demonstrated a positive relationship between pairwise coastal distances
417 and genetic differences for eastern and western Gulf sites based on the mitochondrial (adjusted R^2
418 = 0.2360; $p < 0.05$) and neutral nuclear datasets (adjusted R^2 = 0.5168; $p < 0.01$; Figure 2E). By
419 contrast, no such relationship was observed for Atlantic-eastern Gulf nor Atlantic-western Gulf
420 comparisons (Figure S3).

421 For the adaptive dataset, estimates of pairwise F_{ST} were much larger, but statistically significant
422 comparisons were fewer and predominantly observed between Gulf sites with the greatest
423 latitudinal differences (Table S14). For example, Mobile Bay (the most northern Gulf site) was
424 different from all other Gulf sites except Waccasassa and Galveston Bays; Terra Ceia Bay (the
425 most southern Gulf site) was different from all Gulf sites but Waccasassa, San Antonio, and Corpus
426 Christi Bays. Furthermore, in contrast to the mitochondrial and neutral nuclear datasets, linear
427 regression demonstrated a negative relationship between pairwise genetic differences and coastal
428 distances for eastern and western Gulf sites (adjusted R^2 = 0.4428; $p < 0.01$). Consequently, linear
429 regression was then used to determine if pairwise adaptive F_{ST} increased with latitudinal
430 differences between eastern and western Gulf sites, and a positive relationship was observed
431 (adjusted R^2 = 0.4757; $p < 0.01$; Figure 2H).

432 ***Genetic Diversity and Effective Population Size***

433 Each Atlantic site had fewer mitochondrial haplotypes (3-4) than all but one Gulf site (Mobile
434 Bay; 4), and haplotype and nucleotide diversities were lower in Atlantic sites than in all Gulf sites

435 (Table 2). Though similar numbers of haplotypes were observed within the eastern (4-7) and
436 western Gulf (5-9), haplotype and nucleotide diversities were greater in the western Gulf. For the
437 neutral and adaptive nuclear datasets, gene diversity (H_e) and allelic richness (A_r) differed among
438 the 11 sites ($p < 0.0001$; Table 2). Estimated neutral H_e was lowest in Port Royal Sound (0.1537;
439 Atlantic) and smaller ($p < 0.05$) than all sites except St. Helena Sound (Atlantic); estimated neutral
440 H_e was greatest in San Antonio Bay (0.1584; western Gulf) and greater ($p < 0.05$) than all but three
441 Gulf sites (i.e., Waccasassa, Mobile, and Corpus Christi Bays). Estimated adaptive H_e was lowest
442 in San Antonio (0.1370; western Gulf) and greater in Mobile Bay (0.2076; eastern Gulf) than all
443 other sites ($p < 0.001$.) Estimated neutral A_r was lowest in Port Royal Sound (2.8174; Atlantic)
444 and lower ($p < 0.05$) than three Gulf sites (i.e., Mobile, Galveston, and San Antonio Bays);
445 estimated neutral A_r was greatest in San Antonio Bay (2.8545; western Gulf) and greater ($p < 0.05$)
446 than six sites. Estimated adaptive A_r was lowest in San Antonio Bay (2.8332; western Gulf) and
447 greater in Mobile Bay (3.5343; eastern Gulf) than all other sites ($p < 0.001$).

448 While finite upper and point N_e estimates were obtained for only one and six sites, respectively
449 (Table S15), lower N_e estimates were obtained for all sites and varied considerably, with no
450 obvious pattern among regions (Figure 3).

451 Discussion

452 The results of this study highlight how philopatry can influence genetic population structure at
453 multiple spatial scales by restricting gene flow and facilitating the sorting of adaptive variants by
454 selection. Mitochondrial and neutral genetic structure indicated that blacktip sharks in the U.S.
455 Atlantic and Gulf of Mexico constitute three genetically distinct units with little to no gene flow
456 between them. Structure within Gulf units at putatively adaptive loci, correlated with variation in
457 sea surface temperature and salinity, suggested local adaptation to environmental conditions.

458 Instances of parturition site fidelity were documented based on the sampling of maternally-related
459 siblings, and if this behavior extends across generations (i.e., natal philopatry), it could contribute
460 to the observed patterns of adaptive structure.

461 ***Neutral Genetic Structure***

462 Results from mitochondrial and neutral nuclear data confirm that blacktip sharks in the Atlantic
463 and Gulf are genetically distinct. The first MEM of the spatial RDA grouped Atlantic and Gulf
464 individuals separately (Figure 2A) and genetic structure was also observed between these groups
465 based on hierarchical AMOVA and *post-hoc* estimates of Φ_{ST} and F_{ST} between sites. In addition,
466 genetic diversity was generally lower in the Atlantic than in the Gulf. The finding of genetically
467 distinct blacktip shark units in the Atlantic and Gulf is consistent with previous assessments of
468 mitochondrial DNA (Keeney et al., 2003, 2005) and life history traits such as maximum length
469 and growth rate (Carlson et al., 2006). This observation is also consistent with studies of other
470 marine fishes (Gold et al., 2009; Leidig et al., 2015; Seyoum et al., 2017), including coastal sharks
471 (Dimens et al., 2019; Portnoy et al., 2015, 2016), and aligns with the Florida Vicariance Zone
472 (Neigel, 2009), where constriction of the continental shelf from Miami to West Palm Beach has
473 reduced nearshore habitat (Avise, 1992; Neigel, 2009). Consequently, suitable parturition sites for
474 coastal sharks are lacking in southeastern Florida and may dissuade female movement across the
475 vicariance zone. Although gene flow via males should be less affected, tagging data suggest that
476 male blacktip sharks do not move between the Atlantic and Gulf either (Kohler & Turner, 2019),
477 thus additional factors likely limit connectivity.

478 Neutral genetic structure was also found within the Gulf, but not within the Atlantic. YOY blacktip
479 sharks occupy U.S. Atlantic estuaries from northern Florida to southern North Carolina (Castro,
480 1996; McCallister et al., 2013), so the lack of observed structure in the Atlantic could be due to

481 limited spatial sampling. For the Gulf, single-level AMOVA indicated heterogeneity, and
482 differences in both pairwise Φ_{ST} and F_{ST} were observed between the most eastern and the four
483 western sites. This could indicate an isolation-by-distance effect (Laikre et al., 2005), which is
484 supported by positive relationships between pairwise Φ_{ST}/F_{ST} and coastal distances (Figure 2E).
485 However, the spatial RDA clustered Gulf individuals into eastern and western groups, with
486 individuals from Mobile Bay grouping predominantly with those from Florida (Figure 2A). This
487 division aligns with a biogeographic break in the northern Gulf (McClure & McEachran, 1992),
488 centered on an area of transition from carbonate sediments in the east to mostly terrigenous
489 sediments in the west (McClure & McEachran, 1992; Neigel, 2009). Further, low salinity outflows
490 from the Mississippi and Atchafalaya rivers to the west of Mobile Bay could act as a barrier to
491 gene flow for blacktip sharks. This has been suggested for other stenohaline sharks in the Gulf
492 (Portnoy et al., 2014), as well as a variety of marine species around the world (Rocha, 2003; Volk
493 et al., 2021). In addition, spatial RDA and estimates of pairwise Φ_{ST} and F_{ST} are consistent with
494 the idea that straying by females occurs mostly among neighboring parturition sites, as suggested
495 by other studies of coastal sharks (Duncan et al., 2006; Keeney et al., 2003). Nevertheless, it should
496 be noted that this study did not include sites between Mobile and Galveston Bays because a
497 sufficient number of samples could not be collected in Louisiana. Thus, neutral structure in the
498 Gulf may follow an isolation-by-distance pattern and the lack of samples from Louisiana could
499 have facilitated the finding of discrete eastern and western Gulf groups by spatial RDA. The pattern
500 of neutral structure documented by this study has been observed in multiple marine fishes in the
501 northern Gulf (Karlsson et al., 2009; Portnoy et al., 2014; Seyoum et al., 2018). In particular, the
502 results are similar to those of a genomic assessment of red drum (*Sciaenops ocellatus*; Hollenbeck
503 et al., 2019), which do not give live birth but display spawning site fidelity to estuaries to which

504 juveniles recruit after the larval period (Lowerre-Barbieri et al., 2019; Matlock, 1990). This is in
505 contrast with the patterns seen in genomics studies of two species that spawn offshore, red snapper
506 (*Lutjanus campechanus*; Portnoy et al., 2021) and southern flounder (*Paralichthys lethostigma*;
507 O'Leary et al., 2021), and suggests that habitat use may be an important predictor of genetic
508 structure for fishes of the Gulf of Mexico.

509 A previous assessment of blacktip shark genetic structure found differences among the Atlantic,
510 eastern, and western Gulf in mitochondrial DNA, but not nuclear DNA, and the authors
511 hypothesized that this reflected female regional philopatry and male-mediated gene flow (Keeney
512 et al., 2005). While this study found similar patterns of mitochondrial DNA structure among the
513 Atlantic, eastern, and western Gulf, heterogeneity was also detected among these regions at neutral
514 nuclear loci. Inconsistencies in the observed patterns of neutral nuclear structure are likely due to
515 the greater resolution offered by thousands of SNP-containing loci as compared to the eight
516 microsatellite loci used by Keeney et al. (2005). Thus, it appears that male blacktip sharks also
517 display regional philopatry, or that male-mediated gene flow is insufficient to homogenize allele
518 frequencies among these regions. Evidence of male regional philopatry is noteworthy because it
519 suggests that the widespread notion of male-biased dispersal in elasmobranchs – which developed
520 from mixed-marker studies typically using microsatellites and mitochondrial DNA – may be
521 overemphasized (Phillips et al., 2021). Taken together, the results suggest that regional philopatry
522 by both male and female blacktip sharks has contributed to the formation of genetically distinct
523 units in the Atlantic, eastern Gulf, and western Gulf that align well with the current stock
524 subregions defined by NOAA Fisheries.

525 ***Adaptive Genetic Structure***

526 Genetic structure at putatively adaptive loci was observed on a more localized scale in the Gulf.
527 Environmental RDA structured Gulf individuals into two groups along latitudinal gradients based
528 on minimum annual temperature and mean salinity in June, and in contrast to spatial RDA, Mobile
529 Bay individuals grouped with those from Texas (Figure 2B). These groups correspond with a
530 transition in environmental conditions and a break in the coastal shark assemblage of the northern
531 Gulf that has been described by multiple studies (Bethea et al., 2015; Drymon et al., 2020).
532 Significant pairwise F_{ST} estimates based on adaptive loci were observed between sites within each
533 group, and the greatest F_{ST} values were observed between sites with the greatest latitudinal
534 differences (Figure 2H), indicating local adaptation among parturition sites. Furthermore,
535 estimates of adaptive H_e and A_r were highly elevated in Mobile Bay, which could be related to the
536 spatial and temporal environmental heterogeneity that characterizes this estuary (Kim & Park,
537 2012; Orlando Jnr et al., 1993). However, Mobile Bay is proximal to a marine-suture zone (Portnoy
538 & Gold, 2012), an area of overlap between biotic assemblages (Remington, 1968), so greater
539 diversity could also reflect contact between the eastern and western Gulf.

540 While the lack of a suitable reference genome precludes assessments of putative function, aspects
541 of blacktip shark biology provide potential explanations for the fine-scale adaptive structure
542 observed here. Adaptive differences associated with minimum annual temperature could reflect
543 temporal variation in YOY migration out of parturition sites when waters cool in the autumn. Sea
544 surface temperatures in Gulf estuaries are colder seasonally in the north than in the south and can
545 vary considerably due to a variety of climatic factors. A gradient exists along the Texas coast
546 because temperature differences are predominantly influenced by seasonal heat flux and river
547 discharges (Portela et al., 2018), whereas differences along the Gulf coast of Florida appear less
548 stark. Blacktip sharks born in Terra Ceia Bay were thought to remain until late October to late

549 November, with emigration following dramatic decreases in water temperature (1.5-2°C) to
550 approximately 21°C (Heupel, 2007). However, it now appears that some remain until January
551 because water temperatures do not decrease sufficiently until then (Goldner, 2022). If there is a
552 fitness cost to a shorter residency period, local adaptation could lead to individuals born in estuaries
553 further north being more tolerant of lower temperatures. However, blacktip shark emigration from
554 an Atlantic coast estuary (i.e., Bulls Bay, South Carolina) also coincides with ~21°C (Castro,
555 1996). Therefore, it appears that similar temperature changes stimulate emigration, and blacktip
556 sharks born in more northern Gulf estuaries should migrate earlier in the year when those
557 temperatures are reached. This is observed along the Texas coast where YOY blacktip sharks are
558 found in Corpus Christi Bay until mid-November (Matich et al., 2021), weeks after they have
559 emigrated from Galveston Bay (Matich and Texas Parks and Wildlife unpublished data). Likewise,
560 the species is mostly absent in Mobile Bay after October (Parsons & Hoffmayer, 2007). A similar
561 pattern of migratory timing is seen when Atlantic salmon (*Salmo salar*) leave nurseries in the
562 spring/summer (Hodgson & Quinn, 2002; Hvidsten et al., 1998), with individuals from southern
563 habitats migrating weeks before those born further north because the temperatures that stimulate
564 emigration are reached earlier (Otero et al., 2014; Vollset et al., 2021).

565 Salinities also vary among Gulf estuaries and adaptive differences associated with mean salinity
566 in June – immediately after the peak period of parturition (Baremore & Passerotti, 2013) – could
567 indicate local adaptation based on salinity tolerance. Peninsular Florida estuaries are relatively
568 saline because conditions are predominantly influenced by precipitation, with little freshwater
569 inflow compared with estuaries to the west. Conditions are less saline in the Florida panhandle due
570 to lower evaporation rates and freshwater discharge from the Apalachicola, Chattahoochee, and
571 Flint rivers that flow into Apalachicola Bay (Orlando Jnr et al., 1993). Mobile Bay is relatively

572 hyposaline because of the large freshwater influx via the Mobile River (Orlando Jnr et al., 1993),
573 and June salinities in Texas estuaries are similar to Mobile Bay because precipitation is greatest in
574 May (TexasET, 2022). Also, the major river systems (e.g., Mobile, Mississippi, Rio Grande) that
575 drain into the Gulf are distributed from Alabama to the border with Mexico (USGS, 1990).
576 Nonetheless, a salinity gradient exists along the Texas coast because estuaries in the north receive
577 hyposaline waters from the central Gulf via westerly currents, while isolated freshwater pulses
578 lead to more saline conditions in the south (Orlando Jnr et al., 1993). Consequently, blacktip sharks
579 born in estuaries on the lower Texas coast may experience higher salinities, consistent with the
580 conditions at which individuals have been captured in Corpus Christi (mean: 25.0-33.4) and
581 Galveston Bays (mean: 16.1-22.3; (Matich et al., 2017)). By contrast, the species has been captured
582 in Mobile Bay at salinities as low as 11 (Parsons & Hoffmayer, 2007) and is usually found at
583 salinities of 22.3-34.2 in Florida estuaries (Bethea et al., 2009).

584 A limitation of this study is that the available data sources provide insufficient resolution to
585 describe environmental variation within estuaries. The MARSPEC and BIO-ORACLE databases
586 reflect coastal conditions for which differences are predominantly driven by latitude, and
587 consequently, environmental heterogeneity among the sites is underestimated. Additionally, the
588 environmental measurements are unable to account for habitat usage by blacktip sharks because
589 these individuals are highly mobile, only use a subset of the available estuarine habitat, and move
590 with environmental conditions (Froeschke et al., 2010). Even so, the environmental RDA shows
591 latitudinal gradients in both the eastern and western Gulf, thus the results may reflect local
592 adaptation to conditions that are not described by the environmental data but also vary with
593 latitude.

594 ***Philopatry and Local Adaptation***

595 Two pairs of half-siblings had distinct mitochondrial haplotypes, demonstrating they had different
596 mothers and thus were paternally related. For one pair, both siblings were sampled (born) in the
597 same year, meaning a male reproduced with two females in the same breeding season that each
598 gave birth in Terra Ceia Bay (eastern Gulf). This suggests that breeding sites may be proximal to
599 parturition sites, which is consistent with what is understood about breeding locations in the U.S.
600 Atlantic (Castro, 1996). For the other sibling pair, the individuals were sampled (born) one year
601 apart, providing direct evidence that a male blacktip shark reproduced in the eastern Gulf in
602 consecutive breeding seasons. This observation suggests that male blacktip sharks might display
603 breeding site fidelity and is consistent with the indirect evidence of male regional philopatry based
604 on patterns of neutral genetic structure.

605 Five pairs of half-siblings with the same mitochondrial haplotypes were captured two and four
606 years apart in Terra Ceia Bay, accordant with the biennial reproductive period of female blacktip
607 sharks (Baremore & Passerotti, 2013; Castro, 1996). This implies that five females re-used the
608 habitat for parturition. It is important to note that all randomly sampled siblings were detected in
609 Terra Ceia Bay, and N_e estimates indicated that the number of breeders using this habitat is much
610 smaller than in other sites (Figure 3). Hence, blacktip sharks may exhibit parturition site fidelity
611 to additional estuaries, but the behavior may be easier to detect in Terra Ceia Bay because there is
612 a higher probability of catching siblings. Females that re-use the same estuary for parturition
613 display a strong degree of habitat fidelity, but for this behavior to constitute natal philopatry, the
614 estuary that is re-used must be the habitat in which females were born. Multiple studies have
615 demonstrated that sharks can navigate to their place of birth (Edrén & Gruber, 2005; O'Gower,
616 1995; Sundström et al., 2001), including blacktip sharks (Gardiner et al., 2015; Heupel et al.,
617 2003), and while natal philopatry has been speculated to occur in this species (Hueter et al., 2005),

618 the behavior has been demonstrated directly only in the lemon shark (*Negaprion brevirostris*) in
619 Bimini, The Bahamas (Feldheim et al., 2014). This was possible because lemon sharks in Bimini
620 are captured in a nearly exhaustive manner, relatively few females give birth there, and genetic
621 profiling has been ongoing for decades (Feldheim et al., 2004; Gruber et al., 2001; Postaire et al.,
622 2022). The results presented here indicate that long-term studies focused on identifying kin among
623 blacktip sharks in Terra Ceia Bay may demonstrate a second example of natal philopatry by coastal
624 sharks.

625 While the observation of three genetically distinct units in the Atlantic and Gulf suggests male and
626 female blacktip sharks reproduce in the region of their birth (i.e., regional philopatry), this behavior
627 cannot explain the fine-scale adaptive structure observed within Gulf units. Adaptive variation
628 could sort among neighboring estuaries if alleles adapted to local conditions conferred phenotypes
629 with greater fitness and matrilines carrying these alleles re-used the same estuaries as parturition
630 sites in subsequent generations (i.e., natal philopatry). Under this scenario, YOY with phenotypes
631 locally adapted to their parturition site would have a higher probability of surviving and
632 reproducing. Differential selection pressures among parturition sites would drive selection for
633 locally adapted phenotypes and overcome gene flow of maladapted alleles from neighboring
634 estuaries via patrilines and/or female straying. Given the heterogeneity in conditions like
635 temperature and salinity among Gulf estuaries and the high rates of mortality experienced by YOY
636 blacktip sharks (Heupel & Simpfendorfer, 2002), directional selection and natal philopatry could
637 facilitate the sorting of adaptive alleles (Kawecki & Ebert, 2004), generating the patterns of
638 adaptive structure observed in this study.

639 **Conclusions**

640 The genetic structure found among parturition sites within management units highlights the
641 importance of policies that focus on the preservation of adaptive variation (Funk et al., 2012).
642 Estuaries in which progeny are born and/or reside as juveniles are considered essential because
643 they are fundamental to life cycles (Fluharty, 2000), but if neighboring habitats are
644 environmentally heterogeneous and sources of novel adaptive variants, it may be necessary to
645 individually evaluate their contributions to ensure future persistence (Stiebens et al., 2013). These
646 considerations are particularly important for species displaying fine-scale philopatry because the
647 loss of certain habitats could lead to irreversible declines in recruitment and adaptive potential
648 (Hess et al., 2013; Hueter et al., 2005). Furthermore, as environmental conditions continue to shift
649 with climate change, the capability of organisms to adapt and persist will depend on existing
650 genetic variation and levels of gene flow among habitats.

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665 **References**

666 Allendorf, F. W., Hohenlohe, P. A., & Luikart, G. (2010). Genomics and the future of
667 conservation genetics. *Nature Reviews Genetics*, 11(10), 697–709.
668 <https://doi.org/10.1038/nrg2844>

669 Asahida, T., Ida, H., & Hayashizaki, K. (1995). Karyotypes and Cellular DNA Contents of Some
670 Sharks in the Order Carcharhiniformes. *Japanese Journal of Ichthyology*, 42(1), 21–26.

671 Assis, J., Tyberghein, L., Bosch, S., Verbruggen, H., Serrão, E. A., & De Clerck, O. (2018). Bio-
672 ORACLE v2.0: Extending marine data layers for bioclimatic modelling. *Global Ecology*
673 and Biogeography

674 and Biogeography

675 and Biogeography

676 and Biogeography

677 and Biogeography

678 and Biogeography

679 and Biogeography

680 and Biogeography

681 and Biogeography

682 and Biogeography

683 and Biogeography

684 and Biogeography

Avise, J. C. (1992). Molecular population structure and the biogeographic history of a regional fauna: A case history with lessons for conservation biology. *Oikos*, 63, 62–76.

Baremore, I. E., & Passerotti, M. S. (2013). Reproduction of the Blacktip Shark in the Gulf of Mexico. *Marine and Coastal Fisheries*, 5(1), 127–138.

<https://doi.org/10.1080/19425120.2012.758204>

Beaumont, M., & Nichols, R. (1996). Evaluating loci for use in the genetic analysis of population structure. *Proceedings of the Royal Society of London Series B*, 263, 1619–1626.

Benjamini, Y., & Hochberg, Y. (1995). Controlling the false discovery rate: A practical and powerful approach to multiple testing. *Journal of the Royal Statistical Society, Series B (Methodological)*, 57, 289–300.

685 Bernatchez, L. (2016). On the maintenance of genetic variation and adaptation to environmental
686 change: Considerations from population genomics in fishes. *Journal of Fish Biology*,
687 89(6), 2519–2556. <https://doi.org/10.1111/jfb.13145>

688 Bethea, D. M., Ajemian, M. J., Carlson, J. K., Hoffmayer, E. R., Imhoff, J. L., Grubbs, R. D.,
689 Peterson, C. T., & Burgess, G. H. (2015). Distribution and community structure of
690 coastal sharks in the northeastern Gulf of Mexico. *Environmental Biology of Fishes*,
691 98(5), 1233–1254. <https://doi.org/10.1007/s10641-014-0355-3>

692 Bethea, D. M., Hollensead, L., Carlson, J. K., Ajemian, M. J., Grubbs, R. D., Hoffmayer, E. R.,
693 Del Rio, R., Peterson, G. W., Baltz, D. M., & Romine, J. (2009). *Shark nursery grounds*
694 and *essential fish habitat studies: Gulfspan Gulf of Mexico FY'08—Cooperative Gulf of*
695 *Mexico states shark pupping and nursery survey*. National Fish and Wildlife Service,
696 Sustainable Fisheries Division, Highly Migratory Species Division, National Marine
697 Fisheries Service.

698 Blanchet, F. G., Legendre, P., & Borcard, D. (2008). Forward Selection Of Explanatory
699 Variables. *Ecology*, 89(9), 2623–2632. <https://doi.org/10.1890/07-0986.1>

700 Bosch, S., & Fernandez, S. (2021). *Sdmpredictors: Species Distribution Modelling Predictor*
701 *Datasets. R package version 0.2.10*. <https://cran.r-project.org/package=sdmpredictors>

702 Bowen, B. W., & Roman, J. (2005). Gaia's Handmaidens: The Orlog Model for Conservation
703 Biology. *Conservation Biology*, 19(4), 1037–1043. [1739.2005.00100.x](https://doi.org/10.1111/j.1523-
704 1739.2005.00100.x)

705 Capblancq, T., & Forester, B. R. (2021). Redundancy Analysis (RDA): A Swiss Army knife for
706 landscape genomics. *Methods in Ecology and Evolution*. [210X.13722](https://doi.org/10.1111/2041-
707 210X.13722)

708 Carlson, J. K., Sulikowski, J. R., & Baremore, I. E. (2006). Do differences in life history exist for
709 blacktip sharks, *Carcharhinus limbatus*, from the United States South Atlantic Bight and
710 Eastern Gulf of Mexico? *Environmental Biology of Fishes*, 77(3–4), 279–292.
711 <https://doi.org/10.1007/s10641-006-9129-x>

712 Castro, J. I. (1993). The shark nursery of Bulls Bay, South Carolina, with a review of the shark
713 nurseries of the southeastern coast of the United States. *Environmental Biology of Fishes*,
714 38(1–3), 37–48. <https://doi.org/10.1007/BF00842902>

715 Castro, J. I. (1996). Biology of the Blacktip Shark, *Carcharhinus limbatus*, off the Southeastern
716 United States. *Bulletin of Marine Science*, 59(3), 508–522.

717 Catchen, J. M., Amores, A., Hohenlohe, P., Cresko, W., & Postlethwait, J. H. (2011). Stacks:
718 Building and Genotyping Loci De Novo From Short-Read Sequences. *G3: Genes/Genomes/Genetics*, 1, 171–182. <https://doi.org/10.1534/g3.111.000240>

720 Chapman, D. D., Feldheim, K. A., Papastamatiou, Y. P., & Hueter, R. E. (2015). There and Back
721 Again: A Review of Residency and Return Migrations in Sharks, with Implications for
722 Population Structure and Management. *Annual Review of Marine Science*, 7(1), 547–570.
723 <https://doi.org/10.1146/annurev-marine-010814-015730>

724 Clement, M., Posada, D., & Crandall, K. A. (2000). TCS: a computer program to estimate gene
725 genealogies. *Molecular Ecology*, 9(10), 1657–1659. [294x.2000.01020.x](https://doi.org/10.1046/j.1365-
726 294x.2000.01020.x)

727 Compagno, L., Dando, M., & Fowler, S. (2005). *Sharks of the World*. Princeton Press.

728 Conrath, C. L., & Musick, J. A. (2012). Reproductive Biology of Elasmobranchs. In J. C.
729 Carrier, J. A. Musick, & M. R. Heithaus (Eds.), *Biology of Sharks & their Relatives*. CRC
730 Press.

731 Cormack, R., Hartl, D., & Clark, A. (1990). Principles of Population Genetics. *Biometrics*, 46(2),
732 546–547. <https://doi.org/10.2307/2531471>

733 Danecek, P., Auton, A., Abecasis, G., Albers, C. A., Banks, E., Depristo, M. A., Handsaker, R.
734 E., Lunter, G., Marth, G. T., Sherry, S. T., McVean, G., Durbin, R., Project, G., & 1000
735 Genomes Project Analysis Group (2011). The variant call format and VCFtools.
736 *Bioinformatics*, 27(15), 2156–2158. <https://doi.org/10.1093/bioinformatics/btr330>

737 Dimens, P. V., Willis, S., Grubbs, R. D., & Portnoy, D. S. (2019). A genomic assessment of
738 movement and gene flow around the South Florida vicariance zone in the migratory
739 coastal blacknose shark, *Carcharhinus acronotus*. *Marine Biology*, 166, 86.
740 <https://doi.org/10.1007/s00227-019-3533-1>

741 Do, C., Waples, R. S., Peel, D., Macbeth, G. M., Tillett, B. J., & Ovenden, J. R. (2014).
742 NeEstimator v2: Re-implementation of software for the estimation of contemporary
743 effective population size (N_e) from genetic data. *Molecular Ecology Resources*, 14(1),
744 209–214. <https://doi.org/10.1111/1755-0998.12157>

745 Dray, S., Bauman, D., Blanchet, G., Borcard, D., Clappe, S., Guenard, G., Jombart, T., Larocque,
746 G., Legendre, P., & Madi, N. (2019). *Adespatial: Multivariate multiscale spatial*
747 *analysis. R package version 0.3-7.*

748 Dray, S., Legendre, P., & Peres-Neto, P. R. (2006). Spatial modelling: A comprehensive
749 framework for principal coordinate analysis of neighbour matrices (PCNM). *Ecological*
750 *Modelling*, 196(3), 483–493. <https://doi.org/10.1016/j.ecolmodel.2006.02.015>

751 Drymon, J. M., Dedman, S., Froeschke, J. T., Seubert, E. A., Jefferson, A. E., Kroetz, A. M.,
752 Mareska, J. F., & Powers, S. P. (2020). Defining Sex-Specific Habitat Suitability for a

753 Northern Gulf of Mexico Shark Assemblage. *Frontiers in Marine Science*, 7(February),
754 1–18. <https://doi.org/10.3389/fmars.2020.00035>

755 Dulvy, N. K., Simpfendorfer, C. A., Davidson, L. N. K., Fordham, S. V., Bräutigam, A., Sant,
756 G., & Welch, D. J. (2017). Challenges and Priorities in Shark and Ray Conservation.
757 *Current Biology*, 27(11), 565–572. <https://doi.org/10.1016/j.cub.2017.04.038>

758 Duncan, K., Martin, A., Bowen, B., & De Couet, H. (2006). Global phylogeography of the
759 scalloped hammerhead shark (*Sphyrna lewini*). *Molecular Ecology*, 15(8), 2239–2251.
760 <https://doi.org/10.1111/j.1365-294X.2006.02933.x>

761 Edrén, S. M. C., & Gruber, S. H. (2005). Homing ability of young lemon sharks, *Negaprion*
762 *brevirostris*. *Environmental Biology of Fishes*, 72(3), 267–281.
763 <https://doi.org/10.1007/s10641-004-2583-4>

764 El Mousadik, A., & Petit, R. J. (1996). High level of genetic differentiation for allelic richness
765 among populations of the argan tree [*Argania spinosa* (L.) Skeels] endemic to Morocco.
766 *Theoretical and Applied Genetics*, 92(7), 832–839. <https://doi.org/10.1007/BF00221895>

767 Excoffier, L., & Lischer, H. (2010). Arlequin suite ver 3.5: A new series of programs to perform
768 population genetics analyses under Linux and Windows. *Molecular Ecology Resources*,
769 10, 564–567.

770 Excoffier, L., Smouse, P. E., & Quattro, J. M. (1992). Analysis of molecular variance inferred
771 from metric distances among DNA haplotypes: Application to human mitochondrial
772 DNA restriction data. *Genetics*, 131(2), 479–491.
773 <https://doi.org/10.1093/genetics/131.2.479>

774 Feldheim, K. A., Gruber, S. H., & Ashley, M. V. (2004). Reconstruction of parental
775 microsatellite genotypes reveals female polyandry and philopatry in the lemon shark,
776 *Negaprion brevirostris*. *Evolution*, 58, 2332–2342.

777 Feldheim, K. A., Gruber, S. H., DiBattista, J. D., Babcock, E. A., Kessel, S. T., Hendry, A. P.,
778 Pikitch, E. K., Ashley, M. V., & Chapman, D. D. (2014). Two decades of genetic
779 profiling yields first evidence of natal philopatry and long-term fidelity to parturition sites
780 in sharks. *Molecular Ecology*, 23(1), 110–117. <https://doi.org/10.1111/mec.12583>

781 Fischer, M. C., Foll, M., Excoffier, L., & Heckel, G. (2011). Enhanced AFLP genome scans
782 detect local adaptation in high-altitude populations of a small rodent (*Microtus arvalis*).
783 *Molecular Ecology*, 20(7), 1450–1462. <https://doi.org/10.1111/j.1365-294X.2011.05015.x>

785 Flowers, K. I., Ajemian, M. J., Bassos-Hull, K., Feldheim, K. A., Hueter, R. E., Papastamatiou,
786 Y. P., & Chapman, D. D. (2016). A review of batoid philopatry, with implications for
787 future research and population management. *Marine Ecology Progress Series*, 562, 251–
788 261. <https://doi.org/10.3354/meps11963>

789 Fluharty, D. (2000). Habitat protection, ecological issues, and implementation of the Sustainable
790 Fisheries Act. *Ecological Applications*, 10(2), 325–337. [https://doi.org/10.1890/1051-0761\(2000\)010\[0325:HPEIAI\]2.0.CO;2](https://doi.org/10.1890/1051-0761(2000)010[0325:HPEIAI]2.0.CO;2)

792 Foll, M., Fischer, M. C., Heckel, G., & Excoffier, L. (2010). Estimating population structure
793 from AFLP amplification intensity. *Molecular Ecology*, 19(21), 4638–4647.
794 <https://doi.org/10.1111/j.1365-294X.2010.04820.x>

795 Foll, M., & Gaggiotti, O. (2008). A genome-scan method to identify selected loci appropriate for
796 both dominant and codominant markers: A Bayesian perspective. *Genetics*, 180(2), 977–
797 993. <https://doi.org/10.1534/genetics.108.092221>

798 Forester, B. R., Lasky, J. R., Wagner, H. H., & Urban, D. L. (2018). Comparing methods for
799 detecting multilocus adaptation with multivariate genotype–environment associations.
800 *Molecular Ecology*, 27(9), 2215–2233. <https://doi.org/10.1111/mec.14584>

801 Froeschke, J., Stunz, G. W., & Wildhaber, M. L. (2010). Environmental influences on the
802 occurrence of coastal sharks in estuarine waters. *Marine Ecology Progress Series*, 407,
803 279–292. <https://doi.org/10.3354/meps08546>

804 Funk, W., McKay, J., Hohenlohe, P., & Allendorf, F. (2012). Harnessing genomics for
805 delineating conservation units. *Trends in Ecology & Evolution*, 27, 489–496.

806 Gagnaire, P., Broquet, T., Aurelle, D., Viard, F., Souissi, A., Bonhomme, F., Arnaud-Haond, S.,
807 & Bierne, N. (2015). Using neutral, selected, and hitchhiker loci to assess connectivity of
808 marine populations in the genomic era. *Evolutionary Applications*, 8(8), 769–786.

809 Gagnaire, P.-A., & Gaggiotti, O. E. (2016). Detecting polygenic selection in marine populations
810 by combining population genomics and quantitative genetics approaches. *Current
811 Zoology*, 62(6), 603–616. <https://doi.org/10.1093/cz/zow088>

812 Garant, D., Forde, S. E., & Hendry, A. P. (2007). The multifarious effects of dispersal and gene
813 flow on contemporary adaptation. *Functional Ecology*, 21(3), 434–443.
814 <https://doi.org/10.1111/j.1365-2435.2006.01228.x>

815 Gardiner, J. M., Whitney, N. M., & Hueter, R. E. (2015). Smells like home: The role of olfactory
816 cues in the homing behavior of blacktip sharks, *Carcharhinus limbatus*. *Integrative and
817 Comparative Biology*, 55(3), 495–506. <https://doi.org/10.1093/icb/icv087>

818 Gold, J., Saillant, E., Ebelt, N., & Lem, S. (2009). Conservation genetics of gray snapper
819 (*Lutjanus griseus*) in U.S. waters of the northern Gulf of Mexico and western Atlantic
820 Ocean. *Copeia*, 2009, 277–286.

821 Goldner, V. (2022). *The Effects of Climate Change on the Migration Phenology of Juvenile*
822 *Blacktip Sharks (Carcharhinus limbatus)* (Unpublished thesis). New College of Florida.

823 Goudet, J. (2005). Hierfstat, a package for R to compute and test hierarchical *F*-statistics.
824 *Molecular Ecology Notes*, 5(1), 184–186. <https://doi.org/10.1111/j.1471-8286.2004.00828.x>

825

826 Gruber, S. H., De Marignac, J. R. C., & Hoenig, J. M. (2001). Survival of Juvenile Lemon
827 Sharks at Bimini, Bahamas, Estimated by Mark–Depletion Experiments. *Transactions of*
828 *the American Fisheries Society*, 130(3), 376–384. [https://doi.org/10.1577/1548-8659\(2001\)130<0376:sojlsa>2.0.co;2](https://doi.org/10.1577/1548-8659(2001)130<0376:sojlsa>2.0.co;2)

829

830 Grummer, J. A., Beheregaray, L. B., Bernatchez, L., Hand, B. K., Luikart, G., Narum, S. R., &
831 Taylor, E. B. (2019). Aquatic Landscape Genomics and Environmental Effects on
832 Genetic Variation. *Trends in Ecology and Evolution*, 34(7), 641–654.
833 <https://doi.org/10.1016/j.tree.2019.02.013>

834 Hall, T. A. (1999). BioEdit: A User-Friendly Biological Sequence Alignment Editor and
835 Analysis Program for Windows 95/98/NT. *Nucleic Acids Symposium*, 41, 95–98.

836 Hess, J. E., Campbell, N. R., Close, D. A., Docker, M. F., & Narum, S. R. (2013). Population
837 genomics of Pacific lamprey: Adaptive variation in a highly dispersive species.
838 *Molecular Ecology*, 22(11), 2898–2916. <https://doi.org/10.1111/mec.12150>

839 Heupel, M. R. (2007). Exiting Terra Ceia Bay: Examination of cues stimulating migration from a
840 summer nursery area. *American Fisheries Society Symposium*, 50, 265–280.

841 Heupel, M. R., Carlson, J. K., & Simpfendorfer, C. A. (2007). Shark nursery areas: Concepts,
842 definition, characterization and assumptions. *Marine Ecology Progress Series*, 337, 287–
843 297. <https://doi.org/10.3354/meps337287>

844 Heupel, M. R., Simpfendorfer, C. A., & Hueter, R. E. (2003). Running before the storm:
845 Blacktip sharks respond to falling barometric pressure associated with Tropical Storm
846 Gabrielle. *Journal of Fish Biology*, 63(5), 1357–1363. <https://doi.org/10.1046/j.1095->
847 8649.2003.00250.x

848 Heupel, M. R., Simpfendorfer, C. A., & Hueter, R. E. (2004). Estimation of Shark Home Ranges
849 using Passive Monitoring Techniques. *Environmental Biology of Fishes*, 71(2), 135–142.
850 <https://doi.org/10.1023/b:ebfi.0000045710.18997.f7>

851 Heupel, M. R., & Simpfendorfer, C. A. (2002). Estimation of mortality of juvenile blacktip
852 sharks, *Carcharhinus limbatus*, within a nursery area using telemetry data. *Canadian*
853 *Journal of Fisheries and Aquatic Sciences*, 59, 624–632.

854 Hill, W. G. (1981). Estimation of effective population size from data on linkage disequilibrium.
855 *Genetical Research*, 38, 209–216.

856 Hoban, S., Kelley, J. L., Lotterhos, K. E., Antolin, M. F., Bradburd, G., Lowry, D. B., Poss, M.
857 L., Reed, L. K., Storfer, A., & Whitlock, M. C. (2016). Finding the genomic basis of
858 local adaptation: Pitfalls, practical solutions, and future directions. *American Naturalist*,
859 188(4), 379–397. <https://doi.org/10.1086/688018>

860 Hodgson, S., & Quinn, T. P. (2002). The timing of adult sockeye salmon migration into fresh
861 water: Adaptations by populations to prevailing thermal regimes. *Canadian Journal of*
862 *Zoology*, 80(3), 542–555. <https://doi.org/10.1139/z02-030>

863 Hoey, J. A., & Pinsky, M. L. (2018). Genomic signatures of environmental selection despite
864 near-panmixia in summer flounder. *Evolutionary Applications*, 11(9), 1732–1747.
865 <https://doi.org/10.1111/eva.12676>

866 Hollenbeck, C. M., Portnoy, D. S., & Gold, J. R. (2019). Evolution of population structure in an
867 estuarine-dependent marine fish. *Ecology and Evolution, December 2018*, 3141–3152.
868 <https://doi.org/10.1002/ece3.4936>

869 Hueter, R. E., Heupel, M. R., Heist, E. J., & Keeney, D. B. (2005). Evidence of philopatry in
870 sharks and implications for the management of shark fisheries. *Journal of Northwest
871 Atlantic Fishery Science*, 35(November 2004), 239–247.
872 <https://doi.org/10.2960/J.v35.m493>

873 Hueter, R. E., & Tyminski, J. P. (2007). Species-specific distribution and habitat characteristics
874 of shark nurseries in Gulf of Mexico waters off Peninsular Florida and Texas. *American
875 Fisheries Society Symposium*, 50, 193–223.

876 Hvidsten, N. A., Heggberget, T. G., & Jensen, A. J. (1998). Sea water temperatures at Atlantic
877 salmon smolt entrance. *Nordic Journal of Freshwater Research*, 79–86.

878 Jones, A. T., Ovenden, J. R., & Wang, Y.-G. (2016). Improved confidence intervals for the
879 linkage disequilibrium method for estimating effective population size. *Heredity*, 117(4),
880 217–223. <https://doi.org/10.1038/hdy.2016.19>

881 Kamvar, Z., Tabima, J., & Grünwald, N. (2014). Poppr: An R package for genetic analysis of
882 populations with clonal, partially clonal, and/or sexual reproduction. *PeerJ*, 2, e281.

883 Karl, S. A., Castro, A. L. F., Lopez, J. A., Charvet, P., & Burgess, G. H. (2011). Phylogeography
884 and conservation of the bull shark (*Carcharhinus leucas*) inferred from mitochondrial

885 and microsatellite DNA. *Conservation Genetics*, 12(2), 371–382.

886 <https://doi.org/10.1007/s10592-010-0145-1>

887 Karlsson, S., Saillant, E., & Gold, J. R. (2009). Population structure and genetic variation of lane
888 snapper (*Lutjanus synagris*) in the northern Gulf of Mexico. *Marine Biology*, 156(9),
889 1841–1855. <https://doi.org/10.1007/s00227-009-1217-y>

890 Kawecki, T. J., & Ebert, D. (2004). Conceptual issues in local adaptation. *Ecology Letters*, 7(12),
891 1225–1241. <https://doi.org/10.1111/j.1461-0248.2004.00684.x>

892 Keeney, D. B., Heupel, M. R., Hueter, R. E., & Heist, E. J. (2003). Genetic heterogeneity among
893 blacktip shark, *Carcharhinus limbatus*, continental nurseries along the U.S. Atlantic and
894 Gulf of Mexico. *Marine Biology*, 143(6), 1039–1046. <https://doi.org/10.1007/s00227-003-1166-9>

895 Keeney, D. B., Heupel, M. R., Hueter, R. E., & Heist, E. J. (2005). Microsatellite and
896 mitochondrial DNA analyses of the genetic structure of blacktip shark (*Carcharhinus
897 limbatus*) nurseries in the northwestern Atlantic, Gulf of Mexico, and Caribbean Sea.
898 *Molecular Ecology*, 14(7), 1911–1923. <https://doi.org/10.1111/j.1365-294X.2005.02549.x>

900 Kim, C.-K., & Park, K. (2012). A modeling study of water and salt exchange for a micro-tidal,
901 stratified northern Gulf of Mexico estuary. *Journal of Marine Systems*, 96–97, 103–115.
902 <https://doi.org/10.1016/j.jmarsys.2012.02.008>

903 Kohler, N. E., & Turner, P. A. (2019). Distributions and Movements of Atlantic Shark Species:
904 A 52-Year Retrospective Atlas of Mark and Recapture Data. *Marine Fisheries Review*,
905 81(2).

907 Kraemer, P., & Gerlach, G. (2017). Demerelate: Calculating interindividual relatedness for
908 kinship analysis based on codominant diploid genetic markers using R. *Molecular*
909 *Ecology Resources*, 17(6), 1371–1377. <https://doi.org/10.1111/1755-0998.12666>

910 Laikre, L., Palm, S., & Ryman, N. (2005). Genetic population structure of fishes: Implications
911 for coastal zone management. *Ambio*, 34(2), 111–119.

912 Legendre, P. (1993). Spatial Autocorrelation: Trouble or New Paradigm? *Ecology*, 74(6), 1659–
913 1673. JSTOR. <https://doi.org/10.2307/1939924>

914 Legendre, P., & Legendre, L. (2012). *Numerical Ecology*. Elsevier.

915 Leidig, J. M., Shervette, V. R., McDonough, C. J., & Darden, T. L. (2015). Genetic Population
916 Structure of Black Drum in U.S. Waters. *North American Journal of Fisheries*
917 *Management*, 35(3), 464–477. <https://doi.org/10.1080/02755947.2015.1017123>

918 Leigh, J. W., & Bryant, D. (2015). popart: Full-feature software for haplotype network
919 construction. *Methods in Ecology and Evolution*, 6(9), 1110–1116.
920 <https://doi.org/10.1111/2041-210X.12410>

921 Lowe, C. G. (2002). Bioenergetics of free-ranging juvenile scalloped hammerhead sharks
922 (*Sphyrna lewini*) in Kāne'ohe Bay, O'ahu, HI. *Journal of Experimental Marine Biology*
923 *and Ecology*, 278(2), 141–156. [https://doi.org/10.1016/S0022-0981\(02\)00331-3](https://doi.org/10.1016/S0022-0981(02)00331-3)

924 Lowerre-Barbieri, S. K., Tringali, M. D., Shea, C. P., Walters Burnsed, S., Bickford, J., Murphy,
925 M., & Porch, C. (2019). Assessing red drum spawning aggregations and abundance in the
926 Eastern Gulf of Mexico: A multidisciplinary approach. *ICES Journal of Marine Science*,
927 76(2), 516–529. <https://doi.org/10.1093/icesjms/fsy173>

928 Luikart, G., England, P. R., Tallmon, D. A., Jordan, S., & Taberlet, P. (2003). The power and
929 promise of population genomics: From genotyping to genome typing. *Nature Reviews
930 Genetics*, 4, 981–994.

931 Manire, C. A., & Gruber, S. H. (1993). A preliminary estimate of natural mortality of age-0
932 lemon sharks, *Negaprion brevirostris*. In S. Branstetter (Ed.), *Conservation Biology of
933 Elasmobranchs. NOAA Technical Report NMFS 115* (pp. 65–71).

934 Matich, P., Mohan, J. A., Plumlee, J. D., Tinhan, T., Wells, R. J. D., & Fisher, M. (2017).
935 Factors shaping the co-occurrence of two juvenile shark species along the Texas Gulf
936 Coast. *Marine Biology*, 164(6), 1–16. <https://doi.org/10.1007/s00227-017-3173-2>

937 Matich, P., Plumlee, J. D., & Fisher, M. (2021). Grow fast, die young: Does compensatory
938 growth reduce survival of juvenile blacktip sharks (*Carcharhinus limbatus*) in the
939 western Gulf of Mexico? *Ecology and Evolution*, 11(22), 16280–16295.
940 <https://doi.org/10.1002/ece3.8311>

941 Matlock, G. (1990). The Life History of Red Drum. In G. Chamberlain, R. Miget, & M. Haby
942 (Eds.), *Red Drum Aquaculture* (pp. 1-21).

943 McCallister, M., Ford, R., & Gelsleichter, J. (2013). Abundance and Distribution of Sharks in
944 Northeast Florida Waters and Identification of Potential Nursery Habitat. *Marine and
945 Coastal Fisheries*, 5(1), 200–210. <https://doi.org/10.1080/19425120.2013.786002>

946 McClure, M. R., & McEachran, J. D. (1992). Hybridization between *Prionotus alatus* and *P.
947 paralatus* in the Northern Gulf of Mexico (Pisces: Triglidae). *Copeia*, 1992(4), 1039–
948 1046.

949 Milligan, B. G. (2003). Maximum-Likelihood Estimation of Relatedness. *Genetics*, 1167, 1153–
950 1167.

951 Natanson, L., Deacy, B., Moncrief-Cox, H., & Driggers III, W. B. (2019). *Reproductive*
952 *parameters for blacktip sharks (Carcharhinus limbatus) from the western North Atlantic*
953 *Ocean*. SEDAR65-DW01. SEDAR, North Charleston, SC. 10 pp.

954 National Marine Fisheries Service (NMFS). (2006). Code of Federal Regulations, Atlantic
955 Highly Migratory Species, § 635.27 Quota, 480.
956 https://books.google.com/books?id=n_K2fFVXoSEC

957 Nei, M. (1978). Estimation of Average Heterozygosity and Genetic Distance From a Small
958 Number of Individuals. *Genetics*, 89(3), 583–590.

959 Nei, M. (1987). *Molecular Evolutionary Genetics*. Columbia University Press.
960 <https://doi.org/10.7312/nei-92038>

961 Neigel, J. (2009). Population genetics and biogeography of the Gulf of Mexico. In D. Felder &
962 D. Camp (Eds.), *Gulf of Mexico – Origins, Waters, and Biota: Volume 1, Biodiversity*
963 (pp. 1353–1369). Texas A&M University Press.

964 Nielsen, E. E., Hemmer-Hansen, J., Larsen, P. F., & Bekevold, D. (2009). Population genomics
965 of marine fishes: Identifying adaptive variation in space and time. *Molecular Ecology*,
966 18(15), 3128–3150. <https://doi.org/10.1111/j.1365-294X.2009.04272.x>

967 Nielsen, R. (2001). Statistical tests of selective neutrality in the age of genomics. *Heredity*, 86(6),
968 641–647. <https://doi.org/10.1046/j.1365-2540.2001.00895.x>

969 O’Gower, A. K. (1995). Speculations on a spatial memory for the Port Jackson shark
970 (*Heterodontus portusjacksoni*) (Meyer) (Heterodontidae). *Marine and Freshwater*
971 *Research*, 46(5), 861–871.

972 Oksanen, J., Blanchet, F., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., Minchin, P.,

973 O'Hara, R., Simpson, G., Solymos, P., Stevens, M., Szoecs, E., & Wagner, H. (2018).

974 *vegan: Community Ecology Package*. <https://cran.r-project.org/package=vegan>

975 O'Leary, S. J., Hollenbeck, C. M., Vega, R. R., & Portnoy, D. S. (2021). Disentangling complex

976 genomic signals to understand population structure of an exploited, estuarine-dependent

977 flatfish. *Ecology and Evolution*, 11(19), 13415–13429. <https://doi.org/10.1002/ece3.8064>

978 O'Leary, S. J., Puritz, J. B., Willis, S. C., Hollenbeck, C. M., & Portnoy, D. S. (2018). These

979 aren't the loci you're looking for: Principles of effective SNP filtering for molecular

980 ecologists. *Molecular Ecology*, 27(16), 3193–3206. <https://doi.org/10.1111/mec.14792>

981 Orlando Jnr, S. P., Rozas, L. P., Ward, G., & Klein, C. (1993). *Salinity Characteristics of Gulf of*

982 *Mexico Estuaries* (p. 209). National Oceanic and Atmospheric Administration, Office of

983 Ocean Resources Conservation and Assessment.

984 Otero, J., L'Abée-Lund, J. H., Castro-Santos, T., Leonardsson, K., Storvik, G. O., Jonsson, B.,

985 Dempson, B., Russell, I. C., Jensen, A. J., Baglinière, J.-L., Dionne, M., Armstrong, J. D.,

986 Romakkaniemi, A., Letcher, B. H., Kocik, J. F., Erkinaro, J., Poole, R., Rogan, G.,

987 Lundqvist, H., ... Vøllestad, L. A. (2014). Basin-scale phenology and effects of climate

988 variability on global timing of initial seaward migration of Atlantic salmon (*Salmo salar*). *Global Change Biology*, 20(1), 61–75. <https://doi.org/10.1111/gcb.12363>

990 Palumbi, S. R. (1994). Genetic Divergence, Reproductive Isolation, and Marine Speciation.

991 *Annual Review of Ecology and Systematics*, 25(1), 547–572.

992 <https://doi.org/10.1146/annurev.es.25.110194.002555>

993 Pardini, A. T., Jones, C. S., Noble, L. R., Kreiser, B., Malcolm, H., Bruce, B. D., Stevens, J. D.,

994 Cliff, G., Scholl, M. C., Francis, M. P., Duffy, C. A., & Martin, A. P. (2001). Sex-biased

995 dispersal of great white sharks. *Nature*, 412(6843), 139–140.

996 <https://doi.org/10.1038/35084125>

997 Parsons, G. R., & Hoffmayer, E. (2007). Identification and characterization of shark nursery

998 grounds along the Mississippi and Alabama Gulf coasts. *American Fisheries Society*

999 *Symposium*, 50, 301–316.

1000 Peterson, B. K., Weber, J. N., Kay, E. H., Fisher, H. S., & Hoekstra, H. E. (2012). Double digest

1001 RADseq: An inexpensive method for de novo SNP discovery and genotyping in model

1002 and non-model species. *PLOS One*, 7(5), e37135.

1003 <https://doi.org/10.1371/journal.pone.0037135>

1004 Petrou, E. L., Drinan, D. P., Kopperl, R., Lepofsky, D., Yang, D., Moss, M. L., & Hauser, L.

1005 (2019). Intraspecific DNA contamination distorts subtle population structure in a marine

1006 fish: Decontamination of herring samples before restriction-site associated sequencing

1007 and its effects on population genetic statistics. *Molecular Ecology Resources*, 19(5),

1008 1131–1143. <https://doi.org/10.1111/1755-0998.12978>

1009 Pew, J., Muir, P. H., Wang, J., & Frasier, T. R. (2015). related: An R package for analysing

1010 pairwise relatedness from codominant molecular markers. *Molecular Ecology Resources*,

1011 15(3), 557–561. <https://doi.org/10.1111/1755-0998.12323>

1012 Phillips, N. M., Devloo-Delva, F., McCall, C., & Daly-Engel, T. S. (2021). Reviewing the

1013 genetic evidence for sex-biased dispersal in elasmobranchs. *Reviews in Fish Biology and*

1014 *Fisheries*, 31(4), 821–841. <https://doi.org/10.1007/s11160-021-09673-9>

1015 Portela, E., Tenreiro, M., Pallàs-Sanz, E., Meunier, T., Ruiz-Angulo, A., Sosa-Gutiérrez, R., &

1016 Cusí, S. (2018). Hydrography of the Central and Western Gulf of Mexico. *Journal of*

1017 *Geophysical Research: Oceans*, 123(8), 5134–5149.

1018 <https://doi.org/10.1029/2018JC013813>

1019 Portnoy, D. S., Fields, A. T., Puritz, J. B., Hollenbeck, C. M., & Patterson, W. F. (2021).

1020 Genomic analysis of red snapper, *Lutjanus campechanus*, population structure in the U.S.

1021 Atlantic and Gulf of Mexico. *ICES Journal of Marine Science*.

1022 <https://doi.org/10.1093/icesjms/fsab239>

1023 Portnoy, D. S., & Gold, J. R. (2012). Evidence of multiple vicariance in a marine suture-zone in

1024 the Gulf of Mexico. *Journal of Biogeography*, 39(8), 1499–1507.

1025 <https://doi.org/10.1111/j.1365-2699.2012.02699.x>

1026 Portnoy, D. S., & Heist, E. J. (2012). Molecular markers: Progress and prospects for

1027 understanding reproductive ecology in elasmobranchs. *Journal of Fish Biology*. 80(5),

1028 1120–1140. <https://doi.org/10.1111/j.1095-8649.2011.03206.x>

1029 Portnoy, D. S., Hollenbeck, C. M., Bethea, D., Frazier, B., Gelsleichter, J., & Gold, J. (2016).

1030 Population structure, gene flow, and historical demography of a small coastal shark

1031 (*Carcharhinus isodon*) in US waters of the Western Atlantic Ocean. *ICES Journal of*

1032 *Marine Science*, 73(9), 2322–2332.

1033 Portnoy, D. S., Hollenbeck, C. M., Belcher, C. N., Driggers, W. B., Frazier, B. S., Gelsleichter,

1034 J., Grubbs, R. D., & Gold, J. R. (2014). Contemporary population structure and post-

1035 glacial genetic demography in a migratory marine species, the blacknose shark,

1036 *Carcharhinus acronotus*. *Molecular Ecology*, 23(22), 5480–5495.

1037 <https://doi.org/10.1111/mec.12954>

1038 Portnoy, D. S., McDowell, J. R., McCandless, C. T., Musick, J. A., & Graves, J. E. (2009).

1039 Effective size closely approximates the census size in the heavily exploited western

1062 Ribeiro, J., & Diggle, P. (2001). GeoR: A package for geostatistical analysis. *R-News*, 1(2),
1063 1609–3631.

1064 Rigby, C. L., Carlson, J., Chin, A., Derrick, D., Dicken, M., & Pacourea, N. (2021).
1065 *Carcharhinus limbatus*. The IUCN Red List of Threatened Species.
1066 <https://dx.doi.org/10.2305/IUCN.UK.2021-2.RLTS.T3851A2870736.en>

1067 Rocha, L. A. (2003). Patterns of distribution and processes of speciation in Brazilian reef fishes.
1068 *Journal of Biogeography*, 30(8), 1161–1171. <https://doi.org/10.1046/j.1365->
1069 2699.2003.00900.x

1070 Sambrook, J., Fritsch, E. F. & Maniatis, T. (1989) *Molecular Cloning*. Cold Spring Harbor
1071 Laboratory Press, New York.

1072 Sbrocco, E. J., & Barber, P. H. (2013). MARSPEC: ocean climate layers for marine spatial
1073 ecology. *Ecology*, 94(4), 979. <https://doi.org/10.1890/12-1358.1>

1074 SEDAR. (2018). *Update assessment to SEDAR 29 HMS Gulf of Mexico Blacktip Shark* (pp. 1–
1075 99). National Marine Fisheries Service.

1076 SEDAR. (2020). *SEDAR 65 Atlantic Blacktip Shark Stock Assessment Report* (pp. 1–438).
1077 National Marine Fisheries Service.

1078 Seutin, G., White, B. N., & Boag, P. T. (1991). Preservation of avian blood and tissue samples
1079 for DNA analyses. *Canadian Journal of Zoology*, 69(1), 82–90.

1080 Seyoum, S., McBride, R. S., Puchutulegui, C., Dutka-Gianelli, J., Alvarez, A., & K, P. (2017).
1081 Genetic population structure of sheepshead, *Archosargus probatocephalus* (Sparidae), a
1082 coastal marine fish off the southeastern United States: Multiple population clusters based
1083 on species-specific microsatellite markers. *Bulletin of Marine Science*, 93, 691–713.

1084 Seyoum, S., McBride, R. S., Tringali, M. D., Villanova, V. L., Puchutulegui, C., Gray, S., & Van
1085 Bibber, N. (2018). Genetic population structure of the spotted seatrout (*Cynoscion*
1086 *nebulosus*): Simultaneous examination of the mtDNA control region and microsatellite
1087 marker results. *Bulletin of Marine Science*, 94(1), 47–71.
1088 <https://doi.org/10.5343/bms.2017.1060>

1089 Sievers, F., Wilm, A., Dineen, D., Gibson, T. J., Karplus, K., Li, W., Lopez, R., McWilliam, H.,
1090 Remmert, M., Söding, J., Thompson, J. D., & Higgins, D. G. (2011). Fast, scalable
1091 generation of high-quality protein multiple sequence alignments using Clustal Omega.
1092 *Molecular Systems Biology*, 7(1), 539. <https://doi.org/10.1038/msb.2011.75>

1093 Stiebens, V. A., Merino, S. E., Roder, C., Chain, F. J. J., Lee, P. L. M., & Eizaguirre, C. (2013).
1094 Living on the edge: How philopatry maintains adaptive potential. *Proceedings of the*
1095 *Royal Society B: Biological Sciences*, 280(1763). <https://doi.org/10.1098/rspb.2013.0305>

1096 Sundström, L. F., Gruber, S. H., Clermont, S. M., Correia, J. P. S., de Marignac, J. R. C.,
1097 Morrissey, J. F., Lowrance, C. R., Thomassen, L., & Oliveira, M. T. (2001). Review of
1098 Elasmobranch Behavioral Studies Using Ultrasonic Telemetry with Special Reference to
1099 the Lemon Shark, *Negaprion brevirostris*, Around Bimini Islands, Bahamas.
1100 *Environmental Biology of Fishes*, 60(1), 225–250.
1101 <https://doi.org/10.1023/A:1007657505099>

1102 United States Geological Survey (USGS). (1990). *Largest Rivers in the United States*.
1103 Department of the Interior. <https://pubs.usgs.gov/of/1987/ofr87-242/pdf/ofr87242.pdf>

1104 TexasET. (2022). *Monthly Rainfall (1968-2021)*. The TexasET Network.
1105 <https://etweather.tamu.edu/rainhistory/>

1106 Tyberghein, L., Verbruggen, H., Pauly, K., Troupin, C., Mineur, F., & De Clerck, O. (2012).
1107 Bio-ORACLE: a global environmental dataset for marine species distribution modelling.
1108 *Global Ecology and Biogeography*, 21(2), 272–281. <https://doi.org/10.1111/j.1466-8238.2011.00656.x>

1109

1110 Ulrich, G. F., Jones, C. M., Driggers, W. B., Drymon, J. M., Oakley, D., & Riley, C. (2007).
1111 Habitat Utilization, Relative Abundance, and Seasonality of Sharks in the Estuarine and
1112 Nearshore Waters of South Carolina. *American Fisheries Society Symposium*,
1113 50(December 2015), 125–139.

1114 Van Etten, J. (2017). R Package gdistance: Distances and routes on geographical grids. *Journal*
1115 *of Statistical Software*, 76(13), 1–21. <https://doi.org/10.18637/jss.v076.i13>

1116 Volk, D. R., Konvalina, J. D., Floeter, S. R., Ferreira, C. E. L., & Hoffman, E. A. (2021). Going
1117 against the flow: Barriers to gene flow impact patterns of connectivity in cryptic coral
1118 reef gobies throughout the western Atlantic. *Journal of Biogeography*, 48(2), 427–439.
1119 <https://doi.org/10.1111/jbi.14010>

1120 Vollset, K. W., Lennox, R. J., Lamberg, A., Skaala, Ø., Sandvik, A. D., Sægrov, H., Kvingedal,
1121 E., Kristensen, T., Jensen, A. J., Haraldstad, T., Barlaup, B. T., & Ugedal, O. (2021).
1122 Predicting the nationwide outmigration timing of Atlantic salmon (*Salmo salar*) smolts
1123 along 12 degrees of latitude in Norway. *Diversity and Distributions*, 27(8), 1383–1392.
1124 <https://doi.org/10.1111/ddi.13285>

1125 Wang, J. (2002). An Estimator for Pairwise Relatedness Using Molecular Markers. *Genetics*,
1126 160(3), 1203–1215. <https://doi.org/10.1093/genetics/160.3.1203>

1127 Waples, R. K., Larson, W. A., & Waples, R. S. (2016). Estimating contemporary effective
1128 population size in non-model species using linkage disequilibrium across thousands of
1129 loci. *Heredity*, 117(4), 233–240. <https://doi.org/10.1038/hdy.2016.60>

1130 Waples, R. S. (1998). Separating the wheat from the chaff: Patterns of genetic differentiation in
1131 high gene flow species. *Journal of Heredity*, 89(5), 438–450.
1132 <https://doi.org/10.1093/jhered/89.5.438>

1133 Waples, R. S., & Anderson, E. (2017). Purgling putative siblings from population genetic
1134 datasets: A cautionary view. *Molecular Ecology*, 26(5), 1211–1224.
1135 <https://doi.org/10.1111/ijlh.12426>

1136 Waples, R. S., Punt, A. E., & Cope, J. M. (2008). Integrating genetic data into management of
1137 marine resources: How can we do it better? *Fish and Fisheries*, 9(4), 423–449.
1138 <https://doi.org/10.1111/j.1467-2979.2008.00303.x>

1139 Weber, D. N., Frazier, B. S., Whitney, N. M., Gelsleichter, J., Gorka, S. (2020). Stress response
1140 and postrelease mortality of blacktip sharks (*Carcharhinus limbatus*) captured in shore-
1141 based and charter-boat-based recreational fisheries. *Fishery Bulletin*, 118, 297-314.
1142 <https://doi.org/10.7755/FB.118.3.8>

1143 Weir, B. S., & Cockerham, C. (1984). Estimating F -statistics for the analysis of population
1144 structure. *Evolution*, 38, 1358-1370.

1145 Whitlock, M. C., & Lotterhos, K. E. (2015). Reliable Detection of Loci Responsible for Local
1146 Adaptation: Inference of a Null Model through Trimming the Distribution of F_{ST} . *The*
1147 *American Naturalist*, 186(S1), S24–S36. <https://doi.org/10.1086/682949>

1148 Whitlock, M. C., & McCauley, D. E. (1999). Indirect measures of gene flow and migration:
1149 $F_{ST} \neq 1/(4Nm+1)$. *Heredity*, 82(2), 117–125. <https://doi.org/10.1038/sj.hdy.6884960>

1150 Wickham, H. (2016). *ggplot2: Elegant Graphics for Data Analysis*. Springer-Verlag.
1151 <https://ggplot2.tidyverse.org>
1152 Willis, S. C., Hollenbeck, C. M., Puritz, J. B., Gold, J. R., & Portnoy, D. S. (2017). Haplotyping
1153 RAD loci: An efficient method to filter paralogs and account for physical linkage.
1154 *Molecular Ecology Resources*, 17(5), 955–965. <https://doi.org/10.1111/1755-0998.12647>
1155 Zuur, A. F., Ieno, E. N., & Elphick, C. S. (2010). A protocol for data exploration to avoid
1156 common statistical problems. *Methods in Ecology and Evolution*, 1(1), 3–14.
1157 <https://doi.org/10.1111/j.2041-210X.2009.00001.x>

1158 **Data Accessibility and Benefit-Sharing**

1159 **Genetic Data**

1160 Raw MiSeq and HiSeq reads are available in the NCBI SRA (BioProject PRJNA996573).
1161 Analysis scripts and data are available on GitHub (https://github.com/dgs108/blacktip_philopatry)
1162 and DataDryad (DOI: 10.5061/dryad.vmcvdnczp), respectively.

1163 **Sample Metadata**

1164 Metadata associated with raw reads are available in the NCBI SRA (BioProject PRJNA996573)
1165 and Tables S1-2.

1166 **Author Contributions**

1167 **DG Swift**: conceptualization; funding; sample collection; molecular laboratory work; data
1168 analysis; writing; writing-review and editing.

1169 **SJ O’Leary**: data analysis, writing-review and editing.

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1176 **TR Wiley:** sample collection, writing-review and editing.

1177 **DS Portnoy:** conceptualization; funding; data analysis; writing; writing-review and editing.

1178 **Figure Captions**

1179 **Figure 1:** Haplotype network of the mitochondrial control region and map of sampling locations
1180 in the U.S. Atlantic and Gulf of Mexico for the blacktip shark (*Carcharhinus limbatus*). Dotted
1181 lines denote regions that follow designations by NOAA Fisheries. Mobile Bay, Alabama straddles
1182 the 88th meridian which separates the eastern and western Gulf stock subregions. Mobile Bay was
1183 found to be part of the eastern Gulf in this study. Numbers refer to the sample size for each site
1184 included in the ddRAD-Seq data. Abbreviations of U.S. States: Texas (TX), Louisiana (LA),
1185 Mississippi (MS), Alabama (AL), Florida (FL), Georgia (GA), South Carolina (SC).

1186 **Figure 2:** The differential effects of spatial and environmental differences on genetic population
1187 structure of blacktip sharks (*Carcharhinus limbatus*). A) Biplot showing ordination space loadings
1188 determined by MEM1 and MEM2 from the full spatial redundancy analysis. B) Biplot showing
1189 ordination space loadings determined by minimum annual temperature and mean salinity in June
1190 from the full environmental redundancy analysis. C) Mean \pm one standard deviation value for
1191 MEM1 by site vs. coastal distance. D) Mean \pm one standard deviation value for MEM2 by site vs.

1192 coastal distance. E) Pairwise neutral F_{ST} between eastern and western Gulf sites vs. pairwise
1193 coastal distance between sites. F) Mean \pm one standard deviation value for minimum annual
1194 temperature by site vs. coastal distance. G) Mean \pm one standard deviation value for mean salinity
1195 in June by site vs. coastal distance. H) Pairwise adaptive F_{ST} between eastern and western Gulf
1196 sites vs. absolute difference in latitudinal degrees between sites.

1197 **Figure 3:** Lower 95% confidence interval estimates of contemporary effective population size (N_e)
1198 of blacktip sharks (*Carcharhinus limbatus*) by site. Numbers above the x-axis denote the sample
1199 size per site. Site abbreviations: Bulls Bay (BLB^{*}), St. Helena Sound (SHS^{*}), Port Royal Sound
1200 (PRS^{*}), Terra Ceia Bay (TCB[¥]), Waccasassa Bay (WAB[¥]), Apalachicola Bay (APB[¥]), Mobile Bay
1201 (MOB[¥]), Galveston Bay (GAB[§]), Matagorda Bay (MAB[§]), San Antonio Bay (SAB[§]), and Corpus
1202 Christi Bay (CCB[§]). Symbols denote regions: Atlantic^{*}, eastern Gulf[¥], and western Gulf[§].

1203 **Supporting Information**

1204 Detailed in two documents: supplementary materials and methods (PDF) and supplementary tables
1205 and figures (Excel Workbook).

1206

Table 1: Results from hierarchical and single-level AMOVA using the mitochondrial control region, 4,271 neutral, and 68 putatively adaptive SNP-containing nuclear loci. Underlined *p*-values denote statistically significant heterogeneity. For nuclear datasets, * denotes lower 2.5% of bootstrapped *F*-statistics were greater than zero.

Dataset	Sites	Source of Variation	Variance Components	Percent Variation	Φ/F -statistic	<i>p</i> -value
Mitochondrial Control Region	All	Among groups (i.e., Atlantic and Gulf)	0.0585	9.9700	0.0997	<u><0.05</u>
		Among sites within groups	0.0420	7.1500	0.0795	<u><0.0001</u>
	Atlantic	Among sites	0.0065	2.4600	0.0246	0.1768
		Among individuals within sites	0.2592	97.5400	-	-
	Gulf	Among sites	0.0495	8.2600	0.0826	<u><0.0001</u>
		Among individuals within sites	0.5502	91.7400	-	-
Neutral Nuclear Loci	All	Among groups (i.e., Atlantic and Gulf)	0.5151	0.1542	0.0015	<u><0.0001*</u>
		Among sites within groups	0.1993	0.0597	0.0006	<u><0.001*</u>
	Atlantic	Among sites	0.1167	0.0353	0.0004	0.1871
		Among individuals within sites	330.1258	99.9647	-	-
	Gulf	Among sites	0.2182	0.0652	0.0007	<u><0.0001*</u>
		Among individuals within sites	334.5166	99.9348	-	-
Adaptive Nuclear Loci	All	Among groups (i.e., Atlantic and Gulf)	0.0012	0.0242	0.0002	0.3641
		Among sites within groups	0.0340	0.6858	0.0069	<u><0.0001*</u>
	Atlantic	Among sites	0.0028	0.0568	0.0006	0.4687
		Among individuals within sites	4.9062	99.9432	-	-
	Gulf	Among sites	0.0423	0.8510	0.0085	<u><0.0001*</u>
		Among individuals within sites	4.9233	99.1490	-	-

Table 2: Estimates of genetic diversity based on the mitochondrial control region, 4,271 neutral, and 68 putatively adaptive SNP-containing nuclear loci. Mean estimates are given for all, with \pm one standard deviation estimates for the mitochondrial control region only. n refers to sample size per site. h and π refer to mitochondrial haplotype diversity and nucleotide diversity, respectively. H_e and A_r refer to Nei's gene diversity and allelic richness, respectively. Site abbreviations: Bulls Bay (BLB), St. Helena Sound (SHS), Port Royal Sound (PRS), Terra Ceia Bay (TCB), Waccasassa Bay (WAB), Apalachicola Bay (APB), Mobile Bay (MOB), Galveston Bay (GAB), Matagorda Bay (MAB), San Antonio Bay (SAB), and Corpus Christi Bay (CCB).

Region	Site	Mitochondrial Control Region				n	Neutral Loci		Adaptive Loci	
		n	Haplotypes	h	π		H_e	A_r	H_e	A_r
Atlantic	BLB	30	4	0.4483 ± 0.1021	0.000538 ± 0.000528	49	0.1551	2.8273	0.1434	2.8932
	SHS	29	4	0.5345 ± 0.0725	0.000635 ± 0.000587	47	0.1544	2.8199	0.1445	2.8771
	PRS	12	3	0.3182 ± 0.1637	0.000513 ± 0.000540	16	0.1537	2.8174	0.1466	2.9122
Eastern Gulf	TCB	70	6	0.5706 ± 0.0314	0.000705 ± 0.000615	84	0.1547	2.8234	0.1373	2.8397
	WAB	32	6	0.6492 ± 0.0794	0.001260 ± 0.000930	34	0.1574	2.8448	0.1373	2.8521
	APB	31	7	0.7333 ± 0.0534	0.001208 ± 0.000904	46	0.1568	2.8399	0.1533	2.9057
	MOB	12	4	0.7424 ± 0.0842	0.001275 ± 0.000992	16	0.1578	2.8470	0.2076	3.5343
Western Gulf	GAB	13	5	0.8205 ± 0.0661	0.001429 ± 0.001070	15	0.1572	2.8420	0.1393	2.8943
	MAB	30	6	0.8253 ± 0.0334	0.001455 ± 0.001035	31	0.1562	2.8362	0.1450	2.8881
	SAB	44	9	0.8478 ± 0.0204	0.001606 ± 0.001101	56	0.1584	2.8545	0.1370	2.8323
	CCB	20	6	0.8000 ± 0.0537	0.001427 ± 0.001038	24	0.1578	2.8525	0.1465	2.8988





