

## Otolith stable isotope micro-sampling to discriminate poorly studied stocks: Crevalle Jack in the eastern Gulf of Mexico

Carissa L. Gervasi<sup>1</sup>, James A. Nelson<sup>2</sup>, Peter K. Swart<sup>3</sup>, Rolando O. Santos<sup>1</sup>, Ryan J. Rezek<sup>4</sup>, W. Ryan James<sup>1</sup>, Amanda E. Jefferson<sup>5,6</sup>, J. Marcus Drymon<sup>5,6</sup>, Jessica Carroll<sup>7</sup>, Ross E. Boucek<sup>8</sup>,  
Jennifer S. Rehage<sup>1</sup>

<sup>1</sup>Florida International University  
Institute of Environment  
11200 SW 8<sup>th</sup> St., Miami, FL 33199

<sup>2</sup>University of Louisiana at Lafayette  
Department of Biology  
410 E St Mary Blvd., Lafayette, Louisiana 70503

<sup>3</sup>University of Miami  
Department of Marine Geosciences  
Rosenstiel School of Marine and Atmospheric Science  
4600 Rickenbacker Causeway, Miami, FL 33149

<sup>4</sup>Coastal Carolina University  
Department of Marine Science  
100 Chanticleer Drive East, Conway, SC 29526

<sup>5</sup>Mississippi State University  
Coastal Research and Extension Center  
1815 Popps Ferry Rd., Biloxi, MS 39532

<sup>6</sup>Mississippi-Alabama Sea Grant Consortium  
703 East Beach Dr., Ocean Springs, MS 39564

<sup>7</sup>Florida Fish and Wildlife Conservation Commission  
Fish and Wildlife Research Institute  
100 8th Ave. SE, St. Petersburg, FL 33701

<sup>8</sup>Bonefish & Tarpon Trust  
2937 SW 27th Ave., Suite 203 Miami, FL 33133

Corresponding author: [cgerv002@fiu.edu](mailto:cgerv002@fiu.edu)

1 **Abstract:** Developing conservation and management strategies for species with complex life  
2 histories, broad spatial distributions, and long lifespans is notoriously difficult. Too often  
3 managers cannot identify critical habitats nor vulnerable life stages because of the sheer scale of  
4 migration or uncertainty about connectivity among populations. Advancements in otolith stable  
5 isotope analysis, and specifically sampling of discrete otolith layers, have provided opportunities  
6 to assess lifetime migrations and connectivity without extensive, long-term field sampling. Here,  
7 we compared carbon and oxygen stable isotope values in discrete otolith layers for Crevalle Jack  
8 (an unregulated and data-poor species) captured in two isotopically distinct regions (Alabama  
9 and the Florida Keys). Our goal was to address vital questions about how broad-scale movement  
10 patterns through ontogeny differ between the two regions and whether connectivity occurs  
11 throughout the life history. Our results revealed that Crevalle Jack appear to inhabit inshore  
12 nursery areas at age-0, before migrating to coastal/offshore habitats between age-1 and age-2.  
13 Comparisons between fish collected in northeastern and southeastern Gulf of Mexico regions  
14 revealed significant differences in the patterns of stable isotope values throughout the life history  
15 and in otolith  $\delta^{13}\text{C}$  values. Despite these differences that suggest a separation of the two  
16 populations, individual variability was significant, hindering our ability to determine whether  
17 fish from the two regions represent separate, self-recruiting stocks. Our research illustrates the  
18 potential of otolith stable isotope micro-sampling as a tool for examining broad-scale movement  
19 and migration patterns of fishes, and informing future research and management.

20

21 **Keywords:** stable isotopes, otolith microchemistry, otolith micro-sampling, crevalle jack,  
22 population connectivity

## 23 **1. Introduction**

24           The understanding of population connectivity and dispersal pathways is essential to  
25 effective fisheries management. Knowledge of connectivity patterns in particular is crucial for  
26 delineating appropriate spatial scales of management and for specifying vital subareas to protect  
27 from exploitation (Fogarty and Botsford, 2007). Misaligned stock assessment and population  
28 boundaries can have severe consequences, including over-exploitation and localized depletion  
29 (Berger et al., 2021; Ying et al., 2011). Additionally, successful implementation of marine  
30 protected areas requires that the size and location match the spatial distribution and habitat use of  
31 the species of interest (Kramer and Chapman, 1999; Moffitt et al., 2009). However, population  
32 connectivity (specifically ecological connectivity) can occur over multiple spatiotemporal scales  
33 and can therefore be challenging to assess. Ecological connectivity is defined as the exchange of  
34 individuals among local populations that can affect population dynamics and demographics (Sale  
35 et al., 2010). This exchange can occur over multiple life history stages, and includes larval  
36 dispersal, juvenile recruitment and retention, and large-scale movements of sub-adults and  
37 adults. Ecological connectivity is the basis of metapopulation ecology (Levins, 1968) and  
38 contingent theory (Hjort, 1914; Secor, 1999), paradigms that are increasingly being used to  
39 explain empirical observations of fisheries dynamics (Cadrin and Secor, 2009).

40           Despite increasing recognition of the importance of ecological connectivity and  
41 metapopulation processes in fisheries, many stock assessments are unable to account for  
42 complex population structures, often assuming that a fish stock is a single, spatially  
43 homogeneous population stemming from a single larval pool (Archambault et al., 2016).  
44 Spatially-explicit population models are becoming increasingly common, but studies of  
45 population connectivity mainly focus on early life history stages (e.g., Hinrichsen et al., 2011;

46 Miller, 2007). Juvenile and adult-mediated connectivity is largely understudied, despite research  
47 showing that adult movements can significantly affect metapopulation structure and dynamics  
48 (Frisk et al., 2014). It is therefore crucial that population connectivity of exploited fishes is  
49 assessed at all life history stages, and that this spatial structure is incorporated into population  
50 models and stock assessments.

51         Assessment of population connectivity throughout the life history of a species is  
52 challenging, often requiring extensive, long-term field studies (e.g., routine fisheries-independent  
53 sampling or tag-recapture/animal tracking projects) to examine movement/dispersal patterns of  
54 the different life stages. However, advances in otolith microchemistry provide opportunities to  
55 rapidly examine movement patterns over the entire life history of an individual. Otoliths remain  
56 the most common structure used to age fishes, since clear growth bands are evident for most  
57 species (Campana, 2001). Additionally, because otoliths are inert, trace elements and isotope  
58 ratios are incorporated into otoliths as a fish grows, affected by ambient water concentrations. If  
59 a fish resides in a certain body of water for a portion of its life, the chemical signature of that  
60 water body will be retained in the section of the otolith corresponding with that period of the  
61 fish's life (Walther and Limburg, 2012). Because of these attributes, otoliths have been used in  
62 fisheries science to reconstruct environmental histories (Campana, 1999; Walther and Limburg,  
63 2012).

64         The applications of otolith chemical composition are numerous and include identifying  
65 nursery areas or natal origin (Gerard and Muhling, 2010; Thorrold et al., 2001), reconstructing  
66 migration patterns (Avigliano et al., 2021; Sturrock et al., 2012; Walther and Limburg, 2012),  
67 retrospectively assigning adults to areas of origin (Gerard et al., 2015), and determining stock  
68 structure (Tanner et al., 2016). Furthermore, changes in chemical composition over the lifetime

69 of an individual fish can be discerned via sampling of discrete otolith layers (i.e., micro-  
70 sampling; Jamieson et al., 2004). Stable isotope micro-sampling throughout the entire otolith is a  
71 relatively recent approach that has distinct advantages, including establishing environmental  
72 histories of individual fish, examining partial migration, and comparing life history patterns  
73 among populations (Høie et al., 2004a; Jamieson et al., 2004; Kawazu et al., 2020; Wang et al.,  
74 2021; Weidman and Millner, 2000). More often though, studies concentrate on broad regions of  
75 the otolith (e.g., core vs. edge) corresponding to different life history stages (e.g., birth vs. age at  
76 collection), potentially missing critical transition periods or periods of connectivity throughout  
77 the entire life history.

78         The purpose of this study was to examine lifetime migration patterns of Crevalle Jack  
79 (*Caranx hippos*) captured from two distinct regions of the eastern Gulf of Mexico (the Florida  
80 Keys and Alabama). The Crevalle Jack is an important part of the recreational fishery in both  
81 these areas, yet is currently unregulated and there is evidence that population abundance may be  
82 declining (Gervasi et al., 2022). Specifically, our study aimed to determine whether otolith  
83 micro-sampling could be used to distinguish between fish from each region based on differences  
84 in stable isotope values throughout otoliths. If separation in isotope values is evident throughout  
85 the life history, it would suggest a lack of mixing between the two areas, which has implications  
86 for stock assessment and management. By comparing stable isotope chemistry in discrete otolith  
87 layers between the two groups, our research addresses vital questions about how broad-scale  
88 movement patterns through ontogeny differ between the two regions. In addition to aiding in  
89 conservation and management, our results demonstrate the potential of otolith stable isotope  
90 micro-sampling as a tool for assessing lifetime population connectivity of fishes that can better  
91 inform future population models and stock assessments.

92

## 93 **2. Material & methods**

### 94 *2.1 Study species*

95           The Crevalle Jack is a large marine fish native to the Atlantic coast of North America and  
96 the Gulf of Mexico (GOM; Smith-Vaniz and Carpenter, 2007). The species is targeted by both  
97 commercial and recreational anglers (Kwei, 1978) but is unregulated in all coastal U.S. states  
98 within the species range and is considered data-poor (i.e., there is not enough information  
99 available to estimate relative stock status and appropriate reference points). In recent years, a  
100 decline in population abundance has been observed in the Florida Keys region (Gervasi et al.,  
101 2022) but the extent of the decline is unknown. Knowledge of migration patterns and population  
102 connectivity is, therefore, crucial for conservation and management of the species. However,  
103 little research has been conducted to date on Crevalle Jack biology and ecology, and critical  
104 questions remain about stock structure and life history.

105           Spawning is suspected to occur in subtropical and tropical waters, but in the Western  
106 Atlantic, it has only been observed at Gladden Spit, Belize, a promontory reef that serves as a  
107 multi-species spawning aggregation site (Heyman and Kjerfve, 2008). Although other species in  
108 the Carangidae family have been observed spawning in continental shelf edge habitats  
109 throughout the GOM (Heyman et al., 2019), only indirect evidence (courtship behavior and color  
110 changes) of Crevalle Jack spawning has been observed at a marine sanctuary in the northwestern  
111 GOM (Helies et al., 2016). However, Crevalle Jack larvae have been observed throughout the  
112 GOM, mostly in the spring and summer months (Ditty et al., 2004; Flores-Coto and Sanchez-  
113 Ramirez, 1989). Post-larval fish have also been observed in offshore waters in the summer and  
114 fall (Mohan et al., 2017). Young-of-the-year Crevalle Jack are found in coastal estuaries

115 throughout the Atlantic and GOM (Flaherty et al., 2013; McBride and McKown, 2000; Nelson,  
116 1992), but the linkages between these juvenile nursery habitats and adult populations are  
117 unknown. McBride and McKown (2000) examined seasonal abundance and size structure of  
118 Crevalle Jack from New York to Florida. Their findings suggested that larvae spawning at  
119 subtropical and tropical latitudes are dispersed via ocean currents up the Atlantic coast and into  
120 temperate estuarine habitats, and at least some individuals can migrate back down to the tropics  
121 and return to the spawning sites. This suggests that Crevalle Jack throughout the Atlantic coast of  
122 the U.S. may represent a single stock, but evidence is limited and connectivity with Gulf of  
123 Mexico, Caribbean, and South American populations is unknown. Veteran fishing guides in the  
124 Florida Keys have observed that large, adult Crevalle Jack appear to migrate seasonally into and  
125 out of the south Florida region, but where these individuals go is unknown (Gervasi et al., 2022).  
126 Previous research estimates that Crevalle Jack age at 50% maturity is about 3-4 years, but  
127 maturation data comes from a single population in the Caribbean (Caiafa et al., 2011). Adult  
128 Crevalle Jack are found in a variety of habitats, including coastal flats, coral reefs, artificial reefs,  
129 channels, and canals (Smith-Vaniz and Carpenter, 2007), but again, movement and migration  
130 patterns of adults is currently unknown.

131

## 132 *2.2 Focal populations and alternative hypotheses*

133 Adult Crevalle Jack were collected from two coastal regions: the Florida Keys (FK) and  
134 coastal Alabama (AL; Fig. 1). These regions were selected for several reasons. First, there is  
135 evidence that several ecological and faunal divides exist in the northern GOM. Studies have  
136 suggested that zoogeographic breaks occur at Mobile Bay (Drymon et al., 2020), the DeSoto  
137 Canyon at the eastern edge of the Mississippi River basin (Defenbaugh, 1976; Gallaway, 1981;

138 Ward, 2017), and at Cape San Blas at the eastern end of the Florida panhandle (Estes, 2016;  
139 Zieman and Zieman, 1989). Fish and invertebrate assemblages have been shown to differ on  
140 either side of these boundaries, suggesting that AL and FK Crevalle Jack may represent distinct  
141 populations. Secondly, according to Marine Recreational Information Program data (MRIP;  
142 NOAA, 2021), out of all coastal Atlantic and GOM states, recreational catch of Crevalle Jack  
143 from 2000-2021 was highest in Florida, followed by Alabama. Total catch from Alabama was  
144 more than twice the next highest-ranking state (North Carolina). Assuming recreational landings  
145 are a proxy for fish abundance, the MRIP data suggest that the Florida and Alabama regions may  
146 encompass centers of abundance for the species within the U.S. The Florida Keys was  
147 specifically chosen because previous research found evidence of a decline in Crevalle Jack  
148 abundance in the area, prompting management concerns (Gervasi et al., 2022). To successfully  
149 manage the population in Florida, it is important to understand if connectivity between Florida  
150 and other management jurisdictions exists. Finally, AL and FK represent two isotopically distinct  
151 regions, with both  $\delta^{18}\text{O}$  and  $\delta^{13}\text{C}$  values of water samples, sediments, and primary producers  
152 varying substantially. Coastal areas near the large rivers of the north-central GOM receive  
153 substantial inputs of freshwater depleted in  $^{18}\text{O}$ , with average  $\delta^{18}\text{O}$  of river samples ranging from  
154  $-2.2\text{‰}$  to as low as  $-6.6\text{‰}$  in the Mississippi River (Wagner and Slowey, 2011), and terrestrial  
155 carbon depleted in  $^{13}\text{C}$ , with values near  $-27\text{‰}$  (Fry, 1983). In the Florida Keys, limited  
156 freshwater input combined with high rates of evaporation lead to  $^{18}\text{O}$ -enriched waters ( $1.7\text{‰}$  in  
157 the Upper Florida Keys; Sternberg and Swart, 1987), and abundant  $^{13}\text{C}$ -enriched seagrasses and  
158 macroalgae ( $\delta^{13}\text{C} \sim -10$  to  $-15\text{‰}$ ) grow in nearshore areas and are exported offshore (Fry, 1983).

159 Oxygen stable isotope values in the ocean are influenced by evaporation and freshwater  
160 input, and correlate with salinity (Epstein and Mayeda, 1953). During incorporation into



161 biominerals (i.e., aragonite and calcite in shells and otoliths), stable isotopes of oxygen are  
162 fractionated and the magnitude of fractionation is directly related to water temperature according  
163 to the following equation (Trueman and St. John Glew, 2019):

164

$$165 \quad (1) \quad \delta^{18}O_{otolith} - \delta^{18}O_{water} = B - A(T) ,$$

166

167 where  $T$  is water temperature and  $A$  and  $B$  are species-specific coefficients. Average

168 annual sea surface temperature from NOAA sea surface temperature satellite data from 2010-

169 2020 was about 25°C in the northeastern GOM and about 27°C in the southeastern GOM (Huang

170 et al., 2015). This temperature difference is minor compared to the differences in  $\delta^{18}O_{water}$  values

171 between regions, so we expected to observe higher  $\delta^{18}O_{oto}$  values in FK fish than in AL fish at

172 least in the otolith edge region (point of capture).

173 The carbon isotope composition of fish otoliths is much more complex than the oxygen

174 isotope composition, and is deposited in disequilibrium with the surrounding environment

175 (Martino et al., 2020). Carbon isotopes in otoliths are a combination of dissolved inorganic

176 carbon (DIC) in the aquatic environment and oxidized organic carbon derived from the diet of

177 the fish (Solomon et al., 2006). Studies have estimated that the majority of otolith isotope

178 composition is derived from DIC (~80%) with the remaining 20% derived from the diet (Høie et

179 al., 2003; Nelson et al., 2011; Solomon et al., 2006; Tohse and Mugiya, 2007; Weidman and

180 Millner, 2000). However,  $\delta^{13}C_{DIC}$  values in coastal waters exhibit a fairly limited range

181 (approximately -2‰ to 2‰), with  $^{13}C$  often more depleted in estuarine and freshwater systems

182 than in marine systems (Bouillon et al., 2011). Metabolism also affects  $\delta^{13}C_{oto}$  values, as an

183 increase in respiration and metabolic carbon dioxide increases the proportion of metabolically

184 sourced carbon in the blood, which is deposited into the otolith (Martino et al., 2020). There is an

185 inverse relationship between  $\delta^{13}\text{C}_{\text{oto}}$  values and temperature, as temperature is a primary driver of  
186 metabolic rate (Martino et al., 2020). Due to the minor difference in average temperatures  
187 between GOM regions, we expected that changes in metabolic rate through ontogeny would be  
188 similar among all sampled Crevalle Jack, regardless of the region where they were collected.  
189 And since  $\delta^{13}\text{C}_{\text{DIC}}$  values in coastal waters are rather homogeneous, we therefore expected that  
190 any difference in  $\delta^{13}\text{C}_{\text{oto}}$  values between AL and FK fish would be mainly attributed to the  $\delta^{13}\text{C}$   
191 values of basal organic material. Therefore, we expected that FK Crevalle Jack would have  
192 higher  $\delta^{13}\text{C}_{\text{oto}}$  values than AL Crevalle Jack at least in the otolith edge region.

193 If juveniles exhibited local recruitment (i.e., if the two populations represented separate,  
194 self-recruiting stocks), we expected the  $\delta^{13}\text{C}_{\text{oto}}$  and  $\delta^{18}\text{O}_{\text{oto}}$  values of FK fish to be significantly  
195 different than the values for AL fish throughout the otoliths (H1; Fig. 2). Alternatively, if  
196 recruitment to both adult locations was from common nursery grounds, we would expect an  
197 overlap in the  $\delta^{13}\text{C}_{\text{oto}}$  and  $\delta^{18}\text{O}_{\text{oto}}$  values in the juvenile regions, followed by a divergence to  
198 either the FK or AL adult habitats (H2; Fig. 2). Adult connectivity would be evidenced by an  
199 overlap in the isotope values throughout adult ages (H3; Fig. 2). Otolith micro-chemistry has  
200 been previously used to test similar hypotheses about migratory patterns and connectivity of  
201 other fish species (Avigliano et al., 2021).

202 All FK fish were collected opportunistically in cooperation with local charter boat  
203 captains and recreational fishermen between 2019 and 2021. Otoliths from AL fish were  
204 collected by Mississippi State University (MSU) scientists from fish harvested by recreational  
205 anglers during the annual Alabama Deep Sea Fishing Rodeo from 2017-2019. Exact coordinates  
206 of capture locations were not recorded for most fish, but the broad area of capture was  
207 documented in all cases (Fig. 1). We aimed to analyze otoliths from the oldest fish possible so

208 we could capture the full time series of movements for each population. However, in Florida,  
209 fish over 4-years old were rarely encountered by charter boat captains, so we limited our  
210 analyses to fish 4-5 years old. A total of 12 otolith samples were collected for analysis per  
211 population ( $n = 24$  total; Supplementary, Table S1), with multiple transects milled per otolith for  
212 stable isotope analysis ( $n = 426$  total transects).

213

### 214 *2.3 Otolith sectioning and aging*

215 The left and right otoliths from each fish were embedded, sectioned, and mounted to  
216 microscope slides for further analysis, with the left otoliths used for aging and the right otoliths  
217 used for stable isotope analysis. We first embedded both otoliths in epoxy using a silicone mold  
218 and a two-part epoxy resin. Once solidified, embedded otoliths were sectioned through the  
219 nucleus using an Isomet low-speed diamond bladed saw with four blades separated by spacers to  
220 produce three thin sections per otolith. For the left otoliths, a 0.5 mm spacer was used, such that  
221 resulting thin sections were approximately 0.5 mm thick for aging. For the right otoliths, a 0.6  
222 mm spacer was used to produce thin sections approximately 0.6 mm thick. These thicker sections  
223 were used for stable isotope analysis as they provided a bit more otolith material for analysis. All  
224 thin sections were rinsed with water or 95% ethanol and adhered to clear microscope slides using  
225 a toluene-based mounting medium. Slides were allowed to dry for a minimum of 48 hours before  
226 further analysis.

227 Using the left otolith, opaque zones were counted from the core to the edge using a  
228 stereomicroscope following standard aging protocols (VanderKooy et al., 2020). Age class was  
229 determined based on the number of opaque zones and summer annulus deposition (Snelson,  
230 1992), i.e., age was determined as the number of opaque zones unless the fish was collected

231 between January 1 and July 31 and the margin code was 3 or 4. In this case, age was assigned as  
232 the number of opaque zones plus one. Each fish from the Florida Keys was read with at least two  
233 blind reads at the Florida Fish and Wildlife Conservation Commission Fish and Wildlife  
234 Research Institute (FWRI) in St. Petersburg, FL. If the two reads disagreed, then a third read was  
235 conducted, and the final age was determined from the consensus age of the three reads. Each fish  
236 from Alabama was aged by two readers at MSU. If the two readers disagreed on an age  
237 assignment, a third reader aged the otolith and final age was assigned if two out of the three  
238 readers agreed. If all three readers disagreed, then the first two readers consulted with each other  
239 and either reached an agreement or deemed the otolith unreadable.

240

#### 241 *2.4 Stable isotope analysis*

242 Oxygen and carbon stable isotopes can be used to examine spatial connectivity since  
243 stable isotope values vary predictably across the Gulf of Mexico as a function of climate and  
244 nutrient regimes (McMahon et al., 2013; Trueman et al., 2012). Sectioned otoliths not used for  
245 aging (right sagittal) were first photographed using a high-resolution camera affixed to a  
246 stereomicroscope. Using ImageJ software (Abràmoff et al., 2004), the approximate width of each  
247 growth band in mm was measured. Otoliths were then sampled using a New-Wave micromill in  
248 the University of Miami Stable Isotope Laboratory (Fig. 3). With the micromill system used in  
249 this study, the user manually traces transects across the sample region of interest on an image of  
250 the otolith on the computer screen. The computer then interpolates between two transects at a  
251 fixed distance with the material from each pass being milled on the advancing edge of the cutting  
252 blade. Transects were milled starting from the otolith edge and moving in towards the core in  
253 0.05 mm increments for a maximum of 21 transects per otolith (range = 16-21). Increments of

254 0.05 mm between transects were chosen to maximize temporal resolution while ensuring enough  
255 material was available for analysis (at least 0.04 mg of powder per transect was required).  
256 Powdered material was collected in glass vials for subsequent analysis. The  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$   
257 values of the multiple transects of each analyzed otolith were measured using a ThermoQuest  
258 Finnigan Delta Plus Mass Spectrometer (Thermo Fisher Scientific, Inc., Bremen, Germany)  
259 attached to a Kiel III automated carbonate device. Internal standards (six within each run)  
260 calibrated to NBS-19 (National Bureau of Standards) were processed along with a batch of 40  
261 samples. The measured values were corrected for the usual isobaric interferences and are  
262 reported relative to Vienna Pee Dee Belemnite (V-PDB) using the conventional notation.  
263 Standard deviations determined on the standards were less than 0.05‰ for  $\delta^{13}\text{C}$  and less than  
264 0.1‰ for  $\delta^{18}\text{O}$ .

265

## 266 *2.5 Statistical analysis*

267 According to independent 2-group t-tests for each growth band, there were no significant  
268 differences in growth band widths between the two groups of Crevalle Jack ( $p > 0.05$  for all t-  
269 tests). We therefore assumed growth rates were approximately the same between AL and FK and  
270 transect numbers corresponded to the same time in each fish's life. A series of regression models  
271 were fit to the carbon and oxygen data, with  $\delta^{18}\text{O}_{\text{oto}}$  or  $\delta^{13}\text{C}_{\text{oto}}$  values as the dependent variable.  
272 The goal of the regression modeling was to examine the typical trends in  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$   
273 values throughout the life history of Crevalle Jack and determine if isotopic profiles differed  
274 significantly between fish captured in isotopically distinct regions. Regression models were  
275 employed in a similar study that examined broad scale migration patterns of the Redthroat  
276 Emperor, *Lethrinus miniatus*, using otolith stable isotope chemistry (Currey et al., 2014). The

277 authors used models to test for significant differences in isotope signatures throughout otoliths  
278 by latitude, similar to how our study aimed to test for differences by region.

279 We initially fit several simple generalized linear models (GLMs; Nelder and  
280 Wedderburn, 1972) to the isotope data (linear, logarithmic, quadratic, and cubic), with  $\delta^{18}\text{O}_{\text{oto}}$  or  
281  $\delta^{13}\text{C}_{\text{oto}}$  values as the dependent variable and transect number (proxy for fish age) as the only  
282 independent variable. Residuals and q-q plots were used to assess normality and homogeneity of  
283 variance assumptions. For all models, the residuals did not appear to follow a normal  
284 distribution, thus violating the model assumptions. Linear models assume that there is a linear  
285 relationship between continuous dependent (Y) and independent (X) variables. However, this is  
286 not always the case. Real relationships in nature can have many different shapes that are not  
287 necessarily linear (Marrie et al., 2009). From an ecological point of view, it makes sense that the  
288 relationship between fish age and otolith stable isotope values may not be linear, since  
289 movement patterns of fishes are known to vary at different life stages (Kurth et al., 2019).  
290 Generalized additive models (GAMs; Wood, 2006) provide a robust way to model nonlinear  
291 trends. GAMs are basically extensions of GLMs that allow the expected independent variable to  
292 vary smoothly with a continuous dependent variable (Knappe, 2016). Furthermore, GAMs can be  
293 extended to incorporate random variables in a mixed model format (GAMM). Hanson et al.  
294 (2013) used GAMM models to reconstruct marine life-history strategies of Atlantic salmon,  
295 *Salmo salar*, L., from otolith stable isotopes. Following a similar methodology, GAMM models  
296 were fit in this study to Crevalle Jack  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$  data using the mgcv package in R  
297 (Wood, 2004).

298 Several models were fit and compared using Akaike's Information Criterion (AIC;  
299 Akaike, 1973). First, a simple model was fit with transect as the only independent variable

300 (model 1). Two additional models were fit to the data that incorporated population information;  
301 model 2 with transect and location as independent variables, and model 3 with transect, location,  
302 and a transect by location interaction term. All three models were compared using AIC and the  
303 model with the lowest AIC was selected as the best fitting base model. Data points taken from  
304 the same individual are likely to be correlated. To account for non-independence of the data and  
305 incorporate individual variability into the model, we then fit one additional model (model 4) that  
306 expanded upon the best fitting base model by including fish ID as a random variable in the form  
307 of a random intercept term. All analyses were conducted using R version 4.0 (R Core Team,  
308 2021).

309

### 310 **3. Results**

311 A total of 24 Crevalle Jack otoliths were analyzed, 12 from coastal Alabama and 12 from  
312 the Florida Keys. All fish were collected between 2017 and 2021 and ranged in age from 4-5  
313 years old, with 16 fish age-4 and 8 fish age-5 (Supplementary, Table S1). The number of  
314 transects milled per otolith ranged from 16 to 21, depending on the age of the fish and timing of  
315 collection. The average number of transects milled per otolith was 19.3, and the total sample size  
316 was 426 transects. As there were no significant differences in growth band widths between the  
317 two groups, the average estimated growth band widths from all fish were used to assign the  
318 following approximate ages to each transect: Transects-1-3 were assigned to age-0, transects-4-7  
319 to age-1, transects-8-11 to age-2, transects-12-15 to age-3, transects-16-19 to age-4, and  
320 transects-20-21 to age-5 (Table 1).

321

#### 322 *3.1 Model selection*

323 For both the  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$  data, GAMM models including transect, location, a  
324 transect location interaction term, and fish ID as a random intercept were the best fitting models  
325 according to AIC (Table 2). The model formula therefore took the following form:

326

$$327 \quad (2) \quad y_i = \beta_0 + \beta_1 \text{Location}_i + f_1(\text{Transect}_i) \text{Location}_i + \text{FishID}_i + \varepsilon_i,$$

328

329 where  $y$  is either  $\delta^{18}\text{O}_{\text{oto}}$  or  $\delta^{13}\text{C}_{\text{oto}}$ ,  $\beta_0 + \beta_1 \text{Location}$  is the parametric part of the linear predictor,  
330  $\text{Location}$  and  $\text{Transect}$  are predictor variables,  $f_1$  is a smooth function, and  $\text{FishID}$  is the random  
331 intercept term.

332

### 333 3.2 Lifetime stable isotope patterns

334 The random variable *Fish ID* was significant in both models, revealing significant  
335 individual variability in lifetime stable isotope profiles (Supplementary, Tables 2, 3). Most fish  
336 exhibited relatively low  $\delta^{13}\text{C}_{\text{oto}}$  and  $\delta^{18}\text{O}_{\text{oto}}$  values at age-0 followed by an increase in both  $\delta^{13}\text{C}_{\text{oto}}$   
337 and  $\delta^{18}\text{O}_{\text{oto}}$  values between ages-1 and -2. Across both groups,  $\delta^{13}\text{C}_{\text{oto}}$  and  $\delta^{18}\text{O}_{\text{oto}}$  values  
338 remained relatively stable throughout the remainder of the transects for most individuals.  
339 However, there were individuals that deviated from these common trends. For AL fish, the  
340  $\delta^{13}\text{C}_{\text{oto}}$  and  $\delta^{18}\text{O}_{\text{oto}}$  values for fish #AL 2, AL 7, AL 8, and AL 12 did not increase from age-0 to  
341 age-3, instead remaining relatively constant throughout the life history (Fig. 4). Additionally, fish  
342 #AL 4 and #AL 6 showed much more variability in isotope values over time, with several peaks  
343 and valleys. FK fish showed slightly less individual variability than AL fish, with the general  
344 pattern described above being evident for each fish except fish #FK 12 (Fig. 5). There was  
345 instead a substantial dip in both  $\delta^{13}\text{C}_{\text{oto}}$  and  $\delta^{18}\text{O}_{\text{oto}}$  values at age-3 for this individual.



346

### 347 3.3 Geographic variation

348 The best fitting GAMM models revealed that the two smooth curves for AL and FK fish  
349 were significantly different from zero ( $p < 0.001$ ; Supplementary, Tables 2, 3). In order to test  
350 whether the two curves were significantly different from each other, we fit additional models  
351 using difference curves, where one of the factor levels are taken as a reference level, a smooth is  
352 fit to the reference level, and then difference curves are fit to remaining factor levels (Baayen et  
353 al., 2018). The difference between the AL and FK curves was significant for both the  $\delta^{13}\text{C}_{\text{oto}}$  ( $edf$   
354  $= 4.8$ ,  $F = 58.2$ ,  $p < 0.001$ ) and  $\delta^{18}\text{O}_{\text{oto}}$  ( $edf = 4.4$ ,  $F = 29.9$ ,  $p < 0.001$ ) models. So, despite  
355 significant individual variability, GAMM model results revealed a significant difference in  
356 trends between AL and FK fish (Fig. 6). Furthermore, by plotting the difference smooths (Rij et  
357 al., 2022) between AL and FK for both models, we observed some significant differences  
358 between the regions at certain transect values (Fig. 7). A difference smooth is the difference  
359 between the smooths of two conditions, or levels of a factor. When the difference smooth 95%  
360 confidence interval does not contain zero, it reveals a significant difference between the two  
361 factor levels at the  $\alpha = 0.05$  level. At age-0, the  $\delta^{13}\text{C}_{\text{oto}}$  values for AL fish were significantly  
362 higher than for FK fish. Then from age-2 to age-5,  $\delta^{13}\text{C}_{\text{oto}}$  values were significantly higher for  
363 FK fish than for AL fish. There was much more overlap in  $\delta^{18}\text{O}_{\text{oto}}$  values between the two  
364 regions, but  $\delta^{18}\text{O}_{\text{oto}}$  values were significantly higher for FK fish than for AL fish at the end of the  
365 transects (ages 4 and 5).  $\delta^{18}\text{O}_{\text{oto}}$  values were also generally lower for FK fish than AL fish at age-  
366 0, similar to the  $\delta^{13}\text{C}_{\text{oto}}$  data. AL fish, in general, showed more homogeneous lifetime stable  
367 isotope profiles than FK fish.

368

369 **4. Discussion**

370 Stable isotope micro-sampling of fish otoliths can provide a wealth of information about  
371 habitat use throughout the life-history of a species, which is critical information for fisheries  
372 management and conservation (Compton et al., 2012; Galaiduk et al., 2017; Nagelkerken et al.,  
373 2015). Our research showed the potential of this approach for assessing lifetime migration  
374 patterns and connectivity of data-poor populations with wide-ranging distributions. In the eastern  
375 GOM, there is a substantial difference in environmental stable isotope values of carbon and  
376 oxygen between northern and southern coastal aquatic habitats (Fry, 1983; Vander Zanden et al.,  
377 2015). The main goal of our research was to determine whether otolith stable isotope micro-  
378 sampling could be used to distinguish between individual fish captured from both regions of the  
379 GOM (northeastern and southeastern), and whether fish from the two regions exhibited similar  
380 lifetime migration patterns. Our model results suggested that isotopic profiles of Crevalle Jack  
381 from Alabama and the Florida Keys were significantly different, but followed similar patterns of  
382 increasing  $\delta^{13}\text{C}_{\text{oto}}$  and  $\delta^{18}\text{O}_{\text{oto}}$  values from the otolith core to edge. However, there was overlap in  
383 stable isotope values between some individuals from the two regions, that could be attributed to  
384 inter-annual variability in environmental stable isotope values (as fish were collected over  
385 multiple years), or individual variability in movement patterns of Crevalle Jack, with potential  
386 connectivity between the northeastern and southeastern GOM regions. Our results did not appear  
387 to support any of our original hypotheses (Fig. 2), but instead supported a fourth hypothesis, that  
388 there is a difference in Crevalle Jack habitat use from age-0 to age-5 between the two regions,  
389 but that some level of connectivity may occur due to individual variability. The results of this  
390 study serve as a first step towards elucidating broad-scale movement patterns of an important  
391 sportfish species in the GOM.

392

393 *4.1 Lifetime migration patterns*

394 In our study, most Crevalle Jack otoliths displayed the same trend of increasing  $\delta^{18}\text{O}_{\text{oto}}$   
395 values from age-0 to about age-3 regardless of the region where they were collected (AL or FK).  
396 Among individual fish, the average difference between minimum and maximum  $\delta^{18}\text{O}_{\text{oto}}$  values  
397 throughout the otoliths was 5.27‰ for AL fish and 4.50‰ for FK fish (Supplementary, Table  
398 S4). Research has shown that temperature-induced  $\delta^{18}\text{O}_{\text{oto}}$  varies by approximately 1‰ per 4°C  
399 (Høie et al., 2004b). According to NOAA sea surface temperature satellite data, water  
400 temperature varies by less than 10°C seasonally in the Florida Keys region and by less than 15°C  
401 seasonally in the Alabama region (Huang et al., 2015). These seasonal temperature ranges are  
402 not enough to explain the full range of  $\delta^{18}\text{O}_{\text{oto}}$  values measured in individual Crevalle Jack  
403 otoliths. However, fish migrating among seawater, brackish estuaries, and freshwater  
404 environments can display much larger variations in  $\delta^{18}\text{O}_{\text{oto}}$  values (Hsieh et al., 2019). Open  
405 ocean seawater has measured  $\delta^{18}\text{O}_{\text{water}}$  values close to 0‰, while freshwater is typically depleted  
406 in  $^{18}\text{O}$ , with  $\delta^{18}\text{O}_{\text{water}}$  values as low as -10‰ (Kendall and Coplen, 2001; Lin et al., 2011).  
407 Previous research has additionally shown a positive correlation between  $\delta^{18}\text{O}$  values in fish  
408 otoliths and salinity (Hsieh et al., 2019; Pruell et al., 2012; Shirai et al., 2018). The increase in  
409  $\delta^{18}\text{O}_{\text{oto}}$  values from age-0 to age-3 observed in most Crevalle Jack otoliths likely reflects some  
410 combination of a decrease in water temperature and increase in salinity, which could be  
411 explained by an ontogenetic migration from warm, shallow water inshore habitats to cooler,  
412 higher salinity coastal habitats.

413 Measured  $\delta^{13}\text{C}_{\text{oto}}$  values also increased from age-0 to age-3 for most fish. Among  
414 individual fish, the average difference between minimum and maximum  $\delta^{13}\text{C}_{\text{oto}}$  values

415 throughout the otoliths was 4.68‰ for AL fish and 6.44‰ for FK fish (Supplementary, Table  
416 S4). The increase we observed in Crevalle Jack  $\delta^{13}\text{C}_{\text{oto}}$  values over time is likely predominantly  
417 due to a decrease in size-specific metabolism during ontogeny, which research has shown is the  
418 main driver of changes in  $\delta^{13}\text{C}_{\text{oto}}$  values (Chung et al., 2019; Høie et al., 2004a). For instance,  
419 Chung et al. (2019) showed that changes in metabolism associated with changing foraging and  
420 predation dynamics are clearly encapsulated in  $\delta^{13}\text{C}_{\text{oto}}$  values. As juveniles, fish metabolism and  
421 oxygen consumption are higher than at adult stages, as juveniles have higher predation risk, and  
422 increased pressure to grow (Trueman et al., 2016; Wurster and Patterson, 2003). As such,  
423 lifetime profiles of  $\delta^{13}\text{C}$  in fish otoliths commonly display increases due to ontogenetic changes  
424 in metabolism (Kalish, 1991). Besides metabolic changes, there are other factors that could  
425 contribute to the observed pattern in  $\delta^{13}\text{C}_{\text{oto}}$  values, including ontogenetic migration between  
426 habitats with distinct  $\delta^{13}\text{C}_{\text{DIC}}$  values, a shift in diet, possibly from a more terrestrial source (low  
427  $\delta^{13}\text{C}$  values) to a more marine source (high  $\delta^{13}\text{C}$  values; Peterson and Fry, 1987), and/or a shift  
428 in trophic position, as  $\delta^{13}\text{C}$  values typically increase by 1‰ for every increase in trophic level  
429 (Peterson and Fry, 1987). Additionally, there is an inverse relationship between  $\delta^{13}\text{C}_{\text{oto}}$  values  
430 and temperature, as temperature is a primary driver of metabolic rate (Martino et al., 2020).  
431 Future studies should investigate the external and internal mechanisms that regulate Crevalle  
432 Jack fractionation, to more specifically examine the drivers of changing  $\delta^{13}\text{C}_{\text{oto}}$  values over time.  
433 Due to the complexity of  $\delta^{13}\text{C}$  fractionation in fish otoliths, we cannot make any conclusions  
434 about Crevalle Jack migration through ontogeny solely based on the  $\delta^{13}\text{C}_{\text{oto}}$  data. However, the  
435 observed increase in  $\delta^{18}\text{O}_{\text{oto}}$  values through ontogeny combined with the observed increase in  
436  $\delta^{13}\text{C}_{\text{oto}}$  values do suggest an ontogenetic migration between ages-0 and -3.

437 This ontogenetic migration hypothesis aligns with previous observations of Crevalle Jack  
438 habitat use as well as previous literature. The Florida Fish and Wildlife Conservation  
439 Commission (FWC) conducts regular 183-m seine fisheries-independent monitoring (FIM)  
440 surveys in shallow, coastal estuaries. In all Gulf coast estuaries sampled between 1996 and 2018,  
441 length-frequency distributions of the catch showed that age-0 Crevalle Jack were most abundant,  
442 with few fish captured older than age-1 (FWC, 2021). Previous otolith micro-sampling research  
443 also attributed similar increases in both  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$  values over time observed in Chum  
444 Salmon (*Oncorhynchus keta*) to ontogenetic movement from inshore nurseries to more marine  
445 habitats (Wang et al., 2021). Ontogenetic migrations from inshore, estuarine nursery habitats to  
446 coastal and offshore subadult/adult habitats have been observed for several sportfish species in  
447 the GOM, including Red Drum (*Sciaenops ocellatus*; Winner et al., 2014) and Atlantic Tarpon  
448 (*Megalops atlanticus*; Kurth et al., 2019). Morphology and behavior change as fish grow, which  
449 often necessitates a change in habitat to meet energetic requirements and resource needs  
450 (Huijbers et al., 2015). Movement from juvenile to adult habitats may specifically be associated  
451 with reproduction or habitat shifts reflecting changing ratios of mortality risk and growth rates  
452 (Gillanders et al., 2003). Typically, individuals will reside in areas of low mortality risk as  
453 juveniles, when they are most susceptible to predation, and move to areas with higher mortality  
454 risk but better access to food resources as they grow and become less susceptible to predation.  
455 These ontogenetic habitat shifts therefore often involve trade-offs between fitness and survival  
456 that can impact physiological processes like growth (Higgins et al., 2015). Knowledge of these  
457 changes in habitat niche requirements over the life history of a species is crucial for  
458 understanding how community composition is structured across a mosaic of habitats (Compton

459 et al., 2012; Nagelkerken et al., 2015). This in turn aids in multispecies spatial management  
460 (Galaiduk et al., 2017).

461         Although most Crevalle Jack otoliths exhibited an increase in  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$  values  
462 from age-0 to age-3, there were a few individuals (all AL fish), for which isotope values did not  
463 increase (Fig. 4). Fish #AL 2, AL 7, AL 8, and AL 12 had relatively constant  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$   
464 values from age-0 to age-3. This result does not necessarily preclude ontogenetic migration, as it  
465 is possible these individuals simply moved between nursery and adult habitats that had similar  
466 isotopic profiles. The influence of the large river systems in the northern GOM extends out into  
467 the coastal environment, leading to substantial mixing between inshore and coastal areas that  
468 varies seasonally and annually (Sanial et al., 2019). This mixing could explain the relatively  
469 homogeneous stable isotope profiles of some AL Crevalle Jack even if they engaged in migration  
470 behavior, as oxygen stable isotope profiles would look homogeneous whether the fish was in  
471 inshore or coastal environments. We also noted individual variability in the range of  $\delta^{18}\text{O}_{\text{oto}}$  and  
472  $\delta^{13}\text{C}_{\text{oto}}$  values present throughout the otoliths of fish from both regions. Research has shown that  
473 variability in intrinsic factors such as body condition and sex can influence animal behavior and  
474 lead to substantial individual variability in lifetime movement and migration patterns within  
475 populations (Bolnick et al., 2011; Sih et al., 2004). Understanding the drivers and consequences  
476 of individual variability in movement patterns through ontogeny is important for understanding  
477 how environmental changes affect individual fitness of valuable species (Matich and Heithaus,  
478 2015). Future research that more closely examines individual variability in movement patterns of  
479 Crevalle Jack, using higher resolution methods such as acoustic telemetry, would help determine  
480 the extent and possible causes of individual variability in the population.

481 By age-3, Crevalle Jack appeared to have completed their ontogenetic migration, which  
482 coincides with estimated age at sexual maturity in the Caribbean (Caiafa et al., 2011), though age  
483 at maturation has been shown to vary significantly between populations of the same species, e.g.,  
484 for Bonefish (*Albula vulpes*; Rennert et al., 2019).  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$  values remained relatively  
485 constant to age-5. It is difficult to make any conclusions regarding movement patterns from age-  
486 3 to age-5 based solely on the  $\delta^{13}\text{C}_{\text{oto}}$  values due to the variety of factors that influence carbon  
487 fractionation in fish otoliths. But for several individuals,  $\delta^{18}\text{O}_{\text{oto}}$  values declined from age-3 to  
488 age-5, suggesting movement to an isotopically distinct habitat, perhaps lower salinity and/or  
489 warmer waters. The observed decrease in  $\delta^{18}\text{O}_{\text{oto}}$  values from age-3 to age-4 could feasibly  
490 represent a gradual movement towards the species preferred temperature (Fry, 1947). Adult  
491 Crevalle Jack occupy a range of habitats, including canals, deep-water reefs, and shallow-water  
492 flats (Smith-Vaniz and Carpenter, 2007). High-resolution stable isotope analysis of the older  
493 bands of the otoliths coupled with oceanographic data and other methods like acoustic telemetry  
494 (Ajemian et al., 2020), could increase our understanding of these fine-scale movement patterns  
495 post-ontogeny.

496

#### 497 *4.2 Geographic variation*

498 Overlap in individual stable isotope profiles between AL and FK fish precludes  
499 concluding that fish from the two regions represent distinct, self-recruiting populations with no  
500 mixing. However, there was a difference between the otolith stable isotope values of AL and FK  
501 Crevalle Jack, with GAMM models revealing a significant difference in trends between regions.  
502 The aforementioned substantial mixing between inshore and coastal regions in the northeastern  
503 GOM might explain why we saw different lifetime stable isotope patterns between regions in our

504 GAMM models. We might expect that nursery habitats in the northeastern GOM would have  
505 similar stable isotope values as nearby offshore habitats due to the substantial mixing. But in the  
506 southeastern GOM, nursery habitats may have more distinct stable isotope values than nearby  
507 offshore areas. This could explain why lifetime stable isotope profiles of AL fish were generally  
508 more homogeneous than those of FK fish.

509         At the point of capture (age-4 or age-5)  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$  values were significantly  
510 higher in FK fish than AL fish. In the GOM, numerous studies have examined spatial  
511 heterogeneity of  $\delta^{18}\text{O}$  and  $\delta^{13}\text{C}$  values in coastal water, primary producers, fish muscle tissue,  
512 and fish otoliths. Compared to the southeastern GOM (Florida Keys), where  $\delta^{18}\text{O}_{\text{water}}$  values in  
513 marine environments are relatively high ( $\sim 1.7\text{‰}$ ) due to high rates of evaporation (Sternberg  
514 and Swart, 1987), in the northeastern GOM (Alabama), influxes of freshwater from large river  
515 systems lead to relatively low  $\delta^{18}\text{O}_{\text{water}}$  values ranging from  $-2.2$  to  $-6.6\text{‰}$  (Wagner and Slowey,  
516 2011). This pattern has also been observed in fish otoliths. Gerard and Muhling (2010) examined  
517  $\delta^{18}\text{O}_{\text{oto}}$  values for juvenile Gray Snapper (*Lutjanus griseus*) in the Florida Keys region, with  
518 mean values around  $0\text{‰}$ . Similarly, Patterson et al. (2014) examined  $\delta^{18}\text{O}_{\text{oto}}$  values in juvenile  
519 Red Snapper (*Lutjanus campechanus*) and Lane Snapper (*Lutjanus synagris*) in coastal Alabama  
520 waters, and mean  $\delta^{18}\text{O}_{\text{oto}}$  values were about  $-1.8\text{‰}$  and  $-1.2\text{‰}$ , respectively. Vander Zanden et al.  
521 (2015) created a  $\delta^{13}\text{C}$  value isoscape for the Eastern Gulf of Mexico region using loggerhead  
522 scute tissue that showed clear  $\delta^{13}\text{C}_{\text{scute}}$  value enrichment in south Florida coastal waters ( $\sim -13$  to  
523  $-8\text{‰}$ ) compared to the northern GOM ( $\sim -20$  to  $-15\text{‰}$ ). Fry (1983) compared  $\delta^{13}\text{C}_{\text{muscle}}$  values in  
524 several species of shrimp throughout coastal GOM waters and had similar findings as Vander  
525 Zanden et al. (2015). The authors found a significant difference between the south Florida coast  
526 and the Louisiana and North Texas coasts, with lower  $\delta^{13}\text{C}_{\text{muscle}}$  values in shrimp from the latter



527 regions (~ -13 to -11‰ in Florida compared to ~ -19 to -16‰ in Louisiana). This previous  
528 research aligns with our results and confirms that the spatial heterogeneity of  $\delta^{18}\text{O}$  and  $\delta^{13}\text{C}$   
529 values in the eastern GOM is apparent in the tissues and biominerals of several aquatic  
530 organisms. Our results further highlight that otolith microchemistry can be a useful tool for  
531 assessing fish population dynamics in the region. As metabolic carbon comprises about 20% of  
532 fish otoliths, the difference of around 5 to 10‰ between  $\delta^{13}\text{C}$  values observed in loggerhead  
533 scute tissue and shrimp muscle tissue in the northeastern and southeastern GOM could explain  
534 the approximate 0.6 to 1.5‰ difference between  $\delta^{13}\text{C}_{\text{oto}}$  values of AL and FK Crevalle Jack post-  
535 ontogeny. Since  $\delta^{13}\text{C}_{\text{oto}}$  values for FK fish remained significantly higher than for AL fish from  
536 age-3 to age-5, it suggests most fish in this age range remained in their respective regions  
537 (northeastern or southeastern GOM), but some individuals may have moved between regions.

538         The significant difference in  $\delta^{13}\text{C}_{\text{oto}}$  values between AL and FK fish at age-0 (Fig. 7)  
539 suggests that these two groups of fish may have utilized geographically distinct nursery habitats.  
540 The difference in  $\delta^{13}\text{C}_{\text{oto}}$  values between nurseries could be due to differences in estuarine  
541 vegetation. In the GOM, salt marshes and mangroves are the predominant intertidal habitats, and  
542 there is a north-south transition with salt marshes ranging from 25 to 45°N (IFAS, 2016), and  
543 mangrove habitats dominating latitudes 30°N to 30°S (Giri et al., 2011). Therefore, south Florida  
544 mainly consists of mangrove intertidal habitats while Alabama is dominated by salt marsh  
545 habitats. Mangroves are predominantly  $\text{C}_3$  plants and salt marshes contain predominantly  $\text{C}_4$   
546 plants (Baker et al., 2021). Numerous studies have shown that  $\delta^{13}\text{C}$  values are significantly  
547 higher in  $\text{C}_4$  plants than  $\text{C}_3$  plants (Baker et al., 2013; Bouillon et al., 2008; Cerling et al., 1997;  
548 Currin et al., 1995), with  $\delta^{13}\text{C}$  values of  $\text{C}_4$  plants being around -12‰ and  $\delta^{13}\text{C}$  values of  $\text{C}_3$   
549 plants being around -24‰ to -30‰ (Bouillon et al., 2011). This regional difference in the base of

550 the estuarine food web could explain why the  $\delta^{13}\text{C}_{\text{oto}}$  values were significantly lower in age-0 FK  
551 fish than in age-0 AL fish. If this is the case, it suggests most fish may have recruited from  
552 inshore juvenile habitats to coastal habitats within the same region of the GOM.

553

#### 554 *4.3 Study limitations*

555 An assumption of our study was that growth rates did not vary significantly among all the  
556 sampled Crevalle Jack, and therefore each transect corresponded to the same region of the  
557 otolith. There were slight differences in measured growth band widths among otoliths (Table 1),  
558 but the differences were minor, suggesting that growth rates were similar among the individuals.  
559 Furthermore, there were no significant differences in growth band widths between the two  
560 populations (AL and FK) based on independent 2-group t-tests for each growth band.

561 Nevertheless, the slight differences in growth band widths likely explain some of the variance in  
562  $\delta^{13}\text{C}_{\text{oto}}$  and  $\delta^{18}\text{O}_{\text{oto}}$  values per transect among fish from the same region, and this variation also  
563 means that the groups of transects assigned to each age are approximate. Another potential  
564 source of variation is due to sampling of Crevalle Jack otoliths over multiple years (2017-2021).  
565 Carbon stable isotope values in the ocean can vary annually due to changes in productivity, and  
566 oxygen stable isotope values can vary with changes in runoff, evaporation, and precipitation.

567 This variability can thus effect  $\delta^{13}\text{C}_{\text{oto}}$  and  $\delta^{18}\text{O}_{\text{oto}}$  values (Trueman et al., 2012). Finally, while  
568 lifetime isotopic profiles were only examined through age-5, the Crevalle Jack is a relatively  
569 long-lived species with a maximum recorded age of 20 years (Jefferson et al., 2021). Therefore,  
570 movements beyond age-5 remain unknown. However, a challenge with micromilling otoliths  
571 from very old fish is that growth band width decreases with age as somatic growth slows.

572 Transects at these older bands would therefore likely reflect the combined isotopic environment

573 experienced over an entire year, or even multiple years, possibly obscuring migrations. Studies  
574 that combine otolith stable isotope analysis with other techniques like acoustic telemetry  
575 (Crossin et al., 2017) or genetics (Ovenden, 2013) can help further elucidate lifetime movement  
576 and migration patterns and connectivity of important species.

577

## 578 **5. Conclusions**

579         The results of our research provide previously unknown information about Crevalle Jack  
580 migration patterns through ontogeny and revealed a distinction between individuals captured in  
581 the Florida Keys and coastal Alabama that aligns with known isotopic variations between the  
582 two regions. These findings suggest that some Crevalle Jack may recruit from inshore estuaries  
583 to coastal habitats within the same broad region of the GOM and remain in either the  
584 northeastern or southeastern GOM regions through to at least age-5. However, there was  
585 substantial individual variability, suggesting that some individuals may move between regions  
586 during ontogeny (i.e., recruit to inshore nurseries at age-0 in one region but migrate to sub-adult  
587 habitats in the other region). Alternatively, the individual variability may instead be attributed to  
588 inter-annual differences in environmental stable isotope values within each region, as fish were  
589 collected over a period of several years. Unfortunately, there were no environmental stable  
590 isotope data available over the time period of our study for us to test this theory. Long-term  
591 monitoring of stable isotope values in water, sediment, and primary producer samples over broad  
592 spatial-temporal scales would aid in interpreting the results of studies such as this one, enhancing  
593 the utility of otolith stable isotope micro-sampling in the future. The information gained in this  
594 study about Crevalle Jack movement and migration patterns is critical for spatial management of  
595 the species and will aid in future conservation efforts. Micro-sampling of otoliths allowed us to

596 assess changes in movement and migration throughout the life history quickly and easily, with  
597 only a limited number of samples collected over a short time. Even though we had little *a priori*  
598 knowledge of Crevalle Jack migration patterns and population connectivity, well-known isotopic  
599 gradients in the Gulf of Mexico afforded us the opportunity to examine migration and  
600 connectivity without extensive, long-term sampling. As scientists continue to map isoscapes of  
601 marine and coastal environments and assess seasonal and annual variability, stable isotope  
602 research will further aid in understanding and managing important fish stocks. Our research  
603 serves as a case study for the potential applications of otolith stable isotope micro-analysis to  
604 assessing species ecology, population dynamics, and connectivity. Future work analyzing stable  
605 isotopes of Crevalle Jack juvenile otoliths from a range of nursery habitats could help identify  
606 linkages between juvenile and adult habitats, further elucidating stock structure. Furthermore, the  
607 methods outlined in this manuscript can be applied to any fish species and may be especially  
608 useful for future studies on migration and connectivity of data-poor species with broad spatial  
609 distributions.

610

#### 611 **CRedit authorship contribution statement**

612 **C.L. Gervasi:** Conceptualization, Data curation, Formal analysis, Funding acquisition,  
613 Investigation, Methodology, Project administration, Software, Visualization, Writing – original  
614 draft, Writing – review & editing. **J.A. Nelson:** Conceptualization, Validation, Writing – review  
615 & editing. **P.K. Swart:** Conceptualization, Formal analysis, Methodology, Resources,  
616 Supervision, Validation, Writing – review & editing. **R.O. Santos:** Conceptualization, Writing –  
617 review & editing. **R.J. Rezek:** Conceptualization, Writing – review & editing. **W.R. James:**  
618 Conceptualization, Writing – review & editing. **A.E. Jefferson:** Investigation, Methodology,

619 Resources, Writing – review & editing. **J.M. Drymon:** Investigation, Resources, Writing –  
620 review & editing. **J. Carroll:** Methodology, Resources, Writing – review & editing. **R.E.**  
621 **Boucek:** Conceptualization, Funding acquisition, Resources, Writing – review & editing. **J.S.**  
622 **Rehage:** Conceptualization, Funding acquisition, Project administration, Supervision,  
623 Validation, Writing – review & editing.

624

#### 625 **Declaration of competing interest**

626 The authors declare that they have no known competing financial interests or personal  
627 relationships that could have appeared to influence the work reported in this paper.

628

#### 629 **Acknowledgments**

630 The authors would like to acknowledge the recreational fishing guides who assisted with  
631 sample collection, including N. LaBadie, D. Kilpatrick, J. Hoover, Z. Routman, A. Tipler, J.  
632 Gagliardini, and J. Carlson, as well as the coordinators and participants of the Alabama Deep Sea  
633 Fishing Rodeo. Special thanks go to K. Rynerson and the rest of the age and growth lab at FWRI  
634 for assistance with otolith sectioning and aging. Much appreciation goes to the two anonymous  
635 reviewers for comments that improved the manuscript. This study was developed in  
636 collaboration with the Florida Coastal Everglades Long-Term Ecological Research program  
637 under National Science Foundation Grant No. DEB-1832229 and this work was reviewed and  
638 deemed exempt by Florida International University’s Institutional Animal Care and Use  
639 Committee (IACUC protocol exemption #IACUC-19-005, 8 January 2019). Funding was  
640 provided by the Lower Keys Guides Association, private donation, the Everglades Foundation  
641 FIU ForEverglades Graduate Scholarship, the American Fisheries Society Steven Berkeley

642 Marine Conservation Fellowship, and the FIU University Graduate School Doctoral Evidence  
643 Acquisition Fellowship. This is contribution #xxxx from the Institute of Environment at Florida  
644 International University.

645

## 646 **References**

- 647 Abràmoff, M.D., Magalhães, P.J., Ram, S.J., 2004. Image processing with imageJ. *Biophotonics*  
648 *Int.* 11, 36–41. <https://doi.org/10.1201/9781420005615.ax4>
- 649 Ajemian, M.J., Drymon, J.M., Hammerschlag, N., Wells, R.J.D., Street, G., Falterman, B.,  
650 McKinney, J.A., Driggers, W.B., Hoffmayer, E.R., Fischer, C., Stunz, G.W., 2020.  
651 Movement patterns and habitat use of tiger sharks (*Galeocerdo cuvier*) across ontogeny  
652 in the Gulf of Mexico. *PLoS ONE* 15, 1–24.  
653 <https://doi.org/10.1371/journal.pone.0234868>
- 654 Akaike, H., 1973. Maximum likelihood identification of Gaussian autoregressive moving  
655 average models. *Biometrika* 60, 255–265. <https://doi.org/10.1093/biomet/60.2.255>
- 656 Archambault, B., Le Pape, O., Baulier, L., Vermard, Y., Véron, M., Rivot, E., 2016. Adult-  
657 mediated connectivity affects inferences on population dynamics and stock assessment of  
658 nursery-dependent fish populations. *Fish. Res.* 181, 198–213.  
659 <https://doi.org/10.1016/j.fishres.2016.03.023>
- 660 Avigliano, E., Pisonero, J., Bouchez, J., Pouilly, M., Domanico, A., Sánchez, S., Clavijo, C.,  
661 Scarabotti, P., Facetti, J.F., Caffetti, J.D., del Rosso, F.R., Llamazares Vegh, S., Volpedo,  
662 A.V., 2021. Otolith Sr/Ca ratio complements Sr isotopes to reveal fish migration in large  
663 basins with heterogeneous geochemical landscapes. *Environ. Biol. Fishes* 104, 277–292.  
664 <https://doi.org/10.1007/s10641-021-01074-y>
- 665 Baayen, R.H., van Rij, J., de Cat, C., Wood, S., 2018. Autocorrelated Errors in Experimental  
666 Data in the Language Sciences: Some Solutions Offered by Generalized Additive Mixed  
667 Models, in: Speelman, D., Heylen, K., Geeraerts, D. (Eds.), *Mixed-Effects Regression*  
668 *Models in Linguistics, Quantitative Methods in the Humanities and Social Sciences.*  
669 Springer International Publishing, Cham, pp. 49–69. [https://doi.org/10.1007/978-3-319-](https://doi.org/10.1007/978-3-319-69830-4_4)  
670 [69830-4\\_4](https://doi.org/10.1007/978-3-319-69830-4_4)
- 671 Baker, R., Abrantes, K., Feller, I.C., 2021. Stable isotopes suggest limited role of wetland  
672 macrophyte production supporting aquatic food webs across a mangrove-salt marsh  
673 ecotone. *Estuaries Coasts* 44, 1619–1627. <https://doi.org/10.1007/s12237-021-00895-5>
- 674 Baker, R., Fry, B., Rozas, L.P., Minello, T.J., 2013. Hydrodynamic regulation of salt marsh  
675 contributions to aquatic food webs. *Mar. Ecol. Prog. Ser.* 490, 37–52.  
676 <https://doi.org/10.3354/meps10442>
- 677 Berger, A.M., Deroba, J.J., Bosley, K.M., Goethel, D.R., Langseth, B.J., Schueller, A.M.,  
678 Hanselman, D.H., 2021. Incoherent dimensionality in fisheries management:  
679 consequences of misaligned stock assessment and population boundaries. *ICES J. Mar.*  
680 *Sci.* 78, 155–171. <https://doi.org/10.1093/icesjms/fsaa203>
- 681 Bolnick, D.I., Amarasekare, P., Araújo, M.S., Bürger, R., Levine, J.M., Novak, M., Rudolf,  
682 V.H.W., Schreiber, S.J., Urban, M.C., Vasseur, D.A., 2011. Why intraspecific trait

683 variation matters in community ecology. *Trends Ecol. Evol.* 26, 183–192.  
684 <https://doi.org/10.1016/j.tree.2011.01.009>

685 Bouillon, S., Connolly, R.M., Gillikin, D.P., 2011. Use of stable isotopes to understand food  
686 webs and ecosystem functioning in estuaries. *Treatise Estuar. Coast. Sci.* 7, 143–173.  
687 <https://doi.org/10.1016/B978-0-12-374711-2.00711-7>

688 Bouillon, S., Connolly, R.M., Lee, S.Y., 2008. Organic matter exchange and cycling in  
689 mangrove ecosystems: Recent insights from stable isotope studies. *J. Sea Res.* 59, 44–58.  
690 <https://doi.org/10.1016/j.seares.2007.05.001>

691 Cadrin, S.X., Secor, D.H., 2009. Accounting for spatial population structure in stock assessment:  
692 past, present, and future, in: Beamish, R., Rothschild, B. (Eds.), *The Future of Fisheries  
693 Science in North America*. Springer Science+Business Media B.V., pp. 405–426.  
694 <https://doi.org/10.1007/978-1-4020-9210-7>

695 Caiafa, H.I., Narváez, B.J., Borrero, F.S., 2011. Algunos aspectos de la dinámica poblacional del  
696 jurel *Caranx hippos* (pisces: carangidae) en Bocas de Ceniza, Caribe colombiano. *Rev.  
697 MVZ Cordoba* 16, 2324–2335.

698 Campana, S.E., 2001. Accuracy, precision and quality control in age determination, including a  
699 review of the use and abuse of age validation methods. *J. Fish Biol.* 59, 197–242.  
700 <https://doi.org/10.1006/jfbi.2001.1668>

701 Campana, S.E., 1999. Chemistry and composition of fish otoliths: pathways, mechanisms and  
702 applications. *Mar. Ecol. Prog. Ser.* 188, 263–279.

703 Cerling, T.E., Harris, J.M., MacFadden, B.J., Leakey, M.G., Quade, J., Eisenmann, V.,  
704 Ehleringer, J.R., 1997. Global vegetation change through the Miocene/Pliocene  
705 boundary. *Nature* 389, 153–158. <https://doi.org/10.1038/38229>

706 Chung, M.T., Trueman, C.N., Godiksen, J.A., Holmstrup, M.E., Grønkjær, P., 2019. Field  
707 metabolic rates of teleost fishes are recorded in otolith carbonate. *Commun. Biol.* 2, 1–  
708 10. <https://doi.org/10.1038/s42003-018-0266-5>

709 Compton, T.J., Morrison, M.A., Leathwick, J.R., Carabines, G.D., 2012. Ontogenetic habitat  
710 associations of a demersal fish species, *Pagrus auratus*, identified using boosted  
711 regression trees. *Mar. Ecol. Prog. Ser.* 462, 219–230. <https://doi.org/10.3354/meps09790>

712 Crossin, G.T., Heupel, M.R., Holbrook, C.M., Hussey, N.E., Lowerre-Barbieri, S.K., Nguyen,  
713 V.M., Raby, G.D., Cooke, S.J., 2017. Acoustic telemetry and fisheries management.  
714 *Ecol. Appl.* 27, 1031–1049. <https://doi.org/10.1002/eap.1533>

715 Currey, L.M., Heupel, M.R., Simpfendorfer, C.A., Williams, A.J., 2014. Inferring movement  
716 patterns of a coral reef fish using oxygen and carbon isotopes in otolith carbonate. *J. Exp.  
717 Mar. Biol. Ecol.* 456, 18–25. <https://doi.org/10.1016/j.jembe.2014.03.004>

718 Currin, C. A., Newell, S. Y., Paerl, H. W., 1995. The role of standing dead *Spartina alterniflora*  
719 and benthic microalgae in salt marsh food webs: Considerations based on multiple stable  
720 isotope analysis. *Mar. Ecol. Prog. Ser.* 121, 99–116. <https://doi.org/10.3354/meps121099>

721 Defenbaugh, R., 1976. A Study of the benthic macroinvertebrates of the continental shelf of the  
722 northern Gulf of Mexico. Ph.D. dissertation. Texas A&M University, College Station  
723 TX.

724 Ditty, J.G., Shaw, R., Cope, J.S., 2004. Distribution of carangid larvae (Teleostei: Carangidae)  
725 and concentrations of zooplankton in the northern Gulf of Mexico, with illustrations of  
726 early *Hemicaranx amblyrhynchus* and *Caranx* spp. larvae. *Mar. Biol.* 145, 1001–1014.  
727 <https://doi.org/10.1007/s00227-004-1381-z>

728 Drymon, J.M., Dedman, S., Froeschke, J.T., Seubert, E.A., Jefferson, A.E., Kroetz, A.M.,  
729 Mareska, J.F., Powers, S.P., 2020. Defining sex-specific habitat suitability for a northern  
730 Gulf of Mexico shark assemblage. *Front. Mar. Sci.* 7, 35.  
731 <https://doi.org/10.3389/fmars.2020.00035>

732 Epstein, S., Mayeda, T., 1953. Variation of O<sup>2</sup> content of waters from natural sources. *Geochim.*  
733 *Cosmochim. Acta* 4, 213–224.

734 Estes, C., 2016. Comparison of reef fish communities east and west of Cape San Blas in  
735 northeast Gulf of Mexico. Thesis. Florida State University, Tallahassee, FL.

736 Flaherty, K.E., Matheson, R.E., McMichael, R.H., Perry, W.B., 2013. The influence of  
737 freshwater on nekton community structure in hydrologically distinct basins in  
738 northeastern Florida Bay, FL, USA. *Estuaries Coasts* 36, 918–939.  
739 <https://doi.org/10.1007/s12237-013-9614-3>

740 Flores-Coto, C., Sanchez-Ramirez, M., 1989. Larval distribution and abundance of carangidae  
741 (pisces), from the southern Gulf of Mexico, 1983-1984. *Gulf Res. Rep.* 8, 117–128.  
742 <https://doi.org/10.18785/grr.0802.04>

743 Fogarty, M.J., Botsford, L.W., 2007. Population connectivity and spatial management of marine  
744 fisheries. *Oceanography* 20, 112–123. <https://doi.org/10.5670/oceanog.2007.34>

745 Frisk, M.G., Jordaan, A., Miller, T.J., 2014. Moving beyond the current paradigm in marine  
746 population connectivity: Are adults the missing link? *Fish Fish.* 15, 242–254.  
747 <https://doi.org/10.1111/faf.12014>

748 Fry, B., 1983. Fish and shrimp migrations in the northern Gulf of Mexico analyzed using stable  
749 C, N, and S isotope ratios. *Fish. Bull.* 81, 789–801.

750 Fry, F., 1947. Effects of the environment on animal activity. *Publ. Ont. Fish. Res. Lab.* 68, 1–62.

751 FWC (Florida Fish and Wildlife Conservation Commission), 2021. Fisheries-independent  
752 monitoring using stratified-random sampling [WWW Document]. URL  
753 [https://myfwc.com/research/saltwater/reef-fish/monitoring/fim-stratified-random-](https://myfwc.com/research/saltwater/reef-fish/monitoring/fim-stratified-random-sampling/)  
754 [sampling/](https://myfwc.com/research/saltwater/reef-fish/monitoring/fim-stratified-random-sampling/) (accessed 9.17.21).

755 Galaiduk, R., Radford, B.T., Saunders, B.J., Newman, S.J., Harvey, E.S., 2017. Characterizing  
756 ontogenetic habitat shifts in marine fishes: Advancing nascent methods for marine spatial  
757 management. *Ecol. Appl.* 27, 1776–1788. <https://doi.org/10.1002/eap.1565>

758 Gallaway, B., 1981. An ecosystem analysis of oil and gas development on the Texas-Louisiana  
759 continental shelf. U.S. Fish and Wildlife Service FWS/OBS-81/27.

760 Gerard, T., Malca, E., Muhling, B.A., Mateo, I., Lamkin, J.T., 2015. Isotopic signatures in the  
761 otoliths of reef-associated fishes of southern Florida: Linkages between nursery grounds  
762 and coral reefs. *Reg. Stud. Mar. Sci.* 2, 95–104.  
763 <https://doi.org/10.1016/j.rsma.2015.08.014>

764 Gerard, T., Muhling, B., 2010. Variation in the isotopic signatures of juvenile gray snapper  
765 (*Lutjanus griseus*) from five southern Florida regions. *Fish. Bull.* 104, 98–105.

766 Gervasi, C.L., Santos, R.O., Rezek, R.J., James, W.R., Boucek, R.E., Bradshaw, C., Kavanagh,  
767 C., Osborne, J., Rehage, J.S., 2022. Bottom-up conservation: Using translational ecology  
768 to inform conservation priorities for a recreational fishery. *Can. J. Fish. Aquat. Sci.* 79,  
769 47–62. <https://doi.org/10.1139/cjfas-2021-0024>

770 Gillanders, B.M., Able, K.W., Brown, J.A., Eggleston, D.B., Sheridan, P.F., 2003. Evidence of  
771 connectivity between juvenile and adult habitats for mobile marine fauna: an important  
772 component of nurseries. *Mar. Ecol. Prog. Ser.* 247, 281–295.



- 773 Giri, C., Ochieng, E., Tieszen, L.L., Zhu, Z., Singh, A., Loveland, T., Masek, J., Duke, N., 2011.  
774 Status and distribution of mangrove forests of the world using earth observation satellite  
775 data. *Glob. Ecol. Biogeogr.* 20, 154–159. [https://doi.org/10.1111/j.1466-](https://doi.org/10.1111/j.1466-8238.2010.00584.x)  
776 8238.2010.00584.x
- 777 Hanson, N.N., Wurster, C.M., Todd, C.D., 2013. Reconstructing marine life-history strategies of  
778 wild Atlantic salmon from the stable isotope composition of otoliths. *Mar. Ecol. Prog.*  
779 *Ser.* 475, 249–266. <https://doi.org/10.3354/meps10066>
- 780 Helies, F., Jemison, J., Heyman, W., Gallaway, B., 2016. Prediction and verification of snapper-  
781 grouper spawning aggregation sites on the offshore banks of the Northwestern Gulf of  
782 Mexico, Final Report. NOAA/NMFS Award Number NA 14NMF4270039. Tampa, FL.
- 783 Heyman, W.D., Grüss, A., Biggs, C.R., Kobara, S., Farmer, N.A., Karnauskas, M., Lowerre-  
784 Barbieri, S., Erisman, B., 2019. Cooperative monitoring, assessment, and management of  
785 fish spawning aggregations and associated fisheries in the U.S. Gulf of Mexico. *Mar.*  
786 *Policy* 109, 103689. <https://doi.org/10.1016/j.marpol.2019.103689>
- 787 Heyman, W.D., Kjerfve, B., 2008. Characterization of transient multi-species reef fish spawning  
788 aggregations at Gladden Spit, Belize. *Bull. Mar. Sci.* 83, 531–551.
- 789 Higgins, R.M., Diogo, H., Isidro, E.J., 2015. Modelling growth in fish with complex life  
790 histories. *Rev. Fish Biol. Fish.* 25, 449–462. <https://doi.org/10.1007/s11160-015-9388-8>
- 791 Hinrichsen, H.H., Dickey-Collas, M., Huret, M., Peck, M.A., Vikebø, F.B., 2011. Evaluating the  
792 suitability of coupled biophysical models for fishery management. *ICES J. Mar. Sci.* 68,  
793 1478–1487. <https://doi.org/10.1093/icesjms/fsr056>
- 794 Hjort, J., 1914. Fluctuations in the great fisheries of Northern Europe. *Rapp. Procés-Verbaux* 20,  
795 1–228.
- 796 Høie, H., Andersson, C., Folkvord, A., Karlsen, 2004a. Precision and accuracy of stable isotope  
797 signals in otoliths of pen-reared cod (*Gadus morhua*) when sampled with a high-  
798 resolution micromill. *Mar. Biol.* 144, 1039–1049. [https://doi.org/10.1007/s00227-003-](https://doi.org/10.1007/s00227-003-1275-5)  
799 1275-5
- 800 Høie, H., Folkvord, A., Otterlei, E., 2003. Effect of somatic and otolith growth rate on stable  
801 isotopic composition of early juvenile cod (*Gadus morhua* L) otoliths. *J. Exp. Mar. Biol.*  
802 *Ecol.* 289, 41–58. [https://doi.org/10.1016/S0022-0981\(03\)00034-0](https://doi.org/10.1016/S0022-0981(03)00034-0)
- 803 Høie, H., Otterlei, E., Folkvord, A., 2004b. Temperature-dependent fractionation of stable  
804 oxygen isotopes in otoliths of juvenile cod (*Gadus morhua* L.). *ICES J. Mar. Sci.* 61,  
805 243–251. <https://doi.org/10.1016/j.icesjms.2003.11.006>
- 806 Hsieh, Y., Shiao, J.C., Lin, S., wood, Iizuka, Y., 2019. Quantitative reconstruction of salinity  
807 history by otolith oxygen stable isotopes: An example of a euryhaline fish *Lateolabrax*  
808 *japonicus*. *Rapid Commun. Mass Spectrom.* 33, 1344–1354.  
809 <https://doi.org/10.1002/rcm.8476>
- 810 Huang, B., Banzon, V., Freeman, E., Lawrimore, J., Liu, W., TC, P., Smith, T., Thorne, P.,  
811 Woodruff, S., Zhang, H.-M., 2015. NOAA Extended Reconstructed Sea Surface  
812 Temperature (ERSST), Version 4. [Dataset ID: nceiErsstv4]. NOAA National Centers for  
813 Environmental Information. doi:10.7289/V5KD1VVF (accessed 2.1.22).
- 814 Huijbers, C.M., Nagelkerken, I., Layman, C.A., 2015. Fish movement from nursery bays to coral  
815 reefs: a matter of size? *Hydrobiologia* 750, 89–101. [https://doi.org/10.1007/s10750-014-](https://doi.org/10.1007/s10750-014-2162-4)  
816 2162-4

817 IFAS (University of Florida Institute of Food and Agricultural Sciences Extension, 2016. Gulf  
818 Coast salt marshes [WWW Document]. URL  
819 <https://soils.ifas.ufl.edu/wetlandextension/types/gulfcoastmarsh.htm> (accessed 7.8.21).  
820 Jamieson, R.E., Schwarcz, H.P., Bratney, J., 2004. Carbon isotopic records from the otoliths of  
821 Atlantic cod (*Gadus morhua*) from eastern Newfoundland, Canada. *Fish. Res.* 68, 83–97.  
822 <https://doi.org/10.1016/j.fishres.2004.02.009>  
823 Jefferson, A., Jargowsky, M., Ivec, G., Cooper, P., Carroll, J., Gervasi, C., Rehage, J., Mareska,  
824 J., Powers, S., Drymon, J., 2021. Age, growth, and diet of crevalle jack (*Caranx hippos*)  
825 in the Gulf of Mexico. *Fish. Manag. Ecol.*  
826 Kalish, J.M., 1991. Oxygen and carbon stable isotopes in the otoliths of wild and laboratory-  
827 reared Australian salmon (*Arripis trutta*), *Marine Biology*.  
828 Kawazu, M., Tawa, A., Ishihara, T., Uematsu, Y., Sakai, S., 2020. Discrimination of eastward  
829 trans-Pacific migration of the Pacific bluefin tuna *Thunnus orientalis* through otolith  $\delta^{13}\text{C}$   
830 and  $\delta^{18}\text{O}$  analyses. *Mar. Biol.* 167, 1–7. <https://doi.org/10.1007/s00227-020-03723-9>  
831 Kendall, C., Coplen, T.B., 2001. Distribution of oxygen-18 and deuterium in river waters across  
832 the United States. *Hydrol. Process.* 15, 1363–1393. <https://doi.org/10.1002/hyp.217>  
833 Knape, J., 2016. Decomposing trends in Swedish bird populations using generalized additive  
834 mixed models. *J. Appl. Ecol.* 53, 1852–1861. <https://doi.org/10.1111/1365-2664.12720>  
835 Kramer, D.L., Chapman, M.R., 1999. Implications of fish home range size and relocation for  
836 marine reserve function. *Environ. Biol. Fishes* 55, 65–79.  
837 <https://doi.org/10.1023/a:1007481206399>  
838 Kurth, B.N., Peebles, E.B., Stallings, C.D., 2019. Atlantic Tarpon (*Megalops atlanticus*) exhibit  
839 upper estuarine habitat dependence followed by foraging system fidelity after ontogenetic  
840 habitat shifts. *Estuar. Coast. Shelf Sci.* 225, 106248.  
841 <https://doi.org/10.1016/j.ecss.2019.106248>  
842 Kwei, E.A., 1978. Food and spawning activity of *Caranx hippos* (L.) off the coast of Ghana. *J.*  
843 *Nat. Hist.* 12, 195–215.  
844 <https://doi.org/10.1080/00222937800770081>  
845 Levins, R., 1968. Evolution in changing environments: some theoretical explorations. Princeton  
846 University Press, Princeton, NJ.  
847 Lin, I.T., Wang, C.H., Lin, S., Chen, Y.G., 2011. Groundwater-seawater interactions off the  
848 coast of southern Taiwan: Evidence from environmental isotopes. *J. Asian Earth Sci.* 41,  
849 250–262. <https://doi.org/10.1016/j.jseaes.2011.03.001>  
850 Marrie, R.A., Dawson, N.V., Garland, A., 2009. Quantile regression and restricted cubic splines  
851 are useful for exploring relationships between continuous variables. *J. Clin. Epidemiol.*  
852 62. <https://doi.org/10.1016/j.jclinepi.2008.05.015>  
853 Martino, J.C., Doubleday, Z.A., Chung, M.T., Gillanders, B.M., 2020. Experimental support  
854 towards a metabolic proxy in fish using otolith carbon isotopes. *J. Exp. Biol.* 223.  
855 <https://doi.org/10.1242/jeb.217091>  
856 Matich, P., Heithaus, M.R., 2015. Individual variation in ontogenetic niche shifts in habitat use  
857 and movement patterns of a large estuarine predator (*Carcharhinus leucas*). *Oecologia*  
858 178, 347–359. <https://doi.org/10.1007/s00442-015-3253-2>  
859 McBride, R.S., McKown, K.A., 2000. Consequences of dispersal of subtropically spawned  
860 crevalle jacks, *Caranx hippos*, to temperate estuaries. *Fish. Bull.* 98, 528–538.

861 McMahon, K.W., Hamady, L.L., Thorrold, S.R., 2013. A review of ecogeochemistry approaches  
862 to estimating movements of marine animals. *Limnol. Oceanogr.* 58, 697–714.  
863 <https://doi.org/10.4319/lo.2013.58.2.0697>

864 Miller, T.J., 2007. Contribution of individual-based coupled physical-biological models to  
865 understanding recruitment in marine fish populations. *Mar. Ecol. Prog. Ser.* 347, 127–  
866 138. <https://doi.org/10.3354/meps06973>

867 Moffitt, E.A., Botsford, L.W., Kaplan, D.M., O’Farrell, M.R., 2009. Marine reserve networks for  
868 species that move within a home range. *Ecol. Appl.* 19, 1835–1847.  
869 <https://doi.org/10.1890/08-1101.1>

870 Mohan, J.A., Sutton, T.T., Cook, A.B., Boswell, K.M., Wells, R.J.D., 2017. Influence of  
871 oceanographic conditions on abundance and distribution of post-larval and juvenile  
872 carangid fishes in the northern Gulf of Mexico. *Fish. Oceanogr.* 26, 526–541.  
873 <https://doi.org/10.1111/fog.12214>

874 Nagelkerken, I., Sheaves, M., Baker, R., Connolly, R.M., 2015. The seascape nursery: A novel  
875 spatial approach to identify and manage nurseries for coastal marine fauna. *Fish Fish.* 16,  
876 362–371. <https://doi.org/10.1111/faf.12057>

877 Nelder, J.A., Wedderburn, R.W.M., 1972. Generalized linear models. *J. R. Stat. Soc. Ser. Gen.*  
878 135, 370–384.

879 Nelson, D. (Ed.), 1992. Distribution and abundance of fishes and invertebrates in Gulf of Mexico  
880 estuaries volume I: Data summaries. ELMR Report Number 10. NOAA/NOS Strategic  
881 Environmental Assessments Division, Rockville, MD.

882 Nelson, J., Hanson, C., Koenig, C., Chanton, J., 2011. Influence of diet on stable carbon isotope  
883 composition in otoliths of juvenile red drum *Sciaenops ocellatus*. *Aquat. Biol.* 13, 89–95.  
884 <https://doi.org/10.3354/ab00354>

885 NOAA (National Oceanic and Atmospheric Administration), 2021. About the Marine  
886 Recreational Information Program [WWW Document]. URL  
887 [https://www.fisheries.noaa.gov/recreational-fishing-data/about-marine-recreational-](https://www.fisheries.noaa.gov/recreational-fishing-data/about-marine-recreational-information-program)  
888 [information-program](https://www.fisheries.noaa.gov/recreational-fishing-data/about-marine-recreational-information-program) (accessed 9.9.21).

889 Ovenden, J.R., 2013. Crinkles in connectivity: Combining genetics and other types of biological  
890 data to estimate movement and interbreeding between populations. *Mar. Freshw. Res.* 64,  
891 201–207. <https://doi.org/10.1071/MF12314>

892 Patterson, W.F., Barnett, B.K., Sluis, M.Z., Cowan, J.H., Shiller, A.M., 2014. Interspecific  
893 variation in juvenile snapper otolith chemical signatures in the northern Gulf of Mexico.  
894 *Aquat. Biol.* 21, 1–10. <https://doi.org/10.3354/ab00567>

895 Peterson, B.J., Fry, B., 1987. Stable isotopes in ecosystem studies. *Annu. Rev. Ecol. Syst.* Vol 18  
896 18, 293–320. <https://doi.org/10.1146/annurev.es.18.110187.001453>

897 Pruell, R.J., Taplin, B.K., Karr, J.D., 2012. Spatial and temporal trends in stable carbon and  
898 oxygen isotope ratios of juvenile winter flounder otoliths. *Environ. Biol. Fish* 93, 61–71.  
899 <https://doi.org/10.1007/s10641-011-9890-3>

900 R Core Team, 2021. R: A language and environment for statistical computing. R Foundation for  
901 Statistical Computing. Vienna, Austria. <https://www.R-project.org/>.

902 Rennert, J.J., Shenker, J.M., Angulo-Valdés, J.A., Adams, A.J., 2019. Age, growth, and age at  
903 maturity of bonefish (*Albula species*) among Cuban habitats. *Environ. Biol. Fishes* 102,  
904 253–265. <https://doi.org/10.1007/s10641-018-0836-x>

905 Rij, J. van, Wieling, M., Baayen, R.H., Rijn, H. van, 2022. itsadug: Interpreting Time Series and  
906 Autocorrelated Data Using GAMMs. R package version 2.4.1.

- 907 Sale, P., Van Lavieren, H., Ablan Lagman, M., Atema, J., Butler, M., Fauvelot, C., Hogan, J.,  
908 Jones, G., Lindeman, K., Paris, C., Steneck, R., Stewart, H., 2010. Preserving reef  
909 connectivity: A handbook for marine protected area managers. Connectivity Working  
910 Group, Coral Reef Targeted Research & Capacity Building for Management Program,  
911 UNU-INWEH.
- 912 Sanial, V., Shiller, A.M., Joung, D.J., Ho, P., 2019. Extent of Mississippi River water in the  
913 Mississippi Bight and Louisiana Shelf based on water isotopes. *Estuar. Coast. Shelf Sci.*  
914 226, 106196. <https://doi.org/10.1016/j.ecss.2019.04.030>
- 915 Secor, D.H., 1999. Specifying divergent migrations in the concept of stock: The contingent  
916 hypothesis. *Fish. Res.* 43, 13–34. [https://doi.org/10.1016/S0165-7836\(99\)00064-8](https://doi.org/10.1016/S0165-7836(99)00064-8)
- 917 Shirai, K., Koyama, F., Murakami-Sugihara, N., Nanjo, K., Higuchi, T., Kohno, H., Watanabe,  
918 Y., Okamoto, K., Sano, M., 2018. Reconstruction of the salinity history associated with  
919 movements of mangrove fishes using otolith oxygen isotopic analysis. *Mar. Ecol. Prog.*  
920 *Ser.* 593, 127–139. <https://doi.org/10.3354/meps12514>
- 921 Sih, A., Bell, A., Johnson, J.C., 2004. Behavioral syndromes: An ecological and evolutionary  
922 overview. *Trends Ecol. Evol.* 19, 372–378. <https://doi.org/10.1016/j.tree.2004.04.009>
- 923 Smith-Vaniz, W., Carpenter, K.E., 2007. Review of the Crevalle Jacks, *Caranx hippos* Complex  
924 (Teleostei: Carangidae), with a description of a new species from West Africa. *Fish. Bull.*  
925 105, 207–233.
- 926 Snelson, F., 1992. Biological studies of crevalle jack (*Caranx hippos*) in Florida. Final Report  
927 submitted to Marine Research Institute. Florida Marine Research Institute. Orlando, FL.  
928 Florida Marine Research Institute, St. Petersburg.
- 929 Solomon, C.T., Weber, P.K., Cech, J.J., Ingram, B.L., Conrad, M.E., Machavaram, M.V.,  
930 Pogodina, A.R., Franklin, R.L., 2006. Experimental determination of the sources of  
931 otolith carbon and associated isotopic fractionation. *Can. J. Fish. Aquat. Sci.* 63, 79–89.  
932 <https://doi.org/10.1139/f05-200>
- 933 Sternberg, L. da S.L., Swart, P.K., 1987. Utilization of Freshwater and Ocean Water by Coastal  
934 Plants of Southern Florida. *Ecology* 68, 1898–1905.
- 935 Sturrock, A.M., Trueman, C.N., Darnaude, A.M., Hunter, E., 2012. Can otolith elemental  
936 chemistry retrospectively track migrations in fully marine fishes? *J. Fish Biol.* 81, 766–  
937 795. <https://doi.org/10.1111/j.1095-8649.2012.03372.x>
- 938 Tanner, S.E., Reis-Santos, P., Cabral, H.N., 2016. Otolith chemistry in stock delineation: A brief  
939 overview, current challenges and future prospects. *Fish. Res.* 173, 206–213.  
940 <https://doi.org/10.1016/j.fishres.2015.07.019>
- 941 Thorrold, S.R., Latkoczy, C., Swart, P.K., Jones, C.M., 2001. Natal homing in a marine fish  
942 metapopulation. *Science* 291, 297–299. <https://doi.org/10.1126/science.291.5502.297>
- 943 Tohse, H., Mugiya, Y., 2007. Sources of otolith carbonate: Experimental determination of  
944 carbon incorporation rates from water and metabolic CO<sub>2</sub>, and their diel variations.  
945 *Aquat. Biol.* 1, 259–268. <https://doi.org/10.3354/ab00029>
- 946 Trueman, C.N., Chung, M.-T., Shores, D., 2016. Ecogeochemistry potential in deep time  
947 biodiversity illustrated using a modern deep-water case study. *Philos. Trans. R. Soc. B*  
948 *Biol. Sci.* 371, 20150223. <https://doi.org/10.1098/rstb.2015.0223>
- 949 Trueman, C.N., Mackenzie, K.M., Palmer, M.R., 2012. Identifying migrations in marine fishes  
950 through stable-isotope analysis. *J. Fish Biol.* 81, 826–847. <https://doi.org/10.1111/j.1095-8649.2012.03361.x>  
951

952 Trueman, C.N., St. John Glew, K., 2019. Isotopic tracking of marine animal movement, in:  
953 Hobson, K., Wassenaar, L. (Eds.), *Tracking Animal Migration with Stable Isotopes*.  
954 Elsevier Science & Technology, Cambridge, MA, pp. 137–172.

955 Vander Zanden, H.B., Tucker, A.D., Hart, K.M., Lamont, M.M., Fujisaki, I., Addison, D.S.,  
956 Mansfield, K.L., Phillips, K.F., Wunder, M.B., Bowen, G.J., Pajuelo, M., Bolten, A.B.,  
957 Bjorndal, K.A., 2015. Determining origin in a migratory marine vertebrate: a novel  
958 method to integrate stable isotopes and satellite tracking. *Ecol. Appl.* 25, 320–335.

959 VanderKooy, S., Carroll, J., Elzey, S., Gilmore, J., Kipp, J. (Eds.), 2020. A practical handbook  
960 for determining the ages of Gulf of Mexico fishes, Publication Number 300. Gulf States  
961 Marine Fisheries Commission, Ocean Springs, MS.

962 Wagner, A.J., Slowey, N.C., 2011. Oxygen isotopes in seawater from the Texas-Louisiana shelf.  
963 *Bull. Mar. Sci.* 87, 1–12. <https://doi.org/10.5343/bms.2010.1004>

964 Walther, B.D., Limburg, K.E., 2012. The use of otolith chemistry to characterize diadromous  
965 migrations. *J. Fish Biol.* 81, 796–825. <https://doi.org/10.1111/j.1095-8649.2012.03371.x>

966 Wang, J., Gao, Y., Liu, W., Zhang, H., Dettman, D.L., 2021. The life history and populations of  
967 chum salmon (*Oncorhynchus keta*) in China: An otolith isotopic investigation. *Appl.*  
968 *Geochem.* 127, 104903. <https://doi.org/10.1016/j.apgeochem.2021.104903>

969 Ward, C.H., 2017. Habitats and biota of the Gulf of Mexico: Before the deepwater horizon oil  
970 spill. Volume 1: Water quality, sediments, sediment contaminants, oil and gas seeps,  
971 coastal habitats, offshore plankton and benthos, and shellfish. Springer, New York, NY.

972 Weidman, C.R., Millner, R., 2000. High resolution stable isotope records from North Atlantic  
973 cod. *Fish. Res.* 46, 327–342. <https://doi.org/10.1161/CIRCRESAHA.111.260943>

974 Winner, B.L., Flaherty-Walia, K.E., Switzer, T.S., Vecchio, J.L., 2014. Multidecadal evidence of  
975 recovery of nearshore red drum stocks off west-central Florida and connectivity with  
976 inshore nurseries. *North Am. J. Fish. Manag.* 34, 780–794.  
977 <https://doi.org/10.1080/02755947.2014.920737>

978 Wood, S., 2006. *Generalized Additive Models: An Introduction With R*. CRC Press, Boca Raton,  
979 FL.

980 Wood, S.N., 2004. Stable and Efficient Multiple Smoothing Parameter Estimation for  
981 Generalized Additive Models. *J. Am. Stat. Assoc.* 99, 673–686.  
982 <https://doi.org/10.1198/016214504000000980>

983 Wurster, C.M., Patterson, W.P., 2003. Metabolic rate of late Holocene freshwater fish: evidence  
984 from  $\delta^{13}\text{C}$  values of otoliths. *Paleobiology* 29, 492–505. [https://doi.org/10.1666/0094-8373\(2003\)029<0492:MROLHF>2.0.CO;2](https://doi.org/10.1666/0094-8373(2003)029<0492:MROLHF>2.0.CO;2)

986 Ying, Y., Chen, Y., Lin, L., Gao, T., 2011. Risks of ignoring fish population spatial structure in  
987 fisheries management. *Can. J. Fish. Aquat. Sci.* 68, 2101–2120.  
988 <https://doi.org/10.1139/F2011-116>

989 Zieman, J., Zieman, R., 1989. The ecology of the seagrass meadows of the west coast of Florida:  
990 A community profile. *U.S. Fish Wildl. Serv. Biol. Rep.* 85(7.25).

991

**Table 1.** Mean and standard error of growth band widths measured using ImageJ, approximate number of transects milled within each age band (each transect was 0.05 mm), and transect numbers assigned to each age. Age-4 growth bands were only measured for fish aged to be 5 years old since the entire age-4 band may not have been present at the time of collection for 4-year-old fish. Growth band widths were not measured for age 5, since age-5 fish were culled at varying times during year 5, and thus remaining transects were assigned to age-5.

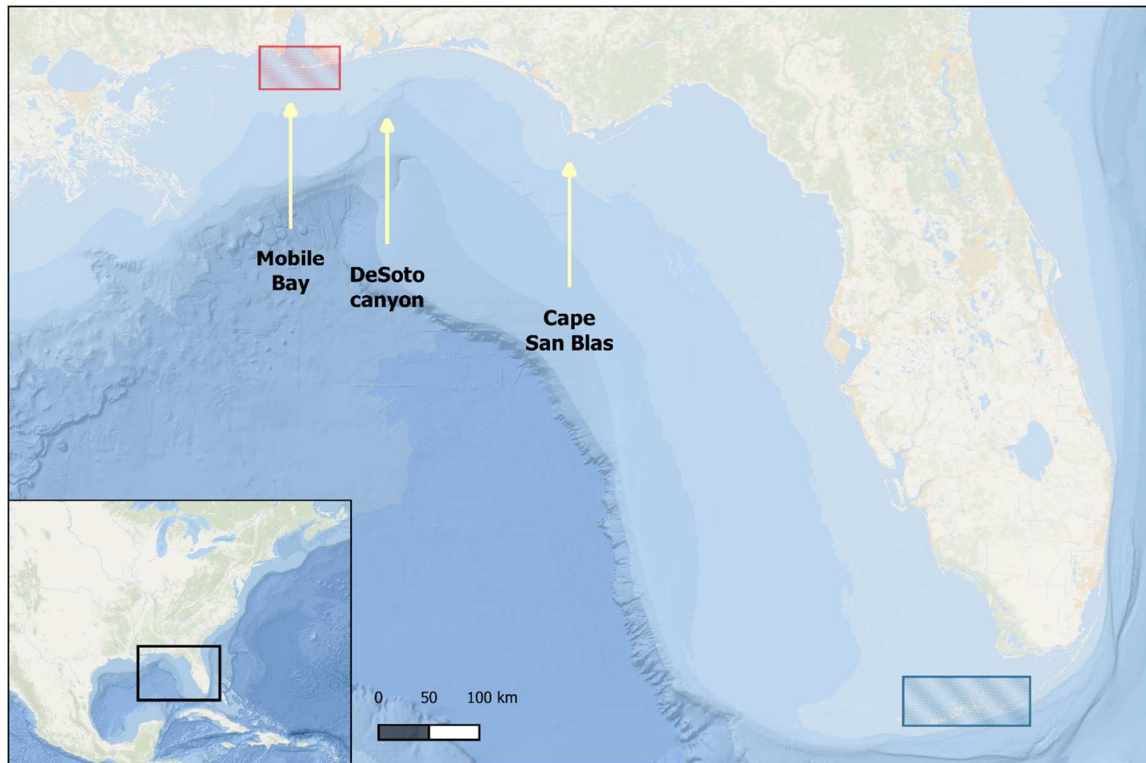
Growth band (age)	Mean width (mm)	Std. error	No. transects	Assigned transects
0	0.148	0.013	3	1-3
1	0.213	0.012	4	4-7
2	0.211	0.021	4	8-11
3	0.190	0.017	4	12-15
4	0.183	0.003	4	16-19
5	-	-	2	20-21

**Table 2.** Adjusted R<sup>2</sup>, AIC, and ΔAIC for each of the GAM models fit to the δ<sup>13</sup>C<sub>oto</sub> and δ<sup>18</sup>O<sub>oto</sub> data. Bolded terms indicate the best models selected.

δ <sup>18</sup> O <sub>oto</sub>	Variables	Adjusted R <sup>2</sup>	AIC	ΔAIC
Model 1	Transect	0.342	1471	86
Model 2	Transect, Location	0.350	1467	82
Model 3	Transect, Location, Transect x Location	0.378	1452	67
Model 4	<b>Transect, Location, Transect x Location, FishID (random intercept)</b>	<b>0.491</b>	<b>1385</b>	<b>0</b>

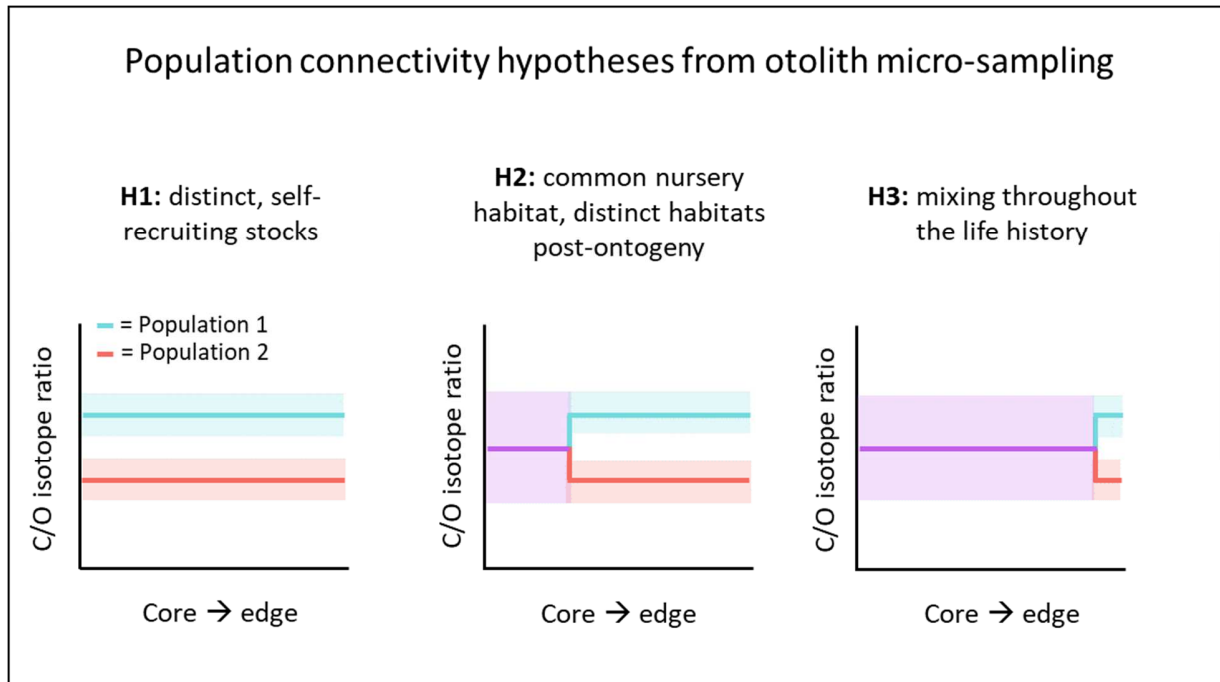
  

δ <sup>13</sup> C <sub>oto</sub>	Variables	Adjusted R <sup>2</sup>	AIC	ΔAIC
Model 1	Transect	0.386	1679	198
Model 2	Transect, Location	0.406	1665	184
Model 3	Transect, Location, Transect x Location	0.456	1632	151
Model 4	<b>Transect, Location, Transect x Location, FishID (random intercept)</b>	<b>0.637</b>	<b>1481</b>	<b>0</b>

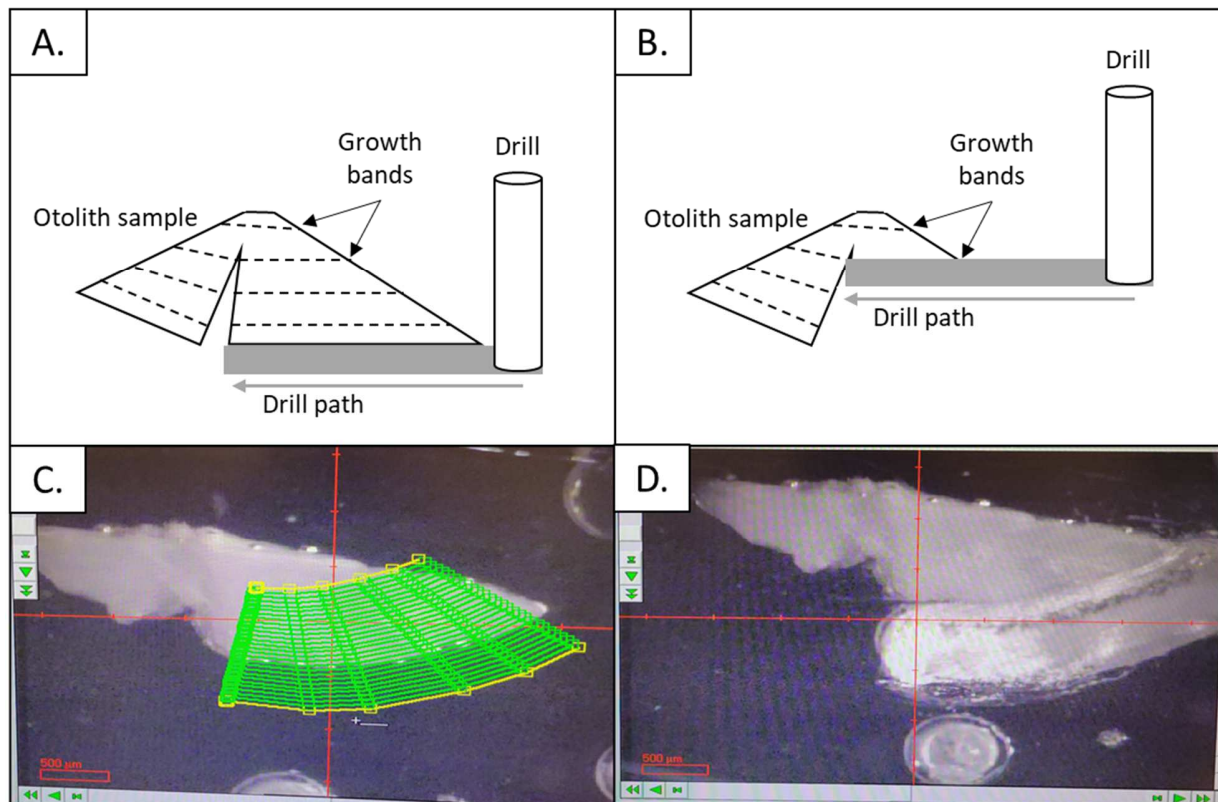


**Fig. 1.** Approximate capture locations of Crevalle Jack in Alabama (red hashed box) and Florida (blue hashed box) sampled for otolith stable isotope analysis. Exact GPS coordinates of capture locations were not recorded. Yellow arrows point to Mobile Bay, the DeSoto canyon, and Cape San Blas, areas that may serve as zoogeographic divides in the northern Gulf of Mexico. Inset map highlights study area in the southeastern United States. Map credit Esri, GEBCO, NOAA, National Geographic, DeLorme, HERE, Geonames.org, and other contributors.

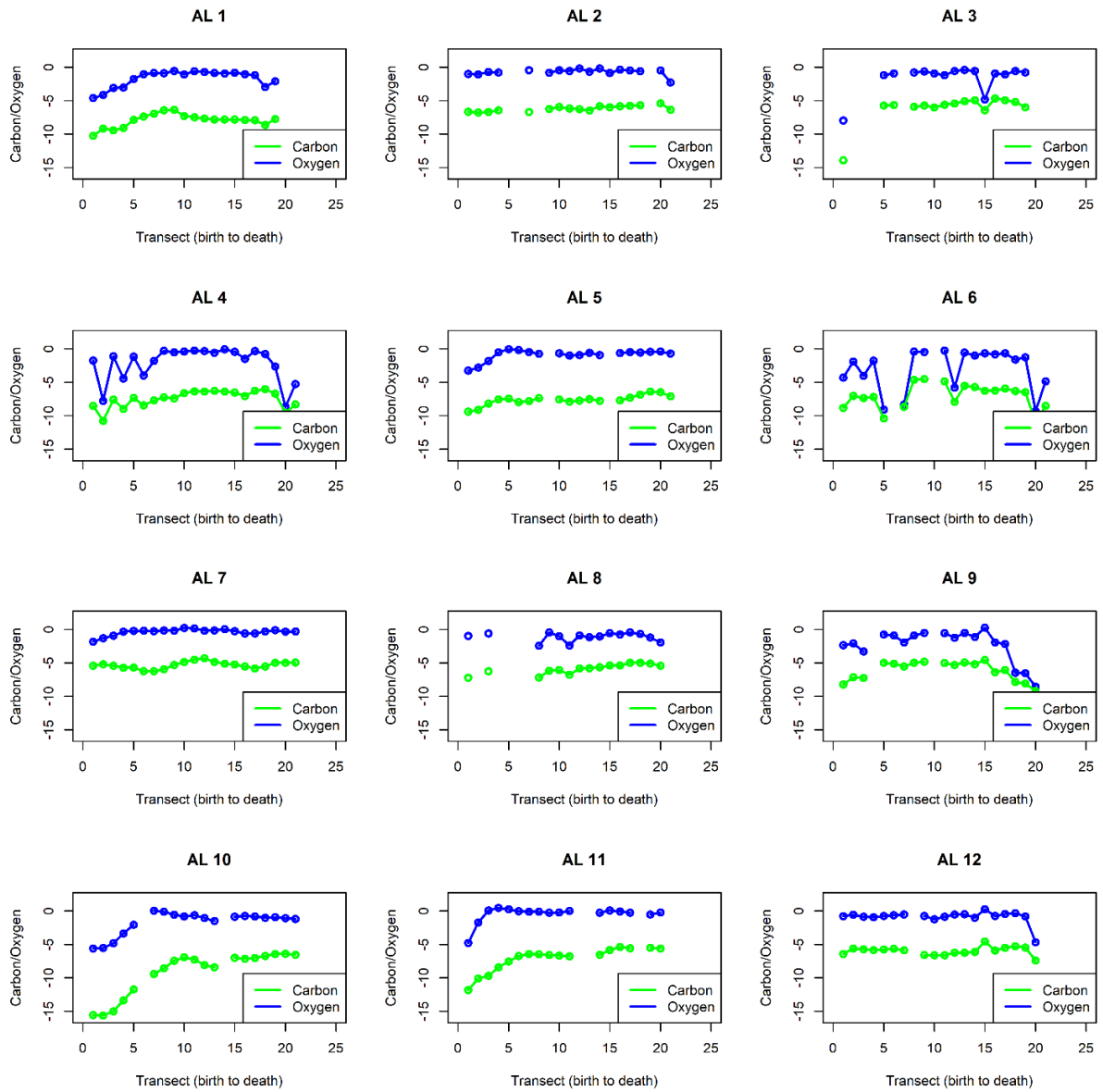




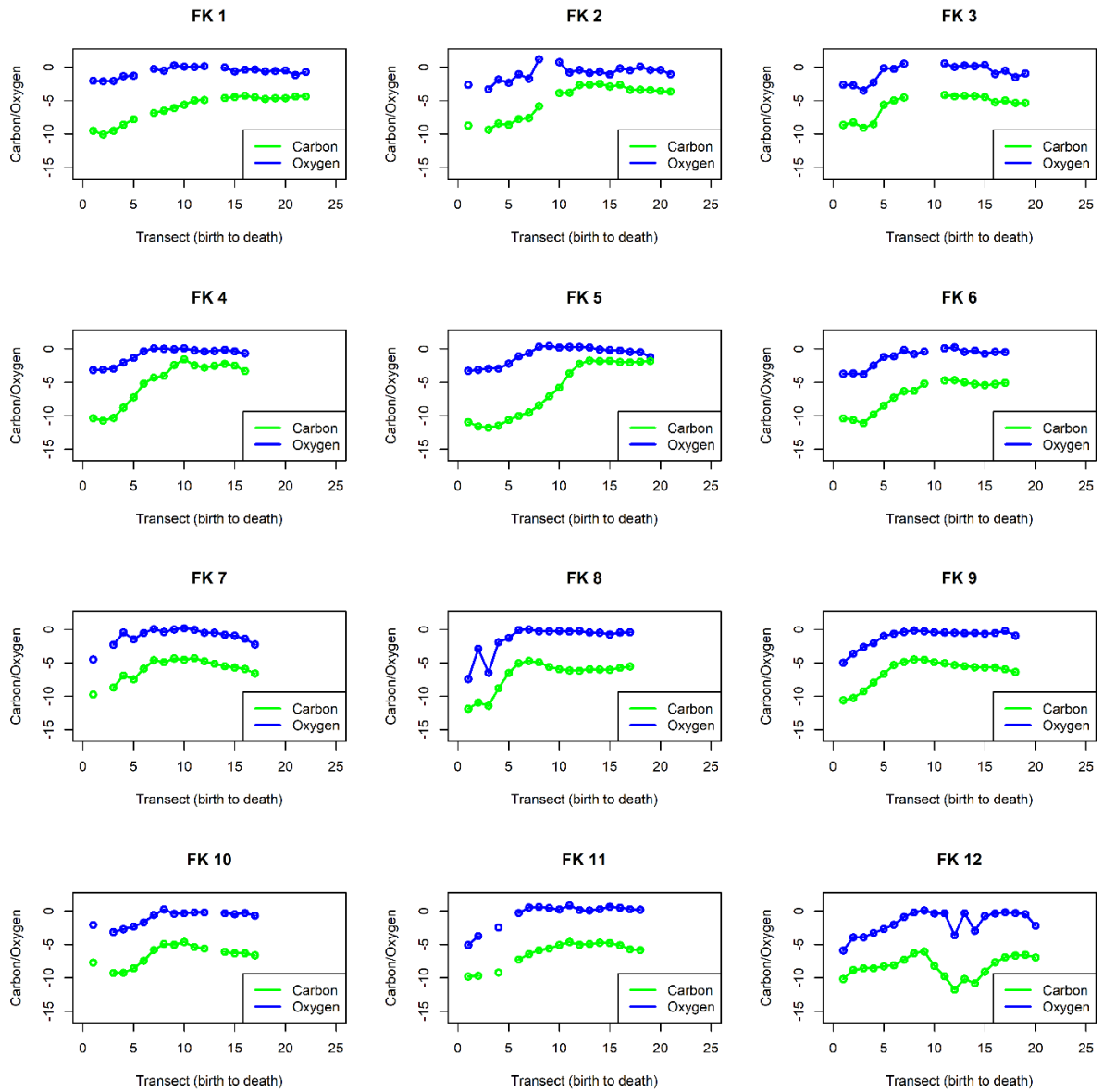
**Fig. 2.** Hypotheses concerning population connectivity between Florida Keys (blue lines) and Alabama (red lines) Creville Jack. Diagrams show simplified expected otolith stable isotope results corresponding to each hypothesis, with shaded regions denoting expected ranges of individual transects. If the two populations represent distinct, self-recruiting stocks (H1), we would expect  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$  values to differ between populations throughout the otolith, regardless of the actual isotope ratio values, with no overlap between individuals from the distinct regions. If there are common nursery habitats but distinct habitats post-ontogeny (H2), we would expect overlap in the  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$  values at earlier ages, with potentially substantial individual variability indicative of random distribution amongst juvenile habitats, followed by a divergence to different values post-ontogeny. If mixing between populations occurs throughout the life history (H3), we would expect overlap in  $\delta^{18}\text{O}_{\text{oto}}$  and  $\delta^{13}\text{C}_{\text{oto}}$  values everywhere except the otolith edge (point of capture), with substantial variability amongst individual fish.



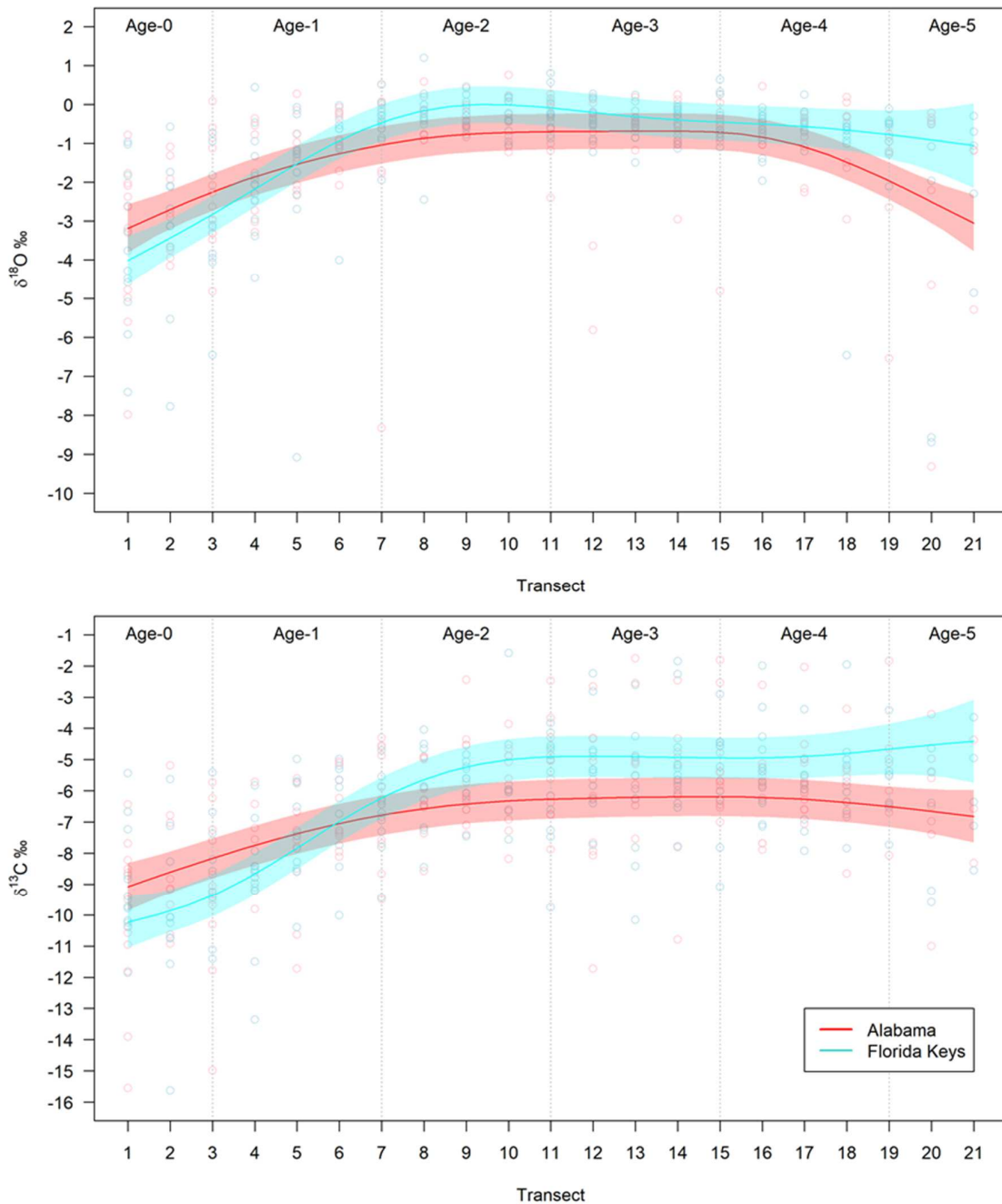
**Fig. 3.** Simplified schematic diagram of the micromilling system used in this study. The initial drill path runs parallel to the edge of the otolith section (A). Powdered sample is collected, and the drill is moved up the sample towards the core in increments of 0.05 mm until the entire otolith has been sampled. Drill path partway through the otolith (B). Image of the user-drawn initial and final drill paths (yellow lines) and computer interpolated drill paths (green lines) (C). Image of the otolith sample with the initial drill path completed (D).



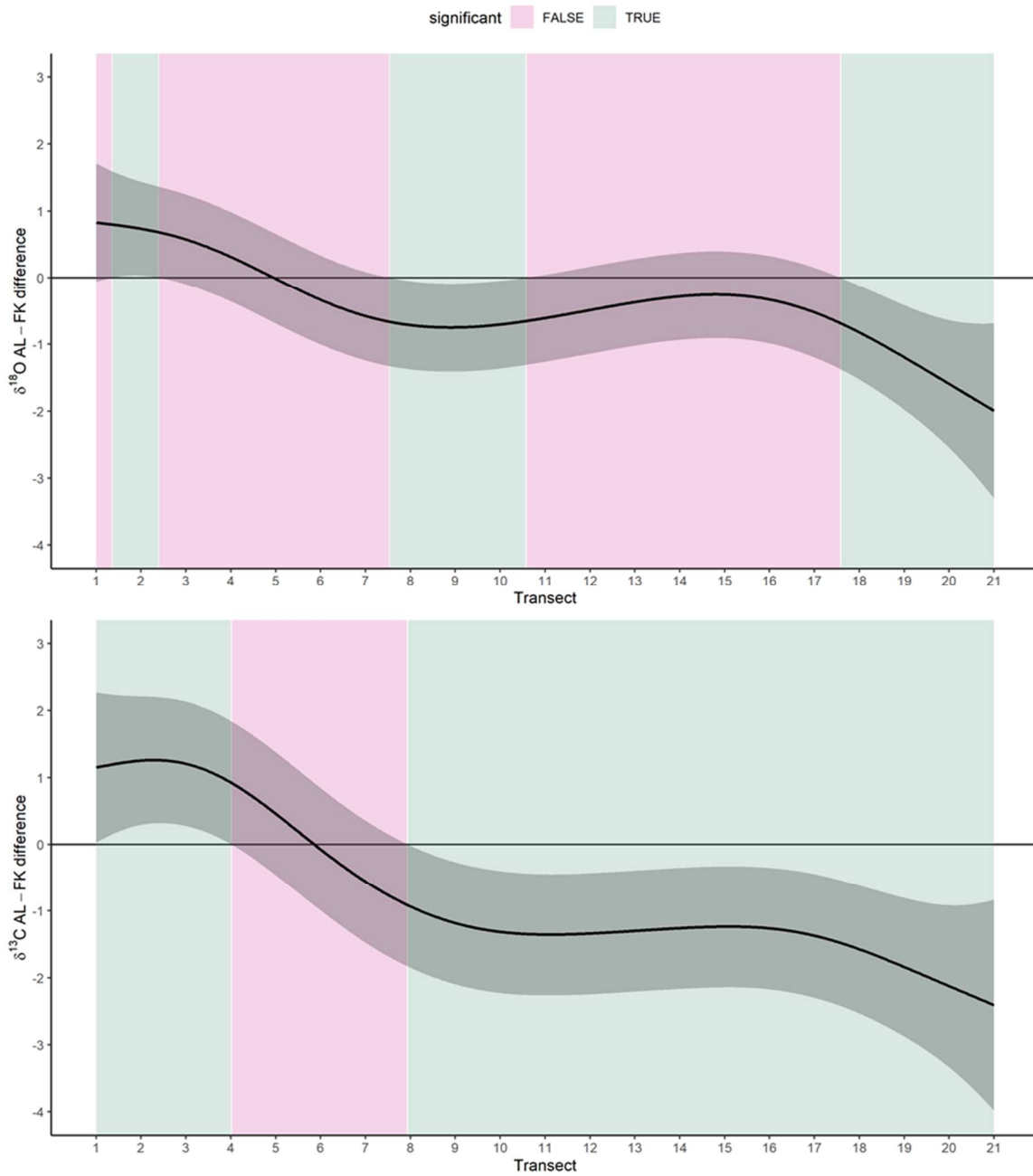
**Fig. 4.** Individual  $\delta^{18}\text{O}_{\text{oto}}$  (oxygen) and  $\delta^{13}\text{C}_{\text{oto}}$  (carbon) transects for each Crevalle Jack from Alabama (AL). Some transects have missing values due to instrument error or small sample size.



**Fig. 5.** Individual  $\delta^{18}\text{O}_{\text{oto}}$  (oxygen) and  $\delta^{13}\text{C}_{\text{oto}}$  (carbon) transects for each Crevalle Jack from the Florida Keys (FK). Some transects have missing values due to instrument error or small sample size.



**Fig. 6.** GAMM model predictions (lines) for  $\delta^{18}\text{O}_{\text{oto}}$  (top) and  $\delta^{13}\text{C}_{\text{oto}}$  (bottom). Red dots are raw data points and shaded red regions are 95% confidence intervals corresponding to Alabama Crevalle Jack (AL). Blue dots and shaded blue regions are 95% confidence intervals corresponding to Florida Crevalle Jack (FK). Gray dotted vertical lines are estimated age transition points, and numbers 0-5 across the top are estimated ages corresponding to each range of transects.



**Fig. 7.** Difference smooths for the  $\delta^{18}\text{O}_{\text{oto}}$  (top) and  $\delta^{13}\text{C}_{\text{oto}}$  (bottom) models between the levels of factor Location (AL and FK). Transects where the difference smooth is significantly different from zero are highlighted in green and transects where the difference smooth is not significantly different from zero are highlighted in pink.