

## RESEARCH ARTICLE

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## Key Points:

- The formaldehyde column depends on the concentrations of OH radical and volatile organic compounds
- Top-down constraints of volatile organic compound emissions should consider variations of OH radical
- The influence of nitrogen oxides on the formaldehyde column includes its feedback on OH radical concentrations

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## The role of OH production in interpreting the variability of CH<sub>2</sub>O columns in the southeast U.S.

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**Abstract** Formaldehyde (CH<sub>2</sub>O), a key atmospheric oxidation intermediate that is detectable from satellite-based UV/visible spectrometers, is primarily formed when hydroxyl radical (OH) reacts with volatile organic compounds (VOC) and is removed by photolysis, reaction with OH or deposition. We investigate the influence of OH and VOC variability on the CH<sub>2</sub>O column using a steady state model and the WRF-Chem regional chemical transport model over the southeast United States for the summer of 2012 (June–August). The steady state model indicates that the CH<sub>2</sub>O column primarily depends on OH production rates ( $P_{OH}$ ) at low concentrations of OH ( $<3 \times 10^6$  molecules cm<sup>-3</sup>), on both  $P_{OH}$  and VOC reactivity ( $VOCR: \sum_i k_i [VOC]_i$ ) at moderate concentrations of OH ( $3 \times 10^6$ – $7 \times 10^6$  molecules cm<sup>-3</sup>) and on VOCR at high concentrations of OH ( $>7 \times 10^6$  molecules cm<sup>-3</sup>). When constrained with WRF-Chem values of boundary layer average  $P_{OH}$  and VOCR, the steady state model of CH<sub>2</sub>O explains most of the daily ( $r^2 = 0.93$ ) and average June–August ( $r^2 = 0.97$ ) spatial variance of the fully simulated cloud-free CH<sub>2</sub>O column. These findings imply that measurements of the CH<sub>2</sub>O column offer the potential to better understand the processes affecting oxidation, particularly in remote regions, where OH concentrations are low. The findings also suggest that the inference of VOC emissions based on measurements of CH<sub>2</sub>O, or any other intermediate oxidation species with a photolytic lifetime that is short relative to removal by reaction with OH (e.g., glyoxal), should carefully account for the influence of OH on the observed pattern, especially where OH concentrations are below  $5 \times 10^6$  molecules cm<sup>-3</sup>, as occurs in remote forests, where OH strongly varies, as occurs downwind of large nitrogen oxide (NO<sub>x</sub>: NO+NO<sub>2</sub>) emission sources, or where OH sources are potentially biased.

### 1. Introduction

The hydroxyl radical (OH) is the most important oxidant in the troposphere, involved in the production of O<sub>3</sub> and particulate matter, and the removal of several greenhouse gases (e.g., CH<sub>4</sub> and HCFCs). Some fraction of reactions between OH and volatile organic compounds (VOC) leads to the formation of formaldehyde (CH<sub>2</sub>O). As the next generation of space-based UV/visible sensors capable of observing CH<sub>2</sub>O undergoes development [Chance *et al.*, 2013; Fishman *et al.*, 2012], there is a need and opportunity to better understand the relationship of the CH<sub>2</sub>O column with OH and its VOC precursors, such as methane and isoprene.

Remotely sensed UV-visible measurements offer the potential to routinely measure the spatial distribution of CH<sub>2</sub>O [e.g., De Smedt *et al.*, 2012; González Abad *et al.*, 2015]. Measurements from space-based spectrometers have provided extensive detail on the variability of the CH<sub>2</sub>O column on daily [e.g., Palmer *et al.*, 2006], seasonal [e.g., Martin *et al.*, 2004; Palmer *et al.*, 2006; Fu *et al.*, 2007] and interannual time scales [e.g., Abbot *et al.*, 2003; De Smedt *et al.*, 2010] at global [e.g., Stavrou *et al.*, 2009a, 2009b; Fortems-Cheiney *et al.*, 2012], regional [e.g., Fu *et al.*, 2007; Millet *et al.*, 2008; Dufour *et al.*, 2009; Marais *et al.*, 2012; Barkley *et al.*, 2013; Jin and Holloway, 2015], and urban spatial scales [e.g., Duncan *et al.*, 2010; Boeke *et al.*, 2011; Witte *et al.*, 2011; Zhu *et al.*, 2014]. The measurements of the CH<sub>2</sub>O column have primarily been used to investigate the emissions of isoprene (E<sub>ISOP</sub>) and other short-lived VOC that give rise to large, localized enhancements of CH<sub>2</sub>O [e.g., Palmer *et al.*, 2003, 2006; Fu *et al.*, 2007; Barkley *et al.*, 2008, 2013; Millet *et al.*, 2008; Stavrou *et al.*, 2009a, 2009b; Curci *et al.*, 2010; Marais *et al.*, 2012, 2014; Zhu *et al.*, 2014]. It is well known, however, that the concentration of CH<sub>2</sub>O also depends on the concentration of OH, both in isoprene-rich regions [e.g., Sumner *et al.*, 2001; Butler *et al.*, 2008; Galloway *et al.*, 2012] and in regions where CH<sub>4</sub> is the dominant VOC [e.g., Song *et al.*, 2010]. We thus hypothesize that variability in the CH<sub>2</sub>O column systematically reflects variations of the OH radical, in addition to reflecting variations of VOC.

To investigate this hypothesis, we first formulate a steady state equation for the CH<sub>2</sub>O column (section 2). We then test the assumptions underlying this steady state model using the WRF-Chem regional model (section 3) before discussing the implications for inferring E<sub>ISOP</sub> from CH<sub>2</sub>O column measurements (section 4). We also explore the potential for daily CH<sub>2</sub>O column measurements (section 5) to provide constraints on the production of OH radicals (P<sub>OH</sub>) and describe the necessary conditions for such constraints (section 6) before concluding (section 7).

## 2. A Steady State Model for the CH<sub>2</sub>O Column: $f(P_{OH}, \text{VOC Reactivity})$

The dominant source of CH<sub>2</sub>O is the reaction of OH with VOC, primarily CH<sub>4</sub>, isoprene or one of their products, with small contributions from ozonolysis and photolysis [Fortems-Cheiney *et al.*, 2012]. In contrast, the CH<sub>2</sub>O lifetime depends less on the concentration of OH as it is determined both by photolysis ( $J_{\text{CH}_2\text{O}} \approx 0.3 \text{ h}^{-1}$ ; typical of summertime northern midlatitudes) and by reaction with OH ( $0.3 \text{ h}^{-1}$  at  $[\text{OH}] = 1 \times 10^7$ ), with much slower loss to deposition. As a result, the concentration of CH<sub>2</sub>O depends on the concentration of OH, both in “background” conditions where most OH reacts with CH<sub>4</sub> or CO [e.g., Song *et al.*, 2010] and in isoprene-rich conditions [e.g., Butler *et al.*, 2008; Galloway *et al.*, 2012].

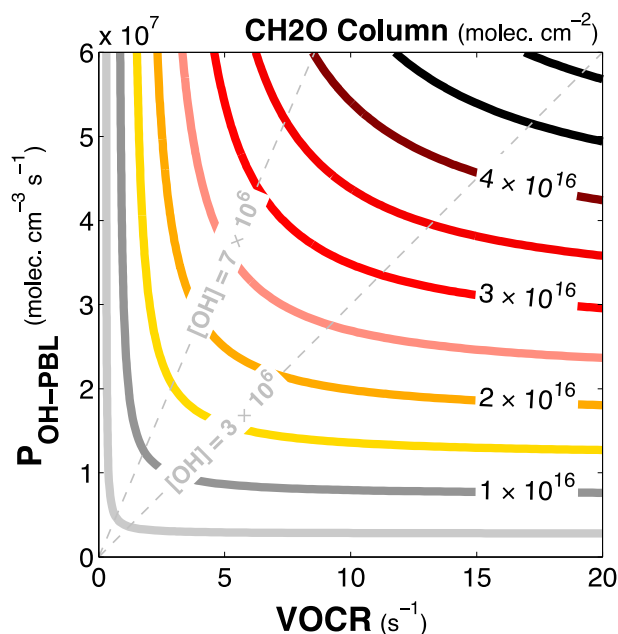
To isolate the influence of OH production (P<sub>OH</sub>) on the CH<sub>2</sub>O column, we write both CH<sub>2</sub>O sources and sinks in terms of P<sub>OH</sub>. P<sub>OH</sub> accounts for all primary (e.g., O<sup>1</sup>D + H<sub>2</sub>O) and secondary (e.g., NO + HO<sub>2</sub>) sources of OH. At steady state, OH production and OH loss are equal (i.e.,  $P_{\text{OH}} = \sum k_{\text{X}i+\text{OH}}[\text{X}_i][\text{OH}]$ ). The rate of CH<sub>2</sub>O production can be related to the rate of OH loss by incorporating the fractional yield of CH<sub>2</sub>O ( $\alpha_{\text{X}i}$ ) from each reaction that consumes OH ( $P_{\text{CH}_2\text{O}} = \sum \alpha_{\text{X}i} k_{\text{X}i+\text{OH}}[\text{X}_i][\text{OH}]$ ). Instead of treating each species and its yield separately [e.g.,  $\alpha_{\text{CH}_4} \leq 1$ ,  $\alpha_{\text{CO}, \text{NO}_2} = 0$ ; e.g., Sumner *et al.*, 2001; Galloway *et al.*, 2012], we define an effective yield ( $\alpha_{\text{eff}}: P_{\text{CH}_2\text{O}} P_{\text{OH}}^{-1}$ ). Thus, the rate of CH<sub>2</sub>O production equals the rate of OH production (P<sub>OH</sub>) multiplied by the effective yield of CH<sub>2</sub>O ( $\alpha_{\text{eff}}: P_{\text{CH}_2\text{O}} P_{\text{OH}}^{-1}$ ). We evaluate this approach by quantifying  $\alpha_{\text{eff}}$  in WRF-Chem (section 3.3) considering both the impacts of individual RO<sub>2</sub> radical pathways on  $\alpha_{\text{eff}}$  (section 3.3.1) and the impacts of VOC composition (section 3.4). The CH<sub>2</sub>O loss rate equals the rate of its loss by photolysis ( $J_{\text{CH}_2\text{O}}$ ) plus its loss by reaction with OH ( $k_{\text{CH}_2\text{O}+\text{OH}}[\text{OH}]$ ). The concentration of OH equals its source (P<sub>OH</sub>) divided by its sink, which we assume is dominated by the reactivity weighted concentration of VOC ( $\text{VOCR}: \sum_i k_i[\text{VOC}_i] + k_{\text{CO}+\text{OH}}[\text{CO}]$ ).

At midday, CH<sub>2</sub>O sources and sinks are both maximum and the CH<sub>2</sub>O lifetime is sufficiently short to assume that the CH<sub>2</sub>O concentration is in steady state. Assuming that alkene ozonolysis and photolysis are negligible sources of CH<sub>2</sub>O, we write a steady state equation for the CH<sub>2</sub>O concentration as

$$[\text{CH}_2\text{O}]_{\text{steady state}} = \frac{\alpha_{\text{eff}} P_{\text{OH}}}{(J_{\text{CH}_2\text{O}} + k_{\text{CH}_2\text{O}+\text{OH}} P_{\text{OH}} / \text{VOCR})} \quad (1)$$

At high OH concentrations (i.e., high P<sub>OH</sub> or low VOCR), loss by reaction with OH is dominant and the concentration of CH<sub>2</sub>O depends on VOCR ( $\sim \alpha_{\text{eff}} \text{VOCR} / k_{\text{CH}_2\text{O}+\text{OH}}$ ). At low OH concentrations (i.e., low P<sub>OH</sub> or high VOCR), loss by photolysis is dominant and the concentration of CH<sub>2</sub>O depends on P<sub>OH</sub> ( $\sim \alpha_{\text{eff}} P_{\text{OH}} / J_{\text{CH}_2\text{O}}$ ). For a given isoprene emission rate, the feedback of P<sub>OH</sub> on VOCR must also be considered because VOCR decreases as P<sub>OH</sub> increases. The precise magnitude of this decrease is uncertain and mechanism dependent [e.g., Karl *et al.*, 2009; Fuchs *et al.*, 2013; Nolscher *et al.*, 2014]. We quantify this effect in section 3.3.2 using WRF-Chem and find that the feedbacks of P<sub>OH</sub> on CH<sub>2</sub>O removal (i.e., via both P<sub>OH</sub> and VOCR; equation (1), denominator) do not offset the impact of P<sub>OH</sub> on CH<sub>2</sub>O production (equation (1), numerator).

Figure 1 shows steady state CH<sub>2</sub>O columns (contours) computed as a function of P<sub>OH</sub> (x axis) and VOCR (y axis) where we specify values for the effective yield ( $\alpha_{\text{eff}} = 0.30 P_{\text{CH}_2\text{O}} P_{\text{OH}}^{-1}$ ) and the loss rate constants ( $J_{\text{CH}_2\text{O}} = 0.3 \text{ h}^{-1}$ ;  $k_{\text{CH}_2\text{O}+\text{OH}} = 0.3 \times 10^{-7} \text{ cm}^3 \text{ molecule}^{-1} \text{ h}^{-1}$ ). To convert steady state CH<sub>2</sub>O concentrations to column values, we assume that VOCR and P<sub>OH</sub> are well mixed below 2 km. Above 2 km, CH<sub>2</sub>O photolysis is fast, VOCR is low, and P<sub>OH</sub> is low, thus resulting in a small contribution to the total CH<sub>2</sub>O column, which we assume here to be a constant value ( $2 \times 10^{15} \text{ molecules cm}^{-2}$ ). Where the concentration of OH is low ( $P_{\text{OH}} / \text{VOCR} < 3 \times 10^6 \text{ molecules cm}^{-3}$ ; Figure 1) and photolysis is dominant, the CH<sub>2</sub>O column is more



**Figure 1.** The steady state  $\text{CH}_2\text{O}$  column (colored curves; molecules  $\text{cm}^{-2}$ ) as a function of VOC reactivity ( $x$  axis;  $\text{VOCR}$ ;  $\text{s}^{-1}$ ) and OH production ( $y$  axis;  $P_{\text{OH}}$ ; molecules  $\text{cm}^{-3} \text{s}^{-1}$ ). The column is computed using equation (1) ( $\alpha_{\text{eff}} = 0.3 P_{\text{CH}_2\text{O}} P_{\text{OH}}^{-1}$ ,  $J_{\text{CH}_2\text{O}} = 0.30 \text{ h}^{-1}$ ,  $k_{\text{CH}_2\text{O}+\text{OH}}[\text{OH}] = 0.03 \text{ h}^{-1}$  at  $[\text{OH}] = 1 \times 10^6$  molecules  $\text{cm}^{-3}$ ) assuming that  $P_{\text{OH}}$  and  $\text{VOCR}$  are uniformly mixed from 0 to 2 km above ground level and are zero above 2 km. Values of constant OH concentration are indicated on the plot (gray dashed lines) under the assumption that  $\text{VOCR}$  is the only sink of OH radicals.

74.7°W–101.3°W) at 36 km horizontal grid resolution (50 N-S  $\times$  70 E-W) for 21 May 2012–31 August 2012. The model has 35 terrain-following hybrid pressure levels varying from  $\sim 85$  m thickness in the first model level to 300 m thickness in the layer 2.5 km above ground level and 3000 m thickness at the model top ( $z \approx 21$  km AGL). We analyze daily 1 P.M. CST model results for June – August 2012 and discard the five outermost grid rows of the domain to avoid excessive influence from the boundary conditions.

The chemical mechanism builds on the RACM2 photooxidation mechanism [Goliff *et al.*, 2013] with recent updates to the isoprene oxidation pathways and extended monoterpene and monoterpene nitrate chemistry [Browne *et al.*, 2014]. The updates to isoprene photochemistry follow those described by Paulot *et al.* [2009a, 2009b], with further updates to peroxy radical isomerization rates [Crouse *et al.*, 2011, 2012] and product yields [Peeters and Muller, 2010]. In addition to standard model diagnostics, we track the rates of different classes of reactions within the mechanism to quantify the processes controlling the production and loss of OH,  $\text{RO}_2$ , and  $\text{CH}_2\text{O}$ .

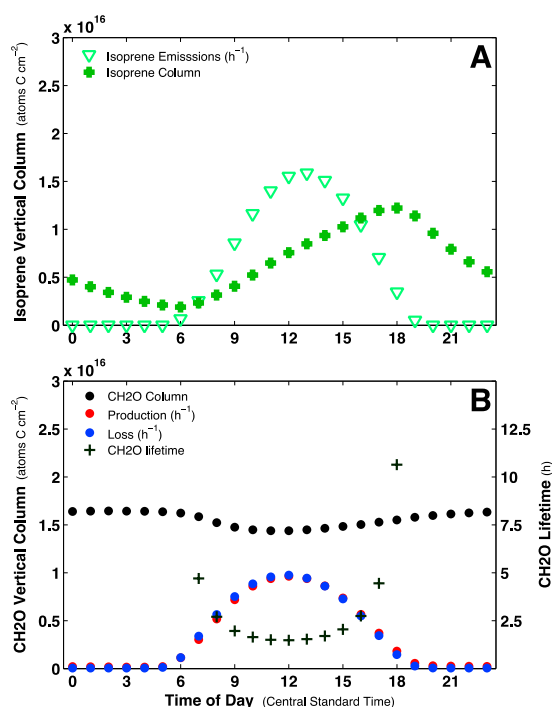
We use ERA-interim reanalysis [Dee *et al.*, 2011] as initial and boundary conditions for meteorological conditions (6-hourly), and MOZART-4/GEOS-5 output for chemical boundary conditions (6-hourly; Emmons *et al.*, 2010; <http://www.acd.ucar.edu/wrf-chem/mozart.shtml>). For emissions, we use the 2005 U.S. National Emission Inventory (NEI2005, available online [ftp://aftp.fsl.noaa.gov/divisions/taq/emissions\\_data\\_2005/](ftp://aftp.fsl.noaa.gov/divisions/taq/emissions_data_2005/)). The anthropogenic nitrogen oxide ( $\text{NO}_x$ :  $\text{NO} + \text{NO}_2$ ) emission rate ( $E_{\text{NO}_x}$ ) integrated over the domain is  $3.65 \text{ Tg N y}^{-1}$ . We also perform a sensitivity simulation in which we uniformly decrease anthropogenic  $E_{\text{NO}_x}$  by a factor of 10. There are no biomass burning or lightning emission sources included in the simulation and any aircraft emissions included in the NEI2005 are allocated to the lower model layers. Biogenic emissions are driven by MEGAN v2.04 [Guenther *et al.*, 2006] (<http://www.acd.ucar.edu/wrf-chem/download.shtml>). Domain total June–August 2012 biogenic  $E_{\text{NO}_x}$  is  $0.31 \text{ Tg N y}^{-1}$ , and  $E_{\text{ISOP}}$  is  $32 \text{ Tg C y}^{-1}$ . The emission rate for isoprene is in line with previous analyses [e.g., Millet *et al.*, 2008] given differences in domain size and interannual variability.

sensitive to changes of  $P_{\text{OH}}$  and less sensitive changes in  $\text{VOCR}$ . The opposite is true when the OH concentration is high ( $P_{\text{OH}}/\text{VOCR} > 7 \times 10^6$  molecules  $\text{cm}^{-3}$ ; Figure 1). At OH concentrations in between these end member cases, the  $\text{CH}_2\text{O}$  column is sensitive to both  $P_{\text{OH}}$  and  $\text{VOCR}$ . For a constant OH concentration, the sensitivity (slope) of the  $\text{CH}_2\text{O}$  column to a perturbation of  $\text{VOCR}$  is steep for high OH concentrations (e.g., a steep increase of  $\text{CH}_2\text{O}$  column contours following along the  $[\text{OH}] = 7 \times 10^6$  molecules  $\text{cm}^{-3}$  gray line; Figure 1) and is shallow for low OH concentrations (e.g., a shallow increase along the  $[\text{OH}] = 3 \times 10^6$  molecules  $\text{cm}^{-3}$ ; gray line; Figure 1).

### 3. Testing the Steady State Model With the Regional WRF-Chem Model

#### 3.1. Model Configuration

We now investigate the relationship of  $P_{\text{OH}}$  and  $\text{VOCR}$  to the  $\text{CH}_2\text{O}$  column using WRF-Chem, a state-of-the-art regional climate chemistry model [Grell *et al.*, 2005]. We simulate the chemistry over the southeast United States (27.7°N–43.8°N;



**Figure 2.** The diurnal profile of (a) isoprene emissions (atom  $\text{C cm}^{-2} \text{h}^{-1}$ ) and the isoprene column (atom  $\text{C cm}^{-2}$ ) and (b) the  $\text{CH}_2\text{O}$  vertical column (atom  $\text{C cm}^{-2}$ ) and the column integrated production rate (atom  $\text{C cm}^{-2} \text{h}^{-1}$ ), loss rate (atom  $\text{C cm}^{-2} \text{h}^{-1}$ ) and lifetime of  $\text{CH}_2\text{O}$  (h) for June–August 2012 over the southeast U.S. ( $32^\circ\text{N}$ – $38^\circ\text{N}$ ;  $75^\circ\text{W}$ – $90^\circ\text{W}$ ), following the approach of *Millet et al.* [2008].

and transport out of the region, in agreement with previous results [*Millet et al.*, 2008]. The increase of the isoprene column through midday, at a rate of  $2 \times 10^{15} \text{ atom C cm}^{-2} \text{h}^{-1}$  or 8% of the emission rate at noon, indicates that  $P_{\text{OH}}$  is insufficient to oxidize all available isoprene, let alone its oxidation products to  $\text{CH}_2\text{O}$ . Thus, we conclude that isoprene is not in steady state with its emissions.

### 3.2.2. Spatial Variability of the $\text{CH}_2\text{O}$ Column

We next explore the spatial patterns of summer average  $E_{\text{ISOP}}$ ,  $\text{VOCR}$ ,  $P_{\text{OH}}$  and the steady state estimate of the  $\text{CH}_2\text{O}$  column at 1 P.M. CST (Figure 3, left column). Unless otherwise stated, we compute the steady state value of the  $\text{CH}_2\text{O}$  column using simulated PBL-average ( $< 2.5 \text{ km}$ ) values for  $P_{\text{OH}}$  and  $\text{VOCR}$  as inputs to equation (1), and we specify the yield of  $\text{CH}_2\text{O}$  per OH consumed ( $\alpha_{\text{eff}} = 0.30 P_{\text{CH}_2\text{O}} P_{\text{OH}}^{-1}$ ) and loss rate constants ( $J_{\text{CH}_2\text{O}} = 0.3 \text{ h}^{-1}$ ;  $k_{\text{CH}_2\text{O}+\text{OH}} = 0.3 \times 10^{-7} \text{ molecules}^{-1} \text{ cm}^3 \text{ h}^{-1}$ ). We convert the PBL steady state value to a column value in all grid cells using the slope ( $3.1 \times 10^5 \text{ cm}$ ) and y intercept ( $2 \times 10^{15} \text{ molecules cm}^{-2}$ ) derived from the reduced major axis linear regression of the fully simulated  $\text{CH}_2\text{O}$  column to the PBL average  $\text{CH}_2\text{O}$  concentration for all daily 1 P.M. CST cloud-free WRF-Chem grid cells over land ( $r^2 = 0.98$ ;  $N = 106,320$ ). The strong correlation supports our assumption that  $\text{CH}_2\text{O}$  concentrations below 2.5 km drive the overall variability of the column, with a smaller and more uniform contribution above 2.5 km.

The June–August 2012 average spatial pattern of isoprene emissions is highly variable, with a maximum over the Ozarks ( $35^\circ\text{N}$ – $38^\circ\text{N}$ ;  $91^\circ\text{W}$ – $95^\circ\text{W}$ ) where emissions exceed all other high-emitting regions by 50%–100% (Figure 3a) [e.g., *Wiedinmyer et al.*, 2005; *Carlton and Baker*, 2011]. The spatial pattern of  $\text{VOCR}$  closely resembles the pattern of  $E_{\text{ISOP}}$  with slight smoothing due to transport ( $r^2 = 0.90$ ; Figure 3a versus 3d), indicating the importance of isoprene emission to total  $\text{VOCR}$  during summer in the region [e.g., *Millet et al.*, 2008]. On average,  $P_{\text{OH}}$  is relatively uniform at a regional scale except for local maxima over large  $\text{NO}_x$  emission sources where  $\text{NO} + \text{HO}_2$  reactions strongly enhance  $P_{\text{OH}}$  (Figure 3g). According to Figure 1, at  $P_{\text{OH}}$  of  $2.5 \times 10^7 \text{ molecules cm}^{-3} \text{ s}^{-1}$  (i.e., the regional background value; Figure 3g), the  $\text{CH}_2\text{O}$  column will linearly

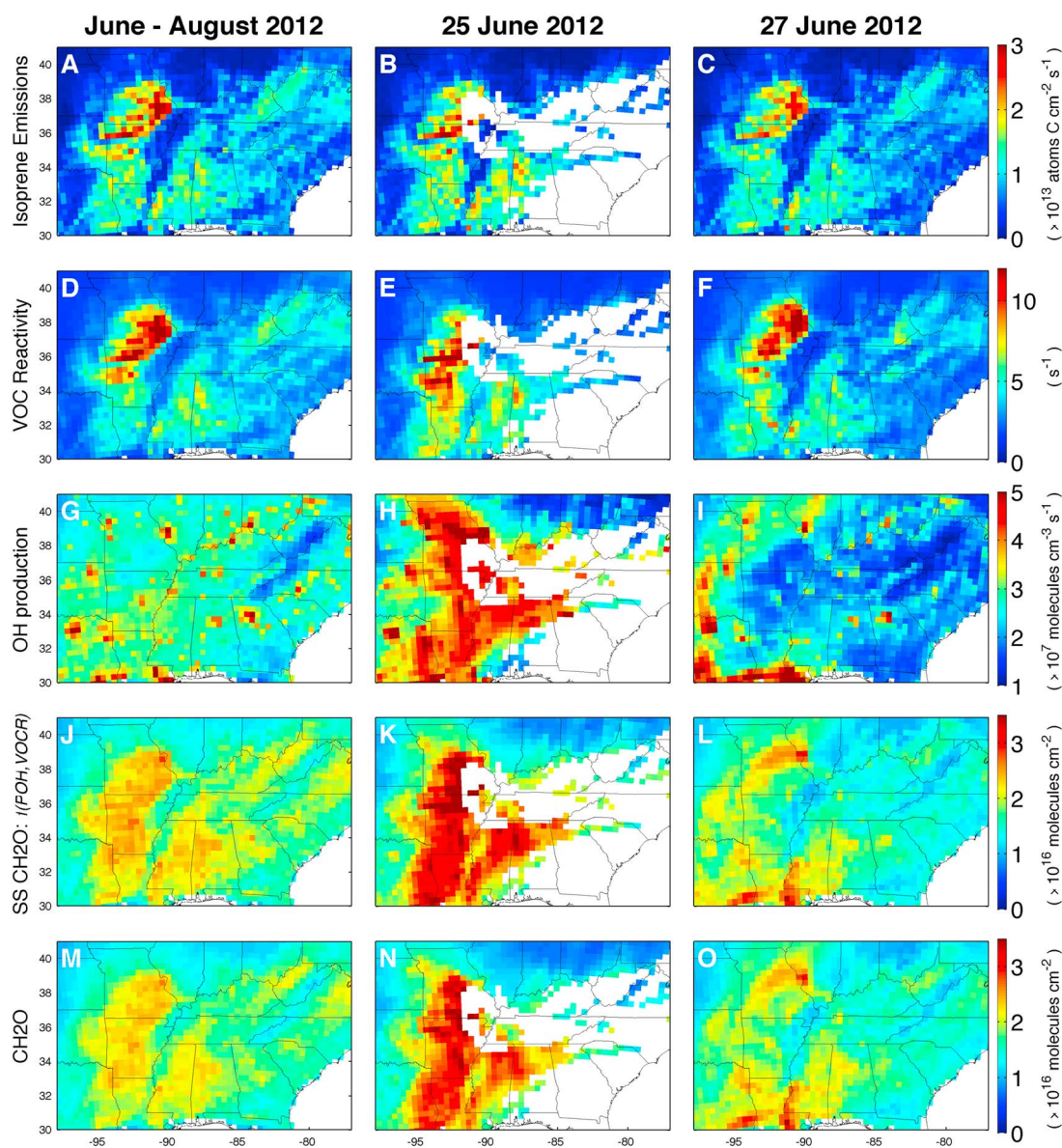
## 3.2. Analysis of $\text{CH}_2\text{O}$ Column Variability and its Drivers

### 3.2.1. Diurnal Variability of the $\text{CH}_2\text{O}$ Column and the Isoprene Columns

At a regional scale ( $32^\circ\text{N}$ – $38^\circ\text{N}$ ;  $75^\circ\text{W}$ – $90^\circ\text{W}$ ), the  $\text{CH}_2\text{O}$  column is near steady state throughout the day (Figure 2b), supporting our approach to understand  $\text{CH}_2\text{O}$  variability (section 2). The column integrated sources and sinks of  $\text{CH}_2\text{O}$  are nearly identical except for a small offset in their rise through early morning and their fall in late afternoon, and a small difference through the nighttime due to production via reactions of  $\text{O}_3$ ,  $\text{NO}_3$ , and OH with VOC (Figure 2b). At a typical wind speed of  $5 \text{ m s}^{-1}$ , the lifetime of  $\text{CH}_2\text{O}$  at midday ( $\sim 2 \text{ h}$ ) corresponds to 36 km, or one model grid cell in our simulation. Thus, it is reasonable to assume that the  $\text{CH}_2\text{O}$  column at 1 P.M. (post meridiem) reflects the balance of local sources and sinks.

While the steady state approximation for the  $\text{CH}_2\text{O}$  column is justified (Figure 2b), column integrated isoprene, a major  $\text{CH}_2\text{O}$  precursor, is not in steady state with its emissions (Figure 2a). The diurnal pattern of isoprene emissions is largest at midday and near zero at nighttime (Figure 2a). The isoprene column increases through the daytime, as emissions exceed removal, and then decreases through the night with an e-folding timescale of 5.5 h due to chemical loss (i.e.,  $\text{O}_3$ , OH, and  $\text{NO}_3$ )

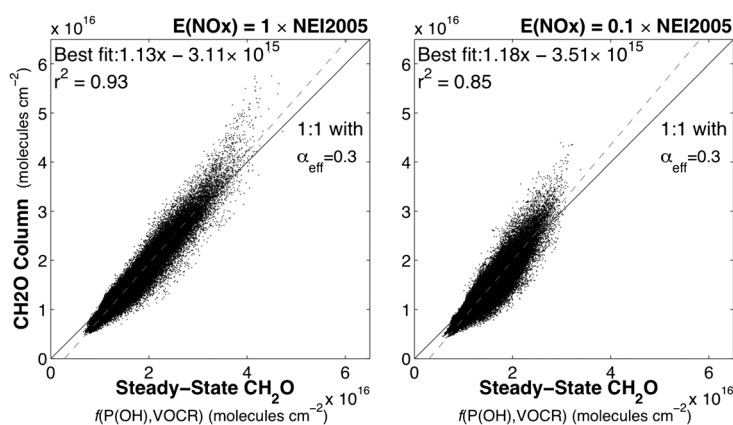




**Figure 3.** WRF-Chem 19 UTC (1 P.M. Central Standard Time; CST) (a–c)  $E_{ISOP}$  (atoms  $C\ cm^{-2}\ s^{-1}$ ), (d–f) PBL VOCR ( $z=0-2.5\ km$ ;  $s^{-1}$ ), (g–i)  $P_{OH}$  ( $z=0-2.5\ km$ ; molecules  $cm^{-3}\ s^{-1}$ ) and the (j–l) steady state  $CH_2O$  column (molecules  $cm^{-2}$ ; equation (1); see text), and the (m–o) fully simulated  $CH_2O$  column (molecules  $cm^{-2}$ ) for the June–August 2012 average (left column), 25 June (middle column) and 27 June 2012 (right column). Only values from cloud-free grid cells over land are included in the results.

respond to VOCR and by extension to  $E_{ISOP}$  for VOCR less than  $\sim 5\ s^{-1}$  but will saturate at larger values of VOCR (i.e.,  $P_{OH}/VOCR \leq 5 \times 10^6\ molecules\ cm^{-3}$ ). The corresponding threshold where the  $CH_2O$  column will saturate with respect to increases in  $E_{ISOP}$  is  $\sim 1 \times 10^{13}\ atom\ C\ cm^{-2}\ s^{-1}$  (Figures 3a and 3d). Approximately 25% of the grid cells over land are above this threshold and contribute 54% of the total  $E_{ISOP}$ .

Thus, the spatial pattern of the June–August average  $CH_2O$  column is described well by a reduced major axis linear regression to  $E_{ISOP}$  (Figures 3a versus 3m,  $r^2 = 0.62$ ) and VOCR (Figures 3d versus 3m,  $r^2 = 0.69$ ) where  $E_{ISOP}$  is less than  $1 \times 10^{13}\ atom\ C\ cm^{-2}\ s^{-1}$  ( $N = 1569$  grid cells) but saturates at  $\sim 2.5 \times 10^{16}\ molecules\ cm^{-2}$  for larger values of  $E_{ISOP}$  (Figures 3a versus 3m). As a result,  $E_{ISOP}$  and VOCR explain less spatial variance of the  $CH_2O$  column where  $E_{ISOP}$  is greater than  $1 \times 10^{13}\ atom\ C\ cm^{-2}\ s^{-1}$  ( $N = 577$  grid cells;  $r^2 = 0.40, 0.34$ , respectively). The steady state model, which accounts for these saturating effects by considering variations of  $P_{OH}$  (Figure 1), explains nearly all of the spatial variance in the  $CH_2O$  columns ( $r^2 = 0.97$ ; Figures 3j versus 3m),



**Figure 4.** The daily 1 P.M. CST June–August 2012 CH<sub>2</sub>O column (y axis; molecules cm<sup>-2</sup>) versus the steady state CH<sub>2</sub>O column (x axis; molecules cm<sup>-2</sup>; equation (1)) for all cloud-free grid cells over land in the (a) standard simulation ( $E_{\text{NO}_x} = 1 \times \text{NEI2005}$ ) and (b) in a simulation with  $E_{\text{NO}_x}$  reduced tenfold ( $E_{\text{NO}_x} = 0.1 \times \text{NEI2005}$ ). The reduced major axis regression (dashed) and one-to-one (solid) line are included.

including where  $E_{\text{ISOP}}$  is large ( $>1 \times 10^{13}$  atom C cm<sup>-2</sup> s<sup>-1</sup>;  $r^2 = 0.93$ ). The threshold at which the CH<sub>2</sub>O column saturates with respect to increases of VO<sub>CR</sub>, or to  $E_{\text{ISOP}}$  by extension, depends on the background average  $P_{\text{OH}}$  (Figure 1). Thus, biases of  $P_{\text{OH}}$  [e.g., *Lelieveld et al.*, 2008; *Butler et al.*, 2008] or failure to accurately segregate isoprene from OH sources [e.g., *Butler et al.*, 2008] will impact the simulated relationship of the CH<sub>2</sub>O column to  $E_{\text{ISOP}}$ , especially in high-VOC, low- $P_{\text{OH}}$  conditions (Figure 1, lower right corner).

### 3.3. The Sensitivity of the CH<sub>2</sub>O Column to $E_{\text{NO}_x}$

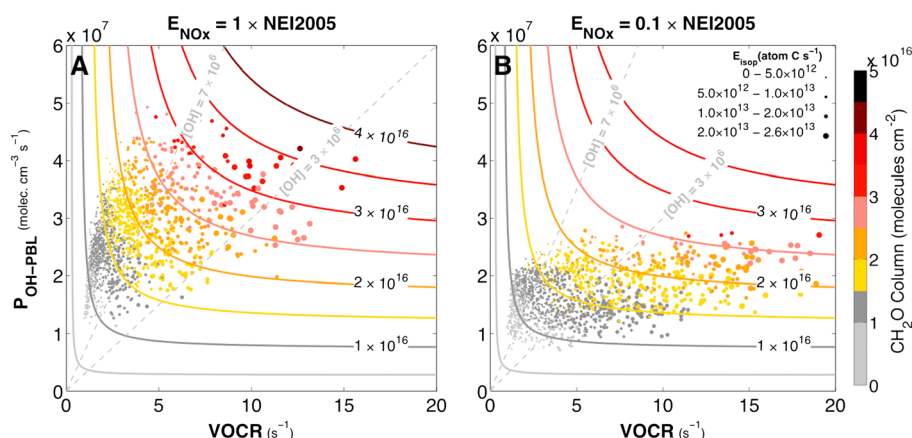
Due to the implementation of successful and ongoing NO<sub>x</sub> emission control strategies, NO<sub>2</sub> columns decreased by ~30% over U.S. cities from 2005 to 2011 [*Russell et al.*, 2012] and by 30–40% over the east U.S. from 2005 to 2013 [*Lamsal et al.*, 2015]. NO<sub>x</sub> is a major source of OH (i.e., NO-HO<sub>2</sub> reactions, O<sub>3</sub> production) and thus we expect variations of  $E_{\text{NO}_x}$  to affect the CH<sub>2</sub>O column. The concentration of NO<sub>x</sub> also affects the rates of the high-yield RO<sub>2</sub>-NO reactions relative to the lower yield RO<sub>2</sub>-HO<sub>2</sub> reactions and thus should affect  $\alpha_{\text{eff}}$  [e.g., *Palmer et al.*, 2003; *Millet et al.*, 2008; *Marais et al.*, 2012]. To disentangle the influences of NO<sub>x</sub> on  $\alpha_{\text{eff}}$  (section 3.3.1) and  $P_{\text{OH}}$  (section 3.3.2), we perform a WRF-Chem simulation identical to our base case except we decrease anthropogenic  $E_{\text{NO}_x}$  by a factor of 10. We evaluate the role of individual reaction pathways (i.e.,  $\alpha_{\text{eff}}$ ) leading to CH<sub>2</sub>O production in both simulations (Figures 4–7 and Table 1).

#### 3.3.1. Dependence of the CH<sub>2</sub>O Yield From RO<sub>2</sub> Chemistry on $E_{\text{NO}_x}$

Until now we have assumed that  $\alpha_{\text{eff}}$  ( $0.3 P_{\text{CH}_2\text{O}} P_{\text{OH}}^{-1}$ ) and  $J_{\text{CH}_2\text{O}}$  ( $0.3 \text{ h}^{-1}$ ) are constant. The production of CH<sub>2</sub>O from RO<sub>2</sub> radicals (i.e.,  $\alpha_{\text{eff}}$ ), however, is expected to depend on the relative concentrations of HO<sub>2</sub>, RO<sub>2</sub> and NO and on the identity of the parent VOC (e.g., CH<sub>4</sub> or isoprene). Our initial findings suggest that in the southeast U.S. these effects are small; the quality of the linear fit of the fully simulated CH<sub>2</sub>O column to the steady state model (Figure 3) suggests that any variability in  $\alpha_{\text{eff}}$  resulting from the details of the RO<sub>2</sub> radical chemistry is small relative to the variability driven by  $P_{\text{OH}}$  or VO<sub>CR</sub>. Furthermore, for a tenfold decrease of anthropogenic  $E_{\text{NO}_x}$ , we find that there is little change in the slope of the fully simulated CH<sub>2</sub>O column to the steady state estimate of the column assuming the same  $\alpha_{\text{eff}}$  in both regressions (Figures 4a versus 4b).

The initial finding that  $\alpha_{\text{eff}}$  does not depend strongly on  $E_{\text{NO}_x}$  motivates us to investigate the parameters affecting  $\alpha_{\text{eff}}$  in more detail. Table 1 reports all known processes affecting  $\alpha_{\text{eff}}$  in WRF-Chem. Instantaneous rates and yields are calculated explicitly at 1 P.M. CST in WRF-Chem for three locations, each with different photochemical characteristics: a high-isoprene, low-NO<sub>x</sub> regime (the Ozarks; 37°N, 92°W); a high-isoprene, high-NO<sub>x</sub> regime (Atlanta; 34°N, 84°W); and a low-isoprene, low-NO<sub>x</sub> regime (central Illinois; 40°N, 89°W).

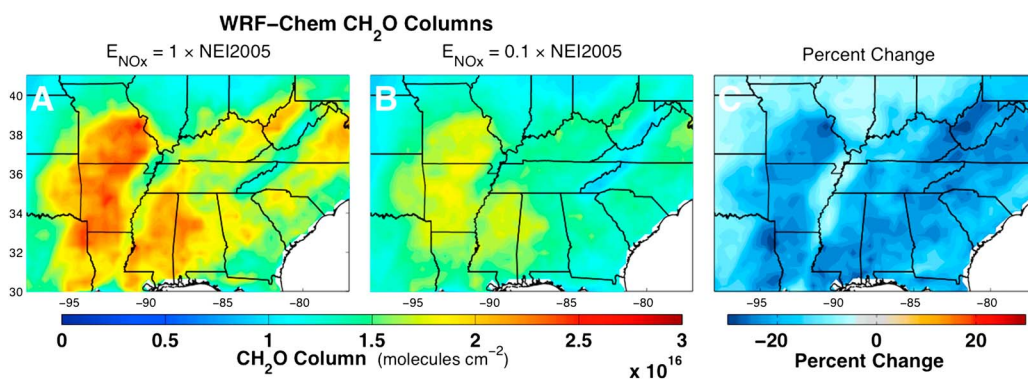
The results presented in Table 1 confirm previous findings: RO<sub>2</sub>-NO reactions are high yield and RO<sub>2</sub>-HO<sub>2</sub> reactions are low yield; these reactions together determine the fate of ~75–95% of all RO<sub>2</sub> [e.g., *Archibald et al.*, 2010, and references therein; *Marais et al.*, 2012; *Mao et al.*, 2013]. Thus, we expect that  $\alpha_{\text{eff}}$  should be larger



**Figure 5.** Daily June–August 2012 1 P.M. CST CH<sub>2</sub>O column (colors; molecules cm<sup>-2</sup>) for boundary layer (≤2.5 km) average P<sub>OH</sub> (y axis; molecules cm<sup>-3</sup> s<sup>-1</sup>) and VOCR (x axis; s<sup>-1</sup>) in WRF-Chem grid cells over land (36 km × 36 km) that have been averaged to a spatial scale of 216 km × 216 km for (a) the standard simulation and for (b) the sensitivity simulation in which anthropogenic E<sub>NOx</sub> is decreased by a factor of 10. The magnitude of E<sub>ISOP</sub> in each grid cell is indicated by the size of the marker. E<sub>ISOP</sub> is identical in the two simulations, but on average, there is 51% more VOCR per unit of E<sub>ISOP</sub> in the lower E<sub>NOx</sub> scenario due to the decrease of P<sub>OH</sub>. We include only values where fewer than 20% of the native 36 km × 36 km grid cells in the spatial average had clouds. Contours from Figure 1 are included for reference.

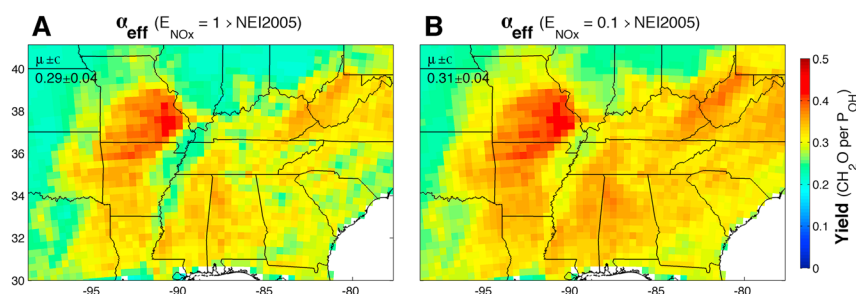
in the high-E<sub>NOx</sub> simulation. However, ~5–25% of all RO<sub>2</sub> react with RO<sub>2</sub>, a high-yield reaction or, in the case of isoprene peroxy radical, undergo isomerization to form hydroperoxyenals [HPALDs; Peeters and Muller, 2010; Crouse et al., 2011; Wolfe et al., 2012] before yielding CH<sub>2</sub>O. These high-yield RO<sub>2</sub> channels increase in importance as NO concentrations decrease (Table 1) and thus buffer α<sub>eff</sub> to the tenfold decrease of anthropogenic E<sub>NOx</sub> in background regions (e.g., Ozarks: 0.38 versus 0.39 P<sub>CH2O</sub> P<sub>OH</sub><sup>-1</sup>; central Illinois: 0.21 versus 0.24 P<sub>CH2O</sub> P<sub>OH</sub><sup>-1</sup>). In cities (e.g., Atlanta: 0.23 versus 0.33 P<sub>CH2O</sub> P<sub>OH</sub><sup>-1</sup>), α<sub>eff</sub> actually increases ~50% for the tenfold reduction of E<sub>NOx</sub>, in part due to the buffering effects of high-yield RO<sub>2</sub> channels discussed above but also due to a large increase in the yield per RO<sub>2</sub>-NO reaction at lower E<sub>NOx</sub> (0.42 versus 0.52 P<sub>CH2O</sub> P<sub>OH</sub><sup>-1</sup> for high- versus low-E<sub>NOx</sub> scenarios, respectively; Table 1). The latter effect is likely due to differences of CH<sub>2</sub>O yields between the earlier generations and later generations of VOC oxidation caused by the nearly threefold decrease of P<sub>OH</sub> (Table 1). Thus, we conclude that CH<sub>2</sub>O yield from RO<sub>2</sub> chemistry is buffered to the effects of NO<sub>x</sub> by high-yield RO<sub>2</sub> channels (Table 1).

OVOC photolysis is a large source of CH<sub>2</sub>O in the Ozarks (up to 25%; Table 1). However, more than 90% of the photolytic source of CH<sub>2</sub>O over the Ozarks is due to the rapid photolysis of HPALDs (τ<sub>HPALD-photolysis</sub> ~ 30 min). Due to the short lifetime of HPALDs, we attribute its production of CH<sub>2</sub>O to its source, the reaction of OH with



**Figure 6.** The June–August 2012 average CH<sub>2</sub>O column simulated with WRF-Chem at 19 UTC (1 P.M. Central Standard Time; CST), including only cloud-free grid cells over land in (a) the standard simulation, (b) the sensitivity simulation in which anthropogenic E<sub>NOx</sub> is decreased by a factor of 10, and (c) the percent difference between the two simulations.





**Figure 7.** The June–August 2012 1 P.M. CST average of  $\alpha_{\text{eff}}$  (colors;  $P_{\text{CH}_2\text{O}} P_{\text{OH}}^{-1}$ ) in the lower troposphere ( $\leq 2.5$  km) calculated explicitly with WRF-Chem for cloud free grid cells over land in both the (a) standard simulation ( $E_{\text{NO}_x} = 1 \times \text{NEI2005}$ ) and (b) in a simulation with  $E_{\text{NO}_x}$  reduced tenfold ( $E_{\text{NO}_x} = 0.1 \times \text{NEI2005}$ ).

isoprene, thus maintaining our assumption that OVOC photolysis and ozonolysis are relatively small sources of  $\text{CH}_2\text{O}$  in comparison to OH-VOC reactions (Table 1).

**3.3.2. Dependence of  $P_{\text{OH}}$  and VOCCR on  $E_{\text{NO}_x}$**

Assuming that  $\alpha_{\text{eff}}$  does not depend on  $E_{\text{NO}_x}$  (see section 3.3.1), we quantify the response of VOCCR,  $P_{\text{OH}}$  and the  $\text{CH}_2\text{O}$  column to the decrease of anthropogenic  $E_{\text{NO}_x}$  (Figures 5 and 6). In response to the tenfold reduction of  $E_{\text{NO}_x}$ , boundary layer  $\text{NO}_x$  concentrations decrease 61% and  $P_{\text{OH}}$  decreases 37%, from  $2.7 \times 10^7$  molecules  $\text{cm}^{-3} \text{s}^{-1}$  (Figure 5a) to  $1.7 \times 10^7$  molecules  $\text{cm}^{-3} \text{s}^{-1}$  (Figure 4b). As a result of the decreased  $P_{\text{OH}}$ , VOC lifetimes increase and the boundary layer average VOCCR increases from  $3.1 \text{ s}^{-1}$  (Figure 5a) to  $4.6 \text{ s}^{-1}$  (Figure 5b), much of it due to an increase of isoprene, which increases from a column concentration of  $1.2 \times 10^{16}$  atoms  $\text{C cm}^{-2}$  to  $3.0 \times 10^{16}$  atoms  $\text{C cm}^{-2}$ .

The decrease of  $P_{\text{OH}}$  has both a direct and indirect effect on the  $\text{CH}_2\text{O}$  removal rate due to its feedback on VOCCR ( $J_{\text{CH}_2\text{O}} + k_{\text{CH}_2\text{O}+\text{OH}} P_{\text{OH}}/\text{VOCCR}$ ; equation (1)). In regions where  $E_{\text{ISOP}}$  is high ( $> 1 \times 10^{13}$  molecules  $\text{cm}^{-2} \text{s}^{-1}$ ), we

**Table 1.** Chemistry Affecting  $\alpha_{\text{eff}}$  Calculated Instantaneously at 1 P.M. CST for June–August 2012 at High (1x) and Low (0.1x)  $E_{\text{NO}_x}$

		High VOCCR, Low $\text{NO}_x$		High VOCCR, High $\text{NO}_x$		Low VOCCR, Low $\text{NO}_x$	
		Ozarks, MO		Atlanta, GA		Central, IL	
		37°N, 92°W		34°N, 84°W		40°N, 89°W	
	$E_{\text{NO}_x\text{-NEI2005}}$	1x	0.1x	1x	0.1x	1x	0.1x
	$[\text{NO}]_{\text{PBL}}$ (ppt)	33	16	350	54	100	47
Fractional Fate of $\text{RO}_2$	$\text{RO}_2\text{-NO}$	0.35	0.22	0.85	0.50	0.60	0.44
	$\text{RO}_2\text{-HO}_2$	0.44	0.52	0.11	0.35	0.24	0.35
	$\text{RO}_2\text{-RO}_2$	0.14	0.15	0.04	0.12	0.15	0.20
Yield ( $\text{CH}_2\text{O}$ per $\text{RO}_2$ reacted)	Isomerization	0.07	0.11	0.00	0.03	0.00	0.01
	$\text{RO}_2\text{-NO}$	0.61	0.63	0.42	0.52	0.53	0.54
	$\text{RO}_2\text{-HO}_2$	0.08	0.09	0.01	0.05	0.01	0.02
$P_{\text{CH}_2\text{O}}$ by source ( $\times 10^6$ molecules $\text{cm}^{-3} \text{s}^{-1}$ )	$\text{RO}_2\text{-NO}$	4.4	2.0	11.4	4.1	4.1	2.2
	$\text{RO}_2\text{-HO}_2$	0.8	0.7	0.0	0.3	0.0	0.1
	$\text{RO}_2\text{-RO}_2$	1.6	1.2	0.7	1.0	1.0	1.0
	$J_{\text{OVOC}}$	1.6	1.8	0.3	0.7	0.2	0.3
	OH-VOC	0.2	0.2	0.6	0.2	0.5	0.3
	$\text{O}_3 + \text{alkene}$	0.9	0.8	0.2	0.3	0.0	0.0
	Total	9.6	6.7	13.4	6.6	5.8	4.0
	$P_{\text{OH}}$ ( $\times 10^6$ molecules $\text{cm}^{-3} \text{s}^{-1}$ )	25.2	17.0	57.5	19.8	27.3	16.6
	$\alpha_{\text{eff}}$ ( $P_{\text{CH}_2\text{O}} P_{\text{OH}}^{-1}$ )	0.38	0.39	0.23	0.33	0.21	0.24
	Fraction of $P_{\text{OH}}$ that reacts with VOC or CO (%)	92.8	93.4	83.7	92.3	80.2	84.8



find that  $P_{OH}$  decreases 40% and VOCR increases 60% for a combined 62% decrease of the OH concentration (i.e.,  $\Delta[OH] \sim \Delta P_{OH}/\Delta VOCR$ ;  $\sim 0.6/1.6$ ). However, the net effect on the  $CH_2O$  lifetime is much smaller (21% increase) because photolysis does not change (2% difference). Thus, the  $\sim 40\%$  reduction in  $CH_2O$  production ( $\alpha_{eff} P_{OH}$ ) is offset by a  $\sim 20\%$  increase in the  $CH_2O$  lifetime, half due to the direct impact of declining  $E_{NOx}$  on  $P_{OH}$  and half due to the indirect feedback on VOCR, resulting in a net 20% decrease of the  $CH_2O$  column (Figure 6c). In regions where  $E_{ISOP}$  is low, the feedbacks of  $P_{OH}$  on the  $CH_2O$  lifetime more fully offset its effects on  $CH_2O$  production, such that the net effect on the  $CH_2O$  column is closer to zero (compare Figures 3a and 6c).

Thus, in response to lower  $E_{NOx}$ , the simulated  $CH_2O$  column decreases everywhere (Figure 6). The decreases are largest ( $-23\%$ , Figure 6c) where  $E_{ISOP}$  is large ( $> 10 \times 10^{12}$  atoms  $C\ cm^{-2}\ s^{-1}$ ) and near zero ( $-5\%$ ) where  $E_{ISOP}$  is small ( $< 1 \times 10^{12}$  atoms  $C\ cm^{-2}\ s^{-1}$ ), a pattern consistent with a decrease in regional background  $P_{OH}$  (Figure 1) and the feedbacks of  $P_{OH}$  on  $CH_2O$  production and loss described above. Due to larger decreases of the  $CH_2O$  column where  $E_{ISOP}$  is large, the spatial correlation simulated between the  $CH_2O$  column and  $E_{ISOP}$  weakens ( $r^2 = 0.59$ ; Figures 3a, and 6b all grid cells;  $r^2 = 0.76$ , Figures 3a and 6a all grid cells).

We perform the same WRF-Chem simulation with low  $E_{NOx}$  ( $0.1 \times NEI2005$ ; Figure 6b) but also decrease  $E_{ISOP}$  by 20% over the Ozarks ( $89.4\text{--}94.9^\circ W$ ;  $35.3\text{--}39.1^\circ N$ ). As expected for low- $P_{OH}$ , high-VOCR conditions (Figure 1), the  $CH_2O$  column over the same region is not very sensitive to changes of  $E_{ISOP}$  and decreases by only 5.2%. Thus, in this high-VOC regime, a 40% perturbation or bias of  $P_{OH}$  can have a twofold larger effect on the  $CH_2O$  column ( $\sim 20\%$ ; Figure 6) than does a 20% perturbation of  $E_{ISOP}$  ( $\sim 5\%$ ), assuming a linear scaling to compare the perturbations.

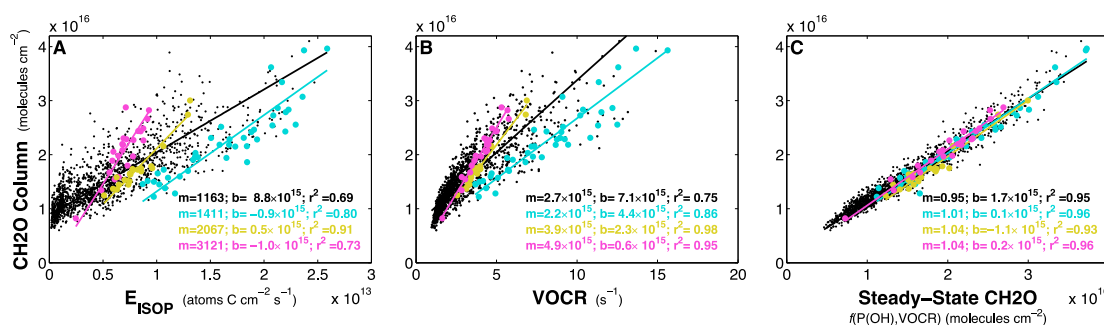
### 3.4. $\alpha_{eff}$ : Dependence on Ambient VOC Mixture

Figure 7 shows the spatial variability of  $\alpha_{eff}$  in both the standard and reduced  $E_{NOx}$  simulations. The value for  $\alpha_{eff}$  ( $P_{CH_2O} P_{OH}^{-1}$ ) is calculated explicitly from 1 P.M. CST June–August 2012 average values for  $P_{CH_2O}$  and  $P_{OH}$ . The spatial pattern and magnitude of  $\alpha_{eff}$  does not depend strongly on the tenfold decrease of  $E_{NOx}$  (Figure 7a versus 7b), again demonstrating that  $\alpha_{eff}$  is relatively buffered to variations of  $E_{NOx}$  (see section 3.3.1). The spatial pattern of  $\alpha_{eff}$  in both scenarios reflects differences in the regionally relevant VOC mixture (i.e., isoprene versus background  $CH_4$ ). In general,  $\alpha_{eff}$  is approximately  $0.25 P_{CH_2O} P_{OH}^{-1}$  in the regional background and is between 0.3 and 0.4  $P_{CH_2O} P_{OH}^{-1}$  in the high-VOCR regions. Because  $\alpha_{eff}$  tends to be large where  $CH_2O$  columns are largest and small where  $CH_2O$  columns are smallest, the steady state calculation of  $CH_2O$  is systematically biased (slope  $> 1$ ; Figure 4) because it assumes that  $\alpha_{eff}$  is spatially constant ( $\alpha_{eff} = 0.30 P_{CH_2O} P_{OH}^{-1}$ ). In background regions,  $CH_4$  ( $\alpha_{CH_4} \leq 1$ ) and  $CO$  ( $\alpha_{CO} = 0$ ) are the most important sinks of OH. Thus,  $\alpha_{eff, background}$  depends on their relative concentration and may change as their abundances change in the future. On the other hand, in isoprene-rich regions, almost every OH radical reacts with isoprene or one of its products, so  $\alpha_{eff}$  in these regions will be buffered to any changes in background CO or  $CH_4$  concentrations. Thus, in isoprene-rich regions, long-term trends of the  $CH_2O$  column should primarily reflect changing  $P_{OH}$  and VOCR assuming that  $\alpha_{eff}$  is buffered (see also section 3.3.1).

## 4. Implications for VOC Emission Inversions From OVOC Measurements

To investigate the implications of our findings for retrieving  $E_{ISOP}$  from measurements of the  $CH_2O$  column, we evaluate the relationship of the  $CH_2O$  column to  $E_{ISOP}$  in the WRF-Chem simulation (Figure 8a) and discuss our findings in the context of VOCR (Figure 8b) and the steady state estimate of the  $CH_2O$  column (Figure 8c). We average model results ( $36\ km \times 36\ km$ ) to a coarser spatial scale ( $216\ km \times 216\ km$ ) similar to the model resolution used in previous analyses [e.g., Palmer *et al.*, 2003, 2006; Millet *et al.*, 2008; Barkley *et al.*, 2008; Marais *et al.*, 2012]. We have highlighted values from individual grid cells over the Missouri Ozarks ( $37^\circ N$ ,  $92.5^\circ W$ ; cyan), Alabama ( $33.5^\circ N$ ,  $85.5^\circ W$ ; yellow), and SW Arkansas ( $33.5^\circ N$ ,  $92.5^\circ W$ ; magenta), to illustrate the regional differences among the domain-wide relationships (Figure 8).

As shown previously [e.g., Palmer *et al.*, 2003, 2006; Millet *et al.*, 2008], there is a strong relationship between the  $CH_2O$  column and  $E_{ISOP}$  when considering all daily values from individual grid cells as a single statistical ensemble ( $r^2 = 0.69$ ; Figure 8a). As expected for lower  $P_{OH}$  conditions, this relationship is weaker in the lower



**Figure 8.** Daily 1 P.M. CH<sub>2</sub>O column versus (a) isoprene emissions, (b) VOCR, and (c) the CH<sub>2</sub>O column derived by assuming steady state (equation (1);  $\alpha_{\text{eff}} = 0.3 P_{\text{CH}_2\text{O}} P_{\text{OH}}^{-1}$ ) for June–August 2012 over the southeast U.S. (29°–39°N, 78°–100°W) in the WRF–Chem model. Values are averaged from the native WRF–Chem grid to a coarser grid of 216 km × 216 km. Columns over the Missouri Ozarks (37°N, 92.5°W; cyan), Alabama (33.5°N, 85.5°W; yellow) and SW Arkansas (33.5°N, 92.5°W; magenta) are highlighted to illustrate regional differences in the relationships. Values with cloud fraction above 0.2 are excluded. Reduced major axis least square regression results are reported in the lower right corner.

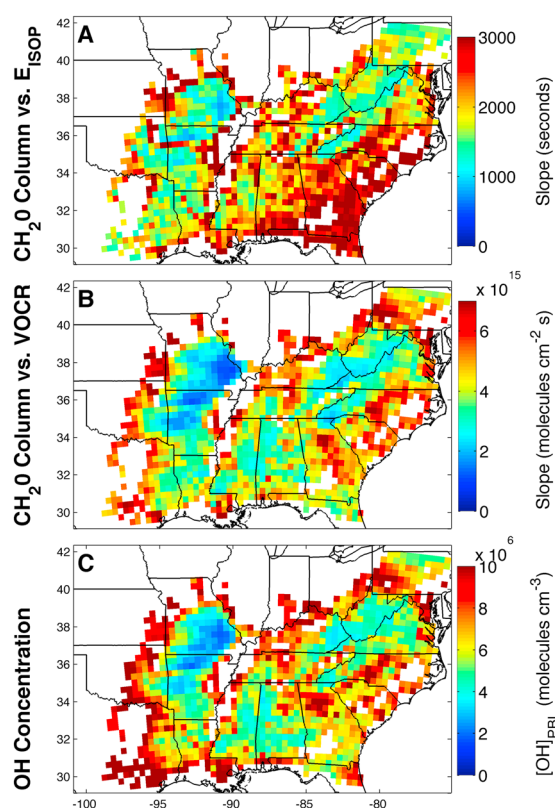
$E_{\text{NO}_x}$  simulation ( $r^2 = 0.49$ ). The relationship is varied locally, with inferred slopes varying by more than a factor of 2 between some locations (Figure 8a, colors), confounding the retrieval of  $E_{\text{ISOP}}$  for a given value of the CH<sub>2</sub>O column (e.g., magenta–Ozarks versus cyan–SW Arkansas), a finding also noted in previous studies [e.g., Millet *et al.*, 2008; Marais *et al.*, 2012].

One hypothesis for the regional differences in slopes of CH<sub>2</sub>O to  $E_{\text{ISOP}}$  is that the transport of isoprene and its products from high-emission grid cells to downwind grid cells diminishes the sensitivity of CH<sub>2</sub>O to  $E_{\text{ISOP}}$  in the high-emission grid cells and inflates it downwind [Millet *et al.*, 2008]. However, VOCR accounts for transport, yet its relationship with the CH<sub>2</sub>O column is equally varied across the regions (Figures 8a versus 8b; e.g., magenta–Ozarks versus cyan–SW Arkansas), thus excluding transport as the primary cause. The steady state model, on the other hand, is able to explain the regional differences in slopes to within 5% of one another (Figure 8c, colors), suggesting that the regional differences in the relationship of CH<sub>2</sub>O to  $E_{\text{ISOP}}$  or VOCR are due to differences of  $P_{\text{OH}}$  (Figures 8a and 8b versus Figure 8c).

Despite regional variability, the local relationships of CH<sub>2</sub>O to  $E_{\text{ISOP}}$  or VOCR are remarkably strong, with correlation coefficients ranging from 0.73 to 0.91 for  $E_{\text{ISOP}}$  (Figure 8a, colors) and from 0.86 to 0.98 for VOCR (Figure 8b, colors). The strong linear relationship between VOCR and CH<sub>2</sub>O requires that the other driving variable,  $P_{\text{OH}}$ , be positively correlated with VOCR such that the day-to-day variation of the OH concentration is small (i.e., daily values of VOCR and  $P_{\text{OH}}$  lie on a line of constant slope in Figure 1). Positive correlation of VOCR and  $P_{\text{OH}}$  suggests that the conditions resulting in higher  $E_{\text{ISOP}}$  and VOCR tend to result in higher  $P_{\text{OH}}$ : high temperatures and clear skies are associated with larger  $E_{\text{ISOP}}$  [Guenther *et al.*, 2006], and high concentrations of water vapor and ozone, both of which are important precursors of OH radicals. Assuming that the day-to-day variation of OH is small in any given location, the sensitivity (i.e., slope) of the CH<sub>2</sub>O column to a perturbation of VOCR in that location is directly proportional to its OH concentration: the slope is steep where the OH concentration is high (e.g., the rise of CH<sub>2</sub>O contours following the gray line with  $[\text{OH}] = 7 \times 10^6$  molecules cm<sup>-3</sup> in Figure 1) and is shallow where the OH concentration is low (e.g., the same for the gray line with  $[\text{OH}] = 3 \times 10^6$  molecules cm<sup>-3</sup> in Figure 1). This same effect will be weaker for  $E_{\text{ISOP}}$  due to the feedbacks of  $P_{\text{OH}}$  on VOCR, but should follow the same general trend (section 3.3.2).

To evaluate this hypothesis, we compare the simulated cloud-free June–August 1 P.M. CST average OH concentration (Figure 9c) to the CH<sub>2</sub>O column sensitivity to VOCR (Figure 9b) and  $E_{\text{ISOP}}$  (Figure 9a) at the native model resolution (36 × 36 km<sup>2</sup>). As expected based on Figure 1, the concentration of OH explains much of the spatial variance of the CH<sub>2</sub>O column sensitivity to VOCR ( $r^2 = 0.64$ ; Figures 9b versus 9c). The CH<sub>2</sub>O lifetime at 1 P.M. is too short (~2 h) for its transport to strongly influence its relationship with VOCR downwind. Furthermore, any transport will similarly affect VOCR, thus not disturbing their ratio.

Performing the same analysis for  $E_{\text{ISOP}}$  instead of VOCR, we find that the quantitative agreement is not as strong ( $r^2 = 0.26$ ; Figures 9a versus 9c), but qualitatively, the results are similar. Previous analyses found anomalously high sensitivities in grid cells where summer average  $E_{\text{ISOP}}$  was low, attributing the enhancement to transport from



**Figure 9.** The scaling factor relating  $\text{CH}_2\text{O}$  columns to (a)  $E_{\text{ISOP}}$  (seconds) and (b)  $\text{VOCCR}$  ( $\text{molecule cm}^{-2} \text{s}$ ) derived as the slope from reduced major axis linear regression within each individual WRF-Chem grid cell over land where June–August 2012 average  $E_{\text{ISOP}} > 5 \times 10^{12} \text{ atom C cm}^{-2} \text{ s}^{-1}$ , screened for cloud-free conditions. Also shown is (c) the June–August 2012 boundary layer average OH concentration ( $\text{molecules cm}^{-3}$ ) in grid cells meeting these criteria.

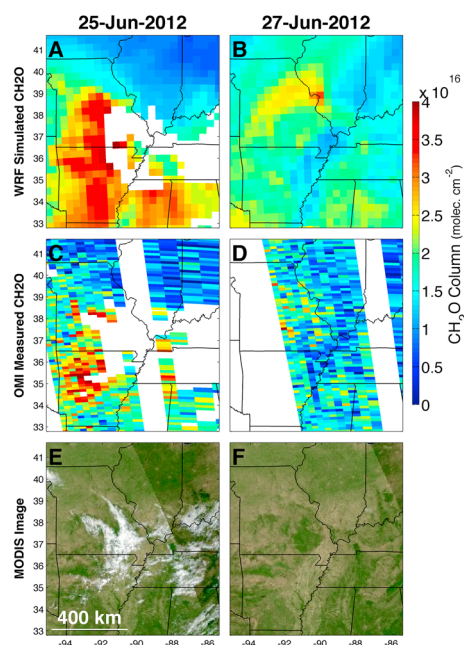
cepts in the same example is also large (Figure 8a;  $\mu \pm \sigma$ ;  $0.1 \times 10^{15} \pm 3.0 \times 10^{15} \text{ molecules cm}^{-2}$ ) and their average is much smaller than the  $y$  intercept derived for the entire ensemble of grid cells (Figure 8a;  $8.8 \times 10^{15} \text{ molecules cm}^{-2}$ ). Thus, the  $y$  intercept does not only represent the background  $\text{CH}_2\text{O}$  column in the absence of isoprene emission [e.g., Palmer *et al.*, 2006] but also reflects nonlinearities in the OH and VOC chemistry that drive  $\text{CH}_2\text{O}$  variability (Figures 1 and 5). As such, the  $y$  intercept should be treated as a free variable along with the scaling parameter (slope).

Based on our findings, we recommend a full formal inversion to derive  $E_{\text{ISOP}}$  from measurements of the  $\text{CH}_2\text{O}$  column that accounts for nonlinearities of the OH and VOC chemistry. However, as long as the day-to-day variability of OH concentrations under cloud free conditions is small (Figure 1; lines of constant slope), our results provide a basis (Figure 9) for performing inversions using locally derived linear formulations, originally introduced by Millet *et al.* [2008] and improved by Marais *et al.* [2012]. These inversions should account for spatial variations of the OH concentration and recognize the limitations of inferring variations of  $E_{\text{ISOP}}$  in low- $P_{\text{OH}}$  environments (e.g., Figure 6b). Marais *et al.* [2012] incorporated information on the variability of the  $\text{NO}_2$  column, an adaptation that implicitly accounts for some, but not all of the variability of OH concentrations, and is also uncertain in regions where  $\text{NO}_2$  columns are near their detection limit. We recommend that a first test of any linear inversion is the comparison of a second iteration of the updated forward model with the  $\text{CH}_2\text{O}$  column measurements.

The influence of OH on the variability of  $\text{CH}_2\text{O}$  is due to the short photolytic lifetime of  $\text{CH}_2\text{O}$  relative to that of reaction with OH (see equation (1)). Thus, the findings that we present here, that variations of the OH radical

larger upwind sources ( $< 2 \times 10^{12} \text{ atom C cm}^{-2} \text{ s}^{-1}$ ) [Millet *et al.*, 2008]. We restrict our analysis to grid cells where average  $E_{\text{ISOP}}$  is greater than  $5 \times 10^{12} \text{ atom C cm}^{-2} \text{ s}^{-1}$ . Transport of isoprene and its oxidation products does affect the sensitivity (slope) of  $\text{CH}_2\text{O}$  to  $E_{\text{ISOP}}$  at the edge of large isoprene sources (e.g., Figure 9a, downwind edge of the Ozarks) [e.g., Millet *et al.*, 2008], but the spatial pattern of OH concentrations also plays a role (Figure 9c).  $\text{CH}_2\text{O}$  columns are most sensitive to  $E_{\text{ISOP}}$  over cities where OH concentrations are highest and are less sensitive in regions where  $E_{\text{ISOP}}$  is largest and OH concentrations are lowest (Figures 3a and 9). We would not expect the relationship between OH concentrations and the  $\text{CH}_2\text{O}$ -to- $E_{\text{ISOP}}$  slopes (Figures 9a versus 9c) to be as strong as that for  $\text{VOCCR}$  (Figures 9b versus 9c) because of VOC transport [e.g., Millet *et al.*, 2008], particularly at this finer spatial scale and in light of feedbacks of OH concentrations on  $\text{VOCCR}$  [e.g., Karl *et al.*, 2009, section 3.3.2].

The slopes and  $y$  intercepts inferred for regressions of the  $\text{CH}_2\text{O}$  column to variations of  $E_{\text{ISOP}}$  depend on the temporal and spatial scale over which they are derived and thus should not be applied to other temporal and spatial scales (Figures 3, 6, 8, and 9). For example, the slope retrieved for the entire ensemble of grid cells is smaller than the slope derived in every individual grid cell (Figure 8a), a gross bias due to combining high  $E_{\text{ISOP}}$ , low-OH locations (i.e., the Ozarks) with lower  $E_{\text{ISOP}}$ , higher-OH locations. The range of the locally derived  $y$  inter-



**Figure 10.** WRF-Chem (a and b) and OMI measurements (c and d) of the  $\text{CH}_2\text{O}$  column at 1 P.M. CST on a day when (Figures 10a, 10c, and 10e)  $P_{\text{OH}}$  is high (25 June 2012) and on a day when (Figures 10b, 10d, and 10f)  $P_{\text{OH}}$  is low (27 June 2012).  $E_{\text{ISOP}}$  varied little between these 2 days over the central U.S. ( $\sim 3\%$  decrease simulated using WRF-Chem/MEGANv2.04). The OMI view geometry is similar on both days ( $\sim 200$  km further east on 27 June 2012). MODIS true color composite images (e and f) recorded  $\sim 15$  min prior to the OMI measurements indicate areas of vegetation and the cloud distribution.

(Figures 3k, 3l, 3n, and 3o). This result again highlights the need to use the appropriate time and space domains for any inversion of  $E_{\text{ISOP}}$  from measurements of the  $\text{CH}_2\text{O}$  column, as the sensitivity of the  $\text{CH}_2\text{O}$  column to  $E_{\text{ISOP}}$  depends on daily variations of  $P_{\text{OH}}$ .

To demonstrate that these daily variations of the  $\text{CH}_2\text{O}$  column are observable, Figure 10 shows the  $\text{CH}_2\text{O}$  columns measured by the Ozone Monitoring Instrument (OMI) from the most recent NASA/SAO product (version 3) [González Abad *et al.*, 2015] and again shows those simulated by WRF-Chem (see also Figure 3), zoomed in over the south-central U.S. on 25 and 27 June 2012. The spatial pattern of the simulated  $\text{CH}_2\text{O}$  column matches the observations at a regional scale (Figure 10). On 25 June 2012, the  $\text{CH}_2\text{O}$  columns are minimum in the northeast corner of the domain but larger throughout the rest of the domain, even exceeding  $4 \times 10^{16}$  molecules  $\text{cm}^{-2}$  in north central Arkansas, an indication that boundary layer average  $P_{\text{OH}}$  ( $>4 \times 10^7$  molecules  $\text{cm}^{-3} \text{s}^{-1}$ ) and VOCR ( $>7.5 \text{s}^{-1}$ ) are both high (i.e., toward the upper right corner of Figure 1). On 27 June 2012, the columns are much lower throughout the domain with a local minimum in the center of the image and in the far northeast corner, indicating that either  $P_{\text{OH}}$  or VOCR is lower than it was on 25 June 2012. Canopy-scale isoprene flux measurements in the region increased from 25 June to 27 June 2012 [Seco *et al.*, 2015] ( $38.74^\circ\text{N}$ ,  $92.20^\circ\text{W}$ ) suggesting that VOCR is unlikely to be lower on 27 June. Consequently, the observed decrease in the  $\text{CH}_2\text{O}$  columns is likely due to the decrease of  $P_{\text{OH}}$  in the region (Figures 3h and 3i), primarily driven by the decrease of water vapor.

The simulated  $\text{CH}_2\text{O}$  columns are biased high relative to the measurements on both days (Figure 10), suggesting that the simulated  $P_{\text{OH}}$  or VOCR is biased high, though it is also possible that the measurements are systematically biased. If the air mass factor (AMF) from version 2 of the retrieval were applied to the

influence the variability of  $\text{CH}_2\text{O}$ , also apply to other oxidized VOC intermediates with short lifetimes against photolysis (e.g., glyoxal;  $\tau_{\text{photolysis}} \sim 2$  h;  $\tau_{\text{OH+glyoxal}} \sim 2.8$  h at  $[\text{OH}] = 1 \times 10^7$  molecules  $\text{cm}^{-2}$ ).

## 5. Daily Variations of the $\text{CH}_2\text{O}$ Column

The daily variability of  $\text{CH}_2\text{O}$  column has been used to infer daily variability of  $E_{\text{ISOP}}$  [e.g., Palmer *et al.*, 2003]. However, the factors affecting  $P_{\text{OH}}$  and thus its influence on the  $\text{CH}_2\text{O}$  column also vary daily. From 25 June 2012 to 27 June 2012 (Figures 3b, 3c, 3e, 3f, 3h, 3i, 3k, 3l, 3n, and 3o), boundary layer average  $P_{\text{OH}}$  decreases from  $4.2 \times 10^7$  molecules  $\text{cm}^{-3} \text{s}^{-1}$  to  $2.4 \times 10^7$  molecules  $\text{cm}^{-3} \text{s}^{-1}$  over Arkansas and Missouri, two U.S. states in the center west of the domain. The decrease of  $P_{\text{OH}}$  is due primarily to a decrease of water vapor in the region (i.e., less  $\text{O}^1\text{D} + \text{H}_2\text{O} \rightarrow 2 \text{OH}$ ). The MODIS near infrared retrieval of the water vapor column indicates that WRF is accurately simulating the regional-scale patterns of the observed water vapor distribution (not shown). On these same days, there is little change in  $E_{\text{ISOP}}$  ( $\sim 3\%$ ,  $1.13 \times 10^{13}$  to  $1.10 \times 10^{13}$  atoms  $\text{C cm}^{-2} \text{s}^{-1}$ ), VOCR ( $\sim 5\%$ ,  $5.8 \text{s}^{-1}$  versus  $5.5 \text{s}^{-1}$ ) or ground-level temperature ( $33.5^\circ\text{C}$  versus  $32.5^\circ\text{C}$ ) over the region. Due to the low- $P_{\text{OH}}$ , high-VOCR conditions, the  $\text{CH}_2\text{O}$  column in the region is sensitive to the decrease of  $P_{\text{OH}}$  (Figure 1) and the statewide average  $\text{CH}_2\text{O}$  column over Missouri and Arkansas decreases from  $2.6 \times 10^{16}$  molecules  $\text{cm}^{-2}$  on 25 June 2012 to  $1.9 \times 10^{16}$  molecules  $\text{cm}^{-2}$  on 27 June 2012, a decrease that is well explained by the steady state model



version 3 slant columns, the vertical columns would be approximately 33% larger, and in better agreement with the simulated values. We are not suggesting that the AMF from one version is better than the other but are rather highlighting the systematic uncertainty of the measurements due to retrieval inputs [e.g., Palmer *et al.*, 2001; Russell *et al.*, 2011; Barkley *et al.*, 2012].

## 6. Variability of the CH<sub>2</sub>O Column: A Constraint on P<sub>OH</sub>?

Despite these potential biases, the observed variability suggests that in addition to reflecting variations of isoprene emissions due to temperature [e.g., Palmer *et al.*, 2003], the CH<sub>2</sub>O column reflects variation of OH sources, such as ozone, NO<sub>x</sub>, water vapor and of particular interest, the magnitude of a potentially large hypothesized source of OH in high-VOC, low-NO<sub>x</sub> environments [Lelieveld *et al.*, 2008; Hofzumahaus *et al.*, 2009; Whalley *et al.*, 2011; Mao *et al.*, 2012; Taraborrelli *et al.*, 2012; Fuchs *et al.*, 2013; Nolscher *et al.*, 2014; Rohrer *et al.*, 2014]. Measurement-based studies suggest that our conventional understanding of OH sources grossly underestimates P<sub>OH</sub> in high-isoprene, low-NO<sub>x</sub> conditions (e.g.,  $1 \times -3 \times$  conventional sources) [Lelieveld *et al.*, 2008; Fuchs *et al.*, 2013]. However, there is also evidence from quantitative, detailed laboratory studies [Paulot *et al.*, 2009a, 2009b; Crouse *et al.*, 2011, 2012] and the field [e.g., Mao *et al.*, 2012] that suggests that the magnitude of the postulated OH sources is not much larger than conventionally understood.

If the postulated source of OH were much larger than what is conventionally understood ( $1 \times -3 \times$  conventional sources) [e.g., Lelieveld *et al.*, 2008; Fuchs *et al.*, 2013], then the CH<sub>2</sub>O column should be less sensitive to variations of P<sub>OH</sub> and more sensitive to variations of VO<sub>CR</sub> (i.e., toward the top of Figure 1;  $[\text{OH}] \geq 5 \times 10^6$  molecules cm<sup>-3</sup>). However, the observations show that the CH<sub>2</sub>O column is sensitive to the variability of conventional sources of OH (i.e., water vapor). The observed variability (Figure 10) is in line with that simulated in WRF-Chem, which we estimate regenerates approximately 33% of conventional OH sources in isoprene-rich conditions [e.g., Paulot *et al.*, 2009a, 2009b; Crouse *et al.*, 2011, 2012], approximately an order of magnitude less than that postulated based on recent OH measurements (~threefold increase) [e.g., Lelieveld *et al.*, 2008]. To make this estimate, we classify OH production in to five classes and assume that all production via photolysis, RO<sub>2</sub>-HO<sub>2</sub> and VOC-OH reactions are unconventional while that due to inorganic reactions and ozonolysis is conventional and is thus an upper bound on the estimated rate. We isolate the effect of low-NO<sub>x</sub> isoprene chemistry by making the estimate for 27 June 2012 over the Ozarks (37.4°N, 91.1°W), where surface level isoprene concentrations are high (18 ppb) and NO concentrations are low (47 parts per trillion (ppt)).

Thus, the observed variability is consistent with an OH source with a magnitude that is closer to that of conventional sources (i.e., toward the bottom of Figure 1). The large uncertainty of any single column measurement [e.g., González Abad *et al.*, 2015] limits the strength with which we can make this conclusion from only two overpasses (25 and 27 June 2012; Figure 10). However, in situ measurements of CH<sub>2</sub>O also indicate that a source of OH far larger than conventionally understood is not consistent with the observed abundance of CH<sub>2</sub>O [e.g., Butler *et al.*, 2008; Galloway *et al.*, 2012].

## 7. Conclusions

We describe the variability of the fully simulated June–August 1 P.M. CST CH<sub>2</sub>O column in WRF-Chem (Figures 2–9) using a steady state framework (equation (1) and Figures 1 and 6) with simulated P<sub>OH</sub> and VO<sub>CR</sub> as the only variables. We find that the CH<sub>2</sub>O column depends on P<sub>OH</sub> most strongly where the midday OH concentration is low and depends on VO<sub>CR</sub> where the midday OH concentration is high (Figure 1). We find that the spatial pattern of E<sub>ISOP</sub> in the east U.S. drives the pattern of VO<sub>CR</sub> (Figure 3). Thus, based on the spatial pattern of the OH concentration simulated in WRF-Chem ( $\sim P_{\text{OH}}/\text{VO}_{\text{CR}}$ ; Figure 9c), we conclude that the CH<sub>2</sub>O column is most sensitive to E<sub>ISOP</sub> where E<sub>ISOP</sub> is small or where P<sub>OH</sub> is large and is least sensitive to E<sub>ISOP</sub> where E<sub>ISOP</sub> is large (Figures 9a versus 9c). The opposite is true for the relationship of P<sub>OH</sub> to the CH<sub>2</sub>O column.

Our findings indicate that variability of the OH concentration should be considered when using measurements of the CH<sub>2</sub>O column to understand the variability of E<sub>ISOP</sub> and other VOC emissions. Our finding has been alluded to in previous work [e.g., Marais *et al.*, 2012] but was not attributed to the variability of P<sub>OH</sub>. The prior attribution to NO<sub>x</sub> [Marais *et al.*, 2012] is consistent with our findings in that NO<sub>x</sub> is a large source

of OH radicals but neglects the impacts of other OH sources such as water vapor, ozone, or possible isoprene-dependent OH production pathways recently discovered [Crouse *et al.*, 2011, 2012] or hypothesized [e.g., Lelieveld *et al.*, 2008; Hofzumahaus *et al.*, 2009]. We examine the influence of NO<sub>x</sub> on RO<sub>2</sub> chemistry by tracking the CH<sub>2</sub>O yield from individual radical reaction pathways and find that the yield of CH<sub>2</sub>O at low NO<sub>x</sub> concentrations is buffered by high-yield RO<sub>2</sub>-RO<sub>2</sub> reactions (Table 1). Thus, we conclude that in isoprene-rich regions, the influence of NO<sub>x</sub> on CH<sub>2</sub>O production is primarily due to its feedback on P<sub>OH</sub>, which controls the rate of RO<sub>2</sub> formation and less so through its effect on the fate of individual RO<sub>2</sub>.

Our findings suggest that measurements of the CH<sub>2</sub>O column have the potential to improve our understanding of OH chemistry. For example, our framework (Figure 1) implies that measurements of the CH<sub>2</sub>O column in regions where VOCR is known to be high [e.g., Lelieveld *et al.*, 2008] (~10 s<sup>-1</sup>) [Hofzumahaus *et al.*, 2009] (~20 s<sup>-1</sup>) can be used to test hypothesized OH sources, insofar as the assumption of a relatively insensitive CH<sub>2</sub>O yield per P<sub>OH</sub> (α<sub>eff</sub>) is valid for any changes to the oxidation mechanism. While this assumption would not hold for all possible isoprene oxidation mechanisms or VOC environments, we show here that the yield of CH<sub>2</sub>O per P<sub>OH</sub> (α<sub>eff</sub>) in a state-of-the-science chemical mechanism is insensitive to a broad range of photochemical conditions in an isoprene-rich environment (Figure 7; Table 1). Regardless of these details, any proposed isoprene oxidation mechanism must reconcile itself with CH<sub>2</sub>O measurements (as well as those of other oxidized VOC), a task which has proven difficult in previous modeling analyses [e.g., Butler *et al.*, 2008; Galloway *et al.*, 2012].

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#### References

- Abbot, D. S., P. I. Palmer, R. V. Martin, K. V. Chance, D. J. Jacob, and A. Guenther (2003), Seasonal and interannual variability of North American isoprene emissions as determined by formaldehyde column measurements from space, *Geophys. Res. Lett.*, *30*(17), 1886, doi:10.1029/2003GL017336.
- Archibald, A. T., M. C. Cooke, S. R. Utembe, D. E. Shallcross, R. G. Derwent, and M. E. Jenkin (2010), Impacts of mechanistic changes on HO<sub>x</sub> formation and recycling in the oxidation of isoprene, *Atmos. Chem. Phys.*, *10*(17), 8097–8118.
- Barkley, M. P., P. I. Palmer, U. Kuhn, J. Kesselmeier, K. Chance, T. P. Kurosu, R. V. Martin, D. Helmig, and A. Guenther (2008), Net ecosystem fluxes of isoprene over tropical South America inferred from Global Ozone Monitoring Experiment (GOME) observations of HCHO columns, *J. Geophys. Res.*, *113*, D20304, doi:10.1029/2008JD009863.
- Barkley, M. P., T. P. Kurosu, K. Chance, I. De Smedt, M. Van Roozendaal, A. Arneeth, D. Hagberg, and A. Guenther (2012), Assessing sources of uncertainty in formaldehyde air mass factors over tropical South America: Implications for top-down isoprene emission estimates, *J. Geophys. Res.*, *117*, D13304, doi:10.1029/2011JD016827.
- Barkley, M. P., et al. (2013), Top-down isoprene emissions over tropical South America inferred from SCIAMACHY and OMI formaldehyde columns, *J. Geophys. Res. Atmos.*, *118*, 6849–6868, doi:10.1002/jgrd.50552.
- Boeke, N. L., et al. (2011), Formaldehyde columns from the Ozone Monitoring Instrument: Urban versus background levels and evaluation using aircraft data and a global model, *J. Geophys. Res.*, *116*, D05303, doi:10.1029/2010JD014870.
- Browne, E. C., P. J. Wooldridge, K. E. Min, and R. C. Cohen (2014), On the role of monoterpene chemistry in the remote continental boundary layer, *Atmos. Chem. Phys.*, *14*(3), 1225–1238.
- Butler, T. M., D. Taraborrelli, C. B. H. Fischer, H. Harder, M. Martinez, J. Williams, M. G. Lawrence, and J. Lelieveld (2008), Improved simulation of isoprene oxidation chemistry with the ECHAM5/MESy chemistry-climate model: Lessons from the GABRIEL airborne field campaign, *Atmos. Chem. Phys.*, *8*(16), 4529–4546.
- Carlton, A. G., and K. R. Baker (2011), Photochemical Modeling of the Ozark Isoprene Volcano: MEGAN, BEIS, and Their Impacts on Air Quality Predictions, *Environ. Sci. Technol.*, *45*(10), 4438–4445.
- Chance, K., X. Liu, R. M. Suleiman, D. E. Flittner, J. Al-Saadi, and S. J. Janz (2013), Tropospheric emissions: Monitoring of pollution (TEMPO), Conference on Earth Observing Systems XVIII, San Diego, Calif.
- Crouse, J. D., F. Paulot, H. G. Kjaergaard, and P. O. Wennberg (2011), Peroxy radical isomerization in the oxidation of isoprene, *Phys. Chem. Chem. Phys.*, *13*(30), 13,607–13,613.
- Crouse, J. D., H. C. Knap, K. B. Orno, S. Jorgensen, F. Paulot, H. G. Kjaergaard, and P. O. Wennberg (2012), Atmospheric Fate of Methacrolein: 1. Peroxy Radical Isomerization Following Addition of OH and O-2, *J. Phys. Chem. A*, *116*(24), 5756–5762.
- Curci, G., P. I. Palmer, T. P. Kurosu, K. Chance, and G. Visconti (2010), Estimating European volatile organic compound emissions using satellite observations of formaldehyde from the Ozone Monitoring Instrument, *Atmos. Chem. Phys.*, *10*(23), 11,501–11,517.
- De Smedt, I., T. Stavrakou, J. F. Muller, R. J. van der A, and M. Van Roozendaal (2010), Trend detection in satellite observations of formaldehyde tropospheric columns, *Geophys. Res. Lett.*, *37*, L18808, doi:10.1029/2010GL044245.
- De Smedt, I., M. Van Roozendaal, T. Stavrakou, J. F. Muller, C. Lerot, N. Theys, P. Valks, N. Hao, and R. van der A (2012), Improved retrieval of global tropospheric formaldehyde columns from GOME-2/MetOp-A addressing noise reduction and instrumental degradation issues, *Atmos. Meas. Tech.*, *5*(11), 2933–2949.
- Dee, D., et al. (2011), The ERA-Interim reanalysis: Configuration and performance of the data assimilation system, *Q. J. R. Meteorol. Soc.*, *137*(656), 553–597.
- Dufour, G., F. Wittrock, M. Camredon, M. Beekmann, A. Richter, B. Aumont, and J. P. Burrows (2009), SCIAMACHY formaldehyde observations: Constraint for isoprene emission estimates over Europe?, *Atmos. Chem. Phys.*, *9*(5), 1647–1664.
- Duncan, B. N., et al. (2010), Application of OMI observations to a space-based indicator of NO<sub>x</sub> and VOC controls on surface ozone formation, *Atmos. Environ.*, *44*(18), 2213–2223.
- Emmons, L. K., et al. (2010), Description and evaluation of the Model for Ozone and Related chemical Tracers, version 4 (MOZART-4), *Geosci. Model Dev.*, *3*(1), 43–67.
- Fishman, J., et al. (2012), The United States' next generation of atmospheric composition and coastal ecosystem measurements—NASA's Geostationary Coastal and Air Pollution Events (GEO-CAPE) Mission, *Bull. Am. Meteorol. Soc.*, *93*(10), 1547–1566.

- Fortems-Cheiney, A., F. Chevallier, I. Pison, P. Bousquet, M. Saunois, S. Szopa, C. Cressot, T. P. Kurosu, K. Chance, and A. Fried (2012), The formaldehyde budget as seen by a global-scale multi-constraint and multi-species inversion system, *Atmos. Chem. Phys.*, *12*(15), 6699–6721.
- Fu, T.-M., D. J. Jacob, P. I. Palmer, K. Chance, Y. X. Wang, B. Barletta, D. R. Blake, J. C. Stanton, and M. J. Pilling (2007), Space-based formaldehyde measurements as constraints on volatile organic compound emissions in east and south Asia and implications for ozone, *J. Geophys. Res.*, *112*, D06312, doi:10.1029/2006JD007853.
- Fuchs, H., et al. (2013), Experimental evidence for efficient hydroxyl radical regeneration in isoprene oxidation, *Nat. Geosci.*, *6*(12), 1023–1026.
- Galloway, M. M., et al. (2012), Observations and modeling of formaldehyde at the PROPHET mixed hardwood forest site in 2008, *Atmos. Environ.*, *49*, 403–410.
- Goliff, W. S., W. R. Stockwell, and C. V. Lawson (2013), The regional atmospheric chemistry mechanism, version 2, *Atmos. Environ.*, *68*, 174–185.
- González Abad, G., X. Liu, K. Chance, H. Wang, T. P. Kurosu, and R. Suleiman (2015), Updated Smithsonian Astrophysical Observatory Ozone Monitoring Instrument (SAO OMI) formaldehyde retrieval, *Atmos. Meas. Tech.*, *8*(1), 19–32.
- Grell, G. A., S. E. Peckham, R. Schmitz, S. A. McKeen, G. Frost, W. C. Skamarock, and B. Eder (2005), Fully coupled “online” chemistry within the WRF model, *Atmos. Environ.*, *39*(37), 6957–6975.
- Guenther, A., T. Karl, P. Harley, C. Wiedinmyer, P. I. Palmer, and C. Geron (2006), Estimates of global terrestrial isoprene emissions using MEGAN (Model of Emissions of Gases and Aerosols from Nature), *Atmos. Chem. Phys.*, *6*, 3181–3210.
- Hofzumahaus, A., et al. (2009), Amplified trace gas removal in the troposphere, *Science*, *324*(5935), 1702–1704.
- Jin, X., and T. Holloway (2015), Spatial and temporal variability of ozone sensitivity over China observed from the Ozone Monitoring Instrument, *J. Geophys. Res. Atmos.*, *120*, 7229–7246, doi:10.1002/2015JD023250.
- Karl, T., A. Guenther, A. Turnipseed, G. Tyndall, P. Artaxo, and S. Martin (2009), Rapid formation of isoprene photo-oxidation products observed in Amazonia, *Atmos. Chem. Phys.*, *9*(20), 7753–7767.
- Lamsal, L. N., B. N. Duncan, Y. Yoshida, N. A. Krotkov, K. E. Pickering, D. G. Streets, and Z. Lu (2015), U.S. NO<sub>2</sub> trends (2005–2013): EPA Air Quality System (AQS) data versus improved observations from the Ozone Monitoring Instrument (OMI), *Atmos. Environ.*, *110*, 130–143.
- Lelieveld, J., et al. (2008), Atmospheric oxidation capacity sustained by a tropical forest, *Nature*, *452*(7188), 737–740.
- Mao, J., et al. (2012), Insights into hydroxyl measurements and atmospheric oxidation in a California forest, *Atmos. Chem. Phys.*, *12*(17), 8009–8020.
- Mao, J., F. Paulot, D. J. Jacob, R. C. Cohen, J. D. Crouse, P. O. Wennberg, C. A. Keller, R. C. Hudman, M. P. Barkley, and L. W. Horowitz (2013), Ozone and organic nitrates over the eastern United States: Sensitivity to isoprene chemistry, *J. Geophys. Res. Atmos.*, *118*, 11,256–11,268, doi:10.1002/jgrd.50817.
- Marais, E. A., et al. (2012), Isoprene emissions in Africa inferred from OMI observations of formaldehyde columns, *Atmos. Chem. Phys.*, *12*(14), 6219–6235.
- Marais, E. A., D. J. Jacob, A. Guenther, K. Chance, T. P. Kurosu, J. G. Murphy, C. E. Reeves, and H. O. T. Pye (2014), Improved model of isoprene emissions in Africa using Ozone Monitoring Instrument (OMI) satellite observations of formaldehyde: Implications for oxidants and particulate matter, *Atmos. Chem. Phys.*, *14*(15), 7693–7703.
- Martin, R. V., A. M. Fiore, and A. Van Donkelaar (2004), Space-based diagnosis of surface ozone sensitivity to anthropogenic emissions, *Geophys. Res. Lett.*, *31*, L06120, doi:10.1029/2004GL019416.
- Millet, D. B., D. J. Jacob, K. F. Boersma, T. M. Fu, T. P. Kurosu, K. Chance, C. L. Heald, and A. Guenther (2008), Spatial distribution of isoprene emissions from North America derived from formaldehyde column measurements by the OMI satellite sensor, *J. Geophys. Res.*, *113*, D02307, doi:10.1029/2007JD008950.
- Nolscher, A. C., T. Butler, J. Auld, P. Veres, A. Munoz, D. Taraborrelli, L. Vereecken, J. Lelieveld, and J. Williams (2014), Using total OH reactivity to assess isoprene photooxidation via measurement and model, *Atmos. Environ.*, *89*, 453–463.
- Palmer, P. I., D. J. Jacob, K. Chance, R. V. Martin, R. J. D. Spurr, T. P. Kurosu, I. Bey, R. Yantosca, A. Fiore, and Q. B. Li (2001), Air mass factor formulation for spectroscopic measurements from satellites: Application to formaldehyde retrievals from the Global Ozone Monitoring Experiment, *J. Geophys. Res.*, *106*(D13), 14,539–14,550, doi:10.1029/2000JD900772.
- Palmer, P. I., D. J. Jacob, A. M. Fiore, R. V. Martin, K. Chance, and T. P. Kurosu (2003), Mapping isoprene emissions over North America using formaldehyde column observations from space, *J. Geophys. Res.*, *108*(D6), 4180, doi:10.1029/2002JD002153.
- Palmer, P. I., et al. (2006), Quantifying the seasonal and interannual variability of North American isoprene emissions using satellite observations of the formaldehyde column, *J. Geophys. Res.*, *111*, D12315, doi:10.1029/2005JD006689.
- Paulot, F., J. D. Crouse, H. G. Kjaergaard, J. H. Kroll, J. H. Seinfeld, and P. O. Wennberg (2009a), Isoprene photooxidation: New insights into the production of acids and organic nitrates, *Atmos. Chem. Phys.*, *9*(4), 1479–1501.
- Paulot, F., J. D. Crouse, H. G. Kjaergaard, A. Kuerten, J. M. S. Clair, J. H. Seinfeld, and P. O. Wennberg (2009b), Unexpected epoxide formation in the gas-phase photooxidation of isoprene, *Science*, *325*(5941), 730–733.
- Peeters, J., and J.-F. Müller (2010), HO<sub>x</sub> radical regeneration in isoprene oxidation via peroxy radical isomerisations. II: Experimental evidence and global impact, *Phys. Chem. Chem. Phys.*, *12*(42), 14,227–14,235.
- Rohrer, F., et al. (2014), Maximum efficiency in the hydroxyl-radical-based self-cleansing of the troposphere, *Nat. Geosci.*, *7*(8), 559–563.
- Russell, A. R., A. E. Perrin, L. C. Valin, E. J. Bucsel, E. C. Browne, K. E. Min, P. J. Wooldridge, and R. C. Cohen (2011), A high spatial resolution retrieval of NO<sub>2</sub> column densities from OMI: Method and evaluation, *Atmos. Chem. Phys.*, *11*(16), 8543–8554.
- Russell, A. R., L. C. Valin, and R. C. Cohen (2012), Trends in OMI NO<sub>2</sub> observations over the United States: Effects of emission control technology and the economic recession, *Atmos. Chem. Phys.*, *12*(24), 12,197–12,209.
- Seco, R., T. Karl, A. Guenther, K. P. Hosman, S. G. Pallardy, L. H. Gu, C. Geron, P. Harley, and S. Kim (2015), Ecosystem-scale volatile organic compound fluxes during an extreme drought in a broadleaf temperate forest of the Missouri Ozarks (central USA), *Global Change Biol.*, *21*(10), 3657–3674.
- Song, C. H., H. S. Kim, R. von Glasow, P. Brimblecombe, J. Kim, R. J. Park, J. H. Woo, and Y. H. Kim (2010), Source identification and budget analysis on elevated levels of formaldehyde within the ship plumes: A ship-plume photochemical/dynamic model analysis, *Atmos. Chem. Phys.*, *10*(23), 11,969–11,985.
- Stavrakou, T., J. F. Müller, I. De Smedt, M. Van Roozendaal, G. R. van der Werf, L. Giglio, and A. Guenther (2009a), Global emissions of non-methane hydrocarbons deduced from SCIAMACHY formaldehyde columns through 2003–2006, *Atmos. Chem. Phys.*, *9*(11), 3663–3679.
- Stavrakou, T., J. F. Müller, I. De Smedt, M. Van Roozendaal, G. R. van der Werf, L. Giglio, and A. Guenther (2009b), Evaluating the performance of pyrogenic and biogenic emission inventories against one decade of space-based formaldehyde columns, *Atmos. Chem. Phys.*, *9*(3), 1037–1060.
- Sumner, A. L., et al. (2001), A study of formaldehyde chemistry above a forest canopy, *J. Geophys. Res.*, *106*(D20), 24,387–24,405, doi:10.1029/2000JD900761.

- Taraborrelli, D., M. G. Lawrence, J. N. Crowley, T. J. Dillon, S. Gromov, C. B. M. Gross, L. Vereecken, and J. Lelieveld (2012), Hydroxyl radical buffered by isoprene oxidation over tropical forests, *Nat. Geosci.*, *5*(3), 190–193.
- Whalley, L. K., et al. (2011), Quantifying the magnitude of a missing hydroxyl radical source in a tropical rainforest, *Atmos. Chem. Phys.*, *11*(14), 7223–7233.
- Wiedinmyer, C., et al. (2005), Ozarks Isoprene Experiment (OZIE): Measurements and modeling of the “isoprene volcano”, *J. Geophys. Res.*, *110*, D18307, doi:10.1029/2005JD005800.
- Witte, J. C., B. N. Duncan, A. R. Douglass, T. P. Kurosu, K. Chance, and C. Retscher (2011), The unique OMI HCHO/NO<sub>2</sub> feature during the 2008 Beijing Olympics: Implications for ozone production sensitivity, *Atmos. Environ.*, *45*(18), 3103–3111.
- Wolfe, G. M., J. D. Crounse, J. D. Parrish, J. M. St. Clair, M. R. Beaver, F. Paulot, T. P. Yoon, P. O. Wennberg, and F. N. Keutsch (2012), Photolysis, OH reactivity and ozone reactivity of a proxy for isoprene-derived hydroperoxyenals (HPALDs), *Phys. Chem. Chem. Phys.*, *14*(20), 7276–7286.
- Zhu, L., D. J. Jacob, L. J. Mickley, E. A. Marais, D. S. Cohan, Y. Yoshida, B. N. Duncan, G. G. Abad, and K. V. Chance (2014), Anthropogenic emissions of highly reactive volatile organic compounds in eastern Texas inferred from oversampling of satellite (OMI) measurements of HCHO columns, *Environ. Res. Lett.*, *9*(11), doi:10.1088/1748-9326/9/11/114004.