

1 Linking land and sea through an ecological-economic model of coral reef
2 recreation

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39 Highlights:

- 40 ● Model integrates social values to simulate coastal management outcomes
- 41 ● Snorkelers prefer sites with better visibility, fish abundance and diversity
- 42 ● Coastal recreation benefits and management priorities vary spatially
- 43 ● Land-sea management provides best strategy overall and at most local beach sites
- 44 ● Some beaches require unique strategies to maximize benefit

45 **Abstract**

46 Coastal zones are popular recreational areas that substantially contribute to social welfare.
47 Managers can use information about specific environmental features that people value, and how

48 these might change under different management scenarios, to spatially target actions to areas
49 of high current or potential value. We explored how snorkelers' experience would be affected by
50 separate and combined land and marine management actions in West Maui, Hawai'i, using a
51 Bayesian Belief Network (BBN) and a spatially explicit ecosystem services model. The BBN
52 simulates the attractiveness of a site for recreation by combining snorkeler preferences for
53 coastal features with expert opinions on ecological dynamics, snorkeler behavior, and
54 management actions. A choice experiment with snorkelers elucidated their preferences for sites
55 with better ecological and water-quality conditions. Linking the economic elicitation to the
56 spatially explicit BBN to evaluate land-sea management scenarios provides specific guidance
57 on where and how to act in West Maui to maximize ecosystem service returns. Improving
58 coastal water quality through sediment runoff and cesspool effluent reductions (land
59 management), and enhancing coral reef ecosystem conditions (marine management) positively
60 affected overall snorkeling attractiveness across the study area, but with differential results at
61 specific sites. The highest improvements were attained through joint land-sea management,
62 driven by strong efforts to increase fish abundance and reduced sediment; however, the effects
63 of management at individual beaches varied.

64

65

66 Keywords: Bayesian Belief Network; Recreational ecosystem service; Management scenario
67 evaluation; Land-sea interactions; Hawai'i

68 1 Introduction

69 The opportunity for recreation is an important coastal ecosystem service, particularly in places
70 where coral reefs support thriving tourism and leisure sectors (Brander et al., 2007; Moberg and
71 Folke, 1999; Spalding et al., 2017). This predominantly non-consumptive service sustains
72 residents living near coral reefs and fuels a multibillion-dollar global tourism industry (Pendleton,
73 1994; Spalding et al., 2017). People directly enjoy reefs when SCUBA diving, snorkeling, and
74 fishing, while activities such as swimming, sunbathing, beachcombing, and surfing at the coast
75 may also be reef-dependent. Particular characteristics of coral reef ecosystems, like complex
76 structure and diverse fauna, directly impact snorkeling, diving, fishing, and even surfing user
77 experiences (Brander et al., 2007; Principe et al., 2012). Globally, a series of studies have
78 documented abiotic, biotic, and social features of reefs that make them valuable to people for
79 recreation (Beharry-Borg and Scarpa, 2010; Cooper et al., 2009; Inglis et al., 1999; Pendleton,
80 1995) including conditions of the reef and fish, presence of charismatic megafauna, water
81 clarity, pollution, and crowding. While visitation, visitor spending, and associated economic
82 impacts may be easier to measure, the recreational attractiveness of reefs may be more difficult
83 to directly measure (Principe et al., 2012).

84

85 Human impacts directly affect the attributes that make reefs most valuable for recreation.
86 Anthropogenic stressors, both global and local, can cause widespread coral mortality that leads
87 to rapid and hard to reverse shifts away from coral dominated systems (Hughes et al., 2007;
88 Nyström et al., 2008), with cascading effects on fish abundance and diversity (Pratchett et al.,
89 2008). Specifically, corals are threatened by extreme sea temperature anomalies that cause
90 coral bleaching, where corals expel their algal symbionts, and if temperatures stay high for too
91 long, this can lead to widespread mortality (Brown and Roughgarden, 1997; Hoegh-Guldberg,

92 1999). Pollution can smother corals (in the case of sediment), exacerbate coral disease (in the
93 case of pathogens from sewage), cause algal outbreaks (in the case of nutrients), have
94 sublethal effects that alter reef genetics, and kill coral outright (in the case of toxins, including
95 sunscreen) (Anthony et al., 2015). Further, unsustainable levels of fish harvest can unbalance
96 the system (Jackson et al., 2001), leading to cascading effects on important ecological
97 processes such as herbivory (Hughes et al., 2010; Mumby and Steneck, 2008). Given the
98 multiple and potentially synergistic and cumulative effects of stressors on reef ecosystems (Ban
99 et al., 2014; Darling and Coté, 2008), research is needed to guide management actions aimed
100 at understanding the boundaries for success, and the tradeoffs that exist among multiple
101 stressors for preventing declines and enhancing recovery that leads to delivery of reef-based
102 recreational ecosystem services (Jouffray et al., 2019; Weijerman et al., 2018).

103

104 A detailed understanding of recreationalists' preferences for coral reef conditions can help
105 managers focus their efforts to preserve or enhance reefs so they can deliver valued ecosystem
106 services. The recreational value of coral reefs has been widely researched in the ecological-
107 economics literature, but, apart from a handful of exceptions where spatial methods were used
108 (Ghermandi and Nunes, 2013; Ruiz-Frau et al., 2013; Spalding et al., 2017; van Riper et al.,
109 2012), studies have predominantly used environmental valuation methods that are point in time
110 estimates with no spatial component. Furthermore, these approaches rarely link values to
111 specific attributes in ways that enable simulation of threats and management scenarios (one
112 exception is van Beukering and Cesar (2004)). Recreational valuation studies have historically
113 relied on methods like contingent valuation, where respondents were asked to state their
114 willingness to pay for certain beach attributes (Ahmed et al., 2007; Loomis and Santiago, 2013;
115 Petrosillo et al., 2007), choice experiments, where respondents were asked to make
116 hypothetical trade-offs amongst attributes (Beharry-Borg and Scarpa, 2010; Nunes et al., 2015;
117 Schuhmann et al., 2013), or travel cost, where respondents' actual recreational behavior was

118 used to model willingness to pay (Ahmed et al., 2007; Ariza et al., 2012; Carr and Mendelsohn,
119 2003; Loomis and Santiago, 2013; Zhang et al., 2015). For a review of valuation studies in
120 islands see Oleson et al. (2018). Despite this effort, most coral reef valuation studies have not
121 been contextualized in a manner that enables place-based management scenario analysis.

122

123 Massive efforts are dedicated to coastal management globally, which raises questions on
124 whether these efforts are targeted at locations and conditions that are most valuable to society.
125 The aim of this study is to develop an applied valuation methodology that provides specific and
126 useful management guidance to coastal managers. Information on the perceived value of
127 specific areas for recreation - and how these might change under different scenarios - could
128 help communities to ensure persistence of important values and services. Specifically, we
129 assess the benefits to recreationalists and recreation-dependent communities of potential land
130 and marine management strategies so that managers can prioritize which actions to take, and
131 where these actions will yield the greatest benefits. To be relevant, our approach needs to
132 include features of the nearshore environment that land and marine management could directly
133 or indirectly affect, as well as physical and social features that influence the value of a site, such
134 as access and crowding. It has to be ecologically sound, based on the best scientific
135 understanding of coral reef dynamics, while also being grounded on the user experience. Our
136 methodology rests on a Bayesian Belief Network (BBN) to integrate multiple types of
137 information, including expert judgment about ecological dynamics, management, and snorkeler
138 behavior, and snorkelers' stated preferences elicited through a choice experiment. While BBNs
139 have been used in studies of coral ecology (Franco et al., 2016; Graham et al., 2008), this is the
140 first study to use BBNs to assess ecosystem services in coral reef systems. An ecosystem
141 services approach is relatable to decision makers, visitors, and residents as it ties ecological
142 conditions to human preferences and wellbeing outcomes (Tallis and Polasky, 2009; Wainger
143 and Mazzotta, 2011; Wainger and Boyd, 2009). The novel ecological-economic method we

144 developed has the advantages of being able to model and provide spatially nuanced and policy-
145 grounded information for conservation and resource management planning. In our spatially
146 explicit case study we identify areas where management returns are highest, as well as specific
147 management measures that would have the largest pay-off for popular beaches on the
148 northwest part of the island of Maui, Hawai'i, USA.

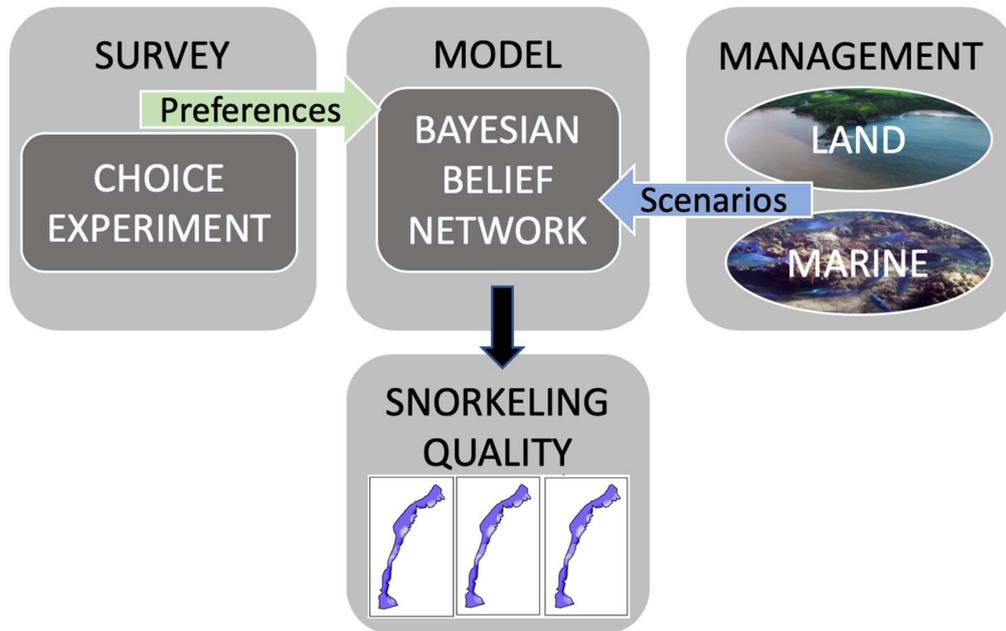
149

150 The rest of the paper is organized as follows. We first present a methodological framework to
151 provide an overview of the methods and models, and how they are linked. We then describe our
152 study site, the survey instrument, choice experiment, and Bayesian Belief Network modeling. In
153 each of these sections we detail the method and the results, as the results are then used as
154 inputs to the subsequent section (i.e., the choice experiment results inform the BBN, which
155 underpin the scenarios). A scenario modeling section follows, describing results of different land
156 and marine management strategies on recreation. Our discussion section focuses on the
157 management implications, modeling innovations, and study limitations.

158 2 Methodological framework

159 Our approach to modeling management effects on the quality of a site for snorkeling integrates
160 different methods and datasets (Figure 1). A survey of snorkelers used a choice experiment to
161 elicit preferences for site attributes. These preferences then helped calibrate a spatial BBN,
162 which connected what snorkelers said they care about to land and marine management actions
163 that affect coral reef ecosystems. The model then outputted maps of snorkeling quality for
164 various land and marine management scenarios.

165



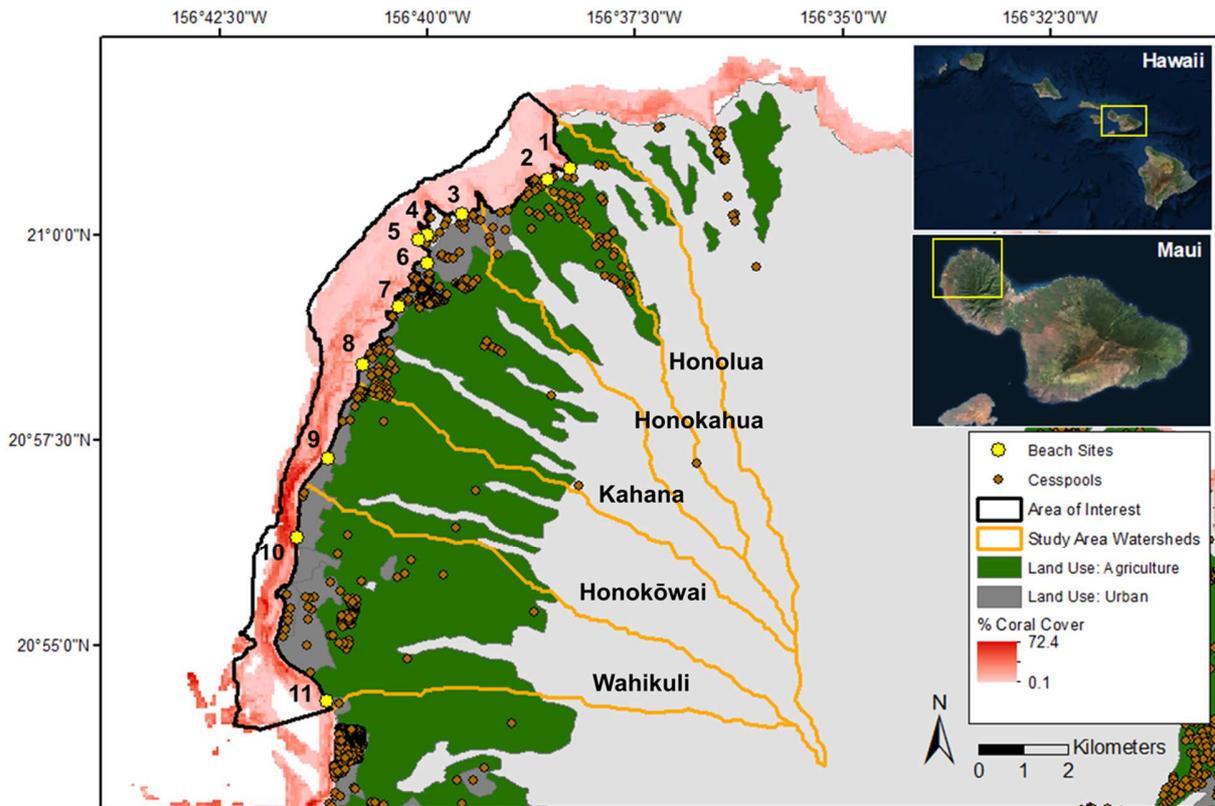
166

167 *Figure 1 Methodological framework*

168 **3 Site characteristics**

169 Over 167,000 people are residents of Maui island, in the state of Hawai'i, USA (U.S. Census
 170 Bureau, 2017). Nearly three million (2.7 million) tourists visited Maui in 2017, spending \$4.68
 171 billion (Hawai'i Tourism Authority, 2016). Our case-study focuses on West Maui (Figure 2). West
 172 Maui's coasts are a popular recreation destination for tourists and residents, many of whom are
 173 attracted to the calm, clear waters and historically high-quality coral reefs. World-famous
 174 beaches in the West Maui region are prime recreation sites. Today, land previously farmed as
 175 sugarcane or pineapple plantations for over a century is kept as fallow or being converted for
 176 residential use, while resort development continues along the coast. Unfortunately, West Maui's
 177 coral reefs have declined in the past fifteen years as a result of fishing and pollution from land
 178 (Sparks et al., 2015).

179



180
 181 *Figure 2 Map of study site with beaches where surveys took place (numbered yellow circles), land use, cesspools*
 182 *(orange dots), and coral reef cover depicted. Watershed boundary and land use from State of Hawai'i Office of*
 183 *Planning (2019) and predicted coral cover from Weijerman et al. (2018).*

184 4 Survey instrument

185 We used a tablet-based survey to collect responses from 290 recreational snorkelers in West
 186 Maui between August and September 2015. We intercepted resident and tourist snorkelers at
 187 beaches and in resort areas (Figure 2), distributing our sampling effort across five watersheds
 188 running north to south (Honolua (5% of respondents), Honokahua (8%), Kahana (22%),
 189 Honokōwai (8%), and Wahikuli (57%) based on visitation, which we estimated using a crowding
 190 model based on social media photo uploads (Wood, Guerry et al. 2013). The survey instrument,
 191 approved by University of Hawai'i's Institutional Review Board (2016-31181), was tested on
 192 beach goers on O'ahu, a nearby island, for convenience; snorkelers on both islands are similar.

193 The survey included questions related to demographics, knowledge, values, experience, and
194 preferences for attributes of snorkeling sites. We focused on snorkelers, as snorkeling is a
195 common activity for both residents and tourists, and snorkelers tend to be aware of
196 environmental conditions. The design enabled us to explore possible differences between
197 residents and tourists. The survey instrument is included as supplementary information (SI_S1).

198

199 Full descriptive statistics are provided in Table SI_T1. Just over half (53%) of the respondents
200 were female. Eighty-one were permanent Maui residents, twenty were seasonal residents, and
201 180 were visitors. The median respondent age was 45, higher than the median in the county
202 (37), the median annual household income was \$87,500, also higher than the average in Maui
203 County (\$72,762), and the sample was more educated than average (26.3% of 167,000
204 residents have a college degree vs. 60% in the sample) (U.S. Census Bureau, 2017). While
205 Maui residents are ethnically diverse, the sample was skewed towards Caucasians (65% vs.
206 35% in Maui (U.S. Census Bureau, 2017)), likely reflecting both the tourists and the
207 demographic who snorkels at the beaches surveyed. Most respondents reported additional
208 snorkeling experience in locations other than Maui (240), and 40 said they had experience
209 snorkeling on Maui. Ten noted they had no snorkeling experience and were planning on going.
210 Snorkelers with experience had a median of 20 events. Nearly a third of all respondents (92)
211 were also SCUBA divers.

212 5 Choice experiment

213 Following examples such as Schuhmann et al. (2013), we used a discrete choice experiment to
214 determine snorkeler preferences for environmental attributes that may be affected by
215 management and/or climate change. Snorkelers were asked to choose among three different
216 beaches characterized by different travel costs and attributes. These attributes represent a
217 subset of those important for snorkeler satisfaction that were cited during interviews with experts

218 and local snorkelers, and reported in the literature (Beharry-Borg and Scarpa, 2010; Loomis and
219 Santiago, 2013; Peng et al., 2017). Due to known cognitive limitations when evaluating trade-
220 offs in choice experiments (Johnston et al., 2017), we restricted the number of environmental
221 attributes included in our choice experiment to: water quality, visibility, fish abundance and
222 diversity, coral cover, and chance of seeing sea turtles, as well as price, which represents both
223 transportation costs to access the beach and the opportunity value of time (Fezzi et al., 2014).
224 We set three levels for each environmental attribute (Table 1), while travel cost had six levels¹.
225 The levels of all attributes were depicted in photos (Figure SI_F1). Each respondent faced 10
226 choice tasks. We validated these levels by asking respondents about their perceptions of
227 snorkeling on Maui.

228

229 A full factorial design for our choice experiments that includes all possible combinations
230 of attributes and levels would use 4,374 choice tasks ($3 \times 3 \times 3 \times 3 \times 3 \times 3 \times 6 = 4,374$). To optimize the
231 discrete choice experiment survey design from the total possible combinations and to reduce
232 fatigue of the respondents, 100 choice tasks with two alternative combinations of attributes and
233 one fixed status quo were generated in a series of ten different choice set versions with ten
234 choice tasks per version in SSI Web 10.0 Sawtooth Software (Sawtooth Software, Utah, USA).
235 We used Sawtooth CBC System for Choice-Based Conjoint Analysis to design the discrete
236 choice experiment. To ensure level balance, and to minimize overlap with levels and correlation
237 between attributes, the choice tasks were generated with an orthogonal experimental design.

238

239 *Table 1 Attributes and levels for choice experiment*

¹ At first glance the price vector may appear quite high, but it actually represents well the travel costs characterizing our case study. In fact, travel time on Maui can be surprisingly long for such a small island, mainly because of the quality of the roads. For example, let us consider a trip from the West side of the island (Kā'anapali) to the East side (Hana). Such a journey can take about three-and-a-half hours with no traffic. In order to calculate the travel cost for such a trip we need to consider both the car running costs (e.g., fuel costs) and the opportunity value of time. Including both components and calculating the value of time as 3/4 of the average wage rate (Fezzi et al, 2014), we obtain a round trip travel cost of more than \$200.

Attribute	Low (Base condition)	Moderate	High	Citation/Justification
Bacterial warning	12 days/year	6 days/year	0 days/year	(Hawai'i Department of Health, 2019) and DOH experts
Visibility	15 feet	30 feet	60 feet	NOAA experts
Coral cover	<15%	26%	>45%	(Sparks et al., 2015)
Fish abundance	75/125m ²	115/125m ²	150/125m ²	(Friedlander et al., 2005; Williams et al., 2008)
Fish diversity	8 species	17 species	28 species	
Turtle sighting	P(sighting) = 0%	<50%	>50%	NOAA experts
Price	\$0, \$10, \$50, \$100, \$175, \$250			Estimate of cost for extra time and transportation

DOH = Department of Health

NOAA = National Oceanic and Atmospheric Administration

240

241 We analyzed the choice experiment data by specifying a random utility model (RUM), following
 242 the method established by McFadden (1974). Under this framework, the utility that respondent j
 243 receives from choosing option i can be written as:

244

$$245 \quad (1) U_{ij} = \sum_{k=1}^5 \theta_{ki} + \gamma cost_i + \beta SQ_i + \varepsilon_{ij}, \quad (1)$$

246

247 Where θ_{ki} indicates the part of utility for each of the five attributes (k) characterizing option i ,
 248 $cost_i$ is the cost of access, γ is the marginal utility of money, SQ_i is a dummy variable indicating
 249 whether the option is the status quo, β is the parameter allowing for "status-quo bias," and ε_{ij} is
 250 the random component encompassing the unobserved (to the researcher) part of the utility that

251 person i associates to option j . The θ_{ki} coefficients illustrate the relative importance of attributes
252 and their levels, and the willingness of respondents to trade one attribute level for another. To
253 allow for maximum modelling flexibility, we model each attribute via dummy variables, with the
254 worst level for each attribute selected as the baseline (for example, for the attribute “bacterial
255 warnings” the baseline level is 12 days per year). The baseline levels were the omitted
256 categories of the dummy coded variables.

257

258 Again following McFadden (1974), by assuming the random error ε_{ij} to be identically and
259 independently distributed as a type I extreme value (i.e., Gumbel), and indicating with V_{ij} the
260 observed portion of the utility (i.e., $V_{ij} = U_{ij} - \varepsilon_{ij}$), we can write the probability of choosing
261 alternative i as:

262

$$263 \quad P_{ij} = \frac{\exp(V_{ij})}{\sum_{h=1}^3 \exp(V_{ih})} \quad (2)$$

264

265 This conditional logit specification includes all the parameters in (1) and can be estimated via
266 maximum likelihood. We analyzed the data using a conditional logit model via the *mcllogit*
267 package in R.

268

269 Results of the choice experiment are summarized in Table 2. All attribute coefficients are
270 significant. Interviewed snorkelers preferred sites with better ecological and water quality
271 conditions, especially high and moderate visibility (coefficients 0.747 and 0.615), followed by
272 high coral cover (0.497), high chance of sighting turtles (0.469), high bacteriological quality
273 (0.465), and finally high fish diversity (0.379) and abundance (0.344). In many cases, most of
274 the value to snorkelers lay in improving conditions to the moderate level from the base level;
275 any additional improvement to the high level was less valued. This diminishing return is

276 particularly strong in the visibility characteristic, suggesting that people were happy with being
 277 able to see 30 feet (+0.615) but the additional gains from visibility up to 60 feet were less valued
 278 (+0.132). In contrast, fish diversity and abundance showed roughly linear preferences from base
 279 conditions through moderate to high. Notably, there were few differences amongst groups.
 280 Residents had similar preferences as tourists and seasonal residents, with one exception
 281 (residents prioritized visibility more), although the low sample size of residents prevents
 282 comparison of many of the attributes (Table SI_T2).

283

284 *Table 2 Choice experiment results. Z-value is the number of standard deviations from the mean value.*

Attribute	Estimate	Std. error	z-value	
Bacteria: 0 days	0.465	0.066	7.046	***
Bacteria: 6 days	0.243	0.063	3.834	***
Visibility: 30 feet (9.14 m)	0.615	0.063	9.707	***
Visibility: 60 feet (18.29 m)	0.747	0.065	11.378	***
Coral cover: high	0.497	0.065	7.628	***
Coral cover: medium	0.304	0.061	4.962	***
Fish number: high	0.344	0.062	5.478	***
Fish number: medium	0.149	0.065	2.27	*
Fish diversity: high	0.379	0.065	5.849	***
Fish diversity: medium	0.144	0.063	2.282	*
Turtles: high	0.469	0.064	7.369	***
Turtles: low	0.234	0.066	3.543	***
Cost	-0.006	0.000	-19.164	***
Status quo	-0.658	0.112	-5.868	***

pseudo R² 0.27

Log likelihood -2281.83

285 Notes: parameters need to be interpreted as differences with the baseline category, which is
286 omitted from the model. For example, for bacteria the baseline category is 12 days in which
287 bathing is unsafe because of potential contamination, for visibility it is 15 feet. All attributes
288 are in Table 1.

289 6 Bayesian Belief Network

290 A BBN graphs the causal structure of variables in an inference or modeling problem, and uses
291 conditional probability distributions to define relationships between variables (Aguilera et al.,
292 2011; Ames et al., 2005). Combining diverse sources of information within a BBN is particularly
293 important when one cannot include all attributes characterizing choices within a stated
294 preference exercise, for well-known issues of cognitive burden (Johnston et al., 2017). BBNs
295 have been used to model ecosystem services (Dee et al., 2017; Landuyt et al., 2013), and as a
296 tool for planning (Gonzalez-Redin et al., 2016), pollution impact assessment (Spence and
297 Jordan, 2013), guiding adaptive management (Nyberg et al., 2006), and assessing ecological
298 water quality (Forio et al., 2015).

299

300 Our BBN model estimates spatially explicit relative snorkeling attractiveness in the West Maui
301 study area by integrating attributes of ecological, water, and social quality such as coral cover,
302 fish richness, pollution, depth, and accessibility. The model area of interest (AOI) consisted of
303 West Maui shoreline from Honolua Bay to south of Black Rock Point, extending to 30m depth
304 (Figure 2). The model variables, structure, and strength of relationships between variables were
305 informed by a literature review, experts (Kuhnert et al., 2010), and the choice experiment
306 described in the section above. Past valuation studies were useful in identifying important

307 attributes for beach users, particularly divers and snorkelers (Grafeld et al., 2016; Parsons and
 308 Thur, 2008; Pendleton, 1994; Schuhmann et al., 2013; Wielgus et al., 2002).

309

310 Ultimately, the BBN had 11 attribute parent nodes that interact, as illustrated by the arrows, in
 311 order to determine snorkeling attractiveness (“Snorkeling Quality” in Figure 3). Each of these
 312 parent nodes have spatial data associated with them (Table 3) (SI, Figure SI_F2A-K). The
 313 current status of each attribute (i.e., prior probabilities) in West Maui is represented by the
 314 colored bars within the parent nodes; these represent the average status across the entire AOI
 315 and are divided into bins (Table 3, Figure 3). Parent nodes are aggregated into four
 316 intermediate nodes (social quality, water quality, visibility, and ecological quality) that determine
 317 snorkeling quality. The grouping of parent nodes into intermediate nodes simplifies the
 318 conditional probabilities of the BBN model and thus reduces the cognitive load required to
 319 determine the relationships. The selection of parent nodes and arrangement of intermediate
 320 nodes constitutes the causal structure of the model. We tested a number of model structures via
 321 interviews with 15 experts, including two marine scientists with the Hawai’i Division of Aquatic
 322 Resources (DAR, the state agency charged with coral reef management), a lifeguard working in
 323 the area, ten avid snorkelers, and two snorkel tour operators.

324

325 *Table 3 Attributes in the Bayesian Belief Network (BBN)*

Attributes	Data source	Measurement & Bins in BBN	Data resolution
Access	(Hawai’i Mapping Research Group, 2016; Wedding et al., 2018)	1-4 (classification)	10m
Exposure	(Wedding et al., 2018)	<5,300, >5,300 (wave energy, J*s/m)	500m
Crowding	(Wood et al., 2013)	<3, 3-6, >6 (Photograph user	60m

Cesspool discharge	Data from Barnes et al. (2019) using methods from Wedding et al. (2018)	0-0.004, 0.004-0.008, >0.008 (kg N/m ²)	500m
Sediment dispersion	Updated, using methods from Wedding et al. (2018)	0-3, 3-10, >10 (ton/ha)	30m
Bathymetry	(Hawai'i Mapping Research Group, 2016)	0-10, >10 (m depth)	5m
Coral cover	(Weijerman et al., 2018)	<20, 20-35, >35 (% cover)	60m
Fish abundance	(Weijerman et al., 2018)	<0.76, 0.76-1.06, >1.06 (count/m ²)	60m
Fish species richness	(Weijerman et al., 2018)	<8, 8-17, >17 (count/grid cell)	60m
Habitat diversity	(Friedlander and Kendall, 2006)	<0.37, 0.37-0.74, >0.74 (ranking)	60m
Turtle chance as a function of habitat	(National Centers for Coastal Ocean Science, 2007)	0-0.35, 0.35-0.99, 0.99-1 (% likelihood of viewing)	50m

Note: Probability of spotting turtles calculated as a function of habitat. High probability - coral dominated hard bottom habitat; Medium probability - algal dominated habitat (including macroalgae, turf, and crustose coralline algae (CCA)), both hard and soft bottom; Low probability - everything else - primarily uncolonized soft bottom or unknown/unclassified.

326

327 The next step was to set the relative importance of each variable via conditional probability

328 tables. The conditional probability distribution defines the relative importance of each parent

329 node. For instance, the intermediate node “water quality” is determined based on the value of

330 two parent nodes, cesspool discharge and sediment dispersion. The water quality outcome is

331 determined by specifying the likelihood that water quality is high, moderate, or low, given levels

332 of cesspool discharge and sediment dispersion (the values of each column always sum to 1).

333 An example conditional probability table for the water quality node is presented in Table 4. The

334 thickness of the arrows in Figure 3, which illustrate each variable’s relative importance to the

335 outcome, denoting average Euclidian influence, are based on the conditional probabilities

336 (Koiter, 2006). Water quality is a relatively simple intermediate node, with only two
 337 determinants; as the relationships become more complicated, the number of columns in the
 338 tables expand very rapidly.

339

340 *Table 4 Water quality (intermediate node) conditional probability table given parent nodes Cesspool discharge and*
 341 *Sediment dispersion.*

Water Quality									
Cesspool Discharge	High			Moderate			Low		
Sediment Dispersion	High	Moderate	Low	High	Moderate	Low	High	Moderate	Low
High	0	0	0.1	0	0.2	0.3	0.4	0.8	0.9
Moderate	0.05	0.1	0.1	0.6	0.6	0.6	0.4	0.2	0.1
Low	0.95	0.9	0.8	0.4	0.2	0.1	0.2	0	0

342

343 We populated the conditional probability tables based on our data from the choice experiment
 344 and additional survey questions, as well as through consultation with coral reef managers and
 345 experts. The choice experiment focused on a limited number of the variables (six) in the BBN to
 346 elicit their relative importance for snorkelers in West Maui. For instance, from the choice
 347 experiment results we understand that snorkelers in West Maui highly valued improved visibility
 348 more than reductions in the probability of bacteriological water quality below recreational water
 349 standards. Features of social quality (like access and crowding) were assessed in the survey.

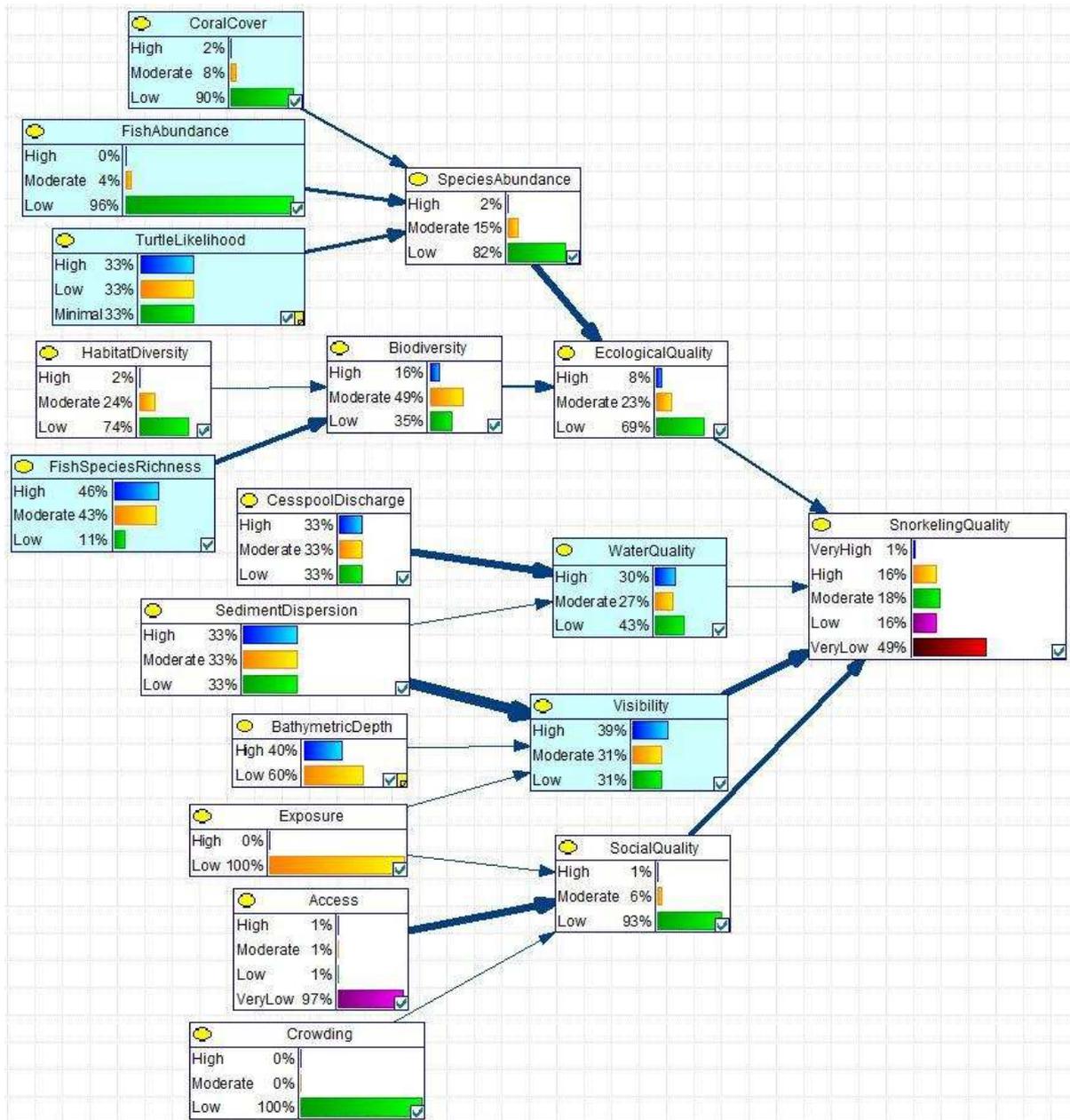
350 Interviews with experts elicited the relative importance of the other variables. Conditional
 351 probability tables for all variables are in Table SI_T4a and strength of influence in Table SI_T4b.

352

353 The model output is a score (from 0 to 100) of the quality or attractiveness of each grid cell for
 354 recreational snorkelers. A score of 100 indicates a very high-quality snorkeling site within the
 355 study area, and 0 very poor. This score range is specific to the AOI and normalized to the range

356 of outcomes and scenarios in this analysis. The score is binned into five levels (0-20 very low;
357 21-40 low; 41-60 moderate; 61-80 high; and 81-100 very high). To explore assumptions of the
358 model, we ran various hypothetical scenarios to see if the results were consistent with
359 expectations. For instance, we set the value of model inputs that the choice experiment or
360 experts told us were highly important (e.g., turtle-sighting likelihood, fish species richness, or
361 visibility) to the highest possible values and evaluated the model's sensitivity to changes in
362 these inputs, as opposed to those deemed to be less important (e.g., crowding or habitat
363 diversity). We generated results for the entire study area, as well as for subsetted areas within
364 the highly and moderately accessible areas surrounding popular beaches. We ran models for
365 current conditions and a set of management scenarios (described below) at 50 m resolution
366 using the Artificial Intelligence for Ecosystem Services (ARIES) modeling platform (Villa et al.,
367 2014).

368



369

370 *Figure 3 Bayesian Belief Network describing a site's snorkeling quality. Nodes shaded in light blue indicate variables*
 371 *included in the choice experiment. Arrow thickness denotes average Euclidian influence per the conditional*
 372 *probability tables (strength of influence for each relationship is included in SI Table SI_T4b). The most influential*
 373 *relationship (Sediment Dispersion on Visibility) is about 10 times the value of the weakest relationship (Crowding on*
 374 *Social Quality). The colored bars indicate current conditions across all pixels in the Area Of Interest in Figure 2.*

375 7 Scenario modeling

376 A primary objective of this paper is to determine what management actions would be most
377 effective and where their implementation would have the strongest effects. Therefore, we
378 modeled a number of land and marine management scenarios. Land management options
379 target sediment and effluent reduction from cesspools. Marine-based management included
380 reducing fishing, and the effect of changes in coral cover and associated fish abundance and
381 richness. Target levels for these reductions were based on the goals stated in official watershed
382 management plans (Group 70, 2015a, 2015b; Sustainable Resources Group International,
383 2012a, 2012b) and telephone, email, and in-person interviews with the watershed management
384 coordinator, environmental consultants who prepared the watershed management plans, the
385 State aquatic resource manager, and a Federal coral reef ecologist familiar with the area. We
386 used four different levels for each scenario to represent increasing levels of investment in each
387 type of management.

388 Land-based management

389 In the watersheds upstream of West Maui's coral reefs, former agricultural lands currently
390 remain fallow and access roads unfixed, stream banks continue to erode, and no cesspools are
391 upgraded (Oleson et al., 2017; Stock et al., 2016; Whittier and El-Kadi, 2014). Land-based
392 management scenarios represent realistic and aspirational levels of local pollution abatement.
393 We modeled the following individually and in combination: reduce sediment input by 10%, 15%,
394 20%, and 25%; reduce cesspool input by 10%, 25%, 50%, and 100%. Notably, cesspools
395 contributed 14% of total groundwater nitrogen flux, but we did not adjust input layers for known
396 cesspool upgrades, and we ignored discharge from the Kahekili wastewater treatment plant
397 (0.3% of groundwater nitrogen flux) (Barnes et al., 2019).

398 Marine-based management

399 We also constructed a second set of management scenarios based on improvements to coral
400 reef benthic habitat and associated changes in coral reef fish communities. Local coral reef
401 experts agreed that increasing coral cover by 5%, 10%, 15%, and 20% above current levels
402 were reasonable aspirations in this area, particularly given historical coral cover levels and
403 improvements in managed areas (Williams et al., 2016). To estimate how fish biomass would
404 change under different marine management scenarios, we draw upon a previously published
405 hierarchical, linear Bayesian model of how multiple biophysical and human population drivers
406 influence fish biomass throughout the main Hawaiian Islands (Gorospe et al., 2018). Data from
407 the same study show that increases in coral cover would also result in increases in reef
408 complexity (Figure SI_F3). Therefore, although reef complexity was not a component of our
409 snorkeler choice experiments, we use both coral cover and complexity to estimate changes in
410 reef fish biomass. Finally, applying a linear model to data from West Maui fish surveys, we
411 translate modeled fish biomass into the more snorkeler-relevant metrics of fish abundance
412 (Figure SI_F4A) and fish species richness (Figure SI_F4B). Overall, this allowed us to derive a
413 complete picture of how the reef attributes in the BBN (coral cover, fish abundance, and fish
414 species richness) collectively changed (Table 5). All data for the above analyses came from fish
415 and benthic surveys conducted by the NOAA Pacific Islands Fisheries Science Center's
416 Ecosystem Science Division in 2012, 2013, and 2015 (Pacific Islands Fisheries Science Center,
417 2019).

418

419 *Table 5 Model-predicted fish biomass, abundance, and species richness based on hypothetical, absolute increases in*
420 *percent coral cover achievable with management. Using field data from throughout the main Hawaiian Islands, a*
421 *hierarchical, linear Bayesian model (Gorospe et al. 2018) was used to predict fish biomass based on increases in*
422 *coral cover and associated increases in reef complexity. Modeled fish abundance and richness outcomes are*
423 *presented for different levels of absolute coral cover change over baseline, where the baseline is the current mean for*

424 the Maui-Lahaina area. When coral reef cover increases over the baseline, the model predicts coral reef complexity
 425 increase (Figure SI_3), fish biomass, fish abundance, and fish richness. For instance, moving from baseline coral
 426 cover and complexity to a scenario where coral cover increases to baseline+5%, fish biomass would increase from
 427 5.89g/m² to 7.10g/m², fish abundance from 0.028 fish/m² to 0.039 fish/m² (scenario is 139% of baseline), and fish
 428 richness from 6.13 to 6.97 species (scenario is 114% of baseline).

Coral Cover (% absolute change over baseline at a site)	Model-linked Fish Biomass	Fish Abundance		Fish Richness	
	(g/m ²)	(# fish/m ²)	(% of baseline)	(# species)	(% of baseline)
Baseline	5.89	0.028	100%	6.13	100%
+5	7.10	0.039	139%	6.97	114%
+10	8.33	0.050	178%	7.83	128%
+15	9.63	0.062	220%	8.74	143%
+20	10.97	0.074	263%	9.68	158%

429

430 Combined marine-land management

431 As a third set of management scenarios, we combined all management outcomes into a single
 432 scenario, where both land-based pollution was reduced and benthic habitat and fish
 433 communities were rehabilitated at increasing levels.

434 Scenario results

435 Baseline snorkeling attractiveness was estimated using the BBN under current conditions and is
 436 mapped in Figure 4. Popular snorkeling destinations such as Kā'anapali Beach have high
 437 snorkeling attractiveness, as expected, due to low exposure, sediment, and cesspool effluent,

438 and good ecological quality. But not all popular beaches score high. For instance Honolulu Bay
439 has a lower than expected score, explained by high sediment, exposure, and crowding, which
440 reduce its attractiveness, despite low cesspool discharge, high fish richness and abundance,
441 and high probability of viewing turtles.

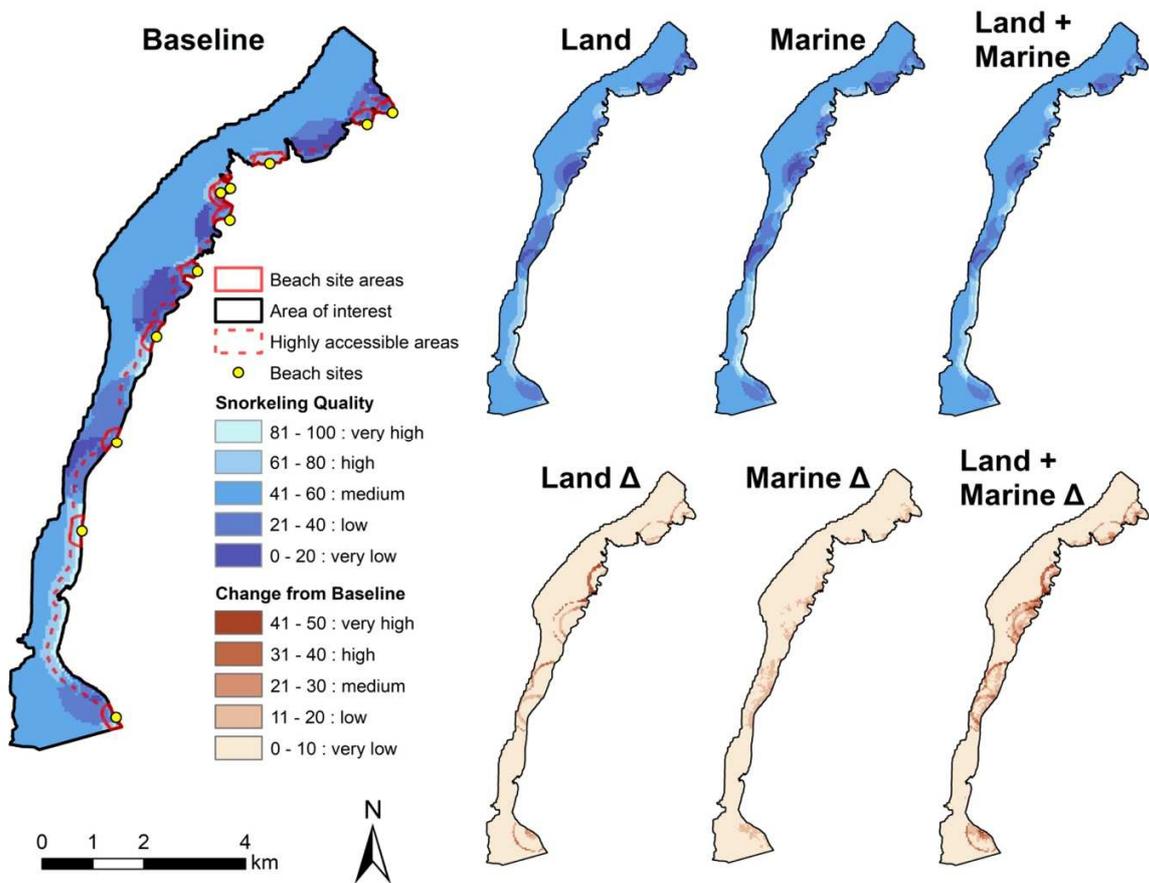
442

443 Using the BBN to estimate the effects of 20 management scenarios on recreation for the entire
444 AOI and a subsetted area of high and moderate accessibility, we found that improving local
445 water quality through controlling sediment and cesspool effluent and enhancing coral reef
446 conditions (i.e., coral cover, fish abundance, fish diversity as “combined marine”) positively
447 affected snorkeling attractiveness across our study AOI (Figure 4; Table SI_T5). Reducing
448 sediment alone had stronger effects on overall attractiveness than cesspool-related pollution
449 reductions. Increasing fish abundance had the strongest effects on snorkeling quality of all
450 ocean-related actions, while combined marine management (coral, fish abundance, and fish
451 richness improvements) resulted in slightly larger quality improvements than combined land
452 management (sediment and cesspool pollution reduction). Results of coral reef restoration
453 scenarios cannot be evaluated independently, as fish abundance and richness estimates are
454 directly tied to coral cover improvements, though we present the 12 decomposed results in
455 Figure 5 to illustrate the relative benefits. The greatest improvements across the entire AOI and
456 the accessible areas came from combining both land- and marine-based management.

457

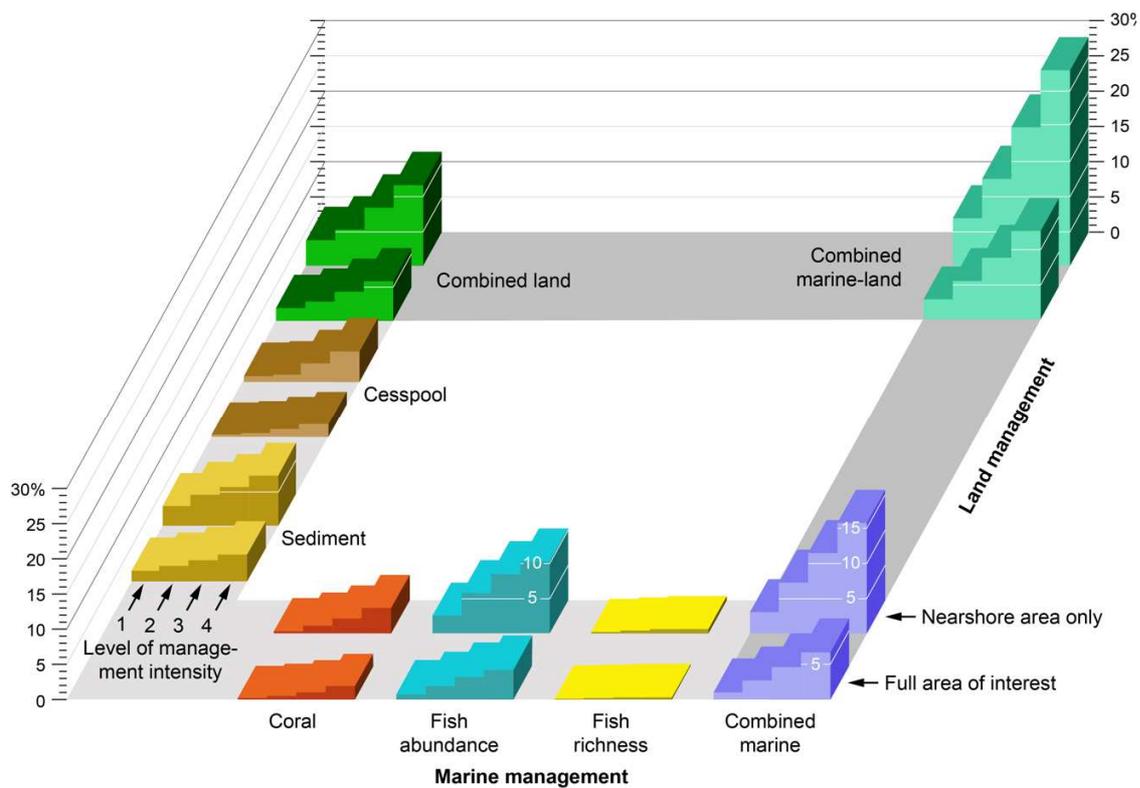
458 Results of land-based scenarios suggest that sediment reductions have the most value to
459 people, more so than cesspool effluent reductions. Reducing sediment by 25% - the highest-
460 level erosion reduction scenario - improved the recreational value more than completely
461 removing cesspools (7.1% vs. 4.3% improvement in the snorkeling attractiveness score for the
462 highly and moderately accessible areas). A coordinated effort to control both sediment and
463 cesspool effluent at the highest levels can improve the value by 11.4% in accessible areas.

464 Increasing coral cover to baseline plus 20%, fish abundance to 263% of baseline, and richness
 465 by 158% of baseline in a combined strategy would increase snorkeling quality by 15.7% in
 466 accessible areas. Combining all land and marine-based management activities at the highest
 467 levels resulted in a 27.7% improvement in snorkeling quality in more accessible areas, 15.7%
 468 from marine management and 11.5% from land management.
 469



470
 471 *Figure 4 Baseline snorkeling quality at current conditions (initial data inputs), binned as 0-20 very low; 21-40 low; 41-*
 472 *60 moderate; 61-80 high; and 81—100 very high (left). State of (top right) and change from baseline (Δ, bottom right)*
 473 *in snorkeling quality based on highest level of land, marine, and combined management. Area of interest, high-*
 474 *moderate access area, and beach site areas depicted. Beach sites indicated by yellow dots and numbers (see beach*
 475 *names in Table 6).*

476



477
 478 *Figure 5 Improvement in snorkeling quality by management action/combination. Results show improvements across*
 479 *the entire area of interest (AOI) in front blocks, and nearshore areas with high to moderate accessibility in back*
 480 *blocks. The sequence of four sets of bars for each management action shows progressively greater improvements for*
 481 *that activity, as described in the methods and Supplemental Information.*

482 Zooming in on popular local beaches illustrates how site-specific conditions determine the
 483 effects of management outcomes within the most accessible areas around those beaches.
 484 While results across the entire AOI and the most accessible areas suggest that reducing
 485 sediment is more impactful than cesspool-related action (Figure 5), this is not always true when
 486 we look at the area around popular beaches individually (Figure 4). The current recreation value
 487 of each beach area, along with results for five of the management scenarios with the largest
 488 improvements in outcomes are summarized in Table 6 for the high-access areas within 300m
 489 around eleven key beaches (see Table SI_T6 for details and Figure SI_F5A-C for maps). In
 490 some beaches, reducing cesspool effluent has more value than reducing sediment, and in

491 others, land management has no effect on recreation. As expected from the overall results,
 492 marine management has the highest outcomes for the majority of examined beaches, higher
 493 even than both land management actions together.

494

495 *Table 6 Snorkeling attractiveness score in highly accessible areas around each beach (listed in order north to south)*
 496 *under baseline conditions, and relative improvements due to high-impact management scenarios: 1. reduce sediment*
 497 *by 25%; 2. eliminate cesspools; 3. do both ["Land"]; 4. improve coral cover to baseline + 20%, fish abundance to*
 498 *263% of baseline, and fish species richness to 158% of baseline ["Marine"]; and 5. do both "Land" and "Marine"*
 499 *simultaneously ["Combined"].*

Beach site #	Beach name	Baseline snorkeling attractiveness score	Snorkeling attractiveness score improvement due to management scenario				
			Sediment	Cesspool	Land	Marine	Combine d
1	Honolua Bay	25.5	1.1	0.0	1.1	7.2	8.3
2	Mokulē'ia Beach	32.5	0.0	0.0	0.0	3.7	3.7
3	Oneloa Bay	66.2	3.3	0.0	3.3	11.3	14.9
4	Hanaka'ō'ō Beach	75.4	3.1	4.2	6.4	5.1	10.8
5	Kapalua Beach	65.4	6.6	0.0	6.6	13.9	20.7
6	Nāpili Bay	36.3	6.9	5.0	10.9	4.3	14.9
7	Keonenui	36.9	6.8	11.1	17.6	16.2	33.3
8	Kahana Beach	39.7	3.7	0.0	3.7	6.3	9.1
9	Honokōwai Beach Park	34.6	0.0	1.3	1.3	7.4	8.8
10	Kā'anapali Beach	78.8	6.0	3.7	9.7	10.9	20.8
11	Wahikuli State Wayside Park	57.0	0.0	10.3	10.3	14.9	26.5

500 8 Discussion

501 *Management implications*

502 State agencies charged with protecting the environment often focus on ecological outcomes, but
 503 the ecosystem services approach used here translates ecological conditions into terms more
 504 relatable to decision makers, visitors, and residents by tying them to human wellbeing and

505 preferences (Tallis and Polasky, 2009; Wainger and Mazzotta, 2011; Wainger and Boyd, 2009).
506 In an era of increasingly scarce management resources and compounding threats, it is all the
507 more important to ensure that management has net benefits. Hawai'i's economy and the
508 Hawaiian lifestyle are tightly linked to ocean recreation, and people have positive willingness to
509 pay for improvements to coastal amenities (Peng et al., 2017; Penn et al., 2016, 2014). Our
510 results underscore and add to the current trend integrating science and management across the
511 land-marine interface to address stressors to the ocean more holistically (Alvarez-Romero et al.,
512 2011; Halpern et al., 2009; Pressey et al., 2007; Tallis et al., 2008; Toft et al., 2013) and
513 efficiently (Klein et al., 2010). We introduce the human dimension to this trend: the benefits of
514 integrated management also apply to maximizing returns to society through recreational
515 ecosystem services.

516

517 Our approach identifies and prioritizes the many opportunities to conserve, improve, and restore
518 recreation quality along West Maui's coast, including which actions yield the greatest
519 improvements in snorkeling attractiveness and where these benefits will occur. Combined
520 efforts to address land and marine problems achieve the best outcomes overall and for most
521 beaches (Figure 5, Table 6). This aligns with recent studies in Hawai'i that have shown that
522 addressing just one or the other (i.e., either land- or marine-based) stressors leads to sub-
523 optimal ecological outcomes, and may even threaten ecological regime shifts (Jouffray et al.,
524 2019; Weijerman et al., 2018). Focusing on particular beaches adds specificity to our
525 management recommendations, highlighting the crucial need for tools to be applied at an
526 appropriate scale. Guided by the broader scale analysis, management recommendations for
527 West Maui as a whole are different than those coming from the local scale analysis. For
528 instance, at some of the beaches, controlling effluent from cesspools would be more impactful
529 than mitigating sediment (Table 6). Fortunately, recent evidence suggests that many of
530 cesspools in West Maui were upgraded by homeowners over the ensuing years since the data

531 were collected (Barnes et al., 2019), but the importance of effluent for recreational quality, and
532 the link between wastewater and coral degradation (Wear and Thurber, 2015), raises the need
533 for future analysis to also consider the effects of various wastewater treatment plants along the
534 coast.

535

536 While the best results will generally come from integrated management, it is notable that marine
537 management had higher payoffs overall than land management (Figure 5), driven by strong
538 preferences for improvements in the various marine attributes, but mainly the modeled
539 improvements in fish abundance (Table 2). The fact that fish abundance can greatly improve the
540 delivery of recreational ecosystem services may help coastal managers, who face challenges
541 managing for coral cover, given bleaching and other hard to mitigate threats, while the tools to
542 manage fishes can be easier to implement. Further, in many places, the jurisdiction of a
543 resource management agency may not cover both land and sea, as in the case of Hawai'i,
544 where the Division of Aquatic Resources has jurisdiction over fisheries but not watershed and
545 land management, which is the responsibility of other divisions within the Department of Land
546 and Natural Resources, as well as other government departments, and water quality is the
547 purview of the Department of Health.

548

549 *Cost benefit analysis*

550 The benefits of the various management actions should ideally be weighed against their costs to
551 determine whether action is justified, and which approaches are the most cost-effective. These
552 benefits likely extend well beyond the recreational benefits measured here, and a full cost-
553 benefit analysis would need to consider all social costs and benefits related to a given
554 management action (De Groot et al., 2013). Other studies have valued the benefits of reef
555 restoration in the U.S. (Brander and Beukering, 2013) and the total economic value of reefs in
556 Hawai'i (Bishop et al., 2011; Cesar and van Beukering, 2004), reporting U.S. household

557 willingness to pay values of ~\$64 per year for restoring 5 acres of reef and \$225 per year for
558 expanding marine protected areas to 25% of Hawaiian reefs. However, a full social cost-benefit
559 analysis is not always required to suggest the need for and direction of action. Our results show
560 strong preferences for improving ecosystem services. Given the scale of recreational use in
561 Hawai'i and the general allure of Hawaiian reefs, willingness to pay is likely more than sufficient
562 to justify taking action. Nevertheless, we do not attempt to estimate the magnitude of social
563 benefit from improved coastal and watershed conditions. Doing so would require the
564 quantification of diverse benefits associated with, for example, improved fisheries, other types of
565 recreation, cultural and spiritual values, shoreline protection, and existence and bequest values.

566

567 We also do not estimate the costs of management, in part because we only present generic
568 categories of land-based and marine-based management. The available tool box for land and
569 marine management is large, with variable ecological effectiveness and implementation costs.
570 Associated costs include many components, including land acquisition, implementation,
571 management, and opportunity costs (Naidoo et al., 2006). Cesspool upgrades in the area could
572 cost millions of dollars and vary depending on site characteristics such as slope and soil type,
573 while sediment reduction efforts could entail tens of millions of dollars of land restoration and
574 infrastructure investments that vary by watershed (Barnes et al., 2019; Group 70, 2015b, 2015a;
575 Powell et al., 2017; Sustainable Resources Group International, 2012a, 2012b). Fisheries
576 management could have high enforcement expenses and opportunity costs for fishers and
577 related businesses. Spatially explicit cost estimates to couple with the ecosystem services
578 benefits modeled here would help decision-makers prioritize the most cost-effective actions and
579 locations.

580

581 *Modeling innovations and limitations*

582 Our efforts contribute to an ongoing research program to evaluate ecosystem services spatially
583 through time using big data techniques and artificial intelligence to inform management (Villa et
584 al., 2014). An increasing number of tools use BBNs in ecosystem services modeling, including
585 plug-ins to GIS (Landuyt et al., 2015) and stand-alone modeling platforms like ARIES, used
586 here (Villa et al., 2014). Our innovation of linking an economic elicitation method to inform the
587 BBN provides additional rigor to the model structure and parameterization. Specifically, we
588 embedded the results of a choice experiment along with an expert elicitation into the BBN's
589 structure and conditional probability tables. This enabled us to model how recreational
590 attractiveness changes with improvements in specific, interrelated conditions. We grounded our
591 management scenarios by eliciting reasonable outcomes for sediment and cesspool reduction
592 and coral reef restoration from land and reef managers, and building an ecological model,
593 based on a Hawaiian archipelago-wide dataset, to evaluate how fish conditions would change
594 given improvements in coral cover.

595

596 The approach has some limitations. Preferences elicited from the choice experiment helped
597 inform the conditional probabilities in the BBN. Because our experimental design did not include
598 an outside option (Johnston et al., 2017) we prefer to interpret the parameters as measures of
599 the relative importance of the attributes, rather than deriving willingness to pay results. Our
600 survey sample likely underrepresented residents and younger snorkelers, although no
601 demographics exist to compare. If managers are interested in examining how different
602 management scenarios would affect different groups (e.g., tourists vs. residents; younger vs.
603 older), then a broader survey could be conducted to build conditional probabilities (and perhaps
604 alternate BBN structures) for these groups. Within a BBN's structure, intermediate nodes can
605 temper or enhance the strength of influence of any given parent node on a subsequent node.
606 For instance, in the choice experiment, snorkelers preferred fish abundance and fish species
607 richness about the same, but in the end, fish abundance had much greater effect on overall

608 snorkeler quality. Examining the arrows in Figure 3 that represent the strength of influence (also
609 Table SI_4b, fish species richness has a strong influence on the biodiversity intermediate node,
610 but the biodiversity node's smaller contribution to the ecological quality diminishes the
611 contribution of fish species richness to the overall snorkeling quality. Intermediate nodes are
612 important for keeping conditional probability tables tractable, but they can have side effects of
613 amplifying or diminishing the importance of other variables. The aim is that the combined
614 structure and conditional probabilities are a faithful representation of the system; validation is
615 important for ensuring this (Marcot et al., 2006). While we used expert opinion and our own
616 intuition to validate and test assumptions of the model based on the chosen conditional
617 probabilities, new capabilities within ARIES for BBN structural learning algorithms would be a
618 useful, additional step (Willcock et al., 2018).

619 9 Conclusion

620 Natural resource managers need to know how potential management strategies are likely to
621 impact people's wellbeing. Ecological-economic models such as the one developed here can
622 help managers choose what actions to take where, based on the outcome's societal value. For
623 recreational ecosystem services, the use of a BBN to combine survey-based data of the relative
624 value of important environmental and socioeconomic features with expert opinion and spatial
625 modeling to enable scenario analysis can provide a new path forward for integrating social and
626 natural science with management. Such integrated modeling of coupled nature-human systems
627 can benefit the management of recreational resources, particularly in settings with complex
628 combinations of stressors and human uses, such as recreation and management at the land-
629 sea interface.

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640 11 Citations

641 Aguilera, P.A., Fernández, A., Fernández, R., 2011. Bayesian networks in environmental
642 modelling. *Environmental Modelling & Software* 26, 1376–1388.

643 Ahmed, M., Umali, G.M., Chong, C.K., Rull, M.F., Garcia, M.C., 2007. Valuing recreational and
644 conservation benefits of coral reefs - The case of Bolinao, Philippines. *Ocean & Coastal*
645 *Management* 50, 103–118. <https://doi.org/10.1016/j.ocecoaman.2006.08.010>

646 Alvarez-Romero, J.G., Pressey, R.L., Ban, N.C., Vance-Borland, K., Willer, C., Klein, C.J.,
647 Gaines, S.D., 2011. Integrated land-sea conservation planning: the missing links. *Annual*
648 *Review of Ecology, Evolution, and Systematics* 42, 381–409.

649 Ames, D.P., Neilson, B.T., Stevens, D.K., Lall, U., 2005. Using Bayesian networks to model
650 watershed management decisions: an East Canyon Creek case study. *Journal of*
651 *Hydroinformatics* | 07, 267.

652 Anthony, K.R.N., Marshall, P.A., Abdulla, A., Beeden, R., Bergh, C., Black, R., Eakin, C.M.,
653 Game, E.T., Gooch, M., Graham, N.A.J., 2015. Operationalizing resilience for adaptive

654 coral reef management under global environmental change. *Global Change Biology* 21,
655 48–61.

656 Ariza, E., Ballester, R., Rigall-I-Torrent, R., Saló, A., Roca, E., Villares, M., Jiménez, J.A.,
657 Sardá, R., 2012. On the relationship between quality, users' perception and economic
658 valuation in NW Mediterranean beaches. *Ocean & coastal management* 63, 55–66.

659 Ban, S.S., Graham, N.A.J., Connolly, S.R., 2014. Evidence for multiple stressor interactions and
660 effects on coral reefs. *Global Change Biology* 20, 681–697.
661 <https://doi.org/10.1111/gcb.12453>

662 Barnes, M.D., Goodell, W., Whittier, R., Falinski, K.A., Callender, T., Htun, H., LeViol, C., Slay,
663 H., Oleson, K.L.L., 2019. Decision analysis to support wastewater management in coral
664 reef priority area. *Marine Pollution Bulletin* 148, 16–29.
665 <https://doi.org/10.1016/j.marpolbul.2019.07.045>

666 Beharry-Borg, N., Scarpa, R., 2010. Valuing quality changes in Caribbean coastal waters for
667 heterogeneous beach visitors. *Ecological Economics* 69, 1124–1139.

668 Bishop, R.C., Chapman, D.J., Kanninen, B.J., Krosnick, J.A., Leeworthy, V.R., Meade, N.F.,
669 2011. Total economic value for protecting and restoring Hawaiian coral reef ecosystems:
670 Final report. US Department of Commerce, National Oceanic and Atmospheric
671 Administration, National Ocean Service.

672 Brander, L., Beukering, P. van, 2013. The total economic value of US coral reefs: a review of
673 the literature.

674 Brander, L.M., Van Beukering, P., Cesar, H.S., 2007. The recreational value of coral reefs: a
675 meta-analysis. *Ecological Economics* 63, 209–218.

676 Brown, G., Roughgarden, J., 1997. A metapopulation model with private property and a
677 common pool. *Ecological Economics* 22, 65–71.

678 Carr, L., Mendelsohn, R., 2003. Valuing coral reefs: A travel cost analysis of the Great Barrier
679 Reef. *Ambio* 32, 353–357. [https://doi.org/10.1639/0044-](https://doi.org/10.1639/0044-7447(2003)032[0353:vcratc]2.0.co;2)
680 [7447\(2003\)032\[0353:vcratc\]2.0.co;2](https://doi.org/10.1639/0044-7447(2003)032[0353:vcratc]2.0.co;2)

681 Cesar, H.S., van Beukering, P., 2004. Economic valuation of the coral reefs of Hawai'i. *Pacific*
682 *Science* 58, 231–242.

683 Cooper, E., Burke, L.M., Bood, N.D., 2009. Coastal capital, Belize: The economic contribution of
684 Belize's coral reefs and mangroves. World Resources Institute.

685 Darling, E.S., Coté, I.M., 2008. Quantifying the evidence for ecological synergies. *Ecology*
686 *Letters* 11, 1278–1286. <https://doi.org/10.1111/j.1461-0248.2008.01243.x>

687 De Groot, R.S., Blignaut, J., Van Der Ploeg, S., Aronson, J., Elmqvist, T., Farley, J., 2013.
688 Benefits of investing in ecosystem restoration. *Conservation Biology* 27, 1286–1293.
689 <https://doi.org/10.1111/cobi.12158>

690 Dee, L.E., Allesina, S., Bonn, A., Eklöf, A., Gaines, S.D., Hines, J., Jacob, U., McDonald-
691 Madden, E., Possingham, H., Schröter, M., 2017. Operationalizing network theory for
692 ecosystem service assessments. *Trends in Ecology & Evolution* 32, 118–130.

693 Fezzi, C., Bateman, I.J., Ferrini, S., 2014. Using revealed preferences to estimate the value of
694 travel time to recreation sites. *Journal of Environmental Economics and Management*
695 67, 58–70. <https://doi.org/10.1016/j.jeem.2013.10.003>

696 Forio, M.A.E., Landuyt, D., Bennetsen, E., Lock, K., Nguyen, T.H.T., Ambarita, M.N.D.,
697 Musonge, P.L.S., Boets, P., Everaert, G., Dominguez-Granda, L., 2015. Bayesian belief
698 network models to analyse and predict ecological water quality in rivers. *Ecological*
699 *Modelling* 312, 222–238.

700 Franco, C., Hepburn, L.A., Smith, D.J., Nimrod, S., Tucker, A., 2016. A Bayesian Belief Network
701 to assess rate of changes in coral reef ecosystems. *Environmental modelling & software*
702 80, 132–142.

703 Friedlander, A., Aeby, G., Brown, E., Clark, A., Coles, S., Dollar, S., Hunter, C., Jokiell, P.,
704 Smith, J., Walsh, B., others, 2005. The state of coral reef ecosystems of the main
705 Hawaiian Islands. The state of coral reef ecosystems of the United States and Pacific
706 freely associated states 222–269.

707 Friedlander, A., Kendall, M., 2006. Fishes - Reef Fish, in: Costa, B., Kendall, M. (Eds.), Marine
708 Biogeographic Assessment of the Main Hawaiian Islands, OCS Study BOEM 2016-035
709 and NOAA Technical Memorandum NOS NCCOS 214. Bureau of Ocean Energy
710 Management National Oceanic and Atmospheric Administration, pp. 156–196.

711 Ghermandi, A., Nunes, P.A.L.D., 2013. A global map of coastal recreation values: Results from
712 a spatially explicit meta-analysis. *Ecological Economics* 86, 1–15.

713 Gonzalez-Redin, J., Luque, S., Poggio, L., Smith, R., Gimona, A., 2016. Spatial Bayesian belief
714 networks as a planning decision tool for mapping ecosystem services trade-offs on
715 forested landscapes. *Environmental Research* 144, 15–26.

716 Gorospe, K.D., Donahue, M.J., Heenan, A., Gove, J.M., Williams, I.D., Brainard, R.E., 2018.
717 Local biomass baselines and the recovery potential for Hawaiian coral reef fish
718 communities. *Frontiers in Marine Science* 5, 1–13.

719 Grafeld, S., Oleson, K., Barnes, M., Peng, M., Chan, C., Weijerman, M., 2016. Divers’
720 willingness to pay for improved coral reef conditions in Guam: An untapped source of
721 funding for management and conservation? *Ecological Economics* 128, 202–213.

722 Graham, N.A.J., McClanahan, T.R., MacNeil, M.A., Wilson, S.K., Polunin, N.V.C., Jennings, S.,
723 Chabanet, P., Clark, S., Spalding, M.D., Letourneur, Y., 2008. Climate warming, marine
724 protected areas and the ocean-scale integrity of coral reef ecosystems. *PLoS One* 3,
725 e3039.

726 Group 70, 2015a. West Maui watershed plan: Kahana, Honokahua, and Honolua watersheds,
727 Strategies and Implementation Report.

728 Group 70, 2015b. Kahana, Honokahua and Honolua watersheds characterization report.

729 Halpern, B.S., Ebert, C.M., Kappel, C.V., Madin, E.M.P., Micheli, F., Perry, M., Selkoe, K.A.,
730 Walbridge, S., 2009. Global priority areas for incorporating land-sea connections in
731 marine conservation. *Conservation Letters* 1–8. [https://doi.org/10.1111/j.1755-](https://doi.org/10.1111/j.1755-263X.2009.00060.x)
732 [263X.2009.00060.x](https://doi.org/10.1111/j.1755-263X.2009.00060.x)

733 Hawai'i Department of Health, 2019. Hawai'i Clean Water Branch (CWB) Beach Water Quality
734 Data.

735 Hawai'i Mapping Research Group, 2016. 5 Meter Bathymetry Synthesis Grid.

736 Hawai'i Tourism Authority, 2016. Fact Sheet: Benefits of Hawaii's Tourism Economy. State of
737 Hawaii, Honolulu, HI.

738 Hoegh-Guldberg, O., 1999. Climate change, coral bleaching and the future of the world's coral
739 reefs. *Marine and Freshwater Research* 50, 839–866. <https://doi.org/10.1071/mf99078>

740 Hughes, T.P., Graham, N.A.J., Jackson, J.B.C., Mumby, P.J., Steneck, R.S., 2010. Rising to the
741 challenge of sustaining coral reef resilience. *Trends in Ecology & Evolution* 25, 633–642.
742 <https://doi.org/10.1016/j.tree.2010.07.011>

743 Hughes, T.P., Rodrigues, M.J., Bellwood, D.R., Ceccarelli, D., Hoegh-Guldberg, O., McCook, L.,
744 Moltschaniwskyj, N., Pratchett, M.S., Steneck, R.S., Willis, B., 2007. Phase shifts,
745 herbivory, and the resilience of coral reefs to climate change. *Current Biology* 17, 360–5.
746 <https://doi.org/10.1016/j.cub.2006.12.049>

747 Inglis, G.J., Johnson, V.I., Ponte, F., 1999. Crowding norms in marine settings: A case study of
748 snorkeling on the Great Barrier Reef. *Environmental Management* 24, 369–381.

749 Jackson, J.B.C., Kirby, M.X., Berger, W.H., Bjorndal, K.A., Botsford, L.W., Bourque, B.J.,
750 Bradbury, R.H., Cooke, R., Erlandson, J., Estes, J.A., 2001. Historical overfishing and
751 the recent collapse of coastal ecosystems. *Science* 293, 629–637.

752 Johnston, R.J., Boyle, K.J., Adamowicz, W., Bennett, J., Brouwer, R., Cameron, T.A.,
753 Hanemann, W.M., Hanley, N., Ryan, M., Scarpa, R., 2017. Contemporary guidance for

754 stated preference studies. *Journal of the Association of Environmental and Resource*
755 *Economists* 4, 319–405.

756 Jouffray, J.-B., Wedding, L.M., Norström, A.V., Donovan, M.K., Williams, G.J., Crowder, L.B.,
757 Erickson, A.L., Friedlander, A.M., Graham, N.A.J., Gove, J.M., 2019. Parsing human and
758 biophysical drivers of coral reef regimes. *Proceedings of the Royal Society B* 286,
759 20182544.

760 Klein, C.J., Ban, N.C., Halpern, B.S., Begger, M., Game, E.T., Grantham, H.S., Green, A., Klein,
761 T.J., Kininmonth, S., Treml, E., 2010. Prioritizing land and sea conservation investments
762 to protect coral reefs. *PLoS One* 5, e12431.

763 Koiter, J., 2006. Visualizing inference in Bayesian networks. Faculty of Electrical Engineering,
764 Mathematics, and Computer Science, Department of Man-Machine Interaction, Delft
765 University of Technology.

766 Kuhnert, P.M., Martin, T.G., Griffiths, S.P., 2010. A guide to eliciting and using expert
767 knowledge in Bayesian ecological models. *Ecology Letters* 13, 900–914.

768 Landuyt, D., Broekx, S., D'Hondt, R., Engelen, G., Aertsens, J., Goethals, P.L.M., 2013. A
769 review of Bayesian belief networks in ecosystem service modelling. *Environmental*
770 *Modelling & Software* 46, 1–11.

771 Landuyt, D., Van der Biest, K., Broekx, S., Staes, J., Meire, P., Goethals, P.L., 2015. A GIS
772 plug-in for Bayesian belief networks: towards a transparent software framework to
773 assess and visualise uncertainties in ecosystem service mapping. *Environmental*
774 *Modelling & Software* 71, 30–38.

775 Loomis, J., Santiago, L., 2013. Economic valuation of beach quality improvements: comparing
776 incremental attribute values estimated from two stated preference valuation methods.
777 *Coastal Management* 41, 75–86.

778 Marcot, B.G., Steventon, J.D., Sutherland, G.D., McCann, R.K., 2006. Guidelines for developing
779 and updating Bayesian belief networks applied to ecological modeling and conservation.
780 *Can. J. For. Res.* 36, 3063–3074. <https://doi.org/10.1139/x06-135>

781 McFadden, D., 1974. Conditional Logit Analysis of Qualitative Choice Behaviour, in: P.
782 Zarembka (Ed.), *Frontiers in Econometrics*. Academic Press, New York.

783 Moberg, F., Folke, C., 1999. Ecological goods and services of coral reef ecosystems. *Ecological*
784 *economics* 29, 215–233.

785 Mumby, P.J., Steneck, R.S., 2008. Coral reef management and conservation in light of rapidly
786 evolving ecological paradigms. *Trends in Ecology & Evolution* 23, 555–563.
787 <https://doi.org/10.1016/j.tree.2008.06.011>

788 Naidoo, R., Balmford, A., Ferraro, P.J., Polasky, S., Ricketts, T.H., Rouget, M., 2006. Integrating
789 economic costs into conservation planning. *Trends in ecology & evolution* 21, 681–687.

790 National Centers for Coastal Ocean Science, 2007. Benthic Habitat Map - Main Hawaiian
791 Islands.

792 Nunes, P.A.L.D., Loureiro, M.L., Piñol, L., Sastre, S., Voltaire, L., Canepa, A., 2015. Analyzing
793 beach recreationists' preferences for the reduction of jellyfish blooms: Economic results
794 from a stated-choice experiment in Catalonia, Spain. *PLoS One* 10, e0126681.

795 Nyberg, J.B., Marcot, B.G., Sulyma, R., 2006. Using Bayesian belief networks in adaptive
796 management. *Canadian Journal of Forest Research* 36, 3104–3116.
797 <https://doi.org/10.1139/x06-108>

798 Nyström, M., Graham, N.A., J., Lokrantz, J., Norström, A.V., 2008. Capturing the cornerstones
799 of coral reef resilience: linking theory to practice 27, 795–809.

800 Oleson, K.L.L., Falinski, K.A., Lecky, J., Rowe, C., Kappel, C.V., Selkoe, K.A., White, C., 2017.
801 Upstream solutions to coral reef conservation: The payoffs of smart and cooperative
802 decision-making. *Journal of Environmental Management* 191, 8–18.
803 <https://doi.org/10.1016/j.jenvman.2016.12.067>

804 Oleson, K.L.L., Grafeld, S., Van Beukering, P., Brander, L., James, P.A.S., Wolfs, E., 2018.
805 Charting progress towards system-scale ecosystem service valuation in islands.
806 Environmental Conservation 45, 212–226.

807 Pacific Islands Fisheries Science Center, 2019. National Coral Reef Monitoring Program:
808 Stratified random surveys (StRS) of reef fish, including benthic estimate data of the
809 Hawaiian Archipelago since 2013.

810 Parsons, G.R., Thur, S.M., 2008. Valuing changes in the quality of coral reef ecosystems: a
811 stated preference study of SCUBA diving in the Bonaire National Marine Park.
812 Environmental and Resource Economics 40, 593–608.

813 Pendleton, L.H., 1995. Valuing coral reef protection. Ocean & Coastal Management 26, 119–
814 131.

815 Pendleton, L.H., 1994. Environmental quality and recreation demand in a Caribbean coral reef.
816 Coastal Management 22, 399–404.

817 Peng, M., Oleson, K.L., others, 2017. Beach Recreationalists' Willingness to Pay and Economic
818 Implications of Coastal Water Quality Problems in Hawaii. Ecological Economics 136,
819 41–52.

820 Penn, J., Hu, W., Cox, L., Kozloff, L., 2016. Values for recreational beach quality in O'ahu,
821 Hawai'i. Marine Resource Economics 31, 47–62. <https://doi.org/10.1086/683795>

822 Penn, J., Hu, W., Cox, L., Kozloff, L., 2014. Resident and tourist preferences for stormwater
823 management strategies in O'ahu, Hawai'i. Ocean & Coastal Management 98, 79–85.
824 <https://doi.org/10.1016/j.ocecoaman.2014.06.002>

825 Petrosillo, I., Zurlini, G., Corliano, M.E., Zaccarelli, N., Dadamo, M., 2007. Tourist perception of
826 recreational environment and management in a marine protected area. Landscape and
827 Urban Planning 79, 29–37.

828 Powell, K.B., Ellsworth, L.M., Litton, C.M., Oleson, K.L., Ammond, S.A., 2017. Toward Cost-
829 Effective Restoration: Scaling up Restoration in Ecosystems Degraded by Nonnative
830 Invasive Grass and Ungulates 1. *Pacific Science* 71, 479–493.

831 Pratchett, M.S., Munday, P.L., Wilson, S.K., Graham, N.A.J., Cinner, J.E., Bellwood, D.R.,
832 Jones, G.P., Polunin, N.V.C., McClanahan, T.R., 2008. Effects of climate-induced coral
833 bleaching on coral-reef fishes - Ecological and economic consequences. *Oceanography*
834 *and Marine Biology: an Annual Review*, Vol 46 46, 251–296.
835 <https://doi.org/10.1201/9781420065756.ch6>

836 Pressey, R.L., Cabeza, M., Watts, M.E., Cowling, R.M., Wilson, K.A., 2007. Conservation
837 planning in a changing world. *Trends in Ecology & Evolution* 22, 583–592.
838 <https://doi.org/10.1016/j.tree.2007.10.001>

839 Principe, P.P., Bradley, P., Yee, S.H., Fisher, W.S., Johnson, E.D., Allen, P., Campbell, D.E.,
840 2012. Quantifying coral reef ecosystem services. U.S. Environmental Protection Agency,
841 Washington, DC.

842 Ruiz-Frau, A., Hinz, H., Edwards-Jones, G., Kaiser, M.J., 2013. Spatially explicit economic
843 assessment of cultural ecosystem services: Non-extractive recreational uses of the
844 coastal environment related to marine biodiversity. *Marine Policy* 38, 90–98.

845 Schuhmann, P.W., Casey, J.F., Horrocks, J.A., Oxenford, H.A., 2013. Recreational SCUBA
846 divers' willingness to pay for marine biodiversity in Barbados. *Journal of Environmental*
847 *Management* 121, 29–36.

848 Spalding, M., Burke, L., Wood, S.A., Ashpole, J., Hutchison, J., zu Ermgassen, P., 2017.
849 Mapping the global value and distribution of coral reef tourism. *Marine Policy* 82, 104–
850 113.

851 Sparks, R., Stone, K., White, D., Ross, M., 2015. Maui and Lānaʻi monitoring report. Hawaiʻi
852 Division of Aquatic Resources, Maui Office, Wailuku, HI.

853 Spence, P.L., Jordan, S.J., 2013. Effects of nitrogen inputs on freshwater wetland ecosystem
854 services—A Bayesian network analysis. *Journal of Environmental Management* 124, 91–
855 99.

856 Stock, J.D., Falinski, K.A., Callender, T., 2016. Reconnaissance Sediment Budget for Selected
857 Watersheds of West Maui, Hawai'i: U.S. Geological Survey Open-File Report 2015–
858 1190. United States Geological Survey, Reston, VA. <https://doi.org/10.3133/ofr20151190>

859 Sustainable Resources Group International, 2012a. Wahikuli-Honokōwai watershed
860 management plan volume 1: Watershed characterization, Ridge to Reef Initiative.

861 Sustainable Resources Group International, 2012b. Wahikuli-Honokōwai watershed
862 management plan volume 2: Strategies and implementation.

863 Tallis, H., Ferdana, Z., Gray, E., 2008. Linking terrestrial and marine conservation planning and
864 threats analysis. *Conservation Biology* 22, 120–130.

865 Tallis, H.T., Polasky, S., 2009. Mapping and valuing ecosystem services as an approach for
866 conservation and natural-resource management. *The Year in Ecology and Conservation*
867 *Biology* 1162, 265–283.

868 Toft, J.E., Burke, J.L., Carey, M.P., Kim, C.K., Marsik, M., Sutherland, D.A., Arkema, K.K.,
869 Guerry, A.D., Levin, P.S., Minello, T.J., 2013. From mountains to sound: modelling the
870 sensitivity of Dungeness crab and Pacific oyster to land–sea interactions in Hood Canal,
871 WA. *ICES Journal of Marine Science* 71, 725–738.

872 U.S. Census Bureau, 2017. QuickFacts Maui County, Hawai'i [WWW Document]. URL
873 <https://www.census.gov/quickfacts/fact/table/mauicountyhawaii/BZA210216> (accessed
874 6.19.19).

875 van Beukering, P., Cesar, H.S., 2004. Ecological economic modeling of coral reefs: Evaluating
876 tourist overuse at Hanauma Bay and algae blooms at the Kihei Coast, Hawai'i. *Pacific*
877 *Science* 58, 243–260.

878 van Riper, C.J., Kyle, G.T., Sutton, S.G., Barnes, M., Sherrouse, B.C., 2012. Mapping outdoor
879 recreationists' perceived social values for ecosystem services at Hinchinbrook Island
880 National Park, Australia. *Applied Geography* 35, 164–173.

881 Villa, F., Bagstad, K.J., Voigt, B., Johnson, G.W., Portela, R., Honzák, M., Batker, D., 2014. A
882 methodology for adaptable and robust ecosystem services assessment. *PLoS One* 9,
883 e91001. <https://doi.org/10.1371/journal.pone.0091001>

884 Wainger, L., Mazzotta, M., 2011. Realizing the potential of ecosystem services: a framework for
885 relating ecological changes to economic benefits. *Environmental Management*.

886 Wainger, L.A., Boyd, J.W., 2009. Valuing ecosystem services, in: McLeod, K.L., Leslie, H.M.
887 (Eds.), *Ecosystem-Based Management for the Oceans*. Island Press, Washington D.C.,
888 pp. 92–111.

889 Wear, S.L., Thurber, R.V., 2015. Sewage pollution: mitigation is key for coral reef stewardship.
890 *Annals of the New York Academy of Sciences* 1355, 15–30.

891 Wedding, L.M., Lecky, J., Gove, J.M., Walecka, H.R., Donovan, M.K., Williams, G.J., Jouffray,
892 J.-B., Crowder, L.B., Erickson, A., Falinski, K., 2018. Advancing the integration of spatial
893 data to map human and natural drivers on coral reefs. *PLoS One* 13, e0189792.

894 Weijerman, M., Veazey, L., Yee, S., Vaché, K., Delevaux, J., Donovan, M., Lecky, J., Oleson,
895 K.L.L., 2018. Managing local stressors for coral reef condition and ecosystem services
896 delivery under climate scenarios. *Frontiers in Marine Science* 5, 425.

897 Whittier, R.B., El-Kadi, A.I., 2014. Human health and environmental risk ranking of on-site
898 sewage disposal systems for the Hawaiian islands of Kaua'i, Moloka'i, Maui, and
899 Hawai'i. University of Hawaii.

900 Wielgus, J., Chadwick-Furman, N.E., Dubinsky, Z., Shechter, M., Zeitouni, N., 2002. Dose-
901 response modeling of recreationally important coral-reef attributes: a review and
902 potential application to the economic valuation of damage. *Coral Reefs* 21, 253–259.

903 Willcock, S., Martínez-López, J., Hooftman, D.A.P., Bagstad, K.J., Balbi, S., Marzo, A., Prato,
904 C., Sciandrello, S., Signorello, G., Voigt, B., Villa, F., Bullock, J.M., Athanasiadis, I.,
905 2018. Machine learning for ecosystem services. *Ecosystem services* 33, 165–174.

906 Williams, I.D., Walsh, W.J., Schroeder, R.E., Friedlander, A.M., Richards, B.L., Stamoulis, K.A.,
907 2008. Assessing the importance of fishing impacts on Hawaiian coral reef fish
908 assemblages along regional-scale human population gradients. *Environmental*
909 *Conservation* 35, 261–272.

910 Williams, I.D., White, D.J., Sparks, R.T., Lino, K.C., Zamzow, J.P., Kelly, E.L.A., Ramey, H.L.,
911 2016. Responses of herbivorous fishes and benthos to 6 Years of protection at the
912 Kahekili Herbivore Fisheries Management Area, Maui. *PLoS One* 11.
913 <https://doi.org/10.1371/journal.pone.0159100>

914 Wood, S.A., Guerry, A.D., Silver, J.M., Lacayo, M., 2013. Using social media to quantify nature-
915 based tourism and recreation. *Scientific Reports* 3. <https://doi.org/10.1038/srep02976>

916 Zhang, F., Wang, X.H., Nunes, P.A.L.D., Ma, C., 2015. The recreational value of gold coast
917 beaches, Australia: An application of the travel cost method. *Ecosystem Services* 11,
918 106–114.

919