

1 Data Weighting for Tagging Data in Integrated Size-Structured Models

2
3 André E. Punt^{1,2}, Roy A. Deng³, M.S.M. Siddeek⁴, Rik. C. Buckworth⁵, and Vicki Vanek⁶

4
5 ¹*School of Aquatic and Fishery Sciences, University of Washington, Seattle, WA 98195-5020, USA*

6 ²*CSIRO Oceans and Atmosphere, Castray Esplanade, Hobart, Tasmania, Australia*

7 ³*CSIRO Oceans and Atmosphere, Queensland Bioscience Precinct, 306 Carmody Road, St Lucia, QLD, 4072,*

8 *Australia*

9 ⁴*Alaska Department of Fish and Game, Division of Commercial Fisheries, 1255 W. 8th Street, Juneau, Alaska*

10 *99811, USA*

11 ⁵*CSIRO Oceans and Atmosphere, Tropical Ecosystems Research Centre, 564 Vanderlin Drive, Berrimah, NT,*

12 *0828 Australia*

13 ⁶*Alaska Department of Fish and Game, Division of Commercial Fisheries, 351 Research Court, Kodiak, AK*

14 *99615, USA*

15 16 Abstract

17 Increasingly, stock assessments for hard-to-age species such as crabs, prawns, rock lobsters,
18 and abalone are being based on integrated size-structured population dynamics models that
19 are fit to a variety of data sources. These data sources include tagging data to inform growth.
20 Diagnostic statistics and plots have been developed to explore how well integrated population
21 models fit the data types typically used for assessment purposes (index data, size- and age-
22 compositions, and conditional age-at-length data). However, such statistics and plots are not
23 available for tagging data, when these data are used to estimate growth. This paper outlines
24 two diagnostic statistics that can be used to evaluate fits to tagging data, and develops a
25 method based on ‘Francis weighting’ for weighting tagging data in integrated models. For
26 illustration, the methods are applied to Aleutian Islands golden king crab (*Lithodes*
27 *aequispinus*) in Alaska, and tiger prawns (*Penaeus semisulcatus* and *P. esculentus*) in
28 Australia’s Northern Prawn Fishery. Some degree of growth model mis-specification was
29 revealed for *P. semisulcatus*, and there were conflicts in the data for the tiger prawns. The
30 standard errors for the estimates of mature male biomass for golden king crab were larger
31 when the tagging data were downweighted based on the proposed weighting method. This
32 serves to emphasise that assessments and their interpretations can be impacted by how
33 tagging data are weighted.

34 Keywords: data weighting; size-structured stock assessment methods; tagging data

35 Corresponding author: A.E. Punt

36 a. Email: aepunt@uw.edu

37 b. Phone: 1-206-221-6319

38 c. Fax : 1-206-685-7471

39 40 Highlights

- 41 • Assessment results can be sensitive to the assumptions for data weighting.
- 42 • Tagging data may be overdispersed relative to the commonly assumed Bernoulli
43 likelihood.
- 44 • ‘Francis weighting’ can be extended to apply to tagging data.

47 1. Introduction

48 There is an increasing trend towards the use of integrated size-structured stock assessments
 49 for species that are difficult to age (Punt et al., 2013). For example, assessments for crab
 50 stocks off Alaska are based on size-structured population dynamics models that often divide
 51 the population into new and old shell crab (i.e. crab that did and did not moult the previous
 52 season; e.g. snow crab *Chionoecetes opilio*; Turnock and Rugulo, 2014; and red king crab
 53 *Paralithodes camtschaticus*; Zheng and Siddeek, 2014) while the assessment of tiger prawns
 54 (*Penaeus semisulcatus* and *P. esculentus*) in Australia's Northern Prawn Fishery (NPF) is
 55 based on a sex- and size-structured population dynamics model (Punt et al., 2010; Buckworth
 56 et al., 2015).

57 Integrated size-structured stock assessment methods make use of several sources of data.
 58 For example, assessments of golden king crab *Lithodes aequispinus* in the Aleutian Islands,
 59 Alaska, include data on landings in numbers, the size-composition of the landings, the size-
 60 composition of observer records for all crab arriving on deck, catch-rate indices for the
 61 retained component of the catch, and tagging data (Siddeek et al., in press). In contrast,
 62 assessments of tiger prawns in the NPF are based on weekly catch and effort data, survey
 63 indices of abundance, survey and commercial size-composition data, and tagging data.
 64 However, it is not uncommon for information in data sources to be in conflict with each other
 65 to some extent (e.g. Richards, 1991). Thus, each data type (and each data point within each
 66 data type) included in a stock assessment needs to be assigned a weight. In principle, this
 67 weight should relate to the deviation between the data point and its expected value (Punt, this
 68 volume), although on occasion weights reflect a subjective evaluation of the reliability of the
 69 data type (e.g. ICCAT, 2013). However, it is not straightforward to objectively select
 70 weights, and history reveals that data weighting can be influential on assessment results (e.g.,
 71 Richards, 1991).

72 The primary purpose of the tagging data in the assessments of Aleutian Islands golden
 73 king crab and of tiger prawns in the Northern Prawn Fishery is to allow growth (rather than
 74 fishing mortality) to be estimated. The component of the likelihood function for the tagging
 75 data (Punt et al., 2009) is:

$$76 \quad L = \prod_i \left(S_{C_i} [\mathbf{X}^{T_i}]_{R_i, C_i} / \sum_j S_j [\mathbf{X}^{T_i}]_{R_i, j} \right) = \prod_i p(C_i | R_i, T_i) \quad (1)$$

77 where T_i is the time-at-liberty for the i^{th} recapture, \mathbf{X} is the size-transition matrix (which
 78 specifies the probability of growing from one size-class to each of the same or larger size-
 79 classes), R_i is size-class in which the i^{th} recapture was when it was released, C_i is the size-
 80 class in which the i^{th} recapture was when it was recaptured, and S_j is the selectivity of an
 81 animal in size-class j (logistic for the example applications reported here). The size-transition
 82 matrix can separate the processes of moulting from those of growth given moult (e.g., Zheng
 83 and Siddeek, 2014) or represent the combined effects of moulting and growth given moult
 84 within a single model (e.g., Punt and Kenndy, 1997; Haist et al., 2009)). The form of the size
 85 transition matrix for the case in which moulting is modelled explicitly is:

$$86 \quad \mathbf{X} = \mathbf{X}'\mathbf{Q} + \mathbf{I}(\mathbf{I} - \mathbf{Q}) \quad (2)$$

87 where \mathbf{Q} is a diagonal matrix with values given by the probability of moulting, \mathbf{X}' is a
 88 matrix where each column is for a size before moult and each entry in each column is the
 89 probability of growing to that size given the size being represented by the column, and \mathbf{I} is
 90 the identity matrix.

Equation 1 treats each recapture as a Bernoulli trial, i.e. each tagged animal is treated as a single data point, independent from all the others. However, there will be overdispersion if tagging is such that some of the tagged animals are pseudoreplicates. This can happen if groups of tagged animals are released together and hence may have moved together and hence been subject to the same environmental conditions and prey fields. Consequently, the growth and probability of recapturing an animal are not independent of those for some of the other tagged animals. To account for overdispersion, the size-composition data used in the current configuration of the assessment methods for the two example fisheries are downweighted, but this is not currently the case for the tagging data. Accounting for this overdispersion requires that the right hand side of Equation 1 is raised to a power (equivalent to multiplying the logarithm of the right hand side of Equation 1 by an overdispersion factor). Several approaches (e.g., McAllister and Ianelli, 1997; Francis, 2011; Punt, this volume) have been developed to estimate overdispersion factors for size-composition data, and these approaches have been used to weight the size-composition data for Aleutian Islands golden king crab and tiger prawns in the NPF. However, methods have not been developed to explore whether the growth model is mis-specified, whether there is overdispersion, and how tagging data used in size-structured stock assessment methods should be weighted.

This paper provides diagnostic statistics for evaluating the fits to tag-recapture data within size-structured integrated assessment models and for estimating an overdispersion factor for weighting tagging data. The approach follows the spirit of the approach of Francis (2011). The proposed diagnostics and weighting factors are illustrated using Aleutian Islands golden king crab and *P. semisulcatus* and *P. esculentus* in the NPF. These two cases were selected because although the assessments are both based on size-structured models, that for Aleutian Islands golden king crab is male-only, has an annual time step, and considers 5 mm size-classes. In contrast, the assessments for *P. semisulcatus* and *P. esculentus* are based on a sex-structured model that has a weekly time-step and 1 mm size-classes.

2. Material and Methods

2.1 Diagnostic statistics

Two diagnostic statistics are considered. Both statistics are computed by time-at-liberty. The first diagnostic statistic is a comparison of frequencies of observed numbers recaptured by size-class versus the model-predicted distribution for size-classes at recapture. The latter distribution is:

$$\hat{P}_j = \sum_i p(j|k_i) \quad (3)$$

where \hat{P}_j is the expected number of recaptures in size-class j , and $p(j|k_i)$ is the probability that the i^{th} individual (which was released in size-class class k) was recaptured in size-class j (see Equation 1).

The second diagnostic statistic involves plotting the observed mean recapture size, \bar{P}_L^{obs} , versus release size-class L , along with the expected distributions of size-at-recapture, as a function of size-class-at-release, characterized by the expected (mean) size-at-recapture \hat{P}_L and the standard error of the observed mean size-at-recapture $SE[\hat{P}_L]$, i.e.:

$$\hat{P}_L = \sum_j \bar{L}_j p(j|L); \quad SE[\hat{P}_L] = \sqrt{\sum_j (\bar{L}_j - \hat{P}_L)^2 / N_L} \quad (4)$$

where \bar{L}_j is the mid-point of size-class j , and N_L is the number of releases of animals in size-class L .

134 2.2 Data weighting

135 Francis weighting (Francis, 2011) involves defining the overdispersion factor for catch size-
 136 composition data as the inverse of the variance of the standardized residuals for the mean size
 137 of the catch. By analogy, the weight W that should be assigned to the tagging composition of
 138 the likelihood is given by:

$$139 \quad W^{-1} = \text{var}[(\bar{P}_L^{obs} - \hat{P}_L) / \text{SE}[\hat{P}_L]] \quad (5)$$

140 In common with the diagnostic statistics, the weighting factors can be computed
 141 separately by time-at-liberty and by sex.

142 Data weighting would entail applying standard methods for weighting compositional data
 143 (e.g. Punt, this volume) and the above method for weighting the tagging data iteratively until
 144 convergence occurs. If the data are in conflict, it may be that this process will not converge
 145 and the weights for some subsets of the data will increase without limit while the weights for
 146 other subsets will be reduced to zero (Punt, this volume).

147 2.3 Applications

148 2.3.1 Aleutian Islands golden king crab

149 Siddeek et al. (in press) outline the stock assessment model used for golden king crab in the
 150 Aluetian Islands region. In relation to the tagging data, rectangular, king crab pots were used
 151 to capture crabs for tagging in all experiments, with the exception of the 1991 experiment
 152 where smaller, conical pots were used. Tagged animals were released during summer (July –
 153 September) before the fishery started. Location, date, and fishing depth were recorded for
 154 each pot retrieved. Upon pot retrieval, the carapace lengths (CL) of crabs were measured to
 155 the nearest millimetre and shell condition (old or new) recorded. Isthmus-loop (“spaghetti”)
 156 tags were used to tag crabs (Gray, 1965), and tagged crabs were released on or adjacent to the
 157 capture location. The majority of tag recaptures were obtained from the fishers during the
 158 commercial pot fishery. The tagging data used in the analyses comprised CL-at-release, CL-
 159 at-recapture, and time-at-liberty. The tagging records were restricted to male golden king crab
 160 releases in the size range 101 – 185 mm CL given the size-range included in the population
 161 dynamics model. There were 27,131 tagged crab releases in this size range, with a 6.33%
 162 overall return rate.

163 The assessment makes use of several data sources in addition to tagging data. Table 1 lists
 164 the specifications for all of the data sources included in the assessments.

165 2.3.2 Northern Prawn Fishery

166 Punt et al. (2010) outline the stock assessment model for the two tiger prawn species in the
 167 Northern Prawn Fishery. Tag-recapture data are available from experiments conducted in the
 168 northwestern Gulf of Carpentaria in 1981 (Kirkwood and Somers, 1984) and in 1983 and
 169 1984 (Somers, 1987; Somers and Kirkwood, 1991). In these experiments, prawns were
 170 captured using chartered commercial prawn trawlers and tagged using numbered streamer
 171 tags (Floy tags in 1981 experiments, Hallprint tags for the later experiments). With date of
 172 release, species, sex, and size (mm CL), and presence/ absence of bopyrid parasites
 173 *Epipenaeon ingens* recorded, prawns were released adjacent to the trawl runs where they
 174 were captured. All recaptures were made by the commercial fleet.

175 In common with Kirkwood and Somers (1984), Somers and Kirkwood (1991) and Wang
 176 et al. (1995), the data used in the assessment were restricted to animals that were at liberty for
 177 at least two weeks and which were not infected (at release or recapture) by the bopyrid
 178 parasites. Only prawns for which species, sex, size-at-release, size-at-recapture, and time-at-
 179 liberty are known, were included in the analyses. Data were excluded for *P. semisulcatus* that
 180 were extreme outliers in the growth curve analyses of Somers and Kirkwood (1991). Those

181 authors ascribed these extreme data to measurement error. The sample sizes by week (the
 182 time-step in the population model) are very small. Size results were thus pooled over time for
 183 the first diagnostic statistic and a 4-week period (“month”) for the second diagnostic statistic.

184 Table 1 lists the full set of information used when fitting the population models for the
 185 two tiger prawn species.

186 3. Results and discussion

187 3.1 Aleutian Islands golden king crab

188 Figure 1 shows the application of the first diagnostic plot (observed and model-predicted
 189 frequencies of recaptured crab by recapture size-class) when the tagging data are assumed to
 190 be Bernoulli trials (“unweighted” tagging data; solid lines). The model is able to mimic the
 191 observed distributions adequately for animals at liberty up to the third year-at-liberty, with
 192 the discrepancy between the observed and model-predicted distributions increasing with
 193 increasing time-at-liberty beyond the third year-at-liberty. This latter result is, however, not
 194 unexpected given sample sizes decline quickly with increasing time at liberty (Fig. 1).

195 The second diagnostic plot (Fig. 2) suggests that the model mimics the mean lengths-at-
 196 recapture very well for animals at liberty for the first year-at-liberty (note that the observed
 197 mean lengths-at-recapture are offset slightly to allow differences to be distinguished). The
 198 expected lengths-at-recapture in Fig. 2 (and the equivalent plot for tiger prawns) are not
 199 linear functions of the lengths-at-release because what is plotted is \hat{P}_L (Equation 4). This
 200 would be a linear function of length-at-release if allowance was made for the possibility of
 201 “shrinkage”, but that is not the case so $p(L|L)$ can be substantial for several size-classes,
 202 particularly the larger size-classes. For times-at-liberty greater than one year, $p(j|L)$ is a
 203 product of \mathbf{X} over time which is not a linear transformation.

204 The discrepancies between the observed and model-predicted mean lengths-at-recapture
 205 increase with increasing time-at-liberty. However, sample sizes get smaller with increasing
 206 time-at-liberty (Fig. 1) so the predicted 95% confidence intervals for mean lengths-at-
 207 recapture also get wider (Fig. 2). As a result, differences between observed and model-
 208 predicted mean lengths-at-recapture appear less consequential when expressed as z-scores.
 209 Application of Equation 5 to the results in Fig. 2 leads to weighting factors of 0.5, 0.2, and
 210 0.35 for times-at-liberty of 1, 2 and 3 years (rounded from 0.536, 0.185, and 0.352). The
 211 weighting factors computed using Equation 5 for times-at-liberty of 4, 5 and 6 years
 212 exceeded 1; “underdispersion” is unlikely so the weighting factors were set to 1 in subsequent
 213 analyses.

214 The fits to the frequencies of tag recaptures are essentially unchanged when the data are
 215 downweighted (Fig. 1, dashed lines). Figure 3 shows time-trajectories of mature male
 216 biomass (the population component used to provide management advice for crab stocks in the
 217 Bering Sea; NPFMC, 2008), with asymptotic 95% confidence intervals when each tag is
 218 assigned equal weight (i.e., default weighting) (upper panel in Fig. 3) and when the tagging
 219 data for animals at liberty for 1, 2 and 3 years are down weighted by factors of 0.5, 0.2 and
 220 0.35 (lower panel in Fig 3). The estimates of mature male biomass are very similar, but the
 221 precision with which the estimates are obtained is lower (by between 0% and 14%) when the
 222 tagging data are downweighted.

223 3.2 Northern Prawn Fishery

224 The solid lines in Figure 4 show the application of the first diagnostic plot to the two prawn
 225 species (and by sex) when the tagging data are assumed to be Bernoulli events. This
 226 diagnostic plot is based on data aggregated over time given low sample sizes. The fits to the
 227 data for *P. semisulcatus* (upper panels in Fig. 4) are fairly poor, with evidence that the model

228 overpredicts growth. In contrast, the fits to the data for *P. esculentus* (for which sample sizes
229 are larger) (lower panels in Fig. 4) are much better.

230 Figure 5 shows the results of applying the second diagnostic approach by prawn species
231 and sex. The over-prediction evident in Figure 4 is clear from the upper two sets of panels in
232 Figure 5, with the model predictions consistently exceeding the observed mean lengths-at-
233 recapture. The model mimics the observed mean lengths-at-recapture well for *P. esculentus*,
234 as might be expected from Fig. 4.

235 The calculated weighting factors for all months, sexes and species are approximately 1 so
236 there is no justifications for reducing the weights. However, the results in Fig. 4 are
237 suggestive of conflicts in the data. Consequently, the assessment was conducted
238 unrealistically increasing the weight assigned to the tagging data by a factor of 1,000 (Fig. 4,
239 dashed lines). The model is now able to mimic the tagging data well. However, the fits to the
240 length-frequency data (Fig. 6) deteriorate quite markedly for some species, sexes, and length-
241 frequency types when the tagging data are upweighted. In addition, the selectivity pattern for
242 *P. semisulcatus* on large individuals is reduced substantially when the tagging data are
243 substantially upweighted.

244 3.3 General comments and thoughts

245 This paper has provided two diagnostic statistics that can be used to evaluate the ability of
246 size-structured assessment models to replicate tagging data and also outlined a way to weight
247 tagging data when these data are included in an assessment along with several other data
248 sources. The results in Figures 4 and 6 show that the fits to length-frequency and tagging data
249 can be in conflict. It is well known that models that fit to size-composition data can be very
250 sensitive to assumptions about growth, especially when the parameters of the growth model
251 are pre-specified (e.g. Aires-da-Silva, and Maunder, 2012). However, this is the first example
252 that shows that the results of size-structured integrated stock assessments can also be
253 sensitive to the weighting of tagging data.

254 While it would be ideal to report both plots when reporting assessment results, plots such
255 as Figs 2 and 5 are more informative than plots such as Figs 1 and 4 because model mis-
256 specification could be missed in the latter plots if, for example, growth of smaller animals
257 was over-estimated while that of larger animals was under-estimated.

258 The statistics and plots outlined in this paper could be extended. For example, Figs 2 and
259 5 indicate measures of uncertainty as ± 1.96 standard errors. However, the matrix \mathbf{X} is usually
260 lower triangular (because most crustaceans do not shrink following a moult). Consequently,
261 for short times-at-liberty, the modal number of size-classes an animal (particularly a large
262 individual) may grow will be 0 (i.e. it is most likely to remain in its size-class-at-release),
263 which suggests a non-symmetric distribution for the size-increment and hence the mean size-
264 at-recapture. Such a distribution could be computed using the quantiles of the distribution for
265 the size-increment from the matrix \mathbf{X} . Figs 2 and 5 (and methods such as those of Francis
266 (2011) and Punt (this volume)) focus on the mean age/length (for composition data) or mean
267 length-at-recapture (for tagging data). However, similar statistics could be developed to
268 explore whether the model is able to adequately mimic the variation in the length-at-
269 recapture. A statistic to make this comparison has been developed for age/length data
270 (“Andre Plots” in the r4ss package [Taylor et al., 2014] that compares the observed and
271 model-predicted variances of the size- or-age-compositions under the assumption that the
272 model-predicted variances are F-distributed). However, unless the number of recaptures is
273 high (which is unlikely to be the case for tagging data), the power of such a statistic is likely
274 to be low, and the statistic consequently uninformative.

275 In common with all diagnostics plots and statistics, the suggestions in this paper indicate
276 potential problems, but cannot prove that the model is adequate if the diagnostics do not

277 provide evidence for model mis-specification and/or overdispersion. Patterns of residuals
 278 (such as those for *P. semisulcatus* in Figs 4 and 5) could be suggestive of model mis-
 279 specification, i.e. in the case that the von Bertalanffy growth curve is an inadequate
 280 representation of growth. In the case of prawns, there is evidence that the growth rate varies
 281 cyclically during the year in some species (e.g., Somers, 1988; Xiao and McShane, 2000;
 282 Lloyd-Jones et al., 2012), but this is not accounted for in the assessment model for NPF tiger
 283 prawns. Whether differences in growth seasonally are sufficient to generate the types of
 284 patterns evident in Figs 4 and 5 is beyond the scope of the present paper. We note, however,
 285 that seasonality in growth was not found to be significant in blue endeavour prawns,
 286 *Metapenaeus endeavouri*, caught in the same fishery (Buckworth, 1992). Moreover, patterns
 287 in residuals may reflect issues with other aspects of the model and be unrelated to the tagging
 288 data and the model of growth.

289 The diagnostic statistics and the approach to weighting outlined in this paper are
 290 heuristics and based on analogy with approaches used to weight age- and size-composition
 291 data. Future work should consider the use of simulation studies to evaluate the performance
 292 of these statistics and the weighting method, particularly in cases in which the model is mis-
 293 specified (e.g., is based on a non-seasonal von Bertalanffy curve when this is not the case).

294 Finally, the assessments considered in this paper are based on the likelihood function
 295 given by Equation 1. However, tagging data can be included in other ways in integrated stock
 296 assessments. For example, the method of Haist et al. (2009) is used in New Zealand to
 297 conduct assessments of rock lobsters and abalone (e.g., Breen et al., 2012; Fu et al., 2010).
 298 This method assumes that the size-increment is governed by a truncated normal distribution
 299 and allows for measurement error when fitting the tag-recapture data. In principle, the
 300 approach of this paper could be applied to create diagnostic statistics to evaluate models
 301 implemented using the method of Haist et al. (2009).

302 **4. Acknowledgements**

303 We would like to thank two anonymous reviewers for their comments on an earlier draft of
 304 this paper. This work was partially funded by NOAA Award NA14NMF4370119, Australian
 305 Fisheries Management Authority (AFMA) project RR2015/0811 and the Joint Institute for
 306 the NOAA Ocean Acidification Program through the Joint Institute of the Study of the
 307 Atmosphere and Ocean (JISAO) under NOAA Cooperative agreement No.
 308 NA10OAR4320148. This is JISAO Contribution No. xxx.

309 **References**

- 310 Aires-da-Silva, A.M., Maunder, M.N., 2012. Status of bigeye tuna in the eastern Pacific Ocean in 2011 and
 311 outlook for the future. Inter-American Tropical Tuna Commission (IATTC), Stock Assessment Report 13,
 312 pp. 18–29.
- 313 Breen, P.A., Haist, V., Starr, P.J., Pomarede, M., 2012. The 2011 stock assessment and management procedure
 314 development for red rock lobsters (*Jasus edwardsii*) in CRA 4. New Zealand Fisheries Assessment Report
 315 2012/09.
- 316 Buckworth, R.C., 1992. Movements and growth of tagged blue endeavour prawns *Metapenaeus endeavouri*
 317 (Schmitt 1926) in the western Gulf of Carpentaria, Australia. Aust. J. Mar. Freshw. Res. 43, 1283–1299.
- 318 Buckworth, R.C., Deng, R.A., Plaganyi, E., Punt, A., Upston, J., Pascoe, S., Miller, M., Hutton, T., Lawrence,
 319 E. B., Venables, B., 2015. Northern Prawn Fishery RAG Assessment 2013-15. AFMA Project R13/0005
 320 170pp.
- 321 Francis, R.I.C.C., 2011. Data weighting in statistical fisheries stock assessment models. Can. J. Fish. Aquat. Sci.
 322 68, 1124–1138.
- 323 Fu, D., McKenzie, A., 2010. The 2010 stock assessment of paua (*Haliotis iris*) for Milford, George, Central, and
 324 Dusky in PAU 5A. New Zealand Fisheries Assessment Report 2010/46.
- 325 Gray, G.W., 1965. Tags for marking king crabs. Progr. Fish-Cult. 27, 221-227.
- 326 Haist, V., Breen, P.A., Starr, P.J., 2009. A multi-stock, length-based assessment model for New Zealand rock
 327 lobster (*Jasus edwardsii*). NZ J. Mar. Freshw. Res. 43, 355-371.

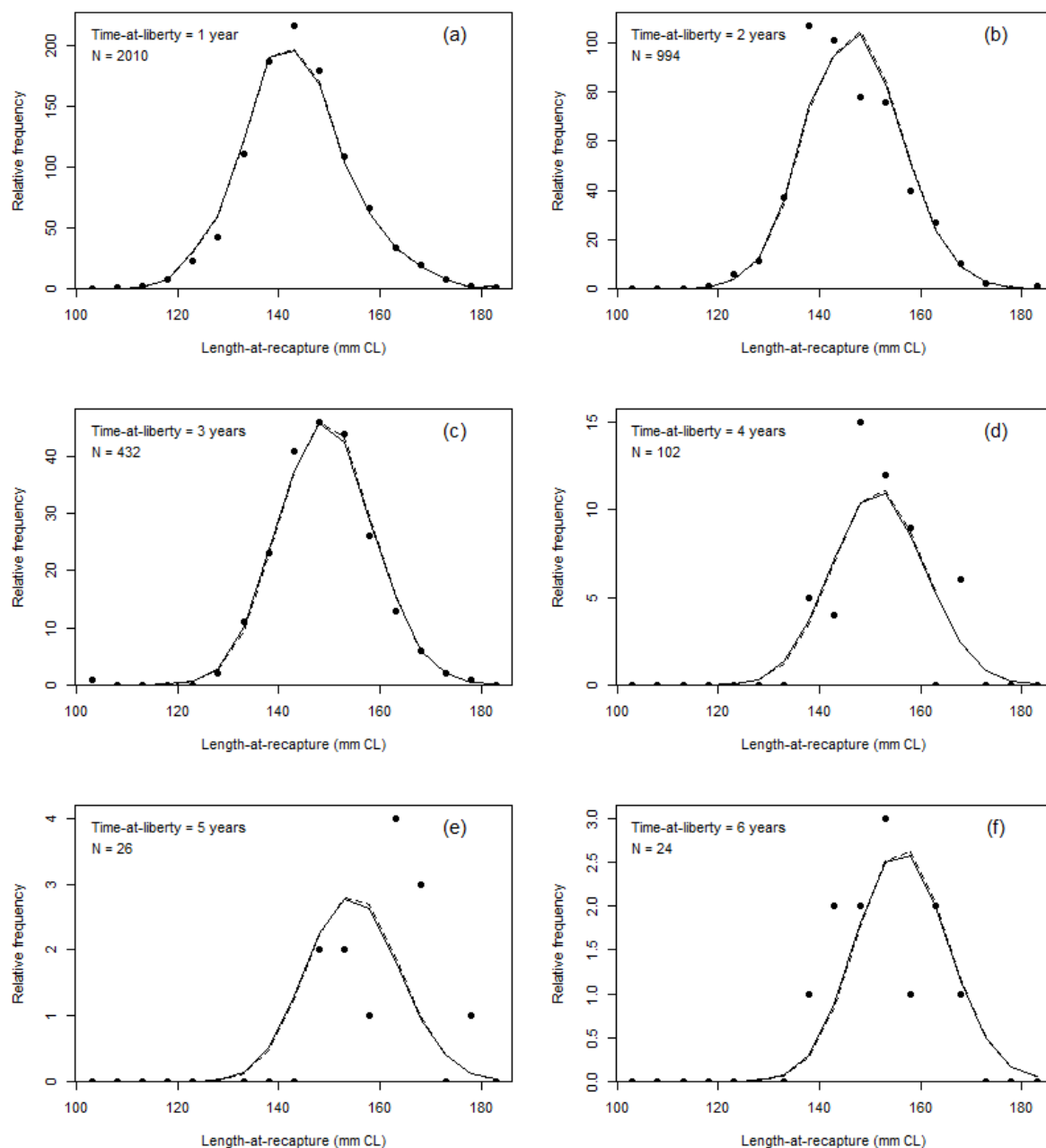
- 328 ICCAT, 2013. Report of the 2012 white marlin stock assessment meeting. Coll. Vol.Sci. Pap. ICCAT 69, 1085–
329 1183.
- 330 Kirkwood, G.P., Somers, I.F., 1984 Growth of two species of tiger prawn, *Penaeus esculentus* and *P.*
331 *semisulcatus*, in the western Gulf of Carpentaria. Aust. J. Mar. Freshw. Res 35, 703-712.
- 332 Lloyd-Jones, L.R., Wang, Y-G, Courtney, A.J., Prosser, A.J., Montgomery, S.S., 2012. Latitudinal and seasonal
333 effects on growth of the Australian eastern king prawn (*Melicertus plebejus*). Can. J. Fish. Aquat. Sci. 69,
334 1525–1538.
- 335 McAllister, M.K., Ianelli, J.N., 1997. Bayesian stock assessment using catch-age data and the
336 sampling/importance resampling algorithm. Can. J. Fish. Aquat. Sci. 54, 284–300.
- 337 North Pacific Fishery Management Council (NPFMC). 2008. Amendment 24. Final Environmental Assessment
338 for amendment 24 to the Fishery Management Plan for Bering Sea/Aleutian Islands King and Tanner Crabs
339 to Revise Overfishing Definitions. North Pacific Fishery Management Council, 605 West 4th Ave,
340 Anchorage, AK 99501.
- 341 Punt, A.E., This volume. Some insights into data weighting in integrated stock assessments. Fish. Res. 00, 00–
342 00.
- 343 Punt, A.E., Kennedy, R.B., 1997. Population modelling of Tasmanian rock lobster, *Jasus edwardsii*, resources.
344 Mar. Freshw. Res. 48, 967–980.
- 345 Punt, A.E., Buckworth, R.C., Dichmont, C.M., Ye, Y., 2009. Performance of methods for estimating size–
346 transition matrices using tag–recapture data. Mar. Freshw. Res. 60, 168–182.
- 347 Punt, A.E., Deng, R.A, Dichmont, C.M. Kompas, T., Venables, W.N., Zhou, S., Pascoe, S., Hutton, T., Kenyon,
348 R., van der Velde, T., Kienzle, M., 2010. Integrating size-structured assessment and bio-economic
349 management advice in Australia’s Northern Prawn Fishery. ICES J. Mar. Sci. 67, 1785–1801.
- 350 Punt, A.E., Haung, T-C., Maunder, M.N., 2013. Review of integrated size-structured models for stock
351 assessment of hard-to-age crustacean and mollusc species. ICES J. Mar. Sci. 70, 16–33.
- 352 Richards, L.J., 1991. Use of contradictory data sources in stock assessments. Fish. Res. 11, 225–238.
- 353 Siddeek, M.S.M., Zheng, J., Punt, A.E., Vanek, V., In press. Estimation of size-transition matrices with and
354 without moult probability for Alaska golden king crab using tag–recapture data. Fish. Res. 00, 00–00.
- 355 Somers, I.F., 1987. Sediment type as a factor in the distribution of commercial prawn species in the western
356 Gulf of Carpentaria, Australia. Aust. J. Mar. Freshw. Res. 38, 133–149.
- 357 Somers I.F., 1988. On a seasonally oscillating growth function. Fishbyte 6, 8–11
- 358 Somers, I.F., Kirkwood, G.P., 1991. Population ecology of the grooved tiger prawn, *Penaeus semisulcatus*, in
359 the north-western Gulf of Carpentaria, Australia: growth, movement, age structure and infestation by the
360 bopyrid parasite *Epipenaeon ingens*. Aust. J. Mar. Freshw. Res. 42, 349–267.
- 361 Taylor, I., Stewart, I., Hicks, A., Garrison, T., Punt, A., Wallace, J., Wetzels, C., Thorson, J., Takeuchi, Y.,
362 Monnahan, C. 2014. Package r4ss. <https://github.com/r4ss>
- 363 Turnock, B.,J., Rugolo, L.J., 2014. Stock assessment of Eastern Bering Sea snow crab. pp. 41-177. Stock
364 Assessment and Fishery Evaluation Report for the king and Tanner crab fisheries of the Bering Sea and
365 Aleutian Islands Regions. North Pacific Fishery Management Council, 605 W. 4th Avenue, #306,
366 Anchorage, AK 9950.
- 367 Wang, Y-G., Thomas, M.R., Somers, I.F., 1995. A maximum likelihood approach for estimating growth from
368 tag–recapture data. Can. J. Fish. Aquat. Sci. 52, 252–259.
- 369 Xiao, Y., McShane P., 2000. Use of age- and time-dependent seasonal growth models in analysis of
370 tag/recapture data on the western king prawn *Penaeus latisulcatus* in the Gulf St. Vincent, Australia. Fish.
371 Res. 49, 85–92.
- 372 Zheng, J., Siddeek, M.S.M. 2014. Bristol Bay red king crab stock assessment in fall 2014. pp. 178-323. Stock
373 Assessment and Fishery Evaluation Report for the king and Tanner crab fisheries of the Bering Sea and
374 Aleutian Islands Regions. North Pacific Fishery Management Council, 605 W. 4th Avenue, #306,
375 Anchorage, AK 9950.
- 376
- 377

378 Table 1. Data available for the stocks considered
 379

| Data type | Year range |
|---|---------------------------------------|
| <i>Aleutian islands golden king crab</i> | |
| Retained catch in numbers in total | 1981 - 1984 |
| Retained catch in numbers by length-class | 1985 - 2014 |
| Total (retained + discarded) catch in numbers | 1990 - 2014 |
| CPUE (observer) | 1995 - 2014 |
| Groundfish bycatch in numbers by length-class | 1989 - 2014 |
| Tagging data (recaptures to 2012) | 1991, 1997, 2000, 2003, 2006 releases |
| <i>P. semisulcatus</i> and <i>P. esculentus</i> | |
| Weekly catch (mass) and effort (vessel days fished) | 1970 - 2014 |
| Fishery length-frequency data | 2002 - 2003 (sex-specific) |
| Spawning survey indices | 2002 - 2014 |
| Spawning survey length-frequencies | 2002 - 2014 (sex-specific) |
| Recruitment survey indices | 2003 - 2014 |
| Recruitment survey length-frequencies | 2003 - 2015 (sex-specific) |
| Tagging data | 1981 - 1984 |

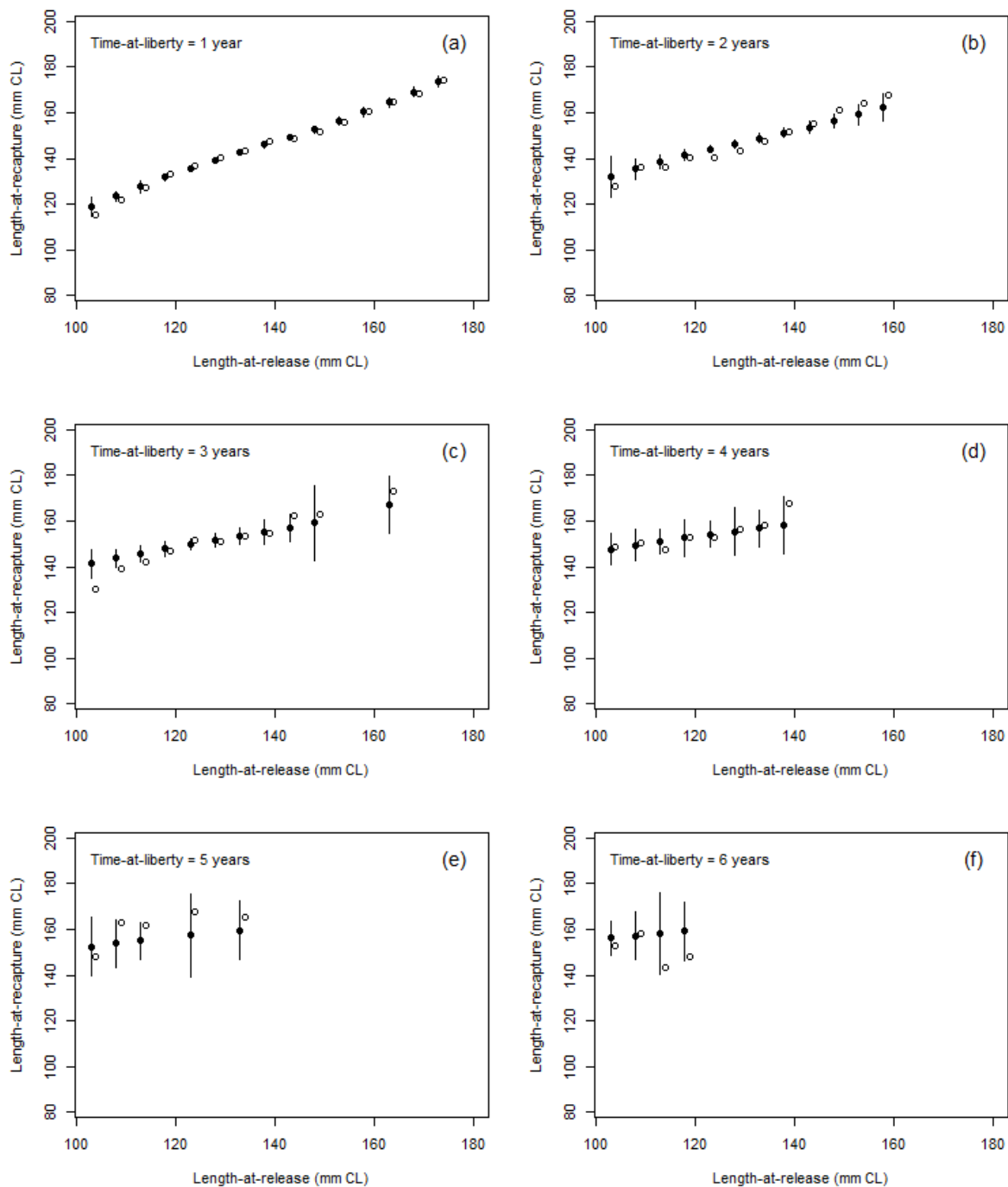
380

381



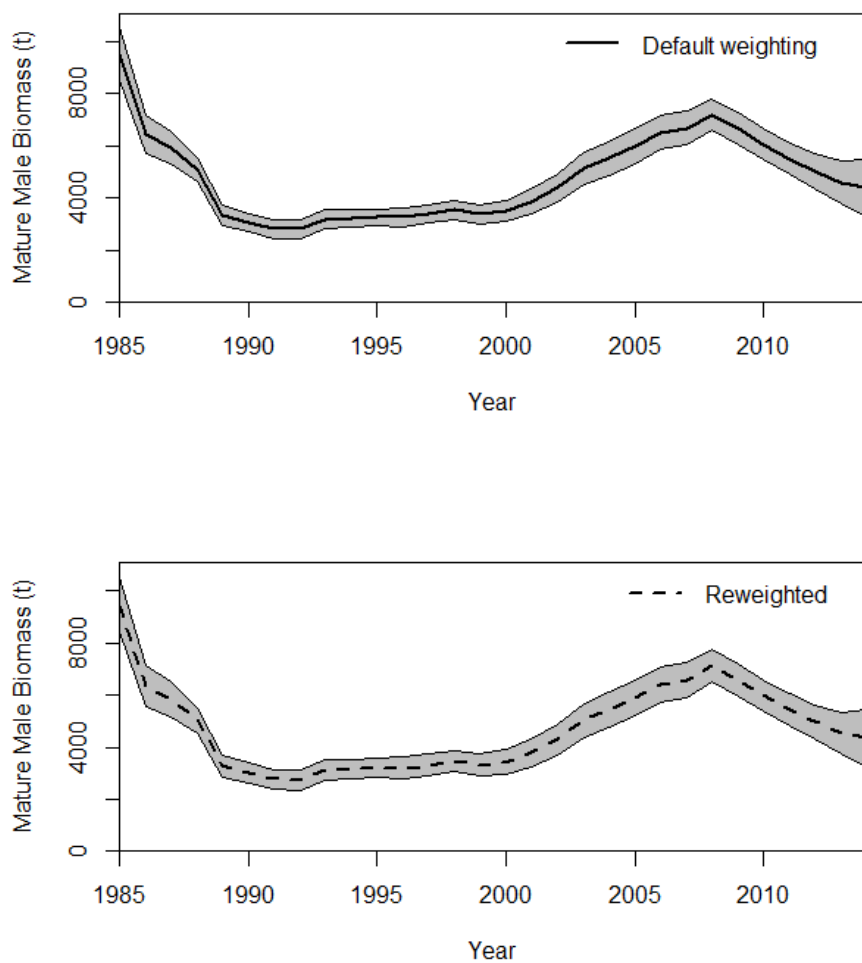
1
 2 Figure 1. Observed tag recaptures (filled circles) versus model-predicted tag recaptures (solid
 3 and dashed lines) by size-class at recapture for male Aleutian Islands golden king crab. Results
 4 are shown for animals that were at liberty for between one and six years. N denotes the sample
 5 size. The solid lines show results for the original weighting and the dashed lines those when
 6 the tagging data are downweighted.

7
 8

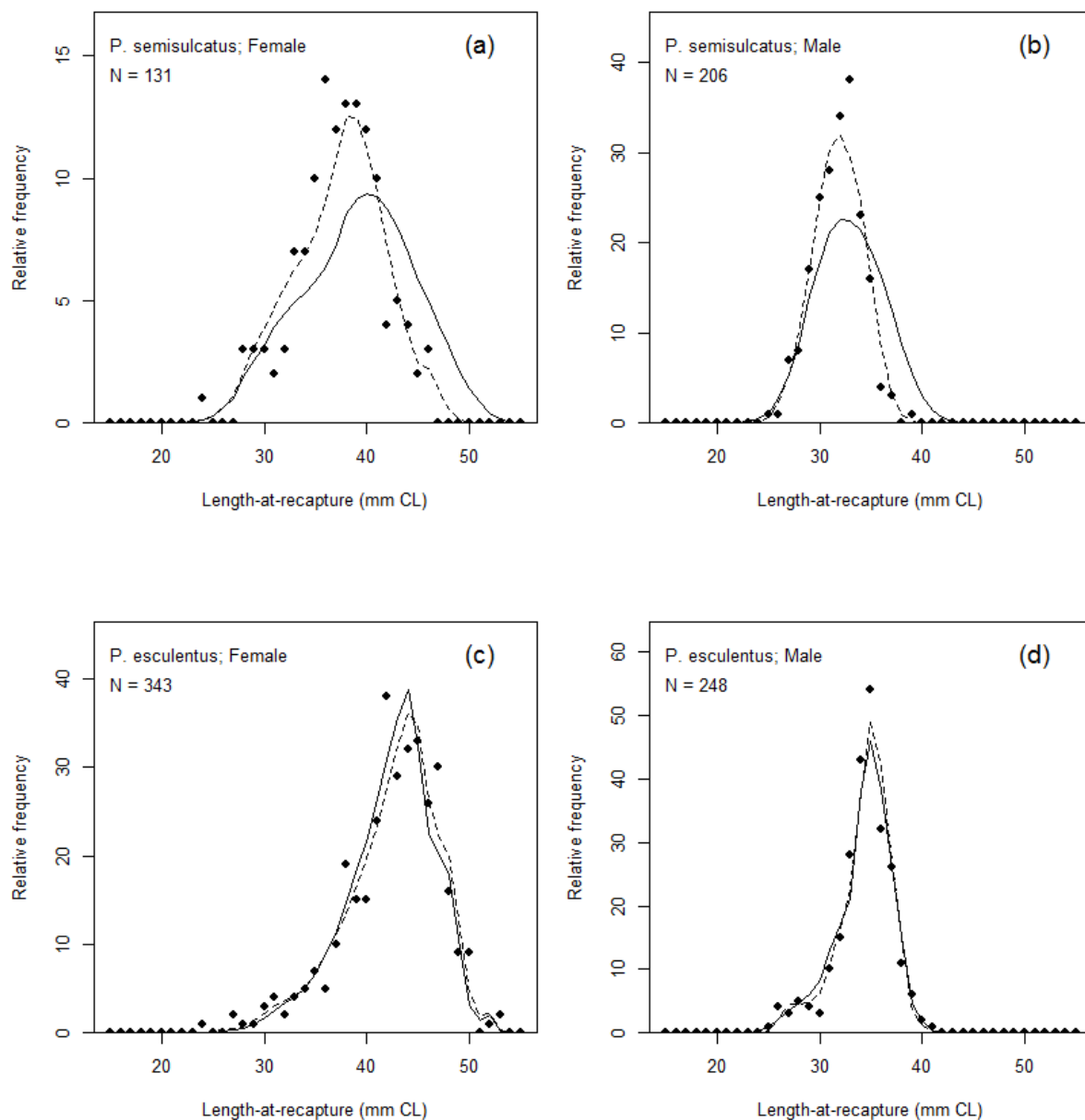


9
10
11
12
13
14
15
16
17

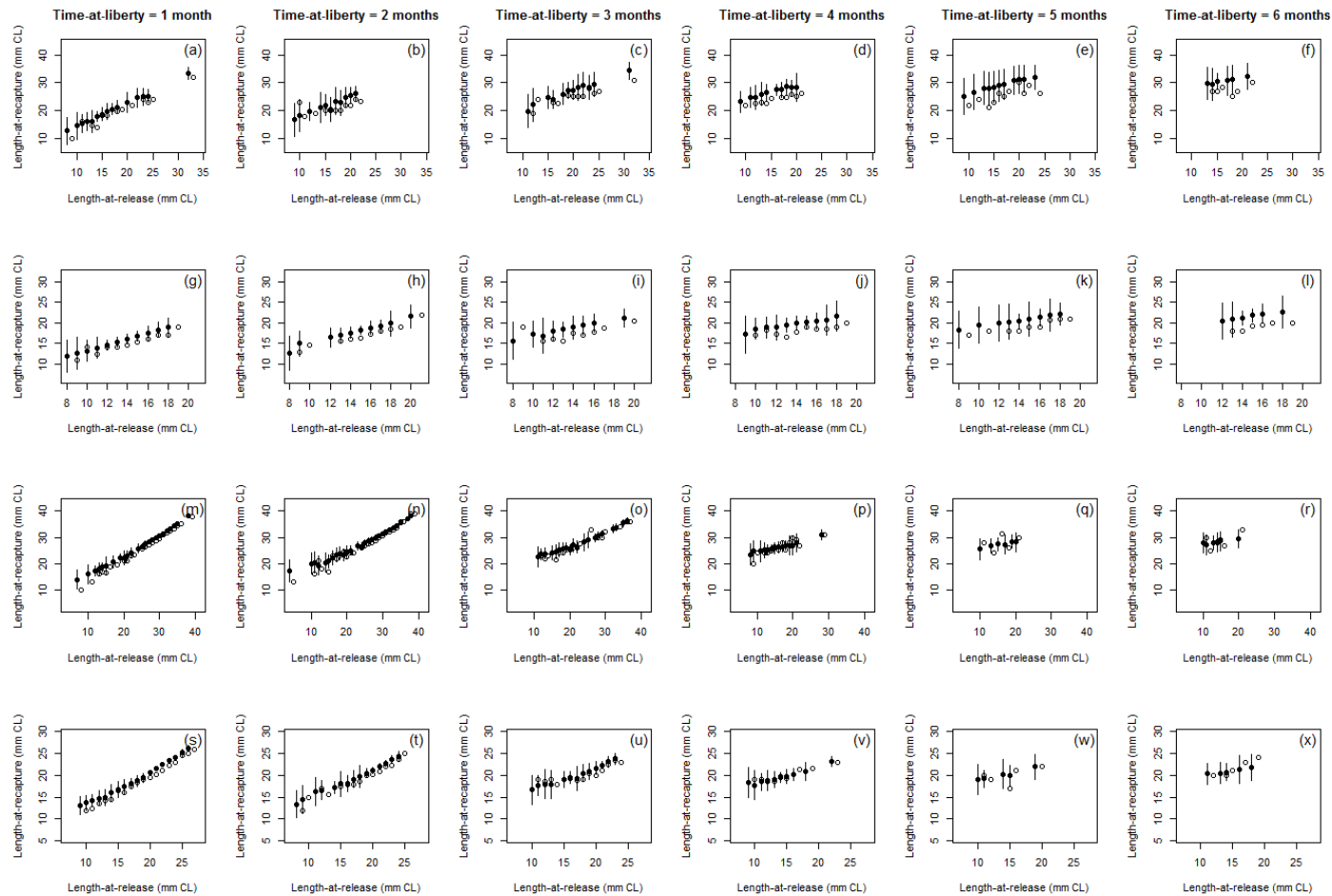
Figure 2. Observed (open circles) and model-predicted (filled circles) mean lengths-at-recapture versus release length for male Aleutian Islands golden king crab. The vertical lines indicate 95% confidence intervals for model-predicted mean lengths-at-recapture. Results are shown for animals that were at liberty for between one and six years. The model predictions correspond to the model configuration in which the tagging data are treated as the outcomes from Bernoulli experiments.



18
19 Figure 3. Time-trajectory of mature male biomass for Aleutian Islands golden king crab.
20 Results are shown when the tagging data are taken to be the results of Bernoulli trials and when
21 the tagging likelihood is downweighted using Equation 5.
22

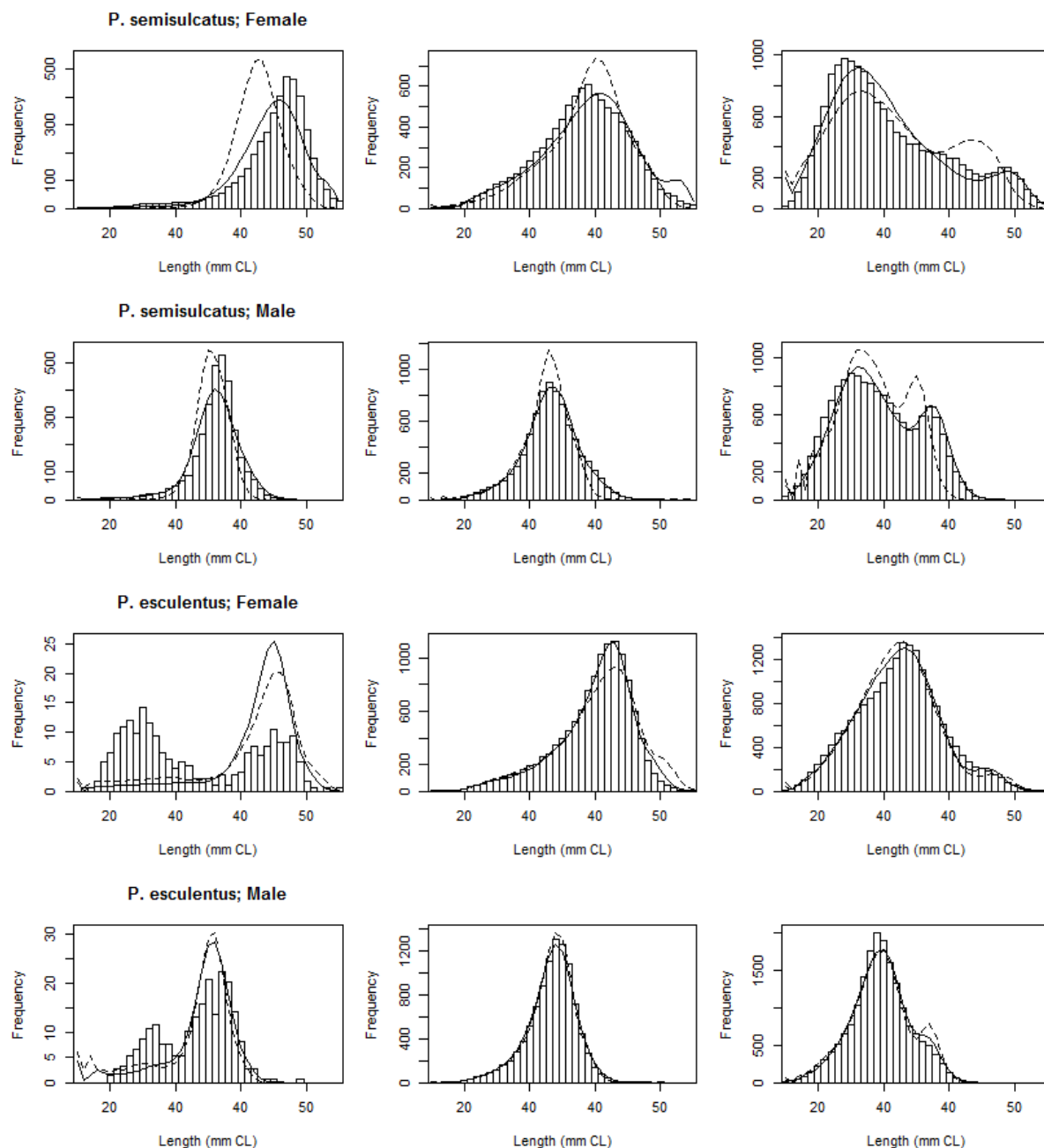


23
 24 Figure 4. Observed tag recaptures (filled circles) versus model-predicted tag recaptures (solid
 25 and dashed lines) by size-class at recapture for two prawn species in Australia's northern prawn
 26 fishery by sex. Data are aggregated over time-at-liberty owing to small sample sizes. N denotes
 27 the sample size. The solid lines show results for the original weighting and the dashed lines
 28 those when the tagging data are upweighted 1000-fold.



29
30

31 Figure 5. Observed (open circles) and model-predicted (filled circles) mean lengths at recapture versus release length for two prawn species in
 32 Australia's northern prawn fishery by sex. The vertical lines indicate 95% confidence intervals for model-predicted mean lengths-at-recapture.
 33 Results are shown for animals that were at liberty for between one and six months. Results for female *P. semisulcatus*, male *P. semisulcatus*,
 34 female *P. esculentus*, and male *P. esculentus* are respectively shown in each row (top to bottom). The model predictions correspond to the model
 35 configuration in which the tagging data are treated as the outcomes from Bernoulli experiments.



36

37 Figure 6. Observed (bars) model-predicted length-frequency distributions when the
 38 observations and model-predictions are aggregated over time (weighted by their assumed
 39 effective sample sizes). Results are shown for two species and two sexes (rows) and for three
 40 types of length-frequencies (catch, left; spawning, middle; recruitment, right). The solid lines
 41 show model predictions for the original weighting and the dashed lines those when the tagging
 42 data are upweighted 1000-fold.