Version of Record: https://www.sciencedirect.com/science/article/pii/S0044848621002155 Manuscript_94deecca591c20a43211dc2fcdd6cad9

1	Running head: predator cues enhance oyster survival
2	
3	Use of predator cues to bolster oyster resilience for aquaculture and reef restoration
4	
5	Benjamin A. Belgrad ^{a,} *, Emily M. Combs ^b , William C. Walton ^c , Delbert L. Smee ^{d,e}
6	
7	
8	^a Dauphin Island Sea Lab, Dauphin Island, AL 36528; Tel: (330)-398-3900; Email:
9	babelgra@eckerd.edu
10	^b College of Science, Florida Atlantic University, Boca Raton, FL 33431; Email:
11	emilymcombs@gmail.com
12	^c Auburn University Shellfish Laboratory 150 Agassiz St, Dauphin Island, AL 36528; Email:
13	billwalton@auburn.edu
14	^d University Programs, Dauphin Island Sea Lab, Dauphin Island, AL 36528; Email: Ismee@disl.org
15	^e Department of Marine Science, University of South Alabama, Mobile, AL 36688
16	
17	*To whom correspondence should be sent

Abstract

Many mollusks alter their shell morphology in response to predator exudates or injured 20 conspecifics to lower their predation risk. However, studies have yet to examine whether this 21 22 predator-avoidance response can be applied to bolster reef restoration, fisheries enhancement, 23 or aquaculture. We tested whether exposure to predator cues under hatchery conditions can increase the survival of oysters, *Crassostrea virginica*, planted in the field on the substrate. 24 Juvenile oysters, set on shells and grown in a flow-through system, were exposed to either 25 26 caged blue crabs, *Callinectes sapidus*, or controls of empty cages for either four or eight weeks 27 then placed in the field for 30 days. We compared oyster shell strength and morphology as well as oyster survival among predator exposure time treatments. Oysters grown in the hatchery for 28 eight weeks were 46% larger and almost 2x stronger than oysters grown for four weeks. 29 However, predator exposure also caused a 50% increase in shell strength for both time periods. 30 31 In the field, oysters suffered relatively little mortality when protected from predators using 32 cages, and virtually all mortality was attributed to predation. Predator cue treatments significantly increased the survival probability of uncaged oysters (as would be done in reef 33 restoration or stock enhancement) compared to unexposed treatments. Early cue exposure 34 35 yielded substantially greater gains in survivorship over time as predator induced oysters nursed 36 for four weeks exhibited 53% higher survival in the field than unexposed oysters while this survivorship gain jumped to 300% for eight weeks of cue exposure. Our findings demonstrate 37 38 that predator cues can be an effective means for the industry to increase the operational efficiency of aquaculture and restoration efforts, and may potentially be applied to other 39 bivalve fisheries. 40

42 Key Words: Survivorship, Crassostrea virginica, shell morphology, phenotypic plasticity

43

1. Introduction

44 Globally, more than 15 million tons of marine bivalves are harvested each year for 45 human consumption, 89% of which comes from aquaculture efforts (Wijsman et al. 2019). 46 Oysters are among the most valued of these species as they not only constitute 33% of this global production (FAO 2019), but also provide a host of ecosystem services. These services 47 range from shoreline protection, water filtration, and habitat creation (Grabowski and Peterson 48 2007) to shaping the cultural identity of regions (Michaelis 2020). Yet, oysters are one of the 49 50 most degraded marine habitats, with ~85% of oyster reefs lost worldwide (Beck et al. 2011). 51 Countries, including the United States, have experienced significant declines in the wild oyster fishery (71% over the past half-century) accompanied by the loss of benefits that oysters 52 provide (Zu Ermgassen et al. 2013, Wijsmanet al. 2019). Consequently, oyster aquaculture 53 54 continues to increase considerably in an effort to both supplement the loss of the wild fishery and to facilitate the restoration of oyster reefs and their ecosystem services. 55

Most oyster aquaculture, restoration, and stock enhancement involves relatively extensive culture methods (rather than intensive operations) where larvae are spawned in a hatchery and juveniles are planted in natural or semi-natural settings to grow to adulthood. One of the greatest challenges to extensive operations is mortality from predation that can decimate populations since even protected stocks on farms can lose 28% of their biomass from predators (Richard et al. 2020) while unprotected regions can lose 94% of planted juveniles

62 within weeks (Mackenzie 1970). Locally, losses to predation within off-bottom containers are 63 typically near zero when maintained properly (Walton, pers. obs.), while predation on-bottom has been observed to inflict losses > 87% (Lappin, 2018). The predominant predators that 64 65 threaten stocks can vary by region and oyster age. For example, in northern latitudes, starfish are frequently considered the most destructive predators to crops (Hancock 1955) while oyster 66 67 drills are a larger threat in the Gulf of Mexico (Butler 1985) to both juveniles and adults. 68 Additionally, mud crabs and a wide number of fish species are common predators of juvenile oysters (McDermott 1960, Anderson and Connell 1999) while fish species like black drum can 69 70 be a major predator to adults (Brown et al. 2008). Consequently, farmers have developed a number of practices to reduce mortality from these different predators (Matthiessen 2001, 71 72 Gosling 2008). Such practices include selecting sites with relatively low predation intensity 73 (Buitrago et al. 2005), mass removal of predators from sites (Calderwood et al. 2016), 74 protecting the bivalves inside some type of container (often suspended or floating, Gosling 75 2008), nursing juvenile bivalves in the hatchery until they reach a size refuge from predation (Wijsman et al. 2019), or some combination of the above. However, many of these techniques 76 are expensive, labor intensive, and/or are not feasible at the desired scale due to conflicts with 77 78 other local economical, ecological, or cultural interests. Similarly, it is not uncommon for oyster 79 reef restoration efforts to fail (Mann and Powell 2007, La Peyre et al. 2014) as yearly age-80 specific mortality rates can exceed 70% in some locations (Mann et al. 2009). Although 81 predators are a common source of mortality in oysters, especially among juveniles (Bisker and 82 Castagna 1987), many of the most effective techniques in farming to prevent predation (e.g. 83 containerized culture) are too labor intensive for large-scale commercial growers or restoration

projects. These large-scale efforts typically use remote setting, where oyster larvae are allowed
to set upon substrate (often oyster shell, and called spat-on-shell) and ultimately stocked into
the target area.

One potential technique to increase the survival of oysters in large-scale aquaculture or 87 reef restoration efforts is early exposure of juvenile oysters to predator cues. Many mollusks, 88 89 including mussels (Leonard et al. 1999), clams (Nakaoka 2000), and oysters (Robinson et al. 90 2014), will strengthen their shells when exposed to predators to reduce their risk of being 91 consumed. Oysters are known to strengthen their shells in response to both crustacean (Newell et al. 2007) and gastropod predators (Lord and Whitlatch 2012, Ponce et al. 2020), which can 92 increase their survival under laboratory settings (Robinson et al. 2014, Ponce et al. 2020). 93 Oysters respond to chemical exudates from injured con— and hetero—specifics as well as 94 95 predator exudates by building thicker shells and altering the composition of shells (Scherer et 96 al. 2018). However, most studies on inducible defenses of bivalves have occurred under closed laboratory conditions which can only induce dozens to hundreds of individuals simultaneously 97 and frequently inflate exposure to predator cues beyond natural conditions. It is unknown 98 whether predator exposure techniques can induce bivalves to grow stronger shells under large-99 100 scale settings that utilize flow-through systems and have the capacity to hold hundreds of 101 thousands to millions of oysters. Additionally, the few studies that have investigated the effects 102 of predator induction on bivalve survival are typically laboratory based and short-term, lasting hours to days (e.g. Robinson et al. 2014, Sherker et al. 2017). Researchers have yet to study the 103 extent to which predator induction enhances survival in the field when encountering a natural 104 105 suite of predators over longer time periods.

We tested the feasibility of using predator cues to increase the survival of oysters in aquaculture and reef restoration operations. We grew eastern oyster (*Crassostrea virginica*) juveniles set on shell under hatchery flow-through conditions, which can raise millions of juveniles per brood, to determine if 1) oysters can be induced to grow thicker shells in mass quantities and 2) predator induction affects survival in the field. We nursed oysters for four weeks (comparable to normal nursery times; Matthiessen 2001) and eight weeks to assess the degree to which cue exposure benefits scale over time, and then assessed survival in the field.

113

2. Methods

114 *2.1 Oyster culturing*

115 Oysters (Crassostrea virginica) were cultured as spat-on-shell at the Auburn University 116 Shellfish Laboratory (AUSL) on Dauphin Island, AL starting in late May 2019 using standard techniques (Congrove et al. 2009). Oysters were ~1.0 mm when the experiment began and 117 housed in four flow-through holding tanks measuring 2.4 m x 0.9 m (length x width) with a 118 119 water depth of 0.4 m (~20,000 spat/tank). Water flow rates in the holding tanks averaged 36.9 L/min. There was immense variation in the number of spat per shell which we elected to 120 121 maintain during the experiment to mimic natural settlement and normal reef restoration 122 practices (5 – 40 spat/shell at four weeks of culturing). Oysters were suspended above the tank bottom in seven oyster aquaculture baskets (64 x 23 x 14 cm with 140 spat covered 123 124 shells/cage; ~80,000 spat total) to prevent sediment buildup from suffocating oysters. These holding containers and shell densities matched normal nursery procedures for spat-on-shell 125 126 (Matthiessen 2001, personal communication, AUSL hatchery manager Scott Rikard).

127 Half of the oysters were exposed to predator exudates by holding four live adult blue crabs, Callinectes sapidus, in two of the flow-through tanks (8 crabs total) while the remaining 128 129 two tanks did not have crabs and served as a control (hereafter known as induced and 130 uninduced oysters respectively). Crabs were held in two partitioned baskets to prevent crabs from consuming the experimental oysters or each other while control tanks had empty crab 131 132 cages. Each crab was fed one adult oyster daily (~5.0 cm in length) to maximize predation risk 133 cues, causing experimental oysters to receive exudates from both crabs and injured oysters as they were being consumed. Oyster cages were rotated daily around crab cages to reduce 134 135 differences in growth due to proximity to cue sources. Crabs were replaced at least every other 136 week to ensure predators remained healthy and to replace crabs that died. After four and eight 137 weeks in the hatchery, subsets of spat-covered shells were taken to the laboratory to measure 138 differences in shell morphology while other subsets were planted in the field to assess effects 139 on survival.

140 2.2 Shell morphology

141 Two shells were taken from every basket and three live spat were selected from each shell for measuring spat shell characteristics after four and eight weeks (number of individuals = 142 84 for each cue exposure x time treatment; 112 shells and 336 spat total). Spat shell 143 144 morphology was assessed by measuring shell size, shell weight, and shell crushing force (sensu 145 Robinson et al. 2014, Scherer et al. 2016). Oysters are roughly round during early life stages, 146 and shell length was measured from the umbo to the outer shell edge to the nearest 0.01 mm using digital calipers. Care was taken to only measure individuals that were not crowded by 147 148 cohorts to reduce any confounding effects on growth due to space limitation, although this was

149 not a common occurrence at these early life stages. We quantified the force needed to break each oyster shell using a penetrometer (Kistler force sensor 9203 and Kistler charge amplifier 150 5995). The force sensor was placed equidistant from the shell edges and perpendicular to the 151 152 shell surface. Gentle, consistent pressure was applied until the shell cracked, and the maximum force needed to break the shell (N) was recorded. This technique is a standard proxy of shell 153 154 hardness (Robinson et al. 2014). We divided shell crushing force by shell length to produce a 155 size-standardized metric of shell strength (i.e. standardized crushing force, N/mm) because larger individuals naturally have a stronger shell as a byproduct of their size. After crushing, 156 oyster shell dry weight was obtained by collecting all the shell fragments and removing any 157 158 soft-tissue before desiccating in an oven at 70 °C for 48 hrs. Only the left oyster valves were 159 weighed as the right valves were bonded to the underlying substrate and because crushing 160 force was applied to just the left valve.

161 We examined the effects of predator cue exposure (present vs absent) and time 162 cultured (4 weeks vs 8 weeks) on standardized shell crushing force, shell length, and shell weight by running three separate generalized linear mixed models with Gamma distributions, 163 one model for each of these three response variables (GLMMs; R package: Ime4). Cue exposure 164 treatment and time were set as fixed effects with an interaction term while shell spat settled 165 166 on, nested in basket, nested in tank were treated as random effects to control for nonindependence among individuals (Bolker et al. 2009). Tukey's multiple comparison test was 167 used to determine pairwise differences in shell morphology (R package: Ismeans). All statistical 168 analyses were conducted using R v3.5.1 (R Development Core Team, 2018). 169

170 2.3 Field survival

171 To quantify the extent that inducing oysters alters survival in the field over time, five to six spat covered shells were selected from each basket after both four and eight weeks in the 172 173 hatchery and placed in the field for 30 days (see Figure 1 for spat sizes and shell strength). Each 174 shell was manually thinned so only 10 spat where present on each shell to standardize predator 175 risk exposure (number of shells used = 80 shells for each cue exposure x time treatment; 320 176 shells and 3,200 spat total). We wished to ensure the experiment had enough replication to 177 detect medium effect sizes on survival (h = 0.5, power = 0.999; Cohen 1988) so we set ~4x more oysters than necessary to achieve this. Four pairs of induced and uninduced oysters were zip 178 179 tied to 1-meter long horizontal PVC frames (20 frames per hatchery cue exposure time; 40 frames total). One pair of shells on each frame was randomly selected to be surrounded by a 180 181 mesh cage to exclude predators and control for mortality events from nonpredatory sources 182 (e.g. disease, abiotic conditions). Initially, cages were composed of a semiflexible mesh, but 183 after predators were repeatedly found within the cages, this setup was replaced with a stiffer 184 inflexible cylindrical plastic cage (diameter = 18 cm, length = 22 cm) and overlain with fine mesh (2 mm pore size). The frames were set at the Point aux Pins Oyster farm (30°23'00.7"N, 185 88°18'46.3"W) approximately 150 m from shore in the same environmental conditions that the 186 187 farm raises its oysters. Oysters cultured commercially on the farm are normally caged within 188 industry baskets suspended above a mudflat that is frequented by oyster drills (Stramonita sp.), 189 black drum (Pogonias cromis), sheepshead (Archosargus probatocephalus), and a variety of 190 brachyuran crabs including mud crabs (Panopeus sp.), stone crabs (Menippe adina), and blue 191 crabs (*Callinectes sapidus*). Here, predators, particularly oyster drills, are most prevalent in the 192 summer months, but species like blue crabs can also be common throughout the year (Laughlin

193 1982, Butler 1985). The oyster frames were designed to keep spat~15 cm above the sediment surface to prevent sediment from covering and suffocating individuals (observations of frames 194 showed that numerous crabs, fish, and oyster drills were still able to reach all spat locations 195 using this setup). Frames were set parallel to the shoreline with at least 0.5 m separating each 196 197 frame. Oysters raised in the hatchery for four weeks were placed in the field on June 25, 2019 198 while oysters raised for eight weeks were planted on July 26, 2019 adjacent to the oysters 199 planted earlier. Once planted, all spat were checked for individual survival approximately every 48 – 72 hrs for 30 days by counting the number of spat still alive on each shell. The experiment 200 was concluded after this timeframe due to the high mortality experienced in the field. 201

202 We assessed whether oyster survival was influenced by the fixed effects of predator cue exposure, culture time, and caging status using a mixed-effects Cox proportional hazards model 203 204 (i.e. a survival analysis; R package: frailtyHL). All interactions were initially included in the model 205 and nonsignificant interactions were removed stepwise, from the most complex interaction 206 terms to the simplest, following the protocol of Crawley (2013) to help resolve the significance of main effects and achieve the lowest Akaike information criterion (AIC) value. Oyster shells, 207 nested in shell pair, nested in PVC frame were treated as random effects to control for 208 209 nonindependence among individuals. This model allowed us to right censor the data to account 210 for spat that were not dead by the end of the trial. A Cox proportional hazards analysis is a 211 statistical model which recognizes that the highest values in a study may simply be the maximum possible value, because a result did not occur by the end of the observation period, 212 so the model weighs the data points accordingly (i.e. the data are right censored). 213

3. Results

215 *3.1 Shell morphology*

216 Oyster spat shells were significantly stronger when grown with predator cues than 217 controls grown without predator cues (estimate = 0.23, t = 3.76, p < 0.0001). After four weeks 218 of cue exposure, shells were on average 41% stronger than comparable control shells and 63% stronger than comparable controls after eight weeks of cue exposure (Figure 1a). Time grown in 219 220 the hatchery also had a significant effect on shell strength. Oysters raised for 8 weeks were 34% 221 stronger than those grown for 4 weeks (estimate = 0.18, t = 4.71, p < 0.001). Thus, oysters 222 grown for four weeks with predator cues had shell strengths comparable to growing oysters 223 eight weeks without cues. There was not a significant interaction between cue exposure and 224 time in the hatchery (estimate = 0.01, t = 0.20, p = 0.840).

Interestingly, shell weight exhibited a significant interaction between cue exposure treatment and growth time (estimate = 5.28, t = 2.29, p = 0.022). Although oysters grown with and without predator cues had the same weight shells after four weeks of growth, shells of oysters grown with predator cues for eight weeks were 15% heavier than those grown without cues (Figure 1b). On average, shells became 2.5x heavier after an additional four weeks of growth (estimate = 6.50, t = 2.80, p = 0.005).

The size of shells also exhibited a significant interaction between cue exposure treatment and growth time (estimate = -0.01, t = -2.80, p = 0.005). Oysters induced with predator cues for four weeks were, on average, 10% larger than controls not exposed to cues, but after eight weeks in the hatchery, predator induced oysters were 10% smaller than controls (Figure 1c). However, there was not a significant difference in shell size between predator cue treatments for either time period (estimate = 0.01, t = 1.94, p = 0.052). Shells, on average, grew 46% larger with an additional four weeks of culture time (estimate = 0.02, t = 27.74, p < 0.001).

238 *3.2 Field survival*

239 In total, only 102 (13%) of caged oyster spat died, while 2124 (88%) of the uncaged 240 oysters died after 30 days in the field (hazard ratio = 28.06, 95% CI = 21.11 - 37.31, z = 22.93, p 241 < 0.001). Most cage mortality could easily be attributed to predators that had breached the 242 cage and were contained therein. Exposure to predator cues in the hatchery significantly affected oyster survivorship, regardless of exposure time (hazard ratio = 1.50, 95% CI = 1.12 -243 2.02, z = 2.71, p = 0.007; analysis of full dataset; Figure 2). However, predator cues only 244 substantially enhanced survival over uninduced oysters when individuals were unprotected. 245 246 Caged oysters exhibited relatively similar survival rates across induction treatments. This difference in survival of uncaged cue induced oysters over uninduced oysters grew 247 248 geometrically over time in the field (Figure 3). Additionally, the survival benefits from cue 249 exposure were more pronounced when oysters were induced with cues for 8 weeks rather than 250 4 weeks. While survivorship of oysters induced with cues for four weeks was ~50% greater than 251 uninduced oysters, eight weeks of cue exposure produced a nearly 300% increase in survival 252 after 30 days in the field. Interestingly, oysters that were grown in the hatchery for eight weeks had 21% greater overall mortality after 30 days in the field than those grown for only four 253 254 weeks in the hatchery (hazard ratio = 3.5495% CI = 2.49 – 5.04, z = 7.02, p < 0.001; Figure 2; Table 1). There was not a significant interaction between cue exposure treatment and time in 255 256 the hatchery on oyster survival (hazard ratio = 1.08, 95% CI = 0.72 - 1.63, z = -0.37, p = 0.710).

4. Discussion

These results demonstrate that oysters can readily be induced to grow stronger shells in 258 mass guantities and that this treatment can substantially increase survival rates in the field. The 259 difference in survival rates between caged and uncaged oysters indicates the primary source of 260 mortality for our oysters was predation. Indeed, most instances of mortality in the cages 261 coincided with predators also being found trapped within the cages. These findings, coupled 262 with the reduced benefits of induction in the caged treatment, also suggest that differences in 263 264 survival rate between induced and uninduced oysters was due to differences in predation rate, 265 consistent with previous laboratory studies (Robinson et al. 2014, Sherker et al. 2017, Ponce et 266 al. 2020).

267 Surprisingly, absolute survivorship was lowest for oysters grown in the hatchery for eight weeks rather than four weeks (Figure 2), despite the larger size and stronger shells of the 268 269 eight-week old oysters (Figure 1). This is likely due to a seasonal shift in the local predator 270 regime. When assessing survival of the eight-week oysters, we frequently observed oyster drills among our samples but rarely encountered them when surveying the four-week oysters that 271 272 had been deployed a month earlier (personal observations). Oyster drills are considered one of the main impediments to profitable oyster aquaculture in many otherwise suitable regions of 273 274 the northern Gulf of Mexico and are generally more abundant later in the summer, after spring rains (Butler 1985). Critically, even in the presence of high levels of this voracious predator, we 275 276 observed a 300% increase in survival of induced oysters over uninduced oysters. However, the 277 oyster drills' sudden appearance here and subsequent drastic increase in overall oyster

278 mortality highlights the importance of extended field assessments when estimating species
279 survival probability or the suitability of a region for aquaculture or restoration.

280 Extremely high juvenile mortality is a common phenomenon among r-selected species, like oysters, which often rely on producing enough offspring so that they can overwhelm 281 predators (Pianka 1970, Bishop and Peterson 2006). Consequently, reef restoration efforts 282 283 frequently involve planting millions to billions of oyster spat to increase the likelihood of 284 establishment of new reefs in regions where recruitment is limited (Brumbaugh and Coen 2009, La Peyre et al. 2014). Although few of our oysters survived longer than one month in the field, 285 the 50-300% greater survivorship of induced oysters over uninduced oysters, coupled with 286 these differences growing progressively larger over time, indicate that applying predator cues in 287 the hatchery can potentially cause dramatic increases in the efficiency of oyster aquaculture, 288 289 particularly when utilized at the scale of commercial bottom production or reef restoration 290 projects. This technique was effective in increasing survival even when predation pressure was 291 intense (Figures 2 and 3). Applying predator cues in the nursery may allow oysters to be grown cost-effectively in some regions which would normally have prohibitively high predation, 292 although more research is necessary to determine the extent to which return on investment for 293 294 cue exposure varies over space and time. Additionally, further assessment is necessary to 295 determine oyster survival when only induced spat are available. While many prey species such 296 as small crabs will likely have trouble breaking toughened shells and will cease feeding on stocks, species like oyster drills that can bore into shells may simply just expend more effort 297 298 consuming induced spat.

299 Interestingly, caging oysters caused the most dramatic increases in survival, highlighting the value of this well-established practice. Although caging oysters and situating operations in 300 301 locations with low predation pressure are common techniques (Matthiessen 2001; Wijsman et 302 al. 2019), these options are not always feasible. Maintaining cages is labor intensive and does 303 not lend itself to large-scale production necessary to meet market demand. Choosing sites with 304 low predation pressure is often a goal of aquaculture and reef restoration but has its own 305 difficulties as such sites may be unavailable or have poor growing conditions. Further, predation pressure within areas can vary substantially among seasons and years making site selection 306 307 challenging. Our results on oyster survival indicate that cue induction may therefore be best suited for these scenarios where oysters are kept uncaged (e.g. restoration projects, on-bottom 308 309 stock supplementation) or when predation pressure is high or unknown.

310 This is one of the first attempts to induce a bivalve species to grow stronger shells under 311 aquaculture conditions. As such, we sought to maximize the potential oyster induction 312 response by feeding oysters to predators daily and using blue crabs. However, a number of different common, noncommercial predator species are known to induce oysters to grow 313 stronger shells, including mud crabs (Robinson et al. 2014), oyster drills (Lord and Whitlatch 314 315 2012), and conchs (Gosnell et al. 2017). Induction responses can also be obtained by feeding 316 predators tissue from a variety of different animals (Scherer et al. 2016). Thus, the cost and 317 efficiency of applying cues to oysters may readily be improved upon by using locally available resources and through additional studies comparing feeding regimes and predator species. 318 Maintaining oysters in a nursery system with predator cues would incur additional economic 319 320 outlay on top of normal farming practices. Nevertheless, many hatcheries maintain spat for

about two weeks before leaving the facility and a number of nursery operations already hold
spat for a month to help oysters reach a size refuge from predation (Matthiessen 2001, Mao et
al. 2019). For these existing time frames, the costs of also providing cues should be minimal,
but cost-benefit analyses are necessary to evaluate the economic viability of this technique,
especially if facility holding times are to be altered as a result.

326 Induced defenses frequently arise at the costs of reduced growth (Kats and Dill 1998, 327 Cronin 2001), slower development (Steiner 2007), and decreased reproductive effort (Lima 328 2009) as resources are shunted towards avoiding predation. Few studies have investigated the 329 amount induced defenses alter oyster somatic tissue production or reproductive output. Gosnell et al. (2017) found that after 58 days of continuous predator exposure, oysters 330 exhibited 20% lower soft tissue mass than controls, but no significant change in the percent 331 332 composition of soft tissue versus shell. Our oysters after both one and two months of cue 333 exposure had the same sized shells, but appeared to be exhibiting slight reductions in growth after two months exposure (Figure 1). As oysters take one to three years to reach harvestable 334 size depending primarily on food availability and water temperature (Matthiessen 2001), any 335 early decreases in soft tissue have a good probability of becoming negligible. However, more 336 research is necessary to quantify the degree cue induction affects oysters at adulthood and 337 338 pinpoint the predator exposure time which maximizes total oyster production.

In conclusion, high mortality from predation plagues the bivalve aquaculture industry (Matthiessen 2001, Gosling 2008, Wijsman et al. 2019) and hinders reef restoration efforts (Mann and Powell 2007). Additionally, many bivalve species commonly cultured by the industry are known to grow stronger shells in the presence of predators (Leonard et al. 1999, Nakaoka

2000, Bishop and Peterson 2006, Robinson et al. 2014). Exposing juvenile bivalves to predator 343 cues in the nursery stage is therefore a promising tool which likely can provide a variety of 344 benefits across the industry as even a small relative increase in survival can change the 345 economics of bivalve aquaculture; causing private operations to be more profitable (or 346 347 profitable at all) as well as improve the return on investment in restoration efforts. Acknowledgements 348 349 We thank Kelly Correia, Randi Cannon, Dom Latona, and Dr. Jessica Lunt for their 350 logistical support in setting up the experiment as well as Scott Rikard and the Auburn University Shellfish Laboratory for their invaluable expertise and facility access. We also thank Steve 351 352 Crockett for providing space at the Point aux Pins oyster farm. This publication was supported by the U.S. Department of Commerce's National Oceanic and Atmospheric Administration 353 354 under NOAA Award NA18OAR4170080, the Mississippi-Alabama Sea Grant Consortium program development grant R/SFA-19-PD, and the National Science Foundation REU Program 355 grant number 1838618. The views expressed herein do not necessarily reflect the views of any 356 357 of these organizations. 358 359 360 361 362

References

364	Anderson, M. J., and Connell, S. D. 1999.	Predation by fish on intertidal oysters. Marine Ecology
365	Progress Series 187:203-211.	

Beck, M.W., Brumbaugh, R.D., Airoldi, L., Carranza, A., Coen, L.D., Crawford, C., et al. 2011.

367 Oyster reefs at risk and recommendations for conservation, restoration, and
 368 management. Bioscience 61:107-116.

369 Bishop, M.J., and Peterson, C. H. 2006. When r-selection may not predict introduced-species

370 proliferation: Predation of a nonnative oyster. Ecological Applications 16:718-730.

Bolker, B., M. Brooks, C. Clark, S. Geange, J. Poulsen, M.H. Stevens, and J.S. White. 2009.

Generalized linear mixed models: A practical guide for ecology and evolution. Trends in
Ecology and Evolution 24:127-135.

Brown, K. M., George, G. J., Peterson, G. W., Thompson, B. A., and Cowan, J. H. 2008. Oyster
 predation by black drum varies spatially and seasonally. Estuaries and Coasts 31:597-

376 604.

Brumbaugh, R.D., and Coen, L.D. 2009. Contemporary approaches for small-scale oyster reef
 restoration to address substrate versus recruitment limitation: a review and comments
 relevant for the Olympia oyster, *Ostrea lurida* Carpenter 1864. Journal of Shellfish
 Research 28:147-161.

381	Buitrago, J., Rada, M., Hernández, H., and Buitrago, E. 2005. A single-use site selection
382	technique, using GIS, for aquaculture planning: choosing locations for mangrove oyster
383	raft culture in Margarita Island, Venezuela. Environmental Management 35: 544-556.
384	Butler, P.A. 1985. Synoptic review of the literature on the southern oyster drill Thais
385	haemastoma floridana. NOAA Technical Report NMFS 35. 1-9.
386	Calderwood, J., O'Connor, N.E., and Roberts, D. 2016. Efficiency of starfish mopping in reducing
387	predation on cultivated benthic mussels (Mytilus edulis Linnaeus). Aquaculture 452:88-
388	96.
389	Cohen, J. 1988. Statistical Power Analysis for the Behavioral Sciences, second ed. New York,
390	New York.
391	Congrove, M.S., Wesson, J.A., Allen, S.K. 2009. A practical manual for remote setting in Virginia.
392	VIMS Marine Resource Report No. 2009-1. Sea Grant Communications. 1-21.
393	Crawley, M.J. 2013. Generalized linear models. In The R book (2nd ed.). Chichester, U.K.: John
394	Wiley. pp. 557-578.
395	Cronin, G. 2001. Resource allocation in seaweeds and marine invertebrates: chemical defense
396	patterns in relation to defense theories. Marine chemical ecology 325-353.
397	FAO. 2019. FAO yearbook. Fishery and Aquaculture Statistics 2017/FAO annuaire. Statistiques
398	des pêches et de l'aquaculture 2017/ FAO anuario. Estadísticas de pesca y acuicultura
399	2017. 109 p. Rome/Roma.

- Gosling, E. (2008). Bivalve molluscs: biology, ecology and culture. John Wiley & Sons. pp. 284327.
- 402 Gosnell, J.S., Spurgin, K., and Levine, E.A. 2017. Caged oysters still get scared: Predator presence
- 403 and density influence growth in oysters, but only at very close ranges. Marine Ecology
 404 Progress Series 568:111-122.
- Grabowski, J.H., and Peterson, C.H. 2007. Restoring oyster reefs to recover ecosystem services.
 Ecosystem Engineers: Plants to Protists 4:281-298.
- 407 Hancock, D. A. 1955. The feeding behaviour of starfish on Essex oyster beds. Journal of the
- 408 Marine Biological Association of the United Kingdom 34: 313-331.
- Kats, L.B., and Dill, L.M. 1998. The scent of death: chemosensory assessment of predation risk
 by prey animals. Ecoscience 5:361-394.
- La Peyre, M., Furlong, J., Brown, L. A., Piazza, B. P., and Brown, K. 2014. Oyster reef restoration
- in the northern Gulf of Mexico: extent, methods and outcomes. Ocean & Coastal
- 413 Management 89:20-28.
- Lappin Jr., D. M. 2018. Remote set of *Crassostrea virginica* as a potential means for public stock
- 415 enhancement in Alabama, and the assessment of larval tank setting distributions.
- 416 Master's Thesis, Auburn University. 1-81.
- 417 Laughlin, R.A. 1982. Feeding habits of the blue crab, *Callinectes sapidus* Rathbun, in the
- 418 Apalachicola Estuary, Florida. Bulletin of Marine Science 32:807-822.

419	Leonard, G.H., Bertness M.D., and Yund P.O. 1999. Crab predation, waterborne cues, and
420	inducible defenses in the blue mussel, Mytilus edulis. Ecology 80:1-14.
421	Lima, S.L. 2009. Predators and the breeding bird: behavioral and reproductive flexibility under
422	the risk of predation. Biological Reviews 84:485-513.
423	Lord, J.P., and Whitlatch, R.B. 2012. Inducible defenses in the eastern oyster Crassostrea
424	virginica Gmelin in response to the presence of the predatory oyster drill Urosalpinx
425	cinerea Say in Long Island Sound. Marine Biology 159:1177-1182.
426	Mackenzie, C.L. 1970. Oyster culture in Long Island Sound 1966-69, in: Edelsberg, E., Lundy, B.
427	(Eds.), Commercial Fisheries Review 32. Fish and Wildlife Service of the United States
428	Department of the Interior, Arlington, pp. 27-40.
429	Mann, R., and Powell, E.N. 2007. Why oyster restoration goals in the Chesapeake Bay are not
430	and probably cannot be achieved. Journal of Shellfish Research 26:905-917.'
431	Mann, R., Southworth, M., Harding, J.M., and Wesson, J.A. 2009. Population studies of the
432	native Eastern oyster, Crassostrea virginica, (Gmelin, 1791) in the James River, Virginia,
433	USA. Journal of Shellfish Research 28:193-220.
434	Mao, Y., Lin, F., Fang, J., Fang, J. Li, J., and Du, M. 2019. Bivalve production in China. In Goods
435	and services of marine bivalves (pp. 51-72). Springer, Cham.

- 436 Matthiessen, G.C. 2001. Oyster Culture (vol 2). Fishing New Books, Oxford, UK. p. 1-162.
- 437 McDermott, J. J. 1960. The predation of oysters and barnacles by crabs of the family Xanthidae.
- 438 Proceedings of the Pennsylvania Academy of Science 34:199-211.

439	Michaelis, A.K., Walton, W.C., Webster, D.W., and Shaffer, L. J. 2020. The role of ecosystem
440	services in the decision to grow oysters: A Maryland case study. Aquaculture
441	529:735633.
442	Nakaoka, M. 2000. Nonlethal effects of predators on prey populations: Predator—mediated
443	change in bivalve growth. Ecology 81:1031-1045.Newell, R.I.E., Kennedy, V.S., and Shaw,
444	K.S. 2007. Comparative vulnerability to predators, and induced defense responses, of
445	eastern oysters Crassostrea virginica and non—native Crassostrea ariakensis oysters in
446	Chesapeake Bay. Marine Biology 152:449-460.
447	Pianka, E.R. 1970. On r-and K-selection. The American Naturalist, 104: 592-597.
448	Ponce, M., Belgrad, B.A., Walton, W., and Smee, D.L. 2020. Nursery exposure of oyster spat to
449	different predators strengthens oyster shells. Gulf and Caribbean Research 31:SC36-
450	SC40.
451	Richard, M., Forget, F., Mignucci, A., Mortreux, S., Le Gall, P., Callier, M. D., et al. 2020. Farmed
452	bivalve loss due to seabream predation in the French Mediterranean Prevost Lagoon.
453	Aquaculture Environment Interactions 12:529-540.
454	Robinson, E.M., Lunt, J., Marshall, C.D., and Smee, D.L. 2014. Eastern oysters Crassostrea
455	virginica deter crab predators by altering their morphology in response to crab cues.
456	Aquatic Biology 20:111-118.

457	Scherer, A.E., Lunt, J., Draper, A.M., and Smee, D.L. 2016. Phenotypic plasticity in oysters
458	(Crassostrea virginica) mediated by chemical signals from predators and injured prey.
459	Invertebrate Biology 135:97-107.
460	Scherer, A.E., Bird, C.E., McCutcheon, M.R., Hu, X., and Smee, D.L. 2018. Two-tiered defense
461	strategy may compensate for predator avoidance costs of an ecosystem engineer.
462	Marine Biology 165:131.
463	Sherker, Z.T., Ellrich, J.A., and Scrosati, R.A. 2017. Predator-induced shell plasticity in mussels
464	hinders predation by drilling snails. Marine Ecology Progress Series 573:167-175.
465	Steiner, U.K. 2007. Investment in defense and cost of predator-induced defense along a
466	resource gradient. Oecologia, 152:201-210.
467	Weissburg, M., Smee, D.L., and Ferner, M.C. 2014. The sensory ecology of nonconsumptive
468	predator effects. The American Naturalist 184:141-157.
469	Wijsman, J.W.M., Troost, K., Fang, J., and Roncarati, A. 2019. Global production of marine
470	bivalves. Trends and challenges. In Goods and services of marine bivalves (pp. 7-26).
471	Springer, Cham.
472	Zu Ermgassen, P.S.E., Gray, M.W., Langdon, C.J., Spalding, M.D., and Brumbaugh, R.D. 2013.
473	Quantifying the historic contribution of Olympia oysters to filtration in Pacific Coast
474	(USA) estuaries and the implications for restoration objectives. Aquatic Ecology, 47:149-
475	161.

Table Legend

478	Table 1: Proportion of spat surviving after 30 days in the field for each experimental treatment.
-----	--

479

Figure Legends

480 **Figure 1:** Oyster spat shell characteristics when reared in the hatchery for four and eight weeks

481 in either the presence (induced) or absence of predator cues (uninduced)(n = 84 per

482 treatment). Mean ± SE A) shell crushing force standardized by shell size (N/mm), B) shell weight

483 (g), and **C)** shell diameter (mm). Letters denote significant differences.

484 **Figure 2:** Survivorship curve of the proportion of individual oysters (*Crassostrea virginica*) which

survived each day in the field as the experiment progressed. Oysters were reared in the

486 hatchery for either four weeks or eight weeks prior to being released into the field. Line color

denotes whether oysters were exposed to predator cues (induced) or no cues (uninduced) in

the hatchery while line shape denotes whether oysters were caged (n = 200 per treatment) or

489 uncaged (n = 600 per treatment) in the field.

Figure 3: Percent increase in survivorship of uncaged induced oysters over uncaged uninduced
oysters after a month in the field. Oysters were exposed (induced) to predator cues for either
one or two months.

493

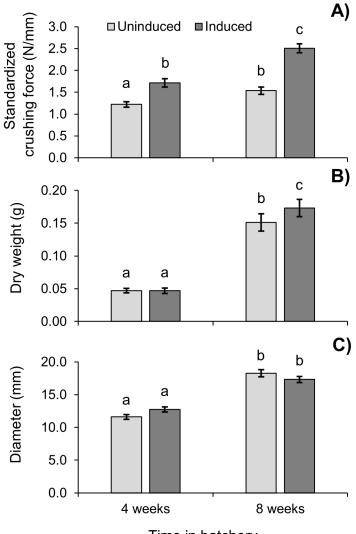
494

495

496

497 Table 1

	4 weeks in hatchery		8 weeks in hatchery	
	induced	uninduced	induced	uninduced
uncaged	0.24	0.15	0.04	0.01
caged	0.81	0.79	0.97	0.96



Time in hatchery

