

1 Title: Macrofaunal communities associated with chemosynthetic habitats from the U.S. Atlantic  
2 margin: a comparison among depth and habitat types

3

4 Authors: Jill R. Bourque<sup>a\*</sup>, Craig M. Robertson<sup>b</sup>, Sandra Brooke<sup>c</sup>, Amanda W.J. Demopoulos<sup>a</sup>,

5 Affiliations: <sup>a</sup> U.S. Geological Survey Wetland and Aquatic Research Center; <sup>b</sup> Bangor

6 University; <sup>c</sup> Florida State University

7

8 \* Corresponding Author; 7920 NW 71<sup>st</sup> ST, Gainesville, FL, 32653, USA; jbourque@usgs.gov

9

10 Abstract:

11 Hydrocarbon seeps support distinct benthic communities capable of tolerating extreme  
12 environmental conditions and utilizing reduced chemical compounds for nutrition. In recent  
13 years, several locations of methane seepage have been mapped along the U.S. Atlantic  
14 continental slope. In 2012 and 2013, two newly discovered seeps were investigated in this  
15 region: a shallow site near Baltimore Canyon (BCS, 366-412 m) and a deep site near Norfolk  
16 Canyon (NCS, 1467-1602 m), with both sites containing extensive chemosynthetic mussel bed  
17 and microbial mat habitats. Sediment push cores, suction samples, and Ekman box cores were  
18 collected to quantify the abundance, diversity, and community structure of benthic macrofauna  
19 (>300 µm) in mussel beds, mats, and slope habitats at both sites. Community data from the deep  
20 site were also assessed in relation to the associated sediment environment (organic carbon and  
21 nitrogen, stable carbon and nitrogen isotopes, grain size, and depth). Infaunal assemblages and  
22 densities differed both between depths and among habitat types. Macrofaunal densities in

1

23 microbial mats were four times greater than those present in mussel beds and slope sediments  
24 and were dominated by the annelid families Dorvilleidae, Capitellidae, and Tubificidae, while  
25 mussel habitats had higher proportions of crustaceans. Diversity was lower in BCS microbial mat  
26 habitats, but higher in mussel and slope sediments compared to NCS habitats. Multivariate  
27 statistical analysis revealed specific sediment properties as important for distinguishing the  
28 macrofaunal communities, including larger grain sizes present within NCS microbial mat  
29 habitats and depleted stable carbon isotopes ( $\delta^{13}\text{C}$ ) in sediments present at mussel beds. These  
30 results suggest that habitat differences in the quality and source of organic matter are driving the  
31 observed patterns in the infaunal assemblages, including high  $\beta$  diversity and high variability in  
32 the macrofaunal community composition. This study is the first investigation of seep infauna  
33 along the U.S. Atlantic slope north of the Blake Ridge Diapir and provides a baseline for future  
34 regional comparisons to other seep habitats along the Atlantic margin.

35

36 **Highlights:**

- 37 ● First investigation of seep infaunal communities in U.S. mid-Atlantic margin north of  
38 Blake Ridge at multiple depths
- 39 ● Microbial mats and mussel bed habitats support locally high densities of infauna
- 40 ● High  $\beta$  diversity over small and large spatial scales
- 41 ● Stable carbon isotopic composition ( $\delta^{13}\text{C}$ ) and mud content explained the most variation  
42 among NCS seep and non-seep habitats

43

44

45 **1. Introduction:**

46 Cold seeps occur worldwide, often where methane or sulfide is forced upward through  
47 the sediment by pressure gradients (Levin, 2005). Anaerobic oxidation of methane and sulfate  
48 reduction results in the formation of carbonates and often high concentrations of hydrogen  
49 sulfide in sediments, which is toxic to most fauna (Vetter et al., 1991). The flow of seep  
50 products through sediments often results in recognizable biogenic habitats, including mussel and  
51 clam beds, microbial mats, and tube worm aggregations (Bernardino et al., 2012), where the  
52 dominant megafauna are dependent on chemoautotrophic endosymbiotic bacteria for nutrition  
53 (Kochevar et al., 1992). In addition, the physical structure created by chemosynthetic organisms  
54 provides heterogeneous habitat for diverse communities (Bergquist et al., 2003; Van Dover and  
55 Trask, 2000); thus these organisms serve as ecosystem engineers (e.g., Jones et al., 1996).

56 Sediment fauna associated with seep communities, including microbial mats and clam  
57 beds, have been studied in many locations worldwide (Cordes et al., 2010b; Levin, 2005; Sibuet  
58 and Olu, 1998); however, sediments associated with mussel habitats have only been examined at  
59 a few locations, including the Blake Ridge Diapir (Robinson et al., 2004) and the Gulf of Guinea  
60 (Menot et al., 2010). Densities of macrofauna in seep sediments are often higher than in  
61 background non-seep sediments, particularly at increasing water depth (Levin, 2005) where food  
62 is often a limited resource and seep-derived carbon provides an additional food source (Levin  
63 and Michener, 2002). Globally, however, density differences among seep habitat types has been  
64 variable (Bernardino et al., 2012), with microbial mat, clam beds, or mussel beds exhibiting

65 similar (Levin et al., 2010) or differing densities in comparison to each another (Levin et al.,  
66 2015; Menot et al., 2010; Robinson et al., 2004; Sahling et al., 2002). At the Blake Ridge Diapir,  
67 macrofaunal densities in sediments near mussels were higher than in microbial mat sediments,  
68 although macrofaunal densities were low for all sampled habitats (0-6,400 ind. m<sup>-2</sup>; Robinson et  
69 al., 2004). High densities found in microbial mat habitats have been attributed to the exploitation  
70 of the chemosynthetically derived food source by seep tolerant taxa, and has been compared to  
71 similar faunal responses from disturbance and sediment organic enrichment events (Bernardino  
72 et al., 2012; Sahling et al., 2002).

73         Macrofaunal diversity patterns among seep and non-seep habitats have also been  
74 variable. Microbial mat habitats often exhibit low diversity and high dominance of a few tolerant  
75 taxa compared to other seep and non-seep habitats due to high sediment sulfide concentrations  
76 (Levin et al., 2003; Sahling et al., 2002). However, low sulfide concentrations in clam beds on  
77 the California slope led to increased macrofaunal diversities by supporting populations of both  
78 ambient and sulfophilic taxa (Levin et al., 2003). In other locations, macrofaunal diversity in  
79 sediments associated with clam beds has been similar (Hydrate Ridge, Sahling et al., 2002) or  
80 lower (Gulf of Guinea, Menot et al., 2010) than non-seep habitats. Differences in the diversity of  
81 mussel-bed sediment macrofauna have also been mixed, and possibly related to site-specific  
82 factors. Higher diversities were found in mussel-associated sediments compared to microbial  
83 mats and non-seep sediments at Blake Ridge (Robinson et al., 2004); in contrast, diversity was  
84 lower in sediments adjacent to mussels than non-seeps in the Gulf of Guinea (Menot et al.,  
85 2010). High  $\beta$  diversity has also been reported for seep habitats, supported by the habitat

86 heterogeneity of foundation species and the varying geochemical environments (Cunha et al.,  
87 2013; Cordes et al., 2010b).

88 Infaunal community assemblages associated with different seep habitats are distinct from  
89 one another (Bernardino et al., 2012; Levin, 2005; Menot et al., 2010) and differ from  
90 background non-seep sediments. Dorvilleid polychaetes are common in seep habitats (Levin,  
91 2005) and are particularly abundant in microbial mat habitats, which is attributed to their broad  
92 environmental tolerance to sulfide concentrations and opportunistic lifestyle (Levin et al., 2013;  
93 Levin et al., 2006; Levin et al., 2003; Robinson et al., 2004; Sahling et al., 2002). Other  
94 characteristic seep macrofauna include the polychaete families Siboglinidae, Capitellidae, and  
95 Ampharetidae, oligochaetes, and thyasirid bivalves (Bernardino et al., 2012; Dando et al., 1991;  
96 Levin et al., 2000; Levin et al., 2003), some of which can benefit from reducing habitats (Levin  
97 et al., 2000). At Blake Ridge, mussel sediment communities were more similar to non-seep  
98 communities (60% similar) than to microbial mat communities (11-54%), suggesting that  
99 mussels help maintain low concentrations of methane and sulfide in sediments, facilitating  
100 communities more similar to non-seep sediments (Robinson et al., 2004). The specific mussel  
101 species present in the sediments has been found to influence the associated faunal community  
102 assemblage (Cordes et al., 2010a). Mussels in the *Bathymodiolus childressi*-complex contain  
103 only methanotrophic symbionts, while those of the *Bathymodiolus boomerang*-complex,  
104 including *Bathymodiolus heckerae* present at Blake Ridge, contain both methanotrophs and  
105 thiotrophs (Olu-LeRoy et al., 2007b). The effect of mussel species composition on epifaunal  
106 communities has been suggested to act as a proxy for habitat chemistry (Cordes et al., 2010a),

107 since the resources (i.e., methane or sulfide) required by the endosymbionts of the different  
108 mussel types are likely indicative of the chemical environment surrounding the mussels. The  
109 extent of endemic infauna in seep habitats globally is still unresolved (Bernardino et al., 2012),  
110 but may be a function of depth (Cordes et al., 2010a; Levin, 2005; Sahling et al., 2003), with  
111 many species occupying seep sediments comprised of the regionally available taxon pool (e.g.,  
112 Levin, 2005).

113         Depth-related patterns have been observed among seep sites worldwide, with  
114 communities at upper bathyal depths (200-1500m) distinct from those at deeper depths (>1500m;  
115 Bernardino et al., 2012). However, there are few comparisons of seeps with depths ranging  
116 >1000m within a geographic region (Cunha et al., 2013; Rodrigues et al., 2013; Cordes et al.,  
117 2010a; Sahling et al., 2003), where other factors structuring deep-sea communities (e.g., food  
118 availability, bottom water oxygen concentrations) are more directly comparable. In the Gulf of  
119 Cadiz, shallow (353-732m) mud volcanoes exhibited higher diversity, lower heterogeneity, and  
120 different community assemblages from deep (1318-3860m) mud volcanoes (Cunha et al., 2013).  
121 Cordes et al. (2010a) observed distinct depth-related patterns in both mussel and tube worm  
122 communities in the Gulf of Mexico (1005-2746m), with mussel beds exhibiting a mid-slope  
123 (~1000m) diversity maximum, consistent with similar trends for soft-sediment communities.  
124 However, the relationship between depth and the sediment communities found within microbial  
125 mats and along the fringe of mussel beds within a particular region is unknown.

126         The distinct epifaunal and infaunal assemblages present in seep habitats are a function of  
127 their proximal sediment geochemical environment (Levin et al., 2003; Sibuet and Olu, 1998),

128 including seepage rates, sulfide concentrations, and biological activity (Cordes et al., 2010a; Olu  
129 et al., 2009; Levin, 2005; Sahling et al., 2002). Microbial mats often form in habitats with high  
130 methane flux rates, with corresponding high sulfide concentrations and low oxygen penetration  
131 into the sediment (Sahling et al., 2002). In contrast, habitats that support clam beds exhibit lower  
132 but variable methane flow through sediments, lower sulfide concentrations, and higher oxygen  
133 penetration through bioturbation (Levin et al., 2003). Comparable data from mussel beds is  
134 limited; however, while these habitats can have similar oxygen penetration profiles, their  
135 associated sediment organic carbon concentrations and methane concentrations in the overlying  
136 water (Menot et al., 2010; Olu-Le Roy et al., 2007a) are often higher than clam beds. Due to  
137 variations in seep activity and fluid flux, the sediment geochemical properties (e.g., organic  
138 carbon and nitrogen, stable carbon and nitrogen isotopes, grain size) often differ between seep  
139 and non-seep habitats (Levin et al., 2000; Levin et al., 2010; Menot et al., 2010; Valentine et al.,  
140 2005). Microbial mats have been documented to contain higher percent carbon content, high  
141 carbon to nitrogen (C:N) ratios, and lower percent nitrogen content than clam beds and non-seep  
142 sediments (Levin et al., 2010). Clam and mussel beds also contain higher organic carbon content  
143 than non-seep sediments at multiple depths (Levin et al., 2000; Levin et al., 2010; Menot et al.,  
144 2010; Valentine et al., 2005).

145         Stable carbon isotopic ( $\delta^{13}\text{C}$ ) composition of sediments and fauna from seep habitats  
146 often reflects the primary nutritional sources available in the environment, where phytoplankton-  
147 derived organic matter typically produce  $\delta^{13}\text{C}$  values ranging from -25‰ to -15‰ (Fry and  
148 Sherr, 1984), very low  $\delta^{13}\text{C}$  values derived from biogenic methane ( $\leq -50\text{‰}$ ; Van Dover, 2007;

149 Whitticar, 1999), and carbon derived from sulfide oxidation with  $\delta^{13}\text{C}$  ranging from -37‰ to -  
150 27‰ (Brooks et al., 1987; Fisher, 1990; Robinson and Cavanaugh, 1995). In the Gulf of Mexico,  
151 sediments near seeps containing bacterial filaments were depleted in both  $^{13}\text{C}$  and  $^{15}\text{N}$  compared  
152 to those with no bacterial filaments present (Demopoulos et al., 2010). Stable isotope values of  
153 seep sediments can vary with seep activity, where higher methane fluxes near mytilid beds were  
154 associated with lower  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  as compared to clam beds and the sediments may contain  
155 different microorganism communities (Cambon-Bonavita et al., 2009; Olu-Le Roy et al., 2007a;  
156 Olu et al., 2009). Thus, light  $\delta^{13}\text{C}$  values can be a useful indicator of seep habitats where  
157 biogenic methane is present. While methane flux and sulfide concentrations are important  
158 mechanistic factors structuring seep faunal communities (Bernardino et al., 2012), stable isotopes  
159 and other sediment parameters also can serve as proxies for and provide insight into the  
160 mechanisms of seep activity occurring within sediments.

161         While originally thought to be unusual on the western Atlantic margin (Van Dover,  
162 2000), increasing numbers of seep areas have been documented since 2011 (Skarke et al.,  
163 2014). Prior to 2011, only two chemosynthetic seep areas were known, the Blake Ridge Diapir  
164 (Paull et al., 1995; Van Dover et al., 2003) and the Cape Fear Diapir (Brothers et al., 2013), both  
165 in deep water (2100-2600m) off of South Carolina, US. However, recent large-scale projects  
166 using high resolution multibeam sonar and backscatter data now document 570 seep areas  
167 between Cape Hatteras and Georges Bank (Skarke et al., 2014), and suggest that tens to  
168 thousands more may be present along the passive Atlantic margin. During this study, two  
169 recently identified chemosynthetic seep areas were examined near Baltimore Canyon (BCS) and



170 Norfolk Canyon (NCS) separated by 90 km. This study addresses the role of geographic setting,  
171 seep habitat type, and sediment geochemistry in determining infaunal densities, community  
172 composition, and diversity of sediment macrofauna (>300 $\mu$ m). We hypothesized that (i)  
173 communities found at seep and non-seep habitats will differ within sites (BCS and NCS) and  
174 between sites; (ii) similar seep habitats at BCS and NCS will exhibit similar community  
175 composition, and (iii) community patterns in seep and non-seep habitats will be related to  
176 sediment geochemical properties. To support our hypotheses, we expect higher macrofaunal  
177 density but lower diversity at shallower BCS than at deeper NCS, similar taxonomic composition  
178 between seep habitat types at BCS and NCS, and distinct sediment geochemical parameters  
179 associated with community assemblages in each habitat type.

180

## 181 **2. Methods:**

### 182 2.1 Study Area

183 Two large cold-seep communities were explored on the U.S. Mid-Atlantic margin in  
184 2012 and 2013 (Figure 1). The first seep, BCS, was located on the slope south of Baltimore  
185 Canyon at depths ranging 366 to 402m. First documented by Hecker et al. (1983) during towed  
186 camera surveys, the exact location was re-discovered in 2012 during this study. The second  
187 seep, NCS, was located south of Norfolk Canyon at depths ranging 1457 to 1602m. The NCS  
188 was identified by the Okeanos Explorer during multibeam mapping activities which detected  
189 active bubble plumes (Skarke et al., 2014). The BCS seep contained large, but patchy,  
190 communities of the bathymodiolid deep-sea mussels (Figure 2a), along with discrete white

191 microbial mats (Figure 2b) and large areas of shell debris. The NCS seep contained extensive  
192 mussel communities (Figure 2c), with areas of filamentous white (Figure 2d) and yellow  
193 microbial mats and shell debris. While the species has yet to be confirmed, initial examination  
194 of the specimens indicate that the mussel species at BCS and NCS are within the *B. childressi*-  
195 complex (Olu-Le Roy et al., 2007b) and will be referred to as *B. childressi*-complex throughout.

196

## 197 2.2 Sampling Procedures

198 Sediment samples were collected from seep habitats on two cruises (Table 1); one in  
199 2012 aboard the NOAA Ship *Nancy Foster* (17 Aug-14 Sep) and one in 2013 aboard the NOAA  
200 Ship *Ronald H. Brown* (2-18 May). Push cores (6.35-cm diameter) were collected in microbial  
201 mats, mussel habitats, and background soft-sediment habitats using the ROV *Kraken* (2012) and  
202 ROV *Jason II* (2013). Background soft-sediments were collected at NCS in the main axis of  
203 Norfolk Canyon using a NIOZ box core, which was sub-sampled with push core tubes (6.35-cm  
204 diameter). Bow wave effects on the box core were minimized by reducing the speed of descent  
205 of the box core as it approached the seafloor. Additionally, the NIOZ box corer completely seals  
206 upon triggering, preventing the loss of surface sediment layers, and only cores that had  
207 undisturbed surface layers were processed in this study. In addition, the sub-coring with push  
208 core tubes provides direct sample-size effort comparisons for our study, which are directly  
209 comparable to other seep studies (Levin and Mendoza, 2007; Levin et al., 2010; Robinson et al.,  
210 2004). Additional cores and non-quantitative suction samples were collected via ROV in 2013  
211 in microbial mats and mussel beds (Table 1). An Ekman corer (0.063 m<sup>2</sup>) was used to collect

212 mussel bed material at both BCS and NCS. Push cores were sectioned vertically (0-2, 2-5 cm)  
213 after recovery for either faunal or sediment geochemistry analysis. Due to time constraints and  
214 the limited number of possible core collections on the ROV, sediments from BCS were only  
215 processed for faunal analysis. Faunal core sections, Ekman samples, and suction samples were  
216 preserved whole in 10% buffered formalin solution until they were returned to the laboratory  
217 where they were stained with rose bengal and washed through a 300- $\mu$ m mesh sieve to retain the  
218 macrofauna portion. Macrofauna were sorted under a dissecting microscope and identified to the  
219 lowest practical taxonomic level, including family level for polychaetes, oligochaetes, peracarid  
220 crustaceans, and molluscs. Sediment geochemistry core fractions were frozen whole at -20°C  
221 until returned to the lab. Subsamples of geochemistry cores were analyzed for the stable isotopes  
222  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$ , and percent carbon and nitrogen. Sediment samples were homogenized prior to  
223 drying and acidified with 1.0 N phosphoric acid before weighing into tin boats. Samples were  
224 analyzed for  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  referenced to Vienna PeeDee Belemnite and atmospheric nitrogen  
225 gas, respectively. Analyses were conducted at Washington State University using a Costech  
226 (Valencia, USA) elemental analyzer interfaced with a GV instruments (Manchester, UK)  
227 Isoprime isotope ratio mass spectrometer. Isotope ratios were expressed in standard delta  
228 notation,  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$ , as per mil (‰). Grain size analysis was performed on fractions of the  
229 sediment geochemistry cores using the Folk method (Folk, 1974).

230

231 2.3 Data Analysis

232 Abundance of individuals and univariate measures of biodiversity were analyzed using  
233 one-way (within sites) and two-way (among sites) analysis of variance (ANOVA) with habitat  
234 type (microbial mat, mussels, background) and site as factors and individual cores as replicates,  
235 followed by post-hoc test Tukey's HSD for multiple comparisons. All data were tested for  
236 normality and heteroscedasticity using Shapiro-Wilk and Levene's tests (Zar, 1999) and  $\log_e$ -  
237 transformed when necessary. If transformation did not achieve normality, a non-parametric  
238 Kruskal-Wallis test was used on univariate measures with a post-hoc pairwise Wilcoxon test  
239 using a Holm correction for multiple comparisons. Depth relationships with abundance and  
240 diversity measures were tested using Spearman's rank correlation. A significance level of  $p <$   
241 0.05 was used in all tests. Univariate statistics were computed with the program R (R  
242 Development Core Team, 2011). Diversity was examined using Pielou's evenness ( $J'$ ), Shannon  
243 diversity ( $H' \log_e$ ), and ES(n) rarefaction based on untransformed abundance data using  
244 DIVERSE in PRIMER Statistical Software version 7 (Clarke and Gorley, 2015).

245 Community structure was assessed by examining the overall contribution of higher level  
246 taxa, composition of polychaete feeding guilds, and multivariate community analysis.  
247 Multivariate analysis of community structure across cores for sites and habitats was performed  
248 on square-root transformed abundance data using Bray-Curtis similarities in PRIMER version 7  
249 (Clarke and Gorley, 2015) with the PERMANOVA+ add on (Anderson et al., 2008). Samples  
250 collected via Ekman cores and suction were analyzed as presence/absence transformed  
251 abundance data within the multivariate analyses, with multiple suction samples from mussel  
252 habitats pooled. Communities were examined using one-way, two-way, and pairwise analysis of

253 variance by permutation (PERMANOVA) with distance-based tests for homogeneity of  
254 multivariate dispersions (PERMDISP). Similarity of percentages (SIMPER) was used to identify  
255 the taxa responsible for discriminating between sites and habitats, and to assess the variability of  
256 the communities within habitats. Variability among Bray-Curtis similarities within site-habitat  
257 combinations was also assessed using multivariate dispersion (MVDISP).

258 To address the relationship of the environmental variables to the multivariate community  
259 data, distance-based linear modeling (DistLM) and distance-based redundancy analysis (dbRDA)  
260 were performed using the PERMANOVA+ add on package to PRIMER 7. DistLM performs  
261 nominal tests of each variables explanatory power on community structure and builds a  
262 multivariate statistical model of explanatory power of a suite of variables when considered  
263 together. Environmental data was only collected at NCS, thus analysis was limited to only the  
264 deep site. Variables included were depth, mud content, stable isotopic composition ( $\delta^{13}\text{C}$  and  
265  $\delta^{15}\text{N}$ ), and organic carbon content. Organic nitrogen content was excluded from the analysis due  
266 to high correlation ( $>0.95$ ) with organic carbon content to reduce redundancy.

267

### 268 **3. Results:**

#### 269 3.1 Density

270 A total of 2,609 individuals were collected from cores in our study, encompassing 86  
271 taxa, including 35 polychaete families, 22 crustacean families, 20 mollusca families, and 9 other  
272 taxa (see Supplemental Table 1 for full list). A further 2,518 individuals were collected from  
273 suction and Ekman grab samples, including 21 additional taxa. Macrofaunal density was

274 significantly higher at BCS than at NCS for all habitat types (Figure 3; Two-way ANOVA,  
275  $F_{1,19}=11.34$ ,  $p=0.003$ ), with the highest densities occurring in microbial mats ( $137,756 \text{ ind. m}^{-2}$ ).  
276 At both sites, the highest densities occurred in microbial mat habitats, followed by mussel  
277 habitats and background habitats. At BCS, macrofaunal density differed among habitats (One-  
278 way ANOVA,  $F_{2,9}=7.58$ ,  $p=0.011$ ), with significantly higher densities in bacterial mats ( $83,649 \pm$   
279  $28,466 \text{ ind. m}^{-2}$ ) than in background soft-sediments ( $15,719 \pm 1,582 \text{ ind. m}^{-2}$ ; Tukey HSD;  $p =$   
280  $0.009$ ). Likewise, at NCS macrofaunal density also differed among habitats (One-way ANOVA,  
281  $F_{2,10} = 10.87$ ,  $p = 0.003$ ), with densities in microbial mats ( $47,962 \pm 13,547 \text{ individuals m}^{-2}$ )  
282 significantly higher than both mussel (Tukey HSD,  $p = 0.007$ ) and background soft-sediments  
283 (Tukey HSD,  $p=0.007$ ). The upper 2 cm of sediments at BCS contained slightly higher  
284 proportions of macrofauna in bacterial mat sediments (79%) as compared to mussel sediments  
285 (76%) and soft sediments (76%). The proportion of macrofauna found in the upper 2cm at NCS  
286 was higher in bacterial mat sediments (84%) as compared to mussel sediments (66%) and soft  
287 sediments (55%).

288

### 289 3.2 Diversity

290 Macrofaunal diversity patterns among habitat types differed between BCS and NCS. At  
291 BCS, diversity ( $H' \log_e$ ; Table 2) was significantly lower in bacterial mat sediments than in both  
292 mussel (Tukey HSD,  $p < 0.0001$ ) and background sediments (Tukey HSD,  $p < 0.0001$ ).  
293 Similarly, taxa evenness ( $J'$ ; Table 3) was significantly lower in bacterial mat sediments than in  
294 both mussel (Tukey HSD,  $p = 0.0001$ ) and background sediments (Tukey HSD,  $p < 0.0001$ ). In

295 contrast at NCS, there was no significant difference in diversity (One-way ANOVA,  $F_{2,10}=0.96$ ,  
296  $p=0.42$ ) or evenness (One-way ANOVA,  $F_{2,10}=4.05$ ,  $p=0.051$ ) among habitat types (Table 2).  
297 Rarefaction analysis within BCS (Figure 4a) and NCS (Figure 4b) indicated similar within-site  
298 patterns as given using Shannon diversity; however, overall diversity of all habitats combined  
299 (Figure 4c) indicated higher diversity at NCS than at BCS.

300         There was high  $\beta$  diversity among habitats. At BCS, 15% of the observed taxa were  
301 shared across all sediment habitats, 23-47% of the taxa were shared between any two habitats,  
302 and 42% of the taxa were unique to a single habitat. Approximately 42% of the taxa in BCS  
303 sediments only occurred in seep habitats. Mussel bed samples (Ekman core) at BCS shared more  
304 taxa with mussel sediment habitats (60%) than with microbial mat (20%) or background  
305 sediments (40%); however, the low number of taxa present in the single mussel bed sample  
306 resulted in low overall diversity compared to mussel sediments (Figure 4a). At NCS, there was  
307 overall greater  $\beta$  diversity than at BCS, with only 13% of taxa shared among all three sediment  
308 habitats and 21-35% occurring in two or more habitats. A high percentage of taxa, 59%,  
309 occurred only in a single habitat at NCS, and 59% of the taxa were only observed in seep  
310 sediments. Similar to BCS, the mussel bed samples at NCS (Ekman core) shared the most taxa  
311 with the mussel cores (56%). The non-quantitative suction samples also shared the most taxa  
312 with their analogous sediment communities; the mat suction sample shared 51% of its taxa with  
313 mat sediments, and the mussel suction samples shared 30% with mussel sediment. Overall, the  
314 mussel bed and mussel suction samples had similar diversity to the mussel sediments, while the  
315 microbial mat suction had higher diversity than microbial mat sediments (Figure 4b). Pooled

316 rarefaction (Figure 3c) for seep habitat push cores combined with Ekman cores and suction  
317 samples resulted in an increase in diversity with each additional habitat at both sites. The high  
318 difference in taxa between the mussel bed samples (Ekman), compared to cores collected  
319 adjacent to the mussel bed suggests high taxonomic turnover on a small (<1m) spatial scale with  
320 minimal taxon overlap.

321

### 322 3.3 Community composition

323 Overall taxonomic composition was similar among habitat types between BCS and NCS  
324 based on push core collections (Figure 5). Polychaetes dominated microbial mat and background  
325 habitats, comprising 63-67% of the communities at BCS and 73-77% at NCS. The polychaete  
326 families Dorvilleidae and Capitellidae composed a large proportion of microbial mat  
327 communities at BCS (66%) and NCS (57%), with the addition of Spionidae and other  
328 polychaetes at NCS. The proportion of oligochaetes was higher in microbial mats at BCS (31%)  
329 than at NCS (4%), while NCS contained higher proportions of Crustacea, Mollusca, and other  
330 taxa. In mussel habitats at both sites, polychaete composition was low (39-47%), with high  
331 proportions of crustaceans (23-50%), specifically amphipods and tanaids. Background  
332 sediments contained the highest proportion of molluscs (BCS: 18%, NCS: 13%). The overall  
333 taxonomic composition of the Ekman cores and suction samples did not resemble the  
334 macrofaunal composition in sediment cores collected from adjacent mussel or mat habitats  
335 (Figure 4). The BCS Ekman core contained a high proportion of isopods (65%), while the NCS  
336 Ekman core contained a lower proportion of amphipods (7%) relative to sediment communities



337 adjacent to mussel beds. The NCS mussel suction contained the highest proportion of  
338 gastropods (46%) while the NCS mat suction contained high proportions of other polychaetes  
339 (33%) and other taxa (31%), specifically Sipuncula (28%), in comparison to mussel and mat  
340 sediment communities. In addition, the Ekman and suction samples were better able to collect  
341 more highly mobile taxa, as indicated by the presence of Nebaliidae and Caridean shrimp.

342 Macrofaunal communities differed both between sites (Figure 6; Two-way  
343 PERMANOVA, Pseudo-F=5.78, p=0.0001) and among habitat types (Two-way PERMANOVA,  
344 Pseudo-F=7.30, p=0.0001) but with a significant interaction of site and habitat (Two-way  
345 PERMANOVA, Pseudo-F=5.04, p=0.0001). Estimates of the source of variation in communities  
346 indicate that differences among habitat types (Estimate=1056) were greater than differences  
347 between sites (Estimate=539). Within each site, community variability among cores was highest  
348 within microbial mat sediments (Table 2, MVDISP). Pairwise analysis of site and habitat  
349 combinations showed significant differences in macrofaunal communities between all  
350 site/habitat combinations (Table 3) except between BCS mussel and background habitats.  
351 Microbial mat communities at BCS and NCS were more similar to each other than they were to  
352 other habitats at their respective sites (Table 3). At BCS, bacterial mats had higher densities of  
353 Capitellidae (Polychaeta), Dorvilleidae (Polychaeta), and Tubificidae (Oligochaeta) than the  
354 background and mussel habitats, contributing 33% of the dissimilarity with mussel habitats and  
355 42% with background habitats. Mussel habitats had higher densities of Tubificidae  
356 (Oligochaeta), Leptocheliidae (Tanaidacea), and Typhlotanaiidae (Tanaidacea) but lower  
357 densities of Opheliidae (Polychaeta) and Yoldiidae (Bivalvia) compared to background soft

358 sediments, contributing 22% of the overall dissimilarity. SIMPER analysis using  
359 presence/absence data (Table 3) indicated the Ekman core collected within the mussel bed at  
360 BCS were more similar to the sediment communities associated with mussels, than to  
361 background sediments, and mat habitats at BCS. However, the taxonomic composition of the  
362 BCS Ekman core was more similar to NCS Ekman and suction samples than to sediment  
363 communities at BCS (Table 3).

364 At NCS, bacterial mats differed from both mussel and background habitats by high  
365 densities of Capitellidae (Polychaeta), Dorvilleidae (Polychaeta), and Spionidae (Polychaeta)  
366 contributing 26% of the dissimilarity with mussel habitats and 27% with background habitats.  
367 Mussel habitats differed from background soft-sediment habitats, with higher densities of  
368 Oedicerotidae (Amphipoda) and Spionidae (Polychaeta), but low densities of Cossuridae  
369 (Polychaeta) and Paraonidae (Polychaeta) contributing 31% of the dissimilarity. At NCS, the  
370 highest community similarities were observed between the NCS Ekman core and mussel  
371 sediment communities (44%, Table 3) and among the Ekman and suction samples (43-54%).

372

### 373 3.4 Relationship to sediment geochemistry

374 Sediment geochemical properties differed among microbial mat, mussel, and background  
375 soft-sediment habitats at NCS (Table 4). Sediment  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  values were significantly lower  
376 in mussel habitats than both microbial mats and background soft-sediments (Tukey HSD,  $\delta^{13}\text{C}$ ,  
377  $p < 0.001$ ;  $\delta^{15}\text{N}$ ,  $p < 0.033$ ). Microbial mat sediments also contained lower  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  values  
378 compared to background soft-sediments (Tukey HSD,  $\delta^{13}\text{C}$ ,  $p < 0.001$ ;  $\delta^{15}\text{N}$ ,  $p = 0.001$ ). In

379 contrast, mussel habitats contained higher percent organic carbon and nitrogen content than both  
380 microbial mat and background soft-sediments (Tukey HSD, %C,  $p < 0.006$ ; %N,  $p < 0.001$ ). There  
381 was no difference in the C:N among habitat types (One-way ANOVA,  $F_{2,7} = 2.37$ ,  $p = 0.16$ ).  
382 Background soft-sediments had the highest mud content, followed by mussel and microbial mat  
383 sediments. It is notable that deeper fractions (2-5 cm) of the microbial mat cores contained  
384 authigenic carbonate rubble that contributed to the higher grain size in those samples.

385 Principal coordinate analysis of macrofaunal communities at NCS (Figure 7) indicates  
386 that two orthogonal axes are capable of explaining 63% of the natural variation among cores.  
387 PCO1 separates mussel from microbial mat and background communities, while PCO2 separates  
388 microbial mat from background communities. Variable correlation with PCO axes indicated that  
389 PCO1 was positively correlated with  $\delta^{13}\text{C}$ ,  $\delta^{15}\text{N}$ , and the C:N ratio, and negatively correlated  
390 with percent organic carbon (%C). Mud content and depth were positively correlated with  
391 PCO2. DISTLM analysis (Table 5) indicated that all environmental variables except C:N can  
392 individually explain a significant portion (24-31%,  $p < 0.019$ ) of the variation in NCS  
393 communities. A combination of  $\delta^{13}\text{C}$  and mud content provide the best explanation of variation  
394 in NCS macrofaunal communities, accounting for 58% of the variation among samples in two  
395 axes.

396

#### 397 **4. Discussion:**

##### 398 4.1 Depth-related patterns

399 Differences between the depths represented by BCS and NCS habitats were apparent for  
400 most of the community parameters measured (e.g., density, diversity, community composition).  
401 Macrofaunal densities along non-seep slope ecosystems generally decrease with depth and  
402 distance from shore, related to decreases in food availability from surface productivity and  
403 transport of organic matter from shelf areas (Rex and Etter, 2010). We observed lower densities  
404 with depth in background sediments, a trend that continues regionally with even lower non-seep  
405 macrofaunal densities at Blake Ridge (Table 6; Robinson et al., 2004). A trend of decreasing  
406 density with depth was also present for seep habitats; however, given the additional nutrition  
407 source provided by the seep this trend cannot be attributed to depth-related patterns alone. Each  
408 habitat at NCS exhibited higher variability in macrofaunal communities (MVDISP) as compared  
409 to BCS habitats, suggesting increased patchiness with depth, consistent with observations from  
410 deep mud volcanoes (Cunha et al., 2013) and general deep-sea ecological paradigms (Rex and  
411 Etter, 2010). The higher variability within seep communities at NCS could be due to the larger  
412 physical distance between the collected individual mussel and microbial mat sediments than at  
413 BCS. However, background sediments at NCS were collected at a finer spatial scale than at  
414 BCS, and we would have expected lower variability at NCS if spatial separation alone was the  
415 controlling factor. Higher community patchiness with depth was also supported by the differing  
416 results from the two diversity analyses (Shannon diversity vs. rarefaction). The higher Shannon  
417 diversity at BCS indicates that diversity was high within cores, but rarefaction suggests there was  
418 a lower overall taxonomic pool present at BCS compared to NCS, although under sampling was  
419 evident for both sites. Overall diversity followed the expected trend and increased with depth

420 (e.g., Rex, 1981), although the opposite pattern was observed for background habitats. Given the  
421 low abundance and limited sampling in background habitats at both sites, our results likely  
422 provide an underrepresentation of background soft-sediment diversity.

423         Community assemblage differences between BCS and NCS may also be depth driven,  
424 consistent with the separation of macrofaunal communities between upper bathyal (200-1500m)  
425 and lower bathyal/abyssal (>1500m) depths worldwide documented by Bernardino et al. (2012).  
426 Similar depth differences have been observed for seep megafaunal communities (Olu et al.,  
427 2010), supporting the interconnection between infaunal and epifaunal communities. Differences  
428 among seep and non-seep sediment communities have been observed to increase with depth  
429 (Levin, 2005), suggesting the greater importance of the additional nutrition source provided by  
430 the seep at increasing depths (Levin and Michener, 2002). Within a geographic region,  
431 comparisons among seeps at different depths have been limited (Cordes et al., 2010a; Cunha et  
432 al., 2013), but consistent with the overall results of Bernardino et al. (2012). In the Gulf of  
433 Mexico, depth was the main driver of communities within seep habitats, with observed  
434 community differences between shallow and deep seeps, but no observed difference across the  
435 large geographic range (~650 km) of the study (Cordes et al., 2010a). Similar depth separation  
436 of communities was observed on a smaller geographic scale (~200 km) for mud volcanoes in the  
437 Gulf of Cadiz (Cunha et al., 2013). Other studies with multiple depths over large geographic  
438 ranges (>425 km) have also found significant community differences, including seeps along the  
439 Pacific margin (525 and 770m; Levin et al., 2010) and the Aleutian margin (3300m and 4400m;  
440 Levin and Mendoza, 2007); however, the distance between seeps in these studies may be

441 confounding depth patterns. The closer proximity between BCS and NCS (90km) than in  
442 previous studies reduces the geographic location effect and community differences likely  
443 highlight depth-related patterns.

444         The depth-related patterns observed at BCS and NCS suggest that depth is a large driving  
445 mechanism that affects both seep and non-seep communities in the mid-Atlantic region. The  
446 complex interaction of continental shelf and slope topography, circulation patterns in the South  
447 Atlantic Bight, and high organic input from enhanced surface productivity (Csanady and  
448 Hamilton, 1988; DeMaster et al., 1994; Rex and Etter, 2010; Schaff et al., 1992; Yoder et al.,  
449 1985) is known to result in regionally high macrofaunal abundance and biomass (Blake and  
450 Grassle, 1994; Blake and Hilbig, 1994; Schaff et al., 1992), with regional patterns of declining  
451 macrofaunal density but increasing diversity with depth (Boesch, 1979). The high macrofaunal  
452 abundance observed at BCS is consistent with regional patterns, suggesting all habitats at this  
453 site are benefiting from the high organic input. Regional patterns are likely to affect seeps  
454 globally, as suggested by similar regional patterns for soft-sediment environments in the Gulf of  
455 Mexico observed in seep communities (Cordes et al., 2010a), providing a framework from which  
456 to assess seep influence on benthic communities.

457

#### 458 4.2 Among habitat comparisons

459         Macrofaunal densities observed in BCS microbial mat sediments (Table 6) were among  
460 the highest recorded for any seep environment worldwide. Locally high densities in seep  
461 habitats have been reported from multiple locations, with the highest densities recorded from

462 microbial mats in the Gulf of Mexico (Table 6; Robinson et al., 2004). High densities have been  
463 recorded in frenulate fields on the Norwegian margin (Decker et al., 2012), microbial mats on the  
464 northern California margin (Levin et al., 2006), and an ampharetid bed in New Zealand (Thurber  
465 et al., 2013), all of which were at deeper depths (Table 6). Macrofaunal density in microbial  
466 mats was also high at NCS compared to microbial mat habitats at similar depths in other  
467 locations (Table 6; Ritt et al., 2011; Robinson et al., 2004). Macrofaunal densities in microbial  
468 mat and mussel sediments at BCS and NCS were greater than those measured at the nearest  
469 previously known seep located 802 km to the southeast at Blake Ridge (Robinson et al., 2004).  
470 Regionally, both seep sites represent localized areas of high densities, as indicated by the lower  
471 densities in background sediments, similar to results for other seep communities worldwide  
472 (Menot et al., 2010). Background sediments at both BCS and NCS also exhibited higher  
473 densities than from other regional and historical sampling efforts north of Cape Hatteras (Table  
474 6; Maciolek et al., 1987; Robertson et al., 2015; Sanders et al., 1965). Cape Hatteras represents a  
475 known zoogeographic barrier (e.g., Cutler, 1975; Hilbig, 1994) due to the convergence of the  
476 northward flowing Gulf Stream and the southward flowing Western Boundary Undercurrent, and  
477 slope densities in this region were both higher (Cape Hatteras, Table 6; Blake and Hilbig, 1994)  
478 and lower (Cape Lookout, Table 6; Blake and Grassle, 1994) than similar background depths at  
479 NCS.

480         The habitats characterized by their dominant faunal component (e.g., microbial mats,  
481 mussel beds) are known to be distinct from one another in other seep locations (Bernardino et al.,  
482 2012; Cordes et al., 2010a; Levin, 2005). While macrofaunal abundances in seep habitats are

483 commonly higher than background soft-sediments (Levin and Mendoza, 2007), differences  
484 between seep habitats (i.e., microbial mats, clam beds, mussel beds) have been variable  
485 (Bernardino et al., 2012). Microbial mat sediments near Costa Rica had macrofaunal densities  
486 two times higher than in clam beds (400-1796m; Levin et al., 2015) while microbial mats on the  
487 Pacific margin (252-770m) had similar (Levin et al., 2010; Levin et al., 2003) or higher densities  
488 than in clam beds (Sahling et al., 2002). The high densities observed in microbial mat habitats at  
489 both BCS and NCS differs from the regional pattern observed at Blake Ridge, where mussel bed  
490 habitats contained higher macrofaunal densities than microbial mats (Robinson et al., 2004).  
491 Low densities in microbial mats at Blake Ridge may also be the result of high sediment sulfide  
492 and methane concentrations that exceed the tolerance limit for most macrofaunal taxa (e.g.,  
493 Bernardino et al., 2012). In addition, the mussel species at Blake Ridge, *B. heckerae*, known to  
494 support both methanotrophs and sulfide oxidizers, differed from the dominant mussels present at  
495 BCS and NCS, *B. childressi*-complex, which are known to support only methanotrophic bacteria  
496 (Olu-Le Roy et al., 2007b). While *B. heckerae* has only been reported from deeper depths (2150-  
497 3300m; Cordes et al., 2010a; Levin and Mendoza, 2007; Robinson et al., 2003), potentially  
498 preventing establishment at either BCS or NCS, the depth range of *B. childressi* (500-2284m;  
499 Cordes et al., 2010a) encompasses that of Blake Ridge. However, the two species were never  
500 collected together in the Gulf of Mexico despite overlapping depth ranges (Cordes et al., 2010a).  
501 As suggested by Cordes et al., (2010a), the specific mussel species present in these seep habitats  
502 may serve as a proxy for the surrounding habitat chemistry, and by extension, the sediment  
503 geochemistry. The different species complexes present at Blake Ridge and our sites suggests an



504 underlying difference in concentrations of methane and sulfide which may help explain the  
505 differing infaunal community patterns observed between sites.

506

#### 507 4.3 Patterns in macrofaunal diversity

508 The low ( $\alpha$ ) diversity observed in microbial mat habitats, particularly at BCS, is  
509 consistent with previous studies which observed lower diversity within microbial mat habitats  
510 compared to nearby clam beds (Bernardino et al., 2012; Levin and Mendoza, 2007; Levin et al.,  
511 2003). Microbial mat sediments at both BCS and NCS were dominated by the annelid families  
512 Capitellidae, Dorvilleidae, and Tubificidae, all of which have been previously observed in seep  
513 habitats (Levin, 2005; Levin et al., 2010; Levin et al., 2003). Dorvilleids are a common  
514 component of seep communities (Levin, 2005), often occurring in high densities in microbial  
515 mat sediments (Robinson et al., 2004; Sahling et al., 2002) where they are likely consuming mat-  
516 forming sulfur bacteria, methanotrophs (Levin and Michener, 2002), and/or archaea (Thurber et  
517 al., 2012). Multiple species of dorvilleids are also known co-occur in seep habitats, with species  
518 exhibiting high trophic and niche partitioning (Levin et al., 2013). Capitellids are known to be a  
519 stress tolerant, opportunistic taxa, with a strong preference for sulfidic environments (Levin et  
520 al., 2000; Levin et al., 2003) and are known to consume methanotrophs at other seeps (Carlier et  
521 al., 2010; Decker and Olu, 2010, 2012). Only the polychaete families Dorvilleidae, Cirratulidae,  
522 and Hesionidae were documented in microbial mat sediments at Blake Ridge (Robinson et al.,  
523 2004), all of which were present in microbial mat sediments at NCS, while Hesionidae were  
524 absent in mat sediments at BCS. In contrast to microbial mats, sediments adjacent to mussels at

525 BCS and NCS contained high proportions of crustaceans, particularly amphipods and tanaids.  
526 Amphipods are known to be sensitive to organic enrichment, increased hydrocarbon  
527 concentrations (Peterson et al., 1996), and oxygen stress (Thurber et al. 2013), and their  
528 distribution at BCS may indicate high methane flux, high sulfide concentrations and low oxygen  
529 concentrations at our microbial mat habitats. For the seeps at Blake Ridge, crustaceans were  
530 only documented in mussel sediments (Robinson et al., 2004), suggesting similarities among  
531 mussel communities across depth regimes. In addition, increased variability in communities has  
532 been used as an indicator of stressed and/or disturbed environments (Fisher et al., 2014; Warwick  
533 and Clarke, 1993). Although fluid flux and sulfide concentrations were not measured, the high  
534 variability (MVDISP) and low diversity in microbial mat sediments, compared to the greater  
535 similarity within mussel and background sediment communities, also suggest a higher stress  
536 environment in microbial mat sediments.

537 High  $\beta$  diversity at seep sites was present over both small (<1 m) and large spatial scales.  
538 High  $\beta$  diversity among seep habitats (mussels and mats) has been documented at seep sites  
539 worldwide (see Cordes et al., 2010b for review) and is suggested to be a result of small-scale  
540 variation in the vertical distribution and concentration of sulfides in sediments (Levin et al.,  
541 2003) and habitat heterogeneity provided by seep megafauna (Cordes et al., 2010b). Hints at  
542 these small-scale variations were observed both in sediment cores collected in mat and  
543 background habitats at BCS and between the Ekman cores collected within the mussel habitat  
544 and cores collected directly adjacent to mussel habitats at both sites. Similar to results observed  
545 in Pacific seeps (Levin et al., 2010), the seep habitats contribute significantly to the regional

546 biodiversity for their specific depth, providing 37-49% of infaunal taxa and high taxon turnover  
547 between seeps and background soft-sediment communities. In addition, while the taxonomic  
548 level applied in this study (family-level) was sufficient to ascertain differences among habitat-  
549 specific communities, further identification (e.g., genus and/or species level) will likely provide  
550 increased separation of habitat-specific communities, biodiversity estimates, identification of  
551 biogeographic boundaries, and insight into seep endemism at these sites. High  $\beta$  diversity among  
552 the mussel habitat, adjacent sediments, and background sediments highlights that habitat  
553 provision of dense mussel communities influences not only the *in situ* macrofaunal communities  
554 found within the beds, but also the communities that occur in the sediments beyond the perimeter  
555 of the mussel bed itself. This ‘reef’ effect has also been, observed for deep-sea coral  
556 communities (Demopoulos et al., 2014). While the effect of seep habitats on sediment  
557 macrofaunal communities has not been detected at distances greater than 250 m from seep  
558 megafauna (Menot et al., 2010), discrete transects from mussel beds to adjacent sediments and  
559 beyond would help quantify the sphere of influence of seep activity and biogenic structures on  
560 adjacent habitats.

561

#### 562 4.4 Relationship to sediment geochemistry

563 The higher proportion of taxa found in the upper 2 cm of sediments in microbial mats  
564 versus deeper sediments, particularly at NCS, may reflect different geochemical settings present  
565 within each habitat. Seeps, along with other reducing environments such as areas of organic  
566 enrichment, large organic falls, and oxygen minimum zones, are often characterized by low

567 oxygen, sulfidic sediments (Levin et al., 2010; Tunnicliffe et al., 2003). The vertical distribution  
568 of taxa in sediments is regulated partly by oxygen and sulfide concentrations (Levin, 2005),  
569 resulting in a trade-off between sulfide tolerance and food availability (Menot et al., 2010). Few  
570 taxa tolerate sulfide concentrations >1 mM, while Dorvilleidae polychaetes can occur in high  
571 densities at concentrations ranging 1 to 6 mM (Levin et al., 2003). The higher proportion of taxa  
572 present in the upper 2 cm of microbial mat sediments suggests these habitats have low oxygen  
573 and potentially high sulfide concentrations that restrict fauna to the surface sediments (Levin et  
574 al., 2003). Whereas, the higher proportion of taxa present in sub-surface sediments (>2cm) in  
575 mussel and background habitats suggests deeper oxygen penetration and lower sulfide  
576 concentrations, allowing more individuals to survive at greater depth within the sediments (Levin  
577 et al., 2001; Levin, 2005). Bioturbation by deeper dwelling taxa in turn facilitates oxygen  
578 penetration and the transfer of organic material, thus also increasing the food availability for  
579 other organisms residing deeper in the sediments. Similar faunal sediment-depth patterns were  
580 reported for microbial mat (Levin et al., 2003) and mussel-associated sediments (Menot et al.,  
581 2010) at other seeps, suggesting that in the absence of specific oxygen and sulfide concentration  
582 measurements, inferences about the geochemical setting based on the faunal composition may be  
583 possible.

584         The high variation observed within NCS microbial mat communities suggests a gradient  
585 among sampling locations in the underlying seep fluid flow, sediment geochemistry, and depth.  
586 Sediments supporting microbial mats are known to sustain high rates of methane emissions, high  
587 concentrations of sulfide, and low oxygen penetration (Bernardino et al., 2012). In contrast,

588 mollusc-dominated habitats (e.g. clam beds) often have lower methane emission rates and lower  
589 sulfide concentrations near the sediment surface (Boetius and Suess, 2004; Levin, 2005; Levin et  
590 al., 2013; Levin et al., 2003; Sahling et al., 2002), while mussel beds are known to have  
591 increased methane concentrations in overlying water (Olu-Le Roy et al., 2007a). The large  
592 continuous fields of mussels present at BCS and NCS suggest regular and diffuse fluid flow  
593 (Olu-Le Roy et al., 2007a), although the patchiness and large areas of shell debris at BCS also  
594 suggest spatially or temporally intermittent flow. Animals occupying sediments below microbial  
595 mats must be tolerant to high levels of sulfide, while those near mussel habitats may not require a  
596 high tolerance, but fall within a tolerance gradient. The high methane flux expected in microbial  
597 mat sediments should contribute to higher sulfate reduction and anaerobic methane oxidation,  
598 while low methane emission rates in mussel sediments may concentrate isotopically depleted  
599 carbon, due to the physical presence of mussels possibly limiting the release of methane. Both of  
600 these processes would yield light carbon isotope values in sediments, and sediment  $\delta^{13}\text{C}$  values  
601 in microbial mats were higher than in mussel bed habitats. Isotopic composition of mussels  
602 collected within these seeps yielded isotopically light  $\delta^{13}\text{C}$  (-64‰ to -61‰; Prouty et al., 2014)  
603 and  $\delta^{15}\text{N}$  values (-2‰ to 6‰; Prouty et al., 2014). The contribution of mussel tissues and/or  
604 waste products to the organic matter pool is indicated by the enriched percent organic carbon  
605 content and depleted  $^{13}\text{C}$  values, and similar organic carbon concentrations were observed in  
606 Gulf of Guinea mussel bed sediments (Menot et al, 2010). Microbial composition may also  
607 influence the stable isotope composition of the microbial mat sediments. Filamentous sulfide  
608 oxidizing bacteria (e.g., *Beggiatoa*, *Thiopluca*) differ from amorphous forms (e.g. *Arcobacter*)

609 and iron-oxidizers, and sediment  $\delta^{13}\text{C}$  values reported here may reflect the microbial  
610 communities supporting the food chain, as well as organic matter contribution from mussels  
611 (Levin and Mendoza, 2007). The presence of mussels from the *B. childressi*-complex, which  
612 contain only methanotrophic symbionts, further suggests methanotrophs are playing a large role  
613 in the transfer of energy at our seeps (e.g., Levin et al., 2015; Levin and Michener, 2002;  
614 Thurber et al., 2012; Thurber et al., 2013). The portion of variation in microbial mat  
615 communities at NCS characterized by mud content, and by extension, the amount of authigenic  
616 carbonate rubble within cores, may reflect the role and duration of anaerobic oxidation of  
617 methane within the sediment, which can influence the macrofaunal community structure.

618         Although we did not measure any sediment pore-water geochemistry at BCS, given the  
619 similar patterns exhibited among microbial mat, mussel, and background sediment communities  
620 in relation to those at NCS, similar sediment geochemical patterns may be structuring infaunal  
621 communities at BCS. Sediment geochemistry for sites within 2 km at shallower (282m) and  
622 deeper (515m) depths on the Baltimore slope indicate lower sediment organic carbon (0.31-  
623 0.43%) and nitrogen (0.1%), C:N ratios (3.1-4.3),  $\delta^{15}\text{N}$  (4.6-4.8‰) values, and mud content (12-  
624 38%, 515m only), but comparable  $\delta^{13}\text{C}$  (-22.3 to -21.9) compared to background sediments  
625 collected at NCS (Mienis et al., 2014). These data suggest a food-limited environment with  
626 increased hydrodynamic flow, as indicated by water column turbidity patterns over the slope  
627 (Mienis et al., 2014). Additional sampling of sediment geochemistry at BCS would allow  
628 regional comparisons between these two discrete seep habitats, and provide further insight into  
629 the mechanisms supporting seep communities in the mid-Atlantic region.

630           There are potential limitations to the comparisons made between seep and background  
631 habitats at both BCS and NCS in our study, including seasonality and inter-annual variation,  
632 location, and sampling methods. At BCS, all of the background sediments were collected in  
633 August 2012, while all but one core from seep habitats were collected in May 2013. Seasonality  
634 in surface productivity and hydrodynamic regimes, as well as disturbance events, promotes shifts  
635 in community assemblages. However, there was no observed difference in the abundance of taxa  
636 in the upper 2 cm of sediments between 2012 and 2013 samples collected at BCS, which might  
637 have been expected if there had been an organic enrichment event during this time period. In  
638 addition, previous temporal studies within the mid-Atlantic region found little inter-annual  
639 variation in macrofaunal communities (Boesch, 1979). Proximity of background, soft-sediment  
640 cores to seep habitats may also affect their observed similarity to seep habitats. Three of the  
641 four background cores were collected within the axis of Baltimore Canyon, while the fourth was  
642 in close proximity (<1m) to microbial mat habitats at the seep on the adjacent slope. The high  
643 similarity among BCS background cores (59%) with the inclusion of the near-mat core suggests  
644 they are an adequate representation of nearby background communities. However, the close  
645 proximity of (<1m) of the 2012 microbial mat core with At NCS, the box cores collected for  
646 background sediments were 18-19 km north from the seep habitats and were located at the base  
647 of the Norfolk Canyon channel. While macrofaunal communities are known to differ between  
648 canyon axis and slope habitats for Norfolk Canyon (Robertson et al., 2015), the samples  
649 examined in this study represent the best comparative information available.

650 Seep habitat-specific communities on the western Atlantic margin exhibit many  
651 similarities to other microbial mat and mollusc-dominated communities worldwide, suggesting  
652 similar environmental controls within these settings. This study is the first to examine seep-  
653 associated infaunal communities at depths <2000 m and in the context of their geochemical  
654 environment in this region of the Atlantic. Discrete differences among seep habitats and sites  
655 indicate that seep community patterns may be driven, in part, by variation in the intricate sub-  
656 seafloor structures supplying methane to the upper sediment/water interface. The potential  
657 ephemeral nature of these seeps and their associated fluid flux (Condon et al., 2015) may  
658 represent a strong driver influencing infaunal communities. Enhanced understanding of the seep  
659 plumbing, methane flux, and associated sediment geochemistry (e.g., pore water sulfide and  
660 methane concentrations) coupled with infaunal community metrics could benefit the  
661 development of generalizations relating to seep environmental controls on infaunal structure and  
662 function.

663

664 **Acknowledgements:**

665 The authors would like to thank the Bureau of Ocean Energy Management (BOEM), NOAA and  
666 USGS DISCOVRE Mid-Atlantic Canyons project teams, the crews of the NOAA Ships *Nancy*  
667 *Foster* and *Ron Brown*, ROV *Jason II* Group (WHOI), ROV *Kraken* group (UConn), S.W. Ross,  
668 J. Chaytor, N. Prouty, U. ten Brink, and C. Morrison. Special thanks go out to Jennifer McClain-  
669 Counts, Mike Rhode, Mark Lavaleye, and the USGS/WARC Benthic Ecology Group for  
670 assistance at sea and laboratory support. We thank Lisa Levin, Howard Jelks, and two



671 anonymous reviewers for their insightful and helpful comments that improved this manuscript.  
672 Funding was provided to A. Demopoulos from the USGS Environments Program through the  
673 Outer Continental shelf study DISCOVRE Mid-Atlantic Canyons, with ship time support  
674 provided by NOAA. Any use of trade, product, or firm names is for descriptive purposes only  
675 and does not imply endorsement by the U.S. Government.

676

677 **References:**

- 678 Anderson, M., Gorley, R.N., Clarke, K.R., 2008. PERMANOVA+ for PRIMER: Guide to  
679 Software and Statistical Methods. PRIMER-E Ltd., Plymouth, UK.
- 680 Bergquist, D.C., Ward, T., Cordes, E.E., McNelis, T., Howlett, S., Kosoff, R., Hourdez, S.,  
681 Carney, R., Fisher, C.R., 2003. Community structure of vestimentiferan-generated habitat  
682 islands from Gulf of Mexico cold seeps. *J. Exp. Mar. Biol. Ecol.* 289, 197-222.
- 683 Bernardino, A.F., Levin, L.A., Thurber, A.R., Smith, C.R., 2012. Comparative Composition,  
684 Diversity and Trophic Ecology of Sediment Macrofauna at Vents, Seeps and Organic Falls.  
685 *PloS One* 7.
- 686 Blake, J.A., Grassle, J.F., 1994. Benthic community structure on the U.S. South Atlantic slope  
687 off the Carolinas: Spatial heterogeneity in a current-dominated system. *Deep Sea Res. Part*  
688 *2 Top. Stud. Oceanogr.* 41, 835-874.
- 689 Blake, J.A., Hilbig, B., 1994. Dense infaunal assemblages on the continental slope off Cape  
690 Hatteras, North Carolina. *Deep Sea Res. Part 2 Top. Stud. Oceanogr.* 41, 875-899.

691 Schaff, T.R., Levin, L., Blair, N., DeMaster, D., Pope, R., Boehme, S., 1992. Spatial  
692 heterogeneity of benthos on the Carolina continental slope: large (100 km)-scale variation.  
693 Mar. Ecol. Prog. Ser. 88, 143-160.

694 Boesch, D.F., 1979. Chapter 6. Benthic ecological studies: macrobenthos, in: E.M, B., Boesch,  
695 D.F., Laird, B.L. (Eds.), Middle Atlantic Outer Continental Shelf Environmental Studies  
696 Volume II-B. Chemical and Biological Benchmark Studies. Virginia Institute of Marine  
697 Science, Gloucester Point, VA, p. 301.

698 Boetius, A., Suess, E., 2004. Hydrate Ridge: a natural laboratory for the study of microbial life  
699 fueled by methane from near-surface gas hydrates. Chem Geol 205, 291-310.

700 Brooks, J.M., Kennicutt, M., Fisher, C.R., Macko, S.A., Cole, K., Childress, J.J., Vetter, R.D.,  
701 1987. Deep-sea hydrocarbon seep communities: Evidence for energy and nutritional  
702 carbon sources. Science 238, 1138-1142.

703 Brothers, L.L., Van Dover, C.L., German, C.R., Kaiser, C.L., Yoerger, D.R., Ruppel, C.D.,  
704 Lobecker, E., Skarke, A.D., Wagner, J.K.S., 2013. Evidence for extensive methane venting  
705 on the southeastern U.S. Atlantic margin. Geology 41, 807-810.

706 Cambon-Bonavita, M.A., Nadalig, T., Roussel, E., Delage, E., Duperron, S., Caprais, J.C.,  
707 Boetius, A., Sibuet, M., 2009. Diversity and distribution of methane-oxidizing microbial  
708 communities associated with different faunal assemblages in a giant pockmark of the  
709 Gabon continental margin. Deep Sea Res. Part 2 Top. Stud. Oceanogr. 56, 2248-2258.

710 Carlier, A., Ritt, B., Rodrigues, C.F., Sarrazin, J., Olu, K., Grall, J., Clavier, J., 2010.  
711 Heterogeneous energetic pathways and carbon sources on deep eastern Mediterranean cold  
712 seep communities. *Mar. Biol.* 157, 2545-2565.

713 Clarke, K.R., Gorley, R.N., 2015. *PRIMER v7: User Manual/Tutorial*. PRIMER-E, Plymouth,  
714 UK.

715 Condon, D., Sahy, D., Ruppel, C., Noble, S., 2015. Tempo and longevity of methane efflux  
716 along the US Atlantic Margin, OS31B-07, AGU Fall Meeting, San Francisco, CA.

717 Cordes, E.E., Becker, E.L., Hourdez, S., Fisher, C.R., 2010a. Influence of foundation species,  
718 depth, and location on diversity and community composition at Gulf of Mexico lower-  
719 slope cold seeps. *Deep Sea Res. Part 2 Top. Stud. Oceanogr.* 57, 1870-1881.

720 Cordes, E.E., Cunha, M.R., Galeron, J., Mora, C., Olu-Le Roy, K., Sibuet, M., Van Gaever, S.,  
721 Vanreusel, A., Levin, L.A., 2010b. The influence of geological, geochemical, and biogenic  
722 habitat heterogeneity on seep biodiversity. *Mar. Ecol.* 31, 51-65.

723 Csanady, G.T., Hamilton, P., 1988. Circulation of slope water. *Continental Shelf Research* 8,  
724 565-624.

725 Cunha, M.R., Rodrigues, C.F., Génio, L., Hilário, A., Ravara, A., Pfannkuche, O., 2013.  
726 Macrofaunal assemblages from mud volcanoes in the Gulf of Cadiz: abundance,  
727 biodiversity and diversity partitioning across spatial scales. *Biogeosciences* 10, 2553-2568.

728 Cutler, E.B., 1975. Zoogeographical barrier on the continental slope off cape lookout, North  
729 Carolina. *Deep Sea Res. Oceanogr. Abs.* 22, 893-901.

730 Dando, P.R., Austen, M.C., Burke, R.A., Kendall, M.A., Kennicutt, M.C., Judd, A.C., Moore,  
731 D.C., O'Hara, S.C.M., Schmaljohann, R., Southward, A.J., 1991. Ecology of a North Sea  
732 pockmark with an active methane seep. *Mar. Ecol. Prog. Ser.* 70, 49-63.

733 Decker, C., Morineaux, M., Van Gaever, S., Caprais, J.-C., Lichtschlag, A., Gauthier, O.,  
734 Andersen, A.C., Olu, K., 2012. Habitat heterogeneity influences cold-seep macrofaunal  
735 communities within and among seeps along the Norwegian margin. Part 1: macrofaunal  
736 community structure. *Mar. Ecol.* 33, 205-230.

737 Decker, C., Olu, K., 2010. Does macrofaunal nutrition vary among habitats at the Hakon Mosby  
738 mud volcano? *Cahiers de Biologie Marine* 51, 361-367.

739 Decker, C., Olu, K., 2012. Habitat heterogeneity influences cold-seep macrofaunal communities  
740 within and among seeps along the Norwegian margin – Part 2: contribution of  
741 chemosynthesis and nutritional patterns. *Mar. Ecol.* 33, 231-245.

742 DeMaster, D.J., Pope, R.H., Levin, L.A., Blair, N.E., 1994. Biological mixing intensity and rates  
743 of organic carbon accumulation in North Carolina slope sediments. *Deep Sea Res. Part 2*  
744 *Top. Stud. Oceanogr.* 41, 735-753.

745 Demopoulos, A.W.J., Bourque, J.R., Frometa, J., 2014. Biodiversity and community  
746 composition of sediment macrofauna associated with deep-sea *Lophelia pertusa* habitats in  
747 the gulf of Mexico. *Deep Sea Res. Part 1 Oceanogr. Res. Pap.* 93, 91-103.

748 Demopoulos, A.W.J., Gualtieri, D., Kovacs, K., 2010. Food-web structure of seep sediment  
749 macrobenthos from the Gulf of Mexico. *Deep Sea Res. Part 2 Top. Stud. Oceanogr.* 57,  
750 1972-1981.

751 Fisher, C.R., 1990. Chemoautotrophic and methanotrophic symbioses in marine invertebrates.  
752 *Rev. Aquat. Sci.* 2, 399-436.

753 Fisher, C.R., Demopoulos, A.W.J., Cordes, E.E., Baums, I.B., White, H.K., Bourque, J.R., 2014.  
754 Coral Communities as Indicators of Ecosystem-Level Impacts of the Deepwater Horizon  
755 Spill. *BioScience* 64, 796-807.

756 Folk, R.L., 1974. Petrology of sediment rocks. Hemphill Publishing Company, Austin, TX.

757 Fry, B., Sherr, E.B., 1984.  $\delta^{13}\text{C}$  measurements as indicators of carbon flow in marine and  
758 freshwater ecosystems. *Contrib. in Mar. Sci.* 27, 13-47.

759 Hecker, B., Logan, D.T., Gandarillas, F.E., Gibson, P.R., 1983. Megafaunal assemblages in  
760 Lydonia Canyon, Baltimore Canyon, and selected slope areas, in: Observatory, L.-D.G.  
761 (Ed.), *Canyon and Slope Processes Study*. Minerals Management Service.

762 Hilbig, B., 1994. Faunistic and zoogeographical characterization of the benthic infauna on the  
763 Carolina continental slope. *Deep Sea Res. Part 2 Top. Stud. Oceanogr.* 41, 929-950.

764 Jones, C.G., Lawton, J.H., Shachak, M., 1996. *Organisms as Ecosystem Engineers*, Ecosystem  
765 Management. Springer New York, pp. 130-147.

766 Kochevar, R.E., Childress, J.J., Fisher, C.R., Minnich, E., 1992. The methane mussel: roles of  
767 symbiont and host in the metabolic utilization of methane. *Mar. Biol.* 112, 389-401.

768 Levin, L., Etter, R.J., Rex, M.A., Gooday, A.J., Smith, C.R., Pineda, J., Stuart, C.T., Hessler,  
769 R.R., Pawson, D., 2001. Environmental influences on regional deep-sea species diversity.  
770 *Annu. Rev. Ecol. Evol. Syst.* 32, 51-93.

771 Levin, L.A., 2005. Ecology of cold seep sediments: interactions of fauna with flow, chemistry  
772 and microbes. *Oceanog. Mar. Biol. Annu. Rev.* 43, 1-46.

773 Levin, L.A., James, D.W., Martin, C.M., Rathburn, A.E., Harris, L.H., Michener, R.H., 2000. Do  
774 methane seeps support distinct macrofaunal assemblages? Observations on community  
775 structure and nutrition from the northern California slope and shelf. *Mar. Ecol. Prog. Ser.*  
776 208, 21-39.

777 Levin, L.A., Mendoza, G.F., 2007. Community structure and nutrition of deep methane-seep  
778 macrobenthos from the North Pacific (Aleutian) margin and the Gulf of Mexico (Florida  
779 Escarpment). *Mar. Ecol.* 28, 131-151.

780 Levin, L.A., Mendoza, G.F., Gonzalez, J.P., Thurber, A.R., Cordes, E.E., 2010. Diversity of  
781 bathyal macrofauna on the northeastern Pacific margin: the influence of methane seeps and  
782 oxygen minimum zones. *Mar. Ecol.* 31, 94-110.

783 Levin, L.A., Mendoza, G.F., Grupe, B.M., Gonzalez, J.P., Jellison, B., Rouse, G., Thurber, A.R.,  
784 Waren, A., 2015. Biodiversity on the Rocks: Macrofauna Inhabiting Authigenic Carbonate  
785 at Costa Rica Methane Seeps. *PloS One* 10, e0131080.

786 Levin, L.A., Michener, R.H., 2002. Isotopic evidence for chemosynthesis-based nutrition of  
787 macrobenthos: the lightness of being at Pacific methane seeps. *Limnol. Oceanogr.* 47,  
788 1336-1345.

789 Levin, L.A., Ziebis, W., F. Mendoza, G., Bertics, V.J., Washington, T., Gonzalez, J., Thurber,  
790 A.R., Ebbe, B., Lee, R.W., 2013. Ecological release and niche partitioning under stress:

791 Lessons from dorvilleid polychaetes in sulfidic sediments at methane seeps. *Deep Sea Res.*  
792 Part 2 Top. Stud. Oceanogr. 92, 214-233.

793 Levin, L.A., Ziebis, W., Mendoza, G.F., Growney-Cannon, V., Walther, S., 2006. Recruitment  
794 response of methane-seep macrofauna to sulfide-rich sediments: An in situ experiment. *J.*  
795 *Exp. Mar. Biol. Ecol.* 330, 132-150.

796 Levin, L.A., Ziebis, W., Mendoza, G.F., Growney, V.A., Tryon, M.D., Brown, K.M., Mahn, C.,  
797 Gieskes, J., Rathburn, A.E., 2003. Spatial heterogeneity of macrofauna at northern  
798 California methane seeps: influence of sulfide concentration and fluid flow. *Mar. Ecol.*  
799 *Prog. Ser.* 265, 123-139.

800 Maciolek, N., Grassle, J.F., Hecker, B., Boehm, P.D., Brown, B., Dade, B., Steinhauer, W.G.,  
801 Baptiste, E., Ruff, R.E., Petrecca, R., 1987. Study of the Biological Processes on the U.S.  
802 Mid-Atlantic Slope and Rise Volume 2 Final Report.

803 Menot, L., Galéron, J., Olu, K., Caprais, J.-C., Crassous, P., Khripounoff, A., Sibuet, M., 2010.  
804 Spatial heterogeneity of macrofaunal communities in and near a giant pockmark area in the  
805 deep Gulf of Guinea. *Mar. Ecol.* 31, 78-93.

806 Mienis, F., Duineveld, G., Robertson, C.M., Demopoulos, A.W.J., Davies, A.J., Prouty, N.,  
807 Roark, B., Ross, S.W., Brooke, S., 2014. Particle transport and deposition in Norfolk and  
808 Baltimore Canyons, NW Atlantic, 2nd International Symposium on Submarine Canyons,  
809 Edinburgh, UK.

810 Olu-Le Roy, K., Caprais, J.C., Fifis, A., Fabri, M.C., Galeron, J., Budzinsky, H., Le Menach, K.,  
811 Khripounoff, A., Ondreas, H., Sibuet, M., 2007a. Cold-seep assemblages on a giant

812 pockmark off West Africa: spatial patterns and environmental control. *Mar. Ecol.* 28, 115-  
813 130.

814 Olu-Le Roy, K., Cosel, R.v., Hourdez, S., Carney, S.L., Jollivet, D., 2007b. Amphi-Atlantic  
815 cold-seep *Bathymodiolus* species complexes across the equatorial belt. *Deep Sea Res. Part*  
816 *1 Oceanogr. Res. Pap.* 54, 1890-1911.

817 Olu, K., Caprais, J.C., Galeron, J., Causse, R., von Cosel, R., Budzinski, H., Le Menach, K., Le  
818 Roux, C., Levache, D., Khripounoff, A., Sibuet, M., 2009. Influence of seep emission on  
819 the non-symbiont-bearing fauna and vagrant species at an active giant pockmark in the  
820 Gulf of Guinea (Congo-Angola margin). *Deep Sea Res. Part 2 Top. Stud. Oceanogr.* 56,  
821 2380-2393.

822 Olu, K., Cordes, E.E., Fisher, C.R., Brooks, J.M., Sibuet, M., Desbruyères, D., 2010.  
823 Biogeography and Potential Exchanges Among the Atlantic Equatorial Belt Cold-Seep  
824 Faunas. *PloS One* 5, e11967.

825 Paull, C.K., Ussler, W., Borowski, W.S., Spiess, F.N., 1995. Methane-rich plumes on the  
826 Carolina continental rise: Associations with gas hydrates. *Geology* 23, 89-92.

827 Peterson, C.H., Kennicutt II, M.C., Green, R.H., Montagna, P., Harper, J.D.E., Powell, E.N.,  
828 Roscigno, P.F., 1996. Ecological consequences of environmental perturbations associated  
829 with offshore hydrocarbon production: a perspective on long-term exposures in the Gulf of  
830 Mexico. *Can. J. Fish. Aquat. Sci.* 53, 2637-2654.



831 Prouty, N., Roark, E.B., Demopoulos, A.W.J., Condon, D., Davis, K., Ross, S.W., Brooke, S.,  
832 2014. Biologic indicators of seabed methane venting along the U.S. Mid-Atlantic margin,  
833 OS21A-1115, AGU Fall Meeting, San Francisco, CA.

834 R Development Core Team, 2011. R: A language an environment for statistical computing. . R  
835 Foundation for Statistical Computing, Vienna, Austria.

836 Rex, M.A., 1981. Community structure in the deep-sea benthos. *Annu. Rev. Ecol. Syst.* 12, 331-  
837 353.

838 Rex, M.A., Etter, R.J., 2010. Deep-sea biodiversity: pattern and scale. Harvard University Press.

839 Ritt, B., Pierre, C., Gauthier, O., Wenzhoefer, F., Boetius, A., Sarrazin, J., 2011. Diversity and  
840 distribution of cold-seep fauna associated with different geological and environmental  
841 settings at mud volcanoes and pockmarks of the Nile Deep-Sea Fan. *Mar. Biol.* 158, 1187-  
842 1210.

843 Robertson, C.M., Bourque, J.R., Demopoulos, A.W.J., Davies, A.J., Mienis, F., Duineveld, G.,  
844 Brooke, S., Ross, S.W., 2015. Unique macrofauna community dynamics in relation to  
845 sediment biogeochemistry and canyon morphology in Baltimore and Norfolk Canyons,  
846 USA, 14th Deep Sea Biology Symposium, Aveiro, Portugal.

847 Robinson, C.A., Bernhard, J.M., Levin, L.A., Mendoza, G.F., Blanks, J.K., 2004. Surficial  
848 hydrocarbon seep infauna from the Blake Ridge (Atlantic Ocean, 2150 m) and the Gulf of  
849 Mexico (690–2240 m). *Mar. Ecol.* 25, 313-336.

850 Robinson, J.J., Cavanaugh, C.M., 1995. Expression of form I and form II Rubisco in  
851 chemoautotrophic symbioses: implications for the interpretation of stable carbon isotopic  
852 values. *Limnol. Oceanogr.* 40, 1496-1502.

853 Rodrigues, C.F., Hilário, A., Cunha, M.R., 2013. Chemosymbiotic species from the Gulf of  
854 Cadiz (NE Atlantic): distribution, life styles and nutritional patterns. *Biogeosciences* 10,  
855 2569-2581.

856 Sahling, H., Galkin, S.V., Salyuk, A., Greinert, J., Foerstel, H., Piepenburg, D., Suess, E., 2003.  
857 Depth-related structure and ecological significance of cold-seep communities - a case study  
858 from the Sea of Okhotsk. *Deep Sea Res. Part 1 Oceanogr. Res. Pap.* 50, 1391-1409.

859 Sahling, H., Rickert, D., Lee, R.W., Linke, P., Suess, E., 2002. Macrofaunal community structure  
860 and sulfide flux gas hydrate deposits from the Cascadia convergent margin, NE Pacific.  
861 *Mar. Ecol. Prog. Ser.* 231, 121-138.

862 Sanders, H.L., Hessler, R.R., Hampson, G.R., 1965. An introduction to the study of deep-sea  
863 benthic faunal assemblages along the Gay Head-Bermuda transect. *Deep Sea Res.*  
864 *Oceanogr. Abs.* 12, 845-867.

865 Schaff, T.R., Levin, L., Blair, N., DeMaster, D., Pope, R., Boehme, S., 1992. Spatial  
866 heterogeneity of benthos on the Carolina continental slope: large (100 km)-scale variation.  
867 *Mar. Ecol. Prog. Ser.* 88, 143-160.

868 Sibuet, M., Olu, K., 1998. Biogeography, biodiversity and fluid dependence of deep-sea cold-  
869 seep communities at active and passive margins. *Deep Sea Res. Part 2 Top. Stud.*  
870 *Oceanogr.* 45, 517-567.

871 Skarke, A., Ruppel, C., Kodis, M., Brothers, D., Lobecker, E., 2014. Widespread methane  
872 leakage from the sea floor on the northern US Atlantic margin. *Nat Geosci* 7, 657-661.

873 Thurber, A.R., Levin, L.A., Orphan, V.J., Marlow, J.J., 2012. Archaea in metazoan diets:  
874 implications for food webs and biogeochemical cycling. *Isme J* 6, 1602-1612.

875 Thurber, A.R., Levin, L.A., Rowden, A.A., Sommer, S., Linke, P., Kröger, K., 2013. Microbes,  
876 macrofauna, and methane: A novel seep community fueled by aerobic methanotrophy. *Limnol.*  
877 *Oceanogr.* 58, 1640-1656.

878 Tunnicliffe, V., Juniper, S.K., Sibuet, M., 2003. Reducing environments of the deep-sea floor, in:  
879 Tyler, P.A. (Ed.), *Ecosystems of the Deep Oceans*. Elsevier, pp. 81-110.

880 Valentine, D.L., Kastner, M., Wardlaw, G.D., Wang, X., Purdy, A., Bartlett, D.H., 2005.  
881 Biogeochemical investigations of marine methane seeps, Hydrate Ridge, Oregon. *Journal*  
882 *of Geophysical Research: Biogeosciences* 110, n/a-n/a.

883 Van Dover, C.L., 2000. *The ecology of hydrothermal vents*. Princeton University Press,  
884 Princeton, NJ.

885 Van Dover, C.L., 2007. Stable isotope studies in marine chemoautotrophically based ecosystems:  
886 An update, in: Michener, R., Lajtha, K. (Eds.), *Stable isotopes in ecology and*  
887 *environmental sciences*. Blackwell Publishing, Second Edition, pp. 202-237.

888 Van Dover, C.L., Aharon, P., Bernhard, J.M., Caylor, E., Doerries, M., Flickinger, W., Gilhooly,  
889 W., Goffredi, S.K., Knick, K.E., Macko, S.A., Rapoport, S., Raulfs, E.C., Ruppel, C.,  
890 Salerno, J.L., Seitz, R.D., Sen Gupta, B.K., Shank, T., Turnipseed, M., Vrijenhoek, R.,

891           2003. Blake Ridge methane seeps: characterization of a soft-sediment, chemo synthetically  
892           based ecosystem. *Deep Sea Res. Part 1 Oceanogr. Res. Pap.* 50, 281-300.

893   Van Dover, C.L., Trask, J.L., 2000. Diversity at Deep-Sea Hydrothermal Vent and Intertidal  
894           Mussel Beds. *Mar. Ecol. Prog. Ser.* 195, 169-178.

895   Vetter, R.D., Powell, M.A., Somero, G.N., 1991. Metazoan adaptations to hydrogen sulphide, in:  
896           Bryant, C. (Ed.), *Metazoan life without oxygen*. Chapman and Hall, London, UK, pp. 109-  
897           128.

898   Warwick, R.M., Clarke, K.R., 1993. Increased variability as a symptom of stress in marine  
899           communities. *J. Exp. Mar. Biol. Ecol.* 172, 215-226.

900   Whiticar, M.J., 1999. Carbon and hydrogen isotope systematics of bacterial formation and  
901           oxidation of methane. *Chem Geol* 161, 291-314.

902   Yoder, J.A., Atkinson, L.P., Stephen Bishop, S., Blanton, J.O., Lee, T.N., Pietrafesa, L.J., 1985.  
903           Phytoplankton dynamics within Gulf Stream intrusions on the southeastern United States  
904           continental shelf during summer 1981. *Continental Shelf Research* 4, 611-635.

905   Zar, J.H., 1999. *Biostatistical analysis*, 4th ed. Prentice Hall, Upper Saddle River, NJ.

906

907   **Figure Captions:**

908   **Figure 1.** Maps showing locations of sampling sites and closest known seeps (a) with detailed  
909   sampling at b) BCS and c) NCS. ■ = Microbial mat habitat push cores; ▲ = Mussel habitat push  
910   cores; ● = Background, soft-sediment habitats; ▲ = Mussel Ekman cores; △ = Mussel suction  
911   samples; □ = Microbial mat suction sample.

912

913 **Figure 2.** Habitats and sample collection at BCS mussel habitats (a), BCS microbial mat habitats  
914 (b), NCS mussel habitats (c), and NCS microbial mat habitats (d). Black lines represent 10-cm  
915 width of lasers on ROV. Images courtesy of Deepwater Canyons 2013 - Pathways to the Abyss,  
916 NOAA-OER/BOEM/USGS.

917

918 **Figure 3.** Mean macrofaunal density (ind. m<sup>-2</sup>) ( $\pm$  1 S.E.) of seep and background soft-sediment  
919 habitats from push core samples collected at BCS and NCS. Letters indicate statistical groupings  
920 ( $p > 0.05$ ) for each site tested separately.

921

922 **Figure 4.** Rarefaction via estimated number of taxa for a) BCS samples; b) NCS samples and c)  
923 pooled by sample type based on per sample untransformed data. For a and b, Mat, Mussel, and  
924 Background include push cores only. Mat = pooled microbial mat cores; Seep Cores = pooled  
925 mussel and microbial mat cores; Seep All = pooled mussel and microbial sediment cores, Ekman  
926 cores, and suction; All = pooled all samples.

927

928 **Figure 5.** Taxonomic composition of dominant macrofauna at BCS and NCS seep and  
929 background habitats collected from a) quantitative push cores b) Ekman cores and suction  
930 samples. Other Taxa includes Halacaridae, Cnidaria, Echinodermata, Nemertea, Sipuncula, and  
931 Turbellaria.

932

933 **Figure 6.** Non-metric multidimensional scaling of Bray-Curtis similarities of square-root  
934 transformed macrofaunal abundance data from push cores collected in BCS and NCS habitats.  
935 Circles and percentages indicate average similarity among cores for each habitat from SIMPER  
936 analysis. Stress value of 0.15 indicates a good to potentially useful confidence in the 2-  
937 dimensional representation (Clarke and Gorley, 2015).

938

939 **Figure 7.** Principal coordinate ordination of Bray-Curtis similarities of square-root transformed  
940 abundance data from sediment push cores collected at NCS habitats with environmental  
941 parameter vectors overlaid.

942

943

944

945

946

947

948

949

950

951

952

953

954 **Table 1.** Number of samples collected at Baltimore and Norfolk seep and background sites, including  
 955 push cores collected for infaunal analysis (Fauna) and sediment geochemistry analysis (SC), Ekman  
 956 cores, and suction samples.

Site	Habitat	Year	Latitude	Longitude	Depth (m)	Fauna	SC	Ekman	Suction
Baltimore BCS	Mat	2012	38.04369	-73.82568	412	1			
		2013	38.04799	-73.82915	366	2			
		2013	38.04995	-73.82191	402	1			
	Mussel	2013	38.04820	-73.82534	372	2			
		2013	38.04848	-73.82193	400	2			
		2013	38.05026	-73.82194	400			1	
	Background	2012	38.09171	-73.80519	446	3			
		2012	38.04369	-73.82568	412	1			
Norfolk NCS	Mat	2013	36.86584	-74.49083	1602	1	1		
		2013	36.86664	-74.49026	1590	1	1		
		2013	36.87140	-74.47758	1487	1	1		
		2013	36.87087	-74.47290	1457	1	1		
		2013	36.87123	-74.47396	1467	1			
		2013	36.86586	-74.49082	1612				1
	Mussel	2013	36.86786	-74.48871	1576	1	1		
		2013	36.86827	-74.48822	1565	1			
		2013	36.86711	-74.48936	1585	1	1		
		2013	36.87139	-74.47633	1482	1	1		
		2013	36.87168	-74.47813	1487	1			
		2013	36.87148	-74.47626	1483			1	
		2013	36.86928	-74.48736	1612				1
		2013	36.87029	-74.48748	1612				1
	Background	2013	36.86934	-74.48698	1612				1
2013		37.03347	-74.45039	1622	1				
2013		37.03343	-74.45032	1619	1				
		2013	37.03343	-74.45030	1620	1	1		

957

958

959

960 **Table 2.** Diversity ( $H' \log_e$ ), evenness ( $J'$ ), and multivariate dispersion (MVDISP) of  
 961 macrofaunal communities collected from cores at Baltimore and Norfolk seep and background  
 962 habitats.

Site	Habitat	N	$J'$		$H'(\log_e)$		MVDISP
Baltimore BCS	Mat	4	0.49	(0.06)	0.96	(0.11)	0.97
	Mussel	4	0.87	(0.03)	2.82	(0.07)	0.75
	Background	4	0.92	(0.02)	2.80	(0.07)	0.52
Norfolk NCS	Mat	5	0.70	(0.07)	1.96	(0.26)	1.56
	Mussel	5	0.85	(0.03)	1.95	(0.12)	0.89
	Background	3	0.92	(0.03)	2.37	(0.19)	1.03

963  
 964  
 965  
 966  
 967  
 968  
 969  
 970  
 971  
 972  
 973  
 974  
 975  
 976  
 977  
 978



979 **Table 3.** Similarity among habitats (above diagonal), within-habitat similarity (diagonal, bold), and PERMANOVA probabilities  
 980 (below diagonal) based on Bray-Curtis similarities of square-root transformed abundance data for the push cores. Comparisons with  
 981 suction and Ekman grab samples were based on Bray-Curtis similarities of presence/absence transformed abundance data.

Site		BCS				NCS					
Habitat		Mats	Mussels	Background	Ekman Mussel	Mats	Mussels	Background	Ekman Mussel	Suction Mussel	Suction Mat
BCS	Mats	<b>54.3</b>	19.9	14.2	14.5	29.9	11.0	16.2	26.1	12.2	16.5
	Mussels	0.002	<b>55.0</b>	49.5	20.1	25.7	16.9	33.1	41.3	41.4	35.3
	Background	0.001	0.056	<b>58.7</b>	17.9	25.4	17.6	33.6	37.1	34.6	36.0
	Ekman Mussel	-	-	-	<b>-</b>	21.0	17.9	20.4	34.3	26.7	26.7
NCS	Mats	0.049	0.021	0.015	-	<b>32.1</b>	15.5	20.1	31.2	30.4	31.8
	Mussels	0.001	0.001	0.001	-	0.002	<b>52.3</b>	20.0	43.9	27.1	14.8
	Background	0.006	0.014	0.013	-	0.031	0.004	<b>50.6</b>	31.5	24.2	31.6
	Ekman Mussel	-	-	-	-	-	-	-	<b>-</b>	53.3	43.3
	Suction Mussel	-	-	-	-	-	-	-	-	<b>-</b>	54.12
	Suction Mat	-	-	-	-	-	-	-	-	-	<b>-</b>

982  
 983  
 984  
 985  
 986

987 **Table 4.** Mean ( $\pm 1$  S.E.) sediment geochemical properties for cores collected at Norfolk seep and background habitats.

<b>Habitat</b>	<b>N</b>	<b><math>\delta^{13}\text{C}</math></b>		<b>%C</b>		<b><math>\delta^{15}\text{N}</math></b>		<b>%N</b>		<b>C:N</b>		<b>%Mud</b>	
Mat	4	-25.41	(0.28)	2.22	(0.32)	5.32	(0.23)	0.30	(0.03)	8.53	(0.29)	61.74	(3.91)
Mussel	3	-39.97	(0.61)	4.41	(0.20)	2.78	(0.22)	0.73	(0.01)	7.01	(0.20)	76.21	(2.39)
Background	3	-21.15	(0.05)	2.36	(0.30)	7.74	(0.97)	0.36	(0.04)	7.62	(0.89)	95.46	(0.52)

988

989

990

991

992

993

994

995

996

997

998

999

1000 **Table 5.** Results from the distance-based linear modeling (DISTLM) of environmental variables  
 1001 with Norfolk microbial mat, mussel, and background soft-sediment communities using the AICc  
 1002 criteria and “best” model selection.

Variable	SS(trace)	Pseudo-F	P	Prop.
$\delta^{13}\text{C}$	8077.3	3.613	0.005	0.311
Percent Carbon	6423.8	2.630	0.014	0.247
$\delta^{15}\text{N}$	6493.7	2.668	0.010	0.250
C:N	4308.6	1.592	0.160	0.166
Mud Content	6815.3	2.847	0.010	0.262
Depth	6266	2.545	0.019	0.241

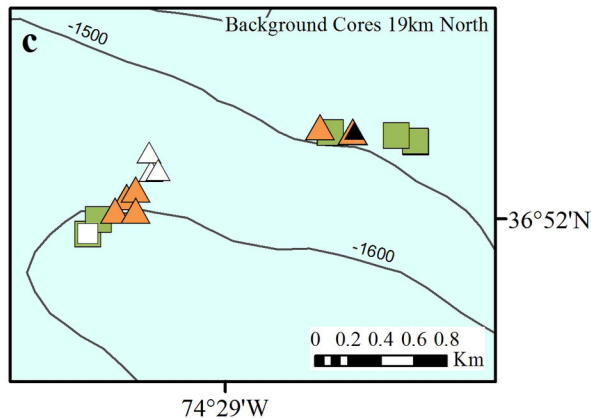
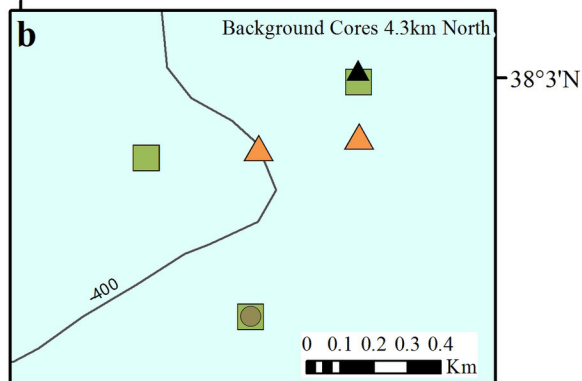
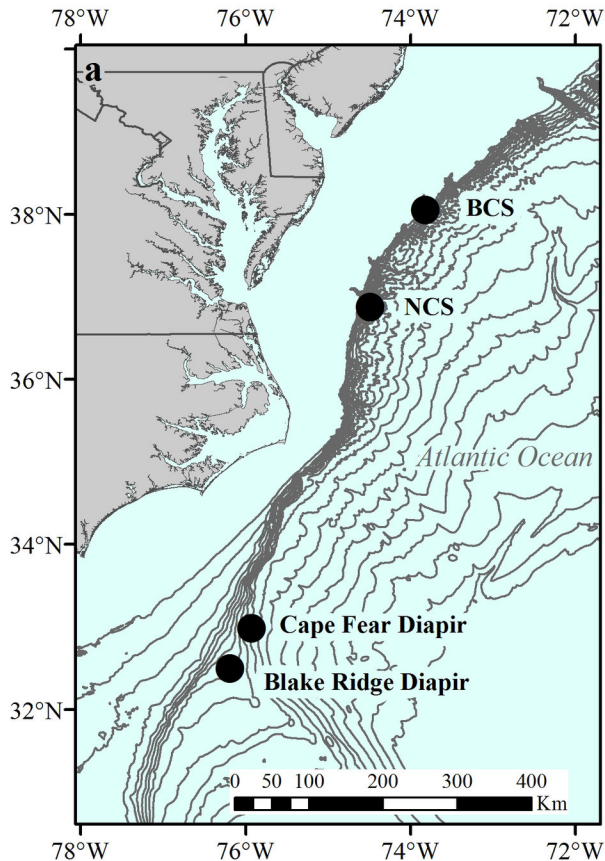
AICc	R <sup>2</sup>	RSS	Selections
80.002	0.57755	10968	$\delta^{13}\text{C}$ , Mud Content
80.606	0.3111	17886	$\delta^{13}\text{C}$
81.083	0.52932	12221	$\delta^{13}\text{C}$ , Depth
81.288	0.26249	19149	Mud Content
81.298	0.51908	12487	$\delta^{15}\text{N}$ , Mud Content
81.455	0.2501	19470	$\delta^{15}\text{N}$
81.491	0.24741	19540	Percent Carbon
81.552	0.50674	12807	Percent Carbon, Mud Content
81.571	0.24133	19698	Depth
81.779	0.49539	13102	Percent Carbon, Depth
Total SS(trace)		25964	

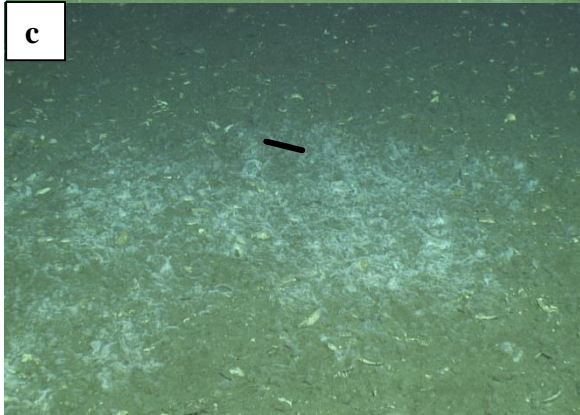
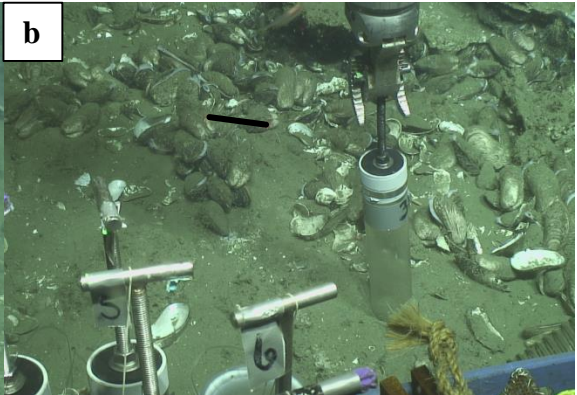
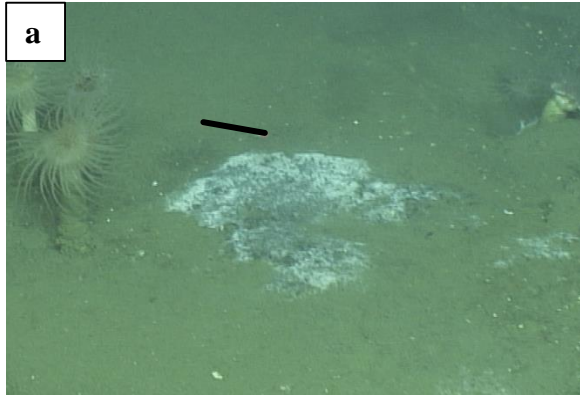
1003  
 1004  
 1005  
 1006  
 1007  
 1008  
 1009

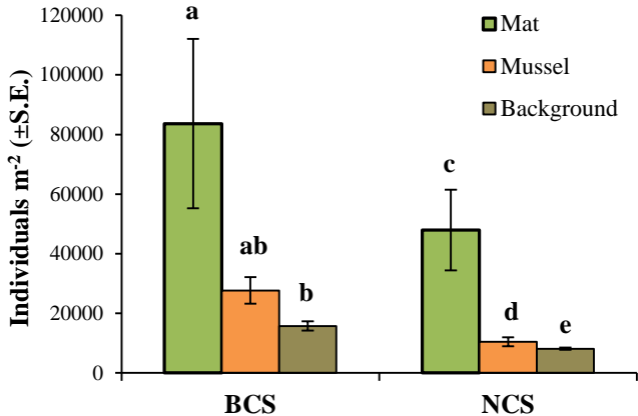
1010 **Table 6.** Summary of macrofaunal seep sediment and regional infaunal studies including closest geographic seeps, comparable depths,  
 1011 and observed high densities.

Study Location	Region	Seep Habitat	Depth (m)	Density individuals m <sup>-2</sup>	Max Density individuals m <sup>-2</sup>	Source
BCS	NW Atlantic	Microbial mat	366-412	83649 ±28466	137757	This study
BCS	NW Atlantic	Mussel beds	372-400	27646 ±4464	40758	This study
NCS	NW Atlantic	Microbial mat	1467-1602	47962 ±13547	78357	This study
NCS	NW Atlantic	Mussel beds	1482-1585	10427 ±1558	15482	This study
Blake Ridge Diapir	NW Atlantic	Microbial mat	2250	800 ±506	2400	Robinson et al., 2004
Blake Ridge Diapir	NW Atlantic	Mussel beds	2250	5000 ±1400	6400	Robinson et al., 2004
Håkon Mosby	NE Atlantic	Frenulate field	1256	92955 ±21617	-	Decker et al., 2012
Gulf of Guinea	SE Atlantic	Mussel beds	3160	22306 -	-	Menot et al., 2010
Costa Rica	SW Atlantic	Microbial mat	376-1854	18060 ±8190	-	Levin et al., 2015
Green Canyon	Gulf of Mexico	Microbial mat	700	198950 ±78150	277100	Robinson et al., 2004
Atwater Canyon,	Gulf of Mexico	Microbial mat	1934	36400 -	-	Robinson et al., 2004
California Margin	E Pacific	Microbial mat	525	62160 -	-	Levin et al., 2006
New Zealand	W Pacific	Ampharetid bed	1057	56728 ±4784	84000	Thurber et al., 2013
Nile Delta	Mediterranean	Microbial mat	1700	2783 ±451	-	Ritt et al. 2011
BCS	NW Atlantic	Background	412-446	15719 ±1582	17694	This study
NCS	NW Atlantic	Background	1619-1622	8110 ±380	8847	This study
Gay Head-Bermuda	NW Atlantic	Background	400	6081 -	-	Sanders et al., 1965
Baltimore Slope	NW Atlantic	Background	550	6546 ±2214	10934	Robertson et al., 2015
Gay Head-Bermuda	NW Atlantic	Background	1500	1719 -	-	Sanders et al., 1965
Mid-Atlantic Slope	NW Atlantic	Background	1613	4953 ±754	6911	Maciolek et al., 1987
Cape Hatteras	NW Atlantic	Background	1410-1535	13219 ±3942	17733	Blake and Hilbig, 1994

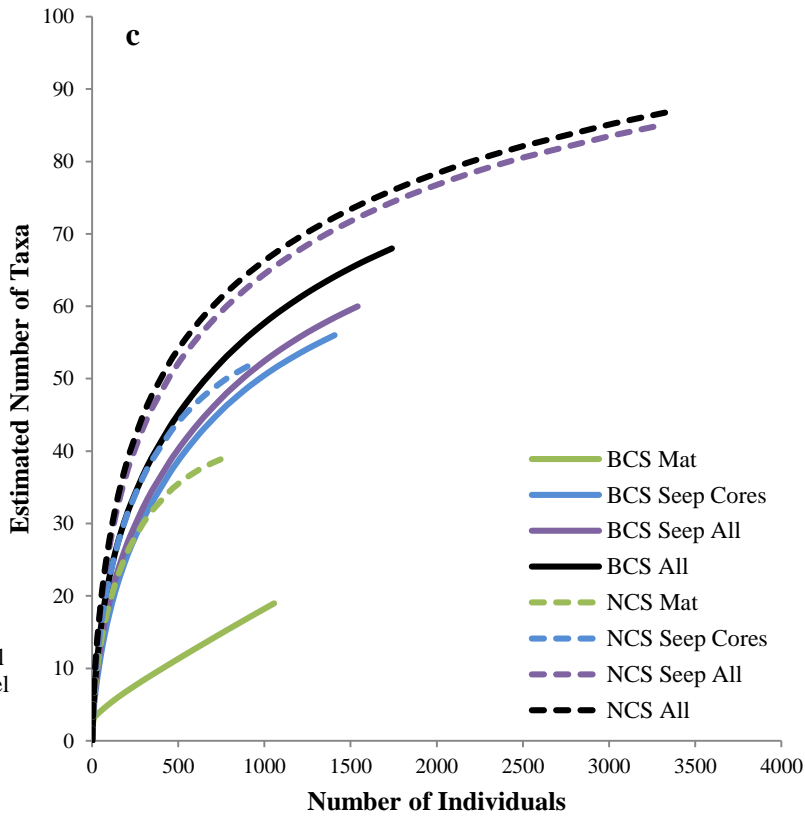
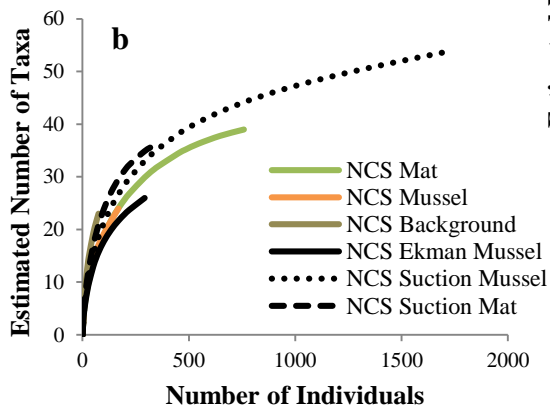
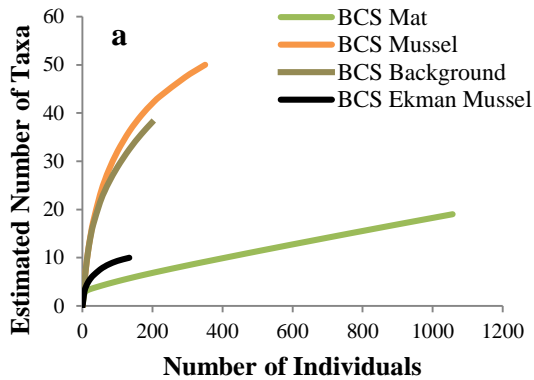
1012 Cape Lookout NW Atlantic Background 1380-1690 5017 - - Blake and Grassle, 1994

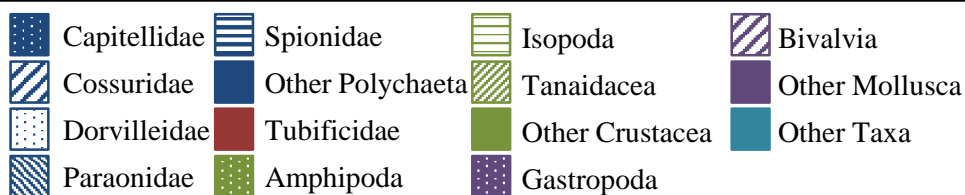
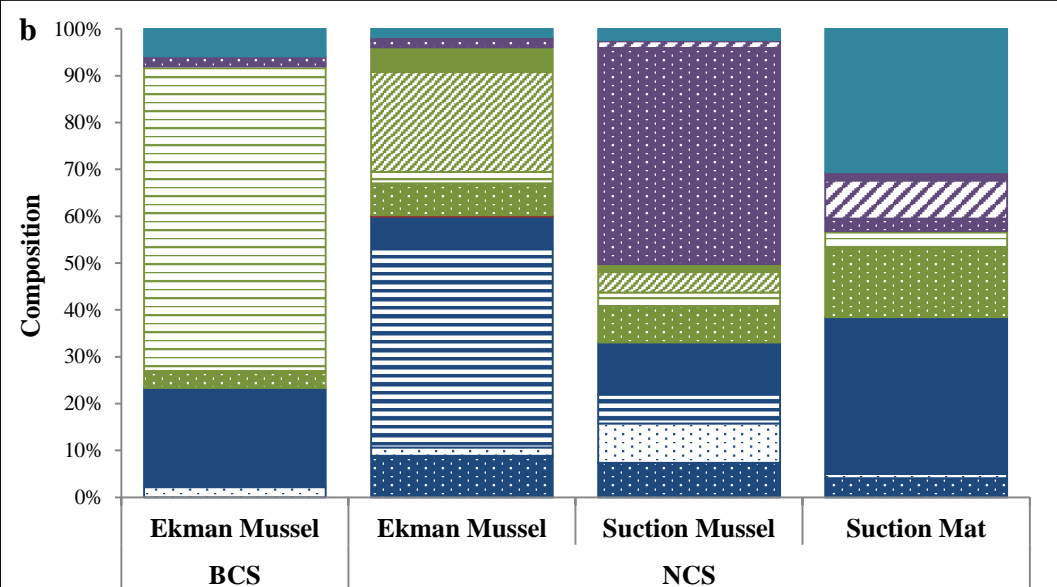
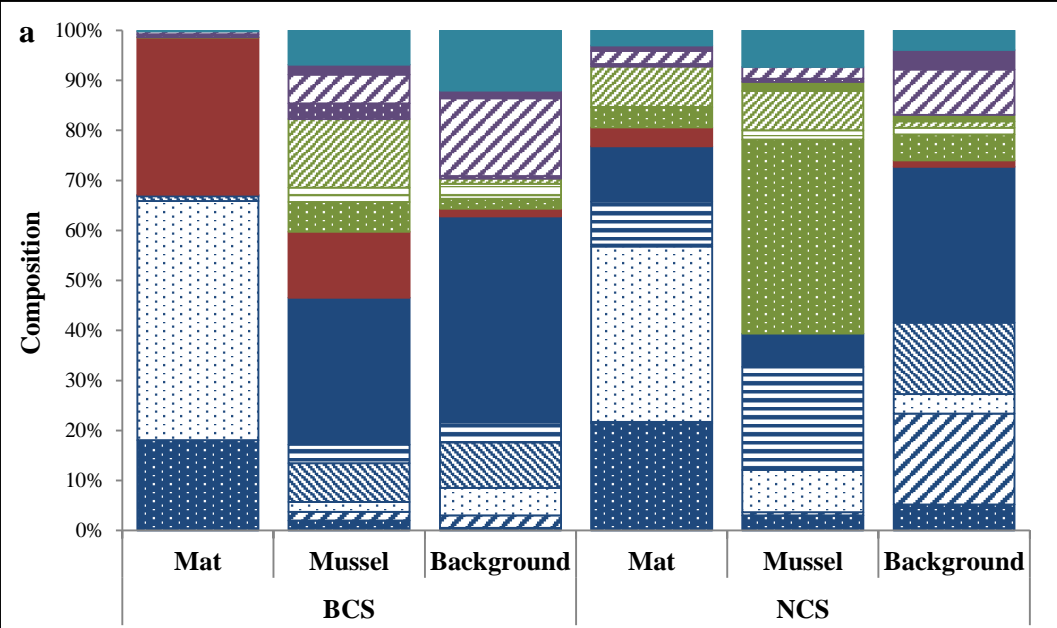




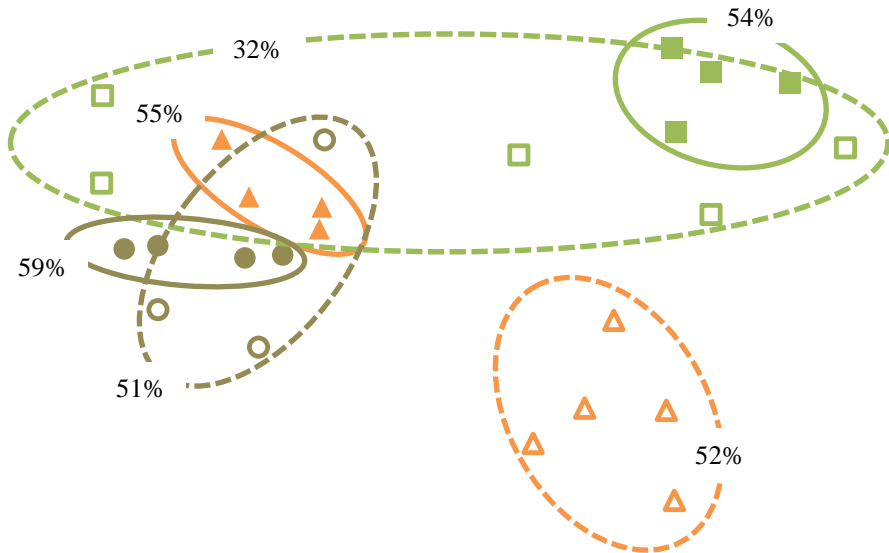








Stress = 0.15



■ BCS Mat

▲ BCS Mussel

● BCS Background

□ NCS Mat

△ NCS Mussel

○ NCS Background

