

Multi-decadal simulation of estuarine sedimentation under sea level rise with a response-surface surrogate model

Matthew W. Brand^{a,*}, Leicheng Guo^b, Eric D. Stein^c and Brett F. Sanders^{a,d}

^aDepartment of Civil and Environmental Engineering, Calit² Building, University of California, Irvine, Irvine, CA 92617

^bState Key Lab of Estuarine and Coastal Research, East China Normal University, Shanghai 200241, China

^cSouthern California Coastal Water Research Project, Costa Mesa, CA 92626, USA

^dDepartment of Urban Planning and Public Policy, University of California, Irvine, Irvine, CA 92617, USA

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ABSTRACT

Multi-decadal prediction of estuarine sedimentation with high-fidelity hydromorphodynamic models presents high computation costs, especially when accounting for stochasticity and uncertainty. A Stochastic model for Multi-decadal Estuarine Sedimentation (SeAMLESS) is formulated here to support a specific decision-need related to resilience planning and coastal management: estimating future sedimentation and dredging within a sedimentation basin for different scenarios of sea level rise and rules for dredging. SeAMLESS combines a reduced-dimension process model and a response-surface surrogate model to yield an ordinary differential equation that can be integrated over stochastic time series of storm events. Applications show that SeAMLESS can predict probabilities and amounts of future basin sedimentation and dredging with minimal loss of accuracy, compared to a high-fidelity model, while delivering $\mathcal{O}(10^4 - 10^5)$ reduction in computational costs.

1. Introduction

Estuaries are embayments open to coastal oceans that receive freshwater runoff (Pritchard, 1967), and are increasingly confronted by climate change and the effects of urban development around the embayment and/or in the watershed such as land reclamation and waste discharges (Lotze et al., 2006). Estuaries represent critical coastal habitats that support ecosystems including birds, fish and invertebrates (McLusky and Elliott, 2004). Additionally, estuaries provide benefits to society (or ecosystem services) including recreational opportunities for coastal communities, pollutant and nutrient processing, support for the shipping, defense and fishing industries, and urban amenities such as access to wildlife, seafood, and open spaces (Barbier et al., 2011). These many benefits are often in competition and need to be balanced, thus posing challenges for management (Elliott and Whitfield, 2011).

Excess sedimentation is one of the costliest and potentially environmentally damaging management challenges of estuaries (Chesapeake Bay Program, 2006; Kiefer et al., 2000). Excess deposition negatively affects navigation and damages ecosystems by submerging wetland habitat and changing inundation regimes. Changes to bathymetry can negatively affect circulation and water quality, and the introduction of non-native or invasive species through dredging operations can reduce or degrade habitat for sensitive marsh animals (Haltiner et al., 1996). Management options such as source control and dredging are costly. Dredging requirements from federal, state and local agencies in the

 mbrand1@uci.edu (M.W. Brand)
ORCID(s):

16 United States are expected to reach \$2 billion/year, with the U.S. Army Corps of Engineers spending \$1.3 billion in
17 2015 alone (US Army Corps of Engineers et al., 2016). Moreover, predicting future dredging requirements is difficult
18 due to uncertain variability in estuarine dynamics and inherent complexities in sedimentation, watershed and tidal
19 dynamics and significant changes due to human influences (Bull et al., 2002). In particular, estuarine sedimentation is
20 driven by watershed runoff and tidal currents which vary with hourly and longer time scales and are also affected by
21 sea level rise (SLR) and land-use/management changes (McLusky and Elliott, 2004; Griggs et al., 2017). High-fidelity
22 deterministic models such as Delft3D are oftentimes used to answer questions on estuarine sedimentation (Yu et al.,
23 2012; Thanh et al., 2019). Arid regions such as southern California experience highly episodic hydrology with stream
24 flows that vary by several orders of magnitude over time scales of hours to a few days. Here, greater than 90% of
25 sedimentation occurs quickly during the most intense storms (Kroll, 1975; Warrick and Milliman, 2003). The wide
26 range of variability in process magnitude and time scales, along with significant uncertainty in model input parameters,
27 makes it virtually impossible to deterministically predict future sedimentation and associated dredging requirements,
28 thereby motivating the need for stochastic modeling approaches.

29 Stochastic modeling, notably with Monte Carlo simulations, has increasingly been used in environmental studies
30 to characterize ranges and likelihoods of system outcomes over time scales of days, months, years and even decades or
31 centuries (Fedra, 1983). Furthermore, stochastic modeling is especially useful for environmental management because
32 probabilities are assigned to a range of outcomes, creating opportunities to enhance dialogue and deliberation among
33 stakeholders towards development of cost effective and fair management measures (Isukapalli et al., 1998). However,
34 a major limitation of stochastic modeling stems from complex, interdependent, environmental process dynamics that
35 require use of mechanistic models with high computational demands (Sparrevik et al., 2012), or so-called *high-fidelity*
36 *models*. Use of high-fidelity models make it very difficult or even impossible to complete the thousands or more Monte
37 Carlo simulations needed to account for stochasticity (Liu et al., 2007).

38 Fast-running surrogate models have emerged as a promising alternative to high-fidelity models with demonstrated
39 ability to radically reduce computational costs (Razavi et al., 2012). One type of surrogate model is a response-surface
40 function, which captures relationships (e.g., through polynomial approximations) between several explanatory vari-
41 ables of a system (Razavi et al., 2012; Koziel and Leifsson, 2013). These response surfaces are developed by running
42 a relatively modest number of high-fidelity model simulations over a carefully selected set of input parameter values
43 (Razavi et al., 2012; Koziel and Leifsson, 2013). Once the surrogate model is trained to quantify the desired system
44 outcomes over a suitable parameter space, thousands or even millions of Monte Carlo simulations can easily be com-
45 pleted with significantly reduced runtimes compared to the mechanistic model (Fedra, 1983; Isukapalli et al., 1998; Liu
46 et al., 2007; Razavi et al., 2012; Koziel and Leifsson, 2013). Examples of response surface surrogate models in water
47 resources include optimization studies (Razavi et al., 2012), improving aquifer management and regulatory support

48 (Schultz et al., 2004; Sreekanth and Datta, 2011; Kourakos and Mantoglou, 2009), and model selection (Mohammadi
49 et al., 2018). Another type of surrogate model is a lower fidelity model, which can be viewed as a lower-resolution
50 versions of a high resolution model (Razavi et al., 2012).

51 Within the field of sediment transport and hydromorphodynamics, Berends et al. (2019) present a low fidelity
52 surrogate modeling approach for characterizing the uncertainty of estuarine sedimentation predicted by a high-fidelity
53 model. The lower resolution model reduced run times by a factor of sixteen compared to the high fidelity model (or
54 6.25%), which allowed for an 85% reduction in the overall compute time needed to map out sedimentation patterns
55 (and uncertainties) using a Monte Carlo approach. Additionally, Mohammadi et al. (2018) describe a response surface
56 surrogate model for a hydro-morphodynamic model (TELEMAC-MASCARET with SISYPHE module) of the lower
57 Rhine river where Bayesian model selection is applied to discern the best choice among empirical sediment transport
58 equations. Here, the response surface surrogate model was applied to quantify the sedimentation uncertainty and
59 model dependence on uncertain parameters. Nevertheless, use of surrogate modeling within hydro-morphodynamic
60 simulation is relatively new.

61 In this paper, we present a new approach to simulate estuarine sedimentation and rule-based dredging over multiple
62 decades by combining a reduced-dimension process model and a response surface surrogate model within a stochastic
63 Monte Carlo simulation framework. We term this framework the StochAstic model for Multi-decadaL Estuarine Sedi-
64 mentation (SeAMLESS). The proposed framework is a response to the engagement of coastal stakeholders in southern
65 California under the SedRISE project (Resilient Infrastructure and Sustainable Environments) funded by the Ecolog-
66 ical Effects of Sea Level Rise Program (EESLR) of the National Centers for Coastal Ocean Science at the National
67 Oceanic and Atmospheric Administration. The EESLR program supports transdisciplinary research projects where
68 stakeholders are engaged with the aim of closing the "usability gap" between what scientists and decision-makers
69 consider useful climate-rated knowledge (DeLorme et al., 2016).

70 In southern California and elsewhere, there is a need to plan for accelerating rates of sea level rise and consider
71 changes to sediment management to yield outcomes favorable to coastal ecosystems, flood risk management, and
72 financial stewardship (Passeri et al., 2015; Morris et al., 2016; Bilskie et al., 2016). Sea level rise threatens submergence
73 of coastal wetlands (Thorne et al., 2018) and increased flood risk (Gallien et al., 2011; Sanders et al., 2020), yet the
74 amounts of sea level rise is highly uncertain and partial mitigation of these impacts may be possible through altered
75 sediment management practices (Sanders and Grant, 2020; Ulibarri et al., 2020).

76 The remainder of the paper is organized as follows: Section 2 presents the theoretical formulation of SeAMLESS
77 including the development of a reduced-dimension process model that can be integrated efficiently with the aid of a
78 response-surface surrogate model. Section 3 presents an application of SeAMLESS to a site in southern California,
79 Newport Bay, where sediment fluxes have been subject to management since the 1980s and there is a need to consider

80 sea level rise in adaptation and resilience planning. Section 4 continues with a second application of SeAMLESS to
 81 a larger, stylized, system for which a high-fidelity model can more easily be run over long time scales. This affords
 82 a more critical examination of the SeAMLESS framework based on side-by-side comparisons of SeAMLESS and
 83 high-fidelity model simulations for the same series of storm events. Section 5 closes the paper with conclusions.

84 2. SeAMLESS Formulation

85 2.1. Decision-Support with High-Fidelity Process Models

High-fidelity models for hydro-morphodynamics such as Delft3D (Lesser et al., 2004) numerically solve process-based equations describing fluid flow, sediment transport and movement of the fluid/bed interface at fine spatial and temporal scales (Lesser et al., 2004). The solution state of a hydro-morphodynamic model can be defined by $\mathbf{U}(\mathbf{x}, t)$ where $\mathbf{x} \in \mathbf{R}^3$ represents the spatial dimensions, t , represents time, and elements of \mathbf{U} include the fluid velocity \mathbf{u} , ground elevation z_b , sediment concentrations c_i for a set of i grain sizes, fluid pressure p , and fluid density ρ . A general representation of hydro-morphodynamic models is a system of partial differential equations for $\mathbf{U}(\mathbf{x}, t)$ that is solved on a spatial domain D and time interval $t = (0, T)$ as follows,

$$\frac{\partial \mathbf{U}}{\partial t} = \mathbf{M}_U(\mathbf{U}, \mathbf{I}) \quad (1)$$

where \mathbf{M}_U is an operator representative of the hydro-morphodynamic simulation model, and \mathbf{I} contains the model inputs that influence the solution, including initial conditions, boundary conditions and model parameters. Once the solution is simulated, additional operations follow to produce information that is used for decision-making. That is, decision-makers likely won't want to know the spatial distributions of fluid velocity, pressure, and density, but rather some integral measures reflective of the level of sedimentation in the system (total volume of sediment) and/or the overall water quality of the system. Hence, there exists a set of decision variables \mathbf{D} that are obtained by operating on the solution state (\mathbf{M}_D) as follows,

$$\mathbf{D} = \mathbf{M}_D(\mathbf{U}) \quad (2)$$

86 Decision variables need to be relatively simple, and hence \mathbf{D} will generally contain orders of magnitude fewer scalar
 87 elements than \mathbf{U} , and maybe even just one or two, such as the maximum height of the sediment bed within a collection
 88 basin or the total volume of sediment in collection basin. Generally, to meet decision-support needs with high-fidelity
 89 models, numerous (computationally demanding) simulations are run (Eq. 1) to generate output describing the spatial
 90 and temporal evolution of the solution state, and then results are post-processed (Eq. 2) to distill decision variables as

91 shown in Fig. 1a.

92 2.2. Reduced Dimension Process Model and Surrogate Model

A common decision variable is the spatially averaged value of a system property, $\bar{\mathbf{U}}(t)$, taken over a subset of the model domain, $D_D \subset D$. The subdomain D_D is chosen to align with the decision-making needs, and in the case of estuarine sedimentation, aligns with the spatial extent of a regulatory sedimentation zone (or basin) where dredging is permitted to occur. A general representation of the time-wise changes of the spatially-averaged solution state within the regulatory basin is given by the following ordinary differential equations (ode) that is solved for the time interval $t = (0, T)$,

$$\frac{d\bar{\mathbf{U}}}{dt} = \bar{\mathbf{M}}_U(\bar{\mathbf{U}}, \mathbf{I}) \quad (3)$$

where the operator $\bar{\mathbf{M}}_U$ represents the bulk effects of hydromorphodynamic processes over the spatial extent of subdomain D_D . In the case of sedimentation and dredging, the key decision-variable is simply the total volume of sediment that has accumulated in a regulatory sedimentation basin, thus Eq. 3 simplifies as follows,

$$\frac{d\bar{z}_b}{dt} = \bar{\mathbf{M}}_z(\bar{\mathbf{U}}, \mathbf{I}) \quad (4)$$

where $\bar{\mathbf{M}}_z$ is a refinement of $\bar{\mathbf{M}}_U$ that supports output of a single scalar describing the rate of change in average bed elevation as a function of the system state and inputs. It is possible to analytically derive the operator $\bar{\mathbf{M}}_z$ from the operator $\bar{\mathbf{M}}_U$ and the subdomain D_D , discretize the resulting process-based terms, and numerically integrate the solution to yield predictions of $\bar{z}_b(t)$. However, this process-based approach is not pursued herein. Rather, we introduce a data-driven approach using a response surface surrogate model, $\tilde{\mathbf{M}}_z$, as follows,

$$\frac{d\bar{z}_b}{dt} = \tilde{\mathbf{M}}_z(\bar{\mathbf{U}}, \mathbf{I}) \quad (5)$$

93 whereby the surrogate model depicts the time rate of change of the basin-average sediment bed height based on model
 94 inputs and system conditions that are found, through a diagnostic process, to represent the primary controls. Further-
 95 more, the response surface surrogate model is quantified by solving the high-fidelity model over a representative range
 96 of the control variables, which essentially organizes a database or library of known responses that can be accessed as
 97 needed to numerically integrate Eq 5. Hence, by combining a reduced-dimension process model and a response sur-
 98 face surrogate model, decision-support needs can be met by numerically integrating an ordinary differential equation
 99 (Eq. 5) while leveraging data produced by a high-fidelity model as shown in Fig. 1b.

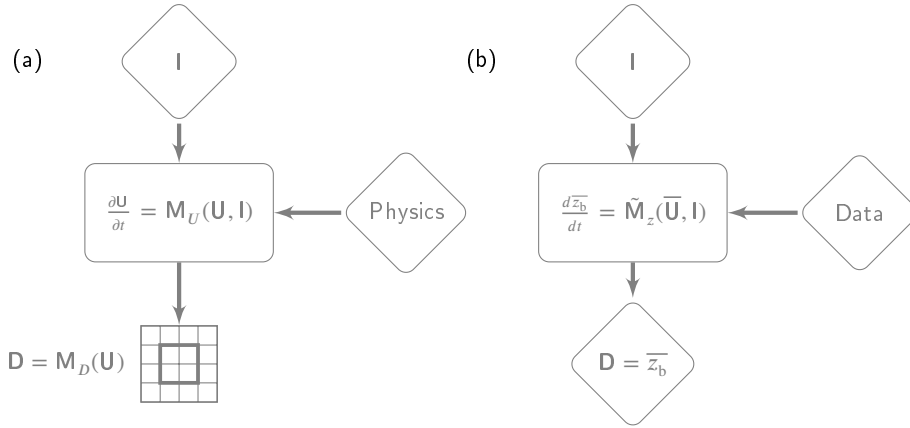


Figure 1: Estimation of decision-support variables, D , using: (a) high-fidelity method involving numerical solution of partial differential equations based on flow physics and post-processing of gridded model output and (b) proposed reduced-dimension, response-surface surrogate modeling approach that uses "Data" generated by a high-fidelity model.

100 Site-specific considerations will influence the design of the response-surface surrogate model, \tilde{M}_z Eq. 5, such as the
 101 size of the estuary, the configuration of the sedimentation basin, the tidal dynamics of the estuary, and the magnitude
 102 and variability of streamflow into the estuary. In southern California, there are many tidally-influenced lagoons, flood
 103 control channels, harbors and embayments subject to sedimentation and dredging. These systems are characterized by
 104 relatively short lengths and negligible tidal amplification. Tides in the region have micro-tidal amplitudes (~ 1 m) with
 105 temporal asymmetry that favors export of coarse and fine sediment to the coastal ocean, although in some systems, tidal
 106 asymmetry may favor the import of coarse or fine sediment (Guo et al., 2018; Nidzicko, 2010). Additionally, inputs of
 107 streamflow and sediment are highly episodic. Sediment loads vary by orders of magnitude with storm events (Warrick
 108 and Milliman, 2003), which last less than a day, and more than 90% of sedimentation occurs during the most intense
 109 storms (Kroll, 1975; Warrick and Milliman, 2003). Combining mass balance considerations and sub-daily time scale
 110 of storm events, we can re-formulate Eq. 5 as follows,

$$\frac{d\bar{z}_b}{dt} = \frac{1}{\rho_s A} \frac{dm}{dt} \quad (6)$$

where m and A are the sediment mass and planform area of the sedimentation basin, respectively and ρ_s is the dry bed density of the sediment. Furthermore, we can integrate over the time scale of a storm event, T , to yield an equation for event-based change in sediment basin elevation as follows,

$$\Delta\bar{z}_b = \int_t^{t+T} \left(\frac{d\bar{z}_b}{dt} \right) dt = \frac{\Delta m}{\rho_s A} \quad (7)$$

and thus we can advance the sediment basin elevation in time from t_i to t_{i+1} as follows,

$$(\bar{z}_b)_{i+1} = (\bar{z}_b)_i + \frac{\Delta m_i}{\rho_s A} \quad (8)$$

subject to an initial condition $(z_b)_0$ at time t_0 which focuses attention on the need for a surrogate model that characterizes the event-based deposition of mass within the sediment basin, Δm_i . Taking the event-based watershed load of sediment to be L_i , we can write the event based deposition as,

$$\Delta m_i = \eta_i L_i \quad (9)$$

where η represents the capture efficiency of the sedimentation basin, i.e., the fraction of the sediment load from a storm event that is deposited in the basin. Moreover, through a sensitivity analysis using a high-fidelity model (described in the following section), we find that the primary factors affecting the capture efficiency of the sediment basin are the peak discharge of the storm event, Q , the average basin elevation \bar{z}_b , and the tidal conditions (which we denote as ϕ). Hence, event-based changes in sediment mass required to update Eq. 8 are computed using a response surface surrogate model for capture efficiency, $\tilde{\eta}$, as follows,

$$\Delta m_i = \tilde{\eta}(Q_i, (\bar{z}_b)_i, \phi_i) L_i \quad (10)$$

111 which points to the need for high-fidelity model simulations over a parameter space defined by ranges in Q , \bar{z}_b , and
 112 ϕ_i to create a library of solutions from which values of η can be estimated by interpolation as (e.g., Razavi et al.,
 113 2012). Further detail on surrogate model parameterization is left for the next section, which presents an application of
 114 SeAMLESS to a site in southern California.

Two final factors must be considered to support multi-decadal simulations of sedimentation: sea level rise and dredging events. The former is approached by assuming that the height of the sediment bed is measured with respect to a tidal datum, mean sea level, which implies that increases in mean sea level correspond to decreases in the bed elevation. Secondly, dredging events are modeled by assuming that removal of sediment occurs when the sediment bed elevation reaches a trigger point, $(\bar{z}_b)_{\text{trig}}$, and that the sediment bed elevation is lowered to the initial height $(\bar{z}_b)_0$. Moreover, we assume that the post-dredging sediment height and trigger height are also measured relative to a tidal datum, which preserves the range of depths that occur in the sedimentation basin as sea level rises. Combining Eqs. 8

and 10, the final update equation is given as follows,

$$(\bar{z}_b)_{i+1} = \begin{cases} (\bar{z}_b)_i + \tilde{\eta} (Q_i, (\bar{z}_b)_i, \phi_i) \frac{L_i}{\rho_s A} - (\Delta z_{SLR})_i, & \text{if } \bar{z}_b \leq (\bar{z}_b)_{\text{trig}} \\ (\bar{z}_b)_0, & \text{if } \bar{z}_b > (\bar{z}_b)_{\text{trig}} \end{cases} \quad (11)$$

115 subject to the initial condition given by $(\bar{z}_b)_0$, and where $(\Delta z_{SLR})_i$ represents the change in (absolute) sea level between
 116 t_i and t_{i+1} . Note that increases in sea level rise act against the effect of sedimentation. Hence, as rates of sea level rise
 117 increase, the amount of sedimentation required to trigger a dredging event increases. Moreover, in the event that the
 118 rate of sea level rise is faster than the rate of sedimentation, it is not possible for the sediment bed elevation to reach
 119 the trigger height and thus no dredging events occur.

120 The SeAMLESS framework described herein represents a significant departure from previous sedimentation mod-
 121 eling frameworks involving surrogate modeling (e.g., Mohammadi et al., 2018; Berends et al., 2019). Whereas previous
 122 work focused on more efficiency estimation of uncertainty in sedimentation (Berends et al., 2019) and improved se-
 123 lection of empirical equations for sediment transport rates (Mohammadi et al., 2018), herein we use a response surface
 124 surrogate model to evaluate the right hand side of reduced dimension model described by an ordinary differential
 125 equation, which in turn is configured to account for human influences on sedimentation (dredging events) as well as
 126 changes in mean sea level due to sea level rise. With this approach, the SeAMLESS framework is positioned to make
 127 multi-decadal predictions of sedimentation and dredging under different sea level rise scenarios and dredging rules,
 128 which is responsive to coastal management decision-making needs in southern California and elsewhere.

129 2.3. Surrogate Model Construction Process

130 Constructing the response surface surrogate model will rely upon modeling expertise coupled with an iterative
 131 process to identify key model sensitivities. It is highly unlikely that two surrogate models constructed for different
 132 regions will ever look the same, or even that key state variable and parameters will remain constant. The general
 133 overview of model construction is provided to the reader below in list format.

- 134 1. Develop, calibrate, and validate high-fidelity hydromorphodynamic model using best available data,
- 135 2. Identify basin with a well-defined geographic boundary and major controls on basin elevation,
- 136 3. Utilize high-fidelity model to explore parameter and state variable space to identify key model sensitivities (flow,
 137 tidal condition, sediment characteristics, dredging, basin elevation) with which to construct response surface
 138 surrogate model,
- 139 4. Compare the surrogate model to high fidelity model output and measured basin elevations to evaluate model
 140 performance,

- 141 5. Utilize Monte-Carlo or Monte-Carlo Markov Chain model to simulate forcing data and use surrogate model to
142 predict future basin elevations.

143 3. SeAMLESS Application to Newport Bay

144 3.1. Site Description

145 SeAMLESS is applied to Newport Bay, a short (10km) estuary in southern California that receives sediment input
146 primarily from San Diego Creek (SDC, shaded blue), as shown in Figure 2. The upper portion of the bay contains
147 protected wetland habitat while the lower portion is used as a recreational harbor. Historically, excessive sedimentation
148 from SDC was the greatest driver of habitat change in Newport Bay, threatening to turn protected wetlands into upland
149 habitat (Trimble, 1997). As a result, a sediment management strategy for a total maximum daily load (TMDL) for SDC
150 was implemented, stipulating a 50% reduction in overall sediment loads in addition to maintaining a minimum depth of
151 2.13 meters below MSL for subtidal habitat in Upper Newport Bay (UNB) (Board, 2014). Newport Bay contains two
152 sediment capture basins (outlined in red, Figure 2), Basin I/III at the mouth of SDC and Basin II, further downstream
153 of SDC, constructed in 2010 at a cost of \$37 million dollars as part of the sediment TMDL with estimated lifespans
154 of 20 years (USACOE, 2011). Hydrodynamic calibration was reported by Guo et al. (2018) and additional calibration
155 and evaluation for sedimentation appears in the Section 3.5.

156 Sedimentation (and the corresponding elevation) of the UNB sediment capture basins is driven by both natural
157 processes and human influences. Natural processes include fluvial input and transport by estuarine currents, which are
158 primarily affected by streamflow and tidal changes in ocean water levels at the mouth of the estuary. Human influences
159 include land uses that affect runoff and the upstream sources of sediment, flood control infrastructure which affects the
160 rate and intensity of streamflow and sediment loads, and dredging which affects tidal circulation. Moreover, human
161 influences are shaped by watershed and estuarine management policies including the sediment TDML and dredging
162 policy. Presently, dredging is required when the sediment capture basins fill to an elevation of 2.13 m below MSL.
163 When that “trigger point” is reached, sediment is dredged to the original basin elevation of 6.65 m below MSL and
164 disposed of offshore, removing them from the system permanently (Board, 2014; USACOE, 2011). Hence, $(\bar{z}_b)_{\text{trig}} =$
165 -2.13 m and $(\bar{z}_b)_0 = -6.65$ m.

166 3.2. Hydrologic, Oceanographic, and Bathymetric Data

167 Stream flow and sediment loading data were acquired for SDC at Campus Drive, a short distance upstream of the
168 connection to Newport Bay. This included a 25-year record of daily peak flow (Figure S10) and a concurrent five-
169 year year record of five-minute interval instantaneous flow data (Figure S1). Sediment transport curves as a function
170 of flow for the same time period were provided by OC Public Works Sediment TMDL Reports ((County of Orange,



Figure 2: Newport Bay is an urban estuary of southern California where sediment is managed for water quality, wetland habitat, recreation, navigation, and the provision of urban amenities. Dredging focuses on sediment capture basins (Basin II and Basin I/III) and a TMDL was implemented to regulate watershed loads from San Diego Creek (highlighted in blue).

171 2016), Figure S2). Oceanic water level measurements for the period of interest are available from the Los Angeles
 172 tide gage (NOAA Gage 9410660). Bathymetric data of UNB for the years 2011, 2012, 2013 and 2015 at 2.1 meter
 173 horizontal resolution (vertical errors not reported) were available as a result of TMDL compliance monitoring by the
 174 Army Corps of Engineers, and dredging excavation depths were available for 2010 (USACOE, 2011). A 2014 survey
 175 of Lower Newport Bay (LNB) bathymetry at 7.6 meter horizontal resolution (vertical accuracy of 15 cm) was used for
 176 LNB (USACE, personal communication), and the 2013 NOAA Coastal Topobathy was used to model land elevation
 177 and offshore bathymetry (horizontal resolution of 1 meter, land vertical accuracy of 4.8 cm RSME, offshore bathymetry
 178 vertical accuracy of 15 cm RSME) (Dewberry, 2013).

Table 1

Model parameters, units, ranges used in sensitivity analysis, and final calibrated value.

Model Parameter	Units	Range	Final Calibrated Value
Chezy Bottom Roughness Coefficient	$\frac{m^{1/2}}{s}$	0 - 1,000	65
Sand Dry Bed Density	$\frac{kg}{m^3}$	1,300 - 2,100	1,600
Mud Dry Bed Density	$\frac{kg}{m^3}$	300 - 500	350
Sand D_{50}	m	Default Value	1×10^{-4}
Critical Stress for Erosion	$\frac{N}{m^2}$	0.16 - 0.75	0.16
Critical Stress for Sedimentation	$\frac{N}{m^2}$	0.1 - 1,000	0.11
Mud Settling Velocity	$\frac{m}{s}$	Default Value	0.00025
Maximum Mud Concentration	$\frac{kg}{m^3}$	10 - 15	12

179 3.3. High-Fidelity Model: Delft3D

180 A two-dimensional (depth-integrated) Delft3D model (version 4.01.01) was constructed with a domain including
 181 the lower part of the San Diego Creek, Newport Bay and a nearshore zone. Delft3D resolves fluid flow, sediment trans-
 182 port, and morphodynamics at fine spatial and temporal scales and has received widespread use for coastal hydromor-
 183 phodynamics (Lesser et al., 2004). Delft3D was configured with an inland inflow boundary where the instantaneous
 184 volumetric flow rate and sediment load is specified, and an open boundary around 8 km offshore where water level is
 185 specified. The model mesh contains approximately 46,519 cells with high resolution in the bay and a minimum cell
 186 size of 5×10 m, with lower resolution offshore and a maximum cell size of roughly 150×350 m. The computational
 187 mesh and flow resistance parameters are based on a hydrodynamic calibration and evaluation described in Guo et al.
 188 (2018). Bed erosion parameters in the Delft3D model were developed from field measurements and computational
 189 experiments (Stein, 2014). Delft3D was calibrated to match morphological change between 2011 and 2012, and then
 190 evaluated for the period 2010-2015. Five-minute interval flow data for 2010 to 2015 was used to specify the freshwater
 191 inflow from SDC (and sediment loading based on sediment transport curves) and six-minute interval tide measure-
 192 ments from Los Angeles were used to specify the water level at the offshore open boundary of the model. Calibration
 193 involved the manual adjustment of several parameters including the sediment curve, erosion/sedimentation thresholds,
 194 and dry density of mud within physically plausible ranges, which are shown in Table 1.

195 A morphological acceleration factor is usually used to achieve decadal to centennial morphodynamic changes
 196 (Roelvink, 2006), but for Newport Bay, spatially distributed bed levels were updated at each hydrodynamic time step
 197 ($= 0.05$ min) which corresponds to a morphological acceleration factor of unity. A sensitivity analysis using the
 198 highest tidal ranges (spring tide in January) and lowest erosion parameters measured in the bay (threshold for erosion
 199 $= 0.16$ N/m²) found that little to no morphological changes occurred throughout the whole estuarine system during dry
 200 weather ($< 1\%$ of yearly morphodynamic change). Based on this analysis, dry weather events river discharge (< 20
 201 cms) were ignored for the calibration and evaluation phases of the Delft3D modeling, and time integration of the Delft

Table 2

Measurements of sediment bed height and 2D and 3D Delft3D model predictions for a one year period, 2010-2011.

Unit I/III Basin Heights			
Year	Measured Height	Modeled Height (2D)	Modeled Height (3D)
2010	-6.654	-6.654	-6.654
2011	-6.294	-6.297	-6.109

Unit II Basin Heights			
Year	Measured Height	Modeled Height (2D)	Modeled Height (3D)
2010	-6.654	-6.654	-6.654
2011	-6.405	-6.333	-6.343

202 3D model was limited to the periods of time when SDC streamflow exceeded 20 cms. The tidal water level from the
 203 LA tide gauge was used for each storm to drive the oceanic boundary condition, with a period between storms driven
 204 by a tidal cycle of 9 to 57 hours. This was done to ensure that the previous storm was not impacting the hydrodynamics
 205 of the proceeding storm.

206 Two sediment fractions, one non-cohesive sand (grain size of 100 μm) and one cohesive mud (settling velocity of
 207 0.25 mm/s), were used to simulate sediment transport, erosion and deposition. The largest source of sediment in this
 208 system is fine material from San Diego Creek. Data collected as part of sediment TMDL found that fines constitute
 209 approximately 56-96% of the total load in San Diego Creek, model calibration found that a 95% mud fraction yielded
 210 the closest agreement between model and data (County of Orange, 2016, 2013). The concentrations of fines are defined
 211 by the sediment rating curves (see Fig. S2) based on 25 years of flow and sediment measurements by OC public works
 212 (County of Orange, 2016). The relatively small sand fraction of the sediment load is modeled under the assumption of
 213 equilibrium concentrations and an unlimited supply of available sediment (Van Rijn et al., 1993; Deltares, 2014).

214 The dry bed densities are 1,600 kg/m³ and 350 kg/m³ for sand and mud, respectively. The Van Rijn et al. (1993)
 215 formula is employed to calculate non-cohesive sand transport in which both bed load transport and suspended sediment
 216 transport are taken into account. Details of the Van Rijn et al. (1993) formulae can be found in the user manual of
 217 Delft3D (Deltares, 2014) and Van Rijn et al. (1993) thus are not repeated here. For cohesive sediment, the Partheniades-
 218 Krone formulation is applied to calculate mud transport (Partheniades, 1965). The critical erosion of the mud used
 219 were set to measured values and are 0.16 N/m² (Stein, 2014). The critical threshold for deposition was calibrated to
 220 0.11 N/m² from 1,000 N/m² (which resulted in a calibration basin height of -6.14 m for the year 2012 compared to
 221 -6.04 m when using 0.11 N/m² and a measured value of -6.047), for the Unit I/III basin. These values suggest that
 222 erosion only happens when the calculated bed shear stress is >0.16 N/m² and deposition occurs when shear stress is
 223 <0.11 N/m² (Deltares, 2014; RMA, 1998).

224 A two-dimensional approach was used over a three-dimensional formulation (using 10- σ layers) based on the results

225 of a 2D vs. 3D comparison for a one year period, 2010-2011. As shown in Table 2, the 2D model performed better
 226 than the 3D model in the Unit I/III basin, with a 3 mm difference in modeled vs measured average basin elevation
 227 compared to the 3D model's 18 cm difference. In the Unit II basin, the 3D model performed slightly better than the
 228 2D model, with only a 1 cm difference between the two. Unit I/III basin is largely well mixed (Trimble, 2003), so the
 229 3D model introduces extra physical processes which increases potential model uncertainty and noise, explaining the
 230 differences between the 2D and 3D model. Other studies have found that a 2D works equivalently well compared to 3D
 231 approach for hydrodynamics /salinity (Sandbach et al., 2018) and sediment transport modeling in estuarine systems
 232 (Achete et al., 2017) Based on the lack of improvement in model accuracy, and very significant gains in computational
 233 speed (roughly 10 times faster), the 2D Delft3D model was chosen to finish model evaluation and develop the surrogate
 234 model.

235 3.4. Response Surface Surrogate Model

The surrogate model can only be built after a numeric model (such as Delft3D) is calibrated and evaluated, at which point numeric model parameters are not changed. Then, the surrogate model can be constructed and evaluated against measured data to ensure surrogate model robustness. To develop the response surface surrogate model, the high fidelity model was configured to simulate 24 hour storm events based on the peak flow into the estuary, Q , a triangular hydrograph shape with a 5 hour and 11 hour time to rise and fall, respectively, a pre-event sediment basin elevation, \bar{z}_b , and a mixed regime tidal boundary (to match the tides at Newport Bay) with a period of 13.3 hours, an amplitude, a , and a phase ϕ . These results showed that the most important variables to consider for estimating the sediment capture efficiency, $\tilde{\eta}$, are the flood peak and the pre-event basin elevation. Furthermore, results showed that the phasing and amplitude of the tide had a secondary effect, with the maximum deposition occurring when the flood peak occurred during the maximum flood currents of a spring tide and the minimum deposition occurring when the flood peak was timed with maximum ebb currents of a spring tide for this system. Capturing all possible combinations of tidal amplitude and phase would introduce two additional independent variables to the response surface surrogate model (in addition to Q and \bar{z}_b), and assuming that m different values of each independent variable would need to be sampled to map out $\tilde{\eta}$, the required number of high-fidelity model runs would increase from m^2 to m^4 , an increase by a factor m^2 . This motivated the formulation of capture efficiency as a weighted average of the minimum and maximum capture efficiency over the tidal cycle, ($\tilde{\eta}_{\min}$ and $\tilde{\eta}_{\max}$, respectively) as follows,

$$\tilde{\eta}(Q, \bar{z}_b, \phi) = [1 - \theta] \times \tilde{\eta}_{\min}(Q, \bar{z}_b) + \theta \times \tilde{\eta}_{\max}(Q, \bar{z}_b) \quad (12)$$

236 where θ is a weighting factor that is treated as a random variable between zero and unity. In southern California,
 237 tidal amplitudes transition between small neap tides to large spring tides with a fortnightly cycle, and also experience

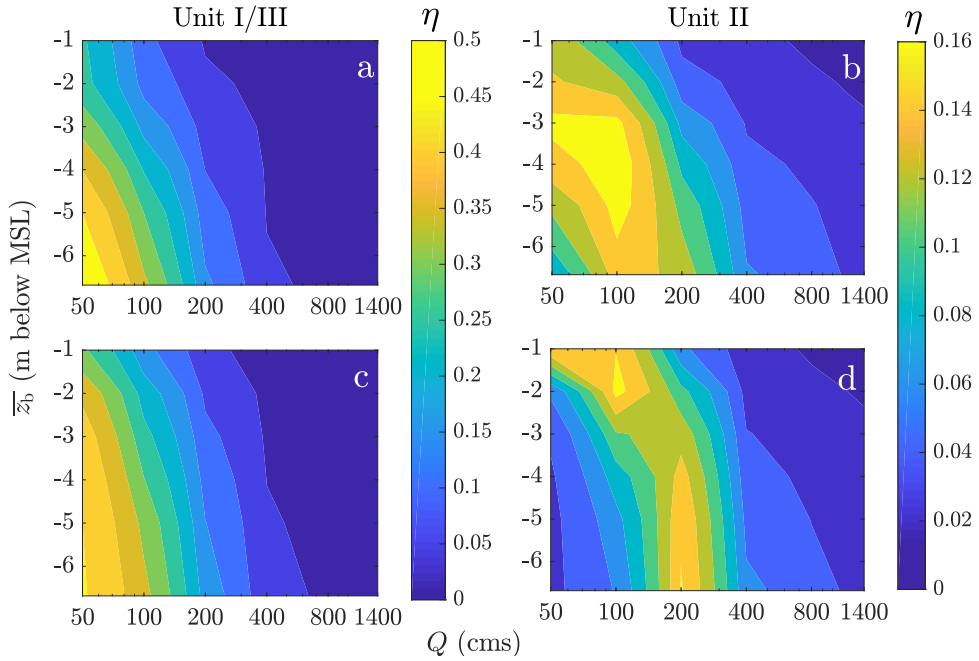


Figure 3: Response surface surrogate models $\tilde{\eta}_{ebb}$ (a,b) and $\tilde{\eta}_{flood}$ (c,d) for Basin I/III (a,c) and Basin II (b,d) of Newport Bay. Note that the x-axis is logarithmic and only shows storms from 50 cms to 1,400 cms for visual clarity.

238 semi-diurnal and diurnal variability. Where a flood peak falls within this cycle is, in fact, random. Hence, treating
 239 θ as a random variable in Eq. 12 with a uniform distribution between zero and unity leads to a surrogate model for
 240 sediment capture efficiency that accounts for both deterministic (Q and \bar{z}_b) and random (tidal conditions) aspects of
 241 the system. In the case of Newport Bay, the maximum capture efficiency was found to occur when the flood peak
 242 was coincident with the peak flood current ($\tilde{\eta}_{max} = \tilde{\eta}_{flood}$), and the minimum capture efficiency was found to occur
 243 when the flood peak was coincident with the maximum ebb current ($\tilde{\eta}_{min} = \tilde{\eta}_{ebb}$). However, we anticipate that this
 244 could vary from site to site which reiterates the site-specific nature of a response-surface surrogate model. Future
 245 studies could investigate a diversity of sites and develop potentially develop generalizations of basin response surface
 246 surrogate model sensitivities to tides, flows, sediment supply and basin elevations.

247 Fig. 3 presents the response surfaces for $\tilde{\eta}_{ebb}$ and $\tilde{\eta}_{flood}$, respectively, for both the Unit I/III and Unit II basins. Basin
 248 I/III has an average η of 12.3% and 13.8% and Unit II with an η of 10.1% and 11.1% for peak ebb and flood currents,
 249 respectively. Additionally, these surfaces show the nonlinear response of capture efficiency to Q and \bar{z}_b characterized
 250 by local maxima and minima. Importantly, mapping out this dependence using the high fidelity model positions the
 251 reduced-dimension surrogate model to efficiently make multi-decadal simulations with a high level of computational
 252 efficiency in two separate basins within the same system.

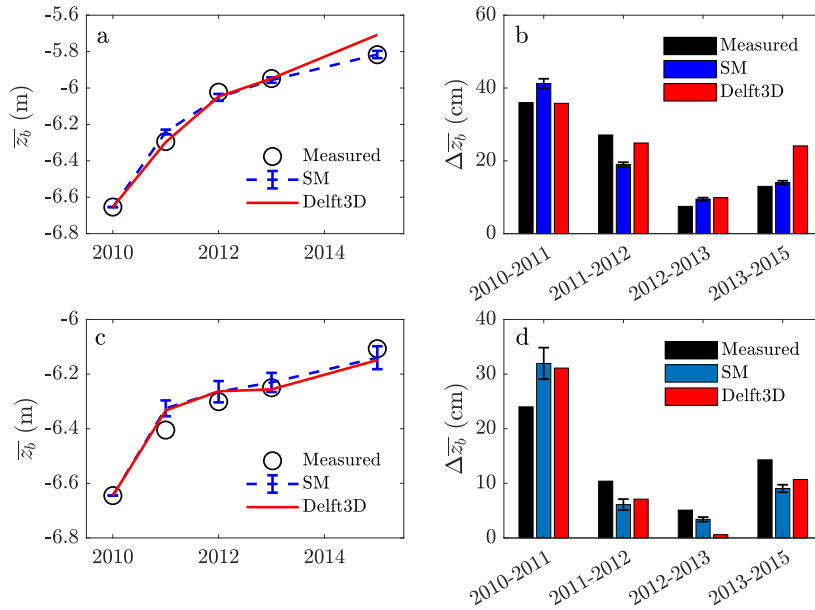


Figure 4: High-fidelity model (Delft3D) and SeAMLESS output (SM) compared to measured quantities of average basin elevation (\bar{z}_b , panels a & c) and yearly change in average basin elevation ($\Delta\bar{z}_b$, panels b & d) for the Unit I/III (panels a & b) and Unit II (panels c & d) basins. The error bars on the surrogate model account for model uncertainty due to tidal influence (θ), which is treated as a random variable.

253 3.5. SeAMLESS Evaluation

254 The high-fidelity model and SeAMLESS were both applied to simulate five years of sedimentation (30 storms
 255 exceeding baseflow threshold during 2010-2015) in Newport Bay which corresponds to a duration when annual bathy-
 256 metric monitoring data are available for validation purposes and gauge measurements are available to specify model
 257 inputs including streamflow, sediment loads and tides. Figure 4 presents comparisons of the high-fidelity and reduced-
 258 dimension surrogate model predictions of \bar{z}_b in Basin I/III and Basin II, which compare favorably to measured changes.

259 The models were quantitatively compared based on the Root Mean Square Error (RSME) in basin elevation.
 260 SeAMLESS performs better than the high-fidelity model for the Unit I/III basin ($RSME_{Delft3D} = 0.00253$ m vs.
 261 $RSME_{SM} = 0.00072$ m) and slightly worse for the Unit II basin ($RSME_{Delft3D} = 0.00166$ m vs. $RSME_{SM} = 0.00180$
 262 m). This is likely due to stronger tidal currents in the Unit II basin which are also reflected by the larger error bars
 263 appearing in Figure 4c-d compared to Figure 4a-b. Furthermore, the SeAMLESS model uncertainty due to tidal influ-
 264 ence bounds the high-fidelity model simulations for most years—the exception being years 2011 and 2015 in Unit I/III.
 265 The maximum yearly error (difference in deposition) was 0.081 m and 0.11 m for the surrogate and Delft3D models,
 266 respectively in the Unit I/III basin.

267 Relative runtimes of SeAMLESS and the high-fidelity model were examined with an 85 year simulation of Newport
 268 Bay. SeAMLESS was run using a single core on a 3.0 Ghz processor, while the high-fidelity model was run on a high-

Table 3

High-fidelity model runtimes and SeAMLESS speedup based on a runtime of 42 seconds.

System	Model Runtime (hrs)	SeAMLESS Speedup
1-core	6,033	5.16×10^5
4-core	3,448	2.95×10^5
8-core	1,675	1.43×10^5
16-core	1,370	1.17×10^5
32-core	944	8.07×10^4
64-core	359	3.07×10^4

269 performance computing cluster (64-core compute nodes with 2.33 GHz AMD processors and 512 GB of RAM) using
 270 1 to 64 cores. We note that SeAMLESS integrates over flood events, and the high fidelity model integrates over periods
 271 with significant streamflow (not dry weather periods) although a short period (<24 hours) is modeled between wet-
 272 weather events for to allow the hydrodynamics to reset to a tidally forced state. SeAMLESS computes this simulation
 273 in 42 seconds, while the high-fidelity model run time varies from hundreds to thousands of hours depending on the
 274 number of cores. Table 3 presents the high-fidelity model runtimes and SeAMLESS speedup, defined by the ratio of
 275 the high-fidelity model runtime to the SeAMLESS runtime (for a single simulation). This shows that individual, multi-
 276 decadal simulations can be completed $\mathcal{O}(10^4 - 10^5)$ times faster using SeAMLESS without any obvious loss of accuracy
 277 in terms of basin-averaged decision variables of interest to stakeholders. Moreover, further gains in computational
 278 efficiency are possible through parallel execution of SeAMLESS simulations. These results show that SeAMLESS
 279 is ideally configured to support Monte Carlo Markov Chain simulations of sedimentation and dredging that account
 280 for a range of possible storm peak sequences, sea level rise trajectories, and dredging rules-modeling that would be
 281 prohibitively expensive using the high-fidelity model.

282 3.6. Multi-decadal SeAMLESS Simulation

283 Stochasticity is introduced to SeAMLESS for multi-decadal simulations with daily SDC peak flow (Q) based
 284 on a Monte-Carlo Markov Chain (MCMC) random sampling (see supplemental text for details on MCMC sampler,
 285 including historical data used to develop the sampler as shown in Fig. S3) and treating variability in tidal conditions
 286 (θ) as a uniformly distributed random variable (Eq. 12, Fig. 3). Sediment loads for each day are based on sediment
 287 rating curves (Fig. S4), and mean sea level is adjusted based on SLR projections (Fig. S5). The surrogate model was
 288 developed to span the range of possible inflows into Newport Bay ($Q = 10$ cms to 1,400 cms), basin elevations ($z_b = 1$
 289 m depth to 6.65 meters depth), and tidal condition θ (peak ebb or peak flood). Note that Fig. 3 only shows the response
 290 surface for $Q > 50$ cms for visual clarity. Storms below 50 cms yield complex surfaces due to a higher influence of
 291 tidal currents and contribute little to overall basin elevation changes.

292 The model accounts for dredging policy by testing for the exceedance of the trigger point (2.13 meters below MSL)

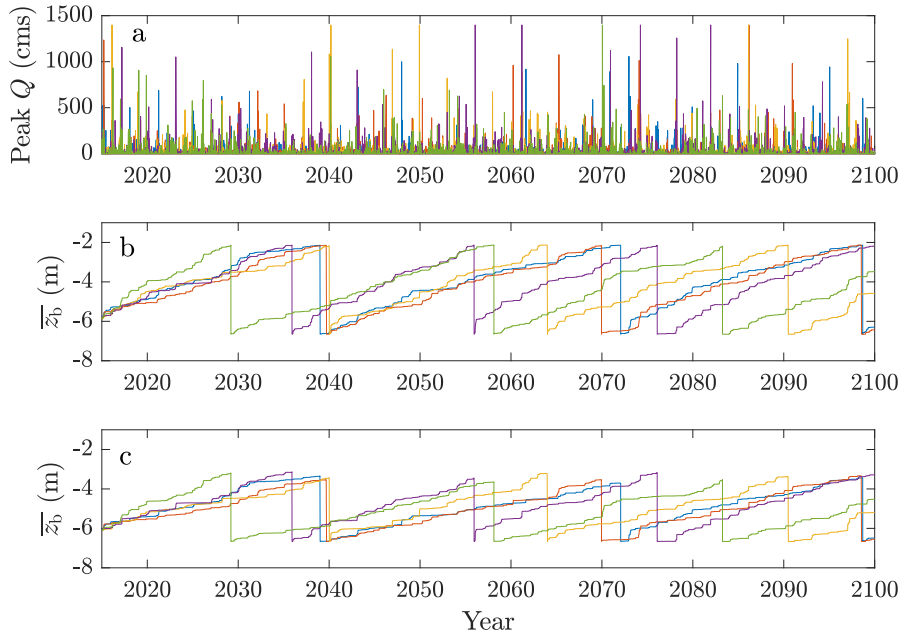


Figure 5: Five examples (color-coded) of multi-decadal MCMC simulations of SDC flow (panel a), Basin I/III elevation (panel b) and Basin II elevation (panel c) using the surrogate model. Dredging events are marked by vertical drop in basin elevation (\bar{z}_b), and differences in timing are linked to the occurrence of large storm events (Q). A total of one thousand MCMC simulations are completed to yield probability distributions in expected sedimentation and dredging over future years.

293 on a daily timescale and resetting the elevation in accordance with the post-dredging elevation. Moreover, various
 294 scenarios involving changes in natural forcing and human influences are considered to directly simulate dredging, the
 295 decision variable of interest (**D**) as in Figure 1. As an example, Figure 5 shows five MCMC simulations of long-term
 296 sediment basin elevation changes for the Unit I/III (b) and Unit II (c) basins given an input of future storm events (a)
 297 generated from the MCMC sampler.

298 Note that sediment basin elevations (\bar{z}_b) increase over time, increase in proportion to the occurrence and magnitude
 299 of storm events (Q), and are lowered with the occurrence of dredging. Also note significant variability in the timing
 300 of dredging which results mainly from the random occurrence of large storm events. Monte Carlo Markov Chain
 301 (MCMC) simulations of daily peak Q were used to drive SeAMLESS were run 1,000 times to capture the future
 302 uncertainty in dredging scenarios. MCMC transition probabilities were kept constant for the purposes this study, as
 303 incorporating climate change is beyond the scope of this work. It is important to note however, that the speed of the
 304 surrogate model allows for quick model re-runs and would allow modelers and stakeholders to easily evaluate impacts
 305 of changing hydro-climatology on dredging, given a hydrologic model of the watershed.

306 The timing and occurrence of future dredging events is of interest to stakeholders due to the considerable expense in
 307 dredging (roughly \$37 million 2010 dollars per cycle). Fig. 6 demonstrates a potential output of SeAMLESS showing

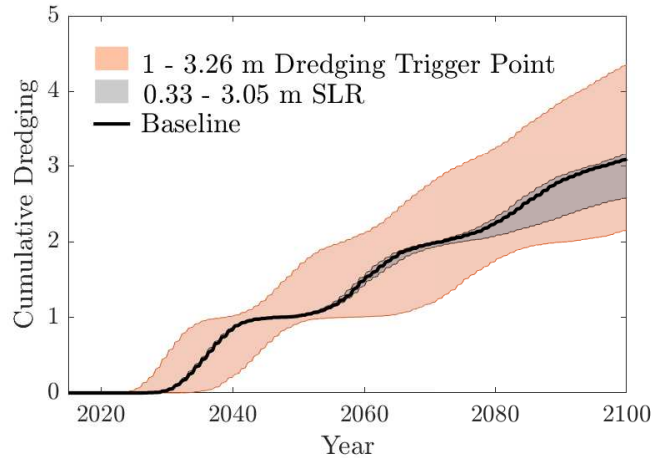


Figure 6: SeAMLESS estimates of the expected number of dredging events through 2100. The baseline scenario (black) corresponds to existing dredging rules and SLR based on the median RCP 4.5 projection (0.61 m by 2100). Colored areas show the range in dredging events based on the range in the dredging trigger point (red) or range in SLR from RCP 2.6 (0.33 m by 2100) and H++ scenario (3.05 m by 2100).

308 the cumulative likelihood of number of dredging events through 2100 for various scenarios. This was calculated by
 309 taking the mean of the MCMC simulations for the cumulative dredging required to maintain minimum basin depth
 310 through 2100.

311 The solid black line shows the cumulative number of required dredging cycles through 2100 under current dredging
 312 requirements ($(\bar{z}_b)_{\text{trig}} = -2.13$ MSL) and SLR based on the median projection RCP 4.5 scenario (0.61 m by 2100)
 313 provided by Griggs et al. (2017). Additionally, the range of dredging cycles corresponding to a range in SLR (through
 314 2100) and a range in the dredging trigger point are also shown. The low and high levels of SLR are based on RCP
 315 2.6 (0.33 m by 2100) and the H++ scenario (3.05 m by 2100) reported for Southern California by Griggs et al. (2017).
 316 High and low value of $(\bar{z}_b)_{\text{trig}}$ corresponding to -1.0 m and -3.26 m MSL, or roughly a 1 m change.

317 Fig. 6 shows that the number of dredging events is not very sensitive to SLR until after 2050, when the differences
 318 in SLR become significant. In this limit, the higher SLR scenario (H++) requires less dredging (mean 2.58 total
 319 dredging cycles) compared to the minimum SLR scenario (RCP 2.6, mean 3.17 dredging cycles). This trend is a result
 320 of SLR adding to basin depth, deepening the basin with respect to the original 2.13 meters below 2015 MSL. Focusing
 321 now on the dredging trigger point, what becomes clear is that the effect on the number of dredging cycles is evident
 322 within the first two decades of the simulation (before 2040). Additionally, the effect on the total number of dredging
 323 cycles is substantial. Maintaining a deeper basin elevation (-3.26 m MSL “trigger point”) mandates more than two
 324 additional dredging cycles through 2100 based on 4.36 total dredging events for a -3.26 m MSL trigger point and 2.15
 325 dredging events for a -1.0 m MSL trigger point.

326 The projections shown in Fig. 6 are responsive to the decision-making needs of coastal sediment management

327 stakeholders: probabilities for how much dredging will need to be done, when it needs to be done, and how trigger
328 points (rules) can be adjusted to manage future dredging. As part of the the SedRISE project, an earlier version of
329 this result was presented at a meeting with stakeholders and received favorably. In particular, it was reported that
330 dredging events need to be planned (and budgeted) well over a decade in advance due to challenges with permitting
331 (e.g., Ulibarri et al., 2020), so a multi-decadal forecast of dredging was highly valued. In addition, stakeholders were
332 very interested on how SLR would impact future dredging events. Simple bathtub approaches for investigating the
333 impacts of SLR on coastlines are simple to use and quick to run, but inadequately characterize the dynamic effects
334 of SLR (Passeri et al., 2015). Using SeAMLESS for coastal management is beyond the scope of this paper, but it's
335 important to emphasize that SeAMLESS makes it possible to quickly simulate how a particular management action (in
336 this case dredging) will be affected by environmental factors like SLR and policy factors like dredging trigger points,
337 which is valuable for dialogue and deliberation by stakeholders (DeLorme et al., 2016; Sanders et al., 2020; Stephens
338 et al., 2020).

339 **4. SeAMLESS Application to a Stylized Estuary**

340 A stylized estuary fed by a river and open to the ocean is used to further validate the efficacy of SeAMLESS
341 for predicting average sediment basin elevation over multi-decadal time scales in systems where sedimentation is
342 controlled by loading during storm events, as in southern California. Moreover, we seek to gain a better understanding
343 of SeAMLESS limitations with a second stylized application.

344 The geometry of the stylized system is presented in Figure 7, and is characterized by a 667 meter wide river from
345 the North at 16 meter depth with respect to mean sea level (MSL). The river opens to a 14×38 km long estuary, with
346 an average depth of 3 meters, and a maintained channel depth of 16 meters. The estuary has a 2×2.5 km sedimentation
347 basin where the river meets the estuary, which is designed to trap riverine material before it enters the lower portions
348 of the estuary, as in Newport Bay. The estuary mouth is 6 km wide with the main channel maintained at 16 meter
349 depth to the open ocean. A high-fidelity Delft3D model is created of this system using a total of 8,993 grid cells, with
350 the highest resolution where the river meets the estuary (100×100 meter spacing) and the coarsest spacing near the
351 oceanic boundaries ($1,000 \times 1,000$ meter spacing).

352 **4.1. High-Fidelity Model**

353 The stylized Delft3D model is forced by streamflow at the northern boundary and a tidal time series at the southern
354 boundary, and all other boundaries were treated as walls. Streamflow was simulated with a 24 hour flood hydrograph,
355 and 12 hour time to rise and fall. The flood peak was randomly generated using Monte-Carlo sampling of peak flows
356 (from 500 - 20,000 cms) which were transformed into daily triangular hydrographs, and sediment load was computed

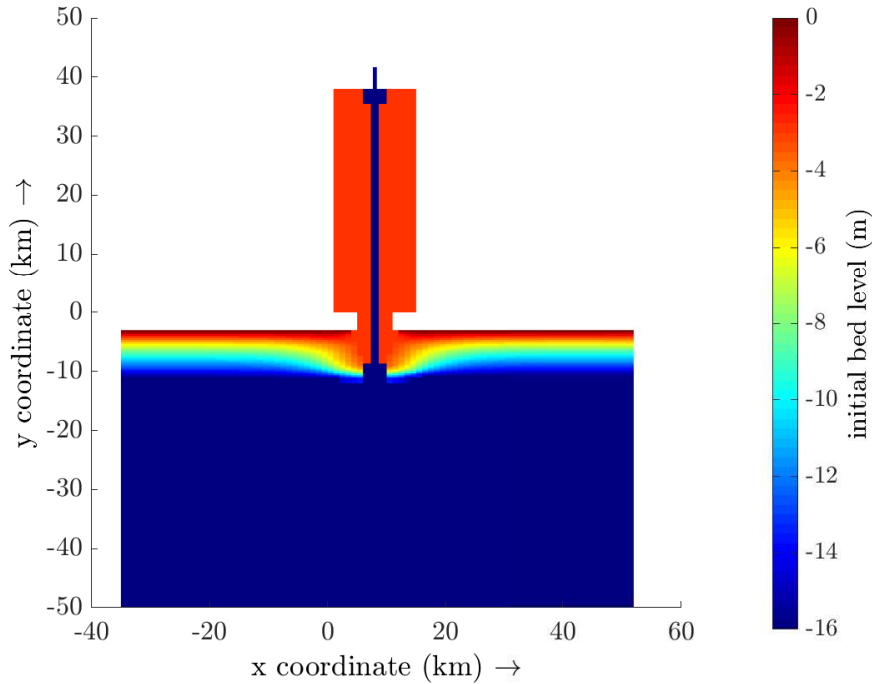


Figure 7: Stylized model domain and bathymetry.

357 with an idealized sediment curve. See Fig. S7 for the peak-flow cumulative distribution function and Fig. S8 for
 358 sediment-concentration curve used in the high-fidelity stylized model. A period with 500 cms baseflow in the high-
 359 fidelity model river lasting a randomly chosen duration between 12-24 hours (uniform distribution) was added between
 360 storms to re-establish tidal control of the hydrodynamics and to avoid storm peaks being phase-locked with the tide.
 361 Tides were modeled by a 1 meter amplitude tide (with respect to MSL), a period of 12 hours, and phase of zero. Rough-
 362 ness was modeled using the Chezy roughness formula ($C=65 \text{ m}^{1/2}/\text{s}$) with a free slip condition for wall roughness.
 363 Two sediment fractions were used, one for sand (dry bed density: $\rho_{\text{sand}} = 1600 \text{ kg/m}^3$) and mud (dry bed density:
 364 $\rho_{\text{mud}} = 500 \text{ kg/m}^3$, critical bed shear stress for sedimentation and erosion was 1,000 and 0.5 N/m^2 , respectively. The
 365 initial bed layer was a 5 meter thick bed of sand. The equilibrium condition for sand transport (Van Rijn and Walstra,
 366 2003) was used at the inflow boundary. A morphological scale factor of 5 was used to speed computation with no loss
 367 in accuracy (Lesser et al., 2004; Ranasinghe et al., 2011).

368 **4.2. Response Surface Surrogate Model**

369 A response surface surrogate model of the form given by Eq. 12 was developed by running the high-fidelity model
 370 for a range of Q and \bar{z}_b values during peak flood and ebb currents and maximum and minimum tides. Peak streamflow
 371 values were 500, 1,000, 2,000, 4,000, 6,000, 8,000, 10,000, 15,000, and 20,000 cms. Average sediment basin elevations

372 ranged from a low of 3 meters, up to 16 meters, spaced at intervals between 1 meter (from 3 to 4 meters) to 2 meters of
 373 elevation change (from 4 to 16 meters). Sediment yield for the surrogate model was computed for a given peak flow (Q)
 374 by integrating the triangular flow hydrograph multiplied by the instantaneous sediment concentration which yielded
 375 the regression equation for sediment load shown in Fig. S9. The surrogate model accounted for the morphological
 376 scaling factor in the high-fidelity model simulations by multiplying all incoming sediment volumes by the equivalent
 377 scaling factor (MF=5). Two variants of the surrogate model were developed based on model simulations which found
 378 that $\tilde{\eta}_{\max}$ and $\tilde{\eta}_{\min}$ were coincident with peak flood ($\tilde{\eta}_{\text{flood}}(Q, \bar{z}_b)$) and ebb ($\tilde{\eta}_{\text{ebb}}(Q, \bar{z}_b)$) currents as in Newport Bay.
 379 However, the authors note that this is site specific, and that high-fidelity model simulations of the system are required
 380 to investigate the major controls on $\tilde{\eta}_{\max}$ and $\tilde{\eta}_{\min}$ for each site.

381 4.3. Multi-Decadal Simulations

382 Multi-decadal simulation of sedimentation in the stylized estuary was completed by configuring both the high
 383 fidelity model and SeAMLESS to depict a sequence of $N=200$ storm events. In southern California, there are roughly
 384 ~ 5 -10 storm events which occur in any given year and thus 200 storms roughly equates to 20-40 years. Integration
 385 over events is straightforward using SeAMLESS (Eq. 11). A sequence of 200 consecutive Monte Carlo samples of
 386 peak flow created time series of storm events, Q_i , $i = 1, \dots, N$ which were input into both SeAMLESS and the high
 387 fidelity model. These storm peaks were simulated until $Q_i = 200$, or until the basin infilled to 3 meters below mean
 388 sea-level, whichever came first. This was repeated for 100 different storm peak sequences (each containing a string of
 389 200 peaks) to sample an exhaustive range of possible storm peak sequences and average out tidal effects. Account of
 390 tides between the high fidelity model and SeAMLESS was slightly different in accordance with the functionality of
 391 each model: whereas the high-fidelity model uses a random variable to vary the phasing between the storm peak and
 392 the tide peak between successive events, SeAMLESS uses a random variable (θ) to compute a weighted average of
 393 $\tilde{\eta}_{\max}$ and $\tilde{\eta}_{\min}$. Moreover, for each storm peak sequence, SeAMLESS was repeated 2,000 times using θ as a random
 394 variable. Hence, the final sequence of $(\bar{z}_b)_i^{\text{SM}}$ for each storm sequence is computed as the average over 2,000 trials and
 395 can be considered to be a tidally-averaged solution at the event time scale.

Solutions for each of the 100 different storm sequences were averaged to compute a time series of $(\bar{z}_b)_i$ represen-
 tative of an average over many different possible storm sequences. Hence, the accuracy of SeAMLESS was measured
 by the mean error (ME) in \bar{z}_b (average over 100 storm sequences) relative to the high-fidelity model (averaged over
 100 storm sequences

$$\text{ME} = \frac{1}{N} \sum_{i=1}^N (\bar{z}_b)_i^{\text{Delft3D}} - (\bar{z}_b)_i^{\text{SM}} \quad (13)$$

396 where N represents the number of events in the storm sequence (when the basin reaches 3 meters below mean sea-level,

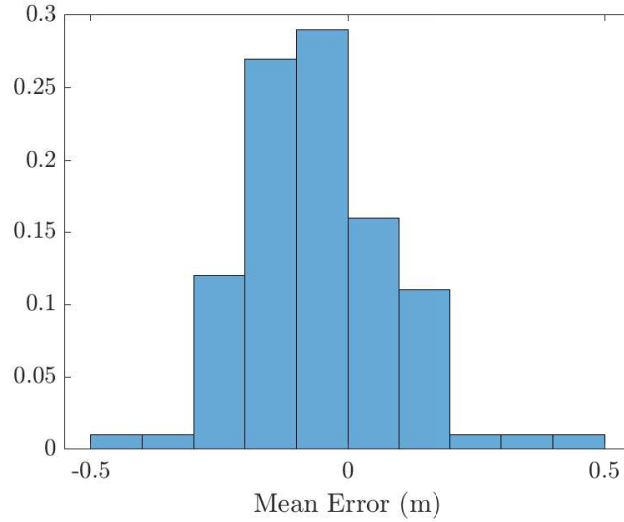


Figure 8: Frequency distribution of Mean Error in \bar{z}_b (Eq. 13) from 100 stochastic storm sequences simulated by SeAMLESS.

397 or $Q_i = 200$ storm events.

398 The distribution of errors from 100 different storm sequences is shown in Fig. 8. This distribution passes the K-S
 399 test for normality ($p=0.47$) with a mean of -0.0607 m and standard deviation (σ_{ME}) of 0.1476 m. Conceptually, the
 400 results in Fig. 8 indicate that the surrogate model (SM) slightly overestimates average basin elevation by roughly 6
 401 cm on a per-storm basis when compared to the high-fidelity model. This equates to only 0.46% when normalized by
 402 overall basin infilling (13 m). Additional insight is obtained by examining the time series predicted by SeAMLESS and
 403 the high-fidelity model for specific storm sequences that generate different levels of error, as measured by Eq. 13. Fig.
 404 9 shows time series of \bar{z}_b predicted by SeAMLESS and the high-fidelity model based on ME at a range of quantiles:
 405 (a) -95% (b), -5% (c), +5%, and (d) +95%. While the surrogate model tends to underestimate deposition (ME =
 406 -0.0607 m), overall, the surrogate model shows nearly equivalent performance compared to Delft3D as the basin is
 407 filled. Additionally, the surrogate model performs with roughly an order of magnitude lower ME for the first 100 storm
 408 sequences (ME = -0.0061 , $\sigma_{ME} = 0.0922$), indicating that SeAMLESS is especially adept at predicting basin elevation
 409 for roughly the first 10-20 years of basin infilling.

410 We attribute the largest source of error in SeAMLESS simulation to arise from the assumption that sediment bed
 411 elevation is uniform across the sedimentation basin, which was used to build the response surface function. Defining
 412 T as the time required for the \bar{z}_b to rise up to the top of the sedimentation basin (3 m), Fig. 10 shows contours of bed
 413 elevation predicted for $t=0$, $0.4T$, $0.8T$, and T . This shows that the sediment basin fills in a non-uniform manner, a
 414 feature that is especially evident for $t \geq 0.8T$.

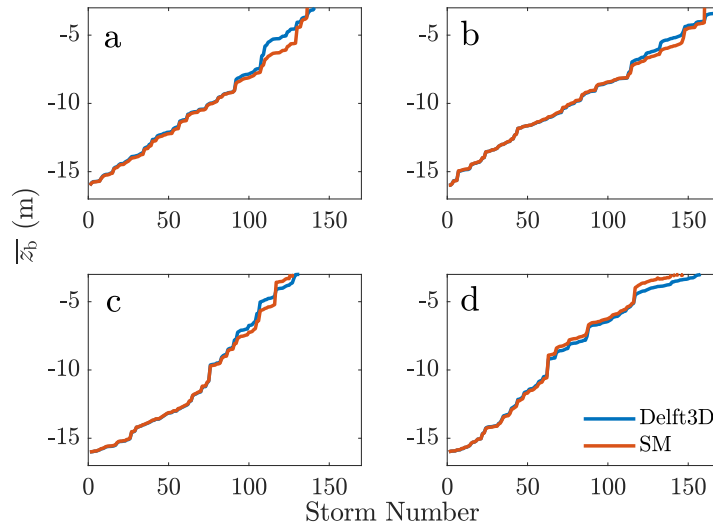


Figure 9: Comparison of high-fidelity model (Delft3D) and SeAMLESS (SM) predictions of \bar{z}_b over 200 sequential storm events for the following quantiles in Mean Error: (a) -95%, (b) -5%, (c) +5%, and (d) +95%.

415 5. Conclusions

416 A reduced-dimension surrogate model for estuarine sedimentation within a managed sediment basin, SeAMLESS,
 417 is formulated herein and shown to support multi-decadal simulation with uncertainty and yield decision-variables
 418 useful for management in Southern California. Useful output includes the number of expected dredging events in
 419 the future, the timing of events, and volumes of sediment associated with these events under different sea level rise
 420 scenarios and rules (or “trigger points”) for dredging.

421 SeAMLESS is shown to yield accuracies comparable to a high-fidelity model, while the wall clock run time is
 422 reduced by a factor of $\mathcal{O}(10^4)$ even after the high-fidelity model is executed in parallel on a high performance computing
 423 cluster with up to 64 cores. The accuracy of SeAMLESS stems from using a calibrated and validated high-fidelity
 424 model to develop and parameterize a response surface surrogate model for the capture efficiency of a sedimentation
 425 basin, $\tilde{\eta}$, defined as the ratio of sedimentation volume to storm load of sediment over the time scale of a storm event.
 426 In essence, SeAMLESS relies on a physics-based model to create a set of data that, in turn, is easily accessed to map
 427 out the response of sediment basins to future storm events that bring sediment into the estuary. The computational
 428 efficiency of the reduced dimension surrogate modeling approach is shown to support Monte Carlo simulation of
 429 future scenarios which provides information about the uncertainty in outcomes.

430 Sea level rise is expected to reduce the need for dredging in many coastal systems, since high sea levels increase
 431 depth. Based on SeAMLESS modeling of Newport Bay, expected dredging through 2100 for the lowest estimate of
 432 RCP 2.6 and highest estimate (H++ scenario) sea level rise trajectories reported by (Griggs et al., 2017) corresponded

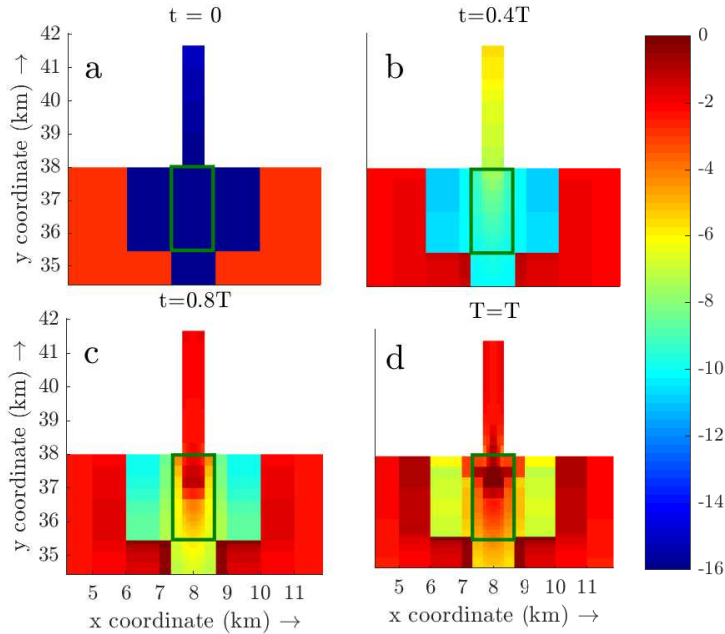


Figure 10: Pattern of sedimentation within the sedimentation basin at times (a) $t=0$, (b) $0.4T$, (c) $0.8T$, and (d) T , where T represents the instant the sedimentation basin reaches its capacity and the green rectangle represents the sediment basin definition used to develop the surrogate model.

433 to 3.2 and 2.6 events, respectively. However, SeAMLESS also shows that moving the trigger height for dredging from
 434 2.13 m (MSL) down to 1 m (MSL) or up to 3.26 m (MSL) could change the expected number of dredging events to 2.1
 435 and 4.4, respectively. This result shows that rules for dredging exert a strong control over the timing and number of
 436 dredging events that are needed to manage critical coastal systems like Newport Bay under sea level rise, and reinforce
 437 the utility of the SeAMLESS framework as a tool for coastal stakeholder groups to develop rules for dredging under
 438 sea level rise.

439 SeAMLESS was formulated based on coastal embayments in Southern California, where tidal basins are relatively
 440 small and experience episodic runoff and sediment loads lasting less than a day, and where dredging is needed to
 441 support recreation, navigation, water equality, and ecosystems needs. While this general approach may be appropriate
 442 at sites elsewhere, it is important to acknowledge that response surface surrogate models need to be custom built
 443 for each site utilizing a high-fidelity numeric model. Hence, a high-fidelity model is important for designing the
 444 response surface function (i.e., choice of independent variables) and for parameterizing the response surface function.
 445 Based on the results of this paper, the authors conclude that a surrogate model works when deposition is 1) spatially
 446 homogenous, 2) the surrogate model area is greater than the numerical model cell size, and 3) confined to a well-defined
 447 basin / region. Further research would benefit surrogate model research by investigating developing surrogate models
 448 for spatially heterogeneous regions (such as wetlands) using statistical methods to describe the spatial structure of

449 deposition. Additional research incorporating ecological and flood impacts of changing dredging trigger points would
450 improve the utility of the model.

451 Consideration of a stylized estuary test case also showed that non-uniform filling of the sedimentation basin may
452 emerge as a source of error in the sediment capture efficiency used by reduced-dimension surrogate model. Neverthe-
453 less, the accuracy of SeAMLESS was found to be comparable to the high-fidelity model for timescales up to several
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465 **CRedit authorship contribution statement**

466 **Matthew W. Brand:** Conceptualization of this study, methodology, software, manuscript. **Leicheng Guo:** Method-
 467 ology, software. **Eric D. Stein:** Revisions to manuscript, overall study guidance. **Brett F. Sanders:** Revisions to
 468 manuscript, overall study guidance.

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- 572 MW Brand completed his Ph.D. at the University of California, Irvine.

Highlights

Multi-decadal simulation of estuarine sedimentation under sea level rise with a response-surface surrogate model

Matthew W. Brand, Leicheng Guo, Eric D. Stein, Brett F. Sanders

- Estuarine sedimentation is a complex physical process
- High-fidelity model runtimes hinder multi-decadal simulation
- A response surface surrogate model estimated multi-decadal basin depths
- Surrogate model was orders of magnitude faster compared to high-fidelity model
- Surrogate model was able to attain equivalent accuracy to high fidelity model