

1 at the
2 **Mt. Bachelor Observatory (MBO) during 2004–2015**

3 **Lei Zhang^{a,*} and Daniel A. Jaffe^{a,b}**

4 ^a School of Science, Technology, Engineering and Mathematics, University of
5 Washington Bothell, Bothell, WA 98011, United States

6 ^b Department of Atmospheric Sciences, University of Washington, Seattle, WA 98195,
7 United States

8 * Corresponding author.

9 *Email address:* lzhang12@uw.edu.

10 **Abstract**

11 In this paper, we report the climatology of tropospheric ozone (O₃) and sub-micron
12 aerosol scattering at the Mt. Bachelor Observatory (MBO, 2.8 km asl) in central
13 Oregon, USA, during 2004–2015. The seasonal cycle for O₃ showed a bimodal
14 pattern with peaks in April and July, while aerosol scattering (σ_{sp}) was lognormally
15 distributed with a very high peak in August and a smaller peak in May. The mean O₃
16 concentrations showed positive and significant trends in all seasons except winter,
17 with a slope of 0.6–0.8 ppbv yr⁻¹. Monthly criteria for isolating free tropospheric (FT)
18 and boundary layer influenced (BLI) air masses at MBO were obtained based on
19 comparison of MBO water vapor (WV) distributions to those of Salem (SLE) and
20 Medford (MFR), Oregon, at equivalent pressure level. In all seasons, FT O₃ was, on
21 average, higher than BLI O₃, but the seasonal patterns were rather similar. For σ_{sp} the
22 FT mean in spring was higher, but the BLI mean in summer was significantly higher,
23 indicating the importance of regional wildfire smoke.

24 To better understand the causes for the seasonal and interannual trends at MBO, we
25 identified four major categories of air masses that impact O₃, carbon monoxide (CO)
26 and aerosols: upper troposphere and lower stratosphere (UTLS) O₃ intrusion, Asian
27 long-range transport (ALRT), Arctic air pollution (AAP) and plumes from the Pacific
28 Northwest region (PNW). ALRT and PNW plumes can be further divided into
29 wildfires (WF), industrial pollution (IP) and mineral dust (MD). Over the 12 years of

30

31 (ERs) and Ångström exponents (AEs) of aerosols were calculated for all events. The
32 lowest slope of $\Delta\sigma_{sp}/\Delta O_3$ is a unique feature of UTLS events. PNW-WF events have
33 the highest averages for $\Delta\sigma_{sp}/\Delta CO$, $\Delta\sigma_{sp}/\Delta O_3$ and $\Delta\sigma_{sp}/\Delta NO_y$ compared to other events.
34 These ERs decrease during long-range transport due to the shorter residence time of
35 aerosols compared to the other pollutants. ALRT-WF events have lower absorption
36 AEs (\mathring{A}_{ap}) than PNW-WF, implying that brown carbon (BrC) is generated from
37 biomass burning but its fraction decreases during long-range transport. Signatures of
38 ERs and AEs are useful tools to identify different plume categories. These results
39 demonstrate the increasing impact of baseline O_3 on US air quality due to both global
40 sources and regional wildfire events.

41 **Keywords:** tropospheric ozone; sub-micron aerosols; western US; enhancement ratios;
42 backward trajectories; long-term trends

43 1. Introduction

44 Tropospheric ozone (O_3) and airborne particles (aerosols), especially fine particulate
45 matter ($PM_{2.5}$ or PM_1), have significant impacts on human health and the environment
46 (Davidson et al., 2005). O_3 plays an important role in atmospheric chemistry and the
47 global climate (IPCC, 2014). Sulfate and other low absorption aerosols have a
48 dominant cooling effect and partially offset warming due to aerosol absorption and
49 greenhouse gases (GHGs) (Charlson et al., 1992; Buseck and Pósfai, 1999; IPCC,
50 2014). There are at least four sources of tropospheric O_3 and aerosols in the western
51 US: upper troposphere and lower stratosphere (UTLS) intrusion, Asian long-range
52 transport (ALRT), Arctic air pollution (AAP) and plumes from the Pacific Northwest
53 region (PNW). In the western US, background sources of O_3 and PM are more
54 important compared to other parts of the US (Jaffe et al., 1999; Heald et al., 2006;
55 Zhang et al., 2009; McDonald-Buller et al., 2011). ALRT can be broadly classified
56 into three categories: biomass burning, mineral dust and industrial pollution (Fischer
57 et al., 2010a; Ambrose et al., 2011).

58 In the troposphere, O_3 comes from secondary photochemical processes involving

59 nitrogen oxides (NO_x nds
60 (VOCs) plus intrusion of stratospheric air. A significant increase in springtime O_3
61 mixing ratios in the western North America has been observed by Cooper et al. (2010)
62 with evidence that it is influenced by increasing precursor emissions from Asia. Lin et
63 al. (2012) argued that much of the recent O_3 increase was driven by variability in
64 circulation. Among the Asian sources of O_3 , Siberian wildfires could also be an
65 important contributor (Gratz et al., 2015). UTLS intrusions have also been linked to
66 high- O_3 events in the western US, especially in springtime (Ambrose et al., 2011;
67 Langford et al., 2012; Lin et al., 2012; Langford et al., 2017). Regional wildfires in
68 the western US also contribute significantly to summertime O_3 enhancement (Pfister
69 et al., 2006; Jaffe et al., 2008; Jaffe et al., 2013; Baylon et al., 2015).

70 Aerosol optical properties, including scattering (σ_{sp}) and absorption (σ_{ap}), are useful
71 indicators for source identification. The wavelength dependence of σ_{sp} is
72 parameterized using the scattering Ångström exponent (\mathring{A}_{sp}), defined as follows
73 (Ångström, 1929):

$$74 \quad \sigma_{sp} = \lambda^{-\mathring{A}_{sp}} \quad (1)$$

75 Because the coefficient σ_{sp} decreases with wavelength for smaller aerosol particles,
76 \mathring{A}_{sp} will be larger for smaller particle distributions (Boren and Huffman, 1983). In a
77 similar way, \mathring{A}_{ap} is the absorption Ångström exponent which is related to the
78 composition, shape and mixing state of the aerosol particles, leading to size variation
79 as well (Bergstrom et al., 2007):

$$80 \quad \sigma_{ap} = \lambda^{-\mathring{A}_{ap}} \quad (2)$$

81 Optical coefficients (σ_{sp} and σ_{ap}) together with their wavelength dependences (\mathring{A}_{sp}
82 and \mathring{A}_{ap}) reflect the differences of diverse aerosol plumes, especially between dust
83 aerosols and combustion-generated carbonaceous aerosols (Fialho et al., 2005; Clarke
84 et al., 2007; Yang et al., 2009; Fischer et al., 2011; Cappa et al., 2012; Cazorla et al.,
85 2013).

86 The slope of a regression curve between two pollutants is known as the
87 enhancement ratio (ER), which is an important tool for plume source apportionment.

89 normalized to CO is often used to identify and characterize the production or loss of
90 the pollutant in smoke plumes, especially wildfire plumes (Goode et al., 2000;
91 Bertschi et al., 2004; Honrath et al., 2004; Bertschi and Jaffe, 2005; Pfister et al., 2006;
92 Val Martin et al., 2006; Real et al., 2007; Paris et al., 2009; Alvarado et al., 2010;
93 Fischer et al., 2010b; Singh et al., 2010; Wigder et al., 2013). This method accounts
94 for air mass dilution given the simple and well understood processing of CO. Jaffe
95 and Wigder (2012) summarized reported $\Delta\text{O}_3/\Delta\text{CO}$ values for wildfire plumes
96 arranged by biome type and by plume age and found that the average $\Delta\text{O}_3/\Delta\text{CO}$ ratio
97 increases with plume age and also that tropical regions tend to have significantly
98 greater ratios compared to boreal and temperate regions. However, they also reported
99 large variability in $\Delta\text{O}_3/\Delta\text{CO}$ ratios. Wigder et al. (2013) focused on the $\Delta\text{PM}_1/\Delta\text{CO}$
100 ratio in wildfire plume events and found evidence for secondary organic aerosol (SOA)
101 production in wildfire plumes. However, they found that with longer transport time
102 PM_1 loss is greater than SOA production. Baylon et al. (2015) looked into wildfire
103 plume events and calculated multiple ERs, including $\Delta\sigma_{sp}/\Delta\text{CO}$, $\Delta\text{O}_3/\Delta\text{CO}$,
104 $\Delta\text{NO}_y/\Delta\text{CO}$, $\Delta\text{NO}_x/\Delta\text{CO}$, $\Delta\text{NO}_x/\Delta\text{NO}_y$ and $\Delta\text{PAN}/\Delta\text{NO}_y$. The correlation between
105 $\Delta\text{O}_3/\Delta\text{CO}$ and $\Delta\text{NO}_x/\Delta\text{NO}_y$ implies that the degree of NO_x oxidation is a key predictor
106 of O_3 production.

107 In this study, we report on seasonal variations and interannual trends of O_3 and
108 sub-micron aerosols at the Mt. Bachelor Observatory (MBO) in central Oregon, USA,
109 for 2004–2015. MBO has one of the longest, continuous records of free tropospheric
110 measurements in North America. Over the 12 years of observations, 177 individual
111 plume events were identified based on elevated σ_{sp} , O_3 or CO. The purpose of this
112 study is to identify background sources of O_3 and aerosols in the western US and
113 characterize each category by a series of signatures including enhancement ratios and
114 aerosol optical properties. We recognize that many plumes have multiple sources
115 associated with them (Cooper et al., 2004), but we can identify the major contributors
116 through gridded emissions for fires and industrial pollution, combined with air mass
117 backward trajectories.

118 2. Methodology

119 2.1. Site description and measurements of air pollutants

120 Mt. Bachelor Observatory (MBO) is located on the summit of a dormant volcano in
121 central Oregon (43°58'39" N 121°41'10" W, 2763 m asl). Due to its elevation and
122 distance from major US source regions, local pollution is relatively rare at MBO.
123 Mixing ratios of O₃ and CO, scattering coefficient of sub-micron aerosols and
124 meteorological parameters (e.g., air temperature, pressure, relative humidity, wind
125 speed and direction, water vapor) have been measured from 2004 to 2015. Mixing
126 ratios of nitrogen oxides (NO_x and NO_y), PAN and mercury as well as sub-micron
127 aerosol scattering and absorption coefficients at multiple wavelengths have been
128 measured during specific campaigns. Detailed instrumentation with method detection
129 limits (MDLs) and estimated total uncertainties (Weiss-Penzias et al., 2006; Fischer et
130 al., 2010a; Fischer et al., 2010b; Virkkula, 2010; Ambrose et al., 2011; Fischer et al.,
131 2011; Chen et al., 2013; Baylon et al., 2015; Gratz et al., 2015; Briggs et al., 2016) is
132 described in Table S1 in the Supporting Information (SI).

133 Both aerosol scattering coefficient (σ_{sp}) and absorption coefficient (σ_{ap}) are reported
134 at ambient temperature and pressure conditions. With σ_{sp} at 450 and 700 nm and σ_{ap} at
135 467 and 660 nm, dimensionless intensive optical properties, including scattering
136 Ångström exponent (\mathring{A}_{sp}) and absorption Ångström exponent (\mathring{A}_{ap}), can be calculated
137 as follows:

$$138 \quad \mathring{A}_{sp} = -\frac{\ln(\sigma_{sp}^{450} / \sigma_{sp}^{700})}{\ln(450 / 700)} \quad (3)$$

$$139 \quad \mathring{A}_{ap} = -\frac{\ln(\sigma_{ap}^{467} / \sigma_{ap}^{660})}{\ln(467 / 660)} \quad (4)$$

140 The coefficients at blue and red wavelengths were utilized in the calculation of the
141 intensive properties, whereas we use the scattering coefficients at green wavelength to
142 calculate plume enhancements and enhancement ratios with other pollutants.

143 2.2. Sounding data analysis

144 Twice daily (0 and 12 UTC) meteorological sounding data were obtained from the

145

146 quantify the vertical distribution of water vapor (WV) concentrations. This was used
147 to help identify free tropospheric (FT) and boundary layer influenced (BLI) air masses
148 at MBO. Soundings from Medford, Oregon (MFR, 42.36° N, 122.86° W, 405 m asl),
149 and Salem, Oregon (SLE, 44.91° N, 123.00° W, 61 m asl), were used to compare with
150 the monthly MBO WV distributions at equivalent pressure level (720–740 mbar). We
151 have used similar techniques previously but at a seasonal resolution (Weiss-Penzias et
152 al., 2006; Fischer et al., 2010b; Ambrose et al., 2011).

153 Ozonesonde data from Trinidad Head, California (THD, 41.05° N, 124.15° W, 107
154 m asl), during 2004–2015 were obtained from the Earth System Research Laboratory
155 (ESRL) of NOAA (<http://www.esrl.noaa.gov/gmd/ozwv/ozsondes>). The THD
156 ozonesonde data are collected in the daytime (16:00–22:00 UTC, or 8:00–14:00 local
157 time).

158 **2.3. Backward trajectory analysis**

159 We computed 240-hour air mass backward trajectories from MBO for every hour of
160 identified plume events using the Hybrid Single-Particle Lagrangian Integrated
161 Trajectory (HYSPLIT) model version 4 (Draxler and Hess, 1998). For 2004 we used
162 meteorological data from the Eta Data Assimilation System (EDAS) (40 km × 40 km)
163 and the Final Global Data Assimilation System (FNL) (191 km × 191 km), and for
164 2005–2015 we used Global Data Assimilation System (GDAS) 1° × 1° gridded
165 meteorological data. The starting height was set to be 1800 m above ground level
166 based on terrain height in the 1° × 1° gridded meteorological data.

167 **2.4. Wildfire maps and gridded industrial CO emission inventories**

168 Daily MODIS fire detection data for North America was obtained from the United
169 States Department of Agriculture (USDA) Forest Service
170 (<http://activefiremaps.fs.fed.us/gisdata.php>). Daily MODIS fire detection data for
171 Eurasia was downloaded from the Fire Information for Resource Management System
172 (FIRMS) of the US National Aeronautics and Space Administration (NASA)
173 (<https://firms.modaps.eosdis.nasa.gov/download>). Gridded maps (0.1° × 0.1°) of

174

175 on the Hemispheric Transport of Air Pollution (HTAP) (<http://www.htap.org>).

176 **2.5. Enhancement ratio analysis**

177 For each plume event, we performed correlation analyses among gaseous pollutants
178 (including O₃, CO, PAN, NO_x and NO_y) and σ_{sp} . Enhancement ratios were calculated
179 by taking the slope ($\Delta Y/\Delta X$) from the reduced major axis (RMA) regression between
180 X and Y (Baylon et al., 2015).

181 **3. Results and Discussion**

182 **3.1. Climatology of tropospheric O₃**

183 3.1.1. Seasonal variation of the O₃ mixing ratio

184 Using the 12-year database (2004–2015), we present the seasonal variation of the O₃
185 mixing ratio at MBO. Fig. 1 shows the distribution of O₃ mixing ratios by month. A
186 bimodal pattern with peaks in April and July was found, which reveals the influence
187 of ALRT, UTLS and regional wildfires in spring and summer. The monthly average
188 O₃ mixing ratio had the highest value (50.4 ppbv) in April and the lowest (40.5 ppbv)
189 in November. O₃ at MBO was highest in the springtime, which was influenced by
190 both ALRT and UTLS intrusion (Weiss-Penzias et al., 2006; Cooper et al., 2010;
191 Ambrose et al., 2011; Lin et al., 2012; Gratz et al., 2015). Enhancements in
192 summertime O₃ at MBO were mainly from regional wildfires in the western US (Jaffe
193 et al., 2013; Wigder et al., 2013; Baylon et al., 2015) and UTLS intrusion in early
194 summer, although Siberian fire plumes also contribute occasionally (Ambrose et al.,
195 2011). The contribution of each source is further discussed in Section 3.4.

196 3.1.2. Trends and interannual variation of the O₃ mixing ratio

197 Fig. 2 shows the interannual variation of O₃ by season from 2004 to 2015. From 2004
198 to 2015, the mean O₃ mixing ratio in spring, summer and fall increased by 0.62 ± 0.25
199 ppbv yr^{-1} ($r^2 = 0.38$, $p < 0.05$), 0.66 ± 0.27 ppbv yr^{-1} ($r^2 = 0.38$, $p < 0.05$) and
200 0.79 ± 0.34 ppbv yr^{-1} ($r^2 = 0.35$, $p < 0.05$), respectively. This trend appears to be driven
201 by Asian pollution in spring (Ambrose et al., 2011) and regional wildfires in summer
202 and early fall (Jaffe et al., 2013; Wigder et al., 2013; Baylon et al., 2015). However,

204 level and 2015 has a high O₃ level (Jaffe and Zhang, 2017). The trends in the three
205 seasons were not significant ($p > 0.05$) for the period of 2005–2014 in contrast to the
206 trends for the entire data record from 2004 to 2015. The springtime O₃ increase was
207 probably linked to the increase of Asian emissions (Gratz et al., 2015). The yearly
208 98th percentile of O₃ at MBO, reflecting the design value, had a more significant
209 positive trend than the median (see Fig. S1 in the SI). Lin et al. (2015) summarized
210 previous O₃ observations from sondes, lidars and aircraft campaigns in the western
211 US during 1995–2014 and showed a significant positive trend. However, the trend
212 significantly slowed after 2005. Interannual variations are also very important.
213 Significant jumps of the average O₃ mixing ratio were found in the summer of 2012
214 and 2015. The 2012 jump was probably related to the more frequent wildfire events as
215 indicated by higher mean CO and aerosols at MBO. It is further verified by source
216 identification in Section 3.4. The 2015 jump was probably linked to summer wildfires
217 and the temperature anomaly in the western US in 2015, especially June (Jaffe and
218 Zhang, 2017).

219 **3.2. Climatology of sub-micron aerosol scattering**

220 3.2.1. Seasonal variation of the aerosol scattering coefficient

221 Fig. 3 displays the distribution of aerosol scattering coefficient (σ_{sp} , green) by month
222 in logarithmic scale. Sub-micron σ_{sp} also has a bimodal seasonal pattern. The highest
223 peak occurred in August followed by a smaller peak in May. Summer plume events
224 had a more significant influence on sub-micron aerosols than spring plume events.
225 The highest value in August was about three orders of magnitude higher than the
226 median value. The springtime aerosols are mainly from ALRT, including biomass
227 burning, mineral dust and industrial pollution (Fischer et al., 2010a; Fischer et al.,
228 2011), while summertime aerosols are mostly generated from regional and Siberian
229 wildfires (Wigder et al., 2013; Laing et al., 2016). In summer, primary aerosols from
230 regional wildfires have more significant impacts on MBO than aerosols from
231 long-range transport, although Siberian wildfires are also important (Laing et al.,

232 2016).

233 3.2.2. Interannual variation of the aerosol scattering coefficient

234 Fig. 4 shows the interannual variation of sub-micron aerosol scattering coefficient by
235 season from 2004 to 2015. For aerosol scattering, we saw an increase only in summer
236 but not statistically significant ($r^2 = 0.32$, $p = 0.06$). This could be driven by recent
237 increases in wildfire activity in the western US (Jaffe et al., 2013; Wigder et al., 2013;
238 Baylon et al., 2015). The level of aerosol scattering in summer 2015 was
239 exceptionally high because of the frequent and large wildfire events (Laing et al.,
240 2016).

241 3.3. Free tropospheric O₃ and aerosols

242 3.3.1. Criteria for FT and BLI air masses at MBO

243 At MBO daily upslope and downslope airflows cause a pronounced diurnal variation
244 in O₃, WV, CO₂ and other pollutants (McClure et al., 2016). We used the 2010–2015
245 sounding data from both MFR and SLE to obtain monthly distributions of the WV
246 concentration at the 720–740 mbar level, which is similar to the MBO pressure. This
247 pressure level at MFR and SLE represent FT air masses as indicated by the lack of
248 diurnal WV variability (Weiss-Penzias et al., 2006; Ambrose et al., 2011). Hourly WV
249 data from MBO at the same period were also processed for monthly distributions. The
250 significant diurnal change of WV suggests an influence of planetary boundary layer
251 (PBL) on the air at MBO in daytime (Reidmiller et al., 2010; McClure et al., 2016).

252 A FT dataset for MBO was defined by the monthly distribution of all WV data at
253 720–740 mbar from the MFR and SLE soundings. We obtained a similar WV
254 distribution from MBO by retaining the drier portion of the distribution such that the
255 monthly averages of the MBO data and the soundings were equivalent. The cut points
256 are referred to as the WV criteria for FT air masses at MBO, as shown in Table 1. The
257 retained data represent FT air masses, while the non-retained data stand for BLI air
258 masses. The BLI air is not purely boundary layer or FT air but represents a mix of the
259 two. MBO is influenced constantly by FT air and occasionally by daytime upslope
260 flow to the summit from the more moist PBL air (McClure et al., 2016). The seasonal

261

262 However with a monthly resolution rather than seasonal, this study contributes to a
263 more accurate FT/BLI isolation.

264 3.3.2. Free tropospheric O₃

265 Based on the WV criteria, the O₃ mixing ratios for FT and BLI were separated and the
266 monthly averages of each are shown in Fig. 5. There is a significant difference (5–10
267 ppbv) between the average O₃ mixing ratios for FT and BLI at MBO. The FT O₃ had a
268 peak in May unlike the BLI case, but the variations and trends for FT and BLI O₃
269 were similar. The most significant discrepancies occurred in May and June, reflecting
270 the influence of ALRT and UTLS O₃ intrusion. We compared seasonal tropospheric
271 O₃ between MBO and THD (see Fig. 5). THD tropospheric O₃ has a bimodal seasonal
272 pattern, 2–8 ppbv higher than MBO. The highest average THD O₃ mixing ratios occur
273 in April. The significant peak of THD O₃ in August suggests a more significant impact
274 of regional industrial pollution and/or wildfires compared to MBO. The difference
275 between THD and MBO could be caused by large-scale transport and local
276 photochemical condition instead of PBL influence. Fig. S2 in the SI shows the
277 climatology of vertical velocity at 700 mbar in the US. THD exhibits a positive
278 overall vertical velocity while MBO has a negative one. Pacific high pressure and
279 subsidence in northern California increase O₃ at THD (Stauffer et al., 2017), while the
280 mean negative vertical velocity indicates greater upward airflow with high WV at
281 MBO.

282 3.3.3. Free tropospheric aerosols

283 With the WV criteria, sub-micron aerosol scattering coefficients for FT and BLI were
284 also separated, and the monthly average σ_{sp} is shown in Fig. 6. The mean FT σ_{sp} was
285 higher in spring, but the mean BLI σ_{sp} was significantly higher in summer. This
286 implies that the regional wildfire smoke (at lower elevation) in summer had more
287 influence on the BLI atmosphere while global sources had more impacts on the FT air
288 masses in the rest of the year. The bimodal pattern was more prominent for the FT air
289 masses than BLI air masses due to springtime ALRT in the free troposphere. The

290

291 et al. (2016) reported that 6 of 19 biomass burning events at MBO during August
292 2015 were influenced by Siberian fires originating near Lake Baikal. The next section
293 provides more evidence on the contribution of Siberian wildfires in summer.

294 **3.4. Plume event identification**

295 To verify the pollution sources that cause the seasonal patterns and interannual trends
296 of O_3 and σ_{sp} , we identified plume events using hourly data of σ_{sp} , O_3 or CO that are
297 above the 97.5 percentile of the annual dataset. The identified hours are called
298 “polluted hours”. A group of no less than eight consecutive polluted hours, plus two
299 hours before and two hours after, was identified as a plume event (i.e., one event is at
300 least 12 hours long). We referred to the pollutants that are enhanced for each event as
301 the trigger pollutants. Based on these criteria, a total of 177 plume events were
302 identified during 2004–2015. It should be noted that most plumes have multiple
303 sources.

304 Based on backward trajectories and correlation between WV and O_3 , plume events
305 at MBO can be broadly classified into four categories: upper troposphere and lower
306 stratosphere (UTLS) O_3 intrusion, Asian long-range transport (ALRT), Arctic air
307 pollution (AAP), and plumes from Pacific Northwest (PNW). UTLS events have
308 trajectories derived from higher elevation (5000–10000 m agl). ALRT events have
309 longer trajectories, often tracing back to Siberia, Mongolia, China or Japan. AAP
310 events have trajectories passing through the Arctic at lower elevations. PNW events
311 have shorter trajectories that originate from the Pacific Northwest region at low
312 elevation. Fig. 7a exhibits the WV- O_3 relationships for the THD sounding data (1997–
313 2017). WV and O_3 are overall negatively correlated. The negative slope of $\Delta O_3/\Delta WV$
314 is steeper at higher elevations. MBO data showed a similar pattern as the THD data
315 (Fig. 7b) which can be used for source identification. UTLS and ALRT events
316 originate from higher elevations, and thus have higher O_3 and lower WV with steeper
317 $\Delta O_3/\Delta WV$ slopes. The WV concentrations of these two categories are usually lower
318 than 3 g kg^{-1} , while those of PNW events are usually higher than 4 g kg^{-1} . UTLS and

319

320 often climb up and back down over the Pacific Ocean entraining UTLS air masses.
321 Besides backward trajectories, we used the CO level to make a rough separation.
322 When the mean CO is in the range of 110–130 ppbv, it is considered as a mixture of
323 UTLS and ALRT. CO level below this range implies the dominant contribution of
324 UTLS while above this range ALRT. In either case, it is difficult to separate source of
325 O₃ for an air parcel transported at high elevations.

326 ALRT and PNW events can be further categorized into wildfires (WF), industrial
327 pollution (IP) and mineral dust (MD). Backward trajectories, wildfire detection maps
328 from MODIS and gridded industrial CO emission inventories from HTAP were used
329 for identifying these sub-categories. For ALRT events, an air mass may pass through
330 fire spots, an industry-intensive region (e.g., China, Japan) or a desert region (e.g.,
331 Mongolia) at low elevation (<3000 m agl). This would carry the air pollution from all
332 of these sources. ALRT events are usually mixed events, but we can identify which is
333 the major contributor among WF, IP and MD. ALRT-MD has elevated σ_{sp} but low CO.
334 However, it is much more difficult to distinguish ALRT-WF from ALRT-IP. To
335 identify the dominant source, we estimate the CO emission rates of WF and IP in the
336 grid boxes through which the trajectories pass to determine their relative contribution.

337 The backward trajectories from MBO back to East Asia have uncertainties of up to
338 1000 kilometers (Draxler and Hess, 1998). Therefore, we partitioned the East Asia
339 region into 10° × 10° grids and identify those grids where the air mass passes through
340 at low altitude (<3000 m agl) as “influential grids”. Based on the industrial CO
341 emission inventory in 2010 from HTAP, we calculated the average CO emission rate
342 of each grid. Fig. S3 in the SI shows the industrial CO emission rates of all grids in
343 East Asia. The grid containing the North China Plain region (Beijing, Tianjin, Hebei,
344 etc.) has the highest industrial emission rate (1024 kg CO s⁻¹). Wildfire maps for the
345 East Asia region seven days (average transport time) before each event were obtained
346 from the MODIS data. We calculated the CO emission rate from wildfires for each
347 fire spot based on the fire radiative power (FRP) using the following equation adopted
348 from Kaiser et al. (2012):

349
$$r = \kappa_l \cdot \beta_l \cdot \text{FRP} \quad (5)$$

350 where r is the emission rate of CO, g CO s^{-1} ; κ_l is the emission factor of CO that is
 351 land cover type dependent, g CO kg^{-1} (dry matter); β_l is the conversion factor that is
 352 also land cover type dependent, $\text{kg (dry matter) MJ}^{-1}$; FRP is the fire radiative power
 353 from the MODIS dataset, MW; l is the land cover type, including extratropical forest
 354 (EF), tropical forest (TF), savannah (SA), agriculture (AG), and so on. κ_l and β_l are
 355 from the study of Kaiser et al. (2012), and land cover information is from the MODIS
 356 data by Global Land Cover Facility (<http://glcf.umd.edu/data/lc>).

357 CO emission rates for each grid were obtained by adding up emission rates of all
 358 fire spots in the grid. We used the ratio of wildfire/industrial (W/I) CO emission rates
 359 in the influential grids to quantify the largest source. Fig. 8 shows the CO emission
 360 rates and W/I ratios of an ALRT-WF event and an ALRT-IP event. If the W/I ratio of
 361 an ALRT event is higher than 2, it is regarded as an ALRT-WF event; if lower than 0.5,
 362 it is regarded as an ALRT-IP event; and if the W/I ratio is in the range of 0.5–2, it is
 363 considered a mixing event (ALRT-WF/IP). Most events with influential grids in
 364 Siberia are ALRT-WF events, while most events with influential grids in China are
 365 ALRT-IP events.

366 Most PNW events are PNW-WF with a high slope of $\Delta\sigma_{sp}/\Delta\text{CO}$. Air masses of
 367 PNW-IP events usually occur when regional wildfires are absent and the air mass
 368 circulates in an industry-intensive area with a long residence time in the western US at
 369 low elevation (<1000 m agl). Air masses of PNW-MD events are from the desert areas
 370 in the southwestern US with elevated σ_{sp} but low CO. PNW-IP and PNW-MD are rare
 371 events at MBO.

372 Fig. 9 shows the distribution of different types of plume events during 2004–2015
 373 by month. UTLS events occur most frequently in spring and early summer. ALRT
 374 events have the highest frequency in spring with a few in summer, mainly Siberian
 375 wildfire events. PNW events mainly occur in summer and early fall.

376 **3.5. Characteristics of typical plume events**

377 Table S2 in the SI shows a summary of all 177 plume events. Six typical plume events

378

379 PNW-IP. Backward trajectories of these events are shown in Fig. 10. WV-O₃ and
380 CO- σ_{sp} relations of these events are shown in Figs. S4–S9 in the SI. Means and
381 maximum values of hourly σ_{sp} , O₃, CO, NO_x, PAN, NO_y and WV of these events are
382 shown in Table 2.

383 3.5.1. Event 138: 21 July 2013 (UTLS)

384 Event 138 occurred at 4:00–20:00 UTC on July 21, 2013. In this event, O₃ was the
385 only elevated compound, while CO and σ_{sp} were both at very low levels. O₃ was
386 negatively correlated with WV ($R^2 = 0.84$). This event had the lowest hourly average
387 of WV (1.20 g kg⁻¹) among the six events. The trajectories came from a very high
388 altitude above the Pacific Ocean (Fig. 10a). Fig. S10 in SI shows the geopotential
389 height at 850 mbar from the National Centers for Environmental Prediction (NCEP)
390 Reanalysis for the East Pacific and North America region on 15 July 2013. We
391 observe the confrontation of high and low pressure zones above the East Pacific
392 region, which is the primary driving force of UTLS intrusion.

393 3.5.2. Event 151: 19 April 2014 (ALRT-WF)

394 Event 151 occurred between 2:00–18:00 UTC on April 19, 2014. The average σ_{sp} rose
395 up to a maximum hourly average of 39.7 Mm⁻¹, and CO and O₃ elevated to high
396 levels. There were significant positive correlations between σ_{sp} and CO ($R^2 = 0.93$)
397 and between σ_{sp} and O₃ ($R^2 = 0.45$). The hourly average WV (2.45 g kg⁻¹) was lower
398 than regional events. Air mass trajectories traveled at low elevation through an
399 intensive wildfire region at the eastern junction of Siberia and China on April 11–13
400 (Fig. 10b). Fig. S11 in the SI shows the CALIPSO lidar vertical profiles one day
401 before arrival at the western US on April 19. Polluted dust and cloud were observed in
402 the profiles. The W/I ratio of this event was 5.28.

403 3.5.3. Event 148: 20–21 November 2013 (ALRT-IP)

404 Event 148 took place in 2013 from 21:00 on November 20 to 11:00 on November 21
405 (UTC). CO and σ_{sp} were enhanced in the plume. O₃ was at a low level. There were
406 significant positive correlations between σ_{sp} and CO ($R^2 = 0.91$) and between O₃ and

407 CO (R^2)
408 which was an industry-intensive region (Fig. 10c). This path avoided wildfire
409 intensive regions and carried with it heavy industrial pollution. The CALIPSO
410 profiles show the arrival of this event in the western US on November 21 (Fig. S12 in
411 SI). ALRT-IP air masses frequently experience cloud processing during transport,
412 which will reduce the amount of aerosols but not the CO. The W/I ratio of this event
413 was 0.06.

414 3.5.4. Event 102: 1 May 2011 (ALRT-MD)

415 Event 102 occurred at 3:00–15:00 UTC on May 1, 2011. The σ_{sp} and O_3 levels were
416 both elevated, but CO remained low. This is typical for ALRT-MD events, which are
417 often mixed with UTLS air. Positive correlation between O_3 and σ_{sp} was observed (R^2
418 = 0.54). Trajectories traveled at low elevation through the Gobi Desert region in South
419 Mongolia and North China, carrying a significant amount of mineral dust (Fig. 10d).
420 Fig. S13 in the SI shows the CALIPSO profiles at the arrival section in the western
421 US on May 1, and the profiles corroborate the presence of mineral dust.

422 3.5.5. Event 123: 25–26 August 2012 (PNW-WF)

423 Event 123 took place in 2012 from 11:00 on August 25 to 22:00 on August 26 (UTC).
424 Extremely high maximum hourly values of σ_{sp} (407 Mm^{-1}) and CO (707 ppbv) were
425 found in this event with a very strong correlation ($R^2 = 0.97$). NO_x and NO_y were also
426 enhanced significantly to maximum hourly values (577 and 4648 pptv, respectively).
427 Fig. 10e shows the local trajectories in the western US. Low-altitude trajectories
428 traveled through two hotspots of wildfires with high FRP values in northern California
429 on August 25–26.

430 3.5.6. Event 53: 25 May 2007 (PNW-IP)

431 Event 53 occurred at 7:00–18:00 UTC on May 25, 2007. In this event, O_3 was
432 enhanced with an hourly average of 64.0 ppbv while CO was not (an hourly average
433 of 124 ppbv). Trajectories circulated at low altitude above the Pacific Ocean and
434 traveled through western Oregon to MBO without impact of wildfires (Fig. 10f). WV
435 was also greater than the FT cutoff value. Urban pollution from industries and mobile

436

437 probably the major contributors. Trajectories for this event were originated from
438 subtropical latitudes where background O₃ is lower, indicating the contribution from
439 local sources.

440 **3.6. Signatures of different event categories**

441 Table 3 shows the mean enhancement ratios (ERs) and the aerosol Ångström
442 exponents of different event types. These are important signatures for source
443 identification. PNW-MD events have little data available and are thus not listed in
444 Table 3.

445 3.6.1. UTLS

446 UTLS events have steep negative $\Delta O_3/\Delta WV$ slopes (-10.05 ± 3.26 ppbv kg g⁻¹), which
447 reveals that an anomalously low WV level is usually linked to a strong O₃ intrusion.
448 UTLS and UTLS/ALRT events also have the only negative average $\Delta \sigma_{sp}/\Delta O_3$ among
449 all event types. The $\Delta O_3/\Delta CO$ slope varies significantly among different UTLS events.
450 UTLS events have higher $\Delta NO_x/\Delta PAN$ (0.96 ± 0.64 pptv pptv⁻¹) than PNW-WF events.
451 However, due to limited available data, this comparison should be interpreted with
452 caution. The high \dot{A}_{sp} in UTLS events should be interpreted with caution as well given
453 the extremely low scattering coefficients.

454 3.6.2. ALRT-WF

455 ALRT-WF events have the highest $\Delta O_3/\Delta CO$ among all event types (0.52 ± 0.64 ppbv
456 ppbv⁻¹), which is caused by the large amount of O₃ precursors generated from
457 wildfires and the long transport time for secondary O₃ generation. ALRT-WF and
458 ALRT-WF/IP events have lower \dot{A}_{ap} levels (1.45 ± 0.02 and 1.54 ± 0.39) than PNW-WF
459 events (1.81 ± 0.59), which implies that brown carbon (BrC) is generated from the
460 biomass burning process but its fraction decreases during the long-range transport due
461 to photobleaching (Laing et al., 2016). O₃ from Siberian wildfires can be an important
462 influence on US surface air quality (Jaffe, 2004).

463 3.6.3. ALRT-IP

464

σ_{sp}

465 due to the high level of CO in industrial pollution. The $\Delta O_3/\Delta CO$ slope of ALRT-IP
466 events (0.21 ± 0.50 ppbv ppbv⁻¹) is higher than PNW-WF but lower than ALRT-WF.
467 The long-range transport gives ALRT-IP events sufficient time for O₃ generation.
468 ALRT-IP events have a less steep negative $\Delta O_3/\Delta WV$ (-7.88 ± 6.26 ppbv kg g⁻¹) than
469 other ALRT events, because ALRT-IP events are usually linked to cloud process
470 during transport resulting in higher WV. ALRT-IP events have the lowest average \dot{A}_{sp}
471 (2.02 ± 0.20) among all types of events, indicating relatively larger particle size. This is
472 possibly due to mixing with mineral dust sources during long-range transport.

473 3.6.4. ALRT-MD

474 ALRT-MD events are usually associated with UTLS. Mineral dust plumes occur in
475 drier air masses (lower WV), and the $\Delta O_3/\Delta WV$ of ALRT-MD is thus steeper
476 (-13.94 ± 6.81 ppbv kg g⁻¹). In recent years, ALRT-MD events are rare compared to
477 large MD events that occurred in prior years (Husar et al., 2001; Jaffe et al., 2003).

478 3.6.5. AAP

479 The average $\Delta \sigma_{sp}/\Delta CO$ of AAP (0.13 ± 0.15 Mm⁻¹ ppbv⁻¹) is lower than ALRT-IP.
480 Arctic pollution reservoir originates mainly from industrial pollution. One signature of
481 AAP events is the high CO level and low oxidant level, resulting in low $\Delta O_3/\Delta CO$
482 (-0.28 ± 0.60 ppbv ppbv⁻¹) and low $\Delta NO_y/\Delta CO$ (2.72 ± 0.26 pptv ppbv⁻¹). However, the
483 number of events is limited to make a conclusive statement.

484 3.6.6. PNW-WF

485 PNW-WF events have the highest $\Delta \sigma_{sp}/\Delta CO$, $\Delta \sigma_{sp}/\Delta O_3$ and $\Delta \sigma_{sp}/\Delta NO_y$ among all
486 types of events (0.59 ± 0.17 Mm⁻¹ ppbv⁻¹, 10.98 ± 13.16 Mm⁻¹ ppbv⁻¹ and 0.10 ± 0.07
487 Mm⁻¹ pptv⁻¹, respectively). This is a distinctive feature of PNW-WF events, where
488 aerosol is a dominant pollutant and has a shorter residence time than the gaseous
489 pollutants (CO, O₃ and NO_y) (Laing et al., 2016). PNW-WF events have a lower level
490 of $\Delta O_3/\Delta CO$ (0.10 ± 0.12 ppbv ppbv⁻¹) than ALRT-WF events, indicating secondary O₃
491 formation during long-range transport and consistent with greater O₃ production after
492 extended transport (Jaffe and Wigder, 2012).

493 3.6.7. PNW-IP

494 Because MBO is near to the coast and has no large cities upwind, PNW-IP events at
495 MBO are relatively rare. These have a steeper negative $\Delta O_3/\Delta WV$ slope (-8.57 ± 0.68
496 ppbv kg g^{-1}) than PNW-WF events, demonstrating a stronger inverse relation likely
497 associated with BLI/FT exchange.

498 4. Conclusions

499 Mixing ratios of O_3 , CO, NO_x , NO_y and PAN and the optical properties of sub-micron
500 aerosols have been measured at MBO from 2004–2015. The monthly average O_3
501 distribution has a bimodal pattern with peaks in April and July. ALRT and UTLS
502 events were the main contributors of O_3 in the western US in spring, while regional
503 wildfires and UTLS contributed the most in summer with a non-negligible
504 contribution from Siberian wildfires. Regional wildfires had more significant impacts
505 on σ_{sp} , resulting in a very high peak in summer especially August. A smaller peak of
506 σ_{sp} is seen in May, likely due to ALRT. Between 2004–2015, the mean O_3 mixing ratio
507 in spring, summer and fall increased by 0.6–0.8 ppbv yr^{-1} driven by Asian pollution in
508 spring and regional wildfires in summer and early fall.

509 We used the 2010–2015 sounding WV data from MFR and SLE at equivalent
510 pressure level with MBO to separate FT and BLI air masses at MBO. At MBO, O_3 is
511 significantly higher (5–10 ppbv) in FT air compared to BLI air. The BLI mean σ_{sp} was
512 higher than the FT mean in summer while lower in other seasons, showing that the
513 regional air pollution in summer had more influence on the BLI atmosphere. O_3 at
514 MBO is slightly lower than an equivalent alternative (THD) over northern California
515 due to the large-scale synoptic patterns. The bimodal pattern of σ_{sp} was more
516 prominent for the FT air masses than BLI air masses due to springtime ALRT in the
517 free troposphere.

518 Over the 12 years of observations, we identified 177 individual plume events.
519 These can be broadly classified into four categories (UTLS, ALRT, AAP and PNW)
520 based on backward trajectories and the WV- O_3 relationship. ALRT and PNW can be
521 further divided into WF, IP and MD based on the CO level and the W/I ratio. Six

522

523 exponents, regarded as signatures of different event categories, were calculated.
524 UTLS events have the lowest $\Delta\sigma_{sp}/\Delta O_3$ slope. PNW-WF events have the highest
525 $\Delta\sigma_{sp}/\Delta CO$, $\Delta\sigma_{sp}/\Delta O_3$ and $\Delta\sigma_{sp}/\Delta NO_y$ among all types of events. The lower averages of
526 ALRT-WF events indicate the sub-micron aerosol has a shorter residence time than
527 the gaseous pollutants. The comparison of \dot{A}_{ap} between PNW-WF and ALRT-WF
528 implies that BrC is generated from biomass burning, but its fraction decreases during
529 the long-range transport, likely due to photobleaching.

530 **Acknowledgements**

531 The Mt. Bachelor Observatory is supported by the National Science Foundation (grant
532 #AGS-1447832) and the National Oceanic and Atmospheric Administration (contract
533 #RA-133R-16-SE-0758). We also acknowledge the critical data used in this analysis
534 provided by NOAA, NASA, USDA, HTAP and the National Weather Service.
535 HYSPLIT data were processed by Jonathan Hee. Editing assistance was provided by
536 Dee Ann Lommers-Johnson. MBO data are permanently archived at the University of
537 Washington data repository (<https://digital.lib.washington.edu/researchworks>).

538 **References**

539 Alvarado, M.J., Logan, J.A., Mao, J., Apel, E., Riemer, D., Blake, D., Cohen, R.C.,
540 Min, K.-E., Perring, A.E., Browne, E.C., Wooldridge, P.J., Diskin, G.S., Sachse,
541 G.W., Fuelberg, H., Sessions, W.R., Harrigan, D.L., Huey, G., Liao, J.,
542 Case-Hanks, A., Jimenez, J.L., Cubison, M.J., Vay, S.A., Weinheimer, A.J.,
543 Knapp, D.J., Montzka, D.D., Flocke, F.M., Pollack, I.B., Wennberg, P.O., Kurten,
544 A., Crouse, J., Clair, J.M.S., Wisthaler, A., Mikoviny, T., Yantosca, R.M.,
545 Carouge, C.C., Le Sager, P., 2010. Nitrogen oxides and PAN in plumes from
546 boreal fires during ARCTAS-B and their impact on ozone: an integrated analysis
547 of aircraft and satellite observations. *Atmos. Chem. Phys.* 10, 9739–9760.
548 Ambrose, J.L., Reidmiller, D.R., Jaffe, D.A., 2011. Causes of high O₃ in the lower
549 free troposphere over the Pacific Northwest as observed at the Mt. Bachelor
550 Observatory. *Atmos. Environ.* 45, 5302–5315.

551
552 the air. *Geogr. Ann.* 11, 156–166.

553 Baylon, P., Jaffe, D.A., Wigder, N.L., Gao, H., Hee, J., 2015. Ozone enhancement in
554 western US wildfire plumes at the Mt. Bachelor Observatory: The role of NO_x.
555 *Atmos. Environ.* 109, 297–304.

556 Bergstrom, R.W., Pilewskie, P., Russell, P.B., Redemann, J., Bond, T.C., Quinn, P.K.,
557 Sierau, B., 2007. Spectral absorption properties of atmospheric aerosols. *Atmos.*
558 *Chem. Phys.* 7, 5937–5943.

559 Bertschi, I.B., Jaffe, D.A., Jaeglé, L., Price, H.U., Dennison, J.B., 2004.
560 PHOBEA/ITCT 2002 airborne observations of trans-Pacific transport of ozone,
561 CO, VOCs and aerosols to the northeast Pacific: impacts of Asian anthropogenic
562 and Siberian Boreal fire emissions. *J. Geophys. Res.* 109, D23S12,
563 doi:10.1029/2003JD004328.

564 Bertschi, I.B., Jaffe, D.A., 2005. Long-range transport of ozone, carbon monoxide and
565 aerosols to the NE Pacific troposphere during the summer of 2003: observations
566 of smoke plumes from Asian Boreal fires. *J. Geophys. Res.* 110, D05303,
567 doi:05310.01029/02004JD005135.

568 Boren, C.F., Huffman, D.R., 1983. *Absorption and Scattering of Light by Small*
569 *Particles*. John Wiley, Hoboken, NJ.

570 Briggs, N.L., Jaffe, D.A., Gao, H., Hee, J.R., Baylon, P.M., Zhang, Q., Zhou, S.,
571 Collier, S.C., Sampson, P.D., Cary, R.A., 2016. Particulate matter, ozone, and
572 nitrogen species in aged wildfire plumes observed at the Mount Bachelor
573 Observatory. *Aerosol Air Qual. Res.*, in press.

574 Buseck, P.R., Pósfai, M., 1999. Airborne minerals and related aerosol particles:
575 Effects on climate and the environment. *Proc. Natl. Acad. Sci. USA* 96, 3372–
576 3379.

577 Cappa, C.D., Onasch, T.B., Massoli, P., Worsnop, D.R., Bates, T.S., Cross, E.S.,
578 Davidovits, P., Hakala, J., Hayden, K.L., Jobson, B.T., Kolesar, K.R., Lack, D.A.,
579 Lerner, B.M., Li, S.-M., Mellon, D., Nuaaman, I., Olfert, J.S., Petäjä, T., Quinn,
580 P.K., Song, C., Subramanian, R., Williams, E.J., Zaveri, R.A., 2012. Radiative

581
582 Science 337, 1078–1081.

583 Cazorla, A., Bahadur, R., Suski, K.J., Cahill, J.F., Chand, D., Schmid, B., Ramanathan,
584 V., Prather, K.A., 2013. Relating aerosol absorption due to soot, organic carbon,
585 and dust to emission sources determined from in-situ chemical measurements.
586 Atmos. Chem. Phys. 13, 9337–9350.

587 Charlson, R.J., Schwartz, S.E., Hales, J.M., Cess, R.D., Coakley, J.A., Hansen, J.E.,
588 Hofmann, D.J., 1992. Climate forcing by anthropogenic aerosols. Science 255,
589 423–430.

590 Chen, H., Karion, A., Rella, C.W., Winderlich, J., Gerbig, C., Filges, A., Newberger,
591 T., Sweeney, C., Tans, P.P., 2013. Accurate measurements of carbon monoxide in
592 humid air using the cavity ring-down spectroscopy (CRDS) technique. Atmos.
593 Meas. Tech. 6, 1031–1040.

594 Clarke, A., McNaughton, C., Kapustin, V., Shinozuka, Y., Howell, S., Dibb, J., Zhou,
595 J., Anderson, B., Brekhovskikh, V., Turner, H., Pinkerton, M., 2007. Biomass
596 burning and pollution aerosol over North America: Organic components and
597 their influence on spectral optical properties and humidification response. J.
598 Geophys. Res. 112, D12S18, doi:10.1029/2006JD007777.

599 Cooper, O.R., Forster, C., Parrish, D., Trainer, M., Dunlea, E., Ryerson, T., Hübler, G.,
600 Fehsenfeld, F., Nicks, D., Holloway, J., de Gouw, J., Warneke, C., Roberts, J.M.,
601 Flocke, F., Moody, J., 2004. A case study of transpacific warm conveyor belt
602 transport: Influence of merging airstreams on trace gas import to North America.
603 J. Geophys. Res. 109, D23S08.

604 Cooper, O.R., Parrish, D.D., Stohl, A., Trainer, M., Nédélec, P., Thouret, V., Cammas,
605 J.P., Oltmans, S.J., Johnson, B.J., Tarasick, D., Leblanc, T., McDermid, I.S., Jaffe,
606 D., Gao, R., Stith, J., Ryerson, T., Aikin, K., Campos, T., Weinheimer, A., Avery,
607 M.A., 2010. Increasing springtime ozone mixing ratios in the free troposphere
608 over western North America. Nature 463, 344–348.

609 Davidson, C.I., Phalen, R.F., Solomon, P.A., 2005. Airborne particulate matter and
610 human health: A review. Aerosol Sci. Technol. 39, 737–749.

611

612 for trajectories, dispersion and deposition. *Aust. Meteorol. Mag.* 47, 295–308.

613 Fialho, P., Hansen, A.D.A., Honrath, R.E., 2005. Absorption coefficients by aerosols

614 in remote areas: A new approach to decouple dust and black carbon absorption

615 coefficients using seven-wavelength Aethalometer data. *J. Aerosol Sci.* 36, 267–

616 282.

617 Fischer, E.V., Jaffe, D.A., Marley, N.A., Gaffney, J.S., Marchany-Rivera, A., 2010a.

618 Optical properties of aged Asian aerosols observed over the U.S. Pacific

619 Northwest. *J. Geophys. Res.* 115, D20209, doi:20210.21029/22010JD013943.

620 Fischer, E.V., Jaffe, D.A., Reidmiller, D.R., Jaegle, L., 2010b. Meteorological controls

621 on observed peroxyacetyl nitrate at Mount Bachelor during the spring of 2008. *J.*

622 *Geophys. Res.* 115, D03302, doi:03310.01029/02009JD012776.

623 Fischer, E.V., Perry, K.D., Jaffe, D.A., 2011. Optical and chemical properties of

624 aerosols transported to Mount Bachelor during spring 2010. *J. Geophys. Res.* 116,

625 D18202, doi:18210.11029/12011JD015932.

626 Goode, J.G., Yokelson, R.J., Ward, D.E., Susott, R.A., Babbitt, R.E., Davies, M.A.,

627 Hao, W.M., 2000. Measurements of excess O₃, CO₂, CO, CH₄, C₂H₄, C₂H₂,

628 HCN, NO, NH₃, HCOOH, CH₃COOH, HCHO, and CH₃OH in 1997 Alaskan

629 biomass burning plumes by airborne Fourier transform infrared spectroscopy

630 (AFTIR). *J. Geophys. Res.* 105, 22147–22166.

631 Gratz, L.E., Jaffe, D.A., Hee, J.R., 2015. Causes of increasing ozone and decreasing

632 carbon monoxide in springtime at the Mt. Bachelor Observatory from 2004 to

633 2013. *Atmos. Environ.* 109, 323–330.

634 Heald, C.L., Jacob, D.J., Park, R.J., Alexander, B., Fairlie, T.D., Yantosca, R.M., Chu,

635 D.A., 2006. Transpacific transport of Asian anthropogenic aerosols and its

636 impact on surface air quality in the United States. *J. Geophys. Res.* 111, D14310,

637 doi:14310.11029/12005JD006847.

638 Honrath, R.E., Owen, R.C., Val Martín, M., Reid, J.S., Lapina, K., Fialho, P., Dziobak,

639 M.P., Kleissl, J., Westphal, D.L., 2004. Regional and hemispheric impacts of

640 anthropogenic and biomass burning emissions on summertime CO and O₃ in the

641

642 doi:24310.21029/22004JD005147.

643 Husar, R.B., Tratt, D.M., Schichtel, B.A., Falke, S.R., Li, F., Jaffe, D., Gassó, S., Gill,
644 T., Laulainen, N.S., Lu, F., Reheis, M.C., Chun, Y., Westphal, D., Holben, B.N.,
645 Gueymard, C., McKendry, I., Kuring, N., Feldman, G.C., McClain, C., Frouin,
646 R.J., Merrill, J., DuBois, D., Vignola, F., Murayama, T., Nickovic, S., Wilson,
647 W.E., Sassen, K., Sugimoto, N., Malm, W.C., 2001. Asian dust events of April
648 1998. *J. Geophys. Res.* 106, 18317–18330.

649 Intergovernmental Panel on Climate Change (IPCC), 2014. *Climate Change 2014:*
650 *Synthesis Report. Contribution of Working Groups I, II and III to the Fifth*
651 *Assessment Report of the Intergovernmental Panel on Climate Change.* IPCC,
652 Geneva, Switzerland.

653 Jaffe, D., Anderson, T., Covert, D., Kotchenruther, R., Trost, B., Danielson, J.,
654 Simpson, W., Berntsen, T., Karlsdottir, S., Blake, D., Harris, J., Carmichael, G.,
655 Uno, I., 1999. Transport of Asian air pollution to North America. *Geophys. Res.*
656 *Lett.* 26, 711–714.

657 Jaffe, D., Snow, J., Cooper, O., 2003. The 2001 Asian dust events: Transport and
658 impact on surface aerosol concentrations in the U.S. *Eos* 84, 501–507.

659 Jaffe, D., 2004. Long-range transport of Siberian biomass burning emissions and
660 impact on surface ozone in western North America. *Geophys. Res. Lett.* 31,
661 L16106.

662 Jaffe, D., Chand, D., Hafner, W., Westerling, A., Spracklen, D., 2008. Influence of
663 fires on O₃ concentrations in the western U.S. *Environ. Sci. Technol.* 42, 5885–
664 5891.

665 Jaffe, D.A., Wigder, N.L., 2012. Ozone production from wildfires: A critical review.
666 *Atmos. Environ.* 51, 1–10.

667 Jaffe, D.A., Wigder, N., Downey, N., Pfister, G., Boynard, A., Reid, S.B., 2013.
668 Impact of wildfires on ozone exceptional events in the western U.S. *Environ. Sci.*
669 *Technol.* 47, 11065–11072.

670 Jaffe, D.A., Zhang, L., 2017. Meteorological anomalies lead to elevated O₃ in the

671
672 Kaiser, J.W., Heil, A., Andreae, M.O., Benedetti, A., Chubarova, N., Jones, L.,
673 Morcrette, J.J., Razinger, M., Schultz, M.G., Suttie, M., van der Werf, G.R., 2012.
674 Biomass burning emissions estimated with a global fire assimilation system
675 based on observed fire radiative power. *Biogeosciences* 9, 527–554.
676 Laing, J.R., Jaffe, D.A., Hee, J.R., 2016. Physical and optical properties of aged
677 biomass burning aerosol from wildfires in Siberia and the Western USA at the
678 Mt. Bachelor Observatory. *Atmos. Chem. Phys.* 16, 15185–15197.
679 Langford, A.O., Brioude, J., Cooper, O.R., Senff, C.J., Alvarez, R.J., Hardesty, R.M.,
680 Johnson, B.J., Oltmans, S.J., 2012. Stratospheric influence on surface ozone in
681 the Los Angeles area during late spring and early summer of 2010. *J. Geophys.*
682 *Res. Atmos.* 117, D00V06.
683 Langford, A.O., Alvarez, R.J., Brioude, J., Fine, R., Gustin, M.S., Lin, M.Y.,
684 Marchbanks, R.D., Pierce, R.B., Sandberg, S.P., Senff, C.J., Weickmann, A.M.,
685 Williams, E.J., 2017. Entrainment of stratospheric air and Asian pollution by the
686 convective boundary layer in the southwestern U.S. *J. Geophys. Res. Atmos.* 122,
687 1312–1337.
688 Lin, M., Fiore, A.M., Cooper, O.R., Horowitz, L.W., Langford, A.O., Levy, H.,
689 Johnson, B.J., Naik, V., Oltmans, S.J., Senff, C.J., 2012. Springtime high surface
690 ozone events over the western United States: Quantifying the role of
691 stratospheric intrusions. *J. Geophys. Res.* 117, D00V22,
692 doi:10.1029/2012JD018151.
693 Lin, M., Horowitz, L.W., Cooper, O.R., Tarasick, D., Conley, S., Iraci, L.T., Johnson,
694 B., Leblanc, T., Petropavlovskikh, I., Yates, E.L., 2015. Revisiting the evidence
695 of increasing springtime ozone mixing ratios in the free troposphere over western
696 North America. *Geophys. Res. Lett.* 42, 8719–8728.
697 McClure, C.D., Jaffe, D.A., Gao, H., 2016. Carbon dioxide in the free troposphere and
698 boundary layer at the Mt. Bachelor Observatory. *Aerosol Air Qual. Res.* 16, 717–
699 728.
700 McDonald-Buller, E.C., Allen, D.T., Brown, N., Jacob, D.J., Jaffe, D., Kolb, C.E.,

701
702 Establishing policy relevant background (PRB) ozone concentrations in the
703 United States. *Environ. Sci. Technol.* 45, 9484–9497.

704 Paris, J.D., Stohl, A., Nedelec, P., Arshinov, M.Y., Panchenko, M.V., Shmargunov, V.P.,
705 Law, K.S., Belan, B.D., Ciais, P., 2009. Wildfire smoke in the Siberian arctic in
706 summer: source characterization and plume evolution from airborne
707 measurements. *Atmos. Chem. Phys.* 9, 9315–9327.

708 Pfister, G.G., Emmons, L.K., Hess, P.G., Honrath, R., Lamarque, J.F., Martin, M.V.,
709 Owen, R.C., Avery, M.A., Browell, E.V., Holloway, J.S., Nedelec, P., Purvis, R.,
710 Ryerson, T.B., Sachse, G.W., Schlager, H., 2006. Ozone production from the
711 2004 North American boreal fires. *J. Geophys. Res.* 111, D24S07,
712 doi:10.1029/2006JD007695.

713 Real, E., Law, K.S., Weinzierl, B., Fiebig, M., Petzold, A., Wild, O., Methven, J.,
714 Arnold, S., Stohl, A., Huntrieser, H., Roiger, A., Schlager, H., Stewart, D., Avery,
715 M., Sachse, G., Browell, E., Ferrare, R., Blake, D., 2007. Processes influencing
716 ozone levels in Alaskan forest fire plumes during long-range transport over the
717 North Atlantic. *J. Geophys. Res.* 112, D10S41, doi:10.1029/2006JD007576.

718 Reidmiller, D.R., Jaffe, D.A., Fischer, E.V., Finley, B., 2010. Nitrogen oxides in the
719 boundary layer and free troposphere at the Mt. Bachelor Observatory. *Atmos.*
720 *Chem. Phys.* 10, 6043–6062.

721 Singh, H.B., Anderson, B.E., Brune, W.H., Cai, C., Cohen, R.C., Crawford, J.H.,
722 Cubison, M.J., Czech, E.P., Emmons, L., Fuelberg, H.E., Huey, G., Jacob, D.J.,
723 Jimenez, J.L., Kaduwela, A., Kondo, Y., Mao, J., Olson, J.R., Sachse, G.W., Vay,
724 S.A., Weinheimer, A., Wennberg, P.O., Wisthaler, A., 2010. Pollution influences
725 on atmospheric composition and chemistry at high northern latitudes: boreal and
726 California forest fire emissions. *Atmos. Environ.* 44, 4553–4564.

727 Stauffer, R.M., Thompson, A.M., Oltmans, S.J., Johnson, B.J., 2017. Tropospheric
728 ozonesonde profiles at long-term U.S. monitoring sites: 2. Links between
729 Trinidad Head, CA, profile clusters and inland surface ozone measurements. *J.*
730 *Geophys. Res. Atmos.* 122, 1261–1280.

731

732 Significant enhancements of nitrogen oxides, black carbon, and ozone in the
733 North Atlantic lower free troposphere resulting from North American boreal
734 wildfires. *J. Geophys. Res.* 111, D23S60, doi:10.1029/2006JD007530.

735 Virkkula, A., 2010. Correction of the calibration of the 3-wavelength Particle Soot
736 Absorption Photometer (3l PSAP). *Aerosol Sci. Technol.* 44, 706–712.

737 Weiss-Penzias, P., Jaffe, D.A., Swartzendruber, P., Dennison, J.B., Chand, D., Hafner,
738 W., Prestbo, E., 2006. Observations of Asian air pollution in the free troposphere
739 at Mount Bachelor Observatory during the spring of 2004. *J. Geophys. Res.* 111,
740 D10304.

741 Wigder, N.L., Jaffe, D.A., Saketa, F.A., 2013. Ozone and particulate matter
742 enhancements from regional wildfires observed at Mount Bachelor during
743 2004-2011. *Atmos. Environ.* 75, 24–31.

744 Yang, M., Howell, S.G., Zhuang, J., Huebert, B.J., 2009. Attribution of aerosol light
745 absorption to black carbon, brown carbon, and dust in China – interpretations of
746 atmospheric measurements during EAST-AIRE. *Atmos. Chem. Phys.* 9, 2035–
747 2050.

748 Zhang, L., Jacob, D.J., Kopacz, M., Henze, D.K., Singh, K., Jaffe, D.A., 2009.
749 Intercontinental source attribution of ozone pollution at western U.S. sites using
750 an adjoint method. *Geophys. Res. Lett.* 36, L11810.

751

752 **Table captions**

753 **Table 1** Water vapor (WV) criteria for free tropospheric/boundary layer influenced
754 (FT/BLI) air masses at MBO and average FT and BLI WV by month. The WV criteria
755 are the cut-off values that make the MBO mean WV similar to the radiosonde data.

756 **Table 2** Means and maximum values of hourly air pollutant concentrations for six
757 typical plume events.

758 **Table 3** Enhancement ratios (mean \pm standard deviation) and aerosol Ångström
759 exponents of different types of plume events based on hourly data at MBO.

760

761 **Table 1**

762 (FT/BLI) air masses at MBO and average FT and BLI WV by month. The WV criteria
763 are the cut-off values that make the MBO mean WV similar to the radiosonde data.

	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec
WV Criteria (g kg ⁻¹)	3.26	2.64	2.46	2.55	3.06	4.25	5.14	5.23	4.60	4.36	3.44	2.97
Average FT WV (g kg ⁻¹)	2.07	1.76	1.85	1.98	2.30	2.94	3.36	3.66	3.05	2.75	2.20	1.96
Average BLI WV (g kg ⁻¹)	3.99	3.46	3.41	3.65	4.45	5.58	6.58	6.41	5.72	5.28	4.42	3.82

764

765 **Table 2** Means and maximum values of hourly air pollutant concentrations for six typical plume events.

Event No.	Event time	Category ^a	σ_{sp} (Mm ⁻¹)		O ₃ (ppbv)		CO (ppbv)		NO _x (pptv)		PAN (pptv)		NO _y (pptv)		WV (g kg ⁻¹)	
			Mean	Max	Mean	Max	Mean	Max	Mean	Max	Mean	Max	Mean	Max	Mean	Max
138	2013/7/21 4:00 – 20:00	UTLS	0.9	1.8	74.3	86.3	84	94	148	171	50	69	–	–	1.20	1.97
151	2014/4/19 2:00 – 18:00	ALRT-WF	26.4	39.7	67.1	79.9	192	226	–	–	–	–	–	–	2.45	3.44
148	2013/11/20 21:00 – 2013/11/21 11:00	ALRT-IP	16.6	28.6	53.3	57.0	215	260	–	–	–	–	–	–	1.53	2.04
102	2011/5/1 3:00 – 15:00	ALRT-MD	4.8	6.0	70.0	77.5	129	134	–	–	–	–	–	–	1.73	3.53
123	2012/8/25 11:00 – 2012/8/26 22:00	PNW-WF	100.6	406.7	60.7	69.4	296	707	283	577	–	–	1418	4648	4.62	5.59
53	2007/5/25 7:00 – 18:00	PNW-IP	3.0	4.3	64.0	76.5	124	130	85	133	–	–	–	–	3.52	5.29

766 ^a UTLS: upper troposphere and lower stratosphere O₃ intrusion; ALRT: Asian long-range transport; PNW: Pacific Northwest pollution; WF: wildfires; IP: industrial pollution; MD:
 767 mineral dust.

768

Table 3 Enhancement ratios (mean \pm standard deviation) and aerosol Ångström exponents of different types of plume events based on hourly data at MBO.

Category ^a	Trigger pollutants	$\Delta O_3/\Delta WV$ (ppbv kg g ⁻¹)	$\Delta \sigma_{sp}/\Delta CO$ (Mm ⁻¹ ppbv ⁻¹)	$\Delta \sigma_{sp}/\Delta O_3$ (Mm ⁻¹ ppbv ⁻¹)	$\Delta \sigma_{sp}/\Delta NO_y$ (Mm ⁻¹ pptv ⁻¹)	$\Delta O_3/\Delta CO$ (ppbv ppbv ⁻¹)	$\Delta NO_y/\Delta CO$ (pptv ppbv ⁻¹)	$\Delta NO_x/\Delta PAN$ (pptv pptv ⁻¹)	Å_{sp}	Å_{ap}
UTLS	O ₃	-10.05±3.26 (23/23) ^b	0.17±0.12 (9/19)	-0.14±0.12 (12/23)	0.04 (1/3)	0.04±1.53 (4/19)	10.08±1.80 (2/3)	0.96±0.64 (4/5)	2.66±0.88 (10)	(0)
UTLS/ALRT	O ₃	-11.06±5.01 (20/20)	0.32±0.36 (9/18)	-0.07±0.20 (5/18)	0.01 (1/5)	0.17±3.17 (5/20)	-0.10±9.52 (2/6)	0.68 (1/2)	2.15±0.18 (9)	0.91±0.74 (4)
ALRT-WF	σ_{sp} , CO	-10.53±3.85 (10/11)	0.42±0.20 (4/9)	0.48±0.71 (5/11)	(0/4)	0.52±0.64 (7/9)	-5.67±21.38 (3/3)	(0/0)	2.19±0.16 (4)	1.45±0.02 (2)
ALRT-WF/IP	σ_{sp} , CO	-10.61±3.66 (6/9)	0.23±0.13 (3/8)	0.66±0.51 (5/8)	0.05±0.03 (2/5)	0.01±0.90 (3/9)	8.14±6.90 (4/5)	(0/1)	2.23±0.30 (4)	1.54±0.39 (4)
ALRT-IP	CO	-7.88±6.26 (15/22)	0.20±0.18 (11/17)	0.51±0.86 (9/18)	-0.12 (1/5)	0.21±0.50 (13/21)	7.69 (1/4)	(0/0)	2.02±0.20 (6)	(0)
ALRT-MD	σ_{sp}	-13.94±6.81 (3/3)	-0.33 (1/3)	0.18±0.01 (2/3)	(0/0)	-1.19±0.60 (3/3)	(0/0)	(0/1)	2.07±0.05 (2)	(0)
AAP	CO	-5.85±3.40 (6/6)	0.13±0.15 (4/6)	1.39±2.40 (2/6)	0.05±0.04 (2/5)	-0.28±0.60 (3/6)	2.72±0.26 (2/5)	(0/0)	(0)	(0)
PNW-WF	σ_{sp} , CO	-5.27±5.99 (42/73)	0.59±0.17 (58/63)	10.98±13.16 (30/70)	0.10±0.07 (19/25)	0.10±0.12 (20/61)	5.83±3.11 (19/25)	0.35±0.25 (6/11)	2.16±0.20 (33)	1.81±0.59 (5)
PNW-IP	CO	-8.57±0.68 (2/3)	(0/1)	1.92 (1/3)	(0/0)	(0/1)	(0/0)	(0/0)	2.25 (1)	1.86 (1)

^a UTLS: upper troposphere and lower stratosphere O₃ intrusion; ALRT: Asian long-range transport; AAP: Arctic air pollution; PNW: Pacific Northwest pollution; WF: wildfires; IP: industrial pollution; MD: mineral dust.

^b The first value in parentheses gives the count of available data ($\Delta Y/\Delta X$) with considerable correlation between X and Y ($R^2 \geq 0.3$), and the second number gives the count of total available data ($\Delta Y/\Delta X$).

774 **Figure captions**

775 **Fig. 1.** Statistical distributions of the O₃ mixing ratio at the Mt. Bachelor Observatory
776 (MBO) by month. The bottom and top of the box represent the 25th and 75th
777 percentiles (the lower and upper quartiles, q₁ and q₃), respectively. The band near the
778 middle of the box represents the 50th percentile (the median). The dot near the middle
779 of the box represents the mean. The ends of the whiskers represent $q_3+1.5(q_3-q_1)$ and
780 $q_1-1.5(q_3-q_1)$, respectively. The dots are outliers.

781 **Fig. 2.** Trends and interannual variations of the O₃ mixing ratio by season. The points
782 show the mean values and the bars show the standard deviations.

783 **Fig. 3.** Statistical distributions of the sub-micron aerosol scattering coefficient at the
784 Mt. Bachelor Observatory (MBO). The bottom and top of the box represent the 25th
785 and 75th percentiles (the lower and upper quartiles, q₁ and q₃), respectively. The band
786 near the middle of the box represents the 50th percentile (the median). The dot near
787 the top of the box represents the mean. The ends of the whiskers represent $q_3+1.5(q_3-$
788 $q_1)$ and $q_1-1.5(q_3-q_1)$, respectively. The dots are outliers.

789 **Fig. 4.** Interannual variation of the aerosol scattering coefficient by season. The points
790 show the mean values and the bars show the standard deviations.

791 **Fig. 5.** Means of the O₃ mixing ratio for MBO free tropospheric air (MBO-FT), MBO
792 boundary layer influenced air (MBO-BLI), MBO overall (MBO) and Trinidad Head
793 (THD) at 720–740 mbar.

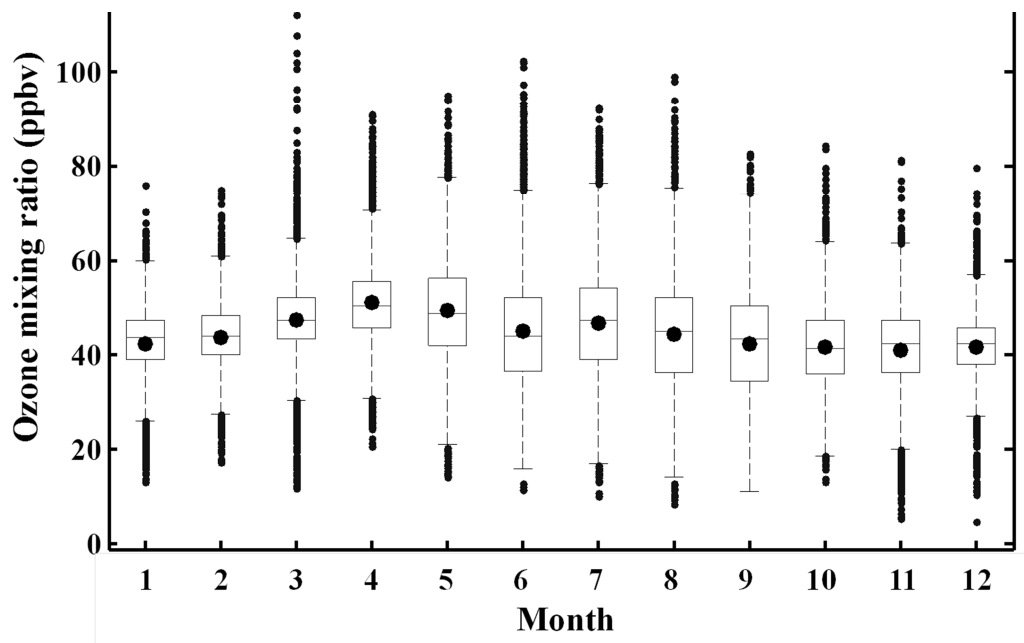
794 **Fig. 6.** Means of the sub-micron aerosol scattering coefficient for free tropospheric
795 (FT) air, boundary layer influenced (BLI) air and overall at MBO.

796 **Fig. 7.** Water vapor-ozone (WV-O₃) relationships for (a) the Trinidad Head (THD)
797 sounding data (1997–2017) at different elevations, and (b) different categories of
798 events at MBO (2004–2015), including upper troposphere and lower stratosphere O₃
799 intrusion (UTLS), Asian long-range transport events (ALRT), the combination of
800 UTLS and ALRT, Arctic air pollution events (AAP), and Pacific Northwest pollution
801 events (PNW).

802 **Fig. 8.**
803 trajectories of (a) an Asian long-range transport wildfire (ALRT-WF) event (Event 36,
804 9 May 2006) and (b) an Asian long-range transport industrial pollution (ALRT-IP)
805 event (Event 148, 20–21 November 2013). Shaded grids are influential grids where
806 trajectories are at low altitude (<3000 m agl); the numbers above and below the line in
807 shaded grids are CO emission rates of wildfire and industry, respectively (unit: kg s⁻¹);
808 The W/I ratio is the ratio of wildfire/industrial CO emissions.

809 **Fig. 9.** Distribution of different types of plume events during 2004–2015 by month.
810 (PNW: Pacific Northwest pollution; IP: industrial pollution; WF: wildfire; MD:
811 mineral dust; AAP: Arctic air pollution; ALRT: Asian long-range transport; UTLS:
812 upper troposphere and lower stratosphere O₃ intrusion)

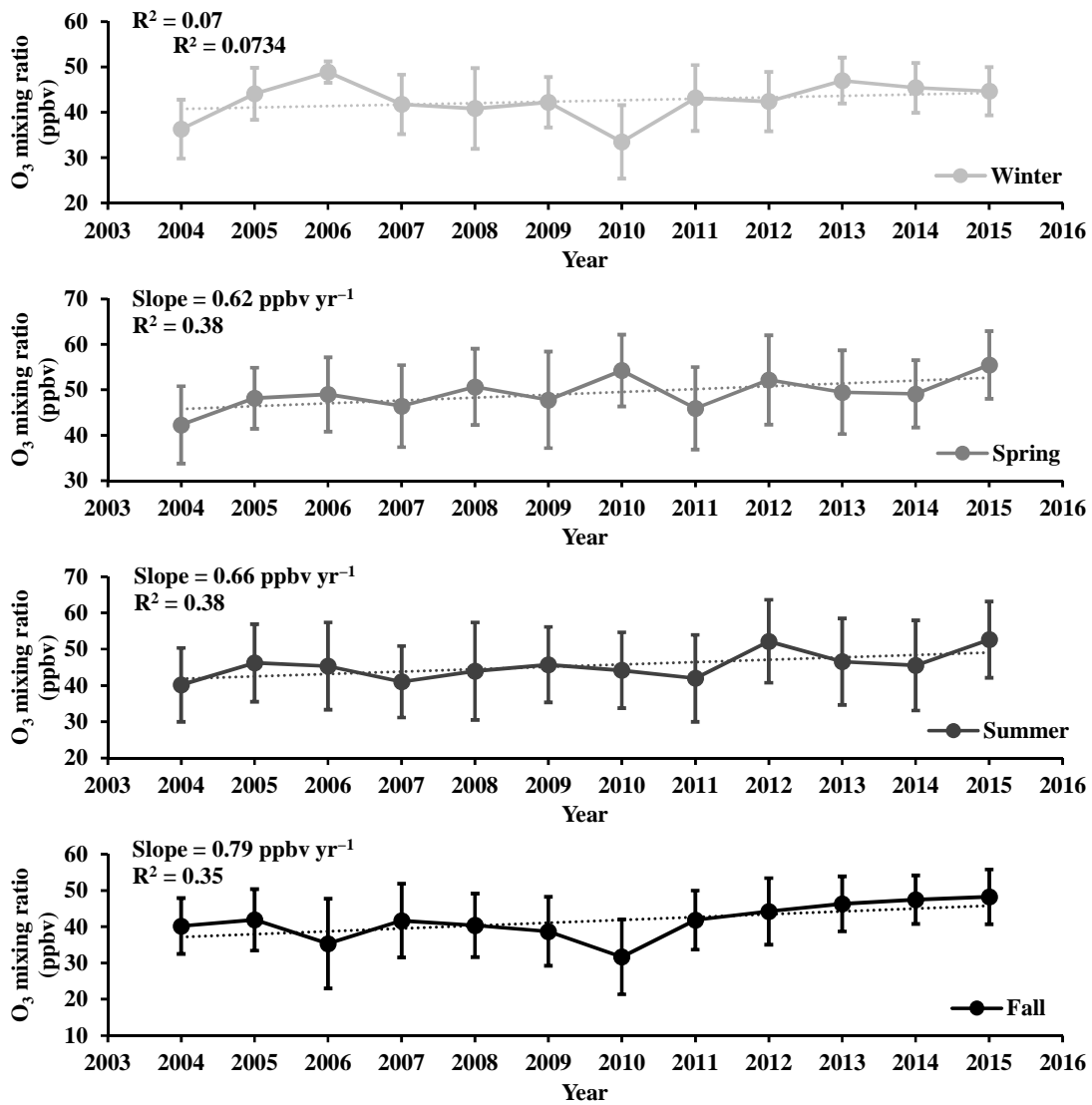
813 **Fig. 10.** Backward trajectories for six typical plume events at MBO: (a) an upper
814 troposphere and lower stratosphere O₃ intrusion (UTLS) event; (b) an Asian
815 long-range transport wildfire (ALRT-WF) event; (c) an Asian long-range transport
816 industrial pollution (ALRT-IP) event; (d) an Asian long-range transport mineral dust
817 (ALRT-MD) event; (e) a Pacific Northwest wildfire (PNW-WF) event; (f) a Pacific
818 Northwest industrial pollution (PNW-IP) event.
819



820

821 **Fig. 1.** Statistical distributions of the O₃ mixing ratio at the Mt. Bachelor Observatory
 822 (MBO) by month. The bottom and top of the box represent the 25th and 75th
 823 percentiles (the lower and upper quartiles, q₁ and q₃), respectively. The band near the
 824 middle of the box represents the 50th percentile (the median). The dot near the middle
 825 of the box represents the mean. The ends of the whiskers represent $q_3 + 1.5(q_3 - q_1)$ and
 826 $q_1 - 1.5(q_3 - q_1)$, respectively. The dots are outliers.

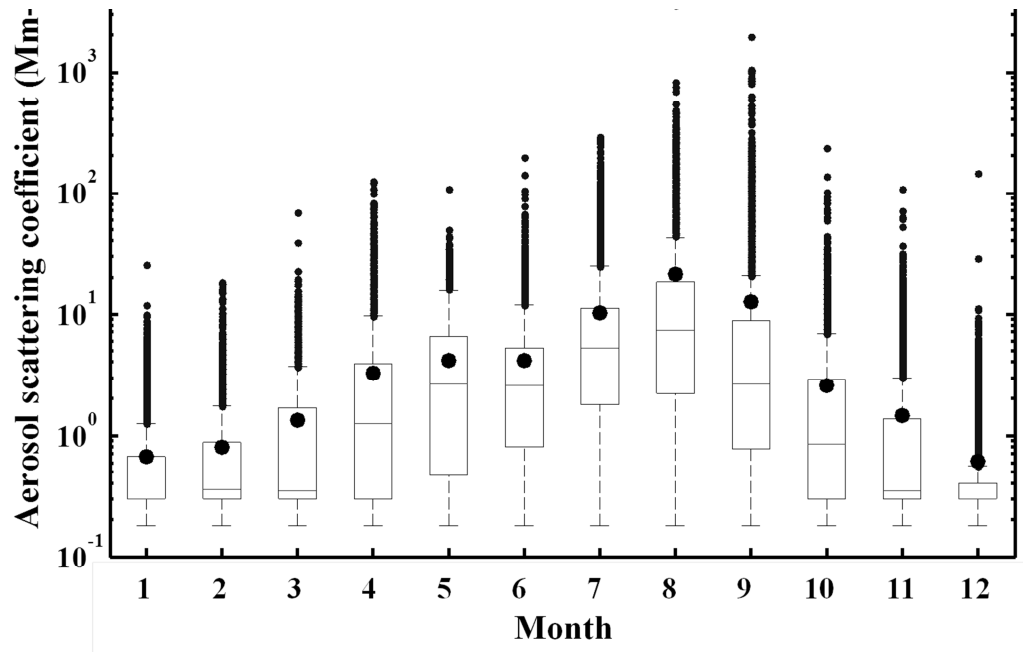
827



828

829 **Fig. 2.** Trends and interannual variations of the O₃ mixing ratio by season. The points
 830 show the mean values and the bars show the standard deviations.

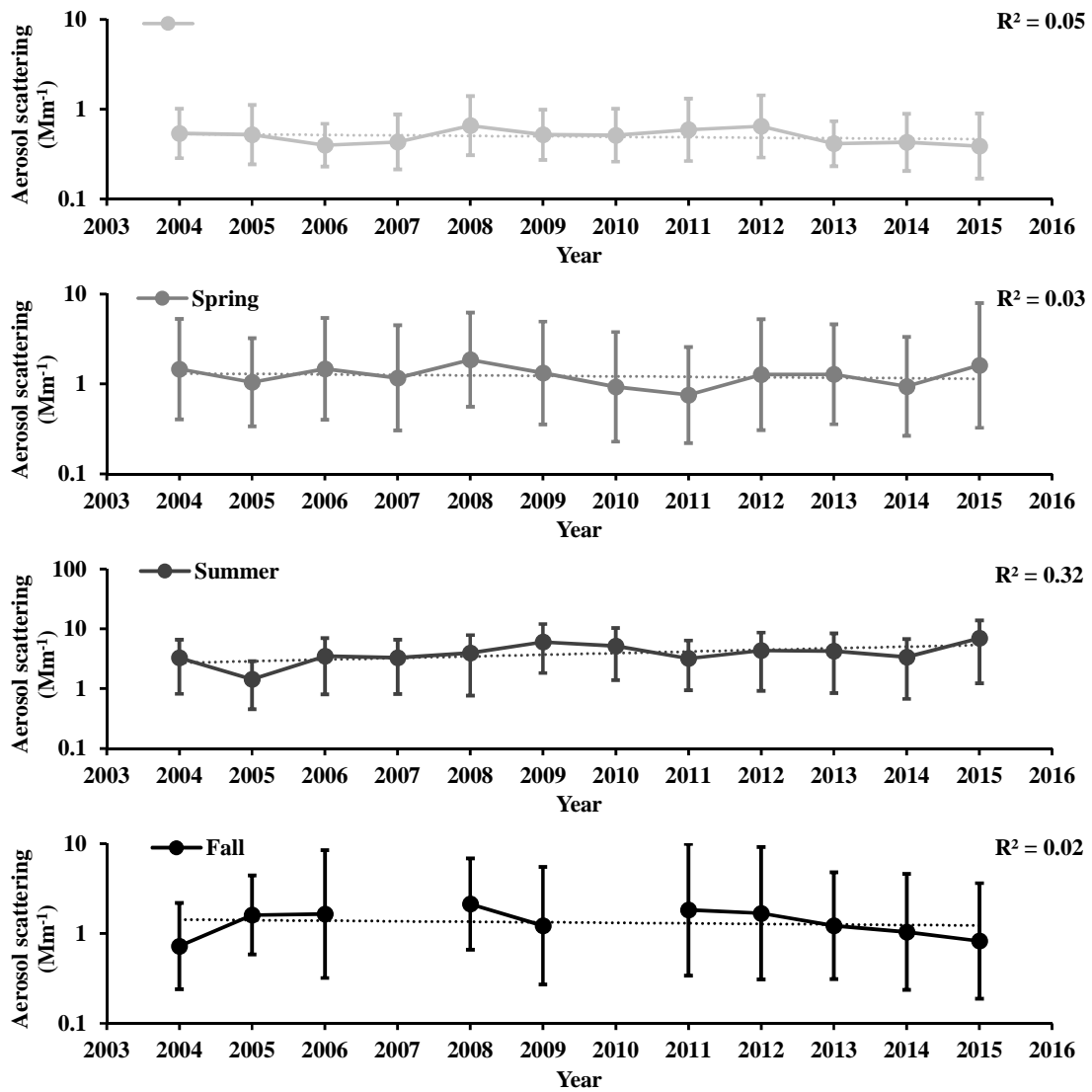
831



832

833 **Fig. 3.** Statistical distributions of the sub-micron aerosol scattering coefficient at the
 834 Mt. Bachelor Observatory (MBO). The bottom and top of the box represent the 25th
 835 and 75th percentiles (the lower and upper quartiles, q_1 and q_3), respectively. The band
 836 near the middle of the box represents the 50th percentile (the median). The dot near
 837 the top of the box represents the mean. The ends of the whiskers represent $q_3+1.5(q_3-q_1)$
 838 and $q_1-1.5(q_3-q_1)$, respectively. The dots are outliers.

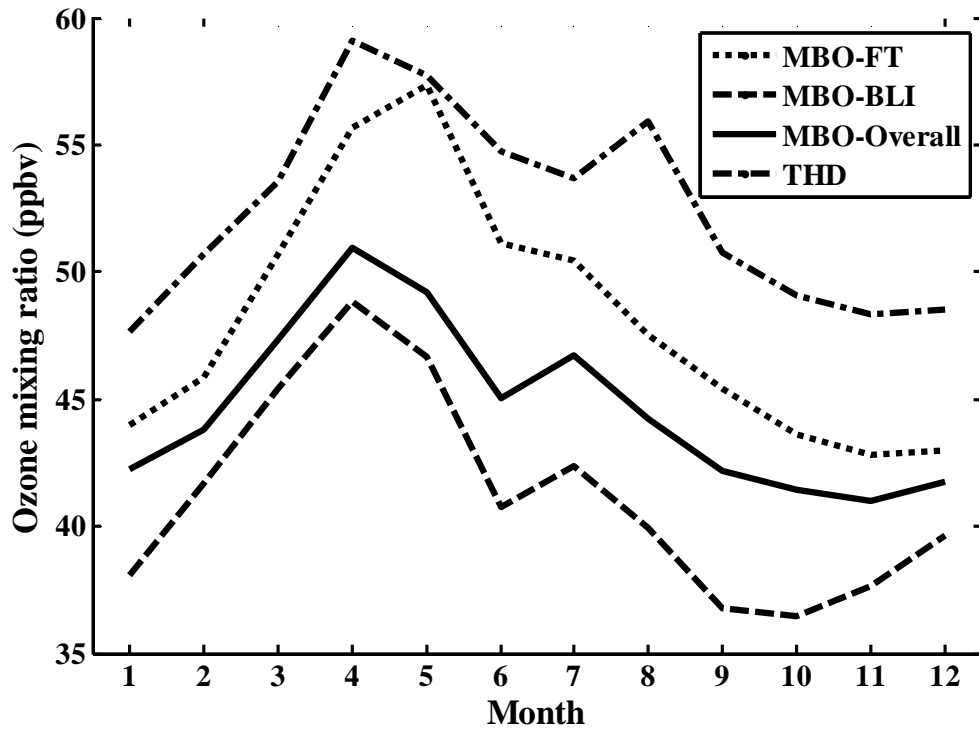
839



840

841 **Fig. 4.** Interannual variation of the aerosol scattering coefficient by season. The points
 842 show the mean values and the bars show the standard deviations.

843



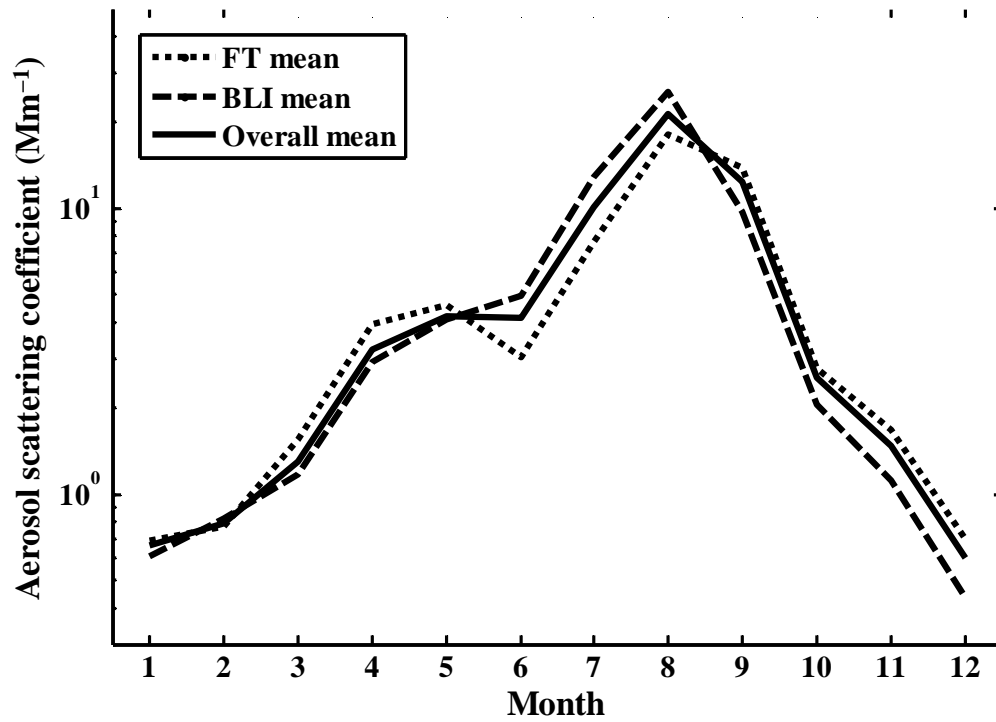
844

845 **Fig. 5.** Means of the O₃ mixing ratio for MBO free tropospheric air (MBO-FT), MBO

846 boundary layer influenced air (MBO-BLI), MBO overall (MBO) and Trinidad Head

847 (THD) at 720–740 mbar.

848

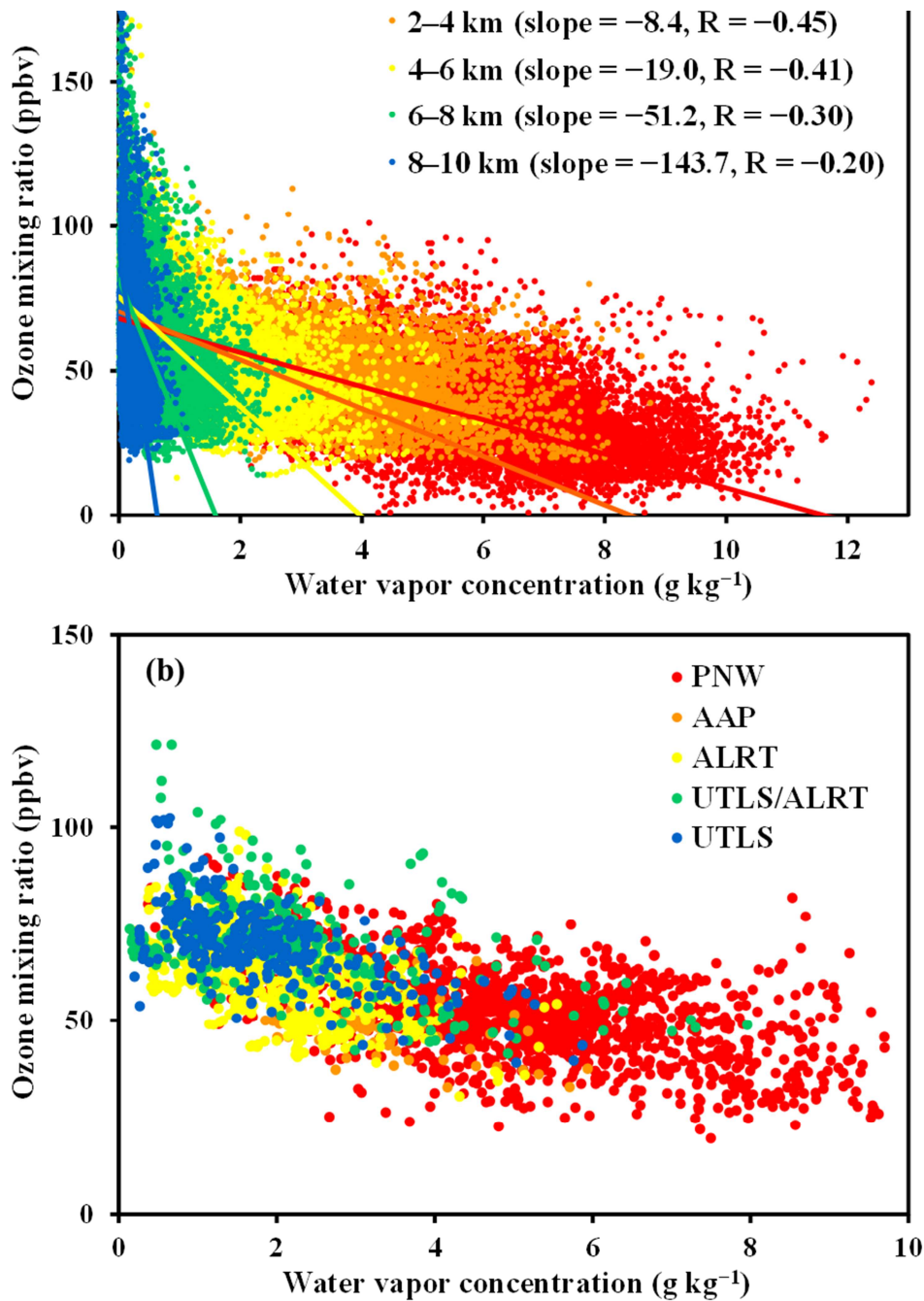


849

850 **Fig. 6.** Means of the sub-micron aerosol scattering coefficient for free tropospheric

851 (FT) air, boundary layer influenced (BLI) air and overall at MBO.

852



853

854 **Fig. 7.** Water vapor-ozone (WV- O_3) relationships for (a) the Trinidad Head (THD)

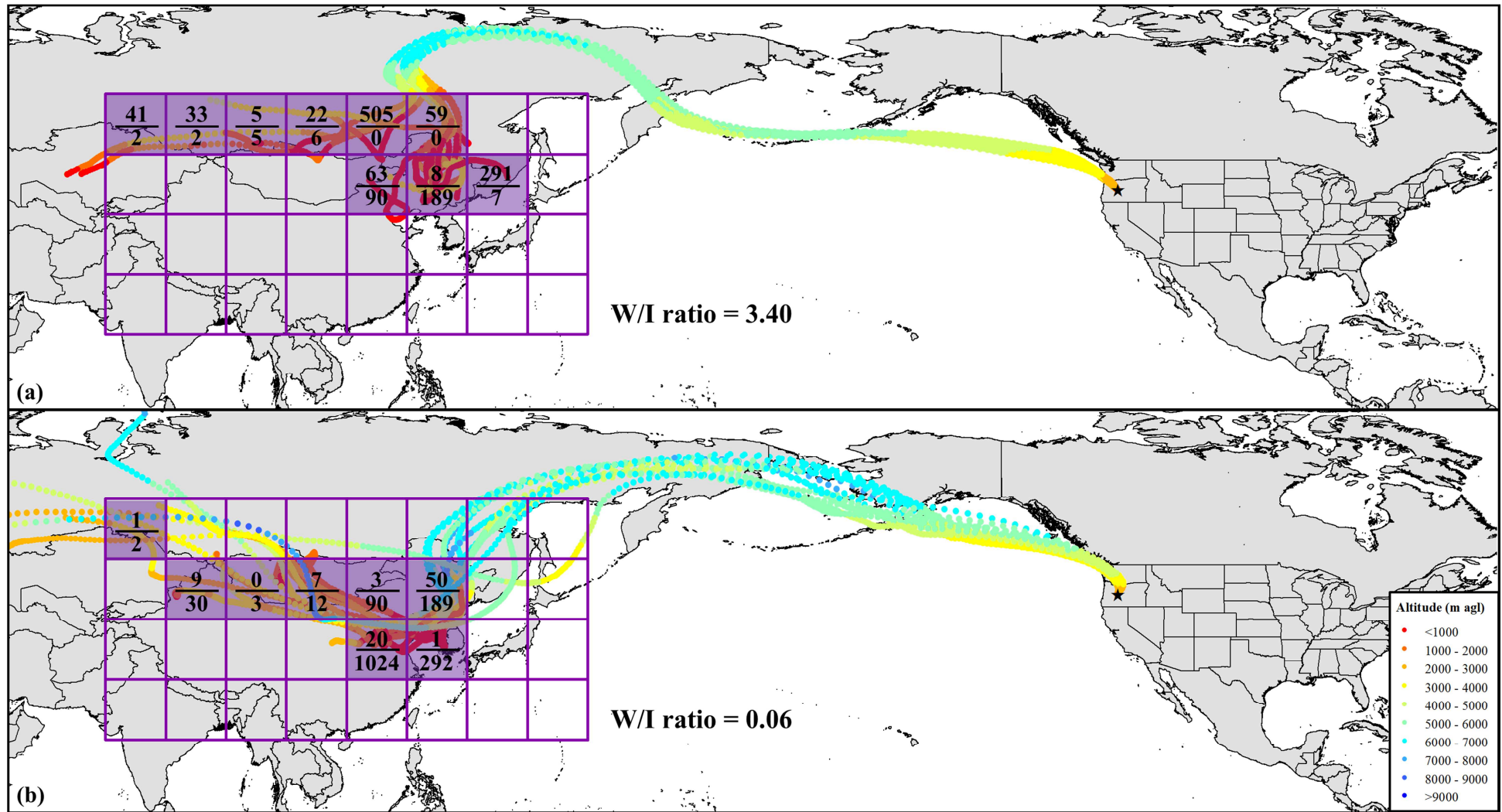
855 sounding data (1997–2017) at different elevations, and (b) different categories of

856 events at MBO (2004–2015), including upper troposphere and lower stratosphere O_3

857 intrusion (UTLS), Asian long-range transport events (ALRT), the combination of

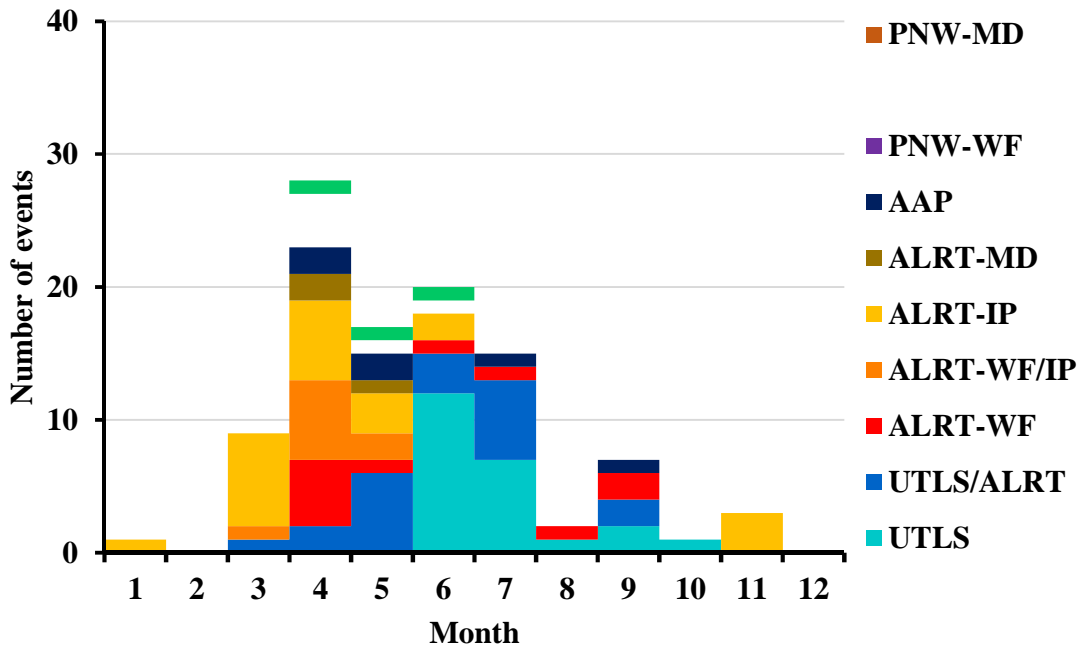
858 UTLS and ALRT, Arctic air pollution events (AAP), and Pacific Northwest pollution

859 events (PNW).



860

861 **Fig. 8.** Gridded East Asian carbon monoxide (CO) emission rates and HYSPLIT trajectories of (a) an Asian long-range transport wildfire (ALRT-WF)
 862 event (Event 36, 9 May 2006) and (b) an Asian long-range transport industrial pollution (ALRT-IP) event (Event 148, 20–21 November 2013). Shaded
 863 grids are influential grids where trajectories are at low altitude (<3000 m agl); the numbers above and below the line in shaded grids are CO emission
 864 rates of wildfire and industry, respectively (unit: kg s^{-1}); The W/I ratio is the ratio of wildfire/industrial CO emissions.



865

866

Fig. 9. Distribution of different types of plume events during 2004–2015 by month. (PNW:

867

Pacific Northwest pollution; IP: industrial pollution; WF: wildfire; MD: mineral dust; AAP:

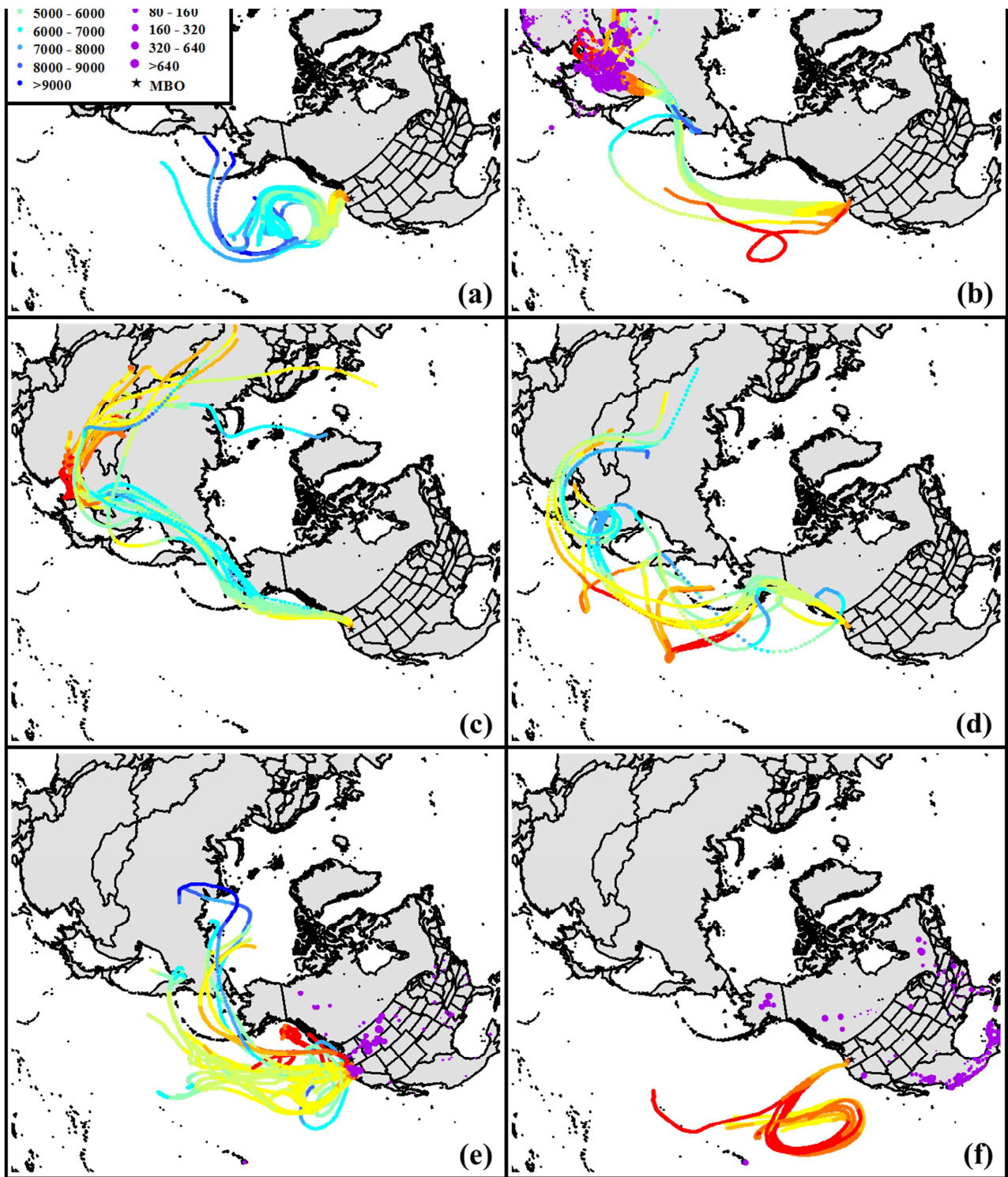
868

Arctic air pollution; ALRT: Asian long-range transport; UTLS: upper troposphere and lower

869

stratosphere O₃ intrusion)

870



871

872 **Fig. 10.** Backward trajectories for six typical plume events at MBO: (a) an upper troposphere and
 873 lower stratosphere O₃ intrusion (UTLS) event; (b) an Asian long-range transport wildfire
 874 (ALRT-WF) event; (c) an Asian long-range transport industrial pollution (ALRT-IP) event; (d) an
 875 Asian long-range transport mineral dust (ALRT-MD) event; (e) a Pacific Northwest wildfire
 876 (PNW-WF) event; (f) a Pacific Northwest industrial pollution (PNW-IP) event.

1. Long-range transport and UTLS enhance tropospheric O₃ in western US in spring.
2. Interannual aerosol increase in summer is driven by regional and Siberian wildfires.
3. CO emission rate helps identify long-range transport wildfire and industrial events.
4. Enhancement ratios and aerosol optical properties provide plume event signatures.