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Article type : Articles

Forecasting the legacy of offshore oil and gas platforms on fish community structure and productivity

Running head: Offshore platform community forecasts

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ABSTRACT

There are currently thousands of offshore platforms in place for oil and gas extraction worldwide, and decommissioning efforts over the next 3 decades are estimated to cost more than \$200 billion USD. As platforms reach the end of their useful lifetime, operators and regulatory agencies will assess the environmental impact of potential decommissioning strategies. Among the many factors that will be weighed in preparation for these major economic and engineering

This is the author manuscript accepted for publication and has undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the [Version of Record](#). Please cite this article as [doi: 10.1002/EAP.2185](https://doi.org/10.1002/EAP.2185)

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30 challenges is the fate of the fish and invertebrate communities that inhabit the structures
31 underwater. Offshore platforms act as inadvertent artificial reefs, and some are recognized
32 among the most productive fish habitats in the global oceans. We present a model for forecasting
33 changes to fish communities surrounding offshore installations following a series of
34 decommissioning alternatives. Using 24 platforms off southern California, we estimate fish
35 biomass and somatic production under three possible decommissioning scenarios: leave in place,
36 partial removal at 26 m depth, and complete removal of the platform and underlying shell
37 mound. We used fish density and size data from scuba and submersible surveys of the platforms
38 from 1995–2013 to estimate biomass and annual somatic production. Bottom trawl surveys were
39 used to characterize future fish assemblages at platform sites under the complete removal
40 decommissioning scenario. Based on a conservatively modeled extrapolation of the survey data,
41 we found that complete removal of a platform resulted in 95% or more reduction in the average
42 fish biomass and annual somatic production at the site, while partial removal resulted in far
43 smaller losses, averaging 10% or less. In the event that all surveyed platforms are completely
44 removed, we estimated a total loss of more than 28,000 kg of fish biomass in the Southern
45 California Bight. Platform habitats, which attract reef-dwelling fish species, had minimal overlap
46 in community composition with the surrounding soft bottom habitat. To best serve the wide
47 range of stakeholder interests, the site-specific biomass, productivity and species composition
48 information provided in this study should be incorporated into strategic decommissioning
49 planning. This approach could be used as a model for informing “rigs to reefs” discussions
50 occurring worldwide.

51
52 Keywords: Offshore platform, Ecological forecasting, Artificial reef, Rigs to Reefs,
53 Decommissioning, Decision-making, Sebastes

54

55

56 INTRODUCTION

57

58 Offshore drilling for oil and gas is a major human use of the oceans, and thousands of oil
59 and gas platforms have been installed on the continental shelves worldwide (Parente et al. 2006).
60 As the underlying hydrocarbon reservoirs are depleted and equipment ages, platform drilling

61 operations become less profitable and are shut down. In this decommissioning process, wells are
62 capped, infrastructure is removed and the seafloor may be cleared of any obstructions (Schroeder
63 & Love 2004). More than 2,600 oil and gas installations are projected to be decommissioned
64 within the next few decades, with costs worldwide from 2010–2040 estimated to reach \$210
65 billion USD (IHS Markit 2016). Environmental impacts associated with platform demolition and
66 removal include a substantial carbon footprint, waste generation and potential release of
67 contaminants (Schroeder & Love 2004, Fowler et al. 2014). The underwater structures also are
68 habitats for marine life, the majority of which is destroyed in the removal process
69 (Basavalinganadoddi & Mount 2004).

70 The “Rigs to Reefs” decommissioning alternative, where all or a portion of the platform
71 is left in the ocean to continue its life as an artificial reef, has been proposed to reduce some of
72 these economic and environmental costs (Reggio 1987, Schroeder & Love 2004, Macreadie
73 2011, Bull & Love 2019). This approach also has drawbacks, however, such as reducing
74 workable area for commercial fishermen (de Wit 2001), and creating ongoing liability for the
75 agency managing the artificial reef (Jagerroos and Krause 2016). Public sentiment on artificial
76 reef conversion is mixed, with some groups preferring sites to be returned to a pristine state (e.g.
77 Jørgensen 2012, Olsen 2016). This decision is multi-faceted, and stakeholders have different
78 environmental, economic and political goals. Ultimately, the decision to pursue an alternative to
79 total structure removal is made by the presiding regulatory agency and the platform operator, in
80 accordance with current laws and international agreements (Osmundsen & Tveteras 2003,
81 Parente et al. 2006, Fowler et al. 2018). Complete removal continues to be the default
82 decommissioning mandate in the USA (Bull & Love 2019), Australia (Techera & Chandler
83 2015), and Europe (Fowler et al. 2018). However, a panel of environmental experts concluded
84 that platform decommissioning decisions should be made on a case-by-case basis, accounting for
85 variations in platform structure, local environment, and ecology, to achieve the best possible
86 environmental outcomes (Fowler et al. 2018).

87 There are currently 27 oil and gas platforms off the coast of southern California (Table
88 1). In part due to their significant vertical extent and position in offshore currents, these artificial
89 reefs are highly productive marine ecosystems (Page 1986, Page & Hubbard 1987, Claisse et al.
90 2014, Santora et al. 2017), and may contribute to the recovery of commercially important reef-
91 dwelling fish species (Love et al. 2005, Love et al. 2006). For example, larval production of two

92 overfished species, cowcod (*Sebastes levis*) and bocaccio (*Sebastes paucispinis*), is
93 proportionately higher on platforms than natural reefs due to the higher densities and larger sizes
94 of spawning adults (Love & Schroeder 2006, Claisse et al. 2019). Seven California platforms
95 were decommissioned in the 1980s and 1990s, with removal of most of the platform structure
96 (Manago & Williamson 1998, Bernstein 2010). The cost and negative impacts of these projects,
97 such as air quality issues, recycling and waste generation, and impacts on marine life, have
98 sparked interest in exploring decommissioning alternatives for the remaining platforms
99 (Schroeder & Love 2004, Macreadie et al. 2011, Pondella et al. 2015). This has become a
100 pressing discussion since several platforms are expected to go through the decommissioning
101 process in the coming years (Bureau of Safety and Environmental Enforcement 2018, 2019, Bull
102 and Love 2019).

103 Three main options are under consideration for decommissioning the remaining
104 California platforms: leave in place, partial removal or complete removal. In the leave-in-place
105 alternative, the wells are capped but the entire underwater portion of the platform is preserved as
106 an artificial reef. Partial removal involves the removal of all platform components near the
107 surface that present a navigational hazard to ships (Stephan 1990; Schroeder & Love 2004). In
108 the complete removal scenario, the platform is severed from the seafloor using explosives or
109 mechanical cutting, removed piecemeal and recycled, reused or discarded (Schroeder & Love
110 2004, Basavalinganadoddi & Mount 2004). The use of explosives and removal of the underwater
111 hard substrate results in the destruction of fishes and invertebrates associated with the structure
112 (Gitschlag et al. 2000, Schroeder & Love 2004), and can pose a hazard to marine mammals and
113 turtles (Klima et al. 1988). Due to the diverse and productive marine communities that can be
114 associated with platforms, California Assembly Bill 2503 states that any analysis evaluating
115 alternatives to complete structure removal should include the “contribution of the proposed
116 structure to protection and productivity of fish and other marine life” (California Marine
117 Resources Legacy Act 2010).

118 Here, we use fish community data from platforms and nearby soft-bottom habitats to
119 examine the ecological impacts of a range of decommissioning scenarios on California
120 platforms. We assessed fish biomass and somatic production on 24 platforms, and forecasted
121 changes to the fish community biomass and production on each platform under the three
122 alternative decommissioning scenarios: leave in place, partial removal to a depth of 26 m and

123 complete structure removal. Predictions of fish biomass and productivity on platforms under the
124 partial removal and leave-in-place scenarios were based on scuba diver and manned submersible
125 surveys conducted from 1995–2013. Only the outer part of each platform was surveyed, and we
126 estimated total platform-associated fish communities in three different ways, first using only the
127 survey data, and then scaling it up to the portions of the water column that were not surveyed,
128 using two alternative approaches.

129 To forecast the fish communities that would inhabit the platform footprint under the
130 complete removal scenario, we used bottom trawl surveys conducted near the platforms to
131 forecast the fish communities that would inhabit the platform footprint upon reversion to soft
132 bottom, building upon previous work that considered only the loss of platform habitat (i.e.
133 Pondella et al. 2015, Claisse et al. 2015). Consideration of both habitat loss and gain under the
134 three decommissioning scenarios reduces the risk of overestimating fish loss following structure
135 removal and presents a more accurate forecast of overall changes to fish abundance, biomass and
136 somatic production. Since oil and gas platforms provide considerable vertical habitat to
137 accommodate fish relative to the surrounding soft bottom, we predicted that structure removal
138 would generally result in an overall net loss of fish biomass and somatic production.

139

140 **METHODS**

141 *Fish surveys*

142 There are several underwater components of an offshore oil and gas platform that provide
143 structural habitat for fish (Schroeder and Love 2004). Each platform consists of a series of large-
144 diameter vertical conductor pipes that span the entire water column to transport the drills and the
145 fossil fuels between the seafloor and the surface. The jacket is a rectangular steel lattice of
146 smaller pipes that surround and support the conductors as well as the platform deck at the
147 surface. Bivalves that grow on the jacket and conductors fall to the seafloor below and form a
148 shell mound that adds additional structural habitat to the site.

149 Fishes around platforms were visually surveyed by scuba divers, manned submersible or
150 ROV around 24 southern California platforms from 1995–2013 (Fig. 1; Love et al. 2017). Belt
151 transects 2 m high and 2 m wide were conducted on the shell mound around the platform, along
152 the perimeter of the platform base, and along each major external horizontal jacket cross beam
153 throughout the midwaters. Fish were counted, identified to species when possible, and total

154 length was estimated to the nearest 5 cm increment. See Appendix S1 for additional detail on
155 survey protocol and effort around platforms.

156 Under the complete removal decommissioning scenario, all of the platform structure and
157 the shell mound will be removed, causing the habitat to revert back to soft bottom. Estimation of
158 fish biomass and somatic production in the complete removal scenario was based on data from
159 bottom trawl surveys of fish communities over surrounding soft substrate. Data from two surveys
160 were used: the NOAA West Coast Groundfish Bottom Trawl survey from 2003–2015 (West
161 Coast Groundfish Bottom Trawl Survey 2017, Keller et al. 2017) and the Southern California
162 Coastal Water Research Project (SCCWRP) Southern California Bight Regional Survey trawls
163 conducted in 2003, 2008 and 2013 (Southern California Bight Regional Survey 2017). Trawls
164 that occurred within a 30 km radius and at an average bottom depth within 25 m of the base
165 depth of an individual platform were associated with that site and used in the complete removal
166 fish community forecast. Trawl survey effort was variable, for example, SCCWRP trawls were
167 not conducted near the most northwestern platforms in the Point Conception region, and
168 WCGBT trawls were rare within a 25 m depth range of the shallowest platforms (Appendix S1:
169 Table S1).

170 The NOAA West Coast Groundfish Bottom Trawl surveys have been conducted
171 biannually with consistent methodologies since 2003 (Keller et al. 2008, 2017). An Aberdeen-
172 type trawl with 31.7 m footrope and cod end mesh size of 14 cm was used (Dickson 1993, Millar
173 1992). Minimum bottom depth was 55 m, limiting the usefulness of this survey for the
174 shallowest platforms. Fish were sorted into taxonomic groups and weighed together. Individual
175 lengths were only recorded for a subsample of the catch. SCCWRP bottom trawl surveys used
176 semi-balloon otter trawls with an 8.8 m footrope and cod end mesh size of 1.25 cm (Allen et al.
177 2011, SCCWRP 2008) and were conducted nearshore with high sampling frequency near the
178 shallowest platforms. Length of all fish was measured to the nearest cm.

179

180 *Abundance, Biomass and Somatic Production Metrics*

181 Fish abundance, biomass and somatic production were calculated and compared for both
182 platform visual surveys and benthic trawl surveys. Comparisons of fish community composition
183 between habitat types are presented at the genus level for ease of interpretation. In platform
184 surveys, the biomass of individual fish was estimated using length-weight relationships from the

185 literature (see Claisse et al. 2014). In trawl surveys, fish were weighed directly. Somatic
186 production was defined as the expected annual mass increase of individual fish that were
187 estimated to survive for at least one year, derived from von Bertalanffy growth functions and
188 survivorship functions (Haddon 2011, Gislason et al. 2010). For platform and SCCWRP surveys,
189 individual fish lengths were recorded and used to estimate somatic production in the following
190 year. Individual fish lengths were not measured in the WCGBT surveys, therefore somatic
191 production was derived from the mean weight per species for each trawl. A detailed explanation
192 of the methods for calculating biomass and somatic production is provided in Appendix S1.

193 A linear model was built to test the effects of site, year, depth and habitat type (platform
194 midwater, platform base, platform shell mound, WCGBT survey and SCCWRP survey) on fish
195 density (Appendix S1). Both species richness and the Shannon-Weiner diversity index were
196 calculated and compared across each of the five habitat types. To examine the effects of site,
197 year, depth and habitat type on species composition, a Bray-Curtis dissimilarity matrix was
198 calculated from species-specific densities (fish m⁻²) at each survey transect conducted on the
199 platforms or soft bottom. Fish densities were first transformed to the fourth root to avoid
200 overrepresentation by species with very high observation counts, similar to previous studies (e.g.
201 Love et al. 2019). A two-dimensional non-metric multidimensional scaling (nMDS) ordination
202 plot was created to visualize the similarity between observed fish communities. Permutational
203 multivariate analysis of variance was performed on the Bray-Curtis distance matrix using 999
204 permutations to test the effects of site, habitat type and observation year as categorical variables
205 and sampling depth as a continuous variable on the assemblages observed (Oksanen 2019, R
206 Core Team 2019). Multivariate dispersion was tested to determine whether variance was
207 homogenous between groups among the dependent variables.

208

209 *Scaling fish metrics up to total platform habitat*

210 The underwater region associated with a single platform provides several unique habitats
211 that may harbor different species and sizes of fishes (Love et al. 2000, 2003, 2019). To account
212 for the community variation between different portions of a single platform, each platform site
213 was divided into several habitat types: the shell mound, the jacket base and a series of depth
214 strata in the jacket midwater corresponding to the positions of every major horizontal beam on
215 the jacket. For each transect surveyed, we calculated species-specific fish abundance, and

216 volumetric densities of biomass and somatic production. These metrics were averaged across all
217 surveys to characterize the difference in abundance, biomass and somatic production between
218 habitat type. Then to characterize site-specific variations in the fish community to inform
219 decommissioning, we calculated the average fish abundance, biomass and somatic production
220 densities at each platform across all surveys conducted within each habitat type. The average fish
221 abundance, biomass and somatic production estimated at each midwater depth strata were then
222 summed at each platform to estimate midwater habitat totals.

223 Since only a small portion of each habitat type is surveyed on each platform, we present
224 three different methods for applying the surveyed volumetric fish densities to characterize the
225 total water column somatic production over each platform footprint. These methods are based on
226 three alternative assumptions about how much of the habitat can be characterized by the
227 surveyed fish densities. The three methods are: (1) beams only, including only fish in surveyed
228 regions; (2) beam slices, extrapolating surveys of horizontal beams into the interior of the
229 platform but only within the 2-m-high vertical strata included in the surveys; and (3) total jacket,
230 extrapolating survey results to the entire volume within the jacket structure (Fig. 2, see Appendix
231 S1 for detail).

232

233 *Decommissioning scenarios*

234 To inform the future decommissioning process, we predicted fish community biomass
235 and somatic production under three potential platform decommissioning scenarios: leave in
236 place, partial removal, and complete removal of the platform and shell mound. Although there
237 are no defined requirements on the portion of structure that must be removed in the partial
238 removal scenario, for the sake of this study, we assumed removal of all structure above a depth
239 of 26 m, which would eliminate the necessity for marking by a lighted buoy based on US Coast
240 Guard guidelines (Stephan et al. 1990, Schroeder & Love 2004). We did not consider potential
241 future environmental or fishing effects on fish populations and community structure in this study,
242 and assumed fish densities from the surveys and trawls represent future populations. The partial
243 removal estimates are identical to the leave-in-place, except all fish from depths shallower than
244 26 m were removed.

245 In the complete removal scenario, we assumed that the fish assemblage will revert back
246 to a soft bottom ecosystem, approximated using data from trawl surveys conducted over soft

247 bottom habitats near each platform site. Biological metrics reflect estimates derived from either
248 SCCWRP or WCGBT when data from only one of the surveys was available, but in most cases
249 SCCWRP and WCGBT estimates were averaged to estimate the complete removal scenario at
250 each platform (Appendix S1: Table S1). Soft bottom densities were calculated per benthic area
251 swept by the trawl. These densities were then scaled to the total footprint area of the platform
252 jacket and surrounding shell mound (Table 1).

253

254 RESULTS

255 *Fish densities and taxa found on platforms and soft bottom*

256 Across all platforms and survey dates for the three habitat types, average fish count
257 densities were typically highest at the jacket base, with a mean density of 1.65 fish/m³ (SE = 0.2)
258 (Fig. 3; Appendix S1: Table S2). Average fish densities in the midwater portion of the jackets
259 and shell mounds were similar, 0.79 fish/m³ (SE = 0.1) and 0.81 fish/m³ (SE = 0.2), respectively
260 (Fig. 3). Rockfishes (genus *Sebastes*) dominated the platform habitats below 26 m depth,
261 including the platform base and shell mound (Fig. 4). The most common rockfish species
262 observed on the platforms were halfbanded (*Sebastes semicinctus*), squarespot (*Sebastes*
263 *hopkinsi*) and widow (*Sebastes entomelas*) rockfishes. The upper 26 m of the platform jackets
264 were typically dominated by blacksmith (*Chromis punctipinnis*) (Fig. 4).

265 Fish densities estimated from both soft bottom trawl surveys were much lower than the
266 average density observed on any of the three platform habitats (Fig. 3; Appendix S1: Table S2).
267 The mean numerical density across all SCCWRP surveys included in this study was 0.03 fish/m³
268 (SE = 0.003) and the mean density across WCGBT surveys was 0.05 fish/m³ (SE = 0.009).
269 Taxonomic compositions of the fishes recorded in the soft bottom trawl surveys differed between
270 the SCCWRP and WCGBT surveys (Fig. 4). The most common species observed in SCCWRP
271 trawls were speckled sanddab (*Citharichthys stigmaeus*), Pacific sanddab (*Citharichthys*
272 *sordidus*) and California lizardfish (*Synodus lucioceps*). The most common species observed in
273 WCGBT surveys were pink seaperch (*Zalembeus rosaceus*), slender sole (*Lyopsetta exilis*), and
274 shortbelly, splitnose and striptail rockfish (*Sebastes jordani*, *S. diploproa* and *S. saxicola*,
275 respectively). There was very little overlap between the fish species observed in the soft bottom
276 surveys and the platform surveys. Although rockfish (*Sebastes*) were still moderately common in

277 the two soft bottom trawl survey types, they were primarily species that prefer sandy and muddy
278 demersal habitats, rather than the reef-dwelling rockfishes that dominate the platforms.

279 Variation in fish species assemblage at each survey was driven primarily by habitat type
280 (Fig. 5). The nMDS ordination plot revealed that the SCCWRP and WCGBT soft bottom
281 communities are similarly structured. In a separate cluster, the three platform habitat types, the
282 midwater, base and shell mound, also exhibited considerable overlap in community structure.
283 The two benthic habitat types surveyed on platforms, the base and shell mound, were more
284 similar to the soft bottom species compositions than the midwater platform communities,
285 however, there was still no overlap between the benthic platform communities and the soft
286 bottom. The PERMANOVA analysis indicated that site, habitat type, year and depth were all
287 significant predictors of the species assemblages observed at each survey, with habitat type
288 having the largest effect on community structure (Table 2). Variances among the tested groups
289 site, habitat type, year and depth were not homogenous, therefore at least a portion of the effect
290 of these variables on community composition is driven by the variance within groups.

291 Differences in diversity between habitat types were statistically significant, with the
292 lowest species richness in the platform midwater habitat (Appendix S1: Table S3). Both species
293 richness and the Shannon-Weiner diversity index were more similar among the remaining four
294 habitat types, the platform base, shell mound, SCCWRP trawls and WCGBT trawls, all of which
295 are on or near the benthos (Appendix S1: Figure S2).

296

297 *Platform biomass and somatic production estimates*

298 Biomass estimates of fishes within the survey areas (beam only) ranged from 15.8 kg (SE
299 = 5.0) on Platform Henry to 577 kg (SE = 54) on Platform Elly (Table 3). Somatic production
300 estimates within the survey areas (beam only) ranged from 3.47 kg/yr (SE unavailable because
301 site was surveyed only once) on Platform Houchin to 100 kg/yr (SE = 21) on Platform Grace
302 (Table 3). These estimates only include fishes that were found within the surveyed habitat
303 directly adjacent to the horizontal beam, which is extremely limited relative to the total volume
304 of water occupied by the platform structure.

305 Using the two approaches for scaling up the survey data resulted in correspondingly
306 higher biomass and somatic production estimates. For the more conservative beam slices
307 method, total platform biomass ranged from 93.4 kg (SE = 11) on Platform Gina to 4440 kg (SE

308 = 610) on Platform Grace. Somatic production ranged from 15.0 kg/yr (SE = 1.7) on Platform
309 Gina to 741 kg/yr (SE = 140) on Platform Grace (Table 3). Using the more liberal total jacket
310 volume scaling method, total platform biomass ranged from 316 kg (SE = 150) on Platform
311 Henry to 32,600 kg (SE = 3300) on Platform Eureka. Somatic production ranged from 80.9 kg/yr
312 (SE unavailable because site was surveyed only once) on Platform Houchin to 5,000 kg/yr (SE =
313 1200) on Platform Grace (Table 3).

314

315 *Biomass and somatic production forecasts under partial and complete removal scenarios*

316 The potential impacts of decommissioning on platform fish communities are
317 demonstrated by forecasting changes in the fish community in the partial and full removal
318 decommissioning scenarios relative to the leave-in-place scenario. Forecasts are presented using
319 the moderate beams slices scaling method; results from the other two scaling methods are
320 included in Data S1.

321 To forecast the partial removal decommissioning scenario, we assumed the loss of all
322 fishes within the upper 26 m of the water column. The partial removal scenario resulted in an
323 average loss of 10% of the fish biomass across all of the surveyed platforms, but varied
324 considerably, ranging from 0% loss on platforms Grace, Harmony, Heritage, Hondo and Irene to
325 44% biomass loss on Platform Gina (Table 4). The partial removal scenario resulted in an
326 estimated average loss of 8% of fish somatic production across all of the surveyed platforms,
327 ranging from 0% loss on platforms Grace, Harmony, Heritage, Hondo and Irene to 48% somatic
328 production loss on Platform Gina (Table 4). The projected loss is 0% on some platforms because
329 there are no major horizontal jacket beams that occur within the upper 26 m of the water column
330 at those sites. Platform Gina, as well as platforms Edith, Henry and Hogan, have relatively high
331 projected losses in the fish assemblage relative to the other platforms because these are the
332 shallowest platforms in the study (Platform Gina sits at a depth of 29 m) and the upper 26 m
333 includes at least half of the horizontal cross beams at these sites (Table 1). For all 24 surveyed
334 platforms, the majority of the fish assemblage was retained following the removal of rig structure
335 above 26 m depth.

336 Soft bottom trawl surveys were used to forecast change in fish biomass and somatic
337 production in the complete removal scenario. We estimated that the complete removal scenario
338 would result in an average loss of 96% of the fish biomass across all of the surveyed platforms,

339 ranging from 83% loss on Platform Heritage to at least a 98% loss on 15 of the surveyed
340 platform sites (Table 4). Similarly, complete removal will result in an average loss of 95% of the
341 somatic production across all surveyed platforms, ranging from 66% loss on Platform Heritage to
342 at least a 98% loss on 14 of the platforms surveyed (Table 4). Losses in somatic production on
343 the large platforms Heritage and Harmony are smaller because the large footprint of these two
344 jackets (Table 1) will result in a substantially larger reclamation of soft bottom habitat relative to
345 the other platform sites.

346 To illustrate the scale of the impact that each decommissioning alternative might have for
347 reef fish offshore of southern California, we aggregated the total fish biomass and somatic
348 production across all 24 platforms included in this study under the three decommissioning
349 scenarios (Fig. 6). In this example, we assume that the same decommissioning alternative is
350 applied to all 24 platform sites and we use the beam slices approach to estimate metrics. If the
351 complete underwater structure at all 24 sites is left in place, the platform-associated habitats will
352 support a total of 29,171 kg (SE=1,741) of fish biomass, and an annual somatic production of
353 4,772 kg/yr (SE=374). If the top 26 m of each platform jacket is removed, but the jacket and
354 shell mound materials below 26 m is retained, the platform-associated habitats will support a
355 total of 27,848 kg (SE=1,737) of fish biomass, and an annual somatic production of 4,584 kg/yr
356 (SE=374). If all 24 platform jackets and shell mounds are completely removed, the new soft
357 bottom habitat will support a total of 518 kg (SE=29) of fish biomass, and an annual somatic
358 production of 112 kg/yr (SE=6).

359

360 **DISCUSSION**

361 The results of this study indicate that less than 5% of the fish biomass and somatic
362 production associated with 24 California offshore oil and gas platform sites will be lost in the
363 event of partial removal of the structures above a depth of 26 m. However, complete removal
364 will result in a net loss of 98% of the biomass even though the artificial-reef-associated fish
365 communities will be replaced with fish characteristic of the surrounding soft bottom
366 environments. Complete removal of the platform and shell mound of all 24 sites included in this
367 study will result in a conservatively estimated net regional biomass loss of 28,653 kg of fish,
368 with an additional loss of 4,659 kg/yr of annual somatic fish production (Fig. 6). To put these
369 numbers into context, total commercial landings in Santa Barbara Harbor in 2017 consisted of

370 35,485 kg of rockfish, valued at \$357,106 (California Department of Fish and Wildlife 2018).
371 This substantial fish loss following complete platform removal occurs because fish densities over
372 soft bottom are much lower than densities found in platform-associated habitats in southern
373 California (Fig. 3).

374 Biological metrics vary substantially between the three methods developed in this study
375 for applying the observed fish assemblages to the entire platform-associated habitat volumes. We
376 recommend using estimates from the beam slices approach for general quantitative description of
377 the platform-associated fish assemblage size because it provides a conservative estimate of the
378 abundance, biomass and somatic production of fishes found at each platform. Since the space
379 between beams is assumed to have no fish, this approach will result in an underestimate of fish
380 present; however, the proportion of the total artificial reef that is not accounted for scales up with
381 platform height since the spacing between major horizontal beams is wider at deeper platforms.
382 The lower metrics provided in the beams-only approach are based on the assumption that there
383 are no fish outside of the targeted survey volume. This approach is useful for quantifying the
384 subset of the fish assemblage that has been directly observed and providing context for the gap
385 between applied survey effort and a more idealized complete habitat monitoring program.
386 However, the beams-only approach does not produce a comprehensive assessment of the total
387 fish community associated with the platform structure since it excludes fish in the jacket interior
388 which may be present at densities higher than those observed on the external jacket beams. In a
389 study of fish surveys in the upper 30 m of 11 California platforms, fish densities were 2.8 times
390 higher surrounding beams spanning the inside of the jacket relative to the external jacket beams
391 that were surveyed in this study (Meyer-Gutbrod et al. 2019b). However, there is no data
392 available to verify whether this trend of higher densities inside the jacket is consistent at other
393 California platforms, or at depths below 30 m where only submersible and ROV surveys are
394 available.

395 The biological metrics estimated in the total jacket volume method provide an estimate of
396 the maximum fish assemblage size at each platform site. Fish densities are often highest in close
397 proximity to structure (Meyer-Gutbrod et al. 2019b), which indicates that using densities
398 observed over the horizontal beams to characterize some of the more open areas inside of the
399 jacket between major beams may lead to an overestimation with the total jacket method.
400 However, none of the three scaling methods account for the halo of fish habitat found outside of,

401 but in close proximity to, the platform jacket. Fish densities just outside the platform jacket have
402 been noted to be considerably higher than in open pelagic environments (Soldal et al. 2002, Scott
403 et al. 2015, Reynolds et al. 2018). Therefore, the total jacket volume method may not necessarily
404 overestimate fish abundance. Future high-resolution survey efforts could be implemented to
405 increase the precision in the total fish community size estimates, which probably fall somewhere
406 between the beam slices and total jacket volume approaches.

407 Removal of the shallowest 26 m of a platform may impact communities in the remaining
408 artificial reef by reducing shell mound formation or recruitment of shallow larvae and juveniles.
409 Since most mussel growth on the platform occurs above 26 m, removal of the shallowest portion
410 of the jacket will curtail the continued formation of the shell mound at the rig base (Page et al.
411 2005, Meyer-Gutbrod et al. 2019a). Persistence of the current shell mound and changes to its
412 biota will depend on local currents and sedimentation rates as well as the reduction in mussel
413 production (Bomkamp et al. 2004, Claisse et al. 2015). Enhancement of the artificial reef by
414 depositing the structure removed from the upper 26 m of the water column alongside the rig base
415 or adding other material such as rock (Holbrook et al. 2000) would create new complex habitat
416 and increase the size of the benthic fish community. Young-of-the-year rockfishes are most
417 abundant on platform regions deeper than 26 m, suggesting that partial removal will not
418 adversely affect larval settlement onto the structure (Nishimoto & Love 2011). Comparisons
419 between young-of-the-year assemblages on platforms, submerged shipwrecks and natural reefs
420 indicated that the extension of structure to the sea surface on platforms did not impact rockfish
421 recruitment to deeper habitats below 26 m (Love et al. 2012). The dominant species lost in the
422 partial removal scenario, *Chromis punctipinnis* (Fig. 4), is regionally abundant and not
423 commercially valuable; however, it may function as an important prey group for higher trophic
424 levels. Although there is significant species overlap between the platform midwaters, base and
425 shell mound (Fig. 5), diversity is lowest in the platform midwaters (Appendix S1: Fig. S2, Table
426 S3), indicating that the partial removal scenario will not result in much decrease in overall
427 diversity at the site.

428 Survey coverage was not sufficient to provide a robust estimate of fish biomass and
429 somatic production at all southern California platforms. Although limited data on platforms
430 Esther and Eva have been reported elsewhere (Martin & Lowe 2010), fish surveys have never
431 been conducted on Platform Emmy. Of the platforms included in this study, Harmony, Heritage,

432 and Houchin have only been surveyed once. The estimates of decommissioning impacts at these
433 sites are not as reliable as the estimates for sites with higher survey effort because there is
434 considerable temporal variability in fish assemblage size and composition (e.g. Love et al. 2019,
435 Claisse et al. 2014, Meyer-Gutbrod et al. 2019a). Additional, more frequent, survey effort would
436 be beneficial to adequately inform the decommissioning process for these and other platforms.

437 Differences between fish behavior, such as attraction or avoidance, associated with the
438 presence of a SCUBA diver, manned submersible, ROV and bottom trawl may vary. These
439 small, but potentially biased discrepancies between survey styles could not be accounted for in
440 this study since surveys were never conducted simultaneously from two or more different
441 platforms. There is a growing body of literature examining bias due to observation platform, and
442 further work in this field should improve accuracy in visual and extractive methods in fish
443 density estimation (Trenkle et al. 2004, Weinberg et al. 2002, Stoner et al. 2008, Stanley and
444 Wilson 1995). Differences in trawl gear and methodologies between the SCCWRP and WCGBT
445 surveys used in this study may also lead to biases in soft bottom fish community size and species
446 composition. Variation in cod end mesh size, footrope length, trawl speed and duration can
447 contribute to these differences (Somerton et al. 2002, Stergiou et al. 1997, Weinberg et al. 2002).
448 Diversity indices are higher in WCGBT surveys compared to SCCWRP surveys (Appendix S1:
449 Fig. S2, Table S3), although species composition is similar (Fig. 5), indicating that the smaller
450 SCCWRP trawls may miss some species.

451 Fish density and community composition are also subject to interannual variability,
452 which could not be fully captured in this study due to differences in survey timing and frequency
453 between sites and habitat types. Studies of the subset of platforms that are surveyed regularly
454 have suggested that much of the observed interannual variability at these sites is driven by
455 rockfish recruitment dynamics and the random nature of sampling large, mobile schools (e.g.,
456 Love et al. 2006, Meyer-Gutbrod et al. 2019, Love et al. 2019). Fish community composition and
457 size may also exhibit some seasonal variability. Although the surveys included in this study were
458 primarily conducted in the summer and fall, many of the rockfish species observed on the
459 platforms are resident throughout the year, reducing the impact of seasonal variability (Lowe et
460 al. 2009, Martin and Lowe 2010). Rockfishes have long reproductive seasons and exhibit
461 significant seasonal variability within and among the different species (Echeverria 1987),
462 indicating that survey timing may not dramatically impact juvenile counts. However, additional

463 surveys conducted in winter and spring would be useful to determine the impact of sampling
464 season on the results of this study.

465 Although fish biomass and somatic production estimates for each platform provide a
466 measure of the main impacts of decommissioning at each site, a holistic analysis could include
467 several additional ecological metrics. Comparison of the invertebrate community above and
468 below 26 m and on the soft bottom may be possible using video recordings of the submersible
469 surveys paired with trawl surveys. Fish collection efforts would be useful for examining the
470 differences in growth rates and egg production on platforms relative to natural reefs. Larger scale
471 comparisons of the value of platform habitat for commercially fished or at risk species such as
472 *Sebastes levis* or *Sebastes paucispinis* with the effect of platforms on soft bottom habitat for
473 commercially fished species such as *Citharichthys stigmaeus* would be valuable. Oceanographic
474 connectivity analyses may reveal the significance of these artificial reefs to the regional
475 distribution and population genetic structure of fish species.

476 Previous studies of fish assemblages on California platforms have focused on the habitat
477 and biological community that would be lost when the platform jacket, conductors and shell
478 mound are removed during the decommissioning process (e.g. Pondella et al. 2015, Claisse et al.
479 2014, Love et al. 2003). This study extends these efforts by considering the habitat and
480 biological communities that would be gained under this scenario as the seafloor reverts back to
481 productive soft bottom habitat. The assumption that this habitat will be devoid of fish following
482 complete removal of the structure and shell mound is problematic, and leads to biased
483 predictions of fish loss under alternative decommissioning scenarios. This study examines the
484 transformation, rather than loss, of habitat to present more accurate predictions of fish
485 community size and composition following a range of decommissioning scenarios.

486 Habitat transformation following platform removal results in the exchange of reef habitat
487 for trawlable soft bottom habitat. Over 500 offshore platforms in the North Sea (Fowler et al.
488 2018) and nearly 2000 installations in the Gulf of Mexico (Kaiser et al. 2019) limit trawlable
489 habitat in ocean basins that are subject to intense fishing pressure. Offshore installations create
490 no-trawl zones, which act as *de facto* marine reserves, possibly increasing productivity similar to
491 a marine protected area (Schroeder & Love 2002, Love et al. 2003). However, no-trawl zones
492 increase competition for space, increase costs for traveling to trawlable areas, and may displace

493 fishing activities into habitats that are more vulnerable to disturbance (Kaiser et al. 2002,
494 Bloomfield et al. 2012).

495 Disagreement between stakeholders on the optimal use of decommissioned platform
496 habitats as either artificial reef or the preceding soft bottom habitat extend to preferences in
497 which species will inhabit these spaces. This study found no evident overlap in community
498 structure between the soft bottom habitat and the platform habitats and it is difficult to compare
499 the value of platform habitat for commercially fished or at risk species such as *Sebastes levis* or
500 *Sebastes paucispinis* with the value of commercially fished soft bottom species such as
501 *Citharichthys stigmaeus*. Since species composition varies by depth on both soft bottom and
502 artificial reef habitats (e.g. Bradburn et al. 2011, Love et al. 2019), the community
503 transformation following structure removal may be examined in more detail at each site.
504 Additionally, fish assemblages in the leave-in-place and partial removal scenarios may differ
505 from those presented in this analysis if fishing near these sites increases after decommissioning.
506 Maximizing fish biomass and production on infrastructure converted to artificial reefs, therefore,
507 will depend on future regulation of fishing at these sites. In the complete removal scenario, the
508 site of the platform footprint will likely become available for fishing, and therefore the species
509 composition and density found in the trawl surveys used in this study are an accurate prediction
510 of the future fish community. Deciding the ultimate fate of these structures and their
511 management with respect to fish habitat will require a careful assessment of the complex costs
512 and benefits associated with each alternative.

513 There are a host of additional factors besides fish communities that must be weighed
514 when considering decommissioning alternatives. For a complete environmental impact analysis,
515 studies must be conducted to assess the effects of pollution related to the dismantling, transport
516 and recycling of structural elements, navigational considerations, resuspension of contaminants
517 trapped under the shell mound, fishing impacts and the effects of explosives on surrounding
518 marine life (Schroeder & Love 2004, Fowler et al. 2014). Beyond environmental impacts,
519 decommissioning strategies will also include considerations for safety, economics, politics and
520 aesthetics. These issues are multi-faceted and decisions should be made following input from a
521 diverse range of stakeholders. However, the fish community forecasts presented here
522 demonstrate that complete removal of California oil platforms will result in a net loss of local

523 and regional fish biomass and production. These results constitute a critical component of the net
524 environmental benefit analysis of California oil and gas platform decommissioning alternatives.

525

526 **CONCLUSIONS**

527 This study forecasts the biomass and somatic production of the fish communities associated with
528 each of 24 oil and gas platforms in the Southern California Bight under three potential
529 decommissioning scenarios to inform net environmental benefit comparisons of
530 decommissioning alternatives (California Marine Resources Legacy Act, CA AB 2503, Pérez
531 2010). The annual somatic production that would be lost if all 24 platforms are completely
532 removed is equal to 13% of the annual Santa Barbara Harbor commercial fishing landings.
533 However, the majority of the fish biomass and annual somatic production at each site will be
534 preserved if the deep portion of the structure is converted to an artificial reef. The fish
535 assemblage forecasts estimated here for most Pacific Outer Continental Shelf platforms can
536 contribute to the development of a decommissioning strategy that minimizes the ecological
537 impacts of decommissioning and maximizes benefit to the marine environment.

538 The substantial loss in fish biomass and somatic production predicted to occur at
539 California offshore platform sites in the scenario where all structure is removed has meaningful
540 implications for platform decommissioning decision-making worldwide. The ecosystems
541 beneath most California platforms have been exceptionally well-surveyed compared to regions
542 that have much higher numbers of offshore platforms, such as the North Sea and the Gulf of
543 Mexico. With thousands of platforms globally eventually facing decommissioning (Parente et al.
544 2006, IHS Markit 2016), survey efforts in each region should be expanded to determine whether
545 platforms in other ocean basins have similar ecological effects. The Gulf of Mexico has the most
546 active rigs to reefs program, where artificial reef conversion has been implemented for more than
547 500 of the nearly 5,000 structures that have been decommissioned (Rigs to Reefs 2017, Bull &
548 Love 2019, Kaiser et al. 2019). Although research documenting biotic assemblages on reefed
549 installations in the Gulf of Mexico (e.g. Ajemian et al. 2015) is sparse, additional survey effort at
550 these sites combined with close monitoring of the ongoing decommissioning process in
551 California would provide critical information for the rigs to reef discussions that are active in
552 regions such as the North Sea, Australia and the Gulf of Thailand.

553

554 **ACKNOWLEDGMENTS**

555 The submersible Delta was piloted by C Ijames and J Lilly and the Dual Deepworker by J
556 Heaton. We thank Captain I Leask and all of the crew of the RV Valero for their able assistance.
557 The ROV was piloted by T Haaland. We also thank Arnold Ammann, Scott Clark, William
558 Golden, Jeff Harding, Eric Hessel, Laird McDonald, Michelle Paddack and helpful volunteers
559 for diving and collecting data. We thank Maria Aiello along with Brittany Munson and
560 volunteers who transcribed archived records into a database. This research was supported by the
561 Bureau of Ocean Energy Management (BOEM) Environmental Studies Program (ESP) through
562 Awards M15AC00014, Synthesis of Pacific Platform Research, and M16AC00025, Net
563 Environmental Benefit Analysis of Pacific Platform Decommissioning Scenarios. Support was
564 also provided by the National Aeronautics and Space Administration Biodiversity and Ecological
565 Forecasting program (Grant NNX14AR62A), the BOEM ESP (Award M15AC00006) and the
566 National Oceanic and Atmospheric Administration in support of the Santa Barbara Channel
567 Marine Biodiversity Observation Network, and by the University of Southern California Sea
568 Grant Program, including a Sea Grant Traineeship to C. M. Williams and A. Roeper.

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814 **Data Availability**

815 The authors declare that all data supporting the findings of this study can be found in the
816 following online repositories. The fish surveys conducted on platform-associated habitats are
817 available on the EDI Data Portal at:
818 <http://dx.doi.org/10.6073/pasta/2dc1e7a1ce14e0f3f070076fc4a85e43>. The NOAA West Coast
819 Groundfish Bottom Trawl data are available from the NOAA Northwest Fisheries Science Center at:
820 <https://www.webapps.nwfsc.noaa.gov/apex/parrdata/inventory/datasets/dataset/131>. The
821 Southern California Coastal Research Project trawl data are available at:
822 <http://sccwrp.org/Data/SearchAndMapData/DataCatalog.aspx>.

823 Table 1. Pacific offshore oil and gas platform structure.

824 *No fish surveys performed

825

| Platform | Region | Jurisdiction | Year installed | Base depth (m) | # Midwater beams | Jacket footprint (m ²) | Jacket volume (m ³) | Shell mound area (m ²) |
|----------|-----------------|--------------|----------------|----------------|------------------|------------------------------------|---------------------------------|------------------------------------|
| A | East SB Channel | Federal | 1968 | 58 | 4 | 1930 | 75653 | 3382 |
| B | East SB Channel | Federal | 1968 | 58 | 5 | 1930 | 75330 | 3122 |
| C | East SB Channel | Federal | 1977 | 58 | 5 | 1930 | 75330 | 3493 |
| EDITH | Orange County | Federal | 1983 | 49 | 2 | 2879 | 112147 | NA |
| ELLEN | Orange County | Federal | 1980 | 80 | 3 | 2511 | 131841 | NA |
| ELLY | Orange County | Federal | 1980 | 77 | 3 | 2949 | 140382 | NA |
| EMMY* | Orange County | State | 1963 | 14 | NA | 669 | 9360 | NA |
| ESTHER* | Orange County | State | 1985 | 10.7 | NA | 669 | 7154 | NA |
| EUREKA | Orange County | Federal | 1984 | 212 | 9 | 4635 | 563814 | NA |
| EVA* | Orange County | State | 1964 | 17 | NA | 669 | 11366 | NA |
| GAIL | East SB Channel | Federal | 1987 | 224 | 9 | 5327 | 671776 | 655 |
| GILDA | East SB Channel | Federal | 1981 | 62 | 3 | 2342 | 97386 | 18290 |
| GINA | East SB Channel | Federal | 1980 | 29 | 1 | 561 | 11305 | 2926 |
| GRACE | East SB Channel | Federal | 1979 | 96 | 4 | 3090 | 199728 | 22754 |
| HABITAT | East SB Channel | Federal | 1981 | 88 | 4 | 2284 | 119889 | 4560 |
| HARMONY | West SB Channel | Federal | 1989 | 363 | 7 | 10606 | 1659349 | NA |
| HARVEST | Pt. Conception | Federal | 1985 | 205 | 7 | 5859 | 683741 | NA |
| HENRY | East SB Channel | Federal | 1979 | 52 | 3 | 1505 | 77220 | 4560 |
| HERITAGE | West SB Channel | Federal | 1989 | 326 | 3 | 10606 | 1490214 | NA |
| HERMOSA | Pt. Conception | Federal | 1985 | 183 | 8 | 5142 | 599639 | 642 |
| HIDALGO | Pt. Conception | Federal | 1986 | 130 | 5 | 4154 | 366288 | 0 |

| | | | | | | | | |
|-----------|-----------------|---------|------|-----|---|------|--------|-------|
| HILLHOUSE | East SB Channel | Federal | 1969 | 58 | 3 | 1960 | 76662 | 4515 |
| HOGAN | East SB Channel | Federal | 1967 | 47 | 3 | 1435 | 67868 | 4932 |
| HOLLY | West SB Channel | State | 1966 | 64 | 4 | 1728 | 69120 | NA |
| HONDO | West SB Channel | Federal | 1976 | 255 | 7 | 4649 | 587722 | 1821 |
| HOUCHIN | East SB Channel | Federal | 1968 | 49 | 3 | 1435 | 70756 | 5721 |
| IRENE | Pt. Conception | Federal | 1985 | 73 | 3 | 2633 | 123884 | 13484 |

826

827 Table 2. Permutational analysis of variance (PERMANOVA) model testing the effects of site, habitat type, observation year and depth
828 on fish community composition. PERMANOVA was performed on a Bray-Curtis dissimilarity matrix of fourth-root transformed fish
829 densities separated by species.

830

831

| | df | F- value | R ² | p- value |
|--------------|----|-------------|----------------|-------------|
| Site | 26 | 12.53 | 0.17 | 0.001 |
| Habitat type | 4 | 111.02 | 0.23 | 0.001 |
| Year | 20 | 4.63 | 0.05 | 0.001 |
| Depth | 1 | 45.15 | 0.02 | 0.001 |

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835 Table 3. Biomass and somatic production estimates for each platform using the three methods: beams only, beam slices and total
836 jacket volume. These values are used to simulate the leave-in-place decommissioning scenario.

837

| Platform | Beams only | | | | Beam slices | | | | Total jacket volume | | | |
|-----------|------------|---------|-----------------------|------------------|-------------|---------|-----------------------|------------------|---------------------|---------|-----------------------|------------------|
| | Biomass | Biomass | Somatic | Somatic | Biomass | Biomass | Somatic | Somatic | Biomass | Biomass | Somatic | Somatic |
| | (kg) | SE | Production (kg/yr) | Production SE | (kg) | SE | Production (kg/yr) | Production SE | (kg) | SE | Production (kg/yr) | Production SE |
| A | 229 | 70 | 40 | 10 | 1297 | 420 | 226 | 60 | 2898 | 537 | 571 | 126 |
| B | 89 | 13 | 24 | 4 | 466 | 68 | 126 | 23 | 1555 | 324 | 453 | 141 |
| C | 77 | 12 | 23 | 4 | 447 | 90 | 126 | 24 | 1360 | 273 | 454 | 122 |
| EDITH | 198 | 32 | 31 | 6 | 938 | 136 | 148 | 33 | 5549 | 1228 | 673 | 114 |
| ELLEN | 403 | 74 | 53 | 21 | 2020 | 400 | 270 | 117 | 18986 | 7058 | 3766 | 2166 |
| ELLY | 577 | 54 | 45 | 7 | 3105 | 313 | 236 | 42 | 18208 | 4266 | 1500 | 426 |
| EUREKA | 480 | 43 | 57 | 6 | 3165 | 295 | 354 | 33 | 32571 | 3253 | 3810 | 380 |
| GAIL | 296 | 30 | 49 | 5 | 2359 | 266 | 393 | 42 | 4097 | 379 | 857 | 112 |
| GILDA | 175 | 48 | 43 | 15 | 2818 | 1362 | 643 | 299 | 4166 | 1394 | 964 | 315 |
| GINA | 36 | 4 | 6 | 1 | 93 | 11 | 15 | 2 | 631 | 81 | 110 | 14 |
| GRACE | 506 | 81 | 100 | 21 | 4443 | 613 | 741 | 137 | 17859 | 4054 | 4998 | 1231 |
| HABITAT | 60 | 11 | 18 | 4 | 343 | 55 | 100 | 21 | 2828 | 836 | 979 | 348 |
| HARMONY | 34 | NA | 5 | NA | 351 | NA | 37 | NA | 2508 | NA | 524 | NA |
| HARVEST | 118 | 18 | 27 | 5 | 730 | 106 | 159 | 29 | 7293 | 1220 | 1723 | 332 |
| HENRY | 16 | 5 | 6 | 2 | 94 | 25 | 35 | 9 | 316 | 151 | 107 | 47 |
| HERITAGE | 30 | NA | 5 | NA | 353 | NA | 55 | NA | 10731 | NA | 1614 | NA |
| HERMOSA | 223 | 33 | 30 | 5 | 1487 | 257 | 189 | 33 | 7802 | 1317 | 1644 | 306 |
| HIDALGO | 154 | 16 | 22 | 3 | 1079 | 118 | 144 | 16 | 3242 | 509 | 609 | 141 |
| HILLHOUSE | 49 | 14 | 12 | 4 | 290 | 76 | 73 | 20 | 1090 | 464 | 345 | 213 |
| HOGAN | 25 | 7 | 4 | 1 | 153 | 44 | 31 | 13 | 765 | 268 | 114 | 45 |
| HOLLY | 105 | 13 | 19 | 2 | 416 | 46 | 77 | 9 | 1416 | 157 | 329 | 51 |
| HONDO | 162 | 45 | 22 | 5 | 891 | 219 | 129 | 29 | 14312 | 4107 | 2098 | 505 |
| HOUCHIN | 17 | NA | 3 | NA | 175 | NA | 38 | NA | 430 | NA | 81 | NA |

| | | | | | | | | | | | | |
|-------|-----|----|----|---|------|-----|-----|----|------|------|------|-----|
| IRENE | 222 | 21 | 57 | 9 | 1660 | 151 | 427 | 59 | 8048 | 1728 | 3118 | 839 |
|-------|-----|----|----|---|------|-----|-----|----|------|------|------|-----|

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841 Table 4. Platform-specific biomass and somatic production model estimates for each platform and the expected percent loss in
842 biomass and somatic production under the scenarios of partial removal at 26 m and complete removal of platform infrastructure and
843 shell mound. The leave in place and partial removal scenarios are based on the beam slices method. The complete removal scenarios
844 are calculated using the biomass and somatic production estimates averaged between the WCGBT and SCCWRP bottom trawl
845 surveys, when both are available.

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847

| Platform | Biomass | | | Somatic production | | |
|----------|--------------------------|-------------------------------|--------------------------------|--------------------------------|-------------------------------|--------------------------------|
| | Biomass on platform (kg) | Loss with partial removal (%) | Loss with complete removal (%) | Production on platform (kg/yr) | Loss with partial removal (%) | Loss with complete removal (%) |
| A | 1297 | 0.08 | 0.99 | 226 | 0.06 | 0.99 |
| B | 466 | 0.14 | 0.98 | 126 | 0.06 | 0.99 |
| C | 447 | 0.03 | 0.98 | 126 | 0.01 | 0.98 |
| EDITH | 938 | 0.46 | 0.99 | 148 | 0.33 | 0.99 |
| ELLEN | 2020 | 0.02 | 0.99 | 270 | 0.02 | 0.98 |
| ELLY | 3105 | 0.03 | 0.99 | 236 | 0.04 | 0.98 |
| EUREKA | 3165 | 0.05 | 0.99 | 354 | 0.09 | 0.98 |
| GAIL | 2359 | 0.02 | 0.99 | 393 | 0.02 | 0.98 |
| GILDA | 2818 | 0.03 | 0.99 | 643 | 0.02 | 0.99 |

| | | | | | | |
|-------------|------|------|------|-----|------|------|
| GINA | 93 | 0.44 | 0.98 | 15 | 0.48 | 0.96 |
| GRACE | 4443 | 0.00 | 0.99 | 741 | 0.00 | 0.99 |
| HABITAT | 343 | 0.05 | 0.92 | 100 | 0.03 | 0.95 |
| HARMONY | 351 | 0.00 | 0.87 | 37 | 0.00 | 0.73 |
| HARVEST | 730 | 0.03 | 0.96 | 159 | 0.03 | 0.98 |
| HENRY | 94 | 0.28 | 0.90 | 35 | 0.22 | 0.94 |
| HERITAGE | 353 | 0.00 | 0.83 | 55 | 0.00 | 0.66 |
| HERMOSA | 1487 | 0.01 | 0.98 | 189 | 0.02 | 0.98 |
| HIDALGO | 1079 | 0.01 | 0.96 | 144 | 0.01 | 0.96 |
| HILLHOUSE | 290 | 0.12 | 0.96 | 73 | 0.09 | 0.96 |
| HOGAN | 153 | 0.37 | 0.96 | 31 | 0.07 | 0.95 |
| HOLLY | 416 | 0.08 | 0.99 | 77 | 0.11 | 0.98 |
| HONDO | 891 | 0.00 | 0.99 | 129 | 0.00 | 0.97 |
| HOUCHIN | 175 | 0.11 | 0.95 | 38 | 0.08 | 0.94 |
| IRENE | 1660 | 0.00 | 0.99 | 427 | 0.00 | 0.99 |
| Mean | 1215 | 0.10 | 0.96 | 199 | 0.08 | 0.95 |

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Figure Legends

Figure 1. Map showing the names, locations and seafloor depths (in meters) of the 27 Southern California offshore oil and gas platforms. Solid circles indicate platform sites that have been surveyed and included in this study.

Figure 2. Schematics show three separate methods for scaling up platform survey fish densities on a simplified platform jacket. Colored polygons indicate the volume associated with each surveyed habitat: midwater beams at three depth levels (yellow, green, pink), the rig base (purple) and the shell mound (brown). (A.) The beams only method accounts only for fish along the beam, base and shell mound survey transects, with 2 m transect widths and heights. (B.) In the beam slices method, fish densities measured along each beam are scaled up to a 2 m high beam slice running through the middle of the jacket. Shell mound densities are scaled up to the total shell mound area. (C.) In the total jacket method, the total volume of the jacket is broken into a stack of truncated pyramids bounded by the legs of the jacket, with horizontal boundaries positioned halfway between adjacent midwater beams. Fish densities measured along each beam are scaled up to the volume of the truncated pyramid intersected by that beam. Shell mound densities are scaled up to the total shell mound area.

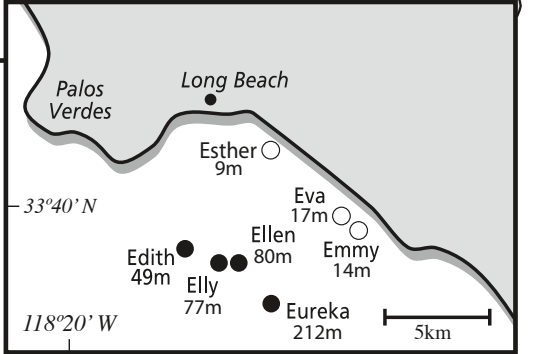
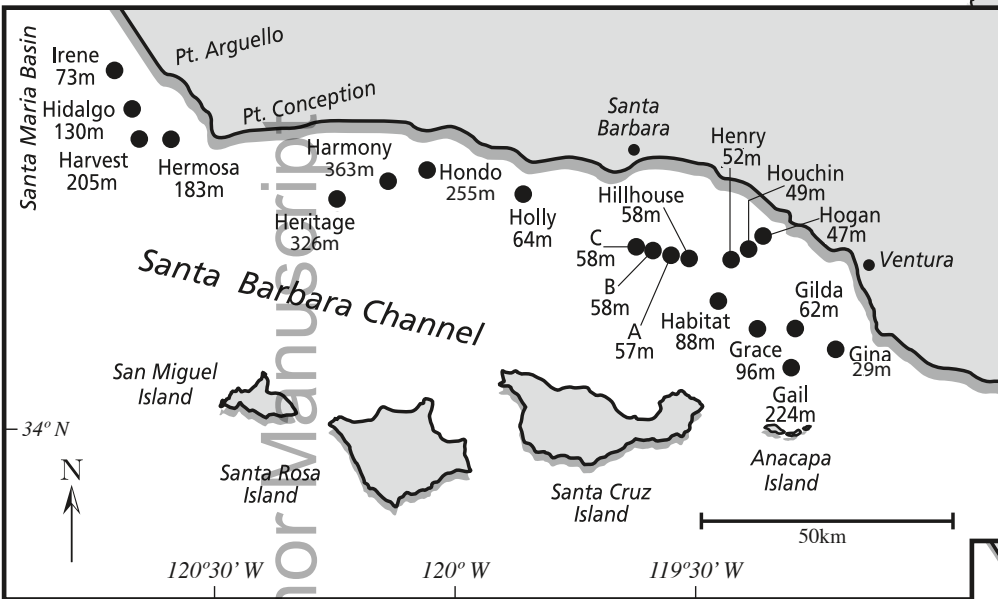
Figure 3. Box plots of fish density (# fish/m³) at each survey by habitat. SCCWRP and WCGBT surveys of soft bottom fish densities are combined. Only trawls that were conducted within 30km of a platform and within 25m depth range of a platform base are included.

Figure 4: Abundance of fish by taxonomic group in all surveys of platforms (top row) and soft bottom trawls (bottom row). Taxonomic groups are categorized by genera; if only one species in the genus was observed, the species name is provided, otherwise the number of species represented by that genus is provided in parenthesis. Platform taxonomic abundance charts are broken into four distinct habitats: jacket midwater surveys above a 26 m depth, jacket midwater surveys below a 26 m depth, jacket base surveys and shell mound surveys. Soft-bottom taxonomic abundance charts are separated into the two trawl survey methods: Southern California Coastal Water Research Project (SCCWRP) trawl surveys and West Coast Groundfish Bottom

Trawl (WCGBT) surveys. Fish taxa representing less than 2% of the total fish abundance for each distinct habitat type or method are aggregated into “Other”.

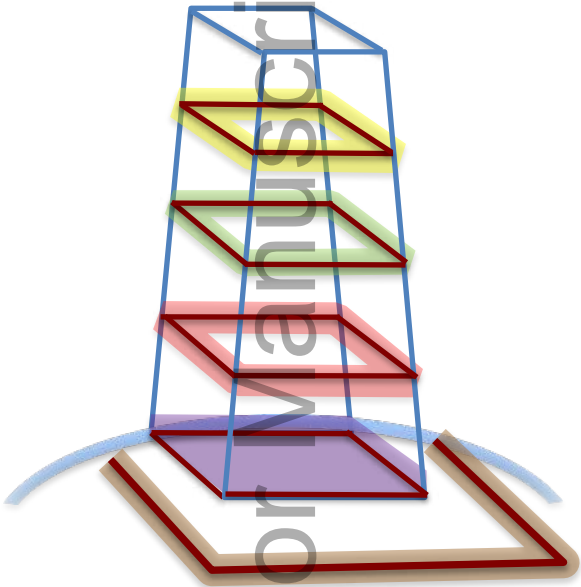
Figure 5. Two-dimensional non-Metric Dimensionless scaling (nMDS) ordination plot demonstrating the two dominant modes of variability in a Bray-Curtis dissimilarity matrix constructed from fourth-root-transformed fish communities surveyed along platforms and soft bottom habitats. Colors indicate the habitat type surveyed, and ellipses represent 95% confidence intervals around the centroid of each habitat type.

Figure 6: Summed total of the mean (A) fish biomass (kg) and (B) annual somatic production (kg/yr) at all 24 platforms included in this study projected under three potential decommissioning scenarios: leave in place, partial removal at 26 m depth, and complete removal. The leave in place and partial removal scenario estimates are based on the beam slices scaling method. Error bars represent the standard error propagated from summing the contributions from each platform site.



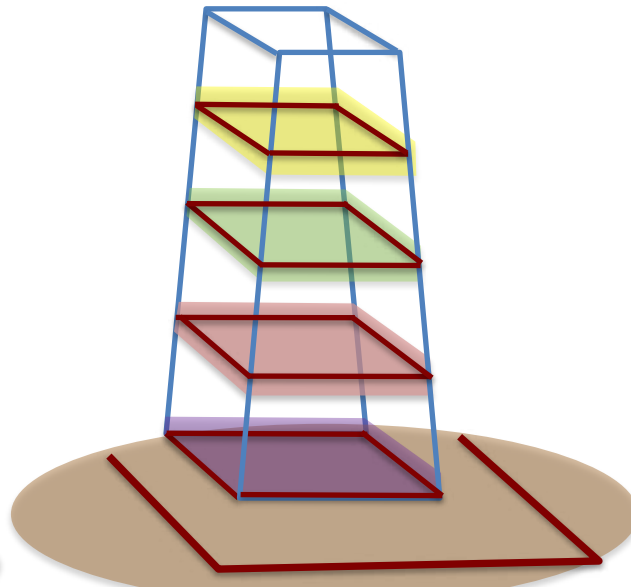
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A



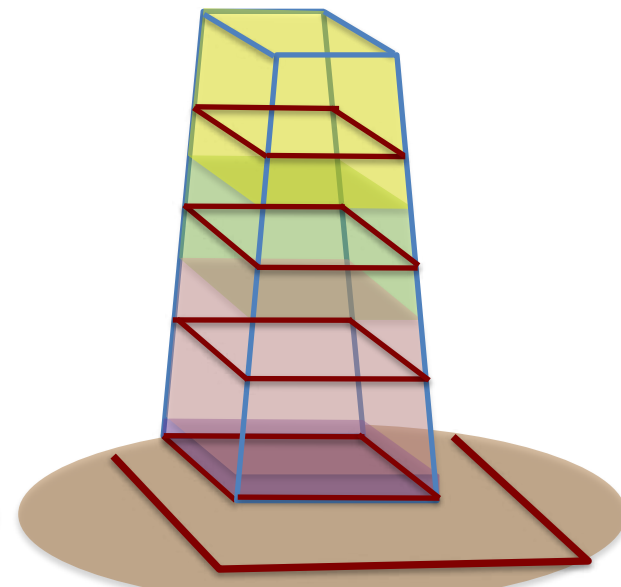
Beams only

B

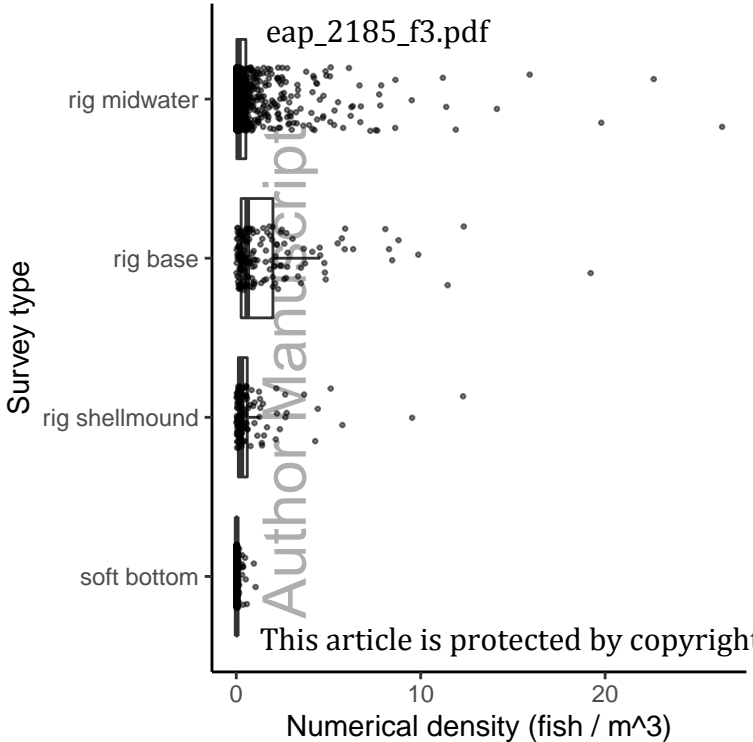


Beam slices

C



Total jacket



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