1	
2	
3	
4	Article type : Article
5	
6	
7	ARTICLE Genetic Divergence of Nearby Walleye Spawning Groups in Central
8	Lake Erie: Implications for Management
9	
10	Carol A. Stepien ¹ *, Matthew R. Snyder ^{2,1} , and Carey T. Knight ³
11	
12	¹ Genetics and Genomics Group, NOAA Pacific Marine Biological Laboratory (PMEL),
13	7600 Sand Point Way NE, Seattle, WA 98115 USA
14	
15 16	² Department of Environmental Sciences, University of Toledo, Toledo, OH 43606 USA
17	³ Ohio Division of Wildlife, Ohio Department of Natural Resources, Fairport Harbor Fisheries Research
18	Station, 1190 High St., Fairport Harbor, OH 44077 USA
19 20	
20	*Corresponding author: <u>carol.stepien@noaa.gov</u> ; <u>carol.stepien@utoledo.edu</u> ; 206-526-6038
21	Abstract
22	Spawning site philopatry may lead to genetic differentiation among reproductive populations,
23	despite their locations being in close proximity in single bodies of water. Identifying and
24	maintaining locally differentiated spawning groups of Walleye Sander vitreus constitute an agency
25	management priority of the multi-agency Walleye Task Group advisory for the Great Lakes
26	Fishery Commission. Although genetically separable spawning groups of Walleye have been
27	identified from several areas in the Great Lakes, those in central Lake Erie were previously
28	unknown. The Ohio Division of Wildlife (ODW) collected Walleye for the present analysis from two
29	spawning groups in Lake Erie's Central Basin, located just 2 km apart – one in the Grand River,
30	Ohio, and the other at the nearby Central Basin Reef. The hypothesis of whether the two spawning
	This is the author manuscript accepted for publication and has undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may

not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the <u>Version of Record</u>. Please cite this article as <u>doi:</u> <u>10.1002/nafm.10176</u>

31 groups (N=147) genetically differ was tested by analyzing variation at 14 nuclear DNA 32 microsatellite loci from Walleye reproducing in 2012 at the two sites, and among three separate 33 years (1996, 2003, and 2012) in the Grand River (to evaluate temporal trends). Results revealed 34 relatively high genetic diversity in both spawning groups, with the reproductive population in the 35 Grand River having significantly greater allelic richness and representation of private alleles. The 36 Grand River runs showed slight temporal decline in allelic richness from 1996-2003, coinciding 37 with ODW reports of declining numbers of Walleye reproducing there. The two spawning groups 38 differed in genetic composition, suggesting that they are closely related yet separable reproductive 39 subpopulations, with both contributing to the overall diversity of Lake Erie Walleye. Their 40 maintenance and integrity thus may bear management attention and further monitoring.

41

42 Running Head: Central Lake Erie Walleye Genetics

43 The Walleye Sander vitreus supports ecologically and economically valuable North American 44 commercial and recreational fisheries (Locke et al. 2005). In the Laurentian Great Lakes, Lake Erie 45 contains the largest population numbers and is known as the "Walleye capital of the world" (Scott and 46 Crossman 1973; Hartig et al. 2009). The combined commercial and recreational Walleye fisheries in Lake 47 Erie are worth ~\$2 billion U.S. per year (Gentner and Bur 2009), with an estimated harvest of 3.078 48 million Walleye in 2016, according to the Lake Erie Walleye Task Group of the Great Lakes Fishery 49 Commission (WTG 2017). The Lake Erie Walleye fishery is managed by binational inter-agency 50 cooperation and advised under the recommendations of the Interagency Walleye Task Group (WTG) to 51 the Lake Erie Committee of the Great Lakes Fishery Commission (WTG 2017). The WTG is guided by 52 the Joint Strategic Plan (JSP) for Management of Great Lakes Fisheries to cooperatively analyze fish 53 communities and fisheries in Lake Erie (Kayle et al. 2015). 54 A key objective of the Lake Erie Walleye Management Plan is to "Maintain and promote genetic 55 diversity by identifying, rehabilitating, conserving, and/or protecting locally adapted stocks" (Kayle et al. 56 2015). Likewise, across the entire Great Lakes, "Identifying and maintaining healthy Walleye genetic 57 stocks" constitutes a Great Lakes Fishery Commission management priority (Ryan et al. 2003). Walleye 58 populations from Lake Erie have been found to possess the greatest overall degree of genetic diversity 59 among all populations analyzed across the species' North American range using nuclear DNA 60 microsatellite loci, containing several genetically differentiated spawning groups (Stepien et al. 2009,

61 2015).

Spatial genetic differentiation and population structure characterize Walleye at their spring spawning
locations, with their returns to natal sites believed to be genetically based (summarized by Stepien et al.
2009, 2015). Otherwise, individual Walleye often travel widely during most of the year and stocks show

65 considerable admixture (Vandergoot et al. 2010). For example, Brenden et al. (2015) analyzed

66 microsatellite loci to determine origins of the Walleye recreational harvest in Saginaw Bay, Lake Huron,

67 estimating that 25% were individuals born in the Lake St. Clair/western Lake Erie system.

68 Walleye reproductive populations across the Great Lakes and throughout other parts of their native 69 range (encompassing much of northeastern and central North America) exhibit both broad-scale and fine-70 scale genetic differences, which appear to trace to their post-glacial recolonization patterns, physical and 71 behavioral barriers to migration, and natal homing (Stepien et al. 2012, 2015; Haponski and Stepien 72 2014a,b). For example, some spawning groups located less than 20 km apart in the Huron–Erie Corridor 73 (spanning the Lake Huron-St. Clair River-Lake St. Clair-Detroit River-western Lake Erie system) 74 significantly differed from each other in allelic composition, whereas others did not (Haponski and 75 Stepien 2014a). Genetic analysis of single nucleotide polymorphism (SNP) variation of Walleye from 76 several lakes and watersheds in Alberta Canada likewise found both broad- and fine-scale stock 77 differentiation among watersheds and lakes, with two nearby lakes housing a single interbreeding 78 population, and others showing genetic demarcation that did not reflect a genetic isolation by geographic 79 distance pattern (Allen et al. 2017). These patterns are very like those discerned for Walleye populations 80 across Lake Erie and the Great Lakes overall using microsatellite loci (summarized by Stepien et al. 81 2015), and as analyzed here to assess two spawning groups in Central Lake Erie for the present study. 82 Using otolith chemical signature assignments, Chen et al. (2017) found that strontium levels in 83 otolith cores significantly differed between Walleye spawning in the Sandusky and Maumee Rivers of 84 western Lake Erie, supporting natal homing. Stepien et al. (2012) identified lower genetic self-85 assignments for Walleve spawning in the Maumee River than for those reproducing in the nearby 86 Sandusky River (or in Van Buren Bay, which is in the Eastern Basin); this was attributed to large number 87 of spawners and the geographic proximity of nearly continuous reef and river spawning locations in 88 Western Lake Erie. This continuity likewise explained close genetic relationships among western Lake 89 Erie spawning groups (Maumee and Sandusky rivers and Western Basin Reefs) found with SNP analysis 90 (Chen 2016) and microsatellite results of Brenden et al. (2015) for the Lake St. Clair/Western Lake Erie 91 system. An acoustic telemetry study by Hayden et al. (2017) uncovered less fidelity (70%) of Walleye 92 tagged in the Maumee River (as opposed to 95% in the Tittabawassee River off Saginaw Bay, Lake 93 Huron), which they postulated is related to the close proximity of other spawning aggregations and sites 94 in Western Lake Erie. Walleye spawning in the Lake St. Clair/Western Lake Erie system significantly 95 differed from those spawning in the Tittabawassee River (Brenden et al. 2015), whose findings were 96 similar to the differentiation elucidated by Haponski and Stepien (2014a) for Walleye spawning in the 97 nearby Flint River, which both flow together through the Saginaw River into Saginaw Bay, Lake Huron. 98 Overall genetic relationships of Walleye spawning groups in Lake Erie showed greater site-specific

differentiation in the Eastern Basin than in the Western Basin (summarized by Stepien et al. 2015), withthe Central Basin spawning groups evaluated here.

Analyses of historic specimens of Lake Erie Walleye from the early to mid-20th century revealed 101 102 lower levels of genetic diversity than found at present, which likely stemmed from habitat loss, pollution, 103 and overexploitation (Haponski and Stepien 2014b). The lakewide Walleye population appears to have 104 largely recovered in genetic diversity during the past several decades (Haponski and Stepien 2014b; 105 Stepien et al. 2015). From the mid-1990s to the present, Walleye spawning groups in the Western and 106 Eastern Basins of Lake Erie were found to have each maintained their respective genetic consistency, 107 composition, and diversity (Stepien et al. 2012, 2015; Haponski and Stepien 2016). Spawning groups 108 from Lake Erie's Central Basin have not previously been investigated for fine-scale differences or genetic 109 continuity over time, largely due to the smaller size of the stocks, stochasticity in the runs (WTG 2017), 110 and corresponding lack of samples. Notably, electrofishing surveys suggest that the Grand River OH 111 spawning stock in the Central Basin now comprises fewer than 3,000 Walleye (C.T. Knight, Ohio 112 Division of Wildlife, unpublished data). This is the primary spawning group in the Central Basin, which 113 has been declining in numbers in comparison to those in the Western and Eastern Basins over the past two 114 decades (Isermann and Knight 2005; C. T. Knight, unpublished data), rendering the current investigation 115 of management interest. Such smaller populations and subpopulations are more likely to be negatively 116 affected by the loss of genetic variation due to harvest, meriting monitoring of genetic variation trends 117 over time (see Allendorf et al. 2013). Two key objectives set by the Great Lakes Fishery Commission 118 (GLFC 2015) in the current "Fishery Research Priorities for the Great Lakes" are: "What are the stock 119 structures of Walleve?" and "How can we identify, rehabilitate, conserve, or protect locally adapted 120 stocks?", indicating the importance of understanding the genetic diversity and relationship of this 121 spawning group to others across Lake Erie. It is essential to weigh genetic data carefully for management 122 decisions on stock designation (see Waples 1999; Allendorf et al. 2013), together with behavioral data 123 from tagging and telemetry studies, and results from ecological habitat and life history analyses. 124 The Ohio Division of Wildlife Fairport Harbor Fisheries Research Station thus requested an 125 evaluation by the Stepien laboratory to determine whether Lake Erie Central Basin Walleye that spawn in 126 the Grand River of Ohio and on the nearby Central Basin Reef (Figure 1) comprise genetically separable 127 stocks. If they differ, conservation of both spawning groups might be subject to the Walleye Task Group's 128 management priorities. Due to available samples and sample sizes, three Walleve spawning runs from 129 three years over a span of 16 years were compared from the Grand River, in order to assess their temporal 130 genetic compositions and relative diversity levels.

131

132 METHODS

133 Samples for the present study comprised Walleve fin clips ($\sim 2 \text{ cm}^2$ of pectoral or caudal fins) 134 collected via electrofishing by the Fairport Harbor Fisheries Research Station during spring spawning 135 runs at the Grand River (Ashtabula, OH at 41.85117° latitude, -81.23746° longitude) in 1996, 2003, and 136 2012, and by gill nets at the Central Basin Reef (just offshore at 41.76960° latitude, -81.22623° longitude) 137 in 2012, totaling 147 individuals (Table 1). The Grand River, Ohio, drains a watershed of 1,844 km², 138 where Walleye can utilize the lower 55 km up to Harpersfield Dam. The sampling location in this study 139 ranged from 2.2 to 7.2 km from the mouth of Lake Erie. Due to the low numbers of spawning Walleye 140 caught in the Grand River in 2012, we analyzed this spawning group both separately and together in 141 comparison to stored samples from two other years of Grand River spawning runs (1996 and 2003), in 142 order to evaluate its relationship to the Central Basin Reef run of 2012. Samples from other time points 143 for the Central Basin Reef run were unavailable. Hardy Point Reef (here termed the Central Basin Reef) is 144 located 4.6 km east of the Grand River, ~80m from shore, and is approximately 1.5 ha, characterized by a 145 mixture of boulder, rock, gravel, and sand with interstitial spaces for egg deposition. All Walleve 146 individuals analyzed were verified as being in spawning condition and when possible, were released after 147 fin-clipping. Fin clips were placed in 95% ethanol in labeled vials by ODW, and stored in the Stepien 148 laboratory.

Genomic DNA was extracted, purified, and amplified using the polymerase chain reaction (PCR) and allelic variation was assessed at 14 nuclear DNA microsatellite loci (See Table 2), following previously published procedures (Stepien et al. 2009, 2012; Haponski and Stepien 2014a, 2016). Amplification products were diluted 1:50, of which 1µl was added to 13µl of formamide and Applied Biosystems (ABI, Fullerton, CA) Gene Scan 500liz size standard in 96-well plates, denatured for 2 min at 95°C, and analyzed on an ABI 3130xl Genetic Analyzer with GENEMAPPER v3.7. Output profiles were checked manually to confirm allelic size variants.

All loci were evaluated for conformance to Hardy-Weinberg equilibrium expectations and linkage disequilibrium, using the Markov Chain Monte Carlo (MCMC) procedure with 10,000 dememorizations, 1,000 batches, and 10,000 iterations per batch in GENEPOP v4.0 (Rousset 2008). Levels of significance were adjusted with standard Bonferroni correction (Zar 1999). MICRO-CHECKER v2.2.3 (van

160 Oosterhout et al. 2004) was used to examine results for possible scoring errors, large allele dropout,

161 stuttering, and/or null alleles at each locus.

162 Per-locus calculations (Table 2) included: number of alleles (N_A) , inbreeding (F_{IS}) , overall genetic

163 deviation across all samples (F_{IT}), and divergence among samples (F_{ST}) in FSTAT v2.9.3.2 (Goudet

164 2002). Genetic diversity comparisons between the spawning sites and sampling years (Table 3) included

165 observed (H_O) heterozygosity +/- standard error (SE) and expected (H_E) heterozygosity (GENEPOP), F_{IS} ,

166 N_A , and allelic richness (A_R) +/- SE, which was adjusted for sample size with rarefaction (FSTAT). Paired

167 *t*-tests in R v3.0.1 (R Core Team 2015) were used to identify whether allelic richness and observed

168 heterozygosity values significantly differed between the spawning groups and sampling years. Number of

169 private alleles (N_{PA}) , i.e., those appearing unique to a given spawning group or sampling year in the

170 present study were identified with CONVERT v1.31 (Glaubitz 2004). Percentage of private alleles (P_{PA})

171 was determined by dividing the number of private alleles for a given sample by its total number of alleles.

172 Due to disparity in sample size, the rarefaction representation of private alleles was evaluated with the

173 program ADZEv1.0 (Szpiech et al. 2008a,b).

174 Pairwise genetic divergences among the spawning groups and sampling years were determined using 175 the F_{ST} analog θ_{ST} (Weir and Cockerham 1984) in FSTAT (Table 4), which is regarded as appropriate for 176 analyzing high gene flow species, small sample sizes, and unknown number of subpopulations 177 (Cockerham and Weir 1993; Waples 1999; Meirmans and Hedrick 2011), and to facilitate comparisons 178 with other studies. Since F-statistic estimates assume a normally distributed data set (Weir and 179 Cockerham 1984) and may be influenced by sample sizes (Raymond and Rousset 1995), we additionally 180 conducted pairwise exact tests of differentiation (χ^2) in GENEPOP, using MCMC chains of 10,000, 1000 181 batches, and 10,000 iterations. Probability values for both types of pairwise comparisons were adjusted 182 using sequential Bonferroni corrections (Rice 1989). This correction is regarded as a very conservative 183 approach that may preclude elucidation of significance when sample sizes are low, leading to type II error 184 (i.e., falsely rejecting the null hypothesis of no significant difference between samples; Cabin and Mitchill 185 2000; Moran 2003; Narum 2006). Thus, we reported significance values after (**) as well as prior to (*) 186 sequential Bonferroni correction, so that results on the borderline could be visualized (which may have 187 been influenced by sample size limitations), in order to aid the design of future studies (see Moran 2003). 188 Relationships among sampling years in the Grand River, combined years, and the Central Basin Reef 189 samples further were examined using three-dimensional factorial correspondence analysis (3d-FCA) 190 (Benzecri 1973) in GENETIX v4.05 (Belkhir et al. 2004), to facilitate visualization of spatial and 191 temporal trends. 192 In addition to the above frequency-based methods that used the sample as the unit of comparison, we

193 employed a Bayesian approach in STRUCTURE v2.3.3 (Evanno et al. 2005), which calculated likelihood 194 assignments of Walleve individuals to K=1-5 hypothetical population groups (the number of sampling 195 events +1) to determine the number of genetic stocks. Each K was run with 10 independent analyses, 196 burn-ins of 50,000 and 100,000 replicates, with and without the LOCPRIOR function. The Evanno et al. 197 (2005) ΔK method in STRUCTURE HARVESTER (Earl and vonHoldt 2012) was used to determine the 198 best-supported K. Individual assignments to each of the spawning groups and year samples additionally 199 were calculated using Bayesian analysis in GENECLASS2 (Piry et al. 2004), with the compute likelihood 200 and enable probability functions, 100,000 simulated individuals, Rannala and Mountain's (1997)

201 criterion, and Paetkau et al.'s (2004) simulation algorithm. GENECLASS was run in three separate

analyses: (A) for the 2012 spawning runs alone, (B) with combined years from the Grand River runs, and

203 (C) for the separate years of Grand River, with all being compared to the Central Basin Reef (2012)

sample.

205

206 **RESULTS**

207 All 14 nuclear DNA microsatellite loci conformed to Hardy-Weinberg equilibrium expectations, 208 except for a single sample at a single locus -Svi14 from the Central Basin Reef, which was attributed to 209 sampling and/or stochastic error. None of the loci exhibited linkage disequilibrium. MICRO-CHECKER 210 findings suggested slight homozygote excess at a few disparate loci in select samples, i.e., Svi6 (Grand 211 River, OH 1996), Svi18 (Grand River 2003 and 2012), Svi7 (Central Basin Reef), Svi14 (Central Basin 212 Reef), and SviL2 (Grand River 2012). However, there were no indications of null alleles or excess 213 homozygotes in other samples at these loci, or across the entire data set. All samples and all loci thus 214 were included in all analyses.

215 Overall, 185 alleles were recovered from 147 Walleye individuals at the 14 microsatellite loci,

216 ranging from 6 (Svi18) to 25 (Svi14) alleles per locus (Table 2). Loci showing the greatest F_{ST}

divergences among the samples were *Svi*33 (0.016) and *Svi*L4 (0.018). Observed heterozygosity (Table 3)

218 appeared slightly higher for the Grand River spawning samples overall (0.76+/-0.03) than in the Central

Basin Reef spawning population (0.73+/-0.03); these values did not significantly differ. For individual

220 years sampled in the Grand River, the earliest sample from 1996 possessed the greatest heterozygosity

221 (0.78+/-0.04), which decreased to 0.76+/-0.03 in 2003 and 0.73+/-0.05 in 2012 (Table 3; these values

222 were not statistically different). Heterozygosity also did not differ between the 2012 samples in the Grand

River versus the Central Basin Reef. Overall number of alleles appeared greater in the Grand River (171)

than in the Central Basin Reef (155), with the earlier to later samples from the Grand River appearing to

decline over time. Allelic richness (A_{R} ; which was adjusted for sample size) was significantly higher for

the Grand River spawning group overall (8.38) than in the Central Basin Reef sample (7.69; $p=0.003^{**}$).

227 Allelic richness comparisons showed some borderline significance (before sequential Bonferroni

correction) between the Central Basin Reef (sample from 2012) and the Grand River spawning population

in 1996 ($p=0.009^*$) and 2003 ($p=0.008^*$), but no difference in 2012. No significant differences in allelic

richness occurred between sampling years for the Grand River population. The percentage of private

- alleles was 18% for the Grand River overall and 9% for the Central Basin Reef; when adjusted for
- rarefaction (Szpiech et al. 2008), the number of private alleles per locus significantly differed (Grand
- River population overall =2.18+/-0.32, Central Basin Reef=1.37+/-0.29; p=0.022*). Comparisons of
- private alleles between sampling years within the Grand River population were not significant, and those

for individual sampling years versus the Central Basin Reef also did not significantly differ. Estimates of

236 F_{IS} suggested slight inbreeding depression in all samples (F_{IS} =0.013–0.066), which were significant

except for the Grand River in 1996.

238 Pairwise F_{ST} analog and exact tests showed significant genetic divergence between Walleve 239 spawning groups in the Grand River overall versus the Central Basin Reef, as well as between the two 240 groups in 2012 (Table 4). Exact tests revealed significant differences between those reproducing at the 241 Central Basin Reef in 2012 versus those from the Grand River for the 1996 and 2003 individual sampling 242 years; the 1996 spawning group also was significant in the $F_{\rm ST}$ comparison. Between Grand River 243 sampling years, the middle (2003) versus the latest (2012) samples showed slight yet insignificant 244 variation with both $F_{\rm ST}$ analog and exact tests, and the 1996 and 2003 samples differed using the exact 245 test alone (Table 4).

Genetic differences between the Walleye spawning in the Grand River and the Central Basin Reef were further depicted by 3d-FCA (Figure 2), with the temporal Grand River samples all clustering closer to one another and separately from the Central Basin Reef sample. Among the Grand River samples, those from 2003 and 2012 diverged the most, as also indicated by the exact tests of differentiation. The three axes of the 3d-FCA explained 100% of the data (Figure 2).

251 STRUCTURE and STRUCTURE HARVESTER analyses indicated that the number of genetic 252 population groups (stocks) was K=2 (Figure 3), supported by delta K results, and other K alternatives 253 were not supported. Results showed genetic difference between individuals spawning in the Grand River 254 (colored dark grey) and at the Central Basin Reef (colored light grey). All of the Grand River individuals showed strongest assignments to the Grand River (dark grey). Self-assignments to the Grand River 255 256 appeared greatest for the individuals spawning in 1996 (averaging ~97%), followed by those in 2003 257 (averaging ~92%), and then 2012 (averaging ~90%). Individuals spawning on the Central Basin Reef 258 averaged about 60% assignment to the Central Basin Reef (light grey) and about 40% assignment to the 259 Grand River (dark grey).

260 GENECLASS assignment tests for the two spawning groups in 2012 discerned 100% self-261 assignment of individuals from the Grand River to the Grand River, with none assigning to the Central 262 Basin Reef (Table 5A). All but two of the 56 individual Walleye sampled in 2012 that spawned at the 263 Central Basin Reef self-assigned to the Central Basin Reef (totaling 97%), with just 3% mis-assigning to 264 the Grand River (Table 5A). When all three spawning run years for the Grand River samples were 265 combined, 100% of the individuals spawning in the Grand River overall self-assigned to the Grand River, 266 with none mis-assigning to the Central Basin Reef (Table 5B). However, when including multiple 267 sampling years for the Grand River (Table 5C), 47% of the Central Basin Reef samples self-assigned to

the Central Basin Reef and 53% mis-assigned to the Grand River (including 3% to the Grand River sample from 2012, 17% to the sample from 2003, and 33% to the sample from 1996).

270

271 **DISCUSSION**

The research objective was to provide genetic assessment to the WTG and the Ohio Division of Wildlife Fairport Fiseries Research Station of the relationship between two nearby Walleye spawning groups in Lake Erie's Central Basin. Our results indicated that Walleye spawning in the Grand River appear to genetically differ from those reproducing at the nearby Central Basin Reef. All Walleye spawning in the Grand River self-assigned to the Grand River, and 97% of individuals sampled on the Central Basin Reef in 2012 self-assigned. These data indicate that the groups likely comprise separable reproductive stocks, which may merit management attention.

279 This research found that mis-assignments were rare for the Central Basin individuals when just the 280 2012 spawners were included, with only 3% mis-assigning to those spawning in the Grand River in 2012. 281 However, when the other temporal runs in the Grand River were considered, over half of the Central 282 Basin Reef samples then mis-assigned to the Grand River – especially to the 1996 spawners from the 283 Grand River. There are several possible explanations for these findings, including: (1) The Central Basin 284 Reef population may have been historically derived from the Grand River population, (2) Walleve that 285 were born in the Grand River may occasionally spawn on the Central Basin Reef, but not vice-versa, 286 and/or (3) Some of the Walleye caught on the Central Basin Reef in spawning condition may have been 287 en route to spawning in the Grand River. It may be that some of the population that once spawned on the 288 Grand River is now spawning on the Central Basin Reef or may reproduce at both locations. 289 Alternatively, some individuals may travel back and forth between these locations before or after 290 spawning (and might have been inadvertently sampled here), for which behavioral data from telemetry 291 studies may be very useful. Further work involving tagging and telemetry, coupled with genetic and 292 otolith signature analyses would help to resolve these questions. Given that the Ohio Division of Wildlife 293 (ODW) has found that the Grand River spawning run has declined in numbers over the past two decades 294 (C. T. Knight, Ohio Division of Wildlife, unpublished data), our baseline genetic information may 295 provide an important gauge to monitor its future success. 296 In a previous study, some of the Walleye spawning groups across the Huron–Erie Corridor were

296 In a previous study, some of the Walleye spawning groups across the Huron–Erie Corridor were 297 discerned to significantly diverge from one another (with levels of genetic difference comparable to that 298 found here between the Grand River and Central Basin Reef samples); these patterns did not correspond 299 to genetic isolation by geographic distance (Haponski and Stepien 2014). Similarly, small yet significant 300 divergences differentiated among closely located populations of the Walleye's congener, the European 301 Pikeperch *S. lucioperca*, in the Baltic Sea (Björkland et al. 2007). Very significant divergences have been

302 discerned among spawning Lake Erie groups of another percid fishery, the related Yellow Perch *Perca*

303 *flavescens*, including Central Lake Erie groups that were located near to the Walleye sampled in the

304 present study (Sepulveda-Villet and Stepien 2011; Kocovsky et al. 2013; Sullivan and Stepien 2015).

Moreover, Yellow Perch spawning adults in Central Lake Erie exhibited morphometric differences
among spawning locations (Kocovsky et al. 2013). Fine-scale divergences of Walleye and Yellow Perch
across their ranges do not correspond to geographic proximity; they instead appear to reflect historic and
behavioral homing patterns (Stepien et al. 2015). Similar to Yellow Perch, populations of the congeneric
European Perch *P. fluviatilis* showed significant divergence between spawning groups located only about
one km apart in Lake Erken, Sweden (Bergek and Olsson 2009).

311 In the present study, Walleye spawning in the Grand River had slightly higher overall genetic 312 diversity (allelic richness and private alleles) than did the Central Basin Reef population. Levels of 313 genetic diversity in these Central Basin Walleye spawning groups were similar to those of other Walleye 314 reproductive groups throughout the Great Lakes, using these same loci in studies also conducted by our 315 laboratory (mean observed heterozygosity (H_0)=0.72+/-0.04; Haponski and Stepien (2014a, b, 2016). 316 Allelic richness of Walleve spawning in the Grand River declined over the timescale of this study 317 (decreasing from 1996 to 2003) and other diversity measures (observed heterozygosity and private 318 alleles) appeared to follow a similar trend but were not significant (likely due to sample size limitations). 319 This decrease appears to coincide with reduced numbers of Walleye spawning in the Grand River, 320 discerned by the Ohio Division of Wildlife Fairport Harbor Fisheries Research Station over the past two 321 decades (C.T. Knight, Ohio Division of Wildlife, unpublished observations). Some factors that may have 322 in the Grand River influenced walleye habitat and populations include a "500-year" flood event in July 323 2006, which moved mobile substrate and re-channelized some areas, along with overall warming 324 temperatures, increasing hypoxia, and increased harmful algal blooms across Lake Erie (see WTG 2017). 325 Influences of these latter factors on Lake Erie tributary habitats, including the Grand River, tend to be 326 more rapid and stochastic than those in Lake Erie proper.

327 In comparison, temporal genetic analyses of other Walleye spawning runs in Lake Erie revealed 328 overall genetic consistency in diversity and allelic composition over time (the last two decades) in the 329 Western Basin (Maumee River and Sandusky Rivers; Stepien et al. 2012; Haponski and Stepien 2016) 330 and the Eastern Basin (Van Buren Bay and Cattaraugus Creek; Stepien et al. 2012; Haponski et al. (2014). 331 Thus, although other spawning stocks in Lake Erie have maintained consistent levels of genetic diversity 332 over the past two decades, this may not be the case for Walleye in the Central Basin. Results of the 333 present study may indicate need for continued monitoring and attention by the Walleye Task Group. 334 A study of genetic variation for 420 homologous single nucleotide polymorphisms (SNPs) described 335 decline in overall genetic diversity of Walleve from Smoke Lake, Alberta Canada in comparisons from

1973 (13 individuals) versus 2005 (19 individuals), which appeared to be linked to high fishery harvest exploitation and the collapse of the fishery (Allen et al. 2017). Like the present results for Walleye spawning in the Grand River, Allen et al. (2017) found a decrease in overall percentage of private alleles and observed heterozygosity. It thus appears warranted to continue monitoring the numbers, genetic diversities, and compositions of the Lake Erie Grand River and Central Basin Reef Walleye spawning groups in the future.

342 In contrast to Walleye, the genetic compositions of Yellow Perch spawning groups in Lake Erie 343 (Sullivan and Stepien 2015) as well as those of European Perch in Lake Erken, Sweden, have varied as 344 much temporally as spatially (Bergek and Olsson 2009). Yellow Perch spawning groups appear to exhibit 345 less fidelity to specific locations from year to year than do Walleye; thus, their genetic divergence 346 patterns fluctuate (Sullivan and Stepien 2015; Stepien et al. 2015). In comparison, large and significant 347 temporal population genetic variation changes in Atlantic Cod Gadus morhua of the North Sea have been 348 associated with extreme stock declines due to overfishing and subsequent increased immigration from 349 other populations (Hutchinson et al. 2003). Such overall stock decline factors do not appear to be the case 350 at present for the Lake Erie Walleye and Yellow Perch fisheries (see Stepien et al. 2015). Whether and 351 how population genetic relationships and stock continuities coincide with management practices for 352 Walleve and Yellow Perch, versus those for other species, is a matter for further investigation.

353 The present study provides new insight into the divergent genetic compositions of Walleye spawning 354 in geographically close but physically different habitats. Overall, both spawning groups of Walleye 355 appear genetically diverse and different; such smaller reproductive subpopulations may significantly 356 contribute to Lake Erie's stock structure as a whole and point to a need for additional surveillance. It is 357 important that future studies investigate these runs over multiple years with larger sample sizes 358 (preferably with non-invasive sampling, such as environmental DNA), and interpret the data in light of 359 Walleye tagging and telemetry studies, life history, and reproductive behavior. In relation to the Lake Erie 360 Walleye Management Plan objective to "Maintain and promote genetic diversity by identifying, 361 rehabilitating, conserving, and/or protecting locally adapted stocks" (Kayle et al. 2015), here we 362 identified two apparently locally differentiated stocks, which merit continued monitoring and possible 363 genetic conservation. Lake Erie Walleye are managed lakewide as a single population, yet improved 364 understanding of the numbers, abundances, and diversity of stocks is critical to managers. Although 365 managers might be hard pressed to limit harvest of an individual stock of lake caught fish, they may be 366 able to further monitor spawning habitat and exploitation in specific areas, where warranted. 367 Identification of stocks needs to be accomplished before management decisions can be considered. Future 368 sampling and analyses of these spawning groups in the Grand River and at the Central Basin Reef should 369 be undertaken to evaluate potential changes to their temporal genetic dynamics and spatial structure.

371

372 ACKNOWLEDGEMENTS

373

374	This is contribution 4640 from NOAA PMEL. Grant awards to CAS from the NOAA Ohio Sea Grant
375	Program R/LR-013, "Temporal and spatial analyses of Walleye and Yellow Perch genetic stock structure:
376	A high-resolution database for fishery management" and USEPA GLRI grants GL-00E01149-0 and GL-
377	00E01898 funded the study, whose data collection was performed at the Stepien laboratory's previous
378	location at the University of Toledo and the analyses and writing at NOAA PMEL. We thank the Ohio
379	Department of Natural Resources' Division of Wildlife, especially J. Deller, for contributing samples.
380	Microsatellite data collection and analyses were aided by Stepien lab members A. Haponski, E. Kramer,
381	S. Yerga-Woolwine, D. Eddins, and A. Elz. In addition, we thank A. Elz for helping to edit the
382	manuscript at PMEL. We appreciate logistic support from R. Lohner, S. McBride, and P. Uzmann.
383	
384	REFERENCES
385	
386	Allen, B., E. Bowles, M. Morris, and S. Rogers. 2017. Loss of SNP genetic diversity following
387	population collapse in a recreational Walleye (Sander vitreus) fishery. Canadian Journal of
388	Fisheries and Aquatic Sciences. Published on the web. Dec 18, 2017. Available:
389	http://www.nrcresearchpress.com/doi/10.1139/cjfas-2017-0164#.WkP1yRNSzUJ
390	(December 2017).
391	Allendorf, F. W., G. Luikart, and S. N. Aitken. 2013. Ch. 18, Exploited populations. Pages 377-395 in
392	Conservation and the genetics of populations. Second edition. Wiley-Blackwell. West Sussex,
393	United Kingdom.
394	Belkhir, K., P. Borsa, L. Chikhi, N. Raufaste, and F. Bonhomme. 2004. GENETIX v.4.05: logiciel sous
395	Windows TM pour la génétique des populations. [GENETIX 4.05: Windows software for
396	population genetics.] Laboratoire Génome, Populations, Interactions, CNRS UMR 5171,
397	Université de Montpellier II, Montpellier, France. Available: <u>http://www.genetix.univ-</u>
398	montp2.fr/genetix/genetix.htm. (December 2017).
399	Benzecri, J. P. 1973. L'analyse des donnees, tome II. L'analyse des correspondances. [The analysis of
400	data, volume II. The analysis of correspondence.] Dunod Press, Paris, France.
401	Bergek, S., and J. Olsson. 2009. Spatiotemporal analysis shows stable genetic differentiation and barriers
402	to dispersal in the Eurasian Perch (Perca fluviatilis L.). Evolutionary Ecology Research 11:827-
403	840.

Björkland, M., T. Aho, and L. C. Larsson. 2007. Genetic differentiation in Pikeperch (Sander lucioperca):
the relative importance of gene flow, drift and common history. Journal of Fish Biology 71:264-
278.
Borer, S., L. M. Miller, and A. R. Kapuscinski. 1999. Microsatellites in Walleye Stizostedion
vitreum. Molecular Ecology 8:336–338.
Brenden, T. O., K. T. Scribner, J. R. Bence, I. Tsehaye, J. Kanefsky, C. S. Vandergoot, and D. G. Fielder.
2015. Contributions of Lake Erie and Lake St. Clair Walleye populations to the Saginaw Bay,
Lake Huron, recreational fishery: Evidence from genetic stock identification. North American
Journal of Fisheries Management 35(3):567–577.
Cabin, R. J., and R. J. Mitchill. 2000. To Bonferroni or not to Bonferroni: when and how are the
questions. Bulletin of the Ecological Society of America. 81(3):246–248.
Chen, KY. 2016. Lake Erie Walleye population structure and stock discrimination methods.
Ph.D. Dissertation, The Ohio State University, Columbus, Ohio. Available:
https://etd.ohiolink.edu/!etd.send_file?accession=osu1469177034&disposition=inline
(December 2017).
Chen, KY., S. A. Ludsin, M. M. Corey, P. D. Collingsworth, M. K. Nims, J. W. Olesik, K.
Dabrowski, J. J. van Tassell, and E. A. Marschall. 2017. Experimental and field evaluation of
otolith strontium as a marker to discriminate between river-spawning populations of Walleye in
Lake Erie. Canadian Journal of Fisheries and Aquatic Sciences. 74:693–701.
Cockerham, C. C. and B. S. Weir. 1993. Estimation of gene flow from G-statistics. Evolution 47:
855-863.
Earl, D. A., and B. M. Vonholdt 2012. STRUCTURE HARVESTER: a website and program for
visualizing STRUCTURE output and implementing the Evanno method. Conservation Genetics
Resources 4:359–361. Available: <u>http://taylor0.biology.ucla.edu/structureHarvester/</u> (December
2017).
Eldridge, W. H., M. D. Bacigalupi, I. R. Adelman, L. M. Miller, and A. R. Kapuscinski. 2002.
Determination of relative survival of two stocked Walleye populations and resident
natural-origin fish by microsatellite DNA parentage assignment. Canadian Journal of
Fisheries and Aquatic Sciences 59:282–290.
Evanno, G., S. Regnaut, and J. Goudet. 2005. Detecting the number of clusters of individuals using the
software STRUCTURE: a simulation study. Molecular Ecology 14:2611–2620.
Available: https://web.stanford.edu/group/pritchardlab/structure.html (December 2017).

437 Excoffier, L., and H. E. L. Lischer. 2010. ARLEQUIN suite v3.5: A new series of programs to perform 438 population genetics analyses under Linux and Windows. Molecular Ecology Resources 10:564-439 567. Available: http://cmpg.unibe.ch/software/arlequin35/ (December 2017). 440 Gentner, B., and M. Bur. 2009. Economic damages of impingement and entrainment of fish, fish eggs, 441 and fish larvae at the Bay Shore Power Plant. Gentner Consulting Group Report. Silver Spring, 442 Maryland. 443 Glaubitz, J. C. 2004. CONVERT: A user-friendly program to reformat diploid genotypic data for 444 commonly used population genetic software packages. Molecular Ecology Notes 4:309–310. 445 Goudet, J. 2002. FSTAT v.2.9.3.2. Available: http://www2.unil.ch/popgen/softwares/fstat.htm. 446 (December 2017). 447 Great Lakes Fishery Commission (GLFC). 2015. Fishery research priorities for the Great Lakes. Great 448 Lakes Fishery Commission. Ann Arbor, Michigan. 14 pp. Available: 449 http://www.glfc.org/pubs/pdfs/research/Basin_Wide_Priorities.pdf (December 2017) 450 Haponski, A. E., and C. A. Stepien. 2014a. Genetic connectivity and diversity of Walleye (Sander 451 vitreus) spawning groups in the Huron-Erie Corridor. Journal of Great Lakes Research 40:89– 452 100. 453 Haponski, A.E., and C.A. Stepien. 2014b. A population genetic window into the past and future of the 454 Walleye Sander vitreus: relation to historic Walleye and the extinct "Blue Pike" S. v. "glaucus". 455 BMC Evolutionary Biology 14:133. 456 Available: https://link.springer.com/content/pdf/10.1186%2F1471-2148-14-133.pdf 457 (December 2017). Haponski, A. E., and C. A. Stepien. 2016. Two decades of genetic consistency in a reproductive 458 459 population in the face of exploitation: Patterns of adult and larval Walleye (Sander vitreus) from 460 Lake Erie's Maumee River. Conservation Genetics 17:1345–1362. 461 Haponski, A. E., H. Dean, B. E. Blake, and C. A. Stepien. 2014. Genetic history of Walleye (Sander 462 *vitreus*) spawning in Lake Erie's Cattaraugus Creek: A comparison of pre– and post–stocking. 463 Transactions of the American Fisheries Society 143:1295–1307. 464 Hartig, J. H., M. A. Zarull, J. J. H. Ciborowski, J. E. Gannon, E. Wilke, G. Norwood, and A. N. Vincent. 465 2009. Long-term ecosystem monitoring and assessment of the Detroit River and Western Lake 466 Erie. Environmental Monitoring and Assessment 158:87–104. 467 Hayden, T. A, T. R. Binder, C. M. Holbrook, C. S. Vandergoot, D. G. Fielder, S. J. Cooke, J. M. 468 Dettmers, and C. C. Krueger. 2017. Spawning site fidelity and apparent annual survival of 469 Walleye (Sander vitreus) differ between a Lake Huron and Lake Erie tributary. Ecology of 470 Freshwater Fish 2017:1–11.

471	Hutchinson, W. F., C. van Oosterhout, S. I. Rogers, and G. R. Carvalho. 2003. Temporal analysis of
472	archived samples indicates marked genetic changes in declining North Sea Cod (Gadus morhua).
473	Proceedings of the Royal Society of London Biological Sciences 270:2125-2132.
474	Isermann, D.A., and C. T. Knight. 2005. Potential effects of jaw tag loss on exploitation estimates for
475	Lake Erie Walleyes. North American Journal of Fisheries Management 25(2):557–562.
476	Kayle, K. C. Murray, J. Francis, and J. Markham. 2015. Lake Erie Walleye management plan (2015–
477	2019). Great Lakes Fishery Commission. Ann Arbor, Michigan. 39 pp. Available:
478	http://www.glfc.org/lakecom/lec/LEC_docs/position_statements/Walleyemanagment_plan.pdf
479	(December 2017).
480	Kocovsky, P., T. J. Sullivan, C. T. Knight, and C. A. Stepien. 2013. Genetic and morphometric
481	differences demonstrate fine-scale population substructure of the Yellow Perch Perca
482	flavescens: Need for redefined management units. Journal of Fish Biology 82:2015–2030.
483	
484	Locke, B., M. Belore, A. Cook, D. Einhouse, R. Kenyon, R. Knight, K. Newman, P. Ryan, and E. Wright.
485	2005. Lake Erie Walleye management plan. Great Lakes Fishery Commission, Ann Arbor,
486	Michigan.
487	Meirmans, P. G., and P. W. Hedrick. 2011. Assessing population structure: F_{ST} and related measures.
488	Molecular Ecology Resources 11:5–18.
489	Moran, M. D. 2003. Arguments for rejecting the sequential Bonferroni in ecological studies. Oikos
490	100(2):403-405.
491	Narum, S. R. 2006. Beyond Bonferroni: less conservative analyses for conservation genetics.
492	Conservation Genetics 7(5):783–787.
493	Paetkau, D., R. Slade, M. Burden, and A. Estoup. 2004. Genetic assignment methods for the direct, real-
494	time estimation of migration rate: a simulation-based exploration of accuracy and power.
495	Molecular Ecology 13:55–65.
496	Piry, S., A. Alapetite, JM. Cornuet, D. Paetkau, L. Baudouin, and A. Estoup. 2004. GENECLASS 2: A
497	software for genetic assignment and first-generation migrant detection. Journal of Heredity
498	95:536–539. Available: <u>http://www1.montpellier.inra.fr/URLB/GeneClass2</u> (December 2017).
499	Raymond, M., and F. Rousset. 1995. An exact test for population differentiation. Evolution 49:1280-
500	1283.
501	R Core Team. 2015. R: a language for statistical computing. R Foundation for Statistical Computing,
502	Vienna, Austria. Available: <u>https://www.r-project.org/</u> (December 2017)
503	Rice, W. R. 1989. Analyzing tables of statistical tests. Evolution 43:223–225.

- Rousset, F. 2008. GENEPOP '007: a complete reimplementation of the GENEPOP software for Windows
 and Linux. Molecular Ecology Research 8:103–106. Available: <u>http://kimura.univ-</u>
 montp2.fr/~rousset/Genepop.htm. (December 2017).
- Ryan, P. A., R. Knight, R. MacGregor, G. Towns, R. Hoopes, and W. Culligan. 2003. Fish-community
 goals and objectives for Lake Erie. Great Lakes Fishery Commission, Special Publication 03–02,
 Ann Arbor, Michigan.
- Scott, W. B., and E. J. Crossman. 1973. Freshwater fishes of Canada. Bulletin of the Fisheries Research
 Board of Canada 184:767–774.
- 512 Sepulveda-Villet, O. J., and C. A. Stepien. 2011. Fine-scale population genetic structure of the Yellow
 513 Perch *Perca flavescens* in Lake Erie. Canadian Journal of Fisheries and Aquatic Sciences
 514 68:1435–1453.
- Stepien, C. A., D. J. Murphy, R. N. Lohner, O. J. Sepulveda-Villet, and A. E. Haponski. 2009. Signatures
 of vicariance, postglacial dispersal and spawning philopatry: population genetics of the Walleye
 Sander vitreus. Molecular Ecology 18:3411–3428.
- Stepien, C. A., J. A. Banda, D. M. Murphy, and A. E. Haponski. 2012. Temporal and spatial genetic
 consistency of Walleye spawning groups. Transactions of the American Fisheries
 Society 141(3):660–672.
- Stepien, C. A., O. J. Sepulveda-Villet, and A. E. Haponski. 2015. Comparative genetic diversity,
 population structure, and adaptations of Walleye and Yellow Perch across North America.
- 523 Chapter 25, Pages 643–690 *in* P. Kestlemont, K. Dabrowski, and R. C. Summerfelt, editors. Biology and
 524 culture of Percid fishes-principles and practices. Springer Associates, Berlin, Germany.
- Sullivan, T. J., and C. A. Stepien. 2015. Temporal population genetic structure of Yellow Perch spawning
 groups in the lower Great Lakes. Transactions of the American Fisheries Society 144:211–226.
- 527 Szpiech, Z.A., M. Jakobson, and N.A. Rosenberg. 2008a. ADZE: A rarefaction approach for counting
 528 alleles private to combinations of populations. Bioinformatics. 24(21):2498-2505.
- 529 Szpiech, Z.A., M. Jakobson, and N.A. Rosenberg. 2008b. ADZE: Allelic diversity analyzer, version 1.0.
 530 Available: <u>https://web.stanford.edu/group/rosenberglab/adze.html</u>
- 531 (December 2017)
- 532 Vandergoot, C. S., H. A. Cook, M. V. Thomas, D. W. Einhouse, and C. Murray. 2010. Status
- 533 of Walleye in western Lake Erie, 1985–2006. Pages 123–150 in Roseman, E., P. Kocovsky, and
- 534 C. Vandergoot, editors. Status of Walleye in the Great Lakes: proceedings of the 2006
- 535 symposium. Great Lakes Fisheries Commission Technical Report. Great Lakes Fishery
- 536 Commission. Ann Arbor, Michigan. 69:123-150.

537	van Oosterhout, C. V., V	V. F. Hutchinson, D. P. M. Wills, and P	. Shipley. 2004. MICRO-CHECKER:
538	software for ide	ntifying and correcting genotyping error	rs in microsatellite data. Molecular
539	Ecology Notes.	4:535–538.	
540	Available: http://	//www.norwichresearchpark.com/ourres	search/researchgroups/elsa/software/micro
541	<u>checker.aspx</u> (D	ecember 2017).	
542	Waples, R. S. 1999. Se	parating the wheat from the chaff: patte	erns of genetic differentiation in high gene
543	flow species. Jo	urnal of Heredity 89(5):438-450.	
544	Weir B. S., and C. C. Co	ockerham. 1984. Estimating F-statistics	for the analysis of population structure.
545	Evolution 38:13	58–1370.	
546	Wirth, T., R. Saint-Laur	ent, and L. Bernatchez. 1999. Isolation	and characterization of microsatellite loci
547	in the Walleye (Stizostedion vitreum), and cross-species	amplification within the family Percidae.
548	Molecular Ecol	ogy 8:1960–1962.	
549			
550	WTG (Walleye Task G	oup of the Lake Erie Committee, Great	Lakes Fishery Commission),
551	2017. Report fo	2016 by the Lake Erie Walleye Task C	Group, 2017. Great Lakes Fishery
552	Commission. Y	psilanti, Michigan. 25 pp.	
553	Available: <u>http</u>	//www.glfc.org/pubs/lake_committees/	erie/WTG_docs/annual_reports/WTG_rep
554	<u>ort_2017</u> .		
555	pdf (December	2017).	
556	Zar, J. H. 1999. Biostati	stical Analysis, 4th edition. Prentice Ha	ll, Inc., Upper Saddle River, New
557	Jersey.		
558			
559	TABLE 1. Collection in	formation including date, location, num	ber of individuals sampled (<i>N</i>), and
560	sample set designation f	or spawning Walleye collected by the C	Dhio Division of Wildlife Fairport Harbor
561	Fisheries Research Stati	on.	
	Collection date	Location	<i>N</i> Spawning group sample
	03/27-04/15/1996	Grand River east of Fairport Harbor,	30 Grand River (GR) 1996
		ОН	
		41.75117°lat., -81.23746°long.	

V	41.75117°lat., -81	.23746°long.		
04/02-06/12/2003	11 66		30	Grand River (GR) 2003
03/28-04/03/2012			16	Grand River (GR) 2012
03/20-05/02/2012	Central Lake Erie	Reef, off	58	Central Basin Reef (RF) 2012

568

563 TABLE 2. Summary of genetic variation per microsatellite locus for Grand River and Central Basin Reef

564 Walleye spawning groups. Table shows primer reference source, annealing temperature (T_A) , number of

alleles (N_A), allelic size range (nucleotides), genetic deviation across all combined samples (F_{IT}), mean

566 genetic divergence (F_{ST}), and inbreeding coefficient (F_{IS} , average divergence within a spawning group),

567 calculated using FSTAT.

Locus	Source	$T_{\rm A}$ (°C)	N_{A}	Size range	$F_{\rm IT}$	$F_{\rm ST}$	$F_{\rm IS}$
Svi4	Borer et al. (1999)	60	8	96-120	-0.052	0.005	-0.057
Svi6		60	14	140-166	0.039	0.015	0.025
Svi17	"	54	10	102-142	0.088	0.005	0.084
Svi18	"	65	6	114-126	0.203	0.000	0.211
Svi33	"	60	14	78-106	0.030	0.016	0.015
SviL2	Wirth et al. (1999)	53	7	263-281	-0.014	0.000	-0.009
SviL3		53	16	233-263	-0.067	0.008	-0.076
SviL4		54	15	121-161	0.115	0.018	0.099
SviL6	"	54	11	108-136	0.014	0.009	0.004
SviL7	"	54	20	174-236	-0.005	0.000	-0.004
Svi2	Eldridge et al. (2002)	60	12	190-220	0.071	0.007	0.064
Svi7	"	60	9	154-172	0.137	0.000	0.139
Svi14	"	54	25	154-214	0.119	0.005	0.115
Svi20	"	50	18	152-190	0.040	0.011	0.043
Total			185		0.049	0.006	0.043

569

570 TABLE 3. Summary of genetic data from spawning Walleye samples in the Grand River and Central

571 Basin Reef including: number of individuals (*N*), observed heterozygosity (H_0) ± standard error (SE),

572 number of alleles (N_A), number of private alleles (N_{PA}), percentage of private alleles (P_{PA}), allelic

573 richness ($A_R \pm SE$; adjusted for sample size), and inbreeding coefficients (F_{IS}) calculated from the 14

574 nuclear DNA microsatellite loci using FSTAT.

Sample		Ν	H ₀ ±SE	$N_{\rm A}$	N_{PA}	$P_{\rm PA}$	$A_{\rm R}\pm {\rm SE}$	$F_{\rm IS}$
Grand River	1996	30	0.78±0.04	147	13	0.05	8.50±0.83	0.013

	2003	30	0.76 ± 0.03	135	5	0.02	8.18±0.69	0.045	
	2012	16	0.73 ± 0.05	116	4	0.02	8.11±0.71	0.066	
Total Grand River		76	0.76 ± 0.03	171	30	0.18	8.38±0.73	0.039	
Central Basin Reef	2012	58	0.73 ± 0.03	155	14	0.09	7.69 ± 0.68	0.051	

576 TABLE 4. Pairwise genetic comparisons between Walleye spawning samples from Grand River (all years

577 combined and separate years (1996, 2003, 2012)) vs. the Central Basin Reef (2012). F_{ST} analog

578 (calculated in FSTAT; below diagonal) and χ^2 (calculated in GENEPOP2; above diagonal). *=significant

579 before, but not after, Bonferroni correction. **= significant difference remained after sequential

580 Bonferroni correction. NS= not significant.

10

Site	GR All	GR 1996	GR 2003	GR 2012	RF 2012
(N)	(76)	(30)	(30)	(16)	(58)
Grand River (GR) All	~	NS	NS	NS	**
GR 1996	0.000	~	*	NS	**
GR 2003	0.000	0.001	~	*	**
GR 2012	0.000	0.004	0.010*	~	*
Central Basin Reef (RF)	0.006**	0.008**	0.004	0.009*	~

582

TABLE 5. GENECLASS assignment test results for Walleye spawning samples from Grand River vs. the
 Central Basin Reef in 2012 (Percentage assignments are in parentheses). A. 2012 spawning runs alone, B.

585 All sampling years combined for the Grand River, and C. Separate Grand River spawning runs (1996,

586 2003, 2012). Self-assignments are *in italics*.

587 588

A. Samples from the 2012 spawning run

	Assigned to				
Sample	GR 2012	RF 2012			
Grand River (GR) 2012	16 (1.00)	~			
Reef (RF) 2012	2 (0.03)	56 (0.97)			
	. ,	,			
B. All sampling years co	mbined for	the Grand			

	Assigned to			
Population	GR	RF		

Grand River (GR)	76 (1.00)	~
Reef (RF) 2012	<u>31 (0.53)</u>	<u>27 (0.47</u>)

- 591
- 592
- 593 C. Separate Grand River spawning run samples and Central Basin Reef 2012
- 594

		Assigned to					
	Sample	GR 1996	GR 2003	GR 2012	RF 2012		
	Grand River (GR) 1996	20 (0.67)	8 (0.27)	2 (0.07)	~		
	GR 2003	11 (0.37)	17 (0.57)	2 (0.07)	~		
	GR 2012	2 (0.13)	4 (0.25)	10 (0.63)	~		
	Reef (RF) 2012	19 (0.33)	10 (0.17)	2 (0.03)	27 (0.47)		
595							
596							
597							
598							
599	Figure Captions						
600							
601	FIGURE 1. Sampling locat	tions of the Wa	alleye spawı	ning groups	in Central L	ake Erie.	
602							
603	FIGURE 2. Three–Dimens	ional Factoria	l Correspond	lence analys	is illustratin	g relationships among	,
604	Walleye spawning	groups per san	npling year f	rom the Gra	nd River (G	R96, GR03, GR12), th	ne
605	Central Basin Reef	(Reef), and the	e combined	Grand River	samples (G	R_All).	
606							
607	FIGURE 3. STRUCTURE	E Bayesian ass	ignment rest	ults showing	individual	Walleye (thin vertical 1	lines)
608	from the two popul	ation groups, f	for which K=	2 population	n groups (lig	ght and dark grey) were	e
609	supported ($K = 1-5$	were tested; Δ	<i>K</i> =6.91).				
610							



