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5	Article type : Management Brief
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9	Spatiotemporal variability of fishery-dependent indices for the declining Louisiana
10	Southern Flounder fishery
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18	Abstract.—In recent years, management agencies across the Gulf of Mexico and southern U.S.
19	Atlantic have recognized Southern Flounder (Paralichthys lethostigma) as a declining fish stock.
20	Population declines in coastal Louisiana are exhibited by indices of recruitment and biomass,
21	which have reached levels that present management concerns. To develop a better understanding
22	of this declining fishery we examined fishery-dependent data collected by the Louisiana
23	Department of Wildlife and Fisheries recreational angler harvest survey, referred to as LA Creel.
24	Data were modeled using generalized additive models (GAMs) to estimate temporal components
	This is the author manuscript accepted for publication and has undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the <u>Version of Record</u> . Please cite this article as <u>doi:</u> 10.1002/NAFM.10701

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25 of recreational Southern Flounder landings in both seasonality and trend. Over the study period 26 (2014–2019), recreational landings exhibited a declining trend statewide. Strong seasonal peaks 27 in the fall were exhibited statewide and regionally among every coastal management zone (i.e., 28 estuary). Understanding the current fishery with the fine-scale resolution provided by the LA 29 Creel survey can be used to help guide future management decisions in the pursuit of a 30 sustainable management strategy inclusive of fishery-dependent information. 31 Indices of abundance developed from fishery-independent data sources are often preferred over 32 indices developed from fishery-dependent sources in determining the status of fish stocks. 33 Fishery-independent data are collected using systematic and random survey designs that attempt 34 to keep spatial, temporal, and effort elements consistent to gather unbiased abundance data that 35 allow for proportionality between survey catch rates and stock abundance to reasonably be 36 assumed (Hilborn and Walters 1992; Hubert and Fabrizio 2007). The non-random aspects of 37 commercial and recreational fisheries, along with any regulatory changes of the fishery through 38 time, can lead to non-proportionality between fishery catch rates and stock abundance that 39 creates biases when interpreting fishery-dependent data as a measure of stock abundance (Grüss 40 et al. 2019; Tate et al. 2020). Additionally, the non-linear relationship between catchability and 41 stock abundance (Crecco and Overholtz 1990; Wilberg et al. 2009) can potentially lead to 42 hyperstability, where indices of stock abundance ostensibly appear stable while a decline in stock 43 size is occurring (Hilborn and Walters 1992).

44 Despite the inferential limitations of fishery-dependent data as a measure of stock 45 abundance, broad applications exist for the use of fishery-dependent data in assessing fish 46 populations. A classic example of applying fishery-dependent data as a source of stock size is 47 through the standardization of catch-per-unit-effort (CPUE) to remove the factors not related to 48 changes in abundance and is often used in stock assessments where fishery-independent data is 49 lacking or unavailable (Winker et al. 2013; Okamura et al. 2017; Grüss et al. 2019; Tate et al. 50 2020). Another popular application of fishery-dependent data includes characterizing the spatial 51 structure of fish distributions and habitats (Pilar-Fonseca et al. 2014; Pennino et al. 2016; Sculley 52 and Brodziak 2020). Fishery-dependent data can also be used to elucidate drivers of fish 53 population fluctuations by evaluating the relationships observed between fishery effort, landings, 54 and stock abundance (Askey and Johnston 2013; van Poorten et al. 2016; Dassow et al. 2020; 55 Feiner et al. 2020). Composition information (e.g., size, age, sex) collected directly from the

56 fishery is also frequently applied to characterize how fish populations and landings within

- 57 specific regions are structured (Ajemian et al. 2016; Bada-Sánchez et al. 2019; Herdter et al.
- 58 2019). Nevertheless, fisheries-dependent data remain underutilized in the management of many
- 59 fisheries, particularly those for which substantial fisheries-independent data exists.

60 In recent years, numerous state management agencies have recognized Southern Flounder 61 (Paralichthys lethostigma) as a declining fish stock throughout much of the Gulf of Mexico and 62 U.S. southeast Atlantic (Froeschke et al. 2011; Powers et al. 2018; Flowers et al. 2019; West et al. 2020; Erickson et al. 2021). In response to this decline, significant regulatory changes have 63 been implemented in recent years for multiple Southern Flounder fisheries. Although Southern 64 65 Flounder regulatory changes have not yet been implemented in Louisiana, the most recent stock 66 assessment by the Louisiana Department of Wildlife and Fisheries (LDWF) stated that 67 management actions will be necessary to recover the depleted stock (West et al. 2020). The 68 decline in the Louisiana Southern Flounder population is underscored by the most recent 69 estimates of spawning stock biomass, abundance of age-one recruits, and the total female stock 70 size, all reaching the lowest levels observed in the terminal year of each time-series reported 71 (West et al. 2020; Figure 1).

72 Fishery-dependent data in Louisiana for marine recreational fisheries is currently 73 collected through the LDWF recreational angler harvest survey (LA Creel), which was 74 developed in 2014 to produce timely recreational landings estimates with a fine temporal 75 resolution (weekly) and a basin-level spatial resolution. Data collected by LA Creel, which 76 utilizes a stratified random survey design, has provided a substantial reference for estimates of 77 recreational fishery harvest in Louisiana. LA Creel became the first Gulf of Mexico recreational 78 state survey to receive full certification from National Oceanic and Atmospheric Administration 79 (NOAA) Fisheries for the estimation of harvest of all recreationally pursued marine fish species 80 (NOAA Fisheries 2018). With fishery-independent abundance indices depicting a declining 81 Southern Flounder stock in Louisiana (West et al. 2020; Erickson et al. 2021), this study 82 evaluated indices developed from LA Creel to better characterize the Louisiana recreational 83 Southern Flounder fishery and determine if the observed decline in fishery-independent indices 84 is reflected within fishery-dependent indices. Due to the biases associated with using fishery-85 dependent data as an index of stock size without a form of standardization, our evaluation did not 86 seek to provide an additional measure of stock abundance. Rather, our study objectives were

confined to evaluating trends within the fishery by answering three primary questions. First, how
have fishery-dependent indices fluctuated annually? Second, how does exploitation of the
species vary on a seasonal basis? Third, what similarities and differences exist regionally
throughout coastal Louisiana when examining the temporal variations in this fishery? Results of
this analysis may inform future policy changes based on the angler harvest behaviors elucidated
within this fishery and provide a useful approach for assessing spatial and temporal aspects of
fishery-dependent data.

94

95 [A]Methods

96 [C]Data.–We confined our study to the recreational Southern Flounder fishery in Louisiana, 97 which is the primary source of directed fishery removals statewide (West et al. 2020). Moreover, 98 our study was focused on recreational landings of Southern Flounder within coastal Louisiana as 99 offshore landings account for a negligible proportion of the total recreational landings (LA Creel, 100 unpublished data). For the purposes of this study, LA Creel estimates of *landings* (the number of 101 Southern Flounder estimated to have been harvested each week) and *effort* (the number of 102 angler-trips estimated to occur each week) were evaluated. LA Creel estimates are regionally 103 stratified into five coastal management zones: Calcasieu, Vermilion, Terrebonne, Barataria, and 104 Pontchartrain (Figure 2). The Calcasieu Zone contains the Calcasieu/Sabine and Mermentau 105 River basins; the Vermilion Zone contains the Vermilion/Teche and western Atchafalaya River 106 basins; the Terrebonne zone contains the eastern Atchafalaya River and Terrebonne basins; the 107 Barataria Zone contains the Barataria and western Mississippi River Delta basins; and the 108 Pontchartrain Zone contains the eastern Mississippi River Delta, Breton Sound, and

109 Pontchartrain basins.

110 The LA Creel survey generates landings estimates of recreationally targeted marine fish 111 species using a complemented survey design, which combines an on-site dockside intercept 112 survey with an off-site effort survey that utilizes both telephone and email contacts. Both the on-113 site and off-site surveys follow a stratified random sampling structure. Access points utilized 114 within the LA Creel survey include all public sites (e.g., boat ramps, piers, beaches) in coastal 115 Louisiana that are utilized by saltwater anglers. The number of weekly assignments is held 116 constant within each coastal management zone and is based upon the diversity and level of 117 angling activity. As more angling activity occurs during the weekend, assignments fall more

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118 frequently during this day-type. During the dockside intercept survey, information on the number 119 of each species harvested is collected from anglers at access points, which are randomly assigned 120 using a probability proportional to size methodology based on the level of angling activity. 121 Angling activity within each site is assessed monthly to optimize site selection. . Species-specific 122 harvest rates are calculated from the information collected in the dockside intercept survey 123 weighted by the day-type (weekend or weekday), site location, and time of each interview (AM 124 or PM). During the off-site effort survey, a random sample of private Louisiana saltwater fishing 125 license holders and Louisiana charter boat license holders are contacted by email or telephone to 126 collect information about the dates and locations of saltwater angling trips. The private angler 127 effort survey sample frame is stratified into five sections based on geographic area, license 128 densities, and license type [north Louisiana, southeast Louisiana, southwest Louisiana, non-129 resident, and Recreational Offshore Landing Permit (ROLP)]. Sixteen hundred private Louisiana 130 saltwater fishing license holders are randomly selected from the sample frame each week for 131 interviews (400 from the ROLP stratum and 300 from the four remaining strata). The weekly call 132 list from this sample frame is randomized and anglers are contacted until a quota of 800 license 133 holders is reached. For the charter boat survey, 30% of ROLP captains and 10% of non-ROLP 134 captains are randomly selected from the sample frame of Louisiana charter boat license holders 135 and contacted for interviews each week. Angler effort within each coastal zone is estimated by 136 determining the mean effort per angler interviewed and applying this estimate to the total 137 population of licensed anglers with a correction factor for license compliance rates. The harvest 138 rates estimated from the on-site dockside intercept survey and the effort levels estimated from 139 the off-site effort survey are combined to develop a complemented weekly harvest estimate for 140 each coastal zone. In addition to the data provided by the LA Creel Survey (2014–2019), our 141 study also analyzed data collected by the earlier federal recreational marine creel programs 142 (Marine Recreational Information Program and the previous Marine Recreational Fishery 143 Statistics Survey: 2000–2013). Due to the differences between estimation procedures of the LA 144 Creel and federal surveys, LA Creel harvest estimates were calibrated to the historic estimates to 145 develop a continuous time-series of estimates in a common currency (West and Zhang 2018). 146

[C]*Generalized Additive Model.*–We chose generalized additive models (GAMs) to assess the
temporal components within the LA Creel data for a variety of reasons, including: 1) GAMs are

149 capable of exhibiting non-linear and complex relationships between response and predictor 150 variables (Wood 2006), 2) GAMs provide easily interpreted visualizations of these complex 151 relationships, and 3) GAMs offer the flexibility to analyze data within a variety of statistical 152 distributions (Guisan et al. 2002). Weekly landings estimates derived from the LA Creel survey 153 within each coastal management zone were best characterized as count data with overdispersion 154 and frequently contained landings values of zero for each week. Zero-inflated models were 155 explored to evaluate if the frequent zero values in our data were adequately modeled. Results 156 using a zero-inflated negative binomial model were virtually indistinguishable from the results 157 using a negative binomial distribution. Due to the prevalence of the negative binomial 158 distribution in the use of modeling fisheries count data with overdispersion and moderate zero 159 inflation (Barry and Welsh 2002; Drexler and Ainsworth 2013; Dance and Rooker 2019) and the 160 potential overutilization of zero-inflated models (Wood 2020), we opted to model LA Creel 161 estimates using the negative binomial distribution. All statistical analyses were performed using 162 R (R Core Team 2020) and the package *mgcv* (Wood 2006).

163 G

164

GAMs were built under the equation:

$$E[y_{ij}] = g^{-1} \left[\beta_0 + \sum_k S_k(x_k) \right]$$

165 where $E[y_{ii}]$ is the expected value of the response variable as the number of Southern Flounder 166 landed within each coastal zone, *i*, during each corresponding weekly period, *i*. The link function is represented by g, β_0 is the intercept, and S_k is the smoothing function of each predictor 167 168 variable, x. Temporal components of landings were included as predictor variables within each 169 model in year (the corresponding year) and season (the corresponding week of the year). 170 Smoothing terms for each predictor variable were represented using thin plate regression splines 171 (year) and cyclic cubic regression splines (season) penalized from a maximum basis dimension 172 (k). The value of k was restricted to six (one for each year of data) for year and 10 for season. 173 Smoothing parameters were selected using maximum likelihood (ML; Wood 2011). The natural 174 logarithm of effort (weekly number of angler-trips within each coastal zone, *j*) was applied as an 175 offset within each model. Models were specified for estimations within two frameworks, one 176 framework for a coastal zone-specific model and one framework for a statewide model. Within 177 the coastal zone-specific model, weekly recreational landings from all coastal zones were pooled 178 as the dependent variable and five smoothing terms were produced for each coastal zone for year 179 and season. A random effect for each coastal zone, *j*, was applied within the coastal zone-180 specific model allowing variation to occur within each smooth in shape. Within the statewide 181 model, weekly recreational landings from all coastal zones were pooled as the dependent 182 variable and a smoothing term was produced for year and season. A random effect for each 183 coastal zone, *i*, was applied within the statewide model to account for the variability within each 184 coastal zone. Significance of each smoothing term was evaluated at $\alpha = 0.05$. Within both 185 statewide and coastal-specific models, substantial positive residual autocorrelation provided 186 evidence of serial correlation, a common issue confounding time-series analysis (Hamilton 187 1994). To account for serial correlation within LA Creel landings estimates, a first-order 188 autoregressive term (Wood 2006) was nested within each coastal zone time-series for both the 189 statewide and coastal zone-specific model frameworks.

190

191 [C]*Synchrony.*—To quantify the similarity of temporal patterns in Southern Flounder landings 192 throughout Louisiana, synchrony of coastal zone-specific model predictions were quantified 193 through the calculation of Spearman correlation coefficients with corresponding asymptotic *p*-194 values approximated using the *t*-value distribution (Harrell 2020). Significant relationships were 195 evaluated at $\alpha = 0.05$. Weekly predictions from 2014–2019 were used in this analysis, making 196 each weekly time-series 313 periods long. With five total time-series (Pontchartrain, Barataria, 197 Terrebonne, Vermilion, Calcasieu), 10 unique correlation coefficients were calculated.

198

199 [A]Results

From 2000–2016, annual landings estimates of Southern Flounder in Louisiana remained
relatively close to the 20-year average (2000–2020) of 200,000 fish per year. During the years
2017–2019, annual landings estimates substantially declined to levels approaching or below 50%
of the 20-year average (Figure 3). Louisiana Southern Flounder landings estimates were driven
by harvest within the Calcasieu Zone where at least 55% of the annual Southern Flounder
landings in Louisiana have been estimated to occur over the LA Creel period (Figure 4).
For the GAM applied to statewide data, the smoothing term for *year* was significant (*p* <

0.001; Table 1) indicating variability in landings over the time-series examined. The statewide
time-series displayed a sharp decline in 2017 with a modest increase in 2019 (Figure 5). Over the
entire time-series, Southern Flounder landings have declined statewide during the LA Creel

210 period. A significant smoothing term (p < 0.05; Table 1) for year indicated variability in landings

211 over the time-series in the Calcasieu, Pontchartrain, Barataria, and Terrebonne Zones (Figure 5).

212 Landings in the Barataria and Terrebonne Zones exhibited declines beginning approximately in

213 2017 with subsequent increases in 2019. The Calcasieu and Pontchartrain Zones exhibited a

- 214 moderate decline over the entire time-series. Within the Vermilion Zone, landings did not
- 215 display significant variability over the time-series.
- The smoothing term for *season* was significant (p < 0.001; Table 1) in the statewide model and displayed a strong peak in fall landings (Figure 6). In the coastal zone-specific model smoothing terms for *season* were significant (p < 0.05; Table 1) within every coastal zone, with strong landings peaks in the fall, specifically during the month of November (Figure 6).

220 The relationships between each coastal zone-specific prediction time-series were all 221 significant (p < 0.001) and positively correlated. Each correlation coefficient indicated 222 relationships ranging from 0.22–0.67 (Figure 7).

223

224 [A]Discussion

This study examined recreational landings in the Louisiana Southern Flounder fishery, indicating that landings have declined statewide in recent years, but more importantly provided a fine-scale spatial and temporal understanding of this decline. While only using data from the six-year LA Creel time-series, the observed decline in landings is significant statewide and consistent with the range-wide declines reported by numerous state agencies (Erickson et al. 2021). Extending the LA Creel time-period to the time-series of hindcast landings estimates places the context of the decline over a 20-year period.

232 Our analysis quantified and confirmed the impact of the fall harvest within the 233 recreational Southern Flounder fishery in coastal Louisiana. Although the fall harvest is well 234 defined by those familiar with the resource (Gulf States Marine Fishery Commission 2015), the 235 seasonality of recreationally harvested Southern Flounder in coastal Louisiana has not been 236 documented in the published literature and our study provides a baseline to understanding future 237 seasonal shifts within this fishery. The peak in fall harvest coincides with the large-scale 238 seasonal migrations that Southern Flounder make to offshore waters from the estuarine 239 environment as water temperatures cool (Craig et al. 2015). As Southern Flounder move through restricted passes to reach offshore spawning grounds, concentrations of fish become increasingly
vulnerable to recreational harvest as indicated by the peaks observed in the fall landings.

242 The resolution of data characterizing specific coastal zones within Louisiana provided the 243 opportunity to evaluate the unique spatial variability that exists in the Louisiana Southern 244 Flounder fishery. Our analysis provided evidence of the substantial influence that the Calcasieu 245 Zone holds on coastal Louisiana recreational Southern Flounder landings. Additionally, the 246 Vermilion Zone did not display significant variability over the LA Creel period in our analysis of 247 annual trend, in contrast with the variability observed within all other coastal zones and 248 statewide. The Vermilion Zone also provided the smallest magnitude of landings within any 249 coastal zone with total landings accounting for less than three percent of statewide landings over 250 the entire LA Creel period. The inferior scale of landings in the Vermilion Zone may be a driving 251 force behind the diverging variability in annual trend. Moreover, our analysis of synchrony 252 displayed significant positive correlations between the Vermilion Zone and all other coastal 253 zones, providing some additional evidence of variability within this zone. A further evaluation of 254 estuary-specific habitat and bay morphology may provide evidence to why differential 255 exploitation in magnitude and annual trend occurs within the Calcasieu and Vermilion Zones. To 256 further evaluate the potential for coastal zone-specific stock structure, age and size composition 257 data could be evaluated within each coastal zone to determine if sub-populations exist within the 258 statewide stock.

259

260 [B]Study Limitations and Strengths

261 This study was reliant upon probabilistic sampling methods that are commonplace in the 262 estimation of recreational effort and landings. While there are limitations to this type of data 263 collection, namely in the inferences that must be made from a fixed amount of interviews over a 264 sampling period (Midway et al. 2020), no substitutable alternatives (e.g., electronic self-265 reporting) currently exist in the collection of recreational fishery-dependent data, specifically 266 within coastal Louisiana. Absent significant advances in non-probabilistic sampling designs and 267 the willingness of users to adopt new technologies (Midway et al. 2020), robust probabilistic 268 sampling methods for fishery-dependent data must continue in order to characterize and 269 understand changes in harvest though time and space.

270 With the availability of only six years of data at the spatial and temporal resolutions of 271 the LA Creel survey, the trends displayed by our models can only be placed in this limited 272 context. However, the fine temporal resolution of the LA Creel survey data allows for inferences 273 to be made over the six-year window. The decline displayed statewide over this period highlights 274 the magnitude of the reduction that is occurring in the Louisiana Southern Flounder fishery. 275 Additionally, the spatial resolution of the LA Creel survey allows coastal zone-specific 276 inferences to be made throughout coastal Louisiana, a spatial distinction that was not previously 277 available with Louisiana recreational fishery estimates prior to the implementation of LA Creel. 278 The LA Creel survey characterizes recreational fisheries with greater resolution, greatly 279 improves the monitoring capabilities within coastal Louisiana fisheries, and allows for real-time 280 and impactful management decisions to be made in preserving the future of significant recreational fisheries. 281

282 The framework offered by a GAM statistical approach provided a significant strength in 283 our evaluation. While other approaches were evaluated, particularly in tree-based machine 284 learning approaches and generalized linear models (GLMs), the GAM approach was recognized 285 as the strongest statistical fit for our evaluation. The flexibility offered by a semi-parametric 286 approach with the ability to provide data-driven response curves (Yee and Mitchell 1991) led to 287 our decision to use a GAM over a GLM. The simplicity offered by fitting the most parsimonious 288 model (rather than risking overfitting with several models; Carvalho et al. 2018) and the 289 improved accuracy in modeling smooth functions (Elith et al. 2008) led to our decision to use a 290 GAM over a tree-based machine learning approach. GAMs were identified as the best statistical 291 tool for this investigation with the ability to provide data-driven response curves of highly non-292 linear and non-monotonic relationships (Yee and Mitchell 1991; Wood 2006), the capacity to 293 account for serial correlation (Wood 2006), and the flexibility to model data in a variety of 294 statistical distributions (Guisan et al. 2002).

295

296 [B]Future of the Fishery

297 The decline in landings of Southern Flounder in Louisiana may be the result of various drivers or

the interactions among multiple drivers, including changing angler behavior or a decline in stock

size, among other possibilities. Angler behavior, in the number of angling trips, has not

300 significantly changed during the LA Creel period within the Pontchartrain or Terrebonne Zones;

301 however, the Barataria, Vermilion, and Calcasieu Zones all displayed significant changes in 302 effort over this period. Effort within the Barataria Zone has increased while effort in the 303 Vermilion and Calcasieu Zones has declined. These outcomes may have a significant effect on 304 Southern Flounder landings, specifically within the Calcasieu Zone as this coastal zone has the 305 strongest influence on statewide landings. The Calcasieu Zone effort declined in mean weekly 306 trip estimates from 7,104 trips in 2014 to 5,753 trips in 2019. Since 2017, the target species from 307 each dockside intercept interview has been recorded by the LA Creel survey. The annual 308 percentage of anglers targeting Southern Flounder in coastal Louisiana has remained relatively low from 2017–2020, annually ranging between 1.4%–1.6% statewide. However, these 309 310 percentages varied seasonally and spatially, particularly within the Calcasieu Zone and during 311 the month of November. During November within the Calcasieu Zone, the percent of anglers 312 targeting Southern Flounder was 21.9% from 2017–2020. This percentage is a substantial 313 increase from all other months during that period, which averaged 4.0%. Moderate increases in 314 the percentage of anglers targeting of Southern Flounder were exhibited in other coastal zones 315 during the month of November; however, these increases were negligible in comparison to the 316 increases in the Calcasieu Zone. The substantial increase in targeting behavior during the month 317 of November is likely a significant factor in the Calcasieu Zone's influence on statewide 318 landings. Continued monitoring of Louisiana angling behaviors are necessary to precisely 319 determine if changing behaviors of anglers are driving fluctuations in Southern Flounder 320 landings.

321 With estimates of recruitment and stock size for Louisiana Southern Flounder reaching 322 the lowest levels recorded (West et al. 2020), it appears that the reduction in landings of 323 Southern Flounder in Louisiana is related to a corresponding decline in stock size. Moreover, an 324 important factor to consider when evaluating trends in fishery-dependent data is that declines in 325 stock size may be masked by hyperstability. There are numerous examples of fisheries where 326 hyperstability has concealed declines among fish stocks that exhibit aggregating behaviors 327 (Erisman et al. 2011; Sadovy de Mitcheson and Erisman 2012; Dassow et al. 2020). The 328 aggregating behavior of Southern Flounder during the fall migration (Gulf States Marine 329 Fisheries Commission 2015) leads to increased vulnerability of the stock as Southern Flounder 330 move through coastal bottlenecks to reach spawning grounds. The fact that statewide recreational 331 estimates of the Louisiana Southern Flounder fishery currently depict a decline in landings is a

332 meaningful indicator of potential overfishing and a depleted stock that warrants management333 attention.

334 Considering this species marked decline, attributes of the recreational Louisiana Southern 335 Flounder fishery demonstrated by the findings of this study can provide insight into future 336 management strategies. The reduction of spawning stock size can significantly impact Southern 337 Flounder populations during the fall migration and has resulted in seasonal restrictions applied to 338 recreational Southern Flounder fisheries throughout much of their range. Our findings indicate 339 that a seasonal restriction during the fall migration would likely produce a reduction in Southern 340 Flounder harvest, an action that may be necessary for the recovery of this declining stock. All 341 coastal zones indicated positively correlated time-series and strong similarities existed within the 342 seasonal trends produced in each coastal zone displaying the converging attributes of each 343 coastal zone. Moreover, the estuarine fidelity that Southern Flounder exhibit as they return to 344 coastal Louisiana is unknown and future research is necessary to examine the level of 345 connectivity among Louisiana coastal zones, as well as the connectivity across state boundaries. 346 For these reasons, the management of Southern Flounder in Louisiana would likely benefit from 347 a statewide strategy rather than a region-specific approach.

348 With Southern Flounder declines occurring throughout their range, extending across 349 multiple jurisdictional boundaries in which various regulatory strategies exist, there is a high 350 capacity for a universal driver in the decline of Southern Flounder. Southern Flounder are 351 susceptible to the effects of a changing climate as this species exhibits environmental sex 352 determination in which during larval development suboptimal water temperatures can lead to 353 increased ratios of phenotypic males (Luckenbach et al. 2009; Honeycutt et al. 2019). 354 Additionally, laboratory studies have also shown that warming water temperatures significantly 355 affect the success of hatching and larval development of Southern Flounder (van Maaren and 356 Daniels 2001). As documented water temperatures have risen over the past few decades within 357 the same locations and times that Southern Flounder develop after hatching (Erickson et al. 358 2021), the role that climate has played as a driver of stock size in Louisiana is likely to be 359 significant. While future management strategies have the potential to help mitigate further 360 declines in Louisiana Southern Flounder stock size, it is important to note that reduced 361 exploitation may not have the potential to fully recover the stock considering the role that 362 climate will continue to play in this species' decline.

363 Recreational fisheries are dynamic systems that provide numerous management 364 challenges as changes occur in climate, environment, and culture (Elmer et al. 2017; 365 Brownscombe et al. 2019; Holder et al. 2020). The approaching difficulties of managing 366 recreational fisheries that will be heavily influenced by these changes underscores the 367 importance in filling the information gap that exists between human dimensions studies and 368 recreational fisheries (Hunt et al. 2013). One way that managers can more closely understand the 369 behaviors of anglers is through an evaluation of fishery-dependent data. Fishery-dependent data 370 provides insights into fisheries that are often underutilized yet commonly collected by 371 management agencies. When management action is required, understanding the various 372 components of recreational fishery harvest can aid in making a management decision that is not 373 only biologically sound but also makes a meaningful impact in preserving the benefits that fisheries provide to anglers. 374

375

376 [A]Acknowledgments

We thank the numerous biologists and agency personnel who have contributed to the LA Creel program over many years. We also thank Kevin Savoie and Drs. Rex Caffey and Jack Isaacs for helpful discussions on the southern flounder recreational fishery in Louisiana.

380 References

- Ajemian, M. J., P. D Jose., J. T. Froeschke, M. L. Wildhaber, and G. W. Stunz. 2016. Was
 everything bigger in Texas? Characterization and trends of a land-based recreational
 shark fishery. Marine and Coastal Fisheries 8(1):553–566.
- Askey, P. J., and N. T. Johnston. 2013. Self-regulation of the Okanagan Lake Kokanee
 recreational fishery: Dynamic angler effort response to varying fish abundance and
 productivity. North American Journal of Fisheries Management 33:926–939.
- 387 Bada-Sánchez, E., J. C. Pérez-Jiménez, L. E. Martínez-Cruz, I. Méndez-Loeza, and E.
- 388 Sosa-Cordero. 2019. Fishery indicators during a predictable aggregation of Atlantic
- 389 Sharpnose Sharks, *Rhizoprionodon terraenovae*, in the southern Gulf of Mexico: An
- 390 alternative to assess a heterogeneous data-poor fishery. Fisheries Management and
- 391 Ecology 26:354–364.

Barry, S., and A. H. Welsh. 2002. Generalized additive models and zero inflated count data.
 Ecological Modelling 157(2):179–188

Brownscombe, J. W., K. Hyder, W. M. Potts, K. L. Wilson, P. A. Danylchuk, S. Cooke, A.

- Clarke, R. Arlinghaus, and J. Post. 2019. The future of recreational fisheries: Advances in
 science, monitoring, management, and practice. Fisheries Research 211:247–255.
- Carvalho, J., J. P. V. Santos, R. T. Torres, F. Santarém, and C. Fonseca. 2018. Tree-based
 methods: Concepts, uses and limitations under the framework of resource selection
 models. Journal of Environmental Informatics 32(2):112–124.
- 400 Craig, J. K., W. E. Smith, F. S. Scharf, and J. P. Monaghan. 2015. Estuarine residency and
 401 migration of Southern Flounder inferred from conventional tag returns at multiple spatial
 402 scales. Marine and Coastal Fisheries 7:450–463.
- 403 Crecco, V., and W. J. Overholtz. 1990. Causes of density-dependent catchability for Georges
 404 Bank Haddock *Melanogrammus aeglefinus*. Canadian Journal of Fisheries and Aquatic
 405 Sciences 47(2): 385–394.
- 406 Dance, M. A., and J. R. Rooker. 2019. Cross-shelf habitat shifts by red snapper (*Lutjanus* 407 *campechanus*) in the Gulf of Mexico. PLoS ONE 14(3):1–22.
- 408 Dassow, C. J., A. J. Ross, O. P. Jensen, G. G. Sass, B. T. van Poorten, C. T. Solomon, and S. E.
 409 Jones. 2020. Experimental demonstration of catch hyperstability from habitat
 410 aggregation, not effort sorting, in a recreational fishery. Canadian Journal of Fisheries
 411 and Aquatic Sciences 77:762–769.
- 412 Drexler, M., and C. H. Ainsworth. 2013. Generalized additive models used to predict species
 413 abundance in the Gulf of Mexico: An ecosystem modeling tool. PLoS ONE, 8(5):1–7.
- 414 Elith, J., J. R. Leathwick, and T. Hastie. 2008. A working guide to boosted regression trees.
 415 Journal of Animal Ecology 77:802–813.
- 416 Elmer, L. K., L. A. Kelly, S. Rivest, S. C. Steell, W. M. Twardek, A. J. Danylchuk, R.
- 417 Arlinghaus, J. R. Bennet, and S. J. Cooke. 2017. Angling into the future: Ten

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418	commandments for recreational fisheries science, management, and stewardship in a
419	good Anthropocene. Environmental Management 60(2):165–175.

- Erickson, K. A., J. West, M. A. Dance, T. M. Farmer, J. C. Ballenger, and S. R. Midway. 2021.
 Changing climate associated with the range-wide decline of an estuarine finfish. Global
 Change Biology 27(11):2520–2536.
- Erisman, B. E., L. G. Allen, J. T. Claisse, D. J. Pondella II, E. F. Miller, and J. H. Murray. 2011.
 The illusion of plenty: hyperstability masks collapses in two recreational fisheries that
 target fish spawning aggregations. Canadian Journal of Fisheries and Aquatic Sciences
 68:1705–1716.

Feiner, Z. S., M. H. Wolter, and A. W. Latzka. 2020. "I will look for you, I will find you, and I
will [harvest] you": Persistent hyperstability in Wisconsin's recreational fishery.
Fisheries Research 230:1–12.

Flowers, A. M., S. D. Allen, A. L. Markwith, and L. M. Lee. 2019. Stock assessment of Southern
 Flounder (*Paralichthys lethostigma*) in the Southern Atlantic, 1989–2017. Joint report of

432 the North Carolina Division of Marine Fisheries, South Carolina Department of Natural

433 Resources, Georgia Coastal Resources Division, Florida Fish and Wildlife Research

434 Institute, University of North Carolina at Wilmington, and Louisiana State University.

Froeschke, B. F., B. Sterba-Boatwright, and G. W. Stunz. 2011. Assessing Southern Flounder
(*Paralichthys lethostigma*) long-term population trends in the northern Gulf of Mexico
using time series analyses. Fisheries Research 108:291–298.

Grüss, A, J. F. Walter III, E. A. Babcok, F. C. Forrestal, J. T. Thorson, M. V. Lauretta, and M. J.
Schirripa. 2019. Evaluation of the impacts of different treatments of spatio-temporal
variation in catch-per-unit-effort standardization models. Fisheries Research 213:75–93.

Guisan, A., T. C. Edwards, Jr., and T. Hastie. 2002. Generalized linear and generalized additive
models in studies of species distributions: Setting the scene. Ecological Modelling
157:89–100.

444	Gulf States Marine Fisheries Commission. 2015. Management profile for the Gulf and Southern
445	Flounder fishery in the Gulf of Mexico. Gulf States Marine Fisheries Commission,
446	Flounder Technical Task Force.

Hamilton, J. D. 1994. Time Series Analysis. Princeton University Press, Princeton, New Jersey. 447

448 Harrell, F. E. 2020. Hmisc: Harrell Miscellaneous. R package version 4.4-0.

- 449 Herdter, E., B. Mahmoudi, E. Peebles, and S. A. Murawski. 2019. Spatial variability in size 450 structure, growth, and recruitment of Spotted Seatrout among six Florida estuaries.
- 451 Marine and Coastal Fisheries: Dynamics, Management, and Ecosystem Science 11:97-111.
- 452

453 Hilborn, R., and Walters, C. J. 1992. Quantitative fisheries stock assessment: choice, dynamics 454 and uncertainty. Chapman and Hall, New York.

455 Holder, P. E., A. L. Jeanson, R. J. Lennox, J. W. Brownscombe, R. Arlinghaus, A. J. Danylchuk, 456 S. D. Bower, K. Hyder, L. M. Hunt, E. P. Fenichel, P. A. Venturelli, E. B. Thorstad, M.

457 S. Allen, W. M. Potts, S. Clark-Danylchuk, J. E. Claussen, J. M. Lyle, J. -i. Tsuboi, R. 458

Brummett, K. M. F. Freire, S. R. Tracey, C. Skov, and S. J. Cooke. 2020. Preparing for a 459 changing future in recreational fisheries: 100 research questions for global consideration

- 460 emerging from a horizon scan. Reviews in Fish Biology and Fisheries 30(1):137–151.
- 461 Honeycutt, J. L., C. A. Deck, S. C. Miller, M. E. Severance, E. B. Atkins, J. A. Luckenbach, J. 462 A. Buckel, H. V. Daniels, J. A. Rice, R. J. Borski, and J. Godwin. 2019. Warmer waters 463 masculinize wild populations of a fish with temperature-dependent sex determination. Scientific Reports 9(1):1–13. 464
- 465 Hubert, W. A., and M. C. Fabrizio. 2007. Relative abundance and catch per unit effort. Pages 279-325 in C. S. Guy, and M. L. Brown, editors. Analysis and interpretation of 466 467 freshwater fisheries data. American Fisheries Society, Bethesda, Maryland.

468 Hunt, L. M., S. G. Sutton, and R. Arlinghaus. 2013. Illustrating the critical role of human 469 dimensions research for understanding and managing recreational fisheries within a

470 social-ecological system framework. Fisheries Management and Ecology 20(2-3):111–
471 124.

Luckenbach, J. A., R. J. Borski, H. V. Daniels, and J. Godwin. 2009. Sex determination in
flatfishes: Mechanisms and environmental influences. Seminars in Cell and
Developmental Biology 20(3):256–263.

- 475 Midway, S. R., J. Adriance, P. Banks, S. Haukebo, and R. H. Caffey. 2020. Electronic
 476 Self-Reporting: Angler Attitudes and Behaviors in the Recreational Red Snapper Fishery.
 477 North American Journal of Fisheries Management 40:1119–1132.
- 478 NOAA (National Oceanic and Atmospheric Administration) Fisheries. 2018. NOAA Fisheries
 479 certifies LA Creel survey design. Available: https://www.fisheries.noaa.gov/feature-
- 480 story/noaa-fisheries-certifies-la-creel-survey-design (October 2020).
- 481 Okamura, H., S. H. Morita, T. Funamoto, M. Ichinokawa, and S. Eguchi. 2018. Target-based
 482 catch-per-unit-effort standardization in multispecies fisheries. Canadian Journal of
 483 Fisheries and Aquatic Sciences 75:452–463.

Pennino, M. G., D. Conesa, A. Lopez-Quilez, F. Munoz, A. Fernandez, and J. M. Bellido. 2016. Fishery-dependent and -independent data lead to consistent estimations of essential habitats. ICES Journal of Marine Science 73(9):2302–2310.

487 Pilar-Fonseca, T., A. Campos, J. Pereira, A. Moreno, S. Lourenço, and M. Afonso-Dias. 2014.
488 Integration of fishery-dependent data sources in support of octopus spatial management.
489 Marine Policy 45:69–75.

Powers, S. P., M. Albins, and J. Mareska. 2018. An assessment of Southern Flounder in Alabama coastal waters. Joint report of Department of Marine Sciences, University of South Alabama, the Dauphin Island Sea Lab, and Alabama Department of Conservation and Natural Resources, Marine Resources Division.

494 R Core Team. 2020. R: A language and environment for statistical computing. R Foundation for
495 Statistical Computing.

496	Sadovy de Mitcheson, Y., and B.E. Erisman. 2012. Fishery and biological implications of fishing
497	spawning aggregations, and the social and economic importance of aggregating fishes.
498	Pages 225–284 in: Y. Sadovy de Mitcheson, and P. L. Colin, editors. Reef fish spawning
499	aggregations: Biology, research and management. Springer, New York.
500	Sculley, M. L., and J. Brodziak. 2020. Quantifying the distribution of swordfish (Xiphias
501	gladius) density in the Hawaii-based longline fishery. Fisheries Research 230:1–10.
502	Tate, A., J. Lo, U. Mueller, G. A. Hyndes, K. L. Ryan, and S. M. Taylor. 2020. Standardizing
503	harvest rates of finfish caught by shore-based recreational fishers. ICES Journal of
504	Marine Science 77(6):2207–2215
505	van Maaren, C. C., and H. Daniels. 2001. Effects of temperature on egg hatch, larval growth and
506	metamorphosis for hatchery cultured Southern Flounder, Paralichthys lethostigma.
507	Journal of Applied Aquaculture 11(1-2):21-33.
508	van Poorten, B. T., C. J. Walters, and H. G. M. Ward. 2016. Predicting changes in the
509	catchability coefficient through effort sorting as less skilled fishers exit the fishery during
510	stock declines. Fisheries Research 183:379–384.
511	West, J., and X. Zhang. 2018. LA Creel/MRIP calibration procedure. Louisiana Department of
512	Wildlife and Fisheries, Office of Fisheries.
513	West, J., X. Zhang, T. Allgood, J. Adriance, K. A. Erickson, and S. R. Midway. 2020.
514	Assessment of Southern Flounder, Paralichthys lethostigma, in Louisiana waters 2020
515	report. Louisiana Department of Wildlife and Fisheries, Office of Fisheries.
516	Wilberg, M.J., J.T. Thorson, B.C. Linton, and J. Berkson. 2009. Incorporating time-varying
517	catchability into population dynamic stock assessment models. Reviews in Fisheries
518	Science 18(1):7–24.
519	Winker, H., S. E. Kerwath, and C. G. Attwood. 2013. Comparison of two approaches to
520	standardize catch-per-unit-effort for targeting behaviour in a multispecies hand-line
521	fishery. Fisheries Research 139:118–131.

- Wood, S. N. 2006. Generalized additive models: an introduction with R. Chapman and Hall,
 Boca Raton, Florida.
- Wood, S. N. 2011. Fast stable restricted maximum likelihood and marginal likelihood estimation
 of semiparametric generalized linear models. Journal of the Royal Statistical Society
 73(1):3–36.
- Wood, S. N. 2020. Inference and computation with generalized additive models and their
 extensions. TEST 29(2):307–339.
- Yee, T. W., and N. D. Mitchell. 1991. Generalized additive models in plant ecology. Journal of
 Vegetation Science 2:587–602.
- 531
- 532 Tables
- 533

Table 1. GAM smoothing term significance within the statewide model and coastal zone-specific model. Significant terms (p < 0.05) are in boldface.

Region	Smoothing Term	<i>p</i> -value
Statewide	time	< 0.001
<u> </u>	season	< 0.001
Calcasieu	time	0.028
	season	< 0.001
Vermilion	time	0.269
	season	< 0.001
Terrebonne	time	< 0.001
_	season	0.008
Barataria	time	< 0.001
	season	< 0.001
Pontchartrain	time	0.008
	season	< 0.001

536

537 Figure Captions

538

539 Figure 1. Estimates of abundance for Louisiana Southern Flounder (1982–2018). Shaded areas

540 represent two asymptotic standard errors above and below the estimate. a) Spawning Stock

- 541 Biomass. b) Abundance of Age-One Recruits. c) Total Female Stock Size (asymptotic standard
- 542 errors were not calculated for this estimate).
- 543

Figure 2. Louisiana coast regionally segmented by the LDWF coastal management zones.

Figure 3. Annual Southern Flounder landing estimates from the LA Creel survey (2014–2019)
and estimates hindcast to the historic Marine Recreational Information Program time-series
(2000–2013).

549

Figure 4. Annual Louisiana Southern Flounder landings within each coastal management zone.

552 Figure 5. Partial effect plots for the smoothing term *year* in the statewide model and in the 553 coastal zone-specific model. Shaded areas represent two standard errors above and below the

smooth curve estimate. Dashed lines mark zero-effect.

555

Figure 6. Partial effect plots for the smoothing term *season* in the statewide model and in the coastal zone-specific model. Shaded areas represent two standard errors above and below the smooth curve estimate. Dashed lines mark zero-effect.

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560 Figure 7. Correlation matrix of weekly recreational LPUE for Southern Flounder within each

561 coastal management zone. Significant relationships (p < 0.05) are indicated by an asterisk.













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