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Relative survival of subyearling chinook salmon at Bonneville Dam,
1992

Coastal Zone and

by Richard D. Ledgerwood, Earl M. Dawley, Lyle G. Gilbreath, L. Ted Parker, Benjamin P. Sandford, and Stephen J. Grabowski

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1992

RELATIVE SURVIVAL OF SUBYEARLING CHINOOK SALMON AFTER PASSAGE THROUGH THE BYPASS SYSTEM AT THE FIRST POWERHOUSE OR A TURBINE AT THE FIRST OR SECOND POWERHOUSE AND THROUGH THE TAILRACE BASINS AT BONNEVILLE DAM, 1992

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by

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Lyle G. Gilbreath L. Ted Parker Benjamin P. Sandford and Stephen J. Grabowski

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INTRODUCTION

By virtue of its position as the lowermost dam, more juvenile salmon

(Oncorhynchus spp.) must pass Bonneville Dam than any other hydroelectric project on

the Columbia River. Hence, improvement in passage survival at Bonneville Dam can

positively influence fishery production. In 1987, the National Marine Fisheries Service

(NMFS), in cooperation with the U.S. Army Corps of Engineers (COE), began a multiyear

study to evaluate relative survival of subyearling fall chinook salmon (0. tshawytscha)

after passage through various routes at Bonneville Dam.

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From 1987 to 1990, this research focused on passage through the Second

Powerhouse turbines, juvenile bypass system, and tailrace, and over the spillway located

between the First and Second Powerhouses (Ledgerwood et al. 1990, 1991a) (Fig. 1). We

compared recovery percentages of juvenile test fish released during those studies and

recaptured in the estuary. These recovery data indicated that fish passing through the

Second Powerhouse bypass system survived at lower rates than fish passing the turbines

or over the spillway. Continuing assessment of these survival differences will be based on

recoveries of tagged adult fish in ocean fisheries, Columbia River fisheries, and Columbia River hatcheries.

Because of the similarity of the First and Second Powerhouse bypass systems, it is

also important to evaluate survival of juvenile salmonids passing through the First

Powerhouse bypass system. Furthermore, it is critical to directly compare the relative

survival of fish passing the First and Second Powerhouse turbines.

Research in 1988 and 1989 (Gessel et al. 1989, 1990) indicated that subyearling

chinook salmon migrating in summer are not effectively guided into the bypass system at

either Bonneville Dam powerhouse: only about 27% were guided at the Second

Powerhouse and a dismal 9% were guided at the First Powerhouse. Thus, the vast

Figure 1.--Release locations for subyearling chinook salmon during the Bonneville Dam survival study, 1992.

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majority of the summer migrants pass through turbines at Bonneville Dam, rather than being intercepted by submersible traveling screens (STS) and shunted into the bypass systems.

Past research on survival of juvenile salmonids through turbines at the First

Powerhouse (Holmes 1952) indicated 85 to 89% survival; similar survival rates were

reported in other studies at low-head Kaplan turbines (Schoeneman et al. 1961, Oligher and Donaldson 1966). Recent studies conducted at Bonneville Dam Second Powerhouse turbines suggested survival through these newer units ranged from 96 to 99% (Ledgerwood et al. 1990, 1991a). Turbines at the Second Powerhouse are more efficient because of an improved design and a 4.3-m deeper average submergence of the blades. Turbine efficiency has been directly correlated to increased juvenile salmon passage survival (Cramer 1965, Oliger and Donaldson 1966, Ruggles 1985). Thus the present

operational criteria that favor juvenile salmonid passage through the First Powerhouse

over passage through the Second Powerhouse at Bonneville Dam may be flawed.

Another important aspect of passage survival at Bonneville Dam is mortality

occurring in the tailrace areas, which is thought to result primarily from predation by

northern squawfish (Ptychocheilus oregonensis). Fish exit the bypass conduit as a point

source release in an area of low velocity, and this likely allows more intensive predation

on bypassed fish than for fish passing through the turbines, where they are broadcast

over a wide area. Indeed, a U. S. Fish and Wildlife Service (FWS) study in 1990

documented that a higher proportion of bypass-released juvenile salmon were consumed

by northern squawfish in the tailrace area of Bonneville Dam than were other groups of

juvenile salmon released at the same time (Thomas Poe, FWS, Columbia River Field

Station, Cook, WA. Pers. commun.). Consequently, the reduced estuarine recovery

percentages of groups that passed Bonneville Dam via the Second Powerhouse bypass

system may be at least in part the result of higher predation in the tailrace. In 1988 and

1989, measures of tailrace mortality at the Second Powerhouse were obtained by

comparing recovery percentages of fish released directly into the tailrace to those of fish

released downstream (mean tailrace mortality 7.6%) (Gilbreath et al. 1993). However,

differences in survival between various passage routes and through the tailrace basins

may change over time due to changes in river conditions and predator populations. For

example, as a result of the system-wide predator control program funded by the

Bonneville Power Administration (BPA), over 100,000 northern squawfish have been

removed from the tailrace areas of Bonneville Dam in 1991 and 1992 (Willis and

Nigro 1993), and removal of these predators has undoubtedly reduced tailrace mortality

for juvenile salmon in this area.

In 1992, the NMFS expanded passage survival research at Bonneville Dam to

include assessment of passage through the turbines and the bypass system at the First

Powerhouse. This assessment was necessary to identify the safest passage routes at

Bonneville Dam for juvenile salmon migrating in the summer. The objective of this study

was to compare relative survival among marked groups of subyearling chinook salmon

released into the bypass system of Bonneville Dam First Powerhouse, the turbines at the

First and Second Powerhouses, and at a site in swift water about 2.5 km downstream

from the dam. Estimates of long- and short-term relative survival- will be developed by comparing recovery percentages of these different groups.

Short-term relative survival is based on recoveries of branded (Mighell 1969)

juvenile fish recovered 157 km downstream from the dam, near the upper boundary of the Columbia River estuary at Jones Beach, Oregon (Fig. 2). Long-term relative survival will

Figure 2.--Columbia River Basin showing the study area.

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be based on coded-wire tags (CWT) (Bergman et al. 1968) from adult fish recovered in

ocean fisheries, Columbia River fisheries, and Columbia River hatcheries. These

comparative survival data are critical for developing dam operation procedures that will

ensure maximum protection for juvenile fish and for assessing the necessity for alternate

bypass-release sites.

A complementary study by the NBS (Thomas Poe, Principal Investigator) assessed

distribution and juvenile salmon consumption rates by northern squawfish in the tailrace

basins. Juvenile salmon CWTs recovered from the stomach contents of captured northern

squawfish assisted in documenting impacts of predation on summer migrants from

different release groups.

METHODS

Experimental Design

In 1992, as in previous years of this study, test dates were chosen to represent the

typical conditions encountered by subyearling fall chinook salmon migrating past

Bonneville Dam in the summer. Release locations at the First Powerhouse turbine and

bypass were new, while those at the Second Powerhouse turbine and at the downstream

sites were the same as in previous years. To provide an unbiased comparison of passage

survival for the two turbine-release groups, water flow through both powerhouses was

equalized for the period of this study. We assume that this provided similar predator

attraction in the tailrace basins of each powerhouse.

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Test Fish

In previous years, upriver bright stock fall chinook salmon reared at the Oregon

Department of Fish and Wildlife Bonneville Hatchery (Fig. 2) were specifically chosen as

test fish because of their similarity to summer migrants, availability, low probability of

straying, and expected high percentage of adult returns. However, availability of upriver

bright stock fish throughout the Columbia River Basin in 1992 was insufficient for the needs of this study. Therefore, tule stock subyearling chinook salmon from Little White Salmon National Fish Hatchery (NFH) were selected as study fish (Fig. 2). By the time the shortage of Bonneville Hatchery fish was confirmed, fish at Little White Salmon NFH had hatched, and disease concerns prevented their transfer to Bonneville Hatchery for rearing. Consequently, test fish were reared and marked at Little White Salmon NFH. Transfer of study fish to Bonneville Hatchery for rearing and marking would have been preferred because of logistics during marking, better expected return rates of adult fish to

Bonneville Hatchery, and less straying of returning adult fish to other locations in the

Columbia River Basin.

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About 1.5 million subyearling chinook salmon were made available by FWS at Little White Salmon NFH. Test fish were the progeny of tule stock fall chinook salmon spawned at Spring Creek NFH and transferred as eyed eggs to Little White Salmon NFH for incubation and rearing. Fish size at release varied from 6.4 to 8.1 g (71.0 to 56.0 fish/lb), and was similar to the size of test fish used in previous years.

Little While Salmon NFH is upstream from Bonneville Dam, and the adult return

rates for tule stock relative to upriver bright stock from Bonneville Hatchery are generally

poor. Tule stock normally migrate earlier in the spring than upriver bright stock, but the

test fish were reared in cold water with reduced rations to provide a size at release

similar to the normal summer migrants.

Marking Procedures

Test fish were marked from 9 June to 6 July, Monday through Friday, using two

marking crews; one crew worked from 0600 to 1400 h and the second from 1430 to 2230 h.

About 60,000 fish were marked each day. The experimental design called for 14 release

blocks for each of four treatment groups, with each group consisting of about 30,000 fish.

Fewer fish than originally estimated were available at the time of marking, so the

number of release blocks was reduced to 13. Each marked group had unique CWTs. Cold

brands were used to visually identify fish from the different treatment groups.

Prior to marking, FWS personnel at Little White Salmon NFH transported

unmarked fish to a holding pond adjacent to the mobile marking trailer. Fish were dip

netted from the pond to the holding tanks in the trailer, apportioned to six marking

stations, anesthetized with tricaine methanesulfonate (MS-222), and marked. Marked fish exited the trailer via 7.6-cm diameter PVC pipes that led to subdivided holding ponds.

The following measures were taken to ensure that marked groups did not differ in

fish size, fish condition, rearing history, or mark quality: 1) the four marked groups

needed for one release block (i.e., a single night's release) were marked simultaneously;

2) differences in mark quality among groups were minimized by rotating marking

personnel between stations, and by alternating marks at each station at 4-hour intervals.

Thus each marking team and each marking station contributed equivalent numbers of

marked fish to each treatment group.

Tag Loss

To assess quality control in the tagging process, samples of about 100 fish from

each marked group were collected and checked for the presence of CWTs. These samples

were taken periodically at the outfall pipe from the marking trailer. In addition, samples

of about 10 fish from each marked group were diverted into a separate holding pond at

2-hour intervals throughout the marking day and held for a minimum of 30 days to determine tag loss and brand retention. Due to space limitations at the hatchery, a single raceway was used to hold this sample. After the holding period, these fish were passed through a tag detector, after which brands (symbol, location, and rotation) were used to assign detection results to particular treatment groups. Estimates of tag loss, based on extended holding of fish from each marked release group, ranged from 0.6 to 9.6% $(\bar{x} = 2.8\%, n = 6,429;$ Appendix Table A1). Tag loss estimates made immediately after marking were low (range 0 to 2.5%). This suggested that study fish continued to lose tags

at a high rate for several days after tagging, and that tag loss may be related to poor tag

placement in the fish (Vreeland 1990). Release data for juvenile and adult recovery

comparisons include an adjustment using estimated tag loss for marked fish held a

minimum of 30 days.

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Release Locations

The specific release locations and rationales for 1992 were as follows:

1) Bypass First Powerhouse: Test fish descended through a 10.1-cm hose and were

released about 1.5 m above the water surface of the downstream migrant collection

channel adjacent to Gatewell B of Turbine 9 (Fig. 3). Released fish encountered a

downwell at elevation 17.7 m, then passed through a 61-cm diameter by 220-m long

conduit discharging them into the tailrace about 90 m downstream from the centerline

Turbine Bypass
lease hose release hose release hose (7.6-cm diameter) Gatewell Downstream

(10.1-cm diameter)

Ice and trash sluiceway

Figure 3.--Cross section of Bonneville Dam First Powerhouse depicting release locations for the turbine and bypass treatment groups.

between Turbines 9 and 10 at elevation 0 m (3 to 7 m below the water surface

depending on tailwater elevation)¹.

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2) Turbine First Powerhouse: Test fish descended through a 7.6-cm hose through

Gatewell A and exited 1 m below the STS water-flow interception line, in the intake of

Turbine 9 (Fig. 3). This site was selected to simulate passage of fish traveling too

deep to be intercepted by an STS.

3) Turbine Second Powerhouse: As in previous years, test fish descended through a hose

through Gatewell A and exited 1 m below the STS water-flow interception line, in the

intake of Turbine 17 (Fig. 4). This site was also selected to simulate passage survival

of fish traveling too deep in the water column to be intercepted by an STS.

4) Downstream: As in previous years, test fish were released at the river surface in mid-

channel adjacent to the Hamilton Island boat launch ramp about 2.5 km downstream

from the dam. This group did not pass through the dam or tailrace basins and was

presumed to be downstream from effects of the dam and away from predators

inhabiting the shoreline. Recoveries of fish released at this site, when compared to

recoveries of fish from other treatment groups, isolate the effects of passage through

the two powerhouses or bypass system and tailraces.

The turbine release groups entered the tailrace from the turbine discharge boil which

dispersed fish over a large area (ca. 700 m^2). These were termed broadcast releases. The

bypass and downstream groups entered the river directly from a pipe or hose; these were

termed point-source releases.

All elevations are referenced to mean sea level.

Figure 4.--Cross section of Bonneville Dam Second Powerhouse depicting release site of turbine treatment groups.

Project Operating Parameters

Powerhouse operating conditions were selected to provide conditions comparable to

those employed for Second Powerhouse tests in 1987 to 1989 when the powerhouse

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operated at about one-half capacity, 53 to 60 kcfs (Appendix Table A2). These conditions

were necessary because it would be impractical for both powerhouses to run at full

capacity (about 250 kcfs) during the summer test period when low flow conditions

normally exist. Turbine Units 1, 2, 5, 8, 9, 10 (First Powerhouse²) and 11, 16, 17, and 18

(Second Powerhouse) were selected for operation. Simultaneous operation of these units

provided similar flows at both powerhouses, minimized tailrace eddies, produced high flow

past the juvenile bypass outlet, and maintained attraction flows to the fishway entrances

for upstream migrant adult salmonids. Turbines used to pass test fish were operated at

full load and maximum efficiency while other turbines were operated within 1% of

maximum efficiency from 0001 to 0800 h on test days. At other times from 18 June to

10 July water flows through the powerhouses were about equal with turbines operated

within 1% of maximum efficiency which provided comparable tailrace flow conditions.

Release Procedures

On 13 days during the period from 18 June to 9 July, test groups of about 30,000

marked fish were released at the four release sites during the early morning darkness

(0200-0300 h). The release days were selected to coincide with the migration period of

juvenile upriver bright fall chinook salmon past Bonneville Dam, and also to provide

sufficient time for marking yet not require more than 15 days holding prior to release.

Uniquely branded fish groups were released at each site during four time series:

18-20 June; 23-25 June; 29 June-2 July; and 7-9 July.

2 Operation of Unit 8 was optional depending on available river flow.

On release days, loading of transport trucks generally began at 2100 h and was

completed by about 2400 h. Fish were crowded from the holding pond into a

funnel-shaped 450-L transfer container. When sufficient fish were inside the container, a

slide gate was closed and the container was lifted over the transport truck. Next, another

gate opened, allowing fish and water to drain through the funnel into the transport truck.

It required about five lifts to load each 30,000-fish treatment group.

Three transport trucks were used on each release night. Two 17,000-L capacity

tank trucks with two compartments were used for releases at the dam. The three

treatment groups released at the dam were rotated nightly between the different tank

compartments. The tank truck used initially for the downstream release had 4,500-L

capacity; however, for the final three releases, a 5,300-L capacity tank truck was used

because test fish had grown to a size which approached maximum desirable loading

density. Fish loading densities were less than 60 g fish/L water (0.5 1b/gal) for all

releases. All releases were made from the transport tanks using smooth-bore plastic

hoses to transfer the fish to the release point; a 7.6-cm diameter by 30-m long hose for the

turbine releases, a 10.1-cm diameter by 30-m long hose for the bypass releases, and a

10.1-cm diameter by 6-m long hose for the downstream releases. Vertical distances from

the transport trucks to the water surface were about 10.7, 6.1, and 1.2 m (35, 20, and 4

ft), respectively, for bypass, turbines, and downstream releases. Bypass and downstream

release groups exited the hoses about 1.5 m above the water surface and turbine release

groups were subsurface.

Hose discharge velocities were calculated to be 5.1, 3.3, and 4.1 m/second,

respectively, for bypass, turbines, and downstream releases. Velocity differences between

water exiting the release hoses and the surrounding water were calculated to be less than

4.5 m/second. The lowest differential velocity shown to cause mortality of juvenile salmonids in laboratory tests was 15 m/second (Groves 1972). Releases were timed such that fish from both powerhouses could traverse their respective tailrace basins and pass the downstream site at about the same time as the downstream groups were released: Second Powerhouse at 0200 h; First Powerhouse at 0230 h; and the downstream release

at 0300 h.

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Sampling at Jones Beach

Short-term survival differences among release groups were assessed from

comparisons of tagged fish recovered near the upper boundary of the Columbia River

estuary at Jones Beach (RKm 75). Recovery methods and sampling site were those

described by Dawley et al. (1985, 1988). In addition to determining recovery differences,

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captured fish were observed for differences in descaling, injuries, size, food consumption,

and migration behavior.

During the period from 15 June through 31 July, sampling was conducted by two

or three crews working 7 days per week for 8 to 12 hours per day, beginning at sunrise

(Appendix Table A3). On 26-27 June, beach- and purse-seine sampling was extended

through the night to determine diel migratory behavior of juvenile salmon. Two stocks of branded/CWT subyearling fall chinook salmon were targeted in estuarine sampling: tule stock used in this study and upriver bright stock used for a concurrent study at release sites near Bonneville Hatchery (Ledgerwood et al. In prep). One group from each stock was released at or just downstream from Bonneville Dam, and within a few hours of one

another. The upriver bright stock group was released at about 2200 h on 19 June, and the tule stock group was released at about 0300 h on 20 June. Releases at the same site

and on the same day provided the opportunity to compare biological characteristics of the

two stocks prior to release and behavioral characteristics after migration to the estuary.

Both purse seines (midstream) and beach seines (Oregon shore) were used to determine

whether study fish were more abundant in midstream or near shore (Fig. 5) and to

maximize effort using the gear type that captured the greatest numbers of study fish.

All captured fish were processed aboard the purse seine vessels. The catch from

each set was anesthetized and enumerated by species. Numbers of dead, injured, or

descaled salmonids were recorded, and subyearling chinook salmon were examined for

excised adipose fins and brands. Marked fish were separated for further processing, while

unmarked fish were returned to the river immediately after counting, evaluation, and

recovery from anesthesia. Descaling was judged rapidly while counting and separating

study fish from non-study fish. Fish were classified as descaled when 25% or more of

their scales on one side were missing.

Freeze brands were used to identify study fish; from these fish we collected CWTs,

obtained biological samples, compared fish size among treatment groups, and adjusted the

daily sampling effort to attain the desired minimum sample size of 0.5% of the number of

fish released. Brand information, biological and associated sampling data (i.e., date,

vessel code, gear code, set number, time of examination, fork length, and incidence of

descaling and mortality) were immediately entered into a computer database and printed.

Fork lengths of marked fish were recorded to the nearest mm. All brand- identified study

fish (including those with illegible brands) were sacrificed to obtain CWTs, which

identified treatment group and day of release.

The heads of branded fish were processed in lots, which were segregated by

recovery day and site of capture. An aqueous solution of 40% potassium hydroxide was

used to dissolve the heads for ease in extracting CWTs. All CWTs were decoded and later

verified; additional details of tag processing followed the methods described by

Ledgerwood et al. (1990).

Data standardization procedures--All catch data obtained from 19 June to

22 July were adjusted to obtain a standard catch per day per group. Purse-seine data

were standardized to a 10-set-per-day effort, while beach-seine data were standardized to

a 5-set-per-day effort. The following formula was used to calculate a standardized catch

per day for each group:

$$
A_i = N_i (S \div P_i)
$$

where:

 A_i = Standardized purse or beach seine catch on day i

 N_i = Actual purse or beach seine catch on day i

S = Constant (weighted daily average number of purse seine sets (10) or

beach seine sets (5) during the sampling period)

 P_i = Actual number of purse or beach seine sets on day i.

On the day when there was no sampling effort for a particular gear type (e.g., beach seine, 25 June), the standardized catch was derived by averaging standardized catches for 1 day prior to and 1 day after the missed day. Dates of median recovery for each marked fish group were determined using the combined standardized data from purse and beach seine catches. Movement rates for each CWT group were calculated as the distance from the downstream release site (RKm 232) to Jones Beach (RKm 75) divided by the travel time

(in days) from release date to the date of the median fish recovery.

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Biological Samples and Assessments--Samples for physiological analyses of

marked fish groups prior to release were collected by NBS personnel (Philip Haner, Cook,

WA) at Bonneville Hatchery (upriver bright stock) and Little White Salmon NFH (tule

stock) and after migration to Jones Beach. At the hatcheries, sample fish $(n = 30)$ were

netted directly from the holding raceways. Physiological samples were obtained at Jones

Beach during the diel sampling period on 26-27 June. Fish were identified as to stock by freeze brand, and immediately placed in a lethal dose of MS-222. Fork length and body weight were recorded and a gill sample frozen for later gill Na⁺-K⁺ ATPase analysis.³ The fish were then videotaped to measure reflectance⁴ and frozen in liquid nitrogen to remove a skin sample for guanine analysis (Beeman et al. 1990). Reflectance is an experimental non-lethal technique to measure silvering of juvenile salmonids as an appraisal of smoltification; the guanine sample was used for confirmation. Stomachs from 116 upriver bright stock and 126 tule stock fish were collected

Haner, P. V., J. C. Faler, R. C. Schreck, and D. W. Rondorf. Unpublished. Skin reflectance as a non-lethal measure of smoltification for juvenile salmonids. (Available from National Biological Survey, MP 5.48L Cook-Underwood Rd. Cook, WA 98605-9701.)

during the diel sample period. Stomachs were excised (esophagus to pyloric caeca) and

cleaned of external fat. A stomach fullness value, based on the proportion of the total

stomach length containing food, was estimated. A scale of 1 to 7 was used to quantify

the fullness as follows: $1 =$ empty, $2 =$ trace of food, $3 =$ one-quarter full, $4 =$ one-half full,

 $5 =$ three-quarters full, $6 =$ full, and $7 =$ distended full (Terry 1977). All stomachs

appearing empty were opened for examination, and a value of 2 was assigned if traces of

³ For details of methodology see Schreck, R. C., J. W. Beeman, D. W. Rondorf, and P. V. Haner. A microassay for gill sodium, potassium-activated ATPase in juvenile salmon. Trans. Am. Fish Soc. In press. (Available from National Biological Survey, MP 5.48L Cook-Underwood Rd, Cook, WA 98605-9701.)

food were observed. Subsamples of collected stomachs were preserved in 10% buffered

formaldehyde solution for weight determination and content analysis as described by Kirn

et al. (1986a). Holding time prior to fullness observations was about 35 minutes.

Statistical Analysis

Differences among recovery percentages for each tagged group at Jones Beach were

evaluated by analysis of variance (ANOVA) using a randomized block design where each

release day was considered a block (Sokal and Rohlf 1981). Transformation of

percentages was not required. Differences among descaling percentages of branded

groups were also evaluated using ANOVA. Fisher's protected least significance difference

procedures were used to rank treatment means for significant F-tests (Petersen 1985).

Chi-square goodness of fit was used to test the hypothesis that different marked groups

released the same day had equal probability of capture through time (Zar 1974). The

mean values of physiological samples (gill Na⁺-K⁺ ATPase, reflectance, and guanine

measurements) were compared using a General Linear Model to test for significant

differences among stocks (P < 0.05).

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RESULTS

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In 1992, a total of 1,540,863 fish were marked with freeze brands, CWTs, and

excision of the adipose fin (Table 1). A total of 5,063 study fish were recovered in the

estuary (ca. 0.3% of those released); most were midriver migrants captured with purse

seines (Appendix Table A4). Handling mortality of all captured juvenile salmon was less

than 0.5% and descaling rates averaged less than 2%. Only four descaled study fish were

captured at Jones Beach, too few for meaningful among-treatment comparison.

Movement Rates and Temporal Catch Patterns

Temporal catch distribution of treatment groups released each day are shown in

Figures 6 and 7, and in Appendix Figures A1-A3. Movement rates of study fish between

the release site at Bonneville Dam and the collection site at Jones Beach, except for fish released during the final series (7-9 July), ranged from 17 to 39 km/day (Table 2). These

rates were somewhat faster than those observed in previous years; however, during the

final release series, fish slowed noticeably (movement rates 12.1 to 15.7 km/day) despite

about a 17% increase in river flow from the previous series. There were no indications of

movement rate differences among treatment groups. Comparisons of fork-length

distributions of study fish at release to those at Jones Beach suggest that all fish grew

during migration; fish from the final release series were largest (Figs. 8-9). There were no

indications of temporal differences in size among treatment groups at recovery

(Figs. 10-11). However, fish from the first three release series (18-20 June, 23-25 June,

and 29 June-2 July) generally increased in mean length during the period of recovery,

while fish from the final release series generally decreased in mean length. Water

temperature at Little White Salmon NFH remained a nearly constant 10°C throughout the marking/holding period of study fish, whereas water temperatures of the Columbia

Table 1.--Summary of releases of marked subyearling chinook salmon, Bonneville Dam survival study, 1992.

Second Powerhouse Turbine-17 releases

First Powerhouse Bypass releases

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Table 1.--Continued.

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- a Brand position (RD = right dorsal, LD = left dorsal), brand used (two-letter combination), and brand rotation (1 or 3).
- Total fish marked; branded, tagged, and adipose fin clipped. $\mathbf b$
- Estimated number of fish released without coded-wire tags. See Appendix Table A1 \mathbf{c} for tag loss sample data.
- d Estimated number of fish released with coded-wire tags.
- AG D1 $D2$ = Agency, Data 1, Data 2. \bullet

Totals 1,540,863 44,062 1,496,801

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Figure 6.--Daily recoveries of test fish by treatment (standardized for enort) at Jones Beach, 1992. Data shown are from the first two release series.

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median

Beach, 1992. Data shown are from the second two release series. Beach, 1992. Data shown are from the second two release series.

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Table 2.--Movement rates from Bonneville Dam to Jones Beach for marked groups of subyearling chinook salmon, Bonneville Dam survival study, 1992.

- a Purse seine plus beach seine recoveries standardized to a constant daily effort (Appendix Table A4). Movement rate = distance from the downstream release site (RKm 232) to recovery site (RKm 75) ÷ travel time in days from release to median fish recovery.
- Fish released during early morning darkness. b
- Average flow through Bonneville Dam within 4 days of the date that the median fish c was captured; by convention, English units were used for river flow volumes $(k[•]ft³/second = 1,000 ft³/second = 28.3 m³/second); flow data courtesy Sonja Dodge,$ COE, Water Management Division, Portland, OR.

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Released 18-20 Jun

Hatchery, n = 609 Jones Beach, n = 385

Figure 8.--Fork length distributions (2 point moving average) of fish at release and after recovery in the estuary, first two release series, 1992.

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Released 7-9 Jul

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Fork length (mm)

Figure 9.--Fork length distributions (2 point moving average) of fish at release and after recovery in the estuary, second two release series, 1992.

Released 18-20 June

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 \mathbf{b} .

Figure 10. Beach, comparing treatments from the first two release series, 1992. Beach, comparing treatments from the first two release series, 1992.

Released 20 June - 2 July

Figure 11. Daily mean fork lengths of subspectively at $\frac{1}{2}$ become two release series, 1992. Beach, comparing treatments from the second two release series, 1992.

River at Jones Beach increased from 17 to 22°C through the recovery period, substantially higher than in previous years (Fig. 12). During the final release series, we speculated that elevated Columbia River water temperatures shocked study fish, or slowed their movement rate, and increased mortality of fish among all treatment groups.

Diel Recovery Patterns

During the diel sampling period, about 12,000 and 10,000 subyearling chinook

salmon (primarily non-study fish) were captured in the beach seine and in purse seines,

respectively (Appendix Table A5). In the purse seines, catches were highest at sunrise,

generally decreased during the day, increased again at dusk, and were lowest at night

(Fig. 13). In the beach seine, catches peaked about 3 hours after sunrise, declined during

the afternoon, increased again in late afternoon, and were also lowest at night. The

pattern of very low catches during darkness for both gear types is similar to patterns

observed in previous years at Jones Beach (Ledgerwood et al. 1991a, b).

During the diel sampling period, a total of 429 branded upriver bright stock and 502

branded tule stock subyearling chinook salmon were captured in beach and purse seines

(Appendix Table A6). There were no apparent differences between stocks in purse seine or beach seine diel catch patterns (Fig. 14). Catch patterns of both stocks followed the

general pattern exhibited by unmarked subyearling chinook salmon except for the

vagaries associated with small sample size.

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Smoltification

During the migration to Jones Beach, tule stock and upriver bright stock both

exhibited significant $(P < 0.05)$ increases in smoltification indicators (gill Na⁺-K⁺ ATPase

activity, reflectance, and guanine) (Table 3). Upriver bright stock had significantly higher

Columbia River temperature at Jones Beach

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Figure 12... Temperature of the Columbia River at Jones Beach and total flow of Columbia River at Bonneville Dam, 1989-1992. By convention, English units were used for river flow volumes (kcfs $= 1,000$ ft σ /second $=$ 28.3 m σ /second).
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Figure 13.--Diel catch patterns for unmarked subyearling chinook salmon captured in beach and purse seines at Jones Beach, 1992.

Figure 14.--Diel catch patterns of tule stock and upriver bright stock subyearling chinook salmon captured in beach and purse seines at Jones Beach, 26-27 June 1992.

Table 3. .-Smoltification parameters measured for tule stock (Little White Salmon National Fish Hatchery) and upriver bright stock (Bonneville Hatchery) subyearling chinook salmon prior to release at the hatcheries and after migration to Jones Beach; mean values \pm SE, n = 27-30 fish per sample.

a Reflectance is measured on a shade of gray ranging from 0 to 10.

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ATPase and guanine values both prior to release and after migration to Jones Beach than

did tule stock (P < 0.05); differences in reflectance between the two stocks were

insignificant. Efforts to establish a relationship between measurements of skin

reflectance and other measures of smoltification are ongoing (Philip Haner, NBS,

Columbia Field Station, Cook, WA. Pers. commun.).

Stomach Fullness and Diet Composition

Based on examination of stomach fullness of subsamples of marked fish, study fish

were feeding by the time they arrived at Jones Beach. Stomachs were generally about

half full in fish collected during daylight hours. Upriver bright stock had slightly higher

fullness values than tule stock sampled concurrently (mean fullness 4.0 and 4.5,

respectively; Fig. 15). During the diel sampling period, mean weights of stomach contents

in upriver bright stock were generally higher than for tule stock (Fig. 16). In both stocks,

mean weights of stomach contents increased during the morning hours, declined

somewhat during the afternoon, and were lowest at night, similar to observations made in

1989 and 1990 (Ledgerwood et al. 1990, 1991a).

Analysis of stomach contents showed that crustaceans and insects were the

dominant prey items in the diet of both upriver bright and tule stock fall chinook salmon

(Fig. 17; Appendix Table A7). Small cladocerans were numerically dominant, although

larger crustaceans (amphipods and mysids) and two orders of insects (Psocoptera and

Diptera) were important dietary components based on their larger size. These principal

dietary components for subyearling chinook salmon were similar to previous years

(Ledgerwood et al. 1990, 1991a; Kirn et al. 1986a). Although numbers of prey items

fluctuated considerably, there were no apparent diel differences in diet composition and,

except for a greater number of cladocerans in beach-seine captured fish, there were no

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Figure 15.--Relative stomach fullness for upriver bright stock and tule stock subyearling chinook salmon captured at Jones Beach, 1992.

Figure 16.-- Mean weight of stomach contents (g) of tule stock and upriver bright stock subyearling chinook salmon captured in beach and purse seines at Jones Beach, 26-27 June 1992.

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Figure 17.-- -- Primary dietary components of tule stock and upriver bright stock subyearling chinook salmon captured at Jones Beach, 1992.

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apparent dietary differences between fish captured along the shoreline (beach seine) and

in midstream (purse seine).

Juvenile Recovery Differences

Estuarine recovery percentages of the 30,000-fish treatment groups released daily at

each site ranged from a high of 0.5000 to a low of 0.1590 (Table 4). Recovery percentages

decreased over time, but proportional differences among treatments were fairly consistent

through the period of testing and provided statistically significant estimates of relative

differences for passage survival. Statistical analysis of migrational timing differences for

treatment groups released on the same day showed no significant difference among any of

the 13 release blocks (α = 0.05), and no difference when blocks were pooled (P = 0.8228;

Appendix B). Thus, there is no evidence to suggest non-homogeneity between treatment

that there were significant differences ($\alpha = 0.05$) in mean recovery percentages among the various treatment groups (Table 4). Rank order (from lowest to highest) was as follows: First Powerhouse bypass and tailrace, Second Powerhouse turbine and tailrace, First Powerhouse turbine and tailrace, and downstream, with mean recovery percentages of 0.31, 0.31, 0.35, and 0.43, respectively. Recovery percentages for the downstream groups were significantly greater than for all other groups, while recovery percentages for the First Powerhouse turbine and tailrace were significantly higher than for the Second Powerhouse turbine and tailrace and the First Powerhouse bypass and tailrace. Recovery

recovery distributions.

Statistical analyses of CWT-fish recoveries at Jones Beach (Appendix B) indicated

difference between the Second Powerhouse turbine and tailrace and the First Powerhouse

bypass and tailrace was not significant. Conclusions regarding differences among mean

Table 4.--Recovery percentages of tagged subyearling juvenile fall chinook salmon at Jones Beach, Bonneville Dam survival study, 1992.

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a Fish were released in early morning darkness.

- ^b Weighted equally by block (i.e., by release day).
- ^e Adjusted for tag loss.
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- Observed catch, purse seine plus beach seine.
- e Compared to downstream release =

(Treatment % - Downstream % + Downstream %) $*$ 100.

recovery percentages derived from the standardized data were similar to those reached with the non-standardized data (Fig. 18.)

Throughout the study, the rank-order of recovery percentages for the various

treatment groups was generally consistent among treatments and between blocks (days of

release) (Table 4). This general consistency contributed to improved overall statistical

power (differences >7.6% were detectable) despite the rather disappointing recovery

percentages of study fish (grand mean = 0.35% recovery) and the forced elimination of 1 of

the 14 proposed release blocks. For example, recovery percentages of

downstream-released groups ranked highest in 12 of 13 release blocks and recovery

percentages of bypass-released groups ranked lowest or next to lowest in 12 of the 13

release blocks. Based on variability observed in juvenile recovery data from 1988 to 1990,

14 release blocks of 30,000 fish/treatment would be needed to detect about an 8.5% annual

difference and about a 4.3% difference for 4 years combined data at $\alpha = 0.05$, $\beta = 0.2$.

DISCUSSION

The 1992 study was conducted under conditions of regional drought, and the

resulting low flows and elevated water temperatures may have affected the comparisons

by severely stressing test fish and contributing to increased predation by northern

squawfish in the tailrace areas of both powerhouses. The overall recovery percentage of

1992 test fish at Jones Beach (mean 0.35%) was lower than the average recovery

percentage from 1987 through 1990 (0.6%).

During the first week of the study (20 June release), another study that used

upriver bright stock was conducted coincidentally near Bonneville Dam. Results from this

study allowed direct comparisons of migration behavior and biological/physiological

Figure 18. --Mean recovery percentages for treatment groups of tagged subyearling chinook salmon, 1992; total catch and total catch standardized for sampling effort.

parameters between tule stock and upriver bright stock. During this period, tule stock were similar in size to upriver bright stock (mean fork lengths at release 83.8 vs. 85.0 mm, respectively) but had lower recovery rates (0.47 vs. 0.60%, resectively). Although tule stock and upriver bright stock had similar diets (Fig. 17), tule stock had lower food consumption rates (Figs. 15-16) and were less smolted (Table 3) than the

upriver bright stock. These differences may be indicative of a generally lower survival

rate for tule stock relative to upriver bright stock. Aberrant, drought-related conditions of

high water temperature (ca. 21°C) and low flows (<150 kcfs) seemed particularly

important for test fish (tule stock) during the final release series (fish released 7, 8, and

9 July). These fish had recovery percentages as low as 0.16%, and passed Jones Beach

when river temperatures generally exceeded 21°C. We believe the low rearing-water

temperature of 10°C, coupled with the elevated river water temperatures during the

migration period of fish in the last release series, severely shocked test fish and increased

mortality and susceptibility to predation among all treatment groups. As a consequence,

we anticipate poor adult recovery percentages for fish from the final release series.

While we speculated that mortalities of test fish related to different routes of

passage at Bonneville Dam were fully expressed in survival differences among marked

groups of juveniles sampled at Jones Beach, we question whether the adult recoveries of

tule stock will provide sufficient data for the supplemental evaluation. However, the

cold-water rearing with reduced rations successfully produced test fish of appropriate size

to simulate the normal summer migrants. We feel that size is the single most important

factor affecting differences in passage survival of test fish at Bonneville Dam. Fish size is

directly related to incidence of physical contact by structures during passage through the

bypass system or turbines and predation rates in the tailrace. These survival differences

should be fully expressed in mark-recovery differences among release groups after their

157-km migration to Jones Beach.

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Effects of River Flow and Powerhouse Discharge

Passage route survival for juvenile salmonids at Bonneville Dam may vary due to

alterations of flow distribution among the two powerhouses and spillway and annual

variations of river flow. Water-flow management at Bonneville Dam is complicated by the required operation of one turbine at the First Powerhouse (for station services) which results in a necessity for additional turbine operation to provide protection flow at the bypass outlet and attraction flow for the fishway entrances. Thus, water flow redistribution to include use of the Second Powerhouse requires a minimum three-turbine operation at the First Powerhouse. The experimental design with about half of each powerhouse in operation is a fairly realistic water distribution probability for summertime use of the Second Powerhouse. However, to include variations of river flow volumes into

the variables being tested, multiple years of testing would be required.

In previous tests at the Second Powerhouse, annual variations in passage survival

were related to differences of river flow. During the multiyear study at the Second

Powerhouse, survival of fish through the bypass increased with increased tailwater

surface elevation (Ledgerwood et al. 1991a). The mechanisms affecting survival difference

were thought to be water velocity in the bypass conduit and shear forces at the outlet of

the bypass pipe. Similarly, survival through the bypass system at the First Powerhouse

may be dependent on tailwater surface elevation which is directly correlated to river flow.

Drought conditions in 1992 may have produced a worst-case survival scenario for summer

migrants passing through the bypass system of the First Powerhouse by creating

conditions of low tailwater surface elevation, increased predation, and greater stress on the fish.

We assume that survival through the turbines at the Second Powerhouse in 1992 was similar to that of previous years (97.0-98.5%) and that the increased difference in recovery percentage between turbine and downstream releases (from the 9% average

difference in previous years to about 26% difference in 1992) was due to increased

predation in the tailrace. In previous years, flows through the Second Powerhouse were

intermittent, generally occurring at night and usually beginning 1 day prior to tests with

no operation between test periods, while flows through the First Powerhouse were

continuous. In 1992, equalized continuous flows may have attracted additional predators to the Second Powerhouse tailrace. Another factor contributing to the increased difference

in recovery percentages may have been increased susceptibility of tule stock test fish to

predation due to high stress resulting from low rearing-water temperature and elevated

river-water temperatures during the test period.

Impacts from Northern Squawfish

Increased abundance of northern squawfish in the lower Columbia River during

recent years (Kirn et al. 1986b) may be severely impacting juvenile salmonids, especially

near Bonneville Dam (Petersen et al. 1990). These impacts were documented by the NBS

during the survival study releases made on 19 June, 25 June, 1 July, and 8 July. On

these dates, beginning about 1 hour after releases of study fish, electrofishing efforts in

the tailrace areas of both the First and Second Powerhouses produced 649 northern

squawfish (Poe et al. 1993).

Of the juvenile salmonids found in the stomachs of these northern squawfish, 441

were CWT fish from the survival study releases (251, 74, and 116 CWTs each, for fish

released into the First Powerhouse bypass, First Powerhouse turbine, and Second

Powerhouse turbine, respectively). These observations of northern squawfish stomach

contents were similar to those made in 1990, when CWT fish released into the bypass at

the Second Powerhouse were more numerous in the stomachs of northern squawfish

collected in the tailrace than turbine- or egress⁵-released fish (Ledgerwood et al. 1991a).

In both years, stomach content analysis supported speculation that predation by northern

squawfish contributed to the apparent lower survival of bypass-released fish compared to

turbine-released fish.

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CONCLUSIONS

The following conclusions are based on 1 year of study. Special operating conditions

of equalized flow through both powerhouses were implemented at Bonneville Dam for this

study to 1) attract predators equally to the two tailrace areas, 2) provide an unbiased

comparison of survival among the various routes of juvenile fish passage as well as to

minimize tailrace eddies, 3) provide high flows past the juvenile bypass outlet, and

4) allow adequate attraction flows to the various fishway entrances for upstream migrant adult salmonids. The regional drought that severely reduced river flow during 1992 may have created a worst-case scenario for salmonid survival due to heavy predation of test fish in the tailrace areas. It is important to consider a wide range of test conditions

before formulating conclusions regarding the safest routes for juvenile salmon passing

Bonneville Dam during the summer.

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The egress-released fish were expelled through a hose into the bypass discharge plume. These releases were designed to introduce fish into the tailrace at the location of the bypass exit, but without having passed through the bypass system.

Several tentative conclusions include:

1) Under the drought conditions of 1992, recoveries of marked subyearling chinook

salmon in the estuary indicated significantly reduced survival of fish released into the

bypass system at the First Powerhouse compared to fish released into the First

Powerhouse turbines or fish released downstream from the tailrace.

- 2) Fish passing through the Second Powerhouse turbines and tailrace had significantly reduced survival compared to fish passing through the First Powerhouse turbines and tailrace.
- 3) The downstream-released fish had significantly higher survival than all other release groups.
- 4) Tule stock subyearling chinook salmon used in this study were subjected to cold-water rearing and reduced rations to maintain a size range at release similar to normal summer migrants (upriver bright stock). However, test fish, particularly those from

the final week of test releases, may have suffered extreme stress due to elevated

Columbia River water temperatures resulting from the regional drought. While the

immediate impacts of dam passage are thought to be fully expressed in mark-recovery

differences among juvenile fish recovered at Jones Beach, the overall survival of test

fish may have been reduced by temperature stress. This potential overall lower

survival of test fish may affect comparisons among treatment groups using CWT data

from adult contributions to the various ocean and river fisheries and returns to the

lower river hatcheries.

5) Because 75 to 90% of the summer migrating juvenile salmon encountering the

powerhouses at Bonneville Dam pass through turbines instead of bypass systems, and

because of the significant difference between turbine plus tailrace passage survivals at

the First and Second Powerhouses, it is extremely important to identify the safest

passage route over a wide range of river flows.

RECOMMENDATIONS

- 1) Tag recovery of adults should be compiled through 1997 to assess passage survival differences adequately.
- 2) The study should be repeated for 3 additional years to bracket a wide range of river

flow and other physical conditions before making final conclusions regarding the

relative survival of summer migrating subyearling chinook salmon through the various

passage routes at Bonneville Dam.

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REFERENCES

Allis-Chalmers Corp. 1978. Bonneville Second Powerhouse model test report. U.S. Army Corps of Engineers, Portland, OR. 400 p.

Beeman, J. W., D. W. Rondorf, J. C. Faler, M. E. Free, and P. V. Haner. 1990. Assessment of smolt condition for travel time analysis. Report to Bonneville Power Administration, 103 p. (Available from U.S. Fish and Wildlife Service, MP 5.48L Cook-Underwood Rd, Cook, WA 98605-9701.)

Bergman, P. K., K. B. Jefferts, H. F. Fiscus, and R. C. Hager. 1968. A preliminary evaluation of an implanted coded wire fish tag. Wash. Dep. Fish., Fish. Res. Pap. 3(1):63-84.

Cramer, F. K. 1965. Fish passage through hydraulic turbines. U.S. Army Corps of Engineers, Walla Walla Dist., Walla Walla, Washington 99362-9265.

Dawley, E. M., L. G. Gilbreath, and R. D. Ledgerwood. 1988. Evaluation of juvenile salmonid survival through the Second Powerhouse turbines and downstream migrant bypass system at Bonneville Dam, 1987. Report to U.S. Army Corps of Engineers, Contract DACW57-87-F-0323, 36 p. plus Appendix. (Available from Northwest Fisheries Science Center, 2725 Montlake Blvd. E., Seattle, WA 98112-2097.)

Dawley, E. M., L. G. Gilbreath, R. D. Ledgerwood, P. J. Bentley, B. P. Sandford, and M. H. Schiewe. 1989. Survival of subyearling chinook salmon which have passed through the turbines, bypass system, and tailrace basin of Bonneville Dam Second Powerhouse, 1988. Report to U.S. Army Corps of Engineers, Contract DACW57-87-F-0323, 78 p. (Available from Northwest Fisheries Science Center, 2725 Montlake Blvd. E., Seattle, WA 98112-2097.)

- Dawley, E. M., R. D. Ledgerwood, and A. L. Jensen. 1985. Beach and purse seine sampling of juvenile salmonids in the Columbia River estuary and ocean plume, 1977-1983. Volume I: Procedures, sampling effort, and catch data. U.S. Dep. of Commer., NOAA Tech. Memo. NMFS N/NWC-74:1-260.
- Fisher, R. A. 1944. Statistical methods for research workers. Ninth edition. Oliver and Boyd Ltd., London 350 p.
- Gessel, M. H., D. A. Brege, B. H. Monk, and J. G. Williams. 1990. Continued studies to evaluate the juvenile bypass system at Bonneville Dam--1989. Report to U.S. Army Corps of Engineers, Contract E8689-95, 20 p. plus Appendix. (Available from Northwest Fisheries Science Center, 2725 Montlake Blvd. E., Seattle, WA 98112-2097.)

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Gessel, M. H., B. H. Monk, D. A. Brege, and J. G. Williams. 1989. Fish guidance efficiency study at Bonneville Dam First and Second Powerhouses--1988. Report to U.S. Army Corps of Engineers, Contract DACW57-87-F-0322, 36 p. (Available from Northwest Fisheries Science Center, 2725 Montlake Blvd. E., Seattle, WA 98112-2097.)

Gilbreath, L. G., E. M. Dawley, R. D. Ledgerwood, P. J. Bentley, and S. J. Grabowski. 1993. Relative survival of subyearling chinook salmon that have passed Bonneville Dam via the spillway or the Second Powerhouse turbines or bypass system: adult recoveries through 1991. Report to U.S. Army Corps of Engineers, Contract E96910013, 25 p. (Available from Northwest Fisheries Science Center, 2725 Montlake Blvd. E., Seattle, WA 98112-2097.)

Groves, A. B. 1972. Effects of hydraulic shearing action on juvenile salmon. Unpubl. manuscr. 7 p. Northwest Fisheries Science Center, 2725 Montlake Blvd. E., Seattle, WA 98112-2097.

Holmes, H. B. 1952. Loss of salmon fingerlings in passing Bonneville Dam as determined by marking experiments. Unpubl. manuscr. 62 p. U.S. Fish and Wildlife Service, Portland, Oregon.

Kirn, R. A., R. D. Ledgerwood, and A. L. Jensen. 1986a. Diet of subyearling chinook salmon (Oncorhynchus tshawytscha) in the Columbia River estuary and changes effected by the 1980 eruption of Mount St. Helens. Northwest Science 60:191-196.

Kirn, R. A., R. D. Ledgerwood, and R. A. Nelson. 1986b. Increased abundance and food consumption of northern squawfish (Ptychocheilus oregonensis) at River Kilometer 75 in the Columbia River. Northwest Science 60:197-200.

Ledgerwood, R. D., E. M. Dawley, L. G. Gilbreath, P. J. Bentley, B. P. Sandford, and M. H. Schiewe. 1990. Relative survival of subyearling chinook salmon which have passed Bonneville Dam via the spillway or the Second Powerhouse turbines or bypass system in 1989, with comparisons to 1987 and 1988. Report to U.S. Army Corps of Engineers, Contract E85890024/E86890097, 64 p. plus Appendixes. (Available from Northwest Fisheries Science Center, 2725 Montlake Blvd. E., Seattle, WA 98112-2097.)

Ledgerwood, R. D., E. M. Dawley, L. G. Gilbreath, P. J. Bentley, B. P. Sandford, and M. H. Schiewe. 1991a. Relative survival of subyearling chinook salmon that have passed through the turbines or bypass system of Bonneville Dam Second Powerhouse, 1990. Report to U.S. Army Corps of Engineers, Contract E86900104, 90 p. (Available from Northwest Fisheries Science Center, 2725 Montlake Blvd. E., Seattle, WA 98112-2097.)

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Ledgerwood R. D., E. M. Dawley, L. G. Gilbreath, L. T. Parker, T. P. Poe, and H. L. Hansen. In prep. Effectiveness of predator removal for protecting juvenile fall chinook salmon released from Bonneville Hatchery, 1992. In C. F. Willis and A. A. Nigro (editors), Development of a system-wide predator control program: stepwise implementation of a predator index, predator control fisheries, and evaluation plan in the Columbia River Basin. Report to Bonneville Power Administration, Contract DE-BI79- 90BP07084.

Ledgerwood, R. D., F. P. Thrower, and E. M. Dawley. 1991b. Diel sampling of migratory juvenile salmonids in the Columbia River Estuary. Fishery Bulletin, 89:69-78.

Mighell, J. H. 1969. Rapid cold-branding of salmon and trout with liquid nitrogen. J. Fish. Res. Board Can. 26:2765-2769.

Oligher, R. C., and I. J. Donaldson. 1966. Fish passage through turbines: tests at Big Cliff hydroelectric plant. U. S. Army Corps of Engineers Walla Walla Dist., Walla Walla, Washington, Progress Report 6, 15 p.

Petersen, J. H., D. B. Jepsen, R. D. Nelle, R. S. Shively, R. A. Tabor, and T. P. Poe. 1990. System-wide significance of predation on juvenile salmonids in Columbia and Snake River reservoirs. Report to Bonneville Power Administration, Project 90-078, 53 p.

Terry, C. 1977. Stomach analysis methodology: still lots of questions, p. 87-92. In C. A. Simenstad and S. J. Lipovsky (editors), Fish food habits studies: First Pacific Northwest Technical Workshop, Proceedings, 13-15 October 1976, University of Washington, Div. Mar. Resources, Wash. Sea Grant, WSG-WO 77-2.

Petersen, R. G. 1985. Design and analysis of experiments. Marcel Dekker. New York, NY, 429 p.

Poe, T. P., M. G. Mesa, P. S. Shively, and R. D. Peters. In press. Development of

biological criteria for siting and operation of juvenile fish bypass systems: implications for protecting juvenile salmonids from predation. U.S. Fish and Wildlife Service, Columbia River Field Station, MP 5.48L Cook-Underwood Road, Cook, WA. Proceedings American Fishery Society symposium in Portland, Oregon in September 1993.

Ruggles, P. C. 1985. Can injury be minimized through turbine design? Hydro Review, Winter:70-76.

Schoeneman, D. E., R. T. Pressey, and C. O. Junge. 1961. Mortalities of downstream migrant salmon at McNary Dam. Trans. Am. Fish. Soc. 90:58-72.

Sokal, R. R., and F. J. Rohlf. 1981. Biometry, 2nd. Edition. W. H. Freeman. San Francisco, CA, 776 p.

Vreeland, R. R. 1990. Random-sampling design to estimate hatchery contributions to fisheries. Am. Fish. Soc. Symposium 7:691-707.

Willis, C. F., and A. A. Nigro (editors). 1993. Development of a system-wide predator control program: stepwise implementation of a predation index, predator control fisheries, and evaluation plan in the Columbia River Basin. Report to Bonneville Power Administration, Contract DE-BI79-90BP07084, 599 p. (Available from Oregon Department of Fish and Wildlife, 2501 SW First Ave. Portland, OR 97207.)

Zar, J. H. 1974. Biostatistical analysis. Prentice-Hall. Englewood Cliffs, NJ, 620 p.

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Appendix A Marking, Release, and Recovery Information

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Appendix Table A1.--Tag loss estimates among branded groups of subyearling chinook salmon after a 30-day holding period; Bonneville Dam Survival Study, 1992.

Release CWT dates Brand^b AGD1D2 AGD1D2 AGD1D2 AGD1D2 NCWT^c Sample^d Turbine 1st Powerhouse

- CWT = coded wire tag; where $AG =$ agency code, $D1 =$ data 1, $D2 =$ data 2. b Brand position RD (right dorsal) or LD (left dorsal) followed by the two-letter brand symbol; the numbers 1 or 3 indicate brand rotation.
- NCWT = Number of branded fish in the sample with no coded-wire tag.
d. Number of branded fish sheeled for the presence of saded wire tage.
	- Number of branded fish checked for the presence of coded-wire tags.

23-25 Jun RDLC3 232742 232743 232744 -- 25 368

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Appendix Table A3.--Daily purse seine and beach seine fishing effort, water temperatures, and Secchi disk transparency measurements at Jones Beach, 1992.

 $a = data not available.$ b First recovery of study fish.

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to a 5 set per day effort. Due to $\sf H$ Ξ Data 2 code. aily purse seine or beach seine catch. code, $\overline{}$ data standardized to Data Agency code,

(RKm 232, downstream release date site minus RKm 75, Jones Beach sampling site) $+$ travel time (in days, from release to date of median fish recovery at Jones Beach).

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= Date that the median fish was captured (total catch, adjusted effort).

Mvmt rate = Movement rate (km/day) = distance traveled (RKm 232, downstream release site minus RKm 75, Jones Beach sampling site) travel time (in days, from release date

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Appendix Table A5.--Diel catch results from purse and beach seine sampling at Jones Beach through a 24-hour period, 26-27 June 1992.

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Purse--Rosa 27 Jun 1056 06 107
Subtotal 17 4,714 -1.66 Subtotal 17 4,714

urse seine 25 9,872 Total purse seine

Appendix Table A6.--Marked recoveries of tule stock and upriver bright stock subyearling chinook salmon in purse seines and beach seines at Jones Beach during the diel sampling period, 26-27 June 1992.

 $N_{\rm{max}}=10^{-10}$

Standardized time intervals used for plotting diel catch curves. b na = data not available (no sampling effort).

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Copepoda

Mysidacea

Mysidacea

Copepoda

C Insecta (Order)

Diptera

Homoptera

Hemiptera Psocoptera

Psocoptera Hymenoptera

Hymenoptera Collumbola Lepidoptera

Collumbola Lepidoptera Amphipoda

Amphipoda Crusta

Amphipoda Copepoda

Mysidacera Copepoda

Cladocera Subtotal Insect
Crustacea (Ord \mathbf{g} Appendix Tab S sample date Set time S time Insecta (Order) Subtotal Insectation Subtotal Crustacea (Order) Subtotal Crusta Upriver bright stock-purse seine Upriver bright stock-beach seine

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Appendix Table
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Sample date
Set time
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Tule stock-beach

Sample date

Set time

Diptera

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Appendix Figure A1.--Daily recoveries of test fish at Jones Beach (standardized for effort) from releases made on 18, 20, and 23 June 1992.

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Appendix Figure A2.--Daily recoveries of test fish at Jones Beach (standardized for effort) from releases made on 25, 29, and 30 June 1992.

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Appendix Figure A3.--Daily recoveries of test fish at Jones Beach (standardized for effort) from releases made on 2, 7, and 9 July 1992.

Appendix B

Statistical Analysis of Juvenile Catch Data

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APPENDIX B

Statistical Analysis of Juvenile Recovery Data

A. Chi-square goodness of fit analysis was used to evaluate differences among observed recoveries (Appendix Table A4) through time for different treatment groups released on the same day (Sokal and Rohlf 1981). A non-significant result indicated that there was equal probability of capture at Jones Beach for each treatment group (i.e., that

the groups were adequately mixed). Results of this analysis are shown below. For

additional details of this procedure see Dawley et al. (1989, Appendix D).

H: There was homogeneity between recovery distributions of treatments in

1992.

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 $Overall Chi-square = 19.3174 26$ $P-value = 0.8228$, non-significant

Appendix B.--Continued.

The 13 tests independently examined the same hypothesis, therefore their results can be combined to obtain an overall test (Fisher 1944). The overall result is:

Blocks 0.2725 12 0.0227
Treatments 0.1209 3 0.0403 Treatments 0.1209 3 0.0403 36.77 0.0000 Error 0.0394 36 0.0011 Total 0.4327 51

Conclusion: No evidence to suggest there is non-homogeneity between treatment recovery distributions.

B. Analysis of treatment effects using a randomized block ANOVA design where each

day was considered a block (Sokal and Rohlf 1981).

Full data set using all release blocks (see Table 4).

H: Mean recovery percents for each treatment are equal.

ANOVA Table

Appendix B.--Continued.

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The H_o is rejected at $\alpha = 0.05$.

 $FPLSD = T_{(a = 0.05)(d0)} \sqrt{2(MSE)/r} = 0.0263$ where:

T = Student's Tabular T value MSE = mean square error term in the ANOVA table $r =$ number of blocks

The treatment means are ranked using Fisher's Protected Least Significance Difference

(FPLSD) test (Petersen 1985).

Bypass/tailrace 1st Powerhouse Turbine/tailrace 2nd Powerhouse Turbine/tailrace 1st Powerhouse Downstream

 0.3061 0.3124 0.3464 0.4272

Any pair of treatment means differing by more than the FPLSD were judged to be

significant. The following shows these differences in rank order, where underlined means

are not significantly different at $\alpha = 0.05$

Treatment mean (%)