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20 DNA metabarcoding - need for robust experimental designs to draw sound ecological 21 conclusions

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111 DNA metabarcoding, especially when coupled with high-throughput DNA sequencing, is 112 currently revolutionizing our capacity to assess biodiversity across a full range of taxa and habitats, from soil microbes (e.g. Thompson et al., 2017) to large marine fish (e.g. Thomsen et al., 2016), and 113 114 from contemporary to tens of thousands year-old biological communities (e.g. Willerslev et al., 2003). 115 The breadth of potential applications is immense and spans surveys on the diversity or diet of species 116 native to specific ecosystems to bioindication (Pawlowski et al., 2018). The approach is also 117 especially cost-effective and easy to implement, which makes DNA metabarcoding one of the tools of choice of the 21<sup>st</sup> century for fundamental research and the future of large-scale biodiversity 118 119 monitoring programs (reviewed in Bohan et al., 2017; Creer et al., 2016; Taberlet, Bonin, Zinger, & 120 Coissac, 2018; Thomsen & Willerslev, 2015). However, as is often the case with any emerging 121 technology, we feel that the rise of DNA metabarcoding is occurring at a pace and in a manner that 122 often loses sight of the challenges in producing high-quality and reproducible data (Baker, 2016). 123 DNA metabarcoding is by essence a multidisciplinary approach building upon many complementary 124 expertises, including field and theoretical knowledge, taxonomic expertise, molecular biology, 125 bioinformatics, and computational statistics. Combining all these within single studies is necessary, 126 not so much for producing and analyzing the data per se, but rather for minimizing and controlling the 127 possible biases that can be introduced at any step of the experimental workflow - i.e. from the 128 sampling to data analysis - and that can lead to spurious ecological conclusions (reviewed in Bálint et 129 al., 2016; Nilsson et al., 2019; Dickie et al., 2018; Taberlet et al., 2018).

130 Whether the starting material consists of DNA from bulk samples (community DNA) and/or 131 from environmental DNA (eDNA), all DNA metabarcoding studies rely on a deceptively simple 132 succession of core experimental steps: (i) sampling and preservation of the starting material, (ii), 133 DNA extraction, (iii) PCR amplification of a taxonomically-informative genomic region, (iv) high-134 throughput DNA sequencing of the amplicons, and (v) sequence analysis using bioinformatic 135 pipelines. Despite this apparent simplicity, each step can potentially introduce its own sources of 136 artifacts and biases (Figure 1). For example, the sampling design might not be effective for capturing 137 the full taxonomic diversity or the ecological processes under study, an undesired bias for studies 138 based on species detection. The availability of DNA in the samples is governed by its production rate, 139 transport and persistence, processes which are all largely dependent on the targeted organisms, their 140 biomass, and the ecosystem considered. A correct assessment of an ecological phenomenon based on 141 DNA metabarcoding require not only implementation of standardized standardized, randomized and 142 repeatable sampling designs and procedures (Dickie et al., 2018), but also consideration of DNA 143 dynamics in the underlying matrix (i.e. in gut, feces, water or soil matrices from tropical or boreal 144 organisms/ecosystems; Barnes & Turner, 2016). Likewise, the community under study can be 145 enriched - on purpose or not - with specific taxa depending on how the sample is collected (e.g. filter 146 size for water samples, removal of roots or not for soils), how it is transported/preserved, and how 147 DNA is extracted (differential extraction efficiencies). PCR amplification is also well known to be an 148 important source of biases, that are now fully revealed with high-throughput DNA sequencing 149 techniques. The preferential amplification of certain taxa over other ones due to inappropriate primers 150 provides one such example of potential bias (Clarke, Soubrier, Weyrich, & Cooper, 2014; Deagle, 151 Jarman, Coissac, Pompanon, & Taberlet, 2014). Primer biases can both skew abundance profiles and 152 lead to false negatives. PCR amplification can produce false negatives too through the presence of e.g. 153 PCR inhibitors, but also many false positives through the introduction of replication errors by the 154 DNA polymerase or the formation of chimeric fragments (reviewed in Taberlet et al., 2018). False 155 positives can also be introduced at any step of the experimental workflow through the presence of 156 reagent contaminants (Salter et al., 2014), or through samples, extractions or PCR crosscontaminations. An even more insidious source of false positives pertains to the occurrence of "tag 157 158 jumps", sometimes referred to as "mistagging", "tag-switching", or "cross-talks" (Carlsen et al., 2012; 159 Edgar, 2018; Esling, Lejzerowicz, & Pawlowski, 2015; Schnell, Bohmann, & Gilbert, 2015). PCR 160 amplicons are indeed often tagged with unique short nucleotide sequences added on the 5'-end of the 161 primers (i.e. "tags"), which allow pooling all PCRs within a single sequencing run and reducing 162 sequencing costs. Each sequence obtained resulting in apparent cross-contaminations can then be 163 bioinformatically assigned back to its sample of origin on the basis of its tag (Schnell et al., 2015). 164 However, the procedures underlying the preparation of DNA libraries and/or the sequencing can 165 introduce these "tag jumps", when the tag assigned to one particular sample is in fact recombined to 166 the sequences belonging to another sample (Taberlet et al., 2018). This introduces additional, non167 negligible levels of sample cross-contaminations, which primarily involve the most abundant taxa and 168 can have a disproportionate impact on samples with low DNA concentrations (Esling et al., 2015; 169 Murray, Coghlan, & Bunce, 2015; Schnell et al., 2015). Similarly, the Illumina index located on the 170 P5 sequencing adaptor can be subjected to "index jumps", resulting in apparent cross-contaminations 171 (Taberlet et al., 2018). This bias happens when several individual Illumina sequencing libraries are 172 pooled and loaded on the same sequencing lane (Kircher, Sawyer, & Meyer, 2012) Finally, high-173 throughput DNA sequencing instruments have their own error rates (Schirmer et al., 2015). The above list of problems is clearly not exhaustive, and the interested reader will find more complete reviews 174 175 elsewhere (e.g. Bálint et al., 2016; Nilsson et al., 2019; Taberlet et al., 2018). Still, it illustrates that 176 any potential bias must be considered carefully when designing an experimental protocol and when 177 interpreting the results. This is crucial to limit their impact on downstream analyses, and to ensure that 178 the conclusion drawn from such data are authentic.

179 There is now an increasingly diverse range of field, laboratory (e.g. Caporaso et al., 2011; 180 Taberlet et al., 2018; Valentini et al., 2009) and bioinformatics (e.g. Boyer et al., 2016; Caporaso et 181 al., 2010; Dumbrell, Ferguson, & Clark, 2016) procedures aiming at reducing the amount of both false 182 negatives (i.e. due to partial sampling, extraction, amplification or sequencing bias) and false positives 183 (i.e. due to contaminations, "tag/index jumps", or PCR and sequencing errors) in DNA metabarcoding 184 experiments. However, using these protocols does not necessarily guarantee that the problem of false 185 positives or negatives is completely under control. These protocols must continuously be 186 reconsidered, especially alongside the emergence of novel DNA sequencing technologies that provide 187 new opportunities, but also new challenges. Additionally, each individual study and each genomic 188 marker comes with its own specificities, and this often requires customization of the above protocols. 189 The sequence clustering threshold to be used to form molecular taxonomic units relevant to the 190 question addressed (e.g. removing intraspecific marker variability when the species level is desired) 191 provides such an example, and will critically depend on both the marker specificities and 192 PCR/sequencing error rates. Bioinformatics tools can further fail to exclude molecular artifacts when 193 the filtering thresholds are relaxed, which inflates sample diversity estimates. Likewise, they can also 194 generate false negatives, for example when a genuine metabarcode is falsely flagged as an error or 195 chimera, or when it is assigned to an incorrect taxon due to incomplete or inappropriate reference 196 databases (Alsos et al., 2018; Coissac, Riaz, & Puillandre, 2012). This can be especially problematic 197 when the question investigated strongly relies on species detection. It is therefore crucial to include 198 several types of experimental controls so as to facilitate the exclusion of spurious signal and support 199 the reliability of the biological conclusions (Figure 1). Amongst these controls, conducting pilot 200 experiments is particularly helpful to assess how appropriate the sampling design is (Dickie et al., 201 2018). We also recommend that both biological replicates (i.e. multiple independent samples) and 202 technical replicates (i.e. multiple extractions/PCR of the same sample and/or extract) are included in

203 the experimental workflow to disentangle the effect of both the biological and technical variances 204 (Ficetola et al., 2015). These replications are necessary because both sampling and PCR can introduce 205 biases in a stochastic manner, especially when the concentration of the target DNA is low. It is also 206 essential to analyze a sufficient number of negative controls at the field sampling, DNA extraction, 207 PCR, and sequencing steps, as well as positive controls consisting of mock communities, known DNA 208 samples, or even synthetic sequences reflecting the attributes of the targeted products (Figure 1). All 209 these controls must be sequenced along the biological samples, as they facilitate the detection of 210 sporadic contaminations and tag or index jumps while helping adjusting filtering and clustering 211 thresholds. Ultimately, they will be a token of the reliability of the whole data curation process (De 212 Barba et al., 2014). We also encourage careful consideration of the bioinformatics workflow itself, 213 since the filtering steps necessary to curate the data will critically depend on the experimental design 214 and the ecological question under study. Typically, sequences of low abundance in a given sample 215 may be genuine or artifacts deriving from PCR/sequencing errors or tag/index jumps. The retained 216 filtering threshold for taxon presence is thus dependent on the underlying rates of artifacts, as well as 217 on the sequencing depth. As the different experimental controls provide direct measurements of these 218 artifacts, they will therefore allow better tuning of the filtering thresholds. All of these technical 219 considerations should be precisely reported within publications together with relevant illustrations and 220 statistics characterizing the workflow, as they are necessary to assess the relevance and quality of the 221 data underpinning specific conclusions. A last, a most obvious example of control consists in 222 assessing the plausibility of the taxonomic composition based on *a priori* knowledge of the system or 223 taxa studied. Such knowledge can be derived from data obtained with complementary sensing 224 approaches such as visual observations. In this case, building exhaustive local reference databases of 225 the genomic marker used from local specimens will secure the taxonomic assignment step (e.g. Alsos 226 et al., 2018). When local information is unavailable, typically when studying microorganisms, it 227 remains possible to assess whether the community is composed of clades that are expected to occur in 228 the system surveyed or not, as e.g. soils, sediments, and gut environments harbour highly different 229 bacterial phyla (e.g. Thompson et al., 2017).

230 As users, readers, referees or editors, we realize that the above-mentioned issues remain too 231 often overlooked. This problematic stance can lead to unsubstantiated claims and undermine scientific 232 advances if not resolved. Inappropriate practices such as estimating species richness from fingerprint 233 profiles (Bent, Pierson, & Forney, 2007), the absence of biological replicates (Prosser, 2010), or that 234 of contaminant controls (Perez-Muñoz, Arrieta, Ramer-Tait, & Walter, 2017) have been repeatedly 235 criticized in the field of microbial ecology, and in the latter case, they contribute to the rising debate 236 about the existence or not of a womb microbiota. Ancient DNA research has also developed rigorous 237 standards to tackle issues related to contamination, sequencing errors, and data reproducibility (Poinar 238 and Cooper (2000). We believe that the community of DNA metabarcoding users has now come of 239 age and learnt from its past errors. At a time when more and more exhaustive guides of best practices 240 on the subject are emerging (Knight et al., 2018; Pollock, Glendinning, Wisedchanwet, & Watson, 241 2018; Taberlet et al., 2018), and where DNA sequencing costs are rapidly decreasing, we should be 242 always mindful of the adage "better safe than sorry". This note does not mean to imply that the 243 systematic use of the highest technical and analytical standards is reasonable nor the universal remedy 244 for all the challenges associated with DNA metabarcoding. Rather, we strongly encourage researchers 245 and end-users to adopt reflective decision-making when designing their experiment and to critically 246 appraise their results, with the ultimate aim to prove the robustness and reproducibility of their conclusions. 247

- 248 References
- Alsos, I. G., Lammers, Y., Yoccoz, N. G., Jørgensen, T., Sjögren, P., Gielly, L., & Edwards, M. E.
  (2018). Plant DNA metabarcoding of lake sediments: How does it represent the contemporary
  vegetation. *PLoS One*, *13*(4), e0195403.
- Baker, M. (2016). 1,500 scientists lift the lid on reproducibility. *Nature News*, 533(7604), 452.
- Bálint, M., Bahram, M., Eren, A. M., Faust, K., Fuhrman, J. A., Lindahl, B., ... Tedersoo, L. (2016).
  Millions of reads, thousands of taxa: microbial community structure and associations analyzed
  via marker genes. *FEMS Microbiology Reviews*, 40(5), 686–700.
- Barnes, M. A., & Turner, C. R. (2016). The ecology of environmental DNA and implications for
   conservation genetics. *Conservation Genetics*, *17*(1), 1–17.
- 258 Bent, S. J., Pierson, J. D., & Forney, L. J. (2007). Measuring species richness based on microbial
- community fingerprints: the emperor has no clothes. *Applied and Environmental Microbiology*,
  73(7), 2399.
- Bohan, D. A., Vacher, C., Tamaddoni-Nezhad, A., Raybould, A., Dumbrell, A. J., & Woodward, G.
  (2017). Next-generation global biomonitoring: large-scale, automated reconstruction of
  ecological networks. *Trends in Ecology & Evolution*, *32*(7), 477–487.
- Boyer, F., Mercier, C., Bonin, A., Le Bras, Y., Taberlet, P., & Coissac, E. (2016). obitools: a unixinspired software package for DNA metabarcoding. *Molecular Ecology Resources*, 16(1), 176–
  182.
- 267 Caporaso, J. G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F. D., Costello, E. K., ...
  268 Knight, R. (2010). QIIME allows analysis of high-throughput community sequencing data.
  269 *Nature Methods*, 7(5), 335–336.
- 270 Caporaso, J. G., Lauber, C. L., Walters, W. A., Berg-Lyons, D., Lozupone, C. A., Turnbaugh, P. J., ...
- 271 Knight, R. (2011). Global patterns of 16S rRNA diversity at a depth of millions of sequences per
- sample. Proceedings of the National Academy of Sciences of the United States of America, 108
  Suppl 1, 4516–4522.
- 274 Carlsen, T., Aas, A. B., Lindner, D., Vrålstad, T., Schumacher, T., & Kauserud, H. (2012). Don't

- make a mista(g)ke: is tag switching an overlooked source of error in amplicon pyrosequencing
  studies? *Fungal Ecology*, 5(6), 747–749.
- Clarke, L. J., Soubrier, J., Weyrich, L. S., & Cooper, A. (2014). Environmental metabarcodes for
  insects: *in silico* PCR reveals potential for taxonomic bias. *Molecular Ecology Resources*, 14(6),
- **279** 1160–1170.
- 280 Coissac, E., Riaz, T., & Puillandre, N. (2012). Bioinformatic challenges for DNA metabarcoding of
  281 plants and animals. *Molecular Ecology*, *21*, 1834–1847.
- Creer, S., Deiner, K., Frey, S., Porazinska, D., Taberlet, P., Thomas, W. K., ... Bik, H. M. (2016). The
  ecologist's field guide to sequence-based identification of biodiversity. *Methods in Ecology and Evolution*, 7(9), 1008–1018.
- Deagle, B. E., Jarman, S. N., Coissac, E., Pompanon, F., & Taberlet, P. (2014). DNA metabarcoding
  and the cytochrome c oxidase subunit I marker: not a perfect match. *Biology Letters*, *10*(9),
  20140562.
- De Barba, M., Miquel, C., Boyer, F., Rioux, D., Coissac, E., & Taberlet, P. (2014). DNA
   metabarcoding multiplexing for omnivorous diet analysis and validation of data accuracy.
   *Molecular Ecology Resources*, 14(2), 306–323.
- Dickie, I. A., Boyer, S., Buckley, H. L., Duncan, R. P., Gardner, P. P., Hogg, I. D., ... Weaver, L.
  (2018). Towards robust and repeatable sampling methods in eDNA-based studies. *Molecular Ecology Resources*, 18(5), 940–952.
- Dumbrell, A. J., Ferguson, R. M. W., & Clark, D. R. (2016). Microbial community analysis by singleamplicon high-throughput next generation sequencing: data analysis from raw output to
  ecology. In T. J. McGenity, K. N. Timmis, & B. Nogales (Eds.), *Hydrocarbon and lipid*
- 297 *microbiology protocols*. Humana Press.
- Edgar, R. (2018). UNCROSS2: identification of cross-talk in 16S rRNA OTU tables. bioRxiv.
  doi:10.1101/400762
- Esling, P., Lejzerowicz, F., & Pawlowski, J. (2015). Accurate multiplexing and filtering for highthroughput amplicon-sequencing. *Nucleic Acids Research*, 43(5), 2513–2524.
- 302 Ficetola, G. F., Pansu, J., Bonin, A., Coissac, E., Giguet-Covex, C., De Barba, M., ... Taberlet, P.
- 303 (2015). Replication levels, false presences and the estimation of the presence/absence from
  304 eDNA metabarcoding data. *Molecular Ecology Resources*, 15(3), 543–556.
- Kircher, M., Sawyer, S., & Meyer, M. (2012). Double indexing overcomes inaccuracies in multiplex
   sequencing on the Illumina platform. *Nucleic Acids Research*, 40. doi:10.1093/nar/gkr771
- 307 Knight, R., Vrbanac, A., Taylor, B. C., Aksenov, A., Callewaert, C., Debelius, J., ... Dorrestein, P. C.
- 308 (2018). Best practices for analysing microbiomes. *Nature Reviews. Microbiology*, *16*(7), 410–
  309 422.
- 310 Murray, D. C., Coghlan, M. L., & Bunce, M. (2015). From benchtop to desktop: important
- 311 considerations when designing amplicon sequencing workflows. *PLoS One*, *10*(4), e0124671.

- 312 Nilsson, R. H., Anslan, S., Bahram, M., Wurzbacher, C., Baldrian, P., & Tedersoo, L. (2019).
- Mycobiome diversity: high-throughput sequencing and identification of fungi. *Nature Reviews*.
   *Microbiology*, *17*(2), 95–109.
- 315 Pawlowski, J., Kelly-Quinn, M., Altermatt, F., Apothéloz-Perret-Gentil, L., Beja, P., Boggero, A., ...
- 316 Kahlert, M. (2018). The future of biotic indices in the ecogenomic era: Integrating (e)DNA
- 317 metabarcoding in biological assessment of aquatic ecosystems. *The Science of the Total*

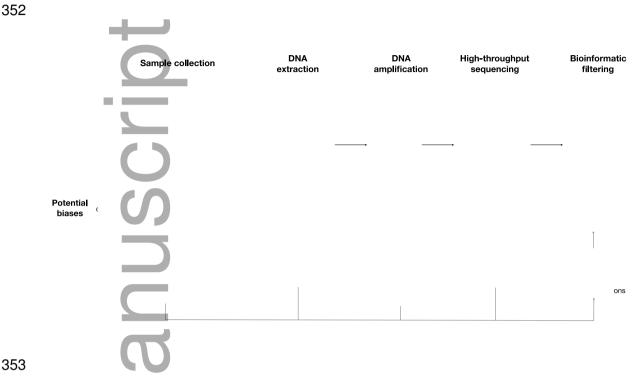
**318** *Environment*, *637-638*, 1295–1310.

- Perez-Muñoz, M. E., Arrieta, M.-C., Ramer-Tait, A. E., & Walter, J. (2017). A critical assessment of
  the ``sterile womb'' and ``in utero colonization" hypotheses: implications for research on the
  pioneer infant microbiome. *Microbiome*, 5(1), 48.
- 322 Poinar, H. N., & Cooper, A. (2000). Ancient DNA: do it right or not at all. Science, 5482, 1139.
- Pollock, J., Glendinning, L., Wisedchanwet, T., & Watson, M. (2018). The madness of microbiome:
  attempting to find consensus "best practice" for 16S microbiome studies. *Applied and Environmental Microbiology*, 84(7), e02627–17.
- 326 Prosser, J. I. (2010). Replicate or lie. *Environmental Microbiology*, *12*(7), 1806–1810.
- 327 Salter, S. J., Cox, M. J., Turek, E. M., Calus, S. T., Cookson, W. O., Moffatt, M. F., ... Walker, A. W.
  328 (2014). Reagent and laboratory contamination can critically impact sequence-based microbiome
  329 analyses. *BMC Biology*, *12*, 87.
- 330 Schirmer, M., Ijaz, U. Z., D'Amore, R., Hall, N., Sloan, W. T., & Quince, C. (2015). Insight into
- biases and sequencing errors for amplicon sequencing with the Illumina MiSeq platform. *Nucleic Acids Research*, 43(6), e37.
- 333 Schnell, I. B., Bohmann, K., & Gilbert, M. T. P. (2015). Tag jumps illuminated reducing sequence334 to-sample misidentifications in metabarcoding studies. *Molecular Ecology Resources*, 15(6),
  335 1289–1303.
- Taberlet, P., Bonin, A., Zinger, L., & Coissac, E. (2018). *Environmental DNA for biodiversity research and monitoring*. Oxford University Press.
- Thompson, L. R., Sanders, J. G., McDonald, D., Amir, A., Ladau, J., Locey, K. J., ... Earth
  Microbiome Project Consortium. (2017). A communal catalogue reveals Earth's multiscale
  microbial diversity. *Nature*, 551(7681), 457–463.
- 341 Thomsen, P. F., Møller, P. R., Sigsgaard, E. E., Knudsen, S. W., Jørgensen, O. A., & Willerslev, E.
- 342 (2016). Environmental DNA from seawater samples correlate with trawl catches of subarctic,
  343 deepwater fishes. *PLoS One*, *11*(11), e0165252.
- Thomsen, P. F., & Willerslev, E. (2015). Environmental DNA as an emerging tool in conservation for
  monitoring past and present biodiversity. *Biological Conservation*, *183*(C), 4–18.
- 346 Valentini, A., Miquel, C., Nawaz, M. A., Bellemain, E., Coissac, E., Pompanon, F., ... Taberlet, P.
- 347 (2009). New perspectives in diet analysis based on DNA barcoding and parallel pyrosequencing:
- 348 the *trn*L approach. *Molecular Ecology Resources*, 9, 51–60.

349 Willerslev, E., Hansen, A. J., Binladen, J., Brand, T. B., Gilbert, M. T. P., Shapiro, B., ... Cooper, A.

350 (2003). Diverse plant and animal genetic records from Holocene and Pleistocene sediments.

351 *Science*, *300*, 791–795.



354 Figure 1 Summarized workflow of DNA metabarcoding and biases in the data production 355 process, with the potential associated controls to assess data quality. Expectations on the local 356 community, either from a priori knowledge on the site or organisms targeted, or obtained through e.g. 357 vizual census, specimen collection, or building of a local reference database, constitute a first 358 assessment of the DNA metabarcoding experiment success. Pilot experiments are essential for 359 optimizing the whole experimental design, from the sampling strategy (e.g. number of biological 360 replicates) to the entire technical approach. Field, extraction, PCR, and tagging-system negative and 361 positive controls should be sequenced along with biological samples. They all aim at identifying (i) 362 potential contaminants that could be introduced at any experimental step, and (ii) potential 363 experimental artifacts due to the DNA extraction, PCR, and sequencing steps. Field negative controls 364 consists of extracting DNA from storage/extraction buffers brought to the field or used to clean 365 sampling instruments. Tagging-system negative controls can only be implemented when amplicons 366 are identified by a unique combination of tags attached to the 5' end of each amplification primer, and 367 where one or several tag combinations remain unused in the experimental design. In such conditions, 368 tagging-system controls can be performed at the bioinformatics analysis step, by monitoring the 369 number of sequences harboring unexpected tag combinations. This number is actually a direct 370 measurement of the tag-jump rate. "Index jumps" are more difficult to evaluate, and can be controlled 371 either by indexing both library adapters (P5 and P7) or when the libraries sequenced together have

372 identifiable sequences that could indicate their origin. The positive controls (constructed using either 373 synthetic DNA with the primer target sequences on both sides, DNA extracted from a mock 374 community, or known environmental samples), as well as prior expectations on the taxa that should 375 occur in the system can be used to evaluate the effectiveness of the data production process, the 376 impact of contaminants on the retrieved ecological signal and the adequacy of bioinformatics filtering

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