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**Title:** Genomic divergence across ecological gradients in the Central African rainforest songbird (*Andropadus virens*)

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44 **Running title:** Genomic divergence across ecological gradients

45

46 **Abstract**

47 The little greenbul, a common rainforest passerine from sub-Saharan Africa, has been the  
48 subject of long-term evolutionary studies to understand the mechanisms leading to  
49 rainforest speciation. Previous research found morphological and behavioral divergence  
50 across rainforest-savanna transition zones (ecotones), and a pattern of divergence with  
51 gene flow suggesting divergent natural selection has contributed to adaptive divergence  
52 and ecotones could be important areas for rainforests speciation. Recent advances in  
53 genomics and environmental modeling make it possible to examine patterns of genetic  
54 divergence in a more comprehensive fashion. To assess the extent to which natural  
55 selection may drive patterns of differentiation, here we investigate patterns of genomic  
56 differentiation among populations across environmental gradients and regions. We find

57 compelling evidence that individuals form discrete genetic clusters corresponding to  
58 distinctive environmental characteristics and habitat types. Pairwise  $F_{ST}$  between  
59 populations in different habitats is significantly higher than within habitats, and this  
60 differentiation is greater than what is expected from geographic distance alone. Moreover,  
61 we identified 140 SNPs that showed extreme differentiation among populations through a  
62 genome-wide selection scan. These outliers were significantly enriched in exonic and  
63 coding regions, suggesting their functional importance. Environmental association analysis  
64 of SNP variation indicates that several environmental variables, including temperature and  
65 elevation, play important roles in driving the pattern of genomic diversification. Results  
66 lend important new genomic evidence for environmental gradients being important in  
67 population differentiation.

68

## 69 **Introduction**

70 Rainforests are heralded for their exceptionally high biological diversity, yet the  
71 evolutionary mechanisms for the generation and maintenance of this diversity have been  
72 debated for decades (Haffer 1969; Mayr & O'Hara 1986; Martin 1991; Smith *et al.* 1997;  
73 Schneider *et al.* 1999; Moritz *et al.* 2000; Ogden & Thorpe 2002; Price 2008; Schluter  
74 2009; Hoorn *et al.* 2010; Ribas *et al.* 2011; Smith *et al.* 2014; Beheregaray *et al.* 2015).  
75 Models of rainforest speciation abound. Some emphasize the importance of neutral  
76 processes, for example genetic drift in allopatric populations isolated by historical refugia  
77 (Haffer 1969), some favor processes such as landscape change (Hoorn *et al.* 2010; Ribas *et al.*  
78 *et al.* 2011) or dispersal (Smith *et al.* 2014), while others point toward a dominant role of  
79 divergent natural selection across ecological gradients and ecotones (Smith *et al.* 1997;  
80 Schneider *et al.* 1999; Ogden & Thorpe 2002; Smith *et al.* 2005; Schluter 2009; Smith *et al.*  
81 *et al.* 2011; Beheregaray *et al.* 2015). Each process is expected to shape the genomes of  
82 natural populations in different ways, leaving a signal that provides insights into the  
83 evolutionary mechanisms that may have led to divergence. Such information is of  
84 importance not only to evolutionary geneticists interested in understanding the processes  
85 involved in speciation, but also to conservation decision makers, who are interested in  
86 preserving biodiversity and prioritizing new regions for protection in the face of rapid  
87 anthropogenic change and climate change.

88            In this study, we explore the roles that population-level processes play in shaping  
89 biodiversity in Central Africa by examining the genomic diversity in a common songbird,  
90 the little greenbul (*Andropadus virens*). The little greenbul provides a particularly useful  
91 taxon for this inquiry because it has a broad geographic distribution across sub-Saharan  
92 Africa where it occurs in ecologically diverse habitats, and has been the subject of long-  
93 term studies of intra-specific diversity and speciation. In the case of *A. virens*, as well as  
94 some other rainforest taxa, the rainforest-savanna transition zones (ecotones) have been  
95 shown to drive phenotypic divergence and likely speciation (Smith *et al.* 1997, 2005;  
96 Kirschel *et al.* 2009; Freedman *et al.* 2010; Mitchell *et al.* 2015; Nadis 2016). Compared  
97 to the central rainforest, ecotone habitats differ dramatically in numerous ways. For  
98 example, ecotones have less tree cover, lower levels of precipitation, and greater intra-  
99 annual variation in environmental variables. These ecological differences may lead to  
100 distinctive food resources, pathogens, acoustic environments and predation levels  
101 (Slabbekoorn & Smith 2002; Smith *et al.* 2005, 2013). Consequently, these differences in  
102 both abiotic and biotic environments are hypothesized to result in divergent selection in  
103 ecotone and rainforest populations, leading to locally adapted populations (Smith *et al.*  
104 1997, 2005; Kirschel *et al.* 2009; Freedman *et al.* 2010; Sehgal *et al.* 2011; Kirschel *et al.*  
105 2011). This hypothesis is supported by the fact that parapatric *A. virens* populations across  
106 rainforest-ecotone gradients have undergone significant divergence in morphological (i.e.  
107 body mass, wing, tail, tarsus and beak length) and vocal characteristics despite significant  
108 levels of gene flow (Smith *et al.* 1997; Slabbekoorn & Smith 2002; Smith *et al.* 2005;  
109 Kirschel *et al.* 2011; Smith *et al.* 2013). This pattern of divergence with gene flow and the  
110 role of ecotones in driving adaptive divergence is further supported by the fact that  
111 allopatric rainforest populations of *A. virens* that were geographically isolated in West and  
112 Central Africa for two million years, had much lower levels of phenotypic divergence in  
113 these traits compared to the level of divergence observed across a narrow (often 100km)  
114 rainforest-ecotone gradient (Smith *et al.* 2005). Together, results for *A. virens* and those  
115 from other species suggest that strong divergent natural selection across the rainforest-  
116 savanna ecotone transition contributes to adaptive phenotypic divergence despite high  
117 levels of ongoing gene-flow (Smith *et al.* 1997, 2001, 2005). Evidence for divergence with  
118 gene flow in *A. virens* is also consistent with models of ecological speciation where

119 natural selection caused by shifts in ecology or invasions of new habitats can result in  
120 divergence in fitness related traits and might play a prominent role in speciation (Orr &  
121 Smith 1998; Schneider *et al.* 1999; Ogden & Thorpe 2002; Rundle & Nosil 2005; Schluter  
122 2009; Beheregaray *et al.* 2015). Opportunities for this kind of divergence are possible  
123 across the little greenbul range, as they occur across a wide diversity of habitats, including  
124 mountains and islands, which are also known as hotspots of diversification and speciation  
125 (Darwin 1859; Myers *et al.* 2000; Orme *et al.* 2005). Previous research has found that,  
126 compared to *A. virens* populations in mainland rainforests, mountain populations and  
127 island populations also show significant divergence in morphological traits typically  
128 related to fitness in birds, including body mass, wing length, tail length, tarsus length and  
129 bill size (Smith *et al.* 2005). Moreover, both habitats have considerable gene flow with  
130 mainland rainforest populations in Lower Guinea (Smith *et al.* 2005), suggesting natural  
131 selection may play an important role in divergence of mountain and island populations in  
132 *A. virens*.

133 To date, the paucity of high-resolution genomic data for rainforest species such as  
134 *A. virens* hinders a full exploration of the evolutionary mechanisms that may be important  
135 for diversification. Previous genetic studies on *A. virens* population structure utilized a  
136 handful of mtDNA markers (Smith *et al.* 2001) and microsatellite loci (Smith *et al.* 2005).  
137 These limited resources were unable to differentiate ecotone and forest populations at  
138 genetic level, therefore debates still remain whether the observed phenotypic divergence  
139 might be the results of plasticity in traits in response to varying environmental conditions,  
140 or strictly genomic divergence between populations in ecotone and rainforest. Recent  
141 development of next generation sequencing techniques (NGS), especially restriction site  
142 associated DNA (RAD) sequencing, enables one to *de novo* assemble hundreds of  
143 thousands of RAD loci across the genome in hundreds of samples without a reference  
144 genome. This cost effective method to produce genomic-wide population data provides  
145 unprecedented opportunities to assess the patterns of diversity with much greater  
146 resolution, to find potential population structure and to identify candidate loci under local  
147 selection in non-model species such as *A. virens*.

148 Here we take a population genomic approach leveraging single nucleotide  
149 polymorphism (SNP) data generated from RAD sequencing to survey the genome-wide

150 diversity of *A. virens* across multiple ecological habitats in Cameroon and Equatorial  
151 Guinea, including rainforests, ectones, mountains, as well as island. Our specific objectives  
152 for this approach were to: 1) estimate overall levels of genetic diversity in *A. virens*; 2)  
153 determine population structure and differentiation across habitats; 3) identify candidate loci  
154 that are potential targets of selection; 4) understand the biological functions of these  
155 candidate loci using transcriptome data; and 5) characterize genetic turnover across  
156 environmental gradients.

157

## 158 **Materials and methods**

### 159 *Sampling, DNA extraction and RADseq library construction*

160 For RAD sequencing, blood samples from adult *A. virens* were collected in Central  
161 Africa and stored in Queens Lysis Buffer (Smith *et al.* 1997, 2005). Overall, 217 samples  
162 were collected from 15 geographically distant sampling sites (Figure 1A), representing 15  
163 populations. Sampling sites were classified into one of four habitat types by researchers in  
164 the field and had previously been confirmed using remote sensing data (Slabbekoorn &  
165 Smith 2002; Smith *et al.* 2005, 2013). Low quality samples were removed by filtering,  
166 resulting in a total of 182 samples included in the final analysis (see *RADseq data*  
167 *bioinformatics processing* below). This included seven rainforest populations (83 samples),  
168 five ecotone populations (59 samples), two mountain populations (18 samples) and one  
169 population from the island of Bioko (12 samples). Each population was represented by 5-  
170 22 samples, with a mean representation of 12 samples (Table S1) (Willing *et al.* 2012;  
171 Nazareno *et al.* 2017).

172 RAD library preparation followed the methods for traditional RAD as described in  
173 Ali *et al.* (2016) that were slightly modified from the original RAD protocol as described in  
174 Baird *et al.* (2008). In short, genomic DNA (50 ng) for each sample was digested with 2.4  
175 units of SbfI-HF (New England Biolabs, NEB, R3642L) at 37 °C for 1 hr in a 12 µl  
176 reaction volume buffered with 1X NEBuffer 4 (NEB, B7004S). Samples were heated to 65  
177 °C for 20 minutes, and then 2 µl indexed SbfI P1 RAD adapter (10 nM) was added to each  
178 sample. Ligation of inline barcoded P1 adaptors was performed by combining 2 µl of the  
179 ligation mix (1.28 µl water, 0.4 µl NEBuffer 4, 0.16 µl rATP (100 mM, Fermentas R0441)  
180 with 0.16 µl T4 DNA Ligase (NEB, M0202M) and heating at 20 °C for 1 hr followed by

181 incubation at 65 °C for 20 min. Following the ligation, half the per sample volume or 5 µl  
182 of each of the 96 samples were pooled into a single tube and cleaned using 1X Agencourt  
183 AMPure XP beads (Beckman Coulter, A63881); the remainder of the sample was stored  
184 for use in an additional library preparation if needed. The pooled DNA was then re-  
185 suspended in 100 µl low TE and sheared to an average fragment size of 500 base pairs  
186 using a Bioruptor NGS sonicator (Diagenode). Sheared DNA was then concentrated to  
187 55.5 µl using Ampure XP beads, and used as the template in the NEBNext Ultra DNA  
188 Library Prep Kit for Illumina (NEB E7370L; version 1.2). The NEBNext protocol for  
189 library prep was followed apart from the fact that we used custom P2 adaptors which were  
190 created by annealing an NEBNext Multiplex Oligo for Illumina (NEB, E7335L) to the  
191 oligo GATCGGAAGAGCACACGTCTGAACTCCAGTCACIIIIIIATCAGAACA\*A (the  
192 \* represents a Phosphorothioated DNA base). In addition, instead of the USER enzyme  
193 step, we used a universal P1 RAD primer  
194 (AATGATACGGCGACCACCGAGATCTACACTCTTTCCCTACACGAC\*G) and a  
195 universal P2 RAD primer (CAAGCAGAAGACGGCATACG\*A) during final  
196 amplification. The final RAD library was cleaned using AMPure XP beads and sequenced  
197 at the UC Berkeley QB3 Vincent J Coates Genome Sequencing Laboratory (GSL) on an  
198 Illumina HiSeq2000 (Illumina, San Diego, CA) using single-end 100 bp sequence reads.

#### 200 *RADseq data bioinformatics processing*

201 We used FastQC (<http://www.bioinformatics.babraham.ac.uk/projects/fastqc/>) to  
202 assess overall data quality of each RADseq sequencing run. To remove the lowest quality  
203 bases, we trimmed all raw sequencing reads (100bp) by 14 bp at the 3' end using *seqtk*  
204 (<https://github.com/lh3/seqtk>). We processed RADseq reads using the *Stacks* pipeline  
205 version 1.32 (Catchen *et al.* 2011, 2013) in the following manner. First, we demultiplexed  
206 the trimmed data by P1 barcodes and removed low-quality reads and those containing  
207 adapter sequences using *process\_radtags*. After demultiplexing, reads were 80 bp in length  
208 (without barcodes) and data from different runs were combined together. These reads were  
209 used to *de novo* assemble RAD loci using *denovo\_map.pl* (parameter settings: m=3 M=4  
210 n=4). The parameters for *de novo* assembly were determined empirically to limit the  
211 impact of over-splitting of loci following methods described in Ilut *et al.* (2014) and

212 Harvey *et al.* (2015). Specifically, we chose one sample that had a depth of coverage close  
213 to the median depth coverage of all samples, and ran the *de novo* assembly over a wide  
214 range of values of M (1-8; n=M) using *ustacks*. The percentage of homozygous and  
215 heterozygous loci plateaued at M=4, suggesting this value appropriately minimized over-  
216 splitting of alleles (Figure S1). Thus, we used M=4 for the final run on all samples. *Stacks*  
217 implements a multinomial-based likelihood model for SNP calling, by estimating the  
218 likelihood of two most frequently observed genotypes at each site and performing a  
219 standard likelihood ratio test using a chi-square distribution (Hohenlohe *et al.* 2010;  
220 Catchen *et al.* 2011). We used the default alpha (chi-square significance level) of 0.05.  
221 Paralogous loci that were stacked together were identified and removed by later quality  
222 control steps built into *Stacks* (e.g. max number of stacks per loci = 3; Ilut *et al.* 2014;  
223 Harvey *et al.* 2015). After the first round of assembly using *denovo\_map.pl*, we ran *stacks*'  
224 correction mode (*rxstacks-cstacks-sstacks*) using the bounded SNP model with a 0.05  
225 upper bound for the error rate (bound\_high = 0.05). The *rxstacks* program made  
226 corrections to genotype and haplotype calls based on population information, rebuilt the  
227 catalog loci, and filtered out loci with average log likelihood less than -8.0  
228 (<http://catchenlab.life.illinois.edu/stacks/>).

229 We then identified a set of high-quality RAD loci for downstream population  
230 genetic analysis using the following steps: 1) We only kept RAD loci if they were present  
231 in at least 80% of all samples, because loci that only assembled in small subset of samples  
232 had limited utility in downstream analyses as well as possibly higher error rates. 2) We  
233 filtered out RAD loci that had more than 40 SNPs along the 80 bp RAD loci sequence, as  
234 these likely represented sequencing errors or over-clustering of paralogous loci. In the final  
235 dataset, the RAD locus that has the most SNPs possessed 25 SNPs. Because the alignments  
236 look reasonable for the RAD loci that have higher number of SNPs, we did not apply any  
237 additional filters to avoid introducing additional biases. 3) We mapped the RAD loci  
238 sequences from *A. virens* to the closest reference genome available, the zebra finch genome  
239 (version 3.24), using BLAT and removed RAD loci that mapped to multiple positions in  
240 the zebra finch genome. 4) We used BLAT to compare RAD loci sequences against each  
241 other, and removed ones that had a match. This step further removes over-splitting RAD  
242 loci.

243           Following these filters, we obtained our final consensus set of RAD loci (Table S2).  
244       Samples that were missing more than 20% of the final consensus RAD loci were identified  
245       in a preliminary run and were removed from final analysis because they likely had low  
246       quality DNA, low quality libraries, or low sequencing coverage. 182 samples were  
247       included in the final dataset (see above). Genotypes were called and filtered using methods  
248       implemented in the *Stacks* pipeline (Hohenlohe *et al.* 2010). We exported genotypes for  
249       the final consensus RAD loci in VCF format using *stacks populations* program (only bi-  
250       allelic SNPs). Additional filters based on genotype calls were performed in *vcftools*  
251       (<https://vcftools.github.io/index.html>) or using custom scripts, which includes removing  
252       SNPs from the last 7 bp of the RAD loci as this part of the locus was enriched for  
253       erroneous SNPs due to the lower sequencing quality at the 3' end of reads, and filtering  
254       sites that have genotyping rate less than 80% of all samples.

255           We used the resulting full SNP dataset with SNPs from all frequencies to estimate  
256       genetic diversity statistics such as number of segregating sites (S), average pairwise  
257       differences ( $\pi$ ) and Waterson's  $\theta$  ( $\theta_w$ ) in each population (Table S1). Rare SNPs that had a  
258       minor allele frequency (MAF) less than 2% in the whole sample set were subsequently  
259       removed using *vcftools*, and the remaining SNPs were used for downstream analyses such  
260       as PCA, pairwise  $F_{ST}$  calculations, *Bayescan* outlier analysis, *gradientForest* analysis.

261

#### 262 *RNA extraction, RNAseq library preparation, and transcriptome de novo assembly*

263           *A. virens* lacks a reference genome. In order to help determine which of the RAD  
264       loci are transcribed, we collected RNAseq data and made a *de novo* assembly of *A. virens*  
265       transcriptome. Fresh tissue samples were collected from 10 live individuals in the field  
266       (five tissue types: blood, brain, breast tissue, heart and liver). Tissue samples were stored  
267       in either PAXgene (Blood RNA Tubes, PreAnalytiX/Qiagen, Switzerland) or Allprotect  
268       (Tissue Reagent, Qiagen, Germany) buffer and shipped to laboratory facilities at UCLA.  
269       RNA was extracted from each sample and tissue type separately using an RNeasy kit  
270       (Qiagen, Germany), and based on quality of extractions (both overall concentration and  
271       260/280 ratio), three RNA samples from three tissue types (brain, heart and liver) were  
272       chosen to perform library preparations. RNAseq libraries were prepared using Illumina  
273       TruSeq RNA Library Prep Kit v.2 (Illumina, San Diego, California) following the

274 manufacture's protocol, and libraries were indexed, normalized, pooled, and sequenced on  
275 a single lane on Illumina HiSeq 2500 (paired end 100 bp reads, Rapid Run mode) at GSL.

276 We obtained one lane of paired-end RNAseq data pooled from three tissue types.  
277 We first removed bases with quality scores lower than 20 and minimum sequence length of  
278 30bp using *trim\_galore* ([http://www.bioinformatics.babraham.ac.uk/projects/trim\\_galore/](http://www.bioinformatics.babraham.ac.uk/projects/trim_galore/)).  
279 We then pooled the remaining paired-end reads from different tissues together for a *de*  
280 *novo* assembly of the transcriptome using the *Trinity* pipeline (Grabherr *et al.* 2011). We  
281 assessed the quality of the assembly using scripts provided in the *Trinity* package, and  
282 predicted the coding regions in the assembled transcriptome using *TransDecoder* in *Trinity*.

283

#### 284 *Detecting population structure using genomic data*

285 To detect the underlying population structure among samples, we performed a  
286 Principal Component Analysis (PCA) using the bioconductor package *SNPrelate* (Zheng *et al.*  
287 *et al.* 2012). 47,482 SNPs with MAF  $\geq 2\%$  were used in PCA. The first six principal  
288 components were visually examined to identify clustering patterns of samples and to  
289 determine whether these genetic clusters tend to segregate with ecological factors or  
290 geography. We also used the program ADMIXTURE (Alexander *et al.* 2009) to estimate  
291 the ancestry of individual genotypes, using only the first SNP of each RAD loci to limit the  
292 impact of linkage disequilibrium. The analysis was run for  $K = 1-15$ .

293 To quantify pairwise population differentiation, we calculated pairwise  $F_{ST}$   
294 between populations using *SNPrelate*. The correlation of population genetic differentiation  
295 (pairwise  $F_{ST}$ ) and geographic distance, in other words, the presence of isolation by  
296 distance (IBD), was estimated by a simple Mantel test with 999,999 permutations using  
297 *vegan2.2-1* in R (Mantel 1967; Oksanen *et al.* 2017). Mantel tests are reported using both  
298 raw  $F_{ST}$  and  $F_{ST}/(1 - F_{ST})$ , as well as both raw Euclidian geographic distance and log-  
299 transformed distances (Slatkin 1995; Rousset 1997).

300 Moreover, we found that pairwise  $F_{ST}$  between populations from different habitats  
301 were higher than pairwise  $F_{ST}$  computed between populations from the same habitat (see  
302 Results). In principle, this pattern could solely be driven by isolation by distance as  
303 populations from the same habitat tend to be located closer geographically to each other  
304 compared to population from different habitats. To determine whether the elevated  $F_{ST}$

305 between populations from different habitats (compared to  $F_{ST}$  between populations from  
306 the same habitat) could be explained simply by the differences in geographic distance, we  
307 performed permutation tests that accounted for the fact that populations from different  
308 habitats tend to be further apart. Specifically, we divided the population pairs into five bins  
309 based upon their geographic distances from each other (i.e. <200km, 200-400km, 400-  
310 600km, 600-800km, >800km). Then within each bin, we permuted whether the pairwise  
311  $F_{ST}$  values are from a within-habitat comparison or a between-habitat comparison. We  
312 generated 10,000 such permutations and, for each permutation, performed a t-test on  
313 whether the  $F_{ST}$  values for between-habitat comparisons were higher than those for within-  
314 habitat comparisons. From the permuted data, we built a null distribution of t-statistics,  
315 which accounted for the effect of geography. Our final empirical p-value for the observed  
316 data was calculated as the percentage of permuted datasets that had a t-statistic as large  
317 or larger than the one seen in the original data. Similar permutation analyses were applied  
318 to the dataset including all habitats as well as to a dataset that only considered rainforest  
319 and ecotone populations. In the null distribution of t-statistics for the test of whether  $F_{ST}$  is  
320 higher between than within habitats, we found that none of the 10,000 permuted datasets  
321 had a t-statistic of  $F_{ST}$  as large or larger than the one seen in the original data, suggesting a  
322 p-value < 1e-04. However, for the null distribution of t-statistics for the test of whether  
323 distance is higher between or within habitats, 1581 of the 10,000 permuted datasets had a  
324 t-statistic as large or larger than the one seen in the original data, suggesting a p-value =  
325 0.158. This suggests that our null distribution of t-statistics accounts for the fact that  
326 populations from similar habitats tend to be closer together geographically.

327 As an alternative method to test whether habitat contributed to the observed pattern  
328 of population differentiation above and beyond geographic distance, we created a binary  
329 matrix that indicated whether a pair of populations was from the same habitat or not. We  
330 tested the correlation of genetic distance matrix and this matrix while controlling for  
331 geographic distance using a partial Mantel test using *vegan* 2.2-1 in R (Mantel 1967;  
332 Oksanen *et al.* 2017). Partial Mantel tests are performed using both raw  $F_{ST}$  and  $F_{ST}/(1-$   
333  $F_{ST})$ , as well as both raw Euclidian geographic distance and log-transformed distances  
334 (Slatkin 1995; Rousset 1997).

335

336 *Identifying outlier SNPs under selection*

337 We used *BayeScan2.1* (Foll & Gaggiotti 2008) to identify highly differentiated  
338 SNPs that are candidates to be under natural selection. This program takes a Bayesian  
339 approach to search for SNPs exhibiting extreme  $F_{ST}$  values that could be due to local  
340 adaptation. Outlier SNPs were identified using SNPs with  $MAF \geq 2\%$  across all samples,  
341 specifying all 15 populations or four habitats (see supplemental Notes). We ran 5000  
342 iterations using prior odds of 10 and assessed the statistical significance of a locus being an  
343 outlier using a false discovery rate (FDR) of 5%.

344 To explore the spatial patterns of population differentiation across chromosomes,  
345 we mapped the consensus RAD loci to the zebra finch genome using BLAT with default  
346 parameters. For the uniquely mapped RAD loci, we plotted the  $F_{ST}$  of each SNP by  
347 genome coordinates to examine spatial patterns of outlier SNPs. To interpret the potential  
348 biological function of the outlier SNPs identified by Bayescan analysis, we used a zebra  
349 finch genome annotation (v3.2.4.78) to identify outlier SNPs mapped to annotated genic  
350 regions.

351 We further examined whether candidate loci under selection were enriched in  
352 exonic (transcribed) or coding regions. To do this, we mapped RAD loci to the *de novo*  
353 assembled *A. virens* transcriptome using BLAT with default parameters. Any RAD locus  
354 that mapped to the transcriptome was considered from exonic regions of the genome, and  
355 the remaining RAD loci were labeled “non-transcribed” regions of the genome. Similarly,  
356 we mapped RAD loci to predicted coding sequences and categorized them into coding and  
357 non-coding RAD loci. We then used a one-sided Fisher’s exact test to examine whether  
358 there was significant enrichment of outlier loci in exon or coding regions of the genome.  
359 Finally, we cross-checked these outliers to see if there was any significant associations  
360 with environment using Latent Factor Mixed Models (Frichot *et al.* 2013) (see  
361 Supplemental Notes for more details).

362

363 *Detecting environmental drivers of genomic variation*

364 In addition to population structure, we also tested whether allele frequencies in  
365 different populations were associated with environmental variables across the range of *A.*  
366 *virens* using the package *gradientForest* (Ellis *et al.* 2012) in the R statistical framework

367 (R Working Group, 2014). Gradient forests are an extension of random forests (Breiman  
368 2001) that treat response variables (in this case, individual SNP minor allele frequencies  
369 within each population) as members of a larger community (the total genome), and  
370 provides summary statistics based on ensembled forest runs to indicate an overall  
371 association of changes in allele frequency to particular environmental variables (Ellis *et al.*  
372 2012; Fitzpatrick & Keller 2015). Gradient forests were run using the following changes to  
373 the default settings: number of trees run for each environmental variable = 500, number of  
374 SNPs included in each bin = 1000. Allelic frequencies across the genome were predicted  
375 for unsampled geographic locations by generating a random set of 100,000 points across  
376 the range of *A. virens*. Then we used our final gradient forest model to predict allele  
377 frequencies at each of those points, given their environmental characteristics. Ordinary  
378 Kriging (Oliver & Webster 1990) was then used within ArcMap (ESRI, Redlands, CA) to  
379 generate a continuous surface across the known range of *A. virens* in Cameroon.

380 We used a suite of 17 environment variables (Table S6), including bioclim  
381 measures of temperature and precipitation (n=9; any variables showing a Pearson's  
382 correlation coefficient > 0.7 were removed) downloaded from the Worldclim database  
383 (www.worldclim.org), measures of vegetation and tree cover captured using the NASA  
384 MODerate-resolution Imaging Spectroradiometer (MODIS, n=4), elevation and slope  
385 captured via the Shuttle Radar Topography Mission (n=2), and surface moisture estimates  
386 measures using the Quick Scatterometer (QuikSCAT, n=2). In addition to these variables,  
387 and to account for purely geographic associations, we also included Euclidean distances  
388 (measured as Latitude and Longitude) as predictor variables in all model.

389

## 390 **Results**

### 391 *SNP discovery and overall genetic diversity*

392 We used RAD sequencing to survey the genome-wide diversity of *A. virens*. The  
393 final sample set included 15 populations from four different habitats, including rainforests,  
394 ecotones, mountains, and an island (Figure 1A; Table S1). After removing low quality  
395 reads and samples, we obtained a total of 916 million reads for 182 *A. virens* samples  
396 (SRA:xxxxxx). The number of raw sequence reads per sample ranged from 1.60 to 20.73  
397 million. On average, 99.2% of these reads were utilized in the *de novo* assembly of the

398 RAD loci. The mean coverage depth ranged from 16× to 136× per sample (mean = 38×,  
399 median = 32×, Figure S2). Using this dataset, we assembled and identified 34,657 high  
400 quality RAD loci that passed our quality filters and were genotyped in more than 80% of  
401 all final samples. On these 34,657 consensus RAD loci, there were a total of 255,290  
402 SNPs. The median number of SNPs per RAD locus is seven. With a minimum minor allele  
403 frequency filter of 2%, we retained 47,482 SNPs that were present on 23,882 RAD tags  
404 (Table S2; Supplemental Notes).

405 The number of segregating sites ranges from 25,936 to 70,598 per population.  
406 Waterson's  $\theta$  ( $\theta_w$ ) was estimated to be 0.0049 - 0.0076/bp (mean = 0.0064/bp), and  $\pi$   
407 ranges from 0.0034 - 0.0037/bp (mean = 0.0036/bp) (Table S1), which is comparable to  $\pi$   
408 estimated from other bird species (Nadachowska-Brzyska *et al.* 2013; Romiguier *et al.*  
409 2014). Overall levels of genetic diversity are comparable in each habitat, including the  
410 island population (Table S1). The finding that  $\theta_w$  is larger than  $\pi$  indicates an excess of  
411 low-frequency variants relative to the standard neutral model which could be driven by  
412 recent population expansions.

#### 413 *Transcriptome assembly and annotation*

414 Transcriptome assembly was performed using 169 million paired-end RNAseq data  
415 from three different tissue types. The assembled transcriptome had a GC content of 45%.  
416 The average contig length was 815 bp and N50 was 1,619 bp. In total, trinity produced  
417 237,226 genes and 286,494 transcripts, and predicted 81,018 coding sequences from these  
418 transcripts (Table S5). Of the 34,657 RAD loci we genotyped, 8412 RAD loci (24.2%)  
419 were mapped to the *de novo* assembled *A. virens* transcriptome, and 3,618 RAD loci  
420 (10.4%) were mapped to the predicted coding sequences (Figure S3). The RAD tags  
421 overlapping coding sequence tend to have fewer SNPs than those that do not overlap with  
422 coding sequences (Figure S4), consistent with the fact that the coding regions are likely  
423 under stronger selective constraint.

424

#### 425 *Population structure*

426 We used PCA to identify population structure in little greenbuls. The first two PCs  
427 revealed a clear clustering pattern of individuals from the same habitats (Figure 1B).  
428 Populations from island, mountains, rainforests, and ecotones formed four discrete clusters,

429 suggesting genomic divergence across ecological gradients and habitats. Island and  
430 mountain populations were most distinct (Figure 1B), however samples from all four  
431 habitats separated on PC1, including those from ecotone and rainforest habitats. PC2  
432 further separated island and mountain samples from rainforest and ecotone samples.  
433 Remarkably, results suggest that, within rainforest and ecotone habitats, individual  
434 populations could be distinguished solely on the basis of genomic markers, mostly by PC1,  
435 with individuals from the same sampling sites clustering together (Figure 1C). The level of  
436 separation of ecotone populations from rainforest populations along PC1 roughly followed  
437 a latitudinal gradient, corresponding to environmental and rainfall gradients that  
438 distinguish ecotone in the north from rainforest in the south of Cameroon (see  
439 environmental analyses below). Specifically, samples collected from sites Wakwa and  
440 Ngaoundaba Ranch, toward the more extreme edge of the ecotone habitat and had the most  
441 extreme ecotone environmental conditions, formed clusters that were most distant from the  
442 rainforest samples, while samples collected from Betare Oya, at lower ecotone that was  
443 closest to the central rainforests, formed a cluster closest to the rainforest (Figure 1C). The  
444 pattern of genomic differentiation across habitats was confirmed using the program  
445 *Admixture* (Figure S5).

446 Pairwise  $F_{ST}$  between populations ranged from 0.017 to 0.078 (mean = 0.038;  
447 Figure 2A-B; Table S3), indicating low overall levels of genomic differentiation across  
448 populations. There was significant correlation between pairwise  $F_{ST}$  and geographic  
449 distance between the populations (Mantel  $r = 0.34$ ; mantel simulated  $p$ -value = 0.003),  
450 suggesting isolation by distance contributes to population differentiation. However,  
451 pairwise  $F_{ST}$  between populations from different habitats was significantly higher than  
452 pairwise  $F_{ST}$  between populations within the same habitat (one tailed  $t$  test,  $p$ -value =  
453  $1.36e-10$ ; Figure 2A). Pairwise geographic distances between populations from different  
454 habitats were also significantly higher than pairwise distances between populations within  
455 the same habitat (one tailed  $t$ -test,  $p$ -value =  $1.015e-06$ ). To account for the fact that  
456 populations from different habitats were also geographically further apart, we performed a  
457 permutation test, where we randomized whether a population pair was from the same or  
458 different habitats in different bins stratified by their geographic distance. Using permuted  
459 datasets, we built a null distribution of these  $t$ -statistics (that already includes the effect of

460 geographic distance), which we used to evaluate the significance of our observed value.  
461 The higher  $F_{ST}$  value for between-habitat comparison was highly significant when  
462 compared to this improved null distribution ( $p$ -value  $< 1e-04$ , Figure 2C and Figure S6),  
463 suggesting that isolation by distance alone cannot explain the higher  $F_{ST}$  between habitats  
464 than within habitats. Similarly, only considering rainforest and ecotone populations,  
465 pairwise  $F_{ST}$  was significantly higher between habitats as compared to within habitats (one  
466 tailed t-test,  $p$ -value =  $9.793e-05$ ). Application of the same permutation test shows the  
467 higher  $F_{ST}$  between ecotone and rainforest populations ( $p$ -value= $0.0055$ ) cannot be  
468 explained by geographic distance alone (Figure 2D and Figure S6).

469 To confirm this finding using an alternative statistical approach, we used partial  
470 Mantel tests to determine the contribution of habitat types of population pairs to their  
471 genetic differentiation ( $F_{ST}$ ), controlling for geographic distance. We found a highly  
472 significant and positive correlation between genetic distance and whether a pair of  
473 population comes from the same habitat, and greater genetic differentiation (higher  $F_{ST}$ )  
474 from between-habitat populations compared to within-habitat populations, while  
475 controlling for geographic distance (Table 1). Taken together these results suggest that  
476 factors other than geographic location, such as local adaptation significantly contribute to  
477 population differentiation between habitats.

478 In addition, mountain and island populations were more diverged from other  
479 populations (Figure 2B). Interestingly,  $F_{ST}$  between two mountain populations were  
480 exceptionally high ( $F_{ST}= 0.060$ ) compared to other within habitat pairwise  $F_{ST}$  (ranging  
481 from 0.017 to 0.040, Figure 2A), despite the fact that the two mountain populations were  
482 geographically very close to each other. The  $F_{ST}$  values between mountain populations and  
483 lowland forest/ecotone populations were larger than pairwise  $F_{ST}$  values between lowland  
484 populations, suggesting mountain populations are highly differentiated both from one  
485 another and from lowland populations.

486

#### 487 *Candidate loci under selection*

488 To further explore potential candidate loci under selection, we identified SNPs with  
489 extreme allele frequency differences across populations, which should be enriched by  
490 targets of local adaptation. We identified 140 outlier SNPs across all populations with a

491 False Discovery Rate of 5% using *Bayescan*. These candidate SNPs are potential targets of  
492 divergent selection across different sampling sites (Figure S7). The 140 outlier SNPs reside  
493 in 119 loci, and 40 of these loci mapped to the zebra finch genome (Figure S9). Of these,  
494 36 mapped to main scaffolds of known chromosomes and four mapped to the Z  
495 chromosome. Only 13 of these outlier loci mapped to annotated genic regions on the zebra  
496 finch genome and nine mapped to genes with functional annotations (Table S4).

497     ■ In order to uncover the functional significance of outlier loci, we used the *de novo*  
498 assembly of greenbul transcriptome to partition the RAD loci and SNPs into different  
499 categories depending on whether they mapped to coding regions or transcribed (exonic)  
500 regions (Figure S10). This enabled us to test for enrichment of outlier SNPs in putatively  
501 functional regions. Of the 47,482 SNPs, 9677 mapped to the transcriptome, and 42 were  
502 outliers based on a Bayescan analysis. Using a one-sided Fisher's exact test, we detected  
503 significant enrichment of outlier loci in exonic regions of the genome ( $p=0.0044$ ; Table S2;  
504 Figure S10). Using the predicted coding sequence from the transcriptome, 3,602 SNPs  
505 mapped to the predicted coding sequences and 21 of these were outliers. We again detected  
506 a significant enrichment of outlier SNPs in protein-coding sequences ( $p=0.002$ ; Table S2;  
507 Figure S10). Taken together, these enrichment results provide additional confidence that  
508 the outlier loci found using Bayescan captured functionally important, biologically relevant  
509 genetic variants, which were not merely loci that fell within the tail of a neutral  
510 distribution.

511

#### 512 *Genomic Turnover Across Environments*

513     Because some environmental adaptation may involve shifts in allele frequency at  
514 many loci across the genome (e.g., polygenic selection involving many genes of small  
515 effect), we used the *gradientForest* approach to look for correlations in allele frequencies  
516 associated with environmental variables. A total of 7238 SNPs, ~15% of all SNPs, had  $R^2$   
517 values above 0 (0.0073-0.83) when testing for a correlation between frequency and an  
518 environmental variable. Of the 19 environmental and geographical variables included in  
519 models (Table S6), variables capturing temperature variation (Min Temp: minimum  
520 temperature of the coldest month, Temp Range: mean diurnal temperature range, and Mean  
521 temp: mean annual temperature) and elevation were most important in explaining

522 environmentally associated variation in SNPs (Figure 3A). In some cases, measures of  
523 surface moisture or tree cover were also important, but axes for these variables largely  
524 overlapped with temperature or elevation measures along PC plot (likely the result of co-  
525 linearity in environmental variables) (Figure 3B). Results from LFMM analyses indicated  
526 these same variables were also associated with differentiation observed at hundreds of  
527 individual loci, although exact functions of these regions remain unknown (see  
528 Supplemental Notes).

529 Geographic variables alone were not as important in explaining variation in allele  
530 frequency, again suggesting that geographic distance cannot fully account for all variation  
531 in SNP frequencies across the range of little greenbuls. Above and beyond neutral  
532 processes, selective pressures imposed by differences in these environments best explains  
533 the observed genomic patterns of variation. Predictions across Cameroon suggest strong  
534 genomic turnover (defined as coordinated shifts in allele frequencies across the genome)  
535 throughout the forest, savannah, and ecotone regions, with diagnostic genomic variation  
536 occurring in each of these habitats (Figure 3). Distinct SNP frequencies at high elevations  
537 (Figure 3B-C), and the fact that elevation explains a large proportion of variation of allele  
538 frequencies in the greenbul genome (largely allied with PC1, Figure 3B) also suggest a  
539 unique genetic signature in populations at elevation.

540

## 541 **Discussion**

542 In this study, we used genome-wide RADseq SNPs to characterize the overall level  
543 of genetic diversity in *A. virens* populations across four different habitats. We found  
544 evidence of population structure of *A. virens* consistent with habitat type and previously  
545 observed phenotypic divergence. We demonstrated that population differentiation across  
546 habitats cannot be explained solely by isolation-by-distance, suggesting local adaptation  
547 further contributes to genomic divergence among habitats. We identified 140 outlier SNPs  
548 that are potential targets of selection and the fact that they are significantly enriched in  
549 exonic and coding regions suggests they are functionally important. Environmental  
550 association analysis further supports this conclusion and shows environmental variables,  
551 including temperature and elevation, are highly associated with patterns of genomic  
552 variation across the range of the little greenbul.

553 In addition to the differences between rainforest and ecotone populations, other  
554 habitats were found to harbor distinct patterns of genetic variation. The population from  
555 Bioko island formed a distinctive genetic cluster based on PCA, and also was identified as  
556 distinct in environmental association models (Figure 1B and 3), consistent with previous  
557 studies (Smith *et al.* 2005). Bioko island is 32 km off the coast of Africa, separated from  
558 mainland ~10,000 years ago and has an area of 2,017 km<sup>2</sup>. Island populations and species  
559 may have smaller effective population sizes than mainland populations or sister taxa  
560 (Robinson *et al.* 2016), due to possible population bottleneck and considerably smaller  
561 ranges. As a result, island populations may have lower genetic diversity compared to their  
562 mainland counterparts (Frankham 1997). However, several recent empirical studies  
563 suggested this may not always be the case, particularly in birds (Francisco *et al.* 2015;  
564 James *et al.* 2016). In our study, the estimates of genetic variation using genome wide SNP  
565 markers in greenbul population on Bioko island are comparable to those from mainland  
566 populations (Table S1). This is consistent with the recent findings that island populations  
567 do not always have lower genetic diversity (Francisco *et al.* 2015; James *et al.* 2016), and  
568 the fact that Bioko island is a large island that only recently separated from the mainland.

569 Tropical mountains are well known to support disproportionately high biodiversity  
570 and are thought to be hotspots for avian speciation (Roy 1997; Myers *et al.* 2000; Smith *et*  
571 *al.* 2000; Orme *et al.* 2005; Fjeldså *et al.* 2007, 2012; Drovetski *et al.* 2013). Little  
572 greenbuls are found at elevations up to 2400 m, where environmental variables,  
573 particularly temperature and vegetation, change rapidly along altitudinal gradients. Our  
574 two mountain populations have high  $F_{ST}$  despite being geographically close and from same  
575 habitat type. Although the Euclidean distance between these two mountain populations is  
576 short, the environmental changes along altitudinal gradients are steep, causing isolation  
577 between populations from different mountains and forming “sky islands”, between which  
578 the level of gene flow probably is much lower than among lowland populations. Moreover,  
579 we found that the two different mountain populations exhibited the lowest within-  
580 population genetic variation among all sampled populations (Table S1). While this  
581 difference was not statistically significant (likely due to small sample sizes), this decreased  
582 level variation can inflate  $F_{ST}$ , the relative measurement of population differentiation. It  
583 also suggests that mountain populations may have overall smaller effective population

584 sizes (consistent with presumably smaller suitable habitat size for mountain populations)  
585 and/or have experienced serial bottleneck/founder effects as range expansions occurred.  
586 These processes can further contributed to divergence due to stronger genetic drift within  
587 each subpopulation leading to faster changes in allele frequency. The idea that elevation  
588 can drive genomic changes is supported by previous estimates of morphological  
589 divergence (Smith *et al.* 2005), and emphasizes the importance of preserving elevational  
590 gradients in tropical ecosystems in general (Thomassen *et al.* 2011).

591 Most of the genes containing outlier SNPs only have annotations predicted from  
592 human homologs, except two that have annotations from bird species. Both of these two  
593 genes are of particular interest. One outlier locus mapped to the 5UTR/coding junction of  
594 *EDIL3*, a calcium-binding protein that has been found to function in avian eggshell  
595 biomineralization (Marie *et al.* 2015). Avian eggshells protect the developing embryo and  
596 keep the egg free from pathogens. Environmental factors such as temperature, humidity,  
597 and partial oxygen pressure have been reported to affect avian eggshell structure, and  
598 previous studies documented rapid evolution of eggshell structure in response to  
599 colonization of novel environments in the house finch (Stein & Badyaev 2011). The  
600 second outlier locus mapped to *MLXIPL* (MLX interacting protein-like), which is co-  
601 activator of the carbohydrate response element binding protein that has been correlated  
602 with fat deposition in caged chicken (Proszkowiec-Weglarz *et al.* 2008; Li *et al.* 2015).  
603 Interestingly, seven more genes that contain outlier SNPs have annotation linked with  
604 metabolic traits or diseases in humans. For example, outlier SNPs were found in *RGS6*  
605 (Sibbel *et al.* 2011 p. 6), *CSAD* (Comuzzie *et al.* 2012) and *UNC13B* intron (Tréguet *et*  
606 *al.* 2008), which were associated with dietary fat intake, food preference, adiposity/obesity  
607 and diabetes in humans. Although metabolic traits were not measured, adult little  
608 greenbuls from the rainforest have significantly smaller body mass and body size  
609 compared to ecotone, mountain, and island populations (Smith *et al.* 1997, 2005), which  
610 could be the result of divergent selection of these genes associated with metabolic traits.  
611 Several recent studies have discussed the limitations of identifying  $F_{ST}$  outlier as loci under  
612 divergent selection, and suggest results should be interpreted carefully, because many other  
613 factors, including demographic history, recombination rate heterogeneity, and background  
614 selection may also create  $F_{ST}$  outliers (Roesti *et al.* 2012; Lotterhos & Whitlock 2014;

615 Cruickshank & Hahn 2014). Current work to model the demographic history of *A. virens*  
616 should help examine these various possibilities in greater detail.

617 Numerous hypotheses have been proposed for how biodiversity is generated in  
618 rainforests (Mayr & O'Hara 1986; Moritz *et al.* 2000). With the rapid advances in  
619 genomics and environmental modeling in the last decade, it is now possible to examine  
620 these mechanisms in greater depth. Using more powerful genome-wide data, we have  
621 shown, for the first time, strong patterns of population structure and genomic  
622 differentiation between rainforest and ecotone habitats in *A. virens* (Figure 1). Previously,  
623 no genetic differentiation was found between morphologically divergent populations in  
624 rainforest and ecotone habitats, leaving open the possibility that the observed  
625 morphological difference could simply be the result of a homogenized meta-population  
626 that differentially responds to environmental gradients. Although identifying the  
627 underlying genetic basis of morphological traits that differ between rainforest and ecotone  
628 populations was beyond the scope of this study, our results complement previous work by  
629 demonstrating that populations along the rainforest - ecotone gradient are diverging at the  
630 genomic level, and raise the possibility that local adaptation could account for the patterns  
631 of morphological variation previously observed across ecotone-rainforest gradients.  
632 Results also complement past research on reproductive behavior, which found differences  
633 in song characteristics along the forest-ecotone gradient, and showed experimentally that  
634 singing males respond more aggressively to male songs from their own habitat, suggesting  
635 incipient reproductive isolation driven by habitat (Slabbekoorn & Smith 2002; Kirschel *et*  
636 *al.* 2011; Smith *et al.* 2013). These patterns of differentiation are consistent with models of  
637 ecological speciation, where natural selection caused by shifts in ecology can promote  
638 speciation (Orr & Smith 1998; Schneider *et al.* 1999; Schluter 2000; Ogden & Thorpe  
639 2002; Rundle & Nosil 2005; Price 2008; Räsänen & Hendry 2008; Schluter 2009;  
640 Beheregaray *et al.* 2015; Hanson *et al.* 2016). However, further research is necessary to  
641 more fully understand the evolutionary significance of divergence across ecological  
642 gradients and ecotones. In particular, studies investigating the underlying genetic basis of  
643 phenotypic differentiation and mate choice experiments would provide additional insights  
644 into their importance in divergence and speciation.

645 **Author Contributions**

646 T.B.S. and K.E.L conceived of and supervised the project. S.L. and R.J.H conducted the  
647 laboratory work. Sequence assemblies, population structure and outlier analysis was  
648 primarily carried out by Y.Z. with assistance from T.N., K.R., E.C.A. and K.E.L.  
649 Environmental association analysis was performed by R.J.H. The manuscript was written  
650 by Y.Z., R.J.H., K.R., K.E.L., and T.B.S., with input from all authors.

651

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660

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848

849 **Data Accessibility:**

850 - RADseq data: NCBI SRA database BioProject ID PRJNA390986

851 - RNAseq data: NCBI SRA database BioProject ID PRJNA390772

852 - Data files including RAD loci consensus sequences, VCF file and sample information  
853 available at Dryad doi:10.5061/dryad.8n8t0

854

855 **Table 1.** Simple Mantel test for IBD (isolation-by-distance) and partial Mantel test for the  
856 effect of habitat.

Simple Mantel test: test for IBD

correlation between	Mantel r	p
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$F_{ST}$	non-transformed distance	0.34	0.003
$F_{ST}$	log-transformed distance	0.29	0.008
$F_{ST}/(1 - F_{ST})$	log-transformed distance	0.28	0.007

857

Partial Mantel test: test for the effect of habitat while controlling for IBD

	correlation between	control for	Mantel r	p
$F_{ST}$	same habitat or not	non-transformed distance	0.48	9.00E-06
$F_{ST}$	same habitat or not	log-transformed distance	0.50	1.00E-06
$F_{ST}/(1 - F_{ST})$	same habitat or not	log-transformed distance	0.50	3.00E-06

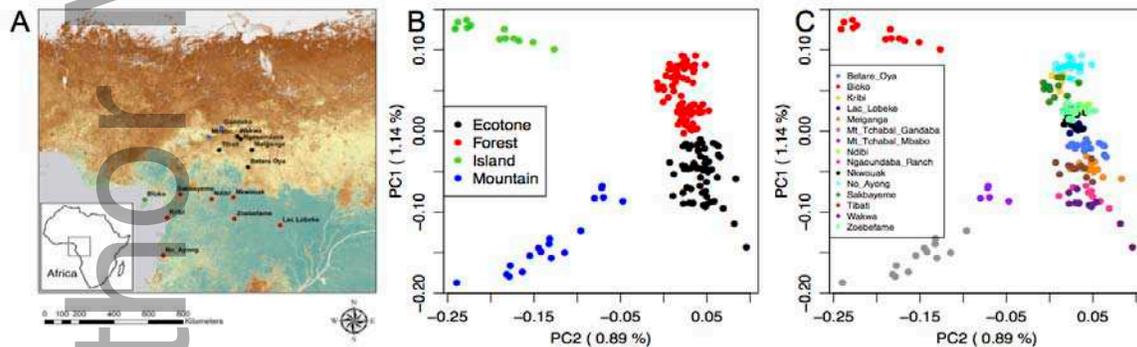
858 P-values were generated by 999,999 permutations.

859 Here “distance” refers to the geographic distance separating the pair of populations on

860 which the  $F_{ST}$  value was computed. **Figure Legends**

861 **Figure 1. Sampling and population structure.** (A), Sampling locations. Each point is a  
 862 sampling location, and habitat types are indicated by color same to (B). (B-C), PCA using  
 863 SNPs that have a minor allele frequency higher than 2%. Each point presents a sample, and  
 864 samples are colored by their habitat types (B) and by populations (C).

865

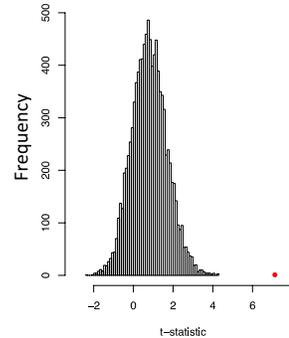


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868 **Figure 2. Pairwise population differentiation.** (A), Pairwise  $F_{ST}$  between populations  
 869 correlates with pairwise geographic distance between populations. Empty circles denote  
 870 pairs of populations from the same type of habitat (shown by the color of the circle). Solid  
 871 circles are pairs of populations from different types of habitats (shown by colors of the  
 872 circle and inside). (B), Heat map of pairwise  $F_{ST}$ . Sampling locations are grouped by  
 873 habitat type in both axes. (C) and (D), The pairwise  $F_{ST}$  of populations from different

874 habitats are greater than the pairwise  $F_{ST}$  of populations from the same habitat, even at the  
875 same geographic distance. (C) includes all populations from four habitats, and (D) includes  
876 only rainforest and ecotone populations. Histogram shows the null distribution of t-  
877 statistics generated by 10000 permutations of habitats within different bins of geographic  
878 distance (see Methods). Red dot shows the observed value.



Frequency  
t-statistic

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880

881 **Figure 3. Environmental drivers of genomic variation.** (A), Environmental and  
882 geographical variables ranked by their importance in explaining SNP allele frequency  
883 variation. (B), PC plot indicates the contribution of the environmental variables to the  
884 predicted patterns of frequency differentiation, with labeled vectors indicating the direction  
885 and magnitude of environmental gradients with greatest contribution. (C), Predicted spatial  
886 variation in population-level genetic composition from SNPs. Red points in (C) are  
887 locations where actual samples were collected in this study. Colors in (B) and (C) represent  
888 gradients in genomic turnover derived from transformed environmental predictors.

889 Locations with similar colors are expected to harbor populations with similar genetic  
890 composition.

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**Table 1.** Simple Mantel test for IBD (isolation-by-distance) and partial Mantel test for the effect of habitat.

Simple Mantel test: test for IBD

correlation between		Mantel r	p
$F_{ST}$	non-transformed distance	0.34	0.003
$F_{ST}$	log-transformed distance	0.29	0.008
$F_{ST}/(1- F_{ST})$	log-transformed distance	0.28	0.007

Partial Mantel test: test for the effect of habitat while controlling for IBD

correlation between		control for	Mantel r	p
$F_{ST}$	same habitat or not	non-transformed distance	0.48	9.00E-06
$F_{ST}$	same habitat or not	log-transformed distance	0.50	1.00E-06
$F_{ST}/(1- F_{ST})$	same habitat or not	log-transformed distance	0.50	3.00E-06

P-values were generated by 999,999 permutations.

Here “distance” refers to the geographic distance separating the pair of populations on which the  $F_{ST}$  value was computed.

