

Combining multiple visual surveys to model the habitat of deep-diving cetaceans at the basin scale

Large scale modelling of deep-diving cetacean habitats

Auriane Virgili ^{1*}, Matthieu Authier ², Oliver Boisseau ^{3,4}, Ana Cañadas ⁵, Diane Claridge ⁶, Tim Cole ⁷, Peter Corkeron ⁷, Ghislain Dorémus ², Léa David ⁸, Nathalie Di-Méglio ⁸, Charlotte Dunn ⁶, Tim E. Dunn ⁹, Isabel García-Barón ¹⁰, Sophie Laran ², Giancarlo Lauriano ¹¹, Mark Lewis ⁹, Maite Louzao ¹⁰, Laura Mannocci ^{12,a}, José Martínez-Cedeira ¹³, Debra Palka ⁷, Simone Panigada ¹⁴, Emeline Pettex ^{2,b}, Jason J. Roberts ¹², Leire Ruiz ¹⁵, Camilo Saavedra ¹⁶, M. Begoña Santos ¹⁶, Olivier Van Canneyt ², José Antonio Vázquez Bonales ⁵, Pascal Monestiez ^{1,17}, Vincent Ridoux ^{1,2}

¹ Centre d'Etudes Biologiques de Chizé - La Rochelle, UMR 7372 CNRS - Université de La Rochelle, Institut du Littoral et de l'Environnement, 17000 La Rochelle, France; ² Observatoire PELAGIS, UMS 3462 CNRS - Université de La Rochelle, Systèmes d'Observation pour la Conservation des Mammifères et des Oiseaux Marins, 17000 La Rochelle, France; ³ Marine Conservation Research, 94 High Street, Kelvedon, CO5 9AA, UK; ⁴ Song of the Whale research team, International Fund for Animal Welfare (IFAW), 87-90 Albert Embankment, London, SE1 7UD, UK; ⁵ Anilam Research and Conservation, Pradillos 29, 28491-Navacerrada, Madrid. Spain; ⁶ Bahamas Marine Mammal Research Organisation, P.O. Box AB-20714, Marsh Harbour, Abaco, Bahamas; ⁷ Protected Species Branch, NOAA Fisheries Northeast Fisheries Science, 166 Water St Woods Hole Massachusetts, 02543 USA; ⁸ EcoOcéan Institut, 34090 Montpellier, France; ⁹ Joint Nature Conservation Committee, Inverdee House, Baxter Street, Aberdeen, AB11 9QA, UK; ¹⁰ AZTI, Herrera Kaia, Portualdea z/g, 20110 Pasaia, Spain; ¹¹ Institute for Environmental Protection and Research – ISPRA, via V. Brancati 60, 00144 Roma, Italy; ¹² Marine Geospatial Ecology Laboratory, Duke University, Durham, North Carolina, USA; ¹³ CEMMA, Camiño do Ceán, n°2, 36350, Nigrán, Pontevedra, Spain; ¹⁴ Tethys Research Institute, Acquario Civico, 20121 Milano, Italy; ¹⁵ AMBAR Elkartea organisation, Mungia Bidea 9, 3B, 48620 Plentzia, Bizkaia, Spain; ¹⁶ Instituto Español de Oceanografía, Centro Oceanográfico de Vigo, 36390 Vigo, Spain; ¹⁷ BioSP, INRA, 84914 Avignon, France.

*Corresponding author: Auriane Virgili, 2 rue Olympe de Gouges 17000 La Rochelle, France
auriane.virgili@univ-lr.fr

^a Current address: UMR MARBEC (IRD, Ifremer, Univ. Montpellier, CNRS), Institut Français de Recherche pour l'Exploitation de la Mer, Av. Jean Monnet, CS 30171, 34203 Sète, France

^b Current address: Cohabys - ADERA, Université de La Rochelle 2, rue Olympe de Gouges, Bâtiment ILE, 17000 La Rochelle

Acknowledgments

This is the author manuscript accepted for publication and has undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the [Version of Record](#). Please cite this article as [doi: 10.1111/geb.12850](https://doi.org/10.1111/geb.12850)

We are grateful to the many observers who participated in the surveys and collected all the data and also to the ships' captains, crews and pilots. We thank Phil Hammond and his team for providing SCANS and CODA survey data. THUNNUS survey was carried out thanks to the collaboration of the General Directorate of Fisheries and Maritime Affairs, Government of Galicia. We thank the *Direction Générale de l'Armement* (DGA), including Odile Gérard and Carole Nahum, for funding Auriane Virgili's doctoral research grant. ML was funded by a Ramón y Cajal (RYC-2012-09897) postdoctoral contract of the Spanish Ministry of Economy, Industry and Competitiveness, whereas IGB was supported by a PhD fellowship (FPI, BES-2014-070597) of the Spanish Ministry of Economy, Industry and Competitiveness. This study is a contribution to the CHALLENGES (CTM2013-47032-R) project of the Spanish Ministry of Economy, Industry and Competitiveness. EcoOcéan Institut acknowledge its partners for the participation in the collection of data at sea: École Pratique des Hautes Études, WWF-France, Swiss Cetacean Society, Cybelle Planète, Participe Futur and Fondation Nicolas Hulot. This work is linked to the program "PELAGIC 2" lead by the CESAB which aims to establish globally relevant areas for the conservation of marine mammals. We finally thank Charlotte Dunn and Tim Dunn for having proofread the English of the manuscript and the anonymous reviewers for their very helpful comments that led to a clearer and much improved manuscript.

Biosketch

This work is part of Auriane Virgili's PhD project which aims to model distributions of rare marine species with a focus on deep-diving cetaceans. These species are rare and difficult to observe at the surface thus it was necessary to assemble datasets from different surveys to model their distribution in the North Atlantic Ocean and the Mediterranean Sea. This required the collaboration of many organisations represented by the different co-authors of this article.

1

2 DR. AURIANE VIRGILI (Orcid ID : 0000-0002-7581-7985)

3

4

5 Article type : Research Papers

6

7

8 **Combining multiple visual surveys to model the habitat of**
9 **deep-diving cetaceans at the basin scale**

10 **Large scale modelling of deep-diving cetacean habitats**

11

12 **Abstract**

13 **Aim:** Deep-diving cetaceans are oceanic species exposed to multiple anthropogenic
14 pressures including high intensity underwater noise, and knowledge of their distribution is
15 crucial to manage their conservation. Due to intrinsic low densities, wide distribution ranges
16 and limited presence at the sea surface, these species are rarely sighted. Pooling data from
17 multiple visual surveys sharing a common line-transect methodology can increase sightings
18 but requires accounting for heterogeneity in protocols and platforms.

19 **Location:** North Atlantic Ocean and Mediterranean Sea

20 **Time period:** 1998 to 2015

21 **Major taxa:** *Ziphiidae*; *Physeteriidae*; *Kogiidae*

22 **Methods:** About 1,240,000 km of pooled effort provided 630 sightings of ziphiids, 836 of
23 physeteriids and 106 of kogiids. For each taxon, we built a hierarchical model to estimate the
24 effective strip width depending on observation conditions and survey types. We then
25 modelled relative densities in a Generalised Additive Modelling framework. Geographical
26 predictions were limited to interpolations identified with a gap analysis of environmental
27 space coverage.

28 **Results:** Deeper areas of the North Atlantic gyre were mostly environmental extrapolation,
29 thereby highlighting gaps in sampling across the different surveys. For the three species
30 groups, the highest relative densities were predicted along continental slopes, particularly in

31 the western North Atlantic Ocean where the Gulf Stream creates dynamic frontal zones and
32 eddies.

33 **Main conclusions:** Pooling a large number of surveys provided the first basin-wide models
34 of distribution for deep-diving cetaceans, including several data-deficient taxa, across the
35 North Atlantic Ocean and the Mediterranean Sea. These models can help the conservation
36 of elusive and poorly known marine megafauna.

37

38 **Keywords:** Beaked whales, Data-assembling, Deep-diving cetaceans, Habitat modelling,
39 Kogiids, Sperm whales

40

41 **1. Introduction**

42 Deep-diving cetaceans, defined here as beaked whales (family *Ziphiidae*; e.g. *Ziphius*
43 *cavirostris*, *Hyperoodon* spp. and *Mesoplodon* spp.) and sperm whales (families
44 *Physeteridae* and *Kogiidae*), are distributed worldwide. They are oceanic species that feed in
45 deep waters during long dives (close to or even longer than an hour; Perrin et al., 2009). Due
46 to their offshore habitat and the short time they remain available at the sea surface, little is
47 known about their synoptic distribution (especially for kogiids and ziphiids). Moreover, these
48 species are threatened by anthropogenic activities, including bycatch, debris ingestion, ship
49 collisions (Carrillo & Ritter, 2010; Madsen et al., 2014; Unger et al., 2016) and any activity
50 producing high intensity noise (e.g. military sonars, seismic guns or techniques used on large
51 maritime construction projects; Stone & Tasker, 2006). Recent studies have demonstrated
52 the sensitivity of deep-diving cetaceans, and particularly beaked whales, to underwater noise
53 pollution, with a number of unusual stranding events associated with the use of military
54 sonars (Fernández et al., 2005; D'Amico et al., 2009). To mitigate the impact of these
55 activities, accurate knowledge of the distribution and abundance of deep-diving cetaceans is
56 crucial to Marine Spatial Planning to inform management measures at a national scale
57 (Douve, 2008). International initiatives, such as Important Marine Mammal Areas (IMMAs,
58 Corrigan et al., 2014), are needed for these highly mobile species. However, any single
59 survey often yields only a handful of sightings that are then restricted to areas too small
60 compared to the large geographical scale needed for effective conservation planning.

61 Data-assembling is increasingly used to model habitat preferences of cetaceans at the
62 basin scale (Roberts et al., 2016; Rogan et al., 2017; Cañadas et al., 2018). Due to the
63 various protocols, platform types and observation heights, species detectability and data
64 quality vary with surveys. In addition, each survey may not collect the same information,
65 particularly with regard to observation conditions. Some surveys only record Beaufort sea-
66 state while others record additional parameters that also influence species detection, such as
67 sun glare, cloud coverage or wave height. In the process of synthesising different datasets,

68 only variables common across all datasets can generally be retained in a broad scale
69 analysis, which nevertheless needs to account for heterogeneity. Finally, to make basin-wide
70 predictions from the assemblage of a number of local surveys, identifying areas of
71 environmental extrapolations is crucial to bolster confidence in predicted maps (Mannocci et
72 al., 2018).

73 Our study aims to understand how deep-diving cetaceans are distributed at a large scale
74 and to highlight areas of high relative densities for conservation purposes. To model the
75 habitats of deep-diving cetaceans at a large scale, we assembled data from different surveys
76 in the North Atlantic Ocean and the Mediterranean Sea from 15 organisations. To take into
77 account heterogeneity in sighting protocols, we built a hierarchical model to estimate the
78 effective strip width across platforms and observation conditions. We then modelled relative
79 densities of three deep-diving cetacean taxa with Generalized Additive Models (GAM).
80 Finally, we performed a gap analysis (Jennings, 2000, Mannocci et al., 2018) to assess the
81 reliability of the predictions outside the surveyed area.

82

83 **2. Methods**

84 **2.1. Data origin**

85 The study area encompassed the North Atlantic Ocean and the Mediterranean Sea from
86 the Guiana Plateau to Iceland, excluding the Baltic and Black Seas, the Gulf of Mexico and
87 the Hudson Bay, both because of an absence of effort data and of ecological and
88 environmental differences (Fig. 1A; Appendix S1 in Supporting Information). Four sub-
89 regions were defined in the study area (Table 1; Fig. 1A): the northeast Atlantic Ocean (NE-
90 ATL), the northwest Atlantic Ocean (NW-ATL), the tropics and the Mediterranean Sea
91 (MED).

92 We assembled visual shipboard and aerial surveys performed by 15 independent
93 organisations in the North Atlantic Ocean and the Mediterranean Sea between 1998 and
94 2015 (Fig. 1; survey-specific information are detailed in Appendix S2 in Supporting
95 Information). Except for the JNCC-ESAS surveys that use a 300m-strip-transect
96 methodology, all surveys used line-transect methodologies that correct for non-detection bias
97 with the estimation of an Effective Strip Width (ESW) from the measurement of the
98 perpendicular distances to the sightings (Buckland et al., 2015; see below).

99 To account for the difficulty in identifying deep-diving cetaceans to the species level (e.g.
100 genera *Mesoplodon*, *Kogia*), we pooled species into three groups: (1) beaked whales,
101 consisting of Cuvier's beaked whales (*Ziphius cavirostris*), mesoplodonts (*Mesoplodon* spp.)
102 and northern bottlenose whales (*Hyperoodon ampullatus*), (2) sperm whales (*Physeter*

103 *macrocephalus*), and (3) kogiids, including pygmy (*Kogia breviceps*) and dwarf sperm whales
104 (*K. sima*).

105

106 **2.3. Data processing**

107 *2.3.1. Data-assembling*

108 All survey datasets were standardised for units and formats (e.g. date, time and
109 coordinates) and aggregated into a single common dataset. A specific coordinate projection
110 encompassing the entire survey area was used for accurate distance computations (Albers
111 equal-area conic defined from <http://projectionwizard.org>). Effort data were linearized and
112 divided into 5 km segments using ArcGIS 10.3 (ESRI, 2016) and the Marine Geospatial
113 Ecology Tools software (Roberts et al., 2010). The segment length represented a trade-off
114 value across varying survey transect lengths, for example aerial surveys had transect lengths
115 of up to 100 km while shipboard surveys were often much shorter. Finally, for each species
116 group, sightings were linked to their respective 5 km segments.

117 Encounter rates were calculated in each sub-region as: (*number of encounters /*
118 *total distance travelled*) * 100.

119

120 *2.3.2. Environmental variables*

121 In habitat models, we tested the static and dynamic variables that were expected to
122 influence the distributions of deep divers (Table 2). All variables were resampled at a 0.25°
123 resolution because of the very large size of the study area and the spatial resolution of the
124 variables (Table 2; Appendix S3 in Supporting Information). Spatial gradients of sea surface
125 temperature (SST) were calculated as the difference between the minimum and maximum
126 SST values in an eight-pixel buffer around a given pixel. Net primary production (NPP) was
127 used as a proxy for prey availability.

128 Dynamic variables, which relate to the movements of water masses or prey availability,
129 were computed at a monthly resolution *i.e.* averaged over the 29 days prior to each sampled
130 day to avoid gaps in remote sensing oceanographic variables. They were used in addition to
131 static variables because they reveal the presence of time-stable structures such as
132 temperature gradients or eddies when variables are averaged.

133

134 *2.3.3. Effective Strip Width estimation*

135 Line-transect surveys are commonly used to estimate cetacean abundance (Hammond
136 et al., 2013; Buckland et al., 2015). A key parameter to estimate this abundance is the
137 effective strip width (ESW) which corrects the decreasing detection of animals with distance

138 from the trackline. ESW is expected to depend on survey platform height, platform type, sea-
139 state, species, etc... (Buckland et al., 2015).

140 ESW estimation was a key step in the data-assembling process to take into account
141 heterogeneity in effort per segment in the models and to directly compare the different
142 surveys (Hedley & Buckland, 2004). ESWs are generally estimated for each survey (*i.e.* no
143 pooling of information) by using the 'Distance' software (Thomas et al., 2010; Buckland et al.,
144 2015). However, for deep-diving cetaceans, the majority of surveys contained insufficient
145 sightings to allow survey-specific detection functions to be fitted. Consequently, for each
146 species group, we pooled sightings from the various surveys, taking into account survey
147 heterogeneity. We built a hierarchical model in which survey identity was included as a
148 random effect.

149 In conventional distance sampling (Marques & Buckland, 2003; Buckland et al., 2015),
150 factors such as the characteristics of the species being surveyed, search methods, search
151 platform, environmental conditions can all affect ESW estimation. However, the different
152 datasets did not always contain this information, especially regarding observation conditions.
153 All surveys recorded environmental data such as Beaufort sea-state, cloud coverage and sun
154 glare, although Beaufort sea-state was the only parameter recorded by all of them. Platform
155 type, observation height and Beaufort sea-state were used as covariates in the hierarchical
156 model.

157 Truncation distance w was first determined as the 95th percentile of the set of
158 perpendicular distances for each species group, *i.e.* the 5% most distant sightings were
159 discarded from the analysis (Buckland et al., 2001, page 16). Then, we created classes to
160 pool the different surveys; namely platform type (plane or boat), observation height (*e.g.* 0-5
161 m; 5-10 m...) and Beaufort sea-state (0-1; 1-2; 2-3 and 3-4; data collected beyond a Beaufort
162 sea-state 4 being removed from the analysis). Hierarchical modelling was then performed in
163 R-3.3.1 (R Core Team, 2016) in a Bayesian framework using JAGS version 4-6 and package
164 'rjags' (jags model in Appendix S4 in Supporting Information; Royle & Dorazio, 2008;
165 Plummer, 2016).

166 For each taxa, perpendicular distances of sightings were used to estimate a detection
167 function with a hazard key. For a sighting i made during survey s at height j under Beaufort
168 sea-state k , let d_{jks}^i denotes the perpendicular distance. The detection probability of sighting
169 i is:

$$\begin{cases} p_{ijk}^s = g_s(d_{ijk}) = 1 - \exp\left(-\left(\frac{d_{ijk}}{\sigma_{jks}}\right)^{-\nu_s}\right) \\ \log(\sigma_{jks}) = \beta_{j0} + \beta_{j1} \times k + \alpha_s \end{cases}$$

170 where β_{j0} and β_{j1} are respectively random intercept and slope parameters for the effect of
171 platform height; and α_s and ν_s are survey random effects. Bivariate random effects were

172 specified with a Cholesky decomposition and using priors for the Cholesky factors from
173 Kinney & Dunson (2008). We used half Student-t distributions with 3 degrees of freedom and
174 scale set to 1.5 as priors for dispersion parameters, and standard normal priors for all other
175 parameters. Four chains were run with a warmup of 10,000 iterations, followed by another
176 10,000 iterations (with a thinning factor of 10). Parameter convergence was assessed with
177 Gelman-Rubin \hat{R} statistics. Posterior inferences were based on the pooled sample of 4,000
178 values (1,000 per chain).

179 The advantage of setting a hierarchical model to estimate detection functions is to
180 borrow strength across the different datasets to increase the precision of estimates. For each
181 combination of survey – platform type – observation height – Beaufort sea-state, estimated
182 detection functions are shrunk towards a common detection function (itself estimated from
183 the data) according to the available data corresponding to this particular combination of
184 survey – platform type – observation height – Beaufort sea-state. If, for a given combination
185 of parameters, there were few sightings, the estimated detection function was very close to
186 the common detection function, whereas if there were enough data, the estimated detection
187 function could deviate from this common function. Upon model fitting and successful
188 parameter estimation, the ESW for each combination of survey – platform type – observation
189 height – Beaufort sea-state was computed:

$$\text{ESW}_{jks} = \int_0^w g_s(x) dx = \int_0^w \left[1 - \exp \left(- \left(\frac{x}{e^{\beta_{j0} + \beta_{j1} \times k + \alpha_s}} \right)^{-v_s} \right) \right] dx$$

190 The posterior mean value of estimated ESW was then allocated to each segment with
191 respect to species group, survey, platform type, sea-state and observation height class.

192

193 2.4. Habitat modelling

194 To model habitat preferences of deep-divers, we fitted Generalised Additive Models
195 (GAMs; Hastie & Tibshirani, 1986; Wood, 2006) with a Tweedie distribution to account for
196 over-dispersion (Foster & Bravington, 2013) with the 'mgcv' R-package (R-3.3.1. version;
197 Wood, 2013). GAMs extend Generalised Linear Models to allow for smooth nonlinear
198 functions of predictor variables (Hastie & Tibshirani, 1986; Wood, 2006). The mean number
199 of individuals per segment μ was modelled with a logarithmic link function:

$$\log(\mu) = \alpha + \sum_p f(X_p)$$

200 where $f(X_p)$ are non-parametric smooth functions (thin plate regression splines) of the
201 covariates and α is the intercept (Hastie & Tibshirani, 1986). To attenuate the scope for
202 over-fitting, the maximum number of knots was limited to 4 (mgcv parameter $k = 4$; Wood,
203 2006). An offset equal to segment length multiplied by twice the ESW was included (except
204 for the JNCC-ESAS surveys in which only one side of the vessel was surveyed). We

205 removed combinations of variables with Spearman partial correlation coefficients higher than
206 |0.7| (Dormann et al., 2013; Mannocci et al., 2014) and tested all models with combinations
207 of one to four variables. A maximum of four covariates per model was used to avoid
208 excessive complexity of models and difficulty in their interpretation (Mannocci et al., 2014).
209 Model selection was done with the Akaike Information Criterion (AIC, the lower the better;
210 Anderson & Burnham, 2002) and Akaike model weight (`akaike.weights` function from 'qpcR'
211 package; Spiess, 2014).

212 A key assumption of line-transect surveys is that animals on the trackline are always
213 detected (Buckland et al., 2001). However, this assumption is not met with diving species
214 and trackline detection probability $g(0)$ needs to be accounted for (Barlow, 2015). Observers
215 on a plane spend less time in a given area and the following inequality is expected:
216 $g^{\text{boat}}(0) > g^{\text{plane}}(0)$. Thus a segment of effort with zero sighting of deep-divers is more likely
217 to be a false absence (non-detection of a diving animal present on the trackline) if that
218 segment comes from a plane survey rather than a boat survey. As detection probability $g(0)$
219 was not available for every survey and is expected to differ between platforms, we calculated
220 the ratio of $g(0)$ between the plane and boat platforms from Roberts et al. (2016) and
221 obtained a ratio of approx. 1/5 for beaked whales, approx. 2/5 for sperm whales and approx.
222 1/3 for kogiids. These crude ratios were then used to weight plane segments with zero
223 sightings when fitting GAMs. While this method does not fully correct for availability bias, it
224 down-weights zeroes from plane surveys.

225 We fitted "year-round" models as the studied taxa have been reported to show little or no
226 seasonal variation in their habitats (e.g. Wimmer & Whitehead, 2004; McSweeney et al.,
227 2007). We did not model yearly variations because of little temporal overlap between
228 surveys. Consequently, the year effect is confounded with survey heterogeneity.

229 Predictions of relative densities (in number of animals per km²) were provided at 0.25°
230 resolution. There was not enough data to fit a model by month or by season (the number of
231 sightings in winter was too low) and we therefore produced averaged maps over the entire
232 time period. These predictive maps provided the expected distribution of beaked whales,
233 sperm whales and kogiids according to static and monthly environmental conditions to
234 highlight relationships with static (canyons and seamounts) and time-stable structures
235 (temperature gradients or eddies).

236 Finally, coefficients of variation (CVs) were estimated for each 0.25° pixel. Coefficients of
237 variations are a measure of the prediction uncertainty per cell, it is a standard error
238 associated with the calculation of the prediction. Therefore, high CVs indicate high model
239 uncertainties due to the lack of detection.

240

241 **2.5. Gap analysis**

242 Even though more than 1,240,000 km of effort was pooled, extensive geographical gaps
243 remained. Predictions in the middle of the Atlantic Ocean are from geographical extrapolation
244 (Fig. 1A) but not necessarily environmental extrapolations. The latter depends on the
245 selected habitat models and covariates therein. We conducted a gap analysis on
246 environmental space coverage to identify areas where habitat models could produce reliable
247 predictions outside survey blocks, *i.e.* geographical extrapolation, whilst remaining within the
248 ranges of surveyed conditions for the combinations of covariates selected by the models, *i.e.*
249 areas of environmental interpolation (Jennings, 2000; Mannocci et al., 2018).

250 From the selected models for each taxa, we estimated the convex hull defined by the
251 environmental data used to fit habitat models (hereafter the calibration data). The convex
252 hull of a set of points is the smallest convex envelop that contains all these points. We then
253 assessed whether a prediction from a set of environmental covariates with a given model fall
254 inside or outside this convex hull (King & Zeng, 2007; Authier et al., 2016). We used
255 climatological predictors instead of monthly predictors to lessen the computational burden.

256 Due to the large number of data (more than 280,000 points in the calibration dataset),
257 convex hulls were estimated by random sub-sampling with the 'WhatIf' R-package (Stoll et
258 al., 2014). We randomly extracted a fraction of the calibration dataset (10,000 points) to
259 estimate a convex hull and assess environmental extrapolation in the prediction dataset. A
260 combination of climatological predictor values that fall inside the convex hull corresponds to
261 an interpolation. Combinations of climatological predictor values that were classified as
262 interpolations were set aside but other combinations were retained and further tested against
263 another random sample of 10,000 points from the calibration data. This procedure was
264 carried out until the complete calibration dataset was examined.

265 The full procedure was conducted twice. In a simple approach, the full range of sampled
266 variables was considered to identify all points of the whole study area where the actual
267 combinations of environmental variables had been sampled in survey blocks. In a more
268 'precautionary approach', we excluded 5% of the extreme values of the sampled
269 environmental variables to include in the interpolation areas only the points whose
270 associated combinations of covariates fell within 95% of the core ranges sampled. This
271 allowed the definition of two levels of confidence (hereafter 'simple' and 'precautionary') in
272 the predictions.

273 Finally, we produced maps delineating the extent of the simple and precautionary
274 interpolation areas and overlaid them with the relative density prediction maps to show areas
275 with greater reliability.

276

277

278 **3. Results**

279 **3.1. Encounter rates**

280 The survey pool represented a total of 1,240,000 km of on-effort transects (*i.e.* following
281 a transect at a specified speed and altitude with a specified level of visual effort) of which
282 58% were carried out by plane and 42% by boat (Fig. 1A, Table 1). Effort data with a
283 Beaufort sea-state higher than 4, which represented 9% of the effort data, were removed
284 from further analysis to only keep sightings collected during good to excellent detection
285 conditions. Most sampling effort was performed in the northeast (37 %) and northwest (45 %)
286 Atlantic Ocean. Surveys in the Mediterranean Sea and in the tropics represented
287 respectively only 16 % and 2 % of total sampling effort.

288 A total of 630 sightings of beaked whales, 836 sightings of sperm whales and 106
289 sightings of kogiids, mainly distributed in the northeast and northwest Atlantic Ocean (north
290 of the 35°N latitude) and in the northwest Mediterranean Sea, were assembled for the
291 present study (Fig. 1B-D).

292 Overall encounter rates were very low with 0.05 sightings·100 km⁻¹ for beaked whales,
293 0.07 sightings·100 km⁻¹ for sperm whales and <0.01 sightings·100 km⁻¹ for kogiids (Table 3).
294 The highest encounter rates were recorded in the tropics for all three species groups,
295 particularly for kogiids. There were no sightings of kogiids in the Mediterranean Sea.

297 **3.2. Effective strip width**

298 Estimated ESWs varied across surveys and platform type and were on average
299 narrower in aerial than shipboard surveys (Fig. 2). This is probably because aerial observers
300 are more restricted to recording animals below the plane while shipboard observers can look
301 further afield. ESWs were generally larger and more consistent between surveys using the
302 same platform type, for sperm whales than for beaked whales. There were not enough kogiid
303 sightings to estimate an ESW for each survey and particularly for shipboard surveys;
304 consequently, we pooled all aerial surveys and estimated an ESW of 1.1 km that was then
305 applied to all surveys (shipboard and aerial).

306 The outcomes from the hierarchical model were consistent with expectations (Fig. S4.1
307 in Supporting Information S4): a decrease in Beaufort sea-state (less wind-sea) resulted in a
308 larger ESW (milder non-detection bias).

310 **3.3. Habitat modelling**

311 For each species group, selected variables, explained deviances and Akaike weights are
312 shown in Table 4.

314 **Beaked whales**

315 Highest relative densities were found in depth *ca.* 1,500 m, high values of slopes and
316 SST and intermediate NPP. This resulted in high predicted relative densities of beaked
317 whales along steep slope areas associated with deep depths and high gradients of
318 temperature, particularly on the western side of the Atlantic Ocean. The lowest relative
319 densities were predicted in the Mediterranean Sea (Fig. 3B).

320 The gap analysis identified areas where the combination of the four variables selected
321 by the best model had not been sampled. Reliable predictions were available for 94% of the
322 study area under the simple approach and only 53% under the precautionary approach (Fig.
323 3B and 3C). This discrepancy was mostly due to low sampling effort in the oceanic zone.
324 Coefficients of (temporal) variation were higher on the continental shelf associated with high
325 gradients of SST, where beaked whales were not sighted in any of the surveys (Fig. S5.2A in
326 Supporting Information S5).

327

328 **Sperm whales**

329 Predicted relative densities of sperm whales increased in deep waters (> 2000 m)
330 associated with high gradients of SST and high NPP. The highest relative densities were
331 also predicted on the western side of the Atlantic Ocean, along the Gulf Stream, although
332 were lowest in the Mediterranean Sea (Fig. 4B).

333 Reliable predictions for sperm whales were available for 84% of the study area under the
334 simple approach and only 30% under the precautionary approach because of low survey
335 effort in deeper areas. The highest predicted relative densities were predicted outside the
336 precautionary interpolation zone (Fig. 4B and 4C). Coefficients of (temporal) variation were
337 highest in non-sampled areas where uncertainty was therefore greatest (Fig. S5.2B in
338 Supporting Information S5).

339

340 **Kogiids**

341 As the Akaike weight was small for kogiids (0.17), we used model-averaging and
342 generated predictions from the five first models (cumulative Akaike weight of 0.63) and
343 because all predictions were very similar (see Appendix S6 in Supporting Information), we
344 only kept the first model for practical reasons. The highest relative densities were found in
345 deep waters associated with fronts, canyons and seamounts (Fig. 5B). The highest relative
346 densities were predicted on the western side of the Atlantic Ocean, along the Gulf Stream
347 (Fig. 5C).

348 Reliable predictions for kogiids were available for 94% of the study area under the
349 simple approach against only 55% under the precautionary approach because of low survey
350 effort in deeper areas (Fig. 5C). Coefficients of (temporal) variation were highest in shallow

351 waters and in the Mediterranean Sea where kogiids were not sighted in any of the surveys
352 (Fig. S5.2C in Supporting Information S5).

353

354 **4. Discussion**

355 Deep-diving cetaceans are species characterised by low sighting rates and modelling
356 their habitats is particularly challenging. Our study pooled different surveys allowing us to
357 capitalise on more than 1,240,000 km of survey effort deployed over the North Atlantic
358 Ocean and the Mediterranean Sea in the past two decades. For each taxon, we built a
359 hierarchical model to estimate the effective strip width depending on observation conditions
360 and surveys. We investigated habitats of deep-divers using GAMs with a focus on
361 quantifying how reliable the predictions were. The selected habitat models of deep-diving
362 cetaceans included static environmental variables such as depth and slope as well as spatial
363 gradients of temperatures, revealing the highest densities in the western North Atlantic
364 Ocean. Deeper areas of the North Atlantic gyre were mostly areas of environmental
365 extrapolation, thereby highlighting gaps in sampling across the different surveys.

366

367 **4.1. Methodological considerations**

368 Over the past few years, data-assembling has been increasingly used for the study of
369 top marine predators (Roberts et al., 2016; Rogan et al., 2017; Cañadas et al., 2018). Due to
370 the very low sighting rates of deep-diving cetaceans, each survey taken separately cannot
371 provide enough data to investigate the habitats of these rare species. In contrast to Rogan et
372 al. (2017), we did not assemble data collected with similar protocols but data collected with
373 different variants of the line transect distance sampling protocol which meant standardising
374 the data according to their core communalities before developing a single spatial model.
375 Ideally, at a time when shared databases are becoming increasingly important (*e.g.* OBIS
376 SEAMAP -- <http://seamap.env.duke.edu/>, EMODnet -- <http://www.emodnet.eu/>),
377 implementing standardised survey methods would greatly improve data compatibility, by
378 enhancing the level of communalities in shared datasets, and helping to describe large-scale
379 habitats and distributions of marine species. However, we realise this can lead to financial
380 and logistical constraints and the work we present here could be a way to embrace and
381 incorporate the diversity of data collection methods.

382 Hierarchical modelling accommodates heterogeneity between surveys; it borrows
383 strength across surveys ('partial pooling') when estimating survey-specific ESWs. The
384 resulting estimates are biased (in proportion to the available data contributed by each
385 survey) toward a common mean, although are more precise than those that would be
386 obtained if each survey was analysed separately ('no pooling' scenario) as it is usually done

387 when the number of sightings per survey is large (Buckland et al., 2015; Laran et al., 2017;
388 Redfern et al., 2017). Results from the hierarchical model were consistent with expectations
389 and showed that a decrease in Beaufort sea-state values resulted in increased ESW
390 estimates.

391 The majority of environmental variables we used in habitat modelling describe the
392 euphotic zone (upper layer) because variables that describe the deep-water column are
393 difficult to obtain or simply do not exist at a basin-wide scale. As deep-diving cetaceans
394 spend most of their time at depth and generally feed on mesopelagic to bathypelagic prey
395 (e.g. Perrin et al., 2009; Spitz et al., 2011), the use of surface variables limits the ability to
396 correctly infer their habitat. The identified relationships between deep-diving cetacean
397 abundance and environmental variables may be indirect rather than causal (Austin, 2006).
398 Although causation may be out of reach, prediction remains a worthy goal, especially for
399 spatial planning and conservation (McShea, 2014).

400 We took care in using appropriate statistical tools for modelling the habitat of species
401 with few sightings (Virgili et al., 2018). Indeed, Virgili et al. (2018) showed that GAMs with a
402 Tweedie distribution generated reliable habitat modelling predictions for rarely sighted marine
403 predators. Here, the habitat models we selected had moderate to high levels of explained
404 deviances (from 20.6% to 55.7%), suggestive of a good fit to the data. Nevertheless, the
405 rather high explained deviance of the kogiid model (55.7%) might indicate some level of
406 model over-fitting due to the small dataset, even if predictions were in general consistent with
407 the known ecology of the species group (McAlpine, 2009).

408

409 **4.2. Large-scale deep-diver habitats**

410 Depth and spatial gradients of sea surface temperature were consistently selected
411 across deep-diving cetaceans, suggesting a major influence of topographic features and
412 thermal fronts in structuring their habitats. As a result, higher relative densities of deep-divers
413 were predicted in areas of strong gradients associated with thermal fronts in which deep-
414 diver prey aggregates (Bost et al., 2009; Woodson & Litvin, 2015). Indeed, deep-divers
415 typically feed on mesopelagic to bathypelagic species, such as pelagic cephalopods and
416 benthic fishes (Spitz et al., 2011) that aggregate along continental slopes where temperature
417 gradients are the strongest. Hence, the Gulf Stream, which is the most active frontal zone in
418 the study area compared to the eastern boundary currents that are broader and much
419 slower, may explain the high predicted relative densities of deep-divers on the western side
420 of the North Atlantic Ocean (Waring et al., 2001; Roberts et al., 2016).

421 Despite commonalities, each studied taxon also showed specificities. Slope appeared to
422 be an important predictor of beaked whale relative density. The prey targeted by beaked

423 whales are more specific than those of sperm whales, which have broader prey size
424 spectrum (Spitz et al., 2011), and their distribution is more driven by dynamic variables than
425 by static features. Accordingly, the selected model for sperm whales included more dynamic
426 variables such as NPP and SSH than for beaked whales. Canyons and seamounts were
427 included in the selected model for kogiids, suggesting a more restricted habitat than for the
428 other two groups of deep-divers conforming Staudinger et al.'s (2014) evidence of how
429 kogiids' feeding areas concentrated on the deeper shelf and slope, particularly in the
430 epipelagic and mesopelagic zones.

431 Overall, our model predictions corroborated species distribution predictions of previous
432 smaller-scale studies. In the Mediterranean Sea, our predictions were consistent with the
433 documented presence of beaked whales and sperm whales in the Alborán, Tyrrhenian and
434 Ligurian Seas (Praca & Gannier, 2008; Arcangeli et al., 2015; Lanfredi et al. 2016; Cañadas
435 et al., 2018) and along the eastern coasts of the Mediterranean Sea (Podestà et al., 2006). In
436 the North Atlantic Ocean, the highest relative densities of beaked whales and sperm whales
437 were predicted along the slope, a result consistent with those of Rogan et al.'s (2017) and
438 Roberts et al.'s (2016). In the Northwest Atlantic Ocean, higher relative densities of kogiids
439 were predicted in warmer and deeper waters, which is consistent with their known ecology
440 (McAlpine, 2009) and the predictions of Mannocci et al. (2017) except for predictions off the
441 coast of Florida. Our predictions could probably be improved by incorporating the NOAA
442 SEFSC surveys of southeast US waters off Florida and Virginia. In contrast to beaked and
443 sperm whales, we were not able to fit a hierarchical model on kogiid sightings and resorted to
444 complete pooling of the plane data to estimate an ESW. This shortcoming probably resulted
445 in a larger bias (with respect to the true density) in predicted relative density of kogiids
446 compared to other deep-diving species. Given the paucity of information on kogiids, we think
447 that our results are tentative but important nonetheless.

448 The gap analysis revealed large gaps in environmental space coverage across the study
449 area, especially in the deeper and less productive waters of the central north Atlantic gyre
450 and in tropical waters. High relative densities of deep-divers were predicted at the margin of
451 the precautionary interpolation zone (Figs. 3-5) in particular because deeper waters and
452 steeper slopes were within the upper 2.5% quantiles of aggregated survey coverage for
453 these two physiographic covariates. This suggests that sampling effort was not sufficient in
454 deeper and steeper areas and more intensive sampling effort performed in these areas could
455 help better describe the habitat used by deep-divers.

456

457 **4.3. Management considerations**

458 The management and conservation of species and ecosystems increasingly relies on
459 habitat models (McShea, 2014; Hazen et al., 2016). The ability of these to predict species
460 occurrence in non-sampled or poorly documented areas is useful (Fleishman et al., 2001;
461 Lumaret & Jay-Robert, 2002) because the implementation of dedicated surveys is
462 sometimes impracticable due to budgetary and logistical challenges. It is logistically
463 challenging to carry out dedicated cetacean surveys in the middle of the North Atlantic
464 Ocean. However, by collecting data on both sides of the Atlantic Ocean, relative density
465 maps were produced and our analyses indicated these predictions may be reliable (Figs. 3D,
466 4D, 5D).

467 Here, we showed that deep-diving cetaceans are closely associated with stable
468 topographic features, thus it could be possible to delineate marine protected areas that cover
469 the principal habitats used by the species (*e.g.* Cañadas et al., 2005). However, these
470 species are also responsive to temporally dynamic structures, such as thermal fronts,
471 implying that protected areas will need to be large enough to capture seasonal variation of
472 such features. In this context, Important Marine Mammal Areas, which are currently being
473 discussed by the Marine Mammal Protected Areas Task Force and incorporate governmental
474 and intergovernmental considerations (Corrigan et al., 2014), could help the delineation of
475 sufficiently large protected areas. In addition, in a Marine Spatial Planning approach
476 (Douvere, 2008), it would be worthwhile to overlay predicted density maps with
477 anthropogenic pressure maps (Halpern et al., 2008) to define areas where pressures could
478 be mitigated.

479

480 **5. Conclusion**

481 Habitat modelling of rare species is particularly challenging because habitat models
482 require large datasets, yet rare species typically yield low numbers of sightings. As a result,
483 combining datasets is a useful strategy to model the large-scale habitats of deep-divers;
484 beaked whales, sperm whales and kogiids, across the North Atlantic Ocean and the
485 Mediterranean Sea. At a local scale, predicted relative densities of deep-diving cetaceans
486 were consistent with previous studies. At a larger scale, a gradient in predicted relative
487 densities emerged, with the highest relative densities predicted on the western side of the
488 study area. This pattern was evidenced thanks to assembling a large dataset and had not
489 been detected previously. It highlighted the pronounced influence of active frontal zones,
490 such as the Gulf Stream, on deep-diving cetaceans. Even though extensive gaps remain at a
491 large scale, we were able to predict the habitats of these taxa throughout the North Atlantic
492 Ocean and adjacent Mediterranean Sea, thus identifying potential habitats, including in non-
493 sampled areas. However, these predictions should be used with caution as most of the study

494 area represented geographical extrapolations and about half (mostly deeper waters)
495 represented environmental extrapolations. Indeed, through an environmental space
496 coverage gap analysis, we identified areas in tropical and deep oceanic waters where
497 sampling effort was insufficient to predict habitats and needs to be increased to improve
498 prediction reliability.

499 References

500 Anderson, D.R., & Burnham, K.P. (2002). Avoiding pitfalls when using information-theoretic
501 methods. *The Journal of Wildlife Management* 66(3): 912–918.

502 Arcangeli, A., Campana, I., Marini, L., & MacLeod, C.D. (2015). Long-term presence and habitat use
503 of Cuvier's beaked whale (*Ziphius cavirostris*) in the Central Tyrrhenian Sea. *Marine Ecology*
504 37: 269–282.

505 Austin, D., Bowen, W.D., McMillan, J.I., & Iverson, S.J. (2006). Linking Movement, Diving, and Habitat
506 to Foraging Success in a Large Marine Predator. *Ecology* 87: 3095–3108.

507 Authier, M., Saraux, C., & Péron, C. (2016). Variable selection and accurate predictions in habitat
508 modelling: a shrinkage approach. *Ecography* 40: 549–560.

509 Barlow, J. (2015). Inferring trackline detection probabilities, $g(0)$, for cetaceans from apparent
510 densities in different survey conditions. *Marine Mammal Science* 31(3): 923–943.

511 Bost, C.A., Cotté, C., Bailleul, F., Cherel, Y., Charrassin, J.B., Guinet, C., ... Weimerskirch, H. (2009).
512 The importance of oceanographic fronts to marine birds and mammals of the southern oceans.
513 *Journal of Marine Systems* 78: 363–376.

514 Buckland, S.T., Anderson, D.R., Burnham, H.P., Laake, J.L., Borchers, D.L. & Thomas, L. (2001).
515 Introduction to Distance Sampling: Estimating Abundance of Biological Populations. Oxford
516 University Press, Oxford.

517 Buckland, S.T., Rexstad, E.A., Marques, T. A., & Oedekoven, C.S. (2015). *Distance sampling:
518 methods and applications*. Springer.

519 Canada Meteorological Center. (2012). GHRSSST Level 4 CMC0.2deg Global Foundation Sea Surface
520 Temperature Analysis (GDS version 2). Ver. 2.0. PO.DAAC, CA, USA.

521 Cañadas, A., Sagarmínaga, R., De Stephanis, R., Urquiola, E., & Hammond, P.S. (2005). Habitat
522 preference modelling as a conservation tool: proposals for marine protected areas for cetaceans
523 in southern Spanish waters. *Aquatic Conservation: Marine and Freshwater Ecosystems* 15(5):
524 495–521.

525 Cañadas, A., de Soto, N.A., Aissi, M., Arcangeli, A., Azzolin, M., B-Nagy, A., ... Roger, Th. (2018).
526 The challenge of habitat modelling for threatened low density species using heterogeneous

- 527 data: The case of Cuvier's beaked whales in the Mediterranean. *Ecological Indicators* 85: 128–
528 136.
- 529 Carrillo, M., & Ritter, F. (2010). Increasing numbers of ship strikes in the Canary Islands: proposals for
530 immediate action to reduce risk of vessel-whale collisions. *Journal of Cetacean Research and*
531 *Management* 11(2): 131–138.
- 532 Corrigan, C.M., Ardron, J.A., Comeros-Raynal, M.T., Hoyt, E., Notarbartolo Di Sciara, G., & Carpenter,
533 K.E. (2014). Developing important marine mammal area criteria: learning from ecologically or
534 biologically significant areas and key biodiversity areas. *Aquatic Conservation: Marine and*
535 *Freshwater Ecosystems* 24(S2): 166–183.
- 536 D'Amico, A., Gisiner, R.C., Ketten, D.R., Hammock, J.A., Johnson, C., Tyack, P.L., & Mead, J. (2009).
537 Beaked whale strandings and naval exercises. *Aquatic Mammals* 35: 452–472.
- 538 Dormann, C.F., Elith, J., Bacher, S., Buchmann, C., Carl, G., Carré, G., ... Münkemüller, T. (2013).
539 Collinearity: a review of methods to deal with it and a simulation study evaluating their
540 performance. *Ecography* 36(1): 27–46.
- 541 Douvère, F. (2008). The importance of marine spatial planning in advancing ecosystem-based sea use
542 management. *Marine policy* 32(5): 762–771.
- 543 ESRI, (2016). ArcGIS - A Complete Integrated System Environmental Systems Research Institute,
544 Inc., Redlands, California. <<http://esri.com/arcgis>>.
- 545 Fernández, A., Edwards, J.F., Rodríguez, F., Espinosa de los Monteros, A., Herráez, P., Castro, P., ...
546 Arbelo, M. (2005). Gas and Fat Embolic Syndrome" Involving a Mass Stranding of Beaked
547 Whales (Family *Ziphiidae*) Exposed to Anthropogenic Sonar Signals. *Veterinary Pathology* 42:
548 446–457.
- 549 Fleishman, E., Nally, R.M., Fay, J.P., & Murphy, D.D. (2001). Modeling and predicting species
550 occurrence using broad-scale environmental variables: an example with butterflies of the Great
551 Basin. *Conservation biology* 15(6): 1674–1685.
- 552 Foster, S.D., & Bravington, M.V. (2013). A Poisson-Gamma model for analysis of ecological non-
553 negative continuous data. *Environmental and ecological statistics* 20: 533–552.
- 554 Hammond, P.S., Macleod, K., Berggren, P., Borchers, D.L., Burt, L., Cañadas, A., ... Vázquez, J.A.
555 (2013). Cetacean abundance and distribution in European Atlantic shelf waters to inform
556 conservation and management. *Biological Conservation* 164: 107–122.
- 557 Halpern, B.S., Walbridge, S., Selkoe, K.A., Kappel, C.V., Micheli, F., D'agrosa, C., ... Watson, R.
558 (2008). A global map of human impact on marine ecosystems. *Science* 319(5865): 948–952.
- 559 Harris, P.T., Macmillan-Lawler, M., Rupp, J., & Baker, E.K. (2014). Geomorphology of the oceans.
560 *Marine Geology* 352: 4–24.

- 561 Hastie, T., & Tibshirani, R. (1986). Generalized Additive Models. *Statistical Science* 3: 297–313.
- 562 Hazen, E.L., Palacios, D.M., Forney, K.A., Howell, E.A., Becker, E., Hoover, A.L., ... Bailey, H. (2017).
563 WhaleWatch: a dynamic management tool for predicting blue whale density in the California
564 Current. *Journal of Applied Ecology* 54(5): 1415–1428.
- 565 Hedley, S.L., & Buckland, S.T. (2004). Spatial models for line transect sampling. *Journal of*
566 *Agricultural, Biological, and Environmental Statistics* 9(2): 181–199.
- 567 Jennings, M.D. (2000). Gap analysis : concepts, methods, and recent results. *Landscape Ecology* 15:
568 5–20.
- 569 Kinney, S.K., & Dunson, D.B. (2008). Bayesian model uncertainty in mixed effects models. *Random*
570 *effect and latent variable model selection*. Springer New York. pp. 37-62.
- 571 King, G., & Zeng, L. (2007). When Can History be Our Guide? The Pitfalls of Counterfactual
572 Inference. *International Studies Quarterly* 51: 183–210.
- 573 Lanfredi, C., Azzellino, A., D'Amico, A., Centurioni, L., Ampolo Rella, M., Pavan, G., & Podestà, M.
574 (2016). Key Oceanographic Characteristics of Cuvier's Beaked Whale (*Ziphius cavirostris*)
575 Habitat in the Gulf of Genoa (Ligurian Sea, NW Mediterranean). *Journal of Oceanography and*
576 *Marine Research* 4: 145.
- 577 Laran, S., Authier, M., Van Canneyt, O., Dorémus, G., Watremez, P., & Ridoux, V. (2017). A
578 comprehensive survey of pelagic megafauna: their distribution, densities and taxonomic
579 richness in the tropical Southwest Indian Ocean. *Frontiers in Marine Science* 4: 139.
- 580 Lumaret, J.P., & Jay-Robert, P. (2002). Modelling the species richness distribution of French dung
581 beetles (*Coleoptera, Scarabaeidae*) and delimiting the predictive capacity of different groups of
582 explanatory variables. *Global Ecology and Biogeography* 11(4): 265–277.
- 583 Madsen, P.T., de Soto, N.A., Tyack, P.L., & Johnson, M. (2014). Beaked whales. *Current*
584 *Biology* 24(16): 728–730.
- 585 Mannocci, L., Catalogna, M., Dorémus, G., Laran, S., Lehodey, P., Massart, W., ...& Ridoux, V.
586 (2014). Predicting cetacean and seabird habitats across a productivity gradient in the South
587 Pacific gyre. *Progress in Oceanography* 120: 383–398.
- 588 Mannocci, L., Roberts, J.J., Miller, D.L., & Halpin, P.N. (2017). Extrapolating cetacean densities to
589 quantitatively assess human impacts on populations in the high seas. *Conservation Biology*. 31:
590 601–614.
- 591 Mannocci, L., Roberts, J.J., Halpin, P.N., Authier, M., Boisseau, O., Bradai, M.N., ... Vella, J. (2018).
592 Assessing cetacean surveys throughout the Mediterranean Sea: a gap analysis in
593 environmental space. *Scientific Reports* 8(1): 3126.

- 594 Marques, F.F., & Buckland, S.T. (2003). Incorporating covariates into standard line transect
595 analyses. *Biometrics* 59(4): 924-935.
- 596 McAlpine, D.F. (2009). *Pygmy and dwarf sperm whales*. Encyclopedia of marine mammals 2nd
597 Edition. pp 936–938. Academic Press.
- 598 McShea, W.J. (2014). What Are the Roles of Species Distribution Models in Conservation Planning?
599 *Environmental Conservation* 41: 93–96.
- 600 McSweeney, D.J., Baird, R.W., & Mahaffy, S.D. (2007). Site fidelity, associations, and movements of
601 Cuvier's (*Ziphius cavirostris*) and Blainville's (*Mesoplodon densirostris*) beaked whales off the
602 island of Hawai'i. *Marine Mammal Science* 23: 666–687.
- 603 Perrin, W.F., Würsig, B., & Thewissen, J.G.M. (Eds.). (2009). *Encyclopedia of marine mammals*.
604 Academic Press.
- 605 Podestà, M., D'Amico, A., Pavan, G., Drougas, A., Komnenou, A., & Portunato, N. (2005). A review of
606 Cuvier's beaked whale strandings in the Mediterranean Sea. *Journal of Cetacean Research and*
607 *Management* 7(3): 251–261.
- 608 Plummer, M. (2016). rjags: Bayesian Graphical Models using MCMC. R package version 4-6.
609 <https://CRAN.R-project.org/package=rjags>
- 610 Praca, E., & Gannier, A. (2007). Ecological niche of three teuthophageous odontocetes in the
611 northwestern Mediterranean Sea. *Ocean Science Discussions* 4: 49–59.
- 612 R Core Team (2016). R: A language and environment for statistical computing. R Foundation for
613 Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>.
- 614 Redfern, J.V., Moore, T.J., Fiedler, P.C., de Vos, A., Brownell, R.L., Forney, K.A., ... Ballance, L.T.
615 (2017). Predicting cetacean distributions in data-poor marine ecosystems. *Diversity and*
616 *Distributions* 23: 394–408.
- 617 Roberts, J.J., Best, B.D., Dunn, D.C., Tremblay, E.A., & Halpin, P.N. (2010). Marine Geospatial Ecology
618 Tools: An integrated framework for ecological geoprocessing with ArcGIS, Python, R, MATLAB,
619 and C++. *Environmental Modelling & Software* 25: 1197–1207.
- 620 Roberts, J.J., Best, B.D., Mannocci, L., Fujioka, E., Halpin, P.N., Palka, D.L., ... Lockhart, G.G. (2016).
621 Habitat-based cetacean density models for the U.S. Atlantic and Gulf of Mexico. *Scientific*
622 *Report* 6.
- 623 Rogan, E., Cañadas, A., Macleod, K., Santos, M.B., Mikkelsen, B., Uriarte, A., ... Hammond, P.S.
624 (2017). Distribution abundance and habitat use of deep diving cetaceans in the North-East
625 Atlantic. *Deep Sea Research Part II: Topical Studies in Oceanography* 141: 8–19.
- 626 Royle, J.A., & Dorazio, R.M. (2008). *Hierarchical modeling and inference in ecology: the analysis of*
627 *data from populations, metapopulations and communities*. Academic Press.

- 628 Spiess, A. (2014). qpcR: Modelling and analysis of real-time PCR data. R package version 1.4-0.
629 <https://CRAN.R-project.org/package=qpcR>
- 630 Spitz, J., Cherel, Y., Bertin, S., Kiszka, J., Dewez, A., & Ridoux, V. (2011). Prey preferences among
631 the community of deep-diving odontocetes from the Bay of Biscay, Northeast Atlantic. *Deep Sea*
632 *Research* 158: 273–282.
- 633 Staudinger, M.D., McAlarney, R.J., McLellan, W.A., & Ann Pabst, D. (2014). Foraging ecology and
634 niche overlap in pygmy (*Kogia breviceps*) and dwarf (*Kogia sima*) sperm whales from waters of
635 the US mid-Atlantic coast. *Marine Mammal Science* 30(2): 626–655.
- 636 Stoll, H., King, G., & Zeng, L. (2014). WhatIf: Software for Evaluating Counterfactuals. R package
637 version 1.5-6. <https://cran.r-project.org/web/packages/WhatIf/index.html>.
- 638 Stone, C.J., & Tasker, M.L. (2006). The effects of seismic airguns on cetaceans in UK waters. *Journal*
639 *of Cetacean Research and Management* 8: 255–263.
- 640 Thomas, L., Buckland, S.T., Rexstad, E.A., Laake, J.L., Strindberg, S., Hedley, S.L., ... Burnham, K.P.
641 (2010). Distance software: Design and analysis of distance sampling surveys for estimating
642 population size. *Journal of Applied Ecology* 47: 5–14.
- 643 Unger, B., Rebolledo, E.L.B., Deaville, R., Gröne, A., IJsseldijk, L.L., Leopold, M.F., ... Herr, H. (2016).
644 Large amounts of marine debris found in sperm whales stranded along the North Sea coast in
645 early 2016. *Marine pollution bulletin* 112(1): 134–141.
- 646 Virgili, A., Authier, M., Monestiez, P., & Ridoux, V. (2018). How many sightings to model rare marine
647 species distributions. *PLoS one* 13(3): e0193231.
- 648 Waring, G.T., Hamazaki, T., Sheehan, D., Wood, G., & Baker, S. (2001). Characterization of beaked
649 whale (*Ziphiidae*) and sperm whale (*Physeter macrocephalus*) summer habitat in shelf-edge
650 and deeper waters off the Northeast U.S. *Marine Mammal Science* 17: 703–717.
- 651 Wimmer, T., & Whitehead, H. (2004). Movements and distribution of northern bottlenose whales,
652 *Hyperoodon ampullatus*, on the Scotian Slope and in adjacent waters. *Canadian Journal of*
653 *Zoology* 82: 1782–1794.
- 654 Wood, S.N. (2006). On confidence intervals for generalized additive models based on penalized
655 regression splines. *Australian & New Zealand Journal of Statistics* 48: 445–464.
- 656 Wood, S. (2013). mgcv: Mixed GAM Computation Vehicle with GCV/AIC/REML smoothness
657 estimation. Retrieved 7 July 2014, from <http://cran.r-project.org/web/packages/mgcv/index.html>.
- 658 Woodson, C.B., & Litvin, S.Y. (2015). Ocean fronts drive marine fishery production and
659 biogeochemical cycling. *Proceedings of the National Academy of Sciences* 112(6): 1710–1715.

660

661 DATA ACCESSIBILITY

662 All sighting and effort data used in this study are available in the OBIS SEAMAP
 663 database: <http://seamap.env.duke.edu/>. All data providers can be contacted via the OBIS
 664 SEAMAP website.

665

666 **Tables**

667 **Table 1. Effort performed by platform type or Beaufort sea-state for all surveys in the North Atlantic**
 668 **Ocean and the Mediterranean Sea.** This table presents the total effort conducted in each sector broken down by
 669 platform type and Beaufort sea-state. Beaufort sea-state values reported with decimals in the surveys were
 670 rounded up. For the analyses, all segments with Beaufort sea-state > 4 were excluded. 'NE-ATL' means northeast
 671 Atlantic Ocean; 'NW-ATL' means northwest Atlantic Ocean and 'MED' means Mediterranean Sea.

Sectors	Total survey effort (km and %)	Total aerial effort (km)	Total shipboard effort (km)	Total effort by Beaufort sea-state class (km)				
				0-1	1-2	2-3	3-4	4-7
NE-ATL	469,000 37 %	70,000	399,000	77,000	118,000	136,000	85,000	53,000
NW-ATL	557,000 45 %	546,000	11,000	43,000	121,000	199,000	132,000	62,000
MED	195,000 16 %	87,000	109,000	92,000	70,000	27,000	6,000	800
TROPICS	19,000 2 %	15,000	4,000	11,000	3,000	4,000	2,000	400
STUDY AREA	1,240,000	718,000 58 %	522,000 42 %	222,000 18 %	312,000 25 %	365,000 30 %	225,000 18 %	116,000 9%

672

673

674

675

676

677

678

679

680

681

682

683

684 **Table 2. Candidate environmental predictors used for the habitat modelling.** All variables were resampled at
 685 a 0.25° resolution. A: Depth and slope were derived from GEBCO-08 30 arc-second database
 686 (<http://www.gebco.net/>); 30 arc-second is approximately equal to 0.008°. B: Surface area per cell was calculated
 687 in ArcGIS 10.3 from the shapefile of canyons and seamounts provided by Harris et al. (2014). C: The mean,
 688 standard error and gradient of Sea Surface Temperature (SST) were calculated from the GHRSSST Level 4 CMC
 689 SST v.2.0 (Canada Meteorological Centre, <https://podaac.jpl.nasa.gov/dataset/CMC0.2deg-CMC-L4-GLOB-v2.0>).

690 D: The Aviso ¼° DT-MADT geostrophic currents dataset was used to compute mean and standard deviation of
 691 Sea Surface Height (SSH) and Eddy Kinetic Energy (EKE; [https://www.aviso.altimetry.fr/en/data/products/sea-](https://www.aviso.altimetry.fr/en/data/products/sea-surface-height-products/global/madt-h-uv.html)
 692 [surface-height-products/global/madt-h-uv.html](https://www.aviso.altimetry.fr/en/data/products/sea-surface-height-products/global/madt-h-uv.html)). E: Net primary production (NPP) was derived from SeaWIFS and
 693 Aqua using the Vertically Generalised Production Model (VGPM; <http://orca.science.oregonstate.edu/1080.by.2160.8day.hdf.vgpm.m.chl.m.sst.php>).
 694
 695

Environmental variable	Original Resolution	Source	Justification
Physiographic			
Depth (m)	30 arc sec	A	Deep-divers feed on squids and fish in the deep water column
Slope (°)	30 sec arc	A	Associated with currents, high slope induce prey aggregation or enhanced primary production
Surface area of canyons and seamounts in a 0.25° cell (km ²)	30 sec arc	B	Deep-divers are often associated with canyons and seamounts structures; the variable indicates the proportion of this habitat in each cell
Oceanographic			
Mean of SST (°C)	0.2°, daily	C	Variability over time and horizontal gradients of SST reveal front locations, potentially associated with prey aggregations or enhanced primary production
Standard error of SST (°C)	0.2°, daily	C	
Mean gradient of SST (°C)	0.2°, daily	C	
Mean of SSH (m)	0.25°, daily	D	High SSH is associated with high mesoscale activity and enhanced prey aggregation or primary production
Standard deviation of SSH (m)	0.25°, daily	D	
Mean of EKE (m ² .s ⁻²)	0.25°, daily	D	High EKE relates to the development of eddies and sediment resuspension induce prey aggregation
Standard error of EKE (m ² .s ⁻²)	0.25°, daily	D	
Mean of NPP (mgC.m ⁻² .day ⁻¹)	9 km, 8 days	E	Net primary production as a proxy of prey availability

696

697

698 **Table 3. Encounter rates in sightings·100 km⁻¹ calculated for the entire study area and each sub-region of**
 699 **the North Atlantic Ocean and the Mediterranean Sea.** 'NE-ATL' means northeast Atlantic Ocean; 'NW-ATL'
 700 means northwest Atlantic Ocean and 'MED' means Mediterranean Sea.

701

	NE-ATL	NW-ATL	MED	TROPICS	STUDY AREA
Beaked whales	0.042	0.058	0.035	0.22	0.051
Sperm whales	0.057	0.067	0.09	0.095	0.067
Kogiids	0.0013	0.01	0.0	0.23	0.0085

702

703
704
705
706

Table 4. Summary of the selected models by species group.

Species group	Selected variables	Explained deviance	Akaike weight	Specific comments
Beaked whale	Depth	33.1 %	0.98	Depth, gradients SST and slope selected in the first 10 models
	Gradients SST			
	Slope			
	NPP			
Sperm whale	Depth	20.6 %	0.76	Depth, gradients SST and SSH mean selected in the first 8 models
	Gradients SST			
	SSH mean			
	NPP			
Kogiids	Depth	55.7 %	0.17	Depth, gradients SST and surface of canyons and seamounts selected in the first 7 models
	Gradients SST			
	EKE mean			
	Surface of canyons and seamounts			

707

708 **Figures**

709 **Fig. 1. Study area divided into sub-regions showing assembled survey effort (A), along with the beaked**
710 **whale (B), sperm whale (C) and kogiid (D) sightings recorded during all surveys.** The blue polygon
711 delineates overall study area and other polygons delineate sub-regions. Surveys were carried out along transects
712 following a line-transect methodology (survey details in Appendix S1 in Supporting Information). Sightings were
713 classified by group sizes with each point representing one group of individuals and point size representing the
714 number of animals in a group.

715

716 **Fig. 2. Beaked whale and sperm whale averaged ESWs estimated for each survey group and each**
717 **platform type.** For each survey group, the boxplot represents the extent of estimated ESWs depending on
718 Beaufort sea-states and observation heights recorded within the group.

719

720 **Fig. 3. Functional relationships for the selected variable (A) and the predicted relative densities of beaked**
721 **whales in individuals-km⁻² (B and C).** A: Solid lines are the estimated smooth functions, and the shaded regions
722 represent the approximate 95% confidence intervals. The y-axes indicate the number of individuals on a log
723 scale, where zero indicates no effect of the covariate. The vertical lines indicate the 2.5th and 97.5th quantiles of
724 the data. Black areas on prediction maps (B: without precautionary approach and C: with a 5% precautionary
725 approach) represent zones where we did not extrapolate the predictions. Percentages represent the proportion of
726 the study area defined as interpolation with the gap analysis.

727

728 **Fig. 4. Functional relationships for the selected variable (A) and the predicted relative densities of sperm**
729 **whales in individuals-km⁻² (B and C).** A: Solid lines are the estimated smooth functions, and the shaded regions
730 represent the approximate 95% confidence intervals. The y-axes indicate the number of individuals on a log
731 scale, where zero indicates no effect of the covariate. The vertical lines indicate the 2.5th and 97.5th quantiles of
732 the data. Black areas on prediction maps (B: without precautionary approach and C: with a 5% precautionary
733 approach) represent zones where we did not extrapolate the predictions. Percentages represent the proportion of
734 the study area defined as interpolation with the gap analysis.

735
736 **Fig. 5. Functional relationships for the selected variable (A) and the predicted relative densities of kogiids**
737 **in individuals-km⁻² (B and C).** A: Solid lines are the estimated smooth functions, and the shaded regions
738 represent the approximate 95% confidence intervals. The y-axes indicate the number of individuals on a log
739 scale, where zero indicates no effect of the covariate. The vertical lines indicate the 2.5th and 97.5th quantiles of
740 the data. Black areas on prediction maps (B: without precautionary approach and C: with a 5% precautionary
741 approach) represent zones where we did not extrapolate the predictions. Percentages represent the proportion of
742 the study area defined as interpolation with the gap analysis.

743

744

745

746 **Supporting Information**

747 **Appendix S1:** Characteristics of the study area.

748 **Appendix S2:** Details of surveys used in the analyses. Total effort represents the total length of transects of each
749 survey (without removing the transects with a Beaufort sea-state > 4). MED: Mediterranean Sea; NE-ATL:
750 Northeast Atlantic Ocean; NW-ATL: Northwest Atlantic Ocean.

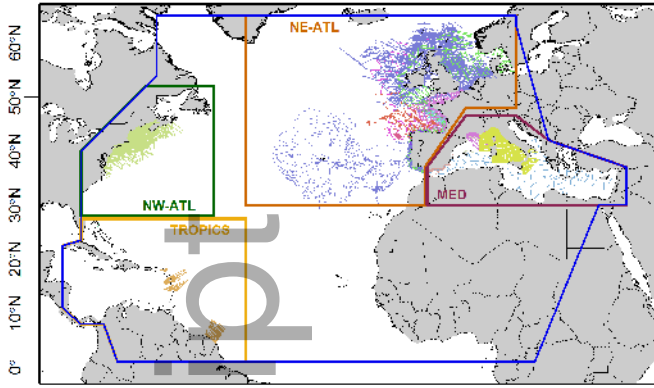
751 **Appendix S3:** Monthly environmental conditions averaged over the study period (from 1998 to 2015).

752 **Appendix S4:** Effective Strip Width estimation methodology.

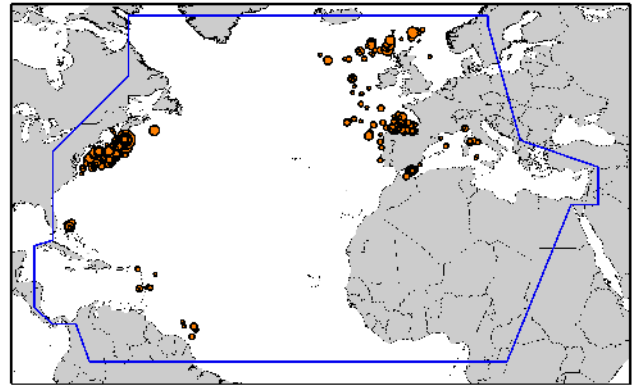
753 **Appendix S5:** Supporting information for the models.

754 **Appendix S6:** Comparison between predictions of the 5 best kogiid models, the average prediction of the 5 best
755 models (Mean) and the average prediction of the 5 best models weighted by the Akaike weight (Weighted).
756 The 5 models are described in the table at the bottom (mod: model; AIC: Akaike Information Criterion).
757 "Mean" is the simple average of the predictions of the 5 best models. To calculate the "Weighted"
758 prediction, we averaged the predictions of the 5 best models by weighting each prediction by the Akaike
759 weight (weighted.mean function of the raster package).

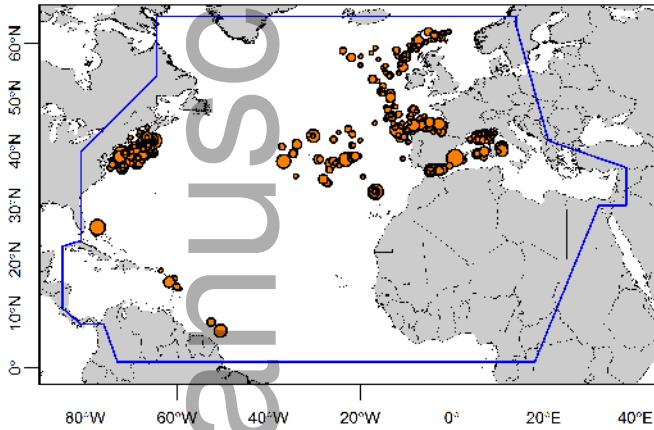
A - Study area



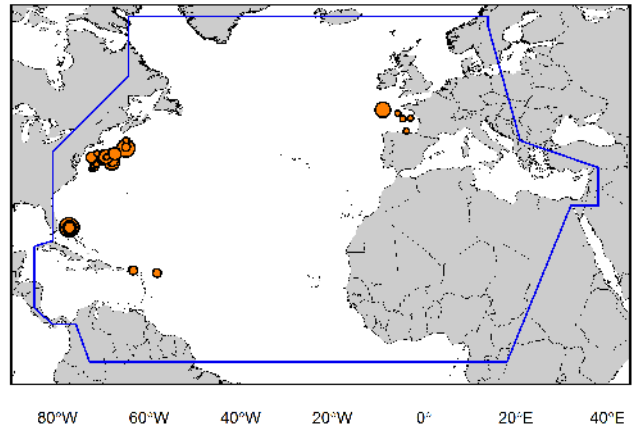
B - Beaked whale sightings



C - Sperm whale sightings



D - Kogiid sightings



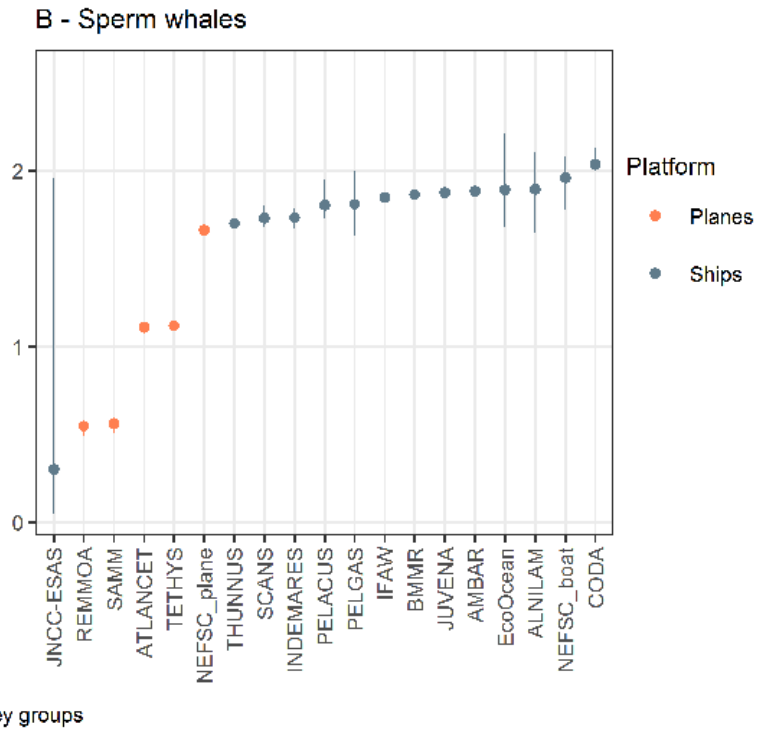
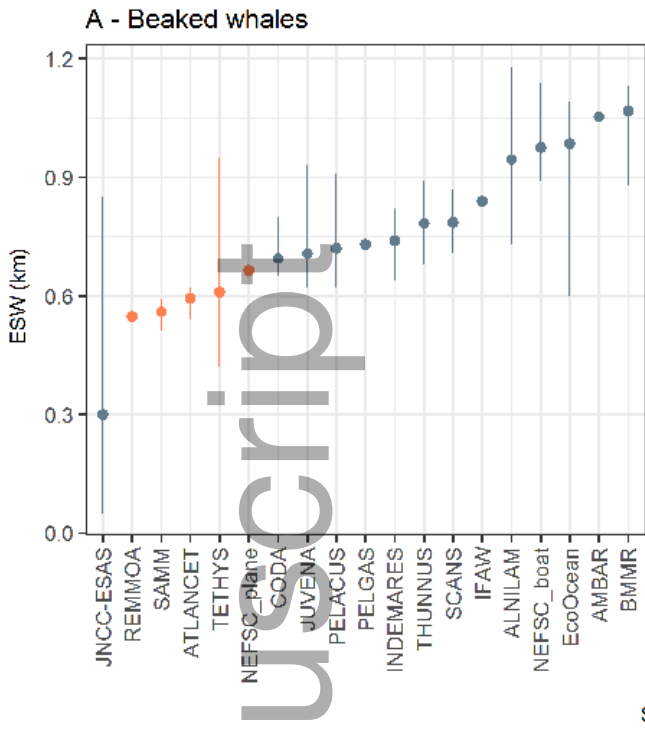
Survey groups

ALNILAM	BMMRO	JNCC-ESAS	JUVENA	PELGAS	SCANS II
AMBAR	CODA	IFAW	NEFSC	REMMOA	TETHYS-ISPRA
ATLANCET	EcoOcean	INDEMARES	PELACUS	SAMM	THUNNUS

Group size (number of individuals)
• 1 • [2:4] • [5:10] • 10+

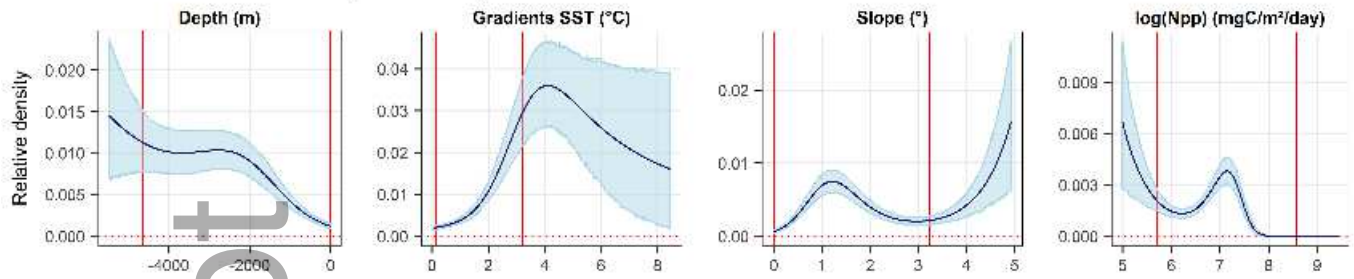
geb_12850_f1.tif

Author

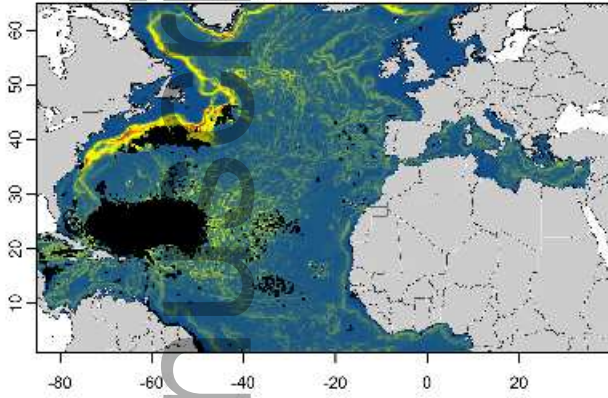


geb_12850_f2.tiff

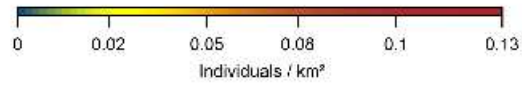
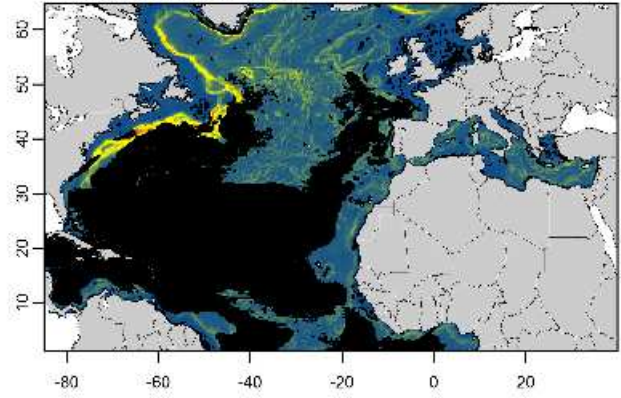
A - Beaked whale model ($D^* = 33.1\%$)



B - Predictions with simple interpolation (interpolation 94%)

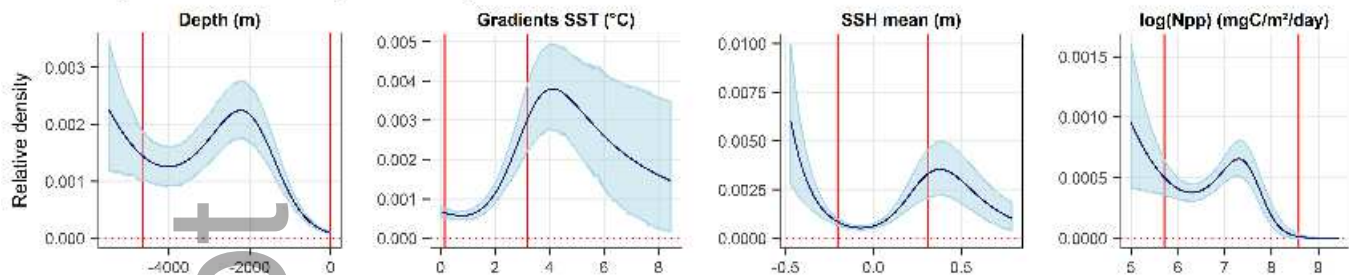


C - Predictions with precautionary interpolation (interpolation 53%)

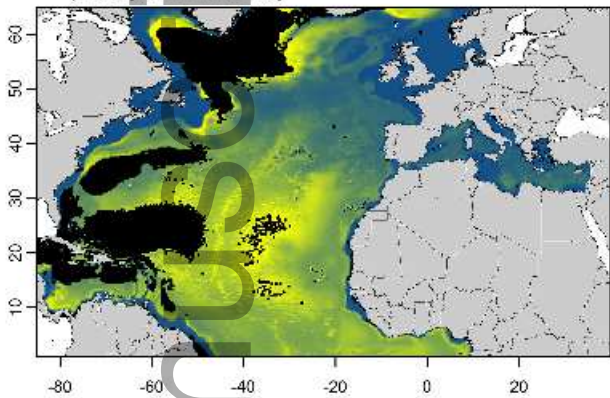


geb_12850_f3.tif

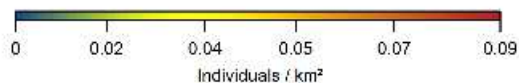
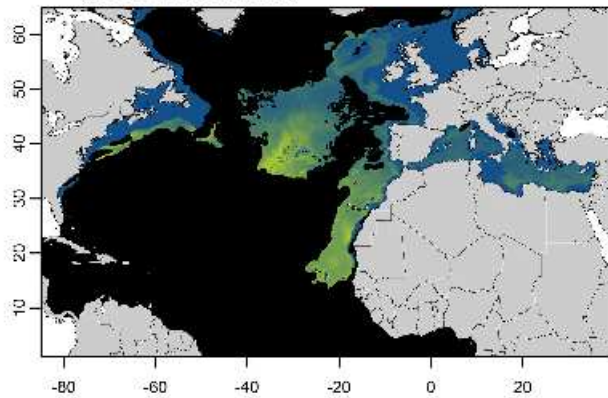
A - Sperm whale model ($D^* = 20.6\%$)



B - Predictions with simple interpolation (interpolation 84%)

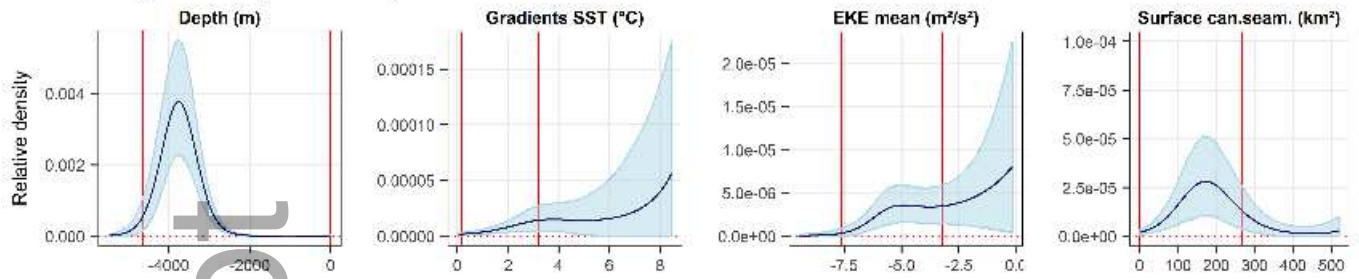


C - Predictions with precautionary interpolation (interpolation 30%)

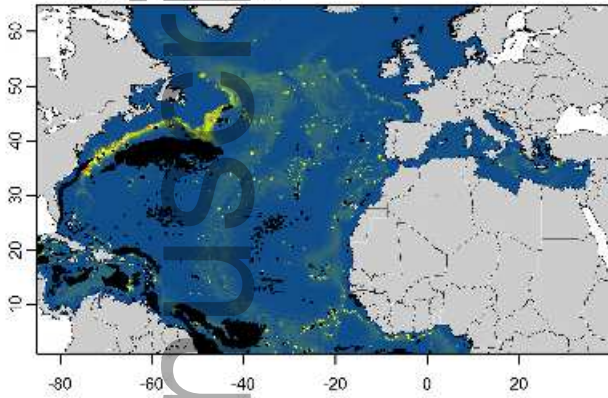


geb_12850_f4.tif

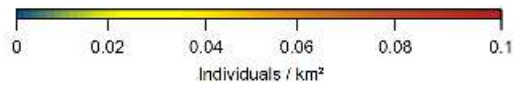
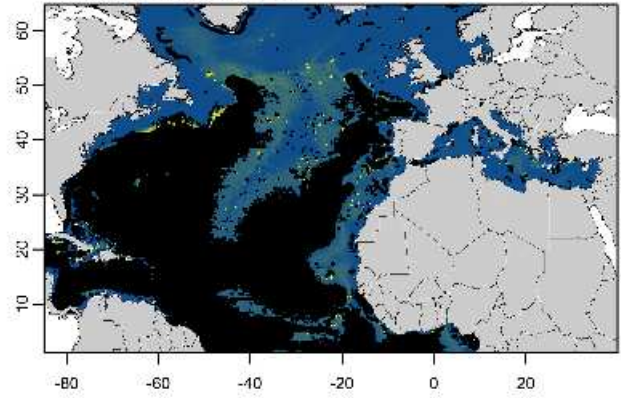
A - Kogiid model ($D^* = 55.7\%$)



B - Predictions with simple interpolation (interpolation 94%)



C - Predictions with precautionary interpolation (interpolation 55%)



geb_12850_f5.tif