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### **A Versatile Rapture (RAD-Capture) Platform for Genotyping Marine Turtles**

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#### **Abstract**

Advances in high-throughput sequencing (HTS) technologies coupled with increased interdisciplinary collaboration are rapidly expanding capacity in the scope and scale of wildlife

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29 genetic studies. While existing HTS methods can be directly applied to address some evolutionary  
30 and ecological questions, certain research goals necessitate tailoring methods to specific study  
31 organisms, such as high-throughput genotyping of the same loci that are comparable over large  
32 spatial and temporal scales. These needs are particularly common for studies of highly mobile  
33 species of conservation concern like marine turtles, where life history traits, limited financial  
34 resources and other constraints require affordable, adaptable methods for HTS genotyping to meet a  
35 variety of study goals. Here, we present a versatile marine turtle HTS targeted enrichment platform  
36 adapted from the recently developed Rapture (RAD-Capture) method specifically designed to meet  
37 these research needs. Our results demonstrate consistent enrichment of targeted regions throughout  
38 the genome and discovery of candidate variants in all species examined for use in various  
39 conservation genetics applications. Accurate species identification confirmed the ability of our  
40 platform to genotype over 1,000 multiplexed samples, and identified areas for future methodological  
41 improvement such as optimization for low initial concentration samples. Finally, analyses within  
42 green turtles supported the ability of this platform to identify informative SNPs for stock structure,  
43 population assignment and other applications over a broad geographic range of interest to  
44 management. This platform provides an additional tool for marine turtle genetic studies and  
45 broadens capacity for future large-scale initiatives such as collaborative global marine turtle genetic  
46 databases.

## 47 **Introduction**

48 Marine turtles are migratory, long-lived megafauna of conservation concern, with  
49 populations of all species classified in high risk categories on the IUCN Red List of Threatened  
50 Species (IUCN 2017). The complex behaviors and life history traits marine turtles exhibit can make  
51 them highly susceptible to human impacts, while also posing challenges to understanding critical  
52 aspects of their biology required for their conservation (Wyneken *et al.* 2013). Over the past several  
53 decades, genetic approaches have provided key insight to important research questions in marine  
54 turtle biology and conservation, including natal homing to breeding grounds, connectivity between  
55 distant foraging grounds and nesting beaches, delineation of broad stocks and DPS's for  
56 management (ESA 1973), and quantifying proportional impacts of fisheries across populations  
57 (reviewed in Jensen *et al.* 2013; Komoroske *et al.* 2017). Yet despite this progress, a diversity of  
58 unresolved research questions persist (Rees *et al.* 2016), many of which are well-suited to being  
59 addressed with emerging genetic and genomic approaches.

60 Genomic technological capabilities, especially high-throughput technologies (HTS), have  
61 rapidly expanded over the past decade to tackle a broader variety of questions in ecology and  
62 evolution (Eklblom & Galindo 2011; Ellegren 2014; Romiguier *et al.* 2014). Whole genome  
63 sequencing (WGS) and reduced representation approaches such as targeted enrichment,  
64 transcriptome (RNA-Seq) and restriction-site associated nuclear DNA (RAD-Seq) sequencing are  
65 becoming increasingly common with the continued decline in HTS costs and improvement of  
66 reference genome availability (Andrews *et al.* 2016; De Wit *et al.* 2015; Jones & Good 2016; Genome  
67 10K 2009; Todd *et al.* 2016). However, resource development and applications in some taxa,  
68 especially many of conservation concern, have lagged behind others (Shafer *et al.* 2015; Garner *et al.*  
69 2016). This is true for marine turtles and other non-mammalian vertebrates, highlighted by the fact  
70 that mammals comprise only 8% of the total number of vertebrate species, but represent over 70%  
71 of existing vertebrate genomes currently on *Ensembl* (Flicek *et al.* 2014). This has been in part due to  
72 limited resources and logistical constraints involved in sampling animals with protected status and  
73 complex life histories, but also because these approaches are not compatible or cost effective with  
74 some of the highest priority research needs for these species. For example, WGS or reduced  
75 representation approaches that can be directly applied with little to no *a priori* genomic resources  
76 (RNA- and RAD-Seq) are well suited to address some research topics like phylogenomics and  
77 adaptive variation (Jarvis *et al.* 2014; Prince *et al.* 2017). However, these methods have limitations for  
78 applications in many wildlife genotyping applications. For example, obtaining high quality RNA  
79 from protected species is often not possible, and conducting RAD-Seq to the depth of coverage  
80 needed to consistently recover the same variants (particularly single nucleotide polymorphisms;  
81 SNPs) for genotyping sample sets over large spatial and temporal scales is usually cost prohibitive.  
82 Targeted enrichment traditionally requires prior knowledge of loci and variants, or the use of coding  
83 or conserved loci that may not yield informative variants for common study goals such as fine-scale  
84 population structure or kinship studies, especially in species with low genomic diversity. In fact,  
85 many applications require cost-effective high-throughput genotype data for specific study organisms  
86 and goals, but do not have high quality reference genomes or other *a priori* genomic information.  
87 This scenario is particularly common in conservation research (Hunter *et al.* 2018) and monitoring of  
88 wide-ranging, long-lived species such as marine turtles, where samples often need to be compared  
89 across regions, continents and generations, such as fisheries bycatch DPS assignment and genetic  
90 capture-recapture studies (Komoroske *et al.* 2017; Shamblin *et al.* 2017; Stewart *et al.* 2016).

91 Several methods have recently emerged to meet these needs, including Genotyping-in-  
92 Thousands by sequencing (GT-Seq; Campbell *et al.* 2015), Rapture (Ali *et al.* 2016), and RADcap  
93 (Hoffberg *et al.* 2016). Each of these approaches has demonstrated utility and strong potential for  
94 future broader application in conservation research under different study objectives and contexts.  
95 Marine turtle conservation researchers frequently need to genotype samples for different species,  
96 sample quantities, numbers of loci (e.g., for stock structure versus relatedness studies), yet have  
97 limited time and financial resources to develop informative markers tailored to each study goal.  
98 Additionally, despite being one of the largest and most threatened vertebrate groups (Shaffer *et al.*  
99 2015), there are currently fewer available genomes or transcriptomes for non-avian reptiles relative  
100 to other classes (but see examples of existing turtle resources: Tzika *et al.* 2015; Shaffer *et al.* 2013;  
101 Wang *et al.* 2013), making it challenging to identify informative SNP loci *a priori*. Finally, researchers  
102 often deal with samples of varying tissue types, storage conditions, quality and quantity due to field,  
103 resource, permitting and other limitations (e.g., samples from decomposing stranded animals, limited  
104 refrigeration in tropical study sites, and international CITES and shipping regulations). Thus, while  
105 no one approach provides an *a priori* solution to all of these research needs, we sought to develop a  
106 robust, flexible platform that could be employed across a variety of research projects by adapting the  
107 Rapture method developed by Ali *et al.* (2016). In particular, we leveraged an existing molecular and  
108 tissue collection to test the utility of our approach with samples spanning the conditions frequently  
109 encountered in marine turtle research, and combined initial RAD-Seq with Rapture to identify  
110 candidate regions prior to enrichment target design. In brief, this entailed conducting RAD-Seq on a  
111 representative subset of samples to perform SNP discovery, followed by designing a custom oligo  
112 bait set to perform sequence capture for a selected subset of RADtags. Here, we present our results  
113 and highlight the strengths, limitations, and future applications of this platform and general  
114 approach in marine turtle biology and conservation research.

115

## 116 **Materials and Methods**

### 117 Sample Selection, Processing and RAD-Sequencing

118 We selected 96 samples from the national Marine Mammal and Sea Turtle Research  
119 Collection (MMASTR) housed at NOAA Southwest Fisheries Science Center (La Jolla, CA) that  
120 collectively were representative of the genetic diversity among and within global leatherback  
121 populations. Samples were collected from 1988-2016, including nesting females, adult males,

122 hatchlings (sex undetermined), as well as in-water foraging, stranded and bycaught animals of both  
123 sexes. Sample selection for this initial phase included only leatherback turtles and was weighted  
124 toward Pacific samples to contribute to a complementary project investigating fine-scale population  
125 structure in the Pacific. As part of previous genetic studies at SWFSC, tissue samples (skin, blood or  
126 muscle) had been preserved in saturated salt when available, shipped, and stored in the NOAA-  
127 National Marine Fisheries Service MMASTR Collection at -20°C. Genomic DNA (gDNA) used in  
128 this study had been either previously isolated from sub-samples of these tissues, or were extracted  
129 during the course of this study using one of the following standard extraction techniques:  
130 phenol/chloroform (Sambrook *et al.* 1989), sodium chloride (Miller *et al.* 1988), a modified DNeasy  
131 Qiagen extraction kit (Qiagen, Valencia, California), or Qiagen reagents on a Corbett CAS-1200  
132 extraction robot (Corbett Robotics, San Francisco, California) or PerkinElmer JANUS robot  
133 (Waltham, MA). After extraction, gDNA was stored at -80°C until use in downstream analyses. All  
134 candidate samples were checked for DNA quantity and quality via Qubit Fluorometry (Thermo  
135 Fisher Scientific, Waltham, MA) and a 4200 TapeStation System (Agilent, Santa Clara, CA),  
136 respectively. Samples with adequate concentrations and the best quality (i.e., highest molecular  
137 weight; targeting those with clear peak  $\geq 15,000$  bp, though samples with lower values were included  
138 if they were the best samples from a high priority location) were normalized and included in the final  
139 sample set for each location. Libraries were prepared following the updated RAD protocol as  
140 described in Ali *et al.* (2016) using *SbfI*-HF and NEBNext Ultra DNA Library Prep Kit for Illumina  
141 (New England Biolabs, Ipswich, MA). The only modification from the NEB-provided protocol was  
142 that the first AMPure XP size selection was performed using 50  $\mu$ l of AMPure XP beads and the  
143 second size selection used an additional 50  $\mu$ l of AMPure XP beads to generate libraries 200 bp-500  
144 bp. We used 12 PCR cycles with conditions as indicated by the NEB protocol. The libraries were  
145 sequenced at the UC Davis Genomics Core Facility for paired-end 100 bp reads in 25% of a lane on  
146 an Illumina HiSeq 3000 instrument.

147

#### 148 RAD Data Analysis & Capture Target Design

149 We demultiplexed samples by assigning only reads with perfect matching barcodes to  
150 samples, and assessed raw sequence data quality with FASTQC (Ali *et al.* 2016; Andrews 2010; see  
151 Data Accessibility section for full details on data analysis parameters and associated scripts). The

152 leatherback turtle genome has not yet been assembled, and the green turtle is the closest related  
153 species with reference genome. Although divergence of the *Dermochelidae* - *Cheloniidae* families is  
154 estimated at approximately 100 million years before present (Duchene *et al.* 2012), given the  
155 evidence for slower rates of DNA evolution among turtles relative to many other vertebrates (Avisé  
156 *et al.* 1992) and the potential benefits of using a common reference genome relative to *de novo*  
157 assembly for our project goals, we aligned the leatherback RAD data to the green turtle genome  
158 (Wang *et al.* 2013) with the Burrows-Wheeler Aligner (BWA v0.7.5; Li & Durbin 2009) and  
159 evaluated mapping performance. We used *SAMtools* (v1.3; Li *et al.* 2009) to sort, filter for proper  
160 pairs and index alignments, remove PCR duplicates, and calculate summary statistics. After  
161 observing high mapping success (see results), we proceeded using these alignments to identify  
162 candidate SNPs and cross-species Rapture target loci. In brief, we employed a *SAMtools* genotype  
163 likelihood model in the program *ANGSD* (Korneliussen *et al.* 2014; Nielsen *et al.* 2012) to infer  
164 major and minor alleles and minor allele frequencies (MAF) for sites with data for at least one  
165 individual, mapping quality score  $\geq 10$  and base quality score  $\geq 20$ . Specifically, we inferred major and  
166 minor alleles and estimated MAF using genotype likelihoods with a fixed major allele and unknown  
167 minor allele (Kim *et al.* 2011), adapted with an expectation-maximization algorithm as implemented  
168 in *ANGSD*. We then identified good candidate regions for targeted enrichment as regions that  
169 consistently had data for the expected locus length ( $\sim 84$  bp) across samples, both up and  
170 downstream of an identified restriction site in a high proportion of total individuals ( $\geq 68\%$  for all  
171 samples;  $\geq 80\%$  for Pacific leatherbacks only), and without any suspected polymorphisms within the  
172 restriction site or unknown nucleotide identity (N) in the reference sequence. Within regions that  
173 passed these criteria, we then randomly selected one of the paired regions (i.e., either up- or  
174 downstream of the restriction site) and created candidate lists for two target types: (1) potential  
175 candidate SNP loci ( $MAF \geq 0.1 \leq 0.4$ , allowing only one variable site within 150bp from the  
176 restriction site; preferentially including those with a SNP within the first 84bp), and (2) no additional  
177 filters, to serve as comparable locus set for genome representation within and across marine turtle  
178 species (e.g., for metrics of sequence diversity). The green turtle genome is not a chromosomal level  
179 assembly so we could only assess loci proximity within scaffolds; however, we manually examined  
180 the resulting list of candidate regions to ensure a high total number of unique scaffolds included, as  
181 well as the physical distance between loci on the large scaffolds. We used corresponding sequences  
182 from the green turtle genome to design a custom MYBaits in-solution DNA target enrichment kit

183 set (120bp baits, Arbor Biosciences, formerly MYcroarray Inc., Ann Arbor, MI) with 1008 targets  
184 for the first category and 999 targets for the second (2007 targets total) according to manufacturer  
185 protocols and quality control filters (e.g., probe compatibility, repeat masking, and melting  
186 temperature filters) with minor modifications to address initial failure of higher GC content baits  
187 (see below and Appendix S1 for details).

188

#### 189 Rapture Sample Selection, Library Preparation & Sequencing

190 We selected DNA samples from the MASTR collection encompassing a cross section of  
191 covariates to examine the versatility of this method for the varied conditions frequently encountered  
192 in our studies (e.g., sample location, sex, life stage, collection method, tissue type, DNA  
193 concentration, DNA quality and collection year; 1342 samples total; see Table S1 for details of  
194 samples and blanks). In particular, we included samples with detectable concentrations at or below 5  
195 ng/ul, which are frequently encountered in minimally invasive sampling of sensitive wildlife species,  
196 but below typical recommended concentrations for many reduced representation genome protocols.  
197 A total gDNA of 50 ng was targeted as starting material for each RAD library across all samples  
198 with a maximum input volume of 10 ul (i.e., samples with initial concentrations < 5 ng/ul had lower  
199 starting input). Although sample selection was again weighted toward leatherbacks for a  
200 complementary study, representative samples from six of the seven extant sea turtle species were  
201 included to evaluate target enrichment success across species and geographic regions, as well as  
202 green turtle samples representative of all currently defined global distinct population segments (DPS;  
203 Seminoff *et al.* 2015) to confirm the consistency of these genome-wide markers with established  
204 management delineations. We prepared RAD libraries as described above (Ali *et al.* 2016; 16 RAD  
205 libraries total; Table S1), with the modification of including samples with initial gDNA  
206 concentrations across the range frequently obtained from wild marine turtle samples (i.e., not  
207 selecting higher concentration samples only). We quantified libraries using a Fragment Analyzer  
208 (Agilent Technologies, Santa Clara, CA), normalized and pooled, followed by targeted enrichment  
209 following manufacturer's protocols (MYbaits version 3.02, July 2016), with the exception of  
210 doubling the capture reaction to include all RAD libraries (i.e., using 2 reactions of a 12 reaction kit  
211 for a pool of 16 RAD libraries, equivalent to 1/8 capture reaction per RAD library). During  
212 amplification steps in RAD and capture enrichment protocols, we used 20% of the template to  
213 perform a PCR test using 15 cycles followed by quantification on a Fragment Analyzer to calculate

214 the minimum number of PCR cycles required for each RAD library or enriched library pool with the  
215 remaining 80% template to minimize PCR clones. The final enriched library pool contained all 16  
216 libraries and 1342 samples. The enriched library pool was sequenced at the UC Davis Genomics  
217 Core Facility on an Illumina HiSeq 3000 instrument in a full lane (paired-end 150-bp reads).

218

#### 219 Rapture Data Quality Assessment & Analyses

220 We demultiplexed samples as described above and assessed assignment error by quantifying  
221 the absolute and proportional number of raw reads (1) assigned to unused Illumina indexes or  
222 blanks (i.e., staggered wells without DNA within each plate/RAD library; Table 1) or (2) had  
223 barcodes on both forward and reverse reads. We assessed sequence data quality with *FASTQC* and  
224 *MultiQC* (Andrews 2010; Ewels *et al.* 2016), and calculated summary statistics in R (R Core Team  
225 2016) to examine depth and evenness of coverage across predictor factors (e.g., RAD library,  
226 species, tissue type, input concentration, sample location, and collection year). We used *BWA* and  
227 *SAMtools* as described above to map sequences and filter alignments. We examined a subset of  
228 samples and loci using the *Integrative Genomics Viewer* (IGV; Robinson *et al.* 2011) as initial checks of  
229 mapping quality and coverage, and quantitatively assessed by locus and sample coverage at a  
230 representative position within target regions (relative position 20) with *Bedtools* (Quinlan & Hall  
231 2010) and R. We combined information from raw read distributions and target loci coverage to  
232 establish quality (success/failure) thresholds, and only samples that passed these thresholds were  
233 included in subsequent data analyses. We used Picard *CollectHsMetrics*  
234 (<http://broadinstitute.github.io/picard/>) and the filtered alignments to estimate the on-target  
235 capture per individual.

236 To examine and compare the success of our approach to generate SNPs within and across  
237 species and populations informative for various genotyping applications, we conducted SNP  
238 discovery, inferred major and minor alleles, and estimated allele frequencies for variable sites using  
239 *ANGSD* (Korneliussen *et al.* 2014; Nielsen *et al.* 2012) on a series of sample sets: (1) all turtle  
240 samples, (2) hardshell (Cheloniid spp.) turtles only, (3) green turtles only, (4) all leatherback samples,  
241 and (5) a representative leatherback population. For each sample set, we employed a genotype  
242 likelihood model and applied quality filters similar to RAD data as described above, additionally only  
243 including samples that passed initial QC thresholds and alignments that were proper pairs and  
244 uniquely mapped. Polymorphic sites were identified and retained in downstream analyses only if



245 there were data for at least 50% of individuals within the group being tested,  $MAF \geq 0.05$ , and p-  
246 value of being variable  $\leq 1e-6$  (Korneliussen *et al.* 2014). To examine relationships of coverage and  
247 predictor variables with genotyping success at multiple stringency levels, we estimated genotype  
248 posterior probabilities for a set of *a priori* candidate SNP positions (identified in RAD analysis  
249 described above) using an allele- frequency based prior and called genotypes with posterior  
250 probability cut-offs of 80, 90, and 95%.

251

## 252 Species Confirmation & Population Structure Analyses

253 To validate our multiplexed, cross-species platform, we first confirmed species identification  
254 with principal components analyses (PCA) by generating a covariance matrix without calling  
255 genotypes using the *ngsCovar* function in *ngsTools* (Fumagalli *et al.* 2014; Fumagalli *et al.* 2013) on all  
256 hardshell turtles, including a small sample set of suspected hybrids (based on morphological  
257 characteristics). To reduce influence of variance in depth of coverage between samples, we used  
258 *SAMtools* to randomly subsample alignments at multiple thresholds to balance information and  
259 sample retention in subsequent analyses (Ali *et al.* 2016). Since the hardshell dataset was weighted  
260 towards green turtle samples that could obstruct distinguishing variation in the other species, we also  
261 repeated these analyses including only loggerhead, olive ridley and Kemp's ridley samples. We also  
262 estimated admixture proportions of individuals using a maximum-likelihood-based clustering  
263 algorithm with the program *NGSAdmix* (Skotte *et al.* 2013) and genetic distances for a representative  
264 subset of samples across species and geographic regions using *ngsDist* (branch support based on  
265 bootstrapping 1000 replicates with 500 SNP blocks; Vieira *et al.* 2016) and plotted as a tree with  
266 *FastME* (BME iterative taxon addition method with NNI tree refinement; Lefort *et al.* 2015) and the  
267 R packages *phangorn* (Schliep 2011) and *ape* (Popescu *et al.* 2012).

268 Secondly, we included green turtle samples from nesting grounds over a geographic range of  
269 interest in order to explore how our platform would perform delineating population structure within  
270 species. Thus, our goal was to evaluate the utility of the identified SNPs with this preliminary dataset  
271 to discern if they were likely to be informative markers in future, larger-scale analyses of stock  
272 structure and population assignment. We employed methods described above for PCA, admixture  
273 and genetic distances, and also estimated allele frequency spectra using *ANGSD* and *realSFS* to  
274 calculate pairwise  $F_{ST}$  values.

275 Finally, we also estimated allele frequency spectra to calculate genetic diversity statistics  
276 (Watterson's estimator,  $\theta_w$ , based on number of segregating sites, and Tajima's estimator,  $\theta_\pi$  or  $\pi$ ,  
277 based on pairwise differences between sequences) in *ANGSD* and *realSFS* among species  
278 (Korneliussen *et al.* 2014; Korneliussen *et al.* 2013; Tajima 1989; Watterson 1975). Unequal sample  
279 sizes, population structure and upstream filtering for SNPs can cause biases in nucleotide diversity  
280 estimations (Lozier 2014; Subramanian 2016; confirmed with subsampling simulations on this  
281 dataset), potentially creating issues in our dataset with variable sample sizes across populations with  
282 likely differing demographic histories and current status (e.g., recovering, declining, etc.). To address  
283 this, we included only the random set of targeted loci as described above with selected subsets of 4-6  
284 QC passed individuals from representative populations from each species, and report results on  
285 evaluation of descriptive statistics only. Thus, although inference from these metrics is constrained,  
286 we include them to demonstrate the utility of this platform for research employing these metrics in  
287 robust sample sets within or across species.

288

## 289 Results

### 290 RAD-Sequencing & Rapture design

291 We recovered 95.7 million total raw sequences, and 89.0% of which were retained based on  
292 sample assignment criteria. *FASTQC* confirmed consistent high sequence quality across the libraries  
293 with no evidence of contamination. After removal of four failed samples (defined as <2% of average  
294 number of sequences assigned to sample), an average of 93.9% ( $\pm 7.3\%$  S.D.) of sequences mapped  
295 to the green turtle genome, an average of 51.2% ( $\pm 4.1\%$  S.D.) of which remained after filtering out  
296 PCR clones. These results of strong concordance supported the use the green turtle genome as a  
297 reference, so we proceeded using these alignments for further Rapture bait development. We  
298 identified a total of 7,282 RAD tags with paired regions that met initial filtering criteria. Of these  
299 regions, a total of 1,379 of these candidate regions further met our SNP criteria, and were included  
300 in bait design, as well as 1,400 additional regions (see Methods). From these 2,779 final candidates,  
301 we were able to design a custom MYBaits kit that met MYcroarray's QC criteria with 2,007 targets  
302 for Rapture genotyping in marine turtles.

303

### 304 Rapture data quality analysis

305 We recovered 396 million total raw sequences, with only 0.38% of these sequences removed  
306 due to assignment to unused Illumina indexes or the presence of barcodes on both forward and  
307 reverse reads. *FASTQC* and *MultiQC* results confirmed high quality scores across and within  
308 libraries and no issues of contamination. Assignment of raw sequences to blanks dispersed across  
309 libraries was extremely low (average= 245, min/max=27/818). Based on sequence count  
310 distributions, we determined an initial sample failure/success threshold of 10,000 raw sequences,  
311 which 1127 samples passed (84%; hereafter referred to as ‘QC passed samples’). Read counts varied  
312 across RAD library and samples, but we did not observe any clear patterns of success or failure  
313 between input factors, particularly among species or DNA input. Samples more recently collected  
314 and with higher DNA initial concentrations more consistently passed initial quality thresholds, but  
315 many low concentration and older samples did as well.

316

317 Rapture target coverage and genotyping success

318 Samples exhibited overall high percentages of mapping to the green turtle genome (average  
319 98.6% reads aligned,  $\pm 1.81\%$  S.D., min/max=100/70.1%) and on-target sequence capture (average  
320 84.2% bases aligned to baited region,  $\pm 6.0\%$  S.D., min/max=97.3/47.7%; Fig. 1A). Mapped filtered  
321 (PCR clones removed) fragments for QC-passed samples were an average of 20.8% ( $\pm 6.9\%$  S.D.) of  
322 the total sequenced fragments per individual, and this was correlated with sample initial gDNA  
323 concentration (Fig. 1B). Average coverage per locus in filtered QC-passed samples was 26.6 ( $\pm 10.1$   
324 S.D.; min/max=0.9/99.1; see Fig. S1 for coverage distributions). Samples generally reached  $\geq 4x$   
325 coverage across loci with approximately 50,000-75,000 filtered alignments (Fig. S2a). However, we  
326 identified samples that passed initial QC thresholds, but had lowered numbers of filtered reads  
327 aligned and few Rapture loci covered at  $\geq 4x$  (Fig. S2b), prompting us to implement an additional  
328 filter of a minimum of 5,000 filtered reads aligned in further downstream analyses. Of these new  
329 QC-passed samples (1097 total), we were able to genotype over 50% of *a priori* identified SNPs in  
330 Rapture loci at all posterior probability thresholds tested (Fig. 2a). Ability to call genotypes increased  
331 with depth of coverage but began reaching saturation at approximately 150,000 sequenced fragments  
332 per individual (depending on posterior probability threshold and sample). However, genotyping  
333 capacity was also clearly affected by the relative position of the SNP within the Rapture locus region  
334 (Fig. 2b), displaying a distinct break at approximately relative position 100, despite the use of longer  
335 150bp paired-end sequencing.

336

### 337 Cross Species Capture Success & SNP discovery

338 We observed consistent success in coverage of Rapture loci across all species tested,  
339 confirming the broad utility of this approach for genotyping studies across marine turtle species. A  
340 reduction in the maximum loci covered regardless of total depth of coverage was observed in non-  
341 green hardshell turtle species (Fig. 3), indicating that a small percentage of selected targets in this  
342 particular enrichment set are not useful for other hardshell species, likely due to polymorphisms in  
343 *SbfI* restriction sites or other compatibility issues. Nevertheless, we identified ample candidate  
344 polymorphic SNPs suitable for within-species genotyping studies (Table 1). However, we emphasize  
345 that because SNP identification is inherently influenced by analysis parameters and input sample  
346 composition, determining informative SNPs within Rapture target regions should be conducted  
347 using samples and filtering thresholds aligned with research goals to avoid ascertainment bias. For  
348 example, the variants identified from individuals across a global distribution may be less informative  
349 for kinship studies within a particular population, so it is advantageous to conduct separate SNP  
350 discovery on a representative sample set (e.g., global vs. St. Croix leatherback groupings in Table 1).

351

### 352 Species Confirmation and Green Turtle Population Structure

353 Individuals strongly separated by species as expected in the first two PC components for all  
354 hardshell species, with the exception of the two ridley species (Fig. 4a) that resolved in further PC  
355 axes in the combined analysis, as well as separate analyses omitting green and hawksbill turtle  
356 samples (Fig. 4b). Clear species separation was similarly observed in admixture proportion results,  
357 but with even more pronounced effects of the unbalanced sample groups when all hardshell samples  
358 were included (i.e., strong breaks in population structure within green turtles began to emerge before  
359 the separation of the ridley species; Fig. 4c,d). Estimated genetic distances among species were  
360 largest as expected between leatherbacks and hardshell turtles, followed by green turtles relative to  
361 other hardshell species (loggerhead, hawksbill, Kemp's ridley, and olive ridley; Fig. S3). Several  
362 hybrids were identified, including three green-loggerhead hybrids and one green-hawksbill hybrid,  
363 however for several other suspected hybrids both PCA and admixture proportion results support  
364 only genetic contributions from olive ridley.

365 In green turtles, pairwise  $F_{st}$  values, genetic distances and PCA discerned strong breaks in  
366 population structure between major ocean regions aligned with previous studies based on mtDNA

367 and microsatellites and green turtle distinct population segment (DPS) designations (Jensen *et al.* in  
368 press; Seminoff *et al.* 2015; Figs. 5 & S4; Table S2). Tree topology branch support of genetic  
369 distances as well as  $F_{st}$  values were higher in the Atlantic compared to the Pacific Ocean. In the  
370 western Pacific, PCA clustering of samples by location for several groups are congruent with  
371 potential finer-scale population structure (Fig. S4b), further supporting the utility of these SNP  
372 markers for future stock structure and population assignment studies.

373

#### 374 Genetic Diversity Estimates

375 Patterns within groups were consistent between  $\theta_w$  and  $\pi$ , and within species, with the  
376 exception of Costa Rica hawksbills that had substantially higher values for both metrics (Fig. 6).  
377 Generally, green turtles exhibited the highest nucleotide diversity, while leatherbacks displayed the  
378 lowest. In particular, all four groups of Pacific leatherbacks had lower levels of variation relative to  
379 the Atlantic population included (Brazil).

380

#### 381 Discussion

382 Technological advances combined with increased interdisciplinary collaboration has rapidly  
383 expanded both the scope and scale of genetic studies over the past decade, yet for many species of  
384 conservation concern such as marine turtles, the realized potential of these advances is only just  
385 beginning (Garner *et al.* 2016; Komoroske *et al.* 2017; Shafer *et al.* 2015). This is in part because life  
386 history traits and protected status of these taxa can create unique research challenges, but also  
387 because the resources required for method development (which often needed to be repeated to  
388 generate informative markers tailored to each species and study goal) often have made it infeasible  
389 for conservation researchers. Our results demonstrate that the adaptation of the Rapture method  
390 developed by Ali *et al.* (2016) provides a flexible platform for marine turtle research. While  
391 limitations and room for further improvement remain, the addition of our platform and general  
392 approach to the marine turtle genetic toolbox opens the door to a diversity of rapid, cost-efficient  
393 genotyping applications. These data can be comparable across laboratories, geographical regions,  
394 and timescales, which can be particularly important in such highly mobile species that can migrate  
395 across entire ocean basins and necessitate international collaboration for effective conservation  
396 (Shamblin *et al.* 2014). Though our specific selected regions for targeted enrichment will not be  
397 suitable for all populations or research questions, our study also demonstrates how initial RAD-

398 Sequencing can be used to develop a Rapture platform suited to specific research needs.  
399 Additionally, these target regions can be adapted to other genotyping platforms that may be better  
400 suited to meet some research needs but require prior knowledge of genomic variants, e.g., GT-Seq  
401 that may have improved performance on lower quality and concentrations samples (Campbell *et al.*  
402 2015).

403 Our results highlight several key strengths of this platform in meeting the diverse needs of  
404 marine turtle genotyping applications. First, researchers often need to analyze few or many samples  
405 at few or many loci, depending on study goals. Our data demonstrate that samples can be combined  
406 and genotyped at the same loci with moderate sequencing coverage using partial capture reactions,  
407 effectively multiplexing samples as has been shown in other targeted enrichment protocols (Rohland  
408 & Reich 2012; Hancock-Hanser *et al.* 2013). This not only facilitates cost-effective, time-efficient  
409 analysis of large sample sets, but also combining samples for different projects. For example,  
410 researchers working on large nesting beaches often have many samples to analyze at the end of the  
411 season (Shamblin *et al.* 2017), while those genotyping samples from fisheries bycaught animals or  
412 some foraging population assessment projects may have smaller sample sets collected intermittently  
413 over the year. In the latter case, it has been particularly problematic to determine how to move from  
414 manual analysis with traditional markers to next-generation sequencing approaches where much of  
415 the reduced cost and time efficiency is related to multiplexing and high-throughput processing.  
416 While genotyping high priority single samples that need to be analyzed in near real-time may still  
417 pose a challenge, the flexibility of the Rapture platform offers options to combine library  
418 preparation and sequencing across projects and species, or multiplex fewer samples and reduce total  
419 sequencing depth (e.g., through the use of a lower output instrument such as an Illumina MiSeq,  
420 MiniSeq or iSeq, or coordinating with other researchers to use different library barcodes and share  
421 sequencing lanes). Additionally, we designed a custom MYBaits enrichment kit with ~2000 targets  
422 to satisfy the needs of a variety of study types, but this approach can be adapted to include fewer or  
423 more loci (see examples in Ali *et al.* (2016) and Margres *et al.* (2018), respectively). For example,  
424 researchers interested in basic population structure and individual assignment may wish to design  
425 kits with a subset of only several hundred informative targets, increasing the per locus depth of  
426 coverage in each sample (Mastretta-Yanes *et al.* 2014; Hoffberg *et al.* 2016). Researchers can also  
427 target regions with multiple SNPs to develop multiallelic microhaplotype markers that provide  
428 greater power per locus (McKinney *et al.* 2017), and can be particularly useful for applications such

429 as kinship studies (Baetscher et al. 2017). Finally, the ability to repeatedly capture the same genomic  
430 regions facilitates studies conducted over broader time periods (e.g., examining trends across many  
431 nesting seasons or even generations) or spatial scales (e.g., collaborating labs can generate and share  
432 data between foraging and nesting grounds).

433         Despite these exciting opportunities, our data also clearly show that our current Rapture  
434 platform has some limitations that are relevant to situations frequently encountered in wildlife  
435 genetics studies. First, although we were able to perform effective sequence capture and genotyping  
436 for samples across tissue types, DNA extraction methods, species, and other co-factors, a portion of  
437 our test samples failed to sequence well. Though no clear patterns emerged with sample age or  
438 molecular weight thresholds, it is likely that highly degraded or contaminated samples (e.g., due to  
439 natural conditions, collection and storage methods) were more likely to fail. While this problem is  
440 often easily circumvented in controlled experimental settings, in many conservation applications  
441 these issues can be unavoidable, such as working with museum collections or opportunistic sampling  
442 of animals that have had substantial exposure to natural elements post-mortem. However, we  
443 emphasize that many samples in our study that exhibited evidence of some degradation were  
444 successful, including those that fall into these sub-optimal categories (e.g., stranded and bycaught  
445 animals). Our results support the initial findings of Ali et al. (2016) that this new RAD protocol is  
446 more robust than previous RAD methods for partially degraded samples, but there may be a point  
447 beyond which it is not a suitable approach. However, it may be possible to generate comparable  
448 genotype data for these samples at a subset of informative Rapture loci with highly-multiplexed PCR  
449 based methods such as GT-Seq (Campbell *et al.* 2015) that amplify short DNA fragments and thus  
450 be more robust to sample degradation. Secondly, we observed a substantial proportion of sequenced  
451 fragments that were PCR clones, and this was correlated with initial sample DNA concentration.  
452 The latter observed effect may be a product of the increased influence of measurement and  
453 pipetting error at low concentrations, which could be targeted for improvement in a future protocol  
454 adaptation. However, since PCR clones are in effect wasted sequences, in practice this currently  
455 means that it is less cost effective to sequence samples with low initial DNA concentrations, and  
456 that calculations of required sequencing to attain a targeted depth of coverage must take these  
457 factors into account. Although sequencing costs are likely to continue to decrease such that  
458 genotyping can still be achieved despite this loss, future efforts to reduce clonality would improve  
459 the efficiency and cost of this approach. Thirdly, we were surprised to detect increased variation of

460 SNP genotyping beyond approximately eighty-five base pairs despite using a longer Illumina reads  
461 (150bp PE), which only became available after we conducted our initial RAD-Seq for target design  
462 (using 100bp PE). It is possible that the discrepancy in read length between the two steps resulted in  
463 some lower confidence SNPs in the extended region. Though not detrimental to the overall  
464 genotyping capability of the platform, as sequencing technologies continue to change within  
465 increased read length capacity, this may be something that researchers need to consider in project  
466 designs. Finally, we estimate the cost per sample to be approximately \$11 per sample (see Table S3  
467 for cost breakdown details). However, this assumes that researchers have access to required  
468 laboratory equipment, as well as the capacity and need to run samples in high-throughput formats.  
469 Although costs and technological accessibility have vastly improved in recent years, access to the  
470 equipment and financial resources to conduct genetic studies is far from universally available. This  
471 makes continued collaboration essential to advancing our understanding of marine turtles,  
472 particularly for researchers with access to such resources to continue efforts to increase capacity  
473 elsewhere, such as through visiting scientist training partnerships and creation of shared genetic  
474 databases. Particularly given the influence that bioinformatics parameters (e.g., filtering criteria,  
475 assembly methodology, genotyping thresholds) can have on results (O'Leary *et al.* 2018), it is  
476 imperative for researchers to include metadata and analysis details to ensure robust and comparable  
477 data across laboratories and over time.

478 We present results of conducting SNP discovery independently for each species and within a  
479 representative leatherback population to demonstrate that substantial variation exists within our  
480 targeted regions to meet a variety of study goals, but also to highlight the importance of appropriate  
481 test data and analyses parameter thresholds to avoid ascertainment bias (i.e., discerning informative  
482 SNPs appropriate for a given study goal; Lachance & Tishkoff 2013). For example, intra-population  
483 questions can require variable SNPs within a target population, which may not be identified in  
484 broader analysis including many populations depending on filtering thresholds and sample sizes  
485 (Andrews *et al.* 2018). One advantage to the flexible Rapture platform is that researchers can  
486 generate data for many genomic regions and then hone in on informative SNPs to genotype without  
487 *a priori* knowledge and the need to develop different markers tailored to each study goal, which can  
488 be cost and time prohibitive. However, as discussed previously, if desired, researchers can also use  
489 preliminary RAD or Rapture data with a representative test dataset to identify the most informative  
490 markers for their study and design a new MYBaits kit or GT-Seq primers to focus exclusively on



491 those targets. The cross-species capacity of our platform also offers flexibility to combine samples  
492 across more potential projects, and there are many other taxonomic groups where this would also  
493 advantageous. To our knowledge there have not yet been any studies examining the potential or  
494 limitations of cross-species Rapture in other taxa, particularly those with faster evolutionary rates  
495 relative to turtles (Awise *et al.* 1992) where it may be more challenging to design targets that are  
496 informative and effective across species. However, other cross-species sequence capture platforms  
497 have been employed across a variety of taxonomic groups (Jones & Good 2016), and may help  
498 inform the design and expectations of future Rapture cross-species studies.

499 Principal components and admixture proportion analyses identified clear separation of all  
500 species examined and our tree depicting relationships among species was in general agreement with  
501 previous research (Duchene *et al.* 2012; Naro-Maciel *et al.* 2008). It is important to note that these  
502 studies were focused on resolving phylogenetic relationships among all marine turtle species, and  
503 thus the methods employed were much more in-depth than our analyses; additionally, we were not  
504 able to include any flatback turtle samples in our study. Thus, clarifying any discrepancies or further  
505 confirmation using our genome-wide markers would require additional studies. However, for the  
506 purpose of our primary study goals, since species were randomized across and within RAD libraries  
507 and we observed low number of sequences assigned to blank wells, our results show that sequences  
508 can be assigned correctly to individuals using this highly-multiplexed approach and our analyses  
509 criteria. Such cross-species, highly multiplexed targeted enrichment may not be as effective in other  
510 taxa with high genomic diversity or for studies that require tens to hundreds of thousands of SNPs,  
511 and researchers working with other marine turtle species may wish to omit targets from our panels  
512 that only yielded coverage in green or leatherback turtles.

513 We identified several hybrids, in agreement with preliminary evaluation of these samples with  
514 three nuclear loci and the mitochondrial control region (Dodge *et al.* 2006), as well as several  
515 suspected hybrids that only displayed genetic contributions from one species. Unbalanced sampling  
516 can mask variants in smaller groups, as we identified in our admixture analyses, but further analyses  
517 including only turtles within these groups still did not detect genomic signatures from multiple  
518 species. This suggests that these could have been misidentified individuals, however, additional  
519 analyses with larger sample sizes from contributing species at the same locations would further  
520 validate these findings and provide insight into the prevalence of hybridization in these populations.

521 Hybridization and complex introgression patterns have been previously documented, primarily  
522 in southeast Atlantic populations (Reis *et al.* 2010; Vilaça *et al.* 2012), but the frequency of such  
523 events elsewhere and hybrid fitness is largely unknown. Given recent concern that increasingly  
524 skewed female-biased sex ratios due to climate change (Jensen *et al.* 2018) and other anthropogenic  
525 pressures (Gaos *et al.* 2018) could cause interspecies mating events to become more prevalent and  
526 further destabilize populations, additional research is needed to better understand these processes  
527 and monitor changes over time; our Rapture platform offers an additional tool for such studies

528 Our exploratory green turtle analyses determined that our platform can also successfully amplify  
529 targeted regions within species across broad geographic locations and identify informative SNPs for  
530 stock structure, population assignment and other management applications. A recent study of green  
531 turtle global phylogeography using mtDNA control region sequences identified eleven divergent  
532 lineages that each encompass a few to many genetically differentiated distinct management units  
533 (MUs) with more recent shared ancestry but deemed to be demographically independent (Jensen *et al.*  
534 *in press*). This comprehensive study builds on previous work within regions documenting  
535 restricted gene flow attributed to female natal philopatry and generally little genetic differentiation  
536 among nesting beaches within 500km (reviewed in Jensen *et al.* 2013; Jensen *et al.* *in press*;  
537 Komoroske *et al.* 2017). While instrumental for our understanding of green turtle evolutionary  
538 history and contemporary stock structure patterns, there is a clear need to complement this work  
539 with studies employing nuclear markers to identify the role of male-mediated gene flow and how  
540 increased marker resolution affects detection of fine-scale patterns. With additional refinement of  
541 the SNPs identified here specifically to meet these goals (e.g., narrower filtering criteria to remove  
542 any biases due to physical linkage or inconsistent coverage), these markers will serve as a valuable  
543 resource for such studies over large spatial and temporal scales, further advancing our understanding  
544 of green turtle population connectivity, MU designation, and human impacts.

545 Finally, comparisons of genetic variation among populations and species can be informative for  
546 a variety of conservation relevant research, such as understanding how genetic diversity may differ  
547 among healthy, recovering, and declining populations (Lozier 2014). While our current sample set  
548 was not designed to address these questions specifically, the ability to consistently amplify over a  
549 thousand regions across the genome for all marine turtles, enables our platform can be effectively  
550 employed for such research goals within or across species. For example, we found that Pacific  
551 leatherbacks exhibited the lowest levels of nucleotide diversity relative to all other groups evaluated,

552 including the (Atlantic) Brazilian nesting stock. While further robust analysis is needed to confirm  
553 this preliminary finding, this could be related to the continued decline of Pacific leatherback  
554 populations in contrast to Atlantic populations.

555 In conclusion, our Rapture platform provides a tool that is complementary to existing traditional  
556 genetic markers as well as other emerging genomic techniques suited to address a broad diversity of  
557 research questions in marine turtle ecology, evolution and conservation (e.g., transcriptome, other  
558 reduced representation, and whole genome sequencing to study adaptive variation and genome-  
559 phenome linkages). Though some limitations still hinder widespread adoption of these techniques,  
560 such as cost and well-assembled and annotated genomic resources, as technologies continue to  
561 advance we anticipate continued application and creative adaptations to meet the challenging needs  
562 of conservation researchers. If realized, this could generate capacity for large-scale initiatives such as  
563 the creation of global genetic databases akin to those that have begun emerging recently for other  
564 taxa (e.g., Deck *et al.* 2017). This would not only expand the scope of research questions that can be  
565 investigated, but also provide traditionally resource-limited marine turtle programs with the ability to  
566 incorporate genetic information in their research and monitoring efforts. Such endeavors will  
567 inevitably present many new challenges, but the successes of analogous initiatives such as the State  
568 of the World's Sea Turtles (SWOT) and the Atlantic-Mediterranean Loggerhead Genetics (LGWG;  
569 Shamblin *et al.* 2014) working groups among others have demonstrated the power of such global  
570 collaborative efforts to answer the major outstanding research questions in these wide-ranging,  
571 complex megafauna.

572

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#### 589 Literature Cited

590

591 Ali O, O'Rourke S, Amish S, Meek M, Luikart G, Jeffres C, Miller M (2016) RAD Capture  
592 (Rapture): Flexible and Efficient Sequence-Based Genotyping. *Genetics* **202**, 389–400.

593 Andrews KR, Adams JR, Cassirer EF, Plowright RK, Gardner C, Dwire M, Hohenlohe PA, Waits  
594 LP (2018) A bioinformatic pipeline for identifying informative SNP panels for parentage assignment  
595 from RADseq data. *Mol Ecol Resour.* <https://doi.org/10.1111/1755-0998.12910>

596 Andrews KR, Good J, Miller MR, Luikart G, Hohenlohe PA (2016) Harnessing the power of  
597 RADseq for ecological and evolutionary genomics. *Nature Reviews Genetics*. **17**, 81–92.

598 Andrews S (2010) FastQC: a quality control tool for high throughput sequence data.  
599 <https://www.bioinformatics.babraham.ac.uk/projects/fastqc/>

600 Avise JC, Bowen BW, Lamb T, Meylan AB, Bermingham E (1992) Mitochondrial DNA evolution at  
601 a Turtle's Pace: evidence for low genetic variability and reduced microevolutionary rate in the  
602 testudines. *Mol Biol Evol*, **9**, 457–473.

603 Baetscher DS, Clemento AJ, Ng TC, Anderson EC, Garza JC (2017) Microhaplotypes provide  
604 increased power from short-read DNA sequences for relationship inference. *Mol Ecol Resour*, **18**,  
605 296-305.

606 Campbell N, Harmon S, Narum S (2015) Genotyping-in-Thousands by sequencing (GT-seq): A cost  
607 effective SNP genotyping method based on custom amplicon sequencing. *Mol Ecol Resour*, **15**, 855-  
608 867.

609 De Wit P, Pespeni M, Palumbi S (2015) SNP genotyping and population genomics from expressed  
610 sequences –current advances and future possibilities. *Mol Ecol*, **24**, 2310–2323.

611 Deck J, Gaither M, Ewing R, Bird C, Davies N, Meyer C, ..., Crandall ED (2017) The Genomic  
612 Observatories Metadatabase (GeOMe): A new repository for field and sampling event metadata  
613 associated with genetic samples. *PLoS Biol*, **15**, e2002925.

614 Dodge K, LeRoux R, Frey A, Dutton P (2006) Confirmation of marine turtle hybrids in northwest  
615 Atlantic waters. In: *Twenty Sixth Annual Symposium on Sea Turtle Biology and Conservation* (eds. Frick M,  
616 Panagopoulou A, Rees AF, Williams K), p. 376 pp. International Sea Turtle Society, Athens, Greece.

617 Duchene S, Frey A, Alfaro-Nunez A, Dutton PH, Gilbert MTP, Morin PA (2012) Marine turtle  
618 mitogenome phylogenetics and evolution. *Mol Phylogenetics Evol*, **65**, 241-250.

619 Ekblom R, Galindo J (2011) Applications of next generation sequencing in molecular ecology of  
620 non-model organisms. *Heredity*, **107**, 1-15.

621 Ellegren H (2014) Genome sequencing and population genomics in non-model organisms. *TREE*,  
622 **29**, 51-63.

623 Endangered Species Act of 1973. U.S. Fish and Wildlife Service Home,  
624 [www.fws.gov/laws/lawsdigest/ESACT.html](http://www.fws.gov/laws/lawsdigest/ESACT.html). Accessed August 24, 2018.

625 Ewels P, Magnusson M, Lundin S, Käller M (2016) MultiQC: Summarize analysis results for  
626 multiple tools and samples in a single report. In: *Bioinformatics*. <http://multiqc.info/>

627 Flicek P, Amode MR, Barrell D, Beal K, Billis K, Brent S, ..., Searle SM (2014) Ensembl 2014.  
628 *Nucleic Acids Res*, **42**, D749-755.

629 Fumagalli M, Vieira F, Linderroth T, Nielsen R (2014) ngsTools: methods for population genetics  
630 analyses from next-generation sequencing data. *Bioinformatics*, **30**, 1486-1487.

631 Fumagalli M, Vieira FG, Korneliussen TS, Linderroth T, Huerta-Sánchez E, Albrechtsen A, Nielsen  
632 R (2013) Quantifying Population Genetic Differentiation from Next-Generation Sequencing Data.  
633 *Genetics*, **195**, 979-992.

634 Gaos AR, Lewison RL, Liles MJ, Henriquez A, Chavarría S, Yañez IL, Stewart K, ..., Dutton PH  
635 (2018) Prevalence of polygyny in a critically endangered marine turtle population. *JEMBE*, **506**, 91-  
636 99.

637 Garner B, Hand B, Amish S, Bernatchez L, Foster J, Miller K, ..., Luikart G (2016) Genomics in  
638 conservation: case studies and bridging the gap between data and application. *TREE*, **31**, 81-83.

639 Genome 10K Community of Scientists (2009) Genome 10K: a proposal to obtain whole-genome  
640 sequence for 10,000 vertebrate species. *J Hered*, **100**, 659–674.

641 Hancock-Hanser, BL, Frey A, Leslie MS, Dutton PH, Archer FI, Morin PA (2013) Targeted  
642 multiplex next-generation sequencing: advances in techniques of mitochondrial and nuclear DNA  
643 sequencing for population genomics. *Mol Ecol Res*, **13**, 254-268.

644 Hoffberg SL, Kieran TJ, Catchen JM, Devault A, Faircloth BC, Mauricio R, Glenn TC (2016)  
645 RADcap: sequence capture of dual digest RADseq libraries with identifiable duplicates and reduced  
646 missing data. *Mol Ecol Res*, **16**, 1264-1278.

647 Hunter ME, Hoban SM, Bruford MW, Segelbacher G, Bernatchez L (2018) Next-generation  
648 conservation genetics and biodiversity monitoring. *Evol Appl*, **11**, 1029-1034.

649 IUCN (2017) The IUCN Red List of Threatened Species. Version 2017-3.  
650 <http://www.iucnredlist.org>. Accessed 05 December 2017.

651 Jarvis ED, Mirarab S, Aberer AJ, Li B, Houde P, Li C, ..., Zhang G (2014) Whole-genome analyses  
652 resolve early branches in the tree of life of modern birds. *Science*, **346**, 1320-1331.

653 Jensen M, FitzSimmons N, Dutton P (2013) Molecular Genetics of Sea Turtles. In: Wyneken J,  
654 Lohmann K, Musick J (eds) *Biology of Sea Turtles* CRC Press. pp. 135-162.

655 Jensen MP, Allen CD, Eguchi T, Bell IP, LaCasella EL, Hilton WA, Hof CAM, Dutton PH (2018)  
656 Environmental Warming and Feminization of One of the Largest Sea Turtle Populations in the  
657 World. *Curr Biol*, **28**, 154-159.

658 Jensen MP, Fitzsimmons NN, Bourjea J, Hamabata T, Reece J, Dutton PH (in press) The  
659 evolutionary history and global phylogeography of the green turtle (*Chelonia mydas*). *J Biogeogr.*

660 Jones M, Good J (2016) Targeted capture in evolutionary and ecological genomics. *Mol Ecol*, **25**, 185-  
661 202.

662 Kim S, Lohmueller K, Albrechtsen A, Li Y, Korneliussen T, Tian G, ..., Nielsen R (2011)  
663 Estimation of allele frequency and association mapping using next-generation sequencing data. *BMC*  
664 *Bioinformatics*, **12**, 231.

665 Komoroske LM, Jensen MP, Stewart KR, Shamblin BM, Dutton PH (2017) Advances in the  
666 Application of Genetics in Marine Turtle Biology and Conservation. *Frontiers in Marine Science*, **4**, 156.

667 Korneliussen TS, Albrechtsen A, Nielsen R (2014) ANGSD: Analysis of Next Generation  
668 Sequencing Data. *BMC Bioinformatics*, **15**, 356.

669 Korneliussen TS, Moltke I, Albrechtsen A, Nielsen R (2013) Calculation of Tajima's D and other  
670 neutrality test statistics from low depth next-generation sequencing data. *BMC Bioinformatics*, **14**, 289.

671 Lachance J, Tishkoff SA (2013) SNP ascertainment bias in population genetic analyses: why it is  
672 important, and how to correct it. *Bioessays*, **35**, 780-786.

673 Lefort V, Desper R, Gascuel O (2015) FastME 2.0: A Comprehensive, Accurate, and Fast Distance-  
674 Based Phylogeny Inference Program. *Mol Biol Evol*, **32**, 2798-2800.

675 Li H, Durbin R (2009) Fast and accurate short read alignment with Burrows-Wheeler Transform.  
676 *Bioinformatics*, **25**, 1754-1760.

677 Li H, Handsaker B, Wysoker A, Fennell T, Ruan J, Homer N, ..., Durbin R, Subgroup GPDP  
678 (2009) The Sequence alignment/map (SAM) format and SAMtools. *Bioinformatics*, **25**, 2078-2079.

679 Lozier JD (2014) Revisiting comparisons of genetic diversity in stable and declining species:  
680 assessing genome-wide polymorphism in North American bumble bees using RAD sequencing. *Mol*  
681 *Ecol*, **23**, 788-801.

682 Margres MJ, Jones ME, Epstein B, Kerlin DH, Comte S, Fox S, ..., Storfer A (2018) Large-effect  
683 loci affect survival in Tasmanian devils (*Sarcophilus harrisii*) infected with a transmissible cancer. *Mol*  
684 *Ecol*, **27**, 4189-4199.

685 Mastretta-Yanes A, Arrigo N, Alvarez N, Jorgensen TH, Piñero D, Emerson BC (2014) Restriction  
686 site-associated DNA sequencing, genotyping error estimation and de novo assembly optimization  
687 for population genetic inference. *Mol Ecol Res*, **15**, 28-41.

688

689 McKinney GJ, Seeb JE, Seeb LW (2017) Managing mixed-stock fisheries: genotyping multi-SNP  
690 haplotypes increases power for genetic stock identification. *Canadian Journal of Fisheries and*  
691 *Aquatic Sciences*, **74**, 429-434.

692 Miller SA, Dykes DD, Polesky HF (1988) A simple salting out procedure for extracting DNA from  
693 human nucleated cells. *Nucl Acids Res*, **16**, 1215.

694 Naro-Maciel E, Le M, FitzSimmons NN, Amato G (2008) Evolutionary relationships of marine  
695 turtles: A molecular phylogeny based on nuclear and mitochondrial genes. *Mol Phylogenet Evol*, **49**,  
696 659-662.

697 Nielsen R, Korneliussen T, Albrechtsen A, Li Y, Wang J (2012) SNP Calling, Genotype Calling, and  
698 Sample Allele Frequency Estimation from New-Generation Sequencing Data. *PLoS ONE*, **7**,  
699 e37558.

700 O'Leary SJ, Puritz JB, Willis SC, Hollenbeck CM, Portnoy DS (2018) These aren't the loci you're  
701 looking for: Principles of effective SNP filtering for molecular ecologists. *Mol Ecol*,  
702 <https://doi.org/10.1111/mec.14792>.

703 Popescu A-A, Huber KT, Paradis E (2012) ape 3.0: New tools for distance-based phylogenetics and  
704 evolutionary analysis in R. *Bioinformatics*, **28**, 1536-1537.



705 Prince DJ, O'Rourke SM, Thompson TQ, Ali OA, Lyman HS, Saglam IK, ..., Miller MR (2017)  
706 The evolutionary basis of premature migration in Pacific salmon highlights the utility of genomics  
707 for informing conservation. *Science Advances*, **3**, e1603198.

708 Quinlan AR, Hall IM (2010) BEDTools: a flexible suite of utilities for comparing genomic features.  
709 *Bioinformatics*, **26**, 841-842.

710 Rees A, Alfaro-Shigueto J, Barata P, Bjorndal K, Bolten AB, Bourjea J, ..., Godley BJ. (2016) Are  
711 we working towards global research priorities for management and conservation of sea turtles?  
712 *Endang Species Res*, **31**, 337-382.

713 Reis EC, Soares LS, Lôbo-Hajdu G (2010) Evidence of olive ridley mitochondrial genome  
714 introgression into loggerhead turtle rookeries of Sergipe, Brazil. *Conserv Genet*, **11**, 1587-1591.

715 Robinson JT, Thorvaldsdóttir H, Winckler W, Guttman M, Lander ES, Getz G, Mesirov JP (2011)  
716 Integrative Genomics Viewer. *Nature Biotechnol*, **29**, 24-26.

717 Rohland N, Reich D (2012) Cost-effective, high-throughput DNA sequencing libraries for  
718 multiplexed target capture. *Gen Res*, **22**, 939-946.

719 Romiguier J, Gayral P, Ballenghien M, Bernard A, Cahais V, Chenuil A, ..., Galtier N (2014)  
720 Comparative population genomics in animals uncovers the determinants of genetic diversity. *Nature*,  
721 **515**, 261-U243.

722 Sambrook J, Fritsch EF, Maniatis T (1989) Molecular cloning: a laboratory manual. Cold Spring  
723 Harbour Laboratory Press, New York.

724 Schliep KP (2011) phangorn: phylogenetic analysis in R. *Bioinformatics*, **27**, 592-593.

725 Seminoff J, Allen C, Balazs G, Dutton P, Eguchi T, Haas H, ..., Waples RS (2015) Status Review of  
726 the Green Turtle (*Chelonia mydas*) under the U.S. Endangered Species Act. NOAA Technical  
727 Memorandum, NOAA-NMFS-SWFSC-539. pp. 1-571, Silver Spring, MD.

728 Shafer ABA, Wolf JBW, Alves PC, Bergström L, Bruford MW, Brännström I, ..., Zieliski P (2015)  
729 Genomics and the challenging translation into conservation practice. *TREE*, **30**, 78-87.

730 Shaffer H, Gidis M, McCartney-Melstad E, Neal K, Oyamaguchi H, Tellez M, Toffelmier E (2015)  
731 Conservation genetics and genomics of amphibians and reptiles. *Annu Rev Anim Biosci*, **3**, 113–138.

732 Shaffer HB, Minx P, Warren DE, Shedlock AM, Thomson RC, Valenzuela N, ..., Wilson RK (2013)  
733 The western painted turtle genome, a model for the evolution of extreme physiological adaptations  
734 in a slowly evolving lineage. *Genome Biol*, **14**, R28.

735 Shamblin BM, Bolten AB, Abreu-Grobois FA, Bjorndal KA, Cardona L, Carreras C, ..., Dutton PH  
736 (2014) Geographic Patterns of Genetic Variation in a Broadly Distributed Marine Vertebrate: New  
737 Insights into Loggerhead Turtle Stock Structure from Expanded Mitochondrial DNA Sequences.  
738 *PLoS ONE*, **9**, e85956.

739 Shamblin BM, Dodd MG, Griffin DB, Pate SM, Godfrey MH, Coyne MS, ..., Nairn CJ (2017)  
740 Improved female abundance and reproductive parameter estimates through subpopulation-scale  
741 genetic capture-recapture of loggerhead turtles. *Mar Biol*, **164**, 138.

742 Skotte L, Korneliussen TS, Albrechtsen A (2013) Estimating Individual Admixture Proportions  
743 from Next Generation Sequencing Data. *Genetics*, **195**, 693-702.

744 Stewart K, LaCasella E, Roden S, Jensen M, Stokes L, Epperly S, Dutton PH (2016) Nesting  
745 population origins of leatherback turtles caught as bycatch in the U.S. pelagic longline fishery.  
746 *Ecosphere*, **7**, e01272.

747 Subramanian S (2016) The effects of sample size on population genomic analyses--implications for  
748 the tests of neutrality. *BMC Genomics*, **17**, 123.

749 Tajima F (1989) Statistical method for testing the neutral mutation hypothesis by DNA  
750 polymorphism. *Genetics*, **123**, 585–595.

751 Todd E, Black M, Gemmell N (2016) The power and promise of RNA-seq in ecology and  
752 evolution. *Mol Ecol*, **25**, 1224-1241.

753 Tzika A, Ullate-Agote A, Grbic D, Milinkovitch M (2015) Reptilian Transcriptomes v2.0: An  
754 Extensive Resource for Sauropsida Genomics and Transcriptomics. *Genome Biol Evol*, **7**, 1827–1841.

- 755 Vieira FG, Lassalle F, Korneliussen TS, Fumagalli M (2016) Improving the estimation of genetic  
756 distances from Next-Generation Sequencing data. *Biol J Linnean Soc*, **117**, 139-149.
- 757 Vilaça ST, Vargas SM, Lara-Ruiz P, Molfetti É, Reis EC, LÔBo-Hajdu G, Soares LS, Santos FR  
758 (2012) Nuclear markers reveal a complex introgression pattern among marine turtle species on the  
759 Brazilian coast. *Mol Ecol*, **21**, 4300-4312.
- 760 Wang Z, Pascual-Anaya J, Zadissa A, Li WQ, Niimura Y, Huang ZY, ..., Irie N (2013) The draft  
761 genomes of soft-shell turtle and green sea turtle yield insights into the development and evolution of  
762 the turtle-specific body plan. *Nature Genetics*, **45**, 701–706.
- 763 Watterson G (1975) On the number of segregating sites in genetical models without recombination.  
764 *Theor Popul Biol*, **7**, 256–276.
- 765 Wyneken J, Lohmann K, Musick J eds. (2013) *Biology of Sea Turtles Volume III*. CRC Press, Boca  
766 Raton, FL.

767

#### 768 **Data Accessibility**

769 Data analyses scripts, documentation and Rapture platform probe sequences are available at  
770 [https://github.com/lkomoro/Marine\\_Turtle\\_Rapture\\_Methods](https://github.com/lkomoro/Marine_Turtle_Rapture_Methods). Illumina raw reads for initial  
771 RAD-Seq of leatherback turtles used to design custom MyBaits kit and Rapture hardshell turtles are  
772 deposited in NCBI Sequence Read Archive (Bioproject PRJNA487648).

773

#### 774 **Author Contributions**

775

776 LMK, MM, SO, MPJ, KRS and PHD contributed to the conceptual design of the project. LMK,  
777 MM and SO conducted laboratory, marker design, and data analyses. LMK, MPJ, KRS and PHD  
778 assessed data interpretation for green turtles, and LMK and PHD wrote the manuscript.

779 Table 1. Initial SNP discovery per species with Rapture data for all QC passed samples (filters of MAF 0.05-0.4 and only sites with data for at least 50%  
780 individuals). Factors such as filtering thresholds, number of input samples, and source population of samples can affect identification of SNPs that are  
781 informative for different study goals.

782

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783	<b>Species</b>	<i>C. mydas</i>	<i>C. caretta</i>	<i>E. imbricata</i>	<i>L. olivacea</i>	<i>L. kempii</i>	<i>D. coriacea</i> <sup>†</sup>	<i>D. coriacea</i> <sup>‡</sup>
784	<b>No. Ind.</b>	47	23	34	6	4	973	203
785	<b>No. SNPs</b>	11042	4502	6514	2048	1542	2835	2710

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786 <sup>†</sup> All QC passed samples, global representation

787 <sup>‡</sup> St. Croix nesting population QC passed samples

788 Figure 1. Panel (A) depicts the on-target proportion per individual after the removal of PCR clones of (1) reads aligned to the green turtle genome (blue  
789 circles) and (2) bases aligned on or near baited Rapture regions (yellow circles). Note that one over-sequenced outlier with >7 million sequenced  
790 fragments was removed to improve visual interpretation. Panel (B) depicts the proportion of filtered mapped alignments/total sequenced fragments per  
791 individual for each category of initial DNA concentration (ng/ul; 'variable' category ranged from <5 to 30 ng/ul).

792  
793 Figure 2. (A) Relationship between the number of sequenced fragments per individual and the number of *a priori* SNP loci genotyped, and (B) the  
794 relationship between the SNP relative position within a Rapture locus and the number of samples genotyped (visualized with 80% posterior probability  
795 threshold). Vertical lines added at relevant thresholds for visual interpretation (see text).

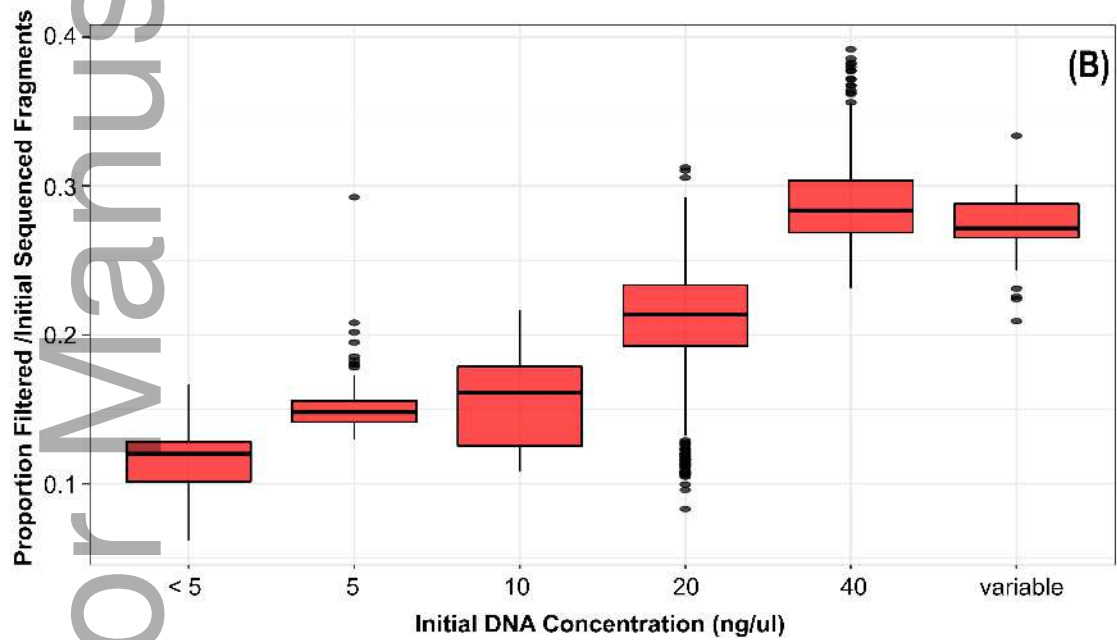
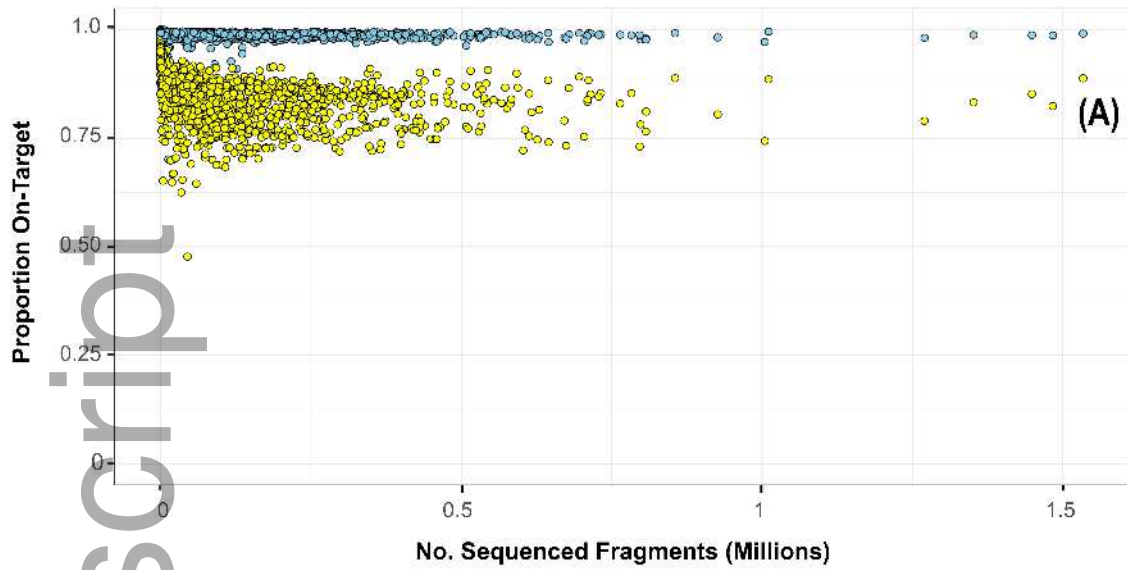
796  
797 Figure 3. (A) Number of Rapture loci covered  $\geq 4x$  for all samples (one over-sequenced outlier with >1 million filtered alignments removed to improve  
798 visual interpretation); (B) depicts hardshell turtles to better visualize that only green turtles and green-hybrids attain coverage at all Rapture loci.

799  
800 Figure 4. Species confirmation in hardshell turtles using principal components analyses (panels A and B) and admixture proportions (panels C and D).  
801 Panels (A) and (C) include all hardshell samples, while (B) and (D) include only of subsets of smaller groups, demonstrating how delineations among  
802 closer-related groups with smaller sample sizes can be masked in larger, disproportionate datasets. Only unresolved hybrids from the complete data set  
803 depicted in Panels A and C are included in Panels B and D.

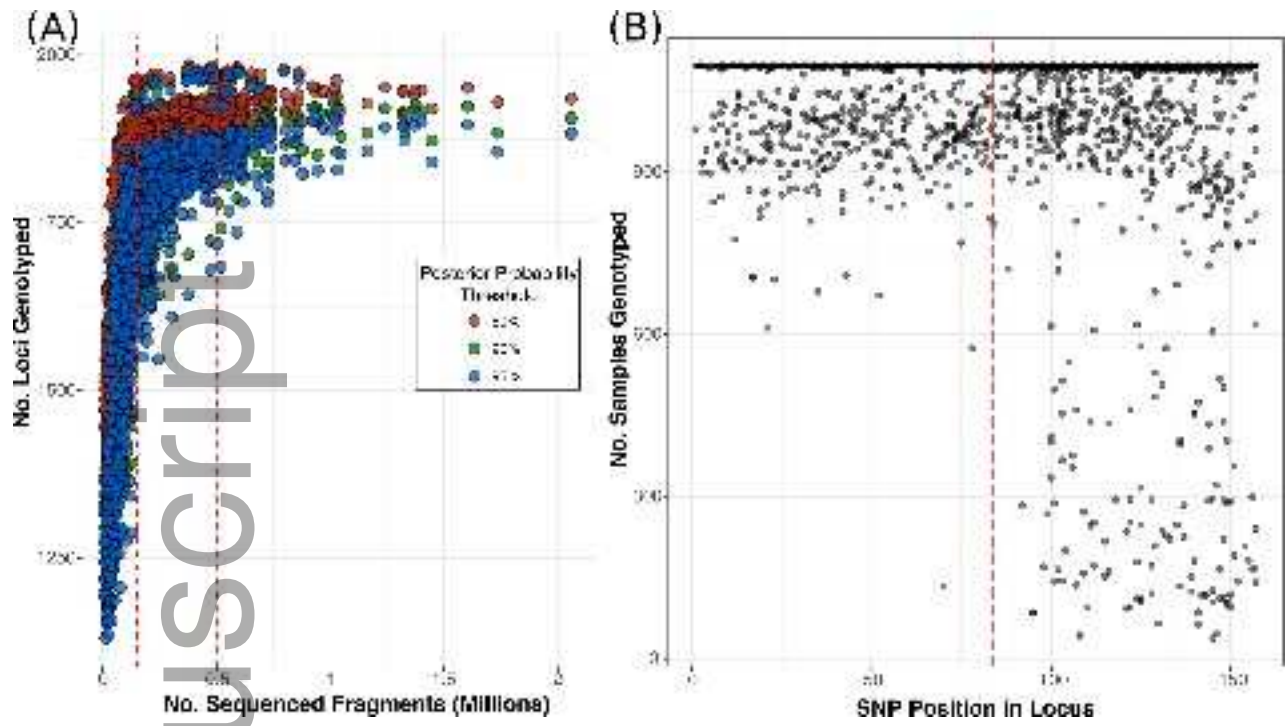
804  
805 Figure 5. (A) Pairwise  $F_{st}$  values between green turtle nesting regions (sample sizes listed in italicized parentheses; black boxes indicates values could not  
806 be reliably calculated due to low sample size and sequencing coverage). (B) *FastME* tree of a representative subset of green turtle samples with topology  
807 and relative branch length based on genetic distances estimated in *ngsDist*. Branch support based on bootstrapping (1000 replicates, blocks of 500  
808 SNPs). Abbreviations: STX=St. Croix, FFS=French Frigate Shoals, RMI= Republic of the Marshall Islands, FSM= Federated States of Micronesia.

809

810 Figure 6. Genetic diversity estimates (top: Watterson's estimator  $\theta_w$ ; bottom: Tajima's estimator  $\theta_\pi$ ) in representative groups for each species. Locations  
811 listed indicate nesting population with the exception of *L. olivacea* for which only bycatch samples with unknown nesting origin were available.

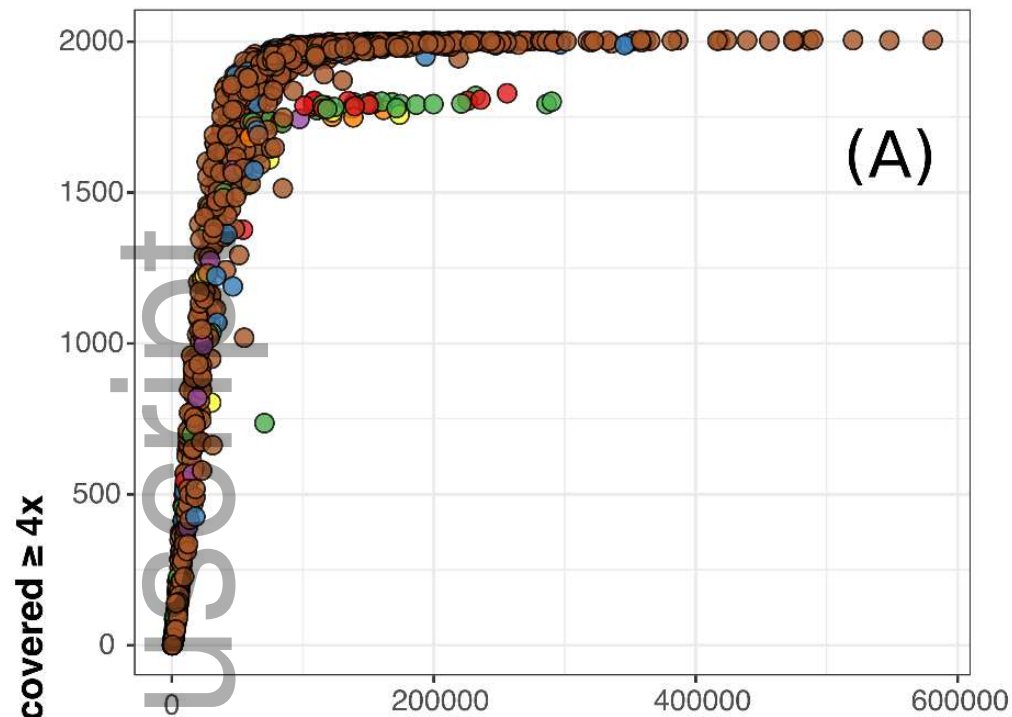


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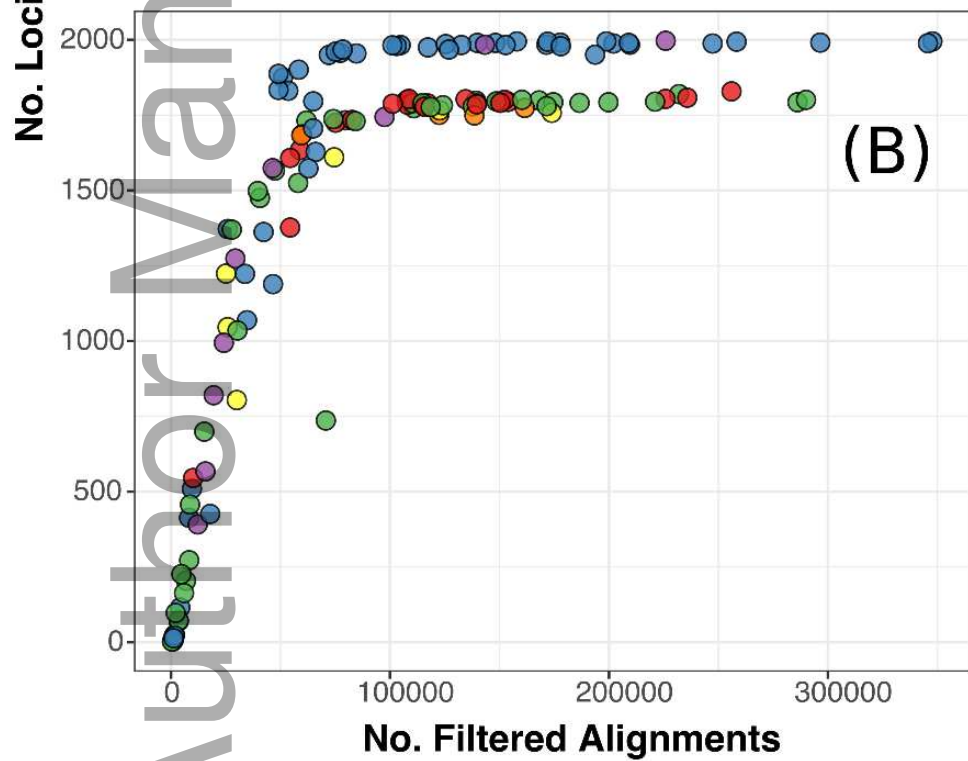
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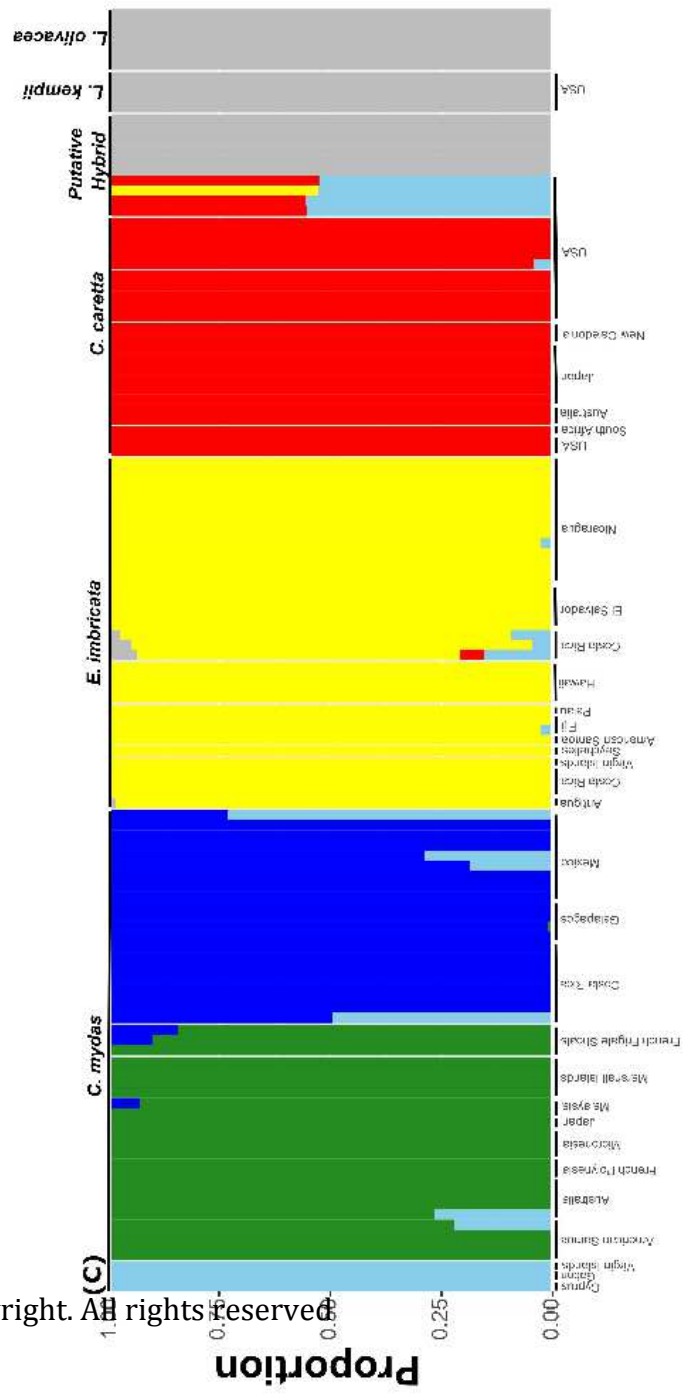
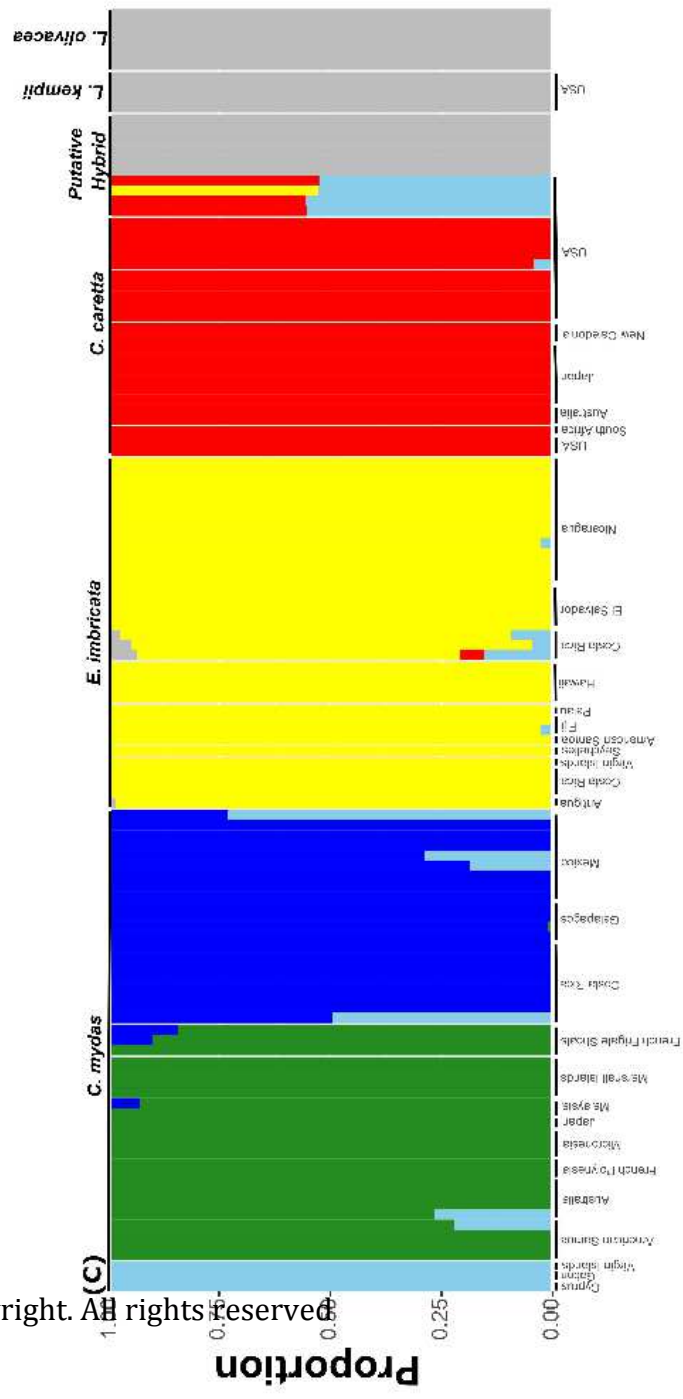
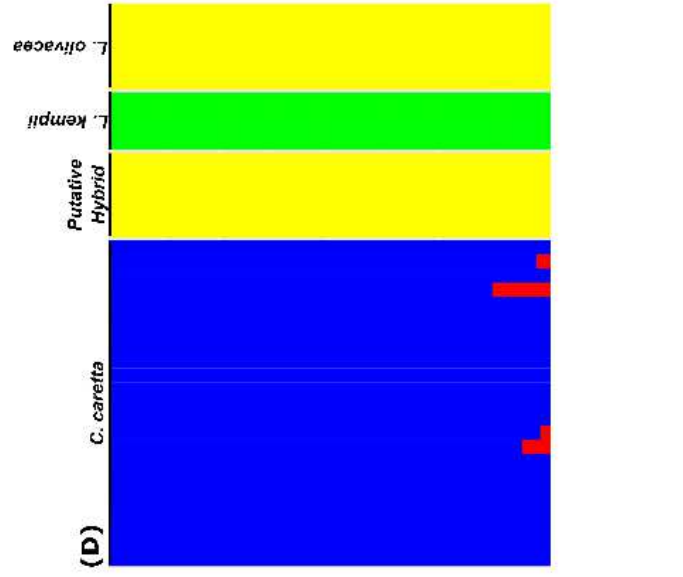
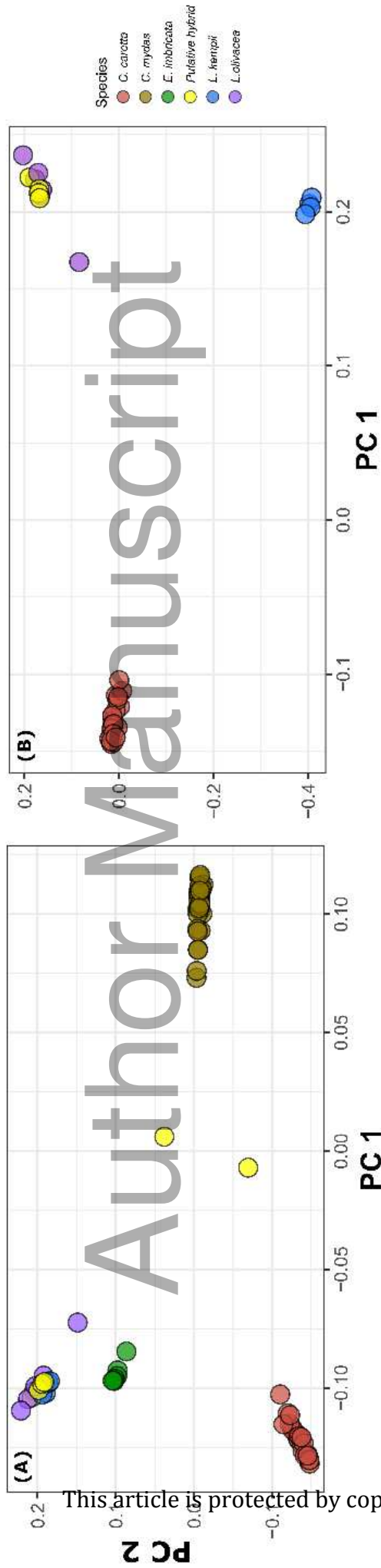


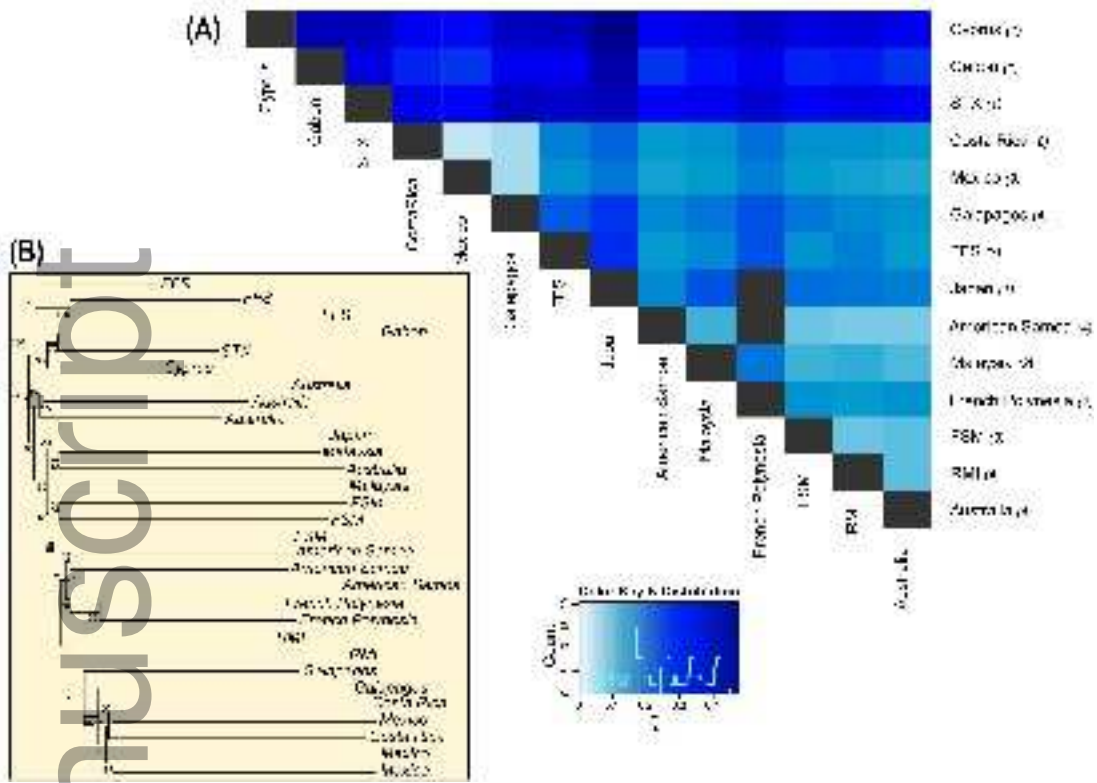
**Species**

- *Loggerhead turtle*
- *Green turtle*
- *Hawksbill turtle*
- *Putative Hybrid*
- *Kemp's ridley turtle*
- *Olive ridley turtle*
- *Leatherback turtle*



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men\_12980\_f5.png

