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Running head: Geographic convergence during stopover

**Range-wide populations of a long-distance migratory songbird converge during stopover in
the tropics**

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9 **Abstract.**

10 Geographic convergence during migration influences the extent to which animal populations
11 may experience carry-over effects across periods of the annual cycle. When most individuals of a
12 population share geographic areas during a given period, carry-over effects are likely stronger
13 than when individuals occupy multiple areas. We used genetic data and stable isotope ($\delta^2\text{H}$)
14 measurements from feathers and claws to describe the likely breeding and wintering geographic
15 origins of a long-distance migratory songbird (Gray-cheeked Thrush, *Catharus minimus*) moving
16 through northern Colombia in spring and fall migration. Furthermore, we used these data
17 coupled with regional occupancy surveys to assess whether individuals from various breeding
18 populations converge during migration, and evaluated whether geographic origin, age, or sex
19 affected stopover strategies. We found that range-wide breeding populations of Gray-cheeked
20 Thrush converged in northern Colombia in an area spanning less than 1% of the breeding range,
21 especially during a prolonged spring stopover in the Sierra Nevada de Santa Marta. Breeding
22 (but not winter) origin, sex, and age, influenced arrival date and body condition upon arrival at
23 stopover sites where populations converged. Birds from more northerly breeding latitudes, males
24 and adults generally arrived earlier and in lower body condition than those with more southerly
25 breeding origins, females and juveniles. Our work and other studies suggest that areas in
26 northern Colombia may function as ecological bottlenecks for Gray-cheeked Thrush because
27 they concentrate individuals from across the breeding range, provide critical resources, and
28 impose constraints during migration. Future studies quantifying the effects of high-convergence
29 areas on fitness and survival of individuals and their demographic consequences are required to
30 assess their roles as ecological bottlenecks.

31 **Key words.-** Bottleneck, *Catharus minimus*, Colombia, deuterium, migratory connectivity, ND2,
32 occupancy, South America, stable isotopes.

33

34 Migratory animals show high geographic convergence when large proportions of populations
35 concentrate in the same area during periods of their annual cycle (Iwamura et al. 2013). For
36 example, geographical constraints may result in large numbers of individuals travelling through
37 and stopping over in certain regions (e.g. small or narrow land masses, low mountain passes), or
38 spatial variation in resource availability may force animals to concentrate in a few high-quality
39 areas to fuel (Buehler and Piersma 2008, Bayly et al. 2017, Cohen et al. 2017). Within these
40 restricted areas, which may be obligatory stopovers, there may be access to high-quality
41 resources or safe flyways, but there may also be heightened competition, predation risk, or
42 exposure to diseases (Buehler and Piersma 2008, Bauer et al. 2016). Therefore, opportunities and
43 constraints at high-convergence areas may enhance positive or negative carry-over effects,
44 significantly influencing individual fitness and population dynamics (Webster and Marra 2005,
45 Boulet and Norris 2006, Betini et al. 2015, Cohen et al. 2018)

46

47 In areas of high geographic convergence, spatial constraints may act simultaneously with
48 temporal constraints. For instance, during spring migration, animals experience strong time
49 constraints if they are to reach their breeding grounds on time to obtain the best territories and
50 mates (Velmalala et al. 2015). Yet, migrants must also invest sufficient time accumulating fuel at
51 stopovers to safely reach their next destination (Hedenström and Alerstam 1997), not become so
52 heavy that mass hinders their ability to avoid predators (Pomeroy 2006), and not arrive too early
53 so as to risk starvation in potentially barren breeding areas (Piersma 2002, Alerstam 2011).
54 Furthermore, because constraints may change over time in relation to food availability or
55 climate, the timing of passage through a convergence area could be limiting for some
56 populations but not for others depending on their migratory strategy (Bauer et al. 2016). Any
57 change disrupting the delicate equilibrium of adaptations involved in migration (e.g. protection
58 or degradation of a high-convergence area imposing multiple selection pressures on individuals)
59 could therefore result in a significant boost or toll on survival through carry-over effects on
60 arrival time to breeding grounds and breeding success (Buehler and Piersma 2008, Iwamura et al.
61 2013, Piersma et al. 2016, Studds et al. 2017). Determining whether species show high
62 geographic convergence in their annual cycle is thus an essential first step to identify potential

63 ecological bottlenecks for populations, and is central to understanding the ecology and guiding
64 the conservation of migratory species.

65

66 In birds, high convergence during migration is common among shorebirds (Rogers et al.
67 2010, Iwamura et al. 2013, Piersma et al. 2016, Studds et al. 2017) and waterfowl (Drent et al.
68 2007, Hübner et al. 2010), with major consequences for population dynamics via carry-over
69 effects (Rogers et al. 2010, Iwamura et al. 2013, Piersma et al. 2016, Studds et al. 2017). For
70 most songbirds, however, we lack information about the range-wide connectivity of populations
71 (Faaborg et al. 2010, Cresswell 2014, O'Connor et al. 2014) and thus we know little about the
72 extent to which periods of the annual cycle may represent bottlenecks resulting from high
73 convergence in particular areas (Rodenhouse et al. 2003, Leyrer et al. 2013, Rockwell et al.
74 2016, Bayly et al. 2017). Not only do we lack knowledge about geographic convergence, but
75 also about the constraints that individuals face in areas of high convergence when they exist.

76

77 Because the distance to be travelled and the time spent fueling largely determine the
78 duration of migration in birds (Alerstam and Lindström 1990, Cohen et al. 2014, Schmaljohann
79 and Eikenaar 2017), stopover behavior is intricately related to migratory ecology (Alerstam et al.
80 2003, Alerstam 2011). Depending on total migration distance, the quality and availability of
81 food, and the energy required to complete migratory flights, birds may attempt either long
82 stopovers and extensive fueling followed by long flights, or short stopovers with minimal fuel
83 replenishment followed by shorter flights (Tsipoura and Burger 1999, Buler et al. 2007,
84 Seewagen et al. 2013). Therefore, individuals from different populations may show different
85 stopover strategies when they converge in particular areas depending on breeding origin or
86 wintering destination (Paxton and Moore 2017). Such differences may be temporal, if individuals
87 from certain populations arrive or leave the stopover before others (Warnock et al. 2004, Bauer
88 et al. 2016). Alternatively, differences in stopover strategy may have a physiological basis, such
89 that individuals from different populations differ in body condition on arrival and in their
90 subsequent fueling behavior (how fast and for how long they gain fuel during stopover) due to
91 conditions experienced on their breeding or wintering grounds (Battley et al. 2012, Paxton and
92 Moore 2017).

93

94 Fuel load on arrival (i.e. the stored energy possessed by birds on arrival at a stopover site)
95 also likely influences the stopover strategy that individuals adopt (Weber and Houston 1997,
96 Alerstam 2011). For instance, birds with low fuel on arrival have likely made longer flights and
97 will require longer stopovers to prepare for the next leg of their journey (Alerstam 2011).
98 Differences in stopover strategies may also exist between males and females or between adults
99 and immature birds given that individuals of different sex or age may face dissimilar constraints
100 or may use areas differing in quality during migration (Cooper et al. 2009, 2011).

101
102 We used an integrative approach to examine geographic convergence of populations
103 across the annual cycle of the Gray-cheeked Thrush (*Catharus minimus*), a Nearctic-Neotropical
104 long-distance migratory songbird in which at least some populations have declined precipitously.
105 Populations of the western breeding subspecies *C. m. aliciae* appear stable whereas populations
106 of the eastern subspecies *C. m. minimus* have shown recent, drastic declines (Whitaker et al.
107 2015, FitzGerald 2017, FitzGerald et al. 2017). We combined regional estimates of occupancy
108 rates based on field surveys with genetic and stable-isotope ($\delta^2\text{H}$) analyses of tissue samples
109 collected at spring and fall stopover sites in northern Colombia to determine the extent of
110 geographic convergence at previously identified stopover sites (Bayly et al. 2013, Gomez et al.
111 2013), and to test the hypothesis that birds may adopt different stopover strategies depending on
112 their breeding origin and wintering destinations. As with other recent studies unraveling
113 connections among populations of migratory species (e.g. Norris et al., 2006; Rushing et al.,
114 2013; Stanley et al., 2014; Hostetler et al., 2015), we used stable isotopes and genes to infer
115 breeding and wintering origins of individuals and then examined potential links between such
116 origins and the stopover behavior of birds.

117 **Methods**

118 **Study system.-** Gray-cheeked Thrush breed in high-latitude taiga and tundra from eastern Siberia
119 to Newfoundland in Canada (Lowther et al. 2001, FitzGerald 2017). Subspecies *C. m. minimus*
120 breeds in Newfoundland and southeastern Labrador and is listed as threatened under the
121 Newfoundland and Labrador Endangered Species Act because of precipitous population declines
122 in recent decades (SSAC 2010, Environment Canada 2014, Whitaker et al. 2015, FitzGerald et
123 al. 2017). In contrast, *C. m. aliciae* breeds from Siberia across North America to central Labrador

124 and is not known to be declining. Gray-cheeked Thrush follow a loop migration, entering South
125 America in the fall mainly via the Isthmus of Panama and the Darién region in the north west,
126 and leaving South America in the spring by crossing the Caribbean Sea from the Sierra Nevada
127 de Santa Marta (Bayly et al. 2013, Gómez et al. 2013, 2014, 2017). Evidence exists for multi-day
128 stopovers in both regions and energy accumulated at spring stopovers in northern Colombia fuels
129 flights in excess of 3000 km (Bayly et al. 2013, Gómez et al. 2017). The winter range of the
130 species is poorly known, but most available records show birds wintering mainly in Amazonia
131 and in the northern Andes (Lowther et al. 2001, eBird 2014, Ungvari-Martin et al. 2016).
132 Furthermore, during 10 years of research along the length of the northern coast of Colombia, we
133 have not had any records of wintering Gray-cheeked Thrush in that region. However, a recent
134 tracking study suggests some individuals from Newfoundland spent the winter in northern
135 Colombia, probably at very low densities (Whitaker et al. *in press*). There is low but significant
136 mitochondrial DNA differentiation between subspecies (Topp et al. 2013, FitzGerald et al.
137 2017), but there is no information regarding potential differences between them in migratory
138 routes, wintering, or stopover areas.

139 ***Fieldwork and sample collection.*** - We captured several Gray-cheeked Thrush during fall
140 migration (September – October 2014 and 2015, n = 499) in the Darién region near the
141 Colombia-Panama border (Reserva Natural Tacarcuna, 08°39'44"N, 77°21'48"W) and during
142 spring migration (April – May 2015 and 2016, n = 877) in the Sierra Nevada de Santa Marta,
143 northern Colombia (Hacienda La Victoria, 11°7'19.84"N, 74°5'34.14"W). All birds were banded
144 and aged according to plumage characteristics as either adult or immature (Pyle 1997, Johnson et
145 al. 2011). We also measured the wing chord (± 1 mm) and body mass (± 0.1 g) of each
146 individual to assess body condition.

147 We used values of stable-hydrogen isotope analyses ($\delta^2\text{H}$) from feathers and claws to
148 infer the breeding and wintering origins of birds migrating through northern Colombia. During
149 spring migration, we clipped the distal ends (~ 1 cm) of the first primary feather (P1) and
150 (~ 2 mm) of both hallux claws from a subset of captured individuals (n = 326). Because P1 is the
151 first feather to be molted by adult Gray-cheeked Thrush during the pre-basic molt after the
152 breeding season (Pyle 1997), it contains the isotopic signature most representative of the
153 breeding grounds in North America (Hobson et al. 2014). Bird claws grow continuously and tips

154 contain the combined isotopic information of the previous 2-4 months pre-capture (Hahn et al.
155 2014); the claw samples from Gray-cheeked Thrush captured during spring migration were
156 therefore expected to contain the isotopic signature of the wintering grounds occupied prior to
157 migration. All isotopic data are available through Dryad (doi:10.5061/dryad.7hk6523).

158 During both migration periods we also collected ~20 µl of blood from the brachial vein of
159 birds and stored samples on filter paper prior to genetic analyses (see below). Because Gray-
160 cheeked Thrush cannot be reliably sexed in the hand outside of the breeding period, blood
161 samples were also used for molecular sex determination (Griffiths et al., 1998; see Appendix S1:
162 Supplementary Methods for details).

163 ***Occupancy surveys during spring and fall migration.***- A first step to evaluate whether a region
164 represents a high-convergence area is to determine whether most individuals concentrate in
165 restricted geographic areas or if they are widely spread across the landscape. To examine this, we
166 carried out repeated surveys for Gray-cheeked Thrush across northern Colombia during the
167 spring and fall migrations of 2016. Repeated surveys (>4 per transect) were carried out every 10
168 days during two months in 200 transects, each 100 m long, spread across 17 sites spanning c.700
169 km from east to west (Fig. 1), for a total effort of 400 person/hours. We analyzed survey data in a
170 dynamic occupancy modelling framework (Fiske and Chandler 2011, Guillera-Aroita 2017)
171 allowing us to estimate the probability of occurrence of Gray-cheeked Thrush throughout the
172 region and to examine the influence of variables such as vegetation cover, precipitation and
173 elevation on the detectability and occupancy rate of the species during both migration seasons.
174 Detailed methods of data collection and occupancy modelling are described in Appendix S1.

175 ***Linking birds from the breeding range and stopover areas using genetics.***- To infer
176 connectivity of Gray-cheeked Thrush populations, we assessed the genetic similarity of birds
177 captured during migration in northern Colombia with individuals from known breeding
178 populations in North America. We compared sequences of the NADH dehydrogenase subunit 2
179 (ND2) gene obtained for 150 individuals captured at our field sites in Colombia with published
180 sequences from breeding areas in Siberia (n = 3), Alaska (n = 7), Quebec (n = 3), southern
181 Labrador (n = 12), western Labrador (n = 11), and Newfoundland (n = 39; FitzGerald et al. 2017;
182 see Appendix S1: Supplementary Methods and Table S1). Sequences were aligned using
183 Geneious version R7.1 (Kearse et al. 2012) and polymorphisms confirmed from the

184 chromatograms. We used two complementary methods to characterize genetic differences
185 between populations. Haplotype networks and pairwise population comparisons were used to
186 assess variation along the entire ND2 sequence (Bandelt et al. 1999), while presence of private
187 alleles was used to identify variation in unique positions among populations (Kekkonen et al.
188 2011).

189 We used POPART (Leigh and Bryant 2015) to estimate a minimum spanning network
190 depicting relationships among haplotypes (Bandelt et al. 1999). In addition, we used estimates of
191 pairwise population differentiation (ϕ_{st}) calculated using the Kimura 2P substitution model
192 (identified as the best fit to the data; FitzGerald et al. 2017) with 1000 permutations of the
193 haplotype matrix in Arlequin v. 3.5 (Excoffier et al. 2005), to evaluate similarities between
194 breeding regions (Newfoundland/southern Labrador [subspecies *C. m. minimus*], western
195 Labrador/Quebec [subspecies *C. m. aliciae*], Alaska/Siberia [subspecies *C. m. aliciae*]) and
196 stopover locations (Darién [fall] and Sierra Nevada de Santa Marta [spring]). We evaluated
197 whether the genetic composition of birds captured in Colombia differed by year or migration
198 period by employing a hierarchical analysis of molecular variance (AMOVA) to assess how
199 much genetic variation was attributed to stopover groupings (i.e. spring vs. fall and 2014 vs.
200 2015) from 20,000 permutations of the dataset. We also estimated allelic frequencies, identified
201 parsimony-informative alleles, and calculated the number of geographic regions in the breeding
202 range where alleles occurred. We then examined whether individuals captured in Colombia
203 shared any private alleles (i.e. those found only in 1-2 breeding regions) with birds of known
204 breeding origin. An assumption of this method is that a random sample of individuals from a
205 population will be roughly representative of the frequency of alleles present in that population
206 (Kekkonen et al. 2011). If most members of a population share a private allele, then it should
207 likely be detected even with relatively small sample sizes. Sharing of private alleles between
208 birds captured in Colombia and the breeding range was therefore taken as an indication of
209 linkage between populations (Kekkonen et al. 2011).

210 ***Assigning individuals to breeding and wintering grounds with stable-isotope analysis of***
211 ***feathers and claws.***- Stable-isotope analyses were carried out at the Stable Isotope Laboratory of
212 Environment and Climate Change Canada, Saskatoon, Canada. Feathers and claws were cleaned
213 of external oils and debris before analysis using a 2:1 chloroform:methanol soak and rinse.

214 Measurements of $\delta^2\text{H}$ in feather ($\delta^2\text{H}_f$) and claw ($\delta^2\text{H}_c$) were obtained through continuous-flow
215 isotope-ratio mass spectrometry (CF-IRMS) following the ‘comparative equilibration’ method
216 which uses calibrated keratin reference materials to correct for exchangeable hydrogen using the
217 keratin standards CBS (-197 ‰) and KHS (-54 ‰; Wassenaar and Hobson 2003). We report $\delta^2\text{H}$
218 values in standard delta (δ) notation, in parts per thousand (‰) relative to Vienna Standard Mean
219 Ocean Water (VSMOW). Based on within-run replicates (n=6) of calibration standards, we
220 estimate analytical error to be of the order of ± 2 ‰.

221 Employing data on species abundance in space and potential distribution models as
222 geographic priors may improve geographic assignments based on stable isotopes (Chabot et al.
223 2012, Fournier et al. 2017, Rushing et al. 2017b). For species like the Gray-cheeked Thrush,
224 which breed and winter in remote locations seldom visited by people, there is considerable
225 uncertainty about true distribution ranges and more so about patterns of spatial abundance. We
226 therefore combined tissue $\delta^2\text{H}$ data with a ‘spatiotemporal exploratory model’ (STEM; Fink et
227 al., 2010) showing relative abundance of Gray-cheeked Thrush during the breeding period based
228 on field observations and, separately with a season-specific (breeding and winter) species
229 distribution model (SDM) generated from field records and climatic variables (FitzGerald 2017,
230 Fournier et al. 2017). Abundance and probability of occurrence estimated by distribution models
231 are usually correlated and considered good proxies for predicting species presence (Weber et al.
232 2016). We evaluated the similarity between geographic assignments based on STEM and SDM
233 priors by estimating the spatial correlation between them. Methods for the construction of the
234 STEM and SDM models are in the Appendix S1. We also present assignments based only on $\delta^2\text{H}$
235 values to assess any biases derived from the priors (Rushing et al. 2017b).

236 ***Surfaces of likely origin based on stable isotopes.***- Isotopic data were used to derive $\delta^2\text{H}_f$ and
237 $\delta^2\text{H}_c$ isoscapes based on $\delta^2\text{H}_p$ during the growing season (Bowen et al. 2005), and a calibration
238 equation estimated for feathers of ground-foraging Neotropical migrants (Hobson et al. 2012):
239 $\delta^2\text{H}_f = 27.09 + 0.95 (\delta^2\text{H}_p)$. Prior to modeling, we removed outliers with $\delta^2\text{H}$ values more
240 than three standard deviations from the mean (feathers = 3, claws = 2; Chabot et al., 2012). Claw
241 $\delta^2\text{H}$ values are closely correlated with feather values but they require a slightly different
242 conversion factor varying by age (Hobson, *unpubl. data*). Therefore, to generate the claw
243 isoscape we first expressed $\delta^2\text{H}_c$ values in terms of equivalent expected $\delta^2\text{H}_f$ values using age-

244 specific claw-to-feather conversion equations derived for *Catharus* thrushes: for immature birds
 245 SY: $\delta^2H_f = -15.7 + 0.94 (\delta^2H_c)$, and for adults ASY: $\delta^2H_f = 24.1 + 1.3 (\delta^2H_c)$ (Hobson,
 246 *unpubl. data*). In addition, we needed to account for the fact that the winter of 2015/2016
 247 coincided with one of the strongest El Niño events ever recorded (Jiménez-Muñoz et al. 2016).
 248 This event produced anomalous droughts in the Amazon, where Gray-cheeked Thrush are
 249 expected to winter, giving rise to δ^2H values enriched by ~16‰ above the long-term average
 250 (Srygley et al. 2010, Brienen et al. 2012, Jiménez-Muñoz et al. 2016). To use our data with the
 251 available long-term δ^2H_p isoscape for the Amazon region (Bowen et al. 2005) while correcting
 252 for the effect of El Niño in 2015, we subtracted 16‰ from our measured δ^2H_c values before
 253 undertaking spatial assignments.

254 The probability of each individual originating from any given ~ 1 km² cell on the
 255 breeding or wintering grounds was estimated based on a normal probability density function:
 256 $f(y^*|\mu_i, \sigma) = \frac{1}{\sqrt{2\pi}\sigma} \exp\left[-\frac{1}{2\sigma^2}(y^* - \mu_i)^2\right]$, where $f(y^*|\mu_i, \sigma)$ is the likelihood that an individual
 257 with feather isotopic values $\delta^2H_f = y^*$ originated from cell i . μ_i is the predicted δ^2H_f value for cell
 258 i , and σ is the standard deviation of the δ^2H_f values within a breeding or wintering site, which
 259 was set at 10.8‰ (Hobson et al. 2012). Likelihood values were transformed into a probability
 260 surface by dividing each likelihood by the total sum of likelihoods (Hobson et al. 2012). A cutoff
 261 of the upper 67% of the probability was defined as the likely area of origin for each individual
 262 (Hobson et al. 2012, Chabot et al. 2012) and cells with values equal to or above this cutoff were
 263 set a probability value of 1 (likely origin), whereas those with lower values were set to 0
 264 (unlikely origin). The 67% cutoff provides a good compromise between assignment precision
 265 and overestimation error (Hobson et al. 2012, Chabot et al. 2012). Finally, using Bayes rule,
 266 $f(b|y) = \frac{f(y|b)f(b)}{\sum_b f(y|b)f(b)}$, we estimated the posterior probability $f(b|y)$ that any 1 km² cell on the
 267 map represented the origin of a bird, given $f(b)$, the prior probability of occurrence based on the
 268 abundance (STEM) or the distribution model (SDM), and $f(y|b)$, the likelihood of the δ^2H
 269 assignment surface (Chabot et al. 2012, Fournier et al. 2017). We finally overlaid the likely
 270 origins of all birds and obtained a cumulative surface for all the individuals in our sample.

271 ***Assessing the influence of breeding and wintering origins on stopover strategies.***- To evaluate
 272 whether stopover strategies of Gray-cheeked Thrushes were related to breeding or wintering

273 origin, we examined how fuel load on arrival to the stopover site varied as a function of date of
274 arrival, breeding or winter origin isotopic values ($\delta^2\text{H}_f$ and $\delta^2\text{H}_c$), age, and sex. We estimated
275 fuel load on arrival as the mass (g) above lean body mass (LBM) of each individual on its first
276 capture: *Fuel load on arrival = Mass on first capture – LBM*. Lean body mass was
277 estimated by regressing wing chords and body masses of all captured individuals with a fat score
278 of 0 (n = 135), which resulted in the equation $LBM = 0.33 \times \text{wing length} - 4.63$ (Gómez et
279 al. 2017). We used AIC values (Burnham and Anderson 2002) to evaluate 15 candidate models
280 plus a null model in which none of the independent variables influenced fuel load on arrival
281 (Table 1). To evaluate whether observed differences in breeding and wintering origin between
282 age and sex classes were significant, we carried out randomization tests, where the observed
283 difference in means between each group (males vs. females or adults vs. immatures) was
284 compared to a random distribution of 999 differences of equivalent datasets selected regardless
285 of age or sex.

286 **Results**

287 ***Occupancy rates during spring and fall migration.***- Occupancy surveys resulted in 159 and 264
288 presence records of Gray-cheeked Thrush during fall and spring migration, respectively, across
289 the 17 sites sampled along the Caribbean coast of Colombia. Spatial predictions of occupancy
290 rates of Gray-cheeked Thrush during spring and fall migration revealed marked concentrations of
291 individuals in the Darién region during fall (sum of area of grid cells with occupancy $\geq 0.5 =$
292 $80,045 \text{ km}^2$, Fig. 1A) and on the northern and western slopes of the Sierra Nevada de Santa
293 Marta in spring ($12,963 \text{ km}^2$, Fig. 1B), with intermediate areas showing very low probabilities of
294 occupancy. These two areas span only 5% and 0.8%, respectively, of areas occupied by Gray-
295 cheeked Thrush during the breeding season ($1,587,463 \text{ km}^2$, according to our isotopic
296 assignments). During fall migration, occupancy increased towards the west, in areas with low-
297 mid elevations, high canopy cover and a dense understory (Appendix S1: Table S2A). During
298 spring, occupancy was predicted to be highest in regions where intermediate levels of annual
299 precipitation coincided with elevations below $< 1700 \text{ m}$ in northernmost Colombia, which
300 largely correspond to the mountain slopes of the Sierra Nevada de Santa Marta (Appendix S1:
301 Table S2B).

302 **Linking birds from the breeding range and stopover areas using genetics.**-We found no
303 significant difference in the genetic composition of birds captured in Colombia between years
304 ($\phi_{st} = 0.010$, $P = 0.073$), with most of the genetic variation (96.6%) found within years ($P =$
305 0.667). Very low but significant differentiation was found between migration seasons (i.e. fall vs.
306 spring $\phi_{st} = 0.047$, $P < 0.001$), yet most of the variation (94.8%) existed within seasons ($P =$
307 0.330). ND2 ϕ_{st} pairwise population comparisons showed that most fall and spring birds in
308 Colombia were more genetically similar to individuals from subspecies *C. m. aliciae* (breeding
309 in Alaska/Siberia) than to *C. m. minimus* (Newfoundland/ S Labrador; Table 2). However, birds
310 captured in Colombia shared at least some haplotypes with birds from throughout the breeding
311 range (Fig. 2A). All Alaska haplotypes and most haplotypes from western Labrador and Quebec
312 were found in Colombia. Conversely, fewer of Newfoundland/S Labrador haplotypes were found
313 in Colombia. A large number of haplotypes found in birds captured in Colombia were not
314 observed in any of the birds of known breeding origin (Fig. 2A), likely reflecting a lack of
315 sampling in north-central Canada and low sample sizes in some of the other breeding areas.
316 However, of those haplotypes shared among birds from Colombia and the breeding grounds,
317 there were 24 informative sites and 14 of these corresponded to private alleles restricted to
318 individuals from only one or two of the sampled breeding populations. Furthermore, 119 out of
319 150 individuals from Colombia had alleles private to birds from all 6 breeding regions; out of
320 these, 8 individuals had alleles private to breeding sites in Newfoundland and southern Labrador
321 (Fig. 2B). Overall, this suggests that although most birds sampled in Colombia were more
322 genetically similar to birds from the western subspecies *C. m. aliciae*, individuals from most
323 breeding regions and both subspecies likely converged in Colombia.

324 **Geographic assignment through stable isotopes and geographic priors.**- In agreement with the
325 genetic data, geographic assignments based on $\delta^2\text{H}$ values and distribution and abundance
326 models suggested that Gray-cheeked Thrush migrating through northern Colombia likely
327 originated from multiple breeding populations (Fig. 3). Geographic assignment of breeding
328 origin using the STEM model as a prior showed a 60% correlation with that based on the species
329 distribution model (SDM), suggesting relatively close agreement between predictions
330 considering geographic variation in abundance and habitat suitability (Weber et al. 2016). The
331 two models predicted high numbers of individuals originating from Alaska and north-central
332 Canada, but the STEM model did not predict as many as the SDM from the eastern extreme of

333 the breeding range. Given the agreement between models, we present only the assignments based
334 on the SDM because they were available for both the breeding and the winter periods.
335 Geographic assignments to breeding areas based solely on isotopes predicted a high probability
336 of origin from three main regions: (1) Alaska, (2) an area bordering Hudson Bay in the provinces
337 of Nunavut and Manitoba, and (3) northern Quebec and Labrador (Appendix S1: Fig. S1).

338 Similarly, isotopic analyses predicted individuals originating from a wide range of
339 wintering locations mostly concentrated in northern Amazonia including portions of Ecuador,
340 Peru, Colombia, Venezuela, and Brazil. A smaller proportion of birds were also predicted to
341 have wintered at mid-elevations in the Andes of Colombia (Fig. 3).

342 ***Effect of date, age, sex and geographic origin, on stopover strategies.***- We found no spatial
343 differences in the inferred breeding or winter origin between adult and immature birds stopping
344 over in northern Colombia (97% and 100% Pearson correlation of breeding and winter surfaces,
345 respectively). However, adults had significantly higher $\delta^2\text{H}_f$ values compared to immature birds
346 (Appendix S1: Fig. S2, $P = 0.02$), suggesting a slightly more southerly breeding/molting origin
347 compared to immature birds. Adults also showed slightly lower values of $\delta^2\text{H}_c$ than immature
348 birds (Appendix S1: Fig. S2, $P = 0.03$). There were no obvious differences in breeding or
349 wintering origin between sexes (97% and 98% Pearson correlation of breeding and wintering
350 surfaces, respectively) although females showed slightly higher values of $\delta^2\text{H}_f$ than males
351 (Appendix S1: Fig. S3, $P = 0.06$), and both sexes showed equivalent winter ranges and values of
352 $\delta^2\text{H}_c$ (Appendix S1: Fig. S3, $P = 0.12$).

353 Date of arrival at the spring stopover site in Colombia varied as a function of breeding
354 but not winter origin (Fig. 4A-B. $\beta_{\text{breeding}} = 0.16 \pm 0.03$, $P < 0.001$; $\beta_{\text{winter}} = -0.05 \pm 0.04$, $P =$
355 0.22), with birds that bred at more northerly latitudes arriving ~10 days earlier on average.
356 Additionally, males arrived in Colombia significantly earlier than females (~5-10 days), during
357 both spring (Fig. 4C, $P = 0.001$) and fall migration (Fig. 4D, $P = 0.01$); similarly, adults arrived
358 in northern Colombia significantly earlier than immature birds during both migration periods
359 (Figs 4E-F, Spring: $P = 0.003$; Fall: $P < 0.001$).

360 Two candidate models explained 84% of the variation in fuel load on arrival at the spring
361 stopover site. Contributing variables included date of arrival, breeding origin, sex, and an
362 interaction between sex and breeding origin (Table 1). Both males and females arriving earlier

363 did so in lower body condition than birds arriving later in the season (Fig 5A, $\beta_{\text{male}} = 0.26 \pm 0.02$,
364 $P < 0.001$; $\beta_{\text{female}} = 0.21 \pm 0.03$, $P < 0.001$). In addition, males with lower $\delta^2\text{H}_f$ values (i.e. those
365 with more northerly breeding origin) arrived earlier and in lower body condition than those from
366 more southerly origins (Fig 5B, $\beta_{\text{males}} = 0.09 \pm 0.04$, $P = 0.02$). This association between
367 breeding origin on fuel loads was not significant for females (Fig 4B, $\beta_{\text{females}} = -0.03 \pm 0.06$, $P =$
368 0.56) and, if anything, showed the opposite tendency than that observed in males.

369 Discussion

370 Our occupancy surveys, genetic comparisons, and stable-isotope analyses suggest that Gray-
371 cheeked Thrush from across the breeding and wintering range funnel through two regions of
372 northern Colombia during migration: the Darién in the fall and the Sierra Nevada de Santa Marta
373 in the spring. Geographic convergence is particularly strong in spring, when individuals from
374 most of the breeding range appear to stopover exclusively in the Sierra Nevada de Santa Marta
375 prior to undertaking non-stop flights (>2500 km) across the Caribbean Sea and Gulf of Mexico
376 (Fig. 1; Bayly et al. 2013, Gómez et al. 2017). Regions occupied in northern Colombia during
377 fall and spring had areas equivalent to only 5% and 0.8%, respectively, of the area used during
378 the breeding season, highlighting the extreme degree of convergence.

379

380 Based on the above results, we hypothesize that the high-convergence area in the Sierra
381 Nevada de Santa Marta likely functions as an ecological bottleneck for populations of the Gray-
382 cheeked Thrush. Given that fat gained in the region may fuel 40% of total migration distance
383 covered by Gray-cheeked Thrush (Bayly et al. 2013) and that events at this spring stopover site
384 carry-over to influence the pace of migration and potentially breeding success (Gómez et al.
385 2017), it is likely that there are fitness and survival consequences associated to a successful
386 spring stopover in northern Colombia (Bayly et al. 2012, 2013, Gómez et al. 2017). Future
387 research should verify whether this is the case for Gray-cheeked Thrushes as well as other
388 migratory species.

389

390 Because our work and other studies (e.g. González-Prieto et al., 2011; Fraser et al., 2013)
391 suggest that stopover regions where range-wide populations converge have the potential to

392 function as ecological bottlenecks, a priority for the conservation of populations of migrants
393 worldwide would be to identify and protect such areas. Most studies on migratory connectivity to
394 date have addressed connections between the breeding and wintering grounds (e.g. Rushing et
395 al., 2013; Hobson et al., 2014; Fraser et al., 2017), and a recent meta-analysis suggested that low
396 migratory connectivity between breeding and wintering populations is common among long-
397 distance migrants (Finch et al. 2017). Fewer studies have examined connectivity during
398 migration (e.g. Hübner et al., 2010; Laughlin et al., 2013; Andueza et al., 2014; Stanley et al.,
399 2014), but these studies consistently found high convergence of populations, particularly at
400 stopover sites before barriers or at obligatory flyways within narrow land masses (Buehler and
401 Piersma 2008, Fort et al. 2009, Piersma et al. 2016). Furthermore, for a few songbird species,
402 mortality rates are known to be higher during migration than any other phase of the annual cycle
403 (Sillett and Holmes 2002, Rockwell et al. 2016, Paxton et al. 2017, Rushing et al. 2017a). This,
404 combined with the recent finding that multi-day stopovers occur in just a few key regions in
405 many species (Bonter et al. 2009, Stanley et al. 2014), means that migration is probably the most
406 vulnerable period in the annual cycle across a range of taxa (Faaborg et al. 2010). Therefore,
407 more research effort should be directed to identify convergence areas during migration and
408 assess their demographic effects on the populations of migrants worldwide.

409
410 A question arising from our study is whether differences in population declines between
411 *C. m. aliciae* and the endangered *C. m. minimus* may be connected with events taking place in
412 the high-convergence area we identified. Given that most variation in $\delta^2\text{H}_f$ values in the Northern
413 Hemisphere is described by latitude, breeding origin isotopic assignments based on hydrogen
414 have low longitudinal precision (Hobson et al. 2012). It is therefore difficult to discriminate with
415 certainty individuals originating from eastern and western populations in northern North
416 America based solely on $\delta^2\text{H}_f$ values (Appendix S1: Fig. S2, but see Cárdenas-Ortiz et al., 2017).
417 However, by combining isotopic assignments with genetic information, we confirmed that birds
418 captured in northern Colombia have likely origins from across the breeding range of the Gray-
419 cheeked Thrush, with most individuals probably belonging to populations of *C. m. aliciae*. The
420 apparently lower proportion of *C. m. minimus* in our samples may reflect either a differential
421 migratory strategy between subspecies, that the range of *C. m. minimus* is smaller, or that there

422 are fewer *C. m. minimus* due to population declines (Whitaker et al. 2015, FitzGerald et al.
423 2017). Spatial or temporal differences in migratory strategy between populations may determine
424 whether a high-convergence site is also a bottleneck for some or all populations (Iwamura et al.
425 2013, Bauer et al. 2016). Tracking of individuals from different subspecies and using genomics
426 to more finely characterize population structure and hence map geographic origins with greater
427 resolution, are promising avenues of research to start testing these hypotheses.

428 Differences in population dynamics may not be related to events occurring on migration
429 but instead depend more on conditions experienced during the winter (Wilson et al. 2011, 2018,
430 Rockwell et al. 2016). We found that breeding but not winter (isotopic) origin influenced the
431 spring migratory strategies of individuals: birds that bred further north migrated earlier and
432 arrived at the spring stopover site in poorer body condition than those from more southern
433 breeding origins. This may reflect chain migration, a pattern where migration occurs in a
434 latitudinal sequence between populations (Lundberg and Alerstam 1986, Kelly 2006, González-
435 Prieto et al. 2011). However, because the isotopic uniformity of Amazonia limits the resolution
436 with which latitudinal patterns can be detected in that region (Bowen et al. 2005), we could not
437 determine whether northern breeding populations indeed winter in more northerly regions.
438 Although our results did not show an effect of winter origin, we cannot rule out that it exists.
439 Other studies have consistently shown that winter habitat use has significant carry-over effects
440 on breeding success by affecting behavior during migration and date of arrival to the breeding
441 grounds (Marra and Holmes 2001, Norris et al. 2004, Norris and Marra 2007, McKinnon et al.
442 2015). Furthermore, given that we only made inferences of likely geographic origin and not of
443 variation in winter habitat quality, we cannot rule out that winter habitat may affect individual
444 performance during stopover in Gray-cheeked Thrush.

445
446 We also documented differences in migration timing between age classes, with adult
447 thrushes migrating earlier than immatures during both fall and spring. Moreover, adults had less
448 negative breeding ground values of $\delta^2\text{H}_f$ than immature birds, suggesting that age classes have
449 either different hydrogen isotope precipitation-to-feather discrimination factors (Hobson and
450 Clark 1992, Rushing et al. 2013) or that Gray-cheeked Thrush exhibit a molt-migration whereby
451 adults disperse from their breeding sites prior to molting – a pattern primarily known for

452 migratory birds from western North America (Rohwer et al. 2005). Differences between age
453 classes in migration phenology have been linked to individual experience because adults may
454 navigate more accurately (Mitchell et al. 2015), fuel up faster (Yong and Moore 1994), make
455 fewer stopovers along the way, or take riskier routes (McKinnon et al. 2014, Dossman et al.
456 2016) than first-year birds. Alternatively, first-year birds may migrate more slowly because they
457 do not benefit from early arrival to the breeding grounds to the same extent as adults (Stewart et
458 al. 2002, Cooper et al. 2009). Even if young birds arrive early they are likely to be outcompeted
459 by adults and still face the perils of variable climatic conditions and increased risk from agonistic
460 interactions with adults early in the season (Cooper et al. 2009). A recent study on Gray-cheeked
461 Thrush found no effect of age on fueling rates (Gómez et al. 2017), but more study is needed to
462 evaluate age differences in navigation skills or route use in this species.

463 We found differences in migration phenology between sexes, with males migrating
464 earlier than females in the spring. Such differences in other species have been attributed to a
465 stronger selection pressure acting on males to arrive earlier than females to the breeding grounds
466 (Kokko et al. 2006, Cooper et al. 2009, 2011, Coppack and Pulido 2009, Schmaljohann et al.
467 2015), such that males that migrate faster arrive earlier and therefore settle in better territories
468 and get the best mates. The ‘mate opportunity hypothesis’ posits that early arrival improves the
469 probability of mate acquisition for males more so than for females, given that most females will
470 breed regardless of arrival date, whereas late-arriving males will likely fail to breed in a given
471 year (Kokko et al. 2006, Cooper et al. 2009, 2011). More study is needed to evaluate sex ratios of
472 Gray-cheeked Thrush on the breeding grounds as well as the strength of sexual selection to see
473 whether this hypothesis is a plausible explanation for our observations. Segregation of sexes by
474 latitude on the winter grounds may also contribute to protandry, when males winter closer to the
475 breeding grounds (Woodworth et al. 2016), yet we did not detect any geographical differences in
476 wintering origin between sexes.

477 Overall, our study emphasizes the importance of northern Colombia and particularly of
478 the Sierra Nevada de Santa Marta for Gray-cheeked Thrush during migration given that
479 individuals from throughout the breeding range converge there during a critical period of the
480 annual cycle. Furthermore, we showed how breeding geographic origin, age and sex influence
481 the migratory strategies of individual Gray-cheeked Thrush during stopover. Verifying whether

482 this high geographic convergence area functions as an ecological bottleneck for Gray-cheeked
483 Thrush populations and potentially other species, should be a research priority (Bayly et al. 2014,
484 Gómez et al. 2015). We thus highlight the need to identify and study other high-convergence
485 regions for migratory songbirds and to quantify their impact on fitness and survival. In areas of
486 high geographic convergence the effects of habitat loss or change may be magnified, potentially
487 making a large contribution to the pronounced population declines recorded in many Neotropical
488 migrants (Robbins et al. 1989, Sauer et al. 2013), or indeed, to the reversal of such long term
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490

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510

511 **References**

- 512 Alerstam, T. 2011. Optimal bird migration revisited. *Journal of Ornithology* 152:5–23.
- 513 Alerstam, T., A. Hedenstrom, and S. Akesson. 2003. Long-distance migration: evolution and
514 determinants. *Oikos* 103:247–260.
- 515 Alerstam, T., and A. Lindström. 1990. Optimal bird migration: the relative importance of time,
516 energy and safety. Pages 331–351 in E. Gwinner, editor. *Bird Migration: The Physiology
517 and Ecophysiology*. Springer, Berlin.
- 518 Andueza, M., E. Barba, J. L. Arroyo, J. Feliu, J. L. Greño, F. Jubete, L. Lozano, J. S. Monrós, R.
519 Moreno-Opo, J. M. Neto, A. Onrubia, P. Tenreiro, T. Valkenburg, C. Zumalacárregui, C.
520 González, A. Herrero, and J. Arizaga. 2014. Connectivity in Reed Warblers *Acrocephalus
521 scirpaceus* between breeding grounds in Europe and autumn stopover sites in Iberia. *Ardea*
522 101:133–141.
- 523 Bandelt, H. J., P. Forster, and A. Röhl. 1999. Median-joining networks for inferring intraspecific
524 phylogenies. *Molecular Biology and Evolution* 16:37–48.
- 525 Battley, P. F., N. Warnock, T. L. Tibbitts, R. E. Gill, T. Piersma, C. J. Hassell, D. C. Douglas, D.
526 M. Mulcahy, B. D. Gartrell, R. Schuckard, D. S. Melville, and A. C. Riegen. 2012.
527 Contrasting extreme long-distance migration patterns in bar-tailed godwits *Limosa
528 lapponica*. *Journal of Avian Biology* 43:21–32.
- 529 Bauer, S., S. Lisovski, and S. Hahn. 2016. Timing is crucial for consequences of migratory
530 connectivity. *Oikos* 125:605–612.
- 531 Bayly, N. J., L. Cárdenas-Ortiz, M. Rubio, and C. Gómez. 2014. Migration of raptors, swallows
532 and other diurnal migratory birds through the Darien of Colombia. *Ornitología Neotropical*
533 25:63–71.
- 534 Bayly, N. J., C. Gómez, and K. A. Hobson. 2013. Energy reserves stored by migrating Gray-
535 cheeked Thrushes (*Catharus minimus*) at a spring stopover site in northern Colombia, are
536 sufficient for a long-distance flight to North America. *Ibis* 155:271–283.
- 537 Bayly, N. J., C. Gómez, K. A. Hobson, A. M. González, and K. V. Rosenberg. 2012. Fall

538 migration of the Veery (*Catharus fuscescens*) in northern Colombia: determining the
539 importance of a stopover site. *The Auk* 129:449–459.

540 Bayly, N. J., K. V. Rosenberg, W. Easton, C. Gómez, J. Carlisle, D. N. Ewert, A. Drake, and L.
541 Goodrich. 2017. Major stopover regions and migratory bottlenecks for Nearctic-Neotropical
542 landbirds within the Neotropics: a review. *Bird Conservation International*:1–26.

543 Betini, G. S., A. Pardy, C. K. Griswold, and D. R. Norris. 2015. The role of seasonality and non-
544 lethal carry-over effects on density-dependent dispersal. *Ecosphere* 6:art272.

545 Bonter, D. N., S. a Gauthreaux, and T. M. Donovan. 2009. Characteristics of important stopover
546 locations for migrating birds: remote sensing with radar in the Great Lakes basin.
547 *Conservation biology : the journal of the Society for Conservation Biology* 23:440–8.

548 Boulet, M., and D. R. Norris. 2006. The past and present of migratory connectivity.
549 *Ornithological Monographs* 61:1–13.

550 Bowen, G. J., L. I. Wassenaar, and K. a Hobson. 2005. Global application of stable hydrogen and
551 oxygen isotopes to wildlife forensics. *Oecologia* 143:337–48.

552 Brien, R. J. W., G. Helle, T. L. Pons, J.-L. Guyot, and M. Gloor. 2012. Oxygen isotopes in tree
553 rings are a good proxy for Amazon precipitation and El Niño-Southern Oscillation
554 variability. *Proceedings of the National Academy of Sciences* 109:16957–16962.

555 Buehler, D. M., and T. Piersma. 2008. Travelling on a budget: predictions and ecological
556 evidence for bottlenecks in the annual cycle of long-distance migrants. *Philosophical
557 transactions of the Royal Society of London. Series B, Biological sciences* 363:247–66.

558 Buler, J. J., F. R. Moore, and S. Woltmann. 2007. A multi-scale examination of stopover habitat
559 use by birds. *Ecology* 88:1789–802.

560 Burnham, K. P., and D. R. Anderson. 2002. *Model Selection and Multinomial Inference: a
561 practical information-theoretic approach*. Springer.

562 Cárdenas-Ortiz, L., N. J. Bayly, G. J. Colorado Z., and K. A. Hobson. 2017. Fall migration and
563 breeding origins of Canada Warblers moving through northern Colombia. *Journal of Field
564 Ornithology* 88:53–64.

- 565 Chabot, A., K. A. Hobson, S. L. Van Wilgenburg, G. J. McQuat, and S. C. Loughheed. 2012.
566 Advances in linking wintering migrant birds to their breeding-ground origins using
567 combined analyses of genetic and stable isotope markers. *PloS one* 7:e43627.
- 568 Cohen, E. B., W. C. Barrow, J. J. Buler, J. L. Deppe, A. Farnsworth, P. P. Marra, S. R.
569 McWilliams, D. W. Mehlman, R. R. Wilson, M. S. Woodrey, and F. R. Moore. 2017. How
570 do en route events around the Gulf of Mexico influence migratory landbird populations?
571 *The Condor* 119:327–343.
- 572 Cohen, E. B., J. A. Hostetler, M. T. Hallworth, C. S. Rushing, T. S. Sillett, and P. P. Marra.
573 2018. Quantifying the strength of migratory connectivity. *Methods in Ecology and*
574 *Evolution* 9:513–524.
- 575 Cohen, E. B., F. R. Moore, and R. a. Fischer. 2014. Fuel stores, time of spring, and movement
576 behavior influence stopover duration of Red-eyed Vireo *Vireo olivaceus*. *Journal of*
577 *Ornithology* 155:785–792.
- 578 Cooper, N. W., M. T. Murphy, and L. J. Redmond. 2009. Age- and sex-dependent spring arrival
579 dates of Eastern Kingbirds. *Journal of Field Ornithology* 80:35–41.
- 580 Cooper, N. W., M. T. Murphy, L. J. Redmond, and A. C. Dolan. 2011. Reproductive correlates
581 of spring arrival date in the Eastern Kingbird *Tyrannus tyrannus*. *Journal of Ornithology*
582 152:143–152.
- 583 Coppack, T., and F. Pulido. 2009. Proximate control and adaptive potential of protandrous
584 migration in birds. *Integrative and comparative biology* 49:493–506.
- 585 Cresswell, W. 2014. Review article Migratory connectivity of Palaearctic – African migratory
586 birds and their responses to environmental change : the serial residency hypothesis:493–
587 510.
- 588 Dossman, B. C., G. W. Mitchell, D. R. Norris, P. D. Taylor, C. G. Guglielmo, S. N. Matthews,
589 and P. G. Rodewald. 2016. The effects of wind and fuel stores on stopover departure
590 behavior across a migratory barrier. *Behavioral Ecology* 27:567–574.
- 591 Drent, R. H., G. Eichhorn, A. Flagstad, A. J. Van der Graaf, K. E. Litvin, and J. Stahl. 2007.
592 Migratory connectivity in Arctic geese: Spring stopovers are the weak links in meeting

593 targets for breeding. *Journal of Ornithology* 148.

594 eBird. 2014. eBird: An online database of bird distribution and abundance (web application
595 www.ebird.org). <http://www.ebird.org>.

596 Environment Canada. 2014. Trend results for Gray-cheeked Thrush. An Environment Canada
597 report. Gatineau, Quebec, Canada. Gatineau, Quebec, Canada.

598 Excoffier, L., G. Laval, and S. Schneider. 2005. ARLEQUIN ver. 3.0: An integrated software
599 package for population genetics data analysis. *Evolutionary Bioinformatics Online* 1:47–50.

600 Faaborg, J., R. T. Holmes, A. D. Anders, K. L. Bildstein, K. M. Dugger, S. A. Gauthreaux, P.
601 Heglund, K. A. Hobson, A. E. Jahn, D. H. Johnson, S. C. Latta, D. J. Levey, P. P. Marra, C.
602 L. Merkord, E. Nol, S. I. Rothstein, T. W. Sherry, T. S. Sillett, F. R. Thompson, and N.
603 Warnock. 2010. Conserving migratory land birds in the new world: do we know enough?
604 *Ecological Applications* 20:398–418.

605 Finch, T., S. J. Butler, A. M. A. Franco, and W. Cresswell. 2017. Low migratory connectivity is
606 common in long-distance migrant birds. *Journal of Animal Ecology* 86:662–673.

607 Fink, D., W. M. Hochachka, D. Winkler, B. Shaby, G. Hooker, B. Zuckerberg, M. A. Munson,
608 D. Sheldon, M. Riedewald, and S. Kelling. 2010. Spatiotemporal Exploratory models for
609 large-scale Survey Data. *Ecological Applications* 20:2131–2147.

610 Fiske, I., and R. Chandler. 2011. **unmarked** : An R Package for Fitting Hierarchical Models of
611 Wildlife Occurrence and Abundance. *Journal of Statistical Software* 43.

612 FitzGerald, A. M. 2017. Division within the North American boreal forest: Ecological niche
613 divergence between the Bicknell's Thrush (*Catharus bicknelli*) and Gray-cheeked Thrush
614 (*C. minimus*). *Ecology and Evolution* 7:5285–5295.

615 FitzGerald, A. M., D. M. Whitaker, J. Ralston, J. J. Kirchman, and I. G. Warkentin. 2017.
616 Taxonomy and distribution of the imperilled Newfoundland Gray-cheeked Thrush, *Catharus*
617 *minimus minimus*. *Avian Conservation and Ecology* 12:art10.

618 Fort, J., W. P. Porter, and D. Grémillet. 2009. Thermodynamic modelling predicts energetic
619 bottleneck for seabirds wintering in the northwest Atlantic. *The Journal of Experimental*

620 Biology 212:2483–2490.

621 Fournier, A. M. V., A. R. Sullivan, J. K. Bump, M. Perkins, M. C. Shieldcastle, and S. L. King.
622 2017. Combining citizen science species distribution models and stable isotopes reveals
623 migratory connectivity in the secretive Virginia rail. *Journal of Applied Ecology* 54:618–
624 627.

625 Fraser, K. C., A. Shave, A. Savage, A. Ritchie, K. Bell, J. Siegrist, J. D. Ray, K. Applegate, and
626 M. Pearman. 2017. Determining fine-scale migratory connectivity and habitat selection for
627 a migratory songbird by using new GPS technology. *Journal of Avian Biology* 48:339–345.

628 Fraser, K. C., B. J. M. Stutchbury, P. Kramer, C. Silverio, J. Barrow, D. Newstead, N. Mickle, T.
629 Shaheen, P. Mammenga, K. Applegate, E. Bridge, and J. Tautin. 2013. Consistent range-
630 wide pattern in fall migration strategy of Purple Martin (*Progne subis*), despite different
631 migration routes at the Gulf of Mexico. *The Auk* 130:291–296.

632 Gómez, C., N. J. Bayly, D. R. Norris, S. A. Mackenzie, K. V. Rosenberg, P. D. Taylor, and C. D.
633 Cadena. 2017. Fuel loads acquired at a stopover site influence the pace of intercontinental
634 migration in a boreal songbird. *Scientific Reports* 7:Art. No.3405.

635 Gómez, C., N. J. Bayly, and K. V. Rosenberg. 2013. Seasonal variation in stopover site use:
636 *Catharus* thrushes and vireos in northern Colombia. *Journal of Ornithology* 154:107–117.

637 Gómez, C., N. J. Bayly, and K. V. Rosenberg. 2014. Fall stopover strategies of three species of
638 thrush (*Catharus*) in northern South America. *The Auk* 131:702–717.

639 Gómez, C., V. Gómez-Bahamón, L. Cárdenas-Ortíz, and N. J. Bayly. 2015. Distribution of
640 Nearctic-Neotropical migratory birds along a South American elevation gradient during
641 spring migration. *Wilson Journal of Ornithology* 127:72–86.

642 González-Prieto, A. M., K. A. Hobson, N. J. Bayly, and C. Gómez. 2011. Geographic origins
643 and timing of fall migration of the Veery in northern Colombia. *The Condor* 113:860–868.

644 Griffiths, R., M. C. Double, and R. J. G. Dawson. 1998. A DNA test to sex most birds.
645 *Molecular Ecology* 7:1071–1075.

646 Guillera-Aroita, G. 2017. Modelling of species distributions, range dynamics and communities

647 under imperfect detection: advances, challenges and opportunities. *Ecography* 40:281–295.

648 Hahn, S., D. Dimitrov, S. Rehse, E. Yohannes, and L. Jenni. 2014. Avian claw morphometry and
649 growth determine the temporal pattern of archived stable isotopes. *Journal of Avian Biology*
650 45:202–207.

651 Hedenström, A., and T. Alerstam. 1997. Optimum fuel loads in migratory birds: distinguishing
652 between time and energy minimization. *Journal of Theoretical Biology* 189:227–234.

653 Hobson, K. A., and R. G. Clark. 1992. Assessing avian diets using stable isotopes II: Factors
654 influencing diet-tissue fractionation. *The Condor* 94:189–197.

655 Hobson, K. A., S. L. Van Wilgenburg, J. Faaborg, J. D. Toms, C. Rengifo, A. L. Sosa, Y. Aubry,
656 and R. Brito Aguilar. 2014. Connecting breeding and wintering grounds of Neotropical
657 migrant songbirds using stable hydrogen isotopes: a call for an isotopic atlas of migratory
658 connectivity. *Journal of Field Ornithology* 85:237–257.

659 Hobson, K. A., S. L. Van Wilgenburg, L. I. Wassenaar, and K. Larson. 2012. Linking Hydrogen
660 ($\delta^2\text{H}$) isotopes in feathers and precipitation: Sources of variance and consequences for
661 assignment to isoscapes. *PLoS ONE* 7:e35137.

662 Hostetler, J. a., T. S. Sillett, and P. P. Marra. 2015. Full-annual-cycle population models for
663 migratory birds. *The Auk* 132:433–449.

664 Hübner, C. E., I. M. Tombre, L. R. Griffin, M. J. J. E. Loonen, P. Shimmings, and I. S.
665 Jonsdottir. 2010. The connectivity of spring stopover sites for geese heading to arctic
666 breeding grounds. *Ardea* 98:145–154.

667 Iwamura, T., H. P. Possingham, I. Chades, C. D. T. Minton, N. J. Murray, E. A. Treml, and R. A.
668 Fuller. 2013. Migratory connectivity magnifies the consequences of habitat loss from sea-
669 level rise for shorebird populations. *Proceedings of the Royal Society B: Biological*
670 *Sciences* 280:20130325.

671 Jiménez-Muñoz, J. C., C. Mattar, J. Barichivich, A. Santamaría-Artigas, K. Takahashi, Y. Malhi,
672 J. A. Sobrino, and G. van der Schrier. 2016. Record-breaking warming and extreme drought
673 in the Amazon rainforest during the course of El Niño 2015–2016. *Scientific Reports*
674 6:33130.

- 675 Johnson, E. I., J. D. Wolfe, T. Brandt Ryder, and P. Pyle. 2011. Modifications to a molt-based
676 ageing system proposed by Wolfe et al. (2010). *Journal of Field Ornithology* 82:422–424.
- 677 Kearse, M., R. Moir, A. Wilson, S. Stones-Havas, M. Cheung, S. Sturrock, S. Buxton, A.
678 Cooper, S. Markowitz, C. Duran, T. Thierer, B. Ashton, P. Mentjies, and A. Drummond.
679 2012. Geneious Basic: an integrated and extendable desktop software platform for the
680 organization and analysis of sequence data. *Bioinformatics* 28:1647–1649.
- 681 Kekkonen, J., P. Seppä, I. K. Hanski, H. Jensen, R. A. Väisänen, and J. E. Brommer. 2011. Low
682 genetic differentiation in a sedentary bird: house sparrow population genetics in a
683 contiguous landscape. *Heredity* 106:183–190.
- 684 Kelly, J. F. 2006. Stable isotope evidence links breeding geography and migration timing in
685 Wood Warblers (Parulidae). *The Auk* 123:1–7.
- 686 Kokko, H., T. G. Gunnarsson, L. J. Morrell, and J. A. Gill. 2006. Why do female migratory birds
687 arrive later than males? *Journal of Animal Ecology* 75:1293–1303.
- 688 Kramer, G. R., D. E. Andersen, D. A. Buehler, P. B. Wood, S. M. Peterson, J. A. Lehman, K. R.
689 Aldinger, L. P. Bulluck, S. Harding, J. A. Jones, J. P. Loegering, C. Smalling, R. Vallender,
690 and H. M. Streby. 2018. Population trends in *Vermivora* warblers are linked to strong
691 migratory connectivity. *Proceedings of the National Academy of Sciences*:201718985.
- 692 Laughlin, A. J., C. M. Taylor, D. W. Bradley, D. LeClair, R. G. Clark, R. D. Dawson, P. O.
693 Dunn, A. Horn, M. Leonard, D. R. Sheldon, D. Shutler, L. a. Whittingham, D. W. Winkler,
694 and D. R. Norris. 2013. Integrating information from geolocators, weather radar, and citizen
695 science to uncover a key stopover area of an aerial insectivore. *The Auk* 130:230–239.
- 696 Leigh, J. W., and D. Bryant. 2015. POPART: Full-feature software for haplotype network
697 construction. *Methods in Ecology and Evolution* 6:1110–1116.
- 698 Leyrer, J., T. Lok, M. Brugge, B. Spaans, B. K. Sandercock, and T. Piersma. 2013. Mortality
699 within the annual cycle: seasonal survival patterns in Afro-Siberian Red Knots (*Calidris*
700 *canutus canutus*). *Journal of Ornithology* 154:933–943.
- 701 Lowther, P. E., C. C. Rimmer, B. Kessel, S. L. W. Johnson, and W. G. Ellison. 2001. Gray-
702 cheeked Thrush (*Catharus minimus*). <https://birdsna.org/Species->

703 Account/bna/species/gychthr.

704 Lundberg, S., and T. Alerstam. 1986. Bird migration patterns: Conditions for stable geographical
705 population segregation. *Journal of Theoretical Biology* 123:403–414.

706 Marra, P. P., and R. T. Holmes. 2001. Consequences of dominance-mediated habitat segregation
707 in American Redstarts during the non-breeding season. *The Auk* 118:92–104.

708 McKinnon, E. A., K. C. Fraser, C. Q. Stanley, and B. J. M. Stutchbury. 2014. Tracking from the
709 tropics reveals behaviour of juvenile songbirds on their first spring migration. *PloS one*
710 9:e105605.

711 McKinnon, E. a., C. Q. Stanley, and B. J. M. Stutchbury. 2015. Carry-Over Effects of
712 Nonbreeding Habitat on Start-to-Finish Spring Migration Performance of a Songbird. *Plos*
713 *One* 10:e0141580.

714 Mitchell, G. W., B. K. Woodworth, P. D. Taylor, and D. R. Norris. 2015. Automated telemetry
715 reveals age specific differences in flight duration and speed are driven by wind conditions in
716 a migratory songbird. *Movement Ecology* 3:19.

717 Norris, D. R., and P. P. Marra. 2007. Seasonal interactions, habitat quality, and population
718 dynamics in migratory birds. *The Condor* 109:535–547.

719 Norris, D. R., P. P. Marra, G. J. Bowen, L. M. Ratcliffe, J. A. Royle, and T. K. Kyser. 2006.
720 Migratory connectivity of a widely distributed songbird, the American Redstart (*Setophaga*
721 *ruticilla*). *Ornithological Monographs* 61:14–28.

722 Norris, D. R., P. P. Marra, T. K. Kyser, T. W. Sherry, and L. M. Ratcliffe. 2004. Tropical winter
723 habitat limits reproductive success on the temperate breeding grounds in a migratory bird.
724 *Proceedings of the Royal Society of London B: Biological Sciences* 271:59–64.

725 O'Connor, C., D. R. Norris, G. T. Crossin, and S. J. Cooke. 2014. Biological carryover effects :
726 linking common concepts and mechanisms in ecology and evolution. *Ecosphere* 5:1–11.

727 Paxton, E. H., S. L. Durst, M. K. Sogge, T. J. Koronkiewicz, and K. L. Paxton. 2017.
728 Survivorship across the annual cycle of amigratory passerine, the Willow flycatcher.
729 *Journal of Avian Biology* 48:1126–1131.

730 Paxton, K. L., and F. R. Moore. 2017. Connecting the dots: Stopover strategies of an
731 intercontinental migratory songbird in the context of the annual cycle. *Ecology and*
732 *Evolution*.

733 Piersma, T. 2002. Energetic bottlenecks and other design constraints in avian annual cycles.
734 *Integrative and comparative biology* 42:51–67.

735 Piersma, T., T. Lok, Y. Chen, C. J. Hassell, H. Y. Yang, A. Boyle, M. Slaymaker, Y. C. Chan, D.
736 S. Melville, Z. W. Zhang, and Z. Ma. 2016. Simultaneous declines in summer survival of
737 three shorebird species signals a flyway at risk. *Journal of Applied Ecology* 53:479–490.

738 Pomeroy, A. C. 2006. Tradeoffs between food abundance and predation danger in spatial usage
739 of a stopover site by western sandpipers, *Calidris mauri*. *Oikos* 3:629–637.

740 Pyle, P. 1997. Identification guide to North American birds, Part I: Columbidae to Ploceidae.
741 Slate Creek Press, Bolinas, California - USA.

742 Robbins, C. S., J. R. Sauer, R. S. Greenberg, and S. Droege. 1989. Population declines in North
743 American birds that migrate to the Neotropics. *Proceedings of the National Academy of*
744 *Sciences of the United States of America* 86:7658–62.

745 Rockwell, S. M., J. M. Wunderle, T. S. Sillett, C. I. Bocetti, D. N. Ewert, D. Currie, J. D. White,
746 and P. P. Marra. 2016. Seasonal survival estimation for a long-distance migratory bird and
747 the influence of winter precipitation. *Oecologia*:1–12.

748 Rodenhouse, N. L., T. S. Sillett, P. J. Doran, and R. T. Holmes. 2003. Multiple density-
749 dependence mechanisms regulate a migratory bird population during the breeding season.
750 *Proceedings of the Royal Society B: Biological Sciences* 270:2105–10.

751 Rogers, D. L., H. Y. Yang, C. J. Hassell, A. N. Boyle, K. G. Rogers, B. Chen, Z. W. Zhang, and
752 T. Piersma. 2010. Red Knots (*Calidris canutus piersmai* and *C. c. rogersi*) depend on a small
753 threatened staging area in Bohai Bay, China. *Emu* 110:307–315.

754 Rohwer, S., L. K. Butler, and D. Froehlich. 2005. Ecology and demography of east-west
755 differences in molt scheduling of neotropical migrant passerines. Pages 87–105 in R.
756 Greenberg and P. P. Marra, editors. *Birds of Two Worlds: The Ecology and Evolution of*
757 *Migratory Birds*. Johns Hopkins University Press, Baltimore.

- 758 Rushing, C. S., J. A. Hostetler, T. S. Sillett, P. P. Marra, J. A. Rotenberg, and T. B. Ryder.
759 2017a. Spatial and temporal drivers of avian population dynamics across the annual cycle.
760 *Ecology* 98:2837–2850.
- 761 Rushing, C. S., P. P. Marra, and C. E. Studds. 2017b. Incorporating breeding abundance into
762 spatial assignments on continuous surfaces. *Ecology and Evolution* 7:3847–3855.
- 763 Rushing, C. S., T. B. Ryder, J. F. Saracco, and P. P. Marra. 2013. Assessing migratory
764 connectivity for a long-distance migratory bird using multiple intrinsic markers. *Ecological*
765 *applications* 24:445–456.
- 766 Sauer, J. R., W. A. Link, J. E. Fallon, K. L. Pardieck, and D. J. Ziolkowski. 2013. The North
767 American Breeding Bird Survey 1966–2011: Summary analysis and species accounts.
768 *North American Fauna* 79:1–32.
- 769 Schmaljohann, H., and C. Eikenaar. 2017. How do energy stores and changes in these affect
770 departure decisions by migratory birds? A critical view on stopover ecology studies and
771 some future perspectives. *Journal of Comparative Physiology A* 203:411–429.
- 772 Schmaljohann, H., C. Meier, D. Arlt, F. Bairlein, H. van Oosten, Y. E. Morbey, S. Åkesson, M.
773 Buchmann, N. Chernetsov, R. Desaever, J. Elliott, M. Hellström, F. Liechti, A. López, J.
774 Middleton, U. Ottosson, T. Pärt, F. Spina, and C. Eikenaar. 2015. Proximate causes of avian
775 protandry differ between subspecies with contrasting migration challenges. *Behavioral*
776 *Ecology* 27:321–331.
- 777 Seewagen, C. L., C. G. Guglielmo, and Y. E. Morbey. 2013. Stopover refueling rate underlies
778 protandry and seasonal variation in migration timing of songbirds. *Behavioral Ecology*
779 24:634–642.
- 780 Sillett, T. S., and R. T. Holmes. 2002. Variation in survivorship of a migratory songbird
781 throughout its annual cycle. *Journal of Animal Ecology* 71:296–308.
- 782 Srygley, R. B., R. Dudley, E. G. Oliveira, R. Aizprúa, N. Z. Pelaez, and A. J. Riveros. 2010. El
783 Niño and dry season rainfall influence hostplant phenology and an annual butterfly
784 migration from Neotropical wet to dry forests. *Global Change Biology* 16:936–945.
- 785 SSAC. 2010. The status of Gray-cheeked Thrush (*Catharus minimus*) in Newfoundland and

- 786 Labrador. Page The species status advisory committee. Report No. 24.
- 787 Stanley, C. Q., E. a McKinnon, K. C. Fraser, M. P. Macpherson, G. Casbourn, L. Friesen, P. P.
788 Marra, C. Studds, T. B. Ryder, N. E. Diggs, and B. J. M. Stutchbury. 2014. Connectivity of
789 Wood Thrush Breeding, Wintering, and Migration Sites Based on Range-Wide Tracking.
790 Conservation Biology 00:1–11.
- 791 Stewart, R. L. M., C. M. Francis, and C. Massey. 2002. Age-Related Differential Timing of
792 Spring Migration Within Sexes in Passerines. *The Wilson Bulletin* 114:264–271.
- 793 Studds, C. E., B. E. Kendall, N. J. Murray, H. B. Wilson, D. I. Rogers, R. S. Clemens, K.
794 Gosbell, C. J. Hassell, R. Jessop, D. S. Melville, D. A. Milton, C. D. T. Minton, H. P.
795 Possingham, A. C. Riegen, P. Straw, E. J. Woehler, and R. A. Fuller. 2017. Rapid
796 population decline in migratory shorebirds relying on Yellow Sea tidal mudflats as stopover
797 sites. *Nature Communications* 8:1–7.
- 798 Topp, C. M., C. L. Pruett, K. G. McCracken, and K. Winker. 2013. How migratory thrushes
799 conquered northern North America: a comparative phylogeography approach. *PeerJ* 1:e206.
- 800 Tsipoura, N., and J. Burger. 1999. Shorebird diet during spring migration stopover on Delaware
801 Bay. *The Condor* 101:635–644.
- 802 Ungvari-Martin, J., C. M. Heckscher, and K. A. Hobson. 2016. Inter-annual site fidelity and
803 breeding origins of Gray-cheeked Thrushes in white sand forests of the Peruvian Amazon.
804 *Journal of Field Ornithology* 87:55–64.
- 805 Velmala, W., S. Helle, M. P. Ahola, M. Klaassen, E. Lehikoinen, K. Rainio, P. M. Sirkiä, and T.
806 Laaksonen. 2015. Natural selection for earlier male arrival to breeding grounds through
807 direct and indirect effects in a migratory songbird. *Ecology and Evolution* 5:1205–1213.
- 808 Warnock, N., J. Y. Takekawa, and M. A. Bishop. 2004. Migration and stopover strategies of
809 individual Dunlin along the Pacific coast of North America. *Canadian Journal of Zoology*
810 82:1687–1697.
- 811 Wassenaar, L. I., and K. A. Hobson. 2003. Comparative equilibration and online technique for
812 determination of non-exchangeable hydrogen of keratins for use in animal migration
813 studies. *Isotopes in environmental and health studies* 39:1–3.

- 814 Weber, M. M., R. D. Stevens, J. A. F. Diniz-filho, and C. E. V Grelle. 2016. Is there a correlation
815 between abundance and environmental suitability derived from ecological niche modelling?
816 A meta-analysis. *Ecography* 40:817–828.
- 817 Weber, T. P., and A. I. Houston. 1997. A general model for time-minimizing avian migration. *J.*
818 *Theor. Biol.* 185:447–458.
- 819 Webster, M. S., and P. P. Marra. 2005. The importance of understanding migratory connectivity
820 and seasonal interactions. Pages 199–209 *Birds of two worlds: the ecology and evolution of*
821 *migration*. Johns Hopkins University Press.
- 822 Whitaker, D. M., P. D. Taylor, and I. G. Warkentin. 2015. Gray-cheeked Thrush (*Catharus*
823 *minimus minimus*) distribution and habitat use in a montane forest landscape of western
824 Newfoundland, Canada. *Avian Conservation and Ecology* 10.
- 825 Whitaker, D. M., I. G. Warkentin, K. A. Hobson, P. Thomas, and R. Boardman. (n.d.). Fall and
826 winter movements of Newfoundland Gray-cheeked Thrushes (*Catharus minimus minimus*).
827 *Animal Migration*.
- 828 Van Wilgenburg, S. L., and K. A. Hobson. 2011. Combining stable-isotope (δD) and band
829 recovery data to improve probabilistic assignment of migratory birds to origin. *Ecological*
830 *applications* 21:1340–1351.
- 831 Wilson, S., S. L. LaDeau, A. P. Tøttrup, and P. P. Marra. 2011. Range-wide effects of breeding-
832 and nonbreeding-season climate on the abundance of a Neotropical migrant songbird.
833 *Ecology* 92:1789–98.
- 834 Wilson, S., J. F. Saracco, R. Krikun, D. T. T. Flockhart, C. M. Godwin, and K. R. Foster. 2018.
835 Drivers of demographic decline across the annual cycle of a threatened migratory bird.
836 *Scientific Reports* 8:1–11.
- 837 Woodworth, B. K., A. E. M. Newman, S. P. Turbek, C. Bryant, K. A. Hobson, L. I. Wassenaar,
838 G. W. Mitchell, T. Nathaniel, and D. R. Norris. 2016. Differential migration and the link
839 between winter latitude , timing of migration , and breeding in a songbird. *Oecologia*
840 181:413–422.
- 841 Yong, W., and F. Moore. 1994. Flight Morphology, Energetic Condition, and Stopover Biology

842 of Migrating Thrushes. The Auk 111:683–692.

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DATA AVAILABILITY

846 Genetic data are available in GenBank (accession numbers in Appendix S1: Table S1). Isotopic

847 data are available from the Dryad Digital Repository: <https://doi.org/10.5061/dryad.7hk6523>

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Tables.

Table 1. Fuel load on arrival of Gray-cheeked Thrush at a spring stopover site in northern Colombia varied as a function of date of arrival, breeding origin ($\delta^2\text{H}$ values of feathers), and sex. The first two models were equally likely and overall explained 84% of the variation in fuel on arrival. Models including age did not rank among the top two.

Model	AICc	ΔAICc	wi
Fuel ~ Date + $\delta^2\text{H}$ feather + $\delta^2\text{H}$ feather:Sex	357.57	0.00	0.54
Fuel ~ Date + $\delta^2\text{H}$ feather + Sex	358.73	1.16	0.30
Fuel ~ Date + $\delta^2\text{H}$ feather + Sex + Age	361.04	3.47	0.09
Fuel ~ Date + $\delta^2\text{H}$ feather + $\delta^2\text{H}$ feather:Sex + $\delta^2\text{H}$ feather:Age	361.70	4.12	0.07
Fuel ~ $\delta^2\text{H}$ feather + Sex	378.30	20.73	0.00
Fuel ~ $\delta^2\text{H}$ feather + Sex + Age	380.37	22.80	0.00
Fuel ~ Date + $\delta^2\text{H}$ feather	1636.68	1279.11	0.00
Fuel ~ Date + $\delta^2\text{H}$ feather + $\delta^2\text{H}$ feather:Age	1640.52	1282.95	0.00
Fuel ~ $\delta^2\text{H}$ feather	1714.82	1357.25	0.00
Fuel ~ Date + Sex	2898.49	2540.92	0.00
Fuel ~ Sex + Age	3015.04	2657.46	0.00
Fuel ~ Sex	3022.43	2664.86	0.00
Fuel ~ Date + Age	8772.38	8414.81	0.00
Fuel ~ Date	8821.76	8464.19	0.00
Fuel ~ Age	9060.65	8703.08	0.00

Table 2. Pairwise ϕ_{st} comparisons between breeding populations and stopover localities, based on mitochondrial ND2. Note that birds from ‘Newfoundland/s. Labrador’ are all designated *C. m. minimus* whereas birds from ‘w. Labrador/Quebec’ and ‘Alaska/Siberia’ are of subspecies *C. m. aliciae*. Significance ($p < 0.05$) is denoted with *.

		Fall stopover	Spring stopover	Newfoundland/s. Labrador	w. Labrador/Quebec	Alaska/Siberia
<i>C.m.minimus</i>	Fall stopover	-				
	Spring stopover	0.047*	-			
<i>C.m.aliciae</i>	Newfoundland/s. Labrador	0.384*	0.330*	-		
	w. Labrador / Quebec	0.381*	0.327*	0.293*	-	
	Alaska / Siberia	0.072*	0.066*	0.301*	0.319*	-

Figure Legends

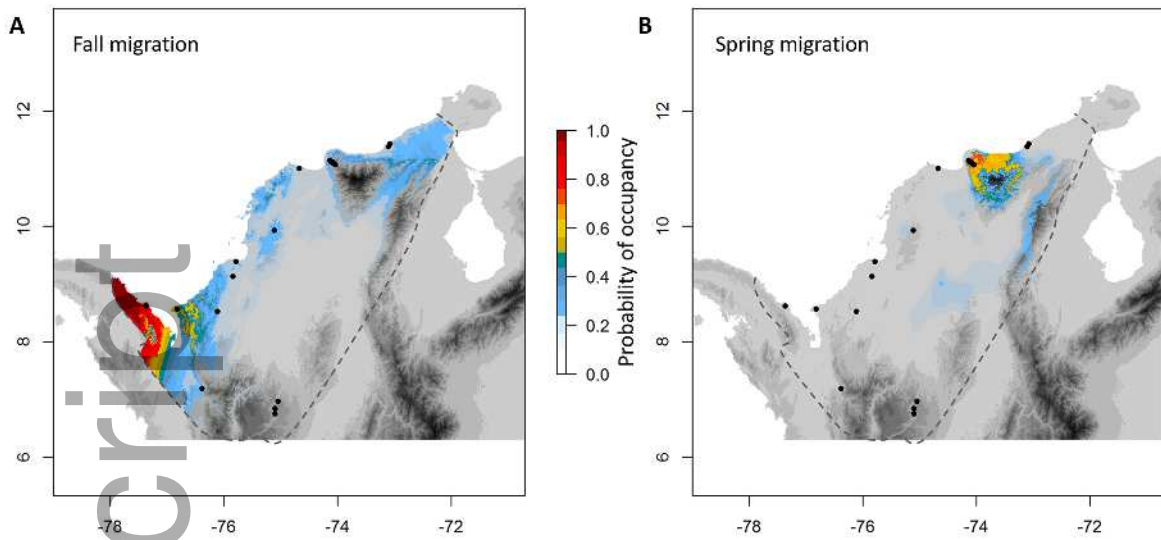
Figure 1. Occupancy surveys carried out at 17 sites (black dots on maps) across the northern coast of Colombia were used to predict probability of occurrence of Gray-cheeked Thrush during fall and spring migration. The dashed lines show the area to which occupancy predictions were limited (see suppl. material for details). High occupancy rates in the Darién in the fall (**A**) and in the Sierra Nevada de Santa Marta in the spring (**B**), suggest a funneling effect where most Gray-cheeked Thrush entering and leaving South America concentrate in these two regions during migration.

Figure 2. A. Low population structure of ND2 among populations of Gray-cheeked Thrush was found, however, birds captured in Colombia (n = 150) shared haplotypes with individuals from throughout the breeding range. The number of individuals sampled from each breeding region are shown on the map, as well as the subspecies to which each population belongs. The dashed line shows the approximate geographic division between subspecies breeding ranges. **B.** 14 out of 24 informative sites within the Gray-cheeked Thrush ND2, were private alleles restricted to individuals from one or two breeding populations. Birds captured in Colombia shared private alleles with all of the sampled breeding regions.

Figure 3A. Temporally explicit breeding and wintering distribution models of the Gray-cheeked Thrush based on monthly climatic variables and presence records. Surfaces reflect the probability of presence based on a maximum entropy algorithm. **B.** Likely breeding and wintering geographic origins of Gray-cheeked Thrush captured during spring migration in northern Colombia based on a species distribution model and on $\delta^2\text{H}$ values of feathers (breeding) and claws (winter). In agreement with genetic data, our assignment suggests that range-wide populations of Gray-cheeked Thrush converge in northern Colombia during stopover. Colored surfaces represent the sum of the upper 67% likelihood probability (2:1 odds) of origin for each bird.

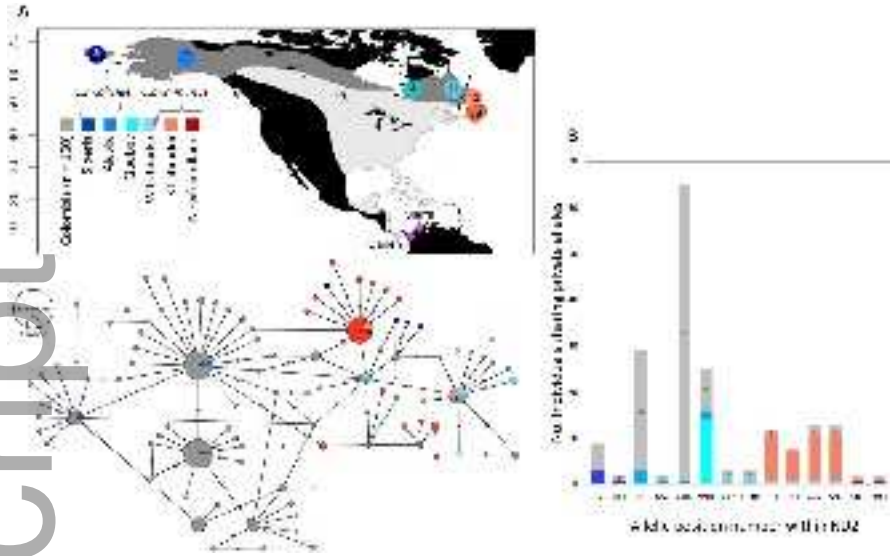
Figure 4. There was a significant effect of breeding origin ($\delta^2\text{H}_f$ values) (**A**) but not of winter origin ($\delta^2\text{H}_c$ values) (**B**) on date of arrival of Gray-cheeked Thrush to the spring stopover in northern Colombia. Males arrived significantly earlier than females both to the spring (**C**) and fall (**D**) stopover sites, and adults arrived significantly earlier than immature birds during spring (**E**) and especially during fall (**F**).

Figure 5. Fuel load on arrival to a spring stopover site for Gray-cheeked Thrush varied as a function of date, breeding origin and sex. **A.** Both males and females captured earlier in the season were in poorer body condition than those that arrived later. **B.** There was no difference in fuel on arrival between females of different breeding origins but males from more northerly breeding origin (more negative $\delta^2\text{H}_f$ values) arrived with lower fuel stores than those from more southerly breeding origins.

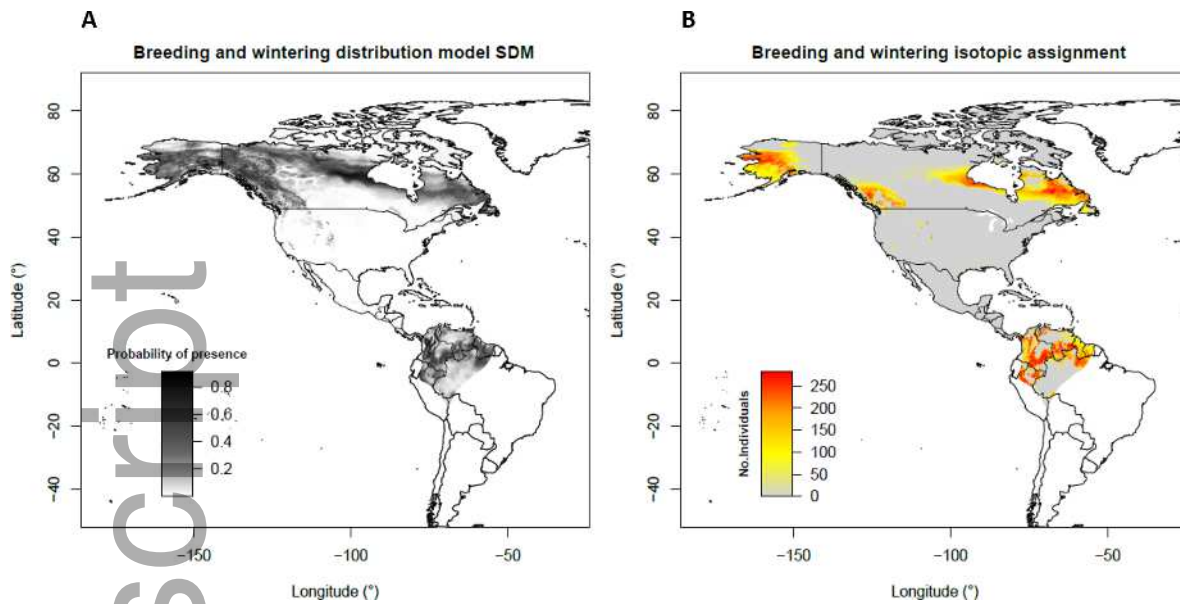


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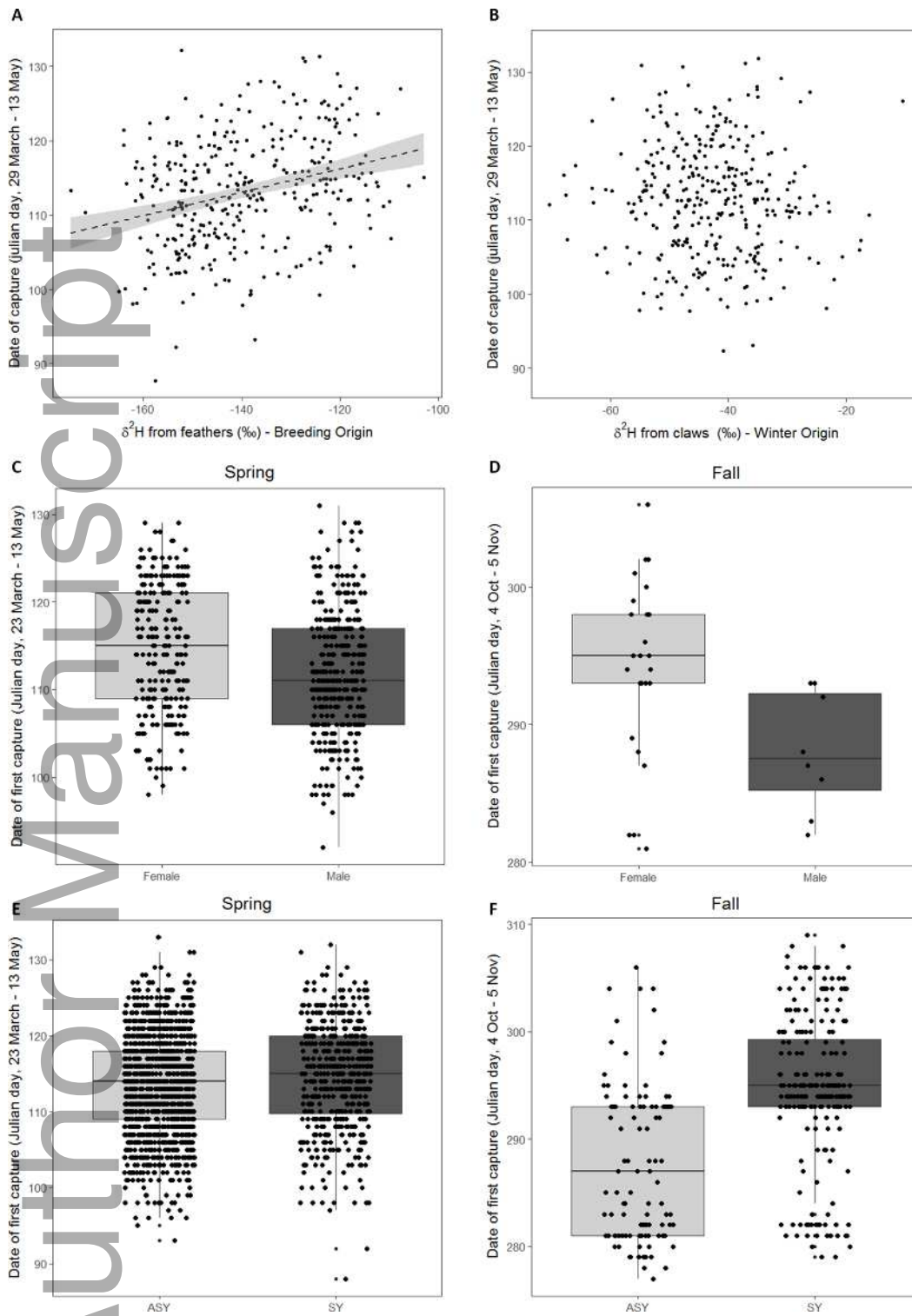
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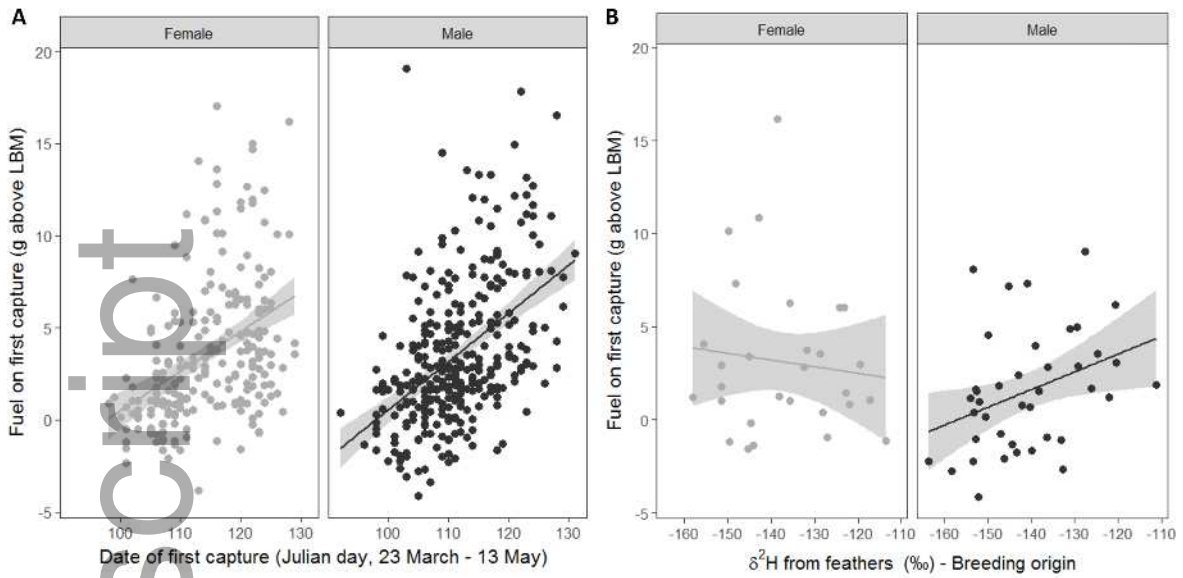
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