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Generic solutions for data-limited fishery assessments are not so simple

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26

27 Abstract

28 The majority of the world's fisheries, by number, are data-poor/limited, and there is a growing body
29 of literature pertaining to approaches to estimate data-limited stock status. There are at least two
30 drivers for assessing of the status of data-limited fisheries. The first is to try to understand and
31 report on the global or regional status of fisheries across many stocks. The second is to attempt to
32 assess individual data-limited stocks, for status reporting and/or guiding management decisions.
33 These drivers have led to attempts to find simple, generic, low-cost solutions, including broad
34 application of generically parameterised models, and the blanket application of a single, or limited
35 number of possible, analytical approach(es). It is unclear that generic methods function as intended,
36 especially when taken out of their original design context or used without care. If the intention is to
37 resolve individual stock status for the purposes of management, there is concern with the
38 indiscriminate application of a single method to a suite of stocks irrespective of the particular
39 circumstances of each. We examine why caution needs to be exercised, and provide guidance on the
40 appropriate application of data-limited assessment methods (DLMs). We recommend: i) obtaining
41 better data, ii) using care in acknowledging and interpreting uncertainties in the results of DLMs, iii)
42 embedding DLMs in harvest strategies that are robust to the higher levels of uncertainty in the
43 output of DLMs by including precautionary management measures or buffers, and iv) selecting and
44 applying DLMs appropriate to specific species' and fisheries' data and context.

45

46 Keywords: data-poor, data-limited stock assessment, stock status

47 Introduction

48 The majority of the world's fisheries, by number, are data-limited (Costello et al. 2012). That is, they
49 have insufficient data (e.g., type, amount, and/or quality of) and/or capacity (e.g., research,
50 institutional, or funding) to enable undertaking a quantitative, model-based stock assessment to
51 estimate time series of biomass and fishing mortality relative to their reference points. There is a
52 growing body of literature aimed at developing approaches to estimate the status of data-limited
53 stocks (e.g. Anderson et al. 2017; Dowling et al. 2008; Dichmont and Brown 2010; Plaganyi et al.
54 2015; Wayte and Klaer 2010; Dowling et al. 2014; Dowling et al. 2016; Pilling et al. 2008), and
55 increasing use of these data-limited assessment methods (DLMs) for management purposes (Bentley
56 2015; Carruthers et al., 2012, 2014, 2016; Dichmont et al. 2017; Geromont and Butterworth
57 2015a,b). Several of the classical, average-length-based methods developed in the 1950s by

58 Beverton and Holt, Gulland, Ricker and other early quantitative fishery scientists are data-limited
59 assessment methods, many of which are being revisited and re-emphasised (eg., Gedamke and
60 Hoenig 2006; Prince et al. 2014; Then et al. 2016). DLMs range from empirical methods, in which
61 performance indicators are based on directly-measured properties, to model-based approaches,
62 where performance indicators are model outputs. **Table 1** provides a summary of types of DLMs,
63 including empirical assessments.

64 There are at least two contexts for undertaking an assessment of the status of data-limited fisheries.
65 The first is to try to understand and report on the global or regional status of fisheries across many
66 stocks. This typically involves application of a single DLM to a large number of stocks (e.g. Rosenberg
67 et al. 2017; Zeller et al. 2016; Costello et al. 2012; Froese et al. 2012; Thorson et al. 2012; FAO 1995),
68 to gain an understanding of the general health of fisheries for the purpose of motivating funding,
69 research and reform. The second is to assess the status of individual data-limited stocks, for status
70 reporting or guiding management decisions, the latter typically within a harvest, or management,
71 strategy (e.g., Sainsbury et al. 2000; Butterworth and Punt 2003, and Fish. Res. Special Issue 94(3)
72 2008) (Figure 1).

73 DLM approaches are also increasingly being used in formal harvest strategies (e.g. Dichmont and
74 Brown 2010; Klaer et al. 2012). The main distinction being highlighted within this paper is the use of
75 assessments for stock status determination, compared to connecting a control rule to the
76 assessment output in order to provide a direct link to a pre-specified management action , as per a
77 formal harvest strategy (e.g., Sainsbury et al. 2000; Butterworth and Punt 2003) (Figure 1). We
78 discuss more fully the stock assessments whose outputs are embedded within harvest strategies
79 below.

80 There are good reasons for management agencies, national governments and international
81 organisations such as the Food and Agricultural Organisation to report on the status of individual
82 stocks. Over the last 20 years, fisheries policy has increasingly required the maintenance of stocks at
83 or around biomass-based target reference points, and the avoidance of biomass-based limit
84 reference points (Oremus et al. 2014; FAO 2016). The stock status reporting requirement associated
85 with managing around target and limit reference points imposes a daunting task for fishery
86 management agencies, leading in many cases to a substantial number of stocks being classified as
87 “uncertain”, and many others being designated as “not overfished” or “overfished”, but with a high
88 risk of misclassification (e.g., Flood et al. 2016; NRC 2012).

89 At the same time, there is, understandably, pressure to resolve the status of “uncertain” stocks (e.g.,
90 Flood et al. 2016) to increase the number of species assessed and decrease the number of species

91 whose status is designated as “uncertain”. While the initial focus of national level stock status
92 reports (e.g., in Australia or the USA) has been on data-rich stocks, there is interest to report on the
93 large number of stocks which are currently listed as uncertain, and to use DLMs to do so. For
94 example, the Status of Australian Fish Stocks (SAFS), the nationally standardised approach to
95 assessing stock status, aims to increase the number of included species from 83 to 200, and to
96 reduce the number of species classified as “undefined” from the current approximately 30%, to less
97 than 10% (Flood et al. 2016). In the United States, there is a legal requirement to assess and set
98 Acceptable Catch Levels for all species, per the 2006 amendment to the Magnuson–Stevens Act
99 (Newman et al. 2015). The interest in resolving stock status extends to developing nations whose
100 legislative requirements are also becoming more stringent, demanding the use of standard target
101 and limit reference points to set management regulations (e.g., Chile, Wiff et al. 2016).

102 The above drivers have led to attempts to find simple, generic, low-cost solutions to assess fishery
103 status. Such solutions include “silver bullet”, or one-size-fits-all approaches, that are simple to
104 implement without much scientific expertise. It is unclear that these generic methods function as
105 intended, especially when taken out of their original design context or used without care.

106 For example, the abovementioned Australian SAFS process has, as an initial step, focused on building
107 capacity around a small number of possible quantitative methods (3) to assess a large number of
108 species, while acknowledging the challenge of appropriate application and interpretation (Haddon et
109 al. in prep). In a developing nation context, the authors have consulted in countries where an
110 outside expert has proposed a method for one species or fishery, and subsequently a blanket
111 application of that approach to other species has occurred, without due critical appraisal of the
112 appropriateness of the method or its assumptions.

113 While all of the above motivations are legitimate, the approach to undertaking an assessment of
114 stock status needs to be appropriate to the objective. If the intention is to resolve individual stock
115 status for the purposes of managing the stock, particularly within a management or harvest strategy,
116 the concern is with the indiscriminate application of a single method to a suite of stocks irrespective
117 of the particular circumstances of each.

118 In this paper, we examine why caution needs to be exercised when using generic approaches to
119 data-limited stock assessment, and provide guidance on the appropriate application of DLMs.

120 What do we mean by “generic assessment approaches”?

121 Generic approaches to stock assessments include two important potentially undesirable
122 applications: i) species-unspecific parameterised models, and ii) blanket application of a single or

123 very few analytical approach(es) to many stocks that do not discriminate among the different
124 circumstances and requirements of each stock. **Table 2** provides a review of regional or global
125 assessment applications, with examples of one-size fits-all approaches, and the rationale or impetus
126 for each. Note that the methods here summarised are not intended as examples of incorrect
127 application; the point is that such approaches exist, and along with them exists the possibility of
128 misapplication or misinterpretation.

129 Generic assessment approaches may include methods that have been tested using data-rich stocks,
130 and are then applied in the same manner to data-limited stocks. We emphasise that we are not
131 against generic approaches *per se*, if they are tuned intelligently and used with care to the specific
132 application, such as with, Froese et al.'s 2017 application of a catch-MSY assessment to a range of
133 data-limited fisheries.

134 i) Broad application of generically parameterised, or data-aggregated, models

135 Generically-applied DLMs may be appropriate (and thus the value of individually-tailored DLMs
136 diminished) if their purpose is not to examine individual stock status, but rather, to undertake a
137 broader (regional, or global) "health check" of overall sustainability (examples include Rosenberg et
138 al. (2017), Costello et al. (2012, 2016), Kleisner et al. (2013), and Thorson et al. (2012)– see **Table 2**
139 for details). Despite the concern of compounding sources of uncertainty, at such scales, it could be
140 argued that the inaccuracy of DLMs due to lack of data/information may be absorbed, rather than
141 compounded by, the lack of accuracy caused by generic parameterization of DLM. That is, if the
142 uncertainty about key input parameters is large (with wide associated prior distributions) then
143 potentially a number of stocks could be included within the range of uncertainty allowed for that
144 particular stock, and the accuracy or lack thereof would be the same whether or not the DLM is
145 generically parameterised. If the assumptions of the method hold for all species to which it is being
146 applied, then this argument may have some validity.

147 Yet there remain examples of misapplication of empirical DLMs – specifically, the use of aggregated
148 time series of catch or catch-per-unit-effort - in the context of obtaining a regional or global estimate
149 of sustainability. Edgar et al.'s (2018) use of aggregated catch time series across over 200 Australian
150 fisheries, each normalised to its maximum value, to infer that Australian fisheries are in decline,
151 acknowledges but fails to account for management intervention and large-scale environmental
152 changes, fails to weight each time series according to (for example) relative biomass, and fails to
153 acknowledge the lack of desirability of maximum catch as a reference point. The now-classic Myers
154 and Worm (2003) claim of rapid worldwide depletion of predatory fish communities, based on

155 analysis of nominal catch-per-unit-effort data combined globally, was rebutted by Hampton et al.
156 (2005), Maunder et al. (2006), and Polacheck (2006).

157 A preferable approach to identifying regional stock status might be to undertake an in-depth analysis
158 of a sample of stocks from the region, rather than attempting to estimate the status of many stocks
159 using a one-size fits all approach. More broadly, if DLMs are to be applied for purposes of fishery-
160 specific management, then generic parameterisation is inadvisable.

161 ii) Blanket application of a single, or limited number of possible, analytical approach(es)

162 Some regional, and, especially, global analyses have applied a single DLM to all stocks (including
163 regression models from data-rich stocks to make predictions for data-poor stocks (Costello et al.
164 2012)), irrespective of whether better information or assessments are available for some. Blanket
165 application of DLMs may be problematic or inappropriate due to a lack of data, problems with data
166 quality, a lack of required inputs, or the violation of assumptions. However, this has typically been
167 done as a practical approach to provide global assessments of stock status or to assess groups of
168 stocks in the face of limited time and resources, and balanced against the relatively low value of
169 some data-limited stocks. We caution that this may not yield meaningful results for individual stocks,
170 but the outcome is preferable to no assessment and no management.

171 It is clear from the literature that, while the application of DLMs may be simple, these methods are
172 very context-specific and each has its own assumptions and caveats, requiring expert guidance
173 and/or local knowledge (Geromont and Butterworth 2015a,b; Dowling et al. 2014; Carruthers et al.
174 2014; Pilling et al. 2008). Blind application of generic assessment packages, be these data-rich or
175 data-limited, may inadvertently result in erroneous assessment outputs and misinformed guidance
176 because of assumption violations and method outputs not fitting management objectives. An
177 example of potential blanket application of a single, or limited number of analytical approaches
178 includes the aforementioned Australian SAFS approach of, as an initial step, advocating a limited
179 number of assessment methods to resolve status for 200 species (Haddon et al. in prep). In
180 developing nations, authors have observed, during capacity building exercises, the misapplication of
181 assessments recommended for a particular species, to others for which it is not appropriate, or
182 where assumptions are violated.

183 The misapplication of any assessment method, be it data-rich or data-limited, can lead to erroneous
184 results and interpretation. This is more likely to be a problem for DLMs, however, since their
185 information requirements are fewer, and, as such, they can be more readily applied. More generally,
186 blanket application of the same models to a suite of species, or species complexes, may result in

187 varying levels of certainty in parameter estimation, according to the amount and quality of data
188 available (Bentley 2015).

189 Why do we need to be so careful about the generic application of DLMSs?

190 While generic methods may be quicker and easier to implement, their broad or blanket application
191 increases the likelihood of violating assumptions, and of not paying due attention to issues of data
192 quality (errors, gaps, bias) or representativeness (of the stock, of the fishery, and in terms of
193 whether data is temporally or spatially consistent) (**Table 3** provides a summary list of typical
194 assumptions, required inputs, and assumed knowledge associated with DLMSs). This compromises
195 the reliability of generic methods.

196 The application of traditional model-based assessment to data-limited fisheries is often further
197 limited (Dowling et al. 2007, 2008a, 2008b) because many such fisheries also have other
198 complicating features, such as high variability in productivity (e.g., squid and scallops), spatial
199 heterogeneities (e.g., sedentary or low-mobility species) or large numbers of interacting species and
200 gears (e.g., tropical multi-species fisheries).

201 For both developed and developing nations, there is the need to counter a legacy from a “quick-and-
202 dirty”-methods era, when stock assessment tools were made readily available with little training
203 support and little awareness of their limitations. This frequently results in the misapplication of
204 DLMSs, whereby assumptions are violated, or the input data are uninformative. For example, yield-
205 per-recruit models are widely available in a variety of user-friendly applications that also promote
206 simple reference points (e.g., $F=M$). This user friendliness allows access to the method, but is
207 sometimes not flexible enough to include important aspects of a fishery. If selectivity is rigidly
208 assumed to be asymptotic, but true selectivity is dome-shaped (often the case in fisheries that fish
209 shallow, but ontogenetic shifts result in larger individuals in deeper water), the interpretation of the
210 fishing rate at maximum yield (F_{max}) will be incorrect. Increased training and bottom-up engagement
211 and support should help overcome the misapplication of DLMSs, and enable the results of user-
212 friendly applications to be correctly interpreted.

213 Although many DLMSs are computationally simple to undertake (e.g., via the use of packages (e.g.,
214 Carruthers and Hordyk 2018; Cope 2018)), caution needs to be applied to interpretation and use of
215 their outcomes, because of the inherent propensity of DLMSs to yield highly uncertain results.

216 Moreover, the application of DLMSs and interpretation of their output are less simple when the
217 assumptions of the DLM are not respected. In addition, many DLMSs can still be resource-demanding
218 - even, at times, approaching the requirements of a data-rich approach. As such, there is a cost-

219 benefit balance between their generic application versus devoting resources to ensuring the most
220 appropriate method is applied. However, the latter is readily facilitated using decision support tools
221 (e.g. FishPath (Dowling et al. 2016)) to guide the choice of appropriate methods.

222 DLMs have an inherent propensity of to yield highly uncertain results

223 Even if a DLM is appropriate for a fishery or stock, in general, data-limited stock assessments have a
224 higher degree of uncertainty and potential bias in their estimates of stock status than data-rich
225 approaches (Fulton et al. 2016). The extent of this uncertainty is highly case specific. It follows that
226 generic application of DLMs further compounds this uncertainty.

227 All stock assessment methods result in some level of uncertainty in status determination:
228 uncertainty in priors and measurement error in the data, and naturally occurring variability in
229 biology result in a lack of certainty in output. Data rich methods usually deal with this uncertainty
230 explicitly –e.g. they can estimate and report the “probability that the stock is below a limit reference
231 point” or undertake sensitivity tests that can be well defined. By contrast, the uncertainty involved in
232 stock assessments conducted using DLMs is rarely estimated or reported (Dichmont et al. 2017 is an
233 exception). However, to adequately define and calculate the uncertainties associated with DLM
234 methods requires only an awareness of the assumptions of the DLM and unknown parameter inputs.
235 In the USA, the Pacific Fishery Management Council has set the uncertainty around a data-limited
236 assessment to be a coefficient of variation of 1.44 (Ralston et al. 2011). This uncertainty is used to
237 describe the distribution around a catch limit derived from a data-limited method. The risk tolerance
238 is then defined as a certain percentile of the distribution, this determining the reduction from the
239 median. A catch-limit based on a data-limited method, therefore, has a larger reduction in catch
240 than other, less data-limited methods (Ralston et al. 2011).

241 Where uncertainty has been reported, data-limited assessments have been often imprecise for
242 status determination of specific stocks. This has been demonstrated by Dichmont et al. (2017), and
243 by studies (e.g., Klaer et al. 2012; Plaganyi et al. 2013; Dichmont et al. 2017) resulting from the
244 Australian government-funded initiative to explore and understand DLM methods through the
245 “Reducing Uncertainty in Stock Status” program (Larcombe et al. 2015). Although it is difficult to
246 generalise because performance is highly species-specific, the comprehensive management strategy
247 evaluation of Dichmont et al. (2017), that considered assessments embracing fully quantitative
248 model-based assessments, catch curves, empirical analysis of average catch-per-unit-effort CPUE
249 relative to target and limit CPUE reference points (Wayte and Klaer 2010), length-based yield-per-
250 recruit (Haddon et al. 2015), SAFE (Zhou et al. 2011), an empirical catch-based trigger system, and an
251 empirical multi-indicator trigger system (Dowling et al. 2008), showed that, given the same inputs

252 and data, stock-specific data-limited assessments generally tended to be highly uncertain and biased
253 towards overestimating population size. In consequence, more precautionary management is
254 required to achieve the same risk outcomes when using data-limited assessment methods, including
255 generic application of outcomes from data-rich situations (Fulton et al. 2016; Dichmont et al. 2017).

256 As such, care needs to be exercised to ensure that the estimates produced by DLMs are
257 appropriately robust given the circumstances. More thought is also required than is usually given to
258 adequately represent the uncertainties in all status determinations given what one knows about the
259 approach and the unknowns due to lack of information. For example, catch-only assessment
260 methods are known to perform poorly or well in different circumstances especially depending on
261 available knowledge of key inputs such as natural mortality (Wetzel and Punt 2011; Caruthers 2014;
262 Arnold and Heppell 2014).

263 Embedding DLMs within harvest strategies

264 Harvest strategies are formal frameworks for managing exploitation of fisheries, usually applied to
265 the target species (e.g., Sainsbury et al. 2000; Butterworth and Punt 2003, and Fish. Res. Special
266 Issue 94(3) 2008) (Figure 1). They comprise a fully-specified set of rules for making tactical
267 management decisions, including specifications for (i) a monitoring program, (ii) the indicators to be
268 calculated from monitoring data (usually via a stock assessment) and (iii) the use of those indicators
269 and their associated reference points to adjust harvest controls (e.g., total allowable catches)
270 through application of harvest control rules (HCRs, Sainsbury et al. 2000; Butterworth and Punt
271 2003; Punt et al. 2002; Rayns 2007; Butterworth 2007). The motivations for using DLMs in a harvest
272 strategy context are usually resource limitations (e.g., sparse or uninformative data, funding
273 limitations, or a lack of scientific capability). Commonly there is limited time or resources to apply
274 the best (tailor-made) assessment for each, often relatively low-value, stock (Bentley 2015).

275 We encourage the use of harvest strategies in many data limited management contexts. DLMs can
276 be embedded within a harvest strategy with control rules that explicitly link the assessment to a
277 specific management action. Moreover, control rules within a harvest strategy can be designed to
278 be conservative and compensate to some extent for imprecision in the assessment. Simulation and
279 real-life experience have shown DLMs to be biased and have higher variance than data-rich
280 assessments (Dichmont et al. 2016), yet can perform adequately in a precautionary HCR setting
281 (Fulton et al. 2016). DLMs linked to precautionary HCRs within a harvest strategy can therefore
282 perform adequately in avoiding overfishing at the expense of less yield, to compensate for poor
283 estimates of stock status by DLMs.

284 DLMs can still be resource-demanding

285 Ironically, some of the more data-limited assessment options (e.g., ratios of fish density inside
286 compared to outside marine protected areas (Babcock and MacCall 2011; McGilliard et al. 2011),
287 hierarchal decision trees based on combinations of multiple empirical indicators (Prince et al. 2011))
288 are labour intensive, and hence quite costly. Undertaking even the more empirical assessments
289 requires investment of a base level of resources and may also require new data to be collected if
290 stock status is determined to be below target levels, demanding a more rigourous assessment
291 (often, “trigger based” empirical assessment frameworks demand that a more quantitative
292 assessment is undertaken, if a trigger is reached that suggests the stock is below target levels
293 (Dowling et al. 2008)). We reiterate the need to balance the costs of generically applying DLMs, and
294 those of ensuring the most appropriate method is applied.

295 Stock-specific knowledge and solutions are required to guide decision making

296 While there are no simple generic solutions to resolve stock status when data are limited, there are
297 simple stock-specific solutions to manage fisheries in data-limited situations. We would argue that
298 the preferred approach, for the purposes of sustainable management using harvest strategies, is to
299 tailor DLMs to individual stocks or fisheries. Overall management success is strongly dependent on
300 the reliability of the stock status estimate (especially when this is not associated with precautionary
301 HCRs), which we argue can be compromised by the application of generic approaches. Tailored
302 application of DLMs is made more feasible by decision support tools that help users easily select the
303 DLM(s) that are most appropriate for their circumstances (Dowling et al. 2016, Carruthers et al.
304 2012, 2014).

305 A fishery’s specific data collection protocols and operational characteristics, the life history
306 characteristics of the species of interest, and the management objectives and capacity all should be
307 explicitly considered in the application of any assessment. It is these aspects of a stock and its fishery
308 that help match appropriate methods and assumptions to inform management. For instance,
309 knowing not just whether data can be collected or are available for a particular method (e.g., a
310 catch-only method) but whether that method performs well for a given life history (e.g., some are
311 better for faster than slower life history types; Wetzel and Punt 2015) and whether the management
312 system can manage the output of that method, are all fundamental to determining the case-specific
313 appropriateness of any given method. Practitioners should apply a DLM that is appropriate to their
314 data and fishery context (in terms of life history and operational characteristics) (Carruthers et al.
315 2012, 2014, 2016; Dowling et al. 2016).

316 What is the way forward for the application of data-limited assessments?

317 Rather than only seeking simple, generic solutions, we recommend dedicating effort to: i) obtaining
318 better (less measurement error, more spatially and temporally continuous and representative, or
319 additional) data, ii) using care in acknowledging and interpreting the uncertainties in the results of
320 DLMs, iii) embedding DLMs in harvest strategies that are robust to the higher levels of uncertainty in
321 the output of DLMs (both stock status and reference points), by including precautionary
322 management measures (e.g., setting size limits that prevent or limit exploitation until after the size
323 at first or second spawning), or precautionary buffers that set more conservative measures to
324 account for assessment uncertainty, and iv) selecting and applying DLMs appropriate to specific
325 species' and fisheries' data and context.

326 i) There is no substitute for better data

327 There is no replacement for data quantity and quality, particularly time series data, to enable
328 reliable assessment of stock status (Bentley 2015). In the absence of a time series of data and/or
329 current (research, or funding) capacity to conduct assessments, it is important to make a
330 commitment to improve the amount and quality of data informing any assessment, for example, by
331 archiving time series of fishery or biological data for later analysis (Dowling et al. 2008).

332 Improved data can be obtained in cost-effective ways (Dowling et al. 2016). These include market,
333 port or processor monitoring programs, interviews (Moore et al. 2010), voluntary logbooks
334 (Breckwoldt and Seidel 2010), participatory community data collection and assessments (Schroeter
335 et al. 2009; Kittinger 2013), and eliciting local ecological knowledge; (Beaudreau and Levin 2014).
336 These nonetheless require some minimum, ongoing financial and capacity commitment.

337 ii) Acknowledge uncertainties and assumptions

338 DLMs should not be applied as a routine, low-risk or technically trivial exercise. The process,
339 uncertainties and outcomes must be critically confronted. Practitioners should be aware of and
340 report on the limitations of the available data inputs, such as their representativeness, uncertainty
341 (in terms of measurement error and reliability), level of contrast (i.e., that a time series of
342 abundance or proxy abundance data embraces both periods of highs and lows), or spatial/temporal
343 continuity.

344 It is good practice to apply more than one DLM method (and this will usually be technically possible)
345 in order to assess consistency of results (e.g., Fitzgerald et al. 2018). The assumptions and limitations
346 of the chosen DLM(s) should be clearly understood and reported. Where appropriate, uncertainties

347 in inputs (prior ranges for parameters) should be considered via Monte Carlo approaches and
348 sensitivity analyses (e.g., Prince et al. 2011).

349 It follows that care should be taken in interpreting the outcomes of DLMs. The uncertainties in the
350 results should be made explicit. For example, typical DLM problems include the inability of
351 sparse/poor data to update priors to give meaningful posteriors, and difficulty estimating MSY.
352 Effort should be dedicated to considering the implications of this uncertainty as it affects
353 management decision efficacy and future monitoring requirements (e.g. Dowling 2011). HCRs should
354 be adequately precautionary (conservative) given the risk associated with the strength of the data
355 (relative to data-rich stocks) and perceived robustness of the DLM. Such measures may provide
356 incentives to resolve uncertainties identified through sensitivity analysis (sensitivities to assumed
357 model inputs and priors).

358 The legislative, policy and/or commercial (through seafood certification requirements) pressures to
359 evaluate stock status of data-limited fisheries, together with a view that inexpensive, generic
360 approaches are available, begs the question: what level of certainty do agencies or organisations
361 regard as acceptable in moving a species from an “uncertain” stock status to a reportable status
362 designation? It may be misleading to assume that a highly uncertain, yet designated stock status
363 category is preferable over an honest categorization of "uncertain". However, this policy question is
364 rarely addressed. The recent SAFS discussion (Flood et al. 2016) and reconsideration of the
365 Australian Fisheries Management Authority (AFMA) tiers (Dichmont et al. 2017; 2016), the Alaskan
366 tier system (North Pacific Fishery Management Council (NPFMC) 2014), and the U.S. Pacific Fishery
367 Management Council stock assessment categories (Ralston et al. 2011) are exceptions. Demanding
368 that data-limited fisheries meet the same precision and accuracy as data rich fisheries is unrealistic.
369 However, data limited methods have value in identifying species at risk of being overfished, even if
370 the assessment is not "certain" enough to, for example, set catch or effort limits as in a data rich
371 fishery. Generic application of outcomes from data-rich situations may also be appropriate in
372 guiding, for example, the setting of broader control rules such as size limits or gear restrictions.
373 Another valid application of DLMs is via the “Robin Hood” approach of Punt et al. (2011) that uses
374 insights from data-rich assessments to inform data-poor assessments.

375 iii) Embed DLMs in harvest strategies

376 We recommend embedding DLMs within data-limited harvest strategies: precautionary HCRs can
377 compensate for poor estimates of stock status by DLMs. At the same time, we caution that this is
378 not a trivial undertaking: data-limited harvest strategies demand expertise and can sometimes be as
379 time consuming to develop as data-rich ones (Dowling et al. 2016; Plaganyi et al. 2013; Dowling et al.

380 2008), and they are frequently difficult to implement and to simulation test (e.g., Carruthers et al.
381 2016; Dowling et al. 2016). Additionally, it is often difficult to provide defensible indicators and
382 reference points for management frameworks based on empirical assessments. For example, what
383 level of catch or catch rate, at what geographical scale, is considered 'good' and can be related to an
384 overall objective, such as optimising human benefits and avoiding recruitment overfishing? Or, in
385 multispecies fisheries, which species should be considered, what is the basis of the proxy reference
386 points for them, and how should the status of many species be reconciled to yield an overall status
387 for a stock complex, or to determine management advice for the fishery? Despite these challenges,
388 DLMs have been successfully embedded within harvest strategies (Dowling et al. 2008; Fulton et al.
389 2016), and we believe that the use of harvest strategies provides the best outcomes for sustainable
390 fisheries management. A process of bottom-up engagement, facilitated by a decision support tool
391 such as FishPath (Dowling et al. 2016), may best enable the design of effective harvest strategies.

392 iv) Context and consequence matter

393 There is a difference between attempts to assess the status of stocks for situations in which dynamic
394 (based feedback from a stock assessment) management protocols exist - typically, developed
395 countries that set total allowable catches (TACs), and situations in which static ("set and forget")
396 management protocols, such as closed seasons, exist. Many fisheries, especially (but not exclusively)
397 small-scale fisheries from developing nations, remain unassessed and largely unmanaged. In those
398 cases, the imperative to estimate stock status is usually less important than establishing some form
399 of management even in the absence of formal stock assessments and harvest control rules (Mahon
400 1997). Ultimately, the resolution of stock status by an appropriately-selected DLM is still paramount,
401 but the introduction of a formal harvest strategy is certainly not essential, at least initially. The
402 immediate priority is to establish interim management controls identified as feasible and affordable
403 within the social-economic and governance contexts and constraints impacting the fishery. Any
404 required reduction in fishing effort, for example, may have to be achieved as a longer-term goal.

405 More broadly, context and consequence must be considered: the same reasons that resulted in the
406 fishery being data-limited may also cause restrictions on assessment and management options. This
407 is typically the case for small-scale coastal fisheries where data limitations go hand in hand with
408 difficulties in enforcement and management more generally as a result of the fishery being quite
409 complex (e.g., multispecies, multisector, spatially dispersed), which preclude certain forms of
410 harvest controls (Parma et al. 2003; Garcia et al. 2008).

411 A common issue is that of migratory, shared and high seas stocks where there is a contrast in
412 information quality from different regions. Practitioners can either agree to using the best available

413 data, regardless of its place of origin, as being representative to inform an assessment, or “manage
414 their own backyard” to the extent that this is possible, given the extent of site fidelity of fish and the
415 available local data. The Australian Eastern Tuna and Billfish Fishery makes local management
416 decisions informed by a hierarchical decision tree based on combinations of empirical indicators
417 (Prince et al. 2011), which draws more broadly from the regional (South Pacific Commission) stock
418 estimates.

419 Conclusion

420 Costs and associated resources for fisheries management are substantial. Management requires a
421 consistent investment and commitment to research, monitoring, assessment and enforcement, and
422 to the collection of data to inform these. Fisheries need to be managed in a demonstrably
423 sustainable manner in the face of limited resources. Even if generic approaches are used, there is a
424 minimum cost associated with stock status resolution for data-limited fisheries. Recognising that
425 there are no cheap, simple, generic, solutions to evaluating stock status for data-limited fisheries
426 does not preclude agencies and other fisheries stakeholders from implementing pragmatic fisheries
427 management. This should involve commitments to cost-effective data collection programs, careful
428 selection of DLMs based on available data, critical evaluation of their results and their associated
429 uncertainties, and embedding DLMs within appropriate harvest strategies to formalise specific
430 management actions.

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434 **References**

- 435 Anderson, S.C., Cooper, A.B., Jensen, O.P., Minto, C., Thorson, J.T., Walsh, J.C., Afflerbach, J.,
436 Dickey-Collas, M., Kleisner, K.M., Longo, C., & Osio, G.C. (2017). Improving estimates of
437 population status and trend with superensemble models. *Fish and Fisheries*.
- 438 Arnold, L., & Heppell, S. (2015). Testing the robustness of data-poor assessment methods to
439 uncertainty in catch and biology: a retrospective approach. *ICES Journal of Marine Science*, 72,
440 243-250.
- 441 Babcock, E.A., & MacCall, A.D. (2011). How useful is the ratio of fish density outside versus inside no-
442 take marine reserves as a metric for fishery management control rules? *Can. J. Fish. Aquat.*
443 *Sci.*, 68 (2), 343–359.

- 444 Basson, M., & Dowling, N.A. (2008). *Development of a robust suite of stock status indicators for the*
445 *Southern and Western and the Eastern tuna and billfish fisheries*. FRDC Project No. 2003/042.
446 348 pp.
- 447 Battista, W., Karr, K., Sarto, N., & Fujita, R. (2017). Comprehensive Assessment of Risk to Ecosystems
448 (CARE): A cumulative ecosystem risk assessment tool. *Fisheries Research*, 185, 115-129.
449 <http://dx.doi.org/10.1016/j.fishres.2016.09.017>
- 450 Beaudreau, A.H., & Levin, P.S. (2014). Advancing the use of local ecological knowledge for assessing
451 data-poor species in coastal ecosystems. *Ecological Applications*, 24, 244-256.
- 452 Bentley, N. (2015). Data and time poverty in fisheries estimation: potential approaches and
453 solutions. *ICES Journal of Marine Science*, 72(1), 186– 193.
454 <https://doi.org/10.1093/icesjms/fsu023>
- 455 Bentley, N., & Langley, A.D. (2012). Feasible stock trajectories: a flexible and efficient sequential
456 estimator for use in fisheries management procedures. *Canadian Journal of Fisheries and*
457 *Aquatic Sciences*, 69, 161-177. <https://doi.org/10.1139/f2011-143>
- 458 Berkson, J., Barbieri, L., Cadrin, S., Cass-Calay, S., Crone, L., Dorn, P., Friess, M., Kobayashi, C., Miller,
459 D., Patrick, T.J., Pautzke, W.S., Ralston, S., & Trianni, S.M. (2011). *Calculating acceptable*
460 *biological catch for stocks that have reliable catch data only (only reliable catch stocks—*
461 *ORCS)*. NOAA Technical Memorandum NMFS-SEFSC-616. 56 pp.
- 462 Branch, T.A., Jensen, O.P., Ricard, D., Ye, Y., & Hilborn, R. (2011). Contrasting Global Trends in Marine
463 Fishery Status Obtained from Catches and from Stock Assessments. *Conservation Biology*,
464 25(4), 777-786.
- 465 Breckwolddt, A., & Seidel, H. (2012). The need to know what to manage - community-based marine
466 resource monitoring in Fiji. *Current Opinion in Environmental Sustainability*, 4, 331-337.
- 467 Butterworth, D.S. (2007). Why a management procedure approach? Some positives and negatives.
468 *ICES J. Mar. Sci.*, 64, 613–617.
- 469 Butterworth, D.S., & Punt, A.E. (2003). The role of harvest control laws, risk and uncertainty and the
470 precautionary approach in ecosystem-based management. In: *Responsible Fisheries in the*
471 *Marine Ecosystem*, pp. 311–319.
- 472 Caddy, J.F. (2004). Current usage of fisheries indicators and reference points, and their potential
473 application to management of fisheries for marine invertebrates. *Can. J. Fish. Aquat. Sci.*, 60,
474 1307-1324. <http://dx.doi.org/10.1139/f04-132>

- 475 Caddy, J.F. (2009). Practical issues in choosing a framework for resource assessment and
476 management of Mediterranean and Black Sea fisheries. *Mediterr. Mar. Sci.*, 10, 3-119.
477 <http://dx.doi.org/10.12681/mms.124>
- 478 Caddy, J.F., Wade, E., Surette, T., Hebert, M., & Moriyasu, M. (2005). Using an empirical traffic light
479 procedure for monitoring and forecasting in the Gulf of St. Lawrence fishery for the snow crab,
480 *Chionoecetes opilio*. *Fish. Res.*, 76, 123-145. <http://dx.doi.org/10.1016/j.fishres.2005.06.003>
- 481 Carruthers, T.R., Walters, C.J., & McAllister, M.K. (2012). Evaluating methods that classify fisheries
482 stock status using only fisheries catch data. *Fish. Res.*, 119, 66-79.
483 <http://dx.doi.org/10.1016/j.fishres.2011.12.011>
- 484 Carruthers, T.R., Punt, A.E., Walters, C.J., MacCall, A., McAllister, M.K., Dick, E.J., & Cope, J. (2014).
485 Evaluating methods for setting catch limits in data-poor fisheries. *Fish. Res.*, 153, 48-68.
486 <http://dx.doi.org/10.1016/j.fishres.2013.12.014>
- 487 Carruthers, T.R., Kell, L.T., Butterworth, D.S., Maunder, M.N., Geromont, H.F., Walters, C., McAllister,
488 M.K., Hillary, R., Levontin, P., Kitakado, T., & Davies, C.R. (2016). Performance review of simple
489 management procedures, *ICES Journal of Marine Science*, 73(2), 464-482,
490 <https://doi.org/10.1093/icesjms/fsv212>
- 491 Carruthers, T.R., & Hordyk, A. (2018). *Data Limited Methods Toolkit (DLM Tool)*. Retrieved from
492 <https://www.datalimitedtoolkit.org/>
- 493 Chapman, D. G., & Robson, D.S. (1960). The analysis of a catch curve. *Biometrics*, 16, 354-368.
- 494 Cope, J. M. (2013). Implementing a statistical catch-at-age model (Stock Synthesis) as a tool for
495 deriving overfishing limits in data-limited situations. *Fisheries Research*, 142, 3-14.
496 <http://dx.doi.org/10.1016/j.fishres.2012.03.006>.
- 497 Cope, J.M., & Punt, A.E. (2009). Length-based reference points for data-limited situations:
498 Applications and restrictions. *Mar. Coast. Fish.*, 1, 169-186. [http://dx.doi.org/10.1577/C08-](http://dx.doi.org/10.1577/C08-025.1)
499 [025.1](http://dx.doi.org/10.1577/C08-025.1)
- 500 Cope, J.M. (2018). *The Shiny DLM application*. Retrieved from
501 https://github.com/shcaba/Shiny_DLMtool
- 502 Costello, C., Ovando, D., Clavelle, T., Kent Strauss, C., Hilborn, R., Melnychuk, M.C., Branch, T.A., Gaines,
503 S.D., Szuwalski, C., Cabral, R.B., Rader, D.N. & Leland, A. (2016). Global fishery prospects under
504 contrasting management regimes. *PNAS*, 113(18), 5125-5129.

505 Costello, C., Ovando, D., Hilborn, R., Gaines, S.D., Deschenes, O., & Lester, S.E. (2012). Status and
506 solutions for the world's unassessed fisheries. *Science*, 338(6106), 517-520.
507 <http://dx.doi.org/10.1126/science.1223389>Dichmont et al. 2017

508 Dichmont, C.M., Fulton, E., Gorton, R., Sporcic, M., Dowling, N., Little, R.L., Punt, A.E., Haddon, M., &
509 Smith, D.C. (2017). From data rich to data-limited harvest strategies – does more data mean
510 better management? *ICES Journal of Marine Science*, doi:10.1093/icesjms/fsw199

511 Dichmont, C.M., Punt, A.E., Dowling, N., De Oliveira, J.A.A., Little, L.R., Sporcic, M., Fulton, E., Gorton, R.,
512 Klaer, N., Haddon, M., & Smith, D.C. (2016). Is risk consistent across tier-based harvest control
513 rule management systems? A comparison of four case studies. *Fish and Fisheries*, 17(3), 731-747.
514 doi: 10.1111/faf.12142

515 Dichmont, C., & Brown, I. (2010). A case study in successful management of a data-poor fishery using
516 simple decision rules: the Queensland Spanner Crab Fishery. *Mar. Coastal Fish.: Dyn. Manage.*
517 *Ecosyst. Sci.*, 2, 1–13.

518 Dick, E.J., & MacCall, A.D. (2011). Depletion-based stock reduction analysis: a catch-based
519 method for determining sustainable yields for data-poor fish stocks. *Fish. Res.*, 110, 331–341.
520 <http://dx.doi.org/10.1016/j.fishres.2011.05.007>

521 Dowling, N.A., & Smith, D.C. (2007). *Development of harvest strategies for AFMA's small fisheries*. In:
522 Final Report for Project 2006/820 to the Australian Fisheries Management Authority,
523 Canberra.

524 Dowling, N.A., Smith, D.C., Knuckey, I., Smith, A.D.M., Domasch, P., Patterson, H.M., & Whitelaw,
525 W. (2008a). Developing harvest strategies for low-value and data-poor fisheries: case studies
526 from three Australian fisheries. *Fish. Res.*, 94, 380–390.
527 <http://dx.doi.org/10.1016/j.fishres.2008.09.033>

528 Dowling, N.A., Smith, D.C., & Smith, A.D.M. (2008b). *Finalisation of Harvest Strategies for AFMA's*
529 *Small Fisheries*. In: Final Report for Project 2007/834 to the Australian Fisheries Management
530 Authority, Canberra.

531 Dowling, N. (2011.). *Management Strategy Evaluation testing of the Management Strategies used*
532 *with North West Slope Trawl Fisheries*. CSIRO, Marine and Atmospheric Research, Hobart. 86p.

533 Dowling, N.A., Dichmont, C., Haddon, M., Smith, D., Smith, T., & Sainsbury, K. (2014). Empirical
534 harvest strategies for data-poor fisheries. A review of the literature. *Fish. Res.*
535 <http://dx.doi.org/10.1016/j.fishres.2014.11.005>

536 Dowling, N.A., Wilson, J.R., Rudd, M.B., Babcock, E.A., Caillaux, M., Cope, J., Fujita, R., Gedamke, T.,
537 Gleason, M., Gutierrez, N.L., Hordyk, A., Maina, G.W., Mous, P., Ovando, D., Parma, A.M.,
538 Prince, J., Revenga, C., Rude, J., Szuwalski, C., Valencia, S. & Victor, S. (2016). *FishPath: A*
539 *Decision Support System for Assessing and Managing Data and Capacity-Limited Fisheries*.
540 Proceedings of the 30th Lowell Wakefield Fisheries Symposium, Anchorage, Alaska, USA
541 (Alaska Sea Grant College Program Report). Fairbanks, Alaska: University of Alaska Sea Grant
542 College Program.

543 Dunn, A., Francis, R.I.C.C., & Doonan, I.J. (2002). Comparison of the Chapman-Robson and regression
544 estimators of Z from catch-curve data when non-sampling stochastic error is present. *Fisheries*
545 *Research*, 59, 149–159.

546 Edgar, G.J., Ward, T., & Stuart-Smith, R.D. (2018). Rapid declines across Australian fishery stocks
547 indicate global sustainability targets will not be achieved without an expanded network of
548 'no-fishing' reserves. *Aquatic Conserv: Mar Freshw Ecosyst*. DOI: 10.1002/aqc.2934

549 FAO (2016). *Implementation of the 1995 FAO Code of Conduct for Responsible Fisheries* - Web site,
550 updated 16 February 2016. Code of Conduct for Responsible Fisheries. FI Institutional
551 Websites. In: FAO Fisheries and Aquaculture Department [online]. Rome.
552 <http://www.fao.org/docrep/005/v9878e/v9878e00.htm>.

553 FAO (1995) *The State of World Fisheries and Aquaculture – 1994 (SOFIA)*, Rome, Italy: Food and
554 Agriculture Organization.

555 Fitzgerald, S. P., Wilson, J. R., & Lenihan, H. S. (2018). Detecting a need for improved management in
556 a data-limited crab fishery. *Fisheries Research*, 208, 133-144.

557 Flood, M. J., Stobutzki, I., Andrews, J., Ashby, C., Begg, G.A., Fletcher, R., Gardner, C., Georgeson, L.,
558 Hansen, S., Hartmann, K, Hone, P., Larcombe, J., Maloney, L., Moore, A., Roach, J., Roelofs, A.,
559 Sainsbury, K., Saunders, T., Sloan, S., Smith, A.D.M., Stewart, J., Stewardson, C., & Wise B.S.
560 (2016). Multijurisdictional fisheries performance reporting: How Australia's nationally
561 standardised approach to assessing stock status compares. *Fisheries Research*, 183, 559–573.

562 Free, C. M., Jensen, O. P., Weidenmann, J., & Deroba, J. J. (2017). The refined ORCS approach: A catch-
563 based method for estimating stock status and catch limits for data-poor fish stocks. *Fisheries*
564 *Research*, 193, 60-70. <http://dx.doi.org/10.1016/j.fishres.2017.03.017>

565 Fox, W. W. Jr. (1970). An exponential surplus-yield model for optimizing exploited fish populations.
566 *Trans.Am.Fish.Soc.*, 99, 80-88.

567 Froese, R., Demirel, N., Coro, G., Kleisner, K. M. & Winker, H. (2017), Estimating fisheries reference
568 points from catch and resilience. *Fish and Fisheries*, 18, 506–526. doi:10.1111/faf.12190

569 Froese, R., Zeller, D., Kleisner, K., & Pauly, D. (2012). What catch data can tell us about the status of
570 global fisheries. *Marine Biology*, 159, 1283-1292.

571 Fulton, E.A., Punt, A.E., Dichmont, C.M., Gorton, R., Sporcic, M., Dowling, N., Little, L.R., Haddon, M.,
572 Klaer, N. & Smith, D.C. (2016). Developing risk equivalent data-rich and data-limited harvest
573 strategies. *Fisheries Research*, 183, 574-587. 10.1016/j.fishres.2016.07.004.

574 Fujita, R., Thornhill, D.J., Karr, K., Cooper, C.H., & Dee, L.E. (2014). Assessing and managing data-limited
575 ornamental fisheries in coral reefs. *Fish and Fisheries*, 15(4), 661-675.
576 <http://dx.doi.org/10.1111/faf.12040>

577 Garcia, S.M., Allison, E.H., Andrew, N.J., Béné, C., Bianchi, G., de Graaf, G.J., Kalikoski, D., Mahon, R.,
578 & Orensanz, J.M. (2008). *Towards integrated assessment and advice in small-scale fisheries:
579 principles and processes*. FAO Fisheries and Aquaculture Technical Paper. No. 515. Rome, FAO.
580 84p.

581 Gedamke, T., & Hoenig, J.M. (2006). Estimating Mortality from Mean Length Data in Nonequilibrium
582 Situations, with Application to the Assessment of Goosefish. *Trans. Am. Fish. Soc.*, 135, 476-
583 487.

584 Geromont, H.F. & Butterworth, D.S. (2015a). Generic management procedures for data-poor
585 fisheries: forecasting with few data. *ICES Journal of Marine Science*, 72(1), 251–261.
586 <https://doi.org/10.1093/icesjms/fst232>

587 Geromont, H.F. & Butterworth, D.S. (2015b). *A review of assessment methods and the development
588 of management procedures for data-poor fisheries*. FAO Report. The Marine Resource
589 Assessment and Management Group (MARAM). University of Cape Town, South Africa 219pp

590 Gulland, J.A. (1971). *The fish resources of the ocean*. Fishing News Books, West Byfleet, UK.

591 Haddon, M. (2010). *Modelling and quantitative methods in fisheries*. Chapter 11. CRC press.

592 Haddon, M. (2010). The tier 4 analyses 1986–2009. In: G.N. Tuck (ed.), *Stock Assessment for the
593 Southern and Eastern Scalefish and Shark Fishery 2009. Part 2* (pp. 319–369). Australian
594 Fisheries Management Authority and CSIRO Marine and Atmospheric Research, Hobart.

595 Halliday, R.G., Fanning, L.P., & Mohn, R.K. (2001). *Use of the traffic light method in fishery
596 management planning*. CSAS Research Document 2001/108, 41 pp. 084

597 Hampton, J., Sibert, J. R., Kleiber, P., Maunder, M. N., & Harley, S. J. (2005). Decline of Pacific tuna
598 populations exaggerated? *Nature*, 434, E1–E2.

599 Hilborn, R., & Walters, C.J. (1992). *Quantitative fisheries stock assessment: Choice, dynamics and*
600 *uncertainty*. Chapman & Hall, New York, pp. 570. [http://dx.doi.org/10.1007/978-1-4615-3598-](http://dx.doi.org/10.1007/978-1-4615-3598-0)
601 0

602 Hinton, M. G., & Maunder, M. N. (2004). Methods for standardizing CPUE and how to select among
603 them. *Col. Vol. Sci. Pap. ICCAT*, 56(1), 169-177.
604 <http://www.iotc.org/sites/default/files/documents/proceedings/2008/wpb/IOTC-2008-WPB->
605 [INF01.pdf](http://www.iotc.org/sites/default/files/documents/proceedings/2008/wpb/IOTC-2008-WPB-INF01.pdf)

606 Hobday, A. J., Smith, A., Webb, H., Daley, R., Wayte, S., Bulman, C., Dowdney, J., Williams, A.,
607 Sporcic, M., Dambacher, J., Fuller, M., & Walker, T. (2007). *Ecological Risk Assessment for the*
608 *Effects of Fishing: Methodology*. Report R04/1072 for the Australian Fisheries Management
609 Authority, Canberra

610 Hobday, A. J., Smith, A. D. M., Stobutzki, I.C., C. Bulman, C., Daley, R., Dambacher, J.M., Deng, R.A.,
611 Dowdney, J., Fuller, M., Furlani, D., Griffiths, S.P., Johnson, D., Kenyon, R., Knuckey, I.A., Ling,
612 S.D., Pitcher, R., Sainsbury, K.J., Sporcic, M., & Zhou, S. (2011a). Ecological Risk Assessment for
613 the Effects of Fishing. *Fisheries Research*, 108, 372–384.
614 <https://doi.org/10.1016/j.fishres.2011.01.013>

615 Hobday, A. J., Bulman, C., Williams, A., & Fuller, M. (2011b). *Ecological risk assessment for effects of*
616 *fishing on habitats and communities* FRDC report 2009/029.
617 http://frdc.com.au/research/Final_Reports/2009-029-DLD.pdf

618 Hordyk, A., Ono, K., Valencia, S., Loneragan, N., & Prince, J. (2015). A novel length-based empirical
619 estimation method of spawning potential ratio (SPR), and tests of its performance, for small-
620 scale, data-poor fisheries. *ICES J. Mar. Sci.*, 72, 217-231.
621 <http://dx.doi.org/10.1093/icesjms/fsu004>

622 Hordyk, A., Ono, K., Sainsbury, K., Loneragan, N., & Prince, J. (2015). Some explorations of the life
623 history ratios to describe length composition, spawning-per-recruit, and the spawning
624 potential ratio. *ICES J. Mar. Sci.*, 71, 204-216. <http://dx.doi.org/10.1093/icesjms/fst235>

625 Klaer, N.L., Wayte, S.E., & Fay, G. (2012). An evaluation of the performance of a harvest strategy that
626 uses an average-length based assessment method. *Fish. Res.*, 134, 42–51.

627 Kittinger, J.N. (2013). Participatory Fishing Community Assessments to Support Coral Reef Fisheries
628 Comanagement. *Pacific Science*, 67, 361-381.

629 Kleisner, K., Zeller, D., Froese, R. & Pauly, D. (2013). Using global catch data for inferences on the
630 world's marine fisheries. *Fish and Fisheries*, 14, 293-311. DOI: 10.1111/j.1467-
631 2979.2012.00469.x

632 Larcombe, J., Noriega, R. & Stobutzki, I. (eds) (2015). *Reducing uncertainty in stock status*.
633 Unpublished report, ABARES, Canberra.

634 Lombardi, L. & Walters, C. (2011). *Stochastic Stock Reduction Analysis (SRA) User Guide*. NOAA
635 Fisheries Service, Southeast Fisheries Science Center, Panama City Laboratory, 3500 Delwood
636 Beach Road, Panama City, Florida 32408. Panama City Laboratory Contribution 11-03. p. 26.

637 Mahon, R. (1997). Does fisheries science serve the needs of managers of small stocks in developing
638 countries? *Canadian Journal of Fisheries and Aquatic Sciences*, 54, 2207-2213.

639 Martell, S. & Froese, R. (2013). A simple method for estimating MSY from catch and resilience. *Fish and*
640 *Fisheries*, 14, 504–514. doi:10.1111/j.1467-2979.2012.00485.x

641 Maunder, M.N., & Punt, A.E. (2004). Standardizing catch and effort data: A review of recent
642 approaches. *Fisheries Research*, 70(2-3), 141-159.
643 <https://doi.org/10.1016/j.fishres.2004.08.002>

644 Maunder, M. N., Sibert, J. R., Fonteneau, A., Hampton, J., Kleiber, P., & Harley, S. (2006). Interpreting
645 catch-per-unit-of-effort data to assess the status of individual stocks and communities. *ICES*
646 *Journal of Marine Science*, 63, 1373–1385.

647 MacCall, A.D. (2009). Depletion-corrected average catch: a simple formula for estimating sustainable
648 yields in data-poor situations. *ICES J. Mar. Sci.*, 66, 2267–2271.
649 <http://dx.doi.org/10.1093/icesjms/fsp209>

650 McClanahan, T. R., Graham, N. A. J., MacNeil, M. A., Muthiga, N. A., Cinner, J. E., Bruggemann, J. H., &
651 Wilson, S. K. (2011). Critical thresholds and tangible targets for ecosystem-based management of
652 coral reef fisheries. *Proceedings of the National Academy of Sciences*, 4-7.
653 www.pnas.org/cgi/doi/10.1073/pnas.1106861108

654 McGarvey, R., & Matthews, J. M. (2001). Incorporating numbers harvested in dynamic estimation of
655 yearly recruitment: onshore wind in interannual variation of South Australian rock lobster (*Jasus*
656 *edwardsii*). *ICES Journal of Marine Science*, 58, 1092.

657 McGarvey, R., Matthews, J. M., & Prescott, J. H. (1997). Estimating lobster recruitment and exploitation
658 rate from landings by weight and numbers, and age-specific weights. *Marine and Freshwater*
659 *Research*, 48, 1001–1008. <https://doi.org/10.1071/MF97209>

660 McGarvey, R., Punt, A. E., & Matthews, J. M. (2005). Assessing the information content of catch-in-
661 numbers: a simulation comparison of catch and effort data sets. In Kruse, G.H., Gallucci, V.F., Hay,
662 D.E., Perry, R.I., Peterman, R.M., Shirley, T.C., & Spencer P.D. et al. (Eds.), *Fisheries Assessment*
663 *and Management in Data-Limited Situations* (pp. 635–653). Alaska Sea Grant College Program,
664 University of Alaska, Fairbanks. <http://nsgl.gso.uri.edu/aku/akuw03002/ak-sg-05-02p635-682.pdf>

- 665 McGilliard, C.R., Hilborn, R., MacCall, A., Punt, A.E., & Field, J.C. (2011). Can information from marine
666 protected areas be used to inform control-rule-based management of small-scale, data-poor
667 stocks? *ICES J. Mar. Sci.*, 68, 201–211.
- 668 Mesnil, B., & Petitgas, P. (2009). Detection of changes in time-series of indicators using CUSUM control
669 charts. *Aquatic Living Resources*, 22(2), 187–192. <https://doi.org/10.1051/alr/2008058>
- 670 Moore, J.E., Cox, T.M., Lewison, R.L., Read, A.J., Bjorkland, R., McDonald, S.L., Crowder, L.B., Aruna,
671 E., Ayissi, I., Espeut, P., Joynson-Hicks, C., Pilcher, N., Poonian, C.N.S., Solarin, B., & Kiszka, J.
672 (2010). An interview-based approach to assess marine mammal and sea turtle captures in
673 artisanal fisheries. *Biological Conservation*, 143, 795-805.
- 674 Myers, R. A., & Worm, B. (2003). Rapid worldwide depletion of predatory fish communities. *Nature*,
675 423, 280–283.
- 676 Newman, D., Berkson, J., & Suatoni, L. (2015). Current Methods for Setting Catch Limits for Data-Limited
677 Fish Stocks in the United States. *Fisheries Research*, 164, 86–93.
- 678 North Pacific Fishery Management Council (NPFMC) (2014). *Stock Assessment and Fishery Evaluation*
679 *Report for the king and Tanner crab fisheries of the Bering Sea and Aleutian Islands Regions*.
680 North Pacific Fishery Management Council, 605 W. 4th Avenue, #306, Anchorage, AK 99501.
- 681 NRC (National research Council) (2013). *Evaluating the effectiveness of fish stock rebuilding plans in*
682 *the United States*. The National Academies Press, Washington DC.
- 683 Oremus, K.L., Suatoni, L. & Sewell, B. (2014). The requirement to rebuild US fish stocks: Is it working?
684 *Marine Policy*, 47, 71-75.
- 685 Parma, A.M., Orensanz, J.M., Elías, I. & Jerez, G. (2003). Diving for shellfish- and data: incentives for
686 the participation of fishers in the monitoring and management of artisanal fisheries around
687 southern South America. In Newman, S.J., Gaughan, D.J., Jackson, G., Mackie, M.C., Molony,
688 B., St John, J. & Kaiola, P. (Eds.), *Towards sustainability of data-limited multi-sector fisheries*
689 (pp. 8-29). Australian Society for Fish Biology Workshop Proceedings, Bunbury, Australia. 23-
690 24 September 2001, Department of Fisheries, Perth, Australia, 186 pp.
- 691 Patrick, W.S., Spencer, P., Link, J., Cope, J., Field, J., Kobayashi, D., Lawson, P., Gedamke, T., Cortes, E.,
692 Ormseth, O., Bigelow, K., & Overholtz, W. (2010). Using productivity and susceptibility indices to
693 assess the vulnerability of United States fish stocks to overfishing. *Fish. Bull.*, 108(3), 305-322.
- 694 Pilling, G.M., Apostolaki, P., Failer, P., Floros, C., Large, P.A., Morales-Nin, B., Reglero, P., Stergiou,
695 K.I., & Tsikliras, A.C. (2008). Assessment and management of data-poor fisheries. In: Payne, A.,

696 Cotter, J. and Potter, T. (Eds.), *Advances in Fisheries science: 50 years on from Beverton and*
697 *Holt* (pp. 280-305). Blackwell Publishing, CEFAS.

698 Pitcher, T.J. & Preikshot, D. (2001). RAPFISH: a rapid appraisal technique to evaluate the sustainability
699 status of fisheries. *Fish. Res.*, 49(3), 255–270. [http://dx.doi.org/10.1016/s0165-7836\(00\)00205-8](http://dx.doi.org/10.1016/s0165-7836(00)00205-8)

700 Pitcher, T.J., Lam, M., Ainsworth, C., Martindale, A., Nakamura, K., Perry, R.I., & Ward, T.
701 (2013). Improvements to Rapfish: a rapid evaluation technique for fisheries integrating ecological and
702 human dimensions. *J. Fish Biol.*, 83, 865–889. <http://dx.doi.org/10.1111/jfb.12122>

703 Plaganyi, E.E., Skewes, T., Murphy, N., Pascual, R., & Fischer, M. (2015). Crop rotations in the sea:
704 Increasing returns and reducing risk of collapse in sea cucumber fisheries. *Proceedings of the*
705 *National Academy of Sciences of the United States of America*, 112, 6760-6765.

706 Plaganyi, E.E., Skewes, T.D., Dowling, N.A., & Haddon, M. (2013). Risk management tools for
707 sustainable fisheries management under changing climate: a sea cucumber example. *Clim.*
708 *Change*, 119(1), 181–197. <http://dx.doi.org/10.1007/s10584-012-0596-0>.

709 Polacheck, T. (2006). Tuna longline catch rates in the Indian Ocean: did industrial fishing result in a
710 90% rapid decline in the abundance of large predatory species? *Marine Policy*, 30, 470–482.

711 Prince, J.D., Dowling, N.A., Davies, C.R., Campbell, R.A., & Kolody, D.S. (2011). A simple cost-effective
712 and scale-less empirical approach to harvest strategies. *ICES Journal of Marine Science*, 68,
713 947-960. <http://dx.doi.org/10.1093/icesjms/fsr029>

714 Prince, J., Hordyk, A., Valencia, S.R., Loneragan, N., & Sainsbury, K. (2014). Revisiting the concept of
715 Beverton–Holt life-history invariants with the aim of informing data-poor fisheries
716 assessment. *ICES J. Mar. Sci.*, 72, 194-203. <http://dx.doi.org/10.1093/icesjms/fsu011>

717 Punt, A.E., Smith, A.D.M., & Cui, G.R. (2002). Evaluation of management tools for Australia’s South
718 East Fishery 3. Towards selecting appropriate harvest strategies. *Mar. Freshwater Res.*, 53,
719 645–660.

720 Punt, A.E., Smith, D.C., & Smith, A.D.M. (2011). Formal use of Inter-Stock and –Species Comparisons
721 for Improving Stock Assessments of Data-Poor Species: the Robin Hood approach. *ICES Journal*
722 *of Marine Science*, 68, 972–981.

723 Rayns, N. (2007). The Australian government’s harvest strategy policy. *ICES J. Mar. Sci.*, 64, 596–598

724 Rosenberg, A. A., Kleisner, K. M., Afflerbach, J., Anderson, S. C., Dickey Collas, M., Cooper, A. B., Fogarty,
725 M. J., Fulton, E.A., Gutierrez, N.L., Hyde, K.J.W., Jardim, E., Jensen, O.P., Kristiansen, T., Longo, C.,
726 Minte-Vera, C.V., Minto, C., Mosqueira, I., Osio, G.C., Ovando, D., Selig, E.R., Thorson, J.T., Walsh,

727 J.C., & Ye, Y. (2017). Applying a new ensemble approach to estimating stock status of marine
728 fisheries around the world. *Conservation Letters*, 11, 1-9. doi: 10.1111/conl.12363

729 Quinn, T.J. & Deriso, R.B. (1999). *Quantitative Fish Dynamics*. Oxford University Press. New York.

730 Ralston, S., Punt, A.E., Hamel, O.S., DeVore, J.D., & Conser, R. (2011). A meta-analytic approach to
731 quantifying scientific uncertainty in stock assessments. *Fishery Bulletin*, 109(2), 217-232.

732 Rudd, M. B., & Thorson, J. T. (0000). Accounting for variable recruitment and fishing mortality in length-
733 based stock assessments for data-limited fisheries. *Can. J. Fish. Aquat. Sci.*, 00, 1–17
734 dx.doi.org/10.1139/cjfas-2017-0143

735 Sainsbury, K.J., Punt, A.E., & Smith, A.D.M. (2000). Design of operational management strategies for
736 achieving fishery ecosystem objectives. *ICES J. Mar. Sci.*, 57, 731–741.

737 Scandol, J.P. (2003). Use of cumulative sum (CUSUM) control charts of landed catch in the management
738 of fisheries. *Fish. Res.*, 64, 19-36. [http://dx.doi.org/10.1016/S0165-7836\(03\)00104-8086](http://dx.doi.org/10.1016/S0165-7836(03)00104-8086)

739 Scandol, J. (2005). Use of quality control methods to monitor the status of fish stocks. In: Kruse, G.H.,
740 Gallucci, V.F., Hay, D.E., Perry, R.I., Peterman, R.M., Shirley, T.C., Spencer, P.D., Wilson, B., &
741 Woodby, D. (Eds.), *Fisheries assessment and management in data-limited situations* (pp. 213-
742 233). Alaska Sea Grant, University of Alaska Fairbanks. <http://dx.doi.org/10.4027/famdis.2005.13>

743 Schaefer, M. (1954). Some aspects of the dynamics of populations important to the management of the
744 commercial marine fisheries. *Bull.I-ATTC/Bol. CIAT*, 1(2), 27-56.

745 Schaefer, M. (1957). A study of the dynamics of the fishery for yellowfin tuna of the eastern tropical
746 Pacific Ocean [in English and Spanish]. *Ibid.*, 2(6), 245-285.

747 Schroeter, S.C., Gutierrez, N., Robinson, M., Hilborn, R., & Halmay, P. (2009). Moving from Data Poor
748 to Data Rich: A Case Study of Community-Based Data Collection for the San Diego Red Sea
749 Urchin Fishery. *Marine and Coastal Fisheries*, 1, 230-243.

750 Smith, M. W., Then, A.Y., Wor, C., Ralph, G., Pollock, K.H., & Hoenig, J.M. (2012). Recommendations
751 for catch-curve analysis. *North American Journal of Fisheries Management*, 32, 956–967.
752 <http://dx.doi.org/10.1080/02755947.2012.711270>

753 Then, A.Y., Hoenig, J.M., Gedamke, T., & Ault, J.S. (2016). Comparison of Two Length-Based
754 Estimators of Total Mortality: A Simulation Approach. *Trans. Am. Fish. Soc.*, 145(6), 1206-1219.

755 Thorson, J.T., Branch, T.A., & Jensen, O.P. (2012). Using model-based inference to evaluate global
756 fisheries status from landings, location, and life history data. *Can. J. Fish. Aquat. Sci.*, 69(4),
757 645-655.

758 Walters, C.J., Martell, S.J.D., & Korman, J. (2006). A stochastic approach to stock reduction analysis.
759 *Can. J. Fish. Aquat. Sci.*, 63, 212-223.

760 Wayte, S.E., & Klaer, N.L. (2010). An effective harvest strategy using improved catch-curves. *Fish.*
761 *Res.*, 106 (3), 310–320.

762 Wetzel, C.R., & Punt, A.E. (2015). Evaluating the performance of data-moderate and catch-only
763 assessment methods for U.S. west coast groundfish. *Fisheries Research*, 171, 170-187.

764 Wetzel, C.R., & Punt, A.E. (2011). Performance of a fisheries catch-at-age model (stocksynthesis) in
765 data-limited situations. *Mar. Freshwater Res.*, 62, 927–936.

766 Wiff, R., Quiroz, J.C., Neira, S., Gacitúa, S., & Barrientos, M.A. (2016). Chilean fishing law, maximum
767 sustainable yield, and the stock-recruitment relationship. *Lat. Am. J. Aquat. Res.*, 44(2), 380-
768 391.

769 Wilson, J.R., Prince, J.D., & Lenihan, H.S. (2010). A management strategy for sedentary nearshore
770 species that uses marine protected areas as a reference. *Mar. Coast. Fish.*, 2, 14-27.
771 <http://dx.doi.org/10.1577/C08-026.1>

772 Zeller D., Palomares M.L.D., Tavakolie A., Ang M., Belhabib D., Cheung W.W.L., Lam V.W.Y., Sy E.,
773 Tsui G., Zylich K. & Pauly D. (2016). Still catching attention: Sea Around Us reconstructed
774 global catch data, their spatial expression and public accessibility. *Marine Policy*, 70, 145-152.

775 Zhou, S. & Griffiths, S.P. (2008). Sustainability assessment for fishing effects (SAFE): a new quantitative
776 ecological risk assessment method and its application to elasmobranch bycatch in an Australian
777 trawl fishery. *Fish. Res.*, 91, 56–68. <http://dx.doi.org/10.1016/j.fishres.2007.11.007>

778 Zhou, S.J., Griffiths, S.P., & Miller, M. (2009). Sustainability assessment for fishing effects (SAFE) on
779 highly diverse and data-limited fish bycatch in a tropical prawn trawl fishery. *Mar. Freshwater*
780 *Res.*, 60, 563–570. <http://dx.doi.org/10.1071/MF08207>

781 Zhou, S., Hobday, A.J., Dichmont, C.M., & Smith, A.D.M. (2016a). Ecological risk assessments for the
782 effects of fishing: A comparison and validation of PSA and SAFE. *Fisheries Research*,
783 <http://dx.doi.org/10.1016/j.fishres.2016.07.015>.

784 Zhou, S., Chen, Z., Dichmont, C.M., Ellis, A.N., Haddon, M., Punt, A.E., Smith, A.D.M., Smith, D.C., & Ye, Y.
785 (2016b). *Catch-based methods for data-poor fisheries*. Report to FAO. CSIRO, Brisbane, Australia.

786 Zhou, S., Punt, A.E., Ye, Y., Ellis, N., Dichmont, C.M., Haddon, M., Smith, D.C., & Smith, A.D.M. (2017).
787 Estimating stock depletion level from patterns of catch history. *Fish and Fisheries*, 18, 742-751.
788 doi: 10.1111/faf.12201

789 **Table 1:** Overview of data-limited assessment methods, from empirical (in which performance
 790 indicators are based on directly-measured properties) to model-based (where performance
 791 indicators are model outputs).

Input-based category	Data-limited assessment	References
Expert Judgement	Move directly to harvest control measures	Dowling et al. (2014)
	Discourse/expert judgement	Dowling et al. (2008a)
	Data exploration via plotting and descriptive statistics	Dowling et al. (2008a)
	Analysis of changes in the spatial distribution of fishing effort	Dowling et al. (2008a)
	Analysis of changes in the spatial distribution of catch	Dowling et al. (2008a)
Empirical Reference Points	Analysis of changes in gear type or manner of deployment	Dowling et al. (2008a)
	Size-based sequential trigger system	Dowling et al. (2008a)
	Sequential effort triggers	Dowling et al. (2008a)
Risk Analysis/Vulnerability	Sequential catch triggers	Dowling et al. (2008a)
	Comprehensive assessment of risk to ecosystems (CARE)	Battista et al. (2017); Fujita et al. (2014)
	Ecosystem threshold analysis	McClanahan et al. (2011)
	RAPFISH (Multi-dimensional scaling)	Pitcher and Preikshot (2001); Pitcher et al. (2013)
	Productivity and Susceptibility Analysis (PSA) to estimate risk of overfishing	Patrick et al. (2010)
Abundance Indicators	Ecological Risk Assessment for the Effects of Fishing (ERAEF)	Hobday et al. (2007, 2011a,b)
	Sustainability Assessment for Fishing Effects (SAFE)	Zhou et al. (2016a); Zhou et al. (2011); Zhou and Griffiths (2008, 2009)
MPA	Analysis of changes in species-composition	Dowling et al. (2008a)
	Use of biomass surveys to inform management	Dowling et al. (2008a); Parma et al. (2003)
	Single-indicator analysis using standardized CPUE	Hinton and Maunder (2004); Maunder and Punt (2004)
	Linear regression to recent time series of CPUE	Haddon (2010); Maunder and Punt (2004); Dichmont and Brown (2010)
Catch Only	Analysis of ratio of density inside and outside marine protected areas (MPAs)	Babcock and MacCall (2011); McGilliard et al. (2011)
	Analysis of length/size-specific catch-rate indicators for fish sampled inside and outside of marine protected areas (MPAs), and per-recruit	Wilson et al. (2010)
Population Dynamics Model	Optimized catch-only method (OCOM)	Zhou et al. (2016b)
	Boosted Regression Tree (BRT) model for stock depletion using catch data	Zhou et al. (2017)
	Only Reliable Catch Series (ORCS)	Berkson et al. (2011); Free et al. (2017)
	Depletion-Corrected Average Catch (DCAC)	MacCall (2009)
	Depletion-Based Stock Reduction Analysis (DB-SRA)	Dick and MacCall (2011)
	Simple Stock Synthesis (SSS)	Cope (2013)
	Stochastic Stock Reduction Analysis (stochastic SRA)	Lombardi and Walters (2011); Walters et al. (2006)
	Catch-MSY/CMSY (MSY = maximum sustainable yield)	Froese et al. (2017); Martell and Froese (2013)
	Feasible stock trajectories	Bentley and Langley (2012)
	Depletion analysis	Hilborn and Walters (1992)
Production model	Fox (1970); Hilborn and Walters (1992); Haddon (2010); Schaefer (1954, 1957)	
Statistical catch-at-age (SCAA)	Hilborn and Walters (1992); Quinn and Deriso (1999)	
qR Method	McGarvey and Matthews (2001); McGarvey et al. (1997); McGarvey et al. (2005)	
Size/Age-Based	Analysis of size relative to size at maturity	Basson and Dowling (2008)
	Analysis of changes in mean length/weight or length/weight percentiles	Dowling et al. (2014); Quinn and Deriso (1999)
	Analysis of sustainability indicators based on length-based reference points (LBRP)	Cope and Punt (2009)
	Catch curve analysis	Chapman and Robson (1960); Dunn et al. (2002); Gulland (1971); Smith et al. (2012)
	Length-based Spawning Potential Ratio (LB-SPR)	Hordyk et al. (2015a, b)
Multiple Indicators	Mortality estimates from length data in non-equilibrium situations	Gedamke and Hoenig (2006)
	Length-based Integrated Mixed Effects (LIME)	Rudd and Thorson (in press)
	Hierarchical decision trees	Dowling et al. (2014); Prince et al. (2011)
assessment	Traffic lights	Caddy (2004, 2009); Caddy et al. (2005); Halliday et al. (2001)
	Cumulative Sum (CUSUM) Control Charts	Mesnil and Petitgas (2009); Scandol (2003, 2005)
	Sequential trigger framework involving catch and/or effort, catch-per-unit-effort (CPUE), size, sex ratio etc.	Dowling et al. (2008a)

800 **Table 2:** Review of
 801 regional or global
 802 assessment

803 applications, with examples of one-size fits-all approaches and the rationale or impetus for each.
 804 Note that the methods here summarised are not intended as examples of incorrect application.

Reference	Method	Rationale/what is estimated	Limitation
Costello et al. 2012	Multivariate regression to identify predictors of stock status from assessed fisheries and use these models to estimate status of unassessed fisheries	Estimates the status of collections (including the global status) of previously unassessed stocks	Per authors, this approach does not produce precise estimates for individual fisheries and therefore is not a substitute for formal assessment.
Costello et al. 2016	Global regression analysis merged with Catch-MSY used to estimate intrinsic population growth rate (r), carrying capacity (K), maximum sustainable yield (MSY), and merged with a microlevel structural bioeconomic model based on global regression analysis	Global MSY and individual stock status based on 4713 fisheries in RAM (Dr. Ransom A. Myers) Legacy and FAO marine capture databases	Emphasis was on investigating alternative approaches to recovering depleted fisheries. Fisheries that failed to meet minimum criteria excluded
Froese et al. 2012	Comparison of maximum catch with MSY and of catch-based analysis with biomass-based analysis.	Evaluates whether maximum catch is correlated with MSY, and whether temporal trends in catch data are consistent with trends in biomass.	Limited to FAO (Food and Agriculture Organization of the United Nations) catch database. Not aiming to resolve stock status for intent of formal management.
Rosenberg et al. 2017	Four catch-only methods were applied, and estimates from these were combined using a superensemble. The 4 catch-only methods included one empirical model (panel regression approach (PRM) developed by Costello et al. (2012)), and three mechanistic models (catch-MSY; catch-only model with sampling-importance resampling (COMSIR), and state-space catch-only model (SSCOM).	Quantitative estimates of exploitation status for 785 FAO fish stocks.	Estimates of stock status were global and within each FAO statistical region. Per authors, there are still many limitations to using this information for stock-specific or even regional advice. These include the high variability of the estimates, the need for longer time series of data, limited life history information for many stocks, and the difficulties of assigning prior distributions.
Branch et al. 2011	Analysed i) simulated random catch data with no trend; ii) stocks classified as collapsed on the basis of catch data to determine whether these stocks actually were collapsed; iii) stock assessments to compare stock status derived from catch data with status derived from biomass data.	How use of catch data affects assessment of fisheries stock status.	Limited to FAO catch database, and stock assessments in RAM Legacy database. Not aiming to resolve stock status for intent of formal management.
Thorson et al. 2012	This model uses logistic regression to extrapolate from assessment results to available landings, life history, and location data. The model classifies stocks into different prediction bins and estimates the probability of collapse in each using cross-validation.	Assesses whether globally available landings, life history, and location information are sufficiently informative to allow model-based predictions regarding stocks for which single-species assessments are not available. An extrapolation model uses stock assessment results to estimate parameters for a model that predicts the probability of collapse.	Estimates probability of collapse. Is a "global extrapolation model" which uses the opportunistic data that are available on a global scale for evaluating fishery questions.
Kleisner et al. 2013	Reviews stock status plots as means to provide a robust overview of fisheries and of the major trends besetting them	Considers challenges, improvements, and uses of stock status plots, and examines their expected performance under alternate scenarios	Uses global catch data; provides broad and indirect overview of stock status. Not intended to guide management of individual stocks. Acknowledges that catch statistics are not a silver bullet when it comes to evaluating stock status, but points out that they are the only means to obtain a global picture of stock status, when analysed with an understanding of the scenarios which may cause misinterpretations.

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808 **Table 3:** A list of common assumptions, required inputs and assumed knowledge associated with data-limited assessments, with some examples of
 809 assessments to which these apply. Acronyms and abbreviations are as defined in Table 1.

Assumption	Example assessments	810
The approach assumes the population is currently in equilibrium (with constant recruitment)	Catch curves; LB-SPR	
The approach requires that unexploited, or theoretical equilibrium, biomass (B_0) is stationary (i.e. meaningful regardless of time)	811 Many	
Selectivity needs to be at least able to be inferred	812 Most	
Selectivity has not changed over time	813 Most	
Asymptotic fishing selectivity	Catch curves; LB-SPR	Figure 1: The fisheries adaptive management cycle, underpinned by a harvest strategy
There have been no major temporal changes in sampling patterns, fishing operational characteristics, management, markets or the environment	Most	
Species is actively and consistently targeted	Many	
Sampling is representative of the stock	Most	
Sampling is representative of the spatial extent of the fleet(s)	Most	
Required inputs		
Estimate of the ratio of fishing mortality at maximum sustainable yield (FMSY) relative to natural mortality (M) (FMSY/M)	DCAC; DB-SRA; SSS	
Estimate of FMSY	Stochastic SRA;	
Estimates of von Bertalanffy growth parameters	LBRP; Size-specific catch rate indicators for fish sampled inside and outside of MPAs, and per-recruit; LB-SPR; SSS; Stochastic SRA; Feasible stock trajectories; Mortality estimates from length data in nonequilibrium situations; SCAA; qR method; LIME	
Estimate of steepness	SSS; Feasible stock trajectories; SCAA, qR method; LIME	
Estimate of natural mortality	Size-specific catch rate indicators for fish sampled inside and outside of MPAs, and per-recruit; Catch curves; SAFE; DCAC; DB-SRA; SSS; Stochastic SRA; Feasible stock trajectories; Mortality estimates from length data in nonequilibrium situations; SCAA; qR method; LIME; OCOM	
Estimate of length-fecundity relationship	SCAA	
Estimate of size at maturity	Hierarchical decision trees; LBRP; DB-SRA; LB-SPR; SSS; Stochastic SRA; Feasible stock trajectories; SCAA; LIME	
Estimate of life-history ratio M/k	LB-SPR	
Estimate of length-weight relationship	SSS; Feasible stock trajectories; SCAA	
Prior estimate for stock status (depletion)	RAPFISH; BRT model for stock-depletion using catch data; ORCS; DCAC; DB-SRA; SSS; Catch-MSY; Feasible stock trajectories	
Assumed knowledge		
Some notion of spatial distribution	SAFE	
The number of times mortality is thought to change, and initial guesses of the years during which mortality is thought to change	Mortality estimates from length data in nonequilibrium situations	

