# Assessment of 11 Northeast Groundfish Stocks through 1999 

# A Report to the New England Fishery Management Council's Multi-Species Monitoring Committee 

 byNorthern Demersal Working Group, Northeast Regional Stock Assessment Workshop
U.S. DEPARTMENT OF COMMERCE

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Northeast Region
Northeast Fisheries Science Center
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## Section 1. Introduction

The Northern Demersal Working Group met on a continuing basis throughout July, 1999 to prepare and review assessments for 11 groundfish stocks. A final review and follow up meeting was held during 27-28 July, 1999. Participants included:
A. Applegate, NEFMC
J. Brodziak, NEFSC
R. Brown, NEFSC
S. Cadrin, NEFSC
S. Clark, NEFSC
S. Correia, MA, DMF
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T.P. Smith, NEFSC
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M. Terceiro, NEFSC
S. Wigley, NEFSC

Each of the stocks had been assessed during 1998 or 1999, and all assessments were peerreviewed during the $27^{\text {th }}, 28^{\text {th }}$, or $29^{\text {th }}$ Stock Assessment Review Committee meetings. For 10 of the 11 stocks, terminal year 1997 fishing mortality and January 1, 1998 survivors had been estimated during the 1998 assessment period at SAW 27 or SAW 28. The Gulf of Maine witch flounder stock was assessed during 1999, and the assessment was reviewed at SAW 29. For this stock, estimates of terminal year 1998 fishing mortality and January 1, 1999 survivors were directly available from the recent SAW.

The Northern Demersal Working Group was requested to meet inter-sessionally between the $29^{\text {th }}$ and $30^{\text {th }}$ Stock Assessment Workshops during July, 1999 to prepare and review updated stock assessments for the remaining 10 stocks through the beginning of calendar year 1999.
In addition, the Working Group was requested to develop medium-term forecasts, taking into account starting stock conditions as of Jan1,2000, with the forecast horizon extending through calendar year 2009. To bridge 1999, at this point during the year (July), the Working Group concluded that, barring any definitive evidence of a change in 1999 fishing effort versus 1998 fishing effort for these stocks, a status quo F scenario would be applied in the forecasts for 1999 (i.e., $\mathrm{F}_{1999}=\mathrm{F}_{1998}$ ).

## Overview

Of the 11 stocks included in the Working Group review. 7 were assessed via VPA and 4 were updated via 1 -year projection of F given the 1998 landings. This report represents a summary of the assessment and forecast results presented to the Working Group at its 27-28 July Final Review Meeting. The report consists of an initial overview of stock projection and forecast methodology, followed by a concise section for each of the 11 stocks in the review. A concluding section contains overall comments by the Working Group. an evaluation of the quality of the data and assessments. and recommendations for future improvement.

The following stocks are addressed in this report:
A. Georges Bank cod
B. Georges Bank haddock
C. Georges Bank yellowtail flounder
D. Southern New England yellowtail flounder
E. Cape Cod yellowtail flounder
F. Gulf of Maine cod
G. Gulf of Maine/Georges Bank witch flounder
H. Gulf of Maine/Georges Bank American plaice
I. Georges Bank winter flounder
J. Southern New England winter flounder
K. Gulf of Maine/Georges Bank white hake

A summary of current stock status, control rule targets, and medium-term projections results for eleven groundfish stocks is given in the following table.

|  |  | target F |  |  |  | year $\mathrm{B}>$ Bmsy |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | stock | B99/Bmsy | on biomass fully-recruited | $50 \%$ prob. | $90 \%$ prob |  |
| SNE yellowtail | 0.11 | 0.00 | 0.00 | 2007 | $>2009$ |  |
| white hake | 0.25 | 0.00 | 0.00 | 2004 | 2005 |  |
| GB winter fl. | 0.30 | 0.00 | 0.00 | 2004 | 2006 |  |
| GOM cod | 0.39 | 0.15 | 0.22 | 2007 | 2009 |  |
| GB cod | 0.41 | 0.13 | 0.19 | 2006 | 2008 |  |
| GB haddock | 0.43 | 0.00 | 0.00 | 2002 | $>2009$ |  |
| CC yellowtail | 0.58 | 0.04 | 0.07 | 2001 | 2002 |  |
| Am. plaice | 0.60 | 0.02 | 0.02 | 2003 | 2004 |  |
| SNE winter fl. | 0.90 | 0.24 | 0.43 | 2001 | $>2009$ |  |
| GB yellowtail | 0.93 | 0.27 | 0.50 | 2002 | $>2009$ |  |
| witch flounder | 1.03 | 0.09 | 0.11 | 1999 | 2001 |  |

## Assessment Issues

## Decreasing Port Sampling Coverage

For many stocks the quantity and the seasonal and market category coverage of the commercial port samples has decreased considerably. This has likely introduced more variability into the estimation of landings at age and, consequently, more uncertainty into the assessment. This issue should be addressed to prevent the sampling coverage from worsening, and to initiate procedures to ensure adequate coverage of the landings.

## Conditioning of $q$ in ASPIC

In some stocks, the catchability (q) in the ASPIC model was fixed to obtain consistency between VPA and ASPIC biomass estimates. This also achieves a more reasonable estimate of r, the intrinsic rate of increase of the stock. However, the adjustment of q may affect the estimates of Bmsy obtained from ASPIC.

## Conversion of fully recruited F to biomass-weighted F

Since the SFA control rule has been developed and specified in terms of mean biomass and biomass-weighted fishing mortality, the estimates of fully recruited F associated with the agestructured VPA results must be converted to biomass-weighted F and vice versa. The translation between these two measures of $F$ will vary depending on the age structure of the stock at the time of the translation. This may cause confusion, and may introduce apparent inconsistencies.

In addition, for some stocks, there is no catch at age 1 or age 2 . If the calculation of biomassweighted $F$ in the SFA control rule reflects all ages, it has been necessary to translate age-specific Fs from the VPA to biomass-weighted Fs by accounting for all ages, including those for which F is 0.0. An adjustment to the control rule F and mean biomass may be required to restore consistency between the two measures of F and biomass.

The Working Group considers this to be an issue to be addressed further.

For the purposes of this analysis. the Working Group has determined the appropriate biomassweighted F based on the SFA control rule for each stock. To execute the 10 -year projections in an age-structured framework, the biomass-weighted F was converted to a corresponding fullyrecruited F . This value of F was then held constant throughout the 10 -year forecast horizon. A more detailed explanation and illustration of the SFA control rule is given in the following section.

## Harvest Control Rules

Many of the harvest control rules in Amendment \#9 were recommended based on deterministic biomass projections (Applegate et al. 1998). Threshold and target fishing mortalities were derived from estimates of minimum biomass that could rebuild to $\mathrm{B}_{\text {MSY }}$ at a constant rebuilding F within a fixed time period (Cadrin 1999).

For the example depicted below, rebuilding targets and thresholds are based on a 5-year rebuilding period. If mean biomass in 1999 was estimated to be 22.000 mt . the target F in 2000 would be 0.18 . The target $F(0.18)$ would be maintained in subsequent years as the stock grows to $\mathrm{B}_{\mathrm{MSY}}$ (which is expected to take 5 years given the historical productivity of the stock).


Therefore target F does not increase during the rebuilding period as stock biomass increases.
Although the control rule targets account for uncertainty in MSY reference points, they do not account for uncertainty in stock status determination. Uncertainty in current F and biomass should be considered to assess the risk of overfishing. because their probability distributions are negatively correlated (i.e., the uncertainty cloud around the point labeled '1999' on the graph above would slope up and to the left).

## References

Applegate, A., S. Cadrin, J. Hoenig, C. Moore, S. Murawski. and E. Pikitch. 1998. Evaluation of existing overfishing definitions and recommendations for new overfishing definitions to comply with the Sustainable Fisheries Act. New England Fishery Management Council Report.

Cadrin. S.X. 1999. A precautionary approach to fishery control rules based on surplus production modeling. National Oceanographic and Atmospheric Administration Technical Memorandum NMFS-F/SPO-40: 17-22.

## Section 2. Projection Methodology

## Stock-Recruitment Data and Analyses for Eleven Northeast Groundfish Stocks.

### 1.0 Background

Time-series of Stock-Recruitment data from VPA's for the eleven groundfish stocks were summarized to determine the suitability of the data for forecasting purposes. A set of exploratory graphs were produced for each stock. These included plots of recruitment (year class strength) vs spawning stock biomass(SSB), recruit per spawner ratios (RSSB) vs spawning stock biomass. and recruitment, SSB, and RSSB vs time. A scatterplot smoother was also applied to these plots to give an indication of trend. Autocorrelation plots for rssb were also generated to evaluate whether there was serial correlation in productivity which might be expected if environmental effects persisted through time. A summary data table for each series and basic statistics for each variable were also produced.

Graphical results (see individual stock sections B-L) showed that the relationship between estimated recruitment and spawning stock varied considerably through the available time series. Some stocks appeared to produce higher recruitment at high spawning stock sizes, while others exhibited no obvious relationship between recruitment and spawning stock. There were time trends in recruitment and spawning stock for some stocks. Some stocks also appeared to have a trend in recruit per spawner ratios. Overall, the graphical results indicated that the available stock-recruitment data were very noisy and that the determination of functional relationships between spawning stock and recruitment is difficult.

### 2.0 Analyses of the Appropriateness of using Stock-Recruitment Relationships for Northeast Groundfish

Several analyses were completed to determine if basic assumptions from stock-recruitment theory were evident in the empirical data for the eleven Northeast groundfish stocks. Myers and Barrowman (1996) analyzed 364 stock-recruitment data sets to examine hypotheses relating recruitment to spawner abundance, concluding that there was strong evidence to support a relationship between spawner abundance and recruitment in general and within several families of fish. We used similar non-parametric rank and ratio estimates to examine hypotheses for the eleven stocks. We were interested in the same set of hypotheses as Myers and Barrowman (1996), (1) if the largest recruitment occurs at the highest $\mathrm{SSB},(2)$ if the smallest recruitment occurs at the lowest SSB, and (3) if recruitment is greater when SSB is above the median SSB (Table 1). In addition we looked at a similar approach for RSSB ratios. We were interested in (1) the rank of the maximum RSSB at high SSB, (2) the rank of the minimum RSSB at low SSB, and (3) the ratio when SSB was above or below the median (Table 1).

Results from these analyses indicated that maximum recruitment had occurred at above average stock size ranking in 6 out of 11 stocks. Similarly, we found that minimum observed recruitment had occurred at below average stock size ranking in 6 out of 11 stocks (Table 1). The ratio of average recruitment when SSB was above its median to average recruitment when SSB was
below its median was greater than 1.0 in 6 out of 11 stocks as well. Together, these measures suggested that extreme values of recruitment were not tightly linked to ssb levels for some stocks. They also suggested that average recruitment levels were not strongly affected by SSB for some stocks.

Results from the analyses on relative productivity (rssb) were more conclusive. In 11 out of 11 stocks, the maximum value of rssb occurred at a below-average stock ranking (Table 1). Similarly, in 8 out of 11 stocks, the minimum rssb value occurred at an above-average stock ranking. The ratio of average rssb when ssb was above its median to its average when ssb was below its median was less than 1 in 10 out of 11 stocks. Overall, these results suggested that maximum values of rssb were associated with below average stock sizes and that minimum values were associated with above average stock sizes. The apparent decline in rssb as ssb increased in 10 out of 11 stocks was consistent with the notion that compensation influenced the relative productivity of these stocks.

We also used contingency tables to compute the probabilities of obtaining various combinations of recruitment and spawner biomass. These were useful for computing the odds of obtaining high or low recruitment at high or low ssb. These 'odds ratios' were also useful for examining the benefits of higher spawning biomass for Northeast groundfish (Table 2). In fact. the odds ratios showed that recruitment likely improves when ssb is above its median in 7 out of 11 stocks. Note that these contingency table analyses were conducted on data available prior to July $15^{\text {th }}$ and that changes to the VPA stock-recruitment series made after this date may affect results.

### 3.0 Stock Recruitment Modeling and Forecasting

A Beverton-Holt ( $\mathrm{B} \& \mathrm{H}$ ) stock-recruitment model with a multiplicative lognormal error structure was previously used in ten year projections for the five New England groundfish stocks (cod. haddock, yellowtail flounder) during SARC 24 (NEFSC 1997; NRC 1998). This form of the model (FORM 1 in equation (1) below)

$$
\begin{equation*}
R=\frac{a S}{b+S} \times e^{\prime \prime} \tag{1}
\end{equation*}
$$

as well as another form of the B\&H model (FORM 3 in equation (2) below)

$$
\begin{equation*}
R=\frac{0.8 \times z \times R \max \times S}{0.2 \times \operatorname{Smax}(1-z)+S(z-0.2)} \tag{2}
\end{equation*}
$$

were thoroughly investigated as potential sources for modeling stock-recruitment in the current exercise. We parameterized B\&H models for the eleven stocks and used results from feasible parameter estimations in preliminary ten year forecasts to determine the applicability of this approach for modeling medium-term stock dynamics of groundfish (Table 3). These trial projections were performed in relation to control rules and required rebuilding schedules currently in place for New England groundfish.

Results from the stock-recruitment model fits were infeasible for Georges Bank winter flounder. Gulf of Maine witch flounder, and American plaice (Table 3. z values >1.0). Each of these stocks appeared to exhibit overcompensation and/or time trending of recruitment and rssb. As a result. the B\&H models were not used for these stocks. Two yellowtail stocks (Cape Cod and Southern New England) showed high resilience (Table 3, z values near 1.0). Of these. Southern New England yellowtail also exhibited high variability about the mean $\mathrm{B} \& H$ curve and also showed evidence of serial correlation in residuals from the model fit. In contrast, Cape Cod yellowtail had relatively low variability and also had residuals that did not appear to be lognormal. Regardless of the model fits, however, the average relationship between recruitment and spawning stock for these two stocks was effectively constant over the observed range of stock sizes. Three stocks had moderate levels of resilience (Table $3, \mathrm{z}$ values between 0.4 and 0.75 ): Georges Bank yellowtail, Gulf of Maine cod, and white hake. Of these, white hake had a pattern of serial correlation in residuals from the model fit. Three other stocks exhibited low resiliency ( $z$ values between 0.15 and 0.35 ): Georges Bank haddock, Georges Bank cod, and Southern New England winter flounder. Of these, Georges Bank haddock exhibited strong departures from B\&H model assumptions. In addition, the point estimate of $z$ of 0.18 for Georges Bank haddock suggested mild depensation in this stock-recruitment curve.

Results suggested that forecasts with these models may have been too optimistic for some stocks (Gulf of Maine cod, Georges Bank haddock, white hake, Southern New England winter flounder), Results for these stocks appeared inconsistent with ASPIC estimates of K and or Bmsy, historic research survey biomass indices, VPA biomass estimates, and feasible increases in biomass from average recruitment. However. forecasts for other stocks such as Georges Bank cod and yellowtail flounder and Southern New England yellowtail flounder using the estimated Beverton-Holt model, were reasonable when compared to other information on stock productivity. For stocks where a stock-recruitment model was not feasible or overly optimistic, the empirical recruitment data were used to generate recruitment over the ten year horizons. It is important to note that this work is currently still in progress and further testing, validation. and software enhancements are necessary. Nonetheless, projection results to date can serve as the foundation for future research and development of approaches for medium and long-term forecasts.

### 4.0 Sources of Uncertainty

## "Prediction is very difficult, especially about the future." Niels Bohr

Projection results may be optimistic and should be interpreted with the following caveats:

- Stock-recruitment data are limited. With the short-time series available. long-term relationships are difficult to discern, especially given the inherent variability in recruitment. Further. there is a general lack of observations of recruitment levels at high spawning stock sizes and this leads to imprecise determination of maximum expected recruitment levels. It is important to note that, with the exception of Georges Bank haddock, the length of the 10-year projection period is roughly half or more of the available time series of stock-recruitment data.
- Measurement error in estimates of spawning stock and recruitment may have obscured patterns in stock-recruitment data. Current stock assessment models may have retrospective patterns in estimates of spawning stock and recruitment that cannot be incorporated in these analysies.
- Several stocks appear to have recent trends in productivity. Modeling nonstationary stock-recruitment relationships is very difficult without prior information on the magnitude and duration of trends.
- The possibility of compensatory changes in weight at age and maturation probability at age are not included in the 10 -year projections. This implies that projection results may be optimistic because at higher stock sizes, individual growth rates may decline. and assumed future values of population mean weights at age may be biased high.
- Genetic diversity of some of these stocks may have been reduced through intensive exploitation. As a result, some stocks may be less fit to compete in portions of their historic range. Similarly, near shore components of some stocks may have been adversely affected by habitat loss and pollution. Recolonization of some near shore and estuarine areas may not be possible without improved habitat. Potential losses in genetic diversity and habitat are impossible to quantify, however, without decadal time series of baseline information.


## Section 3. Stock Assessments and Medium-Term Forecasts

The following sections A-K contain a summary of the assessment results through 1999 and medium-term forecasts through 2009 for 11 Northeast groundfish stocks.

## A. Georges Bank Atlantic Cod by L. O'Brien

### 1.0 Background

This stock was last assessed in 1998 and reviewed by the $27^{\text {th }}$ Northeast Regional SAW. Fully recruited $F$ (ages $4-8, \mathrm{u}$ ) was estimated to be 0.26 , the lowest in the time series (1978-1997) and about $45 \%$ higher than $\mathrm{F}_{0.1}=0.18$. Spawning stock biomass was 36.000 mt in 1997. continuing to increase from the record low estimate of $25,000 \mathrm{mt}$ in 1994. Recruiting year classes continued to decline with the four most recent year classes (1994-1997) being the lowest on record with recruit/SSB survival ratios below the long term average. The NEFSC spring and autumn bottom trawl survey indices continued to decline and remain near record low values. Recruitment indices for the 1994-1996 year classes were among the lowest in the time series.

### 2.0 1999 Assessment

## The Fishery

Total commercial landings of Georges Bank cod (Table A1, Figure A1) in 1998 ( $8,800 \mathrm{mt}$ ) declined $15 \%$ from 1997. USA landings declined about $8 \%$ from $7,500 \mathrm{mt}$ to 7.000 mt and Canadian landings declined about $35 \%$ from 2,900 mt to $1,900 \mathrm{mt}$ (Table A1). No discards estimates were derived for 1998. Recreational landings were estimated at 515 mt , a decline of about $33 \%$ from 1997.

The total number of commercial length samples in 1998 were about equal to the 1997 samples. The number of quarterly samples were adequate for the scrod and market category and were poor for the large market category. The large samples were pooled on a semi-annual basis. Spatial coverage was poor for eastern Georges Bank (SA 561, 562). As in the last assessment. length samples from western Georges Bank and US and Canadian combined age samples from eastern Georges Bank were applied to characterize the landings from eastern Georges Bank. Landings were dominated by age 3 fish in both the US and Canadian fisheries.

## Input data and Analyses

The current assessment is an update of the 1998 assessment (O'Brien and Cadrin 1999) and assumes the same VPA formulation. A slight variation from the previous assessment is that the number of surveys available as tuning indices in the terminal year increases from two to three with the addition of the US spring survey. Catch at age has been updated with total 1998 landings (US and Canadian) and NEFSC 1998 spring and autumn survey indices and NEFSC and Canadian 1999 spring survey indices.

A conditional non-parametric bootstrap procedure (Efron 1982) was used to evaluate the precision of fishing mortality and spawning stock biomass estimates. A retrospective analysis was performed for terminal year fishing mortality, spawning stock biomass, and age 1 recruitment.

### 3.0 Assessment Results

NEFSC spring and autumn survey indices increased in both abundance and biomass in 1998. The 1999 NEFSC spring indices declined to similar values observed in 1997 (Table A2. Figures A2 and A3). The Canadian spring survey index of abundance declined in 1998. then increased in 1999 to a value similar to 1997 (Figure A3). The recruitment index of NEFSC autumn age 1 in 1998 remains well below average.

Fully recruited fishing mortality (age 4-8) is estimated as 0.28 in 1998, an $8 \%$ increase from the 1997 F ( 0.26 ) estimated in the previous assessment (O'Brien and Cadrin 1999). The current assessment now estimates the 1997 fishing mortality as 0.53 (Figure A4). Spawning stock biomass in 1998 was estimated as $28,700 \mathrm{mt}$ an $8 \%$ increase from 1997 and a $27 \%$ increase from the record low SSB in 1994 (Table A3, Figure A5). Recruitment in 1998 is estimated to be similar to the 1992 and 1996 year classes (Table A3, Figure A5). The recruit/SSB survival ratios for the 1995. 1996. and 1998 year classes were above average and the 1997 year class had the lowest survival in the time series. Mean biomass increased from a record low 31.000 mt in 1995 to $38,000 \mathrm{mt}$ in 1997 and declined to $36,000 \mathrm{mt}$ in 1998 (Table A3, Figure A5). Biomass weighted fishing mortality declined from a time series high of 0.63 in 1993 to a time series low of 0.24 in 1998 (Table A3, Figure A4).

## VPA Diagnostics

Stock size estimates for ages $1-8$ were well estimated with CVs ranging from 0.37 to 0.44 . Ages 2. 3. and 4 all have positive residuals in 1999 indicating that all survey estimates may be too low. and ages 6,7 . and 8 all have negative residual indicating that the survey estimates are too high. The distribution of F estimates from the bootstrap analysis ranged from 0.2 to 0.43 with an $80 \%$ probability that $F$ in 1998 was between 0.26 and 0.32 . Estimates of SSB ranged from 20.000 mt to 36.000 mt with an $80 \%$ probability that SSB in 1998 was between 24.000 mt and 32.000 mt .

A retrospective pattern exists in this model formulation back to 1994. In the terminal year, fishing mortality has been underestimated and spawning stock biomass has been overestimated each year since 1994 and recruitment has been underestimated since 1995.

### 4.0 Forecasts

An analysis of the stock-recruit data indicated a relationship of increased recruitment with increased spawning stock biomass. Parameters estimated by a Beverton-Holt stock recruitment model were statistically feasible and this model was used to generate recruitment for forecasts of stock size and landings for 1999-2009. Constraints derived from the $0 \%$, $75 \%$, and $100 \%$ quartiles of the time series of $\mathrm{R} / \mathrm{SSB}$ survival ratios were imposed on the recruitment when the SSB reached a level of 50.000 mt . Input data are presented in Table A4. Catch mean weights, stock mean weights, and the partial recruitment were based on the average of the last 5 years (1994-1998).

Short term forecasts of stock size and landings were performed for 1999-2000 where the $F$ in 1999 was assumed to be equal to the $F$ in 1998 given the lack of complete 1999 landings data (Table A5). SSB is predicted to increase $13 \%$ from $28,700 \mathrm{mt}$ in 1998 to $32,500 \mathrm{mt}$ in 1999. Mean biomass in 1999 is estimated to increase $21 \%$ from $36,300 \mathrm{mt}$ in 1998 to $43,900 \mathrm{mt}$ in 1999.

The SFA control rule for Georges Bank cod is based on $\mathrm{B}_{\text {MSY }}(108,000 \mathrm{mt})$ and states that when the stock biomass is between $1 / 4$ and $1 / 2 \mathrm{~B}_{\mathrm{MSY}}(27,000-54,000 \mathrm{mt})$, the threshold mortality rate is defined by a five year rebuilding time period, and if the stock is between $1 / 2 \mathrm{~B}_{\text {MSY }}$ and $\mathrm{B}_{\text {MSY }}$ the rebuilding time period is 10 years. In 1999 , mean biomass is predicted to be about 44.000 mt . less than $1 / 2 \mathrm{~B}_{\mathrm{MSY}}$. Applying the 1999 mean biomass to the target control rule (Figure A6) indicates that the stock is to be fished at biomass weighted F no higher than 0.125 .

Long term forecasts from 2000-2009 were conducted with the fully recruited F equivalent ( 0.185 ) of the biomass weighted $\mathrm{F}(0.125)$ required to achieve compliance with the SFA control rule. Landings, SSB, and mean biomass trajectories are presented in Table A5. There is a $50 \%$ probability that $\mathrm{B}_{\text {MSY }}$ will be achieved by 2006 (Table A5, Figure A7).

## 5. 0 Sources of Uncertainty

There was poor commercial sampling in 1998 both temporally and spatially. The large market category was not well sampled by quarter, and samples from eastern GB were minimal.

There is a retrospective pattern in the VPA. Fishing mortality has been underestimated since 1994 and recruitment has been underestimated since 1995.

Short term projections are likely optimistic if fishing mortality is underestimated in 1998.
There is inadequate data to characterize both the recreational and discarded catch, particularly if these components increase. The SARC previously rejected using poorly sampled recreational catch since it would only be a scaling factor if the age structure is the same.

### 6.0 References

O`Brien, L. and S.X. Cadrin. 1999. Assessment of the Georges Bank Atlantic Cod Stock for 1998. NEFSC Ref. Doc 99-03.

Efron. B. 1982. The jackknife, the bootstrap and other resampling plans. Phila. Soc. Ind. and Appl. Math. 34: 92 p.

Table A1. Commercial landings (metric tons, live) of Atlantic cod from Georges Bank and South (Division 52 and Subarea 6)., 1960 - 1998.


Table A2. Standardized stratified mean catch per tow in numbers and weight (kg) for Atlantic cod in NEFSC offshore spring and autumn research vessel bottom trawl surveys on Georges Bank (Strata 13-25), 1963-1999. [a,b, c]

[a] During 1963-1984, BMV oval doors were used in spring and autumn surveys; since 1985, Portuguese polyvalent doors have been used in both surveys. Adjustments have been made to the 1963-1984 catch per tow data to standardize these data to polyvalent door equivalents. Conversion coefficients of 1.56 (numbers) and 1.62 (weight) were used in this standardization (NEFC 1991).
[b] Spring surveys during 1980-1982, 1989-1991 and 1994 and autumn surveys during 1977-1981, 19891991, and 1993 were accomplished with the RIV Delaware II; in all other years, the surveys were accomplished using the R/V Albatross IV. Adjustments have been made to the RIV Delaware II catch per tow data to standardize these to RIV Albatross IV equivalents. Conversion coefficients of 0.79 (numbers) and 0.67 (weight) were used in this standardization (NEFC , 1991).
[c] Spring surveys during 1973-1981 were accomplished with a '41 Yankee' trawl; in all other years, spring surveys were accomplished with a '36 Yankee' trawl. No adjustments have been made to the catch per tow data for these gear differences.
[d] Excludes unusually high catch of 1894 cod ( 2558 kg ) at Station 230 (Strata tow 20-4).
[e] Excludes unusually high catch of $1032 \mathrm{cod}(4096 \mathrm{~kg})$ at Station 323 (Strata tow 16-7).
[f] Excludes unusually high catch of $111 \operatorname{cod}(504 \mathrm{~kg})$ at Station 205 (Strata tow 23-4).

Table A3. Estimates of beginning year stock size (thousands of fish), instantaneous fishing mortality ( $F$ ), mean biomass (mt), spawning stock biomass ( mt ), and percent mature of Georges Bank cod, estimated from virtual population analysis (VPA), calibrated using the commercial catch at age ADAPT
formulation, 1978-1998.

## Stock Numbers (Jan 1 ) in thousands

|  | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 27713 | 23514 | 20105 | 41393 | 17471 | 9615 | 27393 | 8674 | 42758 | 16379 | 23460 | 15723 | 9250 | 18016 | 7005 | 9366 | . 7802 | 4247 | 7576 | 8639 | 2280 | 10469 |
| 2 | 4268 | 22688 | 19220 | 16380 | 33865 | 14004 | 7774 | 22354 | 6980 | 34867 | 13386 | 19198 | 12873 | 7567 | 14703 | 5672 | 7664 | 6386 | 3477 | 6202 | 7070 | 1867 |
| 3 | 25526 | 3139 | 16776 | 12319 | 10511 | 19458 | 7588 | 5183 | 12488 | 4515 | 21784 | 9533 | 13829 | 6068 | 4816 | 8259 | 3709 | 5915 | 4874 | 2659 | 4610 | 5120 |
| 4 | 7947 | 13888 | 1755 | 8461 | 6266 | 5145 | 8635 | 3115 | 2033 | 6086 | 2425 | 10577 | 5161 | 6760 | 2033 | 1979 | 2920 | 1656 | 3885 | 3173 | 1599 | 2700 |
| 5 | 2878 | 4422 | 6964 | 985 | 4698 | 2609 | 1990 | 4051 | 1313 | 943 | 3064 | 1069 | 4900 | 2522 | 2566 | 726 | 611 | 739 | 729 | 2064 | 1801 | 926 |
| 6 | 1124 | 1605 | 2524 | 3613 | 594 | 2037 | 1181 | 869 | 1611 | 640 | 520 | 1154 | 576 | 1963 | 746 | 760 | 196 | 144 | 343 | 379 | 972 | 1181 |
| 7 | 1434 | 802 | 900 | 1093 | 1686 | 232 | 965 | 500 | 339 | 752 | 296 | 205 | 455 | 264 | 623 | 245 | 195 | 74 | 78 | 169 | 192 | 581 |
| 8 | 67 | 862 | 587 | 334 | 517 | 772 | 104 | 376 | 212 | 199 | 371 | 97 | 93 | 151 | 102 | 230 | 57 | 36 | 37 | 50 | 63 | 122 |
| 9 | 146 | 12 | 477 | 402 | 162 | 231 | 419 | 45 | 124 | 109 | 106 | 126 | 40 | 44 | 60 | 53 | 59 | 5 | 16 | 27 | 27 | 39 |
| 10+ | 54 | 148 | 28 | 190 | 187 | 148 | 293 | 206 | 76 | 68 | 98 | 45 | 89 | 43 | 18 | 28 | 9 | 2 | 1 | 12 | 18 | 27 |

$1+\quad 71158 \quad 71081 \quad 69336 \quad 85170 \quad 759585425156344453746793464557655105772647267433993267227316232231920321015233761863123032$

## Fishing Mortality

|  | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | -1998 |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 0 | 0 | 0 | 0 | 0.02 | 0.01 | 0 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0.11 | 0.1 | 0.24 | 0.24 | 0.35 | 0.41 | 0.21 | 0.38 | 0.24 | 0.27 | 0.14 | 0.13 | 0.55 | 0.25 | 0.38 | 0.22 | 0.06 | 0.07 | 0.07 | 0.1 | 0.12 |  |  |  |  |  |
| 3 | 0.41 | 0.38 | 0.48 | 0.48 | 0.51 | 0.61 | 0.69 | 0.74 | 0.52 | 0.42 | 0.52 | 0.41 | 0.52 | 0.89 | 0.69 | 0.84 | 0.61 | 0.22 | 0.23 | 0.31 | 0.34 |  |  |  |  |  |
| 4 | 0.39 | 0.49 | 0.38 | 0.39 | 0.68 | 0.75 | 0.56 | 0.66 | 0.57 | 0.49 | 0.62 | 0.57 | 0.52 | 0.77 | 0.83 | 0.97 | 1.17 | 0.62 | 0.43 | 0.37 | 0.35 |  |  |  |  |  |
| 5 | 0.38 | 0.36 | 0.46 | 0.31 | 0.64 | 0.59 | 0.63 | 0.72 | 0.52 | 0.4 | 0.78 | 0.42 | 0.71 | 1.02 | 1.02 | 1.11 | 1.25 | 0.57 | 0.45 | 0.55 | 0.22 |  |  |  |  |  |
| 6 | 0.14 | 0.38 | 0.64 | 0.56 | 0.74 | 0.55 | 0.66 | 0.74 | 0.56 | 0.57 | 0.73 | 0.73 | 0.58 | 0.95 | 0.91 | 1.16 | 0.78 | 0.41 | 0.51 | 0.48 | 0.31 |  |  |  |  |  |
| 7 | 0.31 | 0.11 | 0.79 | 0.55 | 0.58 | 0.6 | 0.74 | 0.66 | 0.33 | 0.51 | 0.92 | 0.58 | 0.9 | 0.75 | 0.79 | 1.26 | 1.5 | 0.5 | 0.24 | 0.8 | 0.25 |  |  |  |  |  |
| 8 | 1.48 | 0.39 | 0.18 | 0.52 | 0.61 | 0.41 | 0.63 | 0.91 | 0.47 | 0.43 | 0.88 | 0.67 | 0.56 | 0.72 | 0.46 | 1.16 | 2.25 | 0.63 | 0.09 | 0.43 | 0.28 |  |  |  |  |  |
| 9 | 0.36 | 0.44 | 0.49 | 0.44 | 0.66 | 0.65 | 0.6 | 0.72 | 0.54 | 0.49 | 0.74 | 0.58 | 0.63 | 0.87 | 0.93 | 1.1 | 1.23 | 0.6 | 0.44 | 0.45 | 0.28 |  |  |  |  |  |
| $10+$ | 0.36 | 0.44 | 0.49 | 0.44 | 0.66 | 0.65 | 0.6 | 0.72 | 0.54 | 0.49 | 0.74 | 0.58 | 0.63 | 0.87 | 0.93 | 1.1 | 1.23 | 0.6 | 0.44 | 0.45 | 028 |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| mn4-8 | 0.54 | 0.35 | 0.49 | 0.47 | 0.65 | 0.58 | 0.64 | 0.74 | 0.49 | 0.48 | 0.79 | 0.59 | 0.65 | 0.84 | 0.80 | 1.13 | 139 | 0.55 | 0.34 | 0.53 | 0.28 |  |  |  |  |  |
| Fwb | 0.31 | 0.29 | 0.39 | 0.32 | 0.47 | 0.52 | 0.41 | 0.53 | 0.29 | 0.33 | 0.42 | 0.35 | 0.53 | 0.55 | 0.56 | 0.63 | 0.49 | 0.26 | 0.25 | 0.27 | 0.24 |  |  |  |  |  |

Table A3 continued. Estimates of beginning year stock size (thousands of fish), instantaneous fishing mortality ( F ), mean biomass ( mt ), spawning stock biomass ( mt ), and percent mature of Georges Bank cod, estimated from virtual population analysis (VPA), calibrated using the commercial catch at age ADAPT formulation, 1978-1998.

## Mean blomass (mt)

| Age | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 17758 | 18931 | 15197 | 33078 | 11990 | 8411 | 26102 | 7071 | 35932 | 10768 | 16709 | 11529 | 6964 | 18162 | 7249 | 7400 | 6406 | 3487 | 6056 | 7468 | 1196 |
| 1 | 4814 | 29258 | 22651 | 19777 | 36452 | 15601 | 10450 | 24028 | 8346 | 41193 | 17253 | 26465 | 14114 | 9905 | 17229 | 7088 | 9851 | 8232 | 4596 | 8464 | 8961 |
| 2 | 5116 | 29981 | 21114 | 20010 | 31666 | 12313 | 7020 | 21795 | 8389 | 36592 | 16166 | 24315 | 9402 | 7860 | 11568 | 5519 | 10117 | 9650 | 4838 | 8218 |  |
| 3 | 47056 | 4858 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 4 | 20860 | 42241 | 4890 | 21843 | 16003 | 10992 | 21920 | 8106 | 5193 | 18438 | 5813 | 27830 | 12980 | 14806 | 4877 | 3881 | 5804 | 4328 | 9753 | 8557 | 4309 |
| 5 | 9451 | 16546 | 28839 | 4029 | 17042 | 8355 | 6880 | 13463 | 5247 | 4127 | 10564 | 4303 | 15708 | 6954 | 6980 | 2009 | 1554 | 2832 | 2629 | 5950 | 6958 |
| 6 | 5522 | 8746 | 11415 | 18262 | 2505 | 9174 | 5217 | 3612 | 8108 | 3448 | 2247 | 5029 | 2534 | 6879 | 2770 | 2648 | 928 | 796 | 1628 | 1655 | 4509 |
| 7 | 8287 | 6326 | 4789 | 6592 | 10954 | 1269 | 5569 | 2720 | 2343 | 4827 | 1565 | 1166 | 2332 | 1266 | 2964 | 968 | 751 | 568 | 530 | 863 | 1262 |
| 8 | 276 | 6708 | 4441 | 2351 | 3515 | 5941 | 712 | 2327 | 1541 | 1477 | 2265 | 692 | 698 | 1040 | 740 | 1162 | 199 | 281 | 268 | 323 | 427 |
| 9 | 1318 | 108 | 2905 | 4201 | 1358 | 1746 | 3261 | 337 | 1112 | 897 | 765 | 1019 | 346 | 261 | 436 | 289 | 306 | 35 | 148 | 238 | 270 |
| $10+$ | 549 | 1382 | 315 | 2602 | 2097 | 1453 | 3099 | 1817 | 755 | 737 | 974 | 532 | 881 | 408 | 205 | 205 | 81 | 25 | 6 | 113 | 206 |



SSB at the start of the spawning season - males and females (mt)

|  |  | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1. | 913 | 1104 | 850 | 1960 | 1200 | 903 | 3124 | 773 | 8514 | 2225 | 3481 | 2481 | 638 | 1979 | 806 | 733 | 215 | 115 | 193 | 256 | 25 |
| $\longmapsto$ | 2 | 1410 | 7539 | 6913 | 5782 | 16138 | 6345 | 4303 | 11651 | 5031 | 25335 | 8903 | 13729 | 6634 | 4243 | 9102 | 3646 | 3643 | 3101 | 1709 | 3063 | 3506 |
| a | 3 | 33844 | 3729 | 22417 | 15929 | 15642 | 26059 | 10500 | 6878 | 18778 | 7106 | 32849 | 14544 | 22040 | 9073 | 7482 | 11649 | 5499 | 8966 | 7986 | 4249 | 7472 |
|  | 4 | 20219 | 38256 | 4297 | 21379 | 15792 | 12648 | 21655 | 8075 | 4842 | 17028 | 6135 | 27198 | 12818 | 16523 | 5305 | 4617 | 6665 | 4162 | 9314 | 8468 | 4159 |
|  | 5 | 8798 | 16585 | 30442 | 3958 | 17473 | 9639 | 7110 | 14906 | 5434 | 3937 | 12381 | 4198 | 18072 | 8437 | 8401 | 2548 | 1923 | 2914 | 2837 | 6788 | 6865 |
|  | 6 | 4882 | 8130 | 12541 | 20323 | 2957 | 10520 | 5655 | 4243 | 8583 | 3705 | 2764 | 5942 | 2957 | 8704 | 3357 | 3317 | 1011 | 775 | 1837 | 1840 | 4401 |
|  | 7 | 8215 | 5550 | 5918 | 7296 | 12172 | 1460 | 6226 | 3166 | 2346 | 5363 | 2024 | 1327 | 2850 | 1546 | 3510 | 1306 | 1049 | 584 | 571 | 1047 | 1249 |
|  | 8 | 367 | 6810 | 5034 | 2696 | 4165 | 6840 | 811 | 2985 | 1705 | 1692 | 2931 | 812 | 770 | 1222 | 784 | 1529 | 317 | 299 | 332 | 386 | 481 |
|  | 9 | 1331 | 112 | 3963 | 4097 | 1561 | 2112 | 3955 | 416 | 1250 | 1033 | 956 | 1192 | 408 | 373 | 563 | 418 | 441 | 42 | 171 | 245 | 259 |
|  | $10+$ | 653 | 1681 | 388 | 3168 | 2710 | 1873 | 3940 | 2384 | 945 | 909 | 1284 | 673 | 1128 | 555 | 283 | 295 | 120 | 32 | 7 | 138 | 239 |
|  | Total | 80633 | 89496 | 92763 | 86588 | 89810 | 78398 | 67279 | 55478 | 57428 | 68331 | 73709 | 72095 | 68315 | 52657 | 39594 | 30058 | 20883 | 20991 | 24959 | 26480 | 28656 |

## Percent Mature (females)

|  | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Age |  |  |  |  |  |  |  |  |  |  |  |  | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 |
| 1 | 7 | 7 | 7 | 7 | 13 | 13 | 13 | 13 | 28 | 28 | 28 | 28 | 12 | 12 | 12 | 12 | 4 | 4 | 4 | 4 | 4 |
| 2 | 34 | 34 | 34 | 34 | 47 | 47 | 47 | 47 | 67 | 67 | 67 | 67 | 52 | 52 | 52 | 52 | 44 | 44 | 44 | 44 | 44 |
| 3 | 78 | 78 | 78 | 78 | 84 | 84 | 84 | 84 | 91 | 91 | 91 | 91 | 90 | 90 | 90 | 90 | 93 | 93 | 93 | 93 | 93 |
| 4 | 96 | 96 | 96 | 96 | 97 | 97 | 97 | 97 | 98 | 98 | 98 | 98 | 99 | 99 | 99 | 99 | 100 | 100 | 100 | 100 | 100 |
| $5-10+$ | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |

Table A4. Input data used for long term stochastic projection of stock biomass for Georges Bank cod, 1999-2009.

## Input for Projections:

Number of Years: 11; Initial Year: 1999; Final Year: 2009
Number of Ages : 10; Age at Recruitment: 1; Last Age: 10
Natural Mortality is assumed Constant over time at: . 20
Proportion of F before spawning: . 1667
Proportion of M before spawning: . 1667
Last age is a PLUS group.

| Age | $\text { Recruitment }^{\text {Proportion }} \frac{\text { Prone }}{\text { Mature }}$ |  | Average Weights |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Catch | Stock |
| 1 | 0.00 | 0.04 | 0.845 | 0.624 |
| 2 | 0.15 | 0.44 | 1.499 | 1.164 |
| 3 | 0.60 | 0.93 | 2.264 | 1.848 |
| 4 | 1.00 | 1.00 | 3.581 | 2.839 |
| 5 | 1.00 | 1.00 | 4.809 | 4.128 |
| 6 | 1.00 | 1.00 | 6.678 | 5.687 |
| 7 | 1.00 | 1.00 | 8.666 | 7.651 |
| 8 | 1.00 | 1.00 | 9.333 | 8.874 |
| 9 | 1.00 | 1.00 | 11.504 | 10.404 |
| $10+$ | 1.00 | 1.00 | 14.203 | 14.203 |

Table A5. Projections of landings, spawning stock biomass, mean biomass, and recruitment with probabilities of exceeding a SSB threshold of $50,000 \mathrm{mt}$ and a mean biomass threshold of 108,000 $\mathrm{mt}, 1999-2009$.

|  | F | Landings | SSB | $\begin{gathered} P(\text { SSB }> \\ 50,000 \mathrm{mt}) \end{gathered}$ | Recruits | Mean <br> Biomass | $\begin{gathered} P(\text { Mean biomass > } \\ 108,000 \mathrm{mt}) \\ \hline \end{gathered}$ | $F$ biomass weighted | $\begin{aligned} & P(F \text { biomass wt'd } \\ & \quad>0.125) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR |  |  |  |  |  |  |  |  |  |
| 1999 | 0.28 | 8.287 | 32.461 | 0.00 | 6778.77 | 43.863 | 0.000 | 0.189 | 1.00 |
| 2000 | 0.185 | 6.199 | 35.769 | 0.00 | 7467.61 | 49.952 | 0.000 | 0.124 | 0.47 |
| 2001 | 0.185 | 7.161 | 43.339 | 0.12 | 9107.23 | 56.415 | 0.000 | 0.127 | 0.62 |
| 2002 | 0.185 | 8.769 | 50.136 | 0.51 | 11027.6 | 65.101 | 0.002 | 0.135 | 0.92 |
| 2003 | 0.185 | 9.679 | 56.486 | 0.81 | 13041.4 | 73.39 | 0.025 | 0.131 | 0.76 |
| 2004 | 0.185 | 10.774 | 63.332 | 0.94 | 15124.8 | 84.909 | 0.127 | 0.127 | 0.60 |
| 2005 | 0.185 | 12.386 | 73.314 | 0.99 | 17607.6 | 99.609 | 0.357 | 0.125 | 0.54 |
| 2006 | 0.185 | 14.272 | 85.538 | 1.00 | 20345.8 | 115.523 | 0.612 | 0.125 | 0.52 |
| 2007 | 0.185 | 16.765 | 99.491 | 1.00 | 23350.4 | 134.695 | 0.809 | 0.125 | 0.53 |
| 2008 | 0.185 | 19.751 | $116.973{ }^{\circ}$ | 1.00 | 27014.2 | 157.698 | 0.924 | 0.126 | 0.56 |
| 2009 | 0.185 | 23.042 | 136.347 | 1.00 | 31024.2 | 183.124 | 0.973 | 0.126 | 0.57 |



Figure A1. Total commercial landings of Georges Bank cod (Division $5 Z$ and Subarea 6), 1893-1998.


Figure A2. Standardized stratified mean catch per tow (kg) of Atlantic cod in NEFSC spring and autumn research vessel bottom trawl surveys on Georges Bank, 1963-1998.


Figure A3. Standardized stratified mean number per tow of Atlantic cod in NEFSC spring and autumn research vessel bottom trawl surveys on Georges Bank, 1963-1998.


Figure A4. Trends in total commercial landings and fishing mortality for Georges Bank cod, 1978-1998.


Figure A5. Trends in spawning stock biomass and recruitment for Georges Bank cod, 1978-1998.


Figure A6. Control rule and recent and projected (1999) stock status for Georges Bank Atlantic cod.


Figure A7. Long-term forecasts of mean stock biomass (ages $1+$ ) with $80 \%$ confidence itervals for Georges Bank cod, 1999-2009, and Bmsy and ASPIC estimated maxiimum biomass.

## B. Georges Bank Haddock by R.W. Brown

### 1.0 Background

The Georges Bank haddock stock was last assessed by the United States in 1998 and reviewed by the Transboundary Resource Assessment Committee and SAW/SARC 27. A Canadian assessment through 1998 was completed in the Spring of 1999 for the $5 \mathrm{Zj} \& m$ portion of the stock. The 1998 U.S. assessment estimated fully recruited fishing mortality (ages 4-7) in 1997 to be 0.11 ( $9 \%$ exploitation rate). which was below the $\mathrm{F}_{0.1}$ rebuilding target in place at the time. Fishing mortality was estimated to have remained between 0.10 and 0.15 between 1995 and 1997. Spawning stock biomass was estimated to have increased from 11.900 mt in 1993 to 40.500 mt in 1997. The age structure of the population was continuing to expand and the age $4+$ biomass was estimated to be at it's highest level since 1983. Although the 1994-1996 year classes appeared moderate relative to recruitment observed over the past decade. recent recruitment was far below average levels observed when the stock was in a healthy condition. The 1996 year class. estimated at 13.8 million fish at age 1 , was expected to result in continued increases in SSB through 1999.

### 2.0 1999 Assessment

## 1998 Fishery

U.S. trip limit regulations were significantly more liberal in 1998 compared to the 1997 calendar year. The trip limit from January-August 1998 was 1,000 pounds day ${ }^{-1}$ up to a maximum of 10.000 pounds $\bullet$ trip ${ }^{-1}$. compared to 1000 pounds $\bullet$ trip ${ }^{-1}$ during the same period in 1997. The trip limit from September-December 1998 was 3.000 pounds $^{-d a y}{ }^{-1}$ up to a maximum of 30.000 pounds $\bullet$ day $^{-1}$, compared to 1,000 pounds $\bullet$ day $^{-1}$ up to a maximum of 10.000 pounds $\bullet$ trip ${ }^{-1}$ during the same period in 1997. As a result of both increased haddock abundance and liberalization of trip limit regulations. U.S. commercial landings of Georges Bank haddock in 1998 were 1.841 mt . a $107 \%$ increase compared to 1997 U.S. landings ( 888 mt ; Table B1; Figure B1). U.S. landings included 1.530 mt of landings from western Georges Bank (statistical areas 521. 522, 525, 526) and 311 mt of landings from eastern Georges Bank (statistical areas $561 \& 562$ ). U.S. catch continues to be displaced inshore as a result of Days at Sea regulations.
U.S. landings from western Georges Bank (statistical areas 521, 522, 525,526), were adequately characterized by 23 port samples. The U.S. landings at age for western Georges Bank was estimated for two market categories (scrod and large) and sampling was pooled separately for quarters 1 and 2, and combined for quarters 3 and 4. Sampling of U.S. landings from eastern Georges Bank ( $\mathrm{n}=1$ ) were insufficient to characterize the size and age distribution of this portion of the U.S. catch.. U.S. landings from eastern Georges Bank represented $17 \%$ of the total U.S. landings. and $5.8 \%$ of the total 1998 catch included in the assessment. Length and age samples from the Canadian fishery was used to characterize the size and age composition of U.S. landings - from eastern Georges Bank.
U.S. discards were estimated based on kept to discard ratios reported in the Vessel Trip Report database following procedures used in previous U.S. haddock assessments. Discarding in the U.S. fishery is estimated to have declined from 626 mt in 1997 to 118 mt in 1998. The proportion of the U.S. catch accounted for by discarding has declined from $51.1 \%$ in 1996 to $41.4 \%$ in 1997. and is estimated to be $6.0 \%$ of the U.S. catch in 1998. Trends in the proportion of U.S. catch accounted for by discarding are consistent with the continuing liberalization of U.S. haddock trip limit regulations.

The size and age distribution of U.S. discards was poorly sampled by the U.S. at-sea observer program. U.S. discards represented $2.2 \%$ of the total catch included in the assessment. Available length samples were used to characterize the length distribution of discards. and age information from U.S. commercial landings and U.S. surveys were used to partition the discard at length data into a discard at age estimate.

Canadian catch from the Georges Bank haddock stock consisted of $3,371 \mathrm{mt}$ of landings (Table B1), approximately $86 \%$ of the allocated 1998 quota of $3,900 \mathrm{mt}$ and $63 \%$ of the total haddock landings from the Georges Bank stock. Canadian landings were monitored at dockside and at-sea monitoring by observers resulted in coverage of almost 400 sea days when approximately $10 \%$ of Canadian haddock landings were taken. Comparison of observer samples with port samples did not reveal any persistent patterns which would indicate that discarding or high grading was occurring. The size and age composition of the 1998 Canadian fishery was characterized by port and at sea samples from all principal gears and all seasons.

The combined catch at age was dominated by age 4 (1994 year class) haddock, although there were significant contributions to the catch by ages 3 (1995 year class), 5 (1993 year class), and 6 (1992 year class). Almost two thirds of the total landings by weight from the stock came from fish that were ages 5 or older.

## Fishery Independent Information

Abundance (stratified mean number $\bullet$ tow $^{-1}$ ) and biomass (stratified mean weight $\bullet$ tow ${ }^{-1}$ ) survey indices in the U.S. Spring survey in 1998 and 1999 remained above levels observed from 19861995. but below levels observed in 1996 and 1997 (Table B2; Figure B2). The 1999 U.S. Spring survey catch of age 1 haddock ( 1998 year class) was the highest level observed since 1979 (1978 year class). The abundance and biomass survey indices in the U.S. Autumn survey in 1998 remained above levels observed since 1986 (Table B2; Figure B2). The 1998 U.S. Autumn survey catch of age 0 haddock (1998 year class) was the highest level observed since 1985 (1985 year class). Aggregate abundance survey indices in the Canadian Spring survey were the highest levels observed since the initiation of the survey in 1986 (Figure B2). The majority of this abundance was comprised of large catches of age 1 haddock (1998 year class) yielding a stratified mean number at age estimate that is 3.7 times higher that the next highest age 1 index in the Canadian survey time series.

## Input Data and Analvsis

The present assessment represents a one-year update to the previous U.S. assessment (Brown 2000. NEFSC 1998). The VPA formulation used for the current assessment was identical to the one used in the 1998 U.S. assessment, except for the addition of current year (1999) Spring survey data. Catch at age data for 1998 and U.S. and Canadian survey abundance indices (stratified mean number • tow ${ }^{-1}$ at age) were updated through 1999. Very minor revisions were made to 1997 catch at age to incorporate revisions to the estimate of 1997 Canadian catch at age.

Precision of the 1999 stock sizes and 1998 fishing mortality and SSB estimates was derived from 1000 bootstrap simulations of the 1999 VPA formulation. A retrospective analysis of terminal year estimates of stock sizes, fully recruited fishing mortality and SSB was carried out back to 1992.

### 3.0 Assessment Results

The current assessment continues to consistently estimate the strength of incoming year classes. indicating that the 1992 ( 15.2 million at age 1), 1993 ( 12.2 million), and 1996 ( 14.9 million) were stronger than other year classes since 1988 (Table B3; Figure B3). The 1997 year class ( 8.3 million at age 1) continues to be estimated as being lower than adjacent year classes. Based on the consistent strength of age $0+$ and 1 survey indices, the 1998 year class is estimated to be 61.9 million fish at age 1. If this estimate is reliable, the 1998 year class would be the third largest year class since 1964. although smaller than the 1975 ( 103.3 million at age 1 ) and 1978 ( 84.0 million) year classes (Table B3). Although all three research surveys that have assessed the 1998 year class indicate a strong year class, the size of this year class remains uncertain until addition fishery dependent and independent information is collected and analyzed. The age distribution of the stock continues to show evidence of broadening, although the numerical abundance of age $4+$ fish has stabilized over the last three years.

The 1994-98 estimates of spawning stock biomass (SSB) have been revised slightly downward by the current assessment. and 1998 SSB is estimated to be 38.100 mt (Table B3; Figure B3). Despite this revision, both the 1998 and 1999 assessments indicate that SSB is continuing to increase. Fully recruited fishing mortality (ages 4-7) in 1998 is estimated to be 0.15 , a slight increase from the fishing mortality estimated for 1997 (Table B3; Figure B4).

## VPA Diagnostics

The sums of squares and mean squared residuals from the VPA were within the range of accepted VPAs from the last three U.S. assessments of Georges Bank haddock. The coefficients of variation on estimated age 1-8 stock sizes (range $0.25-0.49$ ) were slightly to significantly lower than were observed in the previous three U.S. assessments. The reduction of the CV on age 1 in the current assessment ( 0.49 vs. $0.60-0.62$ ) is an artifact of the revised assessment formulation which includes the addition of the U.S. Spring survey in the terminal year. Other VPA diagnostics including the range of CV`s on survey q estimates, the number of large standardized residuals and the maximum partial variance estimates are consistent with previous U.S. haddock
assessments. There were no outstanding residual patterns detected during an analysis of standardized residuals. There was a significant shift in the estimate of $q$ for the Canadian and U.S. Spring indices, raising concerns regarding model performance in estimating the size of the 1998 year class.

Accounting for precision in the current assessment, there is a $90 \%$ probability that fully recruited $F$ in 1998 was between 0.125 and 0.167 , and that SSB in 1998 was between 34.700 and 42.200 mt . There was a zero percent probability that SSB in 1998 had exceeded either the target ( 68.000 mt ) or limit threshold ( $53,000 \mathrm{mt}$ ) biomass level.

The retrospective analysis indicates that before 1994, the VPA calibration had a tendency to significantly overestimate fully recruited $\mathrm{F}(4-7, \mathrm{u})$ and slightly underestimate SSB. After 1994. the pattern shifted to where F is accurately estimated and SSB is slightly overestimated in the terminal year of the assessment.

## Evaluation of the Harvest Control Rule

The Amendment 9 harvest control rule for Georges Bank haddock is based on MSY-based reference point proxies (Figure B5). When SSB is greater than $105,000 \mathrm{mt}$, the overfishing limit is $\mathrm{F}_{0.1}$ (currently estimated to be 0.26 ), and the target F is $75 \%$ of the $\mathrm{F}_{\mathrm{MSY}}$ proxy ( 0.20 ). The limit F decreases linearly from 0.26 at $105,000 \mathrm{mt}$ of SSB to zero at $53,000 \mathrm{mt} \mathrm{SSB}$. and the target F decreases linearly from 0.20 at $105,000 \mathrm{mt}$ of SSB to zero at $68,000 \mathrm{mt}$ of SSB.

To evaluate the stock status relative to the harvest control rule in 1999, a 2-year projection was completed assuming $\mathrm{F}_{1999}$ would be equal to $\mathrm{F}_{1998}(0.15)$. Spawning stock biomass, the biomass reference point proxy in the harvest control rule. was projected to be $44,700 \mathrm{mt}$ in 1999 . Projected SSB levels are below the $1 / 2 \mathrm{SSB}_{\mathrm{MSY}}$ threshold level of $68,000 \mathrm{mt}$. and application of the harvest control rule indicates that fishing mortality should be reduced to zero.

### 4.0 Forecasts

Forecasts for Georges Bank haddock are uncertain both due to the uncertainty of the size of the 19.98 year class and historical patterns in recruitment in relation to SSB. Terminal year stock size estimates from the current assessment were assumed and fishing mortality in 1999 was assumed to be equal to fishing mortality in $1998\left(\mathrm{~F}_{1998}=\mathrm{F}_{1999}=0.15\right)$. Fishing mortality for the years 2000 -2009 was set to the level required by the harvest control rule ( $\mathrm{F}_{2000-2008}=0.00$ ). Projections were based on a partial recruitment vector estimated as the geometric mean of the 1995-1998 F's from the final VPA calibration, arithmetic mean of the 1995-1998 stock and catch weights, and pooled median maturity at age estimates for 1995-1997.

## Projection Methodology

Projections for Georges Bank haddock were completed using AGEPRO software (Brodziak and Rago 1996). Fully recruited fishing mortality in 1999 was assumed to be equal to the 1998 estimate (0.15), and fully recruited fishing mortality for the years 2000-2009 were set at 0.00 as
indicated by the harvest control rule. Projections were based on a partial recruitment vector estimated as the geometric mean of the 1995-1998 F's from the final VPA calibration. arithmetic mean of the 1995-1998 stock and catch weights; and pooled median maturity at age estimates for 1995-1997.

The stock recruitment history for Georges Bank haddock indicates two stanzas of recruitment: a period from 1931-1966 where recruitment was consistently higher, and a period from 1966-1998 when average recruitment was significantly lower. The presence of two recruitment stanzas in the stock recruitment data set for Georges Bank highlights the importance of recruitment assumptions as the stock rebuilds to a level where spawning stock biomass approaches the interface between these recruitment stanzas.

Efforts to develop a stock-recruit relationship for the full stock recruit data set (1931-present) have not been successful because of issues with model significance, convergence and residual problems (Overholtz et al. 1999). In addition, parameters from a non-linear Ricker model produce extremely large estimates of management parameters (Overholtz et al. 1999). Overholtz et al. (1986) used non-parametric methods (probability transition matrices) to provide long-term estimates of productivity for this stock. Overholtz et al. (1999) used a Beverton-Holt model to project recruitment, but recruitment was constrained to values within the observed time series. Initial attempts to complete stock projections using a stock-recruit relationship or R/SSB approaches were largely unsuccessful because of poor diagnostics in the estimation of stockrecruitment relationships, or unrealistic increases in stock numbers and biomass observed during initial projection attempts.

Two sets of projections are reported that use differing approaches to resampling empirical recruitment. A projection was completed by resampling 1966-1999 recruitment values for the entire 11-year projection. Spawning stock biomass increased quickly, exceeding the SSB threshold of 105.000 mt in 2002 (Figure B6). The probability that $\mathrm{SSB}_{\text {MSY }}$ exceeds the threshold SSB ( 105.000 mt ) is $29.7 \%$ in 2001, $64.1 \%$ in 2002, and $74.8 \%$ in 2003 (Table B4). SSB gradually stabilized at a level at approximately 133.000 mt by 2009 . This stabilization may indicate that this level of recruitment over the long term is insufficient to support fishery exploitation and allow for the maintenance of stock sizes at these levels. However, at stock levels approaching $\mathrm{SSB}_{\mathrm{MSY}}$ levels, significant improvement in recruitment above the levels assumed during this project may be realized.

A second projection was completed using a 2-stage resampling approach of empirical recruitment values from the 1931-1965 and 1966-1999 time periods. The more recent recruitment data (19661999) were resampled during the during the first five years of the projection, and the entire time series of recruitment data (1931-1999) was resampled during the remaining 6 years of the projection. Use of the entire time series to model recruitment for the "rebuilt" stock was consistent with the surplus production modeling methods used to develop the control rule. This strategy was employed to simulate a population shifting from the recent recruitment regime, which has been dominated by poor recruitment, to a regime where strong recruitment events occur
with greater frequency. In this projection, spawning stock biomass increased above 105.000 mt by 2002. but began to increase sharply in 2007 as recruitment resampled from the earlier regime began to mature (Table B4; Figure B6). Results from this projection exceed the maximum stock sizes observed for this stock in the 68 year assessment time series. Many of the population compensatory features that affect stocks at high levels of abundance (reduced growth. later maturation, and reduced recruitment) were not accounted for in this modeling effort.

It is important to note that if recruitment improves to the pre-1966 regime. there will be substantial opportunities for fishery harvest, while maintaining stock sizes at levels at or exceeding $\mathrm{B}_{\mathrm{MSY}}$. To achieve recruitments at these levels, the stock must be rebuilt so that it is fully utilizing the two major spawning areas on Georges Bank: the Northeast Peak and the Great South Channel. Currently, the stock appears to be rebuilding in both areas; however. the majority of biomass still remains in the eastern portion of the stock area. Continued rebuilding of haddock stocks in the western portion of the stock area, which is primarily influenced by U.S. management initiatives, is essential to achieve improved recruitment observed from the 1931-1966 regime on a consistent basis.

Each of these projections exhibited an initial sharp increase in SSB between 1999 and 2002. highlighting the importance of incoming recruitment estimated by the VPA (particularly the 1998 year class). These projections demonstrate that SSB will increase to the SSB threshold in 2002. largely as a result of stock size (initial stock size estimates from the VPA) inputs into the projection rather than recruitment assumed within the projection. Indications from research veessel surveys are that the 1998 year class is strong, but there is still considerable uncertainty about it's relative size. If the 1998 year class is either significantly overestimated or underestimated, this will have a significant effect on the accuracy of these stock forecasts.

### 5.0 Sources of Uncertainty

1 Sampling of U.S. landings and discards was insufficient to accurately characterize the size and age distribution of the catch.

2 There is considerable uncertainty regarding the size of the 1998 year class, estimated as age 1 recruits in 1999. Shifts in the estimates of survey q's of the Canadian and U.S. Spring age 1 indices between the 1998 and 1999 stock assessments represent symptoms of a significant change in the weighting of these indices in response to survey indices for this year class.

3 U.S. survey biomass indices do not exhibit the same increases as has been observed in the VPA estimates of biomass and SSB.

4 The VPA assessment continues to exhibit a slight to moderate retrospective pattern in the terminal year estimates of SSB.

## 6.0 Conclusions

The Georges Bank haddock stock continues to rebuild toward historic levels. Continued low exploitation rates have promoted broadening of the age structure and the age $4+$ biomass is at it's highest levels since the early 1980s. Spawning stock biomass has increased almost four fold since 1992 and is expect to exceed $40,000 \mathrm{mt}$ by 1999.

Observed increases in spawning stock biomass of Georges Bank haddock have resulted from conservation of a series of relatively weak year classes. This is a necessary first step in the stock rebuilding process. Stock rebuilding to levels approaching the SSB threshold of 105.000 mt will require recruitment significantly above levels observed during the past two decades. This assessment provides initial indications that the 1998 year class may be significantly larger than year classes observed in the 1980s and 1990s. However, there is considerable uncertainty about the size of the 1998 year class based on diagnostics from the VPA analysis. Projection results over the next five years are highly dependent on the estimate of this year class. If the current estimate of the 1998 year class is reasonably accurate, recruitment of this year class should result in significant stock rebuilding toward the SSB threshold.

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Table B1. Commercial landings (metric tons. live) of haddock from Georges Bank and South (NAFO Division 5Z and Statistical Area 6), 1960-1998. ${ }^{1}$

| Year | USA | Canada | USSR | Spain | Other | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 40800 | 77 | 0 | 0 | 0 | 40877 |
| 1961 | 46384 | 266 | 0 | 0 | 0 | 46650 |
| 1962 | 49409 | 3461 | 1134 | 0 | 0 | 54004 |
| 1963 | 44150 | 8379 | 2317 | 0 | 0 | 54846 |
| 1964 | 46512 | 1162.5 | 5483 | 2 | 464 | 64086 |
| 1965 | 52823 | 14889 | 81882 | $10^{\circ}$ | 758 | 150362 |
| 1966 | 52918 | 18292 | 48409 | 1111 | 544 | 121274 |
| 1967 | 34728 | 13040 | 2316 | 1355 | . 30 | 51469 |
| 1968 | 25469 | 9323 | 1397 | 3014 | 1720 | 40923 |
| 1969 | 16456 | 3990 | 65 | 1201 | 540 | 22252 |
| 1970 | 8415 | 1978 | 103 | 782 | 22 | 11300 |
| 1971 | 7306 | 1630 | 374 | 1310 | 242 | 10862 |
| 1972 | 3869 | 609 | 137 | 1098 | 20 | 5733 |
| 1973 | 2777 | 1563 | 602 | 386 | 3 | 5331 |
| 1974 | 2396 | 462 | 109 | 764 | 559 | 4290 |
| 1975 | 3989 | 1358 | 8 | 61 | 4 | 5420 |
| 1976 | 2904 | 1361 | 4 | 46 | 9 | 4324 |
| 1977 | 7934 | 2909 | 0 | 0 | 0 | 10843 |
| 1978 | 12160 | 10179 | 0 | 0 | 0 | 22339 |
| 1979 | 14279 | 5182 | 0 | 0 | 0 | 19461 |
| 1980 | 17470 | 10017 | 0 | 0 | 0 | 27487 |
| 1981 | 19176 | 5658 | 0 | 0 | 0 | 24834 |
| 1982 | 12625 | 4872 | 0 | 0 | 0 | 17497 |
| 1983 | 8682 | 3208 | 0 | 0 | 0 | 11890 |
| 1984 | 8807 | 1463 | 0 | 0 | 0 | 10270 |
| 1985 | 4273 | 3484 | 0 | 0 | 0 | 7757 |
| 1986 | 3339 | 3415 | 0 | 0 | 0 | 6754 |
| 1987 | 2156 | 4703 | 0 | 0 | 0 | 6859 |
| 1988 | 2492 | $4046{ }^{2}$ | 0 | 0 | 0 | 6538 |
| 1989 | 1430 | 3059 | 0 | 0 | 0 | 4489 |
| 1990 | 2001 | 3340 | 0 | 0 | 0 | 5341 |
| 1991 | 1395 | - 5446 | 0 | 0 | 0 | 6841 |
| 1992 | 2005 | 4061 | 0 | 0 | 0 | 6066 |
| 1993 | 687 | 3727 | 0 | 0 | 0 | 4414 |
| 1994 | 218 | 2411 | 0 | 0 | 0 | 2629 |
| 1995 | 218 | 2064 | 0 | 0 | 0 | 2282 |
| 1996 | 313 | 3643 | 0 | 0 | 0 | 3956 |
| 1997 | 888 | 2622 | 0 | 0 | 0 | 3510 |
| 1998 | 1841 | 3371 | - | 0 | 0 | 5212 |

'All landings 1960-1979 are from Clark et al. (1982): USA landings 1980-1981 are from Overholtz et al. (1983); USA landings 1982-1993 are from NMFS, NEFC Detailed Weighout Files and Canvass data; Canadian landings 1980-1994 from Gavaris and Van Eeckhaute (1996); Canadian landings 1995-1996 from S. Gavaris (Personal Communication). ${ }^{2} 1895$ tons were excluded because of suspected misreporting (Gavaris and Van Eeckhaute 1995).

Table B?. Stratified mean number and mean weight (kg) • tow- of haddock caught in NEFSC Spring and Autumn bottom trawl surveys from 1963-1999. .


Table B3. Beginning year stock size, spawning stock biomass, mean biomass, and fishing mortality of Georges Bank haddock estimated from VPA calibration.

STOCK NUMBERS (Jan 1) in thousands - D: \What 1.6

|  | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 190706 | 471885 | 33154 | 4137 | 12954 | 422 | 988 | 4661 | 369 | 8517 | 19418 | 10547 | 7661 | 103303 |
| 2 | 32266 | 153504 | 377207 | 18457 | 3284 | 9565 | 338 | 807 | 3774 | 301 | 6832 | 13582 | 8593 | 6098 |
| 3 | 32743 | 22756 | 111260 | 194986 | 8920 | 2536 | 5122 | 267 | 51.8 | 1846 | 245 | 3716 | 7211 | 6100 |
| 4 | 45821 | 20096 | 14510 | 50830 | 68425 | 4687 | 1435 | 2657 | 204 | 222 | 1104 | 198 | 2448 | 4217 |
| 5 | 29031 | 27424 | 12131 | 7034. | 24273 | 37321 | 2099 | 770 | 1660 | 131 | 109 | 555 | 160 | 1665 |
| 6 | 9186 | 16351 | 14561 | 5959 | 3254 | 10519 | 17419 | 1127 | 462 | 1097 | 78 | 41 | 391 | 127 |
| 7 | 5595 | 5526 | 8144 | 5868 | 2535 | 1570 | 5446 | 8874 | 729 | 156 | 790 | 37 | 32 | 282 |
| 8 | 2795 | 3309 | 2640 | 3255 | 2694 | 1177 | 682 | 3035 | 5177 | 339 | 57 | 577 | 28 | 22 |
| 9 | 4217 | 4251 | 3258 | 2201 | 2031 | 2163 | 1712 | 1875 | 3245 | 6311 | 1679 | 2702 | 622 | 623 |
| $1+$ | 352360 | 725101 | 576866 | 292727 | 128369 | 69960 | 35241 | 24071 | 16137 | 18919 | 30311 | 31953 | 271.45 | 122438 |
|  | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| 1 | 13809 | 6072 | 83981 | 10135 | 7224 | 2479 | 3106 | 17258 | 1759 | 14733 | 2099 | 16708 | - 1085. | 2640 |
| 2 | 84447 | 11305 | 4971 | 68757 | 8291 | 5914 | 2028 | 2543 | 14129 | 1440 | 12057 | 1718 | 13676 | 888 |
| 3 | 4565 | 51418 | 8567 | 4046 | 28244 | 5211 | 3788 | 1467 | 1998 | 9391 | 1130 | 8066 | 1360 | 10054 |
| 4 | 4497 | 3568 | 29072 | 5452 | 2999 | 13172 | 2789 | 2365 | 932 | 1138 | 5146 | 809 | 4447 | 1036 |
| 5 | 2657 | 3066 | 2645 | 17316 | 3582 | 1703 | 7407 | 1659 | 1279 | 588 | 730 | 2754 | 541 | 2847 |
| 6 | 1168 | 1709 | 1997 | 1691 | 8699 | 2085 | 1041 | 4038 | 999 | 630 | 349 | 487 | 1412 | 313 |
| 7 | 104 | 633 | 931 | 1264 | 847 | 4795 | 1192 | 606 | 1965 | 612 | 359 | 220 | 264 | 832 |
| 8 | 210 | 82 | 392 | 478 | 541 | 394 | 2913 | 808 | 284 | 1133 | 365 | 214 | 130 | 175 |
| 9 | 594 | 390 | 187 | 251 | 319 | 406 | 275 | 1627 | 550 | 254 | 461 | 351 | 208 | 166 |
| $1+$ | 112050 | 78243. | 132744 | 109390 | 60746 | 36158 | 24538 | 32371 | 23896 | 29919 | 22697 | 31327 | 23123 | 18952 |
|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |  |  |  |  |
| 1 | 2389 | 9083 | 15161 | 12175 | 8941 | 8236 | 14897 | 8328 | 61925 |  |  |  |  |  |
| 2 | 2160 | 1951 | 7430 | 12407 | 9967 | 7312 | 6738 | 12170 | 6817 |  |  |  |  |  |
| 3 | 717 | 1363 | 1374 | 5821 | 9914 | 8079 | 5938 | 5359 | 9784 |  |  |  |  | - |
| 4 | 6924 | 505 | 826 | 808 | 4032 | 7578 | 6099 | 4603 | 4012 |  |  |  |  |  |
| 5 | 692 | 3724 | 294 | 406 | 507 | 2888 | 5348 | 4310 | 3315 |  | $\cdot$ |  |  |  |
| 6 | 1546 | 474 | 1668 | 147 | 273 | 361 | 1945 | 3884 | 2903 |  |  |  |  |  |
| 7 | 168 | 895 | 288 | 769 | 57 | 195 | 234 | 1400 | 2704 |  |  |  |  |  |
| 8 | 521 | 71 | 440 | 201 | 493 | 40 | 140 | 175 | 1012 |  | . |  |  |  |
| 9 | 242 | 246 | 208 | 197 | 156 | 58 | 349 | 338 | 363 |  |  |  |  |  |
| 1+ | 15359 | 18312 | 27689 | 32930 | 34341 | 34745 | 41689 | 40567 | 92837 |  |  |  |  |  |

Table B3 (Continued). Beginning year stock size, spawning stock biomass, mean biomass, and fishing mortality of Georges Bank haddock estimated from VPA calibration.

SSB AT THE START OF THE SPANNING SEASON -MALES AND FEMALES (MT) (using SSB mean weights)

|  | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 00 | 00 | 00 | 00 | 00 | 00 | 00 | 00 | 00 | 00 | 00 | 00 | 00 | 00 |
| 2 | 00 | 00 | 00 | - 00 | 00 | 1675 | 61 | 164 | 756 | 67 | 1594 | 3142 | 2253 | 1510 |
| 3 | 24233 | 15655 | 65996 | 91773 | 4934 | 1433 | 3119 | 185 | 411 | 1652 | 273 | 4216 | 7623 | 6068 |
| 4 | 56101 | 23010 | 14892 | 48128 | 60273 | 4294 | 1636 | 3442 | 266 | 304 | 1789 | 359 | 4459 | 6766 |
| 5 | 38629 | 36355 | 15691 | 8788. | 26351 | 41983 | 2731 | 1303 | 3215 | 236 | 189 | 1248 | 342 | 36.94 |
| 6 | 16464 | 25247 | 20964 | 8946 | 5063 | 15410 | 26018 | 2067 | 873 | 2671 | 183 | 116 | 1039 | 316 |
| 7 | 10877 | 10439 | 13799 | 10289 | 4575 | 2780 | 10823 | 17573 | 1590 | 354 | 2.308 | 126 | 113 | 863 |
| 8 | 6533 | 7059 | 5446 | 6850 | 5610 | 2397 | 1526 | 7608 | 12676 | 962 | 170 | 1956 | 105 | 87 |
| 9 | 11435 | 10811 | 8271 | 5784 | 5324 | 5124 | 5278 | 6177 | 10450 | 20678 | 5770 | 10659 | 2455 | 2771 |
| $1+$ | 164272 | 128575 | 145060 | 180558 | 112131 | 75096 | 51190 | 38520 | 30236 | 26923 | 12275 | 2.1821 | 18388 | 22076 |
|  | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| 1 | 00 | 00 | 00 | 00 | 00 | 00 | 00 | 376 | 79 | 1113 | 137 | 1102 | 97 | 117 |
| 2 | 17995 | 2457 | 1134 | 12824 | 1686 | 1074 | 293 | 435 | 4743 | 491 | 4331 | 684 | 502.9 | 338 |
| 3 | 4151 | 45758 | 6799 | 3345 | 20418 | 4054 | 3145 | 1428 | 1786 | 8529 | 1096 | 6673 | 1392 | 9883 |
| 4 | 7098 | 5675 | 44465 | 7304 | 3874 | 17418 | 3990 | 3211 | 1317 | 1592 | 71.13 | 1111 | 6039 | 1504 |
| 5 | 5546 | 6779 | 5353 | 30526 | 6241 | 3137 | 12971 | 2931 | 2273 | 1162 | 1339 | 4731 | 909 - | 4759 |
| 6 | 2927 | 4333 | 5274 | 3784 | 18207 | 4568 | 2278 | 8193 | 2161 | 1445 | 807 | 982 | 2847 | 637 |
| 7 | 351 | 1847 | 2736 | 3438 | 2262 | 12569 | 3239 | 1405 | 4895 | 1525 | 922 | 539 | 671 | 1988 |
| 8 | 725 | 286 | 1233 | 1494 | 1781 | 1315 | 8352 | 2283 | 820 | 3254 | 10.53 | 597 | $376{ }^{\circ}$ | 474 |
| 9 | 2664 | 1797 | 799 | 827 | 1223 | 1445 | 1019 | 5140 | 1816 | 847 | 1708 | 1195 | 706 | 602 |
| $1+$ | 41457 | 68933 | 67794 | 63542 | 55692 | 45579 | 35286 | 25402 | 19890 | 19957 | 18506 | 17613 | 18067 | 20303 |
|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |  |  |  |  |  |  |
| 1 | 93 | 315 | 517 | 246 | 46 | 48 | 77 | 65 |  |  |  |  |  |  |
| 2 | 945 | 827 | 1663 | 2988 | 2112 | 1619 | 1456 | 2488 |  |  |  |  |  |  |
| 3 | 697 | 1586 | 1217 | 5226 | 11119 | 8516 | 7144 | 5693 |  | . |  |  |  |  |
| 4 | 9627 | 665 | 1227 | 1321 | 6766 | 11488 | 9417 | 7642 |  |  | . |  |  |  |
| 5 | 1241 | 6072 | 478 | 878 | 1104 | 5677 | 9963 | 8212 |  |  | - |  |  |  |
| 6 | 3190 | 982 | 3294 | 279 | 683 | 824 | 4360 | 8656 |  |  |  |  |  |  |
| 7 | 366 | 2099 | 680 | 1901 | 156 | 557 | 611 | 3587 |  |  |  |  |  |  |
| 8 | 1313 | 184 | 1122 | 567 | 1358 | 122 | 438 | 492 |  |  |  |  |  |  |
| 9 | 903 | 873 | 700 | 796 | 572 | 195 | 1172 | 1262 |  |  |  |  |  |  |
| $1+$ | 18376 | 13602 | 10898 | 14201 | 23916 | 29046 | 34639 | 38096 |  |  |  |  |  |  |

Table B3 (Continued). Beginning year stock size, spawning stock biomass, mean biomass, and fishing mortality of Georges Bank haddock estimated from VPA calibration.

MEAN BIOMASS (using catch mean weights at age)

|  | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 97717 | 211390 | 14554 | 2142 | 7375 | 223 | 465 | 2983 | 224 | 4739 | 9796 | 6866 | 4247 | 46779 |
| 2 | 23694 | 108938 | 190543 | 9574 | 2026 | 5762 | 235 | 828 | 2777 | 280 | 5270 | 10666 | 7130 | 5242 |
| 3 | 30570 | 20524 | 79454 | 107563 | 6249 | 2031 | 4131 | 286 | 458 | 2511 | 348 | 5529 | 9097 | 7308 |
| 4 | 52683 | 22565 | 13938 | 45273 | 60547 | 4252 | 1805 | 4091 | 287 | 324 | 1700 | 414 | 4489 | 6720 |
| 5 | 37107 | 33322 | 14500 | 8336 | 2.3357 | 41005 | 2735 | 1320 | 3249 | 248 | 167 | 1324 | 315 | 3728 |
| 6 | 15484 | 23667 | 19034 | 8292 | 4737 | 16183 | 25212 | 2183 | 791 | 2731 | 181 | 136 | 980 | 356 |
| 7 | 10228 | 9378 | 12048 | 9296 | 4088 | 2471 | 10392 | 17699 | 1487 | 300 | 2318 | 132 | 107 | 901 |
| 8 | 6577 | 6509 | 4809 | 6534 | 5219 | 2165 | 1497 | 7654 | 12198 | 942 | 169 | 1948 | 96 | 85 |
| 9 | 10122 | 9407 | 6938 | 4896 | 4586 | 4345 | 4563 | 5489 | 9174 | 18665 | 5097 | 9910 | 2.52 | 2513 |
|  |  |  |  |  |  |  |  | . |  |  |  | . |  |  |
| $1+$ | 284182 | 445699 | 355818 | 201907 | 118183 | 78438 | 51035 | 42532 | 30643 | 30739 | 25047 | 36924 | 28713 | 73632 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| 1 | 6633 | 2917 | 40341 | 5050 | 2553 | 494 | 929 | 5162 | 526 | 6008 | 818 | 6359 | 521 | 1531 |
| 2 | 71233 | 9280 | 4493 | 42803 | 5770 | 4629 | 1767 | 2079 | 11483 | 1202 | 8241 | 1502 | 10477 | $776$ |
| 3 | 5826 | 58784 | 8823 | 4231 | 24501 | 5618 | 4139 | 1.557 | 2124 | 9598 | 1374 | 8145 | 1762 | 12420 |
| 4 | 8107 | 6291 | 45831 | - 7702 | 4191 | 18827 | 3981 | 3233 | 1479 | 1682 | 7652 | 1119 | 6422 | 1516 |
| 5 | 5864 | 6958 | 5353 | 27162 | 6362 | 3187 | 12273 | 2863 | 2254 | 1173 | 1351 | 4138 | 921 | 4511 |
| 6 | 2803 | 4087 | 5034 | 3442 | 17828 | 4408 | 2130 | 7686 | 2147 | 1364 | 735 | 893 | 2819 | 595 |
| 7 | 384 | 1695 | 2568 | 3068 | 2195 | 12235 | 3209 | 1258 | 4624 | 1414 | 846 | 508 | 701 | 1865 |
| 8 | 683 | 242 | 1162 | 1257 | 1652 | 1198 | 7489 | 2021 | 810 | 3085 | 1041 | 560 | 370 | 403 |
| 9 | 2411 | 1633 | 711 | 717 | 1071 | 1274 | $897$ | 4442 | 1599 | 755 | 1496 | 1044 | 634 | 532 |
| $1+$ | 103944 | 91887 | 114315 | 95432 | 66124 | 51871 | 36814 | 30300 | 27047 | 26280 | 23554 | 24267 | 2.4627 | 24148 |
|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |  |  |  |  |  |  |
| 1 | 1256 | 4427 | 9053 | 4932 | 3474 | 3403 | 5611 | 3857 |  |  |  |  |  |  |
| 2 | 2079 | 1933 | 7707 | 12148 | 8693 | 7248 | 5986 | 10584 |  |  |  | - |  |  |
| 3 | 793 | 1758 | 1842 | 8005 | 12945 | 10543 | 8881 | 6910 |  |  |  |  |  |  |
| 4 | 9375 | 690 | 1284 | 1427 | 6942 | 11760 | 9765 | 7523 |  |  |  |  |  |  |
| 5 | 1259 | 5595 | 449 | 881 | 1093 | 5546 | 10138 | 8307 |  |  |  |  |  |  |
| 6 | 3149 | 941 | 3051 | 259 | 653 | 745 | 4225 | 8758 |  |  |  |  |  |  |
| 7. | 322 | 1901 | 642 | 1802 | 157 | 568 | 630 | 3620 |  |  |  |  |  |  |
| 8 | 1197 | 170 | 985 | 607 | 1295 | 117 | 402 | $449$ |  |  |  |  |  |  |
| 9 | 790 | 742 | 599 | 714 | 528 | 179 | 1084 | - 1164 |  |  |  |  |  |  |
| $1+$ | 20221 | 18157 | 25610 | 30774 | 35781 | 40108 | 46721 | 51170 |  |  |  |  |  |  |

Table B3 (Continued).
Beginning year stock size, spawning stock biomass, mean biomass, and fishing mortality of Georges Bank haddock estimated from V.PA calibration.

## FISHING MORTALITY - D: $\backslash$ What 1.6.

|  | 1963 | 1964 | 1965 | 1966 | 1967 | 1968 | 1969 | 1970 | 1971 | 1972 | 1973 | 1974 | 1975 | 1976 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.02 | 0.02 | 0.39 | 0.03 | 0.10 | 0.02 | 0.00 | 0.01 | 0.00 | 0.02 | 0.16 | 0.00 | 0.03 | 0.00 |
| 2 | 0.15 | 0.12 | 0.46 | $0.53{ }^{\circ}$ | 0.06 | 0.42 | 0.04 | 0.24 | 0.52 | 0.01 | 0.41 | 0.43 | 0.14 | 0.09 |
| 3 | 0.29 | 0.25 | 0.58 | 0.85 | 0.44 | 0.37 | 0.46 | 0.07 | 0.65 | 0.31 | 0.01 | 0.22 | 0.34 | 0.10 |
| 4 | 0.31 | 0.30 | 0.52 | 0.54 | 0.41 | 0.60 | 0.42 | 0.27 | 0.24 | 0.52 | 0.49 | 0.01 | 0.19 | 0.26 |
| 5 | 0.37 | 0.43 | 0.51 | 0.57 | 0.64 | 0.56 | 0.42 | 0.31 | 0.21 | 0.31 | 0.77 | 0.15 | 0.03 | 0.15 |
| 6 | 0.31 | 0.50 | 0.71 | 0.65 | 0.53 | 0.46 | 0.47 | 0.24 | 0.89 | 0.13 | 0.55 | 0.06 | 0.13 | 0.00 |
| 7 | 0.33 | 0.54 | 0.72 | 0.58 | 0.57 | 0.63 | 0.38 | 0.34 | 0.57 | 0.81 | 0.11 | 0.06 | 0.15 | 0.09 |
| 8 | 0.34 | 0.42 | 0.61 | 0.56 | 0.47 | 0.55 | 0.45 | 0.32 | 0.38 | 0.24 | 0.35 | 0.11 | 0.17 | 0.22 |
| 9 | 0.34 | 0.42 | 0.61 | 0.56 | 0.47 | 0.55 | 0.45 | 0.32 | 0.38 | 0.24 | 0.35. | 0.11 | 0.17 | 0.22 |
|  | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | $0.00{ }^{\text {- }}$ | 0.00 | 0.00 | 0.00 |
| 2 | 0.30 | 0.08 | 0.01 | 0.69 | 0.26 | 0.25 | 0.12 | 0.04 | 0.21 | 0.04 | 0.20 | 0.03 | 0.11 | 0.01 |
| 3 | 0.05 | 0.37 | 0.25 | 0.10 | 0.56 | 0.43 | 0.27 | 0.25 | 0.36 | 0.40 | 0.13 | 0.40 | 0.07 | 0.17 |
| 4 | 0.18 | 0.10 | 0.32 | 0.22 | 0.37 | 0.38 | 0.32 | 0.41 | 0.26 | 0.24 | 0.43 | 0.20 | 0.25 | 0.20 |
| 5 | 0.24 | 0.23 | 0.25 | 0.49 | 0.34 | 0.29 | 0.41 | 0.31 | 0.51 | 0.32 | 0.20 | 0.47 | 0.35 | 0.41 |
| 6 | 0.41 | 0.41 | 0.26 | 0.49 | 0.40 | 0.36 | 0.34 | 0.52 | 0.29 | 0.36 | 0.26 | 0.41 | 0.33 | 0.42 |
| 7 | 0.04 | 0.28 | -0.47 | 0.65 | 0.57 | 0.30 | 0.19 | 0.56 | 0.35 | 0.32 | 0.32 | 0.32 | 0.21 | 0.27 |
| 8 | 0.23 | 0.21 | 0.32 | 0.44 | 0.39 | 0.35 | 0.36 | 0.45 | 0.36 | 0.30 | 0.39 | 0.40 | 0.27 | 0.34 |
| 9 | 0.23 | 0.21 | 0.32 | 0.44 | 0.39 | 0.35 | 0.36 | 0.45 | 0.36 | 0.30 | 0.39 | 0.40 | 0.27 | 0.34 |
|  | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |  |  |  |  |  |  |
| 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  |  |  |  |  |  |
| 2 | 0.26 | 0.15 | 0.04 | 0.02 | 0.01 | 0.01 | 0.03 | 0.02 |  |  |  |  |  |  |
| 3 | 0.15 | 0.30 | 0.33 | 0.17 | 0.07 | 0.08 | 0.05 | 0.09 |  |  |  |  |  |  |
| 4 | 0.42 | 0.34 | 0.51 | 0.27 | 0.13 | 0.15 | 0.15 | 0.13 |  |  |  |  |  |  |
| 5 | 0.18 | 0.60 | 0.50 | 0.20 | 0.14 | 0.20 | 0.12 | 0.20 |  |  |  |  |  |  |
| 6 | 0.35 | 0.30 | 0.57 | 0.74 | 0.14 | 0.23 | 0.13 | 0.16 |  |  |  |  |  | . |
| 7 | 0.65 | 0.51 | 0.16 | 0.24 | 0.17 | 0.13 | 0.09 | 0.12 |  |  |  |  |  |  |
| 8 | 0.40 | 0.54 | 0.51 | 0.27 | 0.14 | 0.16 | 0.13 | 0.14 |  |  |  |  |  |  |
| 9 | 0.40 | 0.54 | 0.51 | 0.27 | 0.14 | 0.16 | 0.13 | 0.14 |  |  |  |  |  |  |

Table B4. Medium term projection results for Georges Bank haddock. Model 1 projected recruitment by resampling empirical age 1 recruitment estimated by the VPA from 1966-1999. Model 2 projected recruitment by resampling empirical age 1 recruitment estimated by the VPA from 1966-1999 for the first five years of the projection and from 1931-1999 for the final 6 years of the projection.

|  | Model 1: Recent (1966-1999) Recruitment Assumed |  | Model 2: 1999-2003: 1966-1999 Recruitment 2004-2009: 1931-1999 Recruitment |  |
| :---: | :---: | :---: | :---: | :---: |
|  | SSB (mt) | Probability of Exceeding SSB Threshold | SSB (mt) | Probability of Exceeding SSB Threshold |
| 1999 | 44.700 | 0.0\% | 44.700 | 0 |
| 2000 | 57.600 | 0.0\% | 57,600 | 0 |
| 2001 | 96,500 | 29.7\% | 96,500 | 29.9\% |
| 2002 | 111.500 | 64.1\% | 111.700 | 64.9\% |
| 2003 | 118,800 | 74.8\% | 119.400 | 75.3\% |
| 2004 | 122.600 | 77.8\% | 123,100 | 76.4\% |
| 2005 | 126.300 | 79.6\% | 126.700 | 79.2\% |
| 2006 | 127.200 | 78.0\% | 136,300 | 84.0\% |
| 2007 | 133,300 | 80.1\% | 171.800 | 92.0\% |
| 2008 | 132.400 | 74.7\% | 199.700 | 94.6\% |
| 2009 | 133.300 | 70.7\% | 227.200 | 95.8\% |



Figure B1. Total commercial landings of haddock from Georges Bank and South, 1904-1998.


Figure B2. NEFSC and Canadian DFO bottom trawl survey abundance (number per tow; Panel A) and biomass (kg per tow; Panel B) for Georges Bank haddock, 1963-1999. Surveys have not been adjusted for catchabilities.


Figure B3. Trends in spawning stock biomass (line) and age 1 recruitment (bars) for Georges Bank (5Z \& 6) haddock, 1931-1998.


Figure B4. Trends in commercial landings (metric tons, live weight) and fully-recruited fishing mortality (mean F, ages 4-7, unweighted) for Georges Bank haddock, 1931-1998.


Figure B5. Harvest control rule for Georges Bank haddock based on proxies of MSY-based reference points and minimum biomass thresholds.


Figure B6. Results of medium term projections of Georges Bank haddock, assuming a fully recruited F1999 = 0.15 and F2000-2008 $=0.00$. Recruitment was modeled using two stage resampling of R/SSB ratios, two stage resampling of empirical recruitment estimates, and resampling of recent (1966-1999) recruitment

## C. Georges Bank Yellowtail Flounder by S.X. Cadrin

### 1.0 Background

In 1997. the Georges Bank yellowtail flounder stock was at a low biomass level ( $55 \%$ of $\mathrm{B}_{\mathrm{MSY}}$ ) and $F$ was low (fully recruited $F$ was 0.13 , Neilson and Cadrin 1998). In 1998, the stock was at $75 \%$ of $\mathrm{B}_{\mathrm{MS} Y}$ with low F (fully recruited F was 0.21 , Neilson et al. 1999). This report updates catch and survey indices, updates estimates of stock size and fishing mortality estimates. and evaluates medium-term projections.

### 2.0 1999 Assessment

2.11998 Landings
U.S. landings were prorated as described in Cadrin et al. (1999; Table C1; Figure C1). Landings from Georges Bank yellowtail (including Canadian landings) increased 70\% from 1997.

Sampling intensity of landings in 1998 was poor. There were no commercial samples of yellowtail flounder in the fourth quarter from Georges Bank, which comprised nearly half of the landings. Landings at length were estimated by half year and market category. Canadian landings at age were estimated from Canadian port sample lengths and NEFSC fall survey ages (Neilson et al. 1999; Table C2a): U.S. landings at age and mean weights at age are reported in . Table C2b.

### 2.2 1998 Discards

Discarded catch was estimated from logbook information on discard to kept ratios by half-year and gear (NEFSC 1998). Discard ratios are similar to those estimated in recent years for Georges Bank yellowtail. which were also based on logbook information. Discards at age were estimated from sea sampled lengths and pooled commercial-survey age-length keys. Discards at age and recent mean weights at age are reported in Table C3.

### 2.3 1998-1999 Survev Indices

Survey abundance and biomass indices are reported in Table C4. Estimates are from valid tows in Georges Bank (offshore strata 13-21; scallop strata 54, 55, 58-72, 74), standardized according to net. vessel. and door changes (NEFSC 1998). All survey indices of total abundance and total biomass increased for Georges Bank yellowtail in 1998 and 1999 (Figure C2).

### 3.0 Assessment Results

### 3.1 Age-Based Analysis

An updated VPA calibration for Georges Bank yellowtail is summarized in Table C5. This analysis is slightly different than that reported by Neilson et al. (1999) in that 1) US total landings in 1998 were revised with recently processed logbook information, 2) 1998 landings at age included some observer lengths to characterize unclassified landings, 3) total discards were estimated from 1998 logbooks and characterized to age with observer lengths and survey ages. rather than assuming 1997 discard at age ratios, and 4) the catch at age was expanded from ages $1-5$ to ages $1-6+.5)$ NEFSC spring indices were included in the calibration. Results indicate that F remained low ( $\mathrm{F}_{4.5}=0.17$ ) and biomass continued to rebuild in $1998(17.000 \mathrm{mt}$ of spawning biomass and $28,000 \mathrm{mt}$ of mean total biomass; Figures 3 and 4). Estimates of $F$ and biomass were similar to those reported by Neilson et al. (1999), but abundance at ages 2 and 3 were substantially greater in the present analysis (the increase appears to be a result of the high spring survey indices). Retrospective analysis indicates a tendency toward underestimating $F$ in the most recent years, but the pattern does not persist further back in time. Bootstrap analysis indicates that abundance was estimated with with moderate to high precision (CV=14-34\%).

### 3.2 Biomass-Based Analysis

Due to continued poor sampling and resulting problems estimating catch at age, surplus production analyses (ASPIC) were updated to provide alternative perspectives on stock status. Results for the Georges Bank stock are similar to those from VP.A; biomass increased to $74 \%$ of $\mathrm{B}_{\text {MSY }}$ in 1998 at low F (Figure C4). The estimate of $\mathrm{B}_{\text {MSY }}(46.850 \mathrm{mt}$ ) was similar to the 1999 TRAC estimate ( 45.050 mt . Neilson et al. 1999).

### 4.0 Forecasts

Age-based, stochastic projections were performed using bootstrap distributions of January 1 , 1999 stock abundance at age, assuming F in 1999 was equal to $115 \% \mathrm{~F}$ in 1998 to account for additional scallop discards, and F in 2000-2009 were from the Amendment \#9 control rule targets. Age-1 abundance in 1999 was estimated from multiple survey indices in the terminal year. Recruitment in subsequent years was estimated from Beverton-Holt stock-recruitment relationships estimated from VPA estimates (Overholtz and Brodziak 1999). Mean weight at age. exploitation pattern, and proportion discard at age was assumed to be equal to the 19941998 average.

Age-based Projection of the Georges Bank yellowtail stock suggests that at status quo F in 1999, the stock is projected to approach $\mathrm{B}_{\mathrm{MSY}}$ (Table C6; Figure C 5 ). At the long-term F target, the stock is projected to fluctuate at levels greater than $\mathrm{B}_{\mathrm{MSY}}$ (approximately $50,000 \mathrm{mt}$ total biomass after 2003) and yield nearly MSY (approximately $13,000 \mathrm{mt}$ total yield after 2006).

Biomass-based. stochastic projections were performed using bootstrap distributions of January 1. 1999 stock biomass, assuming F in 1999 was equal to $115 \% \mathrm{~F}$ in 1998 to account for additional scallop discards), and F in 2000-2009 were from the Amendment \#9 control rule targets. Results for the Georges Bank stock are similar to those from age-based projection (Figure C5).

### 5.0 Sources of Uncertainty

- Estimates of catch at age may not be reliable due to poor sampling intensity. Therefore VPA and age-based projections may be misleading. Odd exploitation patterns and retrospective patterns may indicate inadequate sampling and mis-allocation of catch at age.
- Retrospective patterns indicate that VPA estimates of biomass and F may be overly optimistic. Updated VPAs may indicate that 1998 biomass levels are lower. and 1998 F was greater than reported here.
- Although a historical perspective from production models is valuable. current and projected biomass levels may not be reliable, because recruitment is implicitly assumed to be a function of stock biomass.
- Stock-recruit observations used to derive the relationships assumed in long-term projections are limited to a short time series of relatively low stock sizes. Therefore. long-term forecasts at relatively high stock sizes may be substantially biased.
- Conversion of $F$ on biomass targets to fully-recruited $F$ for age-based projections are imprecise approximations that are conditional on the assumed exploitation pattern. weight at age and future recruitment. Inaccuracies in these assumptions may bias projection results.
- The assumption of status quo proportional discards at age may be inaccurate. because recruitment of the strong 1997 year class may substantially change discard patterns.
- Estimates of prorated landings and discard ratios are based on preliminary logbook data and are subject to change.


### 6.0 References

Cadrin, S.X., W.J. Overholtz, J.D. Neilson, S. Gavaris, and S. Wigley. 1998. Stock assessment of Georges Bank yellowtail flounder for 1997. NEFSC Ref. Doc. 98-06.

Neilson, J.D. and S.X. Cadrin. 1998. Assessment of Georges Bank (5Zjmnh) yellowtail flounder. Canada Department of Fisheries and Oceans Research Document 98/67.

Overholtz. W. and J. Brodziak. 1999. Background stock-recruit data for eleven groundfish stocks. NEFSC NDS Working Paper No. 10 .

Table C1. Catch of Georges Bank yellowtail flounder (thousand mt).

| Year | landings | discards | Canada | Foreign | Total |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1963 | 11.0 | 5.6 | 0.0 | 0.1 | 16.7 |
| 1964 | 14.9 | 4.9 | 0.0 | 0.0 | 19.8 |
| 1965 | 14.2 | 4.4 | 0.0 | 0.8 | 19.4 |
| 1966 | 11.3 | 2.1 | 0.0 | 0.3 | 13.7 |
| 1967 | 8.4 | 5.5 | 0.0 | 1.4 | 15.3 |
| 1968 | 12.8 | 3.6 | 0.0 | 1.8 | 18.2 |
| 1969 | 15.9 | 2.6 | 0.0 | 2.4 | 20.9 |
| 1970 | 15.5 | 5.5 | 0.0 | 0.3 | 21.3 |
| 1971 | 11.9 | 3.1 | 0.0 | 0.5 | 15.5 |
| 1972 | 14.2 | 1.2 | 0.0 | 2.2 | 17.6 |
| 1973 | 15.9 | 0.4 | 0.0 | 0.3 | 16.5 |
| 1974 | 14.6 | 1.0 | 0.0 | 1.0 | 16.6 |
| 1975 | 13.2 | 2.7 | 0.0 | 0.1 | 16.0 |
| 1976 | 11.3 | 3.0 | 0.0 | 0.0 | 14.4 |
| 1977 | 9.4 | 0.6 | 0.0 | 0.0 | 10.0 |
| 1978 | 4.5 | 1.7 | 0.0 | 0.0 | 6.2 |
| 1979 | 5.5 | 0.7 | 0.0 | 0.0 | 6.2 |
| 1980 | 6.5 | 0.4 | 0.0 | 0.0 | 6.9 |
| 1981 | 6.2 | 0.1 | 0.0 | 0.0 | 6.3 |
| 1982 | 10.6 | 1.4 | 0.0 | 0.0 | 12.0 |
| 1983 | 11.4 | 0.1 | 0.0 | 0.0 | 11.4 |
| 1984 | 5.8 | 0.0 | 0.0 | 0.0 | 5.8 |
| 1985 | 2.5 | 0.0 | 0.0 | 0.0 | 2.5 |
| 1986 | 3.0 | 0.0 | 0.0 | 0.0 | 3.1 |
| 1987 | 2.7 | 0.2 | 0.0 | 0.0 | 3.0 |
| 1988 | 1.9 | 0.3 | 0.0 | 0.0 | 2.1 |
| 1989 | 1.1 | 0.1 | 0.0 | 0.0 | 1.2 |
| 1990 | 2.8 | 0.8 | 0.0 | 0.0 | 3.6 |
| 1991 | 1.8 | 0.2 | 0.0 | 0.0 | 2.0 |
| 1992 | 2.9 | 1.9 | 0.0 | 0.0 | 4.7 |
| 1993 | 2.1 | 1.1 | 0.7 | 0.0 | 3.9 |
| 1994 | 1.6 | 0.1 | 2.1 | 0.0 | 3.9 |
| 1995 | 0.3 | 0.0 | 0.5 | 0.0 | 0.8 |
| 1996 | 0.8 | 0.0 | 0.5 | 0.0 | 1.3 |
| 1997 | 1.0 | 0.1 | 0.8 | 0.0 | 1.8 |
| 1998 | 1.8 | 0.1 | 1.2 | 0.0 | 3.1 |
| average | 7.5 | 1.5 | 0.2 | 0.3 | 9.5 |
|  |  |  |  |  |  |
|  |  | 0.0 |  |  |  |

Table C2a. Canadian landings at age (thousands) of Georges Bank yellowtail Flounder (from Neilson et al. 1999).

|  | Age |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ Total |  |
| 1993 | 5 | 85 | 727 | 901 | 27 | 0 | 5 | 0 | 1750 |
| 1994 | 70 | 415 | 2890 | 1701 | 654 | 59 | 29 | 0 | 5818 |
| 1995 | 0 | 100 | 576 | 427 | 66 | 10 | 0 | 0 | 1179 |
| 1996 | 1 | 107 | 655 | 229 | 22 | 4 | 0 | 0 | 1018 |
| 1997 | 9 | 242 | 607 | 614 | 164 | 10 | 15 | 7 | 1668 |
| 1998 | 19 | 447 | 1086 | 642 | 254 | 29 | 6 | 0 | 2482 |
| mean | 17 | 190 | 1091 | 774 | 187 | 17 | 10 | 1 | 2287 |

Table C2b. U.S. landings at age (above) and mean weight at age (below) of Georges Bank yellowtail flounder.

| Landings at age (thousands) |  |  |  | Age |  | - |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | - 6 | 7 | $8+$ |  |
| 1973 | 0 | 3837 | 13076 | 9274 | 3743 | 1259 | 278 | 81 | 31548 |
| 1974 | 180 | 6297 | 7818 | 7397 | 3544 | 852 | 452 | 173 | 26713 |
| 1975 | 427 | 16851 | 6943 | 3391 | 2084 | 671 | 313 | 164 | 30844 |
| 1976 | 43 | 19320 | 5085 | 1347 | 532 | 434 | 287 | 147 | 27195 |
| 1977 | 31 | 6616 | 9805 | 1721 | 394 | 221 | 129 | 124 | 19041 |
| 1978 | 0 | 2140 | 3970 | 1660 | 459 | 102 | 37 | 35 | 8403 |
| 1979 | 17 | 6804 | 3396 | 1242 | 550 | 141 | 79 | 52 | 12281 |
| 1980 | 0 | 2371 | 8696 | 1419 | 321 | 85 | 4 | 10 | 12906 |
| 1981 | 6 | 479 | 5267 | 4555 | 796 | 122 | 4 | 0 | 11229 |
| 1982 | 217 | 13132 | 7061 | 3245 | $1031{ }^{\circ}$ | 62 | 19 | 3 | 24770 |
| 1983 | 239 | 7667 | 16016 | . 2316 | 625 | 109 | 10 | 8 | 26990 |
| 1984 | 244 | 1913 | 4266 | 4734 | 1592 | 257 | 47 | 17 | 13070 |
| 1985 | 371 | 3335 | 816 | 652 | 410 | 60 | 5 | 0 | 5649 |
| 1986 | 90 | 5733 | 978 | 347 | 161 | 52 | 16 | 8 | 7385 |
| 1987 | 15 | 1819 | 2730 | 761 | 132 | 39 | 32 | 41 | 5569 |
| 1988 | 0 | 1650 | 1181 | 624 | 165 | 15 | 20 | 3 | 3658 |
| 1989 | 0 | 1337 | 664 | 262 | 68 | 11 | 8 | 0 | 2350 |
| 1990 | 0 | 735 | 4582 | 738 | 105 | 17 | 3 | 0 | 6180 |
| 1991 | 0 | 27 | 867 | 2256 | 289 | 56 | 4 | 0 | 3499 |
| 1992 | 0 | 3183 | 1891 | 1176 | 502 | 20 | 7 | 0 | 6779 |
| 1993 | 0 | 375 | 1538 | 1392 | 287 | 65 | 4 | 1 | 3662 |
| 1994 | 0 | 129 | 2614 | 853 | 253 | 40 | 8 | 1 | 3897 |
| 1995 | 0 | 12 | 272 | 281 | 70 | 3 | 11 | 3 | 651 |
| 1996 | 0 | 161 | 751 | 482 | 144 | 5 | 5 | 1 | 1550 |
| 1997 | 0 | 205 | 616 | 875 | 175 | 16 | 30 | 12 | 1929 |
| 1998 | 0 | 422 | 1625 | 1156 | 366 | 53 | 14 | 0 | 3636 |
| mean | 72 | 4098 | 4328 | 2083 | 723 | 183 | 70 | 34 | 11592 |


| Landed weight at age (kg) |  |  |  | Age |  |  | 7 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | age-3 | age-4 | age-5 | 6 |  | 8+ |
| 1973 | 0.198 | 0.375 | 0.464 | 0.527 | 0.603 | 0.689 | 1.067 | 1.136 |
| 1974 | 0.200 | 0.378 | 0.500 | 0.609 | 0.680 | 0.725 | 0.906 | 1.249 |
| 1975 | 0.211 | 0.340 | 0.492 | 0.554 | 0.618 | 0.687 | 0.688 | 0.649 |
| 1976 | 0.185 | 0.339 | 0.545 | 0.636 | 0.741 | 0.814 | 0.852 | 0.866 |
| 1977 | 0.197 | 0.364 | 0.527 | 0.634 | 0.782 | 0.865 | 1.036 | 1.013 |
| 1978 | 0.182 | 0.337 | 0.513 | 0.684 | 0.793 | 0.899 | 0.930 | 0.948 |
| 1979 | 0.139 | 0.356 | 0.462 | 0.649 | 0.728 | 0.835 | 1.003 | 0.882 |
| 1980 | 0.138 | 0.354 | 0.495 | 0.656 | 0.813 | 1.054 | 1.256 | 1.214 |
| 1981. | 0.091 | 0.389 | 0.493 | 0.603 | 0.707 | 0.798 | 0.832 | 1.044 |
| 1982 | - 0.213 | 0.313 | 0.487 | 0.650 | 0.748 | 1.052 | 1.024 | 1.311 |
| 1983 | 0.215 | 0.296 | 0.440 | 0.604 | 0.736 | 0.952 | 1.018 | 0.987 |
| 1984 | 0.208 | 0.240 | 0.378 | 0.500 | 0.642 | 0.738 | 0.944 | 1.047 |
| 1985 | 0.236 | 0.363 | 0.497 | 0.647 | 0.733 | 0.819 | 0.732 | 1.044 |
| 1986 | 0.234 | 0.343 | 0.540 | 0.664 | 0.823 | 0.864 | 0.956 | 1.140 |
| 1987 | 0.212 | 0.338 | 0.523 | 0.666 | 0.680 | 0.938 | 0.793 | 0.788 |
| 1988 | -... | 0.351 | 0.557 | - 0.688 | 0.855 | 1.054 | 0.873 | 1.385 |
| 1989 | ---- | 0.355 | 0.543 | 0.725 | 0.883 | 1.026 | 1.254 | 1.044 |
| 1990 | ---- | 0.337 | 0.419 | 0.588 | 0.699 | 0.807 | 1.230 | 1.044 |
| 1991 | ---- | 0.270 | 0.383 | 0.484 | 0.728 | 0.820 | 1.306 | 1.044 |
| 1992 | ---- | 0.341 | 0.381 | 0.528 | 0.648 | 1.203 | 1.125 | 1.044 |
| 1993 | ---- | 0.316 | 0.390 | 0.510 | 0.562 | 0.858 | 1.263 | 1.044 |
| 1994 | ---- | 0.300 | 0.355 | 0.473 | 0.629 | 0.787 | 0.896 | 1.166 |
| 1995 | ---- | 0.309 | 0.379 | 0.465 | 0.583 | 0.778 | 0.785 | 0.531 |
| 1996 | ---- | 0.321 | 0.417 | 0.569 | 0.726 | 0.926 | 1.031 | 1.209 |
| 1997 | ---- | 0.353 | 0.416 | 0.525 | 0.668 | 0.867 | 0.920 | 1.217 |
| 1998 | ---- | 0.360 | 0.468 | 0.540 | 0.664 | 0.819 | 0.879 |  |
| mean | 0.191 | 0.336 | 0.464 | 0.591 | 0.710 | 0.872 | 0.985 | 1.042 |

Table C3. U.S. discards at age (above) and recent mean weights at age (below) of Georges Bank yellowtail flounder.

| Discards at age (thousands) |  |  |  | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ | Total |
| 1973 | 347 | 1053 | 167 | 2 | 0 | 0 | 0 | 0 | 1569 |
| 1974 | 1963 | 2674 | 86 | 1 | 0 | 0 | 0 | 0 | 4724 |
| 1975 | 3945 | 8433 | 114 | 1 | 0 | 0. | 0 | 0 | 12493 |
| 1976 | 572 | 11692 | 61 | 0 | 0 | 0 | 0 | 0 | 12325 |
| 1977 | 299 | 1964 | 112 | 0 | 0 | 0 | 0 | 0 | 2375 |
| 1978 | 9659 | 965 | 64 | 0 | 0 | 0 | 0 | 0 | 10688 |
| 1979 | 216 | 2701 | 49 | 0 | 0 | 0 | . 0 | 0 | 2966 |
| 1980 | 309 | 1201 | 125 | 0 | 0 | 0 | 0 | 0 | 1635 |
| 1981 | 49 | 250 | 84 | 1 | 0 | 0 | 0 | 0 | 384 |
| 1982 | 1846 | 4359 | 61 | 1 | 0 | 0 | 0 | 0 | 6267 |
| 1983 | 457 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 479 |
| 1984 | 184 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 188 |
| 1985 | 279 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 289 |
| 1986 | 68 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 106 |
| 1987 | 125 | 834 | 21 | 0 | 0 | 0 | 0 | 0 | 980 |
| 1988 | 483 | 717 | 10 | 0 | 0 | 0 | 0 | 0 | 1210 |
| 1989 | 185 | 179 | 4 | 0 | 0 | 0 | 0 | 0 | 368 |
| 1990 | 219 | 1196 | 1541 | ¢ 2 | 2 | 0 | 0 | 0 | 3020 |
| 1991 | 412 | 27 | 355 | 174 | 4 | 0 | 0 | 0 | 972 |
| 1992 | 2389 | 5176 | 636 | 93 | 8 | 0 | 0 | 0 | 8302 |
| 1993 | 5189 | 549 | 512 | 99 | 4 | 0 | 0 | 0 | 6353 |
| 1994 | 1 | 317 | 238 | 17 | 3 | 0 | 0 | 0 | 577 |
| 1995 | 14 | 45 | 47 | 7 | 0 | 0 | 0 | 0 | 114 |
| 1996 | 49 | 115 | 103 | 6 | 0 | 0 | 0 | 0 | 273 |
| 1997 | 7 | 148 | 35 | 13 | 1 | 0 | 0 | 0 | 205 |
| 1998 | 7 | 102 | 81 | 26 | 4 | 0 | 0 | 0 | 220 |
| mean | 1126 | 1722 | 173 | 19 | 1 | 0 | 0 | 0 | 3154 |


| Discarded weight at age (kg) |  |  |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ |
| 1994 | 0.130 | 0.238 | 0.287 | 0.417 | 0.512 | 0.622 | --- | -- |
| 1995 | 0.155 | 0.233 | 0.283 | 0.357 | 0.496 | 0.593 | --- | --- |
| 1996 | 0.137 | 0.266 | 0.312 | 0.418 | --- | --- | --- | --- |
| 1997 | 0.162 | 0.250 | 0.315 | 0.442 | 0.544 | 0.671 | 0.792 | 0.895 |
| 1998 | 0.190 | 0.280 | 0.380 | 0.450 | 0.590 | 0.700 | 0.760 | --- |
| mean | 0.155 | 0.254 | 0.315 | 0.417 | 0.536 | 0.647 | 0.776 | 0.713 |

Table C4a. Survey indices of Georges Bank yellowtail abundance and biomass.

| NEFSC Fall Survey |  |  |  |  |  | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ | Total | kg/tow |
| 1963 | 0.000 | 14.722 | 7.896 | 11.226 | 1.858 | 0.495 | 0.281 | 0.034 | 0.233 | 36.746 | 12.791 |
| 1964 | 0.000 | 1.721 | 9.723 | 7.370 | 5.998 | 2.690 | 0.383 | 0.095 | 0.028 | 28.007 | 13.625 |
| 1965 | 0.014 | 1.138 | 5.579 | 5.466 | 3.860 | 1.803 | 0.162 | 0.284 | 0.038 | 18.345 | 9.104 |
| 1966 | 1.177 | 8.772 | 4.776 | 2.070 | 0.837 | 0.092 | 0.051 | 0.000 | 0.000 | 17.775 | 3.989 |
| 1967 | 0.106 | 9.137 | 9.313 | 2.699 | 1.007 | 0.309 | 0.076 | 0.061 | 0.000 | 22.708 | 7.577 |
| 1968 | 0.000 | 11.782 | 11.946 | 5.758 | 0.766 | 0.9 | 0.059 | 0.000 | 0.000 | 31.254 | 10.535 |
| 1969 | 0.135 | 8.106 | 10.381 | 5.855 | 1.662 | 0.553 | 0.149 | 0.182 | 0.000 | 27.023 | 9.278 |
| 1970 | 1.048 | 4.610 | 5.133 | 3.144 | 1.952 | 0.451 | 0.063 | 0.017 | 0.000 | 16.417 | 4.978 |
| 197 | 0.025 | 3.627 | 6.949 | 4.904 | 2.248 | 0.5 | 0.234 | 0.024 | 0.024 | 18.586 | 6.362 |
| 1972 | 0.785 | 2.42 | 6.525 | 4.82 | 2.09 | 0.672 | 0.279 | 0.000 | 0.000 | 17.604 | 6.328 |
| 1973 | 0.094 | 2.494 | 5.497 | 5.104 | 2.944 | 1.216 | 0.416 | 0.171 | 0.031 | 17.967 | 6.600 |
| 1974 | 1.030 | 4.623 | 2.854 | 1.524 | 1.060 | 0.460 | 0.249 | 0.131 | 0.000 | 11.931 | 3.734 |
| 1975 | 0.361 | 4.625 | 2.511 | 0.877 | 0.572 | 0.334 | 0.033 | 0.000 | 0.031 | 9.344 | 2.365 |
| 1976 | 0.000 | 0.336 | 1.929 | 0.475 | 0.117 | 0.122 | 0.033 | 0.000 | 0.067 | 3.079 | 1.533 |
| 1977 | 0.000 | 0.928 | 2.161 | 1.649 | 0.618 | 0.113 | 0.056 | 0.036 | 0.016 | 5.577 | 2.828 |
| 1978 | 0.037 | 4.729 | 1.272 | 0.773 | 0.406 | 0.139 | 0.01 | 0.000 | 0.024 | 7.39 | 2.383 |
| 1979 | 0.018 | 1.312 | 1.999 | 0.316 | 0.122 | 0.138 | 0.038 | 0.064 | 0.007 | 4.014 | 1.520 |
| 1980 | 0.078 | 0.761 | 5.086 | 6.050 | 0.678 | 0.217 | 0.162 | 0.006 | 0.033 | 13.0 | 6.722 |
| 198 | 0.000 | 1.58 | 2.333 | 1.63 | 0.50 | 0.121 | 0.083 | 0.013 | 00000 | 6.264 | 2.621 |
| 1982 | 0.000 | 2.42 | 2.18 | 1.59 | 0.423 | 0.089 | 0.000 | 0.000 | 0.000 | 6.711 | 2.271 |
| 1983 | 0.000 | 0.109 | 2.284 | 1.91 | 0.47 | 0.068 | 0.012 | 0.000 | 0.038 | 4.898 | 2.131 |
| 1984 | 0.012 | 0:661 | 0.400 | 0.306 | 2.428 | 0.090 | 0.029 | 0.000 | 0.018 | 3.944 | 0.593 |
| 1985 | 0.010 | 1.350 | 0.560 | 0.160 | 0.040 | 0.080 | 0.000 | 0.000 | 0.000 | 2.200 | 0.709 |
| 1986 | 0.000 | 0.280 | 1.110 | 0.350 | 0.070 | 0.000 | 0.000 | 0.000 | 0.000 | 1.810 | 0.820 |
| 1987 | 0.000 | 0.113 | 0.390 | 0.396 | 0.053 | 0.079 | 0.000 | 0.000 | 0.000 | 1.031 | 0.509 |
| 1988 | 0.011 | 0.019 | 0.213 | 0.102 | 0.031 | 0.000 | 0.000 | 0.000 | 0.000 | 0.376 | 0.171 |
| 1989 | 0.027 | 0.248 | 1.992 | 0.774 | 0.069 | 0.066 | 0.000 | 0.000 | 0.000 | 3.176 | 0.977 |
| 199 | 0.14 | 0.000 | 0.326 | 1.517 | 0.280 | 0.01 | 0.000 | 0.000 | 0.000 | 2.284 | 0.725 |
| 199 | 0.000 | 2.100 | 0.27 | 0.439 | 0.358 | 0.000 | 0.000 | 0.000 | 0.000 | 3.172 | 0.730 |
| 1992 | 0.000 | 0.151 | 0.396 | 0.712 | 0.16 | 0.14 | 0.027 | 0.000 | 0.000 | 1.592 | 0.576 |
| 1993 | 0.000 | 0.842 | 0.136 | 0.587 | 0.536 | 0.000 | 0.000 | 0.000 | 0.000 | 2.101 | 0.545 |
| 1994 | 0.010 | 1.200 | 0.220 | 0.980 | 0.710 | 0.260 | 0.030 | 0.030 | 0.000 | 3.440 | 0.897 |
| 1995 | 0.070 | 0.280 | 0.120 | 0.350 | 0.280 | 0.050 | 0.010 | 0.000 | 0.000 | 1.160 | 0.354 |
| 1996 | 0.000 | 0.140 | 0.350 | 1.870 | 0.450 | 0.070 | 0.000 | 0.000 | 0.000 | 2.880 | 1.303 |
| 1997 | 0.000 | 1.392 | 0.533 | 3.442 | 2.090 | 1.071 | 0.082 | 0.000 | 0.000 | 8.611 | 3.781 |
| 1998 | 0.050 | 1.900 | 4.817 | 4.202 | 1.190 | 0.298 | 0.055 | 0.019 | 0.000 | 12.531 | 4.347 |
| mean | 0.146 | 2.796 | 3.338 | 2.539 | 1.082 | 0.383 | 0.085 | 0.032 | 0.016 | 10.417 | 3.786 |

Table C4b. Survey indices of Georges Bank yellowtail abundance and biomass.

| NEFSC Spring Survey |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | Age |  |  |  |  |  |  |  | 5 | 6 | 7 | $8+$ | Total | kg/tow |
| 1968 | 0.149 | 3.364 | 3.579 | 0.316 | 0.084 | 0.160 | 0.127 | 0.000 | 7.779 | 2.813 |  |  |  |  |  |  |  |  |
| 1969 | 1.015 | 9.406 | 11.119 | 3.096 | 1.423 | 0.454 | 0.188 | 0.057 | 26.758 | 11.170 |  |  |  |  |  |  |  |  |
| 1970 | 0.093 | 4.485 | 6.030 | 2.422 | 0.570 | 0.121 | 0.190 | 0.000 | 13.911 | 5.312 |  |  |  |  |  |  |  |  |
| 1971 | 0.791 | 3.335 | 4.620 | 3.754 | 0.759 | 0.227 | 0.050 | 0.029 | 13.564 | 4.607 |  |  |  |  |  |  |  |  |
| 1972 | 0.138 | 7.136 | 7.198 | 3.514 | 1.094 | 0.046 | 0.122 | 0.000 | 19.247 | 6.450 |  |  |  |  |  |  |  |  |
| 1973 | 1.931 | 3.266 | 2.368 | 1.063 | 0.410 | 0.173 | 0.023 | 0.020 | 9.254 | 2.938 |  |  |  |  |  |  |  |  |
| 1974 | 0.316 | 2.224 | 1.842 | 1.256 | 0.346 | 0.187 | 0.085 | 0.009 | 6.265 | 2.719 |  |  |  |  |  |  |  |  |
| 1975 | 0.420 | 2.939 | 0.860 | 0.298 | 0.208 | 0.068 | 0.000 | 0.013 | 4.806 | 1.676 |  |  |  |  |  |  |  |  |
| 1976 | 1.034 | 4.368 | 1.247 | 0.311 | 0.196 | 0.026 | 0.048 | 0.037 | 7.268 | 2.273 |  |  |  |  |  |  |  |  |
| 1977 | 0.000 | 0.671 | 1.125 | 0.384 | 0.074 | 0.013 | 0.000 | 0.000 | 2.267 | 0.999 |  |  |  |  |  |  |  |  |
| 1978 | 0.936 | 0.798 | 0.507 | 0.219 | 0.026 | 0.000 | 0.008 | 0.000 | 2.494 | 0.742 |  |  |  |  |  |  |  |  |
| 1979 | 0.279 | 1.933 | 0.385 | 0.328 | 0.059 | 0.046 | 0.041 | 0.000 | 3.072 | 1.227 |  |  |  |  |  |  |  |  |
| 1980 | 0.057 | 4.644 | 5.761 | 0.473 | 0.057 | 0.037 | 0.000 | 0.000 | 11.030 | 4.456 |  |  |  |  |  |  |  |  |
| 1981 | 0.012 | 1.027 | 1.779 | 0.721 | 0.205 | 0.061 | 0.000 | 0.026 | 3.830 | 1.960 |  |  |  |  |  |  |  |  |
| 1982 | 0.045 | 3.742 | 1.122 | 1.016 | 0.455 | 0.065 | 0.000 | 0.026 | 6.472 | 2.500 |  |  |  |  |  |  |  |  |
| 1983 | 0.000 | 1.865 | 2.728 | 0.531 | 0.123 | 0.092 | 0.061 | 0.092 | 5.492 | 2.642 |  |  |  |  |  |  |  |  |
| 1984 | 0.000 | 0.093 | 0.809 | 0.885 | 0.834 | 0.244 | 0.000 | 0.000 | 2.865 | 1.646 |  |  |  |  |  |  |  |  |
| 1985 | 0.110 | 2.198 | 0.262 | 0.282 | 0.148 | 0.000 | 0.000 | 0.000 | 3.000 | 0.988 |  |  |  |  |  |  |  |  |
| 1986 | 0.027 | 1.806 | 0.291 | 0.056 | 0.137 | 0.055 | 0.000 | 0.000 | 2.372 | 0.847 |  |  |  |  |  |  |  |  |
| 1987 | 0.000 | 0.128 | 0.112 | 0.133 | 0.053 | 0.055 | 0.000 | 0.000 | 0.480 | 0.329 |  |  |  |  |  |  |  |  |
| 1988 | 0.078 | 0.275 | 0.366 | 0.242 | 0.199 | 0.027 | 0.000 | 0.000 | 1.187 | 0.566 |  |  |  |  |  |  |  |  |
| 1989 | 0.047 | 0.424 | 0.740 | 0.290 | 0.061 | 0.022 | 0.022 | 0.000 | 1.605 | 0.729 |  |  |  |  |  |  |  |  |
| 1990 | 0.000 | 0.065 | 1.108 | 0.393 | 0.139 | 0.012 | 0.045 | 0.000 | 1.762 | 0.699 |  |  |  |  |  |  |  |  |
| 1991 | 0.435 | 0.000 | 0.254 | 0.675 | 0.274 | 0.020 | 0.000 | 0.000 | 1.659 | 0.631 |  |  |  |  |  |  |  |  |
| 1992 | 0.000 | 2.010 | 1.945 | 0.598 | 0.189 | 0.000 | 0.000 | 0.000 | 4.742 | 1.566 |  |  |  |  |  |  |  |  |
| 1993 | 0.046 | 0.290 | 0.500 | 0.317 | 0.027 | 0.000 | 0.000 | 0.000 | 1.180 | 0.482 |  |  |  |  |  |  |  |  |
| 1994 | 0.000 | 0.621 | 0.638 | 0.357 | 0.145 | 0.043 | 0.000 | 0.000 | 1.804 | 0.660 |  |  |  |  |  |  |  |  |
| 1995 | 0.040 | 1.180 | 4.810 | 1.490 | 0.640 | 0.010 | 0.000 | 0.000 | 8.170 | 2.579 |  |  |  |  |  |  |  |  |
| 1996 | 0.030 | 0.990 | 2.630 | 2.700 | 0.610 | 0.060 | 0.000 | 0.000 | 7.020 | 2.853 |  |  |  |  |  |  |  |  |
| 1997 | 0.019 | 1.169 | 3.733 | 4.081 | 0.703 | 0.134 | 0.000 | 0.000 | 9.837 | 4.359 |  |  |  |  |  |  |  |  |
| 1998 | 0.000 | 2.081 | 1.053 | 1.157 | 0.759 | 0.323 | 0.027 | 0.000 | 5.400 | 2.324 |  |  |  |  |  |  |  |  |
| 1999 | 0.061 | 5.790 | 13.200 | 3.319 | 1.980 | 0.520 | 0.401 | 0.030 | 25.300 | 11.913 |  |  |  |  |  |  |  |  |
| mean | 0.253 | 2.323 | 2.647 | 1.146 | 0.406 | 0.103 | 0.045 | 0.011 | 6.934 | 2.739 |  |  |  |  |  |  |  |  |

Table C4c. Survey indices of Georges Bank yellowtail abundance and biomass.

| Age |  |  |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Canadian Survey <br> Year | 1 | 2 | 3 | 4 | 5 | $6+$ | Total | kg/tow |
| 1987 | 0.12 | 0.68 | 2.00 | 1.09 | 0.06 | 0.00 | 3.95 | 1.26 |
| 1988 | 0.00 | 0.66 | 1.89 | 0.80 | 0.59 | 0.01 | 3.96 | 1.24 |
| 1989 | 0.11 | 0.78 | 0.80 | 0.32 | 0.10 | 0.02 | 2.13 | 0.47 |
| 1990 | 0.00 | 1.27 | 4.62 | 1.12 | 0.43 | 0.01 | 7.45 | 1.58 |
| 1991 | 0.02 | 0.59 | 1.72 | 2.91 | 0.99 | 0.00 | 6.24 | 1.76 |
| 1992 | 0.22 | 10.04 | 4.52 | 1.21 | 0.16 | 0.00 | 16.14 | 2.48 |
| 1993 | 0.33 | 2.16 | 5.04 | 3.47 | 0.62 | 0.00 | 11.63 | 2.64 |
| 1994 | 0.00 | 6.03 | 3.33 | 3.08 | 0.75 | 0.33 | 13.51 | 2.75 |
| 1995 | 0.21 | 1.31 | 4.07 | 2.22 | 1.14 | 0.11 | 9.07 | 2.03 |
| 1996 | 0.45 | 5.54 | 8.44 | 7.49 | 1.37 | 0.16 | 23.45 | 5.30 |
| 1997 | 0.10 | 9.48 | 15.16 | 19.09 | 3.11 | 0.54 | 47.49 | 13.29 |
| 1998 | 0.92 | 3.10 | 3.81 | 5.15 | 2.44 | 0.59 | 16.01 | 4.29 |
| 1999 | 0.20 | 11.66 | 14.96 | 14.40 | 12.99 | 2.87 | 57.07 | 17.67 |
| mean | 0.21 | 4.10 | 5.41 | 4.80 | 1.90 | 0.36 | 16.78 | 4.37 |

## Scallop Survey

| Year | age-1 |
| :---: | :---: |
| 1982 | 0.313 |
| 1983 | 0.140 |
| 1984 | 0.233 |
| 1985 | 0.549 |
| 1986 | 0.103 |
| 1987 | 0.047 |
| 1988 | 0.116 |
| 1989 | 0.195 |
| 1990 | 0.100 |
| 1991 | 2.117 |
| 1992 | 0.167 |
| 1993 | 1.129 |
| 1994 | 1.503 |
| 1995 | 0.609 |
| 1996 | 0.508 |
| 1997 | 1.062 |
| 1998 | 1.872 |
| mean | 0.633 |

Table C5a. Estimates of Georges Bank yellowtail flounder abundance at age. STOCK NUMBERS (Jan 1) in thousands - C: XProgram Eiles $\operatorname{CWHAT} \backslash$ gibyt $99 . ミ$

|  | 1973 | 1974 | 1975 | 1976 | 1977 | 2978 | 290 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 28290 | 50265 | 68516 | 22919 | 15760 | 50823 | 23375 |
| 2 | 23279 | 22848 | 39214 | 52140 | 18208 | 12605 | 329- |
| 3 | 28937 | 14635 | 10589 | 9228 | 14628 | 7144 | $75: 0$ |
| 4 | 16960 | 11709 | 4830 | 2284 | 2899 | 3003 | 2-90 |
| 5 | 6729 | 5492 | 2893 | - 885 | 651 | 816 | 95- |
| 6 | 2859 | 2240 | 1551 | 1417 | 768 | 304 | 465 |


| i- | 107055 | 107189 | 127593 | 88873 | 52914 | 74695 | 67376 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 1 | 22099 | 61066 | 21627 | 5818 | 8620 | 14594 | 6660 |
| 2 | 18927 | 17814 | 49947 | 15840 | 4134 | 6670 | 11361 |
| 3 | 18312 | 12264 | 13925 | 25067 | 6011 | 1650 | 2434 |
| 4 | 3032 | 7011 | 5199 | 4957 | 6031 | 1062 | 613 |
| 5 | 677 | 1198 | 1618 | 1319 | 1962 | 654 | 279 |
| 6 | 206 | 185 | 129 | 264 | 382 | 102 | 129 |


| I+ | 63252 | 99538 | 92445 | 53266 | 27141 | 24732 | 21476 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| 1 | 7023 | 19353 | 8537 | 11736 | 22.157 | 16337 | 13426 |
| 2 | 5310 | 5624 | 15408 | 6822 | 9410 | 17768 | 11214 |
| 3 | 4079 | 1947 | 2462 | 1:243 | 3838 | 7656 | 6983 |
| 4 | 1108 | 851 | 516 | 1412 | 3665 | 2037 | 3981 |
| 5 | 188 | 219 | 132 | 185 | 432 | 802 | 519 |
| 6 | 155 | 49 | 36 | 34 | 86 | 42 | 120 |
| $1+$ | 17863 | 28042 | 27091 | 31432 | 39588 | 44640 | 36243 |
|  | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |  |
| 1 | 15860 | 26676 | 19221 | 37845 | 79474 | 00 |  |
| 2 | 6292 | 12921 | 21828 | 15692 | 30971 | 65.044 |  |
| 3 | 8268 | 4373 | 10437 | 17525 | 12309 | 24478 |  |
| 4 | 3205 | 1574 | 2770 | 7180 | 13210 | 7551 |  |
| 5 | 1095 | 298 | 641 | 1620 | 4519 | 9165 |  |
| 6 | 157 | 58 | 57 | 425 | 743 | 3651 |  |
| $1+$ | 34878 | 45899 | 54955 | 80287 | 141226 | 109889 |  |

Table C5b. Estimates of Georges Bank yellowtail flounder fishing mortality at age.

| EISHING | $\begin{gathered} \text { MORTALITY } \\ 1973 \end{gathered}$ | 1974 | $\begin{gathered} C: \begin{array}{c}  \\ \\ 1975 \end{array}, ~(\text { Program } \end{gathered}$ | $\begin{gathered} \text { Files } \\ 1976 \end{gathered}$ | $\begin{gathered} T \text { \gbyts } \\ 1977 \end{gathered}$ | 1978 | 2979 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.01 | 0.05 | 0.07 | 0.03 | 0.02 | 0.24 | 0.01 |
| $=$ | 0.26 | 0.57 | 1.25 | 1.07 | 0.74 | 0.32 | 0. 39 |
| 3 | 0.70 | 0.91 | 1.33 | 0.96 | 1.38 | 0.98 | 0.72 |
| 4 | 0.93 | 1.20 | 1.50 | 1.05 | 1.07 | 0.94 | 0.98 |
| E | 0.95 | 1.25 | 1.59 | 2.09 | 1.10 | 0.97 | 1.01 |
| E | 0.95 | 1.25 | 1.59 | 1.09 | 1.10 | 0.97 | 1.01 |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 1 | 0.02 | 0.00 | 0.11 | 0.14 | 0.06 | 0.05 | 0.03 |
| 2 | 0.23 | 0.05 | 0.49 | 0.77 | 0.72 | 0.81 | 0.82 |
| 3 | 0.76 | 0.66 | 0.83 | 1.22 | 1.53 | 0.79 | 0.59 |
| 4 | 0.73 | 1.27 | 1.17 | 0.73 | 2.02 | 1.14 | 0.98 |
| 5 | 0.74 | 1.33 | 1.22 | 0.74 | 2.27 | 1.18 | 1.01 |
| 6 | 0.74 | 1.33 | 1.22 | 0.74 | 2.27 | 1.18 | 1.01 |
|  | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| $\square$ | 0.02 | 0.03 | 0.02 | 0.02 | 0.02 | 0.18 | 0.56 |
| こ | 0.80 | 0.63 | 0.12 | 0.38 | 0.01 | 0.73 | 0.10 |
| 3 | 1.37 | 1.13 | 0. 36 | 0.92 | 0.43 | 0.45 | 0.58 |
| 4 | 1.42 | 1.66 | 0.82 | 0.98 | 1. 32 | 1.17 | 1.09 |
| 5 | 1. 50 | 1. 79 | 0.84 | 1.01 | 1.39 | 1.21 | 1.13 |
| E | 1.50 | 1.79 | 0.84 | 1.01 | 1.39 | 1.21 | 1.13 |
|  | 1994 | 1995 | 1996 | 1997 | 1998 |  |  |
| 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  |  |
| 2 | 0.16 | 0.01 | 0.02 | 0.04 | 0.04 |  |  |
| 3 | 1.46 | 0.26 | 0.17 | 0.08 | 0.29 |  |  |
| 4 | 2.18 | 0.70 | 0.34 | 0.26 | 0.17 |  |  |
| 5 | 2.50 | 0.71 | 0.34 | 0.26 | 0.17 |  |  |
| $\bar{\square}$ | 2.50 | 0.71 | 0.34 | 0.26 | 0.17 |  |  |

Average $E$ for 4,5

|  | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4,5 | 0.94 | 1.22 | 1.54 | 1.07 | 1.09 | 0.96 | 0.99 |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 4, 5 | 0.74 | 1.30 | 1.19 | 0.73 | 2.14 | 1.16 | 1.00 |
|  | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| 4,5 | 1.46 | 1.73 | 0.83 | 1.00 | 2.35 | 1.19 | 1.11 |
|  | 1994 | 1995 | 1996 | 1997 | 1998 |  |  |
| 4,5 | 2.34 | $0.70^{\circ}$ | 0.34 | 0.26 | 0.17 |  |  |

Blomass Welgnted $F$

| 2972 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.61 | 0.74 | 0.91 | 0.92 | 0.93 | 0.52 | 0.44 |
| 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 0.48 | 0.36 | 0.62 | 0.96 | 2. 38 | 0.62 | 0.68 |
| 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| 0.94 | 0.54 | 0.18 | 0.62 | 0.31 | 0.61 | 0.49 |
| 1994 | 1995 | 1996 | 1997 | 1998 |  |  |
| 0.84 | 0.11 | 0.10 | 0.10 | 0.10 |  |  |

Table C55．Estimates of Georges Bank yellowtail flounder mean biomass（mt）．

| MEAN | BIOMASS（using 1973 | $\begin{aligned} & \text { catch } \\ & 1974 \end{aligned}$ | mean weights 1975 | $\begin{aligned} & \text { at age) } \\ & .1976 \end{aligned}$ | 1977 | 1978 | 1979 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 2547 | 4451 | 5996 | 2047 | 1412 | 4120 | 2105 |
| 2 | 6462 | 5404 | 6404 | 8972 | 3985 | 3040 | 8088 |
| 3 | 8797 | 4405 | 2648 | 2963 | ． 3847 | 2140 | ここご |
| 4 | 5360 | 3840 | 1288 | 828 | 1042 | 1224 | ミここ |
| 5 | 2408 | 1973 | 832 | 368 | 284 | $38=$ | 405 |
| 6 | 1169 | 858 | 496 | 648 | 371 | 161 | こここ |
| $1+$ | 26743 | 20930 | 17663 | 15826 | 10943 | .11066 | 23934 |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 1 | 1988 | 5532 | 1858 | 493 | 760 | 1291 | 596 |
| 2 | 4883 | 5368 | 10718 | 3003 | 649 | 1526 | 2431 |
| 3 | 5791 | 4034 | 4211 | 5879 | 1079 | 520 | 910 |
| 4 | 1296 | 2218 | 1839 | 1952 | 1210 | 379 | 239 |
| 5 | 356 | 435 | 646 | 629 | 467 | 260 | 133 |
| 6 | 140 | 76 | 73 | 163 | 104 | 45 | 65 |
| $1+$ | 14454 | 17662 | 19346 | 12119 | 4271 | 4021 | 4374 |
|  | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| 1 | 630 | 1731 | 765 | 1053 | 1988 | 1362 | 941 |
| 2 | 1036 | 1221 | 4520 | 1458 | 2194 | 3268 | 2658 |
| 3 | 1073 | 598 | 1024 | 2630 | 1021 | 2023 | 1781 |
| 4 | 365 | 265 | 234 | 475 | 902 | 576 | 1125 |
| 5 | 61 | 81 | 72 | 75 | 157 | 277 | 161 |
| 6 | 70 | 22 | 23 | 16 | 36 | 27 | 57 |
| $1+$ | 3235 | 3919 | 6638 | 5706 | 6297 | 7533 | 6723 |
|  | 1994 | 1995 | 1996 | 1997 | 1998 |  |  |
| 1 | 1434 | 2417 | 1740 | 3429 | 7202 |  |  |
| 2 | 1382 | 3025 | 6056 | 4305 | 8003 |  |  |
| 3 | 1416 | 1288 | 3561 | 6992 | 3876 |  |  |
| 4 | 576 | 481 | 1121 | 3402 | 5665 |  |  |
| 5 | 237 | 114 | 331 | 923 | 2472 |  |  |
| 6 | 43 | 30 | 38 | 297 | 620 |  |  |
| $1+$ | 5089 | 7355 | 12847 | 19348 | 27837 |  |  |

Table C5d．Estimates of Georges Bank yellowtail flounder spawning stock biomass（mt）．
SSE AT THE START OE THE SPAWNING SEASON－MALES AND EEMALES（MT）（Using SSE mean weignts）

|  | 1973 | 1974 | 1975 | 1976 | 1977 | 2978 | 259 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Z | 00 | 00 | 00 | 00 | 00 | 00 | 00 |
| 2 | 2796 | 2530 | 2984 | 4200 | 1870 | 1¢23 | ごと－ |
| $\vdots$ | 8895 | 4500 | 2678 | 3026 | 3883 | 2185 | ころここ |
| 4 | 5531 | 3982 | 1319 | 861 | 1084 | 1275 | Eア |
| 5 | 2509 | 2042 | 848 | 383 | 296 | 397 | 422 |
| $\varepsilon$ | 1218 | 888 | 505 | 673 | 386 | 168 | 234 |
| i＋ | 20949 | ． 13942 | 8334 | 9143 | 7519 | 5438 | 7616 |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 1 | 00 | 00 | 00 | 00 | 00 | 00 | 00 |
| 2 | 2260 | 2678 | 5454 | 1534 | 629 | 1480 | 2358 |
| 3 | 5918 | 4161 | 4347 | 6031 | 1103 | 543 | 947 |
| $\leq$ | 1351 | 2295 | 1908 | 2035 | 1195 | 394 | 248 |
| 5 | 371 | 449 | 670 | 656 | 450 | 270 | 139 |
| 6 | 146 | 78 | 75 | 170 | 101 | 47 | 67 |
| $1+$ | 10047 | 9660 | 12455 | 10426 | 3479 | 2733 | 3760 |
|  | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| 2 | 00 | 00 | 00 | 00 | 00 | 00 | 00 |
| 2 | 1004 | 1183 | 4298 | 1403 | 2072 | 1772 | 1412 |
| 3 | 1106 | 621 | 1058 | 2741 | 1058 | 2099 | 1593 |
| 4 | 375 | 269 | 244 | 495 | 932 | 598 | 1170 |
| 5 | 63 | 82 | 75 | 78 | 162 | 287 | 167 |
| 6 | 72 | 22 | 24 | 17 | 37 | 28 | 59 |
| －－ | 2620 | 2177 | 5700 | 4733 | 4260 | 4784 | 4402 |
|  | 1994 | 1995 | 1996 | 1997 | 1998 |  |  |



Table C6. Summary of results from stochastic projection of Georges Bank yellowtail flounder.



Figure C1. Total catch of Georges Bank yellowtail flounder.


Figure C2. Survey indices of Georges Bank yellowtail flounder biomass.

 SSB Year; Recruitment Yearclass


Figure C3. Summary of Georges Bank yellowtail VPA results.


Figure C4. Status of the Georges Bank yellowtail flounder stock.


Figure C5. Projection of Georges Bank yellowtail flounder biomass at the Amendment \#9 long-term fishing target in 2000-2009.

## D. Southern New England Yellowtail Flounder by S.X. Cadrin

### 1.0 Background

The southern New England yellowtail stock was at low biomass (less than $25 \% \mathrm{~B}_{\text {MSY }}$ ) at low F (fully recruited F was 0.2) in 1997 (NEFSC 1998). This report updates catch and survey indices. updates estimates of stock size and fishing mortality estimates. and evaluates medium-term projections.

### 2.0 1999 Assessment

2.1 1998 Landings
U.S. landings were prorated as described in Cadrin et al. (1999; Table D1; Figure D1). Landings from southern New England increased 48\%.

Sampling intensity of landings in 1998 was poor. The large market category from southern New England was not sampled in the second half of the year, and there were no commercial samples of any market category in the fourth quarter, which comprised nearly half of the annual landings. Landings at length were estimated by half year and market category, where available, but firsthalf large samples were used to characterize second-half large landings. Landings at age and mean weights at age are reported in Table D2.

### 2.21998 Discards

Discarded catch was estimated from logbook information on discard to kept ratios by half-year and gear (NEFSC 1998; Table D4). Discard ratios are similar to those estimated in recent years for the southern New England yellowtail stocks, which were also based on logbook information. Discards at age were estimated from sea sampled lengths and pooled commercial-survey agelength keys. Discards at age and recent mean weights at age are reported in Table D3.

### 2.3 1998-1999 Survev Indices

Survey abundance and biomass indices are reported in Table D4. Estimates are from valid tows in southern New England (offshore strata 5, 6, 9, 10; scallop strata 33-48), standardized according to net. vessel, and door changes (NEFSC 1998). All survey indices of total abundance and total biomass increased for the southern New England yellowtail stocks in 1998 and 1999 (Figure D2).

### 3.0 Assessment Results

### 3.1 Age-Based Analysis

An updated VPA calibration of southern New England yellowtail is summarized in Table D5. This analysis updates the assessment reported in NEFSC (1998) by including 1998 landings and discards. 1998 winter, spring, scallop and fall indices, and 1999 winter and spring indices. Results indicate that F remained low ( $\mathrm{F}_{3-6}=0.20$ ), and biomass slightly increased but remained at low levels in 1998 ( 4.000 mt of spawning biomass and 5.000 mt of mean total biomass: Figures 3 and 4). The number of years.in the retrospective analysis was limited by the short winter survey time series. but results indicate a strong pattern of underestimating F, and overestimating SSB and recruitment. Bootstrap analysis indicates that abundance was estimated with with moderate precision ( $\mathrm{CV}=30-39 \%$ ).

### 3.2 Biomass-Based Analysis

Due to continued poor sampling and resulting problems estimating catch at age, surplus production analyses (ASPIC) were updated to provide alternative perspectives on stock status. Results for the southern New England stock are similar to the VPA for most of the time series: F in 1998 was similarly low, but the estimate of 1998 biomass ( $21 \%$ of $\mathrm{B}_{\mathrm{MSY}}$ ) is substantially greater than the VPA estimate (Figure D4). The estimate of $\mathrm{B}_{\mathrm{MSY}}(62,870 \mathrm{mt})$ was similar to 1998 SARC estimate ( $61,510 \mathrm{mt}$; NEFSC 1998).

### 4.0 Forecasts

Stochastic projections were performed using bootstrap distributions of January 1, 1999 stock abundance at age, assuming $F$ in 1999 was equal to $F$ in 1998, and $F$ in 2000-2009 were from the Amendment 9 control rule targets. Age-1 abundance in 1999 was estimated from multiple survey indices in the terminal year. Recruitment in subsequent years was estimated from Beverton-Holt stock-recruitment relationships estimated from VPA estimates (Overholtz and Brodziak 1999). Mean weight at age, exploitation pattern, and proportion discarded at age was assumed to be equal to the 1994-1998 average.

Projection of the southern New England stock suggests that at status quo F in 1999. the stock will remain less than $25 \% \mathrm{~B}_{\mathrm{MSY}}$ in 1999 (Figure D 4 b ). At the rebuilding targets ( $\mathrm{F}=0$ ), the stock is projected to increase to $\mathrm{B}_{\mathrm{MSY}}$ by 2007.

### 5.0 Sources of Uncertainty

- Estimates of catch at age may not be reliable due to poor sampling intensity. Therefore VPA and age-based projections may be misleading. Retrospective patterns may indicate inadequate sampling and mis-allocation of catch at age.
- Retrospective patterns indicate that VPA estimates of biomass and F may be overly optimistic. Updated VPAs may indicate that 1998 biomass levels are lower, and 1998 F was greater than reported here.
- Although historical perspective from production models are valuable. current biomass levels may not be reliable, because recruitment is implicitly assumed to be a function of stock biomass.
- Stock-recruit observations used to derive the relationships assumed in long-term projections are limited to a short time series of relatively low stock sizes. Therefore. long-term forecasts at relatively high stock sizes may be substantially biased.
- Inappropriate stock delineation may result in underestimated removals (e.g.. from adjacent areas in the mid-Atlantic Bight).
- Medium-term projections may be biased because they implicitly assume constant environmental conditions. Cooling trends offer potential for faster rebuilding.
- Estimates of prorated landings and discard ratios are based on preliminary logbook data and are subject to change.


### 6.0 References

NEFSC (Northeast Fisheries Science Center). 1998. Southern New England yellowtail flounder. NEFSC Ref. Doc. 98-15: 328-350.

Overholtz. W. and J. Brodziak. 1999. Background stock-recruit data for eleven groundfish stocks. NEFSC NDS Working Paper No. 10.

Table D1. Landings of southern New England yellowtail flounder (thousand mt).

| year | US <br> Landings | discards | Industrial landings | Foreign landings | total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1960 | 8.3 | 3.2 | 0.5 |  | 12.0 |
| 1961 | 12.3 | 4.7 | 0.7 |  | 17.7 |
| 1962 | 13.3 | 5.3 | 0.2 | . | 18.8 |
| 1963 | 22.5 | 5.4 | 0.3 | 0.2 | 27.9 |
| 1964 | 19.5 | 9.5 | 0.5 |  | 29.0 |
| 1965 | 21.3 | 6.5 | 1.0 | 1.4 | 27.8 |
| 1966 | 23.3 | 0.3 | 2.7 | 0.7 | 23.6 |
| 1967 | 18.1 | 7.7 | 4.5 | 2.8 | 25.8 |
| 1968 | 21.7 | 6.3 | 3.9 | 3.5 | 28.0 |
| 1969 | 33.2 | 2.4 | 4.2 | 17.6 | 35.6 |
| 1970 | 17.7 | 4.7 | 2.1 | 2.5 | 22.4 |
| 1971 | 8.9 | 3.3 | 0.4 | 0.3 | 12.2 |
| 1972 | 11.5 | 1.7 | 0.3 | 3.0 | 13.2 |
| 1973 | 7.4 | 1.0 | 0.3 | 0.2 | 8.4 |
| 1974 | 6.5 | 8.6 |  | 0.1 | 15.1 |
| 1975 | 3.2 | 1.9 |  |  | 5.1 |
| 1976 | 1.6 | 1.6 |  |  | 3.2 |
| 1977 | 2.8 | 1.9 |  |  | 4.7 |
| 1978 | 2.3 | 5.0 |  |  | 7.3 |
| 1979 | 5.3 | 4.4 |  |  | 9.7 |
| 1980 | 6.0 | 1.7 |  |  | 7.7 |
| 1981 | 4.7 | 1.2 |  |  | 5.9 |
| 1982 | 10.3 | 5.0 |  |  | 15.3 |
| 1983 | 17.0 | 3.5 |  |  | 20.5 |
| 1984 | 7.9 | 1.1 |  |  | 9.0 |
| 1985 | 2.7 | 1.2 |  |  | 3.9 |
| 1986 | 3.3 | 1.1 |  |  | 4.4 |
| 1987 | 1.6 | 0.9 |  |  | 2.5 |
| 1988 | 0.9 | 1.8 |  |  | 2.7 |
| 1989 | 2.5 | 5.5 |  |  | 8.0 |
| 1990 | 8.0 | 9.7 |  |  | 17.7 |
| 1991 | 3.9 | 2.3 |  |  | 6.2 |
| 1992 | 1.4 | 1.1 |  |  | 2.5 |
| 1993 | 0.5 | 0.1 |  |  | 0.6 |
| 1994 | 0.2 | 0.1 |  |  | 0.3 |
| 1995 | 0.2 | 0.1 |  |  | 0.2 |
| 1996 | 0.3 | 0.1 |  |  | 0.4 |
| 1997 | 0.2 | 0.0 |  |  | 0.3 |
| 1998 | 0.4 | 0.1 |  |  | 0.5 |
| average | 8.5 | 3.1 | 1.5 | 2.9 | 11.7 |

Table D2. Landings at age (above) and mean weight at age (below) of southern New England yellowtail flounder.

| Landings at age (thousanas) |  |  |  |  | Age |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | $\bigcirc$ | $8+$ | Total |
| 1973 | 28 | 2570 | 7169 | 4630 | 1716 | 1517 | 257 | 55 | 17942 |
| 1974 | 130 | 1766 | 3922 | 5053 | 2500 | 950 | 1021 | 196 | 15538 |
| 1975 | 170 | 2352 | 1496 | 973 | 1257 | 549 | 308 | 163 | 7268 |
| 1976 | 0 | 1396 | 898 | 245 | 337 | 391 | 167 | 188 | 3622 |
| 1977 | 66 | 2039 - | 3931 | 392 | 205 | 253 | 123 | 160 | 7169 |
| 1978 | 21 | 3209 | 1488 | 1025 | 165 | 34 | 44 | 28 | 6014 |
| 1978 | 19 | 4972 | 8252 | 1033 | 428. | 96 | 24 | 0 | 14824 |
| 1980 | 119 | 4557 | 6324 | 3619 | 472 | 117 | 19 | 12 | 15239 |
| 1981 | 0 | 2732 | 6418 | 2449 | 884 | 128 | 14 | 0 | 12625 |
| 1982 | 56 | 17414 | 12788 | 1741 | 404 | 78 | 7 | 0 | 32488 |
| 1983 | 57 | 13823 | 33242 | 3347 | 376 | 129 | 35 | 7 | 51016 |
| 1984 | 45 | 2624 | 13902 | 6587 | 740 | 244 | 7 | 14 | 24163 |
| 1985 | 166 | 3984 | 1496 | 1312 | 774 | . 135 | 27 | 4 | 7898 |
| 1986 | 39 | 5926 | 2882 | 561 | 324 | 119 | 21 | 1 | 9873 |
| 1987 | 72 | 1370 | 2014 | 803 | 139 | 47 | 8 | 1 | 4454 |
| 1988 | 0 | 1154 | 504 | 407 | 101 | 17 | 6 | 0 | 2189 |
| 1989 | 0 | 5213 | 1269 | 280 | 41 | 3 | 0 | 0 | 6806 |
| 1990 | 0 | 415 | 18476 | 1352 | 68 | 5 | 0 | 0 | 20316 |
| 1991 | 0 | 253 | 2230 | 6606 | 81 | 1 | 17 | 0 | 9188 |
| 1992 | 0 | 301 | 896 | 1687 | 246 | 10 | 3 | 0 | 3143 |
| 1993 | 0 | 211 | 361 | 417 | 124 | 4 | 0 | 0 | 1117 |
| 1994 | 0 | 15 | 187 | 136 | 120 | 48 | 1 | 0 | 507 |
| 1995 | 0 | 154 | 125 | 182 | 18 | 1 | 3 | 0 | 483 |
| 1996 | 0 | 224 | 439 | 122 | 15 | 10 | 5 | 1 | 817 |
| 1997 | 0 | 33 | 319 | 146 | 14 | 2 | 2 | 1 | 518 |
| 1998 | 0 | 300 | 364 | 139 | 25 | 2 | 0 | 0 | 830 |
| mean | 38 | 3039 | 5054 | 1740 | 445 | 188 | 82 | 32 | 10617 |



Table D3. Discards at age of southern New England yellowtail flounder.

| Discards at age (thousands) |  |  |  | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ | Total |
| 1973 | 160 | 2486 | 1130 | 43 | 0 | 0 | 0 | 0 | 3819 |
| 1974 | 728 | 26568 | 793 | 45 | 0 | 0 | 0 | 0 | 28134 |
| 1975 | 8670 | 1427 | 1 | 10 | 0 | 0 | 0 | 0 | 10108 |
| 1976 | 214 | 5203 | 14 | 0 | 0 | 0 | 0 | 0 | 5431 |
| 1977 | 5376 | 2732 | 42 | 0 | 0 | 0 | 0 | 0 | 8150 |
| 1978 | 8677 | 10102 | 7 | 0 | 0 | 0 | 0 | 0 | 18786 |
| 1979 | 185 | 14253 | 119 | 0 | 0 | 0 | 0 | 0 | 14557 |
| 1980 | 869 | 5441 | 18 | 0 | 0 | 0 | 0 | 0 | 6328 |
| 1981 | 38 | 4013 | 319 | 0 | 0 | 0 | 0 | 0 | 4370 |
| 1982 | 113 | 17716 | 905 | 3 | 0 | 0 | 0 | 0 | 18737 |
| 1983 | 2469 | 4607 | 5373 | 17 | 0 | 0 | 0 | 0 | 12466 |
| 1984 | 465 | 3107 | 941 | 74 | 0 | 0 | 0 | 0 | 4587 |
| 1985 | 2064 | 3031 | 20 | 0 | 0 | 0 | 0 | 0 | 5115 |
| 1986 | 423 | 3754 | 39 | 0 | 0 | 0 | 0 | 0 | 4216 |
| 1987 | 1518 | 2034 | 19 | 0 | 0 | 0 | 0 | 0 | 3572 |
| 1988 | 5899 | 896 | 4 | 0 | 0 | 0 | 0 | 0 | 6798 |
| 1989 | 24 | 14002 | 1834 | 131 | 6 | 0. | 0 | 0 | 15996 |
| 1990 | 192 | 1633 | 23709 | 673 | 11 | 0 | 0 | 0 | 26217 |
| 1991 | 445 | 1354 | 2820 | 2883 | 12 | 0 | 0 | 0 | 7514 |
| 1992 | 477 | 1152 | 1086 | 659 | 33 | 0 | 0 | 0 | 3408 |
| 1993 | 13 | 212 | 15 | 9 | 0 | 0 | 0 | 0 | 249 |
| 1994 | 9 | 134 | 35 | 29 | 12 | 2 | 0 | 0 | 221 |
| 1995 | 7 | 94 | 38 | 27 | 12 | 3 | 0 | 0 | 182 |
| 1996 | 21 | 81 | 56 | 29 | 13 | 2 | 0 | 0 | 202 |
| 1997 | 1 | 23 | 32 | 4 | 1 | 0 | 0 | 0 | 61 |
| 1998 | 0 | 88 | 114 | 40 | 9 | 3 | 1 | 0 | 255 |
| mean | 1502 | - 4852 | 1519 | 180 | 4 | 0 | 0 | 0 | 8057 |

Table D4a. Survey indices of southern New England yellowtail abundance and biomass.

| NEFSC Fall Survey |  |  |  |  | Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8+ | Total | kg/tow |
| 1963 | 19.798 | 20.168 | 14.960 | 5.830 | 0.660 | 0.151 | 0.000 | 0.100 | 61.667 | 16.842 |
| 1964 | 22.529 | 31.952 | 5.861 | 8.701 | 3.983 | 1.108 | 0.000 | 0.000 | 74.133 | 19.03 |
| 1965 | 1.3 .231 | 21.390 | 7.771 | 2.140 | 2.167 | 0.155 | 0.000 | 0.090 | 46.944 | 12.675 |
| 1966 | 43.305 | 13.066 | 2.375 | 1.247 | 0.231 | 0.000 | 0.000 | 0.000 | 60.224 | 9.431 |
| 1967 | 22.497 | 31.159 | 13.716 | 1.936 | 0.472 | 0.079 | . 0.160 | 0.000 | 70.019 | 14.057 |
| 1968 | 11.285 | 13.352 | 22.860 | 1.443 | 0.115 | 0.000 | 0.000 | 0.000 | 49.055 | 10.062 |
| 1969 | 14.481 | 11.884 | 33.861 | 6.351 | 0.113 | 0.050 | 0.050 | 0.000 | 66.791 | 14.401 |
| 1970 | 5.157 | 6.736 | 19.936 | 12.961 | 3.067 | 0.520 | 0.089 | 0.000 | 48.466 | 10.965 |
| 1971 | 7.748 | 13.298 | 7.618 | 18.468 | 3.287 | 0.264 | 0.196 | 0.000 | 50.879 | 11.632 |
| 1972 | 5.135 | 20.125 | 24.054 | 22.993 | 14.991 | 2.050 | 0.054 | 0.000 | 89.402 | 20.114 |
| 1973 | 1.726 | 1.590 | 2.224 | 1.640 | 1.241 | 1.057 | 0.212 | 0.000 | 9.689 | 2.264 |
| 1974 | 1.216 | 2.047 | 0.676 | 2.776 | 1.166 | 0.489 | 0.238 | 0.093 | 8.701 | 2.141 |
| 1975 | 1.981 | 0.516 | 0.266 | 0.329 | 0.334 | 0.000 | 0.104 | 0.000 | 3.531 | 0.715 |
| 1976 | 3.632 | 7.331 | 0.877 | 0.088 | 0.139 | 0.361 | 0.423 | 0.189 | 13.041 | 2.962 |
| 1977 | 1.759 | 2.275 | 0.828 | 0.053 | 0:046 | 0.113 | 0.078 | 0.000 | 5.151 | 1.501 |
| 1978 | 3.247 | 7.599 | 0.450 | 0.392 | 0.043 | 0.009 | 0.079 | 0.032 | 11.851 | 3.057 |
| 1979 | 1.794 | 4.533 | 2.537 | 0.388 | 0.043 | 0.041 | 0.000 | 0.000 | 9.335 | 2.565 |
| 1980 | 1.463 | 4.506 | 1.202 | 0.426 | 0.000 | 0.000 | 0.000 | 0.000 | 7.597 | 1.957 |
| 1981 | 4.704 | 8.944 | 1.404 | 0.334 | 0.080 | 0.061 | 0.000 | 0.000 | 15.527 | 3.789 |
| 1982 | 2.610 | 29.372 | 8.673 | 1.025 | 0.409 | 0.000 | 0.000 | 0.000 | 42.088 | 8.126 |
| 1983 | 4.582 | 17.956 | 10.078 | 0.876 | 0.073 | 0.000 | 0.050 | 0.000 | 33.616 | 6.515 |
| 1984 | 0.719 | 2.217 | 2.400 | 0.659 | 0.000 | 0.000 | 0.000 | 0.000 | 5.994 | 1.365 |
| 1985 | 1.018 | 0.447 | 0.161 | 0.122 | 0.000 | 0.000 | 0.000 | 0.000 | 1.748 | 0.438 |
| 1986 | 0.826 | 1.685 | 0.365 | 0.088 | 0.000 | 0.000 | 0.000 | 0.000 | 2.963 | 0.883 |
| 1987 | 1.515 | 0.674 | 0.558 | 0.047 | 0.037 | 0.000 | 0.037 | 0.000 | 2.868 | 0.607 |
| 1988 | 1.261 | 0.388 | 0.173 | 0.195 | 0.048 | 0.000 | 0.000 | 0.000 | 2.065 | 0.496 |
| 1989 | 0.000 | 8.004 | 1.400 | 0.065 | 0.000 | 0.000 | 0.000 | 0.000 | 9.469 | 2.359 |
| 1990 | 0.000 | 0.097 | 2.395 | 0.270 | 0.000 | 0.000 | 0.000 | 0.000 | 2.763 | 0.974 |
| 1991 | 0.865 | 0.219 | 1.709 | 0.453 | 0.000 | 0.000 | 0.000 | 0.000 | 3.247 | 1.013 |
| 1992 | 0.261 | 0.062 | 0.180 | 0.337 | 0.012 | 0.000 | 0.000 | 0.000 | 0.852 | 0.229 |
| 1993 | 0.070 | 0.015 | 0.028 | 0.020 | 0.000 | 0.000 | 0.000 | 0.000 | 0.133 | 0.053 |
| 1994 | 0.754 | 0.553 | 0.198 | 0.192 | 0.085 | 0.011 | 0.000 | 0.000 | 1.793 | 0.374 |
| 1995 | 0.180 | 1.306 | 0.171 | 0.095 | 0.000 | 0.000 | 0.000 | 0.000 | 1.752 | 0.432 |
| 1996 | 0.653 | 0.290 | 0.258 | 0.025 | 0.000 | 0.000 | 0.000 | 0.000 | 1.226 | 0.266 |
| 1997 | 0.889 | 0.716 | 1.687 | 0.373 | 0.037 | 0.000 | 0.000 | 0.000 | 3.702 | 1.041 |
| 1998 | 1.689 | 2.611 | 0.229 | 0.093 | 0.000 | 0.044 | 0.000 | 0.000 | 4.666 | 1.151 |
| mean | 5.683 | 8.030 | 5.393 | 2.595 | 0.913 | 0.182 | 0.049 | 0.014 | 22.860 | 5.180 |

Table D4b. Survey indices of southern New England yellowtail abundance and biomass.


Table D4c. Survey indices of southern New England yellowtail abundance and biomass.

| NEF | er Sur |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ | Total | kg/tow |
| 1992 | 0.000 | 2.884 | 1.881 | 6.418 | 1.295 | 0.000 | 0.000 | 0.000 | 12.478 | 4.402 |
| 1993 | 1.349 | 3.853 | 0.711 | 1.841 | 0.306 | 0.000 | 0.000 | 0.000 | 8.060 | 1.968 |
| 1994 | 0.586 | 17.778 | 1.363 | 2.917 | 1.258 | 0.199 | 0.000 | 0.000 | 24.101 | 6.809 |
| 1995 | 0.368 | 7.615 | 4.474 | 1.317 | 0.493 | 0.123 | 0.036 | 0.000 | 14.426 | 4.059 |
| 1996 | 0.092 | 2.304 | 11.703 | 1.552 | 0.207 | 0.109 | 0.033 | 0.000 | 16.000 | 5.159 |
| 1997 | 0.301 | 3.976 | 9.141 | 2.625 | 0.508 | 0.000 | 0.000 | 0.000 | 16.551 | 5.831 |
| 1998 | 0:326 | 3.855 | 1.476 | 0.445 | 0.000 | 0.000 | 0.050 | 0.000 | 6.152 | 1.640 |
| 1999* | 0.567 | 6.112 | 18.823 | 0.918 | 0.184 | 0.081 | 0.000 | 0.000 | 26.686 | 8.987 |
| mean | 0.449 | 6.047 | 6.197 | 2.254 | 0.531 | 0.064 | 0.015 | 0.000 | 15.557 | 5.4 |

* preliminary, based on unaudited data.


## Scallop Survey

| Year | age-1 |
| :---: | :---: |
| 1982 | 0.584 |
| 1983 | 0.891 |
| 1984 | 0.205 |
| 1985 | 0.647 |
| 1986 | 0.282 |
| 1987 | 0.601 |
| 1988 | 1.343 |
| 1989 | 0.169 |
| 1990 | 0.026 |
| 1991 | 1.060 |
| 1992 | 0.411 |
| 1993 | 0.419 |
| 1994 | 1.265 |
| 1995 | 0.551 |
| 1996 | 0.608 |
| 1997 | 2.744 |
| 1998 | 0.289 |
| mean | 0.711 |

Table D5a. Estimates of abundance at age of southern New England yellowtail flounder.


Table D5b. Estimates of fishing mortality at age of southern New England yellowtail flounder.

|  | 1973 | 1974 | 1975 | 1976 | 1977 | 1978 | 1979 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 0.005 | 0.108 | 0.413 | 0.018 | 0.135 | 0.203 | 0.008 |
| 2 | 0.457 | 2.431 | 0.956 | 0.628 | 0.710 | 0.566 | 0.532 |
| 3 | 0.619 | 1.079 | 1.106 | 0.639 | 1.030 | 0.503 | 0.880 |
| 4 | c.71E | 1.030 | 0.682 | 0.517 | 0.634 | 0.838 | 0.803 |
| 5 | 0.68 c | 1.151 | 0.781 | 0.527 | 1.180 | 0.608 | 1. 105 |
| E | 0.666 | $=.109$ | 0.867 | 0.597 | 1.011 | 0.610 | 0.902 |
| $-$ | 0.666 | 2. 109 | 0.867 | 0.597 | 1.011 | 0.610 | 0.902 |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| - | 0.026 | 0.000 | 0.004 | 0.212 | 0.034 | 0.133 | 0.076 |
| 2 | 0.601 | 0.252 | 0.468 | 0.634 | 1.067 | 0.881 | 1. 395 |
| 3 | 0.968 | 1.135. | 1.242 | 1.615 | 2.051 | 0.958 | 1.270 |
| 4 | 1. 365 | 1.471. | 1.101 | 1.342 | 1.891 | 1.316 | 1.294 |
| 5 | 1.163 | 2.039 | 1.127 | 0.752 | 1.426 | 1.602 | 1. 730 |
| 6 | 1.122 | 2. 305 | 1.274 | 1.680 | 2.195 | . 1.218 | 1. 365 |
| 7 | 1.122 | 1. 305 | 1.274 | 1.680 | 2.195 | 1.218 | 1.365 |
|  | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| 1 | 0.134 | 0.055 | 0.002 | 0.031 | 0.149 | 0.309 | 0.016 |
| 2 | 1.243 | 0.257 | 0.254 | 0.185 | 0.395 | 1.027 | 0.499 |
| 3 | 1.507 | 0.597 | 0.778 | 1.500 | 0.946 | 1.312 | 0.836 |
| 4 | 1.963 | 1.961 | 1.645 | 2.784 | 3.127 | 2.245 | 1.246 |
| 5 | 1.618 | 2.822 | 1.972 | 3.992 | 1.911 | 1.529 | 0.787 |
| 6 | 1.726 | 0.925 | 0.853 | 1.629 | 1.726 | 1.785 | 1.002 |
| 7 | 1.726 | 0.925 | 0.853 | 1.629 | 1.726 | 1.785 | 1.002 |
|  | 1994 | 1995 | 1996 | 1997 | 1998 |  |  |
| 1 | 0.005 | 0.003 | 0.005 | 0.000 | 0.000 |  |  |
| 2 | 0.25 ? | 0.193 | 0.149 | 0.018 | 0.044 |  |  |
| 三 | 0.535 | 0.492 | 0.734 | 0.256 | 0.207 |  |  |
| 4 | $\therefore .205$ | 2.703 | 1.273 | 0.511 | 0.201 |  |  |
| E | 2.776 | 0.732 | 1.416 | 0.374 | 0.204 |  |  |
| $\bar{\varepsilon}$ | 0.863 | 0.815 | 0.846 | 0.305 | 0.204 | - |  |
| - | 0.863 | 0.815 | 0.846 | 0.305 | 0.204 |  |  |
| Age3-E | 1973 | 1974 | 1575 | 1976 | 1977 | 1978 | 1979 |
| 3,6 | 0.67 | 2. 09 | 0.86 | 0.57 | 0.96 | 0.64 | 0.92 |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| $\bar{j}$, | 2. 15 | i. 49 | 1.19 | 1. 35 | 1.89 | 1.27 | 1.41 |
|  | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| 3,6 | 2.70 | 1.58 | 1.31 | 2.48 | 1.93 | 1.72 | 0.97 |
|  | 1994 | 1995 | 1996 | 1997 | 1998 |  |  |
| 3,6 | 1.34 | 0.94 | 1.07 | 0.36 | 0.20 |  |  |
| Biomass$1973$ | $\begin{gathered} \text { Weighted F } \\ 1974 \end{gathered}$ | 1975 | 1976 | 1977 | 1978 | 1979 |  |
|  | 0.39 | 1.39 | 0.62 | 0.41 | 0.37 | 0.38 | 0.62 |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| --------- | $0.4{ }^{-}$ | 0.23 | 0.43 | 1.06 | 1.16 | 0.60 | 1.03 |
|  | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
|  | 0.57 | 0.08 | 0.25 | 1. 08 | 1.32 | 1.35 | 0.71 |
|  | 1994 | 1995 | 1996 | 1997 | 1998 |  |  |
|  | 0.48 | 0.28 | 0.25 | 0.08 | 0.09 |  |  |

Table D5c. Estimates of mean biomass (mt) of southern New England yellowtail flounder.


Table D5d. Estimates of spawning stock biomass of southern New England yellowtail flounder. SSE A- THE START OE THE SPAWNING SEASON -MALES AND FEMALES (MA) (using SSB mean weights

|  | 1973 | 1976 | 1975 | 1976 | 1977 | 1978 | 1979 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| I | 1056 | 214 | 632 | 349 | 1155 | 1347 | 678 |
| - | 2548 | 2592 | 896 | 2476 | $1488^{\circ}$ | 5402 | 4853 |
| $\vdots$ | 5262 | 1623 | 539 | 628 | 1537 | 1225 | 3604 |
| 4 | 2887 | 2243 | 666 | 262 | 339 | 701 | 64 E |
| 5 | 1128 | 1093 | 740 | 357 | 97 | 203 | 241 |
| $\varepsilon$ | 1212 | 430 | 313 | 391 | 128 | 46 | 77 |
| - | $29 ?$ | 588 | 291 | 372 | 178 | 83 | 19 |
| 1+ | 14390 | 8784 | 4077 | 4834 | 4922 | 9007 | 10117 |
|  | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| 1 | 1021 | 2124 | 1434 | 279 | 359 | 410 | 150 |
| 2 | 3631 | 5363 | 15272 | 5921 | 1003 | 1642 | 1535 |
| 3 | 2603 | 2105 | 4029 | 8304 | 2136 | 606 | 795 |
| 4 | 1389 | 843 | 839 | 1314 | 1398 | 449 | 214 |
| 5 | 295 | 271 | 250 | 354 | 270 | 272 | 116 |
| 6 | 98 | 69 | 53 | 66 | 72 | . 72 | 56 |
| 7 | 34 | 05 | 05 | 21 | 07 | 23 | 13 |
| $1+$ | 9073 | 10780 | 21882 | 16259 | 5244 | 3474 | 2880 |
|  | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
| 1 | 390 | 3847 | 609 | 243 | 82 | 35 | 13 |
| 2 | 572 | 1792 | 19481 | 2760 | 821 | 349 | 235 |
| 3 | 499 | 349 | 1603 | 10889 | 1857 | 571 | 200 |
| 4 | 171 | 106 | 142 | 314 | 1123 | 447 | 170 |
| 5 | 48 | 22 | 18 | 12 | 34 | 114 | 107 |
| E | 17 | 18 | 04 | 03 | 01 | 05 | 44 |
| - | 05 | 06 | 00 | 00 | 06 | 02 | 00 |
| - | 1702 | 6140 | 21856 | 14221 | 3922 | 1523 | 769 |
|  | 1994 | 1995 | 1996 | 1997 | 1998 |  |  |
| I | 25 | 44 | 76 | 208 | 119 |  |  |
| $=$ | 143 | 311 | 581 | 700 | 1886 |  |  |
| 亏 | 149 | 139 | 286 | 590 | 932 |  |  |
| 4 | 60 | 59 | 58 | 152 | 485 |  |  |
| 5 | 25 | 29 | 13 | 28 | 119 |  |  |
| $\varepsilon$ | 43 | 05 | 11 | 06 | 20 |  |  |
| 7 | 01 | 03 | 06 | 09 | 04 |  |  |
| $1+$ | 445 | 589 | 1030 | 1694 | 3564 |  |  |

Table D6. Summary of results from stochastic projection of southern New England yellowtail flounder.

| INPUT AS Aaf | I | 2 | 3 | 4 | 5 | 6 | $7+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stock Wt. | 0.130 | 0.318 | 0.398 | 0.473 | 0.636 | 0.785 | 0.850 |
| Landed $W$ t | 0.254 | 0.326 | 0.398 | 0.475 | 0.639 | 0.783 | 0.830 |
| Discard wt. | 0.13 | 0.28 | 0.4 | 0.53 | 0.69 | 0.79 | 0.71 |
| Maturity | 0.13 | 0.74 | 0.98 | 1.00 | 1.00 | 1.00 | 1.00 |
| PF. | 0.01 | 0.12 | 0.53 | 1.00 | 1.00 | 1.00 | 1.00 |
| - Discara | 1.00 | 0.44 | 0.17 | 0.15 | 0.25 | 0.31 | 0.20 |


| E-BESED | PROUECTIONS |
| :--- | :--- |
| TIME-VARYING F |  |
| YEAF | $F$ |
| $29 E G$ | 0.200 |
| 2000 | 0.000 |
| 2001 | 0.000 |
| 2002 | 0.000 |
| 2003 | 0.000 |
| 2004 | 0.000 |
| 2005 | 0.000 |
| 2006 | 0.000 |
| 2007 | 0.000 |
| 2008 | 0.000 |
| 2009 | 0.000 |


| PERCE | ES OF | NG STO | BIOMASS | (000 MT) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | 1. | $5 \frac{5}{\text { ¢ }}$ | 105 | $255^{\circ}$ | 50\% | 75 를 | 90\% | 95 | 99 |
| 1999 | 3.584 | 4.128 | 4.317 | 4.838 | 5.510 | 6.230 | 7.061 | 7.585 | 8.635 |
| 2000 | 3.923 | 4.662 | 4.975 | 5.639 | 6.610 | 7.960 | 9.735 | 11.676 | 16.257 |
| 2001 | 5.245 | 6.184 | 6.893 | 8.159 | 10.526 | 15.002 | 21.196 | 27.714 | 38.847 |
| 2002 | -. 056 | 8.505 | 9.618 | 12.043 | 16.900 | 24.456 | 35.818 | 44.906 | $67.18{ }^{7}$ |
| 2003 | 8.829 | 12. 070 | 12.933 | 16.978 | 24.059 | 34.988 | 49.624 | 61.234 | 95.215 |
| 2004. | -1.174 | 14.198 | 16.743 | 22.701 | 32.157 | 46.327 | 66.811 | 83.769 | 126.580 |
| 2005 | 23.411 | 17.980 | 21.299 | 28.737 | 41.066 | 58.711 | 83.508 | 103.297 | 154.428 |
| 2006 | -5.895 | 21.517 | 25.744 | 34.978 | 48.925 | 70.824 | 9.9 .667 | 125.004 | 180.972 |
| 2007 | 18.701 | 25.369 | 30.290 | 41.663 | 57.758 | 82.208 | 115.411 | 140.787 | 215.165 |
| 2008 | 21.630 | 30.055 | 34.996 | 47.074 | 65.816 | 92.485 | 129.031 | 154.470 | 239.580 |
| 2009 | 24.040 | 33.758 | 39.435 | 52.216 | 71.552 | 100.648 | 138.853 | 168.715 | 262.813 |
| PERCE | Les OE M | STOCK | ASS 1000 | MT) |  |  |  |  |  |
| YEAR | 2 | $5 \frac{5}{\text { \% }}$ | 10ミ | $25 \%$ | $50 \div$ | 75 둘 | $90^{\circ}$ | 95 | 99. |
| 1.95 | 3.802 | 4.556 | 4.934 | 5.591 | 5.607 | 7.718 | 9.865 | 12.760 | 17.970 |
| 2000 | 4.990 | 5.953 | 6.612 | 7.991 | 10.688 | 15.951 | 24.282 | 32.408 | 47.721 |
| 2001 | 6.972 | 8.398 | 9.456 | 12.139 | 17.418 | 26.073 | 38.117 | 47.970 | 75.616 |
| 2002 | 8.660 | 11.636 | 13.288 | 17.710 | 25.277 | 36.841 | 53.125 | 66.824 | 100.042 |
| 2003 | 11.490 | 15.047 | 17.615 | 23.623 | 33.574 | 48.823 | 69.857 | 88.168 | 139.503 |
| 2004 | 14.088 | 18.501 | 21.783 | 29.495 | 42.483 | 60.850 | 86.202 | 106.594 | 163.487 |
| 2005 | 16.296 | 22.407 | 26.687 | 36.625 | 50.713 | 72.936 | 102.197 | 131.214 | 196.289 |
| 2006 | 19.483 | 26.657 | 31.660 | 43.571 | 60.286 | 85.241 | 119.191 | 145.238 | 233.720 |
| 2007 | $22.377^{\circ}$ | 31.514 | 37.153 | 49.986 | 69.502 | 96.973 | 135.045 | 160.800 | 252.330 |
| 2008 | 25.556 | 35.923 | 42.068 | 55.371 | 75.824 | 107.320 | 148.007 | 175.375 | 266.105 |
| 2009 | 28.712 | 39.519 | 46.304 | 60.512 | 82.156 | 115.251 | 153.739 | 190.201 | 301.637 |




Figure D1. Total catch of southern New England yellowtail flounder.


Figure D2. Survey indices of southern New England yellowtail flounder biomass.


Figure D3. Summary of southern New England yellowtail VPA results.


Figure D4. Status of the southern New England yellowtail flounder stock.


Figure D5. Projection ofsouthern New England yellowtail flounder biomass at the Amendment \#9 long-term fishing target in 2000-2009.

## E. Cape Cod Yellowtail Flounder by S.X. Cadrin

### 1.0 Background

The Cape Cod yellowtail flounder stock was at a medium biomass ( $44 \%$ of $\mathrm{B}_{\mathrm{MSY}}$ ) and was overexploited (fully recruited F was 1.01) in 1997 (Cadrin et al. 1999). This report updates catch and survey indices. updates estimates of stock size and fishing mortality and evaluates mediumterm projections.

### 2.0 1999 Assessment

### 2.1 1998 Landings

U.S. landings were prorated as described in Cadrin et al. (1999; Table E1: Figure E1). Landings from the Cape Cod stock increased $12 \%$, and Mid-Atlantic landings decreased by $60 \%$.

Sampling intensity of landings in 1998 was poor. The large market category from the Cape Cod grounds. which comprised nearly half of the stock's landings, was entirely unsampled. Landings at length were estimated by half year categorized as unclassified. Landings at age and mean weights at age are reported in Table E2.

### 2.2 1998 Discards

Discarded catch was estimated from logbook information on discard to kept ratios by half-year and gear (NEFSC 1998; Table E4). However, discards of Cape Cod yellowtail are substantially less than those estimated in recent years, presumably because previous estimates were based on observer data by fishery. Therefore, the level of discards for Cape Cod yellowtail may be underestimated and should be considered preliminary. Discards at age were estimated from sea sampled lengths and pooled commercial-survey age-length keys. Discards at age and recent mean weights at age are reported in Table E3.

### 2.3 1998-1999 Survey Indices

Survey abundance and biomass indices are reported in Table E4. Estimates are from valid tows on the Cape Cod grounds (offshore strata 25. 26; inshore strata 56-66; Massachusetts strata 1736) standardized according to net, vessel, and door changes (NEFSC 1998). Relative change in survey indices of Cape Cod yellowtail were equivocal (Figure E2).

### 3.0 Assessment Results

### 3.1 Age-Based Analysis

An updated VPA calibration of Cape Cod yellowtail is summarized in Table E5. This analysis updates the assessment reported in Cadrin and King (1999) by including 1998 landings and provisional discards, 1998 fall indices, and 1999 spring indices (the provisional Massachusetts age-1 index is based on observed lengths). Results indicate that $F$ decreased ( $F_{4.5}=0.41$ ) and biomass increased in 1998 ( $2,000 \mathrm{mt}$ of spawning biomass and $3,000 \mathrm{mt}$ of mean total biomass: Figures 3.and 4). Retrospective analysis indicates a tendency toward underestimating $F$ in the most recent years. but the pattern does not persist further back in time. Bootstrap analysis indicates that abundance was estimated with with moderate precision ( $\mathrm{CV}=31-40 \%$ ).

### 4.0 Forecasts

Stochastic projections were performed using bootstrap distributions of January 1. 1999 stock abundance at age, assuming $F$ in 1999 was equal to $F$ in 1998, and $F$ in 2000-2009 were from the Amendment 9 control rule targets. Age-1 abundance in 1999 was estimated from multiple survey indices in the terminal year. Recruitment in subsequent years was estimated from Beverton-Holt stock-recruitment relationships estimated from VPA estimates (Overholtz and Brodziak 1999). Mean weight at age, exploitation pattern, and proportion discard at age was assumed to be equal to the 1994-1998 average.

Projection of the Cape Cod stock suggests that at status quo F in 1999, the stock increases to greater than $25 \% \mathrm{~B}_{\mathrm{MSY}}$ (Table E6; Figure E5). At the rebuilding target. the stock is projected to grow to greater than $B_{\text {MSY }}$ within ten years (greater than $6,100 \mathrm{mt}$ total biomass after 2001).

### 5.0 Sources of Uncertainty

- Estimates of catch at age may not be reliable due to poor sampling intensity. Therefore VPA and age-based projections may be misleading. Extreme estimates of mean weights (e.g. ages 2-3), odd exploitation patterns. and retrospective patterns may indicate inadequate sampling and mis-allocation of catch at age.
- Retrospective patterns indicate that VPA estimates of biomass and F may be overly optimistic. Updated VPAs may indicate that 1998 biomass levels are lower, and 1998 F was greater than reported here.
- Stock-recruit observations used to derive the relationships assumed in long-term projections are limited to a short time series of relatively low stock sizes. Therefore, long-term forecasts at relatively high stock sizes may be substantially biased.
- Estimates of prorated landings and discard ratios are based on preliminary logbook data and are subject to change.
- The magnitude of discards in 1998 are probably underestimated.


### 6.0 References

Cadrin, S.. J. King. and L. Suslowicz. 1999. Status of the Cape Cod yellowtail flounder stock for 1998. NEFSC Ref. Doc. 99-04.

Overholtz. W. and J. Brodziak. 1999. Background stock-recruit data for eleven groundfish stocks. NEFSC NDS Working Paper No. 10.

Table E1. Landings of Cape Cod yellowtail flounder (mt).

|  |  | Landings <br> (mt) | Discards <br> (mt) | Percent Discard | Total $(\mathrm{mt})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1960 | 1,500 | 500 | 32 | 2,000 |
|  | 1961 | 1,800 | 600 | 32 | 2,400 |
|  | 1962 | 1,900 | 600 | 32 | 2,500 |
|  | 1963 | 3,600 | 1.000 | 28 | 4.600 |
|  | 1964 | 1,851 | . 600 | 32 | 2.451 |
|  | 1965 | 1,498 | 500 | 33 | 1.998 |
|  | 1966 | 1,808 | 300 | 17 | 2,108 |
|  | 1967 | 1.542 | 800 | 52 | 2,342 |
|  | 1968 | 1,569 | 600 | 38 | 2,169 |
|  | 1969 | 1,346 | 300 | 22 | 1,646 |
|  | 1970 | 1,185 | 400 | 34 | 1,585 |
|  | 1971 | 1,662 | 700 | 42 | 2,362 |
|  | 1972 | 1,364 | 300 | 22 | 1,664 |
|  | 1973 | 1,662 | 0 | 0 | 1.662 |
|  | 1974 | 2,054 | 200 | 10 | 2,254 |
|  | 1975 | 2,027 | 0 | 0 | 2,027 |
|  | 1976 | 3,587 | 100 | 3 | 3.687 |
|  | 1977 | 3,469 | 0 | 0 | 3,469 |
|  | 1978 | 3,683 | 400 | 11 | 4,083 |
|  | 1979 | 4,163 | 500 | 12 | 4,663 |
|  | 1980 | 5,106 | 600 | 12 | 5,706 |
|  | 1981 | 3,149 | 600 | 19 | 3,749 |
|  | 1982 | 3,150 | 400 | 13 | 3,550 |
|  | 1983 | 1,884 | 300 | 16 | 2,184 |
|  | 1984 | 1,121 | 20 | 2 | 1,141 |
|  | 1985 | 967 | 77 | 8 | 1,044 |
|  | 1986 | 1,041 | 305 | 29 | 1,346 |
|  | 1987 | 1,159 | 198 | 17 | 1,357 |
|  | 1988 | 1,085 | 283 | 26 | 1,368 |
|  | 1989 | 909 | 390 | 43 | 1,299 |
|  | 1990 | 2,984 | 1,141 | 38 | 4.125 |
|  | 1991 | 1,472 | 405 | 28 | 1,877 |
|  | 1992 | 828 | 637 | 77 | 1,465 |
|  | . 1993 | 628 | 90 | 14 | 718 |
|  | 1994 | 978 | 192 | 20 | 1.170 |
|  | 1995 | 1,207 | 233 | 19 | 1,440 |
|  | 1996 | 1;064 | 182 | 17 | 1,246 |
|  | 1997 | 1.040 | 257 | 25 | 1.297 |
|  | 1998 | 1.169 | 75 | 6 | 1,244 |
| mean |  | 1,903 | 379 | 23 | 2,284 |

Table E2. Landings at age (above) and mean weight at age (below) of Cape Cod yellowtail flounder.

| Landings at age (thousands) |  |  |  | age |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5. | 6 | 7 | $8+$ |
| 1985 | 5 | 738 | 700 | 522 | 268 | 89 | 3 | 7 |
| 1986 | 0 | 1,998 | 579 | 223 | 32 | 6 | 0 | 1 |
| 1987 | 0 | 609. | 1,786 | 268 | 100 | 29 | 12 | 5 |
| 1988 | 1 | 802 | 1,043 | 625 | 172 | 36 | 0 | 0 |
| 1989 | 0 | 726. | 989 | 231 | 31 | 3 | 2 | 2 |
| 1990 | 0 | 692 | 6,191 | 416 | 32 | 16 | 7 | 3 |
| 1991 | 0 | 311 | 903 | 1,455 | 249 | 33 | 27 | 1 |
| 1992 | 0 | 338 | 807 | 514 | 150 | 6 | 5 | 1 |
| 1993 | 0 | 25 | 684 | 573 | 90 | 24 | 15 | 7 |
| 1994 | 0 | 87 | 1,023 | 650 | 236 | 65 | 38 | 9 |
| 1995 | 0 | 233 | 1,730 | 808 | 152 | 78 | 5 | 0 |
| 1996 | 0 | 150 | 1,097 | 798 | 287 | 11 | 5 | 2 |
| 1997 | 0 | 481 | 1,086 | 702 | 160 | 13 | 0 | 1 |
| 1998 | 0 | 257 | 1,681 | 472 | 141 | 41 | 3 | 0 |
| mean | 0 | 532 | 1,450 | 590 | 150 | 32 | 9 | 3 |


| Landed weight at age $(\mathrm{kg})$ |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ |
| 1985 | 0.19 | 0.32 | 0.37 | 0.49 | 0.60 | 0.73 | 1.20 | 1.39 |
| 1986 | --- | 0.32 | 0.46 | 0.57 | 0.73 | 0.90 | --- | 1.40 |
| 1987 | --- | 0.31 | 0.42 | 0.55 | 0.65 | 0.81 | 1.03 | 1.18 |
| 1988 | 0.11 | 0.31 | 0.37 | 0.53 | 0.70 | 0.85 | --- | --- |
| 1989 | --- | 0.38 | 0.45 | 0.65 | 0.92 | 1.41 | 1.24 | 1.24 |
| 1990 | --- | 0.31 | 0.41 | 0.56 | 0.82 | 0.90 | 0.99 | 1.17 |
| 1991 | --- | 0.35 | 0.39 | 0.54 | 0.74 | 0.99 | 1.06 | 1.01 |
| 1992 | ---- | 0.32 | 0.41 | 0.53 | 0.61 | 0.73 | 1.53 | 1.91 |
| 1993 | ---- | 0.31 | 0.38 | 0.43 | 0.74 | 0.95 | 1.01 | 1.17 |
| 1994 | --- | 0.29 | 0.38 | 0.50 | 0.62 | 0.68 | 1.04 | 1.11 |
| 1995 | ---- | 0.35 | 0.36 | 0.43 | 0.61 | 0.78 | 1.11 | --- |
| 1996 | ---- | 0.32 | 0.42 | 0.50 | 0.53 | 0.91 | 1.19 | 1.18 |
| 1997 | --- | 0.39 | 0.41 | 0.47 | 0.57 | 0.78 | 1.30 | 1.31 |
| 1998 | --- | 0.33 | 0.41 | 0.55 | $0: 63$ | 1.00 | 1.62 | ---- |
| mean | 0.15 | 0.33 | 0.40 | 0.52 | 0.68 | 0.89 | 1.19 | 1.28 |

Table E3. Discards at age (above) and mean weights at age (below) of Cape Cod yellowtail flounder.

| Discards at age (thousands) |  |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | sum |
| 1985 | 340 | 184 | 34 | 0 | 0 | 0 | 558 |
| 1986 | 79 | 1,657 | 75 | 26 | 0 | 0 | 1,837 |
| 1987 | 14 | 877 | 168 | 0 | 0 | 0 | 1,059 |
| 1988 | 360 | 1,328 | 177 | 0 | 0 | 0 | 1,864 |
| 1989 | 114 | 1,405 | 396 | 1 | 0 | 0 | 1,917 |
| 1990 | 81 | 2,047 | 2,501 | 19 | 0 | 0 | 4,648 |
| 1991 | 460 | 895 | 561 | 100 | 7 | 0 | 2,023 |
| 1992 | 1,688 | 3,543 | 731 | 29 | 3 | 0 | 5,994 |
| 1993 | 138 | 324 | 173 | 30 | 0 | 0 | 665 |
| 1994 | 60 | 383 | 279 | 49 | 4 | 1 | 776 |
| 1995 | 453 | 469 | 652 | 50 | 2 | 0 | 1,627 |
| 1996 | 7 | 397 | 327 | 94 | 11 | 0 | 837 |
| 1997 | 1 | 399 | 351 | 117 | 22 | 1 | 891 |
| 1998 | 8 | 39 | 171 | 29 | 6 | 0 | 255 |
| mean | 272 | 996 | 471 | 39 | 4 | 0 | 1,782 |


| Discarded weight at age (kg) |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1 | 2 | 3 | 4 | age |  |
| 1985 | 0.13 | 0.15 | 0.15 |  |  | 6 |
| 1986 | 0.10 | 0.17 | 0.19 | 0.18 |  |  |
| 1987 | 0.06 | 0.19 | 0.19 |  |  |  |
| 1988 | 0.12 | 0.15 | 0.20 |  |  |  |
| 1989 | 0.13 | 0.21 | 0.25 | 0.36 |  |  |
| 1990 | 0.08 | 0.24 | 0.27 | 0.33 |  |  |
| 1991 | 0.12 | 0.19 | 0.27 | 0.37 | 0.54 |  |
| 1992 | 0.05 | 0.11 | 0.22 | 0.31 | 0.36 |  |
| 1993 | 0.09 | 0.15 | 0.27 | 0.33 | 0.63 |  |
| 1994 | 0.08 | 0.20 | 0.29 | 0.32 | 0.38 | 0.34 |
| 1995 | 0.07 | 0.16 | 0.23 | 0.33 | 0.48 |  |
| 1996 | 0.04 | 0.15 | 0.28 | 0.36 | 0.50 |  |
| 1997 | 0.03 | 0.21 | 0.29 | 0.39 | 0.54 | 0.65 |
| 1998 | 0.03 | 0.26 | 0.35 | 0.44 | 0.56 | 0.59 |
| mean | 0.08 | 0.18 | 0.25 | 0.34 | 0.50 | 0.53 |

Table E4a. Survey indices of Cape Cod yellowtail abundance and biomass.

| NEFSC Fall Survey |  |  |  |  | - age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ | sum | kg/tow |
| 1979 | 7.87 | 8.02 | 2.41 | 0.60 | 0.11 | 0.03 | 0.00 | 0.00 | 19.04 | 5.34 |
| 1980 | 20.70 | 17.63 | 8.00 | 3.04 | 0.67 | 0.00 | 0.07 | 0.00 | 50.11 | 13.52 |
| 1981 | 6.34 | 9.64 | 1.74 | 0.45 | 0.29 | 0.00 | 0.00 | 0.00 | 18.46 | 4.11 |
| 1982 | 1.13 | 5.39 | 5.18 | 0.63 | 0.70 | 0.06 | 0.00 | 0.00 | 13.09 | 4.32 |
| 1983 | 0.66 | 0.88 | 0.55 | 0.04 | 0.00 | 0.00 | 0.00 | 0.00 | 2.13 | 0.49 |
| 1984 | 0.64 | 2.25 | 1.04 | 1.31 | 0.93 | 0.30 | 0.15 | 0.15 | 6.77 | 2.79 |
| 1985 | 9.03 | 3.48 | 2.65 | 0.40 | 0.00 | 0.00 | 0.00 | 0.00 | 15.56 | 3.25 |
| 1986 | 2.62 | 7.14 | 0.60 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 10.36 | 1.98 |
| 1987 | 1.08 | . 2.60 | 0.91 | 0.11 | 0.09 | 0.00 | 0.00 | 0.00 | 4.79 | 1.12 |
| 1988 | 6.16 | '9.01 | 0.89 | 0.17 | 0.00 | 0.00 | 0.00 | 0.00 | 16.23 | 2.29 |
| 1989 | 3.53 | 11.39 | 4.19 | 0.74 | 0.00 | 0.00 | 0.00 | 0.14 | 19.99 | 4.70 |
| 1990 | 7.01 | 11.90 | 5.58 | 0.09 | 0.02 | 0.00 | 0.00 | 0.00 | 24.60 | 4.76 |
| 1991 | 3.57 | 3.33 | 2.88 | 0.59 | 0.00 | 0.00 | 0.00 | 0.00 | 10.37 | 2.34 |
| 1992 | 4.82 | 5.29 | 3.68 | 1.52 | 0.36 | 0.27 | 0.00 | 0.00 | 15.94 | 3.81 |
| 1993 | 8.76 | 8.60 | 1.01 | 0.15 | 0.00 | 0.00 | 0.00 | 0.00 | 18.52 | 2.15 |
| 1994 | 4.78 | 14.27 | 5.13 | 1.40 | 0.43 | 0.00 | 0.00 | 0.00 | 26.01 | 5.38 |
| 1995 | 1.18 | 1.64 | 1.57 | 0.34 | 0.08 | 0.00 | 0.00 | 0.00 | 4.81 | 1.49 |
| 1996 | 2.07 | 5.36 | 8.78 | 2.31 | 0.26 | 0.00 | 0.00 | 0.00 | 18.78 | 5.12 |
| 1997 | 2.07 | 4.79 | 5.45 | 2.46 | 1.33 | 0.23 | 0.00 | 0.00 | 16.33 | 4.63 |
| 1998 | 1.96 | 5.60 | 2.35 | 1.71 | 0.48 | 0.00 | 0.00 | 0.00 | 12.10 | 3.24 |
| mean | 4.80 | 6.91 | 3.23 | 0.90 | 0.29 | 0.04 | 0.01 | 0.01 | 16.20 | 3.84 |

NEFSC Spring Survey

| year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ | sum | $\mathrm{kg} / \mathrm{tow}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1979 | 0.55 | 0.71 | 1.33 | 0.85 | 0.04 | 0.03 | 0.00 | 0.00 | 3.51 | 1.20 |
| 1980 | 0.00 | 7.14 | 4.08 | 1.43 | 0.29 | 0.00 | 0.00 | 0.00 | 12.94 | 4.89 |
| 1981 | 0.10 | 6.30 | 4.27 | 0.93 | 1.06 | 0.51 | 0.66 | 0.00 | 13.83 | 4.41 |
| 1982 | 0.08 | 2.79 | 7.23 | 3.71 | 1.00 | 0.57 | 0.63 | 0.16 | 16.17 | 7.16 |
| 1983 | 2.36 | 6.33 | 5.09 | 2.09 | 0.22 | 0.15 | 0.00 | 0.00 | 16.24 | 4.78 |
| 1984 | 0.09 | 2.39 | 1.42 | 0.92 | 0.60 | 0.05 | 0.07 | 0.16 | 5.70 | 1.99 |
| 1985 | 0.13 | 1.86 | 1.81 | 0.43 | 0.25 | 0.10 | 0.00 | 0.00 | 4.58 | 1.37 |
| 1986 | 0.04 | 4.33 | 0.37 | 0.10 | 0.24 | 0.00 | 0.00 | 0.00 | 5.08 | 1.04 |
| 1987 | 0.15 | 3.44 | 5.15 | 0.84 | 1.30 | 1.31 | 1.52 | 0.74 | 14.45 | 7.14 |
| 1988 | 2.13 | 9.11 | 1.87 | 1.22 | 0.47 | 0.18 | 0.08 | 0.00 | 15.06 | 2.51 |
| 1989 | 0.53 | 6.33 | 3.88 | 0.35 | 0.17 | 0.00 | 0.00 | 0.00 | 11.26 | 1.93 |
| 1990 | 0.00 | 5.51 | 13.35 | 0.35 | 0.00 | 0.24 | 0.00 | 0.00 | 19.45 | 4.38 |
| 1991 | 0.96 | 8.23 | 5.67 | 1.80 | 0.42 | 0.00 | 0.11 | 0.00 | 17.19 | 3.76 |
| 1992 | 0.37 | 2.25 | 3.52 | 0.98 | 0.04 | 0.00 | 0.00 | 0.00 | 7.16 | 1.67 |
| 1993 | 0.15 | 1.51 | 1.75 | 0.87 | 0.00 | 0.00 | 0.00 | 0.00 | 4.28 | 0.93 |
| 1994 | 0.80 | 5.64 | 2.33 | 0.90 | 0.33 | 0.19 | 0.00 | 0.00 | 10.19 | 1.79 |
| 1995 | 0.32 | 2.10 | 7.33 | 4.74 | 0.46 | 0.11 | 0.00 | 0.00 | 15.06 | 3.68 |
| 1996 | 0.03 | 0.85 | 1.18 | 0.63 | 0.00 | 0.00 | 0.00 | 0.00 | 2.69 | 0.62 |
| 1997 | 0.05 | 1.98 | 3.15 | 2.54 | 0.56 | 0.00 | 0.00 | 0.00 | 8.28 | 2.43 |
| 1998 | 0.00 | 1.71 | 5.03 | 1.83 | 0.42 | 0.00 | 0.00 | 0.00 | 8.98 | 2.32 |
| 1999 | 0.04 | 1.57 | 5.05 | 3.61 | 0.70 | 0.26 | 0.00 | 0.00 | 11.24 | 3.65 |
| mean | 0.42 | 3.91 | 4.04 | 1.48 | 0.41 | 0.18 | 0.15 | 0.05 | 10.64 | 3.03 |

Table E4b. Survey indices of Capé Cod yellowtail abundance and biomass.

| MADMF Fall Survey |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | $8+$ | sum | kg/tow |
| 1978 | 0.04 | 7.13 | 7.74 | 1.45 | 0.11 | 0.00 | 0.01 | 0.00 | 0.00 | 16.48 | 2.80 |
| 1979 | 0.03 | 24.11 | 22.82 | 1.78 | 0.06 | 0.00 | 0.00 | 0.00 | 0.00 | 48.80 | 7.33 |
| 1980 | 0.03 | 26.54 | 12.38 | 2.70 | 0.35 | 0.00 | 0.00 | 0.00 | 0.00 | 42.00 | 5.90 |
| 1981 | 0.00 | 2.93 | 6.54 | 1.54 | 0.23 | 0.17 | 0.00 | 0.00 | 0.00 | 11.41 | 2.76 |
| 1982 | 0.00 | 9.58 | 3.36 | 5.54 | 0.30 | 0.08 | 0.00 | 0.00 | 0.00 | 18.86 | 4.20 |
| 1983. | 0.00 | 9.68 | 6.68 | 1.60 | 0.13 | 0.00 | 0.00 | 0.00 | 0.00 | 18.09 | 3.39 |
| 1984 | 0.04 | 1.91 | 3.00 | 0.86 | 0.39 | 0.10 | 0.02 | 0.00 | 0.04 | 6.37 | 1.18 |
| 1985 | 0.04 | 5.70 | 1.63 | 1.03 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 8.42 | 1.17 |
| 1986 | 0.01 | 2.60 | . 4.95 | 0.20 | 0.03 | 0.01 | 0.00 | 0.00 | 0.00 | 7.80 | 1.36 |
| 1987 | 0.44 | 5.85 | 2.30 | 0.49 | 0.07 | 0.02 | 0.00 | 0.00 | 0.00 | 9.17 | 1.09 |
| 1988 | 0.00 | 8.96 | 11.24 | 2.27 | 0.15 | 0.00 | 0.00 | - 0.00 | 0.00 | 22.62 | 3.71 |
| 1989 | 0.00 | 2.64 | 5.22 | 0.96 | 0.10 | 0.00 | 0.00 | 0.00 | 0.00 | 8.92 | 1.52 |
| 1990 | 0.00 | 5.20 | 11.93 | 4.84 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 21.98 | 4.16 |
| 1991 | 0.00 | 3.76 | 5.14 | 5.03 | 0.86 | 0.00 | 0.00 | 0.00 | 0.00 | 14.78 | 3.23 |
| 1992 | 0.20 | 7.18 | 3.62 | 2.08 | 0.47 | 0.20 | 0.00 | 0.00 | 0.00 | 13.75 | 2.00 |
| 1993 | 0.00 | 8.39 | 7.29 | 5.80 | 1.43 | 0.00 | 0.00 | 0.00 | 0.00 | 22.91 | 3.99 |
| 1994 | 0.00 | 3.56 | 8.39 | 3.06 | 0.96 | 0.12 | 0.00 | 0.00 | 0.00 | 16.09 | 3.27 |
| 1995 | 0.00 | 11.54 | 11.97 | 4.71 | 1.18 | 0.00 | 0.00 | 0.00 | 0.00 | 29.40 | 5.75 |
| 1996 | 0.01 | 1.87 | 3.94 | 2.18 | 0.17 | 0.00 | 0.00 | 0.00 | 0.00 | 8.17 | 1.56 |
| 1997 | 0.00 | 1.01 | 7.38 | 1.14 | 0.16 | 0.10 | 0.00 | 0.00 | 0.00 | 9.79 | 2.10 |
| 1998 | 0.00 | 7.05 | 7.44 | 1.56 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 16.05 | 2.68 |
| mean | 0.04 | 7.51 | 7.38 | 2.46 | 0.36 | 0.04 | 0.00 | 0.00 | 0.00 | 17.79 | 3.12 |



* preliminary, based on unaudited data.


## Table E5a. Estimates of abundance at age of Cape Cod yellowtail flounder.



Table E5b. Estimates of fishing mortality at age of Cape Cod yellowtail flounder.

| EISHING | $\begin{gathered} \text { MORTALITY } \\ 1985 \end{gathered}$ | - 1986 | $\begin{aligned} C & : \\ & \backslash \text { Program } \\ & 1987 \end{aligned}$ | $\begin{gathered} \text { Files } \backslash W \\ 1988 \end{gathered}$ | $\begin{gathered} T \backslash c c y t 99 \\ 1989 \end{gathered}$ | 1990 | 1991 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.04 | 0.02 | 0.00 | 0.02 | 0.02 | 0.01 | 0.06 |
| 2 | 0.47 | 0.73 | 0.57 | 0.56 | 0.15 | 0.67 | 0.30 |
| 3 | 0.83 | 0.74 | 1.22 | 1.46 | 0.89 | 1.60 | 0.97 |
| 4 | 2.11 | 0.76 | 0.80 | 2.64 | 1.47 | 0.81 | 2.00 |
| 5 | 2.40 | 0.78 | 0.82 | 3.43 | 1.56 | 0.83 | 2.24 |
| 6 | 2.40 | 0.78 | 0.82 | 3.43 | 1.56 | 0.83 | 2.24 |
|  | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| 1 | 0.30 | 0.02 | 0.01 | 0.08 | 0.00 | 0.00 | 0.00 |
| 2 | 0.93 | 0.09 | 0.09 | 0.17 | 0.13 | 0.19 | 0.10 |
| 3 | 0.81 | 0.54 | 0.58 | 0.94 | 0.62 | 0.59 | 0.77 |
| 4 | 1.38 | 0.91 | 1.22 | 1.00 | 1.25 | 0.93 | 0.41 |
| 5 | 1.45 | 0.93 | 1.27 | 1.03 | 1.30 | 0.96 | 0.41 |
| 6 | 1. 45 | 0.93 | 1.27 | 1.03 | 1.30 | 0.96 | 0.41 |
| Average $E$ for 4,5 |  |  |  |  |  |  |  |
|  | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| 4,5 | 2.25 | 0.77 | 0.81 | 3.03 | 1. 52 | 0.82 | 2.12 |
|  | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| 4,5 | 1.41 | 0.92 | 1.25 | 1.02 | 1.27 | 0.95 | 0.41 |
| Biomass Weighted F |  |  |  |  |  |  |  |
|  | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
|  | 0.48 | 0.61 | 0.72 | 0.40 | 0.24 | 1.09 | 0.66 |
|  | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
|  | 0.86 | 0.35 | 0.41 | 0.56 | 0.53 | 0.44 | 0.41 |

Table E5c. Estimates of mean biomass and spawning biomass of Cape Cod yellowtail flounder. MEAN BIOMASS (using catch mean weights at age)

|  | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 2991 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1144 | 423 | 367 | 2288 | 900 | 453 | 969 |
| 2 | 551 | 1267 | 634 | 813 | 3888 | $10^{-7}$ | 926 |
| 3 | 325 | 383 | 656 | 291 | 614 | 2065 | 520 |
| 4 | 126 | 776 | 186 | 131 | 106 | 301 | $\pm 26$ |
| 5 | 70 | 30 | 80 | 37 | 19 | 32 | 8 - |
| 6 | 33 | 09. | 51 | - 09 | 07 | 30 | 28 |
| I+ | 2247 | 2288 | 1974 | 3569 | 5533 | 3957 | 2945 |
|  | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| 1 | 284 | 584 | 439 | 402 | 249 | 119 | 128 |
| 2 | 551 | 608 | 1106 | 905 | 806 | 1441 | 996 |
| 3 | 617 | 583 | 81.8 | 851 | 902 | 943 | 996 |
| 4 | 210 | 290 | 287 | 367 | 358 | 410 | 659 |
| 5 | 66 | 73 | 119 | 93 | 124 | 110 | 226 |
| 6 | 10 | 49 | 74 | 64 | 14 | 12 | 116 |
| 1+ | 1737 | 2187 | 2841 | 2682 | 2454 | 3035 | 3120 |

SSB AT THE START OF THE SPAWNING SEASON -MAIES AND EEMAIES (MT) (USING SSB mean weights)

|  | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| I | 00 | 00 | 00 | 00 | 00 | 00 | 00 |
| 2 | 46 | 106 | 53 | 68 | 319 | 93 | 76 |
| 3 | 275 | 324 | 551 | 243 | 519 | 1705 | 439 |
| 4 | 123 | 183 | 194 | 121 | 108 | 205 | 422 |
| 5 | 66 | 32 | 84 | 30 | 19 | 33 | 84 |
| 6 | 31 | 09 | 53 | 07 | 07 | 32 | 27 |
| $1+$ | 541 | 654 | 935 | 469 | 972 | 2068 | 1048 |
|  | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| $\square$ | 00 | 00 | 00 | 00 | 00 | 00 | 00 |
| 2 | 46 | 50 | 94 | 91 | 66 | 118 | 84 |
| 3 | 521 | 491 | 690 | 698 | 762 | 795 | 842 |
| 4 | 217 | 302 | 297 | 328 | 364 | 418 | 620 |
| E | 68 | 77 | 124 | 95 | 128 | 112 | 234 |
| 6 | 10 | 51 | 76 | 67 | 15 | 12 | 122 |
| -- | 861 | 971 | 1281 | 1279 | 1334 | 1457 | 1901 |

Table E6. Summary of results from stochastic projections of Cape Cod yellowtail flounder.

| INPUT ASSUMPTIONS |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 1 | 2 | 3 | 4 | 5 | $6+$ |
| Stock Wt. | 0.051 | 0.254 | 0.375 | 0.468 | 0.592 | 0.870 |
| Landed Wt. | 0.150 | 0.337 | 0.399 | 0.478 | 0.595 | 0.873 |
| Discard wt. | 0.06 | 0.18 | 0.27 | 0.35 | 0.48 | 0.49 |
| Maturity | 0.00 | 0.08 | 0.81 | 1.00 | 1.00 | 1.00 |
| PR | 0.02 | 0.13 | 0.68 | 1.00 | 1.00 | 1.00 |
| - Discara | 1.00 | 0.56 | 0.21 | 0.09 | 0.05 | 0.03 |
| E-BASED PROJECTIONS |  |  |  |  |  |  |
| TIME-VARYING F |  |  |  |  |  |  |
| YEAR E |  |  |  |  |  |  |
| 19990.410 |  |  |  |  |  |  |
| 2000 . 0.070 |  |  |  |  |  |  |
| 20010.070 |  |  |  |  |  |  |
| 20020.070 |  |  |  |  |  |  |
| 20030.070 |  |  |  |  |  |  |
| 20040.070 |  |  |  |  |  |  |
| 20050.070 |  |  |  |  |  |  |
| $2006 \quad 0.070$ |  |  |  |  |  |  |
| 20070.070 |  |  |  |  | . |  |
| 2008 0.070 |  |  |  |  |  |  |
| 20090.070 |  |  |  |  |  |  |



ANNUAL PROBABILITY THAT MEAN BIOMASS EXCEEDS THRESHOLD: 6.100 THOUSAND MT
YEAR Er MEAN B > Threshold Value!

| 2.900 | 0.005 |
| :--- | :--- |
| 2000 | 0.106 |
| 2002 | 0.573 |
| 2002 | 0.929 |
| 2003 | 0.996 |
| 2004 | 1.000 |
| 2005 | 1.000 |
| 2006 | 1.000 |
| 2007 | 1.000 |
| 2008 | 1.000 |
| 2005 | 1.000 |



Figure E1. Total catch of Cape Cod yellowtail flounder.


Figure E2. Survey indices of Cape Cod yellowtail flounder biomass.
 Year

SSB Year; Recruitment Yearclass


Figure E3. Summary of Cape Cod yellowtail VPA results.


Figure E4. Status of the Cape Cod yellowtail flounder stock.


Figure E5. Projection of Cape Cod yellowtail flounder biomass at the Amendment \#9 long-term fishing target in 2000-2009.

## F. Gulf of Maine Cod by R.K. Mayo

### 1.0 Background

The Gulf of Maine cod stock was last assessed in 1998, and the assessment was reviewed at SAW 27. At that time, fully recruited fishing mortality (ages 4+) in 1997 was estimated to be 0.75, a decline from an average of 1.0 in 1995 and 1996. Spawning stock biomass was estimated to have declined to 8.600 mt in 1997 , a decline from a recent high of 14.300 mt in 1995 . The strength of the most recent recruiting year classes was estimated to be very low. The 1994, 1995 and 1996.year classes were estimated to have been the lowest in the VPA series dating back to 1982 (1980 year class). The recruit/SSB survival ratios for these most recent year classes were also estimated to very low compared to previous year classes. NEFSC spring and autumn research vessel bottom trawl survey indices for Gulf of Maine cod had declined to record low levels in 1994 and 1993, respectively, and have only increased slightly since that time. Recruitment indices for the 1994-1996 year classes derived from the NEFSC and Commonwealth of Massachusetts surveys were also among the lowest in the respective series.

### 2.0 1999 Assessment

## Fishery

Commercial landings of Gulf of Maine cod declined to 4.156 metric tons (mt) in 1998. a 23 \% decline from 1997 (Table F1; Figure F1). No discard estimates were derived for 1998. The estimated recreational catch of Gulf of Maine cod (retained component only) equaled 824 mt in 1998. The number of commercial port samples for this stock declined from 74 in 1997 to 46 in 1998. Sampling was not well distributed among quarters and market categories, requiring substantial pooling over quarter. Market cod samples were pooled on a semi-annual basis, and both Large and Scrod samples were pooled on an annual basis. As has generally been the case. the landings at age in 1998 were dominated by age 3 and 4 cod.

## Input Data and Analyses

The present assessment represents a one-year update to the previous assessment (Mayo et al. 1998: NEFSC 1998). The same VPA formulation used in the previous assessment was employed in the present update, except for the addition of current year (1999) spring survey data. Catch at age data for 1998, and NEFSC and Mass. DMF survey abundance indices (stratified mean number per tow at age) were updated through spring 1999. As in the most recent VPA, commercial CPUE indices were included only through 1993.

Precision of the 1999 stock sizes and 1998 fishing mortality and SSB estimates was derived from 1000 bootstrap simulations of the 1999 VPA formulation. A retrospective analysis of terminal year estimates of stock sizes, fully recruited fishing mortality and SSB was carried out back to 1994.

### 3.0 Assessment Results

NEFSC research vessel bottom trawl survey abundance and biomass indices for Gulf of Maine cod remained relatively low through autumn 1998 and spring 1999 (Table F2: Figure F2). The autumn 1998 indices declined slightly from the 1997 levels. while the spring 1999 indices increased slightly from the 1998 levels, but remain no higher than indices observed in 1996 and 1997. Recruitment indices for the 1994-1997 year classes derived from the NEFSC and Commonwealth of Massachusetts bottom trawl surveys are among the lowest in the respective series, although an index for the 1998 year class appears to be above the recent average.

Fully recruited fishing mortality (ages $4+$ ) in 1998 is estimated to be 0.64 (Table F3a). a decline from 0.75 in 1997, as reported in the previous assessment. However, the 1997 fully recruited F is now estimated to have been 0.85 (Table F3a; Figure F3). Spawning stock biomass is estimated at 8.300 mt in 1998, a decline from 9,900 mt in 1997 (Table F3b; Figure F4). The most recent high level of SSB (14,200 mt) occurred in 1995.

Recent recruiting year classes continue to be poor (Table F3a; Figure F3). The 1994, 1995 and 1996 year classes are still estimated to be the lowest in the VPA series dating back to 1982 (1980 year class), and the recruit/SSB survival ratios for these most recent year classes are also low compared to year classes prior to 1987. Biomass weighted fishing mortality (ages $1+$ ) has declined slightly from 0.43 in 1997 to 0.36 in 1998 (Table F3d; Figure F3), and mean biomass (ages $1+$ ) declined from $28,600 \mathrm{mt}$ in 1991 to 11.800 mt in 1998 (Table F3c; Figure F4).

## VPA Diagnostics

Estimates of 1999 stock sizes at age 2-6 were estimated with reasonable precision (C.Vs ranged from $27-39 \%$ ). There were no outstanding residual patterns in the NEFSC Spring surveys, but the NEFSC Autumn survey residuals suggest that stock sizes at most ages were estimated by the VPA to be higher than indicated by the surveys. Accounting for the precision in the current assessment, there is a $90 \%$ probability that fully recruited F in 1998 was greater than 0.52 , and that SSB in 1998 was less than $9,700 \mathrm{mt}$. There were substantial retrospective patterns in this assessment carrying back to 1994. Average fully recruited fishing mortality on ages 4 and 5 in the terminal year would have been consistently under-estimated since 1994, and recruitment at age 2 would have been under-estimated since 1996. No significant retrospective pattern was detected for SSB.

### 4.0 Forecasts

Forecasts of stock size and landings were performed over the short term (1999-2000) using stochastic projection software. Long-term (1982-1998) mean stock and catch weights, current (1995-1998) average partial recruitment and maturation at age were employed in all forecasts (Table F4).

In contrast to past analyses. forecasts included age 1 as recruits rather than age 2. This was done for consistency between the projected mean biomass and biomass-weighted F and the estimate of Bmsy and Fmsy incorporated in the SFA control rule. To accomplish this. it was necessary to estimate age 1 recruits in 1999 (1998 year class). NEFSC and Mass. DMF autumn surveys conducted in 1998 and and spring surveys conducted in 1999 provided some information on the strength of the 1998 year class. Abundance indices for the 1998 year class from.these 4 surveys were included in a regression model to evaluate the degree to which age 0 and 1 survey indices could be used as a predictor of age 1 year class strength.

Results indicated statistically poor fits for all 4 surveys; no survey explained more than $26 \%$ of the variation in the historical VPA estimates of age 1 recruitment, and the autumn age 0 surveys explained only $8 \%$ of the variation. It was therefore concluded that the survey indices at age 0 and 1 do not contain sufficient information to characterize the potential strength of the 1998 year class For this reason, the strength of the 1998 year class was determined in the same manner as described below for the first three years of the forecast.

Forecasts commenced using 1999 ages 1-7+ survivors as estimated by the 1000 bootstrapped VPA outcomes. Fully recruited fishing mortality in 1999 was assumed equal to that in 1998 ( 0.64 , fully recruited), given that the 1999 annual catch is not directly controlled by current fishing regulations. SSB is estimated to have increased slightly from $8,300 \mathrm{mt}$ in 1998 to 8.800 mt in 1999. Mean biomass (ages 1+) also increased slightly to 12.900 mt in 1999 from an estimated 11.800 mt in 1998.

## 2000-2009 Forecast

For 2000-2009. forecasts utilized a stochastic re-sampling scheme in which observed recruitment estimates derived from the VPA corresponding to the 1980-1996 year classes was used during the 10 years of the projection horizon. The re-sampling scheme was divided into 3 phases. In the first phase, recruitment for 2000 and 2001 was obtained by re-sampling only the most recent 4 estimates.(1993-1996 year classes; GM recruitment $=2.5$ million). In the second phase, recruitment for 2002 and 2005 was obtained by re-sampling the most recent 9 estimates (19881996 year classes: GM recruitment $=3.6$ million. Finally, in the last phase, recruitment for 20062009 was derived by re-sampling all of the observed recruitment estimates (1980-1996 year classes: GM recruitment $=5.3$ million).

The concept underlying this approach is that recruitment to be expected in the most recent years is more likely to be in accordance with the most recent observations when SSB and R/SSB was relatively low. As the forecast moves further away from the current period and SSB increases, the distribution of expected recruitment may be more similar to that observed over a longer period of the past. Thus, as SSB gradually increases, two step functions in the recruitment resampling allow higher recruitment values to be generated in the out years of the projections.

According to the SFA control rule for Gulf of Maine cod (Figure F5). when the mean stock biomass is between $1 / 4$ and $1 / 2 \mathrm{Bmsy}(8,250-16.500 \mathrm{mt}$ ). a 5 -year rebuilding period may be appropriate. According to the 1999 assessment, mean biomass was estimated to be 12.969 mt in 1999. so a biomass-weighted fishing mortality rate ( 0.15 ; equals 0.22 . fully recruited F ). corresponding to this position on the control rule was applied throughout the 10 -year forecast horizon. Biomass trajectories for the 2000-2009 period are given in Table F6 and Figure F6.

The medium-term forecasts suggest that under the control rule F scenario, $\mathrm{B}_{\text {msy }}$ may be achieved with at least a $50 \%$ probability in 2006 , and at least a $90 \%$ probability in 2009.

### 5.0 Sources of Uncertainty

* Poor biological sampling in 1998.

Incomplete seasonal coverage and apparent incomplete sampling of larger cod may have resulted in an underestimate of the number of larger, relatively older cod in the 1998 commercial landings. This would result in an overall lower mean weight, higher numbers landed and a greater dominance of younger fish in the estimated landings. The over-estimate of younger fish may have inflated the size of recruiting year classes in 1998 and 1997.

* Retrospective pattern inVPA.

Fully recruited F has been under-estimated since 1995. Recruitment has been under-estimated since 1994. Therefore, short-term projections are likely to be optimistic if fishing mortality is actually higher in 1998 than initially estimated. As well, the actual estimated recruitment at age 2 appears to have been underestimated in recent terminal years, although recruitment is still estimated to have declined substantially in recent years..

## * Catch at Age

Leaving recreational and discards out of assessment especially if recreational catch becomes a greater percentage of total catch is a source of uncertainty. The SARC previously rejected putting in recreational catch because of very poor sampling of recreational catch. This should only be a scaling factor if age structure is the same.

### 6.0 References

Mayo, R.K., L. O’Brien, and S.E. Wigley. 1998. Assessment of the Gulf of Maine Atlantic Cod Stock for 1998. NMFS/NEFSC, Woods Hole Laboratory Ref. Doc. 98-13.

NEFSC. 1998. $27^{\text {th }}$ Northeast Regional Stock Assessment Workshop ( $27^{\text {th }}$ SAW). Stock Assessment Review Committee (SARC) Consensus Summary of Assessments. NMFS/NEFSC, Woods Hole Laboratory Ref. Doc. 98-15.

Table F1.Commercial landings (metric tons, live) of Atlantic cod the Gulf of Maine (NAFO Division 5Y), 1960 - 1998.

Gulf of Maine


USA 1960-1993 landings from NMFS, NEFSC Detailed Weighout Files and Canvass data.
USA 1994-1998 landings estimated by prorating NMFS, NEFSC Detailed Weighout data by Vessel Trip Reports.

Table F2. Standardized stratified mean catch per tow in numbers and weight (kg) for Atlantic cod from NEFSC offshore spring and autumn research vessel bottom trawl surveys in the Gulf of Maine (Strata 26-30 and 36-40), 1963-1998 [a,b]


| Year | No/Tow | Wt/Tow | No/Tow | Wt/Tow |
| :---: | :---: | :---: | :---: | :---: |
| 1963 | - | - | 5.92 | 17.9 |
| 1964 | - | - | 4.00 | 22.8 |
| 1965 | - | - | 4.49 | 12.0 |
| 1966 | - | - | 3.78 | 12.9 |
| 1967 | - | - | 2.56 | 9.2 |
| 1968 | 5.44 | 17.9 | 4.39 | 19.4 |
| 1969 | 3.25 | 13.2 | 2.76 | 15.4 |
| 1970 | 2.21 | 11.1 | 4.90 | 16.4 |
| 1971 | 1.43 | 7.0 | 4.37 | 16.5 |
| 1972 | 2.06 | 8.0 | 9.31 | 13.0 |
| 1973 | 7.54 | 18.8 | 4.46 | 8.7 |
| 1974 | 2.91 | 7.4 | 4.33 | 9.0 |
| 1975 | 2.51 | 6.0 | 6.15 | 8.6 |
| 1976 | 2.78 | 7.6 | 2.15 | 6.7 |
| 1977 | 3.88 | 8.5 | 3.08 | 10.2 |
| 1978 | 2.06 | 7.7 | 5.75 | 12.9 |
| 1979 | 4.27 | 9.5 | 3.49 | 17.5 |
| 1980 | 2.15 | 6.2 | 7.04 | 14.2 |
| 1981 | 4.86 | 10.8 | 2.42 | 8.1 |
| 1982 | 3.75 | 8.6 | 7.77 | 16.1 |
| 1983 | 3.91 | 10.5 | 4.22 | 8.8 |
| 1984 | 3.40 | 5.8 | 2.42 | 8.8 |
| 1985 | 2.52 | 7.7 | 2.92 | 8.5 |
| 1986 | 1.96 | 3.6 | 1.95 | 5.1 |
| 1987 | 1.68 | 3.0 | 2.98 | 3.4 |
| 1988 | 3.13 | 3.3 | 5.90 | 6.6 |
| 1989 | 2.26 | 2.5 | 4.65 | 4.6 |
| 1990 | 2.36 | 3.1 | 2.99 | 4.9 |
| 1991 | 2.39 | 2.9 | 1.25 | 2.8 |
| 1992 | 2.41 | 8.7 | 1.43 | 2.4 |
| 1993 | 2.50 | 5.9 | 1.23 | 1.0 |
| 1994 | 1.27 | 2.4 | 2.14 | 2.7 |
| 1995 | 1.91 | 2.4 | 2.01 | 3.7 |
| 1996 | 2.46 | 5.4 | 1.32 | 2.4 |
| 1997 | 2.19 | 5.6 | 0.87 | 1.9 |
| 1998 | 1.71 | 4.2 | 0.84 | 1.5 |
| 1999 | 2.30 | 5.1 |  |  |


[a]During 1963-1984, BMV oval doors were used in the spring and autumn surveys; since 1985, Portugeuse polyvalent doors have been used in both surveys. Adjustments have been made to the 1963-1984 catch per tow data to standardize these data to polyvalent door equivalents. Conversion coefficients of 1.56 (numbers) and 1.62 (weight) were used in this standardization (NEFSC 1991).
[b] Spring surveys during 1973-1981 were accomplished with a '41 Yankee' trawl; in all other years, spring surveys were accomplished with a '36 Yankee' trawl. No adjustments have been made to the catch per tow data for these differences.
[c] In the Gulf of Maine, spring surveys during 1980-1982, 1989-1991 and 1994, and autumn surveys during 1977-1978, 1980, 1989-1991 and 1993 were accomplished with the R/V DELAWARE II; in all other years, the surveys were accomplished using the R/V ALBATROSS IV. Adjustments have been made to the R/V DELAWARE II catch per tow data to standardize these to R/V ALBTATROSS IV equivalents. Conversion coefficients 0.79 (number) and 0.67 (weight) were used in this standardization (NEFSC 1991).

Table F3a. Estimates of Stock Size ( 000 s of fish) and Instantaneous Fishing Mortality Rate (F) for Gulf of Maine cod obtained from Virtual Population Analysis (VPA).
 STOCK NUMBERS (Jan 1) in thousands - D:VASSESSIgmcodigmcod99igmcod99.3.out

| Age | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 6162 | 5534 | 7746 | 4914 | 7410 | 9954 | 21648 | 3376 | 3391 |
| 2 | 9108 | 5018 | 4530 | 6339 | 4023 | 6067 | 8148 | 17724 | 2764 |
| 3 | 4328 | 6208 | 3325 | 3306 | 4821 | 3218 | 4772 | 6526 | 14206 |
| 4 | 2666 | 2066 | 2950 | 1600 | 1399 | 1989 | 2096 | 2601 | 3911 |
| 5 | 1661 | 1149 | 734 | 1058 | 413 | 410 | 625 | 854 | 814 |
| 6 | 166 | 787 | 363 | 206 | 296 | 112 | 85 | 145 | 293 |
| 7+ | 547 | 284 | . 250 | 214 | 156 | 132. | 58 | 98 | 182 |
| $1+$ | 24639 | 21046 | 19900 | 17636 | 18519 | 21882 | 37432. | 31323 | 25560 |
| Age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | . 1999 |
| 1 | 5883 | 5309 | 8260 | 3090 | 2912 | 1983 | 2204 | 3490 | 0 |
| 2 | 2776 | 4817 | 4346 | 6763 | 2530 | 2384 | 1624 | 1805 | 2858 |
| 3 | 2077 | 1962 | 3660 | 3490 | 5511 | 1874 | 1893 | 1281 | 1393 |
| 4 | 8532 | 856 | 1127 | 1651 | 1938 | 3716 | 1006 | 1153 | 696 |
| 5 | 1334 | 3220 | 263 | 342 | 325 | 543 | 1469 | 430 | 454 |
| 6 | 277 | 323 | 811 | 98 | 20 | 91 | 131 | 450 | 203 |
| $7+$ | 151 | 131 | 63 | 114 | 55 | 20 | 15 | 23 | 204 |
| $1+$ | 21031 | 16617 | 18529 | 15549 | 13290 | 10611 | 8342 | 8633 | 5807 |

FISHING MORTALITY - D:VASSESSIgmcodlgmcod99igmcod99.3.out

| Age | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2 | 0.18 | 0.21 | 0.12 | 0.07 | 0.02 | 0.04 | 0.02 | 0.02 | 0.09 |
| 3 | 0.54 | 0.54 | 0.53 | 0.66 | 0.69 | 0.23 | 0.41 | 0.31 | 0.31 |
| 4 | 0.64 | 0.83 | 0.83 | 1.15 | 1.03 | 0.96 | 0.70 | 0.96 | 0.88 |
| 5 | 0.55 | 0.95 | 1.07 | 1.07 | 1.10 | 1.37 | 1.26 | 0.87 | 0.88 |
| 6 | 0.61 | 0.90 | 0.89 | 1.16 | 1.08 | 1.05 | 0.82 | 0.97 | 0.90 |
| $7+$ | 0.61 | 0.90 | 0.89 | 1.16 | 1.08 | 1.05 | 0.82 | 0.97 | 0.90 |
|  |  |  |  |  |  |  |  |  |  |
| Avg 4-5 u | 0.60 | 0.89 | 0.95 | 1.11 | 1.07 | 1.17 | 0.98 | 0.92 | 0.88 |
| Avg 4-5 w | 0.61 | 0.88 | 0.87 | 1.12 | 1.05 | 1.03 | 0.83 | 0.94 | 0.88 |
|  |  |  |  |  |  |  |  |  |  |
| Age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |  |
| 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  |
| 2 | 0.15 | 0.07 | 0.02 | 0.00 | 0.10 | 0.03 | 0.04 | 0.06 |  |
| 3 | 0.69 | 0.35 | 0.60 | 0.39 | 0.19 | 0.42 | 0.30 | 0.41 |  |
| 4 | 0.77 | 0.98 | 0.99 | 1.43 | 1.07 | 0.73 | 0.65 | 0.73 |  |
| 5 | 1.22 | 1.18 | 0.78 | 2.65 | 1.08 | 1.22 | 0.98 | 0.55 | 0.64 |
| 6 | 0.84 | 1.18 | 0.98 | 1.66 | 1.11 | 0.80 | 0.85 | 0.64 |  |

Table F3b. Estimates of Spawning Stock Biomass (000s mt) and sexual maturation for Gulf of Maine cod obtained from Virtual Population Analysis (VPA).

SSB AT THE START OF THE SPAWNING SEASON -MALES AND FEMALES (MT)

| Age | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | $\mathbf{1 9 9 0}$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 330 | 297 | 399 | 142 | 214 | 292 | 641 | 108 | 290 |
| 2 | 2144 | 1247 | 1141 | 3096 | 2015 | 3041 | 4025 | 8683 | 725 |
| 3 | 3184 | 4633 | 2503 | 3872 | 6011 | 4218 | 6440 | 8545 | 10617 |
| 4 | 4820 | 3105 | 4650 | 2781 | 2575 | 4029 | 3647 | 5085 | 5317 |
| 5 | 6071 | 2971 | 1738 | 2983 | 1205 | 1184 | 2017 | 2187 | 2260 |
| 6 | 823 | 3496 | 1429. | 739 | 1245 | 511 | 409 | 597 | 1298 |
| $7+$ | 5415 | 2311 | 2125 | 1659 | 1297 | 1096 | 552 | 987 | 2079 |
|  |  |  |  |  |  |  |  |  |  |
| $1+$ | 22786 | 18061 | 13984 | 15272 | 14561 | 14371 | 17732 | 26192 | 22585 |
| $2+$ | 22456 | 17764 | 13585 | 15130 | 14347 | 14079 | 17091 | 26084 | 22295 |


| Age | 1991 | 1992 | 1993 | $\mathbf{1 9 9 4}$ | 1995 | 1996 | 1997 | 1998 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 432 | 424 | 623 | 79 | 74 | 50 | 64 | 102 |
| 2 | 740 | 1514 | 1266 | 2836 | 1115 | 1074 | 741 | 704 |
| 3 | 1300 | 1476 | 3055 | 4463 | 7665 | 2824 | 3011 | 1960 |
| 4 | 12114 | 1174 | 1645 | 3042 | 3604 | 6731 | 2191 | 2521 |
| 5 | 3002 | 6598 | 713 | 626 | 1057 | 1364 | 3323 | 1347 |
| 6 | 1275 | 1142 | 2838 | 372 | 85 | 473 | 452 | 1435 |
| $7+$ | 1451 | 1108 | 570 | 839 | 573 | 196 | 159 | 206 |
| $1+$ | 20313 | 13438 | 10710 | 12258 | 14173 | 12711 | 9940 | 8275 |
| $2+$ | 19881 | 13014 | 10087 | 12179 | 14099 | 12661 | 9876 | 8173 |

PERCENT MATURE (Females)

| Age | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 7 | 7 | 7 | 4 | 4 | 4 | 4 | 4 | 11 |
| 2 | 26 | 26 | 26 | 48 | 48 | 48 | 48 | 48 | 28 |
| 3 | 61 | 61 | 61 | 95 | 95 | 95 | 95 | 95 | 56 |
| 4 | 88 | 88 | 88 | 100 | 100 | 100 | 100 | 100 | 81 |
| 5 | 97 | 97 | 97 | 100 | 100 | 100 | 100 | 100 | 93 |
| 6 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 98 |
| $7+$ | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |
|  |  |  |  |  |  |  |  |  |  |
| Age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 4 |
| 1 | 11 | 11 | 11 | 4 | 4 | 4 | 4 | 4 |  |
| 2 | 28 | 28 | 28 | 38 | 38 | 38 | 38 | 38 |  |
| 3 | 56 | 56 | 56 | 89 | 89 | 89 | 89 | 89 |  |
| 4 | 81 | 81 | 81 | 99 | 99 | 99 | 99 | 99 |  |
| 5 | 93 | 93 | 93 | 100 | 100 | 100 | 100 | 100 |  |
| 6 | 98 | 98 | 98 | 100 | 100 | 100 | 100 | 100 | 100 |
| $7+$ | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100 |  |
| $=======================================================================$ |  |  |  |  |  |  |  |  |  |

Table F3c. Estimates of Mean Stock Biomass (000s mt) for Gulf of Maine cod obtained from Virtual Population Analysis (VPA).

MEAN BIOMASS DURING THE CALENDAR YEAR

| Age | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 5013 | 4514 | 6317 | 4008 | 6044 | 8119 | 17658 | 2754 | 2766 |
| 2 | 8746 | 4788 | 4504 | 6987 | 4701 | 7081 | 9266 | 19827 | 2575 |
| 3 | 5090 | 7269 | 3938 | 3872 | 5877 | 4407 | 6729 | 9071 | 18830 |
| 4 | 4981 | 3186 | 5021 | 2490 | 2355 | 3869 | 3355 | 4604 | 5442 |
| 5 | 5582 | 2577 | 1528 | 2672 . | 1067 | 997 | 1696 | 2025 | 2125 |
| 6 | 766 | 2849 | 1303 | 621 | 1014 | 437 | 360 | 416 | 1358 |
| $7+$ | 4243 | 1685 | 1551 | 1137 . | 905 | 770 | 410 | 708 | 1515 |
| $1+$ | 34422 | 26868 | 24161 | 21786 | 21963 | 25680 | 39474 | 39404 | 34611 |
| 2+ | 29409 | 22354 | 17844 | 17778 | 15919 | 17561 | 21816 | 36650. | 31845 |
| Age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |  |
| 1 | 4799 | 4330 | 6738 | 2521 | 2375 | 1618 | 1798 | 2847 |  |
| 2 | 2651 | 6457 | 5046 | 8867 | 3611 | 3592 | 2505 | 2030 |  |
| 3 | 2160 | 2894 | 4768 | 5126 | 8749 | 2981 | 3334 | 2004 |  |
| 4 | 13694 | 1363 | 1655 | 2571 | 3043 | 5753 | 2038 | 2235 |  |
| 5 | 2947 | 5347 | 728 | 390 | 944 | 1057 | 2751 | 1268 |  |
| 6 | 1258 | 875 | 2943 | 274 | 91 | 423 | 376 | 1282 |  |
| 7+ | 1072 | 757 | 408 | 519 | 397 | 146 | 117 | 160 |  |
| $1+$ | 28580 | 22023 | 22285 | 20268 | 19210 | 15569 | 12919 | 11825 |  |
| 2+ | 23781 | 17693 | 15547 | 17747 | 16835 | 13951 | 11121 | 8978 |  |

Table F3d. Estimates of Biomass (000s mt) weighted F for Gulf of Maine cod obtained from
Virtual Population Analysis (VPA).
=============================================================================================1
BIOMASS WEIGHTED F DURING THE CALENDAR YEAR

| Age | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $1+$ | 0.398 | 0.525 | 0.454 | 0.496 | 0.447 | 0.298 | 0.204 | 0.266 | 0.442 |
| 2+ | 0.464 | 0.631 | 0.615 | 0.608 | 0.617 | 0.435 | 0.370 | 0.286 | 0.481 |
| $3+$ | 0.584 | 0.746 | 0.782 | 0.956 | 0.867 | 0.702 | 0.628 | 0.599 | 0.515 |
| 4+ | 0.598 | 0.891 | 0.887 | 1.122 | 1.062 | 1.045 | 0.879 | 0.938 | 0.886 |
| $5+$ | 0.578 | 0.918 | 0.953 | 1.106 | 1.087 | 1.195 | 1.123 | 0.906 | 0.891 |
| $6+$ | 0.610 | 0.900 | 0.890 | 1.160 | 1.080 | 1.050 | 0.820 | 0.970 | 0.900 |
| 7+ | 0.610 | 0.900 | 0.890 . | 1.160 | 1.080 | 1.050 | 0.820 | 0.970 | 0.900 |
| Age | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |  |
| 1+ | 0.629 | 0.501 | 0.379 | 0.396 | 0.356 | 0.469 | 0.429 | 0.355 |  |
| 2+ | 0.756 | 0.624 | 0.544 | 0.453 | 0.406 | 0.524 | 0.498 | 0.467 |  |
| 3+ | 0.832 | 0.942 | 0.795 | 0.904 | 0.490 | 0.695 | 0.631 | 0.586 |  |
| 4+ | 0.849 | 1.147 | 0.957 | 1.605 | 1.076 | . 0.806 | 0.841 | 0.658 |  |
| $5+$ | 1.052 | 1.180 | 0.944 | 1.986 | 1.090 | '1.073 | 0.960 | 0.598 |  |
| $6+$ | 0.840 | 1.180 | 0.980 | 1.660 | 1.110 | 0.800 | 0.850 | 0.640 |  |
| 7+ | 0.840 | 1.180 | 0.980 | 1.660 | 1.110 | 0.800 | 0.850 | 0.640 |  |

Table F4. Input data used in medium-term forecasts for Gulf of Maine cod.

| Age | Partial F | M | Maturation | Catch Mean Weights | Stock Mean Weights |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.000 | ----- | 0.040 | 0.900 | 0.900 |
| 2 | 0.061 | 0.2 | 0.380 | 1.355 | 1.090 |
| 3 | 0.373 | 0.2 | 0.890 | 1.842 | 1.561 |
| 4 | 0.924 | 0.2 | 0.990 | 2.748 | 2.244 |
| 5 | 1.000 | 0.2 | 1.000 | 4.153 | 3.404 |
| 6 | 1.000 | 0.2 | 1.000 | 6.202 | 5.068 |
| 7+ | 1.000 | 0.2 | 1.000 | 11.129 | 11.129 |

Table F5. Results of medium-term forecasts for Gulf of Maine cod.

| Year | Median Mean Stock Biomass | Prob. of Exceeding Bmsy |
| :---: | :---: | :---: |
| 1999 | 12.969 | 0.000 |
| 2000 | 14.280 | 0.000 |
| 2001 | 16,818 | 0.000 |
| 2002 | 19.089 | 0.000 |
| 2003 | 22.558 | 0.000 |
| 2004 | 26.307 | 0.024 |
| 2005 | 28,899 | 0.139 |
| 2006 | 31.268 | 0.346 |
| 2007 | 34,692 | 0.632 |
| 2008 | 38,340 | 0.829 |
| 2009 | 42.442 | 0.930 |



## GULF OF MAINE COD

USA RESEARCH VESSEL BOTTOM-TRAWL SURVEYS STRATIFIED MEAN CATCH [KG] PER TOW


Figure F2.
Standardized stratified mean catch ( kg ) per tow of Atlantic cod in NEFSC spring and autumn research vessel bottom trawl surveys in the Gulf of Maine, 1963-1999.

Trends in Landings and Fishing Mortality Gulf of Maine Cod


Figure F3. Trends in Landing and Fishing Mortality.

Trends in Spawning Stock Biomass and Recruitment Gulf of Maine Cod


Figure F4. Trends in recruitment and stock biomass


Figure F5. SFA Control rule for Gulf of Maine cod with 1999 mean biomass and biomass-weighted fishing mortality rate indicated.


Figure F6. Medium-term forecast of mean stock biomass for Gulf of Maine Cod at SFA control rule $\mathrm{F}=0.15$

## G. Witch Flounder by S. E. Wigley

### 1.0 Background

Witch flounder was assessed in 1999 and reviewed at SAW 29 (NEFSC 1999).

### 2.0 Summary of Assessment Results

An analytical assessment (VPA) of USA 1982-1998 commercial catch at age data was conducted to estimate fishing mortality rates and abundance at age in 1999. Commercial catch (landings plus discards from the northern shrimp and large-mesh otter trawl fisheries) is given in Table G. 1 and Figure G.1. Information on recruitment and abundance was taken from standardized NEFSC spring and autumn surveys (Table G.2, Figures G.2a and G.2b).

The accepted VPA formulation estimated ages 4-11+ using NEFSC spring and autumn bottom trawl survey indices. The VPA calibration indicated that 1999 age 3 stock size was 131 million fish. approximately 4 -fold higher than any previous estimate for age 3 fish. The negative residual pattern associated with the stock size estimates of age 3 and 4 persisted regardless of alternative formulations, and, given the retrospective pattern associated with age 3 recruitment, it was determined that the best estimates of age 3 and 4 stock sizes in 1999 would come from the survey indices of recruitment. Stock sizes in 1999 for age 3 and 4 were estimated directly from the survey data using regressions of VPA stock sizes on the corresponding survey indices (using the RTC3 program of ICES). The 1999 age 3 stock size was estimated to be 38.706 million fish (1996 year class) and age 4 stock size in 1999 was estimated to be 19.457 million fish ( 1995 year class). To estimate age 3 fish in 1998 (the 1995 year class), the 1995 year class at age 4 was back-calculated (accounting for natural mortality) to be 22.686 million fish (Table G.3).
Assessment results are summarized in Table G. 3 and Figures G. 3 and G.5. For more details see Wigley, Brodziak and Cadrin (1999).

Fishing moratlity: Fishing mortality (ages 7-9. unweighted) increased from 0.21 ( $18 \%$ exploitation) in 1982 to 0.59 ( $42 \%$ exploitation) in 1985, declined to 0.24 ( $20 \%$ exploitation) in 1990. increased to $0.86(54 \%$ exploitation) in 1996. then dropped to 0.37 ( $29 \%$ exploitation) in 1998 (Figure G.3). This trend in F is generally confirmed by the trend of pooled survey Z. There is an $80 \%$ probability that the 1998 F lies between 0.28 and 0.51 . Based on the ADAPT VPA, $3+$ mean biomass in 1998 was 18.934 mt and the 1998 F on $3+$ biomass was 0.13 . Based on catch statistics to-date. the estimated $3+$ biomass in 1999 is projected to be $26,048 \mathrm{mt}$ and F on $3+$ biomass is estimated to be 0.096 (Figure G.5).

Recruitment: Recruitment has been above average since 1992 (Figure G.4).
Mean 3+ biomass: Mean 3+ biomass declined steadily from 27.930 mt in 1982 to $7,742 \mathrm{mt}$ in 1994, then sharply increased to 18.934 mt by 1998. The estimated mean $3+$ biomass in 1999 is $26,048 \mathrm{mt}$ (above the overfishing threshold and near Bmsy; Figure G.4).

Spawning stock biomass: Spawning stock biomass (SSB') declined from 18.000 mt in 1982 to about 4.000 mt in 1995 . Following recruitment and maturation of the 1991-1993 year classes. SSB increased sharply to 8.652 mt in 1998 (Figure G.4). There is an $80 \%$ probability that the 1998 SSB lies between 7.400 mt and $11,000 \mathrm{mt}$. The age composition of the spawning stock biomass revealed that more than half the SSB in 1982 was composed of age $11+$ fish. but by 1998. more than half of the SSB consisted of age 5-7 fish, many of which were first-time spawners. Since the mid-1980's, the age structure remains severely truncated. The low number of older fish may adversely impact the stock's reproductive potential output.

Biological Reference Points: Yield and SSB per recruit analyses updated with an assumed $\mathrm{M}=$ 0.15 indicate that $\mathrm{F}_{0.1}=0.16$ ( $14 \%$ exploitation), $\mathrm{F}_{\max }=0.35$ ( $28 \%$ exploitation), and $\mathrm{F}_{20^{\circ}}=0.37$ (29\% exploitation).

MSY Based Reference Points: A non-equilibrium surplus production analysis (ASPIC) was performed on total catch and survey indices of stock biomass from 1963 to 1998. The model was calibrated with NEFSC spring and autumn biomass indices, where spring indices were lagged back one year to calibrate biomass at the end of the previous year. When q was unconstrained. the intrinsic rate of increase (r) was unreasonably high for this slow growing, long-lived species: therefore, survey q's were fixed according to ADAPT VPA estimates of age $3+$ biomass. Results of the final formulation estimated MSY to be 2.684 mt ; $\mathrm{B}_{\text {msy }}$ to be 25.000 mt and that the corresponding $\mathrm{F}_{\text {msy }}=0.106$.

SFA Control Rule: The Amendment 9 control rule states that when the stock biomass exceeds $B_{\text {msy }}$, the overfishing threshold is $F_{\text {mss. }}$, and target $F$ is the lower $80^{\text {th }}$ percentile (or $10^{\text {th }}$ percentile) of $\mathrm{F}_{\mathrm{msy}}$. When stock biomass is less than $\mathrm{B}_{\text {msy }}$, the overfishing threshold is based on maximum F that would allow rebuilding to $\mathrm{B}_{\text {msy }}$ in five years. When biomass is less than the biomass threshold. $F=0$. The biomass threshold is defined by the minimum stock size that is projected to rebuild to $\mathrm{B}_{\text {msy }}$ in 5 years at $\mathrm{F}=0$.

The Amendment 9 control rule was updated with the revised estimates of $\mathrm{F}_{\text {msy }}(0.106), \mathrm{B}_{\text {msy }}$ ( 25.000 mt ) and the tenth percentile of $\mathrm{F}_{\mathrm{msy}}$ ( 0.090 ; Figure 5). MSY is estimated as 2.684 mt . Based on the ADAPT estimates of age 3+ mean biomass in 1998 ( $18,934 \mathrm{mt}$ ) and 1998 F on $3+$ biomass (0.13), overfishing was occurring in 1998 (Figure 5). Assuming 1999 catches will equal 1998 catches, the 1999 F on fully recruited ages $7+$ is estimated to be 0.20 , and the target fishing mortality prescribed by the control rule for the 1999 stock size is 0.09 on biomass.

### 3.0 Forecasts for 1999-2009

Medium-term stochastic projections were performed to estimate meanl biomass (3+) during 1999-2009 under the control rule ( F on $3+$ biomass $=0.09, \mathrm{~F}$ fully recruited on ages $7+=0.11$ assuming equilibrium age structure) using VPA calibrated stock sizes in 1999 (age 3 and 4 stock sizes were estimated using RCT3, where the mean and standard error were used to generate the
1.000 estimates of age 3 and 4 stock sizes from a log-normal distribution; and ages 5-11+ were estimated by the 1000 bootstrapped VPA outcomes). Catches in 1999 were assumed equal to that in 1998 based on landings to date. The fully-recruited 1999 F was projected to be 0.20 ( F on $3+$ biomass $=0.096$ ). An average 1995-1997 partial recruitment. 1994-1998 maturity ogive, and the average 1982-1998 mean weights at age were used in the projections.

Spawning stock biomass and recruitment (age 3) data for witch flounder were presented in a previous section. Given the negative value for the Beverton-Holt slope parameter, the lack of a reasonable stock-recruitment relationship necessitated that recruitment (age 3 fish) in 1999-2009 be estimated by re-sampling from the empirical observations during 1982-1997 (1979-1994 year classes). Fishing mortality was apportioned among landings and discards based on the proportion landed at age during 1995-1997. The proportion of F and M which occurs before spawning equal 0.1667 (March 1); M was assumed to be 0.15 .

Fishing at the control rule ( F on $3+$ biomass $=0.09, \mathrm{~F}$ on fully recruited ages $7+=0.11$ ). mean biomass (3+) will increase to $40,573 \mathrm{mt}$ in 2005, and then decline to $37,111 \mathrm{mt}$ in $2009.48 \%$ above $B_{\text {msy }}$ of $25,000 \mathrm{mt}$ (Table G.4, Figure G.6).

### 4.0 Sources of Uncertainty

- The bootstrap procedure as applied does not capture the full extent of variability and uncertainty in the VPA results, particularly that which comes from the residual pattern and the retrospective pattern, and, therefore, the percentile distributions of the projected stock parameters understate the extent of the uncertainty in the forecasts.
- Confounding of survey based estimates of discards and use of same surveys as tuning indices for VPA calibration.
- Low frequency of samples across market category and season resulting in variable mean weights at age and estimates of numbers at age.
- Low catchability of standard survey gear leading to highly variable survey indices.
- Lack of data to support direct estimates of discards at age requiring use of various surrogate survey-based methods.
- The simple biomass dynamics model used to derive MSY-reference points does not account for age structure of the stock or current recruitment.
- Estimates of current recruitment are highly variable due to the dependence on catch at younger ages which consists almost entirely of discards and highly variable and imprecise survey indices for recruiting ages.
- High recruitment and survival rates may or may not continue.


### 5.0 References

NEFSC [Northeast Fisheries Science Center]. 1999. Report of the $29^{\text {th }}$ Northeast Regional Stock Assessment Workshop (29th SAW), Stock Assessment Review Committee (SARC) consensus summary of assessments. Northeast Fish. Sci. Cent. Ref. Doc. 99-14. 347 p.

Wigley, S.E., J. K.T. Brodziak, and S.X. Cadrin. 1999. Assessment of the witch flounder stock in Subareas 5 and 6 for 1999. Northeast Fish. Sci. Cent. Ref. Doc. 99-16. 153 p.

Table G1. Witch flounder landings, discards and catch (metric tons, live) from Subareas 5 and 6, by country, 1960-1998 [1960-1963 reported to ICNAF/NAFO (Burnett and Clark, 1983)].

| Year | Landings |  |  |  |  | Total USA Catch (used in VPA) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Canada | USA ${ }^{2}$ | Other ${ }^{1}$ | Total | Discards |  |
| 1960 | - | 1255 | - | 1255 |  |  |
| 1961 | 2 | 1022 | - | 1024 |  |  |
| 1962 | 1 | 976 | - | 977 |  |  |
| 1963 | 27 | 1226 | 121 | 1374 |  |  |
| 1964 | 37 | 1381 | - | 1418 |  |  |
| 1965 | 22 | 2140 | 502 | 2664 |  |  |
| 1966 | 68 | 2935 | 311 | 3314 |  |  |
| 1967 | '63 | 3370 | 249 | 3682 |  |  |
| 1968 | 56 | 2807 | 191 | 3054 |  |  |
| 1969 | - | 2542 | 1310 | 3852 |  |  |
| 1970 | 19 | 3112 | 130 | 3261 |  |  |
| 1971 | 35 | 3220 | 2860 | 6115 |  |  |
| 1972 | 13 | 2934 | 2568 | 5515 |  |  |
| 1973 | 10 | 2523 | 629 | 3162 |  |  |
| 1974 | 9 | 1839 | 292 | 2140 |  |  |
| 1975 | 13 | 2127 | $217^{\circ}$ | 2357 |  |  |
| 1976 | 5 | 1871 | 6 | 1882 |  |  |
| 1977 | 11 | 2469 | 13 | 2493 |  |  |
| 1978 | 18 | 3501 | 6 | 3525 |  |  |
| 1979 | 17 | 2878 | - | 2895 |  |  |
| 1980 | 18 | 3128 | 1 | 3147 |  |  |
| 1981 | 7 | 3422 | - | 3449 |  |  |
| 1982 | 9. | 4906 | - | 4915 | 48 | 4953 |
| 1983 | 45 | 6000 | - | 6045 | 162 | 6162 |
| 1984 | 15 | 6660 | - | 6675 | 100 | 6760 |
| 1985 | 46 | 6130 | - | 6431 | 61 | 6191 |
| 1986 | 67 | 4610 | - | 5216 | 25 | 4635 |
| 1987 | 23 | 3450 | - | 3819 | 47 | 3497 |
| 1988 | 45 | 3262 | - | 3665 | 60 | 3322 |
| 1989 | 13 | 2074 | - - | 2384 | 133 | 2207 |
| 1990 | 12 | 1478 | - | 1492 | 184 | 1662 |
| 1991 | 7 | 1798 | - | 1805 | 95 | 1893 |
| 1992 | 7 | 2246 | - | 2253 | 171 | 2417 |
| 1993 | 10 | 2605 | - | 2615 | 376 | 2981 |
| 1994 | 34 | 2670 | - | 2704 | 422 | 3092 |
| 1995 | 11 | 2212 | - | 2223 | 265 | 2477 |
| 1996 | 10 | 2088 | - | 2098 | 454 | 2542 |
| 1997 | 7 | 1775 | - | 1782 | 393 | 2168 |
| 1998 | * | 1849 | - | 1849 | 334 | 2184 |

${ }^{1}$ Includes West Germany, East Germany, Poland, Spain, Japan, \& the former USSR.
${ }^{2}$ excluding landings from Grand Banks (subarea 3).

* 1998 Canadian landings not available.

Table G2. Stratified mean number, weight ( kg ) and length ( cm ) per tow of witch flounder in NEFSC offshore spring and autumn bottom trawl surveys in Gulf of Maine-Georges Bank region (strata 22-30.36-40). 1963-1999.


Note: During 1963-1984. BMV oval doors were used in the spring and autumn surveys: since 1985. Portuguese polyvalent doors have been used in both surveys. No significant differences in catchability were found for witch flounder, therefore no adjustments have been made (Byrne and Forrester, MS 1991). No significant differences were found between research vessels, and no adjustment have been made (Byrne and Forrester, MS 1991). Spring surveys during 1973-1981 were accomplished with a 41 Yankee trawl: in all other years, a 36 Yankee trawl was used. No adjustments have been made.

Table G3. Estimates of begiming year stock size (thousands of fish), instantaneous fishing mortality ( F ) and spawning stock biomass (mt) for witch flounder estimated from virtual population analysis, 1982-1998. Bold values in 1999 are estimated from RCT3
(regressions of VPA stock sizes and corresponding NEFSC surveys): bold value in 1998 was back-calculated.

| STOC'K NUMBERS (Jan 1) in thousands |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 3 | 15434 | 17862 | 15866 | 7326 | 4876 | 2950 | 9502 | 6359 | 6871 | 8949 | 15279 | 10906 | 13869 | 27833 | 26142 | 20549 | 22686 | 38706 |
| 4 | 12807 | 13107 | 15061 | 13520 | 6191 | 4176 | 2519 | 7622 | 5059 | 5595 | 7293 | 12780 | $9(1) 0$ | 11898 | 23361 | 22390) | 17590) | 19457 |
| 5 | 9766 | 10035 | 10033 | 11603 | 10546 | 4979 | 3426 | 2038 | 6155 | 3764 | 4437 | 5108 | 9322 | 6816 | 9668 | 19224 | 18322 | 14.573 |
| 6 | 7903 | 7285 | 7227 | 6777 | 8022 | 7669 | 3852 | 2704 | 1693 | 4270 | 2430 | 3015 | 3079 | 5461 | 4756 | 6486 | 15185 | 14577 |
| 7 | 4566 | 5433 | 4809 | 4606 | 4037 | 4330 | 5414 | 2705 | 2035 | 1218 | 3135 | 1216 | 1742 | 1453 | 3102 | 2867 | 4296 | 11693 |
| 8 | 2990 | 3313 | 3201 | 2760 | 2550 | 2021 | 2266 | 3377 | 1621 | 1495 | 827 | 2028 | 492 | 731 . | 462 | 1342 | 1525 | 2228 |
| 9 | 2341 | 1965 | 1944 | 1366 | 1218 | 1420 | 931 | 879 | 2086 | 955 | 1057 | 523 | 1201 | 241 | $\cdot 381$ | 153 | 60.5 | 968 |
| 10 | 1372 | 1644 | $10 \times 07$ | 1027 | 613 | 665 | 776 | 429 | 431 | 1483 | 547 | 743 | 247 | 533 | 117 | 128 | 55 | 390 |
| $11+$ | 9014 | 5364 | 4581 | 3459 | 2073 | 1280 | 1728 | 1203 | 938 | 1213 | 1711 | 1034 | 700 | 307 | 230 | 178 | 246 | 179 |
| $3+$ | 66193 | 660008 | 63729 | 52444 | 40126 | $29490$ | 30414 | 27316 | 26889 | 28942 | 36716 | 37353 | 39652 | 55273 | 68219 | 73317 | 80510 | 102771 |
| FISHING MORTALITY |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |  |
| 3 | 0.01 | 0.02 | 0.01 | 0.02 | 0.01 | 0.01 | 0.07 | 0.08 | 0.06 | 0.05 | 0.03 | 0.04 | 0.00 | 0.03 | 0.00 | 0.01 | 0.00 | - |
| 4 | 0.09 | 0.12 | 0.11 | 0.10 | 0.07 | 0.05 | 0.06 | 0.06 | 0.15 | 0.08 | 0.21 | 0.17 | 0.13 | 0.06 | 0.04 | 0.05 | 0.04 |  |
| 5 | 0.14 | 0.18 | 0.24 | 0.22 | 0.17 | 0.11 | 0.09 | 0.04 | 0.22 | 0.29 | 0.24 | 0.36 | 0.38 | 0.21 | 0.25 | 0.09 | 0.08 |  |
| 6 | 0.22 | 0.27 . | 0.30 | 0.37 | 0.47 | 0.20 | 0.20 | 0.13 | 0.18 | 0.16 | 0.54 | 0.40 | 0.60 | 0.42 | 0.36 | 0.26 | 0.11 |  |
| 7 | 0.17 | 0.38 | 0.41 | 0.44 | 0.54 | 0.50 | 0.32 | 0.36 | $0.16{ }^{\text { }}$ | 0.24 | 0.29 | 0.76 | 0.72 | 1.00 | 0.69 | 0.48 | 0.51 |  |
| 8 | 0.27 | 0.38 | 0.70 | 0.67 | 0.44 | 0.62 | 0.80 | 0.33 | 0.38 | 0.20 | 0.31 | 0.37 | 0.57 | 0.50 | 0.95 | 0.65 | 0.30 |  |
| 9 | 0.20 . | 0.52 | 0.49 | 0.65 | 0.45 | 0.45 | 0.62 | 0.56 | 0.19 | 0.41 | 0.20 | 0.60 | 0.66 | 0.57 | 0.94 | 0.88 | 0.29 |  |
| 10 | 0.21 | 0.41 | 0.51 | 0.55 | 0.50 | 0.53 | 0.46 | 0.37 | 0.23 | 0.26 | 0.27 | 0.52 | 0.68 | 0.79 | 0.75 | 0.55 | 0.37 |  |
| $11+$ | 0.21 | 0.41 | 0.51 | 0.55 | 0.50 | 0.53 | 0.46 | 0.37 | 0.23 | 0.26 | 0.27 | 0.52 | 0.68 | 0.79 | 0.75 | 0.55 | 0.37 |  |
| 7-9 | 0.21 | 0.43 | 0.53 | 0.59 | 0.48 | 0.53 | 0.58 | 0.42 | 0.24 | 0.28 | 0.27 | 0.58 | 0.65 | 0.69 | 0.86 | $0: 67$ | 0.37 |  |

Tahle 3 continled
MEAN BIOMASS

| Age | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 2164 | 2447 | 2214 | 863 | 402 | 221 | 384 | 694 | 385 | 502 | 1441 | 1211 | 900 | 1481 | 1477 | 1123 | 1978 |
| 4 | 2750 | 2323 | 3036 | 2837 | 1146 | 724 | 461 | 1167 | 819 | 994 | 1411 | 2214 | 1587 | 1837. | 3290 | 3976 | 3191 |
| 5 | 2785 | 2310 | 2722 | 2959 | 2700 | 1309 | 949 | 597 | 1325 | 1049 | 1395 | 1275 | 2022 | 1763 | 1865 | 4298 | 4896 |
| 6 | 2775 | 2438 | 2450 | 2269 | 2443 | 2804 | 1408 | 1000 | 632 | 1547 | 805 | 1003 | 931 | 1798 | 1586 | 1908 | 5600 |
| 7 | 2148 | 2186 | 1989 | 1965 | 1553 | 1788 | 2322 | 1215 | 1026 | 584 | 1561 | 428 | 622 | 485 | 1164 | 1052 | 1729 |
| 8 | 1775 | 1574 | 1431 | 1303 | 1305 | 965 | 978 | 1827 | 866 | 887 | 491 | 1051 | 243 | 371 | 198 | 581 | 829 |
| 9 | 1747 | 1139 | 1174 | 792 | 780 | 883 | 531 | 515 | 1500 | 612 | 732 | 325 | 684 | 156 | . 198 | 83 | 414 |
| 10 | 1134 | 1231 | 679 | 713 | 441 | 474 | 569 | 324 | 377 | 1183 | 394 | 555 | 152 | 337 | 75 | 95 | 42 |
| 11 | 10652 | 5580 | 4484 | 3306 | 2016 | 1213 | 1713 | 1272 | 1134 | 1412 | 1735 | 1007 | 601 | 247 | 187 | 166 | 253 |
| $3+$ | 27930 | 21228 | 20179 | 17007 | 12786 | 10381 | 9315 | 8611 | 8064 | 8770 | 9965 | 9069 | 7742 | 8475 | 10040 | 13282 | 18934 |
|  |  |  |  |  |  |  |  |  |  | - |  |  |  |  |  |  |  |
| SSB AT THE START OF THE SPAWNING SEASON -MALES AND FEMALES (MT) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| 3 | 0 | 0 | 0 | 5 | 2 | 1 | 4 | 3 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0 | 88 | 107 | 368 | 145 | 79 | 47 | 96 | 109 | 6 | 8 | 17 | 108 | 100 | 172 | 189 | 136 |
| 5 | 55 | 486 | 483 | 1872 | 1729 | 769 | 522 | 328 | 787 | 89 | 114 | 127 | 913 | 719 | 814 | 1638 | 1746 |
| 6 | 423 | 1297 | 1175 | 2238 | 2453 | 2501 | 1255 | 901 | 577 | 534 | 344 | 446 | 854 | 1466 | 1345 | 1493 | 3658 |
| 7 | 1105 | 1859 | 1648 | 2037 | 1719 . | 1858 | 2417 | 1239 | 964 | 460 | 1185 | 415 | 724 | 589 | 1319 | 1184 | 1617 |
| 8 | 1589 | 1708 | 1579 | 1469 | 1429 | 1074 | 1184 | 1888 | 932 | 877 | 486 | 1152 | 265 | 398 | 242 | 693 | 761 |
| 9 | 1807 | 1336 | 1237 | 894 | 846 | 960 | 614 | 577 | 1499 | 659 | 758 | 373 | 780 | 169 | 244 | 101 | 416 |
| 10 | 1206 | 1393 | 772 | 811 | 499 | 543 | 632 | 350 | 375 | 1258 | 438 | 610 | 192 | 410 | 95 | 107 | 46 |
| 11 | 11939 | 6633 | 5492 | 4085 | 2458 | 1491 | 2069 | 1497 | 1280 | 1609 | 1982 | 1235 | 771 | 326 | 244 | 205 | 272 |
| $1+$ | 18124 | 14801 | 12493 | 13779 | 11281 | 9277 | 8743 | 6879 | 6526 | 5492 | 5315 | 4375 | 4608 | 4178 | 4475 | 5611 | 8652 |

Table G4. Input and results of medium-term projections for witch flounder.
Projection input:

| Age | Fish |  |  |  |  | Weights |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Pattern | Mature | Fraction | Catch | Stock | Discards |
| 3 | . 0130 | . 0000 | 1.00 | . 094 | . 056 | 0.030 |
| 4 | . 0730 | . 0800 | 0.89 | . 199 | . 140 | 0.078 |
| 5 | . 2330 | . 4500 | 0.62 | . 299 | . 247 | 0.149 |
| 6. | . 4730 | . 8500 . | 0.12 | . 419 | . 357 | 0.189 |
| 7. | 1.0000 | 1.0000 | 0.00 | . 549 | :484 | 0.235 |
| 8 | 1.0000 | 1.0000 | 0.00 | . 677 | . 615 | 0.235 |
| 9 | 1.0000 | 1.0000 | 0.00 | . 846 | . 764 | 0.235 |
| 10 | 1.0000 | 1.0000 | 0.00 | . 973 | . 907 | 0.235 |
| $11+$ | 1.0000 | 1.0000 | 0.00 | 1.319 | 1.319 | 0.235 |

Results:

| Year M | Median . <br> Mean Biomass (mt) | Probability of Exceeding Bmsy |
| :---: | :---: | :---: |
| 1999 | - 25,701 | 0.55 |
| 2000 | 30,550 | 0.80 |
| 2001 | 1 34,503 | 0.90 |
| 2002 | 37,367 | 0.95 |
| 2003 | 3 39,580 | 0.98 |
| 2004 | 40,416 | 0.99 |
| 2005 | 5 40,573 | 0.99 |
| 2006 | 39,993 | 0.99 |
| 2007 | 39,786 | 0.99 |
| 2008 | 8 38.298 | 0.99 |
| 2009 | - 37,111 | 0.98 |



Figure G.1. Historical USA witch flounder landings (mt), excluding USA landings from the Grand Banks in the mid-1980's. Thin line represents provisional landings data taken from Lange and Lux (1978). Discards from the shrimp and large-mesh otter trawl fishery.


Figure G.2. Stratified mean catch $(\mathrm{kg})$ per tow $(\mathrm{A})$ and mean number per tow $(\mathrm{B})$ of witch flounder in NEFSC spring and autumn research vessel bottom trawl surveys in the Gulf of Maine-Georges Bank region, 1963-1998; 1999 values are preliminary.

## Witch flounder



Figure G.3. Trends in total catch and fishing mortality for witch flounder, 1982-1998.


Figure G.4. Trends in spawning stock biomass, mean biomass (3+), and recruitment for witch flounder, recruitment fo the 1995 and 1996 year classes (hatched bars) estimated from log-log regression of survey and VPA stock sizes estimates.

## Witch flounder



Figure G.5. MSY-based reference points and control rule for witch flounder.


Figure G.6. Meduim-term stochastic projections for witch flounder, median mean biomass 3+ (bold line), with $80 \%$ confidence interval (dot line), and Bmsy and maximum observed mean biomass from ASPIC (thin lines).

## H. Gulf of Maine-Georges Bank American Plaice by L. O'Brien

### 1.0 Background

This stock was last assessed in 1998 (O`Brien et al. 1999) and reviewed by the $28^{\text {th }}$ Northeast Regional SAW. Fully recruited F (ages 5-8, u) in 1997 was estimated to be 0.47 . an increase of $10 \%$ from 1996. Spawning stock biomass was $13,500 \mathrm{mt}$ in 1997 , an increase of $11 \%$ from 1996. Recruitment since 1993 has been near record low values and the 1997 estimate was the lowest in the time series (1980-1997) with a less than average recruit/SSB survival ratio. The NEFSC spring and autumn bottom trawl survey indices have declined each year since 1995 and the MADMF abundance indices have continued to decline since 1994. Recruitment indices were below average for the 1994-1996 year classes.

### 2.0 1999 Assessment Update

## The Fishery

Total commercial landings of Gulf of Maine-Georges Bank American plaice in 1998 were 3.662 mt a $7 \%$ decline from 1997 (Table H1, Figure H1). Discards were not estimated. Estimates of fishing mortality and spawning stock biomass in 1998 were derived by projection of the 1998 landings and are presented in Figures H 4 and H 5 .

## Research Survey Indices

The NEFSC autumn indices of abundance and biomass increased in 1998 and the spring 1999 indices remained stable (Table H2. Figure H2 and H3). Recruitment indices of age 1 fish from the 1998 NEFSC and MADMF autumn survey indicate that the 1997 year class is above average and similar in size to the 1992 year class.

### 3.0 Forecasts

An analysis of stock-recruit data did not indicate an increase in recruitment with an increase in spawning stock biomass. Parameters estimated by a Beverton-Holt stock recruitment model resulted in statistically infeasible parameters and the model was not accepted. The residuals from the model were lognormal. indicating that a lognormal distribution of the recruitment would provide a model to generate recruitment independent of the spawning stock biomass. This model allows for draws of recruitment over the entire distribution. i.e. between observed values, and also for values greater or less than observed. The lognormal distribution was used to generate stochastic recruitment for forecasts of landings and spawning stock size for 1999-2009. Recruitment in 1998 was not available from the previous assessment and was derived by regression analysis (RCT3) of the 1980-1997 VPA estimates of stock size at age 1 and the age 1 recruits from the NEFSC and MADMF spring and autumn survey from 1980-1998. The predicted stock size of age 1 fish in 1998 was estimated at 32,937 million fish. One thousand estimates of recruits in 1998 were then randomly generated from a normal distribution with a mean of $\ln (32,937)$ and $\mathrm{SE}=0.45$.

A projection over the short term was performed to obtain the fishing mortality in 1998-1999. Fishing mortality in 1998 was estimated as 0.32 from the projection of 1998 landings. F in 1999 was then assumed to be equal to status quo 1998 F. Spawning stock biomass in 1998 was 14.436 mt . and in 1999 declined slightly to 13.755 mt (Figures H 4 and H5).

The SFA control rule for Gulf of Maine-Georges Bank American plaice is based on $\mathrm{SSB}_{\text {MSY }}$ which is currently estimated at $24,200 \mathrm{mt}$ (Figure H6). The rule states that $\mathrm{F}_{0,1}$ will be the maximum fishing mortality threshold when the stock is above $\mathrm{SSB}_{\mathrm{MSY}}$ then decrease linearly to zero at $1 / 4 \operatorname{SSB}_{\text {MSY }}(6,050 \mathrm{mt})$. The target F will be $60 \%$ of $\mathrm{F}_{0.1}$ when the stock is above $\mathrm{SSB}_{\text {MS: }}$ and would decrease linearly to zero at $1 / 2$ of $\operatorname{SSB}_{\text {MSY }}(12,100 \mathrm{mt})$. Applying the 1999 SSB $(13.755 \mathrm{mt})$. to the control rule dictates that fishing mortality should be no higher than 0.02 (Figure H6).

Long term forecasts from 2000-2009 were performed with an $\mathrm{F}=0.02$. Landings and SSB trajectories are presented in Table H3. There is a $50 \%$ probability that $\mathrm{SSB}_{\mathrm{MSY}}$ will be achieved between 2002-2003. The projected biomass in 2009 is just above the maximum ( 49.000 mt ) SSB observed in this stock (Figure H7).

### 4.0 Sources of Uncertainty

VPA estimates of 1998 calibrated age 1 stock size were not available. The derived estimates may be optimistic.

Projections of SSB are likely to be optimistic if recruits are overestimated.
From the previous assessment (O'Brien et al. 1999):
Sources of uncertainty included:
1 the effect of not including discards from the small-mesh whiting fishery,
2 . the effect of using number of trips as an effort multiplier versus a more refined measure such as days fished and

3 the effect of expanding discard estimates derived from the small segment covered in the sea sampling program to the whole fleet.

### 5.0 References

O`Brien. L., R.K. Mayo. and C. Esteves. 1999. Assessment of American plaice in the Gulf of Maine-Georges Bank Region for 1998. NEFSC Ref. Doc. 99-09.

|  | Year | Gulf of Maine |  |  | Georges Bank |  |  |  |  | Southern New England |  |  |  | Mid - Atlantic |  |  | Grand Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | USA | Can | Total | USA | Can | SSR | Other | Total | USA | USSR | Other | Total | USA | Other | Total | USA | Other | Total |
|  | 1960 | 620 | 1 | 621 | 689 | - | - | - | 689 | - | - | - | 0 | - | - | 0 | 1309 | 1 | 1310 |
|  | 1961 | 692 | - | 692 | 830 | - | - | - | 830 | - | - | - | 0 | - | - | 0 | 1522 | 0 | 1522 |
|  | 1962 | 694 | - | 694 | 1233 | 44 | - | - | 1277 | - | - | - | 0 | - | - | 0 | 1927 | 44 | 1971 |
|  | 1963 | 693 | - | 693 | 1489 | 127 | 24 | - | 1640 | - | - | - | 0 | - | - | 0 | 2182 | 151 | 2333 |
|  | 1964 | 811 | - | 811 | 2800 | 177 | - | 11 | 2988 | - | - | - | 0 | - | - | 0 | 3611 | 188 | 3799 |
|  | 1965 | 967 | - | 967 | 2376 | 180 | 112 | - | 2668 | - | - | - | 0 | - | - | 0 | 3343 | 292 | 3635 |
|  | 1966 | 955 | 2 | 957 | 2388 | 242 | 279 | 1 | 2910 | - | - | - | 0 | - | - | - 0 | 3343 | 524 | 3867 |
|  | 1967 | 1066 | 6 | 1072 | 2166 | 203 | 1018 | 10 | 3397 | - | - | - | 0 | 4 | - | 4 | 3236 | 1237 | 4473 |
|  | 1968 | 904 | 5 | 909 | 1695 | 173 | 193 | 5 | 2066 | 637 | 145 | - | 782 | 18 | 2 | 20 | 3254 | 523 | 3777 |
|  | 1969 | 1059 | 7 | 1066 | 1738 | 71 | 63 | 17 | 1889 | 505 | 349 | - | 854 | 130 | - | 130 | 3432 | 507 | 3939 |
|  | 1970 | 895 | - | 895 | 1603 | 92 | 927 | 658 | 3280 | 88 | 18 | 40 | 146 | 8 | - | 8 | 2594 | 1735 | 4329 |
|  | 1971 | 648 | 5 | 653 | 1511 | 36 | 228 | 296 | 2071 | 11 | 112 | 206 | 329 | 6 | 2 | 8 | 2176 | 885 | 3061 |
|  | 1972 | 569 | - | 569 | 1222 | 22 | 358 | - | 1602 | 3 | 71 | - | 74 | - | - | 0 | 1794 | 451 | 2245 |
|  | 1973 | 687 | - | 687 | 910 | 38 | 289 | - | 1237 | 5 | 158 | - | 163 | - | - | 0 | 1602 | 485 | 2087 |
|  | 1974 | 945 | 2 | 947 | 1039 | 27 | 16 | 2 | 1084 | 92 | 4 | - | 96 | - | - | 0 | 2076 | 51 | 2127 |
|  | 1975 | 1507 | - | 1507 | 913 | 25 | 148 | - | 1086 | 3 | - | - | 3 | - | - | 0 | 2423 | 173 | 2596 |
|  | 1976 | 2550 | - | 2550 | 948 | 24 | 3 | - | 975 | 10 | - | - | 10 | 1 | - | 1 | 3509 | 27 | 3536 |
| $\stackrel{\sim}{\sim}$ | 1977 | 5647 | - | 5647 | 1408 | 35 | 50 | - | 1493 | 6 | 78 | - | 84 | 7 | - | 7 | 7068 | 163 | 7231. |
| 6 | 1978 | 7287 | 30 | 7317 | 2193 | 77 | - | - | 2270 | 15 | - | - | 15 | 8 | - | 8 | 9503 | 107 | 9610 |
|  | 1979 | 8835 | - | 8835 | 2478 | 23 | - | - | 2501 | 13 | - | 7 | 20 | 4 | - | 4 | 11330 | 30 | 11360 |
|  | 1980 | 11139 | - | 11139 | 2399 | 43 | - | 5 | 2447 | 10 | - | - | 10 | 1 | - | 1 | 13549 | 48 | 13597 |
|  | 1981 | 10327 | 1 | 10328 | 2482 | 15 | - | 2 | 2499 | 26 | - | 2 | 28 | 46 | - | 46 | 12881 | 20 | 12901 |
|  | 1982 | 11147 | - | 11147 | 3935 | 27 | - | 1 | 3963 | 35 | - | 2 | 37 | 9 | - | 9 | 15126 | 30 | 15156 |
|  | 1983 | 9142 | 7 | 9149 | 3955 | 30 | - | - | 3985 | 40 | - | - | 40 | 4 | - | 4 | 13141 | 37 | 13178 |
|  | 1984 | 6833 | 2 | 6835 | 3277 | 6 | - | - | 3283 | 17 | - | - | 17 | 7 | - | 7 | 10134 | 8 | 10142 |
|  | 1985 | 4766 | 1 | 4767 | 2249 | 40 | - | - | 2289 | 12 | - | - | 12 | 2 | - | 2 | 7029 | 41 | 7070 |
|  | 1986 | 3319 | - | 3319 | 1146 | 34 | - | - | 1180 | 4 | - | - | 4 | 3 | - | 3 | 4472 | 34 | 4506 |
|  | 1987 | 2766 | - | 2766 | 1032 | 48 | - | - | 1080 | 2 | - | - | 2 | 1 | - | 1 | 3801 | 48 | 3849 |
|  | 1988 | 2271 | - | 2271 | 1097 | 108 | - | - | 1205 | 13 | - | - | 13 | 1 | - | 1 | 3382 | 108 | 3490 |
|  | 1989 | 1646 | - | 1646 | 703 | 68 | - | - | 771 | 1 | - | - | 1 | 3 | - | 3 | 2353 | 68 | 2421 |
|  | 1990 | 1802 | - | 1802 | 639 | 51 | - | - | 690 | 2 | - | - | 2 | 2 | - | 2 | 2445 | 51 | 2496 |
|  | 1991 | 2936 | - | 2936 | 1310 | - | - | - | 1310 | 15 | - | - | 15 | 0 | - | 0 | 4261 | 0 | 4261 |
|  | 1992 | 4564 | 2 | 4566 | 1838 | - | - | - | 1838 | 10 | - | - | 10 | 4 | - | 4 | 6416 | 2 | 6418 |
|  | 1993 | 3865 | - | 3865 | 1838 | - | - | - | 1838 | 11 | - | - | 11 | 4 | - | 4 | 5718 | 0 | 5718 |
|  | 1994 | 3402 | 29 | 3431 | 1560 | . 2 | - | - | 1562 | 21 | - | - | 21 | $83^{\circ}$ | - | 83 | 5066 | 31 | 5097. |
|  | 1995 | 3123 | 3 | 3126 | 1486 | $\cdot$ | - | - | 1486 | 16 | - | - | 16 | 20 | - | 20. | 4645 | 3 | 4648 |
|  | 1996 | 2920 | 2 | 2922 | 1423 | - | - | - | 1423 | 39 | - | - | 39 | 14 | - | 14 | 4396 | 2 | 4398 |
|  | 1997 | 2331 | 65 | 2396 | 1560 | - | - | - | 1560 | 22 | - | - | 22 | 24 | - | 24 | 3937 | 65 | 4002 |
|  | 1998 |  | * |  |  |  |  |  |  |  |  |  |  |  |  |  | 3662 |  |  |

** 1994-1997 data are spatially distributed based on proportions of landings recorded by area in the VTR database and are considered provisional.

Table H 2 . Stratified mean number' and mean weight per tow ( kg ) of American plaice in NEFSC spring and autumn bottom trawl surveys, adjusted for vessel differences, in the Gulf of Maine Georges Bank area, 1963-1999.

|  | SPRING |  | AUTUMN |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number | Weight | Number | Weight |  |
| 1963 | --- | --- | 14.17 | 5.87 |  |
| 1964 | --- | --- | 8.20 | 2.84 |  |
| 1965 | --- | --- | 11.95 | 3.80 |  |
| 1966 | --- | --- | 17.78 | 4.90 |  |
| 1967 | --- | --- | 11.05 | 2.69 |  |
| 1968 | 11.36 | 3.40 | 8.61 | 2.91 |  |
| 1969 | 8.59 | 2.68 | 7.51 | 2.36 |  |
| 1970 | 5.43 | 1.81 | 6.46 | 2.01 |  |
| 1971 | 3.80 | 1.26 | 7.47 | 1.96 |  |
| 1972 | 4.28 | 1.32 | 7.44 | 1.60 |  |
| 1973 | 7.18 | 1.85 | 6.19 | 1.94 |  |
| 1974 | 8.34 | 1.94 | 6.89 | 1.42 |  |
| 1975 | 5.78 | 1.72 | 8.12 | 2.43 |  |
| 1976 | 11.85 | 3.37 | 9.98 | 2.99 |  |
| 1977. | 14.57 | 5.11 | 11.80 | 3.52 |  |
| 1978 | 10.61 | 3.82 | 15.13 | 4.66 |  |
| 1979 | 9.23 | 3.62 | 9.96 | 4.00 |  |
| 1980 | 18.34 | 4.78 | 14.24 | 5.12 |  |
| 1981 | 18.75 | 5.88 | 13.04 | 5.62 |  |
| 1982 | 11.61 | 3.80 | 5.88 | 2.49 |  |
| 1983 | 16.94 | 4.60 | 9.34 | 3.45 |  |
| 1984 | 4.10 | 1.42 | 7.12 | 2.02 |  |
| 1985 | 4.94 | 1.88 | 6.95 | 2.00 |  |
| 1986 | 3.09 | 0.92 | 5.61 | 1.56 |  |
| 1987 | 3.50 | 0.81 | 4.38 | 1.09 |  |
| 1988 | 3.58 | 0.84 | 9.69 | 1.46 |  |
| 1989 | 4.81 | 0.75 | 9.21 | 1.17 |  |
| 1990 | 5.09 | 0.75 | 15.46 | 2.90 |  |
| 1991 | 5.91 | 1.05 | 7.71 | 1.56 |  |
| 1992 | 4.11 | 1.36 | 6.31 | 1.78 |  |
| 1993 | 5.29 | 1.39 | 11.89 | 2.39 |  |
| 1994 | 4.89 | 0.85 | 18.07 | 2.67 |  |
| 1995 | 9.43 | 1.94 | 11.84 | 2.58 |  |
| 1996 | 7.83 | 1.69 | 7.58 | 2.23 |  |
| 1997 | 7.62 | 1.62 | 6.27 | 1.94 |  |
| 1998 | 4.52 | 1.11 | 9.29 | 2.22 |  |
| 1999 | 4.18 | 1.20 |  |  |  |

Table H3. Projections of landings, spawning stock biomass, and recruitment with probabilities of exceeding a SSB threshold of $24,200 \mathrm{mt}$, for American plaice, 1999-2009.

| Year | $F$ | Landings <br> OOO's mt | SSB <br> OOO's mt | $P(S S B>24,200 \mathrm{mt})$ | Recruits <br> 000 's |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |
| 1998 | 0.32 | 3.645 | 14.436 | 0.000 | 23655 |
| 1999 | 0.32 | 3.753 | 13.755 | 0.000 | 23688 |
| 2000 | 0.02 | 0.254. | 13.866 | 0.000 | 23748 |
| 2001 | 0.02 | 0.313 | 18.572 | 0.011 | 23650 |
| 2002 | 0.02 | 0.388 | 23.106 | 0.357 | 23670 |
| 2003 | 0.02 | 0.466 | 26.980 | 0.781 | 23685 |
| 2004 | 0.02 | 0.555 | 31.267 | 0.959 | 23709 |
| 2005 | 0.02 | 0.641 | 35.312 | 0.993 | 23723 |
| 2006 | 0.02 | 0.744 | 40.933 | 1.000 | 23666 |
| 2007 | 0.02 | 0.814 | 44.616 | 1.000 | 23653 |
| 2008 | 0.02 | 0.871 | 47.638 | 1.000 | 23617 |
| 2009 | 0.02 | 0.917 | 50.032 | 1.000 | 23649 |



Figure H1. Total commercial landings of Gulf of Maine-Georges Bank American plaice (Division 5Zand 6), 1960-1998.

( vessel bottom trawl surveys in the Gulf of Maine-Georges Bank region, 1963-1999.


Figure H3. Standardized stratified mean number per tow of American plaice in NEFSC spring and autumn research vessel bottom trawl surveys in the Gulf of Maine-Georges Bank region, 1963-1999.


Figure H4. Trends in total commercial landings and fishing mortality for Gulf of Maine-Georges Bank American plaice, 1980-1998.



Figure H6. Control rule and recent and projected $(1998,1999)$ stock status for Gulf of Maine-Georges Bank American plaice.


Figure H 7 . Long-term forecasts of spawning stock biomass (ages $1+$ ) with $80 \%$ confidence itervals for Gulf of Maine-Georges Bank American plaice, 1998-2009, and maximum SSB observed.

## I. Georges Bank Winter Flounder by R.W. Brown

### 1.0 Background

The Georges Bank winter flounder stock was last assessed by SAW/SARC 28 in December 1998 (Brown et al. 2000). The 1998 assessment included a catch at age from 1982-1997 and research survey indices through the U.S. and Canadian Spring 1998 surveys. It is not possible to update the stock assessment to include the 1998 catch at age or the Autumn 1998 and Spring 1999 survey numerical catch at age indices because age determinations have not been completed for 1998 commercial fishery samples. Medium term stochastic projections were generated and summarized by assuming 1998 and 1999 fishing mortality, and applying the harvest control rule to determine the corresponding level of fishing mortality for 2000 through 2008.

### 2.0 1998 Assessment

Fishery
Since the late 1960's, U.S. landings have been the dominate component of total commercial landings (Table I1; Figure I1). Canadian landings have averaged $0.1 \%$ to $2.7 \%$ of total fishery landings since 1970. The Canadian industry's interest in the Georges Bank winter flounder resource is increasing, and reported Canadian landings in 1997 reached their highest reported levels since 1966 (Table I1; Figure I1).
U.S. landings of Georges Bank winter flounder were estimated to be 1.178 mt in 1.998 based on prorations performed by the Northeast Regional Office of the National Marine Fisheries Service (Table I1). Canadian landings of Georges Bank winter flounder in 1998 were preliminarily estimated at 151 mt . Discarding of winter flounder occurs at low levels in both the otter trawl and scallop dredge fisheries; however, lack of reliable information to estimate either the magnitude or characterize the size and age distribution of discards precluded the inclusion of discards in the 1998 assessment.

Fishery Independent Information
Two stratified random bottom trawl surveys. the NEFSC Spring survey (April 1968-1998) and the NEFSC Autumn survey (October 1963-1997) were used to estimate changes in abundance (stratified mean number $\cdot$ tow $^{-1}$ ) and biomass (stratified mean weight ( kg ) tow $^{-1}$ ) of winter flounder in the Georges Bank stock area (Table I2; Figure I2). Stratified abundance and biomass indices for Georges Bank winter flounder from the U.S. Spring and Autumn surveys exhibit a considerable amount of variability but generally indicate intermediate levels of abundance from the early 1960s to early 1980s, and declining levels of abundance since the mid-1980s . Stratified mean number at age indices for the NEFSC Spring and Autumn surveys are noisy, but appear to track larger cohorts in the numbers at age matrix for the 1980, 1985, 1987, and 1994 cohorts.

The Department of Fisheries and Oceans. Canada has conducted a stratified random bottom trawl survey on Georges Bank since 1986. Winter flounder captured during the Canadian survey are counted and measured, but no ageing program exists to generate age determinations from this survey. U.S. survey and commercial age keys were used to partition stratified mean numbers at length into stratified mean numbers at age. Sufficient age determinations were available from U.S. Spring survey data to partition stratified mean numbers at length into numbers at age for fish smaller than 40 cm and U.S. commercial age keys from the first quarter of the corresponding year were applied for fish longer than 40 cm . Stratified mean number at age indices from the Canadian survey appear to track larger cohorts corresponding to the 1987, 1992. and 1994 year classes.

## Input Data and Analysis

The VPA calibration from the 1998 assessment estimated ages 2 to 6 and included the U.S. Spring indices (ages 1 to 7), the Canadian Spring indices (ages 1 to 7), and the U.S. Autumn indices (ages 1-6, lagged forward one age and one year). Precision of the 1998 stock sizes and 1997 fishing mortality and SSB estimates was derived from 1000 bootstrap simulations of the 1998 VPA formulation. A retrospective analysis of terminal year estimates of stock sizes, fully recruited fishing mortality and SSB was completed back to 1992.

### 3.0 Assessment Results

The assessment results indicate that stock numbers exceeded 25 million in the early 1980s. gradually declined to reach a low level of approximately 8.8 million in 1993, increased to 13.6 million in 1995, and have again declined to 9.6 million fish in 1997. Age 2 recruitment was relatively stable throughout the time period, but larger 1980, 1985. 1987 and 1994 year classes exceed 5 million fish at age 2 (Figure I3). Recent recruitment, as measured by the 1995 and 1996 year classes, has been among the lowest in the time series (Figure I3). There appears to be little discernable relationship between stock and recruitment over the time period analyzed in this assessment.

Spawning stock biomass declined from levels exceeding $8,000 \mathrm{mt}$ in the early 1980's to less than 2.000 mt in 1994, but increased to almost 3.700 mt in 1996 (Figure I3; Table I3). Spawning stock biomass declined slightly from almost $3,700 \mathrm{mt}$ in 1996 to 3.500 mt in 1997. In the early 1980s. spawning stock biomass consisted of a wide range of ages and the youngest mature ages ( 2 and 3 ) comprised less than $40 \%$ of the total spawning stock biomass. The age structure of the spawning stock biomass became truncated in the mid 1980s to mid 1990s, when the youngest mature ages ( 2 and 3 ) comprised $45 \%$ to $75 \%$ of the spawning stock biomass. Some broadening of the age structure of spawning stock biomass is evident after 1994, but the age structure remains truncated relative to historical levels.

From the early 1980s to the early 1990s, fully recruited fishing mortality (ages 4-6) ranged from approximately 0.5 to as high as 1.4 (Figure I4; Table I3). Fishing mortality declined sharply after 1993 and has fluctuated between 0.3 and 0.5 from 1994 to 1997.

In the terminal year of the assessment (1997). fully recruited fishing mortality was estimated to be 0.41 (Table I3). Patterns in fishing mortality appear to be reasonably well correlated with reported landings.

VPA Diagnostics
In the 1998 assessment VPA calibration, the coefficient of variation on age 2 was relatively high ( 0.52 ). while CV's on older ages were lower ( 0.41 to 0.47 for ages 3 to 6 ). The CV's on the survey q's ranged from 0.22 to 0.27 . Standardized residuals exhibited no discernable patterns. although 8 of 221 standardized residuals had values exceeding 2.0.

Uncertainty and potential bias estimates were assessed using a bootstrap analysis of the VPA calibration. One thousand bootstrap realizations were produced by randomly resampling survey residuals produced by the final ADAPT calibration. The distribution of bootstrap realizations of spawning stock biomass indicates that there is an $80 \%$ chance that the 1997 estimate of SSB is between 3.100 and $4,500 \mathrm{mt}$. The distribution of bootstrap realizations of fishing mortality suggests that there is an $80 \%$ probability that $\mathrm{F}_{1997}$ is between 0.26 and 0.45 .

Retrospective patterns for fishing mortality indicate a pattern of slightly overestimating average fishing mortality in the terminal year. The retrospective patterns for spawning stock biomass indicate that there is a tendency in the most recent years to slightly overestimate spawning stock biomass in the terminal year.

## Evaluation of the Harvest Control Rule

The MSY-based control rule for Georges Bank winter flounder adopted in Amendment 9 was derived from survey-based proxies of biomass and exploitation (Figure I5). The parameters of this control rule were revised during SAW/SARC 28 due to revised estimates of landings and a revision to the strata set used to develop survey indices for the NEFSC Spring and Autumn surveys. The revised control rule defined maximum sustainable yield as 2.700 mt , identified a threshold fishing mortality proxy ( $\mathrm{F}_{\text {MSY }}$ ) as a level of an exploitation index (catch • NEFSC Autumn survey biomass ${ }^{-1}$ ) of 1.125 , and identified a target stock biomass proxy as the NEFSC Autumn survey biomass index value of $2.73 \mathrm{~kg} \cdot \mathrm{tow}^{-1}$. Threshold F is defined as the $\mathrm{F}_{\text {MSY }}$ proxy $\left(\mathrm{F}_{\mathrm{MSY}}=1.125\right.$ ) when the NEFSC Autumn survey index is greater than $2.73 \mathrm{~kg} \cdot$ tow ${ }^{-1 .}$ and declines linearly to zero at $50 \%$ of the $\mathrm{B}_{\mathrm{MSY}}$ proxy (NEFSC Autumn survey index less than $1.37 \mathrm{~kg}^{\circ}$ tow $^{-1}$ ). The target fishing mortality proxy was estimated to be $75 \%$ of the threshold proxy value, and stock biomass proxies were established as $50 \%$ of the target $\mathrm{B}_{\mathrm{MSY}}$ proxy values. Target F is defined as $75 \%$ of the $\mathrm{F}_{\mathrm{MSY}}$ proxy ( $\mathrm{f}_{\mathrm{MSY}}=0.74$ ) when the NEFSC Autumn survey index is greater than $2.73 \mathrm{~kg} \cdot \mathrm{tow}^{-1}$ and declines linearly to zero at $1.37 \mathrm{~kg} \cdot \mathrm{tow}^{-1}$.

Using the harvest control rule to evaluate appropriate fishing mortality for 2000-2009 requires evaluation of stock conditions in the 1999 time period. Landings and stock conditions in 1999 were projected by completing a short-term projection with landings from 1998 to produce an estimate of realize $F$. This $F$ level was then used to complete a two year projection to estimate 1999 landings and 1999 mean biomass. The 1999 estimate of mean biomass ( 3.409 mt ) was
multiplied by the U.S..Autumn survey q estimate from the ASPIC surplus production run to translate the measure of mean biomass into 1999 Autumn' survey biomass index units. Because the mean biomass was available from the VPA, which incorporates recent survey points into the terminal point estimate of biomass, the point estimate rather than a three year average was used to evaluate the harvest control rule. The 1999 biomass index (0.897) was significantly lower than $1 / 2 \mathrm{~B}_{\text {MSY }}$ (Figure I5). This is consistent with the observation that the 1999 projection of mean biomass (3.409) is less than one-half of the $\mathrm{B}_{\text {MSY }}$ estimate from surplus production modeling $(11.400 \mathrm{mt})$. Application of the harvest control rule results in a fishing mortality of 0.00 for 2000-2008.

The availability of an analytical assessment for this stock provides an opportunity to update the harvest control to incorporate information produced by an analytical assessment. A revised control rule which incorporated estimates of mean biomass and F weighted by biomass would eliminate the necessity of translating between mean biomass and Autumn survey units.

### 4.0 Forecasts

Medium-term stochastic projections were performed for 1999-2008. U.S. landings in 1998 were estimated to be $1,178 \mathrm{mt}$ based on Regional Office prorations and Canadian landings in 1998 were approximately 151 mt . Total 1998 landings were used to complete a short projection to produce an estimate of realized F in 1998. This fully recruited $\mathrm{F}(0.41)$ was assumed for 1999 , and the fully recruited F resulting from application of the harvest control rule ( 0.00 ) was projected for 2000-2008.

The projections were based on a partial recruitment vector estimated as the geometric mean of 1994 to 1997 F's at age from the final VPA calibration, 1994 to 1997 arithmetic mean stock and catch weights, and the long-term (1982-1997) maturity schedule for Georges Bank winter flounder. Age 1 recruitment was estimated from the terminal year bootstrap realizations of the VPA in 1998, and recruitment in 1999-2008 was resampled from empirical age 1 recruitment estimated by the VPA calibration from 1982 to 1997.

Projections indicate that $1+$ biomass of winter flounder will increase from 3.300 mt in 1999 to 17.600 mt in 2008 (Figure I6). The biomass levels indicated in the later years of the projection exceed levels estimated through surplus production modeling. Many of the population compensatory mechanisms that would be expected at these large stock sizes (declining growth rates, delayed maturity) were not explicitly accounted for the projection model. The projection indicates that the probability that $1+$ biomass is greater than $\mathrm{B}_{\mathrm{MSY}}$ exceeds $50 \%$ by 2004 (Table I4).

The short time series of stock and recruitment data for this stock over a fairly limited range of stock sizes results in considerable uncertainty about modeling future recruitment. However, the - stock projection indicates that even at current recruitment levels, if fishing mortality were
reduced to 0.00 . stock size would increase at a rapid rate approaching $\mathrm{B}_{\mathrm{MSY}}$ levels in approximately 5-6 years.

### 5.0 Sources of Uncertainty

1 Sampling of U.S. commercial landings may be inadequate to characterize the size and age composition, particularly in the years since 1992. This leads to uncertainty in the age composition of landings in the catch at age matrix.

2 The exclusion of U.S. otter trawl and scallop dredge discards most likely results in an underestimation of fishery removals from the younger age classes (ages 0 to 3 ). Indications from both the sea sample and vessel trip record databases suggests that scallop dredge discards may have increased since the implementation of groundfish retention restrictions resulting in an underestimation of fishery removals of both younger and older age classes.

3 There is some uncertainty about the accuracy of reported Canadian landings because of the non-targeted nature of the Canadian fishery and the tendency to report landings of some flatfish species, including winter flounder, as unclassified flounders.

4 The Canadian fishery has no formal sampling program to estimate the size and age composition of Canadian landings. This assessment assumed that the size and age. composition of Canadian landings was identical to the overall size and age composition in the U.S. fishery. This assumption is sensitive to the possibility that selectivity patterns may be different between the fisheries in each country.

### 6.0 Literature Cited

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Brown. R.W. and W. Gabriel. Winter Flounder. Pages 81-84, in Status of Fishery Resources off the Northeastern United States for 1998. NOAA Technical Memorandum NMFS-NE115.

Table I1. Landings (mt) of Georges Bank winter flounder from 1962-1998 by statistical area and country.

|  | 522-525 | 5 Z (521-543) |  |  |  | 5ZE (521-526.541-543) |  |  |  | Included in <br> Assessment |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | USA | USA | Canada | USSR | Total | USA | Canad | USS | Total |  |
| 1962 |  | 6996 | 26 |  | 7022 |  |  |  |  |  |
| 1963 |  | 6911 | 120 | 19 | 7050 |  |  |  |  |  |
| 1964 | 1371 | 12656 | 146 |  | 12802 |  |  |  |  | 1517 |
| 1965 | 1176 | 10479 | 199 | 312 | 10990 |  |  |  |  | 1687 |
| 1966 | 1877 | 13807 | 164 | 156 | 14127 |  |  |  |  | 2197 |
| 1967 | 1917 | 10815 | 83 | 349 | 11247 | - |  |  |  | 2349 |
| 1968 | 1570 |  | 57 |  |  | 4346 | 59 | 372 | 4777 | 1999 |
| 1969 | 2167 |  | 116 |  |  | 6380 |  | 235 | 6615 | 2518 |
| 1970 | 2615 |  | 61 |  |  | 7020 | 64 | 40 | 7124 | 2716 |
| 1971 | 3092 |  | 62 |  |  | 1400 | 65 | 1029 | 15094 | 4183 |
| 1972 | 2805 |  | 8 |  |  | 1026 | 8 | 1699 | 11973 | 4512 |
| 1973 | 2269 |  | 14 |  |  | 4387 | 14 | 693 | 5094 | 2976 |
| 1974 | 2124 |  | 12 |  |  | 4508 | 12 | 82 | 4602 | 2218 |
| 1975 | 2409 |  | 13 |  |  | 4833 | 13 | 515 | 5361 | 2937 |
| 1976 | 1877 |  | 15 |  |  | 3732 | 11 | 1 | 3744 | 1893 |
| 1977 | 3572 |  | 15 |  |  | 5954 | 15 | 7 | 5976 | 3594 |
| 1978 | 3185 |  | 65 |  |  | 6378 | 65 |  | 6443 | 3250 |
| 1979 | 3045 |  | 19 |  |  | 6293 | 19 |  | 6312 | 3064 |
| 1980 | 3931 |  | 44 |  |  | 9941 | 44 |  | 9985 | 3975 |
| 1981 | 3993 |  | 19 |  |  | 9711 | 19 |  | 9730 | 4012 |
| 1982 | 2961 |  | 19 |  |  | 7347 | 19 |  | 7366 | 2980 |
| 1983 | 3894 |  | 14 |  |  | 8014 | 14 |  | 8028 | 3908 |
| 1984 | 3927 |  | 4 |  |  | 7574 | 4 |  | 7578 | 3931 |
| 1985 | 2151 |  | 12 |  |  | 4758 | 11 |  | 4769 | 2163 |
| 1986 | 1762 |  | 25 |  |  |  |  |  |  | 1787 |
| 1987 | 2637 |  | 32 |  |  |  |  |  |  | 2669 |
| 1988 | 2804 |  | 55 |  |  |  |  |  |  | 2859 |
| 1989 | 1880 |  | 11 |  |  |  |  |  |  | 1891 |
| 1990 | 1898 |  | 55 |  |  |  |  |  |  | 1953 |
| 1991 | 1814 |  | 14 |  |  | . |  |  |  | 1828 |
| 1992 | 1822 |  | 27 |  |  |  |  |  |  | 1849 |
| 1993 | 1662 |  | 21 |  |  |  |  |  |  | 1683 |
| 1994 | 907 |  | 65 |  |  |  |  |  |  | 972 |
| 1995 | 706 |  | 54 |  |  |  |  |  |  | 760 |
| 1996 | 1265 |  | 71 |  |  |  |  |  |  | 1336 |
| 1997 | 1287 |  | 143 |  |  |  |  |  |  | 1430 |
| 1998 | 1178 |  | 151 |  |  |  |  |  |  | 1329 |

flounder from the U.S. NEFSC Spring and Autumn, and Canadian Spring research vessel bottom trawl surveys. U.S. survey strata 01130-01220: Canadian survey strata (5Z1-5Z8). Canadian biomass indices were estimated using the stratified mean number at length and the U.S. survey length-weight regression coefficients. Door standardization coefficients of 1.46 (numbers) and 1.39 (weight) applied to U.S. surve! indices before 1985 to account for differences in catchability between survey doors (NEFSC 1991).

|  | U.S. Spring Survey |  | U.S. Autumn Survey |  | Canada Spring Survey |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Number•tow ${ }^{-1}$ | Weight (kg) tow $^{-1}$ | Number•tow ${ }^{-1}$ | Weight ( kg ) tow $^{-1}$ | $\begin{gathered} \text { Number•to } \\ w^{-1} \end{gathered}$ | Weight (kg) ${ }^{-t o w}$ |
| 1963 |  |  | 1.200 | 1.815 |  |  |
| 1964 |  |  | 1.298 | 1.822 |  |  |
| 1965 |  |  | 2.152 | 2.050 |  |  |
| 1966 |  | . ${ }^{\text {- }}$ | 5.163 | 5.655 |  |  |
| 1967 | 'Spring Surve | initiated in 1968 | 1.791 | 2.074 |  |  |
| 1968 | 2.700 | 3.114 | 1.308 | 1.072 |  |  |
| 1969 | 3.136 | 4.290 | 2.370 | 2.385 |  |  |
| 1970 | 1.864 | 2.294 | 5.620 | 6.490 |  |  |
| 1971 | 1.838 | 2.168 | 1.324 | 1.259 |  |  |
| 1972 | 4.946 | 5.321 | 1.261 | 1.580 |  |  |
| 1973 | 2.946 | 3.507 | 1.218 | 1.195 |  |  |
| 1974 | 6.049 | 5.782 | 1.193 | 1.464 |  |  |
| 1975 | 1.955 | 1.407 | 3.790 | 2.061 |  |  |
| 1976 | 4.672 | 3.012 | 5.987 | 3.925 |  |  |
| 1977 | 3.792 | 1.580 | 4.862 | 3.992 | . |  |
| 1978 | 7.068 | 5.055 | 4.056 | 3.100 |  |  |
| 1979 | 1.736 | 2.206 | 5.065 | 3.829 |  |  |
| 1980 | 3.221 | 2.801 | 1.661 | 1.865 |  |  |
| 1981 | 3.727 | 3.749 | 3.831 | 2.434 |  |  |
| 1982 | 2.295 | 1.523 | 5.301 | 2.692 |  |  |
| 1983 | 8.405 | 7.111 | 2.726 | 2.363 |  |  |
| 1984 | 5.529 | 5.604 | 3.933 | 2.445 |  |  |
| 1985 | 3.837 | 2.650 | 1.979 | 1.119 |  |  |
| 1986 | 2.003 | 1.214 | 3.575 | 2.178 | Canadian Surve | itiated in $198{ }^{-}$ |
| 1987 | 2.803 | 1.247 | 0.762 | 0.889 | 3.73 | 2.83 |
| 1988 | 2.925 | 1.648 | 4.084 | 1.273 | 2.70 | 1.65 |
| 1989 | 1.299 | 0.757 | 1.560 | 1.051 | 3.48 | 1.88 |
| 1990 | 2.803 | 1.573 | 0.498 | 0.346 | 3.29 | 1.74 |
| 1991 | 2.403 | 1.319 | 0.268 | 0.136 | 1.43 | 0.97 |
| 1992 | 1.416 | 0.898 | 0.677 | 0.384 | 2.25 | 1.39 |
| 1993 | 1.018 | 0.570 | 1.166 | 0.663 | 2.78 | 1.45 |
| 1994 | 1.292 | 0.578 | 0.870 | 0.578 | 2.45 | 0.98 |
| 1995 | 2.613 | 1.489 | 2.357 | 1.337 | 3.10 | 1.17 |
| 1996 | 2.314 | 1.504 | 1.539 | 1.756 | 2.20 | 1.12 |
| 1997 | 1.610 | 1.192 | 1.744 | 1.534 | 2.80 | 1.77 |
| 1998 | 0.762 | 0.722 | 1.784 | 1.565 | - 1.42 | 1.08 |
| 1999 | 3.83 | 3.479 | To be conduc | ed in October 1999 | 0.98 | 0.74 |

Table I3. Stock numbers (thousands), fishing mortality, and spawning stock biomass (mt) at age of Georges Bank winter flounder estimated using an ADAPT calibration.

## STOCK NUMBERS

|  | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| - | 4627 | 2725 | 6089 | 5963 | 8027 | 5307 | 9002 | 5243 |
| 1 | 8236 | 3788 | 2222 | 4986 | 4864 | 6572 | 4345 | 7370 |
| 2 | 6532 | 6424 | 2389 | 1564 | 3354 | 3381 | 4210 | 2802 |
| 3 | 3382 | 3803 | 2634 | 1441 | 654 | 1545 | 1247 | 937 |
| 4 | 7263 | 1821 | 1799 | 916 | 445 | 322 | 451 | 258 |
| 5 | 762 | 572 | 992 | 198 | 306 | 158 | 143 | 191 |
| 6 | 822 | 1453. | 1406 | 175 | 204 | 211 | 146 | 106 |
| 7 |  |  |  |  |  |  |  |  |


| $1+$ | 25624 | 20586 | 17532 | 15243 | 17854 | 17496 | 19545 | 16908 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 |
| 1 | 3327 | 4523 | 2441 | 2906 | 4813 | 6944 | 2987 | 946 | 00 |
| 2 | 4293 | 2724 | 3703 | 1998 | 2346 | 3940 | 5447 | 2445 | 774 |
| 3 | 4785 | 3248 | 1694 | 2311 | 1364 | 1439 | 2611 | 3793 | 1568 |
| 4 | 1188 | 2079 | 1510 | 703 | 858 | 590 | 936 | 1625 | 2097 |
| 5 | 307 | 368 | 842 | 578 | 167 | 480 | 313 | 549 | 797 |
| 6 | 78 | 84 | 179 | 266 | 184 | 76 | 324 | 114 | 330 |
| 7 | 28 | 133 | 117 | 75 | 164 | 81 | 258 | 110 | 122 |
| $1+$ | 14005 | 13158 | 10485 | 8837 | 9896 | 13550 | 12876 | 9582 |  |

FISHING MORTALITY

|  | 1982 | 1983 | 1.984 | 1985 | 1986 | 1987 | 1988 | 1989 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| $\ldots$ | -0.0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  |
| 1 | 0.00 | 0.05 | 0.26 | 0.15 | 0.20 | 0.16 | 0.25 | 0.24 | 0.23 |
| 2 | 0.34 | 0.69 | 0.31 | 0.67 | 0.58 | 0.80 | 1.30 | 0.66 |  |
| 3 | 0.42 | 0.55 | 0.86 | 0.97 | 0.51 | 1.03 | 1.37 | 0.92 .1. |  |
| 4 | 0.59 | 0.41 | 2.00 | 0.90 | 0.84 | 0.61 | 0.66 | 0.99 |  |
| 5 | 0.47 | 0.51 | 1.23 | .0 .97 | 0.64 | 0.97 | 1.17 | 0.96 |  |
| 6 | 0.47 | 0.51 | 1.23 | 0.97 | 0.64 | 0.97 | 1.17 | 0.96 |  |
| 7 | 0.49 | 0.49 | 1.36 | 0.95 | 0.66 | 0.87 | 1.07 | 0.96 |  |


|  | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| -0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.04 | 0.00 | 0.00 |
| 1 | 0.08 | 0.28 | 0.27 | 0.18 | 0.29 | 0.21 | 0.16 | 0.24 |
| 2 | 0.63 | 0.57 | 0.68 | 0.79 | 0.64 | 0.23 | 0.27 | 0.39 |
| 3 | 0.97 | 0.70 | 0.76 | 1.23 | 0.38 | 0.44 | $0 . .33$ | 0.51 |
| 4 | 1.09 | 0.52 | 0.95 | 0.94 | 0.59 | 0.19 | 0.81 | 0.31 |
| 5 | 1.03 | 0.69 | 0.84 | 1.13 | 0.42 | 0.32 | 0.44 | 0.41 |
| 6 | 1.03 | 0.69 | 0.84 | 1.13 | 0.42 | 0.32 | 0.44 | 0.41 |
| 7 | 1.03 | 0.64 | 0.85 | 1.10 | 0.46 | 0.32 | 0.53 | 0.41 |

Table I3. (Continued). Stock numbers (thousands). fishing mortality, and spawning stock biomass (mt) at age of Georges Bank winter flounder estimated using an ADAPT calibration.


Table I4. Projected mean biomass (mt) of Georges Bank winter flounder and the probability of exceeding the $\mathrm{B}_{\mathrm{MSY}}$ estimate of 11.400 mt from 1998 to 2009.

| Year | Mean Biomass (mt) | Probability of <br> Exceeding B <br> MSY |
| :---: | ---: | ---: |
| 1998 | 4,300 | $0.0 \%$ |
| $19 \dot{9} 9$ | 4,100 | $0.0 \%$ |
| 2000 | 5,400 | $0.0 \%$ |
| 2001 | 6,800 | $0.0 \%$ |
| 2002 | 8,100 | $3.5 \%$ |
| 2003 | 9,700 | $37.6 \%$ |
| 2004 | 11,800 | $68.7 \%$ |
| 2005 | 13,500 | $91.5 \%$ |
| 2006 | 14,800 | $97.1 \%$ |
| 2007 | 16,000 | $99.0 \%$ |
| 2008 | 16,900 | $99.6 \%$ |
| 2009 | 17.600 | $99.9 \%$ |



Fiyure I 1. Historical landings of Georges Bank winter flounder from the Georges Bank stock area from 1964 to 1998.



Figure 12. U.S. and Canadian research vessel bottom trawl survey abundance (number per tow; Panel A) and biomass (kg per tow; Panel B) for Georges Bank winter flounder, 1963-1999. Canadian weight per tow was estimated using the stratified mean number per tow at length and the U.S. survey length-weight regression equation.


Figure 1 3. Trends in spawning stock biomass (line) and age 2 recruitment (bars) estimated from Virtual Population Analysis for Georges Bank winter flounder from 1980 to 1997.


Figure 14. Trends in commercial landings (mt) and fully-recruited fishing mortality ( $F$, ages 4-6, unweighted) estimated from Virtual Population Analysis for Georges Bank winter flounder from 1982 to 1997.


Fic sre I5. Harvest control rule for Georges Bank winter flounder based on survey equivalents of MSY-based reference points.


Figure 16. Medium term projection (10-year) results for Georges Bank winter flounder. Fishing mortality was assumed to be 0.41 in 1998 and 1999 and projected at $\mathrm{F}=0.00$ for 2000-2008. Age 1 recruitment was estimated from the terminal year bootstrap realizations of the VPA in 1998, and 1999-2008 recruitment was resampled from empirical age 1 recruitment estimated by the VPA calibration from 1982-1997.

## J. Southern New England/Mid-Atlantic Winter Flounder by Mark Terceiro

### 1.0 Background

The Southern New England/Mid-Atlantic stock complex of winter flounder was last assessed by SAW 28 in December 1998, with catches through 1997 (NEFSC 1999). The assessment is for the entire stock complex, which includes several inshore spawning aggregations that individually may not demonstrate the same trend in abundance as the complex. Fully recruited (ages 4-6) fishing mortality in 1997 was estimated at 0.31 (Figure J1), corresponding to a biomass weighted $\mathrm{F}=0.24$ (given current age structure.). Mean stock biomass in 1997 was estimated to be 17.900 mt (Figure J2). In the SAW 28 assessment, $\mathrm{B}_{\text {MSY }}$ was estimated to be 27.810 mt . MSY was estimated to be $10.200 \mathrm{mt}, \mathrm{F}_{\mathrm{MSY}}$ was estimated to be biomass weighted $\mathrm{F}=0.37$. and the FMP Amendment 9 ten year rebuilding target biomass weighted fishing mortality was estimated to be $F_{\text {target10 }}=0.24$ (Figure J3) .

### 2.0 Fishery Catches and Research Survey Indices

Commercial and recreational catch was updated through 1998 (Table J1; Figure J1) and NEFSC spring and autumn survey indices were updated though spring 1999 (Table J2; Figure J4). Commercial landings were taken from NMFS NERO prorated landings reports for 1998, and were reported to be 3.240 mt in 1998. Commercial discards were assumed to be $7 \%$ of the landings, as in SAW 28 projections, and were calculated to be 231 mt for 1998. Recreational landings were taken from the MRFSS, and were estimated to be 564 mt in 1998. Recreational discards were taken from the MRFSS, using the 1997 mean weight of discards. and estimated to be 16 mt in 1998. Total landings were estimated to be 3.804 mt , total discards were estimated to be 247 mt . and total catch was estimated to be 4.051 mt in 1998 .

NEFSC survey indices show an increase in stock biomass since 1993. The NEFSC spring 1999 survey biomass index ( $1.245 \mathrm{~kg} /$ tow $)$ is the highest since 1985 ( $1.983 \mathrm{~kg} /$ tow ), and is about $25 \%$ of the time series maximum in 1981 ( $4.762 \mathrm{~kg} /$ tow). The NEFSC autumn 1998 survey biomass index ( $2.232 \mathrm{~kg} /$ tow) is about $45 \%$ of the time series maximum in 1964 ( $4.894 \mathrm{~kg} / \mathrm{tow}$ ).

### 3.0 Forecasts for 1999-2009

Forecasts were started from the January 1. 1998 stock numbers at age estimated by VPA in the SAW 28 assessment. Mean weights at age in the stock biomass. landings and discards, partial recruitment at age, proportion mature at age, and proportion landed at age were as in the SAW 28 assessment. The reported/estimated 1998 catch was input to estimate fully recruited fishing mortality in 1998. Fully recruited fishing mortality in 1999 was assumed to be the same as in 1998. Fishing mortality in 2000 and later years was input as the biomass weighted target $F(0.24)$ according to the FMP Amendment 9 ten year rebuilding control rule (Figure J3). The fully recruited F corresponding to the biomass weighted F control rule value changes over time in the forecast as stock age structure rebuilds.

Recruitment at age 1 in 1999 and later years was generated randomly from the empirical distribution of recruitment estimated by VPA for 1981-1998 (median recruitment of 23.5 million fish), as in the SAW 28 short term forecast. Attempts to model future recruitment using a Beverton-Holt stock-recruitment model provided unrealistic results. Using the parametric model. spawning stock biomass and recruits increased to abundances well above VPA estimates and corresponding historical survey indices, unless model parameters were tightly constrained ( $\mathrm{R} / \mathrm{SSB}$ ratio constrained to less than or equal to the median observed).

Forecasts indicate that fully recruited $F$ (age 4-6) in 1998 and 1999 was 0.33 . corresponding to a biomass weighted F (ages 1 and older) of 0.19 in 1998 and 1999. Increasing fishing mortality to the FMP Amendment 9 ten year rebuilding control rule biomass weighted $\mathrm{F}=0.24$ during 20002009 (corresponding to decreasing, fully recruited F from 0.43 to 0.40 ) indicates that the stock increases to slightly above $\mathrm{B}_{\mathrm{MSY}}(27,810 \mathrm{mt})$ in 2001 at $29,000 \mathrm{mt}$, stabilizing at 33.400 mt by 2009, about 20\% higher than $\mathrm{B}_{\text {MSY }}$ (Table J3; Figure J5). Landings increase to 6,100 mt in 2001. stabilizing at $7,300 \mathrm{mt}$ (with 400 mt of discards) by 2009. The projected total yield in 2009 of 7.700 mt is about $75 \%$ of MSY ( $10,200 \mathrm{mt}$ ).

### 4.0 References

NEFSC. 1999. $28^{\text {th }}$ Northeast Regional Stock Assessment Workshop ( $28^{\text {th }}$ SAW). Stock Assessment Review Committee (SARC) Consensus Summary of Assessment. NMFS/NEFSC, Woods Hole Laboratory Ref. Doc. 99-08.

Table J1. Total winter flounder recreational and commercial catch for the Southern New England/Mid-Atlantic stock complex in weight ( mt ) and numbers ( 000 s ).

| Year | Commercial <br> Landings |  | Commercial <br> Discards |  | Recreational <br> Landings |  | Recreational Discards |  | Total Catch |  | $\begin{gathered} \% \\ \text { Discards/Total } \end{gathered}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | mt | 000 s | mt | 000 s | mt | 000s | mt | 000s | mt | 000 s | mt | 000 s |
| 1981 | 11,176 | 20,705 | 1,343 | 5,123 | 3,050 | 8,089 | 88 | 437 | 15,657 | 34,354 | 9.1 | 16.2 |
| 1982 | 9,438 | 19,016 | 1,149 | 4,271 | 2,457 | 8,392 | 66 | 341 | 13,110 | 32,020 | 9.3 | 14.4 |
| 1983 | 8,659 | 16,312 | 1,311 | 5,251 | 2,524 | 8,365 | 125 | 399 | 12,619 | 30,327 | 11.4 | 18.6 |
| 1984 | 8,882 | 17,116 | 986 | 3,936 | 5,772 | 12,756 | 148 | 745 | 15,788 | 34,553 | 7.2 | 13.5 |
| 1985 | 7,052 | 14,211 | 1,534 | 4,531 | 5,198 | 13,297 | 230 | 714 | 14,014 | 32,753 | 12.6 | 16.0 |
| 1986 | 4,929 | 9,460 | 1,273 | 4,902 | 2,940 | 6,994 | 66 | 356 | 9,208 | 21,712 | 14.5 | 24.2 |
| 1987 | 5,172 | 10,524 | 950 | 3,545 | 3,141 | 6,899 | 61 | 347 | 9,324 | 21,315 | 10.8 | 18.3 |
| 1988 | 4,312 | 8,377 | 904 | 3,728 | 3,423 | 7,359 | 69. | 416 | 8,708 | 19,880 | 11.2 | 20.8 |
| 1989 | 3,670 | 7,888 | 1,404 | 5,761 | 1,802 | 3,684 | 49 | 335 | 6,925 | 17,668 | 21.0 | . 34.5 |
| 1990 | 4,232 | 7,202 | 673 | 2,567 | 1,063 | 2,485 | 31 | 201 | 5,999 | 12,455 | 11.7 | 22.2 |
| 1991 | 4,823 | 9,063 | 784 | 2,701 | 1,214 | 2,794 | 51 | 230 | 6,872 | 14,788 | 12.2 | - 1.9 .8 |
| 1992 | 3,816 | 6,759 | 511 | 1,811 | 393 | 802 | 15 | 83 | 4,735 | 9,455 | 11.1 | 20.0 |
| 1993 | 3,010 | 5,336 | 457 | 1,580 | 543 | 1,180 | 31 | 155 | 4,041 | 8,251 | 12.1 | 21.0 |
| 1994 | 2,159 | 1,948 | 304 | 344 | 598 | 1,210 | 34 | 93 | 3,095 | 3,595 | 10.9 | 12.2 |
| 1995 | 2,634 | 2,321 | 121 | 107 | 661 | 1,390 | 23 | 69 | 3,439 | 3,887 | 4.2 | 4.5 |
| 1996 | 2,781 | 2,372 | 173 | 149 | 689 | 1,555 | 64 | 168 | 3,707 | 4,244 | 6.4 | 7.5 |
| 1997 | 3,426 | 5,834 | 267 | 1,200 | 618 | 1,204 | 26 | 85 | 4,337 | 8,323 | 6.8 | 15.4 |
| 1998 | 3,240 |  | 231 |  | 564 |  | 16 |  | 4,051 |  | 6.1 |  |
| $155^{\circ}$ |  |  |  |  |  |  |  |  |  |  |  |  |

Table J2. Winter flounder NEFSC survey index stratified mean number and mean weight ( kg ) per tow for the Southern New England- Mid-Atlantic stock complex. strata set (offshore 1-12. 25. 69-76: inshore 1-29. 45-56).


Table J3. Ten year stochastic (median recruitment $=23.5$ million fish) forecast for SNE/MA winter flounder. Forecasts started from January 1, 1998 stock numbers at age estimated by SAW 28 VPA. Mean weights at age in the stock biomass. landings and discards, partial recruitment at age, proportion mature at age, and proportion landed at age were as in the SAW 28 assessment. The reported/estimated 1998 catch was input to estimate fully recruited fishing mortality in 1998. Fully recruited fishing mortality in 1999 was assumed to be the same as in 1998. Fishing mortality in 2000 and later years was input as the biomass weighted target $\mathrm{F}(0.24)$ according to the FMP Amendment 9 ten year rebuilding control rule. Median mean stock biomass in thousands of metric tons. $\mathrm{B}_{\mathrm{MSY}}=27.810 \mathrm{mt}$.

| Year | Median <br> Mean Stock <br> Biomass | Percent <br> Probability <br> Biomass > B |
| :---: | :---: | :---: |
| 1998 |  |  |
| 1999 | 22.3 | 5 |
| 2000 | 27.3 | 30 |
| 2001 | 29.0 | 48 |
| 2002 | 30.4 | 57 |
| 2003 | 31.2 | 64 |
| 2004 | 31.8 | 68 |
| 2005 | 32.4 | 71 |
| 2006 | 32.9 | 74 |
| 2007 | 33.2 | 76 |
| 2008 | 33.3 | 77 |
| 2009 | 33.4 | 77 |
|  |  | 78 |

## SNE/MA Winter Flounder Total Catch and Fishing Mortality



Figure J1. Total catch (landings and discards, thousands of metric tons) and fishing mortality rate (fully recruited F , ages 4-6) for SNE/MA winter flounder.

## SNE/MA Winter Flounder <br> Biomass and Recruitment



Figure J2. Mean stock biomass (B, ages 1-7+, thousands of metric tons), spawning stock biomass (SSB, ages 3-7+, thousands of metric tons) and recruitment (millions of fish at age 1) for SNE/MA winter flounder.

## NEFMC Amendment 9 Control Rule for SNE/MA Winter Flounder



Figure J3. NEFMC FMP Amendment 9 control rule for SNE/MA winter flounder for rebuilding to $\mathrm{B}_{\mathrm{MSY}}$, with current 1997-1999 estimates of biomass weighted F and mean stock biomass.

## SNE/MA Winter Flounder Survey Biomass Indices




Figure J4. Trends in research survey biomass indices for SNE/MA winter flounder.

## SNE/MA Winter Flounder Stochastic Projection Mean Stock Biomass



Figure J5. Ten year stochastic forecast of mean stock biomass (median [ $50^{\text {th }}$ percentile], $10^{\mathrm{th}}$ percentile, and $90^{\mathrm{th}}$ percentile). Horizontal lines are at maximum mean stock biomass estimated from ASPIC (ASPIC max B) and $\mathrm{B}_{\mathrm{MSY}}$.

## K. Georges Bank/Gulf of Maine White Hake by K.A. Sosebee

### 1.0 Background

This stock was last assessed in 1998 and reviewed at SAW 28. Fully recruited fishing mortality (ages 4-8) in 1997 was estimated to be 1.15, an increase from an average of 1.0 in 1995 and 1996. Spawning stock biomass was estimated to have declined to 2.900 mt in 1997. a decline from a recent high of $9,600 \mathrm{mt}$ in 1992. The strength of the 1993-1995 recruiting year classes was low. The 1996 year class was estimated to be an average year class over the time series ( 5.7 million). NEFSC spring and autumn research vessel bottom trawl survey indices had declined to near record low levels in 1997.

### 2.0 1999 Assessment

## Fishery

United States commercial landings of white hake increased to 2.364 metric tons (mt) in 1998, a $6 \%$ increase from 1997 (Table 1; Figure 1). Canadian landings declined to 228 mt ( $21 \%$ decline). No discard estimates were derived for 1998.

## Input Data and Analvses

The present assessment represents a one-year update to the previous assessment (NEFSC 1999). Stochastic projection software was used to estimate fishing mortality and biomass in 1998. Survivors from 1000 bootstrapped VPA outcomes from the previous assessment were used to start the projections. Survey data from the fall of 1998 and the spring of 1999 was aged using seasonal pooled age-length keys from 1982-1999. The age estimates for fall age-1 and spring age-2 were then used to derive an estimate of recruitment for the 1997 year class using RCT3. The estimate and the standard error were used to generate 1000 recruitment estimates for age 1 in the projections. The 1998 year class was estimated from resampling of the previous four years (1993-1996 year classes). Reported landings were used to generate fishing mortality in 1998 and 1999 fishing mortality was assumed to be the same as 1998.

### 3.0 Assessment Results

NEFSC research vessel bottom trawl survey abundance and biomass indices for white hake remained relatively low through autumn 1998 and spring 1999 (Table 2, Figure 2). The autumn 1998 indices declined slightly from the 1997 levels, while the spring 1999 indices increased from the 1998 levels because of the 1996 year class. Recruitment in 1997 was estimated to be 1.9 million fish. the second lowest value in the time series (Figure 4).

Fully recruited fishing mortality (ages 4-8) in 1998 is estimated to be 1.09 (Table 3. Figure 3). a slight decline from 1:15 in 1997, as reported in the previous assessment. Spawning stock biomass is estimated at 2.717 mt in 1998 , a decline from $2,945 \mathrm{mt}$ in 1997. as reported in the previous assessment (Table 3, Figure 4). The most recent high level of SSB ( 9.563 mt ) occurred in 1992. Mean biomass increased slightly to 5,471 in 1998 due to the 1996 year class and to 5.498 in 1999 (Tables 3, Figure 4). Biomass weighted fishing mortality (ages $1+$ ) has declined from 0.8 in 1996 to 0.47 in 1998 (Figure 3). Accounting for the precision in the current assessment, there is a $90 \%$ probability that fully recruited F in 1998 was greater than 0.8 . SSB in 1998 was less than 3.500 mt , and mean biomass was less than $6,800 \mathrm{mt}$.

### 4.0 Forecasts

Forecasts of stock size and landings were performed over the medium term (1999-2009) using stochastic projection software. Forecasts continued from the 1998 projection. Fully recruited fishing mortality in 1999 was assumed equal to that in 1998, given the unpredictability of likely 1999 annual landings. A Beverton-Holt stock-recruitment model was attempted for this stock. but gave results that were infeasible (mean biomass of three times the carrying capacity). Recruitment was estimated using a stochastic resampling of the observed recruitment estimates for the 1984-1996 year classes derived from the VPA. The resampling was partitioned into three time periods. For the 1999-2001 period, recruitment was resampled from only the last four years (1993-1996 year classes: GM recruitment $=2.5$ million). The second time period extended the recruitment values sampled to the 1988 year class (1988-1996 year classes: GM recruitment $=$ 4.3 million). This was used for 2002-2005. For the final time period, the whole set of recruitment values from 1984-1996 was used (1984-1996 year classes: GM recruitment $=4.6$ million). The concept for this approach is similar to that for Gulf of Maine cod. In the near term, given the low recruitments and $\mathrm{R} / \mathrm{SSB}$ ratios in three of the last four years, recruitment is more likely to be similar to these values. As the projection moves farther from the present, the likelihood of achieving different levels of recruitment are greater. The partial recruitment vector, maturity ogive, and mean weights at age are the same as those used in the yield per recruit calculation reported in NEFSC. 1999 (Table 3).

Mean biomass in 1999 was estimated at 5.498 mt . According to the SFA control rule, when mean biomass declines below 6.900 mt . fishing mortality should be zero (Figure 5). Thus. the projections were run with a fishing mortality of zero for 2000-2009. Biomass trajectories with $80 \%$ confidence limits are given in Figure 6. At a fishing mortality of zero, $\mathrm{B}_{\text {msy }}$ can be achieved with a greater than $50 \%$ probability by 2004 (Table 5. Figure 6).

### 5.0 Sources of Uncertainty

*1998 fishing mortality may be uncertain if landings áre not complete and if the PR has changed.

* The time series of recruitment values is short and may not reflect the entire range of values.

From SARC 28:

* Discards are not incorporated into the VPA catch at age.
* Red hake may be mis-identified as white hake and vice versa.
* Missing ages in the survey age/length keys were interpolated.
* White hake may move seasonally into and out of the defined stock area.


### 6.0 References

NEFSC. 1999. 28th Northeast Regional Stock Assessment Workshop (28th SAW). Stock Assessment Review Committee (SARC) Consensus Summary of Assessments. NMFSNEFSC. Woods Hole Laboratory Ref. Doc. 99-08.

Table K1. Total Landings (mt.live) of white hake by country from the Gulf of Maine to Cape
Hatteras (NAFO Subareas 5 and 6), 19641998.

|  | Canada | USA | Other | Grand <br> Total |
| :---: | :---: | :---: | :---: | :---: |
| 1964 | 29 | 3016 | 0 | 3045 |
| 1965 | 0 | 2617 | 0 | 2617 |
| 1966 | 0 | 1563 | 0 | 1563 |
| - 1967 | 16 | 1126 | 0 | 1142 |
| 1968 | 85 | 1210 | 0 | 1295 |
| 1969 | 34 | 1343 | 6 | 1383 |
| 1970 | 46 | 1807 | 280 | 2133 |
| 1971 | 100 | 2583 | 214 | 2897 |
| 1972 | 40 | 2946 | 159 | 3145 |
| 1973 | 117 | 3279 | 5 | 3401 |
| 1974 | 232 | 3773 | 0 | 4005 |
| 1975 | 146 | 3672 | 0 | 3818 |
| 1976 | 195 | 4104 | 0 | 4299 |
| 1977 | 170 | 4976 | 338 | 5484 |
| 1978 | 155 | 4869 | 29 | 5053 |
| 1979 | 251 | 4044 | 4 | 4299 |
| 1980 | 305 | 4746 | 2 | 5053 |
| 1981 | 454 | 5969 | 0 | 6423 |
| 1982 | 764 | 6179 | 2 | 6945 |
| 1983 | 810 | 6408 | 0 | 7218 |
| 1984 | 1013 | 6757 | 0 | 7770 |
| 1985 | 953 | 7353 | 0 | 8306 |
| 1986 | 956 | 6109 | 0 | 7065 |
| 1987 | 555 | 5818 | 0 | 6373 |
| 1988 | 534 | 4783 | 0 | 5317 |
| 1989 | 583 | 4548 | 0 | 5131 |
| 1990 | 547 | 4927 | 0 | 5474 |
| 1991 | 552 | 5607 | 0 | 6159 |
| 1992 | 1138 | 8444 | 0 | 9582 |
| 1993 | 1681 | 7466 | 0 | 9147 |
| 1994 | 955 | 4737 | 0 | 5692 |
| 1995 | 481 | 4333 | 0 | 4814 |
| 1996 | 372 | 3287 | 0 | 3659 |
| 1997 | 290 | 2225 | 0 | 2515 |
| 1998 | 228 | 2364 | 0 | 2592 |

Table K2 Stratified mean catch per tow in numbers and weight ( kg ) for white hake from NEFSC offshore spring and autumn research vessel bottom trawl surveys (strata 21-30.33-40), 1963-1999.

|  |  | Spring |  |  |  |  | Autumn |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
| Year | No/Tow | Wt/Tow | Length |  | No/Tow | Wt/Tow | Length |  |  |
| 1963 |  |  |  |  | 5.00 | 6.31 | 46.2 |  |  |
| 1964 |  |  |  |  | 1.77 | 4.14 | 56.3 |  |  |
| 1965 |  |  |  |  | 4.39 | 6.86 | 50.4 |  |  |
| 1966 |  |  |  |  | 6.79 | 7.67 | 45.1 |  |  |
| 1967 |  |  |  |  | 3.92 | 3.64 | 42.6 |  |  |
| 1968 | 1.60 | 1.74 | 44.1 |  | 4.24 | 4.54 | 44.9 |  |  |
| 1969 | 3.76 | 5.09 | 46.3 |  | 9.24 | 13.09 | 46.8 |  |  |
| 1970 | 5.84 | 11.86 | 52.9 |  | 8.05 | 12.82 | 51.3 |  |  |
| 1971 | 3.31 | 5.14 | 51.3 |  | 10.38 | 12.10 | 43.6 |  |  |
| 1972 | 10.18 | 12.66 | 47.3 |  | 12.52 | 13.10 | 45.2 |  |  |
| 1973 | 9.24 | 12.22 | 49.9 |  | 9.05 | 13.46 | 51.7 |  |  |
| 1974 | 8.08 | 13.99 | 55.0 |  | 5.35 | 11.00 | 54.5 |  |  |
| 1975 | 9.32 | 11.22 | 44.7 |  | 5.28 | 7.23 | 48.5 |  |  |
| 1976 | 9.98 | 17.01 | 52.7 |  | 6.04 | 10.56 | 54.7 |  |  |
| 1977 | 6.13 | 11.01 | 55.5 |  | 9.78 | 13.74 | 47.8 |  |  |
| 1978 | 3.22 | 6.14 | 51.8 |  | 7.87 | 12.54 | 50.2 |  |  |
| 1979 | 5.26 | 4.97 | 43.0 |  | 5.62 | 10.31 | 53.1 |  |  |
| 1980 | 10.38 | 13.96 | 49.7 |  | 10.86 | 16.66 | 48.8 |  |  |
| 1981 | 17.09 | 19.92 | 45.9 |  | 8.70 | 12.16 | 49.9 |  |  |
| 1982 | 6.06 | 8.91 | 51.0 |  | 1.96 | 2.11 | 46.7 |  |  |
| 1983 | 3.23 | 3.12 | 43.7 |  | 8.22 | 10.79 | 48.8 |  |  |
| 1984 | 2.75 | 4.17 | 51.4 |  | 5.32 | 8.23 | 51.9 |  |  |
| 1985 | 4.33 | 5.38 | 48.5 |  | 9.37 | 9.74 | 42.9 |  |  |
| 1986 | 8.24 | 5.61 | 40.0 |  | 14.42 | 11.56 | 41.9 |  |  |
| 1987 | 7.15 | 6.44 | 45.3 |  | 7.59 | 9.62 | 49.2 |  |  |
| 1988 | 4.52 | 3.69 | 41.9 |  | 8.12 | 9.88 | 46.1 |  |  |
| 1989 | 3.65 | 3.22 | 43.0 |  | 11.76 | 9.23 | 40.5 |  |  |
| 1990 | 11.11 | 18.37 | 53.3 |  | 13.09 | 10.58 | 41.5 |  |  |
| 1991 | 8.42 | 6.14 | 41.6 |  | 13.22 | 12.20 | 44.6 |  |  |
| 1992 | 7.59 | 7.11 | 45.1 |  | 10.16 | 11.24 | 47.7 |  |  |
| 1993 | 7.93 | 6.84 | 45.1 |  | 11.35 | 11.66 | 45.2 |  |  |
| 1994 | 4.59 | 3.17 | 40.1 |  | 8.44 | 7.02 | 42.3 |  |  |
| 1995 | 4.38 | 4.02 | 44.1 |  | 9.54 | 8.20 | 40.8 |  |  |
| 1996 | 2.87 | 3.07 | 45.9 |  | 4.52 | 6.35 | 51.2 |  |  |
| 1997 | 1.88 | 0.89 | 38.4 |  | 4.69 | 4.55 | 41.5 |  |  |
| 1998 | 2.25 | 1.09 | 37.7 |  | 4.41 | 4.27 | 44.5 |  |  |
| 1999 | 3.32 | 2.97 | 44.6 |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |

Table K3. Input data and results from stochastic projections at fishing mortality of zero.

| Age | Fish Mort Pattern | Nat Mort Pattern | Proportion Mature | Average Catch | Weights Stock |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.0000 | 1.0000 | 0.0400 | 0.199 | 0.124 |
| 2 | 0.0399 | 1.0000 | 0.2600 | 0.544 | 0.340 |
| 3 | 0.5191 | 1.0000 | 0.7000 | 1.066 | 0.756 |
| 4 | 1.0000 | 1.0000 | 0.8900 | 1.910 | 1.437 |
| 5 | 1.0000 | - 1.0000 | 0.9800 | 3.069 | 2.416 |
| 6 | 1.0000 | 1.0000 | 0.9800 | 4.393 | 3.681 |
| 7 | 1.0000 | 1.0000 | 1.0000 | 6.040 | 5.175 |
| 8 | 1.0000 | 1.0000 | 1.0000 | 7.886 | 6.910 |
| $9+$ | 1.0000 | 1.0000 | 1.0000 | 13.200 | 13.200 |


| Year | Mean Biomass <br> Median | Prob. of Exceeding <br> Bmsy |
| :---: | ---: | :---: |
| 1998 | 5.471 | 0.000 |
| 1999 | 5.498 | 0.000 |
| 2000 | 6.761 | 0.000 |
| 2001 | 10.231 | 0.003 |
| 2002 | 14.286 | 0.029 |
| 2003 | 19.297 | 0.254 |
| 2004 | 25.132 | 0.716 |
| 2005 | 33.252 | 0.985 |
| 2006 | 40.014 | 1.000 |
| 2007 | 47.189 | 1.000 |
| 2008 | 54.347 | 1.000 |
| 2009 | 61.199 | 1.000 |



Figure K1. Total landings of white hake from the Gulf of Maine to Mid-Atlantic region, 1964-1998.


Figure K2. White hake indices of biomass (top panel) and abundance (bottom panel) from the NEFSC bottom trawl spring (open squares) and autumn (solid squares) surveys in the Gulf of Maine to Northern Georges Bank region (offshore strata 21-30, 33-40).

## White Hake

Trends in Landings and Fishing Mortality


Figure K3. Total commercial landings and fishing mortality from the VPA calibration (solid thick lines) and the projection (open circle).


Figure K4. Mean biomass, spawning stock biomass and recuitment of white hake from the Gulf of Maine to Mid-Atlantic region. (Solid lines are VPA estimates. Circles are projection estimates.)


Figure K5. Harvest control rule for white hake.


Figure K6. Results from stochastic projections at fishing mortality of 0 . The dashed lines represent the 10 th and 90 th percentiles.

## Section 4. Working Group Comments and Recommendations

1. Survivorship is increasing for Georges Bank haddock. Georges Bank yellowtail and witch flounder and decreasing for cod stocks. Possible patterns seen in year of maximum recruitment and minimum recruitment among stocks may reflect the impact of environmental factors on recruitment. A suggestion was made to expand research to include more stocks and to begin a collaborative research effort with Canadian scientists.
2. Some stocks such as witch flounder and American plaice do not show a relationship between stock and recruitment, but these stocks have short time-series that do not contain sufficient contrast for these methods. Analyses were done to determine how to incorporate recruitment in the out-years of the projections. Some assessments will use a Beverton-Holt S/R model while others will re-sample observed recruitments from the empirical data.
3. The assessments for some stocks including Gulf of Maine cod and Georges Bank cod have displayed a retrospective pattern in recent years. In particular, for both cod stocks. fishing mortality on the fully recruited ages in the last year of the assessment has been underestimated since 1995. This may result from an overestimation of the stock sizes of these ages. In addition. recruitment has been underestimated somewhat. although the converged estimates still indicate very low recruitment in recent years.
4. Truncated age distributions can have a negative effect on spawning success which in turn. can have an impact on the variability of recruitment, especially the frequency of large year classes. Georges Bank haddock was cited as an example.
