

A review of methods for quantifying spatial predator-prey overlap

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distributions and predator-prey overlap for North Pacific fishery resources” to EH, RS, SB, JS, KH, MP, and MH. We thank Eastern Bering Sea ground trawl survey participants for their hard work in collecting the data used in the case study. We also thank Nick Tolimieri for comments on a previous draft.

Biosketch

Dr. Gemma Carroll is an ecologist with an interest in understanding how predator-prey dynamics are shaped by changing environments.

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A review of methods for quantifying spatial predator-prey overlap

Running title: Quantifying spatial predator-prey overlap

Abstract

Background

Studies that attempt to measure shifts in species distributions often consider a single species in isolation. However, understanding changes in spatial overlap between predators and their prey may give deeper insight into how species redistribution affects food web dynamics.

Predator-prey overlap metrics

Here we review a suite of ten metrics (range overlap, area overlap, the local index of collocation (Pianka's O), Hurlbert's index, biomass-weighted overlap, asymmetrical alpha, Schoener's D, Bhattacharyya's coefficient, the global index of collocation, and the AB ratio) that describe how two species overlap in space, using concepts such as binary co-occurrence, encounter rates, spatial niche similarity, spatial independence, geographic similarity, and trophic transfer. We describe the specific ecological insights that can be gained using each overlap metric, to determine which is most appropriate for describing spatial predator-prey interactions for different applications.

Simulation and case-study

30 We use simulated predator and prey distributions to demonstrate how the ten metrics respond
31 to variation in three types of predator-prey interaction: changing spatial overlap between
32 predator and prey, changing predator population size, and changing patterns of predator
33 aggregation in response to prey density. We also apply these overlap metrics to a case study of
34 a predatory fish (arrowtooth flounder, *Atheresthes stomias*) and its prey (juvenile walleye
35 pollock, *Gadus chalcogrammus*) in the Eastern Bering Sea, Alaska, USA. We show how the
36 metrics can be applied to understand spatial and temporal variation in the overlap of species
37 distributions in this rapidly changing Arctic ecosystem.

38

39 **Conclusions**

40 Using both simulated and empirical data, we provide a roadmap for ecologists and other
41 practitioners to select overlap metrics to describe particular aspects of spatial predator-prey
42 interactions. We outline a range of research and management applications for which each
43 metric may be suited.

44

45 **Key words:** Arrowtooth flounder, climate change, Eastern Bering Sea, predator-prey overlap,
46 species distribution models, ecosystem models, spatial overlap, cold pool, species interactions,
47 walleye pollock

48

49

50 **Introduction**

51

52 Global environmental change is causing species distributions to shift at an accelerating rate,
53 with species moving into new areas and disappearing from their former ranges (Lawing & Polly
54 2011). However, species are not moving one at a time or in isolation. Rather, the distributions
55 of entire communities are shifting, and there is a growing need to understand how these
56 changes affect trophic interactions (Tylianakis et al. 2008; Gilman et al. 2010). For example,
57 climate-induced changes in habitat and phenology may drive differential responses in the
58 distributions of predators and their prey (Durant et al. 2007; Schweiger et al. 2012; Pinsky et al.

59 2016; Siddon et al. 2016). This may result in increased (Vors & Boyce 2009) or reduced
60 (Schweiger et al. 2008; Asch 2015) predation opportunities. These increased matches or
61 mismatches in species distributions may have cascading effects including directional changes in
62 the abundance of predator or prey populations (Durant et al. 2007; Northfield et al. 2017),
63 predators switching prey types (Latham et al. 2013), or changes in competition dynamics within
64 predator guilds (Northfield et al. 2017). Understanding how predator-prey interactions respond
65 to external pressures is therefore essential for predicting how ecosystems will respond to
66 change (Araujo & Luoto 2007), and for making informed ecosystem-based management
67 decisions (Pikitch et al. 2004).

68
69 Indices that summarise spatial overlap between co-occurring species provide simple metrics
70 that can describe the potential strength of their ecological interactions (Hurlbert 1978). When
71 applied to predictions of species distributions, information on changes in overlap can increase
72 our ability to project realistic ecological outcomes for interacting species (Guisan & Thuiller
73 2005; Schweiger et al. 2012), thereby better informing resource management and spatial
74 conservation planning. Overlap metrics can also add time-varying and spatially-explicit
75 attributes to ecological and ecosystem models (e.g. size-spectrum models and multispecies
76 stock assessment models). Within these models, spatial interactions between predator and
77 prey species are often poorly resolved at resolutions most relevant to their ecology, which may
78 complicate interpretations of how spatial overlap influences ecosystem dynamics (Greer &
79 Woodson 2016). Similarly, temporal changes in overlap between life stages of cannibalistic or
80 competitive conspecifics could bias estimates of density-dependent controls on species
81 productivity (Sigler et al. 2016). Incorporating an overlap metric can increase the capacity of
82 these models to more accurately predict the impacts of future environmental change on
83 ecosystem function (Greer & Woodson 2016).

84
85 A number of overlap metrics have been developed for ecological applications. These
86 metrics have been applied to diverse ecological questions, including examining niche
87 equivalency of species in environmental space (Warren et al. 2008; Broennimann et al. 2011),

88 overlap of animal home ranges (Fieberg & Kochanny 2005), overlap in dietary niche among
89 competitors (Woodward & Hildrew 2002), and changes in resource partitioning among species
90 after environmental perturbations (Fattorini et al. 2017). Several reviews of overlap metrics
91 have discussed their mathematical and biological properties, and investigated sources of bias
92 and error (e.g. Hurlbert 1978; Krebs 1989, Rödder & Engler 2011). However, there remains a
93 clear need to understand the specific ecological insights that can be gained using each overlap
94 metric, in order to determine which is most appropriate for describing spatial predator-prey
95 interactions for different applications.

96
97 Here we provide a review of metrics that can be used to quantify spatial overlap between two
98 species. To determine how the metrics resolve spatial predator-prey dynamics, we examine the
99 behaviour of the metrics when applied to simulations of varying predator and prey density
100 distributions. Further, we apply the metrics to spatial interactions between a predatory fish
101 (arrowtooth flounder, *Atheresthes stomias*) and its prey (juvenile walleye pollock, *Gadus*
102 *chalcogrammus*) in the Eastern Bering Sea, Alaska, USA. This case study demonstrates how the
103 metrics can track changes in species overlap driven by differential responses to environmental
104 variability. By summarising the properties of available overlap metrics and illustrating their
105 behaviour in response to various ecological scenarios, we aim to assist ecologists to select
106 appropriate metrics with which to quantify spatial predator-prey overlap for different purposes.

107

108 **Methods**

109

110 **Overlap metric description**

111

112 Here we present ten overlap metrics and their ecological interpretations (Table 1), separating
113 them into descriptive categories to aid with ecological understanding. This is not a
114 comprehensive suite of all available metrics, but rather represents a spectrum of metrics that
115 are commonly used in ecology to measure horizontal overlap between two species
116 distributions, including metrics that are particularly relevant for understanding predator-prey

117 interactions. The metrics we have chosen provide population-level estimates of overlap for
118 non-continuous survey data. Other scales of spatial overlap between species such as patchiness
119 were not included within the scope of this paper (but see e.g. Fauchald et al. 2002; Saraux et al.
120 2014; Greer et al. 2016).

121
122 *Binary co-occurrence*

123 For some applications, a metric of the co-occurrence of two species may be sufficient to
124 describe changes in spatial overlap (Selden et al. 2018). These metrics are particularly useful
125 where data on occurrence but not biomass are available, such as for rare species. Co-
126 occurrence can be described in multiple ways, including the proportion of one species' range
127 where the other species also occurs ("range overlap"; Kernohan et al. 2001; Araújo et al. 2011),
128 or the proportion of a pre-defined study area where both species co-occur ("area overlap";
129 Saraux et al. 2014). These metrics both range from 0 to 1. Binary co-occurrence metrics are
130 simple to interpret but do not discriminate between areas with high and low population
131 density, and therefore provide limited insight into the potential strength of interactions
132 between species where they co-occur.

133
134 *Encounter*

135 Where biomass or abundance data are available from standardised surveys, an alternative
136 approach to the co-occurrence metrics is the "local index of collocation" (also "Pianka's O",
137 Pianka 1973; Bez & Rivoirard 2000). This metric assesses the co-occurrence of two populations
138 using the proportion of their total biomass found at each sampling point, and determines the
139 degree of correlation between two density distributions. This can broadly be thought of as
140 describing the ratio of the probability of interspecific (predator-prey) encounters to the
141 probability of all intraspecific encounters (the product of both predator-predator and prey-prey
142 encounters).

143
144 Similar metrics extend this approach, with the "asymmetrical alpha" reflecting the ratio of the
145 probability of predator-prey encounters to the probability of prey-prey encounters only (Levins

146 1968). This asymmetrical approach gives insight into the amount of pressure exerted by the
147 predator on the prey. Another asymmetrical metric (“biomass-weighted overlap”) is similar to
148 the asymmetrical alpha but uses raw biomass rather than the proportion of total biomass
149 located at each point. These values can be used to give insight into the magnitude of predator
150 biomass that can influence prey, or scaled to the maximum values for predator and prey in a
151 given year, to keep the range between 0 and 1. Biomass-weighted overlap may be useful when
152 estimating potential consumptive demand (Chasco et al. 2017).

153
154 “Hurlbert’s index” (Hurlbert 1978) can be used to assess whether two species use space in
155 proportion to its availability. Accounting for spatial availability is particularly relevant when
156 considering spatial overlap across an arena where spatial units differ in area, as it explicitly
157 accounts for resources of unequal sizes. Hurlbert’s overlap is 0 when species do not share space
158 at all, 1 where both species occupy space in proportion to its availability and > 1 where species
159 demonstrate preferences for particular spatial areas and these preferences coincide. Hurlbert’s
160 index is also the only encounter metric that explicitly accounts for the size of the area occupied
161 by either species (rather than the size of the entire domain), making it sensitive to changes in
162 the total area occupied by predator and prey.

163

164 *Spatial niche similarity*

165 “Schoener’s D” determines whether there is equivalency between the spatial niches occupied
166 by two species (Schoener, 1970). This metric can be used either on modelled probability of
167 occurrence data, or estimates of abundance or density distributions to determine whether
168 species occupy space in a similar way. In contrast to Hurlbert’s index, a potential drawback of
169 Schoener’s D as a predator-prey overlap metric is that it will return a high value of overlap in
170 cases where both species co-occur across large areas at low densities or probabilities of
171 occurrence, where their interaction is unlikely (i.e. their niche similarity is high because the
172 absolute difference between their densities across these locations is low). Equally, this metric
173 returns a low value of overlap in cases where predation pressure might be high, if there are
174 many predators in some areas relative to the number of prey.

175

176 *Spatial independence*

177 The “Bhattacharyya coefficient” is a statistical approach that can quantify the affinity between
178 two probability density functions of spatial use, against the null assumption that the
179 distributions are independent (Bhattacharyya, 1943). This approach is not derived from
180 ecological theory but can be interpreted as assessing whether two populations use space
181 independently of one another (Fieberg & Kochanny 2005). In a similar fashion to Schoener’s D,
182 the Bhattacharyya coefficient can be applied to modelled probability of occurrence data, or on
183 biomass-density values.

184

185 *Geographic similarity*

186 An alternative to measuring overlap as the interaction between populations at the grid cell level
187 is to determine how geographically similar two distributions are across the entire study area
188 (“global index of collocation”; Bez & Revoirard 2000). This can be done using geographic
189 coordinates weighted by biomass, to determine the proximity of the centres of gravity of the
190 two populations given the dispersion or ‘inertia’ of individuals around that point. This approach
191 describes spatial overlap at a regional scale, which may be useful for understanding broad
192 patterns of overlap between two species, rather than fine-scale interactions. For example, a
193 high global index of collocation for two species may not translate to high encounter or
194 consumption rates, if the species do not also co-occur at finer spatial scales (Saraux et al. 2014).

195

196 *Trophic transfer*

197 Variability in predator density that can be explained by local prey density is a useful way of
198 characterising overlap, as it accounts for potential trophic transfer between species (“AB ratio”;
199 Greer & Woodson 2016). An AB ratio of 0 indicates that the mean densities of predator and
200 prey across the whole survey area accurately represent predator-prey encounters, while a
201 value of 1 indicates 100% higher production of the predator as a result of fine-scale spatial
202 overlap with prey, assuming a “Holling Type I” functional response. Negative values can indicate
203 different habitat preferences between species, or predator avoidance by the prey, where -1

204 represents complete avoidance. This metric provides insight into how trophic processes are
205 likely to translate local prey abundance into predator production.

206

207 **Metric responses to simulated predator-prey interactions**

208

209 *Simulated populations*

210 To illustrate how various overlap metrics respond to shifting predator-prey dynamics, we
211 simulated interacting predator and prey populations on a 200 x 100 gridded spatial domain. We
212 included two ways of manipulating predator density in relation to prey density: changing
213 “predator population size” where predator density increased uniformly, and changing
214 “predator aggregation”, where predator biomass shifted from areas where there were no prey
215 into areas of high prey density, keeping total predator abundance constant across the domain.
216 Although these ecological processes occur at different timescales in the real world, we use the
217 resulting distributions as snapshots to determine how different types of spatial interactions
218 between predator and prey influence metric values.

219

220 On the gridded arena, we simulated unconditional Gaussian densities of a prey population
221 ($D_{prey,i}$) with moderate spatial autocorrelation (patchiness), using a spherical variogram model
222 (sill = 0.35, nugget = 0.05, range = 10) in the R package *gstat* (Pebesma 2004). We took the
223 exponent of these values to obtain a lognormal distribution, to match right-skewed densities
224 observed in natural populations. We then allowed predator density ($D_{pred,i}$) to be influenced by
225 three processes:

$$D_{pred,i} = \alpha D_{prey,i} + \beta \bar{D}_{prey,s} + \delta_i$$

226 Where α is a coefficient (0-2) that controls the relationship between predator density and prey
227 density (D_{prey}) in a given cell (i). β is a coefficient (0-2) that controls the response of predator
228 density to mean prey density within a specified search area, representing the “aggregation”
229 response of predators to mean prey density (\bar{D}_{prey}) within a neighbourhood of s grid cells
230 around the cell of interest (i). The spatial variance term (δ_i) is a second log-density value
231 generated using a spherical variogram model (sill = 0.2, nugget = 0.1, range = 9), added to

232 represent spatial processes such as habitat selection that influence predator density
233 independently of prey density.

234

235 *Scenarios of predator-prey interaction*

236 1. Δ Overlap window

237 After generating predator and prey distributions across the whole 200 x 100 spatial domain
238 using the equation above, we clipped the spatial domain into two distinct distributions of
239 predator and prey, with dimensions of 100 x 100 each (Figure 1). We designated species as
240 absent in the portion of the spatial domain outside their distribution. Keeping the size of the
241 predator and prey distributions constant at 100 x 100, we then manipulated the proportion of
242 the prey's distribution that was shared by the predator, from 0 (where the two species shared
243 no common cells) to 1 (where the two species overlapped completely), in 10% spatial
244 increments. While changing the size of this overlap window, we left α ("population size"
245 parameter) and β ("aggregation response" parameter) constant at 1, and fixed s ("search area")
246 to a window of 9 cells (3 x 3), centred on the cell of interest.

247

248 2. Δ Predator population size

249 To simulate the influence of uniformly increasing predator density, we manipulated α from 0 to
250 2 at intervals of 0.2. We left the overlap window constant at 0.5, search area at $s = 9$ and
251 aggregation response at $\beta = 1$.

252

253 3. Δ Aggregation response of predator to prey

254 To simulate a change in the aggregative effect of prey on predator density, we manipulated the
255 predator aggregation response from weaker to stronger (β from 1 to 2 at intervals of 0.1)
256 within the overlap window, where predator and prey interacted directly. Simultaneously, we
257 manipulated β from 1 to 0 in the area outside the overlap window, to ensure that the overall
258 sum of predator densities were unchanged. For example, where predator density increased
259 relative to prey density by a factor of 1.1 in the overlap window, it decreased by a factor of 0.9
260 across the rest of the predator's range, to simulate the predator moving from areas of no prey

261 to areas of high prey density. We fixed the overlap window at 0.5, “bottom up” density
262 response at $\alpha = 1$, and the search window s at 9.

263

264 *Application of overlap metrics*

265 We applied the suite of ten overlap metrics to the predator and prey distributions in each of
266 the three scenarios (changing overlap window, changing predator population size, and changing
267 aggregation response) to characterise patterns of overlap. Because the overlap metrics were
268 influenced by the random variation in biomass distribution in each spatial model run, we
269 calculated the mean value for each overlap metric for each of 500 iterations of predator and
270 prey distributions under each of the three scenarios.

271

272 Functions to implement the predator-prey overlap metric equations in R are included as
273 supplementary material in the online version of this article.

274

275 **Metric responses to predator-prey dynamics in the Eastern Bering Sea**

276

277 To evaluate the performance of the overlap metrics on real ecological data, we present a case
278 study of spatial dynamics between a predator (arrowtooth flounder, *Atheresthes stomias*) and
279 prey (juvenile walleye pollock, *Gadus chalcogrammus*) in the Eastern Bering Sea, Alaska.

280 Arrowtooth flounder is a bottom-dwelling flatfish that has seen an 8-fold increase in abundance
281 on the Eastern Bering Sea shelf over the past 36 years (Wilderbuer et al. 2010). It is a key
282 predator of juvenile walleye pollock, an abundant forage fish in the Eastern Bering Sea
283 ecosystem (Springer 1992). The distributions of both species are influenced by the Eastern
284 Bering Sea “cold pool” (Kotwicki & Lauth 2013), a water mass on the seafloor that is defined by
285 temperatures $< 2^{\circ}\text{C}$ and varies greatly in size between years as a function of sea ice extent
286 during the previous winter (Stabeno et al. 2008). The study species have contrasting responses
287 to the cold pool: juvenile pollock are more tolerant of cold water (Kotwicki & Lauth 2013), and
288 may use the cold pool as a refuge to avoid predation (Hollowed et al. 2012), while arrowtooth

289 flounder are constrained to warmer water and generally avoid the cold pool (Ciannelli et al.
290 2012, Spencer et al. 2016).

291

292 Previous work has demonstrated a potential increase in overlap between these two species as
293 biomass of flounder increases and extent of the cold pool decreases (Hunsicker et al. 2013).
294 This results in increased predation pressure by arrowtooth flounder on juvenile pollock,
295 affecting pollock recruitment and abundance, creating a complex management issue for the
296 Eastern Bering Sea (Hunsicker et al. 2013; Spencer et al. 2016). This relatively well-characterised
297 example of changes in predator-prey dynamics over a 36-year time series provides an
298 opportunity to explore the behaviour of the overlap metrics when estimating real world
299 changes in spatial predator-prey dynamics.

300

301 *Arrowtooth and pollock distributions*

302 Annual summer distributions of juvenile walleye pollock (< 25 cm, approximate age classes 1-2)
303 and adult arrowtooth flounder (> 30 cm) in the Eastern Bering Sea were estimated between
304 1982 and 2017, during annual summer fisheries-independent trawl surveys conducted by the
305 National Oceanic and Atmospheric Administration. These length classes reflect the size above
306 which arrowtooth flounder begins to predate intensively on pollock (Livingston et al. 2017).

307 Catch per unit effort (CPUE; number of fish per km², Alverson & Pereyra 1969) was determined
308 for approximately 376 stations across the survey region using a standard trawl net (83 – 112
309 eastern otter trawl) towing for ~ 30 min at 1.54 m/s. Length classes were determined by
310 measuring the fork length of a subsample of fish from each tow, and expanding this to the
311 entire catch in a given tow based on the ratio of sampled weight to total towed weight for each
312 species.

313

314 We estimated the distribution and density of juvenile pollock and arrowtooth flounder using
315 the Eastern Bering Sea shelf CPUE data, while accounting for sampling effort that was uneven in
316 space and time. To do this, we applied two separate delta-generalised linear mixed models
317 using the Vector Autoregressive Spatiotemporal (VAST) package in R (<https://github.org/james->

318 thorson/VAST). VAST has become a widely used tool for fisheries scientists and managers, and
319 we used default model settings for model parameterisation (Thorson and Barnett 2017;
320 Thorson 2019). The delta model framework jointly estimated the probability of occurrence
321 (binomial distribution) and the positive catch rate (log-normal distribution) for each species for
322 each survey year (1982 – 2017). Model parameters were estimated for 250 locations (“knots”)
323 that were selected by applying a k-means clustering algorithm to the CPUE data to identify
324 geographic locations that reflect survey sampling intensity (Thorson and Barnett, 2017).
325 Environmental covariates were temperature at depth of trawl (°C; measured by trawl gear) and
326 bottom depth (m), and were included as quadratic forms in the model to allow for non-linear
327 responses (Thorson et al. 2017). Year was treated as a fixed effect (default VAST setting), while
328 spatial variation (which does not change among years) and spatio-temporal variation (which is
329 estimated independently in each year) were treated as random effects described by a Gaussian
330 process. These random fields allow modelling of multi-dimensional factors that are not directly
331 included in the model, but that affect the density and distribution of the modelled species.
332 Including spatial variation in the model allowed for correlations in CPUE between nearby
333 locations, with spatial correlation declining with increasing distance. Species density was
334 predicted at each knot by multiplying the probability of occurrence with positive catch rate
335 estimates. Density estimates for each knot were then multiplied by the knot area (km²) to
336 create annual surfaces of species abundance across the entire Eastern Bering Sea shelf.

337

338 In order to create “absences” for the binary cooccurrence metrics, we determined that species
339 were absent at knots where probability of occurrence was less than the lower quartile of
340 probability of occurrence values across the total sampled area for that species in a given year.
341 This approach to defining absences based on the distribution of probability values results in
342 lower bias than using an arbitrary probability threshold such as 0.5 (Liu et al. 2005).

343

344 *Arrowtooth and pollock overlap*

345 To illustrate differences in how the metrics characterise predator-prey overlap, we applied the
346 suite of ten metrics to the estimated density surfaces for arrowtooth flounder and juvenile

347 pollock for 2012 and 2016, years when the cold pool extent was high and low respectively.
348 Given the contrasting preferences of each species for the cold pool, overlap between flounder
349 and pollock might be expected to be low in 2012 and higher in 2016 (Hunsicker et al. 2013). We
350 present spatially-explicit estimates of each overlap metric, by decomposing the global metric
351 values into their grid cell-level components (i.e. we map the value for each cell without
352 integrating or taking means across the whole spatial domain). For the global index of
353 collocation, we map the position of the centre of gravity and inertia axes. Along with spatially-
354 explicit estimates, we show total metric values. We also present the full annual time series
355 (1982 – 2017) of overlap values for each metric.

356

357 To visualise how spatially-explicit overlap related to cold pool extent in 2012 and 2016, we
358 mapped the position of the cold pool (bottom waters $< 2^{\circ}\text{C}$) from the bottom temperatures
359 measured during the trawl surveys. We used ordinary kriging in the R package *gstat* (Pebesma
360 2004) to estimate these temperatures at the same knots as the species abundance data.

361

362 **Results**

363

364 **Metric responses to simulated predator-prey interactions**

365

366 *Sensitivity to changes in spatial overlap*

367 Most overlap metrics demonstrated an increase in response to a larger window of overlap
368 between the predator and prey populations (Figure 2A). These responses were predominantly
369 linear. However, the global index of collocation demonstrated a sigmoidal curve, and Hurlbert's
370 index reached an asymptote as the overlap window shared by the distributions neared one. The
371 AB ratio was by far the most sensitive metric to changes in the size of the overlap window
372 because it has the largest range. This simulation showed that when the distributions reached
373 complete overlap, predator density was four times greater in areas where it overlapped with
374 prey, relative to its mean density across the whole arena (where both species were largely
375 absent).

376

377 *Sensitivity to changes in predator population size*

378 Most overlap metrics increased slightly as predator density increased relative to prey density, in
379 the absence of an aggregative response. These increases generally reached an asymptote as the
380 ratio of predator density to prey density neared one (Figure 2B). Exceptions were the area and
381 range overlaps and the global index of collocation, which remained constant as both the area
382 occupied by each species and their centres of gravity remained the same. Again, the AB ratio
383 was the most sensitive in this scenario as predator density increased in areas where it
384 overlapped with prey, relative to its mean across the domain.

385

386 *Sensitivity to changes in aggregative response of predator to prey*

387 Most overlap metrics increased slightly in response to changes in the aggregative response of
388 predator to prey within the area of overlap (Figure 2C). Unlike in the previous scenario, the
389 global index of collocation increased because the centre of gravity of the predator shifted
390 incrementally towards that of the prey as its biomass in the overlap window increased.
391 Hurlbert's index was sensitive to changes in this parameter, as the predator distribution
392 became increasingly less uniform and coincided more with the spatial niche occupied by the
393 prey. The AB ratio was also sensitive to changes in this parameter because predator density
394 increased in areas where it overlapped with prey, relative to its mean across the domain. The
395 biomass-weighted overlap did not vary in response to this scenario because predator density is
396 scaled to its maximum value across the range, which did not increase. The binary co-occurrence
397 metrics did not change in response to this scenario either, as the size of the overlap window
398 was held constant.

399

400 **Metric responses to predator-prey dynamics in the Eastern Bering Sea**

401

402 *Pollock and arrowtooth overlap*

403 The values of all metrics showed that the overlap between arrowtooth flounder and juvenile
404 walleye pollock was low in 2012, when the cold pool covered most of the shelf area (Figure 3).

405 However, the overlap metrics demonstrated some differences in patterns of spatially-explicit
406 overlap. Centre of gravity and inertia for pollock (key components of the global index of
407 collocation) showed that its distribution was centred in the middle of the shelf, in the cold pool.
408 By contrast, the distribution of arrowtooth flounder was centred on the lower portion of the
409 shelf, outside the cold pool. The co-occurrence metrics showed that during this year, the
410 species co-occurred across the centre of the shelf, but did not co-occur in shallower waters due
411 to the absence of arrowtooth flounder from this area, or in the southeast portion of the shelf
412 due to the absence of pollock from this area. Most other metrics (asymmetrical overlap,
413 Bhattacharyya's coefficient, biomass-weighted overlap, Hurlbert's index and the local index of
414 collocation) showed that the highest area of overlap occurred in the northwest corner of the
415 Eastern Bering Sea shelf, where relatively high densities of both flounder and pollock coincided.
416 The AB ratio returned mostly negative values, indicating that there was general avoidance
417 between juvenile pollock and arrowtooth, probably caused by their different relationships with
418 the cold pool rather than direct avoidance of arrowtooth by juvenile pollock. Values of the AB
419 ratio were lower in areas where arrowtooth density was most negatively associated with
420 pollock density (e.g. where flounder density was high but pollock density was low). The highest
421 values of the AB ratio were around 0, in places where the densities of both species were
422 predicted to be low. Similarly, Schoener's D showed higher values in areas where the difference
423 in the proportion of flounder and pollock abundance was lower (i.e. where niche similarity was
424 high), including large portions of the shelf where both species were present at very low
425 densities.

426

427 During 2016, an unusually warm year in the Eastern Bering Sea, overlap between arrowtooth
428 flounder and juvenile pollock was much higher than in 2012 (Figure 4). The centre of gravity
429 and inertia of the distributions showed a shift westward by pollock toward the reduced cold
430 pool area. Arrowtooth flounder shifted further up onto the shelf as it exploited a greater
431 portion of available shelf habitat due to the absence of the cold pool, and potentially
432 experienced a density-dependent expansion (Spencer 2008). The co-occurrence metrics
433 showed the least change between the two years, with only slightly more overlap in the middle

434 portion of the shelf in 2016, where arrowtooth flounder had expanded its occupation. The
435 encounter metrics showed a small region of intense overlap in the westernmost part of the
436 shelf on the periphery of the cold pool, with spatially-explicit metric values in those cells an
437 order of magnitude greater in 2016 than in 2012. In several cells in this part of the shelf region,
438 the AB ratio indicated that the density of arrowtooth flounder was 2.5 times greater than the
439 mean, potentially as a result of its overlap with juvenile pollock.

440

441 The 36-year time series showed substantial variability in the values of the metrics between
442 years (Figure 5). Perhaps unsurprisingly, the binary cooccurrence metrics showed the smallest
443 range of variation across the time series, with changes of approximately 10% in the amount of
444 the shelf area occupied by both arrowtooth flounder and juvenile pollock. The most sensitive
445 metrics included the global index of collocation, which showed a relative shift in the weighted
446 centres of gravity of both species of approximately 20% of the total metric range. A general
447 increasing trend in overlap was seen in some metrics, including Bhattacharyya's coefficient,
448 Schoener's D and the global index of collocation. For most metrics, the last three years of the
449 time series (2015 – 2017) showed higher overlap than the first three years (1982 – 1984), and
450 overlap was relatively low and high in 2012 and 2016, respectively.

451

452 **Discussion**

453

454 Here we demonstrated the properties of a suite of overlap metrics, and showed how they can
455 describe different types of predator-prey interactions. Below we briefly review the specific
456 ecological insights that we gained from the metrics using the simulated populations and the
457 Eastern Bering Sea case study. Based on these insights, we discuss various applications for
458 which the metrics may be suitable and include a decision tree to help readers select a metric
459 based on data type and ecological question (Figure 6).

460

461 *Insights into ecological interactions*

462 The overlap metrics give a range of insights into spatial interactions between predators and
463 their prey. This is important because there are different types of predator-prey interactions
464 that practitioners may wish to quantify, such as total predation pressure exerted by a predator
465 on a prey population (Spencer et al. 2016), spatial hotspots of predation (Eero et al. 2012), or
466 productivity of a predator population that can be attributed to its overlap with a key prey
467 species (Greer & Woodson 2016). Understanding the specific insights that the metrics give into
468 predator-prey interactions can allow appropriate metrics to be selected for different
469 applications (Figure 6), and for metric values to be interpreted appropriately. Furthermore,
470 tracking changes in metric values through time can give important insight into how shifts in the
471 spatial distribution of interacting species might be altering components of ecosystem function
472 (Tylianakis et al. 2008; Gilman et al. 2010).

473
474 Using our simulation and case study, we show that the binary co-occurrence metrics (area and
475 range overlap) provide a simple and interpretable way of measuring spatial overlap between
476 two species. However, these metrics cannot be used to quantify fine-scale interspecific
477 interactions such as predation, which are a function of factors including species density and
478 aggregation patterns (Hurlbert 1978). Specifically, because species can be designated “present”
479 even at very low densities, the co-occurrence metrics are likely to overestimate the probability
480 of interspecific interactions. In most cases, using habitat models to calculate a spatially-explicit
481 probability of occurrence surface provides more information than simply a “presence” or
482 “absence”, and can deal with biases associated with the detection of species or variability in
483 sampling. Estimated probability values can then be fed into overlap metrics such as Schoener’s
484 D or Bhattacharyya’s coefficient, giving insight into the relative preferences of both species for
485 shared spatial resources (Fieberg & Kochanny 2005).

486
487 Many studies have demonstrated the efficacy of Schoener’s D for understanding niche overlap
488 between species in environmental space (e.g. Warren et al. 2008; Broennimann et al. 2012). We
489 took a purely spatial approach to understanding overlap between predator and prey without
490 explicitly testing the underlying environmental mechanisms driving their distributions. Used in

491 this way, Schoener's D provides insight into whether species share preferences for particular
492 areas, which is important for understanding whether they might be affected in similar ways by
493 anthropogenic or environmental processes. Bhattacharyya's coefficient is not derived from
494 either ecological or spatial theory, however it provides an objective statistical approximation of
495 whether two populations use space independently of one another. Although Schoener's D and
496 Bhattacharyya's coefficient quantify similarities and differences in space use between two
497 populations, they are not designed to give insight into the strength of potential interactions
498 between two species. They may therefore be more appropriate for quantifying overlap in
499 general, rather than as tools to understand specific elements of spatial predator-prey dynamics.

500
501 The encounter metrics (asymmetrical alpha, biomass-weighted overlap, Hurlbert's index, and
502 the local index of collocation) provide the most intuitive definition of overlap as a proxy for
503 predator-prey interactions. The insights from each of the encounter metrics are similar,
504 however Hurlbert's index is sensitive to both the size of the area over which two species occur,
505 as well as variability in the size of spatial sampling units. Explicitly accounting for changes in the
506 size of the area occupied by predator and prey incorporates useful information on species'
507 range expansions or contractions, and better captures the potentially increasing impacts of a
508 predator on its prey as it occupies a larger proportion of the prey's range (Hurlbert 1978).

509 510 *Applications of overlap metrics*

511 There are many applications for which overlap metrics can provide important information
512 about spatial relationships between predators and their prey. The choice of overlap metric for
513 each application depends on both the data type, and the scale of inference that is required
514 (Figure 6).

515
516 The binary co-occurrence metrics describe broad spatial patterns relating to the potential for
517 two populations to be in the same area. This may be desirable for some applications where only
518 occurrence data are available, and a simple approach is required for defining spatial
519 boundaries. For example, binary metrics may be appropriate for conservatively managing

520 interactions with rare species (Hazen et al. 2018), or projecting spatial overlap under future
521 environmental conditions where the precision of estimated distribution shifts is assumed to be
522 low (Selden et al. 2018). Within this category of metrics, the choice of whether the area or
523 range overlap metric is preferred depends on whether a study aims to determine the
524 directional influence of one species on another (range overlap), or the overlap between two
525 species across a given spatial area (area overlap) (Figure 6). For example, the range overlap
526 metric could be used to quantify how much of a predator's future range will overlap with that
527 of its prey, using predictions of species distributions made onto climate forecasts (e.g.
528 Schweiger et al. 2008; Selden et al. 2018). The area overlap metric could be used to quantify
529 the proportion of a pre-defined management area (e.g. park, region, state or continent) that
530 might continue to see the co-occurrence of two species under future conditions.

531
532 The encounter metrics and the AB ratio can be used to add spatial information to non-spatial
533 models. For example, non-spatial ecosystem models (e.g. *Ecosim*) aggregate information about
534 species' biomass, and calculate estimates of consumption and mortality across a whole region.
535 However, these models often assume constant proportions of prey biomass available to a
536 predator (e.g. 100%), which may result in overestimates of consumption rates if two species do
537 not overlap at ecologically relevant spatial or temporal scales (Greer & Woodson 2016).
538 Conversely, consumption rates can be underestimated in cases where prey is highly aggregated
539 and therefore more readily accessible to predators, such as at fronts in the open ocean (Bost et
540 al. 2009), or at water sources in terrestrial systems (de Boer et al. 2010). The encounter metrics
541 and the AB ratio provide information about correlations between the densities of two species
542 across sampling points, which relates to the probability of their interaction or production. These
543 metrics therefore provide a useful time-varying estimate of the potential strength of predator-
544 prey interactions that can be included in ecological models.

545
546 Spatially, the encounter metrics can illustrate areas of more or less intense interaction between
547 predator and prey. In our Eastern Bering Sea case study, the northwest portion of the survey
548 area demonstrated high overlap between arrowtooth flounder and juvenile pollock using the

549 encounter metrics, with overlap increasing in this area under warmer conditions (Figures 3 and
550 4). This information may be useful for managing resources in a spatially-explicit way, in cases
551 where mitigation of the influence of one species on another is desired for ecosystem-based
552 management. For example, targeted culling of a 'problem' predator in areas of high overlap (i.e.
553 high putative predation pressure) may prove an efficient and cost-effective means of boosting
554 the abundance of a prey species that is commercially important, or of conservation concern
555 (Burrows et al. 2003; Persson et al. 2007; Eero et al. 2012).

556
557 The global index of collocation can be applied to define overlap at the broad scale of the ranges
558 of stocks, populations or species; the scales at which spatial conservation and management
559 decisions are usually made. Furthermore, the centre of gravity and inertia from which the
560 global index of collocation is calculated provide simple and interpretable spatial metrics that
561 can aid in understanding the mechanisms underlying changes in patterns of spatial overlap. For
562 example, these can be used to highlight differential rates of poleward shifts by predator and
563 prey in response to climate change (Le Roux & McGeoch 2008). Unlike any of the other metrics
564 examined in this paper, the global index of collocation does not include information on co-
565 occurrence or the correlation of species' biomass at the grid-scale level, making it less useful for
566 understanding interactions between species at scales that are more relevant to their ecology.
567 However, the global index of collocation can provide a useful complement to the encounter
568 metrics, to understand processes governing the overlap of species at nested spatial scales
569 (Saraux et al. 2014).

570

571 ***Overlap between arrowtooth flounder and juvenile pollock***

572 The metrics that we investigated gave new insight into changes in spatial overlap between
573 juvenile walleye pollock and adult arrowtooth flounder, a predator that has been growing in
574 abundance in the Eastern Bering Sea over the past 30 years. We showed that spatial overlap
575 between these two species was very low in 2012, when the presence of an intense Eastern
576 Bering Sea cold pool restricted flounder from moving up onto the shelf. We then showed that in
577 an anomalously warm year (2016), overlap estimated using all 10 metrics was much higher than

578 in 2012, mirroring previous work showing an increase in predation pressure by arrowtooth
579 flounder on juvenile pollock associated with warm conditions (Hunsicker et al. 2013; Spencer et
580 al. 2016).

581
582 By mapping areas of high and low overlap of these species, we show how the metrics can
583 identify important shared habitat. We also highlight areas that may be of management interest
584 during the stanzas of anomalously warm temperatures that have increasingly been impacting
585 the Eastern Bering Sea ecosystem (Stabeno et al. 2017). The full 36-year time series of overlap
586 between juvenile pollock and flounder showed an overall increasing trend in overlap for some
587 metrics. This trend is of concern to managers, as juvenile walleye pollock is a species of great
588 commercial importance for the United States, and predation pressure may have an increasing
589 effect on the population as the Eastern Bering Sea warms (Hunsicker et al. 2013; Spencer et al.
590 2016). This case study highlights how overlap metrics can be used to track species interactions
591 both in space and through time under varying environmental conditions.

592

593 **Conclusions**

594 The diverse suite of overlap metrics examined in this paper quantify spatial predator-prey
595 interactions, and can track how these interactions change through time. Our simulations and
596 case study show that no single metric emerges as being most useful across all scenarios.
597 Instead, we recommend that the overlap metric(s) chosen for a particular study should reflect
598 the type of data available, and the desire to understand particular elements of ecological and
599 spatial relationships between species (Figure 6). In many cases, employing a combination of
600 several metrics may deliver the most comprehensive assessment of spatial predator-prey
601 overlap. For example, the global index of collocation could be chosen to give insight into broad
602 patterns of distribution, Schoener's D to understand niche equivalency across habitat types that
603 may drive overlap, and Hurlbert's index to estimate inter-specific encounter, accounting for
604 variability in spatial resource availability. In such a combination, these metrics provide
605 complementary information regarding shared space use and probability of interaction between
606 predators and prey at nested scales of distribution and behaviour.

607

608 **Data accessibility**

609 We used bottom trawl data collected by the Eastern Bering Sea bottom trawl survey, publicly
610 available at:

611 http://www.afsc.noaa.gov/RACE/groundfish/survey_data/data.htm

612

613 Functions to implement the predator-prey overlap metric equations in R are included as
614 supplementary material in the online version of this article.

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776

777 **Figure and table legends**

778

779 **Table 1:** Ecological description of each predator-prey overlap metric, ecological questions that
780 they are suitable for, equations, data type that they can input, whether the metric returns a
781 symmetrical result for predator and prey, whether the metric ranges from 0 to 1, and citations.
782 In the equations, $pred_i$ and $prey_i$ are densities of predator and prey in a given area, p_{pred_i} and
783 p_{prey_i} are the proportion of the total number of predator and prey in a given area, A_{pred} and
784 A_{prey} are the total area occupied by the predator and prey, $A_{pred,prey}$ is the area occupied by both
785 species, $A_{occupied}$ is the total area occupied by at least one of the species, A_{total} is the total size of
786 the study area, CG is the centre of gravity and I is inertia.

787

788 **Table 2:** Model parameter estimates and significance for the probability of occurrence and
789 positive catch rate of pollock and arrowtooth flounder. μ is the mean and gives the average
790 results for all years between 1982 - 2017; * indicates only some years were significant; σ is the
791 standard deviation of the spatial and spatiotemporal processes. Modelled probability of
792 occurrence and positive catch rate were estimated using a Vector Autoregressive
793 Spatiotemporal (VAST) model, parameterized with a log-normal distribution, with spatial and
794 spatiotemporal variation (FieldConfig=c(1,1,1,1)), and environmental covariates (temperature
795 and depth) as quadratic functions.

796

797 **Figure 1:** Schematic showing three scenarios of spatial interaction between simulated
798 populations of predator and prey: S1) incremental change in the area of overlap between
799 predator and prey (dashed lines), with predator density and aggregation response held
800 constant; S2) incremental change in predator density in response to prey density, with the
801 overlap window and aggregation response held constant; S3) incremental change in the
802 aggregation response of the predator to prey (density of predator increases in relation to prey
803 density only in the overlap window, while decreasing proportionally outside the overlap
804 window), with size of the overlap window and overall predator density held constant. In each
805 panel, the predator is on the left and the prey on the right. The prey is shown in light grey, and
806 the shade of the predator is manipulated to show increases or decreases in density. Areas

807 where the predator and prey overlap are depicted using shades that are intermediate between
808 the two densities, and hatched areas are where both species are absent.

809

810 **Figure 2:** Overlap metric behaviour in response to different scenarios of predator-prey
811 interaction: A) change in area of overlap; B) change in predator density (α is a coefficient (0-2)
812 that controls predator density in relation to prey density), and C) changes in aggregative
813 response (β is a coefficient (0-2) that controls aggregation response of the predator such that
814 predator biomass moves from areas of no prey into areas of high overlap with prey). Metric
815 values are means taken over 500 simulations of predator-prey distribution.

816

817 **Figure 3:** A) Cold pool extent on the Eastern Bering Sea shelf in 2012, a cold year (grey areas
818 reflect temperatures $> 2^{\circ}\text{C}$); B) estimate of arrowtooth flounder (predator) density distribution
819 in 2012 and C) estimate of juvenile walleye pollock (prey) distribution in 2012. Densities at
820 locations where the probability of occurrence was less than the lower quartile across the survey
821 area were deemed to be absences and are greyed out. Spatially explicit overlap calculated by
822 the 10 overlap metrics (AB ratio, asymmetrical overlap, Bhattacharyya coefficient, biomass-
823 weighted overlap, Hurlbert's index, local index of collocation, Schoener's D, area and range
824 overlap) are shown below. Total values of overlap for each metric for this period are displayed
825 in the top right corner of each map.

826

827 **Figure 4:** A) Cold pool extent on the Eastern Bering Sea shelf in 2016, an anomalously warm
828 year (grey areas reflect temperatures $> 2^{\circ}\text{C}$); B) estimate of arrowtooth flounder (predator)
829 density distribution in 2016 and C) estimate of juvenile walleye pollock (prey) distribution in
830 2016. Densities at locations where the probability of occurrence was less than the lower
831 quartile across the survey area were deemed to be absences and are greyed out. Spatially
832 explicit overlap calculated by the 10 overlap metrics (AB ratio, asymmetrical overlap,
833 Bhattacharyya coefficient, biomass-weighted overlap, Hurlbert's index, local index of
834 collocation, Schoener's D, area and range overlap) are shown below. Total values of overlap for
835 each metric for this period are displayed in the top right corner of each map.

836

837 **Figure 5:** 36-year time series (1982 – 2017) of overlap between juvenile walleye pollock and
838 arrowtooth flounder in the Eastern Bering Sea, Alaska, calculated using ten overlap metrics (AB
839 ratio, asymmetrical overlap, Bhattacharyya’s coefficient, biomass-weighted overlap, global
840 index of collocation, Hurlbert’s index, local index of collocation, Schoener’s D, area overlap and
841 range overlap). 2012 and 2016 (highlighted in spatial analyses) are shown with red triangles.

842

843 **Figure 6:** A decision tree to help readers select a predator-prey overlap metric based on
844 considerations such as the type of species distribution data available and the type of predator-
845 prey interactions that are of interest. Colours represent the metric categories, with red =
846 “spatial independence”, dark blue = “niche similarity”, green = “binary co-occurrence”, orange
847 = “geographic similarity”, yellow = “trophic transfer”, light blue = “encounter”.

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Metric	Metric type	Description	Suitability	Equation	Data type	Symmetry	0-1	Reference
Area overlap	Binary co-occurrence	Measures the proportion of all sampled locations across a pre-defined area, where species co-occur	Suitable for estimating area of co-occurrence across a pre-defined region	$A_{pred,prey}/A_{total}$	Occurrence	Yes	Yes	Saraux et al. 2014
Range overlap	Binary co-occurrence	Measures the proportion of one species' range where the other co-occurs	Suitable for estimating influence of one species on another based on their co-occurrence	$A_{pred,prey}/A_{prey}$	Occurrence	No	Yes	Araújo et al. 2011
Local index of collocation	Encounter	Measures co-occurrence by estimating the correlation of predator and prey densities	Suitable for estimating encounter rates	$\frac{\sum_i^n (p_{pred_i} * p_{prey_i})}{\sqrt{\sum_i^n p_{pred_i}^2 \sum_i^n p_{prey_i}^2}}$	Biomass	Yes	Yes	Pianka 1973; Bez & Rivoirard 2000
Asymmetrical alpha	Encounter	Measures the competitive pressure of predator on prey	Suitable for estimating encounter rates relative to prey encounter	$\frac{\sum_i^n (p_{pred_i} * p_{prey_i})}{\sqrt{\sum_i^n p_{prey_i}^2}}$	Biomass	No	Yes	Levin 1968
Biomass-weighted overlap index	Encounter	Measures amount of predator biomass interacting with prey (scaled to maximum prey density)	Useful where relative biomass of predator and prey is of interest	$\frac{\sum_i^n (pred_i/\max(pred) * prey_i/\max(pre))}{\sum_i^n prey_i/\max(pre)}$	Biomass	No	No	J Thorson unpublished
Hurlbert's index	Encounter	Measures interspecific encounter rate between predator and prey	Suitable for estimating encounter rates where resources vary in abundance	$\sum_i^n \left(\frac{p_{pred_i} * p_{prey_i}}{A_i/A_{occupied}} \right)$	Biomass	Yes	No	Hurlbert 1978
Schoener's D	Niche equivalency	Measures how equally predator and prey use space relative to its availability	Suitable to understand niche equivalency in spatial resource use	$1 - 0.5 * \sum_i^n p_{pred_i} - p_{prey_i} $	Biomass or occurrence	Yes	Yes	Schoener 1970
Bhattacharyya's Coefficient	Spatial independence	Measures statistical affinity between two distributions	Suitable for estimating independence in space use between two populations	$\sum_i^n \sqrt{p_{pred_i} * p_{prey_i}}$	Biomass or occurrence	Yes	Yes	Bhattacharyya, 1943

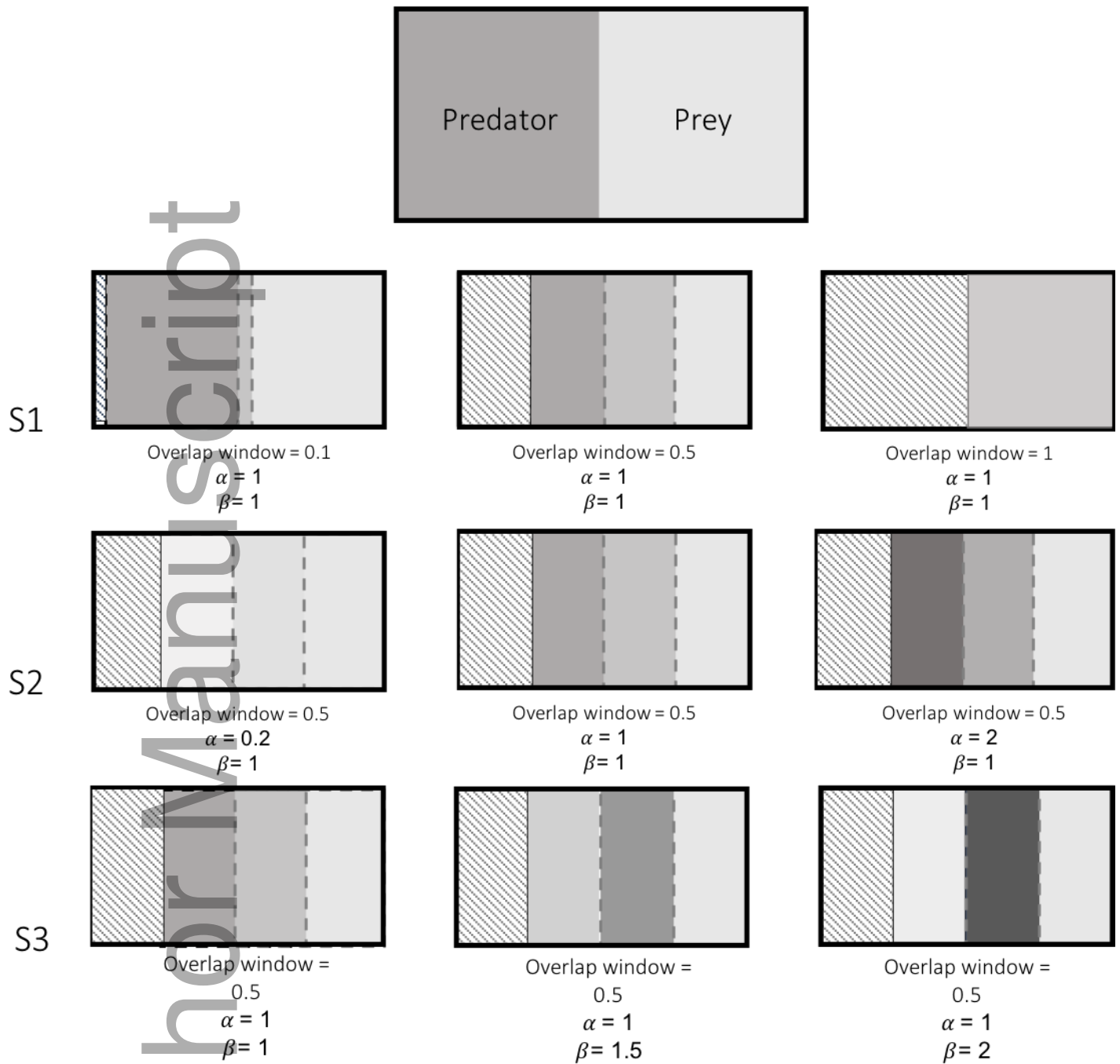
Global index of collocation	Geographic similarity	Measures geographic distinctness by comparing centres of gravity of populations with the dispersion of sampled individuals	Suitable for estimating spatial overlap at a regional scale	$\frac{1 - \Delta CG_{pred,prey}^2}{\Delta CG_{pred,prey}^2 + I_{pred} + I_{prey}}$ <p>where:</p> $CG_{pred} = \frac{\sum_i i * pred_i}{pred_i}$ <p>and:</p> $I_{pred} = \frac{\sum_i^n (i - CG)^2 * pred_i}{pred_i}$	Biomass	Yes	Yes	Bez & Rivoirard 2000
AB ratio	Trophic transfer	Measures predator production that can be attributed to spatial overlap with prey	Suitable for understanding trophic implications of fine-scale predator-prey overlap	$\frac{pred_i - \overline{pred} * prey_i - \overline{prey}}{\overline{pred} * \overline{prey}}$	Biomass	No	No	Greer & Woodson 2016

848 **Table 1**

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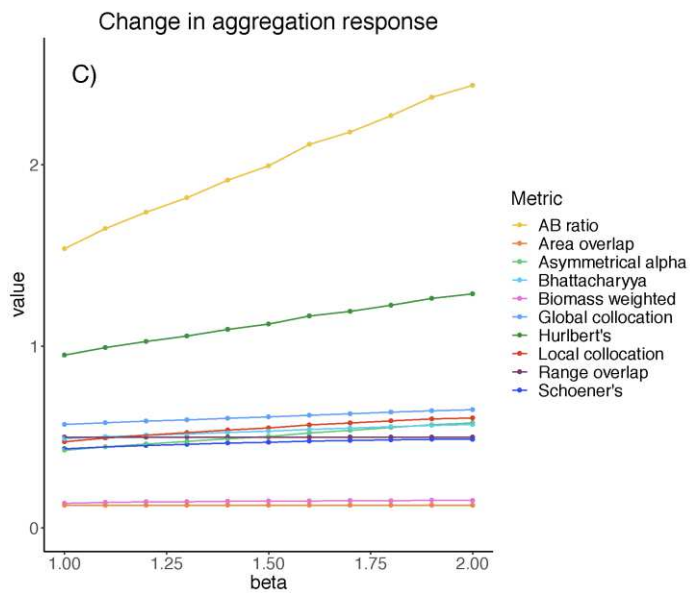
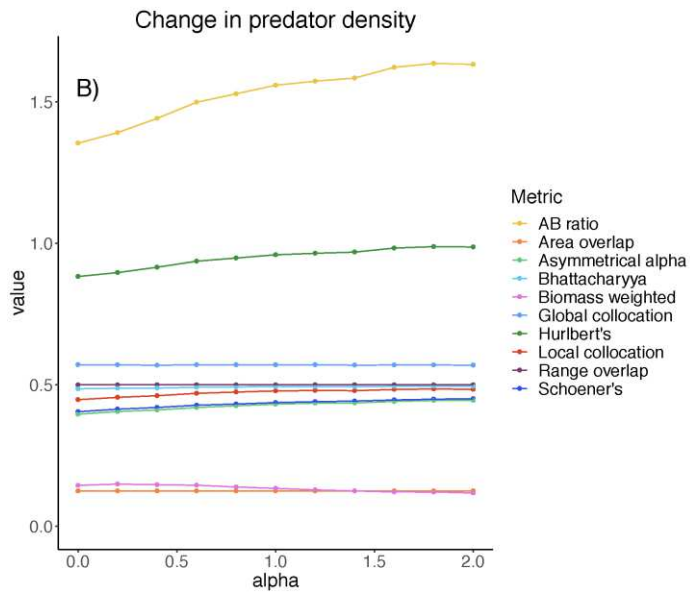
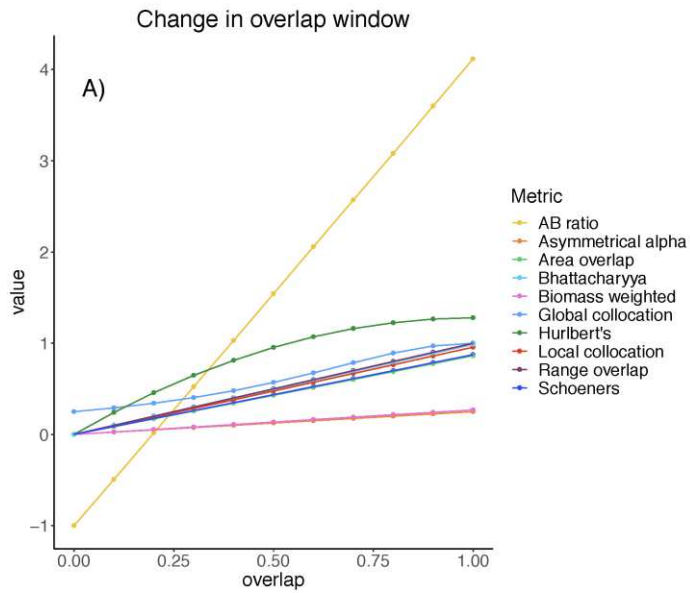
Parameter	Pollock Occurrence		Pollock Catch Rate		Flounder Occurrence		Flounder Catch Rate	
	Estimate	Significance	Estimate	Significance	Estimate	Significance	Estimate	Significance
Year	$\mu = 1.8$	$p < 0.05^*$	$\mu = 2.70$	$p < 0.001$	$\mu = -0.38$	$p < 0.05^*$	$\mu = 1.13$	$p < 0.05^*$
Temperature	-0.06	$p = 0.48$	0.27	$p < 0.001$	2.66	$p < 0.001$	1.40	$p < 0.001$
Temperature ²	0.07	$p = 0.06$	-0.03	$p = 0.2$	-1.14	$p < 0.001$	-0.52	$p < 0.001$
Depth	-0.11	$p = 0.49$	-0.26	$p < 0.05$	3.49	$p < 0.001$	1.28	$p < 0.001$
Depth ²	-0.42	$p < 0.001$	-0.28	$p < 0.001$	-0.89	$p < 0.001$	-0.32	$p < 0.001$
Spatial variation	$\sigma = 1.01$	$p < 0.001$	$\sigma = 0.61$	$p < 0.001$	$\sigma = 1.48$	$p < 0.001$	$\sigma = 0.73$	$p < 0.001$
Spatiotemporal variation	$\sigma = 1.02$	$p < 0.001$	$\sigma = 0.88$	$p < 0.001$	$\sigma = 1.05$	$p < 0.001$	$\sigma = 0.52$	$p < 0.001$



849
850 **Figure 1**

851

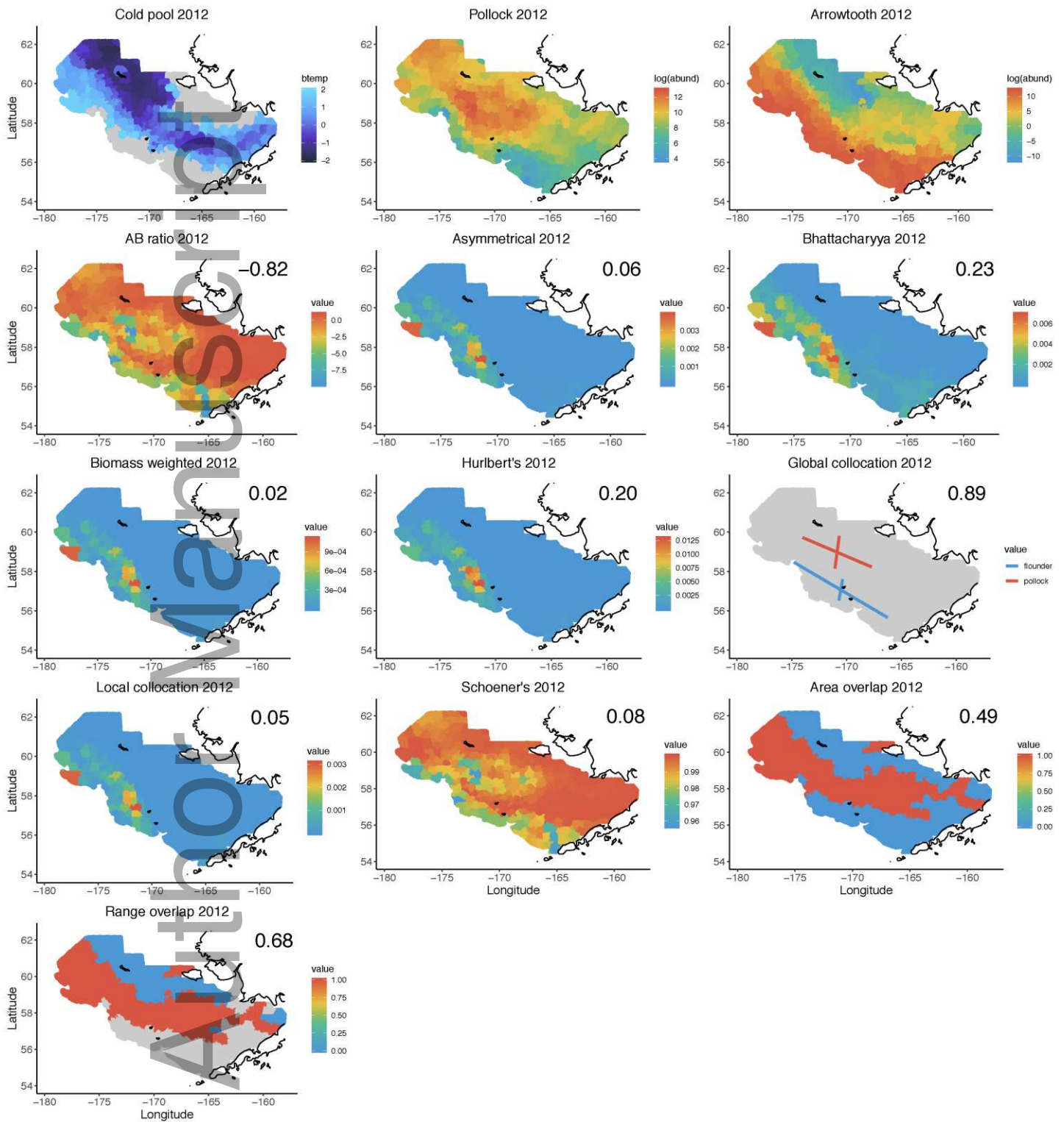
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854

Figure 2

855



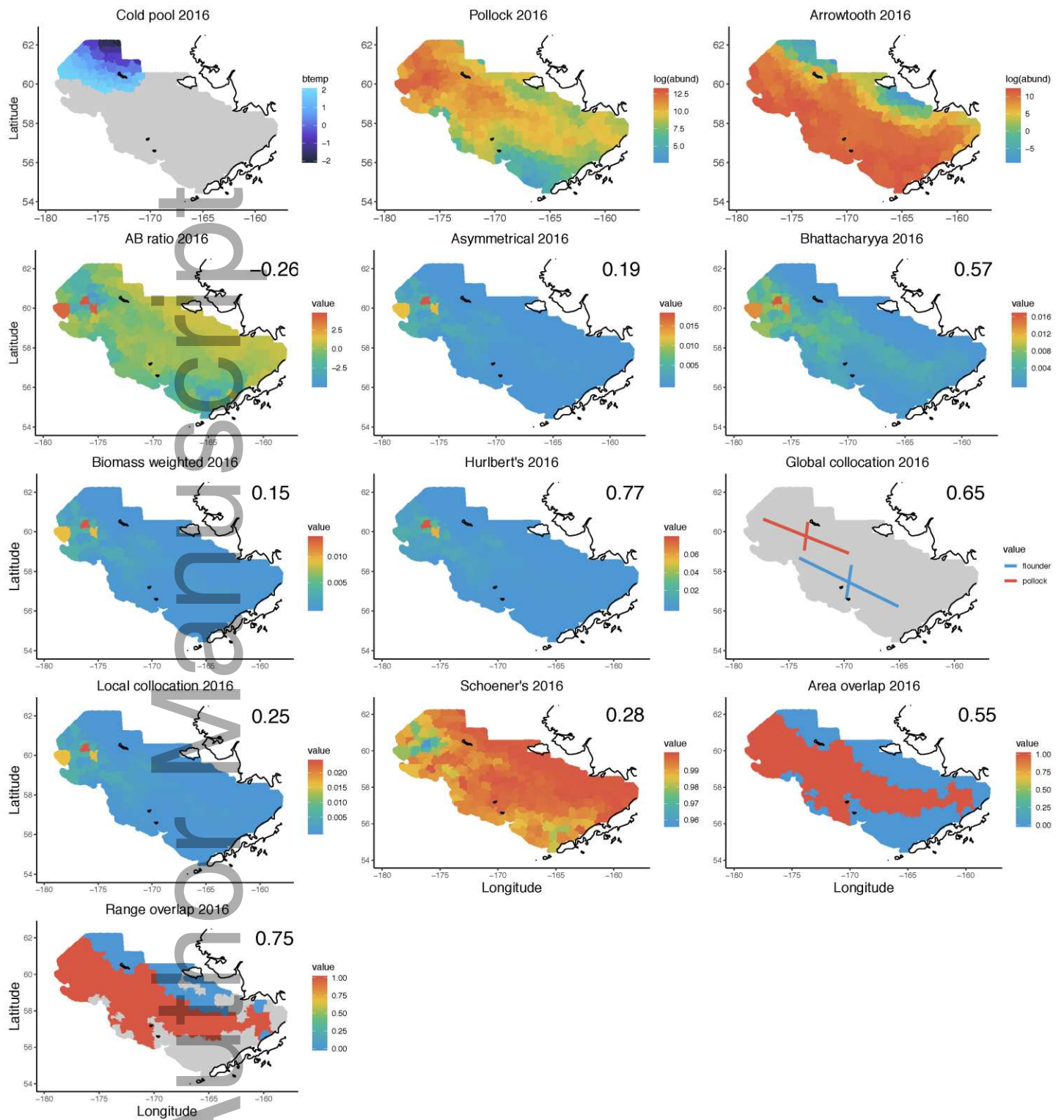
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Figure 3

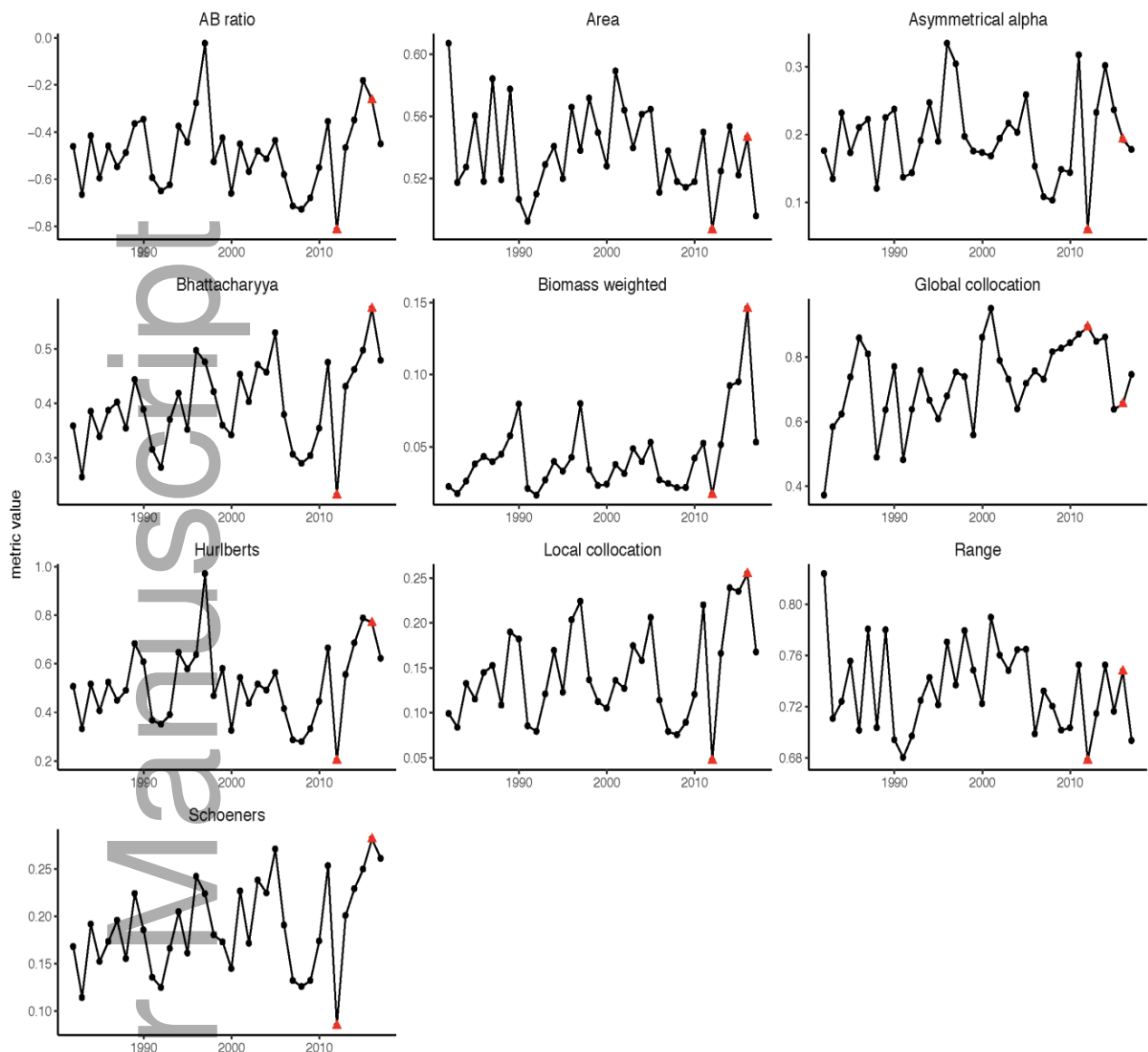
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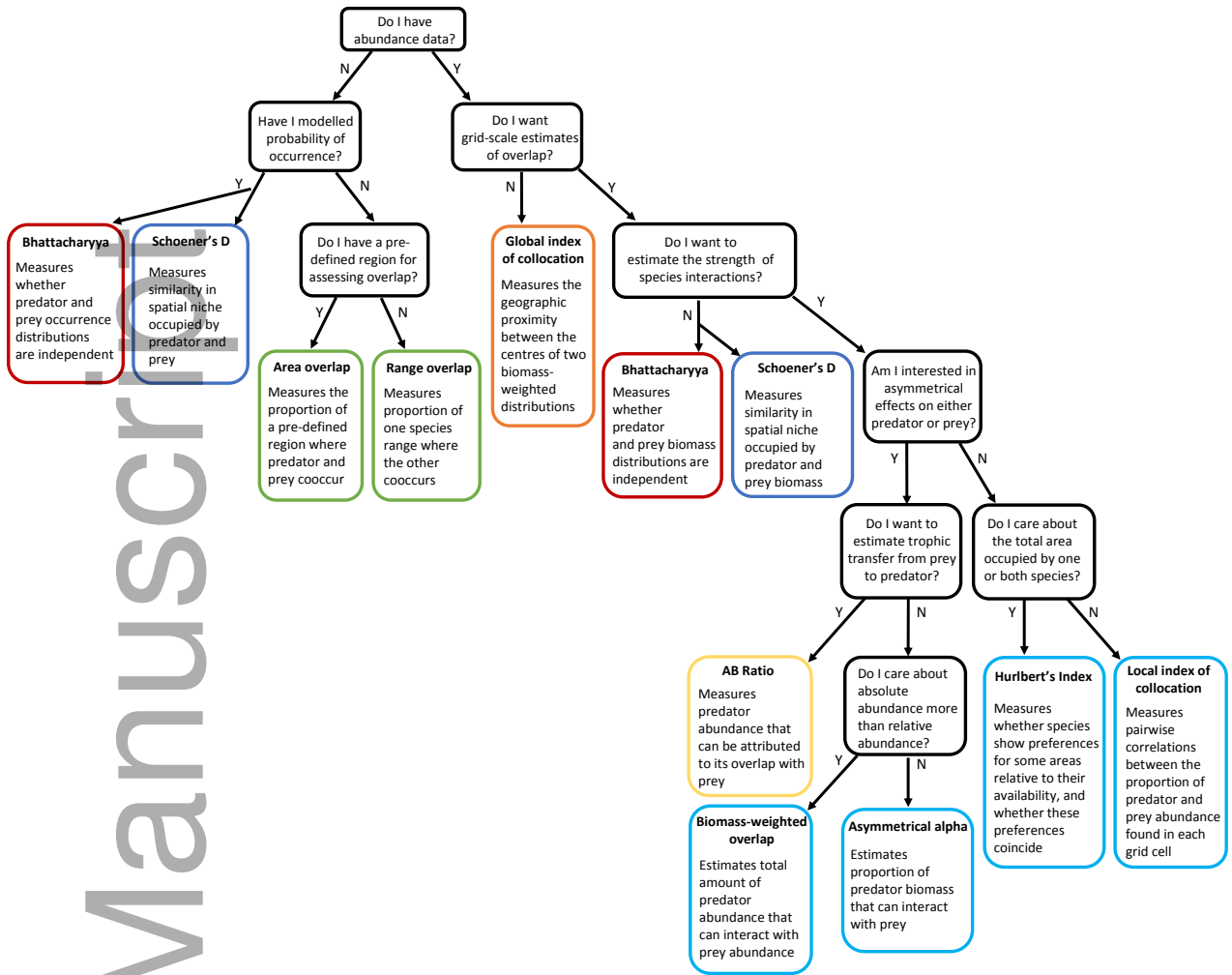
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Figure 4



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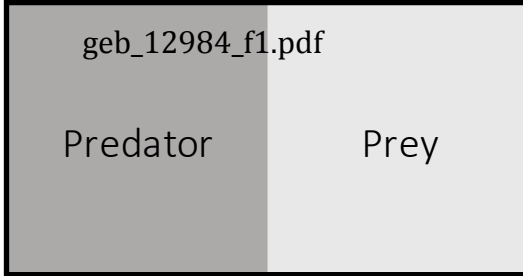
Figure 5



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Figure 6



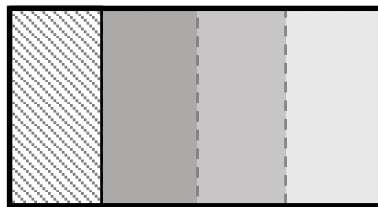
S1



Overlap window = 0.1

$$\alpha = 1$$

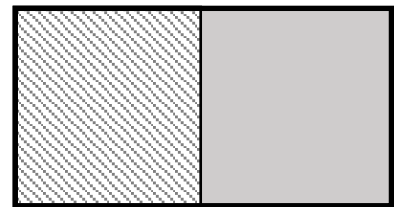
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Overlap window = 0.5

$$\alpha = 1$$

$$\beta = 1$$

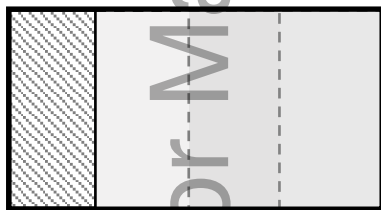


Overlap window = 1

$$\alpha = 1$$

$$\beta = 1$$

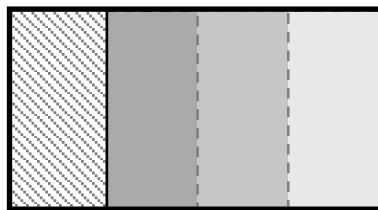
S2



Overlap window = 0.5

$$\alpha = 0.2$$

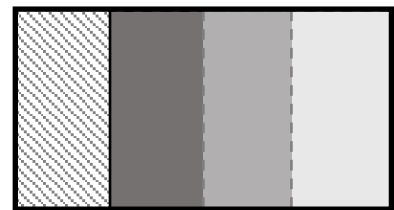
$$\beta = 1$$



Overlap window = 0.5

$$\alpha = 1$$

$$\beta = 1$$

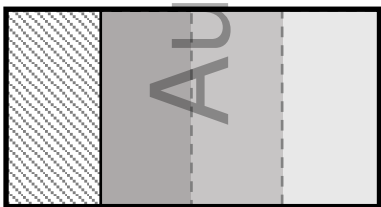


Overlap window = 0.5

$$\alpha = 2$$

$$\beta = 1$$

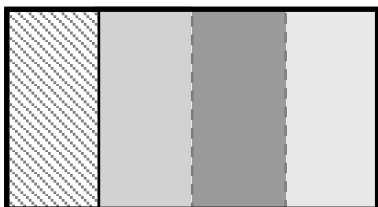
S3



Overlap window = 0.5

$$\alpha = 1$$

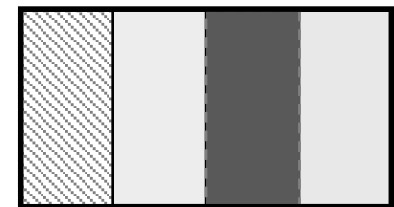
$$\beta = 1$$



Overlap window = 0.5

$$\alpha = 1$$

$$\beta = 1.5$$



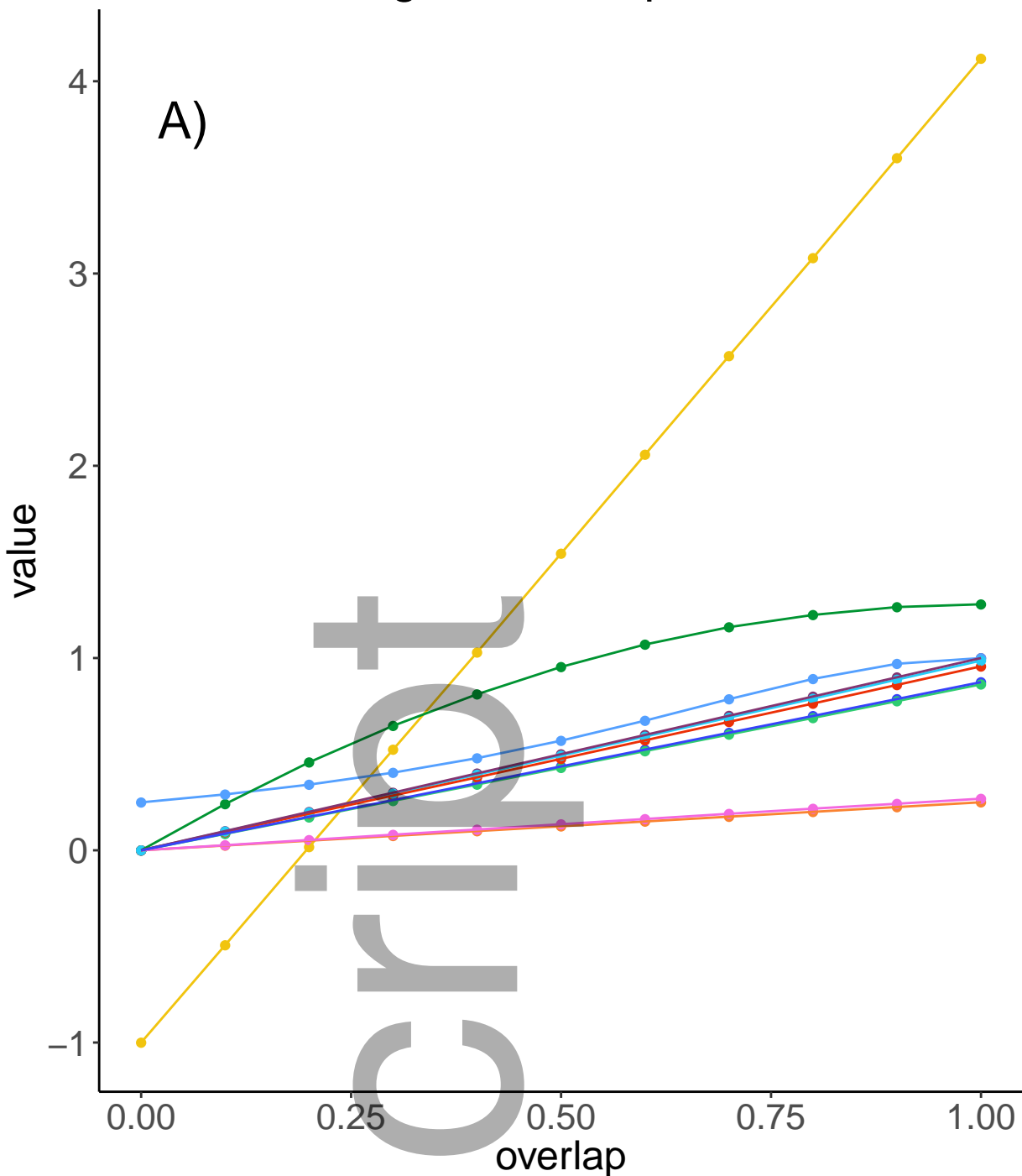
Overlap window = 0.5

$$\alpha = 1$$

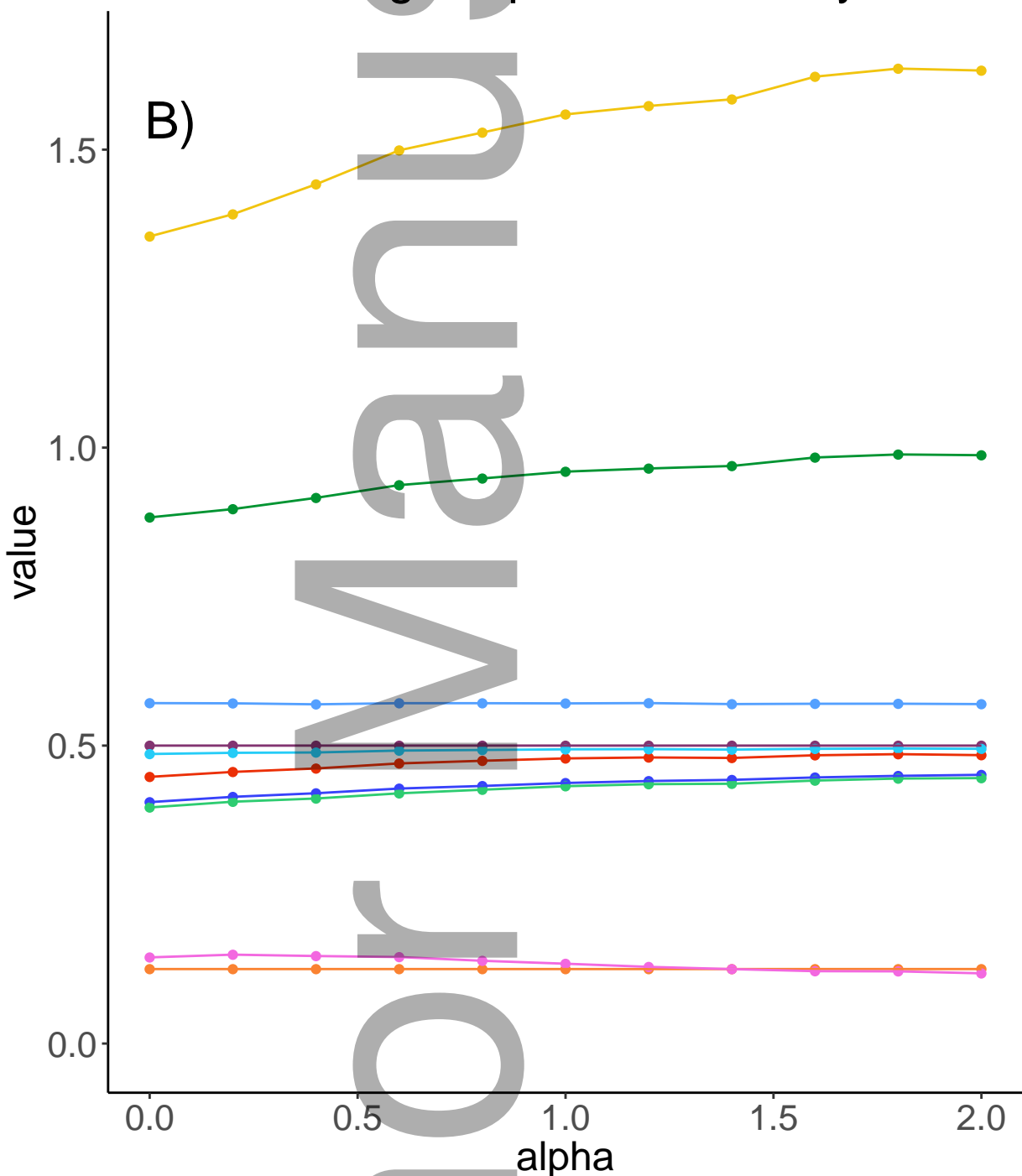
$$\beta = 2$$

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Change in overlap window



Change in predator density



Change in aggregation response

