Contrasting environmental preferences of photosynthetic and nonphotosynthetic soil cyanobacteria across the globe

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BIOSKETCH

Concha Cano-Díaz is a biologist interested in the distribution and ecological drivers of soil cyanobacteria. She is currently studying the effects of climate change and soil formation processes on cyanobacterial communities around the world.

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9 Contrasting environmental preferences of photosynthetic and non 10 photosynthetic soil cyanobacteria across the globe

11 Running title: Global preferences of soil cyanobacteria

12 Abstract

Article type

13 Aim: Cyanobacteria have shaped the history of life on Earth, and continue to play important 14 roles as carbon and nitrogen fixers in terrestrial ecosystems. However, their global distribution 15 and ecological preferences remain poorly understood, particularly for two recently discovered 16 non-photosynthetic cyanobacterial classes (Sericytochromatia and Melainabacteria). 17 Location: 237 locations across six continents encompassing multiple climates (arid, temperate, 18 tropical, continental and polar) and vegetation types (forests, grasslands and shrublands). 19 Time period: Sampling was carried out between 2003 and 2015. 20 Major taxa studied: Photosynthetic and non-photosynthetic cyanobacterial taxa

Methods: We conducted a field survey and used co-occurrence network analysis and structural equation modelling to evaluate the distribution and environmental preferences of soil cyanobacteria across the globe. These ecological preferences were used to create a global atlas (predictive distribution maps) of soil cyanobacteria.

Results: Network analyses identified three major groups of cyanobacteria taxa, which resembled
the three main cyanobacterial classes: the photosynthetic *Oxyphotobacteria*-dominated cluster,
which were prevalent in arid and semiarid areas, and the non-photosynthetic

Sericytochromatia- and Melainabacteria-dominated clusters, which preferred hyperarid
 oligotrophic and acidic/ humid environments, respectively.

Main conclusions: This study provides novel insights into the environmental preferences of nonphotosynthetic cyanobacteria in soils globally. Our findings highlight the contrasting environmental preferences among the three clusters of cyanobacteria and suggest that alterations in environmental conditions linked to climate change may result in important changes in the ecology and biogeography of these functionally important microorganisms.

<u>Keywords</u>: non-photosynthetic Cyanobacteria, Cyanobacteria, global distribution, microbial
 biogeography, microbial network, 16S amplicon sequencing

37 1 INTRODUCTION

38 Cyanobacteria are microorganisms responsible for some of the most important events in Earth's 39 history, including the rise of oxygen levels via oxygenic photosynthesis (Dismukes et al., 2001; 40 Rasmussen et al., 2008) and the formation of plastids through endosymbiosis (Mereschkowsky, 41 1905; Margulis, 1970). Despite being one of the most studied microbial groups (Castenholz et 42 al., 2001; Garcia-Pichel et al., 2003; Garcia-Pichel, 2009; Whitton & Potts, 2012), there are still 43 major gaps of knowledge associated with the diversity and global distribution of these 44 organisms. Recent studies have revealed the existence of two new bacterial clades closely 45 related to cyanobacteria, 4C0d-2 (Melainabacteria) and ML635J-21 (Sericytochromatia), 46 recently proposed as new classes of phylum cyanobacteria (Soo et al., 2014, 2017). These non-47 photosynthetic classes are included in the latest releases of the most commonly used rRNA databases, Silva and Greengenes (DeSantis et al., 2006; Quast et al., 2013). Unlike 48 49 photosynthetic cyanobacteria (hereafter class Oxyphotobacteria), these clades have no genes 50 associated with photosynthesis, and have provided a new perspective on the phylum, 51 broadening our understanding of the functional capabilities of cyanobacteria and their 52 evolutionary origin.

The construction of metagenome-assembled genomes has enabled the assessment of the metabolic potential of these organisms, suggesting that *Melainabacteria* and *Sericytochromatia* are chemoheterotrophs with metabolisms mostly centered on fermentation (Di Rienzi *et al.*, 2013; Soo *et al.*, 2014, 2017; Soo, 2015). Additionally, no genes for phototrophy or carbon (C) fixation have been found in *Melainabacteria* and *Sericytochromatia* (Soo *et al.*, 2017), indicating that oxygenic photosynthesis could be a trait acquired later in *Oxyphotobacteria* by horizontal gene transfer (Raymond *et al.*, 2002). Such physiological and

60 genetic differences might result in contrasting ecological preferences for these novel61 cyanobacterial taxa, but empirical evidence for this is lacking.

62 Soil-borne Oxyphotobacteria are widely distributed on the Earth (Garcia-Pichel et al., 63 2003; Whitton & Potts, 2012; Moreira et al., 2013) but they are specially predominant in hot 64 arid and polar regions with sparse plant cover. They are an important component of biocrusts, 65 soil surface communities dominated by lichens, mosses, cyanobacteria and associated microorganisms (Weber et al. 2016) and play key ecological roles in these environments by 66 67 regulating critical soil processes such as nitrogen (N) and C fixation, soil stabilization and infiltration/runoff (Mager & Thomas, 2011; Sciuto & Moro, 2015). Other terrestrial 68 69 cyanobacterial communities grow on the surface or inside rocks and soil (endolithic and subsoils 70 forms), and are well adapted to dry conditions and high or low irradiation regimes (Warren-71 Rhodes et al., 2006; Domínguez & Asencio, 2011; Puente-Sánchez et al., 2018). The capacity of 72 Oxyphotobacteria to stay dormant during long periods of time is also a fundamental characteristic of these organisms, which allow them to survive in extreme environments 73 74 characterized by high or low temperatures, desiccation regimes or high ultraviolet radiation 75 (Garcia-Pichel, 2009; Quesada & Vincent, 2012; Whitton & Potts, 2012).

76 Local and regional studies show that soil Oxyphotobacteria are generally considered to 77 prefer neutral to alkaline pH for optimum growth (Brock, 1973; Whitton & Sinclair, 1975; Nayak 78 & Prasanna, 2007). However, the global biogeography of soil Oxyphotobacteria has not been 79 fully resolved due to the concentration of cyanobacterial research in particular regions, e.g. 80 studies in western United States or the Antarctic continent (Garcia-Pichel et al. 2001; Namsaraev 81 et al. 2010)(Garcia-Pichel et al., 2003; Moreira et al., 2013; Büdel et al., 2016; Williams et al., 82 2016) and the focus given to key and abundant taxa, such as Microcoleus vaginatus or the genus 83 Chroococidiopsis (Bahl et al., 2011; Dvořák et al., 2012), or specific habitats such as cold 84 ecosystems and deserts (Jungblut et al. 2010; Bahl et al. 2011). There are clear gaps of 85 knowledge of their distribution in certain regions of the world, such as South America (Büdel *et* al., 2016). Despite their wide dispersal ability due to small size, aeolian transport and tolerance 86 87 to desiccation and irradiation (Billi et al., 2000; Kellogg & Griffin, 2006), and their often 88 cosmopolitan distribution (Garcia-Pichel et al., 1996; Taton et al., 2006; Flombaum et al., 2013), 89 current knowledge suggests a more complex biogeography of these microorganisms that is likely 90 to be also influenced by their phylogeny and historical legacies (Garcia-Pichel et al., 1996, 2003; 91 Taton *et al.*, 2006; Nayak & Prasanna, 2007; Flombaum *et al.*, 2013).

The ecology and biogeography of the non-photosynthetic cyanobacteria classes (*Melainabacteria* and *Sericytochromatia*) in soils is poorly known. Available information on these organisms comes from genomes from aphotic environments such as animal guts or subsurface

groundwater and artificial systems such as water treatment facilities and laboratory bioreactors
(Ley *et al.*, 2005; Warnecke *et al.*, 2007; Yagi *et al.*, 2010; Di Rienzi *et al.*, 2013; Soo *et al.*, 2014;
Utami *et al.*, 2018) and the scarce environmental studies correspond only to aquatic ecosystems
such as lakes and algal biofilms (Monchamp *et al.*, 2018, 2019).

99 To advance our understanding of the biogeography and ecological preferences of soil 100 photosynthetic and non-photosynthetic cyanobacteria, we used data from a global soil survey 101 covering a wide diversity of climate, soil and vegetation types (Delgado-Baquerizo et al., 2018). 102 We expected the distinct ecological attributes of photosynthetic and non-photosynthetic 103 cyanobacteria to be associated with very different environmental preferences. For example, we 104 know that some Oxyphotobacteria have developed highly competitive adaptations to thrive in 105 arid soils with low soil organic C and plant productivity (Lund, 1967; Whitton & Sinclair, 1975; 106 Maestre et al., 2015). In these environments, we expect Oxyphotobacteria to dominate due to 107 their capacity to build protective sheath pigments and to fix atmospheric C and N, which can be 108 an important ecological advantage. However, Oxyphotobacteria are also expected to appear in 109 a wide variety of environmental conditions, including low light, low oxygen or even anoxygenic 110 environments due to their enormous functional diversity (Stal & Moezelaar, 1997; Adams & Duggan, 1999; Garcia-Pichel, 2009; Puente-Sánchez et al., 2018). Conversely, non-111 112 photosynthetic cyanobacteria rely on soil organic C pools to grow, which could translate into 113 contrasting preferences related to soil nutrient availability. We expect to find groups of taxa co-114 occurring and sharing similar environmental preferences (hereafter ecological clusters) related 115 to photosynthetic capability, habitat preferences and historical legacies.

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2 MATERIALS AND METHODS

117 2.1 Global survey: Sites, soil collection, soil and molecular analyses

118 We used 16S rRNA gene amplicon sequencing data from a global survey of 237 locations (Fig. S1) across six continents encompassing multiple climates (arid, temperate, tropical, continental 119 and polar) and vegetation types (forests, grasslands and shrublands) (Delgado-Baquerizo et al., 120 121 2018). A composite soil sample (0-7.5 cm depth) was collected under the dominant vegetation 122 at each surveyed location. A fraction of each sample was immediately frozen at -20°C for molecular analyses; the other fraction was air-dried for chemical analyses. Sample collection of 123 124 soils took place between 2003 and 2015. We do not expect differences in the timing of sample 125 collection to largely affect our results for two main reasons. First, at the global scale seasonal 126 variability is expected to be largely overcome by cross-biome variability (e.g., see Carini et al., 127 2020 on the importance of spatial vs. temporal scales when analyzing soil microbial 128 communities). To put it simple, a dryland and a boreal forest are so different that usually harbor

129 distinct microbial communities regardless of their seasonal variability. Second, we are using 130 amplicon sequencing DNA-based analyses (see below), which characterize not only the active 131 portion of cyanobacterial communities but also the dormant one at the moment of sampling (Li 132 et al., 2017). The soils sampled comprise a wide variety of physico-chemical properties, pH 133 ranged from 4.04 to 9.21, texture of the fine fraction (%clay+silt) ranged from 1.4 to 92.0%, soil 134 total organic carbon (OC) from 0.15 to 34.77%, soil total nitrogen (TN) from 0.02 to 1.57, C:N 135 ratio (CN) ranged from 2.12 to 67.52 and soil total phosphorus (TP) from 75.10 to 4111.04 mg P kg⁻¹ soil. These analyses were done using standard laboratory methods described in Delgado-136 137 Baquerizo et al. (2018).

138 Climatic variables (maximum and minimum temperature [MAXT, MINT], precipitation 139 seasonality [inter-annual coefficient of variation in precipitation, PSEA] and mean diurnal 140 temperature range [MDR]) were obtained for each site from the WorldClim database (Hijmans 141 et al., 2005). Aridity Index (precipitation/potential evapotranspiration) was obtained from the 142 Global Potential Evapotranspiration database (Zomer et al., 2008), which uses interpolations 143 from WorldClim. The annual ultraviolet index (UV Index), a measure of the risk of UV exposition 144 ranging from 0 (minimal risk) to 16 (extreme risk), was obtained for each site using data from 145 the Aura satellite (Newman & McKenzie, 2011). Net aboveground primary productivity [ANPP] 146 was estimated with satellite imagery using the Normalized Difference Vegetation Index (NDVI) 147 from the Moderate Resolution Imaging Spectroradiometer (MODIS) aboard NASA's Terra 148 satellites (Justice et al., 1998). This index provides a global measure of the greenness of the Earth 149 for a given period (Pettorelli et al., 2005). Here, we used monthly averaged values for NDVI for 150 the sampling period between 2003 and 2015 (10 km resolution).

151 Microbial DNA was extracted using the PowerSoil DNA Isolation Kit (MoBio Laboratories, 152 Carlsbad, CA, USA) following manufacturer's instructions. DNA extracts were sequenced rRNA gene 153 targeting the bacterial V3-V4 region using 16S primers 341F 154 (CCTACGGGNGGCWGCAG) and 805R (GACTACHVGGGTATCTAATCC) and the Illumina Miseq platform of the Next Generation Genome Sequencing Facility at Western Sydney University 155 (Australia). Bioinformatic analyses were performed with a combination of QIIME (Caporaso et 156 157 al., 2010), USEARCH (Edgar, 2010) and UPARSE (Edgar, 2013). After merging of the reads, the 158 primers were trimmed and sequences of low quality (expected error rate > 1) were discarded. 159 Phylotypes were defined with UCLUST (Edgar, 2010) at an identity level of 97% and taxonomy was assigned using Silva Incremental Alligner Search and classify with Silva database 160 161 (complementing not identified phylotypes with Greengenes database) (DeSantis et al., 2006; 162 Quast et al., 2013). Phylotypes represented by only a single read (singletons) were removed. The 163 final dataset of phylotypes was filtered for phylum Cyanobacteria (excluding Chloroplast) and

the relative abundance each of cyanobacterial phylotype in relation to total bacteria (all 16SrRNA reads) was calculated.

166 2.2 Structure of the community: Network analyses

167 To explore the different patterns of cyanobacterial co-occurrence across our samples, we 168 conducted a network analysis with the CoNet software (Faust & Raes, 2016). This tool detects 169 significant non-random patterns of co-occurrence using multiple correlation and dissimilarity 170 measures. Two correlation coefficients (Pearson and Spearman) and dissimilarity distances 171 (Bray-Curtis and Kullback Leiber) were used to obtain a more reliable network (Faust & Raes, 172 2012). When links were detected by more than one correlation/dissimilarity measure, they were 173 considered as a single link. Samples were standardized prior to network analyses with the 174 "col_norm" function, which divides each column by its sum, converting abundances in column-175 wise proportions. We computed the network with the top 1000 links for each measure and 176 tested the statistical significance of each link with 1000 permutations and the function "shuffle 177 rows" as the resampling strategy. Multiple testing was corrected by using Benjamini-Hochberg's 178 procedure (Benjamini & Hochberg, 1995), keeping links with an adjusted merged p-value below 179 0.05. The final network was visualized with the interactive platform gephi (Bastian et al., 2009). 180 We obtained the ecological clusters with the function "fastgreedy" from the igraph package 181 (Csárdi & Nepusz, 2006) in R version 3.4.0 (Team, 2013), and tested the statistical significance of 182 modularity using 10000 random networks. Network analysis allowed us to divide the community 183 between ecological clusters, that we used for further analysis. The relative abundance of each 184 ecological cluster per sample was calculated by averaging the standardized (z-score) relative 185 abundance of the phylotypes present within each ecological cluster. Thus, we obtained a 186 balanced contribution of each cyanobacterial phylotype to the relative abundance of its 187 ecological cluster. Note that the use of z-score standardization transforms relative abundances, 188 and therefore negative values can be obtained.

189 2.3 Factors determining cyanobacterial global distribution

Environmental effects: We conducted Structural Equation Modelling (SEM, Grace 2006) to evaluate the direct and indirect effects of spatial, climatic, vegetation and soil variables as predictors of the abundance of the main cyanobacterial ecological clusters (See Fig. S2 for our *a priori* model). This approach is useful for simultaneously testing the influence of multiple variables and the separation of direct and indirect effects of the predictors included in the model (Grace, 2006). These included spatial (Latitude, sine Longitude, cosine Longitude), climatic (MDR, MAXT, MINT, PSEA and Aridity [1-Aridity Index]) and vegetation (Grassland, Forest and

197 ANPP) variables, as well as soil properties (CN, soil OC, pH and percentage of clay and silt). Prior 198 to modelling, we transformed them to improve normality: Aridity, OC, PSEA and CN were log-199 transformed and both ANPP and the percentages of clay and silt were square root transformed. 200 We used the chi-square fit test, supplemented with root mean square error of approximation 201 (RMSEA) to test the overall fit of the model. We analysed path coefficients of the model and 202 their associated P values and the total effects of each variable. As some of the variables were 203 not normally distributed despite transforming them, we used 5000 bootstraps to simultaneously 204 test the significance of each path. SEM analyses were conducted using AMOS 24.0.0 (IBM SPSS, Chicago, IL, USA). 205

206 To obtain a prediction of the potential distribution of the main cyanobacterial ecological 207 clusters, we used the regression model Cubist (Quinlan, 2014) as implemented in the R package 208 Cubist (Kuhn et al., 2016). This model uses a linear regression tree analysis that predicts the most 209 important factors affecting the abundance of each ecological cluster based on environmental 210 covariates. Covariates in our models included the same variables used in our SEMs. Global 211 predictions of the distribution of major clusters were done on a 25 km resolution grid. Soil 212 properties for this grid were obtained from SoilGrids (Hengl et al., 2017). Major vegetation types 213 (grasslands and forests) were obtained using Globcover2009 map from the European Space 214 Agency (Bontemps et al., 2013). Information on climate, UV index and net primary productivity 215 were obtained from the WorldClim database and NASA satellites as described above.

216 We conducted multiple analyses to support the validity of our global prediction maps. 217 First, we used kernel density estimations to compare the distribution of key soil and climate 218 variables of our dataset with those from high resolution global maps: SoilGrids (Hengl et al., 219 2017) and Worldclim (Hijmans et al., 2005). Our dataset comprises a large percentage of their 220 global variability (Fig. S3): 78.51% for OC, 94% for pH, 58.25% for Aridity, 45.98% for PSEA, 71.63% for MINT, 47.03% for MAXT and 96.43% for ANPP. These results indicate that our 221 222 sampling covers a large proportion of the environmental variability found on Earth. Second, we found a strong correlation between the relative abundance of our cyanobacterial ecological 223 224 clusters and key microbial environmental factors at the global scale (see results below), which 225 suggests that environmental data can be used to predict their distribution. Finally, predictive 226 maps were cross-validated with an independent dataset obtained from the Earth Microbiome 227 Project (EMP, Thompson et al., 2017), which contains data on soil cyanobacteria from 403 sites 228 worldwide (see Fig. S1). For doing so, we estimated the relative abundance of the three main 229 cyanobacterial clusters for the EMP dataset using the 97% similar EMP phylotypes. We first 230 calculated relative abundance of each cyanobacterial phylotype in relation to total bacteria (all 231 16S rRNA reads of the EMP dataset). Then, the relative abundance of each ecological cluster per

sample was computed by averaging the standardized (z-score) relative abundance of the phylotypes of each ecological cluster, as explained above for our dataset. We then used our predictive maps to extract the predicted relative abundance of each cluster for the EMP locations. These predictive abundances were then compared with the independent results of relative abundance of each cluster calculated with the EMP dataset using Pearson correlations.

We also conducted a Permanova analysis with Bray Curtis distances to evaluate the effect of vegetation type on the abundance of each cyanobacterial cluster with the *adonis* function and 1000 permutations. To test for the differences in the relative abundance of each cluster across vegetation types we first tested the homogeneity of groups dispersions (variances) with *betadisper* function and from the result of that call we performed the post hoc analysis Tukey Honest Significant Differences with *TukeyHSD* function. All these analysis were done with vegan v2.4-2 (Oksanen, 2015) and R version 3.6.0 (Team, 2013).

244 <u>Phylogenetic tree</u>: The phylogenetic tree of cyanobacteria was constructed using the SILVA
245 Alignment, Classification and Tree (ACT) Service (<u>www.arb-silva.de/act</u>). Multiple sequence
246 alignment of the 343 rRNA gene sequences was performed using SINA v1.2.11 (Pruesse *et al.*,
2012). A phylogenetic tree was obtained with their built-in tree computation tool FastTree (Price
248 *et al.*, 2009) using the General Time Reversible Model of nucleotide evolution (Nei & Kumar,
2000) and keeping the default parameters. The display and annotation of phylogenetic tree were
250 made with iTol v5.5 (Letunic & Bork, 2019).

251 3 RESULTS

252 3.1 Global cyanobacterial co-occurrence patterns

253 Despite the common and widespread occurrence of soil cyanobacterial taxa on Earth, we did 254 not find any of the 343 phylotypes present in all samples. The most ubiquitous cyanobacterial 255 phylotype, *Microcoleus vaginatus*, was detected in 113 of the 237 sites surveyed. Moreover, the 256 relative abundance of cyanobacterial phylotypes in our soils ranged from 0.01% to 4.35% of all bacterial 16S rRNA gene sequences (see Table S1). The cyanobacterial orders with the highest 257 258 relative abundances included Oscillatoriales (Oxyphotobacteria), followed by Obscuribacterales 259 (Melainabacteria) and Nostocales (Oxyphotobacteria) (Fig. 1). Non-photosynthetic phylotypes 260 appeared almost in all samples (235/237 samples 99.2%). Photosynthetic cyanobacteria 261 phylotypes appeared in the majority of them (185/237, 78.1%).

262 Our final network had 281 phylotypes and was arranged in 10 ecological clusters. Among 263 these clusters, we identified three major groups of taxa co-occurring and comprising 65% of the

264 cyanobacterial phylotypes identified (Fig. 2a). The remaining seven clusters were minor, 265 encompassing from 8% to 1% of phylotypes. The three main ecological clusters were dominated 266 by either Oxyphotobacteria (82% of 76 phylotypes), Sericytochromatia (52% of 31 phylotypes) 267 or *Melainabacteria* (83% of 76 phylotypes; see Table S1). We focused on these main ecological 268 clusters for the downstream analyses. Our correlation network showed a contrasting node 269 distribution for cyanobacterial phylotypes characterized by photosynthetic and non-270 photosynthetic capabilities (Fig. 2b). Overall, the three ecological clusters identified were 271 strongly dominated by the three extant cyanobacterial classes (Fig. 2c, 2d).

3.2 Environmental preferences of photosynthetic and non-photosynthetic soilcyanobacteria

274 Vegetation type significantly affected the abundance of each of the main cyanobacterial clusters 275 identified (Permanova R^2 =0.28, 0.24 and 0.15 for *Melainabacteria*, *Sericytochromatia* and 276 *Oxyphotobacteria*-dominated clusters, respectively, *p*<0.05 in all cases).

277 Our SEM model indicated that the cluster dominated by Oxyphotobacteria was 278 positively and negatively related to aridity and net aboveground productivity, respectively (Figs. 279 3, 4 and S4a), which explains their high relative abundance in dry grasslands (Fig. 6). We also 280 observed a positive association between the relative abundance of the Oxyphotobacteria dominated cluster and both soil pH and minimum temperature (Fig. 3, 4, and S4a). We predicted 281 282 the distribution of this cluster in a wide range of arid and semiarid areas worldwide (e.g., southern Sahara, southern Africa, northern Australia, India, Arabian Peninsula, areas 283 284 surrounding the Amazon Basin, southwestern US and northwestern Mexico; Fig. 5a).

The cluster dominated by *Sericytochromatia* had a strong preference for arid environments with low soil C content (Fig. 3, 4, 6 and S4b). Taxa within this ecological cluster were also positively associated with locations characterized by high inter-annual rainfall variability (Figs. 3, 4 and S4b). Our global atlas predicts that taxa within this ecological cluster can be found in hyper-arid areas such as the Saharan Desert, central Australia, the Atacama, Gobi and Taklamakan Deserts and the Arabian Peninsula, with almost no areas of intermediate relative abundance (Fig. 5b).

Unlike the other two ecological clusters identified, the *Melainabacteria*-dominated cluster showed a preference for humid and acidic soils, as indicated by the reduced relative abundance of this cluster with increases in aridity and pH (Figs. 3, 4 and. S4c). The vast majority of phylotypes found in our study corresponded to the order Obscuribacterales (1, 2d). This ecological cluster is found mainly in tropical and cold forests and grasslands (which are mostly temperate; see Fig. 6). Prediction maps show high relative abundance values of this cluster in

298 humid areas of the Amazon Basin, central Africa, west Asian coast and Pacific Islands (Fig 5c). 299 Despite the methodological differences between our dataset and the EMP dataset (primer sets 300 used here 341F/805R vs. 515F/806R for the EMP; read lengths here 400bp/sequence vs. <150bp 301 for the EMP and the lack of standardization in the EMP soil sampling protocols and metadata 302 collection) we obtained positive and significant correlations between both results: 303 Melainabacteria dominated cluster Pearson's r=0.28 (P<0.001), Sericytochromatia dominated 304 cluster Pearson's r=0.53 (P<0.001), Oxyphotobacteria dominated cluster Pearson's r=0.35 305 (P<0.001). These results support the validity of our maps as representative of the distribution of 306 the main ecological clusters of cyanobacteria across the globe.

307 4 DISCUSSION

308 The discovery of non-photosynthetic cyanobacteria has expanded one of the currently most 309 diverse bacterial phylum (Castenholz et al., 2001; Garcia-Pichel, 2009; Whitton & Potts, 2012; 310 Dvořák et al., 2017). There is a large body of knowledge about photosynthetic cyanobacteria 311 showing their importance in terrestrial ecosystems, as they are key components of cryptogamic 312 covers, which are estimated to fix 3.9 Pg carbon per year (Elbert et al., 2012). They increase soil 313 fertility by fixing atmospheric N (Cleveland *et al.*, 1999), stabilize soils by producing extracellular 314 polysaccharides (Mazor et al., 1996; Mager & Thomas, 2011), protecting it from erosion and 315 creating suitable habitats for the colonization of mosses and lichens (Zhang, 2005; Lan et al., 316 2015). However we know relatively little about the distribution and environmental drivers of the 317 newly described non-photosynthetic cyanobacteria in soils. Our work provides novel insights 318 into the ecology and biogeography of these key organisms, and advances our understanding of 319 on the potential vulnerabilities of photosynthetic and non-photosynthetic cyanobacteria to 320 changing environmental conditions.

321 Photosynthetic taxa represented by the Oxyphotobacteria-dominated cluster prefer 322 areas with sparse vegetation cover, and therefore greater accessibility to light, such as dry 323 grasslands (Figs. 3,4, 6 and S4a). Accordingly, they are reported as key components of biocrust 324 communities in low productivity ecosystems such as arid environments (Garcia-Pichel, 2009; 325 Belnap et al., 2016), where the ability to fix atmospheric C and N can be an important ecological 326 advantage. As with the remaining bacterial communities (Fierer & Jackson, 2006) soil acidity is 327 a key factor shaping the global distribution of Oxyphotobacteria (Fig. 4). Consistentwith previous 328 studies (Baas-Becking et al., 1960; Brock, 1973; Nayak & Prasanna, 2007) we found that 329 photosynthetic cyanobacteria have a preference for neutral to alkaline soils (Figs. 3,4 and S4a), 330 which are characteristic of drylands (Schlesinger & Bernhardt, 2013). Our analyses further indicate a wide distribution of this cluster in drylands worldwide (Fig. 5), as previously reported 331

332 for members of this taxa in continental-scale distribution studies (Bahl et al., 2011; Garcia-Pichel 333 et al., 2013). Together with temperature, soil moisture plays a key role driving the physiology, 334 small-scale distribution and behaviour of soil photosynthetic cyanobacteria (Garcia-Pichel & 335 Pringault, 2001; Rajeev et al., 2013). The high tolerance and photosynthetic performance of 336 Oxyphotobacteria at high temperatures is one of the reasons why cyanobacterial-dominated 337 biocrusts are so abundant in hyper-arid and arid environments (Grote et al., 2010; Wang et al., 338 2012). Thus, we observed a positive influence of high minimum temperatures and aridity on this 339 cyanobacterial cluster (Figs. 3. and S4a). By moving from local/regional to the global scale, 340 including samples from poorly-studied regions of South America (Garcia-Pichel et al., 2003; Büdel et al., 2016), and considering multiple terrestrial global biomes, our results provide novel 341 342 predictions of the global distribution of Oxyphotobacteria in global soils.

343 Unlike Oxyphotobacteria, non-photosynthetic cyanobacteria require relatively large soil 344 organic C pools for growth. We observed contrasting environmental preferences for each of the 345 non-photosynthetic clusters across the oligotrophic-copiotrophic continuum, such as those 346 reported for other soil heterotrophic organisms (e.g., methanotrophs in Nazaries et al. 2018). A 347 key finding of our study is that the *Melainabacteria*-dominated cluster was especially abundant 348 in mesic forests (tropical and cold forests, Fig. 6) and temperate grasslands, while the 349 Sericytochromatia-dominated cluster is associated with locations with reduced plant cover and 350 high temperatures (e.g., hyperarid deserts in Fig. 5, dry grasslands in Fig. 6). We found very little 351 overlap between the predicted distributions of non-photosynthetic clusters of cyanobacteria 352 (Figs. 5b, 5c) and a negative relationship between the relative abundances of these two nonphotosynthetic clusters (Spearman correlation r= -0.31, p<0.05). Interestingly, a sizable 353 354 percentage of members of Melainabacteria appears in the Sericytochromatia dominated-cluster (38%). We know that members of class Melainabacteria are capable of aerobic respiration 355 356 because they contain respiratory components of the complex III-IV operon, which is adapted to 357 low oxygen conditions, a C-family oxygen reductase and two cytochrome bc oxydases (Soo et al., 2017). However, the Melainabacteria-dominated cluster is dominated by members of the 358 359 order Obscuribacterales (Fig. 2d), for which there is little functional information available in the 360 literature. Genomic analyses of the Candidatus Obscuribacter phosphatis suggest that this 361 particular species is adapted to dynamic environments involving feast-famine nutrient cycles, 362 and has the capacity for aerobic or anaerobic respiration and fermentation (Soo et al., 2014). 363 These features allow it to survive in both oxic and anoxic environments. To our knowledge there 364 is no information available of the contribution of this cyanobacterium to the structure and 365 function of forest ecosystems. However, our results suggest that molecular ecologists and 366 taxonomists targeting taxa in Melainabacteria-dominated cluster should focus mainly on mesic

forests across the globe. We also expect non-photosynthetic cyanobacteria to play a significant role in soil biogeochemical cycles in both high and low productive soils through C degradation and/or H₂ production, as reported for *Melainabacteria* in an alluvial aquifer (Wrighton *et al.*, 2014). However, studies linking non-photosynthetic soil cyanobacteria to carbon degradation in terrestrial environments are still lacking. Future studies are thus needed to identify the relative contributions of non-photosynthetic cyanobacteria to organic matter decomposition and C cycling in soils from contrasting biomes.

The topology of our phylogenetic tree (Fig. 2c) reflects the expected evolutionary 374 375 relationships from previous research with separation of three main clades (Soo et al., 2017); the 376 basal deep branched Sericytochromatia, Melainabacteria and photosynthetic 377 Oxyphotobacteria. As the ecological clusters are related to these classes, their global distribution 378 is likely to be related to past evolutionary events within this ancient phylum (Bahl et al., 2011; 379 Moreira *et al.*, 2013). The ecological diversification observed in the non-photosynthetic clades 380 is particularly noteworthy. We found a niche-differentiation between the basal cyanobacterial 381 clade, Sericytochromatia, which occupies extremely dry environments, and Melainabacteria, 382 which is mostly found in humid forests. Interestingly, the presence of phylotypes from 383 Melainabacteria in the Sericytochromatia-dominated cluster may point to the existence of 384 common ancestral traits between both classes and the later expansion of Melainabacteria into 385 new "humid" niches. Photosynthetic cyanobacteria (Oxyphotobacteria) are known for being 386 extraordinarily ecologically versatile, mostly living in environments with at least some exposure 387 to sunlight, and capable of inactivating their photosynthetic apparatus (Harel et al., 2004) or 388 performing light-independent energy generation (Stal, 2012) when needed. There is still no 389 consensus about the date the acquisition of oxygenic photosynthesis by Oxyphotobacteria; this 390 could have happened either after divergence from other non-photosynthetic clades (Soo et al., 391 2017) or before, sharing a photosynthetic common ancestor (Harel et al., 2015). Regardless, the 392 acquisition of oxygenic photosynthesis was a revolutionary event that allowed cyanobacteria to expand into diverse niches, and also the evolution of algae and terrestrial plants through 393 394 endosymbiosis (Mereschkowsky, 1905; Margulis, 1970).

Our findings represent a starting point towards the understanding of the ecological preferences and global distributions of non-photosynthetic soil cyanobacteria. They highlight the fact that major photosynthetic and non- photosynthetic groups of soil cyanobacteria have contrasting ecological preferences across the globe. However, and given the difficulty of predicting microorganisms at a global scale, conclusions should be viewed as preliminary. The potential distribution maps presented here and the identification of the main environmental drivers of soil cyanobacterial distribution also illustrate how different cyanobacterial lineages

402 might respond to ongoing climate and land use change. For example, the positive influence of 403 aridity on the Sericytochromatia- and Oxyphotobacteria-dominated clusters suggests that the 404 distribution of these taxa could expand under future climate change scenarios (Huang et al., 405 2016). Consequently, our findings advance our understanding of the ecological distributions of 406 these functionally important microbial communities and provide a basis for predicting possible 407 future shifts of cyanobacterial terrestrial communities in a human-dominated, warmer and 408 more arid world. To complement and expand our findings, future studies should further 409 investigate the temporal dynamics of photosynthetic and non-photosynthetic cyanobacteria in 410 terrestrial ecosystems, particularly along multiple temporal scales.

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724 DATA ACCESSIBILITY STATEMENT

Raw data related with this manuscript are available in Figshare,
https://figshare.com/s/82a2d3f5d38ace925492

Author

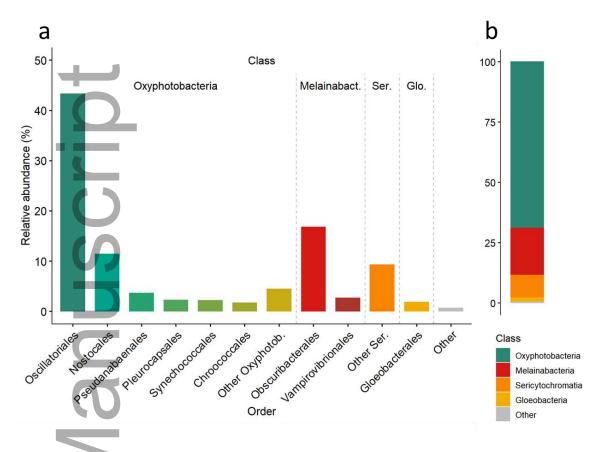


Fig. 1. Taxonomic information on the relative abundance of cyanobacterial orders (a) and classes
(b) across all sites. Ser.= Sericytochromatia (no orders described yet) and Glo. = Gloeobacteria.

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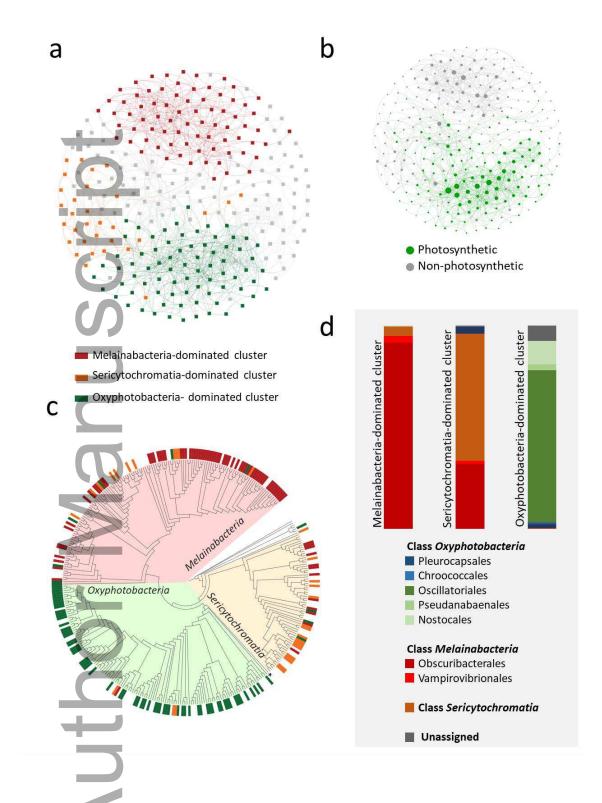
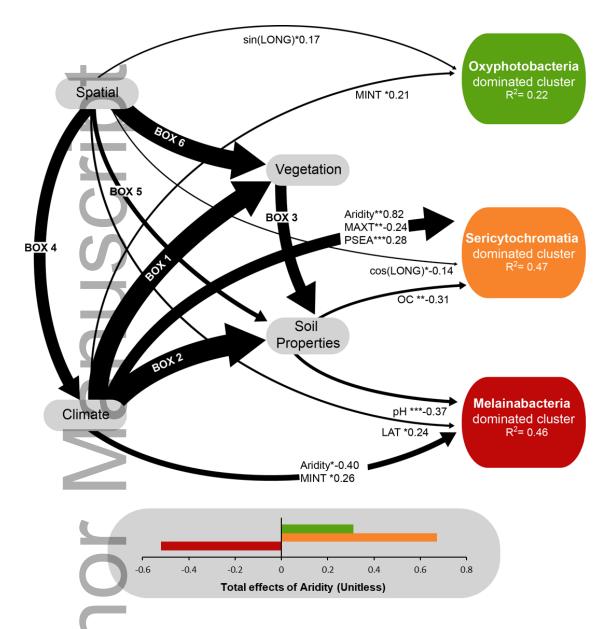
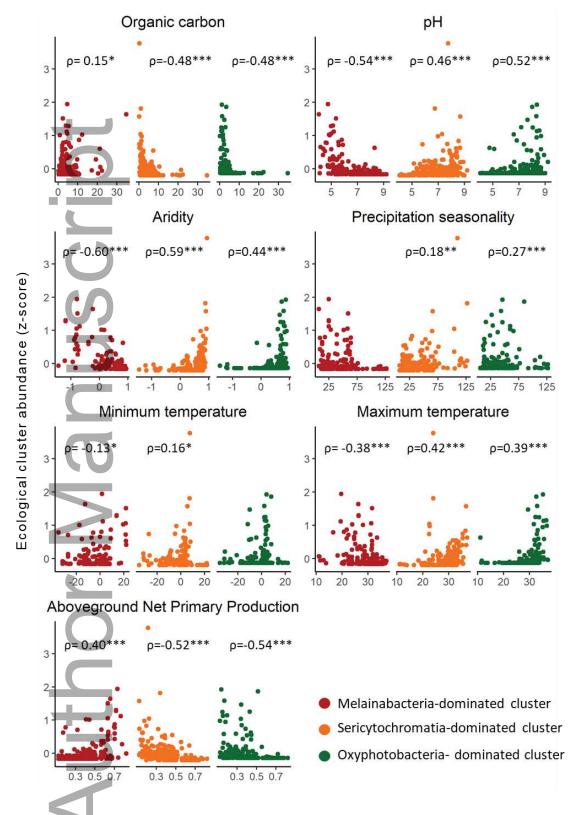


Fig. 2 Global network of co-occurrences within soil cyanobacteria, colored by either main ecological clusters (a) or the photosynthetic capability of taxa (b). The size of the nodes is related to the number of links they contain. The network had 282 nodes (cyanobacterial phylotypes) and 986 significant links (potential ecological interactions between phylotypes) (c) Phylogenetic tree obtained with the main ecological clusters located at the end of the branch. Background

- 738 colored by cyanobacterial class, * for Gloeobacteria class. (d) Taxonomic composition in relation
- to total 16S reads.



741 Fig. 3 Structural equation modelling (SEM) showing the direct effects of spatial (Latitude [LAT], 742 Sine Longitude [sin(LONG)] and Cosine Longitude [cos(LONG)]), climatic (maximum temperature 743 [MAXT], minimum temperature [MINT], precipitation seasonality [PSEA] and aridity, calculated as 1-aridity index) and soil (soil organic carbon [OC] and pH) variables on the abundance of each 744 745 ecological cluster. Numbers in arrows indicate standardized path coefficients, and their width is 746 proportional to the strength of path coefficients. The proportion of variance explained (R^2) 747 appears below every response variable in the model. Significance levels are as follows *P<0.05, **P<0.01, and ***P<0.001. Model X² =2.567, P= 0.463 df= 3, Bootstrap p= 0.254. Information 748 749 on boxes 1-6 is shown in Fig. S2.



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Fig. 4 Relationships between main environmental predictors and the relative abundance (zscore) of each one of the cyanobacterial clusters. Significant (P<0.05) spearman correlation coefficients are shown on the upper part of each panel.

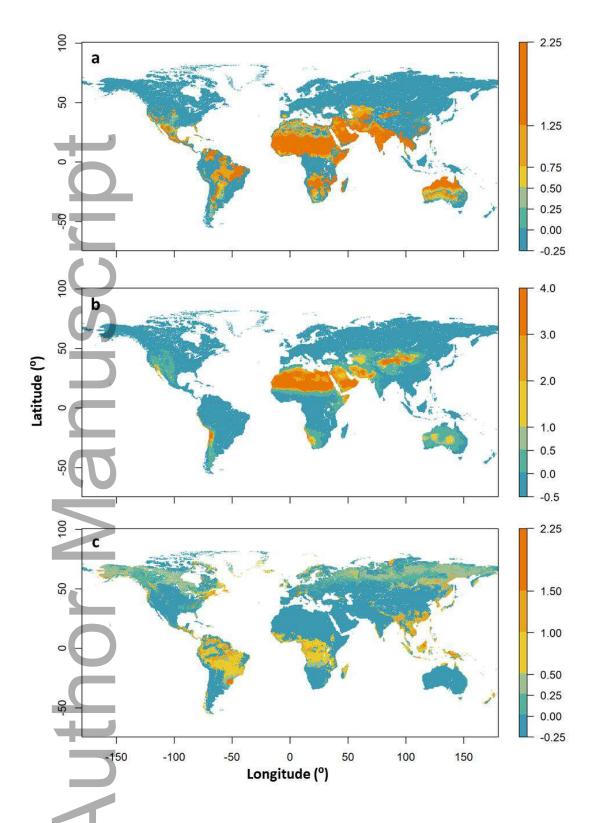
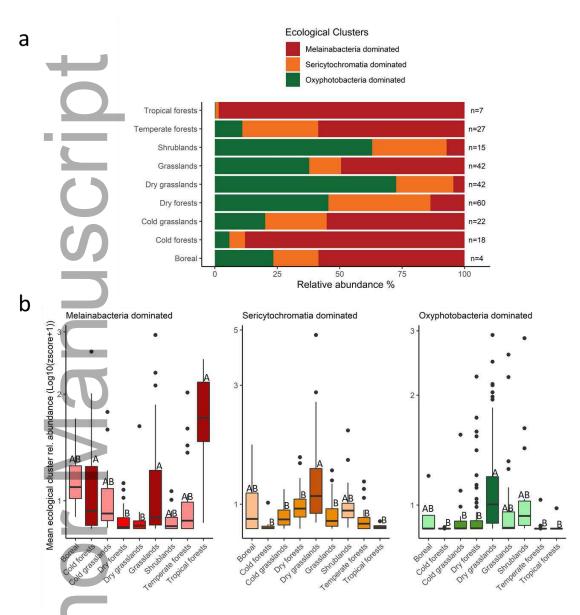


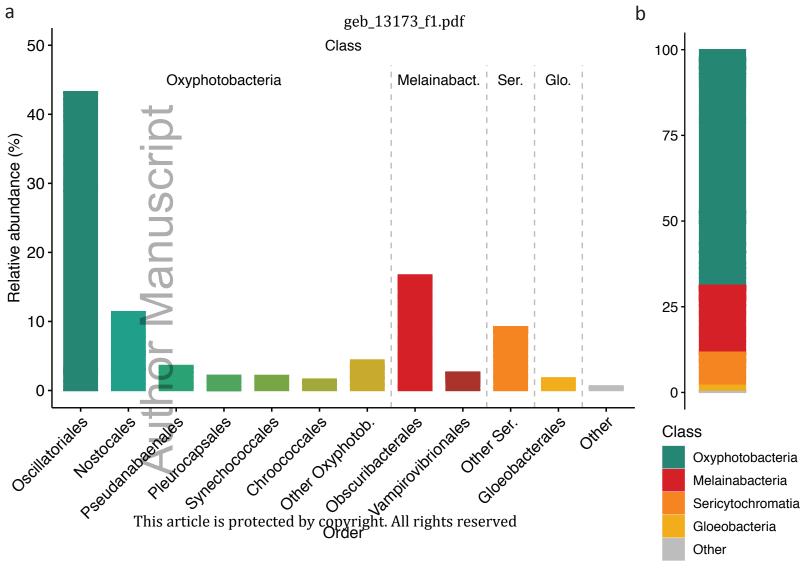
Fig. 5 Predicted global distribution of the relative abundance of the main ecological clusters of soil cyanobacteria. Percentage of variation explained by the models as follows: (a) *Oxyphotobacteria*-dominated cluster $R^2 = 0.28$; P < 0.001, (b) *Sericytochromatia*-dominated cluster $R^2 = 0.66$; P < 0.001, (c) *Melainabacteria*-dominated cluster $R^2 = 0.35$; P < 0.001. The scale bar represents the standardized abundance (z-score) of each ecological cluster. An independent

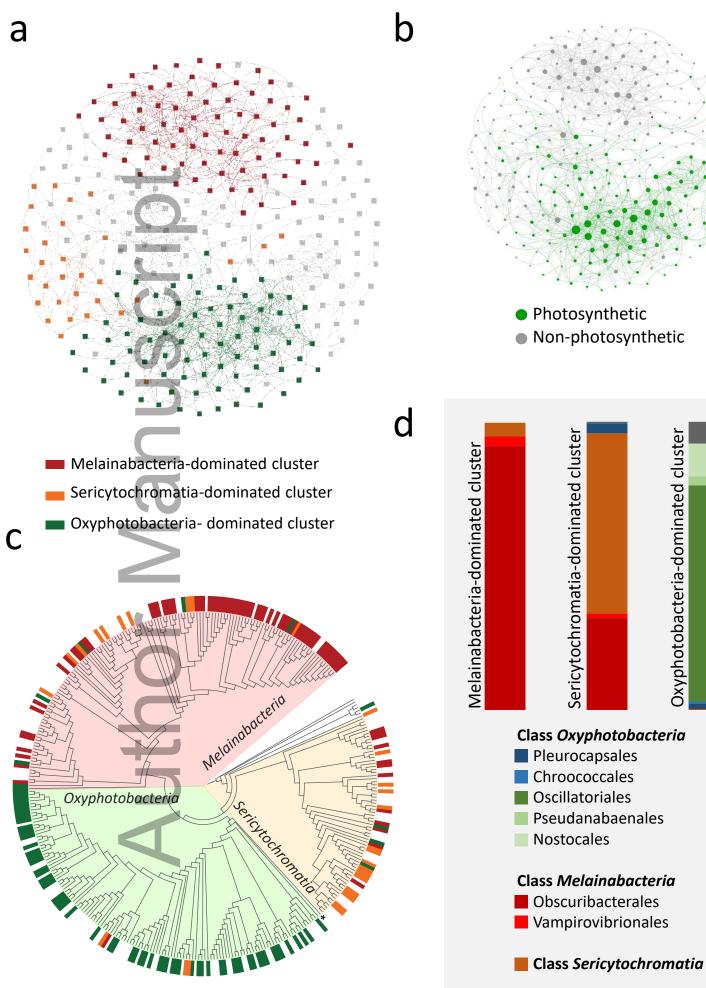
- 761 cross-validation for these maps using data from the Earth Microbiome Project (Thompson et al.,
- 762 2017) is described in the Methods section.



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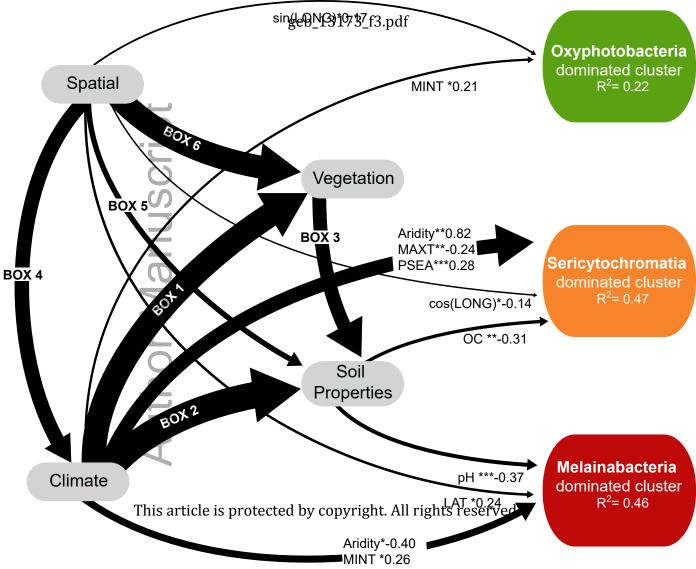
Fig. 6 Relative abundance of cyanobacterial clusters across major vegetation types. A) Stacked
 bars showing the percentage of phylotypes of each ecological cluster per vegetation type.
 n=Number of sites per each vegetation type B) Tukey HSD results testing the differences (letters
 and colour hues) in the relative abundances of each ecological cluster across vegetation types.

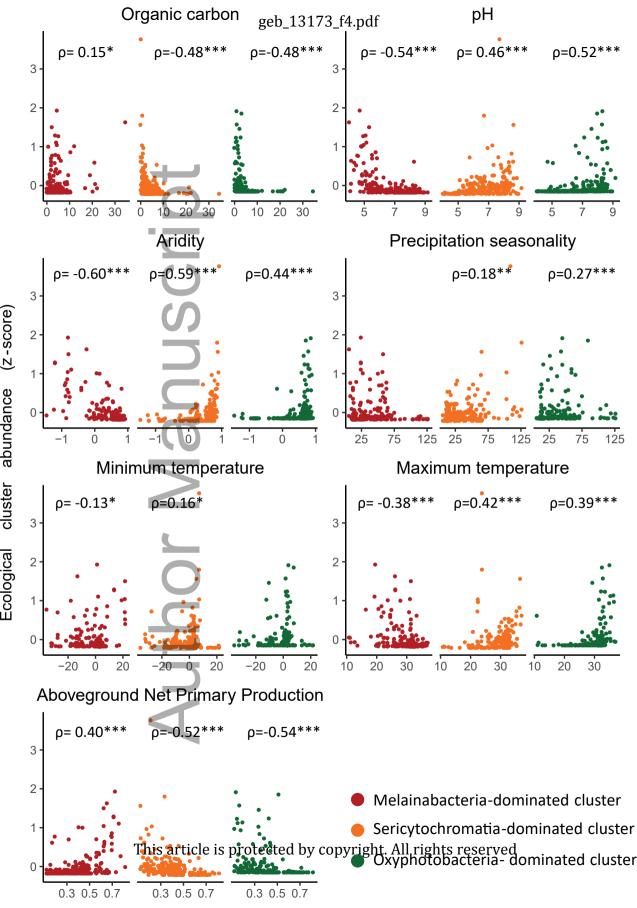


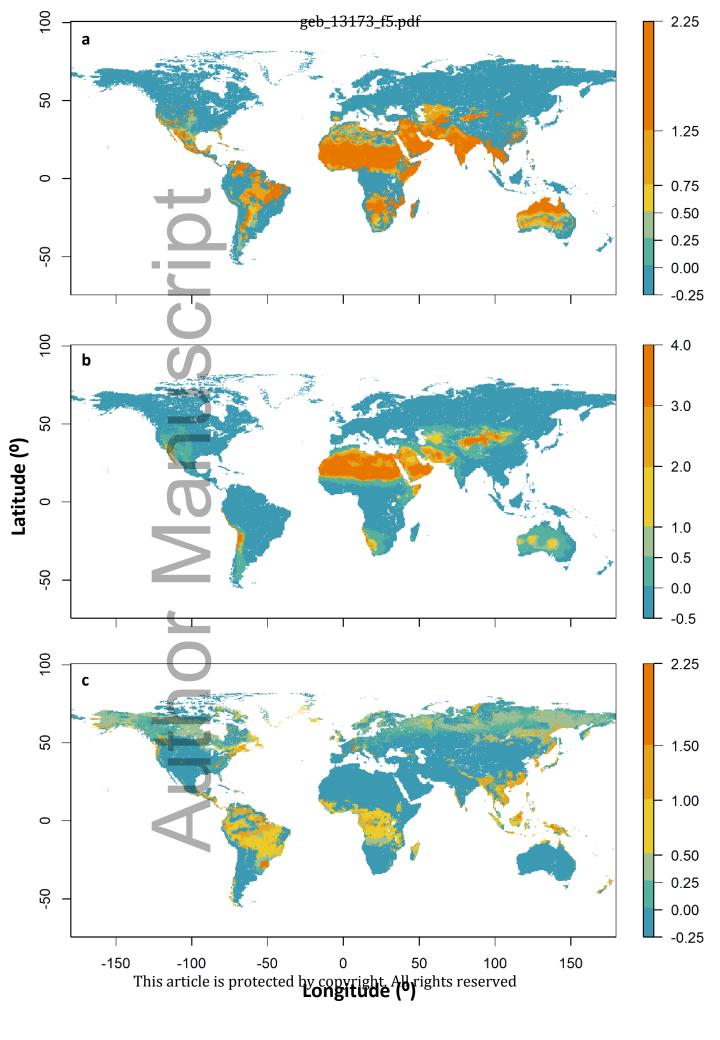


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Unassigned







gebEtalbZBca6Opldisters



Sericytochromatia dominated

Oxyphotobacteria dominated

