

1 **Comparing fishery-independent and fishery-dependent data for analysis of** 2 **the distributions of Oregon shelf groundfishes**

3
4 Rebecca A. Howard^a, howardre@oregonstate.edu*
5 Lorenzo Ciannelli^a, lorenzo.ciannelli@oregonstate.edu
6 W. Waldo Wakefield^a, waldo.wakefield@oregonstate.edu
7 Melissa A. Haltuch^b, melissa.haltuch@noaa.gov
8

9 ^a College of Earth, Ocean and Atmospheric Sciences, Oregon State University, 104 CEAOS
10 Administration Building, Corvallis, Oregon 97331, USA

11 ^b National Marine Fisheries Service, Northwest Fisheries Science Center, 2725 Montlake Blvd.
12 E, Seattle, WA 98112, USA

13 * Corresponding author
14

15 **Abstract**

16 We investigate whether fishery-dependent time series can be used to fill in spatial and temporal
17 data gaps where scientific, fishery-independent data are not available. Limitations in sampling
18 coverage combined with a historical focus on continental slope-dwelling groundfish resulted in a
19 gap in understanding Oregon's nearshore groundfish fishery. Although fisheries-independent
20 surveys have been conducted across most of the fishery's depth range, the data are limited by
21 years and seasons surveyed, as well as the absence of data for areas shallower than 55 meters
22 water depth. Fishery-dependent data are available for those shallow waters and for a broader
23 temporal range. However, these data are self-reported and the coverage was determined by
24 where fishers choose to fish. To investigate the potential for future combined uses for these data
25 sources, we analyzed spatial and temporal changes in catch rates, as well as gaps in fishery
26 (logbook) and scientific (NOAA survey) data, for six flatfishes. We found that more heavily
27 targeted species that live in deeper water, like Dover sole and petrale sole, had more spatial
28 scientific sampling coverage compared to less frequently targeted species that live in shallow
29 water, such as starry flounder and sand sole. Overlap between datasets was variable in space and
30 time but consistently higher near large ports. We identified the winter season, and the pre-2003
31 time-period as having the highest potential to benefit from complementary use of fishery-
32 dependent data. Prior to 2003, the survey design was variable and there was greater spatial and
33 temporal coverage of logbook data compared to post-2003. Integration of these data sets may be
34 useful for future research given their differences in spatial and temporal range. This work
35 provides a new perspective on the value of using fishery-dependent data to understand the spatial
36 distribution of species in habitats that are under sampled in scientific surveys.

37 **Key words:** fishery-independent, fishery-dependent, groundfish, trawl, local index of collocation
38

39 **1 Introduction**

40 Fishery-independent and fishery-dependent data are the two primary sources of information used
41 by fishery scientists for assessing population distributions over time and space. Both have their
42 advantages and disadvantages depending on the methods applied, but their combined use is
43 valuable for the management of many fisheries. For multiple regions, mapping of species
44 distributions are reliable only when using fishery-independent data, as locations are not
45 necessarily recorded correctly or at all by fishery-dependent sources, and because fishery-
46 dependent data are a function of fishery behavior (Hilborn & Walters, 1992). In areas where
47 fishers consistently document location data, visualization or mapping is more difficult due to
48 inherent biases present in fishery-dependent data, whereas there is standardization and scientific
49 sampling design built into fishery-independent data (Hilborn & Walters, 1992; Maunder & Punt,
50 2004). There can also be issues with fishery-independent surveys. These challenges largely stem
51 from a lack of time and funding, which can restrict the spatial and temporal extent of sampling,
52 but inadequate survey design or coverage can also be an issue (Hilborn & Walters, 1992).
53 Region and fishery specific research is needed to determine whether it is possible or necessary to
54 combine fishery-independent and -dependent data to improve species distribution analyses and
55 design future surveys.

56 Preferential collection of fishery-dependent data, meaning that there is no sampling design used
57 to determine where sampling occurs, results in sampling biases, as fishers target areas they
58 believe they will find their target species (or avoid non-target species) and during times of the
59 year that are either optimal or when fishing is allowed. Therefore, both spatial and temporal
60 biases are potential issues for fishery-dependent data. When using data collected
61 opportunistically for species distributions or developing an index, it is necessary to account for
62 preferential sampling when developing the models in order to avoid overestimates of abundance
63 (Pennino et al., 2019) or inaccurate management plans (Grüss et al., 2019). For example, this can
64 be accomplished by using joint (Conn et al., 2017) or marked point process (Pennino et al., 2019)
65 models to reduce bias when predicting species spatial distributions, or through spatiotemporal
66 modeling frameworks when standardizing catch-per-unit-effort data (CPUE) (Ducharme-Barth et

67 al., 2022; Grüss et al., 2019). Applying methods that minimize biases present in fishery-
68 dependent data may allow the use of this type of opportunistic data for answering a greater range
69 of questions.

70 Recent studies examine the differences between fishery-independent and fishery-dependent
71 spatial information and show that it may be possible to use fishery-dependent datasets to
72 visualize or predict species distributions. For example, fishery observer and survey data in the
73 western Mediterranean Sea display similar distribution patterns of elasmobranch bycatch species
74 despite differences in temporal and spatial coverage of the datasets (Pennino et al., 2016).
75 However, the authors also identify differences in predictive ability of each data type for temporal
76 and spatial presence and absence. Earlier studies indicate that the combination of various sources
77 of fishery-independent and -dependent data can provide a more complete picture of population
78 distributions (Pecquerie et al., 2004), allow for a better understanding of bycatch-fishery
79 interactions (Lyons et al., 2013; Murray & Orphanides, 2013), and be valuable for stock
80 assessment models (Booth, 2000). Multiple methods allow for to integrating fishery-independent
81 and -dependent data into the same model, allowing for potentially more accurate species
82 distribution models by using complementary data sources (Bourdad et al., 2017; Sant'Ana et al.,
83 2017; Zhu et al., 2018; Pinto et al., 2019; Rufener et al., 2021).

84 Numerous commercial and recreational fisheries are located in waters off the U.S. West Coast,
85 with Oregon home to the largest portion of the West Coast's groundfish trawl fishery. In 2019,
86 61 vessels participated in the non-whiting groundfish trawl fishery, generating approximately
87 \$18.7 million in ex-vessel value (The Research Group, 2021). This fishery, along with several
88 others, makes up about 8% of Oregon's coastal economy and generates about 9,200 jobs (The
89 Research Group, 2021). Logbook and landing documents from Oregon fishers provide an
90 opportunity for further study given the detailed fishery-dependent data from 1981 to the present.
91 These data include records of set and retrieval locations for fishing gear, catch weight and
92 number of each species caught, and tow duration, amongst other information. Despite the
93 existence of this long time series, these data are infrequently used to analyze population
94 distributions. Additionally, a fishery observer program initiated in 2001 is used to determine the
95 spatial distribution of discards by depth and area. There are also two fishery-independent data
96 sources spanning different time-periods. The Alaska Fisheries Science Center (AFSC) Triennial

97 survey (1977-2004) was superseded by the NOAA West Coast Groundfish Bottom Trawl Survey
98 (WCGBTS, 2003-2021, ongoing). The latter is essential for informing stock assessments in the
99 West Coast groundfish fishery.

100 Fox and Starr (1996) were the first to investigate both West Coast groundfish fishery-
101 independent and -dependent data. In their study, they compared commercial trawl catch data to
102 the AFSC Triennial survey using Geographic Information System mapping. They found that
103 differing results between the datasets included lack of spatial overlap in certain regions and a
104 discrepancy in Dover sole (*Microstomus pacificus*) biomass, likely a result of gear differences,
105 but that overall, the standardized CPUE data could supplement survey data to study fish
106 distribution and abundance (Fox & Starr, 1996). Later work used logbooks and commercial
107 landings data to investigate patterns of fishing effort, as well as visualize spatial and temporal
108 extent of fisheries (Macomber, 2000; Hannah, 2003; Bellman et al., 2005; Bellman & Heppell,
109 2007). Sampson (2011) found that while there are issues with accuracy in the logbooks,
110 information from tows on the continental shelf are typically consistent with landings data and
111 bathymetric maps, but are less accurate on the continental slope.

112 Here, we focus on flatfishes in the Oregon nearshore commercial bottom trawl fishing grounds,
113 defined as water depths of less than or equal to 200 meters off the coast of Oregon and
114 Washington states. This area largely covers the continental shelf in the northern California
115 Current. The shelf inshore of approximately 55 meters depth is understudied compared to the
116 slope and remaining portion of the shelf (Howard et al., 2021). Recently there has been renewed
117 interest in nearshore commercial trawling, which declined following the collapse of the U.S.
118 West Coast groundfish fishery in 2000 (Sjostrom et al, 2021). These nearshore fishing grounds
119 are habitat for many flatfishes, including those most heavily targeted by the groundfish fishery
120 such as petrale sole (*Eopsetta jordani*) and Dover sole, and they are the dominant group fished
121 by Oregon nearshore bottom trawlers (Sjostrom, 2019). Many of these species use the shelf
122 either for their entire lives or to complete their life cycles (Pedersen, 1975; Jacobson & Hunter,
123 1993; Abookire & Bailey, 2007; Toole et al., 2011; Krygier & Pearcy, 1986; Richmond, 1983).
124 Additionally, flatfishes display different behavior near trawl gear compared to rockfishes
125 (*Sebastes* spp.) and roundfishes, so using only flatfish species for analyses allows for a better

126 comparison between species (Ryer, 2008). Historically, flatfishes are also more consistently
127 identified to species in the logbook data compared to rockfishes.

128 With the possibility of increased fishing activity in the nearshore (Sjostrom et al. 2021),
129 increasing our understanding of species distributions in this region will contribute to sustainable
130 management efforts. Therefore, this research assesses the possibility of using the fishery-
131 independent and -dependent data to add to existing knowledge of species distributions and
132 habitat. Integrating data sets into the same model is potentially more useful when the data cover
133 different temporal and spatial ranges, possibly providing a more complete picture of the
134 distribution of a species' biomass. Because the fishery-independent and -dependent data used in
135 this study exhibits those characteristics, it may be useful to combine these data sets in future
136 research. Based on the previous research conducted using logbook and landings data, as well as
137 the limitations of the survey data sets, our first objective was to determine areas of overlap and
138 areas of complementary coverage between fishery-independent and -dependent data. Here, we
139 ascertain whether integrated use of both data types has the potential to improve the
140 characterization of habitats utilized by flatfish species present along the slope and shelf areas.
141 Our second objective was to elucidate which species may be better represented by the
142 complementary use of both types of data. Finally, we determine the potential for use of fishery-
143 dependent data in species distribution models for species poorly represented by available fishery-
144 independent data.

145 **2 Material and methods**

146 **2.1 Fishery-dependent data**

147 Logbook and landings data were obtained from the Oregon Department of Fish and Wildlife
148 (ODFW). Logbook records included information for catch rates of individual species, tow
149 duration, average depth of tow, offload port, gear type, and location of each tow. Catch rates for
150 each species were recorded as the original hail weight and an adjusted weight. Hail weight is the
151 original weight recorded by the vessel captain. For adjusted weight, ODFW applied five
152 classifications for each trip or tow to make any appropriate corrections to the hail weight to
153 obtain these values. These corrections were made using the fish ticket pounds, which are a record
154 of catch documented at the port of landing, and depended on errors identified in the data. For

155 example, this may have included adding information to a logbook record for a species without a
156 haul weight but that was recorded on the fish ticket. Because haul weights were not recorded in
157 early years, the adjusted weights were used for our analyses. The logbook CPUE was calculated
158 as kilograms per hour (kg/hr) using the recorded tow duration and catch weight.

159 To prepare the data for analysis, we limited the logbook records to areas fished by Oregon
160 commercial fishers, between 42 and 47°N (Figure 1), consistent with Sjostrom et al. (2021).
161 Only tows completed using the three most common nearshore groundfish gear types were
162 retained: unspecified bottom trawl, sole net, and selective flatfish trawl (Sjostrom et al. 2021).
163 Tows with missing critical data, such as set latitude or longitude, were removed and locations
164 recorded using the less accurate Loran A and Loran C navigation systems were not used. Tows
165 appearing on landmasses were also removed. Sampson (2011) determined that depth and
166 location consistency was high for catches at depths less than 200 meters, indicating relatively
167 high accuracy for the shelf region, and illustrated that there exist few incentives to misreport
168 data. However, due to numerous previously identified inaccuracies with depth reporting in the
169 logbooks, and the variability in technological capabilities over time, it was deemed best to
170 replace the logbook depth values rather than remove inaccurate records (Sampson, 2011;
171 Sjostrom, 2019). A locally estimated scatterplot smoothing curve (LOESS), a type of
172 nonparametric regression, was used to replace all logbook depth measurements based on
173 recorded position (longitude and latitude). Bathymetric data for the LOESS curve was obtained
174 from the NOAA ETOPO1 global relief model (Amante & Eakins, 2009). The final model (adj. r^2
175 = 0.99), which used a 2nd degree polynomial and 0.01 span, was then used to predict depths for
176 the recorded locations of each logbook tow. Tows with model depths greater than 200 meters or
177 shallower than 10 meters were removed, providing a dataset with 85,143 hauls. Landing and
178 logbook records were joined using individual tow catch ID numbers and the landed weights from
179 the fish tickets were used to calculate total landings over time.

180 **2.2 Fishery-independent data**

181 Both the triennial (1977 – 2004) and WCGBTS annual (2003 – 2018) survey data sets were
182 available for these analyses. For the triennial survey, Alaskan class trawl vessels equipped with a
183 Poly Nor'Eastern trawl were used to sample transects running perpendicular to the coast.
184 Sampling depth, transect spacing, and timing of the survey were variable, but tow duration was

185 consistent at 30 minutes (Dark & Wilkins, 1994; Keller et al., 2017). Until 1995, sampling
186 occurred from July to September or October; after 1995, sampling began in June and ended in
187 October. The annual survey began in 2003 and uses West Coast class vessels equipped with an
188 Aberdeen trawl, towed for 15 minutes at each station. Stations are selected using a stratified
189 random sampling of cells placed along the U.S. West Coast shelf and slope (Keller et al., 2017).
190 The first pass for the survey occurs from May to August and the second pass occurs from August
191 to October. Both the triennial and annual surveys cover the region trawled by the Oregon
192 commercial groundfish fishery, but the sampling density is greater for the annual survey. Survey
193 data were obtained from the NWFSC/FRAM Data Warehouse (url:
194 <https://www.nwfsc.noaa.gov/data/>) as a dataset containing characteristics from each tow and a
195 dataset containing information on species composition. Tow ID numbers were used to match
196 tows to the species composition data. Only data collected at 200 meters depth or shallower was
197 retained, with the shallowest depth at approximately 55 meters. Using the logbook records, the
198 survey data was restricted to waters consistently fished by Oregon commercial fishers, between
199 42 and 47°N. Because complete logbook data was only available between 1981 and 2017, the
200 survey data used below were limited to this period as well, leaving 4,343 tows for the analyses.
201 The survey CPUE was recorded as kilograms per hectare (kg/ha). To avoid overlap of the
202 triennial and annual survey data and ensure there were similar sets of years within the two data
203 sets per period, the 2004 data for the triennial survey was removed.

204 **2.3 Species selection and visualization**

205 Six species of flatfish were chosen for analysis: petrale sole, Dover sole, sand sole (*Psettychthys*
206 *melanostictus*), starry flounder (*Platichthys stellatus*), English sole (*Parophrys vetulus*) and
207 Pacific sanddab (*Citharichthys sordidus*). Petrale sole and Dover sole are the flatfishes most
208 consistently caught by both the survey and commercial fishery, with distributions known to span
209 much of the shelf. Starry flounder and sand sole are two shallow-water species not commonly
210 caught by the surveys or fishery, while English sole and Pacific sanddab are mid-shelf species
211 with some commercial importance (Table 1). Pacific sanddab catch rates were not separated out
212 from speckled sanddab (*Citharichthys stigmaeus*) in logbook records, but speckled sanddab, a
213 similar but smaller species, are assumed to be infrequently encountered by commercial trawl

214 vessels (He et al., 2013). Therefore, the unspecified sanddab grouping was used to assess catch
215 rate of Pacific sanddab.

216 **Table 1:** Species below were selected for analyses. Values represent the number of hauls that contained each
217 species.

Species	Survey	Logbook
Dover sole	3,925	53,460
petrale sole	3,422	56,207
English sole	3,135	54,618
Pacific sanddab	2,678	20,681
sand sole	174	20,608
starry flounder	62	19,315

218 Fishery-independent and -dependent maps and plots for each species were first compared
219 visually before gridding for confidentiality purposes to determine if differences or similarities
220 were present in spatial distribution over time. Dover sole and petrale sole are known to
221 seasonally migrate on and off the shelf, making them useful for investigating seasonal
222 differences in distribution not observed by the NOAA trawl data, which do not capture winter
223 distributions (Ketchen & Forrester, 1966; Pedersen, 1975; Jacobson & Hunter, 1993; Abookire
224 & Bailey, 2007). Therefore, for these species, separate visualizations of the fishery-dependent
225 data were created for months with and without NOAA survey samples.

226 For the spatial visualizations, we gridded maps of species catch and overall fishing effort for four
227 periods (1981 – 1989, 1990 – 2001, 2002 – 2009, 2010 – 2017) using both survey and logbook
228 data, hereafter referred to as the 1980s, 1990s, 2000s, and 2010s respectively. The logbook data
229 were limited to the period of during survey data collections (May to October) to allow for
230 comparison with the NOAA trawl data. We constructed maps with a grid covering 42° to 47°N
231 latitude and 125° to 123.9°W longitude composed of cells with dimensions of 20 north-south
232 intervals (approximately 27.8 kilometers) and 15 east-west intervals (approximately 5.8
233 kilometers), to allow for fishing location confidentiality. To investigate effort distribution, maps
234 of overall fishing effort, measured as average number of tows per year and location, were created
235 for each period. For individual species maps we used scaled catch rates to allow for comparison
236 of species distributions between data sets. First, survey and logbook CPUE was transformed
237 using an $\ln(x+1)$ transformation and each grid cell represented the mean transformed CPUE in
238 that area during the period depicted. Other data transformations were examined but did not affect

239 our conclusions. CPUE was used rather than total catch because tows conducted by fishers can
 240 be of varying length, which typically leads to larger catches for longer tows. Mean survey and
 241 logbook CPUE for each grid cell was calculated as follows, with N representing number of tows:

$$242 \quad CPUE_{avg,lat,lon} = \sum \frac{\ln(CPUE_{lat,lon}+1)}{N_{lat,lon}} \quad (1)$$

243 Then, for comparison between datasets, the logbook data was restricted to the months the survey
 244 operates and the resulting mean survey and logbook CPUE values were scaled from 0 to 1.

245 To examine differences in depth distribution across time and data types, contour plots of density
 246 using two-dimensional kernel density estimation with an axis aligned bivariate Gaussian kernel
 247 evaluated on a square grid were created using the ggplot2 stat_density_2d function (Wickham,
 248 2016). Only points of presence for each species were used and plots were divided into the same
 249 four periods as above.

250 **2.4 Quantitative comparison**

251 To quantitatively assess the level of overlap between the CPUE of the fishery-independent and -
 252 dependent data, we calculated the local index of collocation (LIC). The LIC is a correlation
 253 coefficient that is not sensitive to zero-inflation and assesses the overlap of data across two
 254 sources over the same spatial bounds (Bez & Rivoirard, 2000; Pianka, 1973). The LIC has
 255 previously been used to determine overlap with respect to population abundance (Kotwicki &
 256 Lauth, 2013; Petrik et al., 2015; Carroll et al., 2019), predator-prey interactions (Trenkel et al.,
 257 2005; Carroll et al., 2019), and fishing effort (Bez et al., 2011; Pointin et al., 2019). Values range
 258 between 0 and 1, with lower values indicating less “collocation” (correlation of CPUE) between
 259 the two datasets. The total LIC for a given period and the entire region can be calculated as
 260 follows (Pianka, 1973; Petigas et al., 2017; Carroll et al., 2019):

$$261 \quad LIC_{total,t} = \frac{\sum_i^n (D_{lat,lon,t,i} * I_{lat,lon,t,i})}{\sqrt{\sum_i^n D_{lat,lon,t,i}^2 \sum_i^n I_{lat,lon,t,i}^2}} \quad (2)$$

262
 263
 264 In the above equation for total LIC for a given period, *D* represents the fishery-dependent
 265 untransformed logbook CPUE data, *I* represents the fishery-independent untransformed survey

266 CPUE data, and t represents the period. Measures of uncertainty in the overall LIC were also
267 calculated. This was done by determining overall LIC for each year within a period and
268 calculating the mean and standard deviation (Table S1). To look at spatially explicit overlap, we
269 modified Equation 2 to use CPUE in each grid cell for a given species. This gives the
270 contribution of overlap per cell to the total overlap across the study region, providing a spatially
271 explicit value (Carroll et al., 2019):

$$LIC_{lat,lon,t} = \frac{D_{lat,lon,t} * I_{lat,lon,t}}{\sqrt{\sum_i^n D_{lat,lon,t,i}^2 \sum_i^n I_{lat,lon,t,i}^2}} \quad (3)$$

274
275 Only logbook data are available in areas shallower than 55 meters water depth, so for
276 comparison purposes the fishery-independent mean survey CPUE value for those unsampled grid
277 cells was artificially set to 0, as there was no survey data but logbook data were available. As
278 seen in Equation 3, this results in an LIC of 0, or no overlap, for areas shallower than 55 meters
279 water depth. This allowed us to illustrate the lack of survey data in shallow areas. All analyses
280 and data processing were completed using R statistical software (v. 4.0.4, R Core Team, 2021).

281

282 **3 Results**

283 **3.1 Fishing effort and catch**

284 A linear regression of $\ln(x+1)$ transformed logbook CPUE against vessel length showed little to
285 no relationship between length and catch rate, which indicated that CPUE does not significantly
286 change with vessel length. Overall fishing and sampling effort was distributed differently in
287 space for the fishery-independent and -dependent data, as well as when comparing the triennial
288 and annual surveys (Figure 2). Fishing effort in general on the inner shelf decreased over time.
289 More effort shifted offshore and northward toward Astoria, Oregon over time and there was a
290 decrease in effort across the entire region. Fishing effort in the most recent years (2010-2017)
291 appears to have decreased substantially, with most effort located off the coast near northern
292 Oregon and southern Washington. The survey sampling effort is more evenly distributed, though
293 there was higher average effort and patchy distribution during the triennial survey periods (1981-
294 1989 and 1990-2001). In the most recent two periods (2002-2009 and 2010-2017), sampling

295 effort can be seen in the visualizations to be far more evenly distributed across the shelf, which
296 was expected with the change from transects to stratified random sample design.

297 **3.2 Spatiotemporal visualization of catch**

298 Geographic differences in spatial distribution of catch rates during the survey period were
299 evident between the survey and logbook data types. Maps of average survey and logbook catch
300 rates per year for the two commonly caught species, petrale sole and Dover sole, showed similar
301 patterns of spatial distribution over time within each data type (Figures 3, S1). The highest
302 average logbook catch rate of petrale sole was centered along the 100-meter isobath throughout
303 much of the shelf during the summer, whereas average survey catch values were similar
304 throughout the study area (Figure 3). Dover sole logbook and survey catch rates were highest
305 near the outer edge of the shelf, and in the most recent period, catch rate was highest north of
306 45°N (Figure S1). For the inshore shallow-water species, the logbook data indicate that nonzero
307 catch occurred almost entirely within areas shallower than 100 meters water depth. Sand sole
308 logbook catch rates were consistent across the four periods throughout this depth range (Figures
309 4). Starry flounder logbook catch rates were highest near Astoria, Oregon and decreased
310 southward (Figure S2). Both species were sporadically captured by the survey, and the primary
311 location of tows containing either species were near Newport, Oregon at shallow depths. Survey
312 tows containing Pacific sanddab and English sole were more common across the study area than
313 tows containing the inshore species (Figures 5, S3). Pacific sanddab average survey and logbook
314 catch rates were highest off the shelf near Charleston, Oregon and Newport, though this patch
315 of higher catch rates appeared to shift south over time in visualizations of the logbook data
316 (Figure 5). There were also more tows containing Pacific sanddab north of Astoria. English sole
317 logbook catch rates were highest in the southern region and on the wide section of the shelf near
318 Newport (Figure 5).

319 Depth contour plots and maps revealed the discrepancy in the coverage of spatial sampling
320 between the survey and commercial data. All species commonly caught by the survey (petrale
321 sole, Dover sole, English sole, and Pacific sanddab) had catch rates distributed evenly throughout
322 much of the sampling area (Figures 6, S4-S6). Patches of high presence were visible in areas on
323 the shelf between approximately 44° and 46°N. Petrale sole, Dover sole, and English sole, were
324 present in samples throughout all depths and latitudes sampled by the survey and included in this

325 study but were caught in precise patches by the commercial fishery (Figures 6, S4-S5). Pacific
326 sanddabs were caught consistently across latitudes by the survey but had a clear maximum depth
327 limit at about 100 meters (Figure S6). Recorded logbook catch rates for the infrequently caught
328 shallow-water species, starry flounder and sand sole, occurred almost exclusively inshore of the
329 70-meter isobath (Figures S7-S8). High catch rate areas for English sole and Pacific sanddab in
330 the logbook data also occurred in shallow water, but substantial catch rates did extend farther
331 offshore than for the shallow-water species (Figures S5-S6). Dover sole and petrale sole were
332 found on the shallow portion of the shelf, inshore of 55 meters, but most catch rates were in
333 deeper water (Figures 6, S4). In recent years, the shallow water distribution of these species has
334 been found entirely near Astoria. The 55-meter depth limitation for the survey is clearly visible
335 in the survey contour plots for these two species as well as English sole and Pacific sanddab.
336 With the logbooks, there is obvious movement northward of overall fishing effort.

337 **3.3 Seasonal Variation**

338 The seasonal variation maps for Dover sole indicated a difference between population
339 distribution during the “winter” (November – April) and “summer” (May – October). During the
340 summer, the mapped Dover sole mean logbook catch values for each period were higher when
341 compared to the winter (Figures 7). Most nonzero values for Dover sole mean logbook catch
342 rates were found on the outer shelf during the winter whereas there were shallower catches
343 during the summer. Maps for petrale sole generally showed similar patterns of mean logbook
344 catch rates for each period when comparing the winter and summer (Figure S9). However, mean
345 logbook catch rate values for petrale sole were lower on the inner shelf north of 45°N in the
346 winter and there were higher values on the outer shelf during winters in the 1980s and 1990s.
347 Contour plots for both species showed that there were more catches in deeper water during the
348 winter compared to the summer (Figures 5, S4, S10-S11).

349 **3.4 LIC**

350 Maps of LIC indicated highly variable levels of overlap between fishery-independent and -
351 dependent data, depending on the species. Dover sole had the highest and most consistent degree
352 of total overlap per period, followed by Pacific sanddab and petrale sole except for in the 1980s,
353 when LIC for petrale sole was much lower (Table 2). English sole had an intermediate amount of
354 total overlap, and the total LIC values for sand sole and starry flounder were low in each period.

355 Maps for each species provided a more detailed image of where collocation occurred (Figures 8,
 356 S12-S14). Petrale sole and Dover sole both had spatial overlap between data sources on much of
 357 the shelf between the 55- and 200-meter isobaths (Figures 8, S12). The highest spatially explicit
 358 LIC values for Dover sole were located near the 200-meter isobath, whereas high values for
 359 petrale sole were more evenly dispersed throughout the 55- to 200-meter depth range. Pacific
 360 sanddab had some of the highest total LIC values of all the species, though the spatially explicit
 361 values were largely centered around the mid-latitude area of the shelf similarly to sand sole and
 362 starry flounder (Figures 8, S13). English sole had more variable spatial overlap than Pacific
 363 sanddab and the average total LIC values for each period were lower than those for Pacific
 364 sanddab (Figure S14).

365
 366 **Table 2:** Overall LIC values calculated for each period for each species. Lower values indicate less overlap between
 367 the fishery-independent and -dependent data.

Species	1980s	1990s	2000s	2010s
starry flounder	0.09	0.01	0.12	0.02
petrale sole	0.14	0.46	0.75	0.57
sand sole	0.18	0.19	0.29	0.02
English sole	0.37	0.29	0.50	0.37
Pacific sanddab	0.53	0.59	0.50	0.67
Dover sole	0.67	0.76	0.78	0.53

368

369 **4 Discussion**

370 With new opportunity to revitalize the nearshore fishery, it is crucial for managers to understand
 371 the current and past ecology of Oregon’s continental shelf. Should species like starry flounder
 372 and sand sole become increasingly targeted or caught as bycatch, spatiotemporal data from
 373 logbooks could allow for better future assessment and a more complete historical baseline
 374 dataset. English sole and Pacific sanddab are perhaps more likely than the other species to gain
 375 market interest in Oregon, given their previous importance as part of the nearshore fishery (He et
 376 al., 2013; Cope et al., 2015). In the case of Pacific sanddab, the species has had a moderate
 377 resurgence in popularity in California. While over 90% of the petrale sole quota is caught each
 378 year, this is not true for Dover sole because sablefish is a co-occurring choke species, and there
 379 is limited market interest unless sablefish quota increases. Despite this, there are current efforts
 380 to increase Dover sole quota utilization, particularly for the northern region, so the Oregon

381 nearshore may experience more fishing activity as a result (SaMTAAC, 2019). Additionally,
382 there are ongoing attempts to increase consumption of Oregon seafood through industry-driven
383 groups like Positively Groundfish and Oregon Sea Grant's Eat Oregon Seafood initiative.
384 Utilizing fishery-independent and -dependent data can assist in this endeavor, providing more
385 context for potential bycatch interactions and changes in community composition due to
386 distributional shifts over time. This is not only useful for groundfishes with directed fisheries, but
387 also for those caught less frequently.

388 Lack of survey data (see Table 1) has made it difficult to determine whether there have been
389 spatial shifts within starry flounder and sand sole populations, which could be partially remedied
390 by the use of logbook data. While starry flounder and sand sole are not among the most heavily
391 targeted flatfishes, commercial trawl vessels consistently land them, albeit in smaller amounts
392 than other species (Pearson & McNally, 2005; Ralston, 2005). Both species are considered data-
393 limited by the PFMC, and there is uncertainty about the status of either species (PFMC, 2018).
394 There is not presently a dedicated fishery for starry flounder and sand sole, though they were
395 once more frequently pursued as part of a nearshore mixed flatfish fishery (Alverson et al., 1964;
396 PFMC, 2018). There was very little overlap between fishery dependent and independent data for
397 starry flounder and sand sole catch rates (Table 2), but logbook data indicate both species are
398 present throughout the inshore portion of the shelf (Figures 4, S2). Therefore, logbook data may
399 prove useful for future research on the changes in distribution of starry flounder and sand sole.

400 Data comparison and visualization of petrale sole, Dover sole, English sole, and Pacific sanddab
401 presented very different results from the shallow-water species, in particular, the deeper water
402 species petrale sole and Dover sole. Petrale sole was the only flatfish species previously
403 categorized as overfished by the PFMC and for which nearly all quota was caught each year
404 (Haltuch & Hicks, 2009; PFMC, 2018; Wetzel, 2019). The petrale sole shelf population
405 distribution off Oregon and Washington exhibits decreases in areas that now regularly
406 experience hypoxia but increases off the coast of southern Oregon and on the outer shelf off the
407 coast of northern Oregon and Washington (Howard et al., 2021). There has been an overall
408 increase in presence of Dover sole, especially on the northern shelf (Howard et al., 2021). These
409 changes may explain the increase in mean logbook catch rate per year of Dover sole on the shelf
410 near Astoria seen in the most recent period and the increase in logbook catch rates for petrale

411 sole off the southern Oregon Coast since the 1990s (Figures 3, S1). Although there is not
412 consistent catch of either species in the shallowest portion of the shelf, the depth distribution
413 plots show that the shallow portion of both populations were missed by the NOAA surveys,
414 especially near Astoria (Figures 6, S4). Thus, increased sampling inshore of the 55-meter isobath
415 could provide new information on the distributions of these species and should be considered as
416 an addition to the current sampling protocols for the WCGBTS. This was especially necessary
417 for detecting any population expansion shoreward, which may be occurring for both species
418 given their increase in presence on the shelf and recent higher logbook catch rates inshore
419 (Figures 3, S1). This increase is expected for petrale sole as their population increased in recent
420 periods from low numbers in the 1990s (Haltuch et al., 2013; Wetzel, 2019). Habitat
421 compression, which has the potential to reduce habitat availability and change predator-prey
422 interactions, is also possible due to factors such as hypoxia (Stramma et al., 2010; Stramma et
423 al., 2012). Some species may shift to more oxygenated areas while other hypoxia tolerant
424 species, like Dover sole, may remain.

425 For Dover sole and petrale sole, there is also information missed by the survey due to the timing
426 of sampling. While most fishing occurs during the summer months, fishing happens outside of
427 those months and the winter community composition is different due to seasonal migrations.
428 Therefore, for species like petrale sole, both winter and summer fishery-dependent data have
429 been investigated during the stock assessment process and these data may also be useful for
430 determining change in the distributions of the winter spawning aggregations targeted by fishers.
431 In addition, the winter petrale sole fishery-dependent CPUE indices have been considered more
432 consistent due to fewer changes in spatial management during that time of year (Haltuch et al.,
433 2011; Haltuch et al., 2013). However, this study was limited to the shelf, which explains the
434 minor differences between seasons for petrale sole, as the spawning aggregations occur in deeper
435 water (Haltuch & Hicks, 2009). Future work could investigate inclusion of winter fishery-
436 dependent data in research on seasonally migratory species like petrale sole and Dover sole.

437 The final two species, Pacific sanddab and English sole, portrayed a middle ground between the
438 other four species. Survey data indicated both species are predominantly found on the inner
439 shelf, but this did not necessarily hold true for the logbooks. The narrow latitudinal band of
440 Pacific sanddab commercial catch rates with patches of high catch rates suggests specific

441 geographic areas where either Pacific sanddab are targeted or retained, or where there are high-
442 density clusters (Figure 5). Recent research found a similar distributional pattern for juvenile
443 Pacific sanddab, indicating that these patches in the logbook data are likely groups of juveniles in
444 nursery grounds (Tolimieri et al. 2020). High English sole commercial catch rates were variable
445 and may be associated with either ports or proximity to the estuaries where English sole live as
446 juveniles (Krygier & Percy, 1986; Gunderson et al., 1990). Despite the similar inner shelf
447 spatial distributions for English sole and Pacific sanddab, overlap metrics were surprisingly
448 different. The differing amounts of survey and logbook dataset overlap for Pacific sanddab and
449 English sole, as well as Dover and petrale sole, was unexpected. It was clear that a high total LIC
450 for a given species does not necessarily indicate that there would also be an even distribution of
451 spatially explicit LIC values. For example, in the case of Pacific sanddab and Dover sole there
452 were isolated locations with high spatially explicit LIC values while the petrale sole and English
453 sole spatially explicit LIC values were more evenly distributed throughout the study region
454 (Figures 8, S12-S14). This showed that total overlap does not necessarily correspond to how
455 heavily the species is targeted, as in high LIC does not indicate a heavily fished species, and that
456 fishery-dependent data may be valuable for filling in spatial gaps not just for shallow water
457 populations, but also for species like Dover sole and Pacific sanddab. Evaluation of the utility of
458 combining these data sources for spatiotemporal analyses should therefore be determined on an
459 individual species basis. Species with low spatially explicit or total LIC values, shallow
460 distributions, or seasonal variability would be potential candidates for future work as long as
461 they are reliably identified to species within logbook records. Additionally, this research only
462 examined logbook data from the Oregon fleet and only catches that occurred on the shelf. The
463 flatfishes in this study are distributed throughout the California Current System and are found
464 both on and off the shelf, which necessitates further investigation into additional data sources
465 and locations that are beyond the scope of this project.

466 Fishery-independent and -dependent data have been shown to be complementary when mapping
467 fish distributions in other regions (Pecquerie et al., 2004; Stallings, 2009; Murray & Orphanides,
468 2013; Lyons et al., 2013) and this is the case for the 1980s and 1990s and for starry flounder,
469 sand sole, Pacific sanddab, and English sole in particular. The logbooks provide a large quantity
470 of self-reported data, while the survey data sets are smaller but implement random sampling and
471 have more uniform spatial coverage. Visualization of the fishery-independent and -dependent

472 effort associated with the Oregon nearshore trawl fishing grounds illuminated clear
473 spatiotemporal disparities between the datasets. This lack of agreement happens for several
474 reasons. First, the NOAA surveys do not sample inshore of the 55-meter isobath and there were
475 fewer survey years in the 1980s and 1990s, when the surveys were triennial, than in the 2000s
476 and 2010s. Secondly, even though the fishery annually targets the inshore, total fishing effort
477 during the 2000s and 2010s declined and became concentrated to the north. This leaves a
478 noticeable gap in fishery-dependent data in the south and shoreward of 55 meters in the recent
479 periods. In addition, comparing species distributions from the two surveys requires caution,
480 given the different survey designs. The triennial surveys sampled different depth and latitudinal
481 ranges throughout its duration that differ from those used consistently during the annual surveys
482 that began in 2003.

483 Northerly movement of fishing effort in recent periods, toward Astoria and Newport, took place
484 because the northern region is where most of the fleet is homeported and where factories that
485 process non-whiting groundfish are located (Warlick et al., 2018). This concentration of effort
486 around larger ports or those with processing plants illustrates one reason why the NOAA surveys
487 can provide more spatially consistent and statistically useful data. Although in general there is
488 more fishing near ports, in the 1980s and 1990s these ‘biases’ toward specific ports were less
489 evident. Numerous processing facilities were available to fishers at most ports prior to the
490 intensified regulation and management that followed the fishery’s collapse in 2000 (Hanna,
491 2000; Warlick et al., 2018). Management and policy changes in the early 2000s led to
492 consolidation and a federal buyback of vessels, resulting in only five major processing facilities
493 for non-whiting groundfish. Therefore, the period most suitable for potential use of fishery-
494 dependent data to fill in fishery-independent gaps for flatfish is likely during the 1980s and
495 1990s. For spatial analyses, this likely requires use of preferential sampling models to account
496 for vessels targeting or retaining flatfish (Diggle et al., 2010; Pennino et al., 2016). Caution
497 should be taken when using logbook data for the 1980s and 1990s because the technology
498 available to record fishing locations was less accurate, compliance was variable, and bycatch was
499 not recorded (Sampson, 2011; Sjostrom, 2019). The WCGBTS annual shelf/slope survey that
500 began during 2003 instituted a grid-based stratified random sampling with better spatial coverage
501 (Figure 2), a greater depth range, and a consistent period each year. Logbook data from the
502 2000s and 2010s show the fishery concentrated in the north so fishery-dependent data may be

503 redundant with the WCGBTS for shelf flatfish, such as petrale sole and Dover sole. Additionally,
504 this study focuses only on the shelf portion of the petrale sole and Dover sole range off the
505 Oregon and Washington coasts, so the WCGBTS is better suited for use in stock assessments as
506 their populations extend onto the slope, into Canadian waters, and off the California coast.

507 Fishery-dependent data collected by the Oregon nearshore trawl fleet has potential applications
508 for investigating the distributions of groundfishes that are spatially under sampled by scientific
509 surveys or for species that migrate seasonally. This is crucial for thoroughly understanding
510 habitat use of coastal species throughout the year in light of increased use of nearshore waters.
511 The shelf waters off the Oregon Coast are shared by a number of industries beyond the
512 groundfish trawl fleet, including other commercial fisheries, telecommunications, tourism, and
513 wind and wave energy. This shared use of marine resources is not unique to Oregon, so this
514 research highlights the potential benefits of incorporating additional data sources to better inform
515 future management decisions in areas beyond Oregon where fish habitat is also used by other
516 stakeholders. To effectively utilize fishery-dependent data in these contexts, additional research
517 should focus both on developing appropriate methods for use of opportunistic data to study
518 distributions of under sampled species as well as exploration of methods that can integrate both
519 data types for investigation of seasonally migratory species. These possible directions of future
520 research and the increased incorporation of fishery-dependent data in spatiotemporal analyses
521 have the potential to provide more accurate predictions of species distributions and therefore lead
522 to better informed management decisions.

523 **Acknowledgements**

524 We thank the NOAA Fisheries AFSC triennial and NWFSC West Coast Groundfish Bottom
525 Trawl survey teams, survey volunteers, and the captains and crews of the chartered fishing
526 vessels. We also thank the Oregon Department of Fish and Wildlife for providing access to the
527 logbook and fish ticket databases. This research was funded by Oregon Sea Grant (grant number:
528 NA18OAR4170072).

529

530 **Declaration of interests**

531 The authors declare that they have no known competing financial interests or personal
532 relationships that could have appeared to influence the work reported in this paper.

533

534 **CRedit authorship contribution statement**

535 **Rebecca Howard:** Methodology, Formal analysis, Visualization, Writing – Original draft
536 preparation, Writing – reviewing and editing. **Lorenzo Ciannelli:** Conceptualization,

537 Methodology, Supervision, Funding acquisition, Writing – reviewing and editing. **Waldo**

538 **Wakefield:** Conceptualization, Supervision, Writing – reviewing and editing. **Melissa Haltuch:**

539 Methodology, Writing – reviewing and editing.

540 **Figure Captions**

541 **Figure 1:** Map of the study region off the coasts of Washington (WA) and Oregon (OR), along
542 with relevant ports. Contour lines represent the 50-meter and 200-meter isobaths.

543
544 **Figure 2:** The top four panels depict fishing effort by Oregon commercial fishing vessels. The
545 bottom four panels depict survey sampling effort over the same period. Each grid cell represents
546 average fishing effort per year for each period.

547
548 **Figure 3:** The panels above depict the spatial distribution of catch rates for petrale sole over four
549 periods. The top four panels depict data from logbooks and the bottom four panels depict data
550 from the NOAA surveys. Grid cells indicate scaled catch in each area.

551
552 **Figure 4:** Same as Figure 3, but for sand sole.

553
554 **Figure 5:** Same as Figure 3, but for Pacific sanddab.

555
556 **Figure 6:** Contour maps depicting depth distribution of petrale sole for the logbooks (upper four
557 panels) and for the survey (lower four panels).

558
559 **Figure 7:** The panels above depict the spatial distribution of Dover sole for two different times
560 of the year using logbook data. The bottom four panels depict data from the portion of the year
561 not sampled by the survey, which ranges from October to mid-May. The top four panels depict
562 tows conducted during the same months the survey samples. Grid cells indicate average catch in
563 each area.

564
565 **Figure 8:** The four panels for each species depict average LIC in each grid cell for each period.
566 Darker colors indicate collocation while lighter colors indicate higher LIC. Unfilled grid cells
567 (light blue) indicate areas without both fishery-independent and -dependent data. The contour
568 lines delineate the 200- and 55-meter isobaths. Due to confidentiality reasons, the grid cell size is
569 large enough to cross over the 55-meter isobath, leading to the appearance of grid cells with non-
570 zero LIC in this area.

571

572 **References**

- 573 Abookire, A. A., & Bailey, K. M. (2007). The distribution of life cycle stages of two deep-water
574 pleuronectids, Dover sole (*Microstomus pacificus*) and rex sole (*Glyptocephalus zachirus*),
575 at the northern extent of their range in the Gulf of Alaska. *Journal of Sea Research*, 57(2),
576 198–208. <https://doi.org/10.1016/j.seares.2006.08.004>
- 577 Alverson, D. L., Pruter, A. T., & Ronholt, L. L. (1964). *A study of demersal fishes and fisheries*
578 *of the northeastern Pacific Ocean*. Institute of Fisheries, the University of British
579 Columbia.
- 580 Amante, C., & Eakins, B. W. (2009). *ETOPO1 1 arc-minute global relief model: Procedures,*
581 *data sources, and analysis*. NOAA Technical Memorandum. NESDIS NGDC-24. National
582 Geophysical Data Center, NOAA. doi:10.7289/V5C8276M
- 583 Bellman, M. A., & Heppell, S. A. (2007). *Trawl effort distribution off the U.S. Pacific Coast:*
584 *Regulatory shifts and seafloor habitat conservation* (Biology, Assessment, and
585 Management of North Pacific Rockfishes). Alaska Sea Grant Publication. AK-SG-07-01.
586 Alaska Sea Grant College Program, University of Alaska, Fairbanks.
- 587 Bellman, M. A., Heppell, S. A., & Goldfinger, C. (2005). Evaluation of a US west coast
588 groundfish habitat conservation regulation via analysis of spatial and temporal patterns of
589 trawl fishing effort. *Canadian Journal of Fisheries and Aquatic Sciences*, 62(12), 2886–
590 2900. <https://doi.org/10.1139/f05-180>
- 591 Bez, N., & Rivoirard, J. (2000). Indices of collocation between populations. In D. M. Checkley,
592 J. R. Hunter, L. Motos, & C. D. von der Lingen (Eds.), *Report of a Workshop on the use of*
593 *the Continuous Underway Fish Egg sampler (CUFES) for mapping spawning habitats of*
594 *Pelagic Fish*.
- 595 Bez, N., Walker, E., Gaertner, D., Rivoirard, J., & Gaspar, P. (2011). Fishing activity of tuna
596 purse seiners estimated from vessel monitoring system (VMS) data. *Canadian Journal of*
597 *Fisheries and Aquatic Sciences*, 68(11), 1998–2010. <https://doi.org/10.1139/f2011-114>
- 598 Booth, A. (2000). Incorporating the spatial component of fisheries data into stock assessment
599 models. *ICES Journal of Marine Science*, 57(4), 858–865.
600 <https://doi.org/10.1006/jmsc.2000.0816>
- 601 Carroll, G., Holsman, K. K., Brodie, S., Thorson, J. T., Hazen, E. L., Bograd, S. J., Haltuch, M.
602 A., Kotwicki, S., Samhoury, J., Spencer, P., Willis- Norton, E., & Selden, R. L. (2019). A
603 review of methods for quantifying spatial predator–prey overlap. *Global Ecology and*
604 *Biogeography*, 28(11), 1561–1577. <https://doi.org/10.1111/geb.12984>
- 605 Conn, P. B., Thorson, J. T., & Johnson, D. S. (2017). Confronting preferential sampling when
606 analysing population distributions: Diagnosis and model-based triage. *Methods in Ecology*
607 *and Evolution*, 8(11), 1535–1546. <https://doi.org/10.1111/2041-210X.12803>
- 608 Cope, J., Dick, E. J., MacCall, A., Monk, M., Soper, B., & Wetzel, C. (2015). *Data-moderate*
609 *stock assessments for brown, China, copper, sharpchin, stripetail, and yellowtail rockfishes*
610 *and English and rex soles in 2013*. Pacific Fishery Management Council. 298 pp.
- 611 Dark, T. A., & Wilkins, M. E. (1994). *Distribution, abundance, and biological characteristics of*
612 *groundfish off the coast of Washington, Oregon, and California, 1977-1986*. NOAA
613 Technical Report. NMFS-TR-117. 78 pp.
- 614 Diggle, P. J., Menezes, R., & Su, T. (2010). Geostatistical inference under preferential sampling.
615 *Journal of the Royal Statistical Society: Series C-Applied Statistics*, 59(2), 191–232.
616 <https://doi.org/10.1111/j.1467-9876.2009.00701.x>

617 Ducharme-Barth, N. D., Grüss, A., Vincent, M. T., Kiyofuji, H., Aoki, Y., Pilling, G., Hampton,
618 J., & Thorson, J. T. (2022). Impacts of fisheries-dependent spatial sampling patterns on
619 catch-per-unit-effort standardization: A simulation study and fishery application. *Fisheries*
620 *Research*, 246, 106169. <https://doi.org/10.1016/j.fishres.2021.106169>

621 Fox, D. S., & Starr, R. M. (1996). Comparison of commercial fishery and research catch data.
622 *Canadian Journal of Fisheries and Aquatic Sciences*, 53(12), 2681–2694.

623 Grüss, A., Walter, J. F., Babcock, E. A., Forrestal, F. C., Thorson, J. T., Lauretta, M. V., &
624 Schirripa, M. J. (2019). Evaluation of the impacts of different treatments of spatio-temporal
625 variation in catch-per-unit-effort standardization models. *Fisheries Research*, 213, 75–93.
626 <https://doi.org/10.1016/j.fishres.2019.01.008>

627 Gunderson, D. R., Armstrong, D. A., Shi, Y.-B., & McConnaughey, R. A. (1990). Patterns of
628 estuarine use by juvenile English sole (*Parophrys vetulus*) and Dungeness crab (*Cancer*
629 *magister*). *Estuaries*, 13(1), 59–71. <https://doi.org/10.2307/1351433>

630 Haltuch, M. A., & Hicks, A. (2009). *Status of the U.S. petrale sole resource in 2008*. Pacific
631 Fishery Management Council. 309 pp.

632 Haltuch, M. A., Hicks, A., & See, K. (2011). *Status of the U.S. petrale sole resource in 2010*.
633 Pacific Fishery Management Council. 389 pp.

634 Haltuch, M. A., Ono, K., & Valero, J. (2013). *Status of the U.S. petrale sole resource in 2012*.
635 Pacific Fishery Management Council. 480 pp.

636 Hanna, S. S. (2000). Setting the fishery management stage: Evolution of West Coast groundfish
637 management. *International Institute of Fisheries Economics and Trade*. History of the U.S.
638 West Coast Fishing Industry, Newport, OR.

639 Hannah, R. W. (2003). Spatial changes in trawl fishing effort in response to footrope diameter
640 restrictions in the U.S. West Coast bottom trawl fishery. *North American Journal of*
641 *Fisheries Management*, 23(3), 693–702. <https://doi.org/10.1577/M02-098>

642 He, X., Pearson, D. E., Field, J. C., Lefebvre, L., & Key, M. (2013). *Status of the U.S. Pacific*
643 *sanddab resource in 2013*. Pacific Fishery Management Council. 344 pp.

644 Hilborn, R., & Walters, C. J. (1992). Observing fish populations. In R. Hilborn & C. J. Walters
645 (Eds.), *Quantitative Fisheries Stock Assessment: Choice, Dynamics and Uncertainty*.
646 Springer US.

647 Howard, R.A., Ciannelli, L., Wakefield, W.W., & Fewings, M.R. (2021). The effects of climate,
648 oceanography, and habitat on the distribution and abundance of northern California Current
649 continental shelf groundfishes. *Fisheries Oceanography*, 30(6), 707–725.
650 <https://doi.org/10.1111/fog.12553>

651 Jacobson, L. D., & Hunter, J. R. (1993). Bathymetric demography and management of Dover
652 sole. *North American Journal of Fisheries Management*, 13(3), 405–420.
653 [https://doi.org/10.1577/1548-8675\(1993\)013<0405:BDAMOD>2.3.CO;2](https://doi.org/10.1577/1548-8675(1993)013<0405:BDAMOD>2.3.CO;2)

654 Keller, A. A., Wallace, J. R., & Methot, R. D. (2017). *The Northwest Fisheries Science Center's*
655 *West Coast Groundfish Bottom Trawl Survey: History, design, and description*. NOAA
656 Technical Memorandum. NMFS-NWFSC-136. 37 pp.

657 Ketchen, K. S., & Forrester, C. R. (1966). *Population dynamics of the petrale sole, Eopsetta*
658 *jordani, in waters off western Canada*. Bulletin, 153. Fisheries Research Board of Canada.
659 195 pp.

660 Kotwicki, S., & Lauth, R. R. (2013). Detecting temporal trends and environmentally-driven
661 changes in the spatial distribution of bottom fishes and crabs on the eastern Bering Sea

662 shelf. *Deep Sea Research Part II: Topical Studies in Oceanography*, 94, 231–243.
663 <https://doi.org/10.1016/j.dsr2.2013.03.017>

664 Krygier, E. E., & Percy, W. G. (1986). The role of estuarine and offshore nursery areas for
665 young English sole, *Parophrys vetulus girard*, of Oregon. *Fishery Bulletin*, 84(1), 119–132.

666 Lyons, K., Jarvis, E. T., Jorgensen, S. J., Weng, K., O’Sullivan, J., Winkler, C., & Lowe, C. G.
667 (2013). The degree and result of gillnet fishery interactions with juvenile white sharks in
668 southern California assessed by fishery-independent and -dependent methods. *Fisheries*
669 *Research*, 147, 370–380. <https://doi.org/10.1016/j.fishres.2013.07.009>

670 Macomber, M. F. (2000). *Selecting locations for marine harvest refugia: A GIS study using*
671 *logbook data from the Oregon trawl fishery* [M.S. Thesis]. Oregon State University.

672 Maunder, M. N., & Punt, A. E. (2004). Standardizing catch and effort data: A review of recent
673 approaches. *Fisheries Research*, 70(2), 141–159.
674 <https://doi.org/10.1016/j.fishres.2004.08.002>

675 Murray, K., & Orphanides, C. (2013). Estimating the risk of loggerhead turtle *Caretta caretta*
676 bycatch in the US mid-Atlantic using fishery-independent and -dependent data. *Marine*
677 *Ecology Progress Series*, 477, 259–270. <https://doi.org/10.3354/meps10173>

678 Pearson, D. E., & McNally, S. V. G. (2005). Age, growth, life history, and fisheries of the sand
679 sole, *Psettichthys melanostictus*. *Marine Fisheries Review*, 67(4), 9–18.

680 Pecquerie, L., Drapeau, L., Fréon, P., Coetzee, J. C., Leslie, R. W., & Griffiths, M. H. (2004).
681 Distribution patterns of key fish species of the southern Benguela ecosystem: An approach
682 combining fishery-dependent and fishery-independent data. *African Journal of Marine*
683 *Science*, 26, 115–139.

684 Pedersen, M. G. (1975). Movements and growth of petrale sole (*Eopsetta jordani*) tagged off
685 Washington and Southwest Vancouver Island. *Journal of the Fisheries Research Board of*
686 *Canada*, 32(11), 2169–2177. <https://doi.org/10.1139/f75-255>

687 Pennino, M. G., Conesa, D., López-Quílez, A., Muñoz, F., Fernández, A., & Bellido, J. M.
688 (2016). Fishery-dependent and -independent data lead to consistent estimations of essential
689 habitats. *ICES Journal of Marine Science*, 73(9), 2302–2310.
690 <https://doi.org/10.1093/icesjms/fsw062>

691 Pennino, M. G., Paradinas, I., Illian, J. B., Muñoz, F., Bellido, J. M., López-Quílez, A., &
692 Conesa, D. (2019). Accounting for preferential sampling in species distribution models.
693 *Ecology and Evolution*, 9(1), 653–663. <https://doi.org/10.1002/ece3.4789>

694 Petigas, P., Woillez, M., Rivoirard, J., Renard, D., & Bez, N. (2017). *Handbook of geostatistics*
695 *in R for fisheries and marine ecology*. International Council for the Exploration of the Sea
696 Cooperative Research Report, 338. 177 pp.

697 Petrik, C. M., Duffy-Anderson, J. T., Mueter, F., Hedstrom, K., & Curchitser, E. N. (2015).
698 Biophysical transport model suggests climate variability determines distribution of walleye
699 pollock early life stages in the eastern Bering Sea through effects on spawning. *Progress in*
700 *Oceanography*, 138, 459–474. <https://doi.org/10.1016/j.pocean.2014.06.004>

701 PFMC. (2018). *Status of the Pacific Coast groundfish fishery*. Pacific Fishery Management
702 Council. 323 pp.

703 Pianka, E. R. (1973). The Structure of Lizard Communities. *Annual Review of Ecology and*
704 *Systematics*, 4, 53–74. JSTOR.

705 Pinto, C., Travers-Trolet, M., Macdonald, J. I., Rivot, E., & Vermard, Y. (2019). Combining
706 multiple data sets to unravel the spatiotemporal dynamics of a data-limited fish stock.

707 *Canadian Journal of Fisheries and Aquatic Sciences*, 76(8), 1338–1349.
708 <https://doi.org/10.1139/cjfas-2018-0149>

709 Pointin, F., Daurès, F., & Rochet, M.-J. (2019). Use of avoidance behaviours to reduce the
710 economic impacts of the EU Landing Obligation: The case study of a mixed trawl fishery.
711 *ICES Journal of Marine Science*, 76(6), 1554–1566. <https://doi.org/10.1093/icesjms/fsz032>

712 R Core Team. (2021). *R: A language and environment for statistical computing* (v. 4.0.4)
713 [Computer software]. R Foundation for Statistical Computing. <https://www.R-project.org>

714 Ralston, S. (2005). *An assessment of starry flounder off California, Oregon, and Washington*.
715 Pacific Fishery Management Council. 86 pp.

716 The Research Group, LLC. (2021). *Oregon Commercial and Recreational Fishing Industry*
717 *Economic Activity Coastwide and in Proximity to Marine Reserve Sites for Years 2018 and*
718 *2019*. Prepared for Oregon Department of Fish and Wildlife, Marine Reserve Program and
719 Marine Resource Program. 76 pp.

720 Richmond, N. T. (1983). *The sand sole*. Informational Report 83-1. Oregon Department of Fish
721 and Wildlife. 4 pp.

722 Rufener, M.-C., Kristensen, K., Nielsen, J. R., & Bastardie, F. (2021). Bridging the gap between
723 commercial fisheries and survey data to model the spatiotemporal dynamics of marine
724 species. *Ecological Applications*, 31(8), e02453. <https://doi.org/10.1002/eap.2453>

725 Ryer, C. H. (2008). A review of flatfish behavior relative to trawls. *Fisheries Research*, 90(1),
726 138–146. <https://doi.org/10.1016/j.fishres.2007.10.005>

727 Sampson, D. B. (2011). The accuracy of self-reported fisheries data: Oregon trawl logbook
728 fishing locations and retained catches. *Fisheries Research*, 112(1–2), 59–76.
729 <https://doi.org/10.1016/j.fishres.2011.08.012>

730 SaMTAAC. (2019). *National Marine Fisheries Service (NMFS) report on the purpose and need*
731 *statement*. SaMTAAC Agenda Item D.2. NMFS Report. Sablefish Management and Trawl
732 Allocation Attainment Committee, Pacific Fishery Management Council. 2 pp.

733 Sant’Ana, R., Gerhard Kinas, P., Villwock de Miranda, L., Schwingel, P. R., Castello, J. P., &
734 Paes Vieira, J. (2017). Bayesian state-space models with multiple CPUE data: The case of a
735 mullet fishery. *Scientia Marina*, 81(3), 361. <https://doi.org/10.3989/scimar.04461.11A>

736 Sjostrom, A. J. C., Ciannelli, L., Conway, F., & Wakefield, W. W. (2021). Gathering local
737 ecological knowledge to augment scientific and management understanding of a living
738 coastal resource: The case of Oregon’s nearshore groundfish trawl fishery. *Marine Policy*,
739 131, 104617. <https://doi.org/10.1016/j.marpol.2021.104617>

740 Sjostrom, A. J. C. (2019). *Informing and enhancing scientific and management understanding of*
741 *Oregon’s nearshore groundfish trawl fishery by engaging local ecological knowledge*
742 [M.S. Thesis]. Oregon State University.

743 Stallings, C. D. (2009). Fishery-independent data reveal negative effect of human population
744 density on Caribbean predatory fish communities. *PLOS ONE*, 4(5), e5333.
745 <https://doi.org/10.1371/journal.pone.0005333>

746 Stramma, L., Schmidtko, S., Levin, L. A., & Johnson, G. C. (2010). Ocean oxygen minima
747 expansions and their biological impacts. *Deep Sea Research Part I: Oceanographic*
748 *Research Papers*, 57(4), 587–595. <https://doi.org/10.1016/j.dsr.2010.01.005>

749 Tolimieri, N., Wallace, J., & Haltuch, M. A. (2020). Spatio-temporal patterns in juvenile habitat
750 for 13 groundfishes in the California Current Ecosystem. *PLOS ONE*, 15(8).
751 <https://doi.org/10.1371/journal.pone.0237996>

752 Toole, C., Brodeur, R., Donohoe, C., & Markle, D. (2011). Seasonal and interannual variability
753 in the community structure of small demersal fishes off the central Oregon coast. *Marine*
754 *Ecology Progress Series*, 428, 201–217. <https://doi.org/10.3354/meps09028>

755 Trenkel, V., Pinnegar, J., Dawson, W., du Buit, M., & Tidd, A. (2005). Spatial and temporal
756 structure of predator-prey relationships in the Celtic Sea fish community. *Marine Ecology*
757 *Progress Series*, 299, 257–268. <https://doi.org/10.3354/meps299257>

758 Warlick, A., Steiner, E., & Guldin, M. (2018). History of the west coast groundfish trawl fishery:
759 Tracking socioeconomic characteristics across different management policies in a
760 multispecies fishery. *Marine Policy*, 93, 9–21.
761 <https://doi.org/10.1016/j.marpol.2018.03.014>

762 Wetzel, C. R. (2019). *Status of petrale sole (Eopsetta jordani) along the U.S. West Coast in*
763 *2019*. Pacific Fishery Management Council. 256 pp.

764 Wickham, H. (2016). *ggplot2: Elegant graphics for data analysis* [R]. Springer-Verlag.
765 <https://ggplot2.tidyverse.org>

766 Zhu, M., Yamakawa, T., & Sakai, T. (2018). Combined use of trawl fishery and research vessel
767 survey data in a multivariate autoregressive state-space (MARSS) model to improve the
768 accuracy of abundance index estimates. *Fisheries Science*, 84(3), 437–451.
769 <https://doi.org/10.1007/s12562-018-1190-9>

770
771

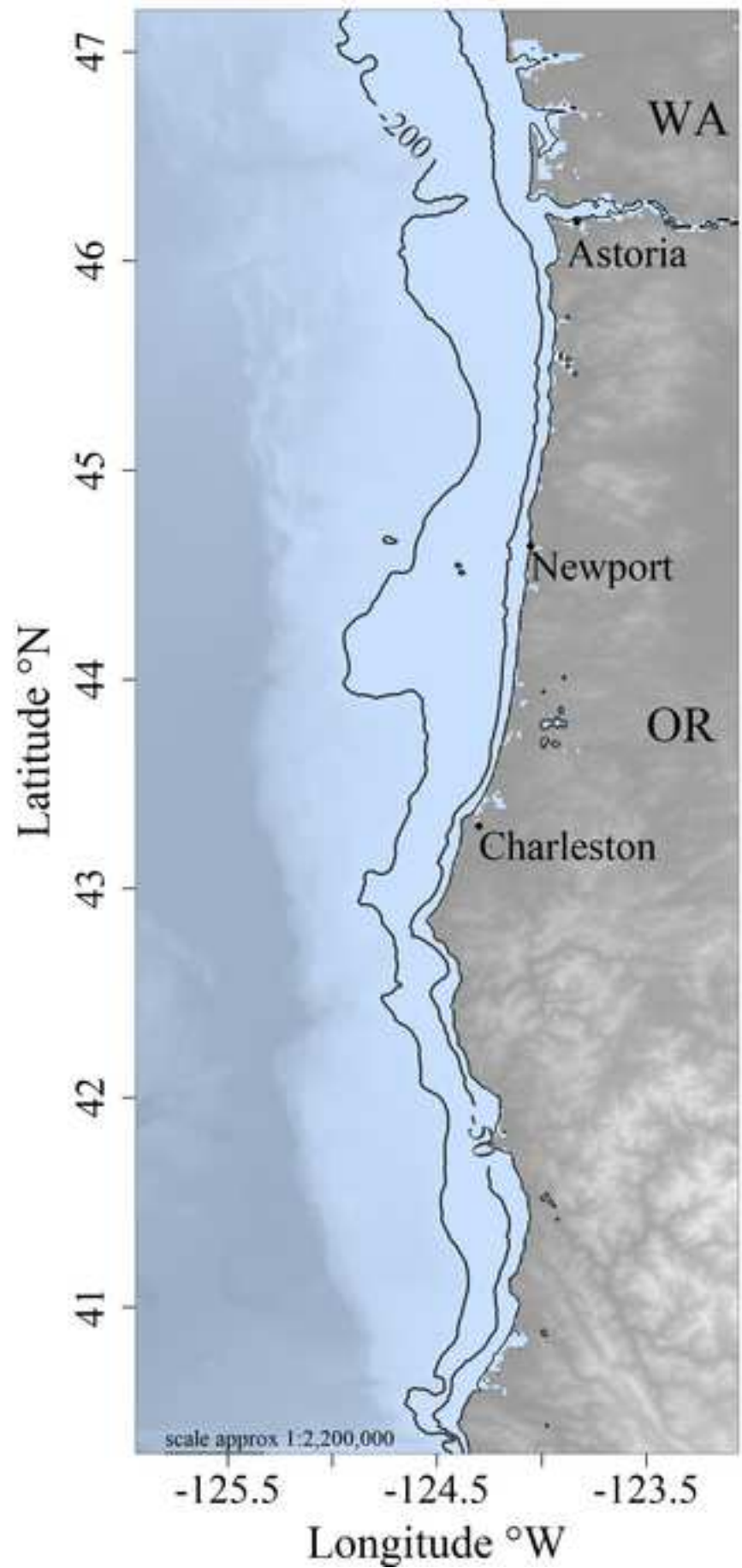


Figure2.jpeg

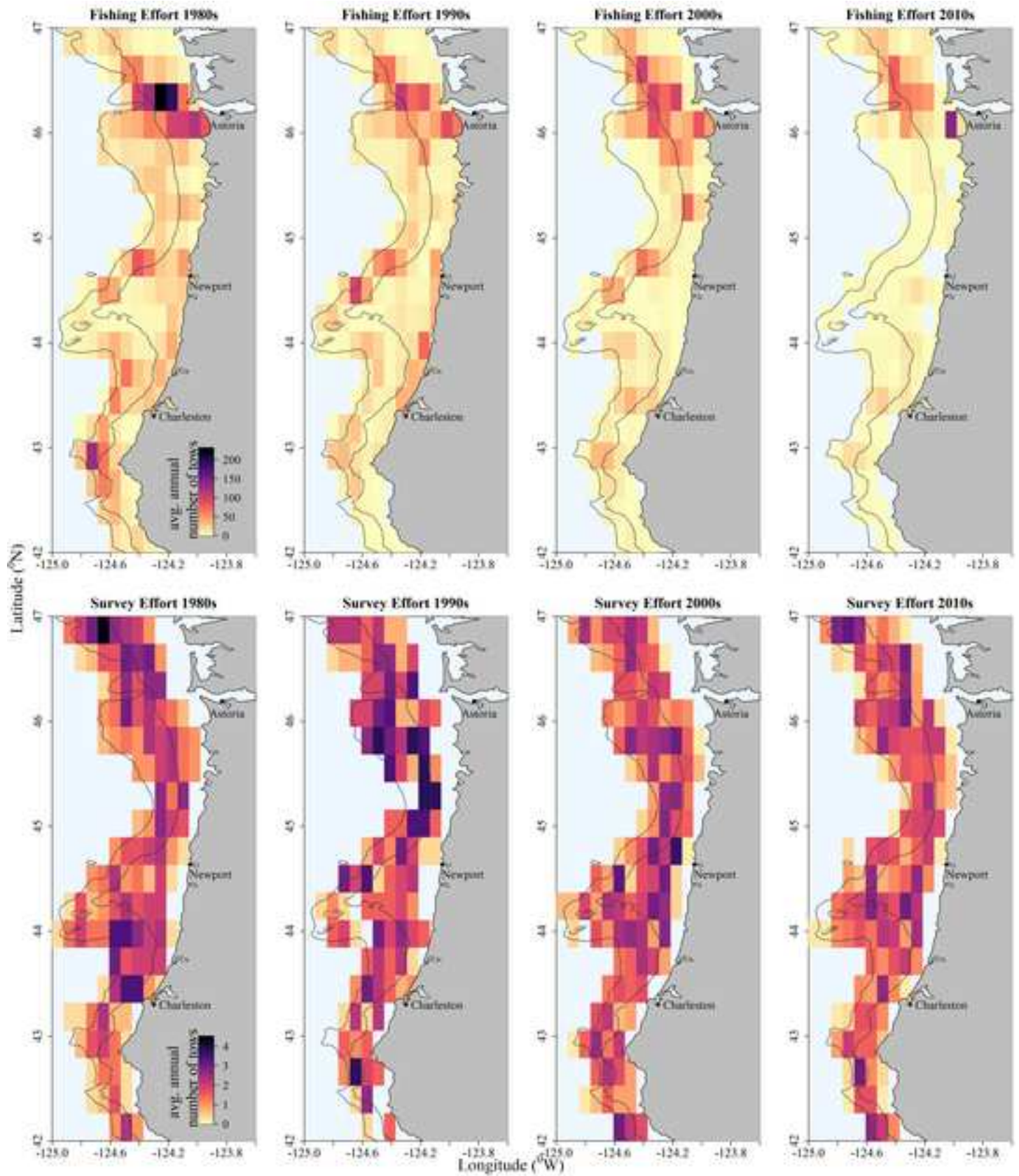


Figure3.jpeg

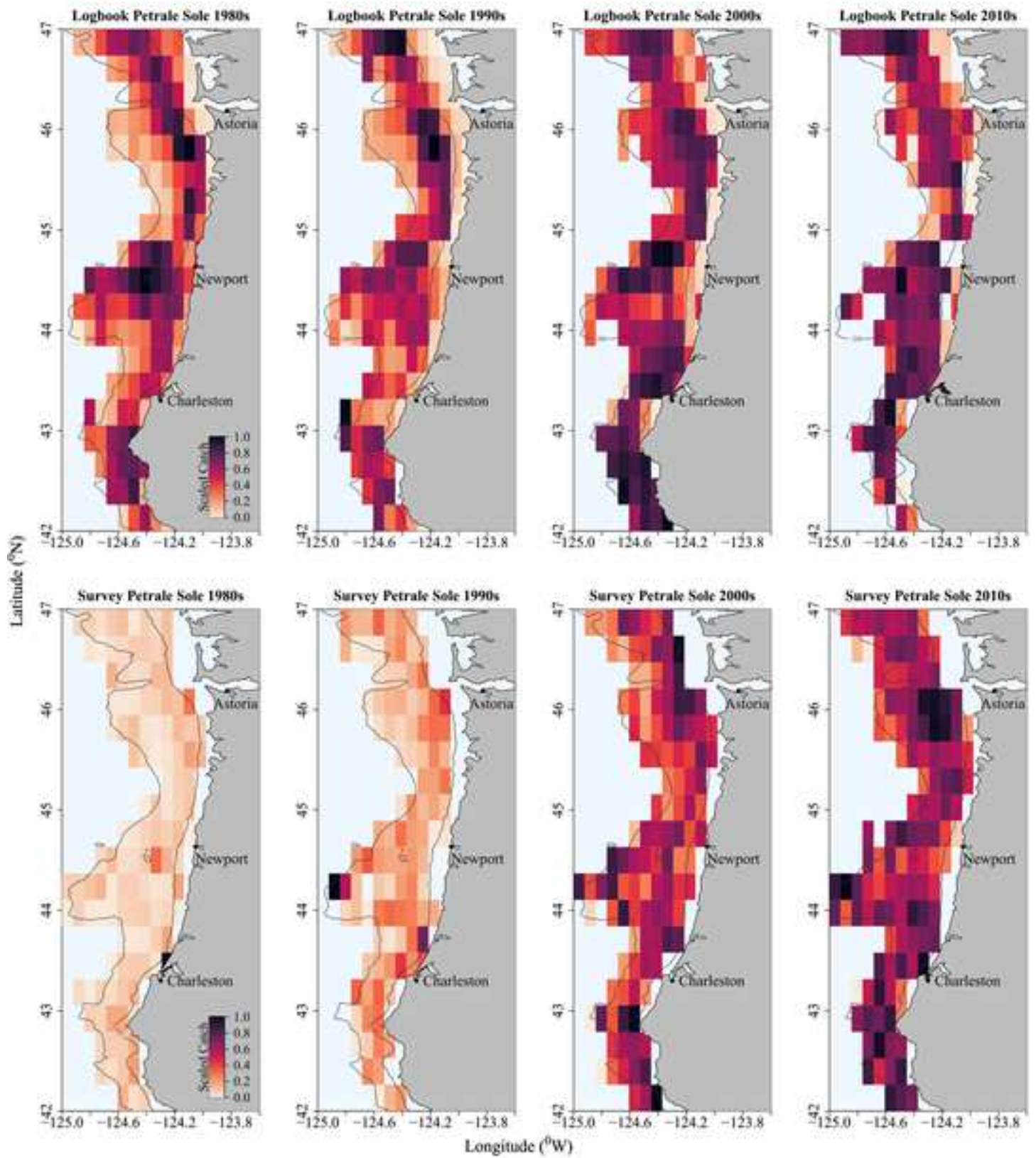


Figure4.jpeg

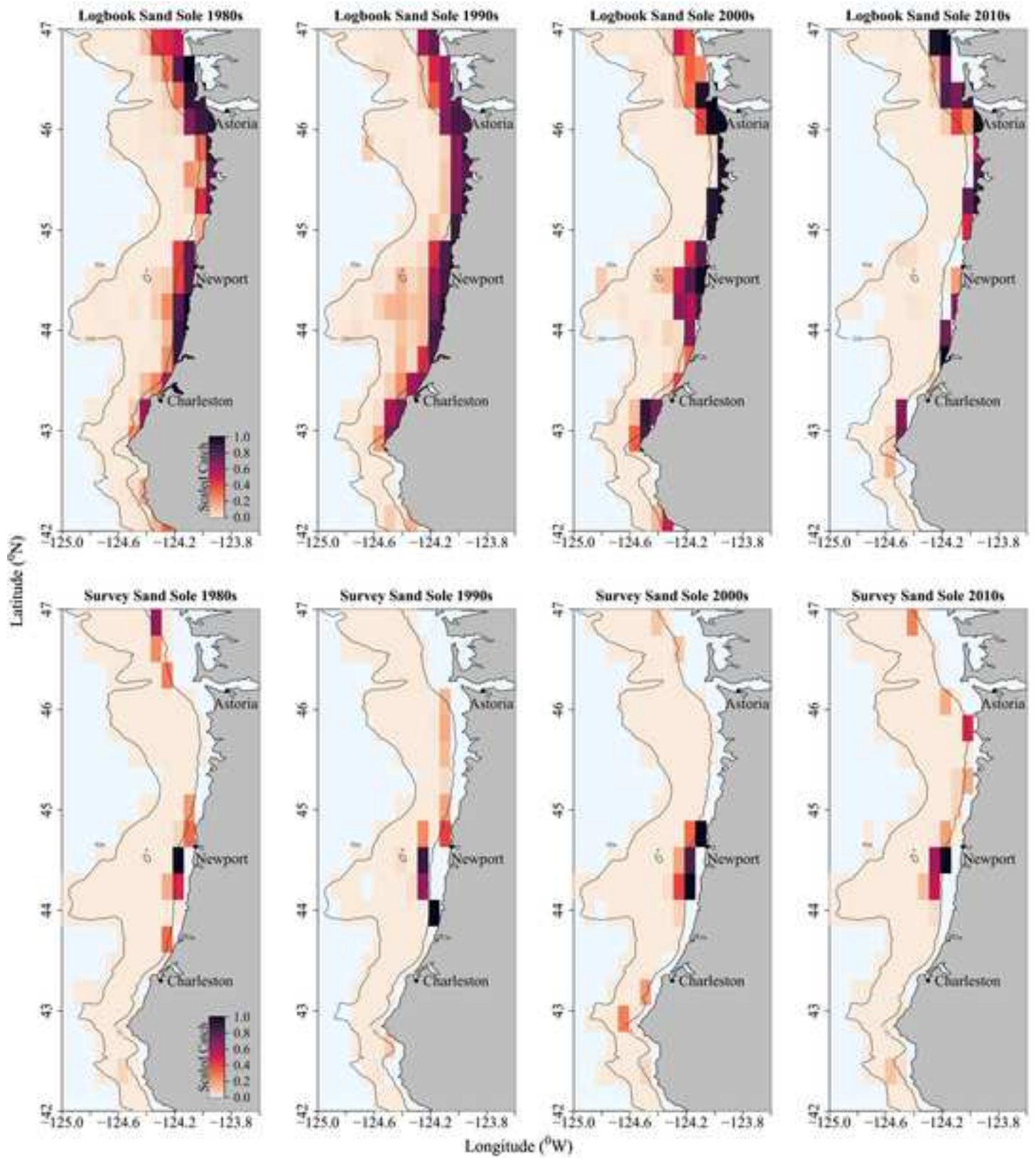
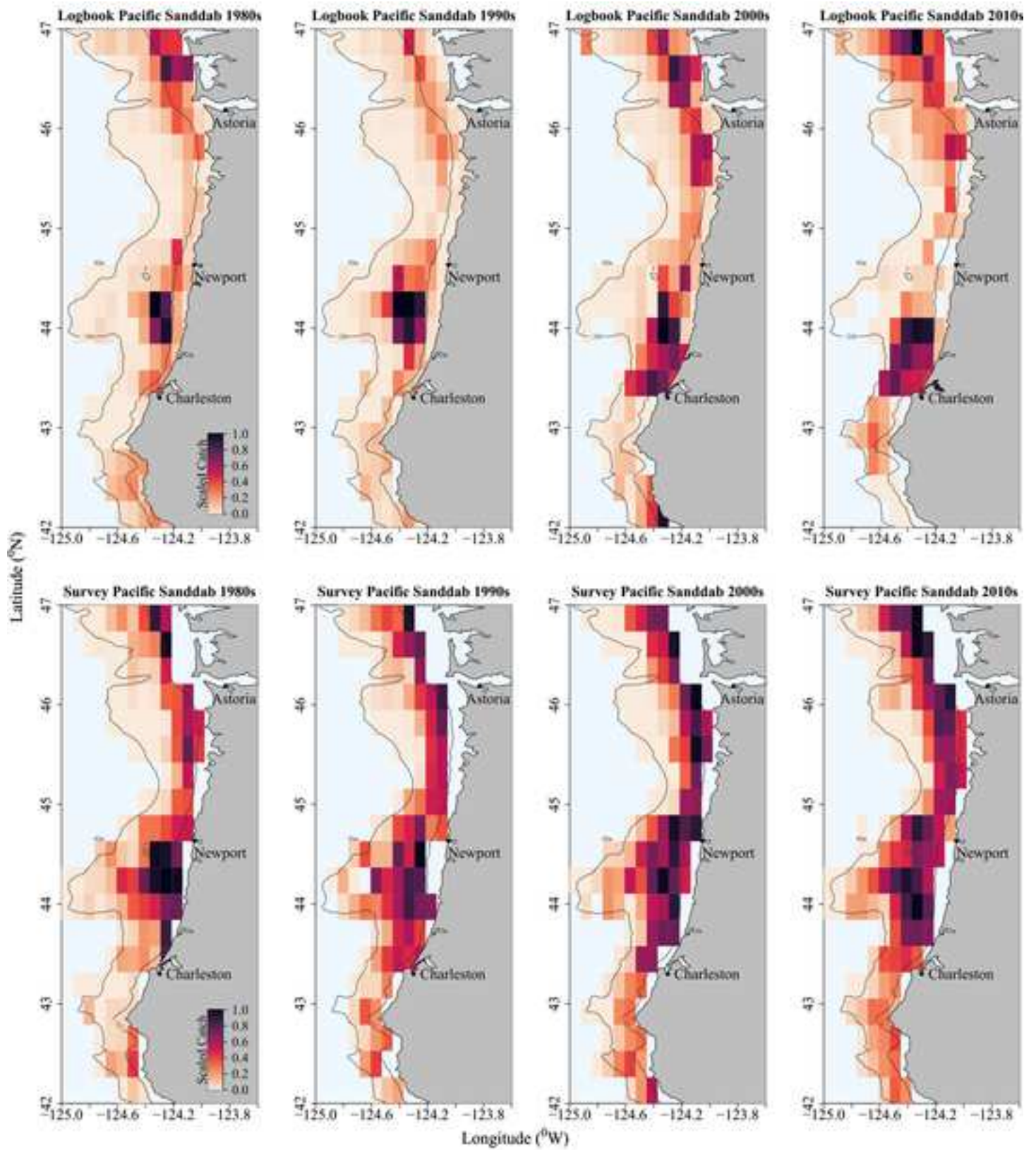


Figure5.jpeg



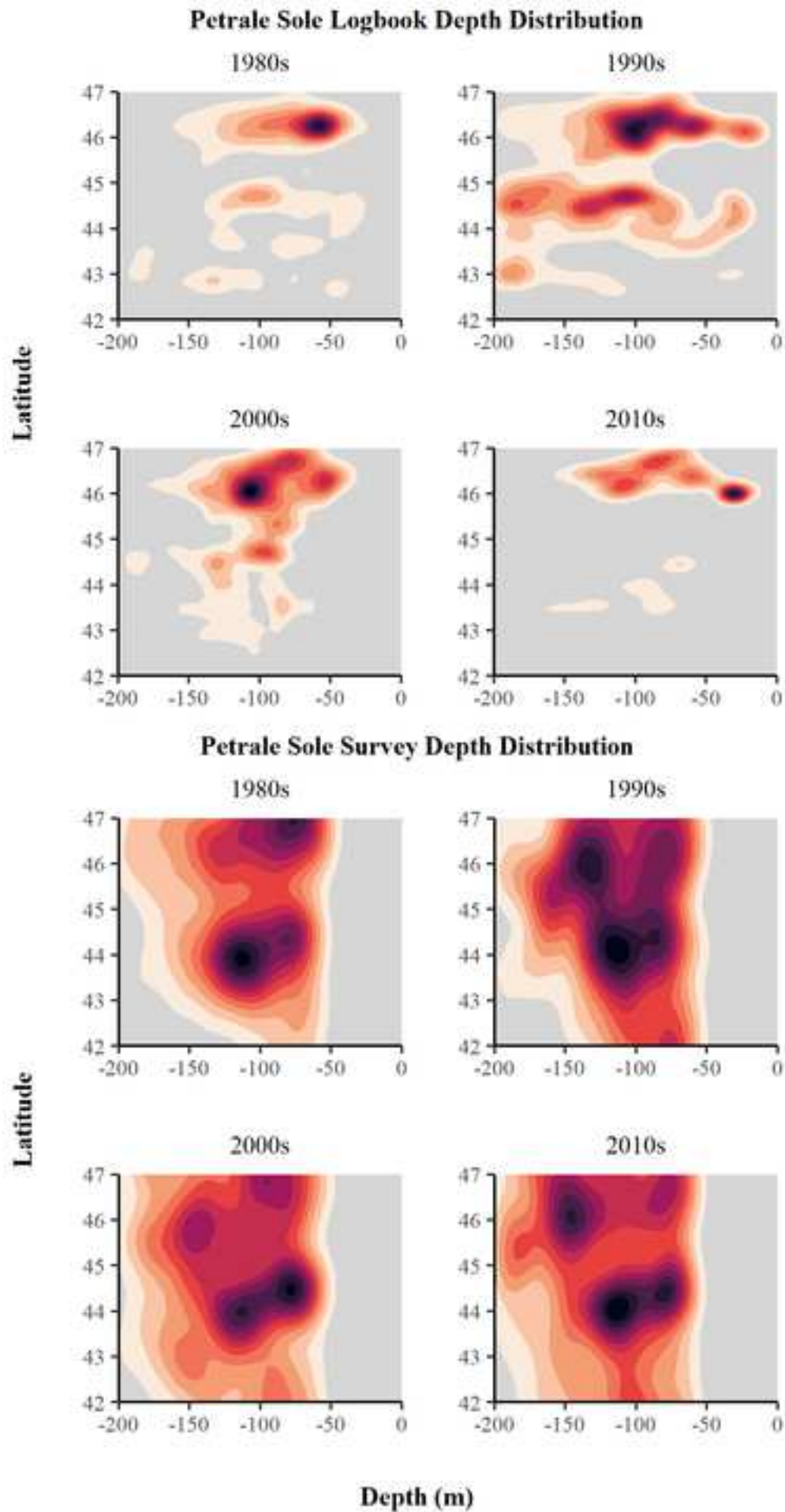


Figure7.jpeg

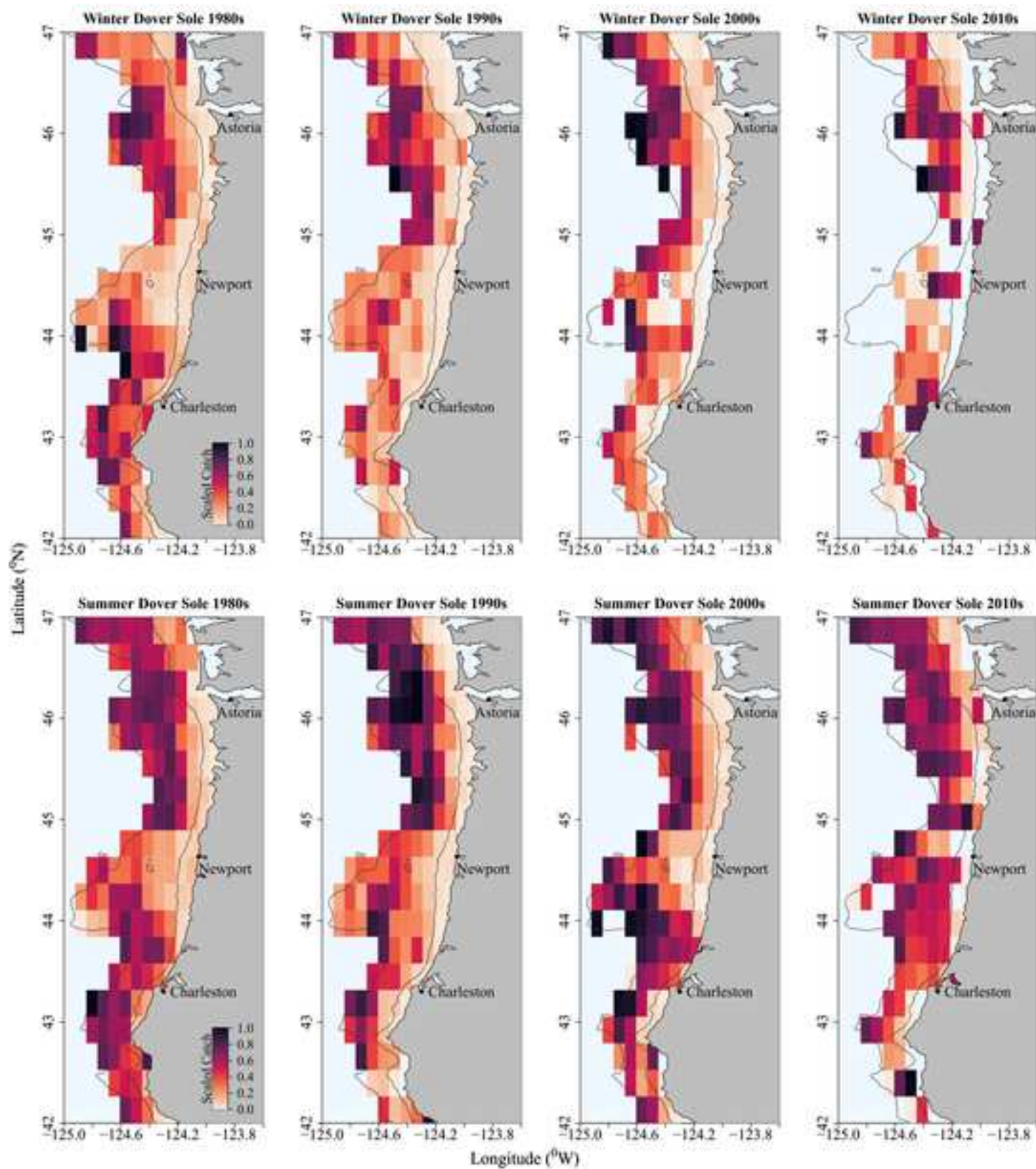


Figure8.jpeg

