

1 **Persistent organic pollutants in female humpback whales *Megaptera novaeangliae* from the**  
2 **Gulf of Maine**

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14  
15 **Abstract**

16 Contaminant studies in cetaceans typically focus on males due to confounding effects of  
17 reproductive status in females and maternal offloading. However, an improved understanding of  
18 contaminant burdens in female cetaceans is needed to better assess potential impacts to  
19 populations. In this study, 36 blubber biopsy samples of female humpback whales (*Megaptera*  
20 *novaeangliae*) from the Gulf of Maine were analyzed to examine contaminant loads across  
21 females of different ages. Sampled individuals were individually-identified from longitudinal  
22 studies and assigned to age class (i.e., adult, subadult, juvenile, calf). Analysis was performed  
23 using gas chromatography/mass spectrometry for persistent organic pollutants (POPs) including  
24 polychlorinated biphenyls (PCBs), dichlorodiphenyltrichloroethanes (DDTs), chlordanes  
25 (CHLDs), polybrominated diphenyl ethers (PBDEs), hexachlorocyclohexanes (HCHs), and other  
26 organochlorine pesticides (OCPs). The most abundant POPs measured were PCB congeners,  
27 with summed values ranging from 380 to 12,300 ng/g, lipid weight, well below the threshold  
28 value for adverse health effects. We found significant differences in mean values between adults  
29 and juveniles and between adults and subadults, with the exception of the less persistent HCHs  
30 for the latter. We also found significant differences in mean levels of  $\Sigma$ HCHs and  $\Sigma$ other OCPs  
31 between the juveniles and subadults. Changes over age are consistent with maternal offloading  
32 and potentially important for evaluating population health and viability.

33

34 **Keywords**

35 **Humpback whale; POPs; Gulf of Maine; PCBs; DDTs; maternal offloading**

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37 **1. Introduction**

38 Persistent organic pollutants (POPs) are anthropogenic toxic chemicals that persist in the  
39 environment, are resistant to metabolism and degradation, and can be circulated globally via  
40 atmospheric transport and ocean currents. Among the different classes of POPs are  
41 polychlorinated biphenyls (PCBs), various organochlorine pesticides (OCPs), including  
42 dichlorodiphenyltrichloroethanes (DDTs), chlordanes (CHLDs), hexachlorocyclohexanes  
43 (HCHs), and the flame retardants, polybrominated diphenyl ethers (PBDEs).

44 Many POPs continue to be measured in environmental samples despite having been  
45 banned for production or open use in the US and many other countries since the 1970s (e.g.  
46 Stockholm Convention, (EPA, 2017; Vijgen et al., 2011). Exposure to POPs has been associated  
47 with adverse health effects such as immune dysfunction, increased susceptibility to disease,  
48 reproductive and endocrine impairment, and neurotoxicity. Such effects have been confirmed in  
49 marine mammal species, including: harbor porpoises (*Phocoena phocoena*), harbor seals (*Phoca  
50 vitulina*), California sea lions (*Zalophus californianus*), beluga whales (*Delphinapterus leucas*),  
51 grey seals (*Halichoerus grypus*) and ringed seals (*Pusa hispida*) (Bergman and Olsson, 1985;  
52 Beland et al., 1993; Hammond et al., 2005; Ylitalo et al., 2005a; Murphy et al., 2015). The  
53 primary route of POP exposure in marine mammals is through diet and these lipophilic  
54 compounds can bioaccumulate to relatively high concentrations in their blubber. As a result,  
55 there is a cause for concern, especially for long-lived marine mammals such as cetaceans.

56 Mysticete cetaceans feed at a lower trophic level than odontocetes and are therefore  
57 assumed to be at a lower risk of adverse health effects from POPs, even when residing in the  
58 same habitats (Aguilar et al., 1999; Pinzone et al., 2015). However, exposure to lower  
59 concentrations may nevertheless be significant in light of the fact that mysticetes are long lived,  
60 have large lipid stores, and can offload POPs from mother to calf during gestation and lactation  
61 (maternal offloading; Aguilar et al., 1999; Rowe 2008). It is important to also assess the  
62 exposure to and impacts of environmental contaminants when evaluating the health of mysticete  
63 populations. However, due to the confounding effects of accumulation through diet and

64 maternal offloading, marine mammal studies commonly focus on measuring contaminant  
65 concentrations in immature animals or adult males and therefore POPs data is more limited for  
66 adult female marine mammals.

67 The Gulf of Maine, off the east coast of North America, is the site of long-term  
68 industrialized activity as well as long-term humpback whale population studies and prior  
69 humpback whale toxicological studies (Elfes *et al.* 2010, Ryan *et al.* 2013). The objectives of  
70 the current study were to extend that work by characterizing for the first time POP  
71 concentrations in blubber of female humpback whales across age classes as well as to better  
72 characterize maternal offloading of these compounds to their offspring.

73

## 74 **2. Materials and Methods**

### 75 **Sample collection**

76 Blubber samples were collected from 36 free-ranging female humpback whales in the  
77 Gulf of Maine (western North Atlantic Ocean) between 2004 and 2012 (Figure 1). Samples were  
78 collected from the lateral flank by biopsy sampling techniques (Palsbøll *et al.* 1991). Samples  
79 were kept on ice in the field and then frozen at -80 degrees Celsius until analysis. A long-term  
80 photo-identification catalog of individual Gulf of Maine humpback whales (Center for Coastal  
81 Studies, Provincetown, MA) was used to determine age class of individuals at the time of  
82 sampling. Calves were dependent offspring within the first year of birth, and juveniles when  
83 independent but no more than four years of age. Subadult females had reached the minimum  
84 documented age at first calving in this population (age 5, Clapham 1992; Robbins 2007) but had  
85 not yet observed with a calf themselves. Females were categorized as adults when known to  
86 have given birth to at least one calf. For three individuals, samples were available from two  
87 different ages.

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### 89 **Chemical analyses**

90 POPs were extracted from samples and analyzed for POPs using gas  
91 chromatography/mass spectrometry (GC/MS) as described in Sloan *et al.*, 2014. Prior to  
92 extraction of POPs with dichloromethane using an automated pressurized solvent extractor,  
93 blubber samples (0.1–0.3 g) were mixed with drying agents (sodium sulfate and magnesium  
94 sulfate) and spiked with a surrogate standard (PCB 103; 75 ng). A single stacked gravity flow

95 silica gel/alumina column was used to remove any highly polar compounds from the sample.  
96 POPs were separated from the bulk lipid and other biogenic material present in each sample  
97 using high performance size exclusion liquid chromatography. After separation of analytes  
98 using a 60 m Agilent DB-5 GC capillary column, they were determined on an Agilent 5973 MS  
99 that was operated in selected ion monitoring and electron impact mode. Specifics of GC  
100 operating conditions and monitored ions can be found in Sloan et al., 2014. The instrument was  
101 calibrated using ten levels of standards of known concentrations. All blubber contaminant  
102 concentrations were reported in ng/g, lipid weight. The concentrations of PCB congeners 17, 18,  
103 28, 31, 33, 44, 49, 52, 66, 70, 74, 82, 87, 95, 99, 101/90, 105, 110, 118, 128, 138/163/164, 149,  
104 151, 153/132, 156, 158, 170, 171, 177, 180, 183, 187/159/182, 191, 194, 195, 199, 205, 206,  
105 208, and 209 were used to calculate summed PCBs ( $\Sigma$ PCBs). The concentrations of *o,p'*-DDD,  
106 *p,p'*-DDD, *o,p'*-DDE, *p,p'*-DDE, *o,p'*-DDT, and *p,p'*-DDT were used to calculate summed DDTs  
107 ( $\Sigma$ DDTs). The concentrations of heptachlor, oxychlordane,  $\beta$ -chlordane, nona-III-chlordane,  $\alpha$ -  
108 chlordane, *trans*-nonachlor, and *cis*-nonachlor were used to calculate summed chlordanes  
109 ( $\Sigma$ CHLDs). The concentrations of PBDE congeners 28, 47, 49, 66, 85, 99, 100, 153, 154, 155,  
110 and 183 were used to calculate summed PBDEs ( $\Sigma$ PBDEs). The concentrations of  $\alpha$ -  
111 hexachlorocyclohexane,  $\beta$ -hexachlorocyclohexane, and  $\gamma$ -hexachlorocyclohexane (lindane) were  
112 used to calculate summed hexachlorocyclohexanes ( $\Sigma$ HCHs). The concentrations of  
113 hexachlorobenzene (HCB), aldrin, mirex, and endosulfan I were used to calculate summed other  
114 OC pesticides ( $\Sigma$ OCPs).

115

## 116 **Lipid content**

117 A 1.5 mL extract subsample was aliquoted for determination of percent lipid gravimetric, as  
118 well as to measure lipid classes using thin-layer chromatography with flame ionization detection  
119 (TLC/FID) (Ylitalo et al., 2005b; Sloan et al. 2014). Lipophilic POP concentrations were  
120 normalized using percent lipid of each blubber sample. Lipid class profiles (i.e., sterol esters/wax  
121 esters, triglycerides, free fatty acids, cholesterol, phospholipids/polar lipids) were evaluated  
122 because the accumulation of lipophilic POPs may be influenced by the proportion of neutral  
123 lipids (e.g., triglycerides) in the blubber (Krahn et al., 2004).

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## 125 **Quality assurance**

126 A solvent (dichloromethane) method blank and a National Institute of Standards and  
127 Technology (NIST) whale blubber Standard Reference Material (SRM 1945) were analyzed with  
128 each humpback whale blubber sample set. No more than five analytes exceeded 2 times the  
129 lower limit of quantitation (LOQ) for each method blank. For each sample set (10-12 blubber  
130 samples), concentrations of  $\geq 70\%$  of individual analytes that were measured in the NIST SRM  
131 1945 were within 30% of the upper and lower ends of the 95% confidence interval range of the  
132 published NIST certified concentrations. The LOQs ranged from  $< 0.37$  to  $< 10$  ng/g wet weight  
133 for the PCBs,  $< 0.37$  to  $< 9.9$  ng/g wet weight for the organochlorine pesticides (DDTs, CHLDs,  
134 HCB, aldrin, mirex, and endosulfan I), and  $< 0.36$  to  $< 2.6$  ng/g wet weight for the PBDEs. The  
135 percent recoveries of the surrogate standard in the field and associated quality assurance samples  
136 ranged from 89% to 109%. Other quality control elements met established laboratory criteria  
137 (Sloan et al., 2019).

138

### 139 **Statistical analyses**

140 Prior to statistical analyses, percent lipid results were arcsine square transformed and  
141 summed concentrations of POPs were log transformed to obtain a more normal distribution and  
142 equal variances. Analysis of variance (ANOVA) and the Tukey-Kramer honestly significance  
143 difference test (HSD) were used to determine if the mean concentrations of the various classes of  
144 POPs were significantly different among the four age classes with a level of significance at  $P \leq$   
145 0.05. Statistical analyses were completed using Statistical analyses were completed using R  
146 (version R-4.0.4).

### 147 **3. Results and Discussion**

148 Concentrations of summed POPs based on whale age class are reported in Table 1. All  
149 POP classes were detected in every individual whale, with concentrations ranging from 6.2  
150 ( $\sum$ HCHs) to 12,300 ng/g, lipid weight ( $\sum$ PCBs). The rank order of POP classes determined in  
151 the whale blubber were PCBs  $>$  DDTs  $>$  PBDEs  $\cong$  CHLDs  $>$  other OCPs  $>$  HCHs. The mean  
152 blubber POP levels (ng/g lipid weight) for the four age classes of the female humpback whales  
153 ranged from 1,700 to 5,900 for  $\sum$ PCBs, 400 to 2,100 for  $\sum$ DDTs, 210 to 980 for  $\sum$ CHLDs, 190  
154 to 980 for  $\sum$ PBDEs, 34 to 170 for  $\sum$ other OCPs, and 20 to 64 for  $\sum$ HCHs (Table 1). The  
155 predominant analytes contributing to each of the corresponding summed values for each age  
156 class were as follows: PCB 153 (16 to 19%) and PCB 138 (13 to 15%) to  $\sum$ PCBs, *p,p'*-DDE (67

157 to 72%) to  $\sum$ DDTs, *trans*-nonachlor (58 to 60%) to  $\sum$ CHLDs, PBDE 47 (66 to 75%) to  
158  $\sum$ PBDEs,  $\alpha$ -hexachlorocyclohexane (46 to 61%) and  $\beta$ -hexachlorocyclohexane (29 to 42%) to  
159  $\sum$ HCHs, and HCB (88 to 97%) to  $\sum$ other OCPs. Mean blubber levels reported previously in  
160 adult males from this humpback whale population (Elfes et al., 2010) were up to 2 to 13 times  
161 higher than the mean values of females in the current study.

162 The lipid content of the blubber samples decreased with animal age class, being highest  
163 in calves (46.4%) and lowest in adults (32.6%, Table 1). The lipids measured in the biopsy  
164 blubber samples consisted primarily of neutral triglycerides (85 to 100% of total lipid). In  
165 addition to triglycerides, blubber samples from a juvenile and nursing adult contained 7.4% and  
166 14.6% phospholipids respectively. The near homogenous lipid class profiles of the female  
167 humpback whale blubber samples across all age classes indicate that POP concentrations were  
168 not biased due to lipid composition.

169

#### 170 **Association between POPs and lipids with age class**

171 Mean summed POP values differed significantly ( $p < 0.05$ ) among the four age classes of  
172 female whales, with juveniles having significantly higher levels than the mean values in adults  
173 (for all 5 POP classes and other OCPs) and subadults (for HCHs and other OCPs). In addition,  
174 subadult females had significantly higher ( $p < 0.04$ ) mean summed POP values than adult  
175 females for all POP classes except HCHs. No other significant differences ( $p > 0.05$ ) in mean  
176 POP values were found among the age classes. The changes in the concentrations of POPs from  
177 one age class to the next are likely due to a combination of factors. Additive factors can be  
178 attributed to accumulation via diet and maternal transfer of contaminants (Aguilar et al. 1999).  
179 Previous cetacean studies have shown that lipophilic compounds are transferred from mother to  
180 calf primarily during lactation and, to a lesser extent during gestation (Cockcroft et al., 1989,  
181 Krahn et al., 2009). Birth order can also influence concentrations of lipophilic contaminants  
182 such as PCBs and DDTs, with first born animals having higher levels than non-first born  
183 (Cockcroft et al., 1989; Ylitalo et al., 2001; Wells et al., 2005). In addition to maternal  
184 offloading, other factors that can decrease POP concentrations include biotransformation as well  
185 as dilution related to growth or changes in blubber lipid content (Aguilar et al., 1999).

186 Comparison of the measured POP values between adults and calves indicated that there  
187 was a 1.1 to 2.8 times greater concentration in calves. The initial summed POP levels of the

188 calves are assumed to be primarily from lactational and to a lesser extent, gestational transfer  
189 from their mothers. The decrease in summed POP concentrations between the juvenile and the  
190 subadult age classes was unexpected. There is a possibility that some individuals classified as  
191 subadults were actually adults that were missed in reproductive years or experienced calf  
192 mortality prior to being seen. Such individuals could therefore have already undergone some  
193 contaminant offloading. Although both adult and subadults likely accumulate POPs through  
194 their diet, the decrease in summed POP levels between these age classes suggests maternal  
195 offloading may be the predominate factor that determines POP levels in adult females. In the  
196 current study, the youngest sampled parous female was at least 7 years and the oldest sampled  
197 subadult was 12 years. The summed concentrations of the five POP classes generally decreased  
198 with age after age 4 and became relatively constant (Figure 2).

199 We also examined the percent contribution of PCB homologs to  $\sum$ PCBs to determine if  
200 there were qualitative differences among the four age classes of whales (Figure 3). PCB  
201 congeners were grouped according to the number of chlorine atoms [e.g., pentachlorinated  
202 congeners (5 Cl atoms), hexachlorinated congeners (6 Cl atoms)]. We found that the  
203 composition of the different PCB homolog groups varied slightly among the four age classes,  
204 with a tendency for the higher chlorinated homologs (i.e., heptachlorobiphenyls,  
205 octachlorobiphenyls, nonachlorobiphenyls) to be proportionately more abundant in adults and  
206 less in the calves. In contrast, hexachlorobiphenyls, the homolog group with the most detected  
207 congeners, were proportionately more abundant in calves and progressively decreased between  
208 the age classes (Figure 3). Previous cetacean studies have shown that higher chlorinated PCBs  
209 are not as readily offloaded from mother to offspring. (Aguilar et al., 1999; Yordy et al., 2010).  
210 The age associated decrease in pentachlorobiphenyls may be a consequence of growth dilution  
211 and/or elimination such as through biotransformation.

212

### 213 **Changes in POPs determined from repeat samples and other studies**

214 Age-related changes in  $\sum$ POPs concentrations were further evaluated based on three  
215 female humpback whales sampled twice, each more than a year apart (Figure 4).

216 These results suggest that POP levels initially increase with age, peaking around age 4-5 and  
217 then continually decrease with age. These results highlight how age and other life history factors

218 at time of sampling are important variables to consider when comparing POP levels in female  
219 humpback whales with prior studies.

220 Blubber POP concentrations reported in the current study were compared with values  
221 reported previously in humpback whales from the western North Atlantic, eastern North Atlantic,  
222 southwestern Indian Ocean, and the western Antarctic waters (Metcalf et al. 2004; Elfes et al.,  
223 2010; Ryan et al., 2013; Dorneles et al., 2015; Das et al., 2017, Table 2). Mean  $\sum$ PCB levels in  
224 adult females from the current study were approximately 1.9 to 630 times higher than levels in  
225 adult females from other populations, with the lowest levels reported in whales from the South  
226 Western Indian Ocean. Mean  $\sum$ PCBs levels determined in two calves in the current study were  
227 2.9 times higher than levels measured in calves from the Gulf of St. Lawrence (Metcalf et al.,  
228 2004). Mean  $\sum$ PCBs levels in adult females were higher than levels in adult males from the  
229 other studies, with the exception of those collected from the same sampling region (Elfes et al.,  
230 2010).

231 The highest observed PCB concentrations in the present study are below the most widely  
232 used threshold of 17,000 ng/g, lipid weight for onset of adverse effects in aquatic mammals  
233 (Kannan et al. 2000). However, a more recent study has suggested the threshold may be lower at  
234 9,000 ng/g, lipid weight (Jepson et al. 2016). This lower threshold value is within the range of  
235 observed values in juvenile whales (Table 1). An added concern is the increased sensitivity of  
236 neonatal and juvenile life stages towards perturbations in thyroid hormone action (Zoeller and  
237 Rovett 2004), which is one of the known effects of PCB exposure in marine mammals (Brouwer  
238 et al. 1989). Estimates of toxicity threshold values integrate results from many studies and  
239 generally are not specific for juvenile life stages nor fully consider potential synergistic  
240 interactions with other POPs present in the whales. Thus, toxicity thresholds for the actual  
241 mixture of POPs may be much lower, especially for juveniles that are potentially more sensitive  
242 and yet have the highest levels of contaminants in humpback whales.

243

#### 244 **4. Conclusion**

245 In summary, POPs varied with age class in blubber samples from female humpback whales  
246 from the Gulf of Maine. Among the five POP classes, mean summed POP levels were highest in  
247 juveniles and subadults and lowest in adults. Among the four age classes, POP mean  
248 concentrations followed the order  $\sum$ PCB >  $\sum$ DDTs >  $\sum$ CHLs >  $\sum$ PBDEs > other OCPs and



249 HCHs. Due to the confounding effects of maternal offloading, POPs data collected from female  
250 marine mammals is less common than that of males, but due to adverse health effects such as  
251 increased susceptibility to disease, immune and reproductive dysfunction and endocrine  
252 impairment, is important in assessing the health of a population. In addition, POP concentration  
253 data, such as those reported in the current study, can be used to develop models that can help  
254 determine the potential impacts of these toxic compounds on population growth of cetaceans  
255 (Hall et al., 2018).

256

### 257 **Declaration of competing interest**

258 The authors declare no potential conflicts of interest with respect to the research, authorship,  
259 and/or publication of this article. Certain commercial equipment or instruments are identified in  
260 the paper in order to adequately specify the experimental procedure. Such identification does not  
261 imply recommendation or endorsement by NOAA Fisheries, nor does it imply the equipment is  
262 the best available for the purpose.

263

### 264 **CRdiT authorship contribution statement**

265 **Keri A. Baugh:** Investigation, formal analysis, data curation, visualization, writing -  
266 original draft. **Jooke Robbins:** Investigation, conceptualization, formal analysis, data curation,  
267 Writing - original draft, review, editing, and funding acquisition. **Irvin R. Schultz:**  
268 Conceptualization, formal analysis, visualization, Writing - original draft, review and editing.  
269 **Gina M. Ylitalo:** Investigation, conceptualization, formal analysis, data curation, visualization,  
270 Writing - original draft, review and editing, project administration, funding acquisition.  
271 .

### 272 **Acknowledgements**

273 We are grateful to David Mattila, Jennifer Tackaberry, Scott Landry for assistance in the field.  
274 We thank the Environmental Chemistry Program staff: Jennie Bolton, Daryle Boyd, Richard  
275 Boyer, Denis da Silva, David Herman, Jonelle Gates, and Catherine Sloan. Samples were  
276 collected under NOAA NMFS permits 633-1483 and 633-1778, issued to the Center for Coastal  
277 Studies.

278

279 **Funding** We appreciate the financial support provided by Dr. Teri Rowles of the NOAA  
280 Fisheries' Office of Protected Resources.

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