

1 **Carotenoid glycosides from cyanobacteria are teratogenic in the zebrafish (*Danio rerio*)**  
2 **embryo model**

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24 **ABSTRACT**

25 Toxigenicity of cyanobacteria is widely associated with production of several well-described  
26 toxins that pose recognized threats to human and ecosystem health as part of both freshwater  
27 eutrophication, and episodic blooms in freshwater and coastal habitats. However, a  
28 preponderance of evidence indicates contribution of additional bioactive, and potentially toxic,  
29 metabolites. In the present study, the zebrafish (*Danio rerio*) embryo was used as a model of  
30 vertebrate development to identify, and subsequently isolate and characterize, teratogenic  
31 metabolites from two representative strains of *C. raciborskii*. Using this approach, three  
32 chemically related carotenoids - and specifically the xanthophyll glycosides, myxol 2'-glycoside  
33 (1), 4-ketomyxol 2'-glycoside (2) and 4-hydroxymyxol 2'-glycoside (3) - which are, otherwise,  
34 well known pigment molecules from cyanobacteria were isolated as potentially teratogenic  
35 compounds. Carotenoids are recognized "pro-retinoids" with retinoic acid, as a metabolic  
36 product of the oxidative cleavage of carotenoids, established as both key mediator of embryo  
37 development and, consequently, a potent teratogen. Accordingly, a comparative toxicological  
38 study of chemically diverse carotenoids, as well as apocarotenoids and retinoids, was undertaken.  
39 Based on this, a working model of the developmental toxicity of carotenoids as pro-retinoids is  
40 proposed, and the teratogenicity of these widespread metabolites is discussed in relation to  
41 possible impacts on aquatic vertebrate populations.

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48 *Keywords:*

49 Cyanobacteria

50 Carotenoid

51 Myxoxanthophyll

52 Zebrafish

53 Teratogenicity

54 Retinoid

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71 **1. Introduction**

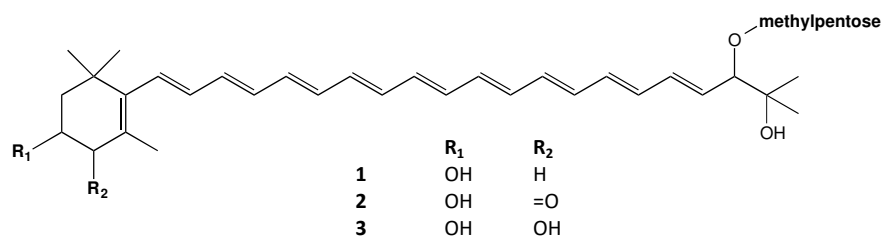
72 Cyanobacteria are ubiquitous photosynthetic prokaryotes that, as the “blue-green algae,” are  
73 perhaps most conspicuous in the environment in association with episodic and seasonal algal  
74 blooms, as well as eutrophication of freshwater and coastal marine habitats. A growing number  
75 of cyanobacterial species are recognized producers of toxic metabolites linked to various human  
76 health concerns, including acute poisoning and chronic health effects (e.g. cancer,  
77 neurodegenerative disease), particularly as part of so-called “harmful algal blooms” (HABs)  
78 (Carmichael, 2008). In addition, a considerable body of knowledge suggests that cyanobacterial  
79 toxins may impact aquatic animal health (e.g. fish kills, poisoning of wildlife) alongside broader  
80 effects on aquatic ecosystems including declines and/or compositional shifts in both invertebrate  
81 and vertebrate populations (Leão et al., 2009; Zanchet and Oliveira-Filho, 2013).

82 Despite a recognized diversity of bioactive compounds from cyanobacteria, studies of  
83 cyanobacterial toxicity have primarily focused on a relatively small number of metabolites  
84 including several water soluble alkaloids and peptides linked to human intoxications as  
85 hepatotoxins (i.e. microcystins, cylindrospermopsin) and neurotoxins (e.g. anatoxin-a, saxitoxin).  
86 However, numerous studies, specifically comparing crude extracts to pure toxins, have  
87 consistently demonstrated contributions of additional, and seemingly unrelated, bioactive  
88 metabolites to the toxic potential of cyanobacteria (e.g. Saker et al., 2003; Nogueira et al., 2006;  
89 Berry et al., 2009; Acs et al., 2013; Jonas et al., 2014). These additional toxic metabolites have  
90 remained, in most cases, uncharacterized both chemically and toxicologically.

91 In previous studies, the zebrafish (*Danio rerio*) embryo was employed as a model of  
92 vertebrate development to evaluate teratogenicity (i.e., developmental toxicity) of metabolites

93 from cyanobacteria (e.g. Berry et al., 2007; Berry et al., 2009; Jaja-Chimedza et al., 2012;  
94 Walton et al., 2014; Jaja-Chimedza et al., 2015). Notable among these, *Cylindrospermopsis*  
95 *raciborskii*, a species frequently (and increasingly) linked to freshwater HABs, was found to  
96 produce teratogenic metabolites (Berry et al., 2009). With respect to toxic potential of the  
97 species, *C. raciborskii* has been most commonly associated with production of toxic alkaloids,  
98 namely cylindrospermopsin (CYN) and saxitoxin (STX), which have been well characterized as  
99 hepatotoxic and neurotoxic compounds, respectively (Kinnear, 2010; Wiese et al., 2010).  
100 Evaluation in the zebrafish embryo system clearly demonstrated that observed teratogenicity,  
101 however, was unrelated to these recognized toxins, and suggested a contribution of otherwise  
102 uncharacterized lipophilic metabolites from the species (Berry et al., 2009).

103 In the present study, the zebrafish embryo teratogenicity assay was employed to isolate (via  
104 bioassay-guided fractionation) and subsequently characterize additional bioactive metabolites  
105 from *C. raciborskii*, and specifically identified a series of carotenoid glycosides (Fig. 1) as  
106 apparently potent teratogens. Indeed, carotenoids including, in particular, a few notable  
107 congeners (e.g.,  $\beta$ -carotene) are well known “pro-retinoids,” serving as metabolic precursors to  
108 retinoic acid (RA) and related retinoids that, in turn, are recognized to be potent teratogens. The  
109 observed teratogenicity of the cyanobacterial carotenoid glycosides, therefore, would be  
110 consistent with a pro-retinoid mechanism of developmental toxicity. In order, to evaluate  
111 whether the observed teratogenicity could be extended to other carotenoids, several chemically  
112 diverse carotenoids (including recognized pro-retinoids), as well as apocarotenoids and retinoids,  
113 were subsequently evaluated in the zebrafish embryo model. A working model of the  
114 teratogenicity of the carotenoids is accordingly proposed.



115

116 **Fig. 1.** Proposed structure of carotenoids (1-3) isolated in the present study.

117

118 **2. Methods and Materials**119 *2.1. C. raciborskii culture material*

120 Carotenoids in the present study were isolated from cultured material of two strains of *C.*  
 121 *raciborskii*. Cultures of *C. raciborskii* AQS were provided by Centro de Investigação Marinha e  
 122 Ambiental (Laboratory of Ecotoxicology) at the University of Porto (Portugal), and originally  
 123 isolated from aquaculture ponds in Australia (Saker and Eaglesham, 1999). Cultures of *C.*  
 124 *raciborskii* 121-1 were derived from samples isolated from Lake Catemaco (Veracruz, Mexico;  
 125 Berry et al., 2012). Cultures of both were maintained as previously described (Berry et al, 2007;  
 126 Gantar et al., 2008; Berry et al. 2009). Algal cells/biomass were harvested after three to four  
 127 weeks by centrifugation, and subsequently freeze-dried for extraction.

128

129 *2.2. Zebrafish embryo teratogenicity assay*

130 The zebrafish (*Danio rerio*) embryo was used as a model of vertebrate development to  
 131 evaluate teratogenicity, and toxicity more generally. Maintenance and breeding of adult  
 132 zebrafish, as well as the zebrafish embryo teratogenicity assay, was done as previously described  
 133 (Berry et al., 2007; Berry et al., 2009; Jaja-Chimedza et al., 2012; Walton et al., 2014). Assays  
 134 were conducted in polypropylene 24-well plates (Evergreen Scientific, Los Angeles, CA) with  
 135 five embryos (4- to 32-cell stage) per well, i.e., replicate (n=3), in E3 medium (Brand et al.,

136 2002). Embryos were observed at 1, 2, 3 and 5 days post fertilization (dpf) with a dissecting light  
137 microscope to assess mortality and developmental toxicity. Mortality was scored by two  
138 complementary criteria including: coagulation of embryos at early stages ( $\leq 2$  dpf), and  
139 deformity sufficiently severe (e.g., lack of somites, lack of heartbeat) to be considered effectively  
140 lethal. Lethal and effective concentrations for 50% mortality and deformity were estimated,  
141 respectively, as  $LC_{50}$  and  $EC_{50}$  by probit analysis. Following the identification of **1-3** as  
142 carotenoids (see below), assays were subsequently conducted in dim light with incubation of  
143 assay plates (0 to 5 dpf) in the dark, and at lowered temperatures (25°C versus optimal 28°C), to  
144 minimize potential degradation of the temperature sensitive and photolabile carotenoids. All  
145 breeding and bioassays involving zebrafish were conducted under protocols approved by the FIU  
146 Institutional Animal Care and Use Committee (IACUC), and performed by trained investigators.

147

### 148 *2.3. Extraction and purification of carotenoids by bioassay-guided fractionation*

149 The zebrafish embryo toxicity assay was used to guide fractionation toward purification of  
150 teratogenic metabolites from both strains of *C. raciborskii*. With no prior knowledge of their  
151 chemical identity, carotenoids were purified from the two strains independently, and specifically  
152 using bioactivity, i.e., developmental toxicity, to guide isolation. Purification from the two  
153 strains, therefore, followed two separate isolation schemes. In both cases, freeze-dried biomass  
154 was twice extracted in chloroform (10 mg biomass/1 mL of extraction solvent), and evaluated for  
155 teratogenicity (i.e. embryo deformities, developmental dysfunction) at several dilutions; the  
156 observed teratogenicity was employed at each subsequent fractionation step to identify relevant  
157 bioactive fractions. Extracts were subsequently fractionated by normal phase (silica gel, 60Å,  
158 Commercial 40–63  $\mu\text{m}$ ) flash column chromatography using two separately developed methods,

159 and specifically either a step-wise gradient of (1) ethyl acetate in hexane, followed by 10%  
160 methanol in ethyl acetate, for *C. raciborskii* AQS; or (2) acetone in hexane, followed by 100%  
161 methanol, for *C. raciborskii* 121-1. The most polar (10% methanol in ethyl acetate) fraction  
162 from AQS was found to be bioactive, and fractionated further (following concentration *in vacuo*)  
163 using reverse-phase (C-18) solid-phase extraction (SPE, Extract-Clean™ C18-HC 10 g/75 mL)  
164 with step-wise gradient of elution by methanol in water. The resulting bioactive fraction (95%  
165 methanol in water) was separated by high-performance liquid chromatography (HPLC), coupled  
166 to photodiode array (PDA) detector. Isocratic elution (65:35 acetonitrile/water) and a reverse-  
167 phase column (Phenomenex Luna C5, 5  $\mu\text{m}$ , 100  $\text{\AA}$ , 100 mm  $\times$  4.6 mm) enabled purification of  
168 two bioactive compounds (**1** and **2**), specifically eluting at 6.7 and 8.6 min. Similarly, the most  
169 polar fraction (100% methanol eluate) from 121-1 was found to be bioactive, and subsequently  
170 separated by HPLC (identically as AQS) affording a single bioactive peak (**3**) eluting at 4.6 min.  
171 Following the purification of initial quantities of the compounds, and tentative identification as  
172 carotenoids, subsequent purification was done in dim yellow light, and all samples were stored in  
173 the dark (at  $-20^{\circ}\text{C}$ ), to minimize degradation of these light- and heat-sensitive compounds.

174

#### 175 2.4. Chemical characterization of isolated carotenoids

176 The bioactive compounds purified from AQS were chemically characterized by UV/Vis  
177 spectroscopy, electrospray ionization mass spectrometry (ESI-MS) and nuclear magnetic  
178 resonance (NMR) spectroscopy. Due to instability of the compounds (see below), UV/Vis  
179 absorbance and MS spectra were primarily obtained with an on-line PDA detector and mass  
180 spectrometer, respectively, coupled to HPLC. Chemical characterization based on UV/Vis  
181 spectroscopy specifically included evaluation of the maximal absorbance ( $\lambda_{\text{max}}$ ) within “three-



182 peak spectra” indicative of carotenoids, and associated *spectral fine structure*. The latter  
183 specifically included calculation of the relative peak height of the second and third highest  $\lambda_{\max}$   
184 within this spectrum (i.e.  $\%III/II = \text{peak height of III}/\text{peak height of II} \times 100\%$ ) that has been  
185 extensively used to determine the structure of carotenoids (Takaichi and Shimada, 1992). Low-  
186 resolution ESI-MS analysis was done using a Thermo TSQ Quantum Access electrospray  
187 ionization (ESI)/triple quadrupole instrument coupled to a Thermo Accela UHPLC. High-  
188 resolution ESI-MS (HRESIMS) was achieved using a Thermo Q Exactive Orbitrap mass  
189 spectrometer, similarly coupled to an Accela UHPLC systems. Chemical characterization by  
190 NMR included one-dimensional (i.e.,  $^1H$ ), and two-dimensional experiments (i.e., homonuclear  
191 COSY) on a Bruker AVANCE 400 MHz instrument.

192 In addition to spectroscopic characterization, purified compounds and bioactive fractions  
193 were evaluated, based on the tentative identification of the bioactive metabolites as carotenoids,  
194 using the previously described (Louda et al., 2008) analytical method specific for algal pigments  
195 including carotenoids. Samples were spiked with relevant carotenoid standards (i.e.,  
196 myxoxanthophyll and aphanizophyll; DHI, Denmark) for confirmatory identification.

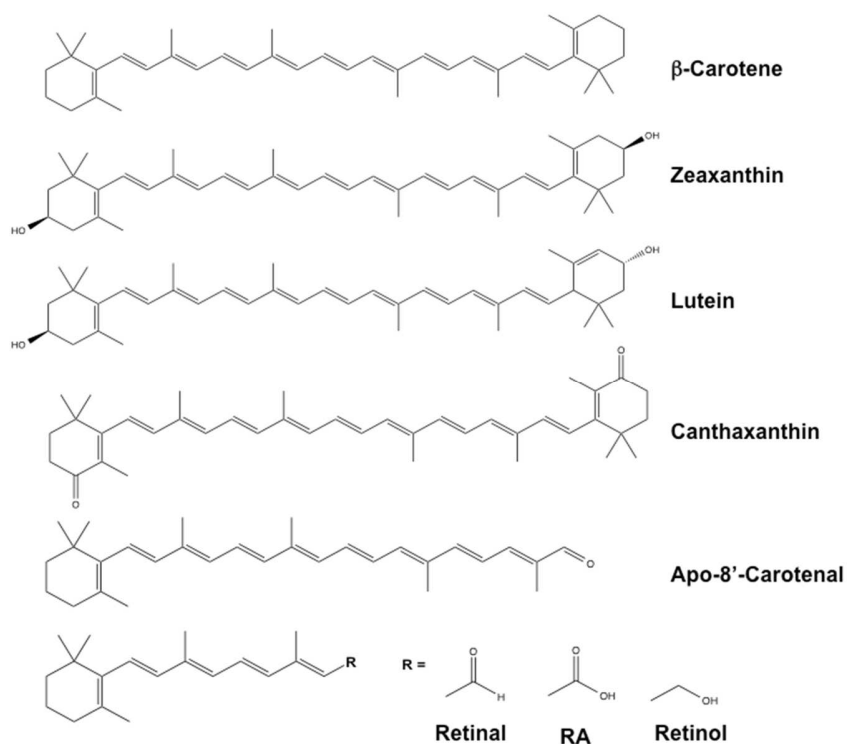
197  
198 **Myxol 2'-glycoside (1)**: red-orange amorphous solid; UV/Vis (PDA)  $\lambda_{\max}$  295, 368, 450, 478,  
199 and 509 nm,  $\%III/II = 58$ ; ESI-MS  $m/z$  567.4 (loss of methylpentoside), 730.5 ( $M^+$ ), 753.5  
200 ( $[M+Na]^+$ ); HRESIMS  $m/z$  730.4801 ( $C_{46}H_{66}O_7$ ,  $[M]^+$ ,  $\Delta m_{mu}$  of -2.24), ring-and-double-bond  
201 equivalents (RDB) = 14.

202

203 **4-Ketomyxol 2'-glycoside (2)**: red-orange amorphous solid; UV/Vis (PDA)  $\lambda_{\max}$  321, 449sh,  
 204 480 and 509 nm, %III/II = 12; ESI-MS  $m/z$  581.4 (loss of methylpentoside), 744.4 ( $M^+$ ), 767.4  
 205 ( $[M+Na]^+$ ). HRESIMS  $m/z$  744.4548 ( $C_{46}H_{64}O_8$ ,  $[M]^+$ ,  $\Delta mmu$  of -2.24), RDB = 15.

206

207 **4-Hydroxymyxol 2'-glycoside (3)**: UV/Vis (PDA)  $\lambda_{\max}$  451, 476, and 507nm, %III/II = 56;  
 208 ESI-MS  $m/z$  583.4 (loss of methylpentoside), 746.5 ( $M^+$ ), 769.5 ( $[M+Na]^+$ ); HRESIMS  $m/z$   
 209 746.4762 ( $C_{46}H_{66}O_8$ ,  $[M]^+$ ,  $\Delta mmu$  of 0.850), RDB = 14.0.



210

211 **Fig. 2.** Carotenoids and retinoids evaluated for teratogenicity in the zebrafish embryo model.

212 Carotenoids include carotenes ( $\beta$ -carotene), and xanthophylls (zeaxanthin, lutein, canthaxanthin),

213 as well as an apocarotenoid (apo-8'-carotenal). Retinoids include the three common products of

214 oxidative cleavage of  $\beta$ -carotene (retinal, retinoic acid [RA] and retinol).

215

216 2.5. Comparative toxicology of carotenoids and retinoids

217 To determine whether the observed teratogenicity was limited to the xanthophyll  
218 glycosides isolated from *C. raciborskii*, or rather common to other members of this chemical  
219 family, and to compare to the teratogenicity of retinoids, several chemically distinct (Fig. 2)  
220 carotenoids including carotenes, xanthophylls and apocarotenoids, along with several retinoids,  
221 were evaluated in the zebrafish embryo teratogenicity assay. Commercially available  
222 compounds tested, in this regard, included: lutein (Cayman Chemical, Ann Arbor, MI, U.S.A.);  
223 canthaxanthin and zeaxanthin (Fluka, Switzerland);  $\beta$ -carotene, trans- $\beta$ -apo-8' carotenal, *all-*  
224 *trans* retinoic acid, *all-trans* retinaldehyde and retinol (Sigma-Aldrich, St. Louis, MO, U.S.A.).  
225 Each compound was tested in the zebrafish teratogenicity assay, alongside **1** and **2**, and relevant  
226 solvent (i.e. methanol) only and untreated controls, in triplicate at 60, 30, 15, 5 and 1  $\mu$ M as  
227 described (see 2.2. *Zebrafish embryo teratogenicity assay*). Toxicity was assessed based on both  
228 mortality and observed developmental deformity; mortality was determined using two  
229 complementary criteria (i.e., coagulation and severe deformity; see 2.2. *Zebrafish embryo*  
230 *teratogenicity assay*). To assess relative toxicity, the percent mortality/developmental deformity  
231 (i.e., number of deformed embryos/number of live embryos x 100%) for each treatment was  
232 calculated. Percent mortality/deformity at each concentration (1, 5, 15, 30 and 60  $\mu$ M) and  
233 observational time point (1, 2, 3, 4 and 5 dpf) were compared by Fisher Exact Test using  
234 Analystsoft software (Walnut, CA, U.S.A.) to determine statistical significance relative to  
235 solvent-only (i.e., methanol) controls. Embryos, including representatives of development  
236 deformities, and relevant controls (i.e., untreated normal embryos), were photographed at each  
237 test concentration (1-60  $\mu$ M) and observational time-point (1-5 dpf).

238

### 239 **3. Results**

#### 240 *3.1. Isolation and chemical characterization of teratogenic carotenoids*

241 Chloroform extracts, and subsequent fractions, from both *C. raciborskii* isolates (AQS  
242 and 121-1) were evaluated by the zebrafish embryo assay, and found to be teratogenic. For AQS,  
243 a consistent pattern of teratogenicity was observed at each sequential fractionation step (i.e. flash  
244 column chromatography, C-18 SPE) enabling the purification of two bioactive components (**1**  
245 and **2**) in adequate quantities for subsequent chemical characterization. The development defects  
246 of embryos exposed to these extracts and subsequent fractions was characterized by a general  
247 lack of differentiation of head and tail, and associated features (e.g., eye), as well as anterior  
248 aggregation (suggesting inhibited migration) of melanophores. Developmental defects were  
249 generally severe enough in all cases, including subsequent bioassay-guided fractionation, to be  
250 effectively lethal to embryos. In parallel, qualitatively similar teratogenicity was observed for  
251 extracts, and subsequent fractions, from *C. raciborskii* 121-1, and used to identify a single  
252 bioactive component (**3**). Compound **3** was not purified in sufficient quantity for extensive  
253 spectroscopic analysis, but chemical characterization based on UV/Vis spectroscopy (Fig. 3) and  
254 mass spectrometry, specifically coupled to analytical separation by HPLC, as well as a  
255 previously validated analytical technique for algal pigments (Louda, 2008), was sufficient for  
256 confirmatory identification of this compound.

257 Initial chemical characterization of purified **1** and **2** included UV/Vis spectroscopy, ESI-MS  
258 and NMR. Both compounds were found to be highly unstable as indicated by, for example, loss  
259 of red-orange color of the compounds following mild heating (to evaporate solvents during  
260 isolation), and upon standing at room temperature and ambient light. Likewise, all attempts to  
261 conduct NMR spectroscopy with the purified compounds indicated clear degradation of the

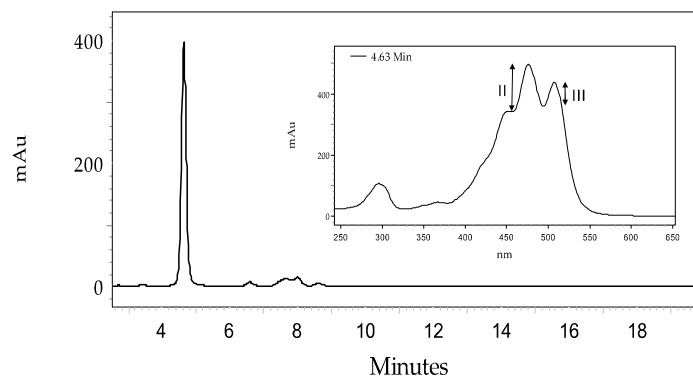
262 compounds such that these analyses were not possible. This instability was, in fact, consistent  
263 with tentative identification of the compounds as carotenoids. Chemical characterization of  
264 carotenoids generally relies on, as the minimal criteria for acceptable identification, UV/Vis  
265 spectroscopy in conjunction with mass spectrometry, and the use of appropriate reference  
266 standards (Britton et al., 2004). The utility of NMR analyses (without derivatization) is largely  
267 limited by this instability, and consequent degradation. In particular, the so-called “three-peak  
268 spectra” of carotenoids - the maximal peaks (I, II and III) of UV/Vis absorbances ( $\lambda_{\max}$ ) between  
269 420-520 nm (e.g., Fig. 3) - is highly indicative of structure, and particularly used to determine  
270 conjugation and cyclization of the variants. Within this, the *spectral fine structure*, and  
271 specifically relative peak heights (i.e., %III/II), is capable of determining conjugation and  
272 cyclization within the carotenoid structure (Takaichi and Shimada, 1992). In conjunction with  
273 UV/Vis spectroscopy, HRESIMS enabled generation of molecular formulae, and consequently  
274 assignment of carotenoid structures.

275 Based on these analyses, **1** was tentatively identified as the previously described xanthophyll,  
276 myxol 2'-glycoside (Fig. 1). The three-peak spectrum of **1** (450, 478 and 509 nm), and  
277 corresponding fine structure (%III/II = 58%), generally enabled tentative assignment as a  $\beta$ -  
278 carotenoid, and specifically a xanthophyll, typical of those reported for cyanobacteria. Mass  
279 spectrometric analysis of **1** identified a molecular mass of 730.4801, and corresponding  
280 molecular formula of C<sub>46</sub>H<sub>66</sub>O<sub>7</sub>. Notably, although ESI-MS was employed, the ionized  
281 molecular ion (i.e., M<sup>+</sup>) rather than the expected protonated molecular ion (i.e., M+H<sup>+</sup>) was  
282 observed. However, the presence of the stable molecular ion in ESI-MS is, in fact, consistent  
283 with previous observations for carotenoids, and has been specifically linked to the high degree of  
284 conjugation, and the resulting ability to withstand loss of an electron, among these neutral

285 analytes (Guaratini et al., 2005). The molecular formula determined by HRESIMS, and  
286 specifically the apparent presence of oxygen, is further suggestive of a xanthophyll (i.e., oxygen-  
287 containing carotenoid). Moreover, fragmentation patterns in the MS analysis, and specifically  
288 the observed loss of 163 amu (i.e.  $m/z$  567.4), is indicative of the presence of a sugar moiety, and  
289 particularly a methylpentoside, typical of many of the xanthophyll glycoside produced by  
290 cyanobacteria (Takaichi et al., 2001; Mohamed et al., 2005; Takaichi et al., 2005). Although the  
291 sugar moiety of **1** was not characterized in the present study, it has been most typically  
292 associated with the methylpentose, fucose (Takaichi et al., 2001; Takaichi et al., 2005).

293 Similar analyses enabled identification of **2** as the congener, ketomyxol 2'-glycoside. The  
294 three-peak spectrum of **2** (449, 480 and 509 nm) and spectral fine structure, and specifically the  
295 small %III/II (12%) of the associated fine structure, was consistent with a carotenoid containing  
296 a carbonyl within the  $\beta$ -ionone ring (effectively extending the conjugation within the  
297 chromophore). The molecular formula ( $C_{46}H_{64}O_8$ ) assigned, based on the molecular ion ( $m/z$   
298 744.4548) identified in HRESIMS analysis, as well as fragmentation patterns which specifically  
299 included, similar to **1**, apparent loss of a methylpentoside (163 amu,  $m/z$  581.4), likewise, were  
300 highly consistent with the proposed assignment as the 4-keto congener of the myxol glycoside  
301 (i.e., "keto-myxoxanthophyll").

302



303  
 304 **Fig. 3.** Bioactive component (**3**) identified from *C. raciborskii* 121-1. Detection by HPLC,  
 305 coupled to PDA detection, identified a single bioactive component with a UV/Vis spectrum  
 306 indicative of aphanizophyll.

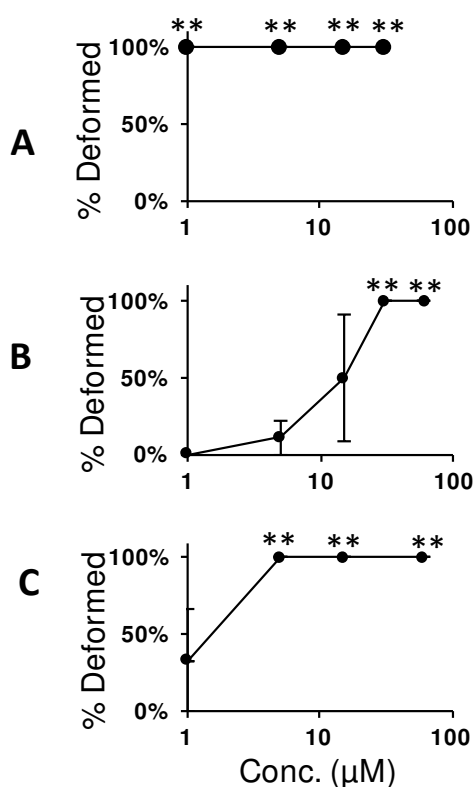
307  
 308 To confirm the identity of **1** and **2**, samples of the purified compounds were evaluated by a  
 309 previously validated (Louda, 2008) technique for analysis of algal pigments (i.e. chlorophyll,  
 310 carotenoids). This ternary HPLC-based method is an effectively confirmatory technique capable  
 311 (based on relative retention time, UV/Vis absorbance and use of relevant analytical standards) of  
 312 unambiguous identification of carotenoids and other pigments (Louda, 2008; Hagerthy et al.,  
 313 2006). Accordingly, the identity of both **1** and **2** were confirmed. This confirmatory technique  
 314 was additionally applied to bioactive fractions from *C. raciborskii* 121-1 (see above), and  
 315 specifically enabled identification of the single bioactive component (**3**) from this strain as the  
 316 previously described 4-hydroxymyxo 2'-glycoside (or "aphanizophyll").

317

### 318 3.2 Teratogenicity of retinoids in the zebrafish embryo model

319 Based on the proposed pro-retinoid mechanism of teratogenicity of **1-3**, three primary  
 320 retinoids, RA, retinal and retinol, were assessed for developmental toxicity. All retinoids

321 evaluated were teratogenic and effectively lethal by 2 dpf with a relative potencies of RA >  
 322 retinal > retinol (Fig. 4). In the present study, RA was tested at concentrations well above the  
 323 LC<sub>50</sub> and EC<sub>50</sub> (for teratogenicity) previously reported in the zebrafish embryo which are  
 324 generally in the nanomolar to sub-nanomolar range, respectively (Hermann, 1995; Selderslaghs  
 325 et al., 2009; Wang et al., 2014). And, indeed, teratogenicity was effectively lethal (Fig. 5A-B) at  
 326 even the lowest concentration of RA tested (1 μM) which is several orders of magnitude above  
 327 previously reported LC<sub>50</sub>/EC<sub>50</sub> values. Retinal and retinol, on the other hand, were quantitatively  
 328 less toxic than RA. Both resulted in developmental deformities which were sufficiently severe to  
 329 be lethal (and as such deformity was equivalent to mortality). However, the corresponding  
 330 EC<sub>50</sub>/LC<sub>50</sub> of retinal and retinol were lower than RA, and specifically calculated as 1.8 ± 0.7 μM  
 331 and 9.2 ± 7.6 μM, respectively.



332



333 **Fig. 4.** Comparison of the teratogenicity of retinoids in the zebrafish embryo model. Shown are  
334 percent deformities (n = 3, 5 embryos per replicate) for concentrations ranging from 1-60  $\mu$ M of  
335 RA (A), retinol (B) and retinal (C) at 3 dpf. Error bars represent one standard error above or  
336 below means. Statistically significant differences, compared to untreated controls, by Fisher  
337 Exact Test: \*\* p < 0.005

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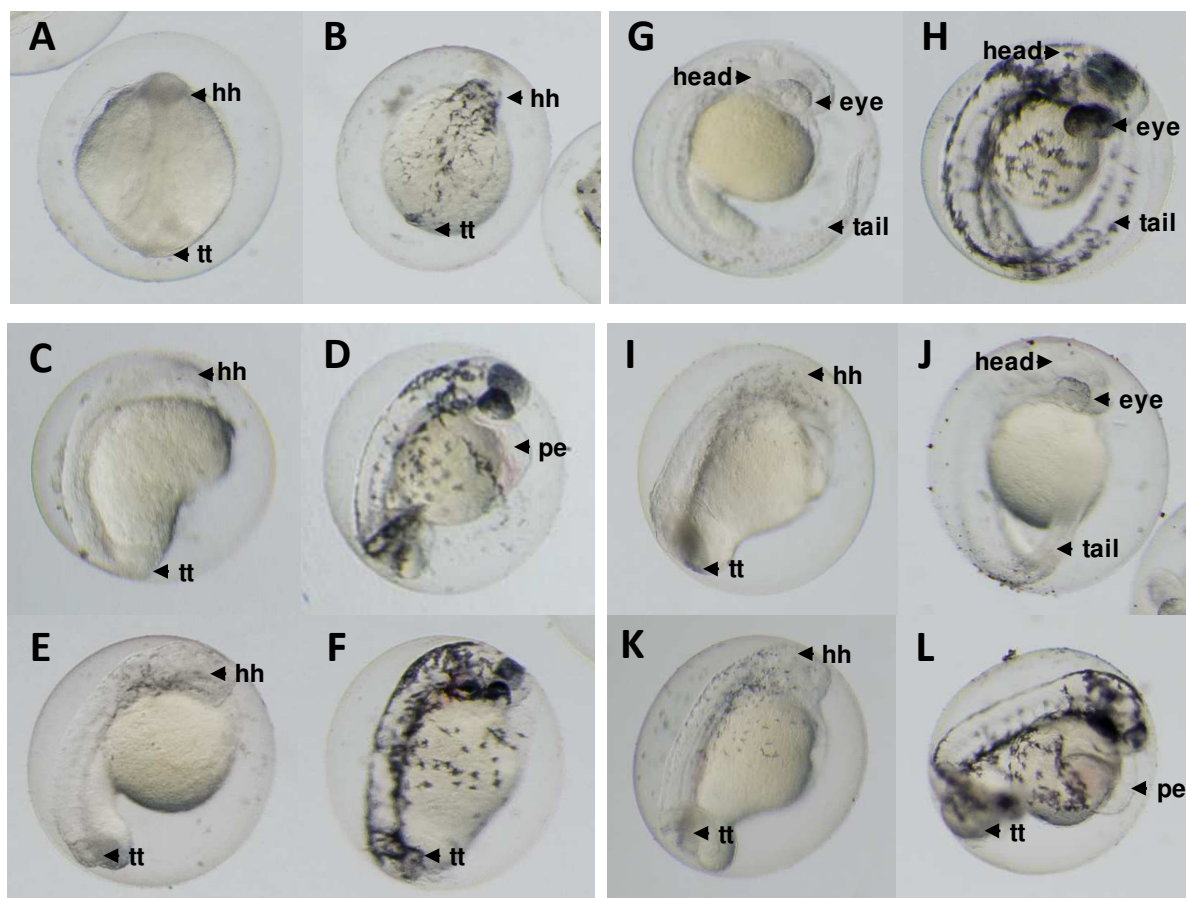
### 339 *3.3. Teratogenicity of carotenoids in the zebrafish embryo model*

340 A comparative toxicology study of several carotenoids in the zebrafish embryo  
341 teratogenicity assay was conducted to assess whether teratogenicity could be extended to other  
342 members of this widespread chemical family.  
343 Aside from teratogenicity of the carotenoids, however, development was generally delayed (see  
344 Fig. 5) in the exposure studies with carotenoids and retinoids (see previous section), specifically  
345 compared to standardized staging series (Kimmel et al., 1995; OECD, 2013), for all embryos  
346 including untreated controls. Under optimal conditions zebrafish embryos typically hatch at  
347 approximately 3 dpf, but embryos in the exposure studies remained unhatched through the 5 days  
348 of the study. Embryos were clearly underdeveloped (Fig. 5) at all stages with respect to  
349 organogenesis and other aspects when compared to typical studies. For example, in the present  
350 study embryos incubated for 3 dpf were equivalent in developmental stage to 31 hpf embryos  
351 typically reported in the literature, while 5 dpf embryos were equivalent to 3 dpf (Kimmel et al,  
352 1995).

353 Delayed development was most likely due to the exposure and rearing of embryos in  
354 constant dark conditions, as well as lowered temperature, which were used to prevent  
355 degradation of the photolabile, and temperature sensitive, carotenoids (see 2.2. *Zebrafish*

356 *embryo teratogenicity assay*). Similar delays in development related to reduced light and  
357 temperature have been previously reported by several authors (Schirone and Gross, 1968;  
358 Billotta, 2000; Villamazar et al., 2014; Di Rosa et al., 2015). Notably, for example, Billotta  
359 (2000) investigated effects of light on various aspects of zebrafish embryo development, and  
360 found that embryos raised in constant dark were generally “developmentally delayed,” and  
361 specifically “were still not hatched by 7 dpf.” Subsequent studies have, likewise, shown  
362 developmental delay due to rearing in constant dark including various consequent post-larval  
363 effects (Villamazar et al., 2014; Di Rosa et al., 2015). Alongside delayed development,  
364 relatively high embryotoxicity within 1-2 dpf was observed for all embryos (including controls)  
365 reared in constant dark, as specifically evidenced by coagulation, and independent of the  
366 generally lethal deformities. Embryo survival within 1-2 dpf was approximately  $61\% \pm 8\%$  for  
367 controls and did not significantly differ for embryos exposed to carotenoids (or retinoids).  
368 However, despite the generally delayed development, and relatively high mortality, comparison  
369 to untreated controls enabled teratogenicity to be clearly distinguished both quantitatively and  
370 qualitatively.

371



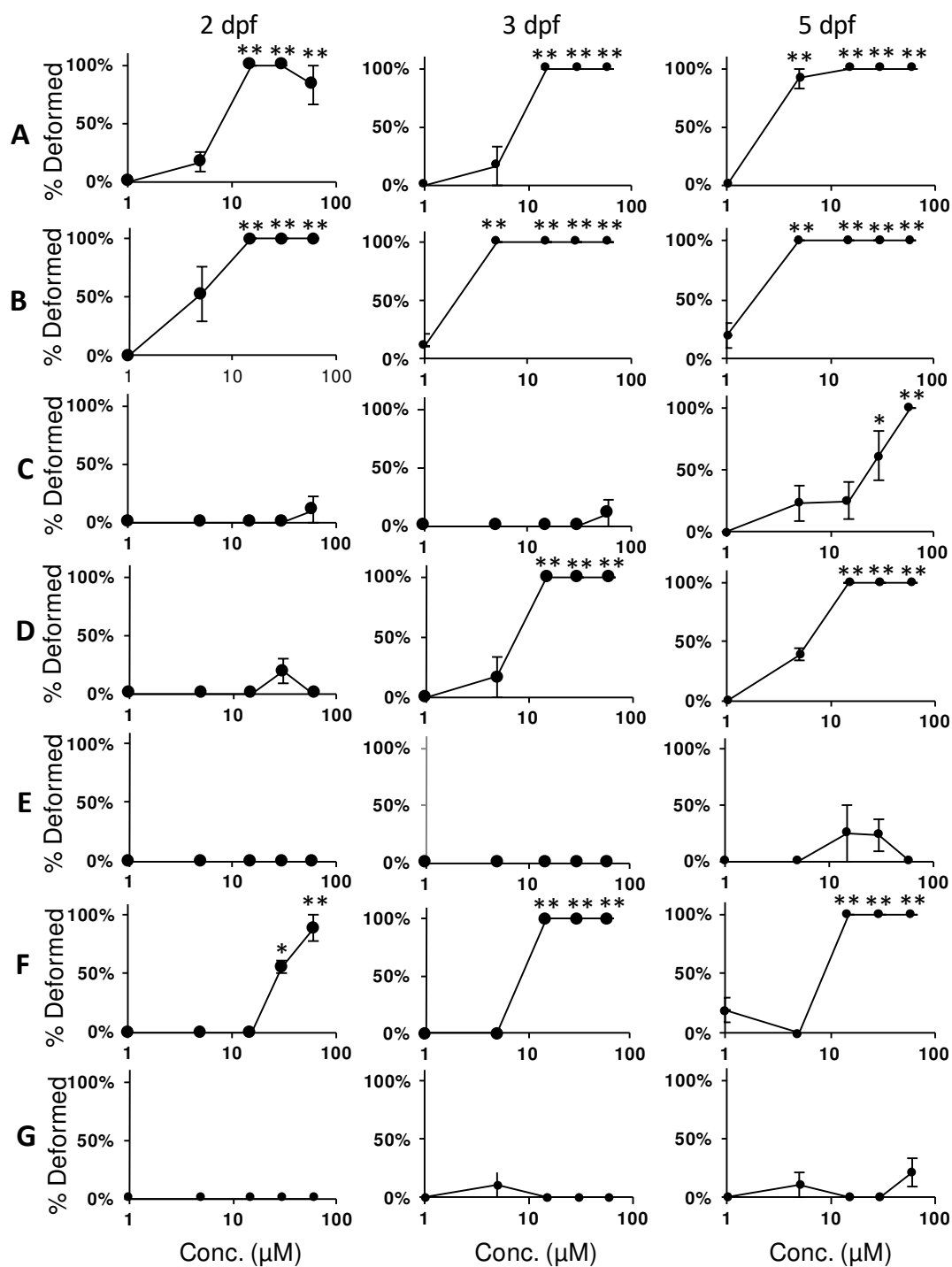
372

373 **Fig. 5.** Teratogenicity of carotenoids and retinoic acid. Shown are examples of the  
 374 teratogenicity of 1  $\mu$ M RA at 3 dpf (A) and 5 dpf (B) with embryos exhibiting severely arrested  
 375 development of the body axis including head and eye hypoplasia and truncated tail; 5  $\mu$ M **1** at 3  
 376 and 5 dpf (C and D, respectively) with head and eye hypoplasia, truncated tail and pericardial  
 377 edema; and **2** (E and F) at 3 and 5 dpf, respectively, with head and eye hypoplasia and truncated  
 378 tail compared untreated controls (G and H) at the same time points. For comparison between  
 379 carotenoid variants, the teratogenicity is shown for embryos exposed to: 30  $\mu$ M lutein (I) with  
 380 head and eye hypoplasia and truncated tail; 60  $\mu$ M zeaxanthin (J) with little effect at 3 dpf; 30  
 381  $\mu$ M canthaxanthin at 3 dpf (K) with head and eye hypoplasia and truncated tail; and 30  $\mu$ M  $\beta$ -

382 carotene at 5 dpf with truncated tail and pericardial edema (L). Abbreviations: hh = head and  
383 eye hypoplasia, tt = truncated tail and pe = pericardial edema.

384

385         Qualitatively, the developmental toxicity varied between carotenoids, and with exposure  
386 concentration (Fig. 5). By 3 dpf, for example, embryos exposed to both **1** and **2** at 5  $\mu$ M and  
387 greater showed pronounced developmental dysfunction (Fig. 5C-F) compared to controls (Fig.  
388 5H-G). Developmental deformities observed were sufficiently severe to be distinguished as  
389 lethal by 3-5 dpf. Lutein (Fig. 5I) and canthaxanthin (Fig. 5K) which reached levels of  
390 teratogenicity equivalent to **1** and **2** by 3 dpf (i.e., 100% deformity at  $\geq 15 \mu$ M; Fig. 6), likewise,  
391 shared several observable features of severe developmental dysfunction with the xanthophyll  
392 glycosides. These features included impaired development along the anterior-posterior axis, and  
393 specifically consequent head and eye hypoplasia and truncation of the tail, as well as pericardial  
394 edemas at 5 dpf (Fig. 5). Inhibition of the anterior-posterior development by carotenoids is  
395 notably reminiscent of, albeit less severe than, the effects of RA exposure (Fig. 5A-B) which  
396 showed severely arrested development along the body axis. In contrast, embryos exposed to  
397 zeaxanthin at the same developmental stage (e.g., 3-5 dpf) - and higher exposure concentrations  
398 (e.g., 60  $\mu$ M) - developed normally, and were indistinguishable from controls (Fig. 5J) as  
399 reflected, likewise, in the quantitative analyses. Teratogenicity of  $\beta$ -carotene which only  
400 manifested at  $\sim$ 5 dpf (Fig. 6) seemingly differed from **1** and **2**, and other xanthophylls (i.e., lutein,  
401 canthaxanthin). Although impairment of head/tail development was observed, it was perhaps  
402 less pronounced, and instead, a range of pericardial - and other - edemas were observed (at 5 dpf;  
403 Fig. 5L).



404

405 **Fig. 6.** Comparison of the teratogenicity of carotenoids in the zebrafish embryo model. Shown

406 are percent deformities (n = 3, 5 embryos per replicate) for concentrations ranging from 1-60 μM

407 of 1 (A), 2 (B), β-carotene (C), lutein (D), zeaxanthin (E), canthaxanthin (F) and β-apo-8'-

408 carotenal (G) at 2, 3 and 5 days post-fertilization (dpf). Structures of carotenoids are shown in  
409 Fig. 2. Error bars represent one standard error above or below means. Statistically significant  
410 differences, compared to untreated controls, by Fisher Exact Test: \*  $p < 0.05$ , \*\*  $p < 0.005$

411

412 Quantitatively, toxicity among carotenoid tested also varied significantly (Fig. 6).  
413 Statistically significant development toxicity (i.e., percent deformity) was observed within 2-3  
414 dpf for **1** and **2** at the three highest concentrations (15, 30 and 60  $\mu\text{M}$ ) tested. Percent deformity  
415 was additionally significant by 3 dpf for embryos exposed to the keto-variant (**2**) at the next  
416 highest concentration (5  $\mu\text{M}$ ). At 2 dpf, no other carotenoids were toxic with the notable  
417 exception of the diketo-xanthophyll, canthaxanthin, for which significant development toxicity  
418 was observed at the highest test concentration (approximately 89% for 60  $\mu\text{M}$ ; Fig. 6). By 3 dpf,  
419 however, teratogenicity of both canthaxanthin and the dihydroxylated xanthophyll, lutein,  
420 reached levels comparable to both **1** and **2** (i.e. 100%) at all three of the highest concentrations  
421 tested (Fig. 6). Developmental toxicity was sufficiently severe, in all cases, such that deformities  
422 were effectively lethal, and  $\text{EC}_{50}$  (for teratogenicity) and  $\text{LC}_{50}$  were, therefore, equivalently  
423 calculated ( $\pm$  standard error) at 5 dpf for myxol glycosides **1** and **2** as  $2.4 \pm 0.9 \mu\text{M}$  and  $1.8 \pm 0.2$   
424  $\mu\text{M}$ , respectively; and for xanthophylls, lutein and canthaxanthin, respectively, as  $4.7 \pm 3.0 \mu\text{M}$   
425 and  $8.7 \pm 1.8 \mu\text{M}$ . Interestingly, no teratogenicity was observed for the chemically related  
426 xanthophyll, zeaxanthin, during the 5-day exposures studies (Fig. 6). Developmental toxicity was  
427 observed for  $\beta$ -carotene, but only at the highest concentrations tested (30 and 60  $\mu\text{M}$ ), and only  
428 after 5 dpf ( $\text{EC}_{50}/\text{LC}_{50} = 23.5 \pm 8.9 \mu\text{M}$ ). Similarly, no teratogenicity was observed for the  
429 apocarotenoid,  $\beta$ -apo-8'carotenal, at 3 dpf; slight developmental toxicity (~22%) was observed at  
430 5 dpf, but this was not significantly different from controls. The relatively limited teratogenicity

431 of  $\beta$ -carotene is particularly notable as it is recognized as the primary substrate for symmetric  
432 cleavage by BCO1, subsequently leading to production of RA (Goodman et al., 1967; Lindqvist  
433 and Andersson, 2002).

434

#### 435 **4. Discussion**

##### 436 *4.2. Isolation and identification of myxol glycosides as teratogenic metabolites*

437 Bioassay-guided fractionation enabled purification of three chemically related xanthophylls  
438 from *C. raciborskii* as teratogens in the zebrafish embryo model, and subsequent chemical  
439 characterization identified the bioactive metabolites as myxol glycoside (**1**), along with its 4-  
440 keto (**2**) and 4-hydroxy (**3**) congeners. Frequently referred to as “myxoxanthophyll” in the  
441 literature (Hertzberg and Liaaen-Jensen, 1969), myxol glycoside (**1**) has been previously  
442 identified from cyanobacteria, including *C. raciborskii* (Varkonyi et al., 2002), and specifically  
443 shown to be required for integrity of thylakoid membranes (Mohamed et al., 2005). In contrast,  
444 4-keto-myxol glycoside (**2**) has not been previously identified from *C. raciborskii*, but has been  
445 identified (Schlüter et al., 2004; Takaichi et al., 2005) in other recognized toxigenic genera of  
446 cyanobacteria (e.g. *Anabaena*, *Nostoc*, *Nodularia*), and even been proposed (Schlüter et al.,  
447 2004) as a proxy of the presence of these toxin-producing species (e.g. *Nodularia spumigena*).  
448 Similarly, the 4-hydroxy variant has, likewise, been found among various cyanobacterial  
449 genera/species including quite recently *C. raciborskii* (Mehnert et al., 2012). Despite widespread  
450 – and, in fact, exclusive – occurrence of the xanthophyll glycosides among cyanobacteria, the  
451 teratogenicity observed in the current study is the first report of any bioactivity, specifically  
452 relevant to potential toxicity, of these metabolite.

453

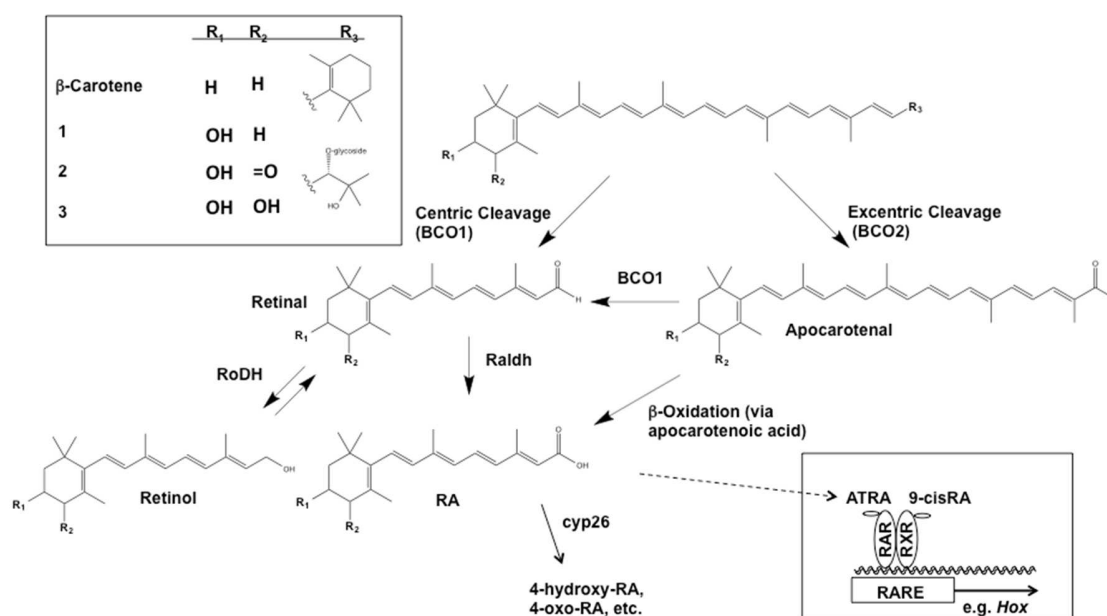
454 4.2. *Proposed pro-retinoid mechanism for teratogenicity of myxol glycosides*

455           Although this is the first report of teratogenicity associated with cyanobacterial (or any)  
456 carotenoids, notable members of the group are well-recognized as “pro-retinoids” serving as  
457 substrates for a biochemical diversity of so-called *carotenoid cleavage enzymes* (CCE) which  
458 convert carotenoids to various apocarotenoids and retinoids. As the primary example,  $\beta$ -carotene  
459 is converted to retinaldehyde via oxidative cleavage (Fig. 7) by a family of oxygenases, and  
460 specifically the so-called  $\beta$ -carotene 15,15'-oxygenases (BCO), which are widely distributed  
461 among vertebrates and invertebrates (von Lintig and Vogt, 2000; Shete and Quadro, 2013).  
462 Although  $\beta$ -carotene is primarily recognized as the substrate for BCOs, studies have  
463 demonstrated the oxidative cleavage of various other carotenoids by these enzymes. In particular,  
464 relatively high substrate specificity for  $\beta$ -carotene has been shown for members of the BCO1  
465 sub-type, whereas the less well characterized BCO2 sub-type is apparently capable of the  
466 oxidative cleavage of diverse carotenoids including carotenes and xanthophylls (Kiefer et al.,  
467 2001; Mein et al., 2011; Hu et al., 2006; Amengual et al., 2011; Sui et al., 2013; de la Seña,  
468 2014). Specifically, BCO2 cleaves carotenoids “eccentrically” as a 9',10'-oxygenase (as  
469 opposed to the symmetrical, or “centric,” cleavage by BCO1), giving rise to an apocarotenoid,  $\beta$ -  
470 apo-10'-carotenal (Fig. 7). The apocarotenoid can be subsequently converted to retinoids by one  
471 of two proposed mechanisms. It has been shown that apocarotenoic acids (via interconversion  
472 from apocarotenal) can undergo  $\beta$ -oxidation to retinoic acid (Wang et al., 1996). Alternatively,  
473 it has been more recently shown that BCO1 can, in fact, cleave a diversity of apocarotenals  
474 (including  $\beta$ -apo-10'-carotenal) to retinal with an efficiency similar to that of  $\beta$ -carotene  
475 (Amengual et al., 2013). Retinaldehyde can, in turn, be converted to other retinoids including



476 retinoic acid (RA) and retinol (i.e., Vitamin A) via retinaldehyde dehydrogenases (Raldh) and  
 477 retinol dehydrogenase (RoDH) steps, respectively (Fig. 7).

478



479

480 **Fig. 7.** Oxidative cleavage of carotenoids to retinoids, and subsequent role in gene expression.

481 Shown is the well-described conversion of  $\beta$ -carotene to retinal, retinol and retinoic acid, and the

482 proposed conversion of the cyanobacterial xanthophyll glycosides (**1-3**) isolated in the present

483 study to hydroxyl- and oxo-retinoids. Oxidative cleavage occurs either via centric cleavage to

484 produce retinals by BCO1, or excentric cleavage to apocarotenoids (e.g. apo-8'-carotenal) by

485 BCO2. Apocarotenoids can be converted to retinoids by either  $\beta$ -oxidation, or cleavage by

486 BCO1, as described. Retinal is converted enzymatically to either retinoic acid (RA) by

487 retinaldehyde dehydrogenase (Raldh), or reversibly to retinol by retinol dehydrogenase (RoDH).

488 Binding of all-trans and 9-cis RA isomers (ATRA and 9-cisRA) to the retinoic acid and "retinoid

489 X" receptors (RAR and RXR, respectively) enables formation of a heterodimer (i.e. RAR/RXR).

490 The dimer, in turn, binds to retinoic acid response element (RARE) sequences in various gene

491 (e.g. *Hox*) promoters. Concentrations of RA in developing cells are controlled via oxidation (and  
492 elimination) by a cytochrome P450, and specifically *cyp26*, to more polar 4-hydroxy and 4-oxo  
493 retinoids.

494

495 Retinoic acid - including all *trans* (ATRA) and 9-*cis* (9-*cis*RA) isomers - are essential  
496 signaling molecules in embryonic development. Binding of ATRA and 9-*cis*RA to so-called  
497 retinoic acid receptors (RAR) and “retinoid X” receptors (RXR), respectively, enables formation  
498 of heterodimers (RAR/RXR) which, in turn, mediate expression of a suite of genes via binding to  
499 *retinoic acid response elements* (RARE) within the genome (Fig. 7). Perhaps most notable  
500 among these genes are the highly conserved homeotic (*Hox*) genes that are fundamental to  
501 anterior-posterior patterning in development of embryos. Despite its role as key signaling  
502 molecule in embryonic development, RA and several of its variants are, therefore, also well-  
503 recognized teratogens (Hermann, 1995; Collins and Mao, 1999; Selderslaghs et al., 2009; Wang  
504 et al., 2014), and production of RA in developing embryos is, thus, tightly regulated. Regulation  
505 at this level specifically occurs via oxidative degradation by a cytochrome P450 (i.e., *cyp26*) that  
506 produces more polar metabolites (e.g. 4-hydroxy-RA, 4-oxo-RA).

507 In the present study, we evaluated a representative series of retinoids including ATRA,  
508 retinol and retinal. Teratogenicity of RA is directly linked (Minucci et al., 1997) to interaction  
509 with RAR (and, following dimerization with RXR, the RARE promoter). In contrast, it has been  
510 shown (Repa et al., 1993) that retinal shows essentially no binding affinity to RAR, whereas  
511 retinol does compete with RA for binding to RAR, but with a much lower (i.e., approximately 4-  
512 to 7-fold) affinity. The differential activity of the retinoids is generally consistent with the well-  
513 described interconversion of the retinoid congeners (Fig. 7; see von Lintig et al., 2005).

514 Specifically, retinal can be directly transformed by Raldh to RA, whereas retinol is reversibly  
515 converted, specifically by RoDH, to retinal that can, in turn, be enzymatically transformed (by  
516 Raldh) to RA. We propose, therefore, that this tripartite interconversion to form teratogenic RA  
517 (i.e. retinol  $\leftrightarrow$  retinal  $\rightarrow$  RA), along perhaps with possible direct binding of retinol, explains the  
518 three-tiered teratogenicity observed in the zebrafish embryo.

519 A number of very recent studies have pointed to the apparent production of retinoids by  
520 cyanobacteria and other freshwater algae (e.g., “green algae,” or Chlorophyta), and have  
521 suggested a role of the teratogenicity of these cyanobacterially derived retinoids in the toxicity of  
522 blooms including impacts on aquatic vertebrate populations (Wu et al., 2012 and 2013; Jonas et  
523 al., 2014 and 2015; Javurek et al., 2015). In a particularly notable study, Wu et al. (2012 and  
524 2013) identified an array of retinoids from algal blooms in the notoriously eutrophic Lake Taihu  
525 (China), as well as thirty-nine algal and cyanobacterial cultures isolated from the lake and  
526 diverse range of other freshwater sources. Evaluation of cyanobacterial and algal (including  
527 chlorophyte, diatom and euglenophyte) strains in this study showed that most (>82%) strains  
528 were capable of producing retinoids including retinal, and various isomers (e.g., 13-*cis*, 9-*cis* and  
529 all *trans*) of RA and 4-oxo-RA. Subsequent studies have, likewise, identified “retinoid-like”  
530 bioactivity of various cyanobacteria including observation of teratogenicity in the zebrafish  
531 embryo model (Jonas et al., 2014 and 2015). In one study, it was observed that retinoid activity  
532 could not be explained by the presence of the detected retinoids, and suggested that other  
533 metabolites contribute to this bioactivity (Javurek et al., 2015).

534 It is perhaps not surprising that cyanobacteria are capable of producing retinoids: prior  
535 studies have, in fact, identified and characterized a family of apocarotenoid oxygenases (ACOs)  
536 from cyanobacteria, and have shown the ability of these enzymes to cleave diverse

537 apocarotenoids including the consequent production of retinal (Sherzinger et al., 2006;  
538 Scherzinger and Al Babili, 2008). Indeed, it has been suggested, but not shown, in these more  
539 recent studies that carotenoids might be the biosynthetic source of retinoids produced  
540 cyanobacteria (Wu et al., 2013).

541 The observed teratogenicity of **1-3** specifically led to the hypothesis that myxol glycosides  
542 act as pro-retinoids, and specifically provide a substrate for BCO cleavage, and subsequent  
543 production of retinoids. Indeed, the observed teratogenicity in zebrafish embryos exposed to **1-3**  
544 bears striking quantitative and qualitative resemblance to the teratogenicity previously observed  
545 (Hermann et al., 1995; Jonas et al., 2014) – and repeated in the current study - for RA in this  
546 model system. Although more severe in the case of RA exposure, embryos exposed to **1, 2** and  
547 RA were characterized by clearly impaired development of the body axis, and consequent  
548 hypoplasia of the head and eye, as well as truncation of the tail (see Fig. 5). And, in fact, this  
549 feature is a hallmark of RA teratogenicity specifically owing to its role in anterior-posterior  
550 patterning via *Hox* gene regulation (Waxman and Yelon, 2009). Whereas retinal, followed by  
551 enzymatic conversion to RA, is the established product of the oxidative cleavage of  $\beta$ -carotene,  
552 oxidative cleavage of xanthophylls is expected to yield hydroxy- and keto-variants of these  
553 retinoids. Hypothesized products of **1** and **2** would thereby include, 3-hydroxy-RA and 3-  
554 hydroxy-4-oxo-RA, respectively. Notably, 4-hydroxy and 4-oxo-RA, as products of the  
555 oxidation of RA by *cyp26*, were previously shown to act as potentially teratogenic retinoids in the  
556 zebrafish and other models (e.g., (i.e. *Xenopus* embryos), and specifically act via interaction with  
557 RAR, and subsequent expression of *Hox* genes (Pijnappel et al., 1993; Hermann, 1995). The  
558 hypothesized products from oxidative cleavage of **1** and **2**, however, remain to be confirmed in  
559 future studies.

560 In further support of pro-retinoid hypothesis, and toward a model of teratogenicity, a series of  
561 carotenoids and apocarotenoids were evaluated in the zebrafish embryo model. Notably, all of  
562 the carotenoids (except **1** and **2**) have been previously detected - along with retinal, and relatively  
563 small quantities of retinol – as endogenous components in the yolk of zebrafish eggs (Lampert et  
564 al., 2003). Moreover, all of the carotenoids (with the exception of lutein, but including **1** and **2**)  
565 have been identified from cyanobacteria. In the present study, variable teratogenicity of several  
566 of the carotenoids was observed. Moreover, the developmental toxicity showed considerable  
567 quantitative and qualitative variation with respect to variant, effective concentration and  
568 embryonic stage at which teratogenicity was observed (Fig. 5 and 6).

569 Further studies are clearly needed, however, we propose based on these observations that the  
570 differential teratogenicity of the carotenoids during development is related, in part, to differential  
571 expression of BCO subtypes during embryo development. Currently two BCO subtypes, i.e.  
572 BCO1 and BCO2, are recognized, and both have been isolated from zebrafish embryos (Lampert  
573 et al., 2003; Lobo et al., 2012a). The latter is compartmentalized within mitochondria, whereas  
574 the former is cytoplasmic (Amengual et al., 2011; Lobo et al., 2012a and 2012b; Palczewski et  
575 al., 2014). The two subtypes have different substrate specificities with BCO1 showing a high  
576 specificity for  $\beta$ -carotene, and a more general requirement of a  $\beta$ -ionone ring (Lindqvist and  
577 Andersson, 2002), whereas BCO2 is able to cleave a relatively wide range of carotenoids notably  
578 including both hydroxylated and 4-oxo-substituted xanthophylls (i.e. lutein, zeaxanthin,  
579 canthaxanthin). In fact, several lines of evidence suggest that BCO2 preferentially cleaves  
580 carotenoids containing 3-hydroxy-ionone ring, and  $\beta$ -carotene has been shown to be a generally  
581 poor substrate for BCO2 (Mein et al., 2011; de la Seña, 2014 and 2016). To date, no study has  
582 directly evaluated either BCO subtype with respect to the xanthophyll glycosides, although

583 cleavage of myxoxanthophyll by both other CCE (e.g., ACO) from cyanobacteria has been  
584 demonstrated (Scherzinger et al., 2006; Scherzinger and Al Babili, 2008).

585       Alongside substrate specificity, we propose that stage dependent occurrence of BCO  
586 subtypes may underlie the observed temporal patterns of the teratogenicity of the carotenoids.  
587 The established mitochondrial localization of BCO2 (Lobo et al., 2012a and 2012b) might, in  
588 particular, imply that this subtype would be present during early stages of embryo development.  
589 Mature oocytes of the zebrafish contain remarkably high densities of maternally derived  
590 mitochondria. In a very recent study (Otten et al., 2016) the number of mitochondrial DNA  
591 (mtDNA) copies, for example, was found to be on the order of  $20 \times 10^6$  copies per cell compared  
592 to five orders of magnitude fewer, i.e., 100-1000 copies, in somatic cells. In the proposed model,  
593 therefore, teratogenicity of the xanthophylls would result from sequential cleavage: during early  
594 stages of development, mitochondrial BCO2 would cleave xanthophylls to corresponding  
595 apocarotenoids, and subsequent CCE (i.e., ACO or BCO) activity, along with possible  $\beta$ -  
596 oxidation, would convert the resulting apocarotenoids to teratogenic retinoids. This general  
597 mechanism has, in fact, been demonstrated recently in a mouse model (Amengual et al., 2013).  
598 On the other hand,  $\beta$ -carotene is a relatively poor substrate for BCO2 (de la Seña, 2014), and this  
599 may explain, therefore, very low teratogenicity of this carotenoid until later stages of  
600 development (i.e., 5 dpf) at which point BCO1 becomes more widely expressed in the zebrafish  
601 embryo (Lampert et al., 2003). Although BCO1 is expressed starting at 14 somite stage  
602 (Lampert et al., 2003), expression is localized to the yolk/embryo interface, and widespread  
603 expression (e.g. gut, liver) is only observed at ~3-4 dpf (Lampert et al, 2003). Also consistent  
604 with this model is the observation of very slight teratogenicity at 5 dpf for  $\beta$ -apo-8'-carotenal

605 which has been shown to be cleaved by BCO1, but with a much lower efficiency than  $\beta$ -carotene  
606 (de la Seña et al., 2013).

607 The most notable discrepancy in this model is the teratogenicity of lutein, but apparent lack  
608 of teratogenicity of its stereoisomer, zeaxanthin. It has been previously demonstrated in  
609 mammalian and avian systems that BCO2 is capable of eccentric cleavage of both zeaxanthin  
610 and lutein, to corresponding 3-hydroxy-apo-10'-carotenals which, in turn, are substrates for  
611 subsequent enzymatic or non-enzymatic (i.e.,  $\beta$ -oxidation) conversion to retinoids (Amengual et  
612 al., 2011; Mein et al., 2011; de la Seña, 2014). That said, studies generally suggest significant  
613 phylogenetic diversification of BCO subtypes which likely relates to differential utilization of  
614 carotenoids in different systems (e.g., fish, avian, mammalian). In particular, recent studies have  
615 specifically suggested an evolutionary divergence of BCO2 isoforms among teleost fish  
616 (Helgeland et al., 2014). As opposed to mammalian systems with two identified subtypes, studies  
617 in a teleost fish model (i.e., Atlantic salmon) – and subsequent evaluation of sequence databases  
618 - identified at least five isoforms, including 2 BCO1 and 3 BCO2, resulting from genome  
619 duplication events, and presumptive subsequent sub-functionalization, in the teleost evolutionary  
620 lineage (Helgeland et al., 2014). Interestingly, these authors specifically proposed a functional  
621 divergence of BCO subtypes among teleost fish related to the high diversity of carotenoids,  
622 particularly from algae and cyanobacteria, present in aquatic systems (Helgeland et al., 2014).  
623 Further characterization of CCE in the zebrafish model might, therefore, clarify the discrepancy  
624 between teratogenicity of lutein and zeaxanthin observed here. In this regard, it is particularly  
625 noteworthy that zeaxanthin, but not lutein, is produced by cyanobacteria and other algae.

626 Finally, in addition to the proposed role of differential substrate specificity of BCO sub-types,  
627 it is also likely that solubility of carotenoids contributes to the variable teratogenicity observed in

628 the present study. Carotenoids are generally lipophilic compounds. Of the carotenoids, the  
629 carotenes, such as  $\beta$ -carotene, are the most lipophilic (and thus least water soluble), whereas  
630 oxygen-containing xanthophylls (owing to the presence of hydroxyl and keto functional groups)  
631 are more polar and, although still lipophilic, relatively more soluble in the aqueous phase.  
632 Extending this trend to the xanthophyll glycosides, the presence of the sugar moiety would  
633 clearly even further increase hydrophilicity, and therefore, the potential availability of these  
634 compound (i.e. **1-3**) in the aqueous assay medium. Indeed, certain carotenoid glycosides, such as  
635 crocin, are the few known water-soluble variants (Háda et al., 2012). Consistent with this, **1-3**  
636 were, in fact, isolated in the most polar fractions from normal-phase chromatography (see 2.3.  
637 *Extraction and purification of carotenoids by bioassay-guided fractionation*). Increased water-  
638 solubility (and, thus, dissolved concentration) of the xanthophyll glycosides may, thereby,  
639 contribute to the particularly rapid and potent toxicity of **1** and **2**. With respect to the  
640 ecotoxicological implications, the increased water-solubility of the xanthophyll glycosides might  
641 similarly support a relatively larger contribution of these carotenoids in aquatic systems, and  
642 specifically as a component of the consequently *dissolved fraction* in these systems.

643

#### 644 4.3. Environmental relevance of the teratogenicity of carotenoids

645 Identification of the cyanobacterial xanthophyll glycosides (**1-3**) as teratogenic in the  
646 zebrafish model, and the teratogenicity of carotenoids more generally, suggest a potential impact  
647 of these compounds on aquatic animal populations including reduced recruitment and survival.  
648 However, relevance of the observed teratogenicity of carotenoids to possible effects on natural  
649 populations remains to be clarified with respect, in particular, to environmental concentrations.



650 Carotenoids associated with *particulate* (i.e., algal/cyanobacterial cell) fractions of aquatic  
651 systems are widely utilized to quantify both algal/cyanobacterial density generally, and as so-  
652 called “pigment markers,” specific taxa (e.g., cyanobacteria, chlorophytes, diatoms, etc.).  
653 Unfortunately, however, very little is known with respect to the concentration of carotenoids in  
654 the *dissolved* fraction of aquatic systems (as a direct route of exposure). Although carotenoids  
655 are generally considered lipophilic, it is hypothesized (as discussed above) that xanthophyll  
656 glycosides will have significantly higher water-solubility, and thus concentrations in dissolved  
657 fractions.

658 In the absence of specific data (i.e., measurements) regarding dissolved concentrations,  
659 potential contribution of carotenoids from cyanobacteria in aquatic systems can be generally  
660 estimated. Typical concentration of carotenoids, specifically including the myxol glycosides,  
661 have been conservatively estimated in the range of 0.1-0.5% dry weight (Montero et al., 2005;  
662 Schagerl and Müller, 2006). Density of cyanobacterial blooms are frequently cited, specifically  
663 using chlorophyll A (Chl A) as a proxy, to be as high as 3,000  $\mu\text{g Chl A L}^{-1}$  (Zohary and  
664 Roberts, 1990); and, using an accepted conversion factor of 0.5-1% Chl A per dry weight of  
665 biomass, can be translated to approximately 0.3 grams cyanobacterial biomass per liter.  
666 Contribution of carotenoids from cyanobacteria can, therefore, be calculated in the range of  
667 approximately 0.4 to 3  $\mu\text{M}$  (using a molar mass range of 537 to 747  $\text{g mol}^{-1}$  for  $\beta$ -carotene and  
668 myxol glycosides, respectively). Although such carotenoid concentrations might be limited to  
669 relatively dense bloom conditions, the calculated range would include concentrations  
670 approaching the lower levels (i.e., 1 and 5  $\mu\text{M}$ ) tested in the current study. While the specific  
671 contribution to carotenoids in the dissolved fraction remains to be seen, it is generally assumed to  
672 be affected by both degradation (or other forms of removal), and conversely, possible

673 accumulation within the dissolved fraction. Realistic concentrations may, therefore, be lower  
674 (due to degradation/removal) or higher (due to accumulation).

675 Finally, it should be noted that exposure concentrations in the present study represent  
676 *relatively high* levels of toxicity. Concentrations in the study were associated with toxicity  
677 specifically on the order of the EC<sub>50</sub>/LC<sub>50</sub> (i.e., 50% severe deformity/lethality) or greater. These  
678 concentrations would not be necessary for carotenoids to contribute to reduced populations.  
679 Otherwise stated, concentrations equivalent to, for instance, the EC<sub>10</sub>/LC<sub>10</sub> (which would reduce  
680 population viabilities by 10%) would be sufficient to have profound effects on the natural  
681 populations. Clearly further study is needed to understand the relevance of the carotenoids,  
682 particularly in relation to effective environmental concentrations.

683

## 684 **5. Conclusions**

685 The present study shows that xanthophyll glycosides, including myxol-2'-glycoside and its 4-  
686 keto and 3,4-dihydroxylated congeners, widely produced by cyanobacteria, can act as potent  
687 teratogens in the zebrafish embryo model. It is specifically hypothesized, based on these  
688 findings, that xanthophyll glycosides act as pro-retinoids similar to other carotenoids which are  
689 well studied in this regard. The comparative toxicology studies suggest that teratogenicity can be  
690 largely, but with exceptions (i.e., zeaxanthin), be potentially explained by the current knowledge  
691 of substrate specificity, and perhaps developmental patterns of the localization and expression, of  
692 the recognized sub-types of CCE associated with oxidative cleavage of carotenoids to retinoids  
693 in the zebrafish model.

694 The teratogenicity of carotenoids observed in the present study suggests a potentially direct  
695 link between cyanobacterial abundance and toxic potential in freshwater systems including, in

696 particular, effects on aquatic vertebrate populations (i.e., fish, amphibians). Unlike other  
697 recognized “cyanotoxins,” carotenoids are not phylogenetically restricted, and produced by all  
698 cyanobacteria which are obligate producers of carotenoids both as accessory pigments, and for  
699 protection against associated photooxidative stress. And, indeed, these algae are the primary  
700 autochthonous source of carotenoids in freshwater systems. Although cyanobacteria produce  
701 several carotenoids, the xanthophyll glycosides are generally recognized as unique to (and  
702 widely distributed within) the phylum (Schlüter et al., 2004). It has been suggested that  
703 carotenoids may be biosynthetic precursors of retinoids recently identified from cyanobacteria  
704 (Wu et al., 2012; Scherzinger et al., 2006). As such, carotenoids arguably represent a two-tiered  
705 source of teratogenic metabolites including both endogenous biosynthesis of retinoids by  
706 cyanobacteria, and *in situ* conversion of carotenoids to retinoids by vertebrate enzymes (i.e.  
707 BCOs, and other CCEs).

708         The toxicity of the carotenoids toward embryonic stages of the zebrafish, as an  
709 environmentally relevant (i.e., freshwater teleost) aquatic vertebrate model specifically  
710 underscores the potential impact of this toxicity to aquatic vertebrate populations. The presence  
711 of similarly teratogenic retinoids in cyanobacterial blooms has, in fact, been recently proposed as  
712 a factor in the local declines, as well as increased frequencies of developmental deformities,  
713 among aquatic vertebrate populations (Wu et al., 2012 and 2013). It is, likewise, possible that  
714 carotenoids may similarly contribute to such effects on freshwater ecosystems either directly  
715 and/or indirectly (as biosynthetic precursors to retinoids). Perhaps most notable, in this regard, is  
716 concurrence between unprecedented declines in amphibian populations worldwide, and  
717 temporally coincident (i.e., decadal) global increases in the proliferation of cyanobacteria in  
718 freshwater and coastal systems. As to the former, a preponderance of evidence points to global

719 decline in amphibian populations since the 1950s, and although several hypothesis have been put  
720 forth to explain this rapid disappearance, there remains no clear explanation (Houlahan et al.,  
721 2000; Blaustein and Kiesecker, 2002; Stuart et al., 2004). On the other hand, an emerging body  
722 of evidence suggests increased frequency and persistence of blue-green algae in freshwater and  
723 coastal systems within an equivalent time-frame. This increased cyanobacterial proliferation in  
724 aquatic is perhaps driven, as it has been argued (e.g., Paerl and Paul, 2012), by accelerated  
725 eutrophication (i.e. nutrient inputs), global climate change (i.e. warming) and various related  
726 factors. Although future study is clearly needed, the potential contribution of teratogenic  
727 carotenoids as a factor in these well documented declines in aquatic vertebrates is a compelling  
728 notion.

729

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740

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