

1 **Termination of the 2018 Florida Red Tide Event: A Tracer Model Perspective**

2
3 Yonggang Liu, Robert H. Weisberg, Lianyuan Zheng

4 College of Marine Science, University of South Florida, St. Petersburg, FL33701

5
6 Cynthia A. Heil

7 Mote Marine Laboratory, Sarasota, FL 34236

8
9 and

10 Katherine A. Hubbard

11 Florida Fish and Wildlife Conservation Commission- Fish and Wildlife Research Institute, St.

12 Petersburg, FL 33701

13
14
15
16 6 May 2022

17
18
19 Corresponding author contact: Yonggang Liu, yliu@usf.edu

20
21
22 Final published version:

23 Liu, Y., Weisberg, R.H., Zheng, L., Heil, C.A., Hubbard, K.A. (2022), [Termination of the 2018 Florida red](#)
24 [tide event: A tracer model perspective](#), *Estuarine, Coastal and Shelf Science*, 272, 107901,
25 <https://doi.org/10.1016/j.ecss.2022.107901>

27 **Abstract**

28 The 2018 *Karenia brevis* harmful algal bloom experienced along the west coast of
29 Florida was the worst red tide occurrence there since 2005. Cell concentrations peaked in early
30 fall of 2018, lessened in winter, and disappeared early in 2019. Here we examine the termination
31 of this red tide event by using hindcast simulations of the West Florida Coastal Ocean Model, a
32 numerical ocean circulation model that downscales from the deep Gulf of Mexico, across the
33 continental shelf and into the estuaries. The underlying hypothesis is that without an offshore
34 source of *K. brevis* cells, a nearshore bloom may quickly dissipate under the influence of a
35 persistent upwelling circulation. To test this hypothesis, we used a passive tracer (without
36 consideration of biological growth or decay) in the model to virtually indicate *K. brevis* cells.
37 The tracer, inputted along the central West Florida coast where highest bloom concentrations
38 were observed, was subsequently transported southward along the coast and offshore,
39 significantly reducing the tracer concentrations over the three-month-long experimental duration,
40 as was observed for the actual *K. brevis* cell concentrations. Whereas modeled tracer
41 concentrations decreased over most of the West Florida coast, relatively higher concentrations
42 remained just south of Sanibel Island, trapped there by the sharp bend in the coastline. Longer
43 residence time for this area has important *K. brevis* implications. Lake Okeechobee nutrient flux
44 through the Caloosahatchee River was thought to contribute to red tide in this region, and while
45 these inputs may be a factor, a persistent upwelling circulation may also play a contributing role.

46

47 **Keywords:** Red tide, West Florida Shelf, tracer model, coastal upwelling, harmful algal bloom,
48 bloom termination

49 **1. Introduction**

50 Blooms of the toxic dinoflagellate *Karenia brevis* (formerly *Gymnodinium breve*) occur
51 frequently on the west coast of Florida (Steidinger, 1975), killing fish and other marine life,
52 threatening public health and adversely impacting local economies throughout Florida (e.g.,
53 Anderson et al., 2021; Stumpf et al., 2022). Mitigating such effects requires improved forecast
54 capabilities and hence better understanding of the mechanisms of red tide initiation,
55 maintenance, and termination over the entire southwest Florida shelf region (e.g., Heil and
56 Steidinger, 2009).

57 *K. brevis* bloom studies have an extensive background, with some 24 hypotheses on
58 bloom initiation and growth reviewed by Vargo (2009). Many of the early thoughts of bloom
59 initiation focused on the nearshore region, ascribing blooms to terrestrial nutrient sources
60 associated with rainfall, runoff and ground water, with more recent examples including Hu et al.
61 (2006), Brand and Compton (2007) and Medina et al. (2020). The concept of offshore initiation
62 and shoreward advection were introduced by Steidinger (1975), and further documented by
63 Steidinger and Haddad (1981) and Tester and Steidinger (1997). Weisberg et al. (2009a)
64 provided a confirmation for such offshore initiation and shoreward advection for the 2005 *K.*
65 *brevis* bloom. Hypotheses of bloom growth and maintenance necessarily involve a suite of
66 complex biological and chemical processes to account for nutrient preferences and competition,
67 growth rates, species competition and mortality (e.g., Liu et al., 2001; Mulholland et al., 2006;
68 Vargo et al., 2008; O'Neil and Heil, 2014; Tilney et al., 2019). These disparate topics began to
69 merge synergistically once interdisciplinary studies of *K. brevis* ramped up around 1998. In a
70 seminal paper, Walsh et al. (2006) provided a process-oriented hypothesis of bloom initiation
71 and development, whereby slow growing *K. brevis* can outcompete faster growing

72 phytoplankton species in silicate deficient, oligotrophic waters by gaining nutrient support from
73 nitrogen fixing *Trichodesmium* (Sipler et al. 2013; Mulholland et al. 2014), whose growth is
74 facilitated by iron-rich Saharan dust (Lenes et al., 2001, 2008). Once the bloom reaches
75 sufficient concentration to dominate the phytoplankton community, it can utilize all available
76 nutrient sources as well as generating its own nutrient supply by utilizing *K. brevis* toxins to kill
77 fish (Walsh et al., 2009; Heil et al 2014).

78 Whether or not the offshore nutrient state is conducive for *K. brevis* outcompeting other
79 species may depend on the circulation. This hypothesis was demonstrated by the companion
80 papers of Weisberg and He (2003) and Walsh et al. (2003). The first of these, using observations
81 and a realistic numerical circulation model simulation for the spring and summer seasons of
82 1998, showed how the Gulf of Mexico Loop Current, when contacting the shelf slope near the
83 Dry Tortugas, set the entire West Florida Shelf (WFS) in an upwelling favorable motion,
84 resulting in the transport of deeper, nutrient-rich waters across the shelf break and shoreward
85 within the bottom Ekman layer. Using the velocity field from the first to drive an ecological
86 model, the second of these accounted for the observed nutrient (nitrate) and phytoplankton
87 (mostly diatoms) distributions offshore as observed in spring and summer 1998.

88 These concepts of how the WFS circulation may give rise to inter-annual variability in *K.*
89 *brevis* bloom intensity were further advanced by Weisberg et al. (2014a), who explained why
90 there was no *K. brevis* bloom in 2010 and by Weisberg et al. (2016b) who compared the
91 relatively pronounced, versus sedate *K. brevis* blooms of 2012 and 2013. In essence, the ocean
92 circulation physics were found to be as important for *K. brevis* ecology as the organism biology.
93 Expanding on these ideas, Liu et al. (2016), through a joint analysis of *K. brevis* cell counts and
94 Loop Current evolution (via satellite altimetry) developed a seasonal prediction scheme for

95 major red tide occurrences (or lack thereof) that was successful in 22 out of 25 years, including
96 the 2018 major red tide event (Weisberg et al., 2019).

97 Compared with *K. brevis* bloom initiation and development studies, there are fewer
98 studies regarding bloom termination. Biological processes potentially contributing to *K. brevis*
99 bloom termination studies that have been studied include grazing from macrozooplankton (e.g.,
100 Speekmann et al., 2006; Breier and Buskey, 2007), zooplankton (e.g., Dagg, 1995; Sutton et al.,
101 2001) and microzooplankton (e.g., Kubanek et al., 2007), lysis from bacterial and/or viral cells
102 (e.g., Paul et al., 2002; Brussaard, 2004; Mayali and Doucette, 2002; Lenes et al., 2013; Patin et
103 al., 2020) and nutrient depletion (Vargo, 2009). Recently, mechanisms of *K. brevis* intrinsic cell
104 loss/death processes have also been investigated with lab experiments (Gao and Erdner, 2022).
105 However, it remains unclear how these complex biological processes may play a role in
106 terminating *K. brevis* blooms on the WFS. Physical processes were suggested to be important for
107 bloom termination by Tester et al. (1991), and transport of *K. brevis* cells from the WFS to the
108 east Florida coast was documented by Tester and Steidinger (1997), Walsh et al. (2009) and
109 Harris et al. (2020). However, no specific studies examining the role of physical processes in
110 termination of blooms presently exist.

111 The 2018 *K. brevis* bloom was the most environmentally destructive harmful algal bloom
112 (HAB) on the west Florida coast in more than a decade. It started quite normally in late-
113 summer/fall of 2017, but then persisted through the winter and subsequent spring, with cell
114 concentrations at high levels (cell counts $> 10^6$ cells/L) southward from Venice, Florida through
115 January 2019. Based on the Liu et al. (2016) “pressure point” prediction scheme, the ocean
116 circulation conditions for 2018 were predicted to be conducive for further offshore development
117 of a new *K. brevis* bloom. The summer of 2018 saw a change in the Loop Current state from one

118 in which the Loop Current penetrated far into the Gulf of Mexico to one in which it retreated
119 back to a more direct inflow outflow state with “pressure point” contact. Whereas this state
120 change occurred too late to mitigate offshore *K. brevis* development in 2018, once the contact
121 occurred in July of 2018, it ensured that what was growing offshore would be transported to the
122 nearshore by the induced upwelling circulation. Consequently, the newly formed 2018 bloom
123 combined together with the existing overwintering 2017 bloom, setting the stage for a persistent
124 and spatially extensive *K. brevis* bloom that lasted through 2018 and into the beginning of 2019.
125 Thus, the coastal ocean circulation played an important role in accounting for the outbreak and
126 intensity of the 2018 bloom (Weisberg et al., 2019).

127 The present work continues the study of the 2017-2019 *K. brevis* bloom event with a
128 focus on the bloom termination. The same numerical model, the West Florida Coastal Ocean
129 Model (WFCOM, Zheng and Weisberg, 2012 and Weisberg et al, 2014b), will be employed
130 here. Different from the applications in Weisberg et al. (2019), which were based on Lagrangian
131 trajectory simulations, the WFCOM with an embedded tracer module is used here instead. The
132 purpose is to investigate how a passive tracer, originating nearshore, may be transported and
133 dissipated by advection and turbulent diffusion, independent of biological or chemical
134 interactions once the tracer source abates. In other words, it examines whether or not the
135 circulation can account for *K. brevis* bloom termination under certain conditions.

136 Note that this WFCOM application is built on top of a series of previous efforts in setting
137 up a proper numerical ocean circulation model for the WFS. A brief review of the relevant
138 modeling work is provided in Section 2, along with the WFCOM settings and the
139 implementation of the tracer model experiment. Model results are analyzed in Section 3,
140 followed by summary and discussion in Section 4.

141
142
143
144
145
146
147
148
149
150
151
152
153
154
155
156
157
158
159
160
161
162
163

2. Methods

2.1. A brief review of the West Florida Coastal Ocean Model

Modeling the WFS circulation requires the inclusion of both local (winds, heat and fresh water fluxes) and deep-ocean forcing (e.g., Weisberg and He, 2003; Weisberg and Liu, 2022). Thus, the WFCOM was designed to downscale from the deep ocean, across the continental shelf and into the estuaries (Zheng and Weisberg, 2012; Weisberg et al., 2014b). The unstructured grid, Finite Volume Community Ocean Model (FVCOM), developed by University of Massachusetts Dartmouth (Chen et al., 2003), was chosen to allow for increasing resolution upon approaching the coast and entering the various estuary inlets. This avoids multiple nesting procedure that is usually required to down scale from a low resolution regional ocean model to a high resolution estuary model. Also, the high resolution triangular grid can better resolve the complicated coastlines. FVCOM, through its many applications (e.g., Hu et al., 2008; Xia et al., 2020), is an established coastal ocean model. To include forcing from the deep ocean, WFCOM nests in the Hybrid Coordinate Ocean Model (HYCOM; e.g., Chassignet et al., 2009). The original application, documented by Zheng and Weisberg (2012), extended landward from about the 200 m isobath and covered a domain from the Desoto Canyon (Pensacola, FL) region in the northwest to just south of the Florida Keys. Recognizing the need to have real time Mississippi River inflows (versus climatology), WFCOM was subsequently modified to extent westward of the Atchafalaya Basin and to nest in the Gulf of Mexico (GOM) HYCOM (e.g., Zamudio and Hogan 2008; Halliwell et al., 2009), which includes tides. The domain of this present version is provided in Figure 1, and an initial application is given by Weisberg et al. (2014b). With its 65,435 nodes and 125,357 triangular elements, WFCOM presently has a horizontal resolution

164 that increases toward the coast from ~12 km at the open boundary to ~150 m in the estuaries.
165 Vertically, the model has 31 sigma layers with higher resolution near the surface and bottom to
166 better resolve the Ekman layers. Surface forcing includes momentum and heat fluxes
167 interpolated from the NOAA NCEP North American Mesoscale Forecast System (NAM) that are
168 served via NOAA Operational Model Archive and Distribution System (NOMADS). WFCOM is
169 one-way nested within the GOM HYCOM. Nesting includes sea surface elevation, 2D and 3D
170 velocity components, inclusive of tides, as well as 3D temperature and salinity, all within with a
171 buffer zone consisting of 10 nodes over which the weighting function varies linearly from 1 to
172 zero. River inflows are added as daily mean river discharges, downloaded from the US
173 Geological Survey (USGS) and South Florida Water Management District (SFWMD) web
174 servers. We also note that WFCOM includes a flooding and drying capability that is essential for
175 simulating scalar concentrations in shallow water that may either dry or result in coastal
176 inundation under strong wind conditions (e.g., Chen et al. 2008a, 2013).

177 Numerous applications over the past decade, each with model/observation comparisons,
178 demonstrate WFCOM's utility in simulating circulation and water property variations of
179 interdisciplinary concern. The first of these (Zheng and Weisberg, 2012) evaluated the
180 performance of the original WFCOM using in situ time series of velocity, temperature and
181 salinity from WFS moorings for calendar year 2007. WFCOM was found to reproduce the
182 circulation and water mass variations reasonably well over this year-long simulation. Spring,
183 2007 was particularly interesting because it contained a period of time when the Loop Current
184 was in contact with the southwest corner of the WFS, later termed the pressure point by Liu et al.
185 (2016). As hypothesized by Hetland et al. (1999) and confirmed by Weisberg and He (2003), the
186 Loop Current when in contact there may set the entire WFS in an upwelling favorable motion.

187 This 2007 event, when coupled with gag grouper juvenile observations, enabled Weisberg et al.
188 (2014b) to solve the gag grouper recruitment conundrum of how larvae, originating at shelf
189 break spawning sites, can arrive at their nearshore and estuarine settlement sites. Several
190 applications of WFCOM to the Deepwater Horizon oil spill event are also notable. Liu et al.
191 (2014) tested the WFCOM simulations using satellite-tracked surface drifter observations during
192 summer 2010 and found the WFCOM had better performance on the WFS than the other models.
193 The question of whether or not Deepwater Horizon hydrocarbons transited to the WFS to
194 account for reef fish lesions found there was addressed by Weisberg et al. (2016c) using a
195 simulated tracer deployed along the northern Gulf coast, where oil was observed to have covered
196 the surface. Weisberg et al. (2017) then used WFCOM to explain how Deepwater Horizon oil
197 arrived on northern Gulf beaches via combination of advection and wave-induced Stokes drift.
198 With regard to the coastal ocean circulation and storm surge response to hurricanes, Liu et al.
199 (2020) compared the WFCOM simulated current velocity with the moored observations at three
200 real-time data buoys on the WFS, and found that the model successfully reproduced the shelf
201 circulation response to Hurricane Irma. The comparisons between simulated and observed sea
202 levels at various coastal tide gauge stations along the west Florida coast were generally good,
203 capturing the initial set downs and subsequent rises of sea levels in Florida Bay and in the
204 estuaries as Hurricane Irma transited northward along the Florida peninsula. All these studies
205 demonstrate fidelity when using WFCOM for a variety of interdisciplinary coastal ocean
206 applications.

207 WFCOM applications to *K. brevis* HABs are also notable. Along with the oil spill, 2010
208 was a year when no bloom event was observed on the WFS. Weisberg et al. (2016a) explained
209 this on the basis of new inorganic nutrients having been upwelled onto the WFS by the persistent

210 upwelling that occurred beginning in May of that year. By abundant upwelled nutrients
211 supporting fast growing phytoplankton species offshore at depth at the expense of slower
212 growing *K. brevis*, we may account for the lack of a red tide. As in the prior studies mentioned
213 above, several time series from WFS moorings, along with data from a glider transect, provided
214 veracity testing. These concepts led to an Occam's Razor approach to *K. brevis* seasonal
215 forecasts (Liu et al., 2016) that was found to be correct for 17 out of 22 years for which joint
216 circulation (via satellite altimetry) and *K. brevis* cell count data were available. Weisberg et al.
217 (2016b) also used WFCOM simulations, moored velocity data and glider data to explain why the
218 2012 red tide was more pronounced than the nominal 2013 event. With the circulation now
219 recognized as being an important contributor to *K. brevis* HABs, and in collaboration with Fish
220 and Wildlife Research Institute (FWRI) colleagues, daily, automated red tide trajectory forecasts
221 are now served to the general public and agency personnel at:
222 http://ocgweb.marine.usf.edu/hab_tracking/. FWRI provides cell counts and WFCOM is used to
223 project these forward in time and space over a 4.5 day interval. Such short-term forecasts were
224 particularly useful to help guide clean-up operations throughout the 2018 *K. brevis* event, and
225 subsequent WFCOM hindcasts enabled us to explain why the event was so intense and why *K.*
226 *brevis* cells were also simultaneously found along the west coast of Florida, the Florida
227 Panhandle and Florida's east coast (Weisberg et al., 2019).

228

229 **2.2. Tracer model experiment**

230 FVCOM includes a dye module that simulates tracer concentration variations by
231 advection and diffusion. The tracer serves as a virtual marker of the water mass in the model
232 domain. It is transported from one place to another by advection and diluted through mixing with

233 ambient water. The dye module conserves tracer mass by using a flux calculation of second-
234 order accuracy to avoid numerical tracer loss (Chen et al., 2008b). Several applications exist in
235 the literature. As examples, Weisberg (2011) adopted it for simulating the flushing of a
236 residential channel under the action of tides and winds. Lai et al. (2013) used it in tracking the
237 radionuclide that flowed out of the Fukushima nuclear power plant. Zhu et al. (2015) applied it in
238 a study of the flushing processes in the Tampa Bay. Weisberg et al. (2016c) employed the virtual
239 tracer for examining the transport of the Deepwater Horizon hydrocarbons to the WFS. Here we
240 use the FVCOM tracer module to study the role of advection and diffusion in terminating the
241 2018 red tide event, independent of more complex, and largely unknown, biological processes.

242 The protracted 2017-2019 *K. brevis* red tide peaked from summer to fall 2018 with
243 sustained cell concentrations exceeding 10^6 cells per liter from September through November
244 (**Figure 2**). Highest concentrations were located along the central WFS coast between Tampa
245 Bay to Charlotte Harbor, attributed to the advection and concentration of new cells advected
246 there from offshore into this region. Lesser regions of *K. brevis* were also found along the
247 Florida Panhandle coast and the Florida east coast, also due to the transport by the ocean currents
248 (Weisberg et al., 2019). Given this evolution, what might happen once the supply of new cells
249 from offshore ceases? To address this question, we released a passive tracer in a coastal zone
250 from latitudes of Tampa Bay to Naples, Florida. The release was limited to a narrow band along
251 the 10 m isobath to the coast (**Figure 1**). Thus, all of the grid nodes within this area were
252 initialized uniformly throughout the water column (i.e., in all the 30 σ layers) with a normalized
253 concentration of 1, whereas all of the other nodes in the WFCOM domain were initialized with a
254 tracer concentration of zero. This initial tracer concentration started at 0000 UTC on 1 October

255 2018, and the tracer concentration experiment was run online as part of a realistic circulation
256 hindcast simulation for the four-month interval: 1 October – 31 December 2018.

257

258 **3. Results**

259 **3.1. Coastal upwelling circulation**

260 The circulation on the inner WFS is mostly wind-driven due to its wide and gently sloping
261 bottom and semi-enclosed coastline (**Figure 1**). The long-term mean circulation is upwelling, with
262 seasonal and synoptic weather band variations as determined from the multiple-year record of
263 moored velocity data (Weisberg et al., 2009b; Liu and Weisberg, 2005, 2007, 2012). The period
264 of October – December corresponds to the fall and winter seasons, and the coastal ocean circulation
265 is predominantly in upwelling patterns, i.e., the currents are generally directed down-shelf
266 (southeastward) with an offshore component in the surface currents and onshore component near
267 the bottom, resulting in an upwelling near the coast due to continuity of the water. This fall-winter
268 upwelling circulation tends to transport coastal materials in the down-shelf direction and spreads
269 them away from the coast near the surface.

270 To better illustrate the coastal upwelling current structures, we use the WFCOM output on
271 13 October 2018 as an example. The daily averaged currents are interpolated from the FVCOM
272 native unstructured grid to a coarser, rectangular mesh for better visualization, and the velocities
273 in the top-most and bottom-most sigma layers are shown as near surface and near bottom currents,
274 respectively (**Figure 3**). In both near surface and bottom layers, the currents were oriented in the
275 down-shelf direction over the inner WFS. The surface currents were southward or southwestward
276 (with an offshore component), with a stronger velocity (maximum 30 – 40 cm/s), while the near-
277 bottom currents were weaker (maximum 10 – 20 cm/s) and southeastward (with an onshore

278 component). From surface down to bottom, the current vector veered to the left over the inner shelf.
279 To further explore the vertical structure of the coastal currents, we show the daily averaged velocity
280 and salinity on the same day (13 October 2018) along a transect across the WFS offshore from
281 Sarasota, Florida. There was a strong coastal upwelling jet with a core (maximum along-shelf
282 velocity > 34 cm/s) at subsurface level between the 50 m and 60 m isobaths, which was
283 accompanied by fully developed bottom Ekman layer with the positive/onshore across-shelf
284 velocity component in the near bottom 10 – 15 m layer and a surface Ekman layer with the
285 negative/offshore across-shelf velocity component in the top 10 – 20 m layer (**Figure 4**). The
286 onshore velocity component extended from 130 km offshore all the way to the coast, with the
287 strongest onshore flow just below the core of the coastal upwelling jet. As a result, the near-bottom
288 water was upwelled onto the inner shelf and near the coast, which is evidenced by the nook of
289 salinity contours near the bottom (higher salinity water was upwelled and mixed with the coastal
290 fresher water). Similarly, surface water was advected away from the coast, which can also be seen
291 from the salinity contours that tend to lean towards offshore in the near surface layer except for
292 the very shallow water area (water depth < 10 m). The coastal upwelling circulation system
293 transported the coastal materials (the tracer or *K. brevis* cells here in this study) down the shelf
294 (southeastward) and spread them offshore. By mixing with deeper water column offshore, the
295 concentration of the materials decreased quickly when advected away from the coast.

296

297 **3.2. Tracer advection and dissipation**

298 Snapshots of WFCOM-simulated surface tracer concentration for the WFS coastal region
299 from Tampa Bay to the Florida Keys at various time intervals (three days, one week, two weeks,
300 and then from one to three months) after initialization are shown in **Figure 5**. With no new
301 tracer added after the first model time step, we see that the tracer concentration quickly decreases

302 nearshore. Thus, if the offshore source abates, then the same upwelling circulation that would
303 have brought new tracer to the shoreline now acts to rid the shoreline of what was initially there.
304 This is particularly evident at the mouths of the Tampa Bay and Charlotte Harbor estuaries
305 where strong tidal currents hasten the flushing of the tracer from these locations. The upper-left
306 hand panel shows that within three days of the tracer release (by 4 October 2018), the tracer
307 concentration is decreased by more than 50% near these two estuarine tidal inlets (**Figure 5a**).
308 Given the persistent upwelling circulation that existed from October to January, the tracer is
309 advected southward along the coast, with steamers extending offshore, resulting in a reduction of
310 the concentration with time. By the end of October 2018 (one month after the tracer release), the
311 normalized tracer concentration was reduced by 70% to 80% (from 1 to 0.2~0.3) for most areas
312 of the coastal region from Tampa Bay to Naples. By the end of November 2018 (two months
313 after the tracer release), the northern half of the original tracer patch was transported to the south
314 and further dissipated, and by the end of the third month (December 2018), the tracer all but
315 disappeared with the exception of the region between from Sanibel Island to the Florida Keys,
316 with much reduced concentration.

317 The lower panel of **Figure 5** shows time series of surface tracer concentration sampled at
318 five locations along the WFS, as indicated in the figure insert, i.e., from north to south, near
319 Clearwater, the Tampa Bay mouth, the Charlotte Harbor mouth, Sanibel Island, and Naples.
320 There was rapid reduction of the tracer concentration during the first several days near the
321 Charlotte Harbor inlet. Within about three weeks of the tracer release, the tracer concentration
322 offshore off Clearwater was reduced by 90%, and those near Naples and the mouths of Tampa
323 Bay and Charlotte Harbor were also largely decreased, by 60% to 80%. The slowest reduction

324 was located just south of Sanibel Island, where the normalized tracer concentration remained
325 above 0.4 until 40 days after the tracer release.

326 The Sanibel Island to the Ft. Meyers Beach area includes a sharp bend to the coastline
327 (**Figure 1**) that tends to trap tracer in its lee under the influence of an upwelling circulation.
328 Thus, the tracer tended to accumulate in this region and farther down coast to the Naples vicinity,
329 as shown by the enlargements of **Figure 6**. The region of highest tracer concentration was
330 located within the coastline nook, with a peak value of 0.6 and above even one month after
331 release (**Figure 6c**). Such relatively higher tracer concentration was in place on 30 November
332 2021, two months after the tracer release (**Figure 6d**). Some of the tracer even entered the
333 Charlotte Harbor estuary behind Sanibel Island, and was maintained in place there through 31
334 December 2018, more than three months after the release (**Figure 6e**). From this we may
335 surmise that abrupt changes in coastline configuration and orientation are important in
336 determining the distribution of HABs. This is consistent with findings by Picher et al. (2010)
337 who considered an array of coastline configurations, including headlands, capes, peninsulas,
338 bays and estuaries, representing regions of increasing isolation and consequently regions of
339 increased retention times for HABs.

340 Whereas the above discussion focused upon surface concentrations (in the top sigma
341 layer), given the shallowness of the nearshore, the near bottom concentrations (in the bottom
342 sigma layer) are very similar to the surface ones (the lower panels **Figures 6g through 6l**). Small
343 differences between the surface and bottom tracer concentration are mainly seen in the deeper
344 waters offshore of the 10 m isobath, where surface tracer was spread offshore and further diluted
345 relative to the bottom tracer that tended to hug the coast consistent with the upwelling
346 circulation.

347

348 **4. Discussion and Conclusions**

349 We examined the termination process of the 2017-2019 Florida *K. brevis* HAB event
350 using a passive tracer deployed in our West Florida Coastal Ocean Model, that was previously
351 shown to provide a realistic hindcast of the coastal ocean circulation for this event (Weisberg et
352 al., 2019). We hypothesized that the same persistent upwelling circulation that supports the
353 advection of an offshore source of red tide to the coast within the bottom Ekman layer, would act
354 to reduce red tide cell concentrations by advection and diffusion once the source of cells was
355 reduced, independent of biological processes or chemical interactions that are largely unknown
356 or lack definite relationships with *K. brevis* cell death/decay. To test this hypothesis, a band of
357 conservative tracer was released (and abruptly stopped) along the coast line between the shore
358 and the 10 m isobath. Without replenishment by new tracer from farther offshore, the tracer
359 concentration quickly decreased by advection and diffusion, particularly near Tampa Bay and
360 Charlotte Harbor estuary mouths where strong tidal currents add to the upwelling circulation.
361 Thus, under the upwelling conditions, the tracer was advected southward along the coast, with
362 steamers dissipating offshore, resulting in a reduction of the concentration over time.
363 Independent of biological growth or decay processes, the upwelling circulation resulted in a
364 substantial reduction in tracer concentration over the three-month period. This simplistic
365 experiment demonstrated that *K. brevis* blooms, initiated offshore and transported to the
366 nearshore along the bottom, can quickly manifest as high concentration blooms along the
367 coastline, but that once the offshore source of cells is cut off, then the circulation can quickly
368 cause bloom concentrations to decrease. Thus we may conclude that the same physical

369 oceanographic processes that play an important role in initiating a *K. brevis* bloom may also
370 account for bloom termination, as demonstrated for the 2017-2019 west Florida *K. brevis* HAB.

371 Additionally, it was found that due to the coastline geometry, an upwelling favorable
372 circulation may trap *K. brevis* cells (or other materials) within the coastline nook just south from
373 Sanibel Island. Whereas nutrient flux of terrestrial and upland origin (i.e., via the Caloosahatchee
374 River draining Lake Okeechobee) is often argued as being related to *K. brevis* in that area (e.g.,
375 Medina et al., 2020), the longer residence time of this region suggests that *K. brevis*
376 accumulation there may also be due to a combination of local hydrography and generation and
377 transport of cells from elsewhere. This interesting finding was obtained through the numerical
378 model experiment, a convenient way to examine the coastal ocean processes. Future in situ
379 observations are warranted to ascertain whether this offshore source is an important contributor
380 to the local red tide.

381

382 **Acknowledgments:**

383 Support was provided by the NOAA ECOHAB program (award NA19NOS4780183),
384 NOAA/IOOS through the Southeast Coastal Ocean Observing Regional Association
385 (SECOORA, award NA21NOS0120097), NOAA Office of Coast Survey through the Center for
386 Ocean Mapping and Innovative Technologies (COMIT, award NA20NOS4000227), the State of
387 Florida through FWC/FWRI (agreement # 20035) and the USF COMPS program. This work
388 benefited from the Extreme Science and Engineering Discovery Environment (XSEDE, award
389 OCE170007), which is supported by National Science Foundation grant ACI-1548562 (Towns et
390 al., 2014). This is NOAA ECOHAB Publication # 1019.

391

392 **References:**

393

394 Anderson, D.M., Fensin, E., Gobler, C.J., Hoeglund, A.E., Hubbard, K.A., Kulis, D.M.,
395 Landsberg, J.H., Lefebvre, K.A., Provoost, P., Richlen, M.L., Smith, J.L., Solow, A.R., Trainer,
396 V.L. (2021). Marine harmful algal blooms (HABs) in the United States: history, current status
397 and future trends. *Harmful Algae* 102, 101975.

398

399 Brand, L.E., & Compton, A. (2007). Long-term increase in *Karenia brevis* abundance along the
400 Southwest Florida Coast. *Harmful Algae* 6, 232-252.

401

402 Breier, C.F., & Buskey, E.J. (2007). Effects of the red tide dinoflagellate, *Karenia brevis*, on
403 grazing and fecundity in the copepod *Acartia tonsa*. *J. Plankton Res.* 29, 115–126.

404

405 Brussaard, C.P.D. (2004). Viral control of phytoplankton populations—a review. *J. Eukaryot.*
406 *Microbiol.* 51, 125–138.

407

408 Chassignet, E. P., Hurlburt, H. E., Metzger, E. J., Smedstad, O. M., Cummings, J., Halliwell, G.
409 R., et al. (2009). U.S. GODAE: Global Ocean Prediction with the HYbrid Coordinate Ocean
410 Model (HYCOM). *Oceanography* 22, 48–59.

411

412 Chen, C. S., Liu, H., & Beardsley, R. C. (2003). An unstructured, finite-volume, three-
413 dimensional, primitive equation ocean model: Application to coastal ocean and estuaries. *J.*
414 *Atmos. Oceanic Technol.* 20(1), 159–186.

415

416 Chen, C., Qi, J., Li, C., Beardsley, R.C., Lin, H., Walker R., & Gates, K. (2008a). Complexity of
417 the flooding/drying process in an estuarine tidal-creek salt-marsh system: an application of
418 FVCOM. *J. Geophys. Res.* 113, C07052, doi:10.1029/2007JC004328.

419

420 Chen, C.S., Xu, Q., Houghton, R., & Beardsley, R.C. (2008b). A model-dye comparison
421 experiment in the tidal mixing front zone on the southern flank of Georges Bank. *J. Geophys.*
422 *Res.* 113, C02005, doi:10.1029/2007jc004106.

423

424 Chen, C., R.C. Beardsley, R.A. Luettich Jr., et al. (2013). Extratropical storm inundation testbed:
425 intermodel comparisons in Scituate, Massachusetts. *J. Geophys. Res. Oceans* 118, 5054–5073,
426 doi:10.1002/jgrc.20397.

427

428 Dagg, M.J. (1995). Copepod grazing and the fate of phytoplankton in the northern Gulf of
429 Mexico. *Con. Shelf Res.* 15, 1303–1317.

430

431 Gao, Y., & Erdner, D.L. (2022). Dynamics of cell death across growth stages and the diel cycle
432 in the dinoflagellate *Karenia brevis*. *J. Eukaryot. Microbiol.* 69, e12874, doi:10.1111/jeu.12874.

433

434 Halliwell, G.R., Jr., A. Barth, R.H. Weisberg, et al. (2009). Impact of GODAE products on
435 nested HYCOM simulations on the west Florida shelf. *Ocean Dynamics* 59 (1), 139–155.

436
437 Harris R.J., Arrington D.A., Porter D., & Lovko V. (2020). Documenting the duration and
438 chlorophyll pigments of an allochthonous *Karenia brevis* bloom in the Loxahatchee River
439 Estuary (LRE), Florida. *Harmful Algae* 97, 101851, doi:10.1016/j.hal.2020.101851.
440
441 Heil, C.A., Dixon, L.K., Hall, E., Garrett, M., Lenes, J.M., O'Neil, J.M., Walsh, B.M., Bronk,
442 D.A., Killberg-Thoreson, L., Hitchcock, G.L., Meyer, K.A. (2014). Blooms of *Karenia brevis*
443 (Davis) G. Hansen & Ø. Moestrup on the West Florida Shelf: Nutrient sources and potential
444 management strategies based on a multi-year regional study. *Harmful Algae* 38, 127–140,
445 doi:10.1016/j.hal.2014.07.016.
446
447 Heil, C.A., Steidinger, K.A. (2009). Monitoring, management, and mitigation of *Karenia* blooms
448 in the eastern Gulf of Mexico. *Harmful Algae* 8, 611-617.
449
450 Hetland, R.D., Hsueh, Y., Leben, R., & Niller, P. (1999). A loop current induced jet along the
451 edge of the west Florida shelf. *Geophys. Res. Lett.* 26, 2239–2242.
452
453 Hu, C., Muller-Karger, F. E., & Swarzenski, P. W. (2006). Hurricanes, submarine groundwater
454 discharge, and Florida's red tides. *Geophys. Res. Lett.* 33, L11601, doi:10.1029/2005GL025449
455
456 Hu, S., Townsend, D.W., Chen, C., Cowles, G., Beardsley, R.C., Ji, R., Houghton, R.W.
457 (2008). Tidal pumping and nutrient fluxes on Georges Bank: A process-oriented modeling
458 study. *J. Mar. Sys.* 74, 528-544, doi:10.1016/j.jmarsys.2008.04.007.
459
460 Kubanek, J., Snell, T.W., & Pirkle, C. (2007). Chemical defense of the red tide dinoflagellate
461 *Karenia brevis* against rotifer grazing. *Limnol. Oceanogr.* 52, 1026–1035.
462
463 Lai, Z., Chen, C., Beardsley, R., Lin, H., Ji, R., Sasaki, J., & Lin J. (2013). Initial spread of ¹³⁷Cs
464 from the Fukushima Dai-ichi Nuclear Power Plant over the Japan continental shelf: a study using
465 high-resolution, global-coastal nested ocean model. *Biogeosciences* 10, 5439–5449.
466
467 Lenes, J.M., Darrow B.P., Cattrall, C., Heil, C.A., et al. (2001). Iron fertilization and the
468 *Trichodesmium* response on the west Florida shelf. *Limnol. Oceanogr.* 46, 1261–1277.
469
470 Lenes J.M., Darrow B.P., Walsh J.J., Prospero J.M., et al. (2008). Saharan dust and phosphatic
471 fidelity: a 3 dimensional biogeochemical model of *Trichodesmium* as a nutrient source for red
472 tides on the West Florida shelf. *Cont. Shelf Res.* 28, 1091–1115.
473
474 Lenes, J.M., Walsh, J.J., Darrow, B.P. (2013). Simulating cell death in the termination of
475 *Karenia brevis* blooms: implications for predicting aerosol toxicity vectors to humans. *Mar.*
476 *Ecol. Prog. Ser.* 493, 71–81. doi:10.3354/meps10515.
477
478 Liu, G., Janowitz, G.S., & Kamykowski, D. (2001). A biophysical model of population dynamics
479 of the autotrophic dinoflagellate *Gymnodinium breve*. *Mar. Ecol. Prog. Sers.* 210, 101-124.

480
481 Liu, Y., & Weisberg, R.H. (2005). Patterns of ocean current variability on the West Florida Shelf
482 using the self-organizing map. *J. Geophys. Res. Oceans* 110, C06003,
483 doi:10.1029/2004JC002786.
484
485 Liu, Y., & Weisberg R.H. (2007). Ocean currents and sea surface heights estimated across the
486 West Florida Shelf. *J. Phys. Oceanogr.* 37(6), 1697-1713, doi:10.1175/JPO3083.1.
487
488 Liu, Y., & Weisberg R.H. (2012). Seasonal variability on the West Florida Shelf. *Progr.*
489 *Oceanogr.* 104, 80-98, doi:10.1016/j.pocean.2012.06.001.
490
491 Liu, Y., Weisberg, R.H., Vignudelli, S., & Mitchum, G.T. (2014). Evaluation of altimetry-
492 derived surface current products using Lagrangian drifter trajectories in the eastern Gulf of
493 Mexico. *J. Geophys. Res. Oceans* 119: 2827–2842, doi:10.1002/2013JC009710.
494
495 Liu, Y., Weisberg, R.H., Lenes, J., Zheng, L., Hubbard, K., & Walsh, J.J. (2016). Offshore
496 forcing on the West Florida shelf “pressure point” and its upwelling influence on harmful algal
497 blooms. *J. Geophys. Res. Oceans* 121, 5501–5515, doi:10.1002/2016JC011938
498
499 Liu, Y., Weisberg, R.H., & Zheng, L. (2020). Impacts of hurricane Irma on the circulation and
500 transport in Florida Bay and the Charlotte Harbor estuary. *Estuaries and Coasts* 43, 1194-1216,
501 doi:10.1007/s12237-019-00647-6.
502
503 Mayali, X., & Doucette, G.J. (2002). Microbial community interactions and population dynamics
504 of an algicidal bacterium active against *Karenia brevis* (Dinophyceae). *Harmful Algae* 1, 277–
505 293.
506
507 Medina, M. Huffaker, R., Jawitz, J.W., & Muñoz-Carpena, R. (2020). Seasonal dynamics of
508 terrestrially sourced nitrogen influenced *Karenia brevis* blooms off Florida's southern Gulf
509 Coast. *Harmful Algae* 98, 101900, doi:10.1016/j.hal.2020.101900.
510
511 Mulholland, M. R., Bernhardt, P. W., Heil, C. A., Bronk, D.A., & Neil, J.M.O. (2006). Nitrogen
512 fixation and release of fixed nitrogen by *Trichodesmium spp.* in the Gulf of Mexico. *Limnol.*
513 *Oceanogr.* 51, 1762–1776. doi:10.4319/lo.2006.51.4.1762.
514
515 Mulholland, M. R., Bernhardt, P. W., Ozmon, I., Procise, L. A., Garrett, M., O’Neil, J. M., Heil,
516 C. A., Bronk, D. A. (2014). Contribution of diazotrophy to nitrogen inputs supporting *Karenia*
517 *brevia* blooms in the Gulf of Mexico. *Harmful Algae* 38, 20–29, doi:10.1016/j.hal.2014.04.004.
518
519 O’Neil, J. M., & Heil, C. A. (Eds.) (2014). Nutrient dynamics of *Karenia brevis* red tide blooms
520 in the eastern Gulf of Mexico. *Harmful Algae* 38, 1–140.
521
522 Patin, N.V., Brown, E., Chebli, G., Garfield, C., Kubanek, J., & Stewart, F.J. (2020). Microbial
523 and chemical dynamics of a toxic dinoflagellate bloom. *PeerJ* 8:e9493, doi:10.7717/peerj.9493.
524

525 Paul, J.P., Houchin, L., Griffin, D., Slifko, T., Guo, M., Richardson, B., & Steidinger, K. (2002).
526 A filterable lytic agent obtained from a red tide bloom that caused lysis of *Karenia brevis*
527 (*Gymnodinium breve*) cultures. *Aq. Microb. Ecol.* 27, 21–27.
528

529 Pitcher, G.C., Figueiras, F.G., Hickey, B.M., & Moita, M.T. (2010). The physical oceanography
530 of upwelling systems and the development of harmful algal blooms. *Prog Oceanogr.* 85(1-2), 5–
531 32, doi:10.1016/j.pocean.2010.02.002.
532

533 Sipler, R.E., Bronk, D.A., Seitzinger, S.P., Lauck, R.J., McGuinness, L.R., Kirkpatrick, G.J.,
534 Heil, C.A., Kerkhof, L.J. and Schofield, O.M. (2013). Trichodesmium-derived dissolved organic
535 matter is a source of nitrogen capable of supporting the growth of toxic red tide *Karenia*
536 *brevis*. *Marine ecology progress series*, 483, 31-45.
537

538 Speekmann, C.L., Hyatt, C.J., & Buskey, E.J. (2006). Effects of *Karenia brevis* diet on
539 RNA:DNA ratios and egg production of *Acartia tonsa*. *Harmful Algae* 5, 693–704.
540

541 Steidinger, K. A. (1975). Implications of dinoflagellate life cycles on initiation of *Gymnodinium*
542 *breve* red tides. *Environmental Letters* 9(2), 129–139, doi:10.1080/00139307509435842.
543

544 Steidinger, K.A., & Haddad, K. (1981). Biologic and hydrographic aspects of red tides.
545 *Bioscience* 31, 814-819.
546

547 Stumpf, R.P., Li, Y., Kirkpatrick, B., Litaker, R.W., Hubbard, K.A., Currier, R.D., et al. (2022).
548 Quantifying *Karenia brevis* bloom severity and respiratory irritation impact along the shoreline
549 of Southwest Florida. *PLoS ONE* 17(1): e0260755, doi:10.1371/journal.pone.0260755.
550

551 Sutton, T.T., Hopkins, T.L., Remsen, A.W., & Burghart, S.E. (2001). Multisensor sampling of
552 pelagic ecosystem variables in a coastal environment to estimate zooplankton grazing impact.
553 *Cont. Shelf Res.* 21, 69–87.
554

555 Tester, P.A., & Steidinger, K.A. (1997). *Gymnodinium breve* red tide blooms: Initiation,
556 transport, and consequences of surface circulation. *Limnol. Oceanogr.* 42, 1039-1051.
557

558 Tester, P.A., Stumpf, R.P., Vukovich, F.M., Fowler, P.K., & Turner, J.T. (1991). An expatriate
559 red tide bloom: Transport, distribution, and persistence. *Limnol. Oceanogr.* 36, 762–767.
560

561 Tilney, C.L., Shankar, S., Hubbard, K.A., Corcoran, A.A. (2019). Is *Karenia brevis* really a low-
562 light-adapted species? *Harmful Algae* 90, 101709, doi:10.1016/j.hal.2019.101709.
563

564 Towns, J., Cockerill, T., Dahan, M., Foster, I., Gaither, K., Grimshaw, A., et al. (2014). XSEDE:
565 Accelerating scientific discovery. *Computing in Science and Engineering* 16(5), 62–74,
566 doi:10.1109/mcse.2014.80.
567

568 Vargo, G.A. (2009). A brief summary of the physiology and ecology of *Karenia brevis* Davis (G.
569 Hansen and Moestrup comb. nov.) red tides on the West Florida Shelf and of hypotheses posed
570 for their initiation, growth, maintenance, and termination. *Harmful Algae* 8, 573-584.
571

572 Vargo, G. A., Heil, C. A., Fanning, K. A., Dixon, L. K., Neely, M. B., Lester, K., *et al.* (2008).
573 Nutrient availability in support of *Karenia brevis* blooms on the central West Florida Shelf:
574 What keeps *Karenia* blooming? *Cont. Shelf Res.* 28(1), 73–98, doi:10.1016/j.csr.2007.04.008.
575

576 Walsh, J. J., et al. (2003). The phytoplankton response to intrusions of slope water on the West
577 Florida shelf: Models and observations. *J. Geophys. Res.* 108(C6), 3190,
578 doi:10.1029/2002JC001406.
579

580 Walsh, J. J., Jolliff, J., Darrow, B. P., Lenes, J. M., Milroy, S. P., Remsen, D., *et al.* (2006). Red
581 tides in the Gulf of Mexico: Where, when, and why? *J. Geophys. Res.* 111, C11003,
582 doi:10.1029/2004JC002813.
583

584 Walsh, J.J, Weisberg, R.H., Lenes, J.M. Chen, F.R. Dieterle, D.A., Zheng, L., *et al.*
585 (2009). Isotopic evidence for dead fish maintenance of Florida red tides, with implications for
586 coastal fisheries over both source regions of the West Florida Shelf and within downstream
587 waters of the South Atlantic Bight. *Progr. Oceanogr.*, 80, 51-73.
588

589 Weisberg, R.H. (2011). Coastal ocean pollution, water quality and ecology: a commentary. *Mar.*
590 *Technol. Soc. J.* 45 (2), 35-42.
591

592 Weisberg, R.H., & He, R. (2003). Local and deep-ocean forcing contributions to anomalous
593 water properties on the West Florida Shelf. *J. Geophys. Res.* 108, C6, 15,
594 doi:10.1029/2002JC001407.
595

596 Weisberg, R.H., Barth, A., Alvera-Azcárate, A., & Zheng, L. (2009a). A coordinated coastal
597 ocean observing and modeling system for the West Florida Shelf. *Harmful Algae* 8, 585-598.
598

599 Weisberg, R.H., Liu, Y., & Mayer, D.A. (2009b). Mean circulation on the west Florida
600 continental shelf observed with long-term moorings. *Geophys. Res. Lett.* 36, L19610,
601 doi:10.1029/2009GL040028.
602

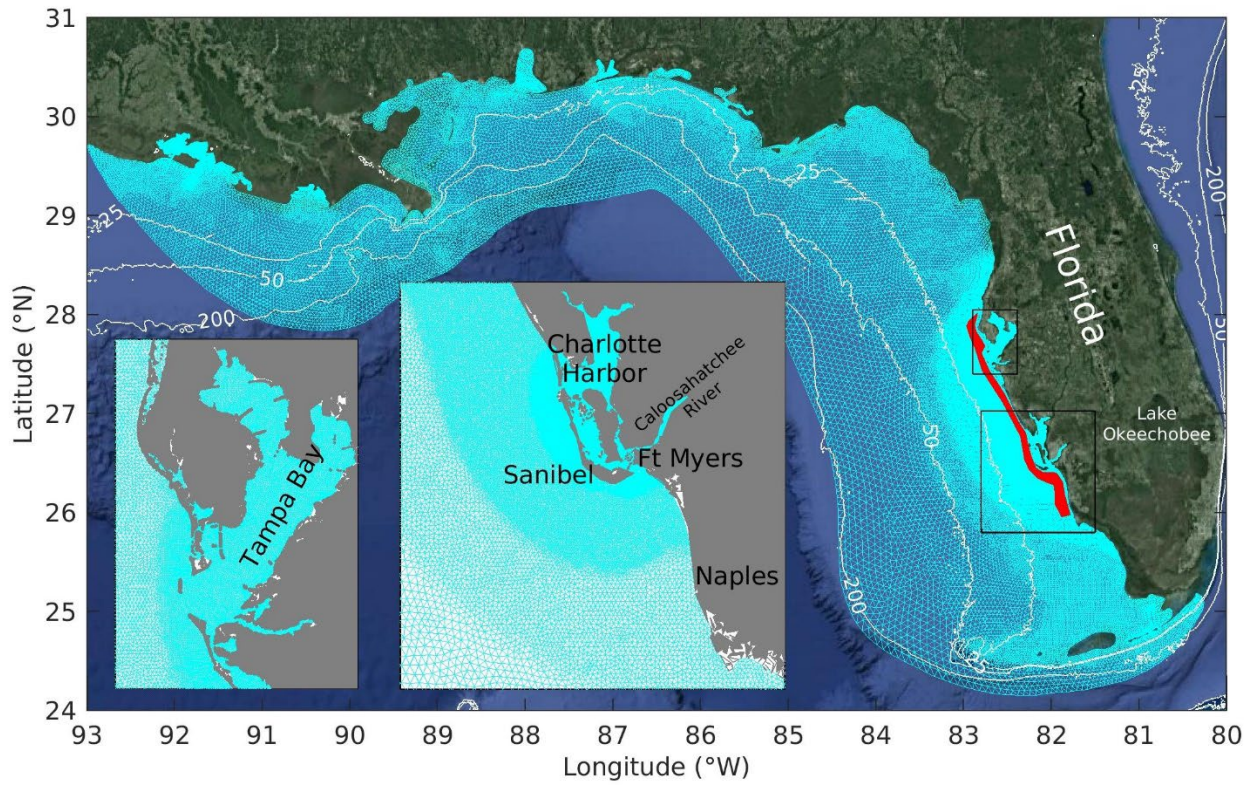
603 Weisberg, R.H., Zheng, L., Liu, Y., Lembke, C., Lenes, J. M., & Walsh, J.J. (2014a). Why a red
604 tide was not observed on the West Florida continental shelf in 2010. *Harmful Algae* 38, 119–126,
605 doi:10.1016/j.hal.2014.04.010.
606

607 Weisberg, R.H., L. Zheng, L. & Peebles, E. (2014b). Gag grouper larvae pathways on the West
608 Florida Shelf, *Cont. Shelf Res.*, 88, 11-23, Doi:10.1016/j.csr.2014.06.003.
609

610 Weisberg, R.H., Zheng, L., & Liu, Y. (2016a). West Florida Shelf upwelling: Origins and
611 pathways. *J. Geophys. Res. Oceans* 121, 5672-5681, doi:10.1002/2015JC011384.

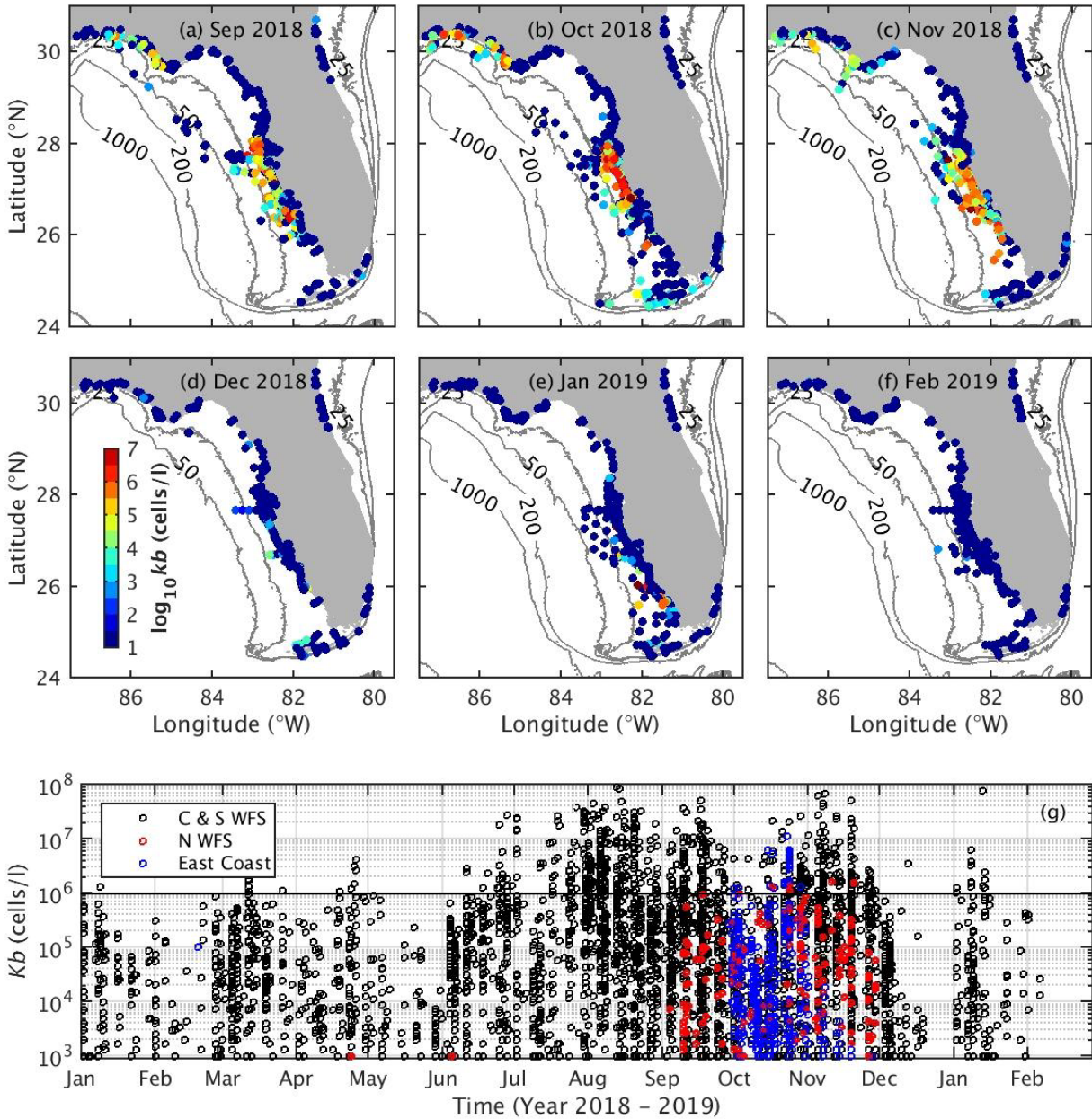
612
613 Weisberg, R. H., Zheng, L., Liu, Y., Corcoran, A., Lembke, C., Hu, C., *et al.* (2016b). *Karenia*
614 *brevis* blooms on the West Florida shelf: A comparative study of the robust 2012 bloom and the
615 nearly null 2013 event. *Cont. Shelf Res.* 120, 106–121, doi:10.1016/j.csr.2016.03.011.
616
617 Weisberg, R.H., Zheng, L., Liu, Y., Murawski, S., Hu, C., & Paul, J. (2016c). Did Deepwater
618 Horizon hydrocarbons transit to the west Florida continental shelf? *Deep Sea Res. II*, 129, 259-
619 272, doi:10.1016/j.dsr2.2014.02.002.
620
621 Weisberg, R.H., Zheng, L., & Liu, Y. (2017). On the movement of Deepwater Horizon Oil to
622 northern Gulf beaches. *Ocean Modell.*, 111, 81-97, doi:10.1016/j.ocemod.2017.02.002.
623
624 Weisberg, R.H., Liu, Y., Lembke, C., Hu, C., Hubbard, K., Garrett, M. (2019). The coastal ocean
625 circulation influence on the 2018 West Florida Shelf *K. brevis* red tide bloom. *J. Geophys. Res.*
626 *Oceans* 124, 2501-2512, doi:10.1029/2018JC014887.
627
628 Weisberg, R.H., & Liu, Y. (2022). Local and deep-ocean forcing effects on the West Florida
629 Continental Shelf circulation and ecology. *Frontiers in Marine Science*,
630 doi:10.3389/fmars.2022.863227.
631
632 Xia, M., Mao, M., & Niu, Q. (2020). Implementation and comparison of the recent three-
633 dimensional radiation stress theory and vortex-force formalism in an unstructured-grid coastal
634 circulation model. *Est. Coast & Shelf Sci.* 240, 106771, doi:10.1016/j.ecss.2020.106771.
635
636 Zamudio, L., & Hogan, P.J. (2008). Nesting the Gulf of Mexico in Atlantic HYCOM:
637 oceanographic processes generated by Hurricane Ivan. *Ocean Modelling* 21, 106–125.
638
639 Zheng, L., & Weisberg, R. H. (2012). Modeling the West Florida Coastal Ocean by downscaling
640 from the Deep Ocean, across the continental shelf and into the estuaries. *Ocean Modell.* 48, 10–
641 29, doi:10.1016/j.ocemod.2012.02.002.
642
643 Zhu, J., Weisberg, R.H., Zheng, L., & Han, S. (2015). On the flushing of Tampa Bay. *Estuaries*
644 *and Coasts* 38, 118-131, doi:10.1007/s12237-014-9793-6.
645

646



647

648 **Figure 1.** West Florida Coastal Ocean Model (WFCOM) domain and grid mesh (cyan). Both
649 Tampa Bay and Charlotte Harbor regions are zoomed in as inserted maps. Also shown are tracer
650 initiation area (red) and bathymetry contours 25, 50 and 200 m.



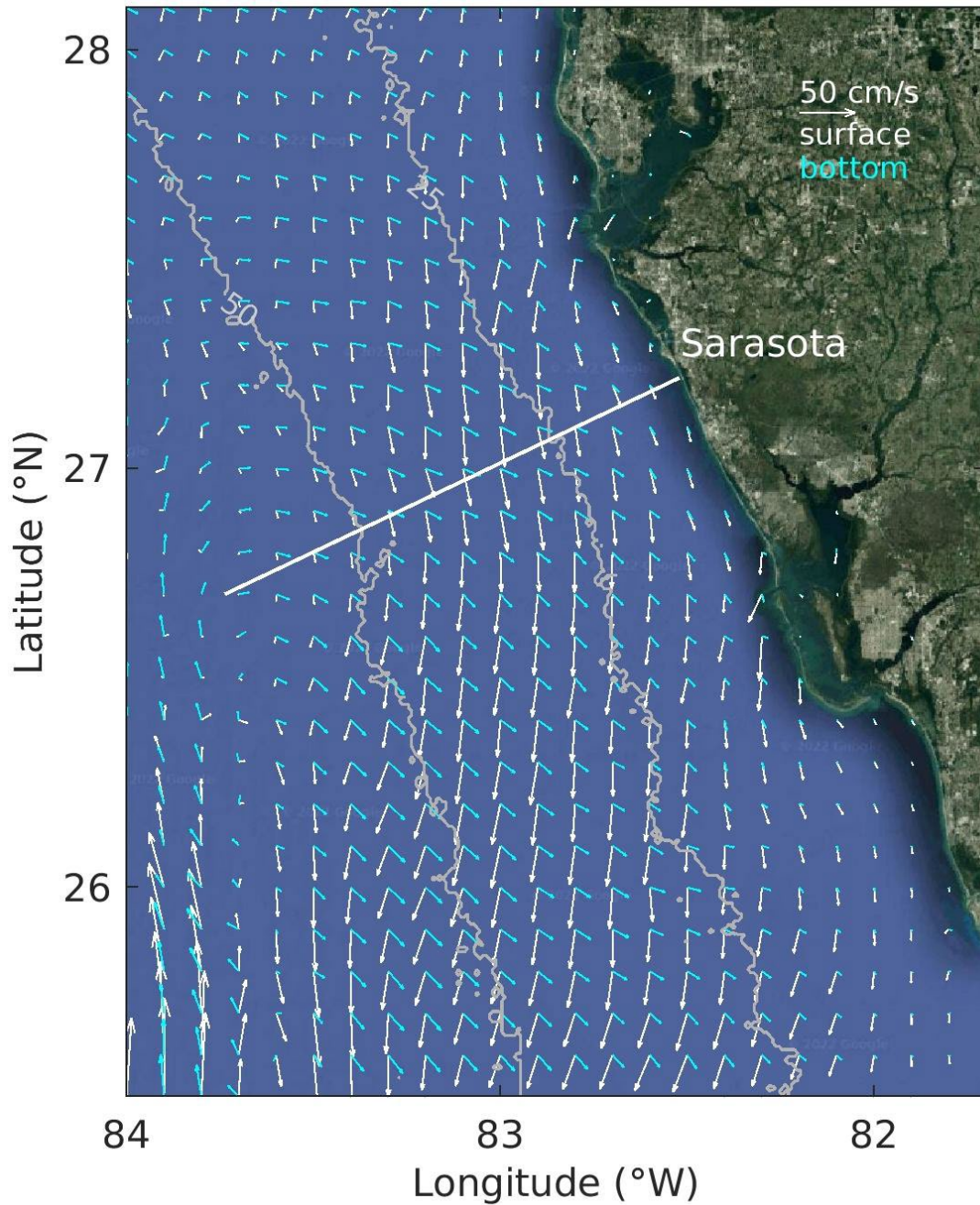
651

652

653 **Figure 2.** Monthly maps of *Karenia brevis* observations around Florida coast during September
 654 2018 – February 2019 (a – f). The *K. brevis* cell counts (cells/liter) are shown as log₁₀ (color
 655 coded). Bathymetry units in m. Time series of the *K. brevis* bloom conditions (g) in the central
 656 and southern West Florida Shelf (WFS) coastal region (C & S WFS), the Florida Panhandle (N
 657 WFS) and on the east coast.

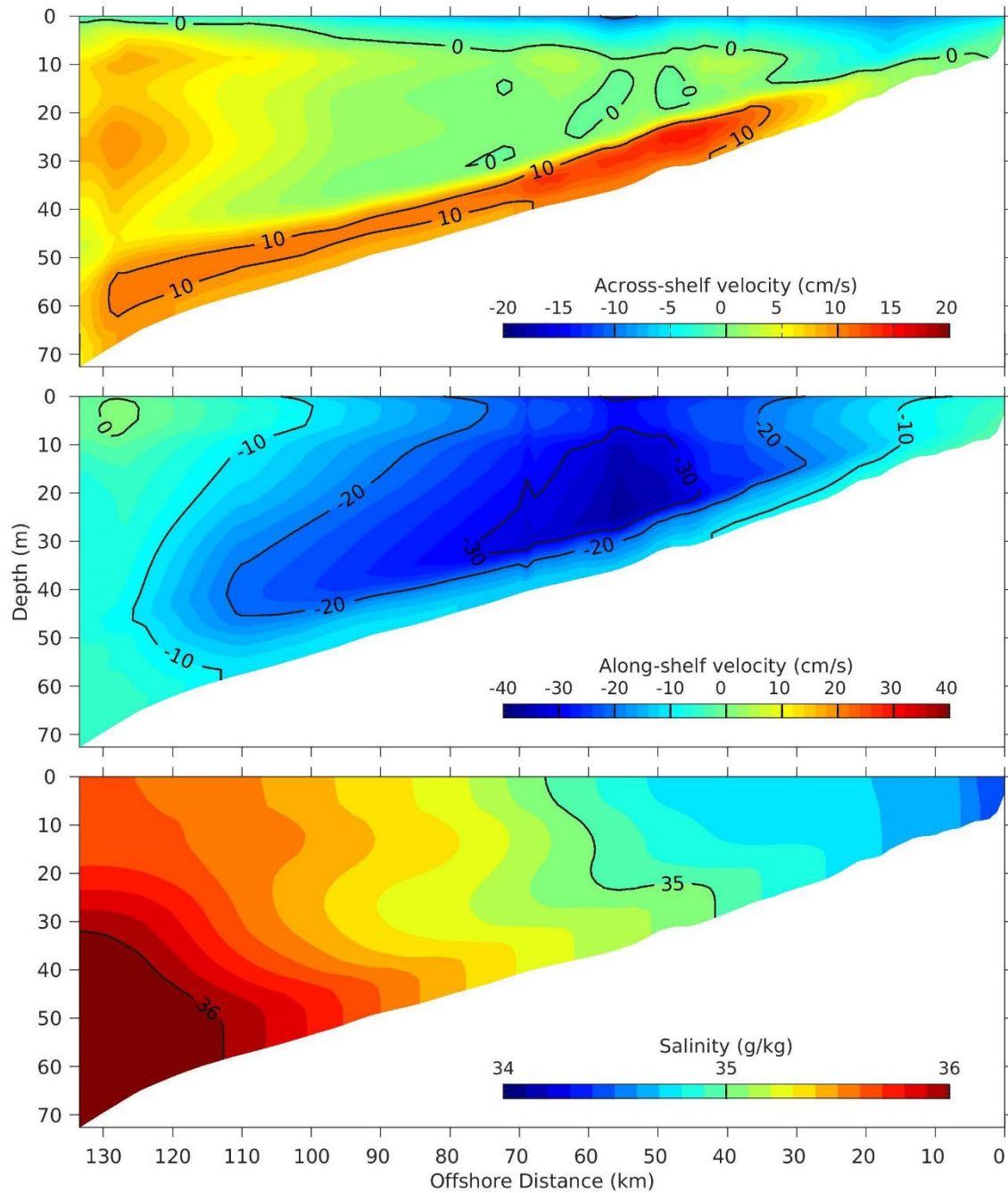
658

659



660

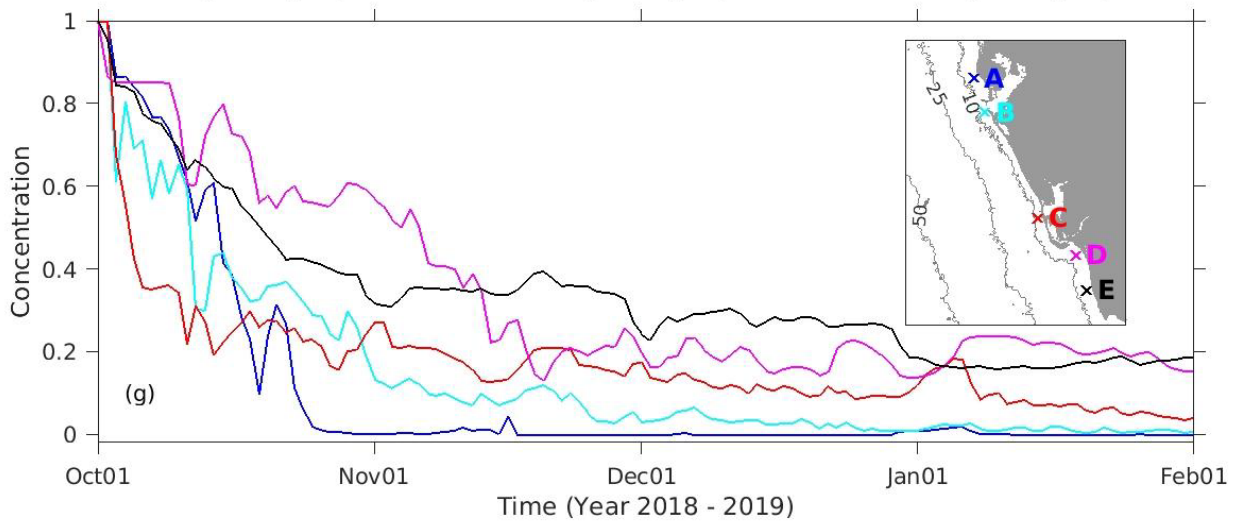
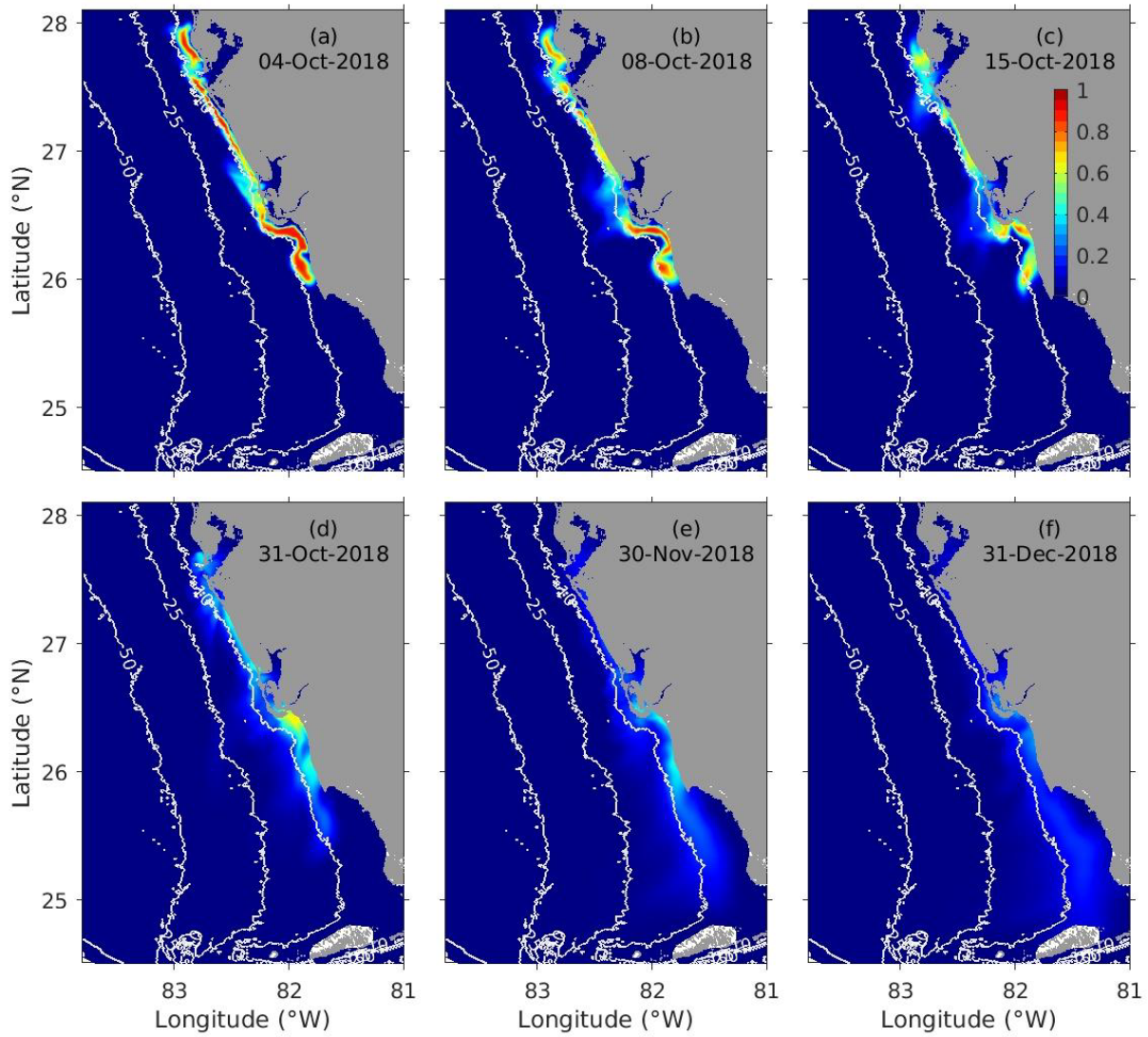
661 Figure 3. An example of coastal upwelling circulation pattern on the West Florida Shelf as
 662 simulated by the WFCOM for 13 October 2018. Daily averaged near surface and bottom currents
 663 are shown as white and cyan vectors, respectively. The current velocities are interpolated to a
 664 coarse rectangular grid for better visualization. The straight line shows an across-shelf transect
 665 offshore from Sarasota, Florida. Also shown are 25 m and 50 m isobaths.



666

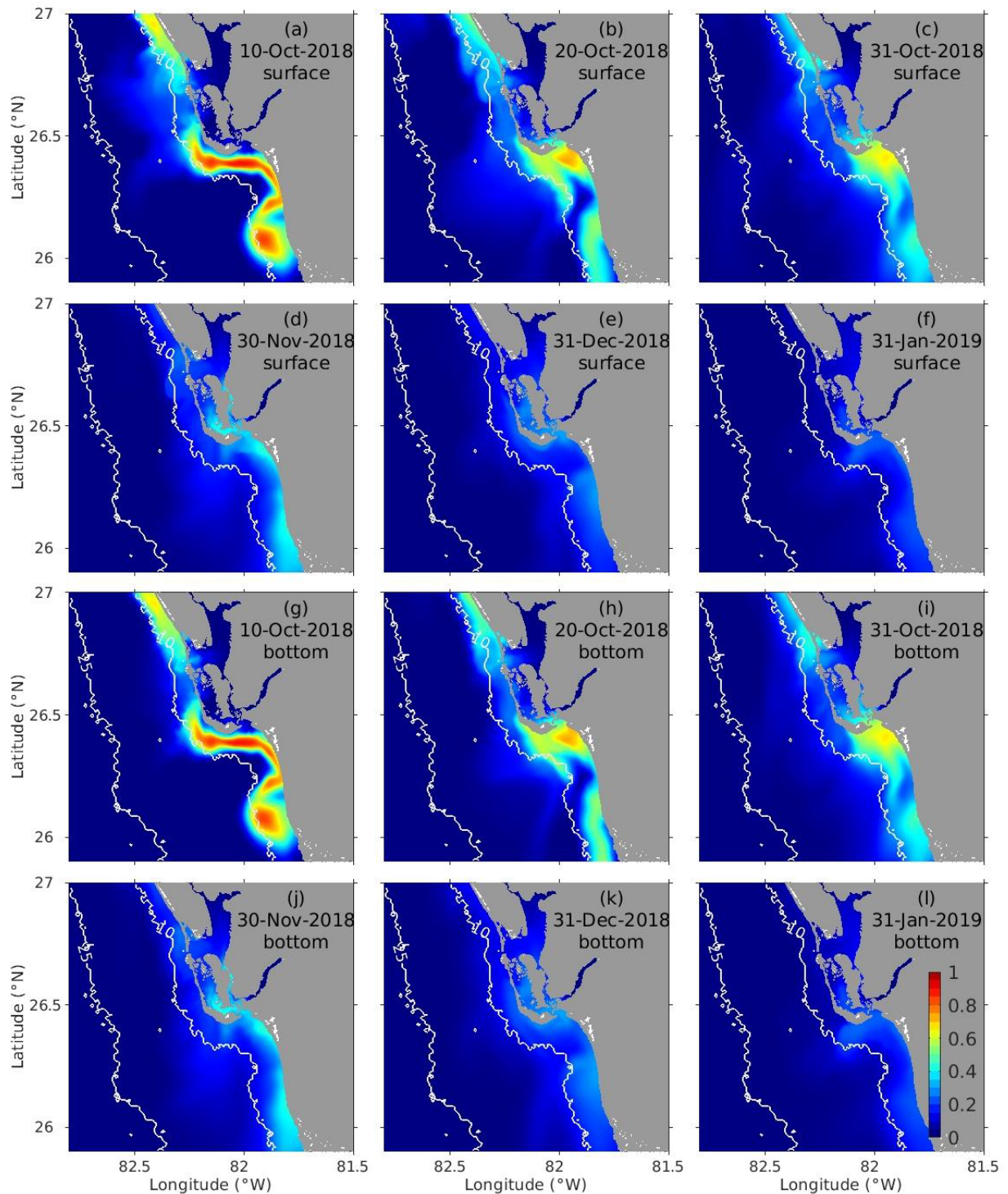
667 Figure 4. Across-shelf distribution of the currents and salinity in an upwelling event as simulated
 668 by WFCOM. From top to bottom, daily averaged across-shelf current velocity component (+
 669 onshore, - offshore) and along-shelf current velocity component (+ northwestward, -
 670 southeastward) and salinity along the Sarasota transect on 13 October 2018. Note that the across-
 671 shelf velocity component shows both a bottom Ekman layer and a surface Ekman layer (top panel),
 672 and along-shelf current shows a strong coastal upwelling jet with a core at subsurface (middle
 673 panel).

674



676 **Figure 5.** Snapshots of model simulated surface tracer concentration for the central – south
677 West Florida Shelf. Snapshots of surface tracer concentration three days after release (a), one and
678 two weeks after release (b and c), and one, two and three months after release (d – f). Time series
679 of the tracer concentration (g) sampled at five locations along the coast, from north to south, as
680 shown in an inserted map: Clearwater coast (A), Tampa Bay mouth (B), Charlotte Harbor mouth
681 (C), Sanibel Island (D), and Naples coast (E). Bathymetry units in m.

682



684

685 **Figure 6.** Snapshots of model simulated surface (top 6 panels, a – f) and bottom (bottom 6
 686 panels, g – l) tracer concentration for Sanibel Island – Ft. Myers coastal region on 10, 20 and 31
 687 October 2018, 30 November 2018, 31 December 2018, and 31 January 2019. Bathymetry units
 688 in m.