1 Biogeography of reef water microbes from within reef to global scales

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12 Abstract

Seawater microorganisms play an important role in coral reef ecosystem functioning and can be influenced by biological, chemical, and physical features of reefs. As coral reefs continue to respond to environmental changes, the reef seawater microbiome has been proposed as a conservation tool for monitoring perturbations. However, the spatial variability of reef seawater microbial communities is not well studied, limiting our ability to make generalizable inferences across reefs. In order to better understand how

19 microorganisms are distributed at multiple spatial scales, we examined seawater microbial 20 communities in Florida Reef Tract and U.S. Virgin Islands reef systems using a nested 21 sampling design. On three reefs per reef system, we sampled seawater at regular spatial 22 intervals close to the benthos. We assessed the microbial community composition of these 23 waters using ribosomal RNA gene amplicon sequencing. Our analysis revealed that reef water microbial communities varied as a function of reef system and individual reefs, but 24 25 communities did not differ within reefs and were not significantly influenced by benthic 26 composition. For the reef system and inter-reef differences, abundant microbial taxa were 27 found to be potentially useful indicators of environmental difference due to their high 28 prevalence and variance. We further examined reef water microbial biogeography on a 29 global scale using a secondary analysis of five studies, which revealed that microbial communities were more distinct with increasing geographic distance. These results suggest 30 31 that biogeography is a distinguishing feature for reef water microbiomes, and that 32 development of monitoring criteria may necessitate regionally-specific sampling and 33 analyses.

35 **1 Introduction**

Coral reefs are currently experiencing significant challenges due to global and local factors 36 37 (Hughes et al. 2017). Among them, climate change and ocean acidification affect corals 38 worldwide while stressors such as human impacts and disease outbreaks are more 39 localized. These crises are driving the development of new management and conservation 40 strategies to preserve and monitor reef biodiversity. Awareness of the coral as a holobiont an assemblage of a host and all of its associated symbiotic microorganisms (Knowlton & 41 42 Rohwer 2003, Rosenberg et al. 2007) - has spurred research into establishing microbial 43 solutions to reef stress, such as coral probiotics and microbiome-based monitoring (Peixoto et al. 2017, Glasl et al. 2017). In particular, a holistic characterization of microbes 44 in coral reefs will aid in predicting reef resilience and environmental threats (Kelly et al. 45 2018). 46

47 Reefs harbor many distinct niches for bacterial and archaeal communities, including corals, sponges, sediments, and the water-column itself (Tout et al. 2014, McDevitt-Irwin et al. 48 49 2017). Free living water column microbes, residing above the reef substrate, are influenced 50 by hydrological conditions (Sweet et al. 2010, Becker et al. 2020), general benthic 51 community composition (Haas et al. 2011, Kelly et al. 2014), local nutrient regimes (van 52 Duyl & Gast 2001, Nelson et al. 2011), and temporal dynamics (Weber & Apprill 2020, 53 Becker et al. 2020). When combined, these influences cause reef-associated seawater 54 microbiomes to be readily distinguishable between reefs, as well as between zones within a 55 reef (Jeffries et al. 2015, Salerno et al. 2016, Frade et al. 2020). Microbial communities can 56 be powerful indicators of reef health and environmental conditions (Glasl et al. 2017).

57 Indeed, as reefs transition from coral to algae dominated, the exudates released from the benthos also likely shift, causing increased heterotrophy and decreased oligotrophy in the 58 59 seawater microbiome (Haas et al. 2011, Nelson et al. 2013). In a proccess called 60 microbialization, the heterotroph-dominated microbial community further depresses 61 growth of coral and encourages the growth of algae (Haas et al. 2016, Kelly et al. 2018). 62 This microbial phase shift may be an important process to monitor in at-risk reefs. 63 Additionally, reef microorganisms respond rapidly to nutrient and temperature 64 fluctuations, potentially providing a sensitive and non-invasive diagnostic or predictive 65 tool for perturbations that may provide knowledge prior to visible reef changes (Glasl et al. 2019. Becker et al. 2020). 66

67 Implementation of large scale reef water monitoring efforts for reef microorganisms is 68 partially limited by our understanding of reef seawater microbial diversity across spatial 69 scales (Bourne et al. 2016, Glasl et al. 2017). Biogeographic patterns of coral reef microbial 70 assemblages have been found at a variety of spatial scales. Small scale patterns such as 71 within a coral skeleton (Marcelino et al. 2018), in the boundary layer overlying the coral 72 mucus (Weber et al. 2019), and in micro-habitats generated by coral structures (Schöttner 73 et al. 2012) highlight potential mechanisms affecting reef microbial composition, but may 74 not represent the state of an entire reef. On the other hand, studies and models of marine microbial distribution at the scale of oceans (Amend et al. 2012, Hellweger et al. 2014) 75 76 provide insight into the global drivers of microbial abundance, but are not specific to the 77 unique environments of reefs. Therefore, a better understanding of the biogeography of 78 coral reef seawater microbes across distinct spatial scales is warranted.

79 The goal of this study is to understand the variability of coral reef seawater microbial 80 communities across different spatial scales. We examined this question in two parts. For 81 the first part, we examined reef water microbial communities within and between two reef 82 systems to understand the influence of both reef and reef benthic composition on microbial 83 diversity (Figure 1A). Secondly, in order to quantify the impact of larger geographic 84 distances on reef water microbial communities, we conducted a secondary analysis of 85 aggregated 16S rRNA gene sequences from five studies that used similar sampling methodology (Figure 1B, Table S1). We predicted that microbial community structure 86 87 would differ primarily on the scale of individual reefs but and secondarily on underlying 88 benthic structure. Additionally, we expected the secondary analysis to recapitulate the individual reef and reef system-based biogeographic patterns seen in the Fl and VI systems 89 90 on a more global scale.

91 2 Materials and Methods

92 2.1 FL and VI transects: Sampling

93 The first part of this study took place in two reef systems, the Florida Reef Tract (Fl) in June 94 2019 and off the southern coast of St. Thomas in the U.S. Virgin Islands (VI) in February 95 2020. A total of three reefs were sampled in the Florida Keys; the northernmost reef was 96 Biscayne, located within the boundaries of Biscayne National Park, the reef Grecian was 97 located at the Grecian Rocks reef off the coast of Key Largo, and the reef Dry Tortugas was 98 located within Dry Tortugas National Park. All reefs in the Florida Reef Tract (Fl) were 99 forereefs within the barrier reef. Similarly, three reefs were sampled in St. Thomas (VI);

Brewers Bay and Black Point were forereef zones on fringing reefs a few hundred meters from the coast while Flat Cay was a fringing reef located near an uninhabited island named Flat Cay about two kilometers off the coast (Table 1). Average sampling depth was between 5.0 and 7.1 meters, with the exception of the Dry Tortugas, which was deeper with an average depth of 18.0 meters. Due to the difference in season of sampling, the average temperature in the Florida reefs was slightly higher than in the Virgin Islands reefs (28.7 °C and 26.9 °C respectively).

107 At each reef, three 10 meter transects were taken by laying down a 10 m weighted line that 108 was marked every meter. Water samples were taken by a diver using a 60 or 100 mL 109 syringe positioned approximately 5 cm above the benthos at each meter line. The transects 110 were laid haphazardly, but did not intersect with each other. Because of inclement 111 conditions, only 1 transect was collected at the Biscavne reef. At the Fl reefs, benthic 112 composition - represented by percent cover of coral skeleton, crustose coralline algae, 113 cyanobacteria, hard coral, macroalgae, non biological, other invertebrates, soft coral, 114 sponge, and turf algae - was determined using large-area imagery collected from 10 by 10 115 m area plots. All transects were placed within these 100 m² plots. Stratified random points 116 (2500) were dropped across the reef area and classified to generate reef-wide cover 117 estimates - see full methods in Fox et al. (2019). At the VI reefs, benthic composition was 118 recorded at the precise location of each syringe sample using a video survey of the transect 119 line as well as noted in writing by a diver during sampling. Video and written record were 120 cross-referenced and each sample was then classified into a single category from algae. 121 dead coral, live coral, rock, sand, sponge, and undetermined.

To capture the seawater microbial community, 60 mL of the seawater was filtered through
a 0.22 µm Supor filter (25 mm; Pall Corporation). The water volume of 60 mL has
previously been found to be comparable to larger volumes (1-2L) for characterizing
seawater microbial communities using amplicon sequencing (Weber et al. 2019). Filters
were placed in 2 mL cryovials, flash frozen in a liquid nitrogen dry shipper, and processed
upon returning to Woods Hole, MA.

128 2.2 FL and VI transects: DNA extraction, PCR amplification, and sequencing

129 DNA was extracted from the filters using the DNeasy PowerBiofilm Kit (Qiagen) according 130 to manufacturer protocols. Seven DNA extraction controls, consisting of unused 0.22 µm 131 filters, were processed alongside samples. Extracted DNA was quantified using the Qubit 132 2.0 fluorometer HS dsDNA assay (ThermoFisher Scientific). Primers 515FY (Parada et al. 133 2016) and 806RB (Apprill et al. 2015) containing Illumina overhang adapter sequences 134 were used to amplify the V4 region of the small subunit rRNA gene in bacteria and archaea. 135 PCR reactions contained 14.75 µL molecular grade water, 5 µL GoTag Flexi 5X buffer 136 (Promega Corporation), 2.5 µL of 25 mM MgCl₂, 1 µL of 10 mM dNTPs, 1 µL of 10 mM 137 forward and reverse primers, 0.5 µL GoTaq DNA polymerase (Promega) and 1 µL of DNA 138 template. Three PCR controls consisting of 1 µL of PCR-grade water as template were also 139 included, as well as microbial genomic DNA from a Human Microbiome Project mock 140 community (BEI Resources, NIAID, NIH as part of the Human Microbiome Project: Genomic 141 DNA from Microbial Mock Community B (Even, Low Concentration), v5.1L, for 16S rRNA 142 Gene Sequencing, HM-782D). The first stage PCR conditions were: 28 cycles (95°C 20s, 143 55°C 20s, 72°C 5 min) with a 2 min 95°C hot start and 10 min 72°C final elongation. PCR

144 products were screened for quality using gel electrophoresis and purified using the 145 MinElute PCR purification kit (Oiagen). PCR products were then barcoded using the 146 Nextera XT Index Kit v2 set A primers (Illumina) using the following conditions: 8 cycles (95°C 30s, 55°C 30s, 72°C 30s) with 3 min 95° hot start and 5 min 72°C final elongation. 147 148 Barcoded products were purified as above and concentrations of the purified products 149 were assessed using the HS dsDNA assay on the Qubit 2.0 fluorometer (ThermoFisher 150 Scientific). Products were diluted with Tris HCl to 5 nM before being pooled randomly into 151 two libraries. The libraries were diluted to a final loading concentration of 50 pM with a 5% 152 spike-in of 50 pM PhiX. The libraries were then sequenced on the iSeq 100 System 153 (Illumina) using paired-end 150 bp reads. Data are accessible in the NCBI Sequence Read Archive under bioproject PRJNA733652. 154

155 **2.3 FL and VI transects: Data analysis**

156 All code used to generate the figures and analyses in this paper is publicly available on 157 GitHub (https://github.com/microlei/AME biogeography 2021). Sequences were 158 processed using the DADA2 package (v. 1.12.1) in R (v. 3.6.2) (Callahan et al. 2016). Due to 159 the short length of iSeq reads (150bp), it was not possible to merge the reads and therefore 160 only forward reads were used in analysis. Forward reads were filtered using the default 161 parameters of the function filterAndTrim in DADA2 except: trimLeft=20 (to remove the 162 primer), truncLen=125, truncQ=2, maxEE=1. The parameter of truncLen was determined 163 after observing quality dropping during the last 5 bp of the fast reads. Chimera removal 164 and amplicon sequence variant (ASV) generation was also done by DADA2. Taxonomy was 165 assigned, without percent identity clustering, using the naive Bayesian classifier method of

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166	Wang et al. (2007) trained on the Silva SSU rRNA database (Version 138) (Pruesse et al.
167	2007). Putative contaminant reads were identified using the prevalence method in the R
168	package decontam (v. 1.4.0) (Davis et al. 2018) by using the negative controls to identify
169	contaminants; contaminant reads were subsequently removed. Reads matching the
170	Kingdom Eukaryota or Order Chloroplast were also removed.
171	Data analysis was completed in RStudio (v 1.2.5.001) (RStudio Team 2019) using,
172	primarily, the packages phyloseq (v 1.28.0) and vegan (v 2.5-7) (McMurdie & Holmes 2013,
173	Oksanen et al. 2020). Graphics were generated using ggplot2 (v 3.3.3) (Wickham 2016).
174	Alpha diversity metrics were estimated using the <i>estimate_richness</i> function in vegan with
175	unrarefied read counts. Differences in alpha diversity metrics were assessed using pairwise
176	t tests corrected for multiple comparisons using the holm method (Holm 1979). To
177	understand the variability of microbial community composition across all samples, ASV
178	counts were transformed to relative abundances and Bray-Curtis dissimilarity was
179	calculated between each sample pair. Dissimilarity values were plotted using non-metric
180	multidimensional scaling ordination (NMDS). Taxa that were differentially abundant were
181	identified using the package corncob (v 0.2.0)(Bryan D Martin et al. 2021), with a false
182	discovery rate cutoff of 0.05 using Benjamini-Hochberg correction.
183	In order to account for differences in read counts arising from variances in sequencing
184	depth and read quality as well as to improve the quality of the distance-decay analyses, the
185	Aitchison distance was used instead of Bray-Curtis when comparing community similarity
186	across studies in the secondary analysis and, for consistency, between samples in the

187 transect-based study (Gloor et al. 2017, Clark et al. 2021). The package zCompositions (v

188 1.3.4) (Palarea-Albaladeio & Martín-Fernández 2015) was used to impute zeroes before 189 performing a centered-log-ratio (CLR) transformation on the count data. Taking the 190 euclidean distances of the CLR transformed data generated the Aitchison distances. 191 Geographic distances were calculated using the *qdist* function in the package Imap (y 1.32) 192 (John R Wallace 2012), which uses the Vincenty inverse formula for ellipsoids. The adonis2 function in the vegan package was used to perform PERMANOVA analysis (999 iterations) 193 194 on the dissimilarity indices at the scales of transect (within reef), reef, and reef system. 195 Differences in dispersion at various spatial scales were calculated using the vegan function 196 *betadisper* and tested using the vegan function *permutest*, which performs a permutation 197 test (999 permutations) of multivariate homogeneity of groups dispersions. The *mantel* function in the vegan package was used (999 permutation) to test for correlation between 198 199 the geographic distance matrix and the community similarity for both the secondary 200 analysis and transect-based study.

201 2.4 Secondary analysis: Sample information

202 The methods for sample collection, DNA extraction, PCR amplification, and sequencing 203 used by the five studies in the secondary analysis are highly similar with small variations. 204 The studies collected seawater from reefs at a variety of depths, ranging from surface (0.3 205 m) to benthic (13 m). Seawater sampling was done by filtering replicate 2 liter volumes of 206 seawater through 0.22 µm pore size, 25 mm Supor® filters using a peristaltic pump. For 207 DNA extraction, all studies used bead beating followed by spin column purification, 208 although the DNA extraction reagents differed. Neave et al. (2017) used the PowerPlant Pro 209 DNA isolation kit (Qiagen) while Weber et al. (2020) used a sucrose-lysis with bead beating

method followed by column purification with the Qiagen DNeasy Blood and Tissue Kit
(Santoro et al. 2010) as well as a phenol chloroform protocol (Urakawa et al. 2010) and
pooled the extracts. Unpublished data from Becker et al. used the DNeasy PowerBiofilm kit
(Qiagen). Full methods for these unpublished data are included in the Supplemental Text
S1. The remaining two studies used the extraction method described in Santoro et al.
(2010).

216 All studies amplified the V4 hypervariable region of the 16S rRNA gene using the reverse 217 primer described in Apprill et al. (2015), but two studies (Neave et al. 2017, and Apprill et 218 al. 2021) used the forward primer not optimized for Thaumarchaeota (Caporaso et al. 219 2011) while the others used the forward primer described in Parada et al. (2016). All 220 studies used the 250 bp paired-end Illumina MiSeq platform, although Weber et al. (2020) 221 used the Fluidgm® platform (Fluidgm Corporation) for library preparation while others 222 followed the methods described in Kozich et al. (2013). Primer choice, sequencing 223 technology, and DNA extraction method are known to influence downstream 16S rRNA 224 gene sequence analysis, such as in marine biofilms and seawater (Urakawa et al. 2010, 225 Corcoll et al. 2017). A comparison of different DNA extraction techniques on aquatic samples concluded that rare taxa are more affected by differing extraction technique, 226 227 driving small but significant differences in Bray-Curtis distances (Liu et al. 2019). However, 228 the secondary analysis is based on pairwise Aitchison distances, which are less influenced 229 by presence/absence of individual taxa (Gloor et al. 2017) and does not seek to compare 230 groups of samples based on distances.

231 **2.5 Secondary Analysis: Data acquisition and processing**

Raw sequence data from the five studies in the secondary analysis were collected from the 232 233 NCBI Sequence Read Archive (SRA) and for the unpublished study, with consent from the 234 authors. Because the transect comparisons collected in Fl and VI were sequenced with 235 shorter reads, these samples were excluded from the secondary analysis. Using metadata 236 from the studies, sequence files were filtered such that only samples taken from reef 237 associated seawater (and not controls) were included. Samples were classified based on 238 the reef that was sampled as well as the overall reef system (Table S1). Primer sequences 239 were removed using cutadapt (Martin 2011). Sequences were processed in DADA2 as 240 above with the parameters trimLeft=(20,20), truncLen=(205,205), truncQ=(2,2), 241 maxEE=(1,1), and error estimation was performed by pooling all sequences into one error 242 model. Paired forward and reverse reads were assembled into one contig and trimmed to 243 230 bp. Chimera removal, ASV generation, and taxonomy assignment were performed as in 244 section 2.3. Four samples with fewer than 10,000 reads were removed. Because negative 245 controls are specific to each study, contaminant reads were not identified or removed, but 246 reads matching the Kingdom Eukaryota and Order Chloroplast were removed. Data 247 analysis was performed as described in section 2.3. Briefly, the package zCompositions 248 (Palarea-Albaladejo & Martín-Fernández 2015) was used to impute zeroes before using the 249 center-log-ratio transform to normalize the read counts. The Aitchison distance was then 250 plotted against the geographic distance between samples to examine the distance-decay 251 relationship between samples.

252 **3 Results**

253 **3.1 Fl and VI transects: Site characteristics**

In the Florida reef system, Dry Tortugas and Grecian were dominated by macroalgae (56-

255 67% of cover), while Biscayne was dominated by turf algae (45% of cover). Hard coral was

256 more abundant at Dry Tortugas (21%) and Biscayne (20%), but only comprised 3% of

257 Grecian (Figure S1A). In the Virgin Islands system, live coral predmoninated at Brewer's

Bay and Flat Cay (40-43%), but algae was slightly more prevalent at Black Point (40% of

cover) (Figure S1B).

260 At the time of sampling, all reefs with the exception of Dry Tortugas had been experiencing

261 outbreaks of Stony Coral Tissue Loss Disease (SCTLD) to varying degrees of severity and

duration (Precht et al. 2016, Brandt et al. 2021).

263 **3.2 Fl and VI transects: Sequence output**

After quality control of the Fl and VI 16S rRNA gene amplicons from reef water transects, a
total of 13,382,051 reads were retained and the number of reads per sample ranged from
19,686 to 186,683 with a median of 78,976. A total of 20,488 amplicon sequence variants
(ASVs) were identified over 156 samples. Per sample unique ASVs averaged 461. The
abundance matrices of the ASV counts per sample were very sparse, comprising of 97.7%
zeros, indicating that a small number of taxa comprised the majority of the dataset.
Specifically, only 1,228 ASVs make up the top 90% of observations across all samples.

271 3.3 Fl and VI transects: Alpha and beta diversity metrics

272 Alpha diversity metrics of the reef water microbiomes, including observed ASV richness, 273 Shannon index (a measure of evenness), and Simpson's index (a measure of dominance) 274 measured at the transect (within reef), individual reef, and reef system level in Fl and VI 275 showed comparable values with some notable differences. At the transect level (within 276 reefs), observed ASVs were most variable (highest and lowest values) at the Dry Tortugas 277 reefs, with some outliers at both Fl and VI reefs. Simpson's index was most variable at the 278 Dry Tortugas and Brewer's Bay reef (Figure S2, pairwise t-test: p<0.0001 for all significant 279 comparisons involving Dry Tortugas and Brewer's Bay). At the individual reef level, 280 significantly lower Simpson's index values were detected at Dry Tortugas (mean of 0.93) 281 and Brewer's Bay (mean of 0.94) compared to the other reefs (mean of 0.97). Both reefs 282 also displayed significantly lower Shannon index values than three other reefs, except 283 Biscayne (Figure S2, pairwise t-test: p<0.002 for significant differences, p>0.1 for 284 nonsignificant differences). At the reef system level (Fl compared to VI), each of the 285 diversity metrics were significantly different (Figure S2). 286 An NMDS ordination of Bray-Curtis dissimilarities between reef water microbiomes 287 showed that the VI reefs clustered together along with the Dry Tortugas reef water 288 microbial communities, and Fl reefs Grecian and Biscayne were more separated (Figure 289 2A). Bray-Curtis dissimilarities for reef system (Fl and VI) and individual reefs differed 290 significantly, but transects within a reef were not significantly different from each other 291 (PERMANOVA)(Table 2).

292 A comparison of microbial community beta dispersion, calculated as the distance from the 293 centroid of each reef's reef microbial communities in principal coordinate space, showed 294 significant differences at the level of reef system (permutest; F=30.12, p=0.001), individual 295 reef (F=4.87, p=0.001), and within-reef transects (permutest; F=1.90, p=0.026). However, a 296 post hoc test (Tukey's Honest Significant Differences) on the reef-based and transect-based 297 beta dispersions found that the difference was driven solely by the comparisons between 298 the Dry Tortugas reef versus Grecian and Black Point (Figure 2B). The Dry Tortugas reef 299 had low variance among its samples while Grecian and Black Point had a larger variance 300 among samples. When Dry Tortugas was removed from the analysis, there was no longer a 301 significant effect of individual reef or transect on beta dispersion.

302 While only reef-wide benthic composition was recorded for the Fl reefs, we recorded the 303 underlying benthic composition for each sample in the VI reefs to enable comparison 304 between substrate type and the overlying seawater microbiome. When all samples from VI 305 were considered together, Bray-Curtis dissimilarities weakly correlated with benthic 306 substrate type (PERMANOVA; R²=0.093, Pseudo-F=1.64, p=0.042). However, a comparison 307 of live coral compared to other categories (grouped together) did not show a correlation 308 (PERMANOVA; R²=0.0091, Pseudo-F= 0.77, p=0.62), and when samples were nested by 309 their respective reef, the correlation with benthic substrate was no longer significant 310 (PERMANOVA; Pseudo-F=1.64, p= 0.23). Group dispersions were not different between 311 benthic substrate classes or between live coral and other substrates and no taxa were 312 found to be differentially abundant in either of these contrasts (permutest; F=2.32). 313 p=0.063).

314 **3.4 Fl and VI transects: Differentially abundant taxa**

315 To investigate which taxa may be driving differences in community composition with 316 respect to individual reefs, a differential abundance (DA) analysis was performed on the 317 dataset. DA analysis revealed 138 taxa that were significantly differentially abundant 318 (p<0.05, false discovery rate corrected using the Benjamini-Hochberg method) across the319 six reefs sampled. These significant taxa included common oligotrophic marine groups, 320 such as the SAR11 and SAR86 clade, *Cyanobiaceae*, SAR116, and the Archaean Marine 321 Group II. Opportunistic copiotrophs, such as *Flavobacteriaceae*, *Rhodobacteraceae*, and 322 *Vibrionaceae* were also well represented. All significant taxa were among the most 323 abundant and most variable (displayed the highest variance in their relative abundances 324 across samples) in the dataset (Figure S3). A list of the significant taxa along with their 325 sequences is provided in Supplemental Table 2.

326 **3.5 Secondary analysis: Sequence output and methods analysis**

327 After assembly and quality control of the raw sequence reads from the five studies 328 comprising the secondary analysis, a total of 8,761,462 reads were retained. The number of 329 reads per sample ranged from 10,441 to 159,388 with a median of 35,967. A total of 15,005 330 ASVs were identified across the samples of all studies. Per sample unique ASVs averaged 331 272. Although the Fl and VI transect samples had on average greater sequencing depth and 332 ASV count per sample than the studies in the secondary analysis, the relationship between 333 sequencing depth and observed ASVs does not appear to have been saturated in either case 334 (Figure S4).

Because two studies in the analysis used the 806R primer while the other three used the
806RB primer, we evaluated the impact of this difference on the study. The group
dispersions between the two primer sets are not significantly different (permutest; F=2.11,
p=0.14), indicating that primer choice did not significantly contribute to community
variability. While DNA extraction methods were generally similar across studies, this was
not similarly tested as a factor because only two studies shared the same method.

341 3.6 Distance-Decay relationship

342 In order to investigate the impact of geographic distance on the microbial community, a 343 geographic distance matrix was generated using the samples' physical location and 344 compared to the Aitchison distances calculated between microbial communities (a 345 measure of dissimilarity). The comparisons of samples within transects at an individual 346 reef spanned <10 m, while individual reefs were 1-3 km apart for VI reefs and 34-279 km 347 for Fl reefs. Mantel tests of these two matrices revealed that physical distance was 348 significantly correlated with the Aitchison distance between samples within each (Fl and 349 VI) reef system, with a stronger relationship in Florida (r=0.45, p=0.001 for Fl and r=0.10, 350 p=0.004 for VI) (Table 3). There was also a relationship between geographic distances and 351 microbial communities for both Fl and VI reef systems combined, with comparisons 352 spanning 0-1,978 km (r=0.34, p=0.001).

Incorporating the additional reef water microbiomes from the secondary analysis provided us with the opportunity to extend this study to thousands of kilometers. Distances between reefs within a reef system ranged from 1-2775 km while reef systems were separated by

421-16,874 km. The Aitchison distance between these secondary analysis samples showed
a significant relationship to geographic distance (r=0.28, p=0.001)(Table 3).

- 358 In the Fl and VI study, distance-decay plots showed no relationship between geographic
- distance and microbial community similarity within reefs (at the transect level) at the scale
- of meters (p>0.05) (Figure 3A). However, at the scale of kilometers, there was a negative
- 361 relationship between geographic distance and microbial community similarity for the Fl
- and VI study (Figure 3B) as well as for the more expansive secondary analysis (Figure 3C),
- 363 with microbial communities becoming less similar with increasing distance ($R^2 = 0.11$,
- 364 p<0.001; R² = 0.08, p<0.001, respectively). The slope of the negative distance-decay
- 365 relationship increased in magnitude as the geographic extent of the samples increased
- 366 (Table 3, Figure 3C).

367 **3.7 Drivers of distance-decay relationship**

368 In the secondary analysis, the effects of collection depth, temperature, and reef type were 369 examined as potential drivers of community similarity. A PERMANOVA assessing the 370 marginal impacts of these abiotic factors as well the effect of study found that study 371 accounted for the most variation ($R^2=0.14$, Pseudo-F=9.32, p=0.001), distantly followed by 372 reef type (R²=0.03, Pseudo-F=2.06, p=0.001), collection depth (R²=0.017, Pseudo-F=4.64, 373 p=0.001), and finally temperature (R²=0.017, Pseudo-F=4.49, p=0.001). Despite explaining 374 the least amount of variation, difference in temperature was significantly correlated with 375 Aitchison distance (Mantel: r=0.18, p=0.001), meaning communities that were more 376 different (distant) in temperature were also more dissimilar. A similar correlation for 377 depth was not found (Mantel: r=-0.00035, p=0.49).

378 **4 Discussion**

379 In this study we used a nested distance sampling design to examine how reef seawater 380 microbiomes vary at multiple spatial scales, including within reefs, between individual 381 reefs in a reef system as well as across northern Caribbean (Fl and VI) forereef systems. 382 Overall, we found that individual reef and reef system-related features had the largest 383 influence on microbial community diversity and composition. No differences in microbial 384 community diversity or composition were detected within different locations on individual 385 reefs, and there was a weak correlation with the benthic substrate underlying the sample. 386 Despite the large number of observed microbial taxa in the transect-based study, just over 387 one hundred of the most abundant were identified as differentially abundant between 388 reefs, suggesting that these abundant taxa may be useful indicators of reef change. We also 389 used data from five previous studies in a secondary analysis to understand the 390 biogeography of more distant reef seawater microbiomes, and this revealed that microbial 391 communities are more distinct with increasing geographic distance.

392 4.1 Microbial communities differentiate by individual reef (>1 km) and reef

393 system (>100 km), but not within each reef (<10 m)

Counter to expectations, the benthic substrate did not have a strong influence on the
composition of the seawater microbiome. Although there is evidence that substrate type
influences the surrounding seawater microbial community (Schöttner et al. 2012, Tout et
al. 2014), we did not find a correlation in these data, nor did we find differentially abundant
microbial taxa between substrate types. It is likely that differential hydrodynamic

conditions, which have not yet been measured in the context of coral reef benthic-pelagic
microbial interactions, may play a role in these differential results. As such, additional
research concerning benthic-pelagic exchange on coral reefs is needed to understand the
impact of substrate seawater microorganisms close to the reef surface.

403 All six reefs were distinguishable in terms of the composition of their reef water microbial 404 communities, despite the three VI reefs being separated by only 1 to 3 kilometers. These 405 results align with our expectation that microbes in the water column above reefs would 406 display reef-specific signatures because marine microbial communities are reflective of 407 their physical and chemical environment (Azam & Malfatti 2007, Kelly et al. 2018). While 408 benthic substrate had a weak influence on the seawater microbes, there were other 409 influences that varied between the reefs, including season, time of sampling, depth, and 410 reef type, but these generally were consistent within VI or Fl and therefore difficult to 411 statistically examine. Indeed, in our secondary analysis, which included a larger number of 412 reef sites and more geographic locations, temperature, depth, and reef type were small but 413 significant contributors to the community variation. It must be noted, however, that all 414 variables examined were highly confounded by the specific sampling scheme of each study. 415 For example, Becker et al. (2020) only sampled reef seawater from four forereefs in the 416 Virgin Islands at 0.3 meter depth, and Weber et al. (2020) contains a disproportionate 417 number of samples at cooler temperatures (67 out of 82 samples below average 418 temperature of the secondary analysis), all from Cuban reefs. In addition to our study, 419 other studies of the Indian Ocean and Northwestern Hawaiian islands have shown strong 420 microbial biogeographic signatures (Jeffries et al. 2015, Salerno et al. 2016). 421 Hydrodynamics likely plays a major role in explaining some of the biogeographical

portioning between reef water microbial communities because it impacts distribution and
transport of nutrients and facilitates dispersal of pelagic microorganisms. Previous studies
have suggested links between water masses and microbial community composition (Varela
et al. 2008, Galand et al. 2010, Jeffries et al. 2015). Comparison of reef water microbial
communities within and between hydrographic regimes and current systems could help us
better understand this influence.

428 Surprisingly, we did not identify consistent differences in the microbial communities 429 within each reef (at the transect level). Samples collected within meters of each other were 430 indistinguishable in the VI and Fl, but trends in beta dispersion did suggest some within-431 reef variability, indicating differences in beta diversity among one or more groups 432 (Anderson et al. 2006). Dispersion was greatest at Grecian and lowest at Dry Tortugas, both 433 Florida reefs. Grecian reef is located in the Upper Keys, which on average has elevated nutrients, organic carbon, and turbidity compared to the Lower Keys (Lirman & Fong 434 435 2007). In contrast, the Dry Tortugas reef is located within a marine protected zone (US 436 National Park), is more distant from the shore, was the deepest reef sampled (60 feet), and 437 was the only reef not experiencing active outbreaks of Stony Coral Tissue Loss Disease at 438 the time of sampling. These factors could contribute to the relative homogeneity of the 439 samples collected at Dry Tortugas. Reef depth and coastal influence may be among the regional geographic conditions that influence the variability of microbial communities in 440 441 reef associated seawater (Weber et al. 2020, Frade et al. 2020).

442 4.2 Community similarity decays with distance beginning at the kilometer scale

The distance-decay of community similarity - a widely studied relationship in ecology 443 444 (Soininen et al. 2007) - quantifies the decrease in community similarity with increased 445 geographic distance. Typically, communities that are closer geographically are also more 446 similar to each other compositionally (Soininen et al. 2007). One mechanism that drives this relationship is spatial structuring, where locations closer together have more similar 447 448 environments, thus leading to selection of more similar communities. In the absence of 449 selection (e.g., in a homogeneous environment), neutral drift interacts with dispersal 450 limitation to differentiate communities over space (Soininen et al. 2007, Hanson et al. 451 2012). These mechanisms represent two hypotheses for what drives species distributions: 452 environmental selection and historical contingency (Martiny et al. 2006). A number of 453 studies have examined the distance-decay relationship in both soil and marine 454 environments and found that microorganisms tend to display a weaker (i.e., less negative) relationship compared to macroorganisms on the same scale, a phenomenon attributed to 455 456 the small size and large populations of microorganisms leading to greater dispersal (Martiny et al. 2006, Green & Bohannan 2006, Meyer et al. 2018). 457 458 The sampling pattern in this study allowed us to assess this biogeographic pattern in the 459 context of coral reef associated seawater. Within transects in a reef (<10 m scale), no 460 distance-decay relationship was found, likely due to high mixing rates on the reef. 461 However, there was a significant correlation between community similarity and geographic 462 distance beginning at the reef level (1 km scale), and the steepness of the relationship 463 increased with an increase in geographic extent (10,000 km scale) (Table 3). Differences in

464 the steepness and strength of correlation of the distance-decay relationship may reflect 465 different mechanisms driving the decay at multiple spatial scales (Martiny et al. 2011). The 466 larger correlation and slope observed in the Fl reefs compared to VI reefs may reflect the 467 orientation of Fl reefs in a north-south line along the Florida current, with the most distant 468 reef upstream of the two closer reefs, while the VI reefs were closer together and not 469 oriented in relation to the surrounding Caribbean current. The autocorrelation of distance 470 and environmental similarity (Lirman & Fong 2007) along the Florida Reef Tract likely 471 drives the stronger correlation compared to the VI reefs. The steep slope and weaker 472 correlation found in the secondary analysis likely reflects historical factors such as 473 dispersal limitation and drift as distant reefs recruit from different metapopulations 474 (Hellweger et al. 2014, Clark et al. 2021).

475 **4.3** Abundant taxa are most variable and more likely to differentiate individual

476 **reefs**

477 Although we recovered a total of over twenty thousand microbial ASVs from the Fl and VI 478 transect sampling, the vast majority were rare and samples were dominated by just over one thousand highly abundant taxa. This is a common occurance in microbiome 479 480 sequencing, especially with the advent of high throughput deep sequencing, and there is 481 debate over the importance of these rare taxa (McMurdie & Holmes 2014, Cao et al. 2021). 482 Abundant taxa tend to be the most prevalent, and in this study, also displayed the highest variance in their relative abundance values between samples. The ASVs that were 483 484 identified as differentially abundant between reefs were also among the most abundant 485 taxa (Figure S3). Glasl et al. (2019) found that the relative abundances of indicator taxa in

486 coral reef seawater that best predict environmental conditions range from 0.5-20%, and
487 those taxa were also prevalent throughout that study's sampling period. Overlap in
488 taxonomic assignment between these indicator taxa and differentially abundant taxa found
489 within this study include *Synechococcus, Prochlorococcus,* Rhodobacteraceae, unclassified
490 Alphaproteobacteria, and others. Because extremely rare taxa can be difficult to reliably
491 detect, more frequent, shallower sequencing may be more important for capturing the
492 salient variability of a reef.

493 **4.4 Caveats**

494 Although the nested design surveyed seawater microbial communities across multiple 495 spatial scales, the temporal scale of seawater variability was not considered in this study. 496 The Florida samples were taken in June, while the St. Thomas samples were taken in 497 February. Each reef was only sampled once and not throughout the day. Seasonal as well as 498 diurnal/tidal cycles in reef seawater microbial communities are well documented (Weber 499 & Apprill 2020, Glasl et al. 2020, Frade et al. 2020, Becker et al. 2020). The differences 500 between the Florida and VI microbial communities may be in part due to the different 501 seasons in which the samples were taken. Temporal differences in sampling can make 502 direct comparisons between distant reefs challenging, even within the same study (Weber 503 & Apprill 2020). While microbial communities are sensitive to environmental conditions, 504 coral reef seawater remains distinct from other seawater habitats (Becker et al. 2020) and 505 variation of the microbial community is better explained by reef-level environmental 506 parameters rather than seasonal differences (Glasl et al. 2019). Within a reef, repeated sampling throughout the calendar year may be needed to establish the baseline variability. 507

In contrast to many other seawater microbiome studies, including those in the secondary analysis, this study filtered a small volume of seawater (60 mL) for each sample rather than the more typical 1-2 L. Weber et al. (2019) directly examined the effect of sampling volume and found that while species richness modestly increased with larger volumes due to sampling rare taxa, beta diversity and overall community composition was not influenced by sampling volume.

514 **4.5 Conclusions**

515 In conclusion, we suggest that due to reef and reef system-level influences, the 516 development of reef water microbiome monitoring criteria may need to be regionally 517 tailored. We found that the community composition of reef seawater microbiomes are 518 distinguishable even when reefs are a few kilometers apart and that there can be large 519 differences in the beta dispersion within a reef. Detecting a shift in the community 520 composition as a whole will necessitate an understanding of each reef region's variability. 521 Individual reefs within a reef system may also be experiencing different regional stressors, 522 such as varying degrees of anthropogenic influence. Such differences may be reflected in 523 both the baseline microbial community composition and variability as reef conditions 524 change. Additionally, microbial taxa common between reef regions are vastly outnumbered 525 by taxa that are unique, making it difficult to develop a generalized database of indicator 526 microbial taxa for reef environental conditions. Overall, we found that the seawater 527 microbial communities of reefs closer together are more similar, and that the local 528 oceanographic conditions which differentiate these communities are important to 529 investigate.

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730 Figures and tables



733 **Figure 1. Locations of reefs in the present study and the secondary analysis** A.

734 Sampling locations and method for the transect-based study. Six reefs were surveyed

across two reef systems. At each reef, three 10 meter transects were laid and divers used a

- syringe to sample seawater just above the benthos at 1 meter intervals. B. The locations of
- the five studies in the secondary analysis span many major reef systems across the globe.
- 738 Collections for all studies were performed by the same lab group using nearly identical
- techniques in the field and in the lab.



Figure 2. Reefs have distinct microbial communities and variable group dispersions
A. An NMDS ordination using Bray-Curtis dissimilarities. Tries=20, 2D stress=0.1006136. B.
The group dispersions calculated as distance to centroid using the Bray-Curtis dissimilarity
metric. The vertical line in each boxplot indicates the median, the box hinges represent the
first and third quartiles while the whiskers extend to 1.5*IQR (interquartile range). Colors
correspond to the individual reef and letters indicate significance groups (pairwise t test;
p-adjusted<0.05).



Figure 3. Reef seawater communities exibit a distance-decay relationship at the scale
of 100 km but not at <10 m scale A. The pairwise geographic distances between samples
within each transect in the Fl/VI-based study is plotted on the x-axis and the corresponding
Aitchison distances are plotted on the y axis. Between transect distances are not known
and therefore not included. B. Distance-decay plot of all pairwise distances (i.e., not just
within a transect) between samples collected in the Fl/VI-based study. C. Distance-decay
plot using the samples from the five studies included in the secondary analysis.

Poof System	Poof	Transacta	Samples	Ave Dopth (m)	CDS	Data	
iteel System	neer	Transects	Samples	Avg. Deptii (iii)	GIS	Date	
Flands Varia	Dry Tortugas	3	30	17.98	24.722 N 82.828 W	7-Jun-19	
Florida Keys	Grecian	3	30	6.71	$25.110~{\rm N}~80.303~{\rm W}$	14-Jun-19	
	Biscayne	1	9	5.18	$25.386~{\rm N}~80.162~{\rm W}$	17-Jun-19	
	Flat Cay	3	30	7.12	$18.316~{\rm N}~64.987~{\rm W}$	12-Feb- 20	
St. Thomas, USVI	Black Point	3	30	7.00	$18.344~{\rm N}~64.986~{\rm W}$	12-Feb-20	
	Brewer's Bay	3	27	5.92	$18.343~{\rm N}~64.980~{\rm W}$	12-Feb- 20	

Table 1: Florida reef tract (Fl) and St. Thomas (VI) Sampling locations

Table 2: Summary of PERMANOVA results based on Bray-Curtis dissimilarties of microbial community abundances grouped at the spatial scales of within reefs, between reefs, and between reef regions

Scope	$\mathbf{D}\mathbf{f}$	$\mathbf{Sum} \ \mathbf{Sq}$	$\mathbf{R2}$	Pseudo-F	Р
Transect (within reef)	15	8.366559	0.6524673	17.52265	0.412
Residual	140	4.456397	0.3475327		
Total	155	12.822956	1.0000000		
Site (between reefs)	5	8.050737	0.6278378	50.61002	0.001
Residual	150	4.772219	0.3721622		
Total	155	12.822956	1.0000000		
Reef system (between reef regions)	1	3.284216	0.2561200	53.02264	0.001
Residual	154	9.538740	0.7438800		
Total	155	12.822956	1.0000000		

Note: Df - degrees of freedom; R2 - R-squared value; Pseudo-F values derived from 999 permutations. Values in bold are significant at P<0.05.

Table 3:	Distance	decay	relationship	os at	multip	le spati	al scales
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Scope	Spatial scale	Correlation (Mantel r)	Mantel p value	Slope of linear fit
Within transects	0-9 m	NA^{*}	NA^{*}	-0.002
(Reef system) USVI	$0-3 \mathrm{~km}$	0.1009	0.004	-0.008^{\dagger}
(Reef system) Florida Keys	$0-279 \mathrm{~km}$	0.4492	0.001	-3.208^\dagger
All samples in Fl/VI-based study	$0-1,978 { m \ km}$	0.3358	0.001	-16.246^{\dagger}
All samples in meta-analysis	$0-16,874 { m \ km}$	0.2840	0.001	-113.128^{\dagger}

* Mantel test not performed because distances exist for samples within transects but not between, therefore matrix had >50% missing values † Linear fit significant at p < 0.01