# Selected Reef Fish Visual Census Studies Conducted by the Miami Laboratory Reef Resources Team, 1985-2002 

David B. McClellan and Douglas E. Harper

(Editors)

U.S. Department of Commerce

National Oceanic and Atmospheric Administration
National Marine Fisheries Service
Southeast Fisheries Science Center
75 Virginia Beach Drive
Miami, Florida 33149

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## Foreword

## G. Todd Kellison and James A. Bohnsack

This document is a compilation of unpublished reports describing six unrelated fish survey projects conducted by the NOAA Southeast Fisheries Science Center (SEFSC) Reef Resources Team during the period 1985-2002. The projects span the geographical range from Puerto Rico and Little Cayman in the Caribbean to the Florida middle grounds (West Florida Shelf), Dry Tortugas region, and Florida Keys. The projects described in this compilation represent early research and monitoring efforts.

The purpose of this document is to make previously unpublished research available for potential use by researchers, managers, and other parties interested in the fish communities associated with the study areas described in the studies herein. For example, researchers or managers might use data collected and summarized in this compilation as quantitative or qualitative baselines against which to assess changes over time for the study areas of interest.

In general, the reports contained in this compilation are as initially written following the original study. Thus, the report format is not consistent throughout the compilation, and with minor exceptions reports have not been edited following their original completion. For each of the reports, nomenclature was as listed in the then-current edition of the American Fisheries Society publication "Common and Scientific Names of Fishes from the United States and Canada".

For most of the reports, fish data are summarized in tabular format. For readers interested in obtaining more detailed data, inquiries should be sent to the address below.

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# Chapter 1. Reef Fish Survey, Mona Island, Puerto Rico, October 1985 

James A. Bohnsack and Douglas E. Harper

## Purpose

The purpose of this study was to document reef fish community structure in damaged and undamaged reef areas near the grounded vessel A Regina at Mona Island, Puerto Rico (Fig. 1). The specific objectives were to quantitatively document reef fish abundances and frequencies near the grounded vessel using standard visual survey methods.

## Methods

Standard visual stationary sampling methodology was used (Bohnsack and Bannerot, 1986) to assess reef fish community structure. A series of visual samples were conducted on damaged reef areas within 50 m of the vessel (inside damage stations, $\mathrm{n}=15$; and adjacent to damage stations, $\mathrm{n}=$ 9). A series of control stations were made on reef areas showing no obvious signs of damage between 50-300 m away from the vessel (inshore control stations, $n=6$; and offshore control stations, $\mathrm{n}=11$ ). The percent composition of bottom substrates was recorded at each station as viewed by the diver from one central point. All stations were taken between 11 and 15 October 1985.

Data were summarized and analyzed to provide estimates of percent frequency and mean abundance for each species for future comparisons. The mean number of individuals and species per sample were compared for damaged and control areas of Mona Island. Also the size frequency and mean abundance of the ocean surgeonfish, Acanthurus bahianus, were compared between damaged and control areas. The number of individuals per sample was first transformed by a $\log _{10}$ transform before analysis. Abundance of A. bahianus was transformed by a $\log _{10}(x+1)$. Parametric tests were used if data appeared normally distributed and variances were approximately equal. Non-parametric U-tests were used if either of these assumptions were violated.

Three swimming transects were conducted in which all observed predators were censused in 15 min random swims (Bohnsack, 1982). Additional samples were not taken and these data were not analyzed because so few predators were observed.

## Results

Bottom composition varied considerably between sample areas (Table 1). Control areas had considerably more coral coverage and damage areas had considerably more rubble coverage. The damage areas varied in morphology between offshore and inshore areas. The inshore area was primarily composed of intact coral colonies both living and dead. The dead colonies appeared to have been killed by sediments or abrasion. The offshore damage areas (seaward of the A Regina) differed in that physical damage caused by the grounded vessel had destroyed most coral colonies leaving an irregular carbonate rock and rubble substrate with little relief. Depths of stations ranged between 3 and 6 m for inside damage, 5 and 7 m for outside damage, 3 and 6 m inshore control, and 6 and 9 m for offshore control stations.

A total of 65 fish species were observed in 41 stationary samples (Table 2). The mean number of species per sample from damaged areas (mean $=15.3 \pm 1.6595 \% \mathrm{CI}, \mathrm{n}=24$ ) was less than that found from control areas (mean $=18.2 \pm 1.6995 \% \mathrm{CI}, \mathrm{n}=17$ ). The difference was highly significant ( $\mathrm{p}<0.01$, t -test). The mean number of individuals per sample was also significantly less ( $\mathrm{p}<0.05, \mathrm{t}$-test) in damaged areas (transformed mean $=2.1217 \pm 0.134895 \% \mathrm{CI}, \mathrm{n}=24$ ) than in undamaged control areas (transformed mean $=2.3500 \pm 0.125195 \% \mathrm{CI}, \mathrm{n}=17$ ).

Statistical descriptions were made for each species observed from the four areas (Appendix A). Abundance and frequency-of occurrence patterns for each species were compared from the four sample areas (Table 3). Statistical tests for differences were not made for every species although graphical comparisons of patterns of abundance (Fig. 2) and frequency-of-occurrence (Fig. 3) were made. The abundance of grunts (Haemulidae) and snapper (Lutjanidae) were very low from all areas. Species conspicuously absent were yellowtail snapper, Ocyurus chrysurus, and the threespot damselfish, Pomacentrus planifrons. The redlip blenny Ophioblennius atlanticus and black durgon Melichthys niger were conspicuously abundant.

Differences in abundance of $A$. bahianus between damaged areas and control areas of Mona Island were significant ( $p<0.05$, one tailed $t$-test) with damaged sites having significantly more individuals (transformed mean $=1.1 .1778 \pm 0.224595 \% \mathrm{CI}, \mathrm{n}=24$ ) than undamaged control sites (transformed mean $=0.9036 \pm 0.127895 \% \mathrm{CI}, \mathrm{n}=17$ ). The size distribution of A. bahianus was also significantly different between control and damaged areas ( $\mathrm{p}<0.05$, Kolmogorov-Smirnov test) with smaller individuals observed in damaged areas (Fig. 4).

## Discussion and Conclusion

The fish fauna at Mona Island was generally depauperate compared to other reefs we have examined along the North American continent from Florida to Belize. This phenomenon is most likely the result of the isolation of Mona Island from other reef habitats, the surrounding oceanic conditions, and the small amount of living reef and shelf area at Mona Island. Starck (1968) observed few grunts and snapper from islands in the Caribbean with little shelf area. This is apparently the result of limited available foraging area to support fishes. The impacts of local fishing pressure on the biota are unknown but may have influenced the observed pattern.

As would be expected in a major ship grounding, bottom substrate data indicated significant differences between damage and control areas in terms of substrate composition. More coral cover was present in undamaged areas and more rubble and bare rock was present in damaged areas. The reduced habitat diversity and profile probably accounts for the observed reduced mean number of species and individuals in stations from damaged areas. Observations of the Wellwood grounding off Key Largo, Florida (Bohnsack, unpubl. data) indicated significant increases in numbers of the herbivores A. bahianus and Scarus croicensis on damaged areas. Also, the average size of $A$. bahianus was significantly smaller on damaged areas because of a large number of recruits. Results from damaged and undamaged areas of Mona Island show the same patterns for A. bahianus. Too few S. croicensis were present at Mona Island to support a statistical comparison. Although the mean number of all scarids were greater in damaged areas (mean $=6.25 \pm 1.9495 \% \mathrm{CI}$ ) than in control areas ( $5.38 \pm 1.706$ ), the difference was not statistically significant ( $\mathrm{p}>0.05$ ).

The greater abundance of the herbivorous $A$. bahianus in damaged areas is probably the result of greater availability of early successional algae colonizing the newly exposed bare rock and dead coral. Smaller individuals probably occur in the damaged areas because of greater attraction for juveniles, greater recruitment of settling larvae, or better juvenile survival from reduced predation or competition. The exact mechanism has not been demonstrated.

Detailed comparisons of abundance patterns between the various sample areas were not made on a species by species basis. However, the brown chromis, Chromis multilineatus, showed a clear pattern of greater abundance in undamaged areas. The brown chromis, a highly mobile diurnal planktivore, normally schools around prominent outcrops and feeds on the passing plankton. Damaged areas may lack sufficient relief for brown chromis or the presence of the grounded ship may somehow alter current patterns in a way that was unfavorable for maintaining their presence.

## Acknowledgements

We thank the Department of Natural Resources, Government of Puerto Rico for providing the use of the vessel Jean $\underline{A}$, food, and logistical support for this study. Special assistance was provided by Gilberto Cintron.

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Table 1. Substrate composition. Percentage of substrate viewed by a stationary diver within a 7.5 m radius.

| Bottom Type | Sample Locations |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Inshore Damage | Offshore Damage | Inshore <br> Control | Offshore Control |
| ${ }^{1}{ }^{1}$ | 20.4 | 5.8 | 42.3 | 69.0 |
| Sand 2 | 24.0 | 7.8 | 5.2 | 20.5 |
| Rock | 36.6 | 71.4 | 50.8 | 5.5 |
| Rubble | 19.0 | 15.0 | 1.7 | 5.0 |
| ```1 Includes palmata (liv by macro-bio 2 Includes micro-biota.``` | ving co and dead <br> posed | colonies, gorgonian <br> carbonate | tact stand nd reef | of Acro ers domin <br> dominated |

Table 2. Alphabetical listing of fishes at the Mona Island grounding site, October 11-15, 1985, 41 samples recorded. * indicates that the species was observed during sampling, but after the initial 5 minute sampling period and no abundance estimates were recorded. Numbers identify species in Figures 1 and 2.

| SPECIES CODE |  | RANDOM POINT SAMPLES |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | SCIENTIFIC NAME | COMMON NAME | Freq. | Tot. Abund. |
| 1 ABU SAXA | Abudefduf saxatilis | Sergeant major | 7 | 53 |
| 2 ACA BAHI | Acanthurus bahianus | Ocean surgeon | 40 | 848 |
| 3 ACA COER | Acanthurus coeruleus | Blue tang | 36 | 439 |
| 4 ADI VEXI | Adioryx vexillarius | Dusky squirrelfish | 1 | 1 |
| 5 AMB PINO | Amblycirrhitus pinos | Redspotted hawkfish | 1 | 1 |
| 6 AUL MACU | Aulostomus maculatus | Trumpetfish | 1 | 1 |
| 7 BOD RUFU | Bodianus rufus | Spanish hogfish | 21 | 38 |
| 8 BOT LUNA | Bothus lunatus | Peacock flounder | 3 | 3 |
| 9 CAN PULL | Cantherhines pullus | Orangespotted filefish | 1 | 1 |
| 10 CAR RUBE | Caranx ruber | Bar jack | 18 | 34 |
| 11 CHA CAPI | Chaetodon capistratus | Foureye butterflyfish | 3 | 5 |
| 12 CHA STRI | Chaetodon striatus | Banded butterflyfish | 8 | 10 |
| 13 CHR CYAN | Chromis cyaneus | Blue chromis | 1 | 2 |
| 14 CHR MULT | Chromis multilineatus | Brown chromis | 12 | 322 |
| 15 EPI FULV | Epinephelus fulvus | Coney | 18 | 65 |
| 16 GNA THOM | Gnatholepis thompsoni | Goldspot goby | 1 | 1 |
| 17 GOB EVEL | Gobiosoma evelynae | Sharpnose goby | 3 | 8 |
| 18 HAE AURO | Haemulon aurolineatum | Tomtate | 1 | 15 |
| 19 HAE CARB | Haemulon carbonarium | Caesar grunt | 10 | 12 |
| 20 HAE CHRY | Haemulon chrysargyreum | Smallmouth grunt | 1 | 16 |
| 21 HAL BIVI | Halichoeres bivittatus | Slippery dick | 31 | 131 |
| 22 HAL GARN | Halichoeres garnoti | Yellowhead wrasse | 17 | 68 |
| 23 HAL MACU | Halichoeres maculipinna | Clown wrasse | 29 | 99 |
| 24 HAL PICT | Halichoeres pictus | Rainbow wrasse | 4 | 23 |
| 25 HAL POEY | Halichoeres poeyi | Blackear wrasse | 16 | 44 |
| 26 HAL RADI | Halichoeres radiatus | Puddingwife | 34 | 100 |
| 27 HOL ADSC | Holocentrus adscensionis | Squirrelfish | 2 | 2 |
| 28 HOL RUFU | Holocentrus rufus | Longspine squirrelfish | 5 | 6 |
| 29 HOL TRIC | Holacanthus tricolor | Rock beauty | 8 | 14 |
| 30 KYP SECT | Kyphosus sectatrix | Bermuda chub | 6 | 103 |
| 31 LAC BICA | Lactophrys bicaudalis | Spotted trunkfish | 1 | 1 |
| 32 LAC POLY | Lactophrys polygonia | Honeycomb cowfish | 1 | 1 |
| 33 LAC TRIQ | Lactophrys triqueter | Smooth trunkfish | 3 | 3 |
| 34 LUT APOD | Lutjanus apodus | Schoolmaster | 4 | 4 |
| 35 LUT JOCU | Lutjanus jocu | Dog snapper | 2 | 2 |
| 36 LUT MAHO | Lutjanus mahogoni | Mahogany snapper | 5 | 15 |
| 37 MAL AURO | Malacoctenus aurolineatus | Goldline blenny | 2 | 3 |
| 38 MAL PLUM | Malacanthus plumieri | Sand tilefish | 4 | 10 |
| 39 MAL SPE. | Malacoctenus sp. | Unidentified blenny | 1 | 2 |
| 40 MAL TRIA | Malacoctenus triangulatus | Saddled blenny | 12 | 24 |

Table 2 (cont.)

| SPECIES |  | RANDOM POINT SAMPLES |  |  |
| :---: | :---: | :---: | :---: | :---: |
| CODE | SCIENTIFIC NAME | COMMON NAME | Freq. | Tot. Abund. |
| 41 MEL NIGE | Melichthys niger | Black durgon | 38 | 575 |
| 42 MIC CHRY | Microspathodon chrysurus | Yellowtail damselfish | 36 | 207 |
| 43 MUL MART | Mulloidichthys martinicus | Yellow goatfish | 9 | 21 |
| 44 MYR JACO | Myripristis jacobus | Blackbar soldierfish | 1 | 1 |
| 45 OPH ATLA | Ophioblennius atlanticus | Redlip blenny | 34 | 355 |
| 46 POM FUSC | Pomacentrus fuscus | Dusky damselfish | 34 | 226 |
| 47 POM PART | Pomacentrus partitus | Bicolor damselfish | 11 | 284 |
| 48 POM PARU | Pomacanthus paru | French angelfish | 1 | 1 |
| 49 PSE MACU | Pseudupeneus maculatus | Spotted goatfish | 4 | 5 |
| 50 RYP SAPO | Rypticus saponaceus | Greater soapfish | 1 | 1 |
| 51 SCA CROI | Scarus croicensis | Striped parrotfish | 6 | 12 |
| 52 SCA TAEN | Scarus taeniopterus | Princess parrotfish | 3 | 8 |
| 53 SCA VETU | Scarus vetula | Queen parrotfish | 5 | 5 |
| 54 SPA AURO | Sparisoma aurofrenatum | Redband parrotfish | 26 | 63 |
| 55 SPA CHRY | Sparisoma chrysopterum | Redtail parrotfish | 9 | 25 |
| 56 SPA RUBR | Sparisoma rubripinne | Yellowtail parrotfish | 18 | 65 |
| 57 SPA VIRI | Sparisoma viride | Stoplight parrotfish | 28 | 72 |
| 58 SPH BARR | Sphyraena barracuda | Barracuda | 4 | 4 |
| 59 SPH PICU | Sphyraena picudilla | Southern sennet | 4 | 920 |
| 60 THA BIFA | Thalassoma bifasciatum | Bluehead | 40 | 3,141 |
| (Additional species observed) |  |  |  |  |
| BAL VETU | Balistes vetula | Queen triggerfish | * |  |
| CAR BART | Caranx bartholomaei | Yellow jack | * |  |
| CLE PARR | Clepticus parrai | Purple reeffish | * | * |
| DAS AMER | Dasyatis americana | Southern stingray | * | * |
| URO JAMA | Urolophus jamaicensis | Yellow stingray | * | * |

Table 3. Statistical comparison of species from four sample areas.

| No. SPECIES | REGINA DAMAGED AREA |  |  |  |  |  | REGINA CONTROL AREA |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | INSHORE |  |  | OUTSID |  |  | INSHORE |  |  | HORE |  |  |
|  |  |  |  | \%FREQ | MEAN |  | \%FREQ |  |  | \%FREQ |  |  |
|  | ( $\mathrm{N}=15$ ) | ( $\mathrm{N}=15$ ) | 95\% CI | ( $\mathrm{N}=9$ ) | ( $\mathrm{N}=9$ ) | 95\% Cl | ( $\mathrm{N}=6$ ) | ( $\mathrm{N}=6$ ) | 95\% CI | ( $\mathrm{N}=11$ ) | ( $\mathrm{N}=11$ ) | 95\% CI |
| 1 ABU SAXA | 26.7\% | 0.47 | 0.53 | 11.1\% | 0.44 | 0 | 16.7\% | 0.50 | 0 | 9.1\% | 3.55 | 0 |
| 2 ACA BAHI | 93.3\% | 21.13 | 20.43 | 100.0\% | 43.00 | 31.4 | 100.0\% | 7.83 | 8.005 | 100.0\% | 8.82 | 3.96 |
| 3 ACA COER | 93.3\% | 18.80 | 31.98 | 66.7\% | 2.89 | 2.368 | 83.3\% | 6.67 | 6.758 | 100.0\% | 8.27 | 6.758 |
| 4 ADI VEXI | 0.0\% | 0.00 | 0 | 11.1\% | 0.11 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 |
| 5 AMB PINO | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 9.1\% | 0.09 | 0 |
| 6 AUL MACU | 6.7\% | 0.07 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 |
| 7 BOD RUFU | 46.7\% | 0.73 | 0.435 | 44.4\% | 0.67 | 0.435 | 50.0\% | 0.67 | 0.605 | 63.6\% | 1.55 | 1.443 |
| 8 BOT LUNA | 6.7\% | 0.07 | 0 | 0.0\% | 0.00 | 0 | 16.7\% | 0.17 | 0 | 9.1\% | 0.09 | 0 |
| 9 CAN PULL | 0.0\% | 0.00 | 0 | 11.1\% | 0.11 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 |
| 10 CAR RUBE | 40.0\% | 0.67 | 0.286 | 22.2\% | 0.33 | 0.533 | 50.0\% | 0.83 | 1.208 | 63.6\% | 1.45 | 1.208 |
| 11 CHA CAPI | 6.7\% | 0.07 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 18.2\% | 0.36 | 0 |
| 12 CHA STRI | 20.0\% | 0.27 | 0.319 | 11.1\% | 0.11 | 0 | 0.0\% | 0.00 |  | 36.4\% | 0.45 | 0.335 |
| 13 CHR CYAN | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 9.1\% | 0.18 | 0 |
| 14 CHR MULT | 6.7\% | 0.27 | 0 | 0.0\% | 0.00 | 0 | 33.3\% | 10.33 | 7.421 | 81.8\% | 23.27 | 11.91 |
| 15 EPI FULV | 6.7\% | 0.27 | 0 | 55.6\% | 2.56 | 2.9 | 33.3\% | 0.67 | 0 | 90.9\% | 3.09 | 0.906 |
| 16 GNA THOM | 6.7\% | 0.07 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 |
| 17 GOB EVEL | 6.7\% | 0.13 | 0 | 0.0\% | 0.00 | 0 | 16.7\% | 0.67 | 0 | 9.1\% | 0.18 | 0 |
| 18 HAE AURO | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 16.7\% | 2.50 | 0 | 0.0\% | 0.00 | 0 |
| 19 HAE CARB | 53.3\% | 0.67 | 0.256 | 22.2\% | 0.22 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 |
| 20 HAE CHRY | 6.7\% | 1.07 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 |
| 21 HAL BIVI | 86.7\% | 3.87 | 1.188 | 88.9\% | 4.22 | 2.443 | 66.7\% | 1.67 | 1.817 | 54.5\% | 2.27 | 1.303 |
| 22 HAL GARN | 13.3\% | 0.27 | 0 | 22.2\% | 0.56 | 0.533 | 33.3\% | 0.83 | 2.226 | 100.0\% | 4.91 | 3.547 |
| 23 HAL MACU | 66.7\% | 2.73 | 1.151 | 66.7\% | 3.44 | 3.352 | 66.7\% | 1.50 | 1.004 | 81.8\% | 1.64 | 0.671 |
| 24 HAL PICT | 6.7\% | 0.07 | 0 | 33.3\% | 2.44 | 1.569 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 |
| 25 HAL POEY | 40.0\% | 0.80 | 0.35 | 88.9\% | 3.22 | 1.392 | 16.7\% | 0.33 | 0 | 9.1\% | 0.09 | 0 |
| 26 HAL RADI | 66.7\% | 1.60 | 0.65 | 100.0\% | 2.33 | 1.131 | 83.3\% | 2.83 | 0.938 | 90.9\% | 3.45 | 1.813 |
| 27 HOL ADSC | 0.0\% | 0.00 | 0 | 11.1\% | 0.11 | 0 | 0.0\% | 0.00 | 0 | 9.1\% | 0.09 | 0 |
| 28 HOL RUFU | 0.0\% | 0.00 | 0 | 11.1\% | 0.11 | 0 | 16.7\% | 0.17 | 0 | 27.3\% | 0.36 | 0.387 |
| 29 HOL TRIC | 13.3\% | 0.13 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 54.5\% | 1.09 | 0.6 |
| 30 KYP SECT | 6.7\% | 0.07 | 0 | 11.1\% | 4.44 | 0 | 50.0\% | 4.33 | 12.16 | 9.1\% | 3.27 | 0 |
| 31 LAC BICA | 6.7\% | 0.07 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 |
| 32 LAC POLY | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 9.1\% | 0.09 | 0 |
| 33 LAC TRIQ | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 16.7\% | 0.17 | 0 | 18.2\% | 0.18 | 0 |
| 34 LUT APOD | 13.3\% | 0.13 | 0 | 0.0\% | 0.00 | 0 | 33.3\% | 0.33 | 0 | 0.0\% | 0.00 | 0 |
| 35 LUT JOCU | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 33.3\% | 0.33 | 0 | 0.0\% | 0.00 | 0 |
| 36 LUT MAHO | 33.3\% | 1.00 | 1.918 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 |
| 37 MAL AURO | 6.7\% | 0.07 | 0 | 0.0\% | 0.00 | 0 | 16.7\% | 0.33 | 0 | 0.0\% | 0.00 | 0 |
| 38 MAL PLUM | 20.0\% | 0.53 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 9.1\% | 0.18 | 0 |
| 39 MAL SPE. | 0.0\% | 0.00 | 1.152 | 11.1\% | 0.22 | 0 | 0.0\% | 0.00 |  | 0.0\% | 0.00 | 0 |
| 40 MAL TRIA | 13.3\% | 0.13 | 0 | 22.2\% | 0.56 | 1.599 | 16.7\% | 0.17 | 0 | 63.6\% | 1.45 | 0.507 |
| 41 MEL NIGE | 86.7\% | 13.73 | 19.62 | 88.9\% | 14.67 | 10.39 | 100.0\% | 13.00 | 5.593 | 100.0\% | 14.45 | 8.19 |
| 42 MIC CHRY | 86.7\% | 3.40 | 1.184 | 66.7\% | 2.22 | 1.628 | 100.0\% | 7.83 | 4.473 | 100.0\% | 8.09 | 2.069 |
| 43 MUL MART | 33.3\% | 0.67 | 0.553 | 44.4\% | 1.22 | 0.948 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 |
| 44 MYR JACO | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 9.1\% | 0.09 | 0 |
| 45 OPH ATLA | 60.0\% | 2.40 | 0.875 | 88.9\% | 12.22 | 7.833 | 100.0\% | 18.33 | 6.072 | 100.0\% | 9.00 | 4.108 |
| 46 POM FUSC | 93.3\% | 5.07 | 1.813 | 88.9\% | 6.78 | 2.609 | 100.0\% | 9.67 | 4.233 | 54.5\% | 2.82 | 2.318 |
| 47 POM PART | 6.7\% | 0.73 | 0 | 11.1\% | 0.11 | 0 | 0.0\% | 0.00 | 0 | 81.8\% | 24.73 | 8.115 |
| 48 POM PARU | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 9.1\% | 0.09 | 0 |
| 49 PSE MACU | 6.7\% | 0.07 | 0 | 22.2\% | 0.33 | 0.533 | 0.0\% | 0.00 | 0 | 9.1\% | 0.09 | 0 |
| 50 RYP SAPO | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 9.1\% | 0.09 | 0 |
| 51 SCA CROI | 33.3\% | 0.67 | 0.391 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 9.1\% | 0.18 | 0 |
| 52 SCA TAEN | 6.7\% | 0.13 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 18.2\% | 0.55 | 0 |

Table 3 (cont.)

| No. SPECIES | REGINA DAMAGED AREA |  |  |  |  |  |  | REGINA CONTROL AREA |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | INSHORE |  |  | OUTSIDE |  |  | \| | INSHOR |  |  | OFFSHO | ORE |
|  | \%FREQ $(\mathrm{N}=15)$ | $\begin{array}{r} \text { MEAN } \\ (\mathrm{N}=15) \\ \hline \hline \end{array}$ | $\begin{array}{r} \text { + or - } \\ 95 \% \\ \hline \end{array}$ | \%FREQ $(\mathrm{N}=9)$ | $\begin{aligned} & \text { MEAN } \\ & (\mathrm{N}=9) \\ & \hline \end{aligned}$ | $\begin{array}{r} + \text { or }- \\ 95 \% \\ \hline \end{array}$ | \%FREQ $(\mathrm{N}=6)$ | $\begin{array}{r} \text { MEAN } \\ (\mathrm{N}=6) \\ \hline \end{array}$ | $\begin{array}{r} + \text { or }- \\ 95 \% \end{array}$ | \%FREQ $(\mathrm{N}=11)$ | $\begin{array}{r} \text { MEAN } \\ (\mathrm{N}=11) \\ \hline \end{array}$ | $\begin{array}{r} + \text { or }- \\ 95 \% \mathrm{CI} \\ \hline \end{array}$ |
| 53 SCA VETU | 6.7\% | 0.07 | 0 | 22.2\% | 0.22 | 0 | 16.7\% | 0.17 | 0 | 9.1\% | 0.09 | 0 |
| 54 SPA AURO | 66.7\% | 1.73 | 0.467 | 77.8\% | 1.56 | 1.056 | 16.7\% | 0.33 | 0 | 72.7\% | 1.91 | 1.562 |
| 55 SPA CHRY | 20.0\% | 0.87 | 1.279 | 55.6\% | 1.11 | 0.754 | 16.7\% | 0.33 | 0 | 0.0\% | 0.00 | 0 |
| 56 SPA RUBR | 46.7\% | 1.27 | 1.419 | 22.2\% | 1.00 | 3.732 | 100.0\% | 4.67 | 0.857 | 27.3\% | 0.82 | 1.163 |
| 57 SPA VIRI | 73.3\% | 2.47 | 0.793 | 44.4\% | 0.78 | 0.721 | 66.7\% | 1.83 | 1.004 | 81.8\% | 1.55 | 0.783 |
| 58 SPH BARR | 13.3\% | 0.13 | 0 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 | 18.2\% | 0.18 | 0 |
| 59 SPH PICU | 6.7\% | 14.00 | 0 | 33.3\% | 78.89 | 82.71 | 0.0\% | 0.00 | 0 | 0.0\% | 0.00 | 0 |
| 60 THA BIFA | 93.3\% | 32.20 | 30.29 | 100.0\% | 45.00 | 38.49 | 100.0\% | 86.67 | 111.2 | 100.0\% | 157.55 | 73.13 |

Figure 1. Location of $\underline{\text { A Regina grounding at Mona Island, Puerto Rico. }}$


Figure 2. Comparison of mean abundance by species for the four sample locations. Species are numbered alphabetically (see Table 2).


Figure 3. Comparison of frequency-of-occurrence by species for the four sample locations. Species are numbered alphabetically (see Table 2).


Figure 4. Comparison of size frequency distributions for Acanthurus bahianus from damaged (top) and control (bottom) stations.


Appendix A.1. Inside damage area, REGINA grounding, Mona Island, Puerto Rico.

| No. Species | Total <br> Indiv. | $\begin{array}{r} \text { Frequ } \\ (\mathrm{N}=15) \\ \hline \end{array}$ | ncy | Mean Abund. | Stand. <br> Dev. | Range High | Low | Var/Mean $\qquad$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 ABU SAXA | 7 | 4 | 26.7\% | 0.47 | 0.92 | 3 | 0 | 1.96 |
| 2 ACA BAHI | 317 | 14 | 93.3\% | 21.13 | 36.03 | 142 | 0 | 64.39 |
| 3 ACA COER | 282 | 14 | 93.3\% | 18.80 | 55.90 | 220 | 0 | 177.42 |
| 4 ADI VEXI | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 5 AMB PINO | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 6 AUL MACU | 1 | 1 | 6.7\% | 0.07 | 0.26 | 1 | 0 | 0.00 |
| 7 BOD RUFU | 11 | 7 | 46.7\% | 0.73 | 0.96 | 3 | 0 | 0.84 |
| 8 BOT LUNA | 1 | 1 | 6.7\% | 0.07 | 0.26 | 1 | 0 | 0.00 |
| 9 CAN PULL | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 10 CAR RUBE | 10 | 6 | 40.0\% | 0.67 | 0.90 | 2 | 0 | 0.40 |
| 11 CHA CAPI | 1 | 1 | 6.7\% | 0.07 | 0.26 | 1 | 0 | 0.00 |
| 12 CHA STRI | 4 | 3 | 20.0\% | 0.27 | 0.59 | 2 | 0 | 1.25 |
| 13 CHR CYAN | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 14 CHR MULT | 4 | 1 | 6.7\% | 0.27 | 1.03 | 4 | 0 | 0.00 |
| 15 EPI FULV | 4 | 1 | 6.7\% | 0.27 | 1.03 | 4 | 0 | 0.00 |
| 16 GNA THOM | 1 | 1 | 6.7\% | 0.07 | 0.26 | 1 | 0 | 0.00 |
| 17 GOB EVEL | 2 | 1 | 6.7\% | 0.13 | 0.52 | 2 | 0 | 0.00 |
| 18 HAE AURO | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 19 HAE CARB | 10 | 8 | 53.3\% | 0.67 | 0.72 | 2 | 0 | 0.32 |
| 20 HAE CHRY | 16 | 1 | 6.7\% | 1.07 | 4.13 | 16 | 0 | 0.00 |
| 21 HAL BIVI | 58 | 13 | 86.7\% | 3.87 | 2.53 | 8 | 0 | 1.19 |
| 22 HAL GARN | 4 | 2 | 13.3\% | 0.27 | 0.70 | 2 | 0 | 0.00 |
| 23 HAL MACU | 41 | 10 | 66.7\% | 2.73 | 2.60 | 8 | 0 | 1.58 |
| 24 HAL PICT | 1 | 1 | 6.7\% | 0.07 | 0.26 | 1 | 0 | 0.00 |
| 25 HAL POEY | 12 | 6 | 40.0\% | 0.80 | 1.08 | 3 | 0 | 0.50 |
| 26 HAL RADI | 24 | 10 | 66.7\% | 1.60 | 1.50 | 5 | 0 | 0.86 |
| 27 HOL ADSC | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 28 HOL RUFU | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 29 HOL TRIC | 2 | 2 | 13.3\% | 0.13 | 0.35 | 1 | 0 | 0.00 |
| 30 KYP SECT | 1 | 1 | 6.7\% | 0.07 | 0.26 | 1 | 0 | 0.00 |
| 31 LAC BICA | 1 | 1 | 6.7\% | 0.07 | 0.26 | 1 | 0 | 0.00 |
| 32 LAC POLY | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 33 LAC TRIQ | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 34 LUT APOD | 2 | 2 | 13.3\% | 0.13 | 0.35 | 1 | 0 | 0.00 |
| 35 LUT JOCU | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 36 LUT MAHO | 15 | 5 | 33.3\% | 1.00 | 2.36 | 9 | 0 | 12.00 |
| 37 MAL AURO | 1 | 1 | 6.7\% | 0.07 | 0.26 | 1 | 0 | 0.00 |
| 38 MAL PLUM | 8 | 3 | 20.0\% | 0.53 | 1.36 | 5 | 0 | 8.13 |

Appendix A.1. (cont.)


Appendix A.2. Outside damage area, REGINA grounding, Mona Island, Puerto Rico.

| No. Species | Total Indiv. | Frequency$(\mathrm{N}=9)$ |  | Mean Abund. | Stand. <br> Dev. | Ran High | Low | Var/Mean Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 ABU SAXA | 4 | 1 | 11.1\% | 0.44 | 1.33 | 4 | 0 | 0.00 |
| 2 ACA BAHI | 387 | 9 | 100.0\% | 43.00 | 41.66 | 120 | 5 | 40.35 |
| 3 ACA COER | 26 | 6 | 66.7\% | 2.89 | 3.30 | 8 | 0 | 3.42 |
| 4 ADI VEXI | 1 | 1 | 11.1\% | 0.11 | 0.33 | 1 | 0 | 0.00 |
| 5 AMB PINO | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 6 AUL MACU | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 7 BOD RUFU | 6 | 4 | 44.4\% | 0.67 | 0.87 | 2 | 0 | 0.50 |
| 8 BOT LUNA | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 9 CAN PULL | 1 | 1 | 11.1\% | 0.11 | 0.33 | 1 | 0 | 0.00 |
| 10 CAR RUBE | 3 | 2 | 22.2\% | 0.33 | 0.71 | 2 | 0 | 1.50 |
| 11 CHA CAPI | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 12 CHA STRI | 1 | 1 | 11.1\% | 0.11 | 0.33 | 1 | 0 | 0.00 |
| 13 CHR CYAN | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 14 CHR MULT | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 15 EPI FULV | 23 | 5 | 55.6\% | 2.56 | 3.64 | 11 | 0 | 5.79 |
| 16 GNA THOM | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 17 GOB EVEL | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 18 HAE AURO | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 19 HAE CARB | 2 | 2 | 22.2\% | 0.22 | 0.44 | 1 | 0 | 0.00 |
| 20 HAE CHRY | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 21 HAL BIVI | 38 | 8 | 88.9\% | 4.22 | 3.42 | 10 | 0 | 2.49 |
| 22 HAL GARN | 5 | 2 | 22.2\% | 0.56 | 1.13 | 3 | 0 | 0.90 |
| 23 HAL MACU | 31 | 6 | 66.7\% | 3.44 | 4.36 | 14 | 0 | 5.74 |
| 24 HAL PICT | 22 | 3 | 33.3\% | 2.44 | 3.81 | 9 | 0 | 1.77 |
| 25 HAL POEY | 29 | 8 | 88.9\% | 3.22 | 2.11 | 7 | 0 | 1.06 |
| 26 HAL RADI | 21 | 9 | 100.0\% | 2.33 | 1.50 | 6 | 1 | 0.96 |
| 27 HOL ADSC | 1 | 1 | 11.1\% | 0.11 | 0.33 | 1 | 0 | 0.00 |
| 28 HOL RUFU | 1 | 1 | 11.1\% | 0.11 | 0.33 | 1 | 0 | 0.00 |
| 29 HOL TRIC | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 30 KYP SECT | 40 | 1 | 11.1\% | 4.44 | 13.33 | 40 | 0 | 0.00 |
| 31 LAC BICA | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 32 LAC POLY | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 33 LAC TRIQ | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 34 LUT APOD | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 35 LUT JOCU | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 36 LUT MAHO | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 37 MAL AURO | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 38 MAL PLUM | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |

Appendix A. 2. (cont.)


Appendix A.3. Inshore control area, REGINA grounding, Mona Island, Puerto Rico.

| No. Species | Total Frequency <br> Indiv. (N=6) |  |  | Mean Abund. | Stand. Dev. | Range High | Var/Mean |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Low |  |  | Ratio |
| 1 ABU SAXA | 3 | 1 | 16.7\% |  | 0.50 | 1.22 | 3 | 0 | 0.00 |
| 2 ACA BAHI | 47 | 6 | 100.0\% | 7.83 | 7.63 | 23 | 2 | 7.43 |
| 3 ACA COER | 40 | 5 | 83.3\% | 6.67 | 5.57 | 16 | 0 | 3.83 |
| 4 ADI VEXI | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 5 AMB PINO | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 6 AUL MACU | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 7 BOD RUFU | 4 | 3 | 50.0\% | 0.67 | 0.82 | 2 | 0 | 0.50 |
| 8 BOT LUNA | 1 | 1 | 16.7\% | 0.17 | 0.41 | 1 | 0 | 0.00 |
| 9 CAN PULL | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 10 CAR RUBE | 5 | 3 | 50.0\% | 0.83 | 0.98 | 2 | 0 | 0.40 |
| 11 CHA CAPI | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 12 CHA STRI | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 13 CHR CYAN | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 14 CHR MULT | 62 | 2 | 33.3\% | 10.33 | 16.32 | 36 | 0 | 4.84 |
| 15 EPI FULV | 4 | 2 | 33.3\% | 0.67 | 1.03 | 2 | 0 | 0.00 |
| 16 GNA THOM | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 17 GOB EVEL | 4 | 1 | 16.7\% | 0.67 | 1.63 | 4 | 0 | 0.00 |
| 18 HAE AURO | 15 | 1 | 16.7\% | 2.50 | 6.12 | 15 | 0 | 0.00 |
| 19 HAE CARB | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 20 HAE CHRY | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 21 HAL BIVI | 10 | 4 | 66.7\% | 1.67 | 1.86 | 4 | 0 | 1.80 |
| 22 HAL GARN | 5 | 2 | 33.3\% | 0.83 | 1.60 | 4 | 0 | 5.40 |
| 23 HAL MACU | 9 | 4 | 66.7\% | 1.50 | 1.38 | 3 | 0 | 0.61 |
| 24 HAL PICT | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 25 HAL POEY | 2 | 1 | 16.7\% | 0.33 | 0.82 | 2 | 0 | 0.00 |
| 26 HAL RADI | 17 | 5 | 83.3\% | 2.83 | 1.60 | 4 | 0 | 0.28 |
| 27 HOL ADSC | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 28 HOL RUFU | 1 | 1 | 16.7\% | 0.17 | 0.41 | 1 | 0 | 0.00 |
| 29 HOL TRIC | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 30 KYP SECT | 26 | 3 | 50.0\% | 4.33 | 8.73 | 22 | 0 | 31.00 |
| 31 LAC BICA | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 32 LAC POLY | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 33 LAC TRIQ | 1 | 1 | 16.7\% | 0.17 | 0.41 | 1 | 0 | 0.00 |
| 34 LUT APOD | 2 | 2 | 33.3\% | 0.33 | 0.52 | 1 | 0 | 0.00 |
| 35 LUT JOCU | 2 | 2 | 33.3\% | 0.33 | 0.52 | 1 | 0 | 0.00 |
| 36 LUT MAHO | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 37 MAL AURO | 2 | 1 | 16.7\% | 0.33 | 0.82 | 2 | 0 | 0.00 |
| 38 MAL PLUM | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |

Appendix A. 3. (cont.)


Appendix A.4. Offshore control area, REGINA grounding, Mona Island, Puerto Rico.

| No. Species | Total Frequency <br> Indiv. ( $\mathrm{N}=11$ ) \% |  |  | Mean Abund. | Stand. <br> Dev. | Range High | Low | Mean Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 ABU SAXA | 39 | 1 | 9.1\% | 3.55 | 11.76 | 39 | 0 | 0.00 |
| 2 ACA BAHI | 97 | 11 | 100.0\% | 8.82 | 5.90 | 23 | 3 | 3.94 |
| 3 ACA COER | 91 | 11 | 100.0\% | 8.27 | 10.06 | 36 | 1 | 12.24 |
| 4 ADI VEXI | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 5 AMB PINO | 1 | 1 | 9.1\% | 0.09 | 0.30 | 1 | 0 | 0.00 |
| 6 AUL MACU | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 7 BOD RUFU | 17 | 7 | 63.6\% | 1.55 | 2.07 | 7 | 0 | 2.99 |
| 8 BOT LUNA | 1 | 1 | 9.1\% | 0.09 | 0.30 | 1 | 0 | 0.00 |
| 9 CAN PULL | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 10 CAR RUBE | 16 | 7 | 63.6\% | 1.45 | 1.81 | 6 | 0 | 2.23 |
| 11 CHA CAPI | 4 | 2 | 18.2\% | 0.36 | 0.81 | 2 | 0 | 0.00 |
| 12 CHA STRI | 5 | 4 | 36.4\% | 0.45 | 0.69 | 2 | 0 | 0.55 |
| 13 CHR CYAN | 2 | 1 | 9.1\% | 0.18 | 0.60 | 2 | 0 | 0.00 |
| 14 CHR MULT | 256 | 9 | 81.8\% | 23.27 | 19.60 | 60 | 0 | 13.51 |
| 15 EPI FULV | 34 | 10 | 90.9\% | 3.09 | 1.64 | 6 | 0 | 0.59 |
| 16 GNA THOM | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 17 GOB EVEL | 2 | 1 | 9.1\% | 0.18 | 0.60 | 2 | 0 | 0.00 |
| 18 HAE AURO | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 19 HAE CARB | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 20 HAE CHRY | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 21 HAL BIVI | 25 | 6 | 54.5\% | 2.27 | 2.57 | 7 | 0 | 1.66 |
| 22 HAL GARN | 54 | 11 | 100.0\% | 4.91 | 5.28 | 20 | 1 | 5.68 |
| 23 HAL MACU | 18 | 9 | 81.8\% | 1.64 | 1.21 | 4 | 0 | 0.61 |
| 24 HAL PICT | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 25 HAL POEY | 1 | 1 | 9.1\% | 0.09 | 0.30 | 1 | 0 | 0.00 |
| 26 HAL RADI | 38 | 10 | 90.9\% | 3.45 | 2.81 | 9 | 0 | 2.11 |
| 27 HOL ADSC | 1 | 1 | 9.1\% | 0.09 | 0.30 | 1 | 0 | 0.00 |
| 28 HOL RUFU | 4 | 3 | 27.3\% | 0.36 | 0.67 | 2 | 0 | 0.92 |
| 29 HOL TRIC | 12 | 6 | 54.5\% | 1.09 | 1.22 | 3 | 0 | 0.73 |
| 30 KYP SECT | 36 | 1 | 9.1\% | 3.27 | 10.85 | 36 | 0 | 0.00 |
| 31 LAC BICA | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 32 LAC POLY | 1 | 1 | 9.1\% | 0.09 | 0.30 | 1 | 0 | 0.00 |
| 33 LAC TRIQ | 2 | 2 | 18.2\% | 0.18 | 0.40 | 1 | 0 | 0.00 |
| 34 LUT APOD | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 35 LUT JOCU | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 36 LUT MAHO | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 37 MAL AURO | 0 | 0 | 0.0\% | 0.00 | NA | NA | NA | NA |
| 38 MAL PLUM | 2 | 1 | 9.1\% | 0.18 | 0.60 | 2 | 0 | 0.00 |

Appendix A. 4. (cont.)


# Chapter 2. A limited survey of reef fish abundance and species composition at the proposed Aquarius site, Conch Reef, Florida. 

Douglas E. Harper, James A. Bohnsack, and Stephania Bolden

A reef fish survey of Conch Reef was conducted by members of the Reef Resources Team, Miami Laboratory, Southeast Fisheries Center on June 7, 1991. The survey utilized standardized visual sampling methods (Bohnsack and Bannerot, 1986). Fishes observed within 7.5 m of a stationary SCUBA diver were censused at randomly selected locations in the reef area adjacent to the proposed site of the Aquarius Underwater Habitat. Data collected provide fish species presence, abundance, frequency, and average size and size range. This information should be of interest to researchers desiring to conduct studies on reef fishes using the underwater habitat as a base for operations.

Six visual samples were performed by three observers at a mean depth of 53.3 feet (range $=49$ to 58 feet). A summary of survey results is presented in Table 1. A total of 1,079 individual fishes representing 53 species ( 17 families) were recorded during the six samples. The mean number of fish per sample was 179.8 and the mean number of species per sample was 24.8 . Five species were observed in all samples. These fishes along with total number observed were: bicolor damselfish, Pomacentrus partitus ( $\mathrm{n}=466$ ); bluehead, Thalassoma bifasciatum $(\mathrm{n}=160)$; blue chromis, Chromis cyaneus $(\mathrm{n}=$ 91); redband parrotfish, Sparisoma aurofrenatum (n = 32); and ocean surgeon, Acanthurus bahianus (n $=30$ ). In addition to the fishes recorded during the regular 5 minute observational interval, two species; bluelip parrotfish, Cryptotomus roseus and tobaccofish, Serranus tabacarius; were observed during the enumeration phase of the sampling procedure.

Reef fish populations demonstrate a high degree of variability both temporally and spatially. Although limited in scope, the results of this survey indicate that the reef fish fauna near the proposed site of the Aquarius Underwater Habitat is abundant and complex. Additional studies should be conducted to further quantitatively assess the dynamics of fish populations near this site.

## Acknowledgements

We thank Dave Ward, Glen Taylor, and Tom Potts at the Key Largo facility of the National Undersea Research Program for their excellent assistance and support provided during the field work involved in this study. This paper was previously unpublished Miami Laboratory Contribution MIA-90/91-60.

## Literature Cited

Bohnsack, J.A. and S.P. Bannerot. 1986. A stationary visual census technique for quantitatively assessing community structure of coral reef fishes. NOAA Technical Report 41, 15 pp .

Table 1. Summary of reef fish visual censusing at Conch Reef on June 7, 1991


# Chapter 3. Protected and unprotected reefs in John Pennekamp Coral Reef State Park, Florida: a comparative analysis 1992-1995. 

David B. McClellan, James A. Bohnsack, Douglas E. Harper, and Stephania K. Bolden

## Introduction

Coral reefs are one of the Earth's most complex ecosystems and an important resource for commercial, recreational, scientific, and educational use. In recent years these ecosystems have received increased exploitation and usage. Fishing impacts reef structure and fauna. Many reef fish stocks are fully or over exploited (Plan Development Team, 1990). The effects of fishing on reef fish populations can only be effectively evaluated by comparing areas with no fishing with fished areas. For example, rapid biomass build-up on the Caribbean Saba reef has recently been quantified following the establishment of a marine sanctuary (Roberts, 1995). Increases in size and number of fishery species in protected marine areas have been identified (reviews by Roberts and Polunin, 1991; Dugan and Davis, 1993) in many studies.

Marine reserves, areas with no consumptive usage, have been proposed as a viable management measure to protect reef fish stocks and increase net yield (Plan Development Team, 1990). The newly mandated Florida Keys National Marine Sanctuary (FKNMS) is a holistic approach to managing fisheries - one which protects the ecosystem rather than an individual species. In order to wisely manage coral reef resources and evaluate marine reserves, there is need for monitoring and comparing fish populations in areas with and without fishing activity.

The Reef Resources Team of the Southeast Fisheries Science Center, National Marine Fisheries Service, NOAA (SEFSC/NMFS/NOAA) has developed resource survey techniques to provide baseline data on reef fish abundance and composition for long-term resource monitoring. The method uses standard, non-destructive, fishery independent, visual sampling methodology (Bohnsack and Bannerot, 1986). The method has been used by the National Marine Fisheries Service, the National Park Service, and agencies of various governments. The method also has been used extensively on reefs in southeastern Florida, including Biscayne National Park, Key Largo National Marine Sanctuary, Looe Key National Marine Sanctuary, and the Dry Tortugas.

This status report is the fourth of a series which have been submitted yearly to the John Pennekamp Coral Reef State Park (JPCRSP) complying with the statement of work in the Memorandum of Understanding between the Florida Department of Environmental Protection, Division of Recreation and Parks (FDEP) and SEFSC/NMFS/NOAA.

## Purpose

The ultimate purpose of this research is to evaluate the effects of fisheries on coral reef fish populations. The specific purpose of this research is to evaluate the effect of prohibiting fishing in small areas by monitoring the reef fish population at reefs "open" and "closed" to fishing. Baseline data for subsequent fluctuations in fish composition are being collected and future monitoring may determine if closing small areas affect fish size structure and abundance. This report summarizes fish populations, size structure and abundance at reefs in JPCRSP from 15 May 1992 to 27 June 1995.

## Methods

This study involved monitoring of reef fishes on six small patch reefs within JPCRSP. The reef environment in JPCRSP has been described by Jaap (1984) and the study sites by Bohnsack and Harper (1992), McClellan et al. (1993), and Bolden (1994). Two patch reefs were closed to public access (protected) in August 1991 while nearby patch reefs were open to public access for fishing
and diving (unprotected). Study sites included protected and unprotected patch reefs at Basin Hill Shoals $(\mathrm{BC}=$ closed, $\mathrm{BO}=$ open, and $\mathrm{BN}=$ new open $)$ and Mosquito $\mathrm{Bank}(\mathrm{MC}=$ closed, $\mathrm{MO}=$ open, and $\mathrm{MN}=$ new open) (Fig. 1).

Initially (May 1992, November 1992, and May 1993) the survey assessed two patch reefs per area: one "open" to fishing and one "closed." An additional "new" site which was open to fishing was added in May 1994 to equate for large areal differences between the protected and unprotected reefs. Areal coverage of censused reefs was calculated by a planimeter on aerial maps and is presented with mean depth in Table 1.

Annual sampling at each reef site was conducted during the Spring (April-June) between May 15, 1992 and June 27, 1995. A Fall survey was conducted during October-November 1992. An average 12 samples per reef (range: 11-17) were assessed (Table 2). Sampling used standardized stationary diver, non-destructive, fishery independent, visual sampling methods as described in Bohnsack and Bannerot (1986).

Two-sample t-tests were utilized to examine both the mean number of species per sample and mean abundance in 1992 and 1993. Comparisons were made between adjacent patch reefs only. Mean abundances were used to detect potential differences between study sites in 1994 and 1995. Both overall similarity and paired inter-reef differences were sought. Analysis of variance first examined mean abundance at all reefs. Second, inter-reef differences were sought by Fisher's least-significant-differences test. This comparison examines type I comparison wise error rates (not experimental).

An index of relative abundance was calculated for each species in order to provide a standard for comparison of species composition between study patch reefs (Greenfield and Johnson, 1989). This index gave equal weight to abundance and frequency-of-occurrence. The index of relative dominance (IRD), of a species, $i$, was calculated by:

$$
I R D_{i}=\left(R A_{i} * F_{i}\right)
$$

where $R A$ is relative abundance and $R F$ is relative frequency for species $i$. The relative abundance (RA) of species $i$ is the total number of individuals of species $i$ expressed as a percentage of the sum of the total individuals censused. The relative frequency (RF) is the number of times a species was observed in a census sample expressed as a percentage of the sum of the total number of samples.

Numerical classification technique (cluster analysis) were used to compare similarity based upon species assemblages (mean species abundance) for study reefs and sampling periods. Similarity relationships were depicted using dendrograms generated by an interactive computer program which analyzes community data from ecological studies (Wolfe and Chester, 1991). Similarity was measured by the Bray-Curtis index using a flexible sorting strategy with $B=-0.25$ (Clifford and Stephenson, 1975). Additionally, only those species with overall IRD values greater than 0.5 were included in these analyses because rare species provide little information on the basic patterns of community structure (Ludwig and Reynolds, 1988; Sedberry and von Dolah, 1984).

Biomass estimates were calculated for each species using length-weight relationships reported by Bohnsack and Harper (1988) and unpublished data (SEFSC/NMFS/NOAA, Miami Lab). Bohnsack et al. (1994) reported that in 1992, reef fishes comprised $28 \%$ of commercial landings in the Florida Keys of which $56 \%$ were dominated by snappers, groupers, grunts and hogfish. In addition, porgies comprised an important component of recreational landings. Economically important families and groups were then examined for shifts in biomass and individual numbers as a result of fishing pressure and to identify potential differences between areas opened and closed to fishing (Figs. 3 and 4).

## Results and Discussion

A total of 24,338 fishes representing 30 families, 50 genera and 109 species were observed in 298 visual censuses conducted at the six study patch reefs within JPCRSP from May 15, 1992 to June 27, 1995 (Table 3). Seven species (including one unidentified species) accounted for nearly $70 \%$ total number of fishes observed. These fishes along with percentage of total individuals censused were: white grunt, Haemulon plumeri, $16.9 \%$; tomtate, H. aurolineatum, $14.5 \%$; striped parrotfish, Scarus croicensis, $11.9 \%$; unidentified specie, $8.8 \%$; gray snapper, Lutjanus griseus, $6.4 \%$; bluestriped grunt, H. sciurus, $5.6 \%$; and yellowtail snapper, Ocyurus chrysurus, $5.2 \%$. Five species accounted for nearly $60.0 \%$ of total biomass: great barracuda, Sphyraena barracuda, ( $432.5 \mathrm{~kg}, 24.5 \%$ ) ; gray snapper, ( $196.8 \mathrm{~kg}, 11.2 \%$ ); yellow jack, Caranx bartholomaei, (161.9kgf $9.2 \%$ ) ; gray angelfish, Pomacanthus arcuatus ( $161.6 \mathrm{~kg}, 9.2 \%$ ) and white grunt, ( $102.0 \mathrm{~kg}, 5.8 \%$ ).

JPCRSP is located in close proximity to Hawk Channel which is an area of high mixing caused by wind and current, coupled with significant boat traffic. Hawk Channel probably affects water area clarity within sampling sites. Sample depths ranged from 1.2-3.5m with a mean of 2.3 m (Table 1). Visibility ranged from $4-8.5 \mathrm{~m}$ during sampling days which may have affected the number of fishes assessed.

Total abundance, mean abundance/standard deviation per sample, frequency-of-occurrence, percent frequency-of-occurrence, biomass, and mean, minimum, and maximum sizes were calculated for each species by individual study reefs (Tables 4-9). The total number of species for pooled samples varied by reef, ranging from 51 (BN) to 73 (BC) (Table 2). Two species, white grunt and tomtate, consistently ranked first and second by mean abundance at four of the six study reefs and no species was observed in all samples at any individual reef.

Mean number of species per sample was highest at BO (15.6, Spring 94) and lowest at MN (8.8, Spring 95) per sample for any sampling unit (site and sample) during the study (Fig. 2 and Table 2). Significant ( $\mathrm{P}>0.05$ ) differences in number of species per sample were identified by ANOVA during 1992 with MO greater than MC and BO greater than BC in 1993. Overall the species richness per count seemed to remain constant at Basin Hill (open > closed) with a slight decrease in 1995. At Mosquito Bank there was an increase between 1992 and 1994 (open > closed) with a decrease also in 1995.

Mean fish abundance per sampling unit (site and season) was highest for BC (246.1, Spring95) and lowest for MN (22.3 Spring 95) (Table 2 and Fig. 2). T-test comparison of mean abundances showed significant $(\mathrm{P}>0.05)$ differences per sample during 1992 with BC over BO. 1993 data revealed MO supported significantly greater mean abundances than nearby MC. ANOVA showed no reefs were significantly different from one another during 1994 but both "new open" study reefs were significantly different ( $\mathrm{P}>0.05$ ) from the other sites (open and closed) in 1995; BN had significantly greater mean abundance and MN had significantly less. The new open to fishing reefs was added to compensate for the areal differences between the closed and open fishing sites. One would expect them to have similar effects relative to mean abundance, but obviously other factors are influencing mean fish abundance. $\mathrm{R}^{2}$ values accounted for less than $1 \%$ of mean abundance variability in these analyses ( $1994 \mathrm{R}^{2}=0.005$; $1995 \mathrm{R}^{2}=0.004$ ).

Fishes were combined by family (snappers, Lutjanidae; groupers, Serranidae; grunts and porgies, Haemulidae and Sparidae; surgeonfish, Acanthuridae; hogfish, Labridae; and parrotfish, Scaridae) to assess biomass changes in number/sample and biomass/sample changes per year and reef (Figs. 2 and 3). Commercially and recreationally important groups (snappers, groupers, hogfish, jacks (Carangidae), grunts and porgies, permit, Trachinotus falcatus, and barracuda) and herbivores (surgeonfish and parrotfish) were also combined to analyze any differences (Figs. 2 and 3). Roberts (1995) showed significant increases in fish biomass per count at the family and group level after the

Saba Bay (Caribbean) closure.
Basin Hill showed a significant increase in total biomass per sample from 1993-1995 for the closed site and consistently ranked higher than the open or new site except for 1993 (Fig. 2). This marked increase was also apparent in the combined commercial and the combined herbivore species groups (Fig. 2). At the family level, the snappers showed the most obvious increase. Parrotfishes had a consistently higher biomass at closed reefs but the means did not increase.

Total biomass per sample was greater at closed reefs except for 1994 at Mosquito Bank but did not show a marked increase in total biomass over the four years (Fig. 2). The commercial group showed a great increase from 1992-1993 at the closed site, the open sites greater in 1994-1995 (Fig. 3). Snappers increased dramatically from 1992-1994 at the open sites and were consistently greater than the closed sites which only differed slightly over the time period (Fig. 3).

Paired reef comparisons by year (Fisher's least-significant-differences test) revealed significant differences ( $\mathrm{P}>0.05$ ) in abundance between specific reefs. BN was significantly different from MO in both 1994 and 1995. In 1994, MC was significantly different from both BN and BO. 1995 censuses also revealed that MN was statistically different from both BC and BN; BN also separated from BO.

The striped parrotfish and white grunt consistently ranked among the top four IRD fishes at all Mosquito Bank study patch reefs, and at ten of twelve Basin Hill study patch reefs (Table 10). Many of the commercially and recreationally valuable foodfishes - as managed by the South Atlantic Fishery Management Council - were at some period or site ranked among the top 10 IRD species. These important fishes include: white grunt, tomtate, gray snapper, yellowtail snapper, bluestriped grunt, French grunt (H. flavolineatum), cottonwick (H. melanurum), bar jack (Caranx ruber), lane snapper (Lutjanus synagris), hogfish (Lachnolaimus maximus), and yellow jack (Caranx bartholomaei).

Bray-Curtis similarity cluster analysis showed a separation between two groups of study reefs when pooled mean abundances for 48 fishes with IRD scores greater than 0.5 were analyzed (Fig. 5). Mosquito Bank open clustered with the three Basin Hill sites (Group A), while Mosquito Bank closed and new clustered separately (Group B). When sampling efforts were further partitioned into units by season and study reef, again two group complexes could be identified (Fig. 6). Group complex A was composed of primarily Basin Hill sampling units, with two of the three 1992 Mosquito Bank sampling units (MC92s and MC92F), and represented the majority of Basin Hill sampling efforts ( 10 of $12,83 \%$ ). Group complex B was composed only of Mosquito Bank sampling units. This similarity analysis suggests that study reefs, and not season or open or closed to fishing/diving, plays a major role in influencing fish community structure, although the relative contribution of each factor cannot be determined based on sampling conducted to date. A longer time series of data is needed to test the persistence of these patterns impacting fish community structure relationships. Future monitoring will determine if closing small areas affect fish abundance patterns by testing the hypothesis that changes observed at closed reefs are no different than control reefs.

## Summary

Reef fish were assessed at six patch reefs in 298 censuses at John Pennekamp Coral Reef State Park for the time period of May 15, 1992 to June 27, 1995. The rich ichthyofauna is comprised of 30 families, 50 genera, and 109 species of 24,338 observed fishes. It is premature to discuss community patterns. Based on these data, the mean number of species appears to be correlated with the mean number of individuals observed. Within this area, north (Basin Hill) - south (Mosquito

Bank) geographic differences, reef structure (size and openness), average visibility and usage patterns (Mosquito Bank has higher fishing/diving pressure) may impact on fish assemblages. Clearly, Basin Hill had greater mean number of individuals, mean number of species, and mean biomass than Mosquito Bank. Year class differences relative to environmental factors such as Florida Bay may have also affected the study.

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Table 1. Area and mean depth (meters) of censused patch reef in Basin Hill ( BH ) and Mosquito Bank (MB), John Pennekamp Coral Reef State Park, Florida. Reefs are characterized by open (O) or closed (C) to fishing, with numbers designating historic (1) or recent reef addition (2) to survey.

|  |  |  |
| :--- | :---: | :---: |
| REEF | HECTARES | XDEPTH |
| BH01 | 0.125 | 2.26 |
| BHC | 0.449 | 2.27 |
| BHO2 | 0.267 | 2.16 |
|  |  |  |
| MBO1 | 0.757 | 2.55 |
| MBC | 0.169 | 2.52 |
| MB02 | 0.206 | 2.79 |

Table 2a. Comparison of species richness for open, new open, and closed study patch reefs in John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995.

| Season | Patch Reef | Cumulative Number of Species | Number of Samples | Mean \# of Species per sample | Std, Dev. | Std, <br> Error | Two-sample t-test |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | t-statistic | df | p |
| Spring 92 | Basin Hill Closed | 47 | 17 | 14.23 | 4.52 | 1.10 |  |  |  |
|  | Basin Hill Open | 41 | 16 | 12.94 | 2.97 | 0.74 | 0.96 | 31.00 | ns |
|  | Mosquito Bank Closed | 39 | 13 | 9.77 | 2.35 | 0.65 |  |  |  |
|  | Mosquito Bank Open | 38 | 12 | 13.75 | 3.60 | 1.04 | -3.25 | 18.70 | p<01 |
| Fall 92 |  |  |  |  |  |  |  |  |  |
|  | Basin Hill Closed | 43 | 12 | 13.17 | 2.69 | 0.78 |  |  |  |
|  | Basin Hill Open | 38 | 12 | 14.67 | 2.39 | 0.69 | -1.44 | 22.00 | ns |
|  | Mosquito Bank Closed | 40 | 11 | 10.82 | 1.72 | 0.52 |  |  |  |
|  | Mosquito Bank Open | 42 | 12 | 12.17 | 1.95 | 0.56 | 1.75 | 21.00 | ns |
| Spring 93 | Basin Hill Closed | 40 | 12 | 11.75 | 2.38 | 0.69 |  |  |  |
|  | Basin Hill Open | 41 | 12 | 14.92 | 2.02 | 0.58 | -3.51 | 22.00 | p< 01 |
|  | Mosquito Bank Closed | 42 | 11 | 13.09 | 3.33 | 1.00 |  |  |  |
|  | Mosquito Bank Open | 40 | 12 | 14,00 | 1,71 | 0.49 | 0.81 | 14.60 | ns |
| Spring 94 | Basin Hill Closed | 43 | 12 | 13.33 | 3.45 | 0.99 |  |  |  |
|  | Basin Hill Open | 40 | 12 | 15.58 | 2.57 | 0.74 |  |  |  |
|  | Basin Hill New | 37 | 12 | 14.08 | 3.55 | 1.03 |  |  |  |
|  | Mosquita Bank Closed | 47 | 12 | 13.33 | 3.45 | 0.99 |  |  |  |
|  | Mosquito Bank Open | 42 | 12 | 12.92 | 2.02 | 0.58 |  |  |  |
|  | Mosquito Bank New | 45 | 12 | 14.00 | 3.91 | 1.13 |  |  |  |
| Spring 95 | Basin Hill Closed | 48 | 12 | 11.50 | 3.71 | 1.07 |  |  |  |
|  | Basin Hill Open | 41 | 12 | 12.08 | 3.48 | 1.00 |  |  |  |
|  | Basin Hill New | 46 | 12 | 14.75 | 2.63 | 0.76 |  |  |  |
|  | Mosquito Bank Closed | 37 | 12 | 10.50 | 3.21 | 0.93 |  |  |  |
|  | Mosquito Bank Open | 38 | 14 | 9.21 | 3.04 | 0.81 |  |  |  |
|  | Mosquito Bank New | 38 | 12 | 8.75 | 2.60 | 0.75 |  |  |  |
| Combined (AlL POOLED) | Basin Hill Closed | 73 | 65 | 12.91 : | 3.59 | 0.45 | \% |  |  |
|  | Basin Hill Open | 63 | 64 | 13.97. | 2.97 | 0.37 |  |  |  |
|  | Basin Hill New | 51 | 24. | 14.42 . | 3.08 | 0.63 |  |  |  |
|  | Mosquito Bank Closed | 70 | 59 | 11.46 | 3.15 | ... 0.41 | A. |  |  |
|  | Mosquito Bank Open. | 69 | -62 | 1069 | 5.52 | - 0.70 |  |  |  |
|  | Mosquito Bank New . | 56 | - $\times 24$ | -11,38 | 4.21 | 0.86 |  |  |  |

Table 2 b . Comparison of fish abundance for open, new open, and closed study patch reefs in John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995.

| Season | Patch Reef | Total Number of Individuals | Mean <br> Number of Abundance <br> Samples per sample |  | Std. Dev. | Std. <br> Error | Two-sample t-test |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | t-statistic |  | df | p |
| Spring 92 | Basin Hill Closed | 1905 | 17 | 112.06 |  | 66.43 | 16.11 |  |  |  |
|  | Basin Hill Open | 1171 | 16 | 73.19 | 28.32 | 7.08 | 2.21 | 21.90 | p< 05 |
|  | Mosquito Bank Closed Mosquito Bank Open | $\begin{array}{r} 935 \\ 1404 \end{array}$ | 13 12 | $\begin{array}{r} 71.92 \\ 117.00 \end{array}$ | $\begin{aligned} & 127.20 \\ & 152.10 \end{aligned}$ | $\begin{aligned} & 35.28 \\ & 43.90 \end{aligned}$ | 0.81 | 23.00 | ns |
| Fall 92 | Basin Hill Closed | 768 | 12 | 64.00 | 17.82 | 5.14 |  |  |  |
|  | Basin Hill Open | 1046 | 12 | 87.17 | 36.31 | 10.48 | -1.98 | 16.00 | ns |
|  | Mosquito Bank Closed | 599 | 14 | 54.45 | 27.39 | 8.26 |  |  |  |
|  | Mosquito Bank Open | 978 | 12 | 81.50 | 44.60 | 12.87 | 1.73 | 21.00 | ns |
| Spring 93 | Basin Hill Closed | 913 | 12 | 67.75 | 15.44 | 4.46 |  |  |  |
|  | Easin Hill Open | 1076 | 12 | 89.67 | 35.95 | 10.38 | -1.94 | 14.90 | ns |
|  | Mosquito Bank Closed | 526 | 11 | 47.82 | 22.70 | 6.84 |  |  |  |
|  | Mosquito Bank Open | 1236 | 12 | 103.00 | 40.25 | 11.62 | 4.09 | 17.60 | p<. 01 |
| Spring 94 | Basin Hill Closed | 907 | 12 | 75.58 | 43.07 | 12.43 |  |  |  |
|  | Basin Hill Open | 1158 | 12 | 96.50 | 55.18 | 15.93 |  |  |  |
|  | Basin Hill New | 1197 | 12 | 99.75 | 53.37 | 15.41 |  |  |  |
|  | Mosquito Bank Closed | 795 | 12 | 66.25 | 25.64 | 7.40 |  |  |  |
|  | Mosquito Bank Open | 895 | 12 | 74.08 | 20.33 | 5.87 |  |  |  |
|  | Mosquito Bank New | 1150 | 12 | 95.83 | 58.70 | 16.95 |  |  |  |
| Spring 95 | Basin Hill Closed | 2953 | 12 | 246.08 | 574.35 | 165.80 |  |  |  |
|  | Basin Hill Open | 526 | 12 | 43.83 | 14.06 | 4.06 |  |  |  |
|  | Basin Hill New | 1029 | 12 | 85.75 | 39.41 | 11.38 |  |  |  |
|  | Mosquito Bank Closed | 578 | 12 | 48.17 | 18.77 | 5.42 |  |  |  |
|  | Mosquito Bank Open | 425 | 14 | 30.36 | 14.33 | 3.83 |  |  |  |
|  | Mosquito Bank New | 288 | 12 | 22.33 | 10.52 | 3.04 |  |  |  |
| Combined (ALi PơiEe) | Basin Hill Closed | 7346 | 65 | 113.02 | 250.26 | 31.05 |  |  |  |
|  | Basin Hill Open | 4213 | 64 | 77.77 | 39.35 | 4.92 |  |  |  |
|  | Basin Hill New | 2226 | 24 | 92.75 | 46.43 | 9.48 |  |  |  |
|  |  |  | 59. | 58.19. | 62.10 | 8.09 |  |  |  |
|  | Mosquito Eank Closed! | 3433 | 62 | 74.74 | 80.60 | 10.24 |  |  |  |
|  | Mosquito Bank Open | 4938 | 24 | 69.08 | 55.77 | 11.38 |  |  |  |
|  | Mosquito Bank New | 1418 |  |  |  |  |  |  |  |

Table 2c. Comparison of biomass for open, new open, and closed study patch reefs in John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995.

| $\frac{\text { Season }}{}$ | Patch Reef | Total <br> Biomass (g) | Number of Samples | Mean Biomass per sample | Std. <br> Dev. | Std. <br> Error | Two-sample t-test |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | t-statistic | df | p |
|  | Basin Hill Closed | 286467.3 | 17 | 16851.02 | 28222.30 | 6844.90 |  |  |  |
| Spring 92 | Basin Hill Open | 59538.8 | 16 | 3721.18 | 2792.90 | 677.40 |  |  |  |
|  | Mosquito Bank Closed Mosquito Bank Open | $\begin{aligned} & 33798.8 \\ & 26604.1 \end{aligned}$ | 13 12 | $\begin{aligned} & 2599.91 \\ & 2217.01 \end{aligned}$ | $\begin{aligned} & 3219.20 \\ & 2423.20 \end{aligned}$ | $\begin{aligned} & 892.90 \\ & 140.60 \end{aligned}$ |  |  |  |
| Fall 92 | Basin Hill Closed | 94349 | 12 | 7862.42 | 6551.30 | 1588.90 |  |  |  |
|  | Basin Hill Open | 28999 | 12 | 2416.58 | 1614.40 | 391.60 |  |  |  |
|  | Mosquito Bank Closed Mosquito Bank Open | $\begin{array}{r} 75583 \\ 29504.2 \end{array}$ | $\begin{aligned} & 11 \\ & 12 \end{aligned}$ | $\begin{aligned} & 6871.18 \\ & 2458.68 \end{aligned}$ | $\begin{array}{r} 10248.00 \\ 1415.00 \end{array}$ | $\begin{array}{r} 3089.90 \\ 150.60 \end{array}$ |  |  |  |
| Spring 93 | Basin Hill Closed | 65175.7 | 12 | 5431.31 | 3836.00 | 930.40 |  |  |  |
|  | Basin Hill Open | 85764 | 12 | 7147.00 | 8316.40 | 2017.00 |  |  |  |
|  | Mosquito Bank Closed | 69602.3 | 11 | 6327.48 | 5797.60 | 1748.00 |  |  |  |
|  | Mosquito Bank Open | 63156.2 | 12 | 5263.02 | 2866.10 | 827.40 |  |  |  |
| Spring 94 | Basin Hill Closed | 106136.7 | 12 | 8844.73 | 6758.80 | 1639.20 |  |  |  |
|  | Basin Hilll Open | 39551.7 | 12 | 3295.98 | 2170.60 | 526.50 |  |  |  |
|  | Basin Hill New | 98880.6 | 12 | 8240.05 | 5868.60 | 1423.30 |  |  |  |
|  | Mosquito Bank Closed | 25206.8 | 12 | 2100.57 | 1579.50 | 456.00 |  |  |  |
|  | Mosquito Bank Open | 72354.4 | 12 | 6029.53 | 6136.00 | 1771.30 |  |  |  |
|  | Mosquito Bank New | 109965.6 | 12 | 9163.80 | 7794.70 | 2250.10 |  |  |  |
| Spring 95 | Basin Hill Closed | 239945.6 | 12 | 19995.47 | 23315.70 | 5654.90 |  |  |  |
|  | Basin Hill Open | 52219.5 | 12 | 4351.63 | 4612.60 | 1118.70 |  |  |  |
|  | Basin Hill New | 65123.2 | 12 | 5426.93 | 4017.70 | 974.40 |  |  |  |
|  | Mosquito Bank Closed | 44640.4 | 12 | 3720.03 | 4287.4 | 1237.70 |  |  |  |
|  | Mosquito Bank Open | 16704.7 | 14 | 1193.19 | 933.80 | 249.60 |  |  |  |
|  | Mosquito Bank New | 57898.6 | 12 | 4824.88 | 4286.50 | 1237.40 |  |  |  |
| Combined (ALI PoOlED) | Basin Hill Closed | 792074.2 | 65 | 12185.76 | 18937.70 | 4593.10 |  |  |  |
|  | Basin Hill Open | 266073 | 64 | 4157.39 | 4766.30 | 1156.00 |  |  |  |
|  | Basin Hill New | 164003.9 | 24 | 6833.50 | 5222.00 | 1266.50 |  |  |  |
|  | Mosquito Bank Closed | 248831.4 | 59 | 4217.48 | 6004.80 | 781.80 |  |  |  |
|  | Mosquito Bank Open | 2083235 | 62 | 3360.06 | 3761.50 | 477.70 |  |  |  |
|  | Mosquito Bank New | 167864.2 | 24 | 6994.34 | 6653.7 | 1658.2 |  |  |  |

Table 3. Summary of all reef fish censused at John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995.

| SPECIES | COMMON NAME | NUMBER |  |  | Sample $\mathrm{N}=298$ |  | Number of species $=109$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total | Mean | Stand. | OCCU | ENCE |  | STH | cmi) | BIOMASS (gms) |
|  |  | Abund. | Abund. | Dev. | Freq. | \% | Mean | Min | Max | Total |
| Abudefduf saxatilis | Sergeant major | 684 | 2.30 | 8.05 | 84 | 28.2\% | 6.0 | 1 | 15 | 6,714.7 |
| Acanthurus bahianus | Ocean surgeon | 100 | 0.34 | 0.78 | 63 | 21.1\% | 7.5 | 2 | 30 | 2,693.3 |
| Acanthurus chirurgus | Doctorfish | 102 | 0.34 | 0.95 | 57 | 19.1\% | 10.0 | 3 | 30 | 4,938.9 |
| Acanthurus coeruleus | Blue tang | 249 | 0.84 | 1.52 | 138 | 46.3\% | 13.4 | 2 | 35 | 31,672.5 |
| Aluterus schoepfi | Orange filefish | 2 | 0.01 | 0.12 | 1 | 0.3\% | 34.0 | 34 | 34 | 722.0 |
| Aluterus scriptus | Scrawled filefish | 3 | 0.01 | 0.10 | 3 | 1.0\% | 42.7 | 23 | 60 | 2,443.8 |
| Anisotremus virginicus | Porkfish | 256 | 0.86 | 1.90 | 114 | 38.3\% | 6.6 | 1 | 30 | 6,756.6 |
| Archosargus rhomboidalis | Sea bream | 47 | 0.16 | 1.02 | 12 | 4.0\% | 17.7 | 10 | 30 | 8,066.7 |
| Aulostomus maculatus | Trumpetfish | 4 | 0.01 | 0.12 | 4 | 1.3\% | 31.5 | 30 | 33 | 312.1 |
| Bodianus pulchellus | Spotfin hogfish | 1 | 0.00 | 0.06 | 1 | 0.3\% | 8.0 | 8 | 8 | 8.3 |
| Bodianus rufus | Spanish hogfish | 1 | 0.00 | 0.06 | 1 | 0.3\% | 3.0 | 3 | 3 | 0.4 |
| Calamus calamus | Saucereye porgy | 25 | 0.08 | 0.31 | 22 | 7.4\% | 19.6 | 6 | 36 | 5,679.8 |
| Canthigaster rostrata | Sharpnose puffer | 5 | 0.02 | 0.13 | 5 | 1.7\% | 5.2 | 4 | 6 | 13.8 |
| Caranx bartholomaei | Yellow jack | 91 | 0.31 | 2.59 | 9 | 3.0\% | 41.7 | 10 | 65 | 161,902.5 |
| Caranx crysos | Blue runner | 2 | 0.01 | 0.12 | 1 | 0.3\% | 20.0 | 20 | 20 | 331.8 |
| Caranx ruber | Bar jack | 722 | 2.42 | 16.21 | 49 | 16.4\% | 15.2 | 3 | 50 | 79,392.9 |
| Chaetodon capistratus | Foureye butterflyfish | 92 | 0.31 | 0.61 | 74 | 24.8\% | 5.7 | 3 | 10 | 641.0 |
| Chaetodiperus faber | Atlantic spadefish | 28 | 0.09 | 0.58 | 13 | 4.4\% | 31.3 | 22 | 42 | 28,387.1 |
| Chaetodon ocellatus | Spotfin butterflyfish | 18 | 0.06 | 0.34 | 10 | 3.4\% | 9.3 | 6 | 14 | 495.8 |
| Chaetodon sedentarius | Reef butterflyfish | 2 | 0.01 | 0.08 | 2 | 0.7\% | 5.0 | 4 | 6 | 8.0 |
| Chaetodon striatus | Banded butterflyfish | 2 | 0.01 | 0.12 | 1 | 0.3\% | 12.0 | 12 | 12 | 108.3 |
| Coryphopterus dicrus | Colon goby | 17 | 0.06 | 0.34 | 11 | 3.7\% | 4.0 | 2 | 6 | 15.7 |
| Coryphopterus glaucofraenum | Bridled goby | 147 | 0.49 | 1.30 | 58 | 19.5\% | 3.6 | 1 | 6 | 102.2 |
| Coryphopterus personatus | Masked goby | 16 | 0.05 | 0.37 | 7 | 2.3\% | 2.7 | 2 | 3 | 4.3 |
| Coryphopterus species | Unknown goby | 6 | 0.02 | 0.22 | 3 | 1.0\% | 5.3 | 5 | 7 | 12.0 |
| Cryptotomus roseus | Bluelip parrotfish | 8 | 0.03 | 0.37 | 2 | 0.7\% | 4.8 | 4 | 5 | 58.9 |
| Diodon holocanthus | Balloonfish | 5 | 0.02 | 0.13 | 5 | 1.7\% | 15.0 | 14 | 17 | 532.1 |
| Diodon hystrix | Porcupinefish | 3 | 0.01 | 0.13 | 2 | 0.7\% | 20.0 | 10 | 25 | 1,721.0 |
| Diplodus holbrooki | Spottail pinfish | 1 | 0.00 | 0.06 | 1 | 0.3\% | 16.0 | 16 | 16 | 85.0 |
| Echeneis naucrates | Sharksucker | 1 | 0.00 | 0.06 | 1 | 0.3\% | 20.0 | 20 | 20 | 71.1 |
| Epinephelus adscensionis | Rock hind | 1 | 0.00 | 0.06 | 1 | 0.3\% | 20.0 | 20 | 20 | 124.1 |
| Epinephelus cruentatus | Graysby | 2 | 0.01 | 0.08 | 2 | 0.7\% | 13.0 | 11 | 15 | 66.8 |
| Epinephelus morio | Red grouper | 16 | 0.05 | 0.28 | 13 | 4.4\% | 28.3 | 16 | 50 | 7,230.9 |
| Epinephelus striatus | Nassau grouper | 1 | 0.00 | 0.06 | 1 | 0.3\% | 38.0 | 38 | 38 | 820.8 |
| Equetus punctatus | Spotted drum | 1 | 0.00 | 0.06 | 1 | 0.3\% | 6.0 | 6 | 6 | 2.7 |
| Gerres cinereus | Yellowfin mojarra | 42 | 0.14 | 0.92 | 14 | 4.7\% | 18.9 | 10 | 47 | 9,461.2 |
| Ginglymostoma cirratum | Nurse shark | 1 | 0.00 | 0.06 | 1 | 0.3\% | 15.0 | 15 | 51 | 915.2 |
| Gnatholepis thompsoni | Goldspot goby | 5 | 0.02 | 0.21 | 2 | 0.7\% | 3.6 | 2 | 5 | 3.1 |
| Gobiosoma oceanops | Neon goby | 81 | 0.27 | 0.81 | 43 | 14.4\% | 3.0 | 1 | 5 | 23.0 |
| Goby-like fish | Goby-like fish | 1 | 0.00 | 0.06 | 1 | 0.3\% | 4.0 | 4 | 4 | 0.6 |
| Gymnothorax funebris | Green moray | 1 | 0.00 | 0.06 | 1 | 0.3\% | 8.0 | 8 | 8 | 1.6 |
| Gymnothorax moringa | Spotted moray | 2 | 0.01 | 0.08 | 2 | 0.7\% | 35.0 | 30 | 40 | 161.3 |
| Haemulon aurolineatum | Tomtate | 3,529 | 11.84 | 28.56 | 134 | 45.0\% | 6.6 | 1 | 16 | 21,642.9 |
| Haemulon carbonarium | Caesar grunt | 6 | 0.02 | 0.20 | 4 | 1.3\% | 12.8 | 8 | 20 | 276.4 |
| Haemulon chrysargyreum | Smallmouth grunt | 2 | 0.01 | 0.08 | 2 | 0.7\% | 9.0 | 8 | 10 | 92.2 |
| Haemulon flavolineatum | French grunt | 398 | 1.34 | 4.69 | 78 | 26.2\% | 8.0 | 3 | 28 | 8,430.2 |
| Haemulon macrostomum | Spanish grunt | 10 | 0.03 | 0.24 | 7 | 2.3\% | 13.9 | 4 | 30 | 1,886.4 |
| Haemulon melanurum | Cottonwick | 80 | 0.27 | 2.48 | 4 | 1.3\% | 13.0 | 10 | 15 | 3,510.9 |
| Haemulon parra | Sailor's choice | 5 | 0.02 | 0.19 | 3 | 1.0\% | 17.4 | 14 | 31 | 793.1 |
| Haemulon plumieri | White grunt | 4,102 | 13.77 | 18.13 | 250 | 83.9\% | 9.9 | 2 | 30 | 102,008.2 |
| Haemulon sciurus | Bluestriped grunt | 1,380 | 4.63 | 7.68 | 198 | 66.4\% | 13.1 | 2 | 30 | 87,828.4 |
| Haemulon sp. | Unidentified grunt | 533 | 1.79 | 20.93 | 10 | 3.4\% | 3.2 | 3 | 7 | 289.5 |
| Halichoeres bivittatus | Slippery dick | 142 | 0.48 | 1.66 | 54 | 18.1\% | 7.0 | 3 | 15 | 745.0 |
| Halichoeres garnoti | Yellowhead wrasse | 6 | 0.02 | 0.16 | 5 | 1.7\% | 9.0 | 5 | 15 | 89.5 |
| Halichoeres maculipinna | Clown wrasse | 14 | 0.05 | 0.26 | 11 | 3.7\% | 8.9 | 4 | 14 | 177.2 |
| Halichoeres radiatus | Puddingwife | 3 | 0.01 | 0.13 | 2 | 0.7\% | 17.3 | 4 | 40 | 968.7 |
| Hemipteronotus martinicensis | Rosy razorfish | 1 | 0.00 | 0.06 | 1 | 0.3\% | 6.0 | 6 | 6 | 4.5 |
| Holacanthus bermudensis | Blue angelfish | 81 | 0.27 | 0.69 | 62 | 20.8\% | 14.5 | 3 | 36 | 11,758.9 |
| Holacanthus ciliaris | Queen anglefish | 43 | 0.14 | 0.49 | 32 | 10.7\% | 12.7 | 3 | 30 | 4,171.8 |
| Hyploplectrus nigricans | Black hamlet | 2 | 0.01 | 0.08 | 2 | 0.7\% | 4.0 | 4 | 4 | 0.4 |
| Hypoplectrus puella | Barred hamlet | 2 | 0.01 | 0.08 | 2 | 0.7\% | 8.0 | 6 | 10 | 3.3 |

Table 3．（cont．）

| SPECIES | COMMON NAME | NTMMBER Sample $\mathrm{N}=298$ Number of species $=109$ |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total Abund． | Mean Abund． | Stand． Dev． | OCCURENCE |  | LENGTH（cm） |  |  | $\frac{\text { BIOMASS } \text { (gms) }}{\text { Total }}$ |
|  |  |  |  |  | Freq． | \％ | Mean | Min | Max |  |
|  |  | ＝＝ |  |  |  |  |  |  |  | ニニニニニニ＝＝＝＝＝＝＝ |
| Hypoplectrus unicolor | Butter hamlet | 21 | 0.07 | 0.27 | 20 | 6．7\％ | 6.5 | 4 | 10 | 18.7 |
| Kyphosus sectatrix | Bermuda chub | 66 | 0.22 | 2.48 | 13 | 4．4\％ | 21.0 | 10 | 64 | 21，180．2 |
| Lactophrys bicaudalis | Spotted trunkfish | 2 | 0.01 | 0.08 | 2 | 0．7\％ | 20.0 | 14 | 26 | 530.2 |
| Lachnolaimus maximus | Hogfish | 129 | 0.43 | 0.95 | 75 | 25．2\％ | 25.7 | 7 | 51 | 68，340．5 |
| Lactophrys polygonia | Honeycomb cowfish | 1 | 0.00 | 0.06 | 1 | 0．3\％ | 19.0 | 19 | 19 | 98.0 |
| Lutjanus analis | Mutton snapper | 30 | 0.10 | 0.35 | 26 | 8．7\％ | 43.3 | 6 | 70 | 53，675．3 |
| Lutjanus apodus | Schoolmaster | 69 | 0.23 | 0.84 | 44 | 14．8\％ | 21.5 | 10 | 42 | 17，226．6 |
| Lutjanus buccanella | Blackfin snapper | 1 | 0.00 | 0.06 | 1 | 0．3\％ | 4.0 | 4 | 4 | 1.3 |
| Lutjanus cyanopterus | Cubera snapper | 2 | 0.01 | 0.12 | 1 | 0．3\％ | 43.0 | 41 | 46 | 3，159．1 |
| Lutjanus griseus | Gray snapper | 1，553 | 5.21 | 7.43 | 216 | 72．5\％ | 17.7 | 4 | 60 | 196，781．4 |
| Lutjanus jocu | Dog snapper | 1 | 0.00 | 0.06 | 1 | 0．3\％ | 24.0 | 24 | 24 | 270.9 |
| Lutjanus mahogoni | Mahogany snapper | 20 | 0.07 | 0.52 | 6 | 2．0\％ | 15.4 | 12 | 22 | 1，540．0 |
| Lutjanus synagris | Lane snapper | 140 | 0.47 | 2.55 | 24 | 8．1\％ | 11.1 | 6 | 25 | 4，250．9 |
| Malacoctenus macrops | Rosy blenny | 3 | 0.01 | 0.10 | 3 | 1．0\％ | 4.0 | 4 | 4 | 1.6 |
| Mycteroperca bonaci | Black grouper | 22 | 0.07 | 0.29 | 20 | 6．7\％ | 29.6 | 16 | 45 | 9，625．2 |
| Mycteroperca microlepis | Gag | 2 | 0.01 | 0.12 | 1 | 0．3\％ | 12.0 | 10 | 15 | 61.5 |
| Mycteroperca phenax | Scamp | 1 | 0.00 | 0.06 | 1 | 0．3\％ | 16.0 | 16 | 16 | 57.9 |
| Ocyurus chrysurus | Yellowtail snapper | 1，271 | 4.27 | 9.17 | 199 | 66．8\％ | 10.7 | 4 | 30 | 41，800．5 |
| Odontoscion dentex | Reef crocker | 7 | 0.02 | 0.15 | 7 | 2．3\％ | 10.1 | 6 | 14 | 90.7 |
| Ophioblennius atlanticus | Redlip blenny | 1 | 0.00 | 0.06 | 1 | 0．3\％ | 5.0 | 5 | 5 | 1.5 |
| Pomacanthus arcuatus | Gray angelfish | 305 | 1.02 | 1.37 | 150 | 50．3\％ | 23.1 | 4 | 50 | 161，624．7 |
| Pomacentrus fuscus | Dusky damselfish | 74 | 0.25 | 0.99 | 33 | 11．1\％ | 5.5 | 2 | 9 | 422.4 |
| Pomacentrus leucostictus | Beaugregory | 130 | 0.44 | 0.92 | 76 | 25．5\％ | 5.0 | 1 | 10 | 517.5 |
| Pomacentrus partitus | Bicolor damselfish | 38 | 0.13 | 0.60 | 22 | 7．4\％ | 4.0 | 3 | 8 | 68.2 |
| Pomacanthus paru | French angelfish | 41 | 0.14 | 0.45 | 28 | 9．4\％ | 26.0 | 6 | 41 | 27，898．0 |
| Pomacentrus planifrons | Three spot damselfish | 629 | 2.11 | 3.58 | 170 | 57．0\％ | 5.9 | 2 | 12 | 4，627．6 |
| Pomacentrus variabilis | Cocoa damselfish | 235 | 0.79 | 1.19 | 119 | 39．9\％ | 5.9 | 2 | 10 | 1，346．0 |
| Pseudupeneus maculatus | Spotted goatfish | 33 | 0.11 | 0.52 | 20 | 6．7\％ | 10.8 | 4 | 19 | 1，008．7 |
| Scarus coelestinus | Midnight parrotfish | 57 | 0.19 | 1.03 | 31 | 10．4\％ | 29.3 | 10 | 50 | 34，842．1 |
| Scarus coeruleus | Blue parrotfish | 45 | 0.15 | 0.54 | 33 | 11．1\％ | 19.5 | 3 | 40 | 9，406．4 |
| Scarus croicensis | Striped parrotfish | 2，897 | 9.72 | 8.63 | 236 | 79．2\％ | 5.4 | 1 | 16 | 9，742．2 |
| Scarus guacamaia | Rainbow parrotfish | 13 | 0.04 | 0.24 | 11 | 3．7\％ | 35.7 | 14 | 60 | 16，993．7 |
| Scarus taeniopterus | Princess parrotfish | 388 | 1.30 | 4.53 | 76 | 25．5\％ | 7.1 | 1 | 30 | 5，452．8 |
| Scarus vetula | Queen parrotfish | 3 | 0.01 | 0.10 | 3 | 1．0\％ | 28.3 | 25 | 30 | 1，256．9 |
| Sparisoma atomarium | Greenblotch parrotfish | 1 | 0.00 | 0.06 | 1 | 0．3\％ | 4.0 | 4 | 4 | 0.8 |
| Sparisoma aurofrenatum | Redband parrotfish | 248 | 0.83 | 1.71 | 103 | 34．6\％ | 7.9 | 3 | 20 | 2，842．5 |
| Sparisoma chrysopterum | Redtail parrotfish | 37 | 0.12 | 0.86 | 14 | 4．7\％ | 15.3 | 7 | 26 | 3，080．2 |
| Sparisoma radians | Bucktooth parrotfish | 8 | 0.03 | 0.23 | 5 | 1．7\％ | 5.8 | 3 | 10 | 24.2 |
| Sparisoma rubripinne | Yellowtail parrotfish | 11 | 0.04 | 0.35 | 4 | 1．3\％ | 6.2 | 2 | 14 | 87.6 |
| Sparisoma viride | Stoplight parrotfish | 248 | 0.83 | 1.92 | 135 | 45．3\％ | 10.6 | 2 | 35 | 16，815．2 |
| Sphyraena barracuda | Great barracuda | 150 | 0.50 | 1.40 | 67 | 22．5\％ | 59.9 | 10 | 170 | 432，489．7 |
| Sphoeroides spengleri | Bandtail puffer | 3 | 0.01 | 0.10 | 3 | 1．0\％ | 10.7 | 10 | 12 | 75.8 |
| Synodus intermedius | Sand diver | 1 | 0.00 | 0.06 | 1 | 0．3\％ | 13.0 | 13 | 13 | 21.7 |
| Thalassoma bifasciatum | Bluehead | 111 | 0.37 | 0.70 | 81 | 27．2\％ | 7.8 | 2 | 15 | 594.5 |
| Trachinotus falcatus | Permit | 1 | 0.00 | 0.06 | 1 | 0．3\％ | 61.0 | 61 | 61 | 4，034．9 |
| Tylosurus crocodilus | Houndfish | 1 | 0.00 | 0.06 | 1 | 0．3\％ | 46.0 | 46 | 46 | 863.0 |
| Unidentified sp． | Unidentified species | 2，151 | 7.22 | 116.01 | 3 | 1．0\％ | 2.9 | 2 | 3 |  |
| Urolophus jamaicensis | Yellow stingray | 4 | 0.01 | 0.12 | 4 | 1．3\％ | 29.5 | 26 | 35 | 1，025．1 |
| 24，338 |  |  |  |  |  |  |  |  |  | 1，763，256．0 |

Table 4. Summary of reef fish censused at Basin Hill (closed to fishing) in John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995.

| SPECIES | COMMON NAME | NUMBER |  |  | Sample $N=65$ |  | Number of species $=73$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total | Mean Abund. | $\begin{gathered} \text { Stand. } \\ \text { Dev. } \end{gathered}$ | OCCURENCE |  | LENGTH (cm) |  |  | $\begin{gathered} \text { BIOMASS (gms) } \\ \text { Total } \\ \hline \end{gathered}$ |
|  |  |  |  |  | Freq. | \% | Mean | Min | Max |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Abudefduf saxatilis | Sergeant major | 257 | 4.0 | 12.22 | 16 | 24.6\% | 6.5 | 2 | 11 | 2,364.5 |
| Acanthurus bahianus | Ocean surgeon | 27 | 0.4 | 0.92 | 14 | 21.5\% | 6.1 | 3 | 18 | 324.3 |
| Acanthurus chirurgus | Doctorfish | 27 | 0.4 | 1.20 | 12 | 18.5\% | 7.8 | 3 | 25 | 893.0 |
| Acanthurus coeruleus | Blue tang | 74 | 1.1 | 2.38 | 32 | 49.2\% | 18.4 | 4 | 35 | 21,381.8 |
| Aluterus scriptus | Scrawled filefish | 2 | 0.0 | 0.17 | 2 | 3.1\% | 52.5 | 45 | 60 | 2,201.1 |
| Anisotremus virginicus | Porkfish | 52 | 0.8 | 1.86 | 27 | 41.5\% | 7.4 | 2 | 24 | 1,164.2 |
| Archosargus rhomboidalis | Sea bream | 33 | 0.5 | 2.02 | 5 | 7.7\% | 18.8 | 10 | 30 | 6,878.4 |
| Calamus calamus | Saucereye porgy | 2 | 0.0 | 0.17 | 2 | 3.1\% | 27.0 | 25 | 29 | 887.7 |
| Caranx bartholomaei | Yellow jack | 84 | 1.3 | 5.44 | 6 | 9.2\% | 41.8 | 10 | 65 | 148,560.0 |
| Caranx crysos | Blue runner | 2 | 0.0 | 0.25 | 1 | 1.5\% | 20.0 | 20 | 20 | 331.8 |
| Caranx ruber | Bar jack | 377 | 5.8 | 31.78 | 13 | 20.0\% | 19.9 | 3 | 50 | 57,763.9 |
| Chaetodon capistratus | Foureye butterflyfish | 13 | 0.2 | 0.47 | 11 | 16.9\% | 6.0 | 4 | 8 | 94.5 |
| Chaetodiperus faber | Atlantic spadefish | 5 | 0.1 | 0.62 | 1 | 1.5\% | 30.0 | 30 | 30 | 4,257.1 |
| Coryphopterus dicrus | Colon goby | 3 | 0.0 | 0.21 | 3 | 4.6\% | 4.3 | 4 | 5 | 3.2 |
| Coryphopterus glaucofraenum | Bridled goby | 24 | 0.4 | 1.05 | 10 | 15.4\% | 3.1 | 1 | 5 | 11.8 |
| Coryphopterus personatus | Masked goby | 1 | 0.0 | 0.12 | 1 | 1.5\% | 3.0 | 3 | 3 | 0.3 |
| Diodon holocanthus | Balloonfish | 1 | 0.0 | 0.12 |  | 1.5\% | 14.0 | 14 | 14 | 89.4 |
| Diodon hystrix | Porcupinefish | 2 | 0.0 | 0.25 | 1 | 1.5\% | 25.0 | 25 | 25 | 1,620.3 |
| Echeneis naucrates | Sharksucker | 1 | 0.0 | 0.12 | 1 | 1.5\% | 20.0 | 20 | 20 | 71.1 |
| Gerres cinereus | Yellowfin mojarra | 13 | 0.2 | 1.25 | 4 | 6.2\% | 18.1 | 10 | 47 | 3,943.6 |
| Ginglymostoma cirratum | Nurse shark | 1 | 0.0 | 0.12 | 1 | 1.5\% | 200.0 | 200 | 200 | 47,606.3 |
| Gobiosoma oceanops | Neon goby | 24 | 0.4 | 0.78 | 14 | 21.5\% | 3.2 | 2 | 4 | 7.7 |
| Goby-like fish | Goby-like fish | 7 | 0.0 | 0.12 | 1 | 1.5\% | 4.0 | 4 | 4 | 0.6 |
| Haemulon aurolineatum | Tomtate | 477 | 7.3 | 12.69 | 30 | 46.2\% | 6.3 | 1 | 12 | 2,409.5 |
| Haemulon carbonarium | Caesar grunt | 1 | 0.0 | 0.12 | 1 | 1.5\% | 20.0 | 20 | 20 | 139.3 |
| Haemulon flavolineatum | French grunt | 123 | 1.9 | 7.77 | 16 | 24.6\% | 7.3 | 3 | 25 | 1,305.3 |
| Haemulon macrostomum | Spanish grunt | 1 | 0.0 | 0.12 |  | 1.5\% | 5.0 | 5 | 5 | 3.2 |
| Haemulon parra | Sailor's choice | 1 | 0.0 | 0.12 | 1 | 1.5\% | 14.0 | 14 | 14 | 53.6 |
| Haemulon plumieri | White grunt | 787 | 12.1 | 16.17 | 53 | 81.5\% | 10.8 | 4 | 22 | 22,160.6 |
| Haemulon sciurus | Bluestriped grunt | 431 | 6.6 | 11.28 | 44 | 67.7\% | 12.2 | 4 | 26 | 19,181.7 |
| Haemulon sp. | Unidentified grunt | 43 | 0.7 | 2.73 | 4 | 6.2\% | 3.8 | 3 | 7 | 56.1 |
| Halichoeres bivittatus | Slippery dick | 19 | 0.3 | 0.96 | 11 | 16.9\% | 8.0 | 5 | 11 | 134.2 |
| Halichoeres garnoti | Yellowhead wrasse | 2 | 0.0 | 0.17 | 2 | 3.1\% | 12.5 | 10 | 15 | 60.7 |
| Halichoeres maculipinna | Clown wrasse | 2 | 0.0 | 0.17 | 2 | 3.1\% | 10.5 | 7 | 14 | 50.8 |
| Halichoeres radiatus | Puddingwife | 1 | 0.0 | 0.12 | 1 | 1.5\% | 40.0 | 40 | 40 | 963.0 |
| Holacanthus bermudensis | Blue angelfish | 14 | 0.2 | 0.54 | 11 | 16.9\% | 13.1 | 4 | 28 | 1,225.7 |
| Holacanthus ciliaris | Queen anglefish | 6 | 0.1 | 0.29 | 6 | 9.2\% | 10.5 | 7 | 20 | 289.6 |
| Kyphosus sectatrix | Bermuda chub | 7 | 0.1 | 0.53 | 4 | 6.2\% | 25.7 | 15 | 35 | 3,271.8 |
| Lachnolaimus maximus | Hogfish | 14 | 0.2 | 0.80 | 9 | 13.8\% | 23.9 | 12 | 50 | 5,855.1 |
| Lutjanus analis | Mutton snapper | 10 | 0.2 | 0.48 | 8 | 12.3\% | 53.1 | 35 | 70 | 29,307.3 |
| Lutjanus apodus | Schoolmaster | 14 | 0.2 | 0.67 | 9 | 13.8\% | 23.6 | 18 | 30 | 3,431.8 |
| Lutjanus cyanopterus | Cubera snapper | 2 | 0.0 | 0.25 | 1 | 1.5\% | 43.0 | 41 | 46 | 3,159.1 |
| Lutjanus griseus | Gray snapper | 512 | 7.9 | 9.50 | 50 | 76.9\% | 18.3 | 5 | 60 | 70,791.5 |
| Lutjanus jocu | Dog snapper | 1 | 0.0 | 0.12 | 1 | 1.5\% | 24.0 | 24 | 24 | 270.9 |
| Lutjanus synagris | Lane snapper | 4 | 0.1 | 0.30 | 3 | 4.6\% | 10.8 | 7 | 14 | 106.9 |
| Malacoctenus macrops | Rosy blenny | 1 | 0.0 | 0.12 | 1 | 1.5\% | 4.0 | 4 | 4 | 0.5 |
| Mycteroperca bonaci | Black grouper | 6 | 0.1 | 0.29 | 6 | 9.2\% | 28.3 | 18 | 34 | 2,050.2 |
| Ocyurus chrysurus | Yellowtail snapper | 272 | 4.2 | 5.01 | 51 | 78.5\% | 11.9 | 4 | 30 | 13,247.3 |
| Odontoscion dentex | Reef crocker | , | 0.0 | 0.12 | 1 | 1.5\% | 6.0 | 6 | 6 | 2.3 |
| Pomacanthus arcuatus | Gray angelfish | 74 | 1.1 | 1.39 | 37 | 56.9\% | 25.2 | 6 | 35 | 45,606.3 |
| Pomacentrus fuscus | Dusky damselfish | 20 | 0.3 | 0.90 | 9 | 13.8\% | 5.9 | 4 |  | 129.0 |
| Pomacentrus leucostictus | Beaugregory | 26 | 0.4 | 0.90 | 15 | 23.1\% | 4.9 | 3 | 8 | 89.9 |
| Pomacentrus partitus | Bicolor damselfish | 7 | 0.1 | 0.53 | 4 | 6.2\% | 4.0 | 3 | 5 | 11.3 |
| Pomacanthus paru | French angelfish | 7 | 0.1 | 0.44 | 4 | 6.2\% | 26.3 | 6 | 36 | 4,982.7 |
| Pomacentrus planifrons | Three spot damselfish | 62 | 1.0 | 1.60 | 24 | 36.9\% | 4.7 | 2 | 9 | 266.9 |
| Pomacentrus variabilis | Cocoa damselfish | 40 | 0.6 | 0.93 | 24 | 36.9\% | 6.0 | 4 | 10 | 223.3 |
| Pseudupeneus maculatus | Spotted goattish | 2 | 0.0 | 0.17 | 2 | 3.1\% | 12.0 | 12 | 12 | 58.7 |
| Scarus coelestinus | Midnight parrotfish | 27 | 0.4 | 2.02 | 10 | 15.4\% | 34.5 | 12 | 50 | 24,136.2 |
| Scarus coeruleus | Blue parrotish | 24 | 0.4 | 0.89 | 15 | 23.1\% | 20.6 | 6 | 40 | 5,461.1 |
| Scarus croicensis | Striped parrotfish | 721 | 11.1 | 10.00 | 49 | 75.4\% | 6.1 | 3 | 12 | 3,080.3 |
| Scarus guacamaia | Rair bow parrotfish | 6 | 0.1 | 0.34 | 5 | 7.7\% | 28.8 | 14 | 45 | 3.842 .9 |

Table 4. cont.)

| SPECIES | COMMON NAME | NUMBER |  |  | Sample $\mathrm{N}=65$ |  | Number of species $=73$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total Abund. | Mean Abund. | Stand. Dev. | OCCURENCE |  | LENGTH (cm) BIOMASS (gms) |  |  |  |
|  |  |  |  |  | Freq. | \% | Mean | Min | Max | Total |
| Scarus taeniopterus | Princess parrotfish | 146 | 2.2 | 7.44 | 13 | 20.0\% | 5.3 | 2 | 30 | 1,140.7 |
| Scarus vetula | Queen parrotfish | 2 | 0.0 | 0.17 | 2 | 3.1\% | 27.5 | 25 | 30 | 767.3 |
| Sparisoma aurofrenatum | Redband parrotfish | 69 | 1.1 | 2.35 | 26 | 40.0\% | 8.4 | 4 | 20 | 991.4 |
| Sparisoma chrysopterum | Redtail parrotfish | 4 | 0.1 | 0.30 | 3 | 4.6\% | 11.0 | 7 | 20 | 158.6 |
| Sparisoma radians | Bucktooth parrotfish | 6 | 0.1 | 0.46 | 3 | 4.6\% | 5.3 | 3 | 8 | 11.1 |
| Sparisoma viride | Stoplight parrotfish | 39 | 0.6 | 0.98 | 26 | 40.0\% | 13.1 | 3 | 32 | 4,537.0 |
| Sphyraena barracuda | Barracuda | 109 | 1.7 | 2.51 | 38 | 58.5\% | 56.9 | 10 | 152 | 255,699.3 |
| Sphoeroides spengleri | Bandtail puffer | 1 | 0.0 | 0.12 | 1 | 1.5\% | 10.0 | 10 | 10 | 19.9 |
| Thalassoma bifasciatum | Bluehead | 22 | 0.3 | 0.67 | 16 | 24.6\% | 7.9 | 3 | 13 | 131.0 |
| Tylosurus crocodilus | Houndfish | 1 | 0.0 | 0.12 | 1 | 1.5\% | 46.0 | 46 | 46 | 863.0 |
| Unidentified sp. | Unidentified species | 2,151 | 33.1 | 248.16 | 4 | 6.2\% | 2.9 | 2 | 3 |  |
| Urolophus jamaicensis | Yellow stingray | 2 | 0.0 | 0.17 | 2 | 3.1\% | 30.5 | 26 | 35 | 580.7 |
|  |  | 7,346 |  |  |  |  |  |  |  | 832,773.6 |

Table 5. Summary of reef fish censused at Basin Hill (open to fishing) in John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995.

| SPECIES | COMMON NAME | NUMBER Sample $\mathrm{N}=24$ Number of species $=51$ |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total Mean <br> Abund. Abund. |  | Stand. Dev. | OCCURENCE |  | LENGTH (cm) BIOMASS (gms) |  |  |  |
|  |  |  |  | Freq. | \% | Mean | Min | Max | Total |
| Abudefduf saxatilis | Sergeant major | 38 | 1.6 |  | 4.90 | 8 | 33.3\% | 5.5 | 3 | 10 | 242.8 |
| Acanthurus bahianus | Ocean surgeon | 8 | 0.3 | 0.48 | 8 | 33.3\% | 11.8 | 4 | 30 | 872.0 |
| Acanthurus chirurgus | Doctorfish | 5 | 0.2 | 0.51 | 4 | 16.7\% | 19.6 | 15 | 25 | 888.7 |
| Acanthurus coeruleus | Blue tang | 29 | 1.2 | 1.06 | 18 | 75.0\% | 13.0 | 3 | 26 | 2,795.4 |
| Anisotremus virginicus | Porkfish | 23 | 1.0 | 1.37 | 11 | 45.8\% | 8.2 | 1 | 20 | 538.1 |
| Bodianus pulchellus | Spotfin hogfish | 1 | 0.0 | 0.20 | 1 | 4.2\% | 8.0 | 8 | 8 | 8.3 |
| Calamus calamus | Saucereye porgy | 7 | 0.3 | 0.46 | 7 | 29.2\% | 20.6 | 6 | 36 | 2,101.6 |
| Caranx ruber | Bar jack | 67 | 2.8 | 5.85 | 7 | 29.2\% | 12.9 | 4 | 30 | 4,622.0 |
| Chaetodon capistratus | Foureye butterflyfish | 6 | 0.3 | 0.68 | 4 | 16.7\% | 7.2 | 6 | 8 | 67.0 |
| Chaetodiperus faber | Atlantic spadefish | 2 | 0.1 | 0.28 | 2 | 8.3\% | 32.0 | 28 | 36 | 2,096.3 |
| Chaetodon ocellatus | Spotfin butterflyfish | 1 | 0.0 | 0.20 | 1 | 4.2\% | 6.0 | 6 | 6 | 6.7 |
| Coryphopterus glaucofraenum | Bridled goby | 6 | 0.3 | 0.44 | 6 | 25.0\% | 4.5 | 3 | 6 | 8.4 |
| Coryphopterus personatus | Masked goby | 2 | 0.1 | 0.41 | 1 | 4.2\% | 3.0 | 3 | 3 | 0.7 |
| Epinephelus morio | Red grouper | 8 | 0.3 | 0.76 | 5 | 20.8\% | 33.9 | 21 | 50 | 5,610.3 |
| Gnatholepis thompsoni | Goldspot goby | 5 | 0.2 | 0.72 | 2 | 8.3\% | 3.6 | 2 | 5 | 3.1 |
| Gobiosoma oceanops | Neon goby | 4 | 0.2 | 0.38 | 4 | 16.7\% | 3.0 | 2 | 4 | 1.2 |
| Gymnothorax moringa | Spotted moray | 1 | 0.0 | 0.20 | 1 | 4.2\% | 30.0 | 30 | 30 | 46.3 |
| Haemulon aurolineatum | Tomtate | 529 | 22.0 | 37.43 | 13 | 54.2\% | 6.4 | 2 | 9 | 2,593.8 |
| Haemulon carbonarium | Caesar grunt | 1 | 0.0 | 0.20 | 1 | 4.2\% | 8.0 | 8 | 8 | 8.5 |
| Haemulon flavolineatum | French grunt | 15 | 0.6 | 1.13 | 7 | 29.2\% | 8.6 | 8 | 9 | 176.7 |
| Haemulon plumieri | White grunt | 486 | 20.3 | 29.63 | 15 | 62.5\% | 11.1 | 4 | 29 | 15,350.9 |
| Haemulon sciurus | Bluestriped grunt | 162 | 6.8 | 10.57 | 16 | 66.7\% | 14.3 | 5 | 30 | 11,579.1 |
| Halichoeres maculipinna | Clown wrasse | 2 | 0.1 | 0.28 | 2 | 8.3\% | 8.0 | 8 | 8 | 11.9 |
| Holacanthus bermudensis | Blue angelfish | 11 | 0.5 | 0.66 | 9 | 37.5\% | 14.2 | 7 | 30 | 1,227.0 |
| Holacanthus ciliaris | Queen anglefish | 9 | 0.4 | 0.88 | 5 | 20.8\% | 12.0 | 6 | 20 | 538.0 |
| Hypoplectrus puella \# | Barred hamlet | 1 | 0.0 | 0.20 | 1 | 4.2\% | 10.0 | 10 | 10 | 2.6 |
| Hypoplectrus unicolor | Butter hamlet | 4 | 0.2 | 0.38 | 4 | 16.7\% | 8.3 | 7 | 10 | 6.4 |
| Lachnolaimus maximus | Hogfish | 26 | 1.1 | 1.59 | 12 | 50.0\% | 24.9 | 7 | 51 | 14,578.6 |
| Lutjanus analis | Mutton snapper | 6 | 0.3 | 0.44 | 6 | 25.0\% | 44.5 | 35 | 53 | 9,403.2 |
| Lutjanus apodus | Schoolmaster | 7 | 0.3 | 0.86 | 4 | 16.7\% | 29.6 | 13 | 40 | 3,413.1 |
| Lutjanus griseus | Gray snapper | 162 | 6.8 | 7.74 | 19 | 79.2\% | 23.5 | 10 | 41 | 40,203.1 |
| Malacoctenus macrops | Rosy blenny | 1 | 0.0 | 0.20 | 1 | 4.2\% | 4.0 | 4 | 4 | 0.5 |
| Mycteroperca bonaci | Black grouper | 3 | 0.1 | 0.34 | 3 | 12.5\% | 26.7 | 20 | 40 | 1,136.2 |
| Ocyurus chrysurus | Yellowtail snapper | 152 | 6.3 | 6.18 | 20 | 83.3\% | 10.4 | 5 | 30 | 5,018.5 |
| Pomacanthus arcuatus | Gray angelfish | 43 | 1.8 | 1.59 | 17 | 70.8\% | 23.3 | 6 | 37 | 25,056.0 |
| Pomacentrus leucostictus | Beaugregory | 2 | 0.1 | 0.41 | 1 | 4.2\% | 5.0 | 5 | 5 | 6.3 |
| Pomacentrus partitus | Bicolor damselfish | 2 | 0.1 | 0.28 | 2 | 8.3\% | 5.5 | 3 | 8 | 13.4 |
| Pomacanthus paru | French angelfish | 5 | 0.2 | 0.51 | 4 | 16.7\% | 25.8 | 20 | 35 | 2,904.8 |
| Pomacentrus planifrons | Three spot damselfish | 15 | 0.6 | 0.82 | 11 | 45.8\% | 5.5 | 3 | 8 | 93.9 |
| Pomacentrus variabilis | Cocoa damselfish | 32 | 1.3 | 1.43 | 15 | 62.5\% | 5.8 | 3 | 8 | 160.0 |
| Pseudupeneus maculatus | Spotted goatfish | 2 | 0.1 | 0.41 | 1 | 4.2\% | 5.0 | 4 | 6 | 4.7 |
| Scarus coelestinus | Midnight parrotfish | 12 | 0.5 | 0.78 | 8 | 33.3\% | 20.5 | 13 | 35 | 2,476.2 |
| Scarus coeruleus | Blue parrotfish | 2 | 0.1 | 0.28 | 2 | 8.3\% | 18.5 | 15 | 22 | 243.2 |
| Scarus croicensis | Striped parrotfish | 243 | 10.1 | 7.85 | 20 | 83.3\% | 5.3 | 2 | 10 | 651.1 |
| Scarus guacamaia | Rainbow parrotfish | 1 | 0.0 | 0.20 | 1 | 4.2\% | 35.0 | 35 | 35 | 832.1 |
| Scarus taeniopterus | Princess parrotfish | 20 | 0.8 | 3.09 | 4 | 16.7\% | 8.1 | 5 | 25 | 469.3 |
| Sparisoma aurofrenatum | Redband parrotfish | 8 | 0.3 | 0.64 | 6 | 25.0\% | 8.9 | 4 | 16 | 150.5 |
| Sparisoma chrysopterum | Redtail parrotfish | 11 | 0.5 | 2.25 | 1 | 4.2\% | 22.0 | 20 | 23 | 1,944.8 |
| Sparisoma viride | Stoplight parrotfish | 26 | 1.1 | 1.21 | 15 | 62.5\% | 10.1 | 2 | 32 | 1,628.1 |
| Sphyraena barracuda | Barracuda | 7 | 0.3 | 0.62 | 5 | 20.8\% | 30.7 | 10 | 50 | 2,193.3 |
| Thalassoma bifasciatum | Bluehead | 5 | 0.2 | 0.41 | 5 | 20.8\% | 7.0 | 4 | 10 | 19.2 |
|  |  | 2,226 |  |  |  |  |  |  |  | 164,003.9 |

Table 6. Summary of reef fish censused at Basin Hill (open to fishing) in John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995.

| SPECIES | COMMON NAME | NUMBER |  |  | Sample $\mathrm{N}=64$ |  | Number of species $=63$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total Abund. | Mean Abund | $\begin{aligned} & \text { Stand. } \\ & \text { Dev. } \end{aligned}$ | DCCURENCE |  | LENGTH (cm) |  |  | $\begin{gathered} \text { BIOMASS (gms) } \\ \text { Total } \end{gathered}$ |
|  |  |  |  |  | Freq. | \% | Mean | Min | Max |  |
|  | $===========$ | $===$ | ==== | === | ==== | $===$ | $===$ | $=$ | = | ============ |
| Abudefduf saxatilis | Sergeant major | 210 | 3.3 | 7.98 | 30 | 46.9\% | 5.0 | 1 | 15 | 1,582.4 |
| Acanthurus bahianus | Ocean surgeon | 22 | 0.3 | 0.84 | 13 | 20.3\% | 8.1 | 3 | 25 | 872.0 |
| Acanthurus chirurgus | Doctorfish | 31 | 0.5 | 1.23 | 15 | 23.4\% | 8.3 | 3 | 22 | 812.4 |
| Acanthurus coeruleus | Blue tang | 48 | 0.8 | 0.99 | 30 | 46.9\% | 10.9 | 3 | 20 | 2,569.1 |
| Anisotremus virginicus | Porkfish | 47 | 0.7 | 0.96 | 31 | 48.4\% | 8.3 | 1 | 30 | 2,962.9 |
| Calamus calamus | Saucereye porgy | 5 | 0.1 | 0.32 | 4 | 6.3\% | 18.4 | 16 | 22 | 769.7 |
| Canthigaster rostrata | Sharpnose puffer | 2 | 0.0 | 0.18 | 2 | 3.1\% | 5.5 | 5 | 6 | 6.4 |
| Caranx bartholomaei | Yellow jack | 4 | 0.1 | 0.39 | 2 | 3.1\% | 26.5 | 22 | 40 | 1,805.9 |
| Caranx ruber | Bar jack | 68 | 1.1 | 3.61 | 10 | 15.6\% | 11.3 | 5 | 48 | 8,439.1 |
| Chaetodon capistratus | Foureye butterflyfish | 21 | 0.3 | 0.64 | 16 | 25.0\% | 5.0 | 3 | 7 | 92.7 |
| Chaetodon ocellatus | Spotfin butterflyfish | 5 | 0.1 | 0.37 | 3 | 4.7\% | 9.2 | 7 | 10 | 118.1 |
| Coryphopterus dicrus | Colon goby | 6 | 0.1 | 0.53 | 2 | 3.1\% | 5.0 | 3 | 6 | 8.9 |
| Coryphopterus glaucofraenum | Bridled goby | 70 | 1.1 | 1.81 | 26 | 40.6\% | 3.9 | 2 | 6 | 62.4 |
| Coryphopterus personatus | Masked goby | 7 | 0.1 | 0.62 | 2 | 3.1\% | 2.6 | 2 | 3 | 1.7 |
| Coryphopterus species | Unknown goby | 6 | 0.1 | 0.46 | 3 | 4.7\% | 5.3 | 5 | 7 | 12.0 |
| Epinephelus morio | Red grouper | 3 | 0.0 | 0.21 | 3 | 4.7\% | 28.7 | 20 | 34 | 1,108.2 |
| Gerres cinereus | Yellowfin mojarra | 8 | 0.1 | 0.55 | 4 | 6.3\% | 15.9 | 15 | 17 | 732.1 |
| Gobiosoma oceanops | Neon goby | 29 | 0.5 | 1.11 | 14 | 21.9\% | 2.9 | 1 | 5 | 7.9 |
| Haemulon aurolineatum | Tomtate | 1,244 | 19.4 | 27.24 | 40 | 62.5\% | 6.2 | 1 | 14 | 5,498.8 |
| Haemulon carbonarium | Caesar grunt | 3 | 0.0 | 0.38 | 1 | 1.6\% | 13.0 | 10 | 15 | 111.9 |
| Haemulon chrysargyreum | Smallmouth grunt | 1 | 0.0 | 0.13 | 1 | 1.6\% | 10.0 | 10 | 10 | 57.0 |
| Haemulon flavolineatum | French grunt | 67 | 1.0 | 3.87 | 18 | 28.1\% | 7.2 | , | 14 | 564.5 |
| Haemulon macrostomum | Spanish grunt | 1 | 0.0 | 0.13 | 1 | 1.6\% | 10.0 | 10 | 10 | 26.1 |
| Haemulon parra | Sailor's choice | 1 | 0.0 | 0.13 | 1 | 1.6\% | 31.0 | 31 | 31 | 578.7 |
| Haemulon plumieri | White grunt | 675 | 10.5 | 12.25 | 54 | 84.4\% | 9.7 | 3 | 30 | 16,818.1 |
| Haemulon sciurus | Bluestriped grunt | 333 | 5.2 | 6.82 | 52 | 81.3\% | 14.3 |  | 30 | 30,008.4 |
| Haemulon sp. | Unidentified grunt | 38 | 0.6 | 4.75 | 1 | 1.6\% | 3.0 | 3 | 3 | 14.2 |
| Halichoeres bivittatus | Slippery dick | 1 | 0.0 | 0.13 | 1 | 1.6\% | 9.0 | 9 | 9 | 8.6 |
| Holacanthus bermudensis | Blue angelfish | 32 | 0.5 | 1.14 | 21 | 32.8\% | 16.0 | 3 | 36 | 7,219.6 |
| Holacanthus ciliaris | Queen anglefish | 12 | 0.2 | 0.66 | 7 | 10.9\% | 10.7 | 3 | 30 | 1,208.7 |
| Hypoplectrus unicolor | Butter hamlet | 6 | 0.1 | 0.29 | 6 | 9.4\% | 6.2 | 4 | 8 | 4.6 |
| Lactophrys bicaudalis | Spotted trunkfish | 1 | 0.0 | 0.13 | 1 | 1.6\% | 14.0 | 14 | 14 | 113.7 |
| Lachnolaimus maximus | Hogfish | 45 | 0.7 | 1.16 | 22 | 34.4\% | 25.2 | 12 | 50 | 20,103.0 |
| Lutjanus analis | Muttori snapper | 5 | 0.1 | 0.37 | 3 | 4.7\% | 27.8 | 6 | 51 | 3,802.2 |
| Lutjanus apodus | Schoolmaster | 22 | 0.3 | 1.45 | 9 | 14.1\% | 16.1 | 10 | 30 | 2,565.8 |
| Lutjanus griseus | Gray snapper | 442 | 6.9 | 8.52 | 51 | 79.7\% | 16.4 | 4 | 46 | 46,551.6 |
| Lutjanus synagris | Lane snapper | 38 | 0.6 | 1.85 | 9 | 14.1\% | 11.0 | 7 | 25 | 1,204.5 |
| Malacoctenus macrops | Rosy blenny | 1 | 0.0 | 0.13 | 1 | 1.6\% | 4.0 | 4 | 4 | 0.5 |
| Mycteroperca bonaci | Black grouper | 2 | 0.0 | 0.18 | 2 | 3.1\% | 34.5 | 34 | 35 | 1,163.2 |
| Mycteroperca microlepis | Gag | 2 | 0.0 | 0.25 | 1 | 1.6\% | 12.0 | 10 | 15 | 61.5 |
| Ocyurus chrysurus | Yellowtail snapper | 229 | 3.6 | 5.27 | 42 | 65.6\% | 9.7 | 4 | 25 | 6,391.3 |
| Odontoscion dentex | Reef crocker | 3 | 0.0 | 0.21 | 3 | 4.7\% | 9.3 | 9 | 10 | 26.3 |
| Ophioblennius atlanticus | Redip blenny | 1 | 0.0 | 0.13 | 1 | 1.6\% | 5.0 | 5 | 5 | 1.5 |
| Pomacanthus arcuatus | Gray angelfish | 48 | 0.8 | 1.05 | 28 | 43.8\% | 20.4 | 4 | 50 | 21,298.2 |
| Pomacentrus fuscus | Dusky damselfish | 39 | 0.6 | 1.80 | 13 | 20.3\% | 5.3 | 3 | 7 | 199.3 |
| Pomacentrus leucostictus | Beaugregory | 32 | 0.5 | 0.96 | 21 | 32.8\% | 4.8 | , | 8 | 115.6 |
| Pomacentrus partitus | Bicolor damselfish | 14 | 0.2 | 0.72 | 9 | 14.1\% | 3.6 | 3 | 5 | 16.2 |
| Pomacanthus paru | French angelfish | 8 | 0.1 | 0.42 | 6 | 9.4\% | 26.3 | 15 | 37 | 5,266.8 |
| Pomacentrus planifrons | Three spot damselfish | 199 | 3.1 | 5.44 | 39 | 60.9\% | 6.2 | 2 | 10 | 1,662.0 |
| Pomacentrus variabilis | Cocoa damselfish | 70 | 1.1 | 1.31 | 34 | 53.1\% | 5.7 | 2 | 8 | 376.3 |
| Pseudupeneus maculatus | Spotted goattish | 6 | 0.1 | 0.34 | 5 | 7.8\% | 10.8 | 9 | 12 | 134.2 |
| Scarus coelestinus | Midnight parrotish | 7 | 0.1 | 0.36 | 6 | 9.4\% | 31.3 | 26 | 35 | 4,119.6 |
| Scarus coeruleus | Blue parrotish | 4 | 0.1 | 0.24 | 4 | 6.3\% | 20.0 |  | 32 | 997.7 |
| Scarus croicensis | Striped parrotfish | 543 | 8.5 | 8.47 | 54 | 84.4\% | 5.4 | 2 | 14 | 1,938.5 |
| Scarus taeniopterus | Princess parrotish | 60 | 0.9 | 2.36 | 18 | 28.1\% | 10.3 | 4 | 20 | 1,570.6 |
| Scarus vetula | Queer parrotish | 1 | 0.0 | 0.13 |  | 1.6\% | 30.0 | 30 | 30 | 489.6 |
| Sparisoma aurofrenatum | Redband parrotfish | 43 | 0.7 | 1.14 | 22 | 34.4\% | 6.8 |  | 15 | 259.4 |
| Sparisoma chrysopterum | Redtail parrotfish | 4 | 0.1 | 0.30 | 3 | 4.7\% | 17.3 | 11 | 26 | 441.2 |
| Sparisoma rubripinne | Yellowtail parrotfish | 2 | 0.0 | 0.25 |  | 1.6\% | 7.0 | 6 | 8 | 12.9 |
| Sparisoma viride | Stoplight parrottish | 44 | 0.7 | 0.77 | 32 | 50.0\% | 12.3 | 4 | 35 | 4,045.6 |
| Sphyraena barracuda | Barracuda | 14 | 0.2 | 0.63 | 9 | 14.1\% | 74.9 | 40 | 120 | 56,760.0 |
| Sphoeroides spengleri | Bandtail puffer | 2 | 0.0 | 0.18 | 2 | 3.1\% | 11.0 | 10 | 12 | 55.9 |
| Thalassoma bifasciatum | Bluehead | 40 | 0.6 | 0.86 | 28 | 43.8\% | 8.3 | 3 | 15 | 245.5 |
|  |  | 4,213 |  |  |  |  |  |  |  | 266,073.0 |

Table 7. Summary of reef fish censused at Mosquito Bank (closed to fishing) in John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995.

| SPECIES | COMMON NAME | NUMBER |  |  | Sample $\mathrm{N}=59$ |  | Number of species $=70$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total | Mean | Stand. | OCCU | ENCE |  | GTH | cm) | BIOMASS (gms) |
|  |  | Abund. | Abund. | Dev. | Freq. | \% | Mean | Min | Max | Total |
| Abudefduf saxatilis | Sergeant major | 41 | 0.7 | 2.17 | 9 | 15.3\% | 10.4 | 8 | 15 | 1,631.1 |
| Acanthurus bahianus | Ocean surgeon | 9 | 0.2 | 0.41 | 8 | 13.6\% | 7.8 | 3 | 12 | 131.0 |
| Acanthurus chirurgus | Doctorfish | 8 | 0.1 | 0.47 | 6 | 10.2\% | 15.8 | 10 | 22 | 686.9 |
| Acanthurus coeruleus | Blue tang | 30 | 0.5 | 0.84 | 20 | 33.9\% | 10.6 | 6 | 16 | 1,201.5 |
| Aluterus schoepfi | Orange filefish | 2 | 0.0 | 0.26 | 1 | 1.7\% | 34.0 | 34 | 34 | 722.0 |
| Anisotremus virginicus | Porkfish | 37 | 0.6 | 1.88 | 12 | 20.3\% | 5.8 | 2 | 25 | 1,026.9 |
| Archosargus rhomboidalis | Sea bream | 5 | 0.1 | 0.43 | 3 | 5.1\% | 15.6 | 15 | 17 | 456.1 |
| Aulostomus maculatus | Trumpetfish | 4 | 0.1 | 0.25 | 4 | 6.8\% | 31.5 | 30 | 33 | 312.1 |
| Calamus calamus | Saucereye porgy | 8 | 0.1 | 0.43 | 6 | 10.2\% | 15.8 | 6 | 20 | 864.5 |
| Canthigaster rostrata | Sharpnose puffer | 2 | 0.0 | 0.18 | 2 | 3.4\% | 5.0 | 4 | 6 | 5.2 |
| Caranx ruber | Bar jack | 92 | 1.6 | 9.20 | 6 | 10.2\% | 8.4 | 6 | 24 | 1,054.6 |
| Chaetodon capistratus | Foureye butterflyfish | 22 | 0.4 | 0.74 | 16 | 27.1\% | 5.4 | 4 | 9 | 146.1 |
| Chaetodiperus faber | Atlantic spadefish | 2 | 0.0 | 0.18 | 2 | 3.4\% | 26.0 | 22 | 30 | 1,221.8 |
| Chaetodon striatus | Banded butterflyfish | 2 | 0.0 | 0.26 | 1 | 1.7\% | 12.0 | 12 | 12 | 108.3 |
| Coryphopterus dicrus | Colon goby | 5 | 0.1 | 0.43 | 3 | 5.1\% | 3.0 | 3 | 3 | 1.7 |
| Coryphopterus glaucofraenum | Bridied goby | 15 | 0.3 | 1.04 | 5 | 8.5\% | 3.1 | 2 | 4 | 5.4 |
| Coryphopterus personatus | Masked goby | 4 | 0.1 | 0.37 | 2 | 3.4\% | 2.5 | 2 | 3 | 0.9 |
| Cryptotomus roseus | Bluelip parrotfish | 8 | 0.1 | 0.82 | 2 | 3.4\% | 4.8 | 4 | 5 | 58.9 |
| Epinephelus cruentatus | Graysby | 2 | 0.0 | 0.18 | 2 | 3.4\% | 13.0 | 11 | 15 | 66.8 |
| Epinephelus morio | Red grouper | 1 | 0.0 | 0.13 | 1 | 1.7\% | 24.0 | 24 | 24 | 189.5 |
| Gerres cinereus | Yellowfin mojarra | 3 | 0.1 | 0.29 | 2 | 3.4\% | 16.0 | 15 | 18 | 293.0 |
| Gobiosoma oceanops | Neon goby | 3 | 0.1 | 0.29 | 2 | 3.4\% | 2.7 | 2 | 5 | 1.6 |
| Haemulon aurolineatum | Tomtate | 628 | 10.6 | 45.64 | 24 | 40.7\% | 7.9 | 3 | 16 | 6,540.3 |
| Haemulon flavolineatum | French grunt | 78 | 1.3 | 3.03 | 18 | 30.5\% | 12.8 | 5 | 28 | 5,870.0 |
| Haemulon macrostomum | Spanish grunt | 3 | 0.1 | 0.22 | 3 | 5.1\% | 27.3 | 26 | 30 | 1,673.7 |
| Haemulon melanurum | Cottonwick | 80 | 1.4 | 5.48 | 4 | 6.8\% | 13.0 | 10 | 15 | 3,510.9 |
| Haemulon plumieri | White grunt | 658 | 11.2 | 12.22 | 55 | 93.2\% | 10.6 | 3 | 30 | 21,031.6 |
| Haemulon sciurus | Bluestriped grunt | 98 | 1.7 | 2.81 | 30 | 50.8\% | 13.7 | 5 | 30 | 7,623.7 |
| Halichoeres bivittatus | Slippery dick | 47 | 0.8 | 1.40 | 19 | 32.2\% | 6.5 | 3 | 15 | 228.6 |
| Halichoeres garnoti | Yellowhead wrasse | 1 | 0.0 | 0.13 | 1 | 1.7\% | 12.0 | 12 | 12 | 22.8 |
| Halichoeres maculipinna | Clown wrasse | 5 | 0.1 | 0.34 | 4 | 6.8\% | 10.0 | 8 | 12 | 79.0 |
| Hemipteronotus martinicensis | Rosy razorfish | 1 | 0.0 | 0.13 | 1 | 1.7\% | 6.0 | 6 | 6 | 4.5 |
| Holacanthus bermudensis | Blue angelfish | 3 | 0.1 | 0.22 | 3 | 5.1\% | 13.3 | 9 | 20 | 241.1 |
| Holacanthus ciliaris | Queen anglefish | 8 | 0.1 | 0.39 | 7 | 11.9\% | 16.5 | 10 | 25 | 1,221.4 |
| Hyploplectrus nigricans | Black hamlet | 1 | 0.0 | 0.13 | 1 | 1.7\% | 4.0 | 4 | 4 | 0.2 |
| Hypoplectrus puella | Barred hamlet | 1 | 0.0 | 0.13 | 1 | 1.7\% | 6.0 | 6 | 6 | 0.6 |
| Hypoplectrus unicolor | Butter hamiet | 5 | 0.1 | 0.28 | 5 | 8.5\% | 6.2 | 6 | 7 | 3.4 |
| Kyphosus sectatrix | Bermuda chub | 59 | 1.0 | 5.52 | 9 | 15.3\% | 20.5 | 10 | 27 | 11,898.8 |
| Lachnolaimus maximus | Hogfish | 15 | 0.3 | 0.58 | 12 | 20.3\% | 23.6 | 10 | 35 | 5,403.7 |
| Lactophrys polygonia | Honeycomb cowfish | , | 0.0 | 0.13 | 1 | 1.7\% | 19.0 | 19 | 19 | 98.0 |
| Lutjanus analis | Mutton snapper | 2 | 0.0 | 0.18 | 2 | 3.4\% | 47.0 | 45 | 49 | 3,521.4 |
| Lutjanus apodus | Schoolmaster | 11 | 0.2 | 0.43 | 10 | 16.9\% | 21.2 | 14 | 38 | 2,913.3 |
| Lutjanus griseus | Gray snapper | 71 | 1.2 | 1.69 | 31 | 52.5\% | 18.1 | 7 | 38 | 8,660.0 |
| Lutjanus mahogoni | Mahogany snapper | 5 | 0.1 | 0.53 | 2 | 3.4\% | 18.0 | 16 | 22 | 572.7 |
| Lutjanus synagris | Lane snapper | 29 | 0.5 | 3.52 | 3 | 5.1\% | 10.1 | 8 | 15 | 602.7 |
| Mycteroperca bonaci | Black grouper | 3 | 0.1 | 0.29 | 2 | 3.4\% | 36.7 | 30 | 45 | 2,341.9 |
| Ocyurus chrysurus | Yellowtail snapper | 171 | 2.9 | 5.15 | 31 | 52.5\% | 10.6 | 5 | 20 | 4,903.1 |
| Odontoscion dentex | Reef crocker | 3 | 0.1 | 0.22 | 3 | 5.1\% | 12.3 | 11 | 14 | 62.2 |
| Pomacanthus arcuatus | Gray angelfish | 73 | 1.2 | 1.63 | 34 | 57.6\% | 25.0 | 8 | 40 | 46,322.7 |
| Pomacentrus fuscus | Dusky damselfish | 3 | 0.1 | 0.29 | 2 | 3.4\% | 8.0 | 6 | 9 | 47.2 |
| Pomacentrus leucostictus | Beaugregory | 14 | 0.2 | 0.70 | 8 | 13.6\% | 5.4 | 4 | 10 | 70.8 |
| Pomacentrus partitus | Bicolor damselfish | 12 | 0.2 | 0.91 | 4 | 6.8\% | 4.1 | 3 | 5 | 19.3 |
| Pomacanthus paru | French angelfish | 10 | 0.2 | 0.53 | 6 | 10.2\% | 24.8 | 13 | 36 | 6,813.1 |
| Pomacentrus planifrons | Three spot damselfish | 168 | 2.8 | 3.59 | 43 | 72.9\% | 6.6 | 3 | 12 | 1,608.9 |
| Pomacentrus variabilis | Cocoa damselfish | 36 | 0.6 | 1.35 | 14 | 23.7\% | 6.8 | 3 | 10 | 300.3 |
| Pseudupeneus maculatus | Spotted goatfish | 16 | 0.3 | 0.98 | 6 | 10.2\% | 10.9 | 5 | 19 | 577.4 |
| Scarus coelestinus | Midnight parrotfish | 1 | 0.0 | 0.13 | 1 | 1.7\% | 25.0 | 25 | 25 | 292.1 |
| Scarus coeruleus | Blue parrotfish | 8 | 0.1 | 0.57 | 5 | 8.5\% | 13.8 | 6 | 24 | 566.2 |
| Scarus croicensis | Striped parrotfish | 536 | 9.1 | 7.82 | 47 | 79.7\% | 5.4 | 1 | 16 | 1,942.5 |
| Scarus guacamaia | Rainbow parrotfish | 5 | 0.1 | 0.34 | 4 | 6.8\% | 39.2 | 30 | 60 | 7,982.8 |
| Scarus taeniopterus | Princess parrotfish | 76 | 1.3 | 2.24 | 29 | 49.2\% | 8.9 | 3 | 26 | 1,737.8 |

Table 7 (cont.)

| SPECIES | COMMON NAME | NUMBER Sample $\mathrm{N}=59$ |  |  |  |  | Number of species $=70$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total Abund. | Mean Abund. | Stand. Dev. | OCCURENCE |  | LENGTH (cm) |  |  | $\begin{aligned} & \text { BIOMASS (gms) } \\ & \hline \text { Total } \end{aligned}$ |
|  |  |  |  |  | Freq. | \% | Mean | Min | Max |  |
| Sparisoma aurofrenatum | Redband parrotfish | 71 | 1.2 | 2.07 | 24 | 40.7\% | 8.3 | 3 | 16 | 878.3 |
| Sparisoma chrysopterum | Redtail parrotfish | 2 | 0.0 | 0.18 | 2 | 3.4\% | 10.0 | 10 | 10 | 29.3 |
| Sparisoma radians | Bucktooth parrotfish | 1 | 0.0 | 0.13 | 1 | 1.7\% | 4.0 | 4 | 4 | 0.5 |
| Sparisoma rubripinne | Yellowtail parrotfish | 8 | 0.1 | 0.73 | 2 | 3.4\% | 5.0 | 2 | 8 | 23.9 |
| Sparisoma viride | Stoplight parrotfish | 73 | 1.2 | 3.47 | 27 | 45.8\% | 10.0 | 2 | 30 | 4,097.6 |
| Sphyraena barracuda | Barracuda | 11 | 0.2 | 0.54 | 8 | 13.6\% | 71.9 | 16 | 170 | 71,991.4 |
| Thalassoma bifasciatum | Bluehead | 15 | 0.3 | 0.63 | 10 | 16.9\% | 8.3 | 5 | 13 | 92.8 |
| Trachinotus falcatus | Permit | 1 | 0.0 | 0.13 | 1 | 1.7\% | 61.0 | 61 | 61 | 4,034.9 |
| Urolophus jamaicensis | Yellow stingray | 1 | 0.0 | 0.13 | 1 | 1.7\% | 27.0 | 27 | 27 | 186.4 |
|  |  | 3,433 |  |  |  |  |  |  |  | 248,831.4 |

Table 8. Summary of reef fish censused at Mosquito Bank (open to fishing) in John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995.

| SPECIES | COMMON NAME | W. NUMBER Sample $\mathrm{N}=62$ Number of species $=69$ |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\begin{gathered} \text { Total } \\ \text { Abund. } \end{gathered}$ | Mean Abund. | Stand. Dev. | OCCURENCEFreq. $1 \%$ |  | LENGTH (cm) <br> Mean Min Max |  |  | $\begin{aligned} & \text { B[OMASS (gms) } \\ & \frac{\text { Total }}{} \end{aligned}$ |
|  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |
| Abudefduf saxatilis | Sergeant major | 122 | 2.0 | 8.35 | 18 | 29.0\% | 5.6 | 2 | 10 | 760.2 |
| Acanthurus bahianus | Ocean surgeon | 34 | 0.5 | 1.00 | 20 | 32.3\% | 7.0 | 2 | 14 | 494.0 |
| Acanthurus chirurgus | Doctorfish | 26 | 0.4 | 0.92 | 15 | 24.2\% | 10.2 | 3 | 30 | 1,376.3 |
| Acanthurus coeruleus | Blue tang | 45 | 0.7 | 1.33 | 26 | 41.9\% | 10.2 | 2 | 22 | 1,894.3 |
| Aluterus scriptus | Scrawled filefish | , | 0.0 | 0.13 | 1 | 1.6\% | 23.0 | 23 | 23 | 242.7 |
| Anisotremus virginicus | Porkfish | 79 | 1.3 | 2.64 | 27 | 43.5\% | 4.5 | 1 | 17 | 458.2 |
| Archosargus rhomboidalis | Sea bream | 2 | 0.0 | 0.18 | 2 | 3.2\% | 16.5 | 16 | 17 | 215.9 |
| Bodianus rufus | Spanish hogfish |  | 0.0 | 0.13 | 1 | 1.6\% | 3.0 | 3 | 3 | 0.4 |
| Calamus calamus | Saucereye porgy | , | 0.0 | 0.13 | 1 | 1.6\% | 23.0 | 23 | 23 | 279.4 |
| Canthigaster rostrata | Sharpnose puffer |  | 0.0 | 0.13 | 1 | 1.6\% | 5.0 | 5 | 5 | 2.2 |
| Caranx bartholomaei | Yellow jack | 3 | 0.0 | 0.38 | 1 | 1.6\% | 60.0 | 60 | 60 | 11,536.6 |
| Caranx ruber | Bar jack | 113 | 1.8 | 9.82 | 9 | 14.5\% | 8.8 | 6 | 36 | 6,414.3 |
| Chaetodon capistratus | Foureye butterflyfish | 24 | 0.4 | 0.58 | 21 | 33.9\% | 6.0 | 4 | 10 | 191.3 |
| Chaetodiperus faber | Atlantic spadefish | 18 | 0.3 | 1.05 | 7 | 11.3\% | 32.5 | 25 | 42 | 20,231.9 |
| Chaetodon ocellatus | Spotif butterflyfish | 12 | 0.2 | 0.62 | 6 | 9.7\% | 9.7 | 8 | 14 | 371.1 |
| Chaetodon sedentarius | Reef butterlyfish | 2 | 0.0 | 0.18 | 2 | 3.2\% | 5.0 | 4 | 6 | 8.0 |
| Coryphopterus dicrus | Colon goby | 3 | 0.0 | 0.22 | 3 | 4.8\% | 3.3 | 2 | 5 | 2.0 |
| Coryphopterus glaucofraenum | Bridled goby | 29 | 0.5 | 1.35 | 10 | 16.1\% | 3.2 | 2 | 5 | 13.2 |
| Diodon holocanthus | Balloonfish | 3 | 0.0 | 0.22 | 3 | 4.8\% | 14.7 | 14 | 15 | 300.3 |
| Diodon hystrix | Porcupinefish | 1 | 0.0 | 0.13 | 1 | 1.6\% | 10.0 | 10 | 10 | 100.6 |
| Diplodus holbrooki | Spottail pinfish | 1 | 0.0 | 0.13 | 1 | 1.6\% | 16.0 | 16 | 16 | 85.0 |
| Equetus punctatus | Spotted drum | 1 | 0.0 | 0.13 | 1 | 1.6\% | 6.0 | 6 | 6 | 2.7 |
| Gerres cinereus | Yellowfin mojarra | 2 | 0.0 | 0.18 | 2 | 3.2\% | 28.0 | 26 | 30 | 1,087.4 |
| Gobiosoma oceanops | Neon goby | 11 | 0.2 | 0.59 | 6 | 9.7\% | 3.0 | 3 | 3 | 2.8 |
| Gymnothorax moringa | Spotted moray | 1 | 0.0 | 0.13 | 1 | 1.6\% | 40.0 | 40 | 40 | 114.9 |
| Haemulon aurolineatum | Tomtate | 338 | 5.5 | 10.59 | 19 | 30.6\% | 5.7 | 2 | 15 | 1,700.9 |
| Haemulon flavolineatum | French grunt | 84 | 1.4 | 3.76 | 13 | 21.0\% | 6.1 | 3 | 11 | 424.1 |
| Haemulon macrostomum | Spanish grunt | 5 | 0.1 | 0.45 | 2 | 3.2\% | 8.4 | 4 | 15 | 183.4 |
| Haemulon parra | Sailor's choice | 3 | 0.0 | 0.38 | 1 | 1.6\% | 14.0 | 14 | 14 | 160.8 |
| Haemulon plumieri | White grunt | 1,241 | 20.0 | 23.34 | 55 | 88.7\% | 8.7 | 2 | 25 | 21,365.0 |
| Haemulon sciurus | Bluestriped grunt | 318 | 5.1 | 5.72 | 46 | 74.2\% | 12.1 | 3 | 30 | 16,701.0 |
| Haemulon sp. | Unidentified grunt | 446 | 7.2 | 45.42 | 4 | 6.5\% | 3.2 | 3 | 4 | 213.5 |
| Halichoeres bivittatus | Slippery dick | 55 | 0.9 | 2.76 | 16 | 25.8\% | 7.1 | 4 | 12 | 277.6 |
| Halichoeres garnoti | Yellowhead wrasse | 2 | 0.0 | 0.25 | 1 | 1.6\% | 5.0 | 5 | 5 | 2.4 |
| Halichoeres maculipinna | Clown wrasse | 3 | 0.0 | 0.28 | 2 | 3.2\% | 8.7 | 7 | 11 | 29.0 |
| Halichoeres radiatus | Puddingwife | 2 | 0.0 | 0.25 | 1 | 1.6\% | 6.0 | 4 | 7 | 5.7 |
| Holacanthus bermudensis | Blue angelfish | 14 | 0.2 | 0.49 | 12 | 19.4\% | 13.5 | 4 | 25 | 1,360.0 |
| Holacanthus ciliaris | Queen anglefish | 5 | 0.1 | 0.27 | 5 | 8.1\% | 11.2 | 3 | 18 | 266.0 |
| Hypoplectrus unicolor | Butter hamlet | 3 | 0.0 | 0.28 | 2 | 3.2\% | 5.7 | 5 | 6 | 1.6 |
| Lactophrys bicaudalis | Spotted trunkfish | 1 | 0.0 | 0.13 | 1 | 1.6\% | 26.0 | 26 | 26 | 416.5 |
| Lachnolaimus maximus | Hogfish | 10 | 0.2 | 0.41 | 9 | 14.5\% | 31.6 | 12 | 50 | 8,812.2 |
| Lutjanus analis | Mutton snapper | 7 | 0.1 | 0.32 | 7 | 11.3\% | 38.4 | 27 | 55 | 7,641.2 |
| Lutjanus apodus | Schoolmaster | 7 | 0.1 | 0.41 | 5 | 8.1\% | 22.9 | 14 | 30 | 1,820.9 |
| Lutjanus griseus | Gray snapper | 263 | 4.2 | 6.08 | 45 | 72.6\% | 15.9 | 4 | 40 | 22,236.2 |
| Lutjanus mahogoni | Mahogany snapper | 10 | 0.2 | 0.91 | 2 | 3.2\% | 14.8 | 12 | 17 | 682.0 |
| Lutjanus synagris | Lane snapper | 69 | 1.1 | 3.93 | 9 | 14.5\% | 11.5 | 6 | 16 | 2,336.8 |
| Mycteroperca bonaci | Black grouper | 7 | 0.1 | 0.37 | 6 | 9.7\% | 29.1 | 16 | 40 | 2,861.5 |
| Ocyurus chrysurus | Yellowtail snapper | 392 | 6.3 | 17.33 | 43 | 69.4\% | 10.8 | 5 | 18 | 10,881.3 |
| Pomacanthus arcuatus | Gray angelfish | 39 | 0.6 | 1.01 | 24 | 38.7\% | 20.7 | 7 | 35 | 13,766.9 |
| Pomacentrus fuscus | Dusky damselfish | 12 | 0.2 | 0.51 | 9 | 14.5\% | 4.7 | 2 | 7 | 46.9 |
| Pomacentrus leucostictus | Beaugregory | 35 | 0.6 | 0.99 | 21 | 33.9\% | 4.9 | 2 | 8 | 130.5 |
| Pomacentrus partitus | Biccilor damselfish | 3 | 0.0 | 0.22 | 3 | 4.8\% | 4.7 | 4 | 6 | 8.0 |
| Pomacanthus paru | Frerich angelfish | 7 | 0.1 | 0.45 | 4 | 6.5\% | 26.1 | 8 | 35 | 4,498.4 |
| Pomacentrus planifrons | Three spot damselfish | 125 | 2.0 | 3.15 | 37 | 59.7\% | 5.2 | 2 | 8 | 625.1 |
| Pomacentrus variabilis | Cocoa damselfish | 49 | 0.8 | 1.01 | 27 | 43.5\% | 5.7 | 4 | 10 | 240.3 |
| Pseudupeneus maculatus | Spoited goattish | 7 | 0.1 | 0.37 | 6 | 9.7\% | 11.7 | 9 | 18 | 233.6 |
| Scarus coelestinus | Midnight parrotfish | 3 | 0.0 | 0.22 | 3 | 4.8\% | 27.3 | 21 | 39 | 1,508.6 |
| Scarus coeruleus | Blue parrotish | 6 | 0.1 | 0.30 | 6 | 9.7\% | 21.3 | 3 | 32 | 1,648.5 |
| Scarus croicensis | Striped parrotfish | 678 | 10.9 | 8.12 | 50 | 80.6\% | 4.9 | 1 | 12 | 1,435.3 |
| Scarus taeniopterus | Princess parrotfish | 38 | 0.6 | 2.49 | 8 | 12.9\% | 6.9 | 1 | 15 | 340.1 |
| Sparisoma aurofrenatum | Redband parrottish | 44 | 0.7 | 1.38 | 19 | 30.6\% | 7.3 | 4 | 16 | 399.5 |

Table 8 （cont．）

| SPECIES | COMMON NAME | NUMBER ${ }^{\text {a }}$ Sample $\mathrm{N}=62$ |  |  |  |  | Number of species $=69$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total Abund． | Mean Abund． | Stand． Dev． | OCCURENCE <br> Freq．$\quad \%$ |  | LENGTH（cm） |  |  | $\begin{array}{c\|} \hline \text { BIOMASS (gms) } \\ \hline \text { Total } \end{array}$ |
|  |  |  |  |  |  |  | Mean | Min | Max |  |
| ＝＝＝＝＝＝＝－ |  | － | ＝＝ | $=$ | ＝ | ＝ | ＝＝ | $=$ | ＝：＝ | ＝ニニニニニ＝＝＝＝＝＝＝ |
| Sparisoma chrysopterum | Redtail parrotfish | 2 | 0.0 | 0.18 | 2 | 3．2\％ | 14.0 | 12 | 16 | 91.2 |
| Sparisoma radians | Bucktooth parrotfish | 1 | 0.0 | 0.13 | 1 | 1．6\％ | 10.0 | 10 | 10 | 12.6 |
| Sparisoma rubripinne | Yellowtail parrotfish | 1 | 0.0 | 0.13 | 1 | 1．6\％ | 14.0 | 14 | 14 | 50.8 |
| Sparisoma viride | Stoplight parrotfish | 32 | 0.5 | 0.80 | 23 | 37．1\％ | 9.7 | 3 | 30 | 1，424．3 |
| Sphyraena barracuda | Barracuda | 7 | 0.1 | 0.41 | 5 | 8．1\％ | 77.1 | 40 | 120 | 34，936．0 |
| Synodus intermedius | Sand diver | 1 | 0.0 | 0.13 | 1 | 1．6\％ | 13.0 | 13 | 13 | 21.7 |
| Thalassoma bifasciatum | Bluehead | 26 | 0.4 | 0.69 | 20 | 32．3\％ | 6.6 | 2 | 10 | 81.9 |
| Urolophus jamaicensis | Yellow stingray | 1 | 0.0 | 0.13 | 1 | 1．6\％ | 30.0 | 30 | 30 | 257.9 |
|  |  | 4，938 |  |  |  |  |  |  |  | 208，323．5 |

Table 9. Summary of reef fish censused at Mosquito Bank new (open to fishing) in John Pennekamp Coral Reef State Park, Florida, May 31, 1994 - June 27, 1995.

| SPECIES | COMMON NAME | NUMBER Sample $\mathrm{N}=24$ Number of species $=56$ |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total Abund. | Mean Stand. <br> Abund. Dev. |  | OCCURENCE |  |  | TGTH | (cm) | BIOMASS (gms) |
|  |  |  |  |  | Freq. | \% | Mean | Min | Max | Total |
|  |  |  |  |  |  |  |  |  |  |  |
| Abudefduf saxatilis | Sergeant major | 16 | 0.7 | 2.24 | 3 | 12.5\% | 6.1 | 2 | 10 | 133.7 |
| Acanthurus chirurgus | Doctorfish | 5 | 0.2 | 0.41 | 5 | 20.8\% | 12.6 | 6 | 20 | 281.6 |
| Acanthurus coeruleus | Blue tang | 23 | 1.0 | 1.68 | 12 | 50.0\% | 13.2 | 3 | 20 | 1,830.4 |
| Anisotremus virginicus | Porkfish | 18 | 0.8 | 2.11 | 6 | 25.0\% | 8.6 | 1 | 18 | 606.4 |
| Archosargus rhomboidalis | Sea bream | 7 | 0.3 | 1.00 | 2 | 8.3\% | 14.6 | 13 | 15 | 516.4 |
| Calamus calamus | Saucereye porgy | 2 | 0.1 | 0.28 | 2 | 8.3\% | 25.0 | 20 | 30 | 777.0 |
| Caranx ruber | Bar jack | 5 | 0.2 | 0.51 | 4 | 16.7\% | 16.2 | 5 | 38 | 1,099.0 |
| Chaetodon capistratus | Foureye butterflyfish | 6 | 0.3 | 0.44 | 6 | 25.0\% | 6.2 |  | 8 | 49.5 |
| Chaetodiperus faber | Atlantic spadefish |  | 0.0 | 0.20 | 1 | 4.2\% | 26.0 | 26 | 26 | 579.9 |
| Coryphopterus glaucofraenum | Bridled goby | 3 | 0.1 | 0.61 | 1 | 4.2\% | 3.0 | 3 | 3 | 1.0 |
| Coryphopterus personatus | Masked goby | 2 | 0.1 | 0.41 | 1 | 4.2\% | 3.0 | 3 | 3 | 0.7 |
| Diodon holocanthus | Balloonfish | 1 | 0.0 | 0.20 | 1 | 4.2\% | 17.0 | 17 | 17 | 142.4 |
| Epinephelus adscensionis | Rock hind | 1 | 0.0 | 0.20 | 1 | 4.2\% | 20.0 | 20 | 20 | 124.1 |
| Epinephelus morio | Red grouper | 4 | 0.2 | 0.38 | 4 | 16.7\% | 17.8 | 16 | 22 | 322.8 |
| Epinephelus striatus | Nassau grouper | 1 | 0.0 | 0.20 | 1 | 4.2\% | 38.0 | 38 | 38 | 820.8 |
| Gerres cinereus | Yellowfin mojarra | 16 | 0.7 | 2.26 | 2 | 8.3\% | 20.5 | 15 | 25 | 3,405.2 |
| Gobiosoma oceanops | Neon goby | 10 | 0.4 | 1.35 | 3 | 12.5\% | 2.6 | 1 | 4 | 1.9 |
| Gymnothorax funebris | Green moray | 1 | 0.0 | 0.20 | 1 | 4.2\% | 8.0 | 8 | 8 | 1.6 |
| Haemulon aurolineatum | Tomtate | 313 | 13.0 | 25.66 | 8 | 33.3\% | 8.1 | 4 | 14 | 2,899.6 |
| Haemulon carbonarium | Caesar grunt | 1 | 0.0 | 0.20 | 1 | 4.2\% | 10.0 | 10 | 10 | 16.7 |
| Haemulon chrysargyreum | Smallmouth grunt | 1 | 0.0 | 0.20 | 1 | 4.2\% | 8.0 | 8 | 8 | 35.2 |
| Haemulon flavolineatum | French grunt | 31 | 1.3 | 3.41 | 6 | 25.0\% | 5.3 | 4 | 8 | 89.5 |
| Haemulon plumieri | White grunt | 255 | 10.6 | 13.84 | 18 | 75.0\% | 8.9 | 4 | 24 | 5,282.1 |
| Haemulon sciurus | Bluestriped grunt | 38 | 1.6 | 3.57 | 10 | 41.7\% | 14.0 | 5 | 24 | 2,734.6 |
| Haemulon sp. | Unidentified grunt | 6 | 0.3 | 1.22 | 1 | 4.2\% | 4.0 | 4 | 4 | 5.7 |
| Halichoeres bivittatus | Slippery dick | 20 | 0.8 | 2.46 | 7 | 29.2\% | 7.3 | 3 | 10 | 96.1 |
| Halichoeres garnoti | Yellowhead wrasse | 1 | 0.0 | 0.20 | 1 | 4.2\% | 7.0 | 7 | 7 | 3.7 |
| Halichoeres maculipinna | Clown wrasse | 2 | 0.1 | 0.41 | 1 | 4.2\% | 6.0 | 4 | 8 | 6.4 |
| Holacanthus bermudensis | Blue angelfish | 7 | 0.3 | 0.55 | 6 | 25.0\% | 13.6 | 7 | 17 | 485.6 |
| Holacanthus ciliaris | Queen anglefish | 3 | 0.1 | 0.45 | 2 | 8.3\% | 19.3 | 12 | 24 | 648.2 |
| Hyploplectrus nigricans | Black hamlet | 1 | 0.0 | 0.20 | 1 | 4.2\% | 4.0 | , | 4 | 0.2 |
| Hypoplectrus unicolor | Butter hamlet | 3 | 0.1 | 0.34 | 3 | 12.5\% | 6.3 | 4 | 9 | 2.8 |
| Lachnolaimus maximus | Hogfish | 19 | 0.8 | 1.06 | 11 | 45.8\% | 27.7 | 10 | 51 | 13,587.9 |
| Lutjanus apodus | Schoolmaster | 8 | 0.3 | 0.56 | 7 | 29.2\% | 24.6 | 13 | 42 | 3,081.7 |
| Lutjanus buccanella | Blackfin snapper | 1 | 0.0 | 0.20 | 1 | 4.2\% | 4.0 | 4 | 4 | 1.3 |
| Lutjanus griseus | Gray snapper | 103 | 4.3 | 4.69 | 20 | 83.3\% | 16.0 | 8 | 30 | 8,339.2 |
| Lutjanus mahogoni | Mahogany snapper | 5 | 0.2 | 0.72 | 2 | 8.3\% | 13.8 | 12 | 15 | 285.3 |
| Mycteroperca bonaci | Black grouper | 1 | 0.0 | 0.20 | 1 | 4.2\% | 18.0 | 18 | 18 | 72.2 |
| Mycteroperca phenax | Scamp | 1 | 0.0 | 0.20 | 1 | 4.2\% | 16.0 | 16 | 16 | 57.9 |
| Ocyurus chrysurus | Yellowtail snapper | 55 | 2.3 | 3.42 | 12 | 50.0\% | 10.1 | 6 | 20 | 1,359.0 |
| Pomacanthus arcuatus | Gray angelfish | 28 | 1.2 | 1.55 | 10 | 41.7\% | 20.0 | 7 | 32 | 9,574.5 |
| Pomacentrus leucostictus | Beaugregory | 21 | 0.9 | 1.30 | 10 | 41.7\% | 5.5 | 3 | 8 | 104.4 |
| Pomacanthus paru | French angelfish | 4 | 0.2 | 0.38 | 4 | 16.7\% | 27.8 | 20 | 41 | 3,432.1 |
| Pomacentrus planifrons | Three spot damselfish | 60 | 2.5 | 2.65 | 16 | 66.7\% | 5.6 | 2 | 8 | 370.7 |
| Pomacentrus variabilis | Cocoa damselfish | 8 | 0.3 | 0.87 | 5 | 20.8\% | 6.0 | 4 | 8 | 45.7 |
| Scarus coelestinus | Midnight parrotfish | 7 | 0.3 | 0.91 | 3 | 12.5\% | 24.0 | 10 | 30 | 2,309.4 |
| Scarus coeruleus | Blue parrotfish | 1 | 0.0 | 0.20 | 1 | 4.2\% | 30.0 | 30 | 30 | 489.6 |
| Scarus croicensis | Striped parrotish | 176 | 7.3 | 8.72 | 16 | 66.7\% | 5.6 | 2 | 14 | 694.5 |
| Scarus guacamaia | Rainbow parrotish | 1 | 0.0 | 0.20 | 1 | 4.2\% | 60.0 | 60 | 60 | 4,336.0 |
| Scarus taeniopterus | Princess parrotish | 48 | 2.0 | 7.24 | 4 | 16.7\% | 5.7 | 3 | 9 | 194.3 |
| Sparisoma atomarium | Greenblotch parrotish | 1 | 0.0 | 0.20 | 1 | 4.2\% | 4.0 | 4 | 4 | 0.8 |
| Sparisoma aurofrenatum | Redband parrotish | 13 | 0.5 | 1.22 | 6 | 25.0\% | 8.7 | 5 | 15 | 163.4 |
| Sparisoma chrysopterum | Redtail parrotfish | 14 | 0.6 | 1.84 | 3 | 12.5\% | 11.7 | 9 | 20 | 415.1 |
| Sparisoma viride | Stoplight parrotfish | 34 | 1.4 | 2.89 | 12 | 50.0\% | 8.1 | 3 | 28 | 1,082.5 |
| Sphyraena barracuda | Barracuda | 2 | 0.1 | 0.28 | 2 | 8.3\% | 90.0 | 80 | 100 | 10,909.8 |
| Thalassoma bifasciatum | Bluehead | 3 | 0.1 | 0.45 | 2 | 8.3\% | 10.0 | 10 | 10 | 24.2 |
|  |  | 1,418 |  |  |  |  |  |  |  | 83,932.1 |

Table 10. Comparison of reef fish at patch reefs based on Index of Relative Dominance (IRD) values for the top ten species censused at John Pennekamp Coral Reef State Park, Florida, Florida, May 15, 1992 - June 27, 1995. IRD ranks are presented as overall for the park for all years and patch reef by year.

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Table 10 （cont．）

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Figure 1. Map depicting study sites at John Pennekamp Coral Reef State Park, Florida. Mosquito Bank (MB) and Basin Hill ( BH ) each contain three study reefs, with reefs either open (O) or closed (C) to fishing. Numbers depict historic (1) or recent reef additions (2) to the study. Latitude and longitude are presented for each patch reef. Exclamation marks represent buoys.


Figure 2. Mean number of censused species, individuals, ( + or $-95 \% \mathrm{CI}$ ) and biomass per sample ( + or - SE) at study reefs. Stars show significant differences between paired locations. N shows size per reef. 2,151 individuals of an unknown species was not included in SP95. $\mathrm{N}=$ number of samples per site.


Figure 2 (cont.)


Figure 3. Mean biomass per sample of major reef fish families and groups sampled at Basin Hill reef in John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995. Bars indicate + or -SE .


Figure 3 (cont.)


Figure 4. Mean biomass per sample of major reef fish families and groups sampled at Mosquito Bank reef in John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995. Bars indicate + or -SE .


Figure 4 (cont.)


Figure 5. Dendrogram from Bray-Curtis similarity matrix of reef fish visual census samples for John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995. Data analyzed were pooled mean species abundance data (IRD $>0.05, \mathrm{n}=48$ ) for study reefs by season.


Figure 6. Dendrogram from Bray-Curtis similarity matrix of reef fish visual census samples for John Pennekamp Coral Reef State Park, Florida, May 15, 1992 - June 27, 1995. Data analyzed were pooled mean species abundance data (IRD $>0.05, \mathrm{n}=48$ ) for study reefs by season.


# Chapter 4. Summary of Dry Tortugas Research 

## James A. Bohnsack and David B. McClellan

## Fishery Landings Trends

In 1994 we summarized commercial and recreational fishery trends for Monroe County, including the Dry Tortugas (Bohnsack, et al., 1994). In that publication headboat landings were reported by weight (Table 3) and numbers (Table 6) from the Dry Tortugas. The Dry Tortugas accounted for $23 \%$ of the total $5.8 \times 10^{3}$ fish landed by headboats. Reef fishes accounted for $97 \%$ of Dry Tortugas headboat landings. Interestingly, between 1989 through 1991, more grouper were landed from the Dry Tortugas, despite the smaller total area, than the rest of the Florida Keys.

## Visual Reef Fish Assessments

In 1978, data were collected on fish assemblages on isolated coral heads and observations made on the effects of a severe cold snap on corals in the Dry Tortugas (Bohnsack 1983).

With support from NOAA, NURC, and BRD of USGS, we have been involved in various cruises conducted annually in the Dry Tortugas between 1994 and 1997 where data were collected to assess reef fish community structure on 30 reef sites inside and outside the Dry Tortugas National Park (DTNP) (Fig. 1). Data were collected using a stationary point sampling technique (Bohnsack and Bannerot 1986) and a 15 min swimming predator search (Bohnsack 1982). A total of 518 stationary samples were collected from 9 reefs in DTNP, 20 reefs in FKNMS, and 1 reef, Sherwood Forest, outside the FKNMS boundary in the Gulf of Mexico (Table 1). A total of 162 species (76,408 individual fish) were observed in this effort (Table 2) and are statistically summarized in Table 3.

A total of 129 predator searches were also conducted on 11 reefs in DTNP and 6 reefs in FKNMS between 1994 and 1997 (Table 4). During these surveys, 39 piscivorous predatory species were observed (Table 5), including 11,794 individuals. Surveys are statistically summarized in Table 6.

In 1998, divers observed 88 species ( 6,961 individuals) in 80 visual point samples from seven reefs. During 12 predator searches, NMFS divers observed 20 predatory species ( 367 individuals) on 3 reefs.

## Comparison of Reef Fish Assemblages inside DTNP and outside DTNP

Cluster analysis was used to distinguish fish assemblages between reefs inside versus outside Dry Tortugas National Park boundaries (Fig. 2). A comparison was made of fishes on reefs inside and outside DTNP with data collected through1997 (Table 7, Fig. 3). There were obvious habitat differences between sites inside and outside DTNP. Sites inside DTNP tended to be shallower (Fig. 4), more turbid, and more likely to have sand substrate and seagrasses. The substrate had higher cover of coral or 'rock' outside DTNP (Fig. 4). Performance of various parameters were analyzed by reef and compared between sites inside and outside DTNP. Reefs outside DTNP tended to have a higher average number of individuals (mostly planktivorous damselfishes, Fig. 3), species, total biomass (Fig. 5) as well as being deeper and with more hard substrate (Table 8, Fig, 4). Performance analyses showed total snapper were more variable outside DTNP (Fig. 6A) primarily because of large schools of yellowtail snapper (Ocyurus chryurus) (Fig. 8C). Mean total grouper varied similarly inside and outside DTNP (Fig. 6B). Hogfish (Lachnolaimus maximus) were more consistently observed and less variable inside DTNP but at lower abundances (Fig. 6C). Among
individual species, black grouper (Mycteroperca bonaci) were more frequently seen in DTNP than outside (Fig. 7C). As with yellowtail, the mean abundance of mutton snapper (Lutjanus apodus) was more variable outside DTNP (Fig. 7B).

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Bohnsack, J.A., and S.P. Bannerot. 1986. A stationary visual census technique for quantitatively assessing community structure of coral reef fishes. NOAA Tech. Rep. NMFS 41:1-15.
Bohnsack, J.A., D.E. Harper, and D.B. McClellan. 1994. Fisheries trends from Monroe County, Florida. Bull. Mar. Sci. 54:982-1018.

Table 1. Summary of Dry Tortugas stationary point sampling effort (1994-1997).


Table 2. Cumulative species listing for all 1994-1997 stationary point sampling. SPCODE is species code derived from first 3 letters of genera and first four letters from species. NUM is number of individuals observed.

| SPCODE | NUM SPECIES | COMMON NAME | FAMILY | FAMILY NAME |
| :---: | :---: | :---: | :---: | :---: |
| 1 ABU SAXA | 10 Abudefduf saxatilis | Sergeant major | POMACENTRIDAE | Damselfishes |
| 2 ACA BAHI | 30 Acanthurus bahianus | Ocean surgeon | ACANTHURIDAE | Surgeonfishes |
| 3 ACA CHIR | 50 Acanthurus chirurgus | Doctorfish | ACANTHURIDAE | Surgeonfishes |
| 4 ACA COER | 60 Acanthurus coeruleus | Blue tang | ACANTHURIDAE | Surgeonfishes |
| 5 AET NARI | 66 Aetobatus narinari | Spotted eagle ray | MYLIOBATIDAE | Eagle rays |
| 6 ALE CILI | 70 Alectis ciliaris | African pompano | CARANGIDAE | Jacks |
| 7 ALU SCRI | 90 Aluterus scriptus | Scrawled filefish | BALISTIDAE | Leatherjackets |
| 8 ANI VIRG | 120 Anisotremus virginicus | Porkfish | HAEMULIDAE | Grunts |
| 9 APO BINO | 125 Apogon binotatus | Barred cardinalfish | APOGONIDAE | Cardinalfishes |
| 10 AUL MACU | 180 Aulostomus maculatus | Trumpetfish | AULOSTOMIDAE | Trumpetfishes |
| 11 BAL VETU | 200 Balistes vetula | Queen triggerfish | BALISTIDAE | Leatherjackets |
| 12 BLE CRIS | 210 Blennius cristata | Molly miller | BLENNIIDAE | Combtooth blennies |
| 13 BOD RUFU | 220 Bodianus rufus | Spanish hogfish | LABRIDAE | Wrasses |
| 14 CAL BAJO | 240 Calamus bajonado | Jolthead porgy | SPARIDAE | Porgies |
| 15 CAL CALA | 260 Calamus calamus | Saucereye porgy | SPARIDAE | Porgies |
| 16 CAN MACR | 280 Cantherhines macrocerus | Whitespotted filefish | BALISTIDAE | Leatherjackets |
| 17 CAN PULL | 290 Cantherhines pullus | Orangespotted filefish | BALISTIDAE | Leatherjackets |
| 18 CAN ROST | 300 Canthigaster rostrata | Sharpnose puffer | TETRAODONTIDAE | Puffers |
| 19 CAN SUFF | 310 Canthidermis sufflamen | Ocean triggerfish | BALISTIDAE | Leatherjackets |
| 20 CAR BART | 320 Caranx bartholomaei | Yellow jack | CARANGIDAE | Jacks |
| 21 CAR CRYS | 330 Caranx crysos | Blue runner | CARANGIDAE | Jacks |
| 22 CAR HIPP | 340 Caranx hippos | Crevalle jack | CARANGIDAE | Jacks |
| 23 CAR RUBE | 350 Caranx ruber | Bar jack | CARANGIDAE | Jacks |
| 24 CHA CAPI | 370 Chaetodon capistratus | Foureye butterflyfish | CHAETODONTIDAE | Butterflyfishes |
| 25 CHA OCEL | 390 Chaetodon ocellatus | Spotfin butterflyfish | CHAETODONTIDAE | Butterflyfishes |
| 26 CHA SEDE | 400 Chaetodon sedentarius | Reef butterflyfish | CHAETODONTIDAE | Butterflyfishes |
| 27 CHA STRI | 410 Chaetodon striatus | Banded butterflyfish | CHAETODONTIDAE | Butterflyfishes |
| 28 CHR CYAN | 430 Chromis cyaneus | Blue chromis | POMACENTRIDAE | Damselfishes |
| 29 CHR ENCH | 435 Chromis enchrysurus | Yellowtail reeffish | POMACENTRIDAE | Damselfishes |
| 30 CHR INSO | 440 Chromis insolatus | Sunshinefish | POMACENTRIDAE | Damselfishes |
| 31 CHR MULT | 450 Chromis multilineatus | Brown chromis | POMACENTRIDAE | Damselfishes |
| 32 CHR SCOT | 460 Chromis scotti | Purple reeffish | POMACENTRIDAE | Damselfishes |
| 33 CLE PARR | 470 Clepticus parrai | Creole wrasse | LABRIDAE | Wrasses |
| 34 COR DICR | 480 Coryphopterus dicrus | Colon goby | GOBIIDAE | Gobies |
| 35 COR EIDO | 485 Coryphopterus eidolon | Pallid goby | GOBIIDAE | Gobies |
| 36 COR GLAU | 490 Coryphopterus glaucofraenum | Bridled goby | GOBIIDAE | Gobies |
| 37 COR PERS | 493 Coryphopterus personatus | Masked goby | GOBIIDAE | Gobies |
| 38 COR SPE. | 495 Coryphopterus species | Unknown goby | GOBIIDAE | Gobies |
| 39 CRY ROSE | 510 Cryptotomus roseus | Bluelip parrotfish | SCARIDAE | Parrotfishes |
| 40 DAS AMER | 530 Dasyatis americana | Southern stingray | DASYATIDAE | Stingrays |
| 41 DEC MACA | 540 Decapterus macarellus | Mackerel scad | CARANGIDAE | Jacks |
| 42 DIO HYST | 570 Diodon hystrix | Porcupinefish | DIODONTIDAE | Porcupinefishes |
| 43 DIP ARGE | 577 Diplodus argenteus | Silver porgy | SPARIDAE | Porgies |
| 44 ECH NAUC | 590 Echeneis naucrates | Sharksucker | ECHENEIDAE | Remoras |
| 45 EPI ADSC | 650 Epinephelus adscensionis | Rock hind | SERRANIDAE | Sea basses |
| 46 EPI CRUE | 660 Epinephelus cruentatus | Graysby | SERRANIDAE | Sea basses |
| 47 EPI FULV | 675 Epinephelus fulvus | Coney | SERRANIDAE | Sea basses |
| 48 EPI GUTT | 680 Epinephelus guttatus | Red hind | SERRANIDAE | Sea basses |
| 49 EPI ITAJ | 685 Epinephelus itajara | Jewfish | SERRANIDAE | Sea basses |
| 50 EPI MORI | 690 Epinephelus morio | Red grouper | SERRANIDAE | Sea basses |
| 51 EPI STRI | 710 Epinephelus striatus | Nassau grouper | SERRANIDAE | Sea basses |
| 52 EQU ACUM | 720 Equetus acuminatus | High-hat | SCIAENIDAE | Drums |
| 53 GIN CIRR | 760 Ginglymostoma cirratum | Nurse shark | ORECTOLOBIDAE | Carpet sharks |
| 54 GNA THOM | 770 Gnatholepis thompsoni | Goldspot goby | GOBIIDAE | Gobies |
| 55 GOB OCEA | 790 Gobiosoma oceanops | Neon goby | GOBIIDAE | Gobies |
| 56 GOB RAND | 793 Gobiosoma randalli | Yellownose goby | GOBIIDAE | Gobies |
| 57 GOB SPE. | 795 Goby-like fish | Goby-like fish | GOBIIDAE | Gobies |
| 58 GYM MORI | 830 Gymnothorax moringa | Spotted moray | MURAENIDAE | Morays |
| 59 HAE AURO | 870 Haemulon aurolineatum | Tomtate | HAEMULIDAE | Grunts |
| 60 HAE CARB | 880 Haemulon carbonarium | Caesar grunt | HAEMULIDAE | Grunts |
| 61 HAE CHRY | 890 Haemulon chrysargyreum | Smallmouth grunt | HAEMULIDAE | Grunts |

Table 2 (cont.)

| $\begin{aligned} & \text { SPCODE } \\ & ========= \end{aligned}$ | NUM SPECIES $================================$ | COMMON NAME $\qquad$ | FAMILY | FAMILY NAME $\qquad$ |
| :---: | :---: | :---: | :---: | :---: |
| 62 HAE FLAV | 900 Haemulon flavolineatum | French grunt | HAEMULIDAE | Grunts |
| 63 HAE MACR | 910 Haemulon macrostomum | Spanish grunt | HAEMULIDAE | Grunts |
| 64 HAE MELA | 920 Haemulon melanurum | Cottonwick | HAEMULIDAE | Grunts |
| 65 HAE PARR | 930 Haemulon parrai | Sailor's choice | HAEMULIDAE | Grunts |
| 66 HAE PLUM | 940 Haemulon plumieri | White grunt | HAEMULIDAE | Grunts |
| 67 HAE SCIU | 950 Haemulon sciurus | Bluestriped grunt | HAEMULIDAE | Grunts |
| 68 HAE SPE. | 955 Haemulon sp. | Unidentified grunt | HAEMULIDAE | Grunts |
| 69 HAE STRI | 960 Haemulon striatum | Striped grunt | HAEMULIDAE | Grunts |
| 70 HAL BIVI | 970 Halichoeres bivittatus | Slippery dick | LABRIDAE | Wrasses |
| 71 HAL CYAN | 975 Halichoeres cyanocephalus | Yellowcheek wrasse | LABRIDAE | Wrasses |
| 72 HAL GARN | 980 Halichoeres garnoti | Yellowhead wrasse | LABRIDAE | Wrasses |
| 73 HAL MACU | 990 Halichoeres maculipinna | Clown wrasse | LABRIDAE | Wrasses |
| 74 HAL POEY | 1010 Halichoeres poeyi | Blackear wrasse | LABRIDAE | Wrasses |
| 75 HAL RADI | 1020 Halichoeres radiatus | Puddingwife | LABRIDAE | Wrasses |
| 76 HEM BRAS | 1030 Hemiramphus brasiliensis | Ballyhoo | EXOCETIDAE | Flyingfishes/Halfbeaks |
| 77 HEM MART | 1035 Hemipteronotus martinicensis | Rosy razorfish | LABRIDAE | Wrasses |
| 78 HEM SIMU | 1050 Hemiemblemaria simulus | Wrasse blenny | CLINIDAE | Clinids |
| 79 HOL ADSC | 1070 Holocentrus adscensionis | Squirrelfish | HOLOCENTRIDAE | Squirrelfishes |
| 80 HOL BERM | 1080 Holacanthus bermudensis | Blue angelfish | POMACANTHIDAE | Angelfishes |
| 81 HOL CILI | 1090 Holacanthus ciliaris | Queen angelfish | POMACANTHIDAE | Angelfishes |
| 82 HOL RUFU | 1120 Holocentrus rufus | Longspine squirrelfish | HOLOCENTRIDAE | Squirrelfishes |
| 83 HOL TOWN | 1128 Holacanthus (bermudensis x ciliaris) | Townsend angelfish | POMACANTHIDAE | Angelfishes |
| 84 HOL TRIC | 1130 Holacanthus tricolor | Rock beauty | POMACANTHIDAE | Angelfishes |
| 85 HOL VEXI | 1140 Holocentrus vexillarius | Dusky squirrelfish | HOLOCENTRIDAE | Squirrelfishes |
| 86 HYP BERM | 1150 Hypleurochilus bermudensis | Barred blenny | BLENNIIDAE | Combtooth blennies |
| 87 HYP GEMM | 1160 Hypoplectrus gemma \# | Blue hamlet | SERRANIDAE | Sea basses |
| 88 HYP GUTT | 1162 Hypoplectrus guttavarius \# | Shy hamlet | SERRANIDAE | Sea basses |
| 89 HYP HYBR | 1165 Hypoplectrus (hybrid) \# | Hybrid hamlet | SERRANIDAE | Sea basses |
| 90 HYP INDI | 1166 Hypoplectrus indigo \# | Indigo hamlet | SERRANIDAE | Sea basses |
| 91 HYP NIGR | 1170 Hyploplectrus nigricans \# | Black hamlet | SERRANIDAE | Sea basses |
| 92 HYP PUEL | 1180 Hypoplectrus puella \# | Barred hamlet | SERRANIDAE | Sea basses |
| 93 HYP TANN | 1195 Hypoplectrus (tan) \# | Tan hamlet | SERRANIDAE | Sea basses |
| 94 HYP UNIC | 1190 Hypoplectrus unicolor | Butter hamlet | SERRANIDAE | Sea basses |
| 95 INE VITT | 1200 Inermia vittata | Boga | EMMELICHTHYIDAE | Bonnetmouths |
| 96 IOG CALL | 1210 loglossus calliurus | Blue goby | GOBIIDAE | Gobies |
| 97 IOG HELE | 1215 loglossus helenae | Hovering goby | GOBIIDAE | Gobies |
| 98 KYP SECT | 1230 Kyphosus sectatrix | Bermuda chub | KYPHOSIDAE | Sea chubs |
| 99 LAC BICA | 1240 Lactophrys bicaudalis | Spotted trunkfish | OSTRACIIDAE | Boxfishes |
| 100 LAC MAXI | 1250 Lachnolaimus maximus | Hogfish | LABRIDAE | Wrasses |
| 101 LAC TRIQ | 1290 Lactophrys triqueter | Smooth trunkfish | OSTRACIIDAE | Boxfishes |
| 102 LUT ANAL | 1310 Lutjanus analis | Mutton snapper | LUTJANIDAE | Snappers |
| 103 LUT APOD | 1320 Lutjanus apodus | Schoolmaster | LUTJANIDAE | Snappers |
| 104 LUT GRIS | 1350 Lutjanus griseus | Gray snapper | LUTJANIDAE | Snappers |
| 105 LUT JOCU | 1360 Lutjanus jocu | Dog snapper | LUTJANIDAE | Snappers |
| 106 LUT MAHO | 1370 Lutjanus mahogoni | Mahogany snapper | LUTJANIDAE | Snappers |
| 107 LUT SYNA | 1385 Lutjanus synagris | Lane snapper | LUTJANIDAE | Snappers |
| 108 MAL MACR | 1410 Malacoctenus macrops | Rosy blenny | CLINIDAE | Clinids |
| 109 MAL PLUM | 1420 Malacanthus plumieri | Sand tilefish | MALACANTHIDAE | Tilefishes |
| 110 MAL TRIA | 1430 Malacoctenus triangulatus | Saddled blenny | CLINIDAE | Clinids |
| 111 MAL VERS | 1440 Malacoctenus versicolor | Barfin blenny | CLINIDAE | Clinids |
| 112 MEG ATLA | 1460 Megalops atlanticus | Tarpon | ELOPIDAE | Tarpons |
| 113 MIC CHRY | 1480 Microspathodon chrysurus | Yellowtail damselfish | POMACENTRIDAE | Damselfishes |
| 114 MON TUCK | 1500 Monacanthus tuckeri | Slender filefish | BALISTIDAE | Leatherjackets |
| 115 MUL MART | 1510 Mulloidichthys martinicus | Yellow goatfish | MULLIDAE | Goatfishes |
| 116 MYC BONA | 1540 Mycteroperca bonaci | Black grouper | SERRANIDAE | Sea basses |
| 117 MYC MICR | 1550 Mycteroperca microlepis | Gag | SERRANIDAE | Sea basses |
| 118 MYC PHEN | 1560 Mycteroperca phenax | Scamp | SERRANIDAE | Sea basses |
| 119 OCY CHRY | 1600 Ocyurus chrysurus | Yellowtail snapper | LUTJANIDAE | Snappers |
| 120 ODO DENT | 1610 Odontoscion dentex | Reef crocker | SCIAENIDAE | Drums |
| 121 OPH ATLA | 1630 Ophioblennius atlanticus | Redlip blenny | BLENNIIDAE | Combtooth blennies |
| 122 OPI AURI | 1650 Opistognathus aurifrons | Yellowhead jawfish | OPISTOGNATHIDAE | Jawfishes |

Table 2 (cont.)


Table 3. Statistical summary of Dry Tortugas reef fish visual censuses (1994-1997).

| Species | SAMPLE FREQUENCY |  |  |  | SAMP. FREQ. RANGE\| |  |  | FISH LENGTH (cm) |  |  | BIOMASS <br> Total (gms) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | \% | Mean <br> Abund. | Stand. Dev. | High | Low | Mean | Min. | Max. |  |
| 1 ABU SAXA | 1,426 | 122 | 23.6\% | 2.8 | 8.33 | 56 | 01 | 9.3 | 2 | 13 | 41,321.6 |
| 2 ACA BAHI | 434 | 160 | 30.9\% | 0.8 | 1.95 | 18 | 01 | 10.9 | 3 | 23 | 17,062.2 |
| 3 ACA CHIR | 257 | 100 | 19.3\% | 0.5 | 1.50 | 17 | 01 | 12.2 | 2 | 35 | 18,471.4 |
| 4 ACA COER | 879 | 340 | 65.6\% | 1.7 | 2.87 | 45 | 01 | 12.7 | 2 | 33 | 82,459.2 |
| 5 AET NARI | 1 | 1 | 0.2\% | 0.0 | 0.04 | 1 | 01 | 90.0 | 90 | 90 | 238.1 |
| 6 ALE CILI | 9 | 3 | 0.6\% | 0.0 | 0.25 | 5 | 01 | 82.2 | 70 | 100 | 92,523.9 |
| 7 ALU SCRI | 4 | 4 | 0.8\% | 0.0 | 0.09 | 1 | 01 | 39.3 | 22 | 55 | 2,788.9 |
| 8 ANI VIRG | 82 | 60 | 11.6\% | 0.2 | 0.51 | 5 | 01 | 14.6 | 4 | 36 | 12,445.1 |
| 9 APO BINO | 1 | 1 | 0.2\% | 0.0 | 0.04 | 1 | 01 | 5.0 | 5 | 5 | 2.2 |
| 10 AUL MACU | 19 | 18 | 3.5\% | 0.0 | 0.20 | 2 | 01 | 36.1 | 16 | 60 | 2,738.1 |
| 11 BAL VETU | 2 | 2 | 0.4\% | 0.0 | 0.06 | 1 | 01 | 32.5 | 30 | 35 | 1,803.2 |
| 12 BLE CRIS | 4 | 2 | 0.4\% | 0.0 | 0.14 | 3 | 01 | 4.8 | 4 | 5 | 4.9 |
| 13 BOD RUFU | 60 | 51 | 9.8\% | 0.1 | 0.38 | 3 | 01 | 19.5 | 3 | 36 | 12,068.6 |
| 14 CAL BAJO | 5 | 3 | 0.6\% | 0.0 | 0.15 | 3 | 01 | 28.8 | 12 | 38 | 3,585.7 |
| 15 CAL CALA | 204 | 141 | 27.2\% | 0.4 | 0.75 | 4 | 01 | 17.3 | 3 | 39 | 36,644.5 |
| 16 CAN MACR | 3 | 3 | 0.6\% | 0.0 | 0.08 | 1 | 01 | 24.7 | 15 | 30 | 967.8 |
| 17 CAN PULL | 11 | 9 | 1.7\% | 0.0 | 0.17 | 2 | 01 | 10.9 | 6 | 16 | 407.3 |
| 18 CAN ROST | 122 | 86 | 16.6\% | 0.2 | 0.65 | 8 | 01 | 5.4 | 2 | 10 | 568.0 |
| 19 CAN SUFF | 11 | 9 | 1.7\% | 0.0 | 0.17 | 2 | 01 | 41.6 | 26 | 60 | 19,919.6 |
| 20 CAR BART | 105 | 19 | 3.7\% | 0.2 | 2.31 | 50 | 01 | 33.2 | 18 | 65 | 86,514.6 |
| 21 CAR CRYS | 82 | 12 | 2.3\% | 0.2 | 1.53 | 25 | 01 | 20.4 | 15 | 40 | 17,251.3 |
| 22 CAR HIPP | 1 | 1 | 0.2\% | 0.0 | 0.04 | 1 | 01 | 75.0 | 75 | 75 | 6,944.5 |
| 23 CAR RUBE | 1,002 | 78 | 15.1\% | 1.9 | 14.27 | 300 | 01 | 20.2 | 2 | 70 | 181,564.1 |
| 24 CHA CAPI | 295 | 152 | 29.3\% | 0.6 | 1.02 | 6 | 01 | 7.6 | 2 | 12 | 4,625.3 |
| 25 CHA OCEL | 238 | 135 | 26.1\% | 0.5 | 0.88 | 6 | 01 | 10.0 | 3 | 16 | 7,919.0 |
| 26 CHA SEDE | 116 | 64 | 12.4\% | 0.2 | 0.70 | 7 | 01 | 8.1 | 2 | 13 | 2,190.2 |
| 27 CHA STRI | 19 | 14 | 2.7\% | 0.0 | 0.23 | 2 | 01 | 8.7 | 2 | 12 | 500.1 |
| 28 CHR CYAN | 1,192 | 136 | 26.3\% | 2.3 | 7.27 | 75 | 01 | 5.6 | 1 | 15 | 9,030.8 |
| 29 CHR ENCH | 38 | 8 | 1.5\% | 0.1 | 0.80 | 11 | 01 | 2.3 | 1 | 4 | 13.2 |
| 30 CHR INSO | 5 | 5 | 1.0\% | 0.0 | 0.10 | 1 | 01 | 4.2 | 2 | 6 | 11.1 |
| 31 CHR MULT | 907 | 34 | 6.6\% | 1.8 | 13.09 | 195 | 01 | 8.9 | 4 | 12 | 19,775.6 |
| 32 CHR SCOT | 4,288 | 189 | 36.5\% | 8.3 | 24.87 | 350 | 01 | 4.2 | 1 | 10 | 11,156.8 |
| 33 CLE PARR | 1,135 | 46 | 8.9\% | 2.2 | 12.61 | 200 | 01 | 12.9 | 2 | 30 | 65,402.9 |
| 34 COR DICR | 61 | 31 | 6.0\% | 0.1 | 0.64 | 9 | 01 | 3.4 | 2 | 6 | 39.6 |
| 35 COR EIDO | 4 | 3 | 0.6\% | 0.0 | 0.11 | 2 | 01 | 2.3 | 2 | 3 | 0.7 |
| 36 COR GLAU | 1,093 | 212 | 40.9\% | 2.1 | 4.40 | 35 | 01 | 3.2 | 1 | 8 | 602.2 |
| 37 COR PERS | 12,437 | 212 | 40.9\% | 24.0 | 55.70 | 600 | 01 | 2.6 | 1 | 6 | 3,345.4 |
| 38 COR SPE. | 50 | 1 | 0.2\% | 0.1 | 2.20 | 50 | 01 | 2.0 | 2 | 3 | 5.4 |
| 39 CRY ROSE | 4 | 3 | 0.6\% | 0.0 | 0.11 | 2 | 01 | 4.8 | 3 | 8 | 47.6 |
| 40 DAS AMER | 9 | 8 | 1.5\% | 0.0 | 0.14 | 2 | 01 | 153.3 | 80 | 200 | 210,638.3 |
| 41 DEC MACA | 20 | 1 | 0.2\% | 0.0 | 0.88 | 20 | 01 | 8.0 | 6 | 9 | 116.6 |
| 42 DIO HYST | 2 | 2 | 0.4\% | 0.0 | 0.06 | 1 | 01 | 30.5 | 26 | 35 | 2,628.4 |
| 43 DIP ARGE | 2 | 2 | 0.4\% | 0.0 | 0.06 | 1 | 01 | 13.0 | 9 | 17 | 116.6 |
| 44 ECH NAUC | 4 | 4 | 0.8\% | 0.0 | 0.09 | 1 | 01 | 10.0 | 9 | 12 | 66.9 |
| 45 EPI ADSC | 7 | 6 | 1.2\% | 0.0 | 0.13 | 2 | 01 | 26.3 | 15 | 35 | 2,449.0 |
| 46 EPI CRUE | 128 | 82 | 15.8\% | 0.2 | 1.14 | 23 | 01 | 19.6 | 5 | 35 | 19,255.9 |
| 47 EPI FULV | 1 | 1 | 0.2\% | 0.0 | 0.04 | 1 | 01 | 26.0 | 26 | 26 | 270.6 |
| 48 EPI GUTT | 15 | 14 | 2.7\% | 0.0 | 0.18 | 2 | 01 | 23.9 | 12 | 43 | 5,005.0 |
| 49 EPIITAJ | 3 | 3 | 0.6\% | 0.0 | 0.08 | 1 | 01 | 135.3 | 6 | 200 | 320,814.3 |
| 50 EPI MORI | 88 | 74 | 14.3\% | 0.2 | 0.45 | 3 | 01 | 40.2 | 7 | 75 | 102,876.8 |
| 51 EPI STRI | 3 | 3 | 0.6\% | 0.0 | 0.08 | 1 | 01 | 48.3 | 40 | 60 | 5,972.9 |
| 52 EQU ACUM | 6 | 3 | 0.6\% | 0.0 | 0.19 | 4 | 01 | 9.2 | 3 | 12 | 84.4 |
| 53 GIN CIRR | 1 | 1 | 0.2\% | 0.0 | 0.04 | 1 | 01 | 100.0 | 100 | 100 | 6,414.4 |
| 54 GNA THOM | 103 | 20 | 3.9\% | 0.2 | 1.35 | 17 | 01 | 3.1 | 2 | 5 | 31.1 |
| 55 GOB OCEA | 356 | 140 | 27.0\% | 0.7 | 1.58 | 16 | 01 | 2.6 | 1 | 5 | 70.9 |
| 56 GOB RAND | 1 | 1 | 0.2\% | 0.0 | 0.04 | 1 | 01 | 3.0 | 3 | 3 | 0.3 |
| 57 GOB SPE. | 4 | 1 | 0.2\% | 0.0 | 0.18 | 4 | 01 | 3.0 | 2 | 3 | 0.8 |
| 58 GYM MORI | 1 | 1 | 0.2\% | 0.0 | 0.04 | 1 | 01 | 40.0 | 40 | 40 | 114.9 |
| 59 HAE AURO | 5,648 | 68 | 13.1\% | 10.9 | 48.19 | 500 | 01 | 5.0 | 1 | 15 | 14,407.4 |
| 60 HAE CARB | 44 | 12 | 2.3\% | 0.1 | 0.80 | 12 | 01 | 19.8 | 4 | 30 | 8,849.9 |
| 61 HAE CHRY | 1 | 1 | 0.2\% | 0.0 | 0.04 | 1 | 01 | 16.0 | 16 | 16 | 157.0 |
| 62 HAE FLAV | 1,099 | 109 | 21.0\% | 2.1 | 11.09 | 150 | 01 | 10.7 | 3 | 25 | 41,152.6 |
| 63 HAE MACR | 28 | 10 | 1.9\% | 0.1 | 0.64 | 13 | 01 | 14.0 | 5 | 40 | 4,907.0 |
| 64 HAE MELA | 6 | 2 | 0.4\% | 0.0 | 0.20 | 4 | 01 | 6.0 | 5 | 6 | 25.0 |
| 65 HAE PARR | 9 | 3 | 0.6\% | 0.0 | 0.23 | 3 | 01 | 28.3 | 25 | 34 | 4,226.4 |
| 66 HAE PLUM | 1,487 | 305 | 58.9\% | 2.9 | 5.83 | 55 | 01 | 13.1 | 2 | 40 | 147,127.9 |
| 67 HAE SCIU | 109 | 35 | 6.8\% | 0.2 | 1.78 | 37 | 01 | 20.8 | 6 | 38 | 23,689.7 |

Table 3. (cont.)

| Species | SAMPLE FREQUENCY |  |  | Mean Abund. | Stand. <br> Dev. $=====$ | SAMP. FREQ. RANGE \| |  | FISH LENGTH (cm) |  |  | $\begin{gathered} \hline \text { BIOMASS } \\ \text { Total (gms) } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Indiv. |  | \% |  |  | High | Low \| | Mean | Min. | Max. |  |
| 68 HAE SPE. | 1,132 | 17 | 3.3\% | 2.2 | 20.35 | 400 | 01 | 2.5 | 1 | 5 | 376.1 |
| 69 HAE STRI | 4 | 1 | 0.2\% | 0.0 | 0.18 | 4 | 01 | 22.0 | 17 | 25 | 994.2 |
| 70 HAL BIVI | 3,345 | 298 | 57.5\% | 6.5 | 10.37 | 80 | 01 | 5.7 | 1 | 15 | 9,465.6 |
| 71 HAL CYAN | 6 | 4 | 0.8\% | 0.0 | 0.15 | 3 | 01 | 5.3 | 3 | 8 | 12.0 |
| 72 HAL GARN | 1,026 | 253 | 48.8\% | 2.0 | 3.22 | 24 | 01 | 6.7 | 1 | 15 | 5,100.7 |
| 73 HAL MACU | 469 | 132 | 25.5\% | 0.9 | 2.09 | 15 | 01 | 6.0 | 2 | 22 | 1,680.2 |
| 74 HAL POEY | 6 | 4 | 0.8\% | 0.0 | 0.14 | 2 | 01 | 6.3 | 4 | 9 | 22.1 |
| 75 HAL RADI | 39 | 34 | 6.6\% | 0.1 | 0.30 | 3 | 01 | 7.1 | 2 | 40 | 1,797.3 |
| 76 HEM BRAS | 30 | 1 | 0.2\% | 0.1 | 1.32 | 30 | 01 | 14.0 | 14 | 14 | 1,571.1 |
| 77 HEM MART | 2 | 2 | 0.4\% | 0.0 | 0.06 | 1 | 01 | 9.5 | 6 | 13 | 52.9 |
| 78 HEM SIMU | 29 | 4 | 0.8\% | 0.1 | 1.14 | 26 | 01 | 3.1 | 3 | 4 | 16.9 |
| 79 HOL ADSC | 113 | 27 | 5.2\% | 0.2 | 2.30 | 37 | 01 | 15.8 | 12 | 27 | 12,472.6 |
| 80 HOL BERM | 342 | 197 | 38.0\% | 0.7 | 1.09 | 6 | 01 | 21.3 | 3 | 42 | 110,445.9 |
| 81 HOL CILI | 70 | 49 | 9.5\% | 0.1 | 0.50 | 4 | 01 | 18.0 | 4 | 38 | 17,439.0 |
| 82 HOL RUFU | 79 | 31 | 6.0\% | 0.2 | 0.81 | 9 | 01 | 20.8 | 7 | 35 | 17,683.9 |
| 83 HOL TOWN | 1 | 1 | 0.2\% | 0.0 | 0.04 | 1 | 01 | 15.0 | 15 | 15 | 82.1 |
| 84 HOL TRIC | 57 | 39 | 7.5\% | 0.1 | 0.42 | 3 | 01 | 14.2 | 4 | 24 | 5,761.2 |
| 85 HOL VEXI | 4 | 3 | 0.6\% | 0.0 | 0.11 | 2 | 01 | 10.0 | 9 | 12 | 129.6 |
| 86 HYP BERM | 5 | 5 | 1.0\% | 0.0 | 0.10 | 1 | 01 | 3.0 | 2 | 4 | 2.0 |
| 87 HYP GEMM | 326 | 159 | 30.7\% | 0.6 | 1.51 | 21 | 01 | 6.4 | 3 | 13 | 1,619.4 |
| 88 HYP GUTT | 3 | 1 | 0.2\% | 0.0 | 0.13 | 3 | 01 | 7.0 | 5 | 10 | 23.9 |
| 89 HYP HYBR | 1 | 1 | 0.2\% | 0.0 | 0.04 | 1 | 01 | 8.0 | 8 | 8 | 8.2 |
| 90 HYP INDI | 4 | 4 | 0.8\% | 0.0 | 0.09 | 1 | 01 | 7.3 | 6 | 9 | 25.9 |
| 91 HYP NIGR | 42 | 36 | 6.9\% | 0.1 | 0.32 | 3 | 01 | 6.1 | 3 | 11 | 197.9 |
| 92 HYP PUEL | 156 | 119 | 23.0\% | 0.3 | 0.62 | 3 | 01 | 6.2 | 3 | 11 | 724.8 |
| 93 HYP TANN | 29 | 25 | 4.8\% | 0.1 | 0.27 | 3 | 01 | 5.8 | 3 | 11 | 121.1 |
| 94 HYP UNIC | 466 | 253 | 48.8\% | 0.9 | 1.23 | 10 | 01 | 6.1 | 2 | 12 | 2,090.4 |
| 95 INE VITT | 514 | 5 | 1.0\% | 1.0 | 21.97 | 500 | 01 | 9.2 | 8 | 20 | 5,013.1 |
| 96 IOG CALL | 22 | 9 | 1.7\% | 0.0 | 0.37 | 6 | 01 | 7.8 | 2 | 10 | 113.6 |
| 97 IOG HELE | 5 | 2 | 0.4\% | 0.0 | 0.18 | 4 | 01 | 5.2 | 4 | 10 | 11.3 |
| 98 KYP SECT | 274 | 30 | 5.8\% | 0.5 | 3.31 | 40 | 01 | 28.3 | 10 | 70 | 201,319.3 |
| 99 LAC BICA | 3 | 3 | 0.6\% | 0.0 | 0.08 | 1 | 01 | 16.0 | 7 | 21 | 532.9 |
| 100 LAC MAXI | 91 | 60 | 11.6\% | 0.2 | 0.61 | 8 | 01 | 34.2 | 2 | 60 | 94,072.2 |
| 101 LAC TRIQ | 11 | 9 | 1.7\% | 0.0 | 0.17 | 2 | 01 | 16.7 | 11 | 22 | 1,934.8 |
| 102 LUT ANAL | 75 | 44 | 8.5\% | 0.1 | 0.65 | 10 | 01 | 49.8 | 30 | 75 | 176,636.3 |
| 103 LUT APOD | 22 | 10 | 1.9\% | 0.0 | 0.37 | 6 | 01 | 22.6 | 12 | 35 | 6,634.0 |
| 104 LUT GRIS | 657 | 121 | 23.4\% | 1.3 | 3.76 | 31 | 01 | 26.6 | 8 | 50 | 268,226.0 |
| 105 LUT JOCU | 2 | 2 | 0.4\% | 0.0 | 0.06 | 1 | 01 | 42.5 | 35 | 50 | 3,002.3 |
| 106 LUT MAHO | 2 | 2 | 0.4\% | 0.0 | 0.06 | 1 | 01 | 25.5 | 24 | 27 | 576.5 |
| 107 LUT SYNA | 21 | 3 | 0.6\% | 0.0 | 0.68 | 15 | 01 | 26.7 | 12 | 30 | 6,757.6 |
| 108 MAL MACR | 157 | 57 | 11.0\% | 0.3 | 1.13 | 9 | 01 | 3.6 | 2 | 6 | 68.5 |
| 109 MAL PLUM | 4 | 2 | 0.4\% | 0.0 | 0.12 | 2 | 01 | 9.5 | 6 | 15 | 54.1 |
| 110 MAL TRIA | 321 | 135 | 26.1\% | 0.6 | 1.94 | 33 | 01 | 3.9 | 1 | 8 | 202.4 |
| 111 MAL VERS | 5 | 2 | 0.4\% | 0.0 | 0.16 | 3 | 01 | 5.0 | 4 | 5 | 7.0 |
| 112 MEG ATLA | 3 | 3 | 0.6\% | 0.0 | 0.08 | 1 | 01 | 161.7 | 135 | 200 | 148,902.3 |
| 113 MIC CHRY | 64 | 41 | 7.9\% | 0.1 | 0.49 | 4 | 01 | 9.8 | 4 | 14 | 2,104.1 |
| 114 MON TUCK | 4 | 4 | 0.8\% | 0.0 | 0.09 | 1 | 01 | 5.5 | 4 | 7 | 18.8 |
| 115 MUL MART | 151 | 29 | 5.6\% | 0.3 | 2.25 | 40 | 01 | 19.7 | 3 | 42 | 38,792.5 |
| 116 MYC BONA | 43 | 37 | 7.1\% | 0.1 | 0.32 | 2 | 01 | 48.0 | 12 | 100 | 112,387.0 |
| 117 MYC MICR | 1 | 1 | 0.2\% | 0.0 | 0.04 | 1 | 01 | 30.0 | 30 | 30 | 389.0 |
| 118 MYC PHEN | 33 | 20 | 3.9\% | 0.1 | 0.39 | 6 | 01 | 28.3 | 4 | 50 | 15,817.4 |
| 119 OCY CHRY | 3,213 | 317 | 61.2\% | 6.2 | 18.02 | 250 | 01 | 16.9 | 2 | 50 | 369,271.1 |
| 120 ODO DENT | 6 | 4 | 0.8\% | 0.0 | 0.14 | 2 | 01 | 15.0 | 11 | 18 | 235.5 |
| 121 OPH ATLA | 16 | 9 | 1.7\% | 0.0 | 0.28 | 5 | 01 | 3.4 | 3 | 4 | 9.7 |
| 122 OPI AURI | 593 | 109 | 21.0\% | 1.1 | 3.28 | 31 | 01 | 6.5 | 2 | 10 | 1,687.6 |
| 123 PAR FURC | 25 | 1 | 0.2\% | 0.0 | 1.10 | 25 | 01 | 15.0 | 10 | 20 | 1,314.5 |
| 124 PAR MARM | 52 | 20 | 3.9\% | 0.1 | 0.60 | 6 | 01 | 3.3 | 1 | 7 | 50.4 |
| 125 POM ARCU | 214 | 145 | 28.0\% | 0.4 | 0.78 | 6 | 01 | 23.3 | 8 | 45 | 108,845.4 |
| 126 POM DIEN | 16 | 6 | 1.2\% | 0.0 | 0.35 | 6 | 01 | 4.7 | 2 | 8 | 63.2 |
| 127 POM FUSC | 805 | 108 | 20.8\% | 1.6 | 4.71 | 43 | 01 | 5.1 | 2 | 8 | 3,487.1 |
| 128 POM LEUC | 807 | 190 | 36.7\% | 1.6 | 3.23 | 35 | 01 | 4.0 | 1 | 8 | 1,670.1 |
| 129 POM PART | 3,358 | 237 | 45.8\% | 6.5 | 11.14 | 62 | 01 | 4.5 | 1 | 10 | 8,356.2 |
| 130 POM PARU | 49 | 37 | 7.1\% | 0.1 | 0.39 | 4 | 01 | 26.7 | 4 | 50 | 38,894.6 |
| 131 POM PLAN | 1,631 | 234 | 45.2\% | 3.1 | 5.78 | 54 | 01 | 5.6 | 2 | 10 | 9,418.7 |
| 132 POM VARI | 1,992 | 382 | 73.7\% | 3.8 | 4.70 | 40 | 01 | 4.8 | 1 | 10 | 6,517.1 |
| 133 PRI AREN | 11 | 2 | 0.4\% | 0.0 | 0.40 | 9 | 01 | 30.2 | 21 | 35 | 4,779.9 |
| 134 PSE MACU | 151 | 78 | 15.1\% | 0.3 | 1.16 | 15 | 01 | 9.7 | 3 | 25 | 4,748.3 |
| 135 SCA COEL | 19 | 9 | 1.7\% | 0.0 | 0.34 | 5 | 01 | 28.8 | 11 | 45 | 11,251.2 |

Table 3. (cont.)


Table 4. Summary of Dry Tortugas predator count effort (1994-1997).


Table 5. Species listing for all 1994-1997 predator searches.

| SPCODE | NUM SPECIES | COMMON NAME | FAMILY | FAMILY NAME |
| :---: | :---: | :---: | :---: | :---: |
| 1 ALE CILI | 70 Alectis ciliaris | African pompano | CARANGIDAE | Jacks |
| 2 AUL MACU | 180 Aulostomus maculatus | Trumpetfish | AULOSTOMIDAE | Trumpetfishes |
| 3 CAR BART | 320 Caranx bartholomaei | Yellow jack | CARANGIDAE | Jacks |
| 4 CAR CRYS | 330 Caranx crysos | Blue runner | CARANGIDAE | Jacks |
| 5 CAR HIPP | 340 Caranx hippos | Crevalle jack | CARANGIDAE | Jacks |
| 6 CAR LATU | 345 Caranx latus | Horse-eye jack | CARANGIDAE | Jacks |
| 7 CAR RUBE | 350 Caranx ruber | Bar jack | CARANGIDAE | Jacks |
| 8 DAS AMER | 530 Dasyatis americana | Southern stingray | DASYATIDAE | Stingrays |
| 9 EPI ADSC | 650 Epinephelus adscensionis | Rock hind | SERRANIDAE | Sea basses |
| 10 EPI CRUE | 660 Epinephelus cruentatus | Graysby | SERRANIDAE | Sea basses |
| 11 EPI FULV | 670 Epinephelus fulvus | Coney | SERRANIDAE | Sea basses |
| 12 EPI GUTT | 680 Epinephelus guttatus | Red hind | SERRANIDAE | Sea basses |
| 13 EPI ITAJ | 685 Epinephelus itajara | Jewfish | SERRANIDAE | Sea basses |
| 14 EPI MORI | 690 Epinephelus morio | Red grouper | SERRANIDAE | Sea basses |
| 15 EPI STRI | 710 Epinephelus striatus | Nassau grouper | SERRANIDAE | Sea basses |
| 16 GIN CIRR | 760 Ginglymostoma cirratum | Nurse shark | ORECTOLOBIDAE | Carpet sharks |
| 17 GYM MORI | 830 Gymnothorax moringa | Spotted moray | MURAENIDAE | Morays |
| 18 GYM SAXI | 840 Gymnothorax saxicola | Ocellated moray | MURAENIDAE | Morays |
| 19 HYP GEMM | 1160 Hypoplectrus gemma \# | Blue hamlet | SERRANIDAE | Sea basses |
| 20 HYP NIGR | 1170 Hyploplectrus nigricans \# | Black hamlet | SERRANIDAE | Sea basses |
| 21 HYP PUEL | 1190 Hypoplectrus puella \# | Barred hamlet | SERRANIDAE | Sea basses |
| 22 HYP TANN | 1180 Hypoplectrus (tan) | Tan hamlet | SERRANIDAE | Sea basses |
| 23 HYP UNIC | 1200 Hypoplectrus unicolor | Butter hamlet | SERRANIDAE | Sea basses |
| 24 LAC MAXI | 1260 Lachnolaimus maximus | Hogfish | LABRIDAE | Wrasses |
| 25 LUT ANAL | 1320 Lutjanus analis | Mutton snapper | LUTJANIDAE | Snappers |
| 26 LUT APOD | 1330 Lutjanus apodus | Schoolmaster | LUTJANIDAE | Snappers |
| 27 LUT GRIS | 1360 Lutjanus griseus | Gray snapper | LUTJANIDAE | Snappers |
| 28 LUT JOCU | 1370 Lutjanus jocu | Dog snapper | LUTJANIDAE | Snappers |
| 29 LUT MAHO | 1380 Lutjanus mahogoni | Mahogany snapper | LUTJANIDAE | Snappers |
| 30 LUT SYNA | 1385 Lutjanus synagris | Lane snapper | LUTJANIDAE | Snappers |
| 31 MEG ATLA | 1465 Megalops atlanticus | Tarpon | ELOPIDAE | Tarpons |
| 32 MYC BONA | 1545 Mycteroperca bonaci | Black grouper | SERRANIDAE | Sea basses |
| 33 MYC MICR | 1560 Mycteroperca microlepis | Gag | SERRANIDAE | Sea basses |
| 34 MYC PHEN | 1570 Mycteroperca phenax | Scamp | SERRANIDAE | Sea basses |
| 35 OCY CHRY | 1610 Ocyurus chrysurus | Yellowtail snapper | LUTJANIDAE | Snappers |
| 36 SCO REGA | 1970 Scomberomorus regalis | Cero mackerel | SCOMBRIDAE | Mackerels/Tunas |
| 37 SPH BARR | 2097 Sphyraena barracuda | Barracuda | SPHYRAENIDAE | Barracudas |
| 38 SYN INTE | 2190 Synodus intermedius | Sand diver | SYNODONTIDAE | Lizardfishes |
| 39 TRA FALC | 2203 Trachinotus falcatus | Permit | CARANGIDAE | Jacks |

Table 6a. Statistical summary of Dry Tortugas predator searches (1994-1997).

| Species |  | SAMPLE FREQUENCY |  |  | SAMP. FREQ. RANG \| |  |  |  | FISH LENGTH (cm) |  |  | BIOMASS <br> Total (gms) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Total Indiv. | $(\mathrm{N}=126)$ | ------ | Mean Abund. | Stand. <br> Dev. | High | Low | Mean | Min. | Max. |  |
|  | 1 ALE CILI | 2 | 1 | 0.8\% | 0.0 | 0.18 | 2 | 0 \| | 100.0 | 100 | 100 | 33,981.0 |
|  | 2 AUL MACU | 20 | 14 | 11.1\% | 0.2 | 0.53 | 4 | 01 | 41.1 | 20 | 60 | 3,976.8 |
|  | 3 CAR BART | 61 | 5 | 4.0\% | 0.5 | 3.53 | 30 | 01 | 25.7 | 15 | 70 | 32,081.2 |
|  | 4 CAR CRYS | 141 | 7 | 5.6\% | 1.1 | 9.20 | 101 | 01 | 20.9 | 15 | 40 | 30,125.1 |
|  | 5 CAR HIPP | 20 | 1 | 0.8\% | 0.2 | 1.78 | 20 | 01 | 34.0 | 32 | 36 | 15,977.8 |
|  | 6 CAR LATU | 10 | 3 | 2.4\% | 0.1 | 0.57 | 5 | 0 | 53.5 | 35 | 100 | 41,157.9 |
|  | 7 CAR RUBE | 1,211 | 50 | 39.7\% | 9.6 | 25.58 | 165 | 01 | 20.5 | 3 | 40 | 205,862.5 |
|  | 8 DAS AMER | 5 | 5 | 4.0\% | 0.0 | 0.20 | 1 | 01 | 92.0 | 60 | 110 | 19,029.6 |
|  | 9 EPI ADSC | 1 | 1 | 0.8\% | 0.0 | 0.09 | 1 | 01 | 28.0 | 28 | 28 | 353.6 |
|  | 0 EPI CRUE | 49 | 27 | 21.4\% | 0.4 | 0.87 | 4 | 01 | 20.0 | 7 | 31 | 8,248.5 |
|  | 1 EPI FULV | 3 | 2 | 1.6\% | 0.0 | 0.20 | 2 | 01 | 21.7 | 18 | 25 | 498.9 |
|  | 2 EPI GUTT | 1 | 1 | 0.8\% | 0.0 | 0.09 | 1 | 0 \| | 25.0 | 25 | 25 | 248.5 |
|  | 3 EPI ITAJ | 4 | 3 | 2.4\% | 0.0 | 0.22 | 2 | 01 | 185.0 | 170 | 200 | 525,085.8 |
|  | 4 EPI MORI | 92 | 49 | 38.9\% | 0.7 | 1.34 | 8 | 01 | 35.6 | 9 | 75 | 79,500.6 |
|  | 5 EPI STRI | 1 | 1 | 0.8\% | 0.0 | 0.09 | 1 | 01 | 32.0 | 32 | 32 | 471.2 |
|  | 6 GIN CIRR | 9 | 7 | 5.6\% | 0.1 | 0.31 | 2 | 01 | 179.6 | 125 | 230 | 334,261.1 |
|  | 7 GYM MORI | 3 | 3 | 2.4\% | 0.0 | 0.15 | 1 | 01 | 46.0 | 30 | 60 | 664.3 |
|  | 8 GYM SAXI | 1 | 1 | 0.8\% | 0.0 | 0.09 | 1 | 01 | 50.0 | 50 | 50 | 232.5 |
|  | 9 HYP GEMM | 25 | 14 | 11.1\% | 0.2 | 0.69 | 5 | 01 | 6.4 | 4 | 9 | 112.0 |
|  | 0 HYP NIGR | 4 | 3 | 2.4\% | 0.0 | 0.22 | 2 | 01 | 6.8 | 6 | 7 | 17.3 |
|  | 1 HYP PUEL | 11 | 9 | 7.1\% | 0.1 | 0.34 | 2 | 01 | 6.3 | 5 | 9 | 46.6 |
|  | 2 HYP TANN | 1 | 1 | 0.8\% | 0.0 | 0.09 | 1 | 0 | 5.0 | 5 | 5 | 1.8 |
|  | 3 HYP UNIC | 29 | 14 | 11.1\% | 0.2 | 0.74 | 4 | 01 | 6.3 | 3 | 14 | 184.9 |
|  | LAC MAXI | 179 | 69 | 54.8\% | 1.4 | 2.03 | 11 | 01 | 32.5 | 10 | 75 | 184,452.3 |
|  | 5 LUT ANAL | 24 | 16 | 12.7\% | 0.2 | 0.58 | 3 | 01 | 50.8 | 22 | 70 | 65,649.6 |
|  | 6 LUT APOD | 357 | 30 | 23.8\% | 2.8 | 8.30 | 53 | 0 | 23.5 | 10 | 55 | 103,713.9 |
|  | 7 LUT GRIS | 2,952 | 80 | 63.5\% | 23.4 | 37.05 | 170 | 01 | 25.5 | 10 | 55 | 906,632.0 |
|  | 8 LUT JOCU | 6 | 4 | 3.2\% | 0.0 | 0.28 | 2 | 01 | 54.3 | 22 | 65 | 19,475.5 |
|  | 9 LUT MAHO | 161 | 12 | 9.5\% | 1.3 | 7.02 | 53 | 01 | 20.4 | 10 | 40 | 34,088.9 |
|  | 0 LUT SYNA | 21 | 2 | 1.6\% | 0.2 | 1.32 | 11 | 01 | 32.4 | 20 | 35 | 11,140.5 |
|  | 1 MEG ATLA | 7 | 7 | 5.6\% | 0.1 | 0.23 | 1 | 0 | 155.0 | 70 | 250 | 371,842.5 |
|  | MYC BONA | 113 | 55 | 43.7\% | 0.9 | 1.45 | 9 | 0 | 46.1 | 15 | 150 | 406,213.8 |
|  | 3 MYC MICR | 5 | 5 | 4.0\% | 0.0 | 0.20 | 1 | 0 | 33.6 | 22 | 60 | 4,463.0 |
|  | 4 MYC PHEN | 37 | 20 | 15.9\% | 0.3 | 0.78 | 4 | 01 | 27.2 | 10 | 50 | 15,304.7 |
|  | 5 OCY CHRY | 6,114 | 115 | 91.3\% | 48.5 | 84.22 | 524 | 0 | 19.7 | 4 | 50 | 993,487.7 |
|  | SCO REGA | 12 | 9 | 7.1\% | 0.1 | 0.43 | 4 | 0 | 44.3 | 20 | 96 | 14,528.1 |
|  | 7 SPH BARR | 100 | 51 | 40.5\% | 0.8 | 2.03 | 20 | 01 | 69.2 | 17 | 150 | 495,262.5 |
|  | SYN INTE | 1 | 1 | 0.8\% | 0.0 | 0.09 | 1 | 0 | 18.0 | 18 | 18 | 57.7 |
|  | TRA FALC | 1 | 1 | 0.8\% | 0.0 | 0.09 | 1 | 01 | 60.0 | 60 | 60 | 3,845.5 |
|  | NO. SAMPLES | = | 126 |  | SAMPLE A |  | - | PRED | RCES |  |  |  |
|  | NO. SPECIES | = | 39 |  | NO. OBSE | ERS | - | 4 |  |  |  |  |
|  | TOT.INDIVIDU TOT. BIOMASS | S $=$ | $\begin{array}{r} 11,794 \\ 4,962,278.1 \end{array}$ |  | SAMPLING | ONDITI | - | 15-25 |  |  |  |  |

Table 6b. Statistical summary of Dry Tortugas predator searches (1998).


Table 7. Summary comparison of fishes inside and outside DTNP based on stationary sample data.

|  | Sanctuary | Park | Sanctuary | Park |
| :---: | :---: | :---: | :---: | :---: |
| DATE | 1994-1997 | 1994-1997 | 1994-1997 | 1994-1997 |
| Total samples | 183 | 335 |  |  |
| Total species | 135 | 140 |  |  |
| Mean species/sample | 20.10 | 17.68 |  |  |
| Total individuals | 33,302 | 43,106 |  |  |
| Mean individuals/sample | 181.98 | 128.67 |  |  |
| Total biomass (g) | 2,109,001.2 | 2,328,141.9 |  |  |
| Mean biomass (g)/sample | 11524.60 | 6949.68 |  |  |
| INDIVIDUALS | Mean | Mean | \% | \% |
| Barracuda | 0.14 | 0.10 | 0.08\% | 0.08\% |
| Damselfishes | 57.26 | 18.06 | 31.47\% | 14.04\% |
| Grunts and Porgies | 8.03 | 25.04 | 4.41\% | 19.46\% |
| Other | 54.06 | 25.12 | 29.71\% | 25.14\% |
| Parrotfishes | 15.31 | 23.87 | 8.41\% | 18.55\% |
| Serranids | 3.55 | 2.63 | 1.95\% | 2.04\% |
| Lutjanids | 9.78 | 6.58 | 5.37\% | 5.11\% |
| Surgeonfishes | 2.67 | 3.23 | 1.47\% | 2.51\% |
| Wrasses | 31.17 | 16.82 | 17.13\% | 13.07\% |
| BIOMASS | Mean (g) | Mean (g) | \% | \% |
| Barracuda | 705.49 | 1095.30 | 6.12\% | 15.76\% |
| Damselfishes | 360.28 | 140.28 | 3.13\% | 2.02\% |
| Grunts and Porgies | 519.00 | 607.80 | 4.50\% | 8.75\% |
| Other | 4113.35 | 2381.36 | 35.69\% | 34.27\% |
| Parrotfishes | 525.14 | 616.16 | 4.56\% | 8.87\% |
| Serranids | 2261.57 | 529.34 | 19.62\% | 7.62\% |
| Lutjanids | 2511.31 | 1109.06 | 21.79\% | 15.96\% |
| Surgeonfishes | 177.99 | 254.99 | 1.54\% | 3.67\% |
| Wrasses | 350.46 | 215.39 | 3.04\% | 3.10\% |
| TOTAL NUMBER OF SPECIES | OUTSIDE | INSIDE |  |  |
| Barracuda | 1 | 1 | 1 |  |
| Damselfishes | 12 | 13 | 13 |  |
| Grunts and Porgies | 11 | 13 | 14 |  |
| Other | 62 | 63 | 80 |  |
| Parrotfishes | 13 | 12 | 13 |  |
| Serranids | 19 | 19 | 22 |  |
| Snappers | 6 | 7 | 7 |  |
| Surgeonfishes | 3 | 3 | 3 |  |
| Wrasses | 8 | 9 | 9 |  |
| TOTAL | 135 | 140 | 162 |  |

Table 8. Reef characteristics and summaries for DTNP and FKNMS study sites.

|  | Mean Total Individuals | Mean <br> Total Species | Mean <br> Total Biomass | Average Depth | Average Percent Coral | Mean <br> Total <br> Snapper | Mean <br> Total Grouper |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GG (12) | 108.83 | 18.00 | 2217.96 | 33.50 | 51.66\% | 2.90 | 4.72 |  |
| LA (51) | 103.45 | 17.06 | 7642.28 | 13.04 | 36.75\% | 3.88 | 3.22 | 0.14 |
| LB (74) | 128.14 | 17.73 | 7924.97 | 30.88 | 43.28\% | 6.23 | 1.49 | 0.04 |
| LG (43) | 122.81 | 17.00 | 4898.15 | 25.91 | 21.44\% | 5.95 | 1.65 | 0.30 |
| PS (61) | 146.93 | 20.75 | 11228.31 | 38.77 | 38.62\% | 15.84 | 3.11 | 0.30 |
| TX (37) | 151.03 | 18.54 | 4826.30 | 33.81 | 72.59\% | 9.00 | 0.57 | 0.05 |
| WS (57) | 126.49 | 14.74 | 4407.05 | 15.26 | 18.68\% | 6.92 | 4.50 | 0.14 |
| Inside (335) | 128.67 | 17.68 | 6949.68 | 26.72 | 37.99\% | 6.58 | 2.63 | 0.15 |
| Outside (183) | 181.98 | 20.10 | 11524.60 | 56.74 | 61.26\% | 9.78 | 3.55 | 0.23 |
| BL (8) | 250.38 | 18.88 | 50464.13 | 66.25 | 86.88\% | 17.75 | 5.25 | 0.13 |
| CC (3) | 190.67 | 18.33 | 2024.87 | 48.67 | 83.33\% | 3.00 | 2.33 |  |
| CZ (19) | 78.53 | 15.63 | 1767.86 | 55.84 | 44.74\% | 0.47 | 1.42 | 0.11 |
| DV (8) | 126.00 | 18.75 | 10750.875 | 46.50 | 43.38\% | 6.00 | 3.75 |  |
| EF (8) | 222.88 | 20.88 | 23144.68 | 61.88 | 82.50\% | 21.13 | 2.75 |  |
| GA (6) | 248.17 | 19.33 | 14910.70 | 56.50 | 84.17\% | 1.67 | 1.33 | 0.50 |
| H1 (5) | 223.20 | 19.80 | 5671.22 | 51.40 | 88.00\% | 11.17 | 3.33 |  |
| H2 (3) | 519.33 | 17.33 | 3352.80 | 56.67 | 68.33\% | 1.60 | 2.20 |  |
| H3 (6) | 409.83 | 22.50 | 20165.75 | 52.17 | 95.83\% | 1.67 | 1.33 |  |
| HO (12) | 245.75 | 25.75 | 7246.33 | 47.67 | 59.08\% | 47.83 | 2.83 |  |
| JH (32) | 106.50 | 19.63 | 8442.73 | 62.88 | 51.81\% | 3.09 | 3.88 | 0.47 |
| LI (8) | 150.75 | 23.00 | 22529.20 | 58.88 | 67.50\% | 35.50 | 5.75 | 1.38 |
| MH (4) | 159.50 | 14.25 | 925.33 | 64.75 | 42.50\% | 0.00 | 3.75 |  |
| MV (9) | 207.11 | 21.44 | 6706.88 | 27.00 | 23.89\% | 11.00 | 2.89 | 0.44 |
| RZ (8) | 121.50 | 21.25 | 6383.81 | 65.00 | 60.00\% | 1.25 | 3.25 | 0.25 |
| SF (8) | 362.63 | 18.63 | 10294.05 | 80.63 | 92.50\% | 2.00 | 4.75 |  |
| TB (21) | 157.38 | 21.62 | 9320.30 | 55.05 | 75.67\% | 25.00 | 0.46 | 0.05 |
| TF (5) | 97.40 | 16.40 | 3315.00 | 65.20 | 18.00\% | 0.00 | 1.60 | 0.20 |
| TP (10) | 209.20 | 23.00 | 19844.24 | 44.00 | 69.30\% | 10.00 | 5.14 |  |

Table 9. Summary comparison of fishes inside and outside DTNP based on predator search data.

|  | Sanctuary | Park | Sanctuary | Park |
| :---: | :---: | :---: | :---: | :---: |
| DATE | 1994-1997 | 1994-1997 | 1994-1997 | 1994-1997 |
| Total samples | 21 | 105 |  |  |
| Total species | 22 | 37 |  |  |
| Mean species/sample | 5.67 | 5.52 |  |  |
| Total individuals | 3,166 | 8,628 |  |  |
| Mean individuals/sample | 150.76 | 82.17 |  |  |
| Total biomass (g) | 1000386.08 | 3961892.04 |  |  |
| Mean biomass (g)/sample | 47637.43 | 37732.31 |  |  |
| INDIVIDUALS | Mean | Mean | \% | \% |
| Barracuda | 0.8 | 0.8 | 0.51\% | 0.97\% |
| Serranids | 3.6 | 2.8 | 2.46\% | 3.45\% |
| Other | 37.8 | 6.8 | 25.05\% | 8.24\% |
| Lutjanids | 106.2 | 70.5 | 70.47\% | 85.81\% |
| Hogfish | 2.3 | 1.2 | 1.52\% | 1.52\% |
| BIOMASS | Mean (g) | Mean (g) | \% | \% |
| Barracuda | 4274.31 | 3861.92 | 8.97\% | 10.24\% |
| Serranids | 6443.99 | 8623.12 | 13.53\% | 22.85\% |
| Other | 13724.90 | 7803.82 | 28.81\% | 20.68\% |
| Lutjanids | 20585.23 | 16208.56 | 43.21\% | 42.96\% |
| Hogfish | 2609.00 | 1234.89 | 5.48\% | 3.27\% |
| TOTAL NUMBER OF SPECIES | OUTSIDE | INSIDE |  |  |
| Barracuda | 1 | 1 |  |  |
| Serranids | 8 | 14 |  |  |
| Other | 8 | 14 |  |  |
| Lutjanids | 4 | 7 |  |  |
| Hogfish | 1 | 1 |  |  |
| TOTAL | 22 | 37 |  |  |

Figure 1. Summary or reef sites sampled.


Figure 2. Cluster analysis of fish assemblage similarity from Dry Tortugas Reefs.


Figure 3. Family comparisons of fishes observed in stationary point samples inside and outside DTNP.


Figure 4. Average depth and percentage or coral and rock substrate at reef sites inside and outside DTNP.


Figure 5. Scattergrams showing performance of mean total individuals, mean total species, and mean total biomass from individual reef sites.


Figure 6. Scattergrams showing performance of mean total snapper (lutjanids), grouper (serranids), and hogfish from individual reef sites.


Figure 7. Scattergrams showing performance of mean total abundances of hogfish, mutton snapper, black grouper and yellowtail from individual reef sites.


Figure 8. Family comparisons of fishes observed in predator samples inside and outside DTNP.


# Chapter 5. Reef fish abundance and species composition from the Florida Middle Grounds: R/V Suncoaster cruise August 18-23, 2000. 

David B. McClellan and Michael T. Judge

## Introduction

A limited reef fish survey of the proposed Habitat Area of Particular Concern (HAPC) in the Florida Middle Ground (FMG) was conducted by two members of the Rapid Fish Assessment Team, Reef Resources Team, Protected Resources and Biodiversity Division, SEFSC, NOAA Fisheries, from August 18-23, 2000. The primary purpose of this R/V Suncoaster cruise, PI Dr. David Mallison, was to map the bottom of the area utilizing side-scan sonar and multibeam bathymetry survey methods. At sites selected by bottom features on the sonar, dives were conducted to collect sediment samples and videotape of the habitat, as well as survey the fish fauna (Figure 1).

## Results and Discussion

Data were collected using a stationary point sampling technique (Bohnsack and Bannerot 1986) which utilizes standard visual sampling methods. Twelve stationary samples were collected from six sites (two samples per site) at a mean depth of 92 feet (range 79 to 100 feet). Only one count per diver could be finished per dive because of depth and bottom time constraints. Data collected provide fish species presence, abundance, frequency, biomass, average size, and size range. This information is important for accessing fish composition of the Florida Middle Ground and for management of the proposed HAPC.

Summary information of the August 18-23, 2000 FMG cruise data are provided in Tables 1 and 2. A total of 4,340 fishes representing 54 species ( 19 families) were recorded from the 12 samples. The mean number of fish observed per sample was 362 (range 87 to 1167) and the mean number of species per sample was 21 (range 14 to 25 ). Five species were observed from all samples; the purple reef fish Chromis scottii ( $\mathrm{n}=2,988$ ), striped parrotfish Scarus croiscensis ( $\mathrm{n}=196$ ), scamp Mycteroperca phenax ( $\mathrm{n}=145$ ), three unknown porgies Calamus $s p . \mathrm{n}=96$ ), and blue angelfish Holocanthus bermudensis $(\mathrm{n}=59)$.

The purple reef fish was the most prevalent species seen from all the sites ( $\mathrm{n}=2,988$ ), followed by the striped parrotfish ( $\mathrm{n}=196$ ), scamp ( $\mathrm{n}=145$ ), clown wrasse Halichoeres maculipinna ( $\mathrm{n}=135$ ), and cocoa damselfish Pomacentrus variabilis ( $\mathrm{n}=131$ ). The most common predator species seen was the scamp grouper ( $\mathrm{n}=145$ ) which also had the highest biomass ( 47.5 kg total, average weight 1.1 kg ). The mangrove snapper Lutjanus griseus ( $\mathrm{n}=68,45.2 \mathrm{~kg}$ total, average weight 0.67 kg ) had the next highest biomass, followed by the red grouper Epinephelus morio ( $\mathrm{n}=25,26.8 \mathrm{~kg}$, average weight 1.1 kg ), and combined porgy species Calamus $s p$. ( $\mathrm{n}=96,27.5 \mathrm{~kg}$ total, 0.29 kg ). Only three gag grouper Mycteroperca microlepis and two Gulf red snapper Lutjanus campechanus were observed during the counts. An additional 10 species were recorded after five minutes; the two-spot cardinalfish Apogon pseudomaculatus, trumpetfish Aulostomus maculatus, smooth puffer Canthigaster rostrata, bar jack Caranx ruber, sunshine fish Chromis insolatus, goldspot goby Gnatholepis thompsoni, spotted moray Gymnothorax moringa, red porgy Pagrus pagrus, spotted goatfish Pseudopeneus maculatus, and almaco jack Seriola rivolina.

Habitat information was collected from each site using the criteria developed by the Rapid Assessment Team (Smith et al. 2000). Five of the six sites were prominent outcrops dominated by the blade fire coral Millepora $s p$. and unidentified gorgonian species. The FMG region is characterized by steep-profile limestone escarpments and knolls rising 10-15 meters above the surrounding sand, sand-shell substrate (Smith et al. 1975), with a zone of minor reef building on
some banks in the Millepora zone (18-52 meters) (Rezak et al. 1990).
The reef fish census conducted during this cruise, although limited in scope, demonstrates the complexity and abundance of the reef fish fauna associated with the Florida Middle Ground. Smith et al. (1975) recorded 128 fish species representing 49 families from all habitats in the FMG using numerous census methods. The 54 species from 19 families observed during our censuses represent species associated with a single type reef habitat. Species such as the purple reef fish, the porgies, the red grouper, and the scamp were considered abundant in the 1970's (Smith et al. 1975), as they are today. The red snapper and gag grouper, considered common in the 1970's (Smith et al. 1975), were rare during our study. An average of 12.1 scamp per count (range four to 30) were recorded, representing a much higher number than seen elsewhere in waters of the Florida Keys and Dry Tortugas (Bohnsack et al. 2000). A 16 inch ( 41 centimeter) total length commercial and recreational size limit presently occurs for the scamp in Federal waters of the Gulf of Mexico, and most of the scamp observed during this study fall below this size. Additional reef fish censusing is needed to establish a more precise baseline, as well as supply data for future assessments of the reef fish complex in the Florida Middle Ground Habitat Area of Particular Concern.

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Table 1. Analysis of reef fish visual sampling during the R/V Suncoaster cruise to the Florida Middle Ground Habitat Area of Particular Concern, August 18-23, 2000.


Table 2. Analysis of reef fish visual sampling during the R/V Suncoaster cruise to the Florida Middle Ground Habitat Area of Particular Concern, August 18-23, 2000.

| SPCODE | NTM SPECIES | COMHON NaHE | Family | FAMILY NAME |
| :---: | :---: | :---: | :---: | :---: |
| 1 ACA BAHI | 30 Acanthurus bahianus | Ocean surgeon | ACANTHURIDAE | Surgeonfishes |
| 2 ACA CHIR | 50 Acanthurus chirurgus | Doctorfish | ACANTHURIDAE | Surgeonfishes |
| 3 ACA COER | 60 Acanthurus coeruleus | Blue tang | aCanthuridae | Surgeonfishes |
| 4 ADI VEXI | 63 Adiorys vexillarius | Dusky squirrelfish | HOLOCENTRIDAE | Squirrelfishes |
| 5 APO MACJ | 130 Apogon maculatus | Flamefish | apoconidaE | Cardinalfishes |
| 6 BAL CAPR | 190 Balistes capriscus | Gray triggerfish | BaLISTIDAE | Leatherjackets |
| 7 CAL BAJO | 240 Calamus bajonado | Jolthead porgy | SParidaE | Porgies |
| 8 CAL CALA | 260 Calamus calamus | Saucereye porgy | SParidaE | Porgies |
| 9 CAL SPE . | 276 Calamus spe. | Unknown porgy | SParidaE | Porgies |
| 10 CAN SJFF | 310 Canthidermis suff lamen | Ocean triggerfish | BaLISTIDAE | Leatherjackets |
| 11 CAR BART | 320 Carans bartholonaei | Yellow jack | CaRaNGIDAE | Jacke |
| 12 CHA CAPI | 370 Chaetodon capistratus | Foureye butterflyfish | CHAETODONTIDAE | Butterflyfishes |
| 13 CHA OCEL | 390 Chaetodon ocellatus | Spotfin butterflyfish | CHAETODONTIDAE | Butterflyfishes |
| 14 CHA SEDE | 400 Chaetodon sedentarius | Reef butterflyfish | CHAETODONTIDAE | Butterflyfishes |
| 15 CHR SCOT | 460 Chromis scotti | Purple reeffish | POHaCEITRIDAE | Damselfishes |
| 16 COR GLAT | 490 Coryphopterus glaucofraenus | Eridled goby | GOBIIDAE | Gobies |
| 17 CRY FOSE | 510 Cryptotonus roseus | Bluelip parrotfish | SCARIDAE | Parrotfishes |
| 18 EFI ADSC | 650 Epinephelus adscensionis | Rock hind | SERRANIDAE | Sea basses |
| 19 EPI CRUE | 660 Epinephelus cruentatus | Graysby | SERRANIDAE | Sea basses |
| 20 EFI MORI | 690 Epinephelus norio | Red grouper | SERRaNTIDAE | Sea basses |
| 21 EQU ACUH | 720 Equetus acuninatus | High-hat | SCIaENIDAE | Drums |
| 22 GHA THOH | 770 Gnatholepis thompsoni | Goldspot goby | GOBIIDAE | Gobies |
| 23 GOB OCEA | 790 Gobiosoma oceanops | Neon goby | GOBIIDAE | Gobies |
| 24 HAE PLUM | 940 Haenulon plunieri | White grunt | HaEmidilda | Grunts |
| 25 HAL BIVI | 970 Halichoeres bivittatus | Slippery dick | LabridaE | Wrasses |
| 26 HAL GARN | 980 Halichoeres garnoti | Yellowhead wrasse | LabridaE | Wrasses |
| 27 HAL MACJ | 990 Halichoeres maculipima | Clown wrasse | LabridaE | Wrasses |
| 28 HOL ADSC | 1070 Holocentrus adscensionis | Squirrelfish | HOLOCEHTRIDAE | Squirrelfishes |
| 29 HOL BERM | 1080 Holacanthus bermudensis | Blue angelfish | POHaCalthidaE | Angelfishes |
| 30 HOL CILI | 1090 Holacanthus ciliaris | Queen anglefish | POHACANTHIDAE | Angelfishes |
| 31 HOL MARI | 1110 Holocentrus marianus | Longjan squirrelfish | HOLOCEFTRIDAE | Squirrelfishes |
| 32 HOL RUFU | 1120 Holocentrus rufus | Longspine squirrelfish | HOLOCENTRIDAE | Squirrelfishes |
| 33 HYP PUEL | 1190 Hypoplectrus puella \# | Barred hamlet. | SERRANIDAE | Sea basses |
| 34 IOG CALL | 1215 Ioglossus calliurus | Blue goby | GOBIIDAE | Gobies |
| 35 LAC MAXI | 1260 Lachnolainus naximus | Hogf ish | LabridaE | Wrasses |
| 36 LUT CAMP | 1340 Lutjanus campechamus | Red snapper | LUTJantida | Snappers |
| 37 LUT GRIS | 1360 Lutjanus griseus | Gray snapper | LJTJakIDaE | Snappers |
| 38 MYC MICR | 1560 Mycteroperca sicrolepis | Gag | SERRaNIDAE | Sea basses |
| 39 MYC PHEN | 1570 Mycteroperca phenas | Scamp | SERRANIDAE | Sea basses |
| 40 OPI AJRI | 1680 Opistognathus aurifrons | Yellowhead jawfish | OPISTOGNaTHIDAE | Jawfishes |
| 41 FOM ARCO | 1745 Pomacanthus arcuatus | Gray angelfish | POHaCanthidae | Angelfishes |
| 42 POM DIEN | 1760 Pontacentrus diencaeus | Longf in danselfish | POHACENTRIDAE | Damselfishes |
| 43 POM PART | 1790 Pomacentrus partitus | Bicolor danselfish | POHACENTRIDAE | Damselfishes |
| 44 POM PLAN | 1810 Ponacentrus planifrons | Three spot damselfish | POHACEHTRIDAE | Damselfishes |
| 45 POM WARI | 1815 Pomacentrus variabilis | Cocoa danselfish | POHACENTRIDAE | Damselfishes |
| 46 PRI AREN | 1830 Priacanthus arenatus | Bigeye | PRIACASTHIDAE | Bigeyes |
| 47 FSE MACU | 1850 Pseudupeneus naculatus | Spotted goatfish | MULLIDAE | Goatfishes |
| 48 SCA CROI | 1900 Scarus croicensis | Striped parrotfish | SCARIDAE | Parrotfishes |
| 49 SER PHOE | 1950 Seriola phoebe | Tattler bass | SERRAMIDAE | Sea basses |
| 50 SER DUME | 2000 Seriola dumerili | Greater anberjack | CaRangidaE | Jacks |
| 51 SPA ATOH | 2050 Sparisoma atomariun | Greenblotch parrotfish | SCARIDAE | Parrotfishes |
| 52 SPA AURO | 2060 Sparisona aurofrenatum | Redband parrotifish | SCARIDAE | Parrotfishes |
| 53 SPA RADI | 2080 Sparisoma radians | Bucktooth parrotfish | SCARIDAE | Parrotfishes |
| 54 THA BIFA | 2200 Thalassoma bifasciatum | Bluehead | LabridaE | Wrasses |

Figure 1: Area of reef fish sampling in the Florida Middle Ground Habitat of Particular Concern, August 18-23, 2000. (Adapted from Smith et al. 1975)

## Sampling Sites

Dive 1: 28 32.353N 84 18.422W Dive 4: 2833.270 N 8420.343 W
Dive 2: 2836.366 N 8420.966 W Dive 5: 2830.350 N 8417.280 W
Dive 3: 28 32.364N 84 16.514W Dive 6: 2830.485 N 8417.334 W


# Chapter 6. Nassau grouper distribution and habitat characteristics at Little Cayman, Cayman Islands, British West Indies, December 2002. 

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## Introduction

Surveys made by the Reef Environmental Education Foundation (REEF) in January 2002 revealed a large spawning aggregation (SPAG) of Nassau grouper at the southwest corner of Little Cayman, Cayman Islands, BWI (Whaylen et al., 2004). Approximately 5,200 Nassau grouper, exhibiting courtship behavior and color change, were present at the spawning site and significant spawning events were witnessed and documented. Since there was an active fishery at the time, 1,934 groupers were captured from this aggregation, with fishers and divers vying for propriety of the aggregation site. The previous year, when the aggregation was discovered, fishers harvested approximately 2,000 Nassau groupers from this site (Whaylen et al., 2004). Nassau grouper SPAGs are particularly susceptible to overfishing with very low incidence of recovery or reformation (Olsen and LaPlace, 1979; Sadovy, 1994).

A team from the NOAA Fisheries, Southeast Fisheries Science Center (NOAA SEFSC) traveled to Little Cayman to observe spawning activity of Nassau grouper during the period surrounding the December full moon of $2002(12 / 19 / 02)$. The objectives for this research trip were to (1) document, if present, spawning behavior of Nassau grouper at Little Cayman Island spawning sites, (2) take fishery dependent samples from Nassau grouper landed during the SPAG to determine sex ratio, fecundity, and age and (3) characterize the associated fish assemblage and habitat. Sampling of fishes were conducted in December because of a pending government mandate to close fishing at spawning aggregation sites every other year, beginning in January 2003.

## Methods

We conducted underwater visual censuses by both roving and stationary techniques. Research methodologies represented an evolution of various census techniques used by the Reef Fish Research Team at NOAA SEFSC. Our research team of four was divided into two dive pairs. Each pair conducted Reef Visual Census (RVC) point counts at specific locations on the reef, and/or conducted focused predator counts during exploratory drift dives along the edge of the wall. Twenty one paired research dives were conducted during the five day research trip, resulting in sixteen RVC point counts and eleven predator counts. Almost all diving activity was done adjacent to the Cayman Island Shelf, an area commonly referred to as "the wall" (Figure 1).

The RVC technique, a standardized stationary point count method, was the dominant research tool utilized (Bohnsack and Bannerot, 1986). A diver attempts to count all individuals and species within five minutes in an imaginary, 7.5 meter ( 24 ft .) radius cylinder extending from the seafloor to the water's surface. New species are enumerated while rotating in ones cylinder. After the initial five minutes, divers systematically record data for each species seen (numbers and sizes) working from last to first observed. This method is highly versatile and is useful in both large scale and small scale surveys. Information collected provides a quantitative snapshot for future research on fish assemblages and spawning aggregations in this region. Prior research in Little Cayman (Pattengill-Semmens and Semmens, 2002; Whaylen et al., 2004) relied on the REEF Roving Diver Technique which, while more accessible and inclusive of both divers engaged and species observed, is less quantitative than the RVC technique (Bohnsack, 1996). The RVC technique was augmented by the additional recording of benthic habitat information (McClellan and Miller, 2003). RVC
research dives were conducted on a Cayman Island Department of Environment research vessel or a recreational SCUBA diving vessel provided by and run by the Southern Cross Club.

Alternatively, the predator counts allowed researchers to quantitatively scan the underwater landscape for conspicuous piscivores, particularly groupers (Eklund et al., 2000). Predator counts consisted of paired SCUBA divers swimming along a transect adjacent to the shelf edge, usually drifting with the prevailing current. The depth and length of these transects varied depending on the depth of the shelf edge and current speed and direction. One diver recorded numbers and sizes of conspicuous predatory fish, including time seen, while a second diver documented the habitat, fish assemblage, and fish behavior using digital photography or videography. Notable behaviors, color patterns, interactions, and physical conditions were additionally recorded. Surface support recorded starting and ending positions and times using a handheld GPS unit. These locations were later downloaded into a Geographic Information System (GIS) software package (Arcview 3.1) for mapping of diver effort in relation to aggregation sites (Figure 1). Information on the exact time each fish was seen helped map locations of these fish along the transect.

These quantitative samples characterized the distribution and abundance of grouper populations, species specific spawning activities, underlying habitat, and baseline fish assemblage information. Dives were conducted throughout the day and often included early evening or dusk dives. Nassau grouper spawning activities have been previously and almost exclusively documented at dusk. Extensive photographs and video were taken to help characterize habitat and document behaviors during these dives. Data collected during all research dives was entered into a computer software package designed by NOAA SEFSC and the University of Miami Rosenstiel School of Marine and Atmospheric Science (RSMAS). Summary statistics were produced using a bio-analysis program. Metadata descriptions of the Cayman Island research trip is in FGDC format and stored at the SEFSC clearinghouse (www.sefsc.noaa.gov).

Fishery dependent sampling was also to occur using Nassau grouper landed by local fishermen. They were to be weighed and measured, otoliths and dorsal fin spines and rays kept for age and growth analysis, and gonads extracted and preserved using a $10 \%$ formalin solution to determine sex, reproductive stage, and fecundity.

## Results and Discussion

## a. Fish Censuses

Overall, 100 fish species representing 30 families were observed in Little Cayman waters. All species had been previously reported at Little Cayman by Pattengill-Semmens and Semmens (2002). Hawksbill sea turtles, Eretmochelys imbricate; spiny lobster, Panulirus argus; and queen conch, Strombus gigas were also observed. Additional observations of almaco jack, Seriola rivoliana, and the tripletail, Lobotes surinamensis, observed floating with a piece of flotsam beyond the edge of the shelf was not included in our species list.

The predator count methodology is a search technique that allows for a quantitative scan of underwater habitats and can be used to cover long distances in a short period of time. It is useful when looking for a particular or conspicuous group of predatory fish, Nassau grouper in our case (Eklund et al., 2000). Our team conducted eleven predator counts along the Cayman Island shelf edge. We observed 23 predator species, 10 of which had not been documented during RVC point counts (Tables 1 and 5). While we did not locate a spawning aggregation, Nassau grouper were present in $100 \%$ of our predator searches (along with the schoolmaster snapper, Lutjanus apodus). Several species of jacks (Carangidae) were also conspicuous in predator counts since these fast moving fish tend to travel along or over the edge of the wall. Our RVC counts were done adjacent to or on top of this habitat, and therefore, tended to underestimate the numbers of jacks present.

Predator counts are similar in technique to REEF Roving Diver Surveys (RDT). In 2002, REEF reported sighting frequencies for four grouper species: tiger grouper (79.1\%), Nassau grouper (62.6\%), yellowfin grouper (Mycteroperca venenosa; 41.7\%), and black grouper (M. bonaci; 13.4\%) (Whaylen et al., in press). Our counts reported sighting frequencies of $72.3 \%, 100 \%, 36.36 \%$, and $72.73 \%$ respectively. Additionally, we also observed yellowmouth grouper (M. interstitalis; $27.27 \%$ ), graysby (E. cruentatus; 45.45\%), coney (E. fulvus; 36.36\%), and red hind (E. guttatus; 18.18\%).

A total of 16 RVC samples were accomplished and 68 species representing 21 families and 4,636 individuals were observed (Tables 1 and 2 ). The creole wrasse, Clepticus parrai $(\mathrm{N}=1,324$, $28.6 \%$ of total) was the most abundant species observed followed by the fairy basslet, Gramma loreto ( $\mathrm{N}=524,11.3 \%$ of total); blue chromis, Chromis cyanea $(\mathrm{N}=401,8.6 \%$ of total); and masked goby, Coryphoptrus personatus ( $\mathrm{N}=351,7.6 \%$ of total). These four planktivorous species represented over $56.1 \%$ of the total number observed. Creole wrasse; fairy basslet; black durgon, Melichthys niger; and bicolor damselfish, Pomacentrus partitus, were observed in $100 \%$ of the samples. Biomass estimates were derived for all species using length-to-weight comparisons compiled in Bohnsack and Harper (1988). Yellowtail snapper, Ocyurus chrysurus, represented the most biomass ( $69.5 \mathrm{~kg}, 23.2 \%$ of total) followed by Bermuda chub, Kyphosus sectatrix ( 41.9 kg , $14.0 \%$ of total), Nassau grouper, E. striatus ( $25.9 \mathrm{~kg}, 8.6 \%$ of total), and tiger grouper, Mycteroperca tigris ( $20.3 \mathrm{~kg}, 6.8 \%$ of total). Yellowtail snapper, Nassau grouper, and tiger grouper are conspicuous higher trophic level predators and important commercial species. Large grouper species appear at high densities and biomass, paralleling results by Chaippone et al. (2000) showing that areas experiencing light or no fishing have higher densities, biomass and species diversity of large groupers than areas with heavy fishing and no protection. The large biomass of Bermuda chub, an herbivore, is notable because these fish are occasionally observed in large numbers at Nassau grouper spawning sites in the Bahamas ( ${ }^{1}$ D.B. Eggleston, pers. comm.; Figure 4b).

Descriptive habitat data was collected in conjunction with the fish assemblage information during RVC fish counts. Habitat was consistent across sampling areas since all diving activity was conducted adjacent to the Cayman Island Shelf. Figure 2 depicts the benthic habitat composition for sites near the grouper aggregation site and compares it to a composite of sites along the wall. Habitat composition appears to be similar; however, additional samples must be collected in order to determine the degree of similarity. Figure 3 displays some images of typical habitat encountered during our research dives at Little Cayman.

Colin (1992) wrote, "differences in some physical or biological factors would be expected between spawning and other non-spawning locations, if indeed spawning sites were measurably superior locations". Domeier and Colin (1997) proposed objective criteria for characterizing spawning aggregations and their discrete locations in specific terms, rather than simply describing SPAGS as "spectacularly high densities of spawning size fish". They also called for a quantitative comparison of densities of aggregating fish between non-reproductive and reproductive periods. We feel, by using the RVC methodology, we can make an important contribution to this classification. Although we did not witness spawning, we can characterize the fish assemblage at a particular site using the RVC methodology and compare it to other sites over time. Tables 3 and 4 summarize the RVC data collected from both the Nassau grouper aggregation site and other sites along the wall. Table 6 compares summary data from these two groups of data to the total. While our research only presents a snapshot of this idea, hypothetically, if the sites were continuously censused, one would

[^0]expect to see significant differences such as a spike in the numbers of spawning fish at a spawning site versus a non-spawning site during a particular spawning season. Additionally, one might document the utilization of the site by other species during other times of the year. Due to limited time and personnel, we could not conduct enough samples to adequately analyze the differences in fish assemblages between the grouper aggregation site and other reference points along the Little Cayman Shelf. Continuing this data comparison over time could lead to better understandings of spawning aggregation site dynamics.

## b. Characterization of SPAG Location

The environmental characteristics common to spawning aggregation sites generally include a description of lunar cycle, light conditions, geomorphology, water temperature, currents, learned behavior, and geography. Reef promontories near drop off and near the most seaward point of islands can be preferred habitat for spawning (Colin 1992, Colin et al. 1987). The Little Cayman spawning aggregation site was in approximately 30 m of water on the edge of the Cayman Shelf. Here, high relief spur and groove formations run perpendicular to the wall edge as it curves around the contour of the southwest point of the island. While we did not experience the high currents common to this area, anecdotal reports indicate that currents between 1 and 3 knots are often present here. Dives were conducted during the appropriate lunar phase (full moon) for Nassau grouper spawning, however, Nassau grouper prefer water temperatures of about $26^{\circ} \mathrm{C}$ for spawning to occur (Tucker et al. 1993; Colin 1992; Carter et al. 1994), and water temperatures were consistently between $26.7^{\circ} \mathrm{C}$ and $28.3^{\circ} \mathrm{C}$.; a further indicator of less-than-optimal spawning conditions.

During diving activities, we noted uncharacteristically bold behavior by dog snapper and ocean triggerfish and high incidences of tiger and yellowfin groupers (no courtship behavior or coloration was documented). The physical and oceanographic conditions that make a site ideal for grouper spawning are also probably suitable for a suite of other gregarious spawners (Heyman, 2001). Whaylen et al. (2004) reported on the presence of ten additional species demonstrating courtship behavior and/or spawning at the Nassau grouper spawning site in January 2002: tiger grouper; yellowfin grouper; black grouper; horse-eye jack, Caranx latus; bar jack, C. rubber; black jack, C. lugubris; yellow jack, C. bartholomaei; mackerel scad, Decapterus macarellus; dog snapper, Lutjanus jocu; and ocean triggerfish, Canthidermis sufflamen. Eggleston (pers. com.) described spawning by the smooth trunkfish, Lactophrys triqueter, and horse-eye jacks, and high abundances of yellowtail snapper, Bermuda chub, and creole wrasse during the January 2003 spawning event. Carter et al. (1994) noted at a single site the presence of courting or spawning dog snapper, black grouper, yellowfin grouper, and coneys in Belize. The presence and behavior of these fishes (especially the non-fished ones) could indicate undiscovered or previously exploited spawning aggregation sites.

Notably, Nassau grouper were not present in sufficient numbers (nor were they exhibiting characteristic behavior or coloration) to define a spawning aggregation. The Nassau grouper observed along the wall both during RVC point counts ( $\mathrm{N}=9$ ) and predator searches $(\mathrm{N}=31$ ) ranged in size from 15 to 72 cm (mean $=43.3 \mathrm{~cm}$ ). Nassau grouper are thought to reach maturity between $40-45 \mathrm{~cm}$, which corresponds to 4-7 years of age (Sadovy and Eklund 1999). Colin (1992) describes four color phases for spawning Nassau grouper: white belly, bicolor, dark, and barred (normal). The Nassau groupers observed were generally solitary and were not exhibiting any courtship behavior or color changes. Only two Nassau groupers were observed with a very faint white belly color pattern (see Figure 4a), and two very small Nassau groupers ( $<40 \mathrm{~cm}$ ) were observed exhibiting some sort of mock territoriality or courtship behavior, blanching white over the
white sand as they circled each other tightly. However, several of the larger fish appeared to have somewhat distended abdomens.

## c. Fishery Dependent Sampling

Ripe Nassau grouper are collected from December through February (Sadovy and Eklund, 1999). However, most spawning activity for Nassau grouper in Cayman Islands occurs between the months of November and February, with the greatest activity during January and February (Tucker et al., 1993). If the full moon occurs late in the month (as in the case with 2002), spawning could be in December and January (Tucker et al., 1993; Colin et al., 2003). Because of the impending fishery closure, our team sampled the aggregation site on the full moon in mid-December to capitalize on the opportunity to take biological samples from fish caught at the aggregation site. The legislation enacted by the Cayman Islands Government in February 2002 protects Nassau grouper by closing all spawning aggregations to fishing in alternate years (i.e. closed 2003, open 2004 etc..), limiting the size and number of fish that can be taken in the open years (size limit of $>12$ inches ( 30.48 cm ), maximum of 12 fish per boat per day), and prohibiting traps within one nautical mile of any designated grouper spawning area ( ${ }^{2}$ Phillipe Bush, pers comm.)

Phillipe Bush, research manager at the Cayman Islands Department of Environment, has been working with the government to institute fishing closure for Nassau grouper aggregation sites during the spawning season. Last January's successful Grouper Moon project, lead by REEF, brought tremendous media attention to not only the fascinating ecology of these groupers, but also the imminent threat they face from overfishing. Pressure on the Cayman Island Government lead to the aggregation closures, effective in January 2003. Phillipe Bush hosted an informational forum for fishermen during the December 2002 full moon to explain the rationale and the rules for the closure, listening to their concerns, misconceptions and answering their questions.

Our research team expected to work with local fishermen to obtain reproductive and age and growth samples from fish captured during the last month of open Nassau grouper fishing prior to the January 2003 closure. In January 2002, approximately 1,934 Nassau grouper were landed at Little Cayman, many of which were captured at the spawning aggregation site (in the previous year, an estimated 2000 fish were taken). Cayman Island Department of the Environment scientists reported an average size of 61.9 cm for landed fish, and a female to male sex ratio of 1:1.6 (Whaylen et al., 2004, based on 275 fish measured, and 431 fish sexed). Other fished aggregation sites in the Caribbean indicates sex ratios skewed to females, which generally indicates the overharvesting of spawning fish (Colin et al., 1987; Colin, 1992; Carter et al., 1994). Thus, this Little Cayman aggregation may be relatively healthy in comparison to other historically fished aggregations in the Caribbean, despite the two years of take from this recently discovered site.

Most of the fishermen targeting the aggregation were from a visiting island, Cayman Brac, and many were insensitive to the impact that large groupers have on the recreational dive community on Little Cayman. Many Little Caymanians decided to boycott grouper meat during that time, and the harvested meat was either exported to Grand Cayman or consumed on Cayman Brac. Some locals reported that some of the harvested meat spoiled before it could be sold. Because of this incident and because of the pending legislation, Nassau grouper fishing in December 2002 was greatly curtailed, and we were unable to collect any fishery dependent samples. However, L. Alan Collins, a fishery biologist and reproductive histology expert from NOAA Fisheries' Panama City Laboratory, presented a talk to fisheries officers from Grand Cayman on the importance of collecting

[^1]age and growth and fecundity data, and our group will hopefully return to Grand Cayman in January of 2004 to collaboratively conduct this work. Very significant contacts were made during our visit to Little Cayman, and we will be able to capitalize on these relationships during future research opportunities.

## Conclusion

Little Cayman displays a very healthy, dynamic reef environment, with only artisanal level fishing pressure. However, with Nassau grouper SPAGs consistently under attack from these fishers, it is imperative to continue to monitor the health of these reefs and their ability, or more probably, their inability to sustain this level of take. Of the five Nassau grouper SPAGs documented in the Cayman Islands, three have disappeared or are commercially extinct (Whaylen et al., 2004). Since it is widely believed that Cayman Island aggregations are self-recruiting because of the expanse of deep water separating the islands, protection of these aggregations is paramount (Colin et al.,1987). Furthermore, Little Cayman's aggregations appear to be more resilient and larger than aggregations in other areas of the Caribbean (Whaylen et al., 2004) and may provide an exceptional opportunity for conservation and research. Unfortunately, the size limit enacted by the Cayman Island Department of Environment may be insufficient to ensure sub-adult Nassau grouper have an opportunity to reach maturity as immature fish will recruit into the fishery well before attaining sexual maturity. This has been a common theme throughout the Caribbean (Sadovy et al., 2000).

While these aggregations have been the focus of research efforts by the Cayman Island Department of the Environment, the surrounding habitat and fish assemblage must be characterized as well in order to provide a robust assessment of changes to the landscape. In 2002, Cayman Island Fisheries Officers visited Little Cayman monthly to monitor the spawning site. However, these reconnaissance dives did not quantify the conditions or species present. Use of the RVC

Methodology may provide a cost effective, efficient way to census the fish biota at spawning aggregation sites during and beyond the spawning season. The RVC and predator count methodologies are standardized and fairly easy to learn, and may offer an efficient, cost effective way for Caribbean fisheries managers to begin collecting statistically sound baseline data on overall reef health (Bohnsack, 1996). During the winter full moons of 2004, when the fishery is open, we hope to conduct similar surveys in conjunction with sampling landed fish from the region to further characterize the islands' fish fauna, habitat, and spawning aggregation dynamics; and to compare these results with prior studies both in the Cayman Islands and the greater Caribbean.

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Table 1. Phylogenic listing of families and species observed in visual samples from Little Cayman Island, December 2002. Names are according to Robins et al. (1991), with the exception that Hypoplectrus species (denoted by \#) which were listed as H. unicolor. The species codes were derived from the first three and four letters, respectively, of the genus and trivial species name. Trophic level codes: B browser, F piscivore, H herbivore, Ma macroinvertivore, Mi microinvertivoore, P planktivore. Predominate adult trophic mode indicated in bold (Bohnsack et al. 1999). $*=$ seen in predator searches only. $* *=$ seen at other times.

| FAMILY NAME | Scientific name | Family common name | Species common name | Trophic Level | Species <br> Code |
| :---: | :---: | :---: | :---: | :---: | :---: |
| RHINCODONTIDAE |  | Carpet sharks |  |  |  |
|  | Ginglymostoma cirratum |  | Nurse shark | Ma,F | GIN CIRR ** |
| MYLIOBATIDAE |  | Eagle rays |  |  |  |
|  | Aetobatus narinari |  | Spotted eagle ray | Ma | AET NARI ** |
| MURAENIDAE |  | Morays |  |  |  |
|  | Gymnothorax funebris |  | Green moray | F,Ma | GYM FUNE * |
| HOLOCENTRIDAE |  | Squirrelfishes |  |  |  |
|  | Holocentrus adscensionis |  | Squirrelfish | Ma, Mi | HOL ADSC |
|  | Holocentrus marianus |  | Longjaw squirrelfish | Ma,Mi | HOL MARI |
|  | Holocentrus rufus |  | Longspine squirrelfish | Ma, Mi | HOL RUFU |
|  | Myripristis jacobus |  | Blackbar soldierfish | P | MYR JACO |
| AULOSTOMIDAE |  | Trumpetfishes |  |  |  |
|  | Aulostomus maculatus |  | Trumpetfish | F | AUL MACU |
| SERRANIDAE |  | Sea basses |  |  |  |
|  | Epinephelus cruentatus |  | Graysby | F,Ma | EPI CRUE |
|  | Epinephelus fulvus |  | Coney | F,Ma | EPI FULV |
|  | Epinephelus guttatus |  | Red hind | Ma, F | EPI GUTT |
|  | Epinephelus striatus |  | Nassau grouper | F,Ma | EPI STRI |
|  | Hypoplectrus puella \# |  | Barred hamlet | Mi | HYP PUEL |
|  | Hypoplectrus unicolor\# |  | Butter hamlet | Mi | HYP UNIC |
|  | Liopropoma mowbrayi |  | Cave basslet | Mi | LIO MOWB ** |
|  | Liopropoma rubre |  | Peppermint bass | Mi | LIO RUBE |
|  | Mycteroperca bonaci |  | Black grouper | F,Ma | MYC BONA *** |
|  | Mycteroperca interstitialis |  | Yellowmouth grouper | F,Ma | MYC INTE *** |
|  | Mycteroperca tigris |  | Tiger grouper | F,Ma | MYC TIGR |
|  | Mycteroperca venenosa |  | Yellowfin grouper | F,Ma | MYC VENE |
|  | Rypticus saponaceus |  | Greater soapfish | F,Ma | RYP SAPO ** |
|  | Serranus tabacarius |  | Tobaccofish | Mi | SER TABA** |
|  | Serranus tigrinus |  | Harlequin bass | Mi | SER TIGR |
| GRAMMATIDAE |  | Basslets |  |  |  |
|  | Gramma loreto |  | Fairy basslet | Mi | GRA LORE |
|  | Gramma melacara |  | Blackcap basslet | Mi | GRA MELA |

Table 1 (cont.)

| FAMILY NAME | Scientific name | Family common name | Species common name | Trophic Level | Species Code |
| :---: | :---: | :---: | :---: | :---: | :---: |
| POMACANTHIDAE |  | Angelfishes |  |  |  |
|  | Holacanthus ciliaris |  | Queen angelfish | B | HOL CILI |
|  | Holacanthus tricolor |  | Rock beauty | B | HOL TRIC |
|  | Pomacanthus arcuatus |  | Gray angelfish | B | POM ARCU |
|  | Pomacanthus paru |  | French angelfish | B | POM PARU |
| POMACENTRIDAE |  | Damselfishes |  |  |  |
|  | Chromis cyanea |  | Blue chromis | P | CHR CYAN |
|  | Chromis insolata |  | Sunshinefish | P | CHR INSO |
|  | Chromis multilineata |  | Brown chromis | P | CHR MULT |
|  | Pomacentrus fuscus |  | Dusky damselfish | H | POM FUSC |
|  | Pomacentrus leucostictus |  | Cocoa damselfish | H | POM LEUC ** |
|  | Pomacentrus partitus |  | Bicolor damselfish | P | POM PART |
|  | Pomacentrus planifrons |  | Three spot damselfish | H | POM PLAN |
| CIRRHITIDAE |  | Hawkfishes |  |  |  |
|  | Amblycirrhitus pinos |  | Redspotted hawkfish | Mi | AMB PINO |
| LABRIDAE |  | Wrasses |  |  |  |
|  | Clepticus parrae |  | Creole wrasse | P | CLE PARR |
|  | Halichoeres garnoti |  | Yellowhead wrasse | $\mathrm{Ma}, \mathrm{Mi}$ | HAL GARN |
|  | Halichoeres maculipinna |  | Clown wrasse | Mi,Ma | HAL MACU |
|  | Lachnolaimus maximus |  | Hogfish | Ma | LAC MAXI |
|  | Thalassoma bifasciatum |  | Bluehead | P,Mi,Ma | THA BIFA |
| SPHYRAENIDAE |  | Barracudas |  |  |  |
|  | Sphyraena barracuda |  | Great barracuda | F,Ma | SPH BARR * |
| SCARIDAE |  | Parrotfishes |  |  |  |
|  | Scarus croicensis |  | Striped parrotfish | H | SCA CROI |
|  | Scarus taeniopterus |  | Princess parrotfish | H | SCA TAEN |
|  | Scarus vetula |  | Queen parrotfish | H | SCA VETU |
|  | Sparisoma aurofrenatum |  | Redband parrotfish | H | SPA AURO |
|  | Sparisoma chrysopterum |  | Redtail parrotfish | H | SPA CHRY |
|  | Sparisoma rubripinne |  | Redfin parrotfish | H | SPA RUBR ** |
|  | Sparisoma viride |  | Stoplight parrotfish | H | SPA VIRI |
| OPISTOGNATHIDAE |  | Jawfishes |  |  |  |
|  | Opistognathus aurifrons |  | Yellowhead jawfish | P | OPI AURI |
| CLINIDAE |  | Clinids |  |  |  |
|  | Malacoctenus boehlkei |  | Diamond blenny | Mi, P | MAL BOEH ** |
|  | Malacoctenus triangulatus |  | Saddled blenny | Mi, P | MAL TRIA ** |

Table 1 (cont.)

| FAMILY NAME | Scientific name | Family common name | Species common name | Trophic Level | Species Code |
| :---: | :---: | :---: | :---: | :---: | :---: |
| GOBIIDAE |  | Gobies |  |  |  |
|  | Coryphopterus glaucofraenum |  | Bridled goby | H | COR GLAU ** |
|  | Coryphopterus personatus |  | Masked goby | P | COR PERS |
|  | Gobiosoma evelynae |  | Sharknose goby | Mi , | GOB EVEL ** |
|  | Gobiosoma genie |  | Cleaning goby | Mi , | GOB GENI ** |
|  | Goby-like fish |  | Goby-like fish | Mi, H | GOB SPE./COR SPE. ** |
|  | Priolepis hipoliti |  | Rusty goby | P | PRI HIPO ** |
| ACANTHURIDAE |  | Surgeonfishes |  |  |  |
|  | Acanthurus bahianus |  | Ocean surgeon | H | ACA BAHI |
|  | Acanthurus chirurgus |  | Doctorfish | H | ACA CHIR |
|  | Acanthurus coeruleus |  | Blue tang | H | ACA COER |
| BALISTIDAE |  | Leatheriackets |  |  |  |
|  | Aluterus scriptus |  | Scrawled filefish | H,B | ALU SCRI ** |
|  | Balistes vetula |  | Queen triggerfish | Ma | BAL VETU ** |
|  | Canthidermis sufflamen |  | Ocean triggerfish | Ma, P | CAN SUFF |
|  | Melichthys niger |  | Black durgon | P | MEL NIGE |
| MONACANTHIDAE |  | Filefishes |  |  |  |
|  | Cantherhines macrocerus |  | Whitespotted filefish | B,H | CAN MACR ** |
| OSTRACIIDAE |  | Boxfishes |  |  |  |
|  | Lactophrys quadricornis |  | Scrawled cowfish | B | LAC QUAD |
|  | Lactophrys triqueter |  | Smooth trunkfish | B | LAC TRIQ ** |
| TETRAODONTIDAE |  | Puffers |  |  |  |
|  | Canthigaster rostrata |  | Sharpnose puffer | H,B,Mi | CAN ROST |
|  | Diodon holocanthus |  | Balloonfish | Ma | DIO HOLO |

Table 2. Statistical summary by species at Little Cayman Island, December 2002. Species are listed alphabetically by species code. Scientific names for codes are given in Table 1.

| Species | Total Indiv. | SampleFrequency $(\mathrm{N}=16)$ | Mean SAMP. FREQ. RANGE |  |  | FISH LENGTH (cm) |  |  | BIOMASS <br> Total(gms) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Abund. | High | Low | Mean | Min. | Max. |  |
| ACA BAHI | 20 | 37.5\% | 1.3 | 8 | 0 | 12.9 | 8 | 20 | 1,207.9 |
| ACA CHIR | 5 | 12.5\% | 0.3 | 3 | 0 | 17.6 | 14 | 25 | 761.7 |
| ACA COER | 32 | 93.8\% | 2.0 | 6 | 0 | 11.7 | 6 | 15 | 1,536.8 |
| AMB PINO | 1 | 6.3\% | 0.1 | 1 | 0 | 10.0 | 10 | 10 | 6.9 |
| AUL MACU | 3 | 18.8\% | 0.2 | 1 | 0 | 38.3 | 35 | 40 | 413.6 |
| CAN ROST | 10 | 31.3\% | 0.6 | 3 | 0 | 5.4 | 2 | 10 | 52.3 |
| CAN SUFF | 20 | 31.3\% | 1.3 | 8 | 0 | 29.0 | 20 | 40 | 11,822.1 |
| CAR LUGU | 5 | 31.3\% | 0.3 | 1 | 0 | 36.6 | 30 | 40 | 4,406.3 |
| CAR RUBE | 8 | 25.0\% | 0.5 | 4 | 0 | 17.3 | 7 | 40 | 1,779.0 |
| CHA ACUL | 4 | 12.5\% | 0.3 | 3 | 0 | 8.0 | 6 | 8 | 51.6 |
| CHA CAPI | 16 | 56.3\% | 1.0 | 2 | 0 | 8.4 | 6 | 12 | 389.9 |
| CHA OCEL | 7 | 25.0\% | 0.4 | 2 | 0 | 9.4 | 8 | 12 | 175.1 |
| CHA SEDE | 10 | 6.3\% | 0.6 | 10 | 0 | 6.0 | 4 | 15 | 156.1 |
| CHA STRI | 4 | 18.8\% | 0.3 | 2 | 0 | 9.5 | 8 | 10 | 106.8 |
| CHR CYAN | 401 | 93.8\% | 25.1 | 50 | 0 | 6.1 | 3 | 10 | 2,536.0 |
| CHR INSO | 2 | 6.3\% | 0.1 | 2 | 0 | 2.0 | 2 | 2 | 0.3 |
| CHR MULT | 65 | 25.0\% | 4.1 | 27 | 0 | 5.2 | 3 | 7 | 233.1 |
| CLE PARR | 1,324 | 100.0\% | 82.8 | 250 | 5 | 8.1 | 4 | 17 | 11,182.8 |
| COR PERS | 351 | 37.5\% | 21.9 | 120 | 0 | 1.9 | 1 | 4 | 40.2 |
| DIO HOLO | 1 | 6.3\% | 0.1 | 1 | 0 | 18.0 | 18 | 18 | 163.3 |
| EPI CRUE | 17 | 50.0\% | 1.1 | 6 | 0 | 15.6 | 8 | 30 | 1,373.5 |
| EPI FULV | 5 | 18.8\% | 0.3 | 2 | 0 | 13.2 | 12 | 16 | 193.2 |
| EPI GUTT | 1 | 6.3\% | 0.1 | 1 | 0 | 24.0 | 24 | 24 | 218.9 |
| EPI STRI | 9 | 37.5\% | 0.6 | 3 | 0 | 50.4 | 30 | 72 | 25,861.3 |
| GOB SPE. | 1 | 6.3\% | 0.1 | 1 | 0 | 4.0 | 4 | 4 | 0.8 |
| GRA LORE | 524 | 100.0\% | 32.8 | 150 | 2 | 5.9 | 2 | 8 | 1,426.3 |
| GRA MELA | 210 | 25.0\% | 11.9 | 100 | 0 | 6.9 | 3 | 10 | 650.0 |
| HAE CARB | 50 | 6.3\% | 3.1 | 50 | 0 | 23.0 | 20 | 25 | 10,662.9 |
| HAE FLAV | 26 | 43.8\% | 1.6 | 12 | 0 | 11.4 | 10 | 30 | 1,186.0 |
| HAE PARR | 25 | 6.3\% | 1.6 | 25 | 0 | 25.0 | 20 | 30 | 7,671.6 |
| HAE PLUM | 23 | 18.8\% | 1.4 | 15 | 0 | 21.6 | 12 | 30 | 5,411.4 |
| HAE SCIU | 55 | 31.3\% | 3.4 | 25 | 0 | 21.6 | 14 | 30 | 11,614.6 |
| HAL GARN | 32 | 75.0\% | 2.0 | 6 | 0 | 7.7 | 4 | 15 | 326.4 |
| HAL MACU | 2 | 12.5\% | 0.1 | 1 | 0 | 6.0 | 6 | 6 | 4.1 |
| HOL ADSC | 15 | 25.0\% | 0.9 | 6 | 0 | 15.7 | 10 | 22 | 1,649.4 |
| HOL CILI | 1 | 6.3\% | 0.1 | 1 | 0 | 18.0 | 18 | 18 | 147.3 |
| HOL MARI | 7 | 25.0\% | 0.4 | 3 | 0 | 8.9 | 6 | 12 | 174.7 |
| HOL RUFU | 9 | 31.3\% | 0.6 | 3 | 0 | 16.0 | 14 | 20 | 722.3 |
| HOL TRIC | 3 | 18.8\% | 0.2 | 1 | 0 | 12.3 | 10 | 15 | 181.1 |
| HYP PUEL | 6 | 25.0\% | 0.4 | 2 | 0 | 7.7 | 6 | 10 | 47.0 |
| HYP UNIC | 1 | 6.3\% | 0.1 | 1 | 0 | 8.0 | 8 | 8 | 8.2 |
| KYP SECT | 77 | 62.5\% | 4.8 | 25 | 0 | 26.4 | 12 | 50 | 41,949.7 |
| LAC MAXI | 2 | 12.5\% | 0.1 | 1 | 0 | 57.5 | 55 | 60 | 7,400.3 |
| LAC QUAD | 1 | 6.3\% | 0.1 | 1 | 0 | 35.0 | 35 | 35 | 546.4 |
| LIO RUBE | 3 | 6.3\% | 0.2 | 3 | 0 | 2.0 | 2 | 2 | 27.0 |
| LUT APOD | 78 | 50.0\% | 4.9 | 44 | 0 | 17.2 | 10 | 40 | 9,528.5 |
| LUT JOCU | 2 | 12.5\% | 0.1 | 1 | 0 | 45.0 | 30 | 60 | 4,226.9 |
| LUT MAHO | 70 | 18.8\% | 4.4 | 35 | 0 | 16.6 | 10 | 25 | 6,950.5 |
| MEL NIGE | 108 | 100.0\% | 6.8 | 17 | 1 | 14.3 | 8 | 25 | 7,634.1 |
| MUL MART | 4 | 18.8\% | 0.3 | 2 | 0 | 17.8 | 15 | 20 | 334.0 |
| MYC TIGR | 15 | 43.8\% | 0.9 | 7 | 0 | 37.3 | 12 | 80 | 20,342.4 |
| MYC VENE | 2 | 12.5\% | 0.1 | 1 | 0 | 30.0 | 30 | 30 | 778.0 |
| MYR JACO | 19 | 18.8\% | 1.2 | 16 | 0 | 18.3 | 8 | 20 | 3,045.3 |
| OCY CHRY | 230 | 93.8\% | 14.4 | 50 | 0 | 25.2 | 14 | 50 | 69,536.8 |
| OPI AURI | 20 | 6.3\% | 1.3 | 20 | 0 | 6.0 | 6 | 6 | 39.4 |

Table 2 (cont.)

| Species | Total Indiv. | Sample <br> Frequency ( $\mathrm{N}=16$ ) | Mean SAMP. FREQ. RANGE |  |  | FISH LENGTH (cm) |  |  | BIOMASS <br> Total(gms) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Abund. | High | Low | Mean | Min. | Max. |  |
| POM ARCU | 4 | 18.8\% | 0.3 | 2 | 0 | 26.5 | 20 | 35 | 2,650.8 |
| POM FUSC | 12 | 18.8\% | 0.8 | 8 | 0 | 5.7 | 4 | 10 | 77.4 |
| POM PART | 238 | 100.0\% | 14.9 | 34 | 1 | 4.7 | 2 | 10 | 827.9 |
| POM PARU | 4 | 18.8\% | 0.3 | 2 | 0 | 31.3 | 20 | 35 | 4,335.8 |
| POM PLAN | 20 | 37.5\% | 1.3 | 7 | 0 | 6.4 | 4 | 12 | 202.3 |
| SCA CROI | 70 | 93.8\% | 4.4 | 20 | 0 | 8.8 | 4 | 20 | 1,125.9 |
| SCA TAEN | 33 | 56.3\% | 2.1 | 6 | 0 | 11.6 | 4 | 34 | 1,483.0 |
| SCA VETU | 1 | 6.3\% | 0.1 | 1 | 0 | 28.0 | 28 | 28 | 395.0 |
| SER TIGR | 5 | 31.3\% | 0.3 | 1 | 0 | 8.0 | 6 | 10 | 47.4 |
| SPA AURO | 23 | 56.3\% | 1.4 | 6 | 0 | 9.5 | 4 | 20 | 745.1 |
| SPA CHRY | 2 | 12.5\% | 0.1 | 1 | 0 | 24.5 | 24 | 25 | 503.1 |
| SPA VIRI | 27 | 62.5\% | 1.7 | 8 | 0 | 20.6 | 5 | 40 | 8,104.8 |
| THA BIFA | 255 | 87.5\% | 15.9 | 32 | 0 | 5.6 | 1 | 15 | 570.4 |
| NO. SAMPLES $=16$ |  |  |  |  |  |  |  |  |  |
| NO. SPECIES = |  | 68 |  |  |  |  |  |  |  |
| TOT.INDIVIDUALS = |  | 4,636 |  |  |  |  |  |  |  |
| TOT. BIOMASS (g) = |  | 299,961.10 |  |  |  |  |  |  |  |

Table 3. Analysis of RVC samples taken at the Little Cayman Island Nassau grouper SPAG, December 2002.


Table 4. Analysis of RVC samples taken at the Little Cayman Island Bloody Wall site, December 2002.

| Species | Total SAMPLE FREQ |  | Mean Abund. | SAMP.FREQ. RANGE \| |  | FISH LENGTH (cm) |  |  |  | BIOMASS <br> Total(gms) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Indiv. | ( $\mathrm{N}=12$ ) |  | High | Low \| | Mean | Min. | Max. |  |  |
| ACA BAHI | 7 | 25.0\% | 0.6 | 4 | 0 \| | 12.6 | 8 | 20 |  | 457.2 |
| ACA COER | 15 | 83.3\% | 1.3 | 3 | 0 \| | 11.8 | 6 | 15 |  | 772.9 |
| AUL MACU | 3 | 25.0\% | 0.3 | 1 | 0 \| | 38.3 | 35 | 40 |  | 413.6 |
| CAN ROST | 9 | 33.3\% | 0.8 | 3 | 0 \| | 5.4 | 2 | 10 |  | 50.1 |
| CAN SUFF | 5 | 16.7\% | 0.4 | 3 | 01 | 20.0 | 20 | 20 |  | 831.7 |
| CAR LUGU | 2 | 16.7\% | 0.2 | 1 | 01 | 40.0 | 40 | 40 |  | 2,265.4 |
| CAR RUBE | 6 | 16.7\% | 0.5 | 4 | 01 | 11.3 | 7 | 18 |  | 200.0 |
| CHA ACUL | 4 | 16.7\% | 0.3 | 3 | 0 I | 8.0 | 6 | 8 |  | 51.6 |
| CHA CAPI | 10 | 50.0\% | 0.8 | 2 | 0 \| | 8.4 | 6 | 12 |  | 233.1 |
| CHA OCEL | 7 | 33.3\% | 0.6 | 2 | 0 \| | 9.4 | 8 | 12 |  | 175.1 |
| CHA SEDE | 10 | 8.3\% | 0.8 | 10 | 01 | 6.0 | 4 | 15 |  | 156.1 |
| CHA STRI | 4 | 25.0\% | 0.3 | 2 | 01 | 9.5 | 8 | 10 |  | 106.8 |
| CHR CYAN | 318 | 91.7\% | 26.5 | 50 | 01 | 6.3 | 4 | 10 |  | 2,249.5 |
| CHR INSO | 2 | 8.3\% | 0.2 | 2 | 01 | 2.0 | 2 | 2 |  | 0.3 |
| CHR MULT | 65 | 33.3\% | 5.4 | 27 | 01 | 5.2 | 3 | 7 |  | 233.1 |
| CLE PARR | 849 | 100.0\% | 70.8 | 150 | 51 | 7.8 | 4 | 15 |  | 6,337.5 |
| COR PERS | 351 | 50.0\% | 29.3 | 120 | 01 | 1.9 | 1 | 4 |  | 40.2 |
| DIO HOLO | 1 | 8.3\% | 0.1 | 1 | 01 | 18.0 | 18 | 18 |  | 163.3 |
| EPI CRUE | 13 | 50.0\% | 1.1 | 6 | 0 \| | 12.5 | 8 | 18 |  | 415.2 |
| EPI FULV | 3 | 16.7\% | 0.3 | 2 | 01 | 13.3 | 12 | 16 |  | 121.2 |
| EPI STRI | 9 | 50.0\% | 0.8 | 3 | 01 | 50.4 | 30 | 72 |  | 25,861.3 |
| GOB SPE. | 1 | 8.3\% | 0.1 | 1 | 01 | 4.0 | 4 | 4 |  | 0.8 |
| GRA LORE | 470 | 100.0\% | 39.2 | 150 | 21 | 6.1 | 2 | 8 |  | 1,376.6 |
| GRA MELA | 210 | 33.3\% | 15.8 | 100 | 01 | 6.9 | 3 | 10 |  | 650.0 |
| HAE CARB | 50 | 8.3\% | 4.2 | 50 | 01 | 23.0 | 20 | 25 |  | 10,662.9 |
| HAE FLAV | 10 | 33.3\% | 0.8 | 5 | 01 | 13.0 | 10 | 30 |  | 841.9 |
| HAE PARR | 25 | 8.3\% | 2.1 | 25 | 01 | 25.0 | 20 | 30 |  | 7,671.6 |
| HAE PLUM | 20 | 16.7\% | 1.7 | 15 | 01 | 22.8 | 14 | 30 |  | 5,251.8 |
| HAE SCIU | 34 | 25.0\% | 2.8 | 25 | 0 I | 22.6 | 14 | 30 |  | 8,356.4 |
| HAL GARN | 15 | 66.7\% | 1.3 | 4 | 01 | 8.2 | 4 | 15 |  | 186.1 |
| HAL MACU | 1 | 8.3\% | 0.1 | 1 | 01 | 6.0 | 6 | 6 |  | 2.1 |
| HOL ADSC | 15 | 33.3\% | 1.3 | 6 | 0 \| | 15.7 | 10 | 22 |  | 1,649.4 |
| HOL CILI | 1 | 8.3\% | 0.1 | 1 | 01 | 18.0 | 18 | 18 |  | 147.3 |
| HOL MARI | 7 | 33.3\% | 0.6 | 3 | 01 | 8.9 | 6 | 12 |  | 174.7 |
| HOL RUFU | 8 | 33.3\% | 0.7 | 3 | 01 | 16.0 | 14 | 20 |  | 649.8 |
| HOL TRIC | 2 | 16.7\% | 0.2 | 1 | 01 | 12.5 | 10 | 15 |  | 129.1 |
| HYP PUEL | 5 | 25.0\% | 0.4 | 2 | 01 | 7.6 | 6 | 10 |  | 38.9 |
| HYP UNIC | 1 | 8.3\% | 0.1 | 1 | 01 | 8.0 | 8 | 8 |  | 8.2 |
| KYP SECT | 58 | 50.0\% | 4.8 | 25 | 01 | 28.9 | 18 | 50 |  | 37,826.5 |
| LAC MAXI | 2 | 16.7\% | 0.2 | 1 | 01 | 57.5 | 55 | 60 |  | 7,400.3 |
| LIO RUBE | 3 | 8.3\% | 0.3 | 3 | 01 | 2.0 | 2 | 2 |  | 27.0 |
| LUT APOD | 72 | 41.7\% | 6.0 | 44 | 01 | 16.5 | 10 | 40 |  | 7,150.4 |
| LUT MAHO | 70 | 25.0\% | 5.8 | 35 | 01 | 16.6 | 10 | 25 |  | 6,950.5 |
| MEL NIGE | 73 | 100.0\% | 6.1 | 17 | 11 | 14.6 | 10 | 20 | 1 | 5,357.9 |
| MUL MART | 4 | 25.0\% | 0.3 | 2 | 01 | 17.8 | 15 | 20 | \| | 334.0 |
| MYC TIGR | 13 | 41.7\% | 1.1 | 7 | 01 | 35.1 | 12 | 55 | I | 11,117.1 |
| MYC VENE | 1 | 8.3\% | 0.1 | 1 | 0 I | 30.0 | 30 | 30 | 1 | 389.0 |
| MYR JACO | 1 | 8.3\% | 0.1 | 1 | 01 | 8.0 | 8 | 8 | \| | 17.8 |
| OCY CHRY | 160 | 91.7\% | 13.3 | 50 | 01 | 26.2 | 14 | 50 | 1 | 53,489.3 |
| POM ARCU | 3 | 16.7\% | 0.3 | 2 | 0 I | 28.7 | 20 | 35 | - | 2,400.6 |
| POM FUSC | 11 | 16.7\% | 0.9 | 8 | 01 | 5.3 | 4 | 6 | \| | 49.6 |
| POM PART | 168 | 100.0\% | 14.0 | 25 | 61 | 4.7 | 2 | 10 | \| | 605.0 |

Table 4 (cont.)

| Species | Total SAMPLE FREQ |  | Mean <br> Abund | SAMP. FREQ. RANGE । |  | FISH LENGTH (cm) |  |  | BIOMASS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Indiv. | ( $\mathrm{N}=12$ ) |  | High | Low \| | Mean | Min. | Max. | Total(gms) |
| POM PARU | 1 | 8.3\% | 0.1 | 1 | 0 \| | 20.0 | 20 | 20 | 237.5 |
| POM PLAN | 20 | 50.0\% | 1.7 | 7 | 01 | 6.4 | 4 | 12 | 202.3 |
| SCA CROI | 57 | 100.0\% | 4.8 | 20 | 11 | 8.4 | 4 | 20 | 747.4 |
| SCA TAEN | 25 | 58.3\% | 2.1 | 6 | 01 | 12.4 | 4 | 34 | 1,360.7 |
| SCA VETU | 1 | 8.3\% | 0.1 | 1 | 01 | 28.0 | 28 | 28 | 395.0 |
| SER TIGR | 2 | 16.7\% | 0.2 | 1 | 01 | 8.0 | 6 | 10 | 19.6 |
| SPA AURO | 15 | 50.0\% | 1.3 | 6 | 01 | 9.7 | 5 | 20 | 460.7 |
| SPA CHRY | 2 | 16.7\% | 0.2 | 1 | 01 | 24.5 | 24 | 25 | 503.1 |
| SPA VIRI | 22 | 66.7\% | 1.8 | 8 | 01 | 18.5 | 5 | 40 | 5,423.5 |
| THA BIFA | 169 | 83.3\% | 14.1 | 32 | 01 | 5.6 | 1 | 15 | 419.0 |
| NO. SAMPLES = |  | 12 |  |  |  |  |  |  |  |
| NO. SPECIES = |  | 62 |  |  |  |  |  |  |  |
| TOT.INDIVIDUALS |  | 3,521 |  |  |  |  |  |  |  |
| BIOMASS ( g ) = |  | 221,816.20 |  |  |  |  |  |  |  |

Table 5. Alphabetical summary of piscivores observed during drift predator counts along the Little Cayman Island's shelf edge. Table 1 defines species code.

| Species | $\begin{array}{c}\text { Number of } \\ \text { Individuals }\end{array}$ | $\begin{array}{l}\text { Sample } \\ \text { Frequency }\end{array}$ | Mean | $\begin{array}{c}\text { FISH LENGTH (cm) } \\ \text { Min }\end{array}$ |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
| CAR BART | 115 | $18.18 \%$ | 26.5 | 18 | 35 |
| Max |  |  |  |  |  |$]$

Table 6. Comparison of fish fauna between the southwest Little Cayman Island Nassau grouper SPAG and similar sites along the little Cayman Island shelf.

|  | Overall | Aggregation Site | Little Cayman Island Shelf |
| :---: | :---: | :---: | :---: |
| Number of RVC |  |  |  |
| Counts | 16 | 4 | 12 |
| Total Number of |  |  |  |
| Species Observed | 68 | 43 | 62 |
| Mean Biomass |  |  |  |
| (grams/individual) | 64.7 | 70.1 | 63.0 |
| Most Abundand 5 |  | Creole Wrasse, Bluehead Wrasse, Blue |  |
| species in declining | Creole Wrasse, Fairy Basslet, Blue | Chromis, Yellowtail Snapper (tie), Bicolored | Creole Wrasse, Fairy Basslet, Masked |
|  | Chromis, Masked Goby, Bluehead Wrasse | Damselfish (tie), Fairy Basslet | Goby, Blue Chromis, Blackcap Basslet |
| Species Present in $100 \%$ of Samples |  | Blue Chromis, Blue Tang, Creole Wrasse, Fairy Basslet, Yellowhead Wrasse, Bermuda | Creole Wrasse, Fairy Basslet, Black |
|  | Creole Wrasse, Fairy Basslet, Black Durgon, Bicolored Damselfish, | Chub, Black Durgon, Yellowtail Snapper, Bicolored Damselfish, Bluehead Wrasse | Durgon, Bicolored Damsel, Striped Parrotish |
| Five species |  |  |  |
| representing the most | Yellowtail Snapper (23.2\%), Bermuda Chub | Yellowtail Snapper (20.5\%), Ocean Trigger | Yellowtail Snapper (24.1\%), Bermuda |
| biomass in declining order | (14.0\%), Nassau Grouper (8.6\%), Tiger Grouper (6.8\%), Ocean Trigger (3.9\%) | (14.1\%), Tiger Grouper (11.8\%), Creole Wrasse (6.2\%), Bermuda Chub (5.3\%) | Chub (30.8\%), Nassau Grouper <br> (21.1\%), Caesar Grunt (8.7\%) |

Figure 1. Little Cayman Island, Cayman Islands, British West Indies. Triangles represent predator counts and Red circles represent RVC point counts. The aggregation site was located at the southwest point of the island.


Figure 2. Habitat analysis of Little Cayman RVC sampling sites, December 2002.


Figure 3. Representative photographs of habitat sampled at Little Cayman Island. All habitats are adjacent to the shelf edge. 3 a - Nassau grouper aggregation site located adjacent to the shelf edge on the southwest corner of Little Cayman Island. The visible mooring line and float were installed by the Cayman Island Department of Environment during our research trip. 3b-Diver conducting RVC point count. A dog snapper swims in the foreground. 3c - Diver conducting roving predator search along the edge of the wall.

$3 b$.


3c


Figure 4. Representative photographs of fauna encountered during research dives on Little Cayman Island. $4 \mathrm{a}-$ Solitary Nassau grouper, exhibiting faint "white belly" coloration, swimming adjacent to the wall; note distended abdomen. 4 b - Large aggregation of Bermuda chub encountered during predator search. $4 \mathrm{c}-$ Typical assemblage of fish swarming above the reef consisting of predominantly of black durgon, creole wrasse, and blue chromis.


4b.


4c.



[^0]:    ${ }^{1}$ Eggleston, David B. 2002. North Carolina State University, Dept. of MEAS, Raleigh, NC 27695-7840. Personal commun.

[^1]:    ${ }^{2}$ Bush, Phillipe. 2003. Cayman Island Department of Environment P.O. Box 486 GT. Grand Cayman, BWI. Personal commun.

