# DRAFT <br> STATUS OF RED GROUPER IN UNITED STATES WATERS OF THE GULF OF MEXICO DURING 1986-2001 revised <br> Southeast Fisheries Science Center Staff 

December 14, 2002
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Sustainable Fisheries Division Contribution No. SFD-01/02-175rev

## Introduction

The primary objective of the analyses documented in this report was to update the assessment of Schirripa et al. (1999) as used by the Gulf of Mexico Fishery Management Council's Reef Fish Stock Assessment Panel (RFSAP) for developing management advice (RFSAP 2000). Sensitivity of the assessment to new biological information and other aspects was also examined.

At the Reef Fish Stock Assessment Panel meeting 17-20 September 2002 in Miami, FL the panel asked for clarification on some aspects of the 6 September 2002 version of this document and additional information and analyses were presented. In this revised document clarifications have been added and an addendum with information presented to the Panel during and after the meeting has been added. The primary modifications to the original report included: (1) clarification that yield is reported in gutted weight, (2) clarification of the approach used in calculating sampled age composition, (3) inclusion of relevant equations in the description of the estimation of derived age composition and associated dead discards, and (4) standardization of the labels (FECvsTL, FECvsAge, etc.) for the per capita fecundity functions to maintain consistency with the labeling used with additional analyses developed at the meeting. The addendum includes: estimated landings and harvest in mt, further comparisons of derived and sampled age composition, the weighting factors used in ASAP, additional analyses of per capita fecundity, and additional stock assessments and projections.

## BIOLOGICAL INFORMATION

## size information

Weights were tabulated and reported in kg and mt . All weights presented in the report are gutted weights.

The equations used to convert fork and standard length to total length, total length to gutted weight and whole weight to gutted weight were from Goodyear and Schirripa (1993) and are presented in Table 1. Those equations were in inches and pounds, so observations in metric units were converted as necessary.

Mean weights were used to calculate (1) the recreational yield from observed numbers caught and (2) the number of fish caught by the commercial sector (see below). When available, observed weights were used, otherwise weights were calculated from length. If both length and weight were available for a fish, then the data were examined to determine whether the information appeared reasonable. Following the approach used by Schirripa et al. (1999), if a predicted weight was less than $75 \%$ or greater than $150 \%$ of an observed weight, then the weight was rejected and the observed length was retained.

Length was used in determining the age composition of the catch. Following Schirripa et al. (1999) the lengths from four recreational modes (shore, headboat, charter and private) in inches were used. The observed annual length frequencies by mode are presented in Table 2; shore mode is not shown because there were less than 25 observed lengths in any year. Headboat was the only mode with more than 100 observed lengths in every year; for both charter and private modes there were less than 100
observed lengths in six years. Overall the private mode had the fewest observed lengths.. Also, following Schirripa et al.(1999), the lengths from two commercial groups (longline and handline plus other) were used and the associated annual size frequencies are shown in Table 3. For the longline fishery there were roughly 1,000 to 40,000 fish measured annually while for the handline and other fisheries there were about 1,000 to 10,000 measured annually.

## sampled age composition

Two types of age composition information were used in the various assessments reported in this document. One type, referred to as derived age composition, was calculated primarily from the observed length samples by use of the probabilistic method of Goodyear (1997). Derived age composition was the only type of age composition used in the previous assessment.

The second type of age composition used in the assessments was based on ages read from otoliths (Lombari-Carlson et al. 2002) and will be referred to as sampled age composition. In a review of the 1999 red grouper assessment Sullivan (1999) recommended that age composition from samples be incorporated in the assessment. This section briefly notes a few aspects of the samples provided by the NMFS Panama City Laboratory and the development of effective weights applied to these samples in ASAP.

Lombardi-Carlson et al. (2002) provided information on the age composition of the red grouper catch in the Gulf of Mexico. Their Table 1 showed that the sampling levels differed among sectors (commercial and recreational) and among commercial gears and recreational modes; almost $80 \%$ of the samples came from the commercial sector and most were from longline. The majority of the samples were from central and northwestern Florida.

The assessment program (ASAP, Legault and Restrepo 1998, Schirripa et al. 1999) measures the disparity between the observed and model-predicted frequency distribution of age by use of a multinomial negative log-likelihood function,

$$
L=\sum_{a, i} n_{i, y} o_{i, a, y} \ln \left(p_{i, a, y}\right)
$$

where $o_{i, a, y}$ is the observed relative frequency of age class $a$ in the catch of fishery $i$ during year $y, p_{i, a, y}$ is the corresponding model prediction and $n$ is the effective sample size. Inasmuch as the available age data were randomly sampled from various strata within a fishery, but not necessarily in proportion to the catch corresponding to those strata, the age composition of the fishery was computed as a catch-weighted average:
$o_{i, a, y}=\frac{\sum_{s} C_{i, s, y} f_{i, s, a, y}}{n_{i, y} \sum_{s} c_{i, s, y}}$
$n_{i, y}=\sum_{a}\left(\frac{\sum_{s} C_{i, s, y} f_{i, s, a, y}}{\sum_{s} C_{i, s, y}}\right)$
where $f_{i, s, a, y}$ is the observed frequency of age $a$ in the sample and $C_{i, s, y}$ is the total catch from strata $s$ of fishery $i$ in year $y$.

Weighted age compositions were computed for each of the three fishery categories included in the assessment (longline, handline plus other commercial gears and recreational) with geographic (northwest FL, central west FL and southwest FL) and gear/mode strata. For longline there were three geographic strata; for handline and other gears there were six strata (three geographic and two gears: handline and trap), and for recreational there were three strata (headboat, charter and private: geographic stratification could not be used because MRFSS estimates were for the entire west Florida)..

The weighted, sampled age composition for the three fishery categories with the effective sample sizes are presented in Table 4. The only category with effective sample sizes over 50 in most years was longline. Examination of the annual relative frequencies for longline indicates that the modal age ranged from 6 to 8 and most often was 7 (Figure 1).

## growth

The growth equation used in the previous assessment (Goodyear 1994) was used (Table 1). The catches and catch rates were tabulated on an annual basis, so the age composition for the assessment was also tabulated on an annual basis. The growth curve used was based on otoliths which were assumed to form a mark 1 June each year; to convert the timing of the growth curve to a calendar year, the fraction of a year from Jan. 1 to June 1 was added to $t_{0}$; therefore age 1 fish in the assessment were actually aged 0.6 to 1.59 .

Lombardi-Carlson et. al. (2002) presented growth equations derived from the 1992-2001 ageing analyses. Their Richards function was compared to the equation used in the previous assessment (Goodyear 1994) and found to be similar within the range of ages for which observations existed (Figure 2) but to diverge outside of that range, particularly at younger ages. The implications of alternative growth curves have not been investigated.

## reproduction

The previous assessment (Schirripa et al. 1999) modeled the per capita fecundity of the red grouper population as the product of gonad weight and the proportion of each age class that was female. Gonad weight (GW) was expressed as a power function of total length (TL), GW=2.073E-07 TL. ${ }^{6.092}$, which was converted to a function of age via the growth equation. The proportion female was expressed by a logistic function of age fit to data from Moe (1969) and Koenig (1993). Females older than two years were assumed to be mature and the abundance of males was assumed not to be limiting.

Collins et al. (2002) have recently analyzed over 2,000 red grouper gonads sampled from the eastern Gulf of Mexico during 1992 to 2001. They expressed batch fecundity (BF) in number of eggs as an exponential function of total length, $\mathrm{BF}=6742.1 \mathrm{e}^{0.0058 \mathrm{TL}}$, or total age, $\mathrm{BF}=67085 \mathrm{e}^{0.1616 \mathrm{age}}$. Inasmuch as the assessment requires fecundity at age, the batch fecundity at length curve must be converted to batch fecundity at age by use of a growth equation (just as Shirripa et al. 1999 did with gonad weight). The converted length curve, however, is rather different from the age curve- having a much slower ascent
(Figure 3). For statistical reasons one would normally prefer a curve fit directly to age over the use of an imprecise growth equation to convert a curve fit to length. However the region of the batch fecundity at age curve with the steepest ascent is largely beyond the range of the data and is heavily influenced by a single large batch fecundity observation at age 21 (Figure 3). For these reason we have preferred the ageconverted batch fecundity at length curve for this assessment.

Collins et al. (2002) also provided age-specific data on the sex ratio and proportion of the female population that is actively spawning, which were used along with the equivalent data from Moe (1969) and Koenig (1993) to develop age-specific ogives (see Figure 4). The product of the fitted sex-ratio (proportion female), proportion active and batch fecundity estimates was taken as a measure of the per capita number of eggs produced by each age class (spawning frequency did not vary demonstrably with age).

Two per capita fecundity curves, one using the batch fecundity curve based on age FECvsAGE) and the other using the age-converted batch fecundity curve based on length (FECvsTL), are compared with the gonad weight curve (GWTvsTL) developed by Schirripa et al. (1999) in Table 5 and Figure 5. The FECvsTL and GWTvsTL curves are fairly similar, largely because the relative relationship between batch fecundity and length is similar to that between gonad weight and length (Figure 3). The main differences between the FECvsTL and GWTvsTL curves are the former's use of activity at age information (the latter assumes that all age classes older than 2 are equally active) and its prediction of relatively more production from very young animals. The FECvsAGE curve is fundamentally different from the other two inasmuch as it predicts that per capita fecundity continues to increase with age despite the decreasing fraction of females. This is because the fitted batch fecundity at age curve increases much faster than the age-converted batch fecundity at length curve as discussed above.

Additional analyses related to per capita fecundity were conducted by the RFSAP at the September meeting and are documented in the addendum.

## REMOVALS

## Yield, harvest and catch

Catch is generally used to refer to the number of fish caught and often includes live releases. Harvest is used to refer to the number of fish killed in the fishing process and may include fish landed, discarded dead at sea and fish used for bait; this is particularly used for the AB1 estimates from MRFSS (Marine Recreational Fisheries Statistics Survey, A=observed kill, B1=unobserved kill). Yield is used to refer to landings in weight for the commercial fishery and the headboat fishery; it is also used to refer to weight of the harvest (MRFSS). Yields are reported in gutted weight.

## Commercial Yield

Commercial yields were tabulated from the Accumulated Landings System (ALS) data base maintained at the Southeast Fisheries Science Center (SEFSC). Yields in the ALS data base are generally recorded in whole units. For groupers a conversion factor of 1.18 was used to convert gutted weight to whole weight in the ALS (Goodyear and Schirripa 1993); that same conversion factor was used to
reconvert to gutted weight for use in this assessment.
The Accumulated Landings System consists of subsets with different degrees of resolution. The primary data base consists of dealer reports of landed yield by year, month and species; for some states that information also includes gear and fishing area (water body), but not for Florida in 1986-1996 nor Louisiana and Texas since the early 1990's. Information on grouper landings by species is available only from 1986 and later; prior to that time only unclassified groupers were recorded; since 1986 unclassified grouper have been recorded, but the quantity has declined substantially. A portion of the unclassified groupers were assumed to be red grouper; that proportion was calculated by dividing the west Florida landing of red grouper by the west Florida landings of all identified groupers (excluding goliath and warsaw) as had Schirripa et al. (1999).

All reported commercial landings of red grouper from west Florida through Texas were tabulated by year as done in the previous assessment. Landings recorded from Florida inland counties were not included. Reported red grouper, the proportion of unclassified grouper assumed to be red grouper and the total calculated red grouper (used for assessment) are presented in Table 6. The total calculated commercial landings are shown by state in Table 7.

Yield for each of two gear categories was used in the program used to create derived age composition and in the assessment; those categories were (1) longline and (2) handline combined with other gears. For 1986 through 1989 the annual proportion of the total commercial landings by longline was assumed from Schirripa et al. (1999). For 1990-2001 the annual proportions by gear category were calculated from reef fish log books; the calculated landings by the two gear categories are presented in Table 8. The 1990-1997 proportions by gear calculated for this assessment were moderately different from the proportions calculated from the inputs by gear to the 1999 assessment (Figure 6).

Additionally the number of commercially caught fish by gear and region for 1992-2001 was used in the assessment runs which included the sampled age composition. As with the yield by gear, the distribution of yield by gear and region was calculated from log books. Three gear categories (longline, trap, and handline combined with others) and three regions (southwest FL -shrimp grids 1-4, central west FL - grids 5-6 and northwest FL and west - grids 7+) were used. Average weights for converting yield to number were from TIP (trip interview program) data by the same gear categories and regions; if there were less than 25 observed sizes for calculating a mean weight by year, gear group and region, then a mean for that year and gear group was used, and if there were insufficient observations for that strata then an overall mean (all years, gears and regions was used).

## Recreational Catch and Yield

Recreational catches (harvests and releases) were tabulated from the MRFSS, the SEFSC Headboat Survey and the Texas Parks and Wildlife (TPWD) data sets. MRFSS data was available through 2001, headboat catch estimates were available through 1999 and TPWD estimates were available through 2000. For MRFSS both harvests (A+B1) and live releases (B2) were tabulated; estimated weight of the harvest was calculated from mean weights derived from observed weights and weights predicted from lengths if weight was not recorded. For the headboat data the numbers of fish landed and the associated yields were tabulated from the Headboat Survey estimates. For the headboat survey the average catch and yield from 1995-1999 was used to estimate the 2000 and 2001 values. Red grouper were not recorded in the

TPWD data set during 1986-2000, though small numbers were recorded as landed in the headboat survey.

The total recreational harvests by state and by mode are presented in Tables 9 and 10. Schirripa et al. (1999) used total annual harvests by mode; to provide an indication of the variability in estimated harvests, coefficients of variation about the MRFSS estimates for west Florida harvests by year and mode are provided in Table 11.

The estimated weight of the recreational harvest is presented in Addendum-Table 1.

Estimates of live releases are also provided by the MRFSS and these are presented by state and mode in Tables 12 and 13. The ageing approach used by Schirripa et al. (1999) incorporated a method of estimating numbers released dead by all parts of the recreational fishery which used annual ratios of live releases to total catch (harvests plus releases) estimated from the MRFSS data. The annual proportions of the total catch which was released is presented in Table 14; especially in 1986-1988 those proportions differed from the proportions used for the 1999 assessment (Figure 7).

## DERIVED AGE COMPOSITION OF REMOVALS

The second type of age composition information used in the assessments, derived age composition, was calculated primarily from the observed length samples, the annual probabilistic age composition derived from the length samples using a growth curve (Goodyear 1997) and the annual harvest or yield. No independent abundance information was incorporated in the procedure (the independent index was assumed constant for all year classes), mortality rates (M and F) were constant and the ageing procedure was not iterated (Schirripa et al. 1999).

For each sampled length $(\mathrm{L})$ the probability of age $\left(\mathrm{p}_{\mathrm{a}, \mathrm{L}}\right)$ is assumed to be from a normal distribution with mean $\bar{L}_{a}$ [computed from the von Bertalanfy growth equation using an adjustment to $t_{0}$ in that curve to shift the curve from the marking time (assumed 1 June) to January 1 (Table 1) and taking into account the date when the sample was taken], and the standard deviation of length at age, $\mathrm{s}_{\mathrm{a}}$ (derived from the product of a constant coefficient of variation of length at age ( 0.16 ), and the mean length at each age). The value of $\mathrm{p}_{\mathrm{a}, \mathrm{L}}$ is obtained from the difference in cumulative probabilities of age for lengths slightly above and below the sampled length:

$$
\mathrm{p}_{\mathrm{a}, \mathrm{~L}}=\frac{\left[\operatorname{prob}\left(L+0.05 s_{a}-\bar{L}_{a}\right)-\operatorname{prob}\left(L-0.05 s_{a}-\bar{L}_{a}\right)\right] W_{y a}}{\sum_{a=j}^{20}\left[\operatorname{prob}\left(L+0.05 s_{a}-\bar{L}_{a}\right)-\operatorname{prob}\left(L-0.05 s_{a}-\bar{L}_{a}\right)\right] W_{y a}}
$$

where age j is the youngest age considered vulnerable to the fishery (age 1 for recreational fisheries and age 2 for commercial fisheries) and $\mathrm{W}_{\mathrm{ya}}$ is a weighting factor derived from an annual recruitment index (I) and fishing $\left(\mathrm{F}_{\mathrm{ya}}\right)$ and the natural mortality $(\mathrm{M})$ rates (the index values and mortality rates were all assumed constant for this assessment):

$$
\mathrm{W}_{\mathrm{ya}}=I_{y-a} \exp -\left(F_{y a}+M\right)
$$

The frequency at age was then calculated from all the length samples:

$$
\mathrm{f}_{\mathrm{yak}}=\sum_{a=j}^{20} \sum_{L} p_{a: L} m_{L k}
$$

where $m_{L k}$ is the number of fish measured at each length, for gear (or mode), $k$, and $L$ represents each length in the samples.

The landed catch at age for the commercial fishery is calculated as:

$$
\mathrm{C}_{\mathrm{yak}}=f_{y a k} \frac{Y_{y k}}{w_{y k}}
$$

where $\mathrm{Y}_{\mathrm{yk}}$ is the landed yield by year and gear group (longline and handline plus other) and $\mathrm{w}_{\mathrm{yk}}$ is the total weight of the of the measured fish (from the observed or estimated weight of a measured fish) by year and gear group. If less than 25 lengths were available to compute $f_{y a k}$, the overall age composition for that year and sector (commercial or recreational) was used.

For the recreational fishery the harvested (A+B1 for MRFSS and landings for the Texas and headboat data sets) catch at age is calculated as:

$$
\mathrm{C}_{\mathrm{yak}}=\frac{f_{y a k}}{\sum_{a=0}^{20} f_{y a k}} H_{y k}
$$

where $\mathrm{H}_{\mathrm{yk}}$ refers to the number of fish harvested by year and mode.

The number of fish discarded at age was calculated in the ageing procedure as was done for the previous assessment. The average probability ( g ) that the minimum size (msize) in a year ( y ) is greater than the mean length $(\mathrm{L})$ for a given age on a given day ( d ) is computed as

$$
\mathrm{g}_{\mathrm{ya}}=\frac{\sum_{d=1}^{n} \operatorname{prob}\left(\text { msize }_{y}>\bar{L}_{a d}\right)}{n}
$$

where n is the number of days over which the probabilities are computed and the mean length is computed from the growth equation.

Commercial discards

The number estimated to have been discarded dead from the commercial fishery, $\mathrm{D}_{\text {yak }}$, was initially modeled as:

$$
\mathrm{D}_{\mathrm{yak}}=G_{y a k}\left(\frac{1-g_{y a}}{g_{y a}}\right) d_{k}
$$

where $d_{k}$ is the discard mortality rate for gear or mode, $k$, ( 0.33 for handline combined with other gears, 0.33 or 0.90 for longline, and 0.1 for all recreational modes) and $G_{\text {yak }}$ is landed catch of fish greater than or equal to the minimum size in year $y$, at age a by gear k. For all commercial discard estimates the youngest age considered vulnerable to fishing was age 2 .

For commercial longline catches, the modeled number of dead discards, $\mathrm{D}_{\text {yag }}$, was further modified because at-sea observers recorded a release fraction, r, in 1994 which was lower than that estimated through modeling. The observed release fraction used in the previous assessment was 0.469 (1446/3080) and it was assumed for all years 1990 and later when the minimum size for catches in federal waters was in place. The revised longline discards, D', was calculated as:

$$
\mathrm{D}_{\mathrm{yak}}^{\prime}=D_{y a k}\left(\frac{r}{\frac{D_{y a k}}{\left(G_{y k}+D_{y a k}\right)}}\right) d_{k}
$$

Recreational discards

For the recreational fishery an estimate of the discards (B2) was available from MRFSS estimates. That information was used in estimating the total dead discards. The total number of recreational fish released alive was then calculated as:

$$
\mathrm{R}_{\mathrm{y}}=H_{y}\left(\frac{r_{y}}{1-r_{y}}\right)
$$

where $\mathrm{H}_{\mathrm{y}}$ is number of fish harvested (MRFSS A+B1, plus headboat landings, plus Texas landings) and $\mathrm{r}_{\mathrm{y}}$ is the proportion released alive from MRFSS (B2/(A+B1+B2)) (Table 14).

The average probability that the minimum size is smaller than the mean length at age $\left(\mathrm{s}_{\mathrm{ya}}\right)$ is the reciprocal of $g_{y a}$ given above. It is then weighted by the an input selectivity at age ( $\mathrm{S}_{\mathrm{ya}}$, with values of 0.67 , 0.92 and 0.99 for ages 1-3 years, then 1.0 for the ages 4 and older) and re-normalized

$$
\mathrm{s}_{\text {yа }}=\frac{S_{y a} S_{a}}{\sum_{a=1}^{20} S_{y a} S_{a}}
$$

Total recreational discards by year and age were then computed as

$$
\mathrm{D}_{\mathrm{yak}}=R_{y} s_{y a}^{\prime} d_{k}
$$

Two sets of derived age composition were created: one for 1986-1997 to examine the effects of differences in the proportions of recreational fish released (Figure 7) and the other with information for 1986-2001. The alternative, derived age composition for the $33 \%$ mortality assumption for longline discards is shown in Table 15 for 1986-1997 and Table 16 for 1986-2001.

The derived age composition for longline is shown in Figure 8 and can be compared to the sampled longline age composition shown in Figure 1. The modal age in the derived age composition ranged from 3 to 6 and was 5 in 1991through 2001 which was about 2 years younger than the modal age from the sampled age composition. An alternative figure with these age compositions compared was requested by the RFSAP at the meeting in September 2002 is shown in Addendum Figure A1.

## INDICES OF ABUNDANCE

Several indices of abundance were developed based on observations of catch-per-unit-effort (CPUE) and limited fishery-independent surveys. The data were standardized using essentially the same methods as outlined by Schirripa et al (1999). The results are summarized in Table 17 and Figure 9.

## Commercial operating units.

Schirripa et al. (1999) developed a CPUE index based on the total commercial landings and estimates of the U.S. fleet effort derived from the NMFS operating units file. Due to time limitations, no attempt was made to update this index to 2001. However, this should be of little consequence because the recent time period is amply covered by the logbook indices discussed below.

## Reeffish Logbooks.

Data available from the Reeffish Logbook Program were used to develop standardized CPUE series for commercial fish traps, handlines and bottom longlines from August 1990 to December 2001. The Reeffish Logbook Program was initiated in 1990. At that time the program required all vessels holding reeffish permits in Alabama, Mississippi, Louisiana and Texas, as well as all trap fishermen in Florida, to report on each fishing trip made. For other Florida permitted vessels, only a randomly selected sub-sample ( $20 \%$ ) were required to report until 1993, when mandatory reporting for all Florida vessels began. Only landings (in pounds) were recorded; releases were not and discards were generally not recorded. Thus, these CPUE series are really harvest per unit effort (HPUE) measures and are affected to varying degrees by size regulations.

The standardization procedures were intended as an update to the analyses conducted for the last assessment and therefore followed the methods detailed in Schirripa et al. (1999). Only trips landing red grouper were used. Catch was defined as total pounds landed in whole weight. Effort was defined as hook*hours for handline and days-at-sea for bottom longline and trap gears (the description by Schirripa et al. 1999 mistakenly indicates otherwise). Standardized indices were developed using generalized log-linear models where catch rates were modeled as a function of the factors YEAR, MONTH, and GRID (fishing area). (Much of the data were inadvertently excluded during the trap CPUE analysis conducted for the previous assessment. This has been corrected for the current, updated analysis).

## MRFSS recreational.

The catches of recreational vessels fishing in the Gulf of Mexico and Atlantic Ocean are monitored by the National Marine Fisheries Service's Marine Recreational Fisheries Statistical Survey (MRFSS). Anglers intercepted at fishing access sites are interviewed to determine, among other things, the number of each species that were landed and observed by a sampler (type A), the number of fish killed not seen by an sampler (B1, including dead discards) and the number released alive (B2). They are also asked the number of hours they spent fishing. Thus, it is possible to construct catch per unit effort (CPUE) indices of abundance by dividing the total catch $(A+B 1+B 2)$ by the number of angler-hours (number of anglers in the party times the number of hours fished). One may also compute harvest per unit effort (HPUE), using the type A and B1 catches, but this measure is directly affected by size and bag limit regulations.

A GLM standardization procedure was adopted following the methods used for the previous assessment, where the CPUE of private vessels was expressed as a linear function of the factors YEAR COUNTY (the description given by Schirripa et al. mistakenly asserts that both private and charter vessels were used and attributed significant effect to year, month, area and mode).

## Fishery independent.

Fishery independent trap and video surveys were conducted as part of the Southeast Area Monitoring Program (SEAMAP) during the months of June, July and August from 1992-1997 (data courtesy of C. T. Gledhill). An abbreviated video survey was conducted in 2001, and a complete survey in 2002, however the data are still being quality assured. The trap survey was not continued in 2001 or 2002 and therefore not updated. Accordingly, no updates were possible and the indices from the previous assessment were retained.

## Tag-recapture index.

Estimates of the loss rate of red grouper tagged by the Mote Tagging program were derived using the methods discussed in Legault et al. (1999). As was done in the previous assessment, these estimates were converted to a relative abundance index by solving the catch equation for numbers of fish (where the catches were the MRFSS estimates of the number of fish released alive and the natural mortality rate was assumed to be $0.2 \mathrm{yr}^{-1}$ ). This conversion approach makes the implicit assumptions that tag shedding is negligible and that the tagged fish mix randomly with the untagged population, neither of which appear to be true of red grouper. However, we found the relative abundance index to be relatively insensitive to the assumed levels of tag-shedding and incomplete mixing (Figure 10). No doubt the relative index would be
more sensitive to systematic changes in the level of mixing or tag shedding, but this is akin to asserting that CPUE indices are sensitive to systematic changes in factors not included in the standardization process. Accordingly, while further investigation is warranted, we can offer little basis for rejecting it over some of the CPUE indices used in this assessment. (In principle, a more appropriate use of the loss rate estimates would be to treat them as an index of relative mortality rate in the manner of Porch 2001, but there was insufficient time to incorporate this approach into the ASAP model for this assessment.)

## POSSIBLE CURRENT CONDITION OF THE STOCK

The condition of the Gulf of Mexico red grouper stock was evaluated using the ASAP model (Legault and Restrepo, 1998) applied as described in the previous assessment (Schirripa et al., 1999), but with modifications introduced during the 2000 meeting of the RFSAP (see Table A2 in the addendum). The analyses were conducted in three phases. In the first phase, the two runs judged by the RFSAP (2000) to provide the best scientific advice were re-run using the revised derived catch at age, revised discard estimates and trap CPUE series discussed earlier. The purpose of these runs was to determine if substantially different advice would have been given had these revisions been made in 2000 . The second phase of the analysis includes catches estimated using the revised commercial yield ratio (Figure 6) and updates information to the year 2001. The third phase expands on the phase 2 model by including age composition data. The effects of a change in the fecundity-at-age vector and various alternative treatments of the CPUE series are also examined.

## Phase 1: Revised catch, discard and trap CPUE information

The 2000 RFSAP based its advice largely on the results from two ASAP models (one assuming a steepness value of 0.7 for the stock-recruitment relationship and the other assuming a steepness value of 0.8 ) with the following specifications:

- "short" time series (1986-1997) of catch, discard and CPUE estimates.
- discards estimated via the "probabilistic" method tuned so that the predicted discard rate in 1994 matched the observed discard rate from 1994 observed program (NMFS 1995).
- $33 \%$ release mortality rate for commerical fisheries

These runs were repeated exactly as specified in 2000 except using the revisions to the derived catch at age, discard, and commercial trap CPUE discussed earlier. The effect of assuming $90 \%$ release mortality of longline caught fish was also examined (a total of four runs- two levels of steepness and two levels of longline release mortality).

The results are summarized in Table 18 and Figure 11. The estimates of MSY remain near 3000 mt . The estimates of $\mathrm{F}_{\text {current }}$ and $\mathrm{F}_{\text {MSY }}$ have both increased relative to the previous assessment, whereas the estimates of $\mathrm{SS}_{1997}$ and $\mathrm{SS}_{\mathrm{MSY}}$ have decreased. The net effect is that the population is estimated to be overfished to a lesser extent than formerly indicated. Nevertheless, the estimated date of recovery to $\mathrm{SS}_{\text {MSY }}$ under no fishing $\left(\mathrm{t}_{\text {min }}\right)$ is the same (2002-2004).

As noted in the previous assessment, the estimates of stock status are somewhat less optimistic under the higher ( $90 \%$ ) release mortality assumption.

## Phase 2: Data updated to 2001

The runs in this phase were conducted as described for phase 1 with the addition of using the derived catch at age using the revised commercial yield ratio, and catch and cpue updated to 2001. The estimated trends in SS and $\mathrm{SS}_{\text {/SS }}^{\text {MSY }}$ are contrasted with those from phase 1 (data to 1997) and the projections results examined by RFSAP (2000) in Figure 12. The estimates of SS are nearly identical for all runs, however the projections from the previous assessment predicted a downturn after 1999 in contrast to the upswing estimated with the latest data. The difference is largely attributable to estimates of recruitment that exceeded the expectation of the Beverton and Holt spawner-recruit relationship used in the projections (Figure 13). The estimated trends in $\mathrm{SS} / \mathrm{SS}_{\mathrm{MSY}}$ are more sensitive to the various model treatments than are the trends in SS (Figure 12).

Applications of ASAP to the updated (1986-2001) catch and CPUE information led to somewhat less optimistic SS/SS $_{\text {MSY }}$ ratios for 1997 than did applications to the shorter time series (1986-1997). The status in 2001 is estimated to have improved relative to the status in 1997 for all models, but remains overfished (Figure 14, Table 19). Higher levels of steepness and lower levels of longline release mortality favor somewhat more optimistic estimates of stock status. The estimates of MSY are 10-20\% higher than estimated from the 1986-1997 data ( $3300-3600$ MT). The estimates of $\mathrm{F}_{\text {current }}$ and $\mathrm{F}_{\text {MSY }}$ are similar to the previous assessment, whereas the estimates of $\mathrm{SS}_{2001}$ and $\mathrm{SS}_{\text {MSY }}$ have both increased. The catch expected to permit the population to recover to $\mathrm{SS}_{\text {MSY }}$ by 2012 is between 2600 and 3100 MT (about $20 \%$ less than MSY).

## Phase 3: Addition of sampled age composition data and sensitivity analyses

The runs in phase 3 were conducted exactly as described for phase 2 except for the additional use of age composition data derived from otoliths taken from a sample of the catch (see Lombardi-Carlson et al., 2002). Also sensitivity runs were made to examine the effects of a new fecundity at age vector (see section on reproduction), down-weighting the indices of abundance to account for variations in the degree to which they reflect population abundance $\left(\mathrm{CV}_{\text {index }}=\operatorname{SQRT}\left(0.2 * 0.2+\mathrm{CV}_{\text {GLM }} * \mathrm{CV}_{\text {GLM }}\right)\right.$, adjusting the CPUE indices downwards by $25 \%$ to account for increases in efficiency associated with increased use of GPS plotters, and splitting the historical CPUE series into pre- and post-size limit (1990) eras.

## Base run

Estimates were made using the two different levels of steepness ( 0.7 and 0.8 ) and longline release mortality ( $33 \%$ and $90 \%$ ). The estimated levels of spawning stock (in collective gonad weight) are similar for all five models (Figure 15), as are the abundance and fishing mortality rate estimates (Figure 16, one case shown). However, the estimates of $\mathrm{SS}_{\text {MSY }}$ are sensitive to the level of steepness assumed- generally leading to more optimistic perceptions of stock status (no longer overfished) as steepness increases. As was true of the previous assessment, there is insufficient contrast in the SS and recruitment series to allow steepness to be reliably estimated, although the model does favor the higher steepness (i.e., the objective function attains a lower value with the higher steepness).

The phase 3 base run results are contrasted with the corresponding phase 2 results in Figure 17.

The use of the sampled age composition data causes the estimates of SS and $\mathrm{SS} / \mathrm{SS}_{\text {MSY }}$ to increase faster than when such data are excluded, resulting in a somewhat more optimistic picture of stock status.

## Sensitivity runs

The results of the sensitivity runs are contrasted with those of the base runs in Figure 18 and Table 20. The use of the new fecundity at age vector has the most pronounced effect, moving the estimated stock status from overfished to merely below $\mathrm{SS}_{\text {MSY }}$ (but above MSST=0.8 $\mathrm{SS}_{\text {MSY }}$ ) with a steepness of 0.7 and above $\mathrm{SS}_{\text {MSY }}$ with a steepness of 8 . However, adjusting the CPUE series downwards to account for increased use of GPS plotters moved the estimated status back to overfished, even with a steepness of 0.8. Downweighting the indices of abundance and splitting the historical CPUE index had relatively little effect.

## Recommendations

A fuller evaluation of the inclusion of longer catch and catch rate series may provide additional contrast in the data to provide a stronger basis for stock status evaluations.

Additional research on index standardization for red grouper and application of indices of abundance in the assessment model is warranted.

Fishery independent indices of abundance with amble spatial and temporal coverage are needed

Research into improved models of fecundity at age and methods to incorporate uncertainty in these estimates is warranted given the sensitivity of the assessment results to these estimates.

Information is needed on fish released at sea including amounts for most sectors, size composition from all sectors and their fate.

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Table 1. Size conversion equations used to convert red grouper sizes to total length, gutted weight and age.

| to | from | equation | source |
| :---: | :---: | :---: | :---: |
| TL (in) | FL (in) | TL $=-0.134+\mathrm{FL}$ * 1.052 | Goodyear and Schirripa (1993) |
| TL (in) | SL (in) | TL $=0.819+\mathrm{FL}$ * 1.185 | Goodyear and Schirripa (1993) |
| gutted wt (lb) | TL (in) | gwt $=0.00112$ * FL ** 2.76 | Goodyear and Schirripa (1993) |
| gutted wt (lb) | FL (in) | gwt $=0.000506$ * FL ** 3.04 | Goodyear and Schirripa (1993) |
| whole wt (lb) | TL (in) | wwt $=0.000399$ * TL ** 3.1 | Goodyear and Schirripa (1993) |
| gutted wt | whole wt | $\mathrm{gwt}=\mathrm{wwt} / 1.048$ | Goodyear and Schirripa (1993) |
| age (years) | TL (in) | $\mathrm{TL}=31.81$ * (1-exp (-0.210 * (age + 0.30) ) | Goodyear (1994) |
| age (years) | TL (in) | $\mathrm{TL}=31.81$ * (1-exp (-0.210 * (age + 0.30 + 0.416) ) | Goodyear (1994) with calendar year adjustment ${ }^{1}$ |

${ }^{1}$ The calendar year adjustment is used to shift the curve from the apparent marking date of 1 June to 1 January so that in the assessment all fish in a cohort would be assigned to the same year class; thus the assessment age of 1 year corresponds to an approximate true age of 0.6 to 1.59 years.

Table 2a. Length samples used for calculating age composition of the recreational landings by year and mode. Headboat mode.Lengths are in inches.

| headboat m | ode |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |  |
| 10 | 14 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 11 | 11 | 0 | 0 | 10 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 12 | 11 | 13 | 5 | 36 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 |  |
| 13 | 35 | 37 | 21 | 117 | 7 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |  |
| 14 | 65 | 83 | 56 | 117 | 12 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |  |
| 15 | 76 | 78 | 64 | 89 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 1 | 0 | 0 |  |
| 16 | 66 | 84 | 36 | 53 | 7 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 17 | 55 | 68 | 37 | 71 | 13 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 18 | 59 | 55 | 41 | 52 | 12 | 1 | 1 | 2 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |  |
| 19 | 46 | 72 | 38 | 76 | 22 | 4 | 3 | 5 | 5 | 3 | 12 | 16 | 26 | 14 | 15 |  |
| 20 | 39 | 65 | 23 | 37 | 25 | 25 | 14 | 23 | 33 | 24 | 50 | 58 | 50 | 45 | 38 |  |
| 21 | 44 | 67 | 42 | 58 | 73 | 24 | 26 | 20 | 32 | 40 | 47 | 40 | 48 | 67 | 45 |  |
| 22 | 18 | 31 | 21 | 23 | 31 | 11 | 9 | 19 | 18 | 20 | 42 | 29 | 42 | 28 | 25 |  |
| 23 | 8 | 19 | 12 | 13 | 19 | 12 | 5 | 4 | 5 | २2 | 24 | 23 | 14 | 16 | 15 |  |
| 24 | 22 | 26 | 25 | 34 | 42 | 10 | 8 | 10 | 10 | 11 | 15 | 38 | 35 | 40 | 24 |  |
| 25 | 20 | 17 | 12 | 18 | 20 | 9 | 3 | 9 | 8 | 8 | 23 | 13 | 8 | 12 | 8 |  |
| 26 | 13 | 15 | 13 | 23 | 19 | 9 | 6 | 6 | 7 | 8 | 13 | 6 | 4 | 15 | 8 |  |
| 27 | 10 | 12 | 8 | 25 | 17 | 6 | 5 | 4 | 1 | 7 | 11 | 6 | 14 | 8 | 4 |  |
| 28 | 8 | 9 | 5 | 4 | 8 | 3 | 4 | 3 | 7 | 1 | 9 | 7 | 8 | 2 | 1 |  |
| 29 | 5 | 4 | 7 | 10 | 6 | 4 | 6 | 0 | 1 | 3 | 2 | 10 | 2 | 4 | 0 |  |
| 30 | 2 | 1 | 3 | 3 | 6 | 2 | 2 | 4 | 2 | 0 | 6 | 3 | 5 | 4 | 0 |  |
| 31 | 5 | 4 | 1 | 8 | 4 | 2 | 5 | 0 | 0 | 0 | 0 | 8 | 0 | 3 | 0 |  |
| 32 | 2 | 2 | 1 | 3 | 2 | 7 | 1 | 5 | 0 | 1 | 1 | 0 | 0 | 0 | 1 |  |
| 33 | 1 | 2 | 0 | 1 | 3 | 9 | 0 | 0 | 0 | 1 | 0 | 4 | 0 | 0 | 0 |  |
| 34 | 0 | 0 | 2 | 0 | 1 | 4 | 2 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 35 | 1 | 1 | 0 | 1 | 0 | 2 | 0 | 1 | 4 | 0 | 0 | 0 | 0 | 1 | 0 |  |
| 36 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 37 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 39 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 42 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 44 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 45 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| total | 639 | 765 | 474 | 882 | 359 | 145 | 100 | 121 | 135 | 149 | 256 | 270 | 259 | 259 | 185 |  |

Table 2 b . Length samples used for calculating age composition of the recreational landings by year and mode. Charter mode. Lengths are in inches.

| charter mod |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 12 | 1 | 0 | 0 | 1 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 3 | 6 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 14 | 2 | 2 | 3 | 1 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 |
| 15 | 4 | 3 | 3 | 5 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 4 | 0 |
| 16 | 4 | 1 | 9 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 0 | 0 | 0 |
| 17 | 10 | 3 | 3 | 12 | 0 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 |
| 18 | 4 | 1 | 4 | 5 | 0 | 14 | 5 | 0 | 0 | 0 | 4 | 0 | 14 | 0 | 2 | 0 |
| 19 | 4 | 1 | 6 | 6 | 1 | 9 | 5 | 2 | 18 | 3 | 13 | 4 | 13 | 11 | 24 | 24 |
| 20 | 2 | 3 | 5 | 10 | 3 | 20 | 19 | 13 | 64 | 4 | 51 | 25 | 43 | 40 | 97 | 67 |
| 21 | 2 | 2 | 1 | 2 | 1 | 11 | 31 | 7 | 57 | 20 | 59 | 46 | 52 | 55 | 72 | 79 |
| 22 | 1 | 4 | 5 | 3 | 3 | 14 | 19 | 11 | 32 | 9 | 62 | 38 | 36 | 54 | 37 | 79 |
| 23 | 1 | 0 | 2 | 3 | 2 | 4 | 18 | 13 | 33 | 13 | 53 | 40 | 55 | 29 | 37 | 54 |
| 24 | 1 | 1 | 1 | 0 | 1 | 4 | 17 | 22 | 16 | 3 | 36 | 31 | 54 | 34 | 22 | 31 |
| 25 | 0 | 0 | 2 | 0 | 2 | 8 | 6 | 13 | 7 | 8 | 19 | 22 | 57 | 35 | 14 | 21 |
| 26 | 0 | 1 | 3 | 0 | 0 | 0 | 8 | 12 | 17 | 6 | 35 | 32 | 27 | 22 | 21 | 15 |
| 27 | 0 | 1 | 0 | 1 | 0 | 8 | 14 | 20 | 32 | 7 | 14 | 13 | 24 | 17 | 15 | 11 |
| 28 | 0 | 0 | 1 | 1 | 0 | 4 | 18 | 48 | 14 | 6 | 12 | 30 | 9 | 12 | 25 | 10 |
| 29 | 0 | 1 | 1 | 0 | 0 | 4 | 18 | 32 | 12 | 6 | 14 | 4 | 25 | 7 | 9 | 6 |
| 30 | 0 | 1 | 0 | 0 | 0 | 0 | 24 | 20 | 8 | 0 | 9 | 9 | 8 | 2 | 6 | 4 |
| 31 | 0 | 0 | 0 | 0 | 0 | 4 | 5 | 20 | 8 | 1 | 8 | 12 | 9 | 7 | 4 | 4 |
| 32 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 16 | 8 | 0 | 0 | 0 | 4 | 0 | 1 | 5 |
| 33 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 | 12 | 0 | 8 | 8 | 0 | 4 | 2 | 0 |
| 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 4 | 1 | 0 | 0 | 4 | 4 | 2 | 1 |
| 35 | 0 | 0 | 0 | 0 | 0 | 12 | 0 | 0 | 0 | 0 | 0 | 4 | 4 | 0 | 0 | 0 |
| 36 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 |
| 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 44 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 45 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| total | 39 | 32 | 49 | 59 | 13 | 152 | 221 | 253 | 342 | 87 | 401 | 318 | 458 | 334 | 395 | 411 |

Table 2c. Length samples used for calculating age composition of the recreational landings by year and mode. Private mode. Lengths are in inches.

| private mode |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10 | 0 | 4 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 0 | 6 | 2 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12 | 4 | 6 | 10 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 1 | 9 | 12 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 14 | 4 | 16 | 20 | 6 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 15 | 3 | 9 | 17 | 19 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 16 | 6 | 10 | 31 | 30 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 3 | 3 | 0 |
| 17 | 2 | 10 | 17 | 12 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 0 |
| 18 | 3 | 5 | 7 | 15 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 2 | 2 | 0 |
| 19 | 2 | 11 | 11 | 7 | 4 | 3 | 9 | 12 | 9 | 8 | 3 | 0 | 4 | 2 | 4 | 2 |
| 20 | 2 | 11 | 11 | 18 | 3 | 13 | 41 | 24 | 16 | 17 | 4 | 1 | 14 | 13 | 16 | 22 |
| 21 | 4 | 3 | 15 | 12 | 7 | 22 | 44 | 25 | 23 | 14 | 6 | 9 | 22 | 26 | 31 | 21 |
| 22 | 2 | 6 | 11 | 2 | 5 | 16 | 36 | 21 | 16 | 17 | 5 | 16 | 15 | 21 | 17 | 22 |
| 23 | 1 | 3 | 8 | 2 | 3 | 13 | 23 | 17 | 21 | 17 | 13 | 7 | 7 | 16 | 15 | 15 |
| 24 | 0 | 1 | 8 | 7 | 6 | 7 | 28 | 9 | 16 | 15 | 7 | 6 | 4 | 18 | 10 | 10 |
| 25 | 0 | 1 | 5 | 3 | 1 | 10 | 21 | 11 | 5 | 18 | 3 | 4 | 10 | 15 | 9 | 8 |
| 26 | 0 | 2 | 3 | 2 | 6 | 3 | 14 | 6 | 5 | 9 | 4 | 3 | 7 | 5 | 9 | 7 |
| 27 | 0 | 1 | 0 | 0 | 3 | 2 | 8 | 2 | 4 | 6 | 2 | 0 | 5 | 6 | 6 | 3 |
| 28 | 0 | 1 | 0 | 0 | 0 | 0 | 3 | 1 | 2 | 5 | 4 | 0 | 3 | 6 | 4 | 4 |
| 29 | 0 | 0 | 1 | 0 | 0 | 1 | 1 | 0 | 2 | 1 | 0 | 0 | 2 | 4 | 6 | 3 |
| 30 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 3 | 3 | 1 |
| 31 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 2 | 1 | 0 | 1 | 0 | 5 | 1 |
| 32 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 1 |
| 33 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
| 34 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 44 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 45 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| total | 34 | 118 | 189 | 155 | 43 | 92 | 232 | 134 | 120 | 129 | 52 | 49 | 97 | 142 | 146 | 121 |

Table 3a. Length samples used for calculating age composition of the commercial landings by year and gear. Longline. Lengths are in inches.

|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| longline |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 2 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 3 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 4 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 5 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 6 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 7 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 8 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 9 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 10 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 11 | 6 | 4 | 15 | - | - | - | - | - | - | - | - | - | 2 | - | - | - |
| 12 | 70 | 28 | 63 | 6 | - | - | - | - | - | - | - | - | - | - | - | - |
| 13 | 339 | 149 | 157 | 17 | 1 | - | - | - | - | - | - | - | - | - | - | - |
| 14 | 420 | 186 | 109 | 31 | - | - | - | - | - | - | - | 1 | - | - | - | - |
| 15 | 383 | 193 | 99 | 34 | - | - | 1 | - | - | - | - | 1 | - | 2 | - | - |
| 16 | 269 | 172 | 96 | 39 | - | - | - | - | - | - | - | - | 2 | 2 | - | - |
| 17 | 331 | 166 | 112 | 42 | 4 | - | 3 | 1 | 1 | 1 | 3 | 3 | 6 | 2 | 1 | 1 |
| 18 | 386 | 178 | 103 | 35 | 12 | 9 | 4 | 5 | 2 | 29 | 34 | 34 | 31 | 68 | 24 | 21 |
| 19 | 420 | 175 | 113 | 68 | 86 | 96 | 103 | 124 | 138 | 288 | 333 | 407 | 722 | 1,151 | 634 | 437 |
| 20 | 366 | 184 | 94 | 101 | 444 | 676 | 611 | 663 | 713 | 1,322 | 1,126 | 1,627 | 2,733 | 4,775 | 2,800 | 2,158 |
| 21 | 278 | 142 | 76 | 86 | 751 | 1,156 | 947 | 889 | 946 | 1,507 | 1,395 | 2,049 | 3,089 | 5,580 | 3,510 | 2,855 |
| 22 | 364 | 158 | 44 | 159 | 1,176 | 1,903 | 1,129 | 1,029 | 880 | 1,212 | 1,164 | 1,723 | 2,994 | 5,139 | 3,123 | 2,399 |
| 23 | 364 | 128 | 48 | 125 | 906 | 1,076 | 762 | 908 | 733 | 1,023 | 1,015 | 1,353 | 2,798 | 4,678 | 3,087 | 2,217 |
| 24 | 303 | 142 | 35 | 185 | 1,074 | 1,325 | 814 | 1,034 | 814 | 927 | 944 | 1,208 | 2,749 | 3,618 | 2,318 | 1,609 |
| 25 | 307 | 132 | 54 | 110 | 818 | 802 | 675 | 1,107 | 840 | 978 | 823 | 1,056 | 3,110 | 3,744 | 2,589 | 1,640 |
| 26 | 223 | 131 | 29 | 113 | 1,056 | 1,105 | 752 | 936 | 818 | 902 | 736 | 992 | 2,958 | 3,624 | 2,646 | 1,655 |
| 27 | 163 | 64 | 34 | 83 | 762 | 798 | 553 | 697 | 590 | 750 | 596 | 746 | 2,428 | 3,044 | 2,209 | 1,353 |
| 28 | 177 | 95 | 30 | 91 | 958 | 1,054 | 669 | 698 | 524 | 714 | 516 | 662 | 2,006 | 2,982 | 2,317 | 1,273 |
| 29 | 168 | 64 | 16 | 62 | 577 | 699 | 412 | 469 | 407 | 543 | 386 | 559 | 1,276 | 2,247 | 1,695 | 994 |
| 30 | 121 | 56 | 16 | 85 | 726 | 792 | 403 | 440 | 297 | 374 | 237 | 505 | 820 | 1,366 | 1,188 | 579 |
| 31 | 89 | 29 | 15 | 77 | 547 | 484 | 302 | 362 | 223 | 262 | 167 | 319 | 578 | 1,019 | 863 | 368 |
| 32 | 66 | 22 | 5 | 25 | 328 | 204 | 158 | 175 | 111 | 144 | 97 | 201 | 328 | 614 | 538 | 204 |
| 33 | 60 | 22 | 3 | 48 | 255 | 199 | 116 | 107 | 73 | 74 | 55 | 121 | 201 | 306 | 337 | 115 |
| 34 | 46 | 13 | 4 | 16 | 139 | 98 | 43 | 46 | 41 | 42 | 22 | 41 | 83 | 170 | 168 | 73 |
| 35 | 28 | 10 | 4 | 10 | 87 | 66 | 35 | 32 | 19 | 11 | 9 | 21 | 43 | 66 | 69 | 49 |
| 36 | 12 | 6 | 1 | 4 | 43 | 18 | 15 | 9 | 4 | 6 | 6 | 7 | 10 | 18 | 17 | 8 |
| 37 | 1 | 1 | - | 4 | 27 | 9 | 7 | 6 | 2 | 1 | 1 | 1 | 2 | 5 | 2 | 7 |
| 38 | 1 | 1 | - | - | 7 | 1 | 1 | - | - | 2 | - | - | - | - | - | 1 |
| 39 | - | - | - | 2 | 6 | - | - | - | - | - | - | 1 | 1 | - | - | - |
| 40 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 41 | - | - | - | - | 1 | 1 | - | - | - | - | - | - | - | - | - | - |
| 42 | - | 1 | - | - | 3 | - | - | - | - | - | - | - | - | - | - | - |
| 43 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 44 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 45 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 46 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 47 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 48 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 49 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 50 | - | - | - | - | - | - | - | - | - | - | - | - | - | 1 | - | - |
| total | 5,750 | 2,634 | 1,374 | 1,103 | 10,438 | 12,536 | 8,453 | 9,540 | 8,092 | 10,966 | 9,655 | 13,626 | 28,933 | 44,202 | 30,069 | 19,388 |

Table 3b. Length samples used for calculating age composition of the commercial landings by year and gear. Handline and other gears. Lengths are in inches

|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| handline+ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 2 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 3 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 4 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 5 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 6 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 7 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 8 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 1 |
| 9 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 10 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 11 | 1 | 2 | 2 | 2 | - | - | - | - | - | - | - | - | - | - | - | - |
| 12 | 104 | 71 | 122 | 70 | 9 | 2 | - | - | - | - | - | - | - | - | - | - |
| 13 | 213 | 236 | 218 | 140 | 19 | 7 | 7 | - | - | 1 | - | - | 2 | - | 5 | - |
| 14 | 165 | 189 | 129 | 144 | 10 | 11 | 9 | 1 | - | 2 | 2 | - | 1 | - | 20 | 1 |
| 15 | 139 | 214 | 119 | 85 | 8 | 15 | 14 | 1 | - | 3 | 3 | 1 | 1 | 1 | 24 | 6 |
| 16 | 100 | 164 | 93 | 102 | 15 | 9 | 3 | 1 | - | - | 13 | 3 | - | - | 11 | 7 |
| 17 | 95 | 137 | 71 | 82 | 9 | 17 | 6 | 3 | 2 | 7 | 11 | 10 | 4 | 2 | 9 | 9 |
| 18 | 123 | 144 | 74 | 77 | 11 | 24 | 6 | 9 | 16 | 24 | 51 | 31 | 16 | 6 | 28 | 15 |
| 19 | 124 | 122 | 57 | 52 | 30 | 69 | 70 | 51 | 139 | 169 | 168 | 129 | 152 | 242 | 387 | 348 |
| 20 | 97 | 108 | 38 | 41 | 96 | 322 | 369 | 267 | 398 | 449 | 587 | 618 | 530 | 968 | 1,227 | 1,398 |
| 21 | 97 | 94 | 45 | 36 | 79 | 297 | 382 | 280 | 450 | 439 | 561 | 549 | 599 | 1,034 | 1,280 | 1,613 |
| 22 | 80 | 79 | 42 | 38 | 116 | 325 | 369 | 302 | 361 | 404 | 480 | 513 | 492 | 1,060 | 1,060 | 1,458 |
| 23 | 58 | 76 | 30 | 30 | 112 | 256 | 283 | 260 | 292 | 343 | 362 | 405 | 430 | 992 | 1,021 | 1,221 |
| 24 | 77 | 88 | 30 | 24 | 98 | 295 | 321 | 323 | 288 | 352 | 311 | 330 | 401 | 763 | 839 | 845 |
| 25 | 69 | 93 | 27 | 21 | 81 | 224 | 281 | 343 | 321 | 288 | 319 | 323 | 411 | 809 | 940 | 739 |
| 26 | 65 | 69 | 13 | 10 | 89 | 171 | 259 | 231 | 296 | 263 | 247 | 273 | 408 | 717 | 899 | 734 |
| 27 | 47 | 69 | 20 | 12 | 81 | 144 | 212 | 239 | 235 | 226 | 174 | 211 | 343 | 550 | 755 | 657 |
| 28 | 30 | 63 | 17 | 5 | 120 | 159 | 216 | 187 | 186 | 199 | 175 | 197 | 276 | 548 | 697 | 655 |
| 29 | 21 | 53 | 11 | 3 | 90 | 87 | 133 | 133 | 139 | 113 | 138 | 114 | 164 | 375 | 467 | 474 |
| 30 | 19 | 27 | 8 | - | 94 | 79 | 149 | 128 | 126 | 90 | 95 | 94 | 104 | 248 | 301 | 305 |
| 31 | 9 | 34 | 7 | 2 | 61 | 55 | 79 | 81 | 110 | 80 | 64 | 72 | 77 | 148 | 211 | 202 |
| 32 | 11 | 11 | 4 | - | 35 | 42 | 42 | 37 | 51 | 51 | 46 | 50 | 56 | 92 | 124 | 123 |
| 33 | 6 | 14 | 3 | - | 20 | 26 | 39 | 29 | 28 | 22 | 25 | 16 | 28 | 42 | 54 | 56 |
| 34 | 5 | 16 | 2 | - | 15 | 10 | 26 | 17 | 12 | 16 | 6 | 8 | 10 | 27 | 27 | 24 |
| 35 | 9 | 8 | 2 | - | 9 | 9 | 10 | 7 | 1 | 3 | 2 | 5 | 4 | 15 | 7 | 12 |
| 36 | 4 | 6 | - | - | 4 | - | 4 | 1 | 1 | 4 | 1 | 1 | - | 1 | 2 | 2 |
| 37 | 2 | 4 | - | - | 3 | 1 | 4 | 1 | - | - | - | - | - | 1 | 2 | 1 |
| 38 | - | - | - | - | - | - | 1 | - | 1 | - | - | - | - | - | - | - |
| 39 | - | - | - | - | 1 | - | - | - | 1 | - | - | 1 | - | - | - | - |
| 40 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 41 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | 1 | - |
| 42 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 43 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 44 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 45 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 46 | - | - | - | - | 1 | - | - | - | - | - | - | - | - | - | - | - |
| 47 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 48 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 49 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| 50 | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - | - |
| total | 1,763 | 2,168 | 1,183 | 968 | 1,316 | 2,626 | 3,251 | 2,709 | 3,182 | 3,499 | 3,761 | 3,896 | 4,452 | 8,512 | 9,500 | 9,622 |

Table 4. Weighted age samples and total effective sample size for the three fishery categories used in the assessment (handline- HL, longline- LL and recreational- rec).

| fishery | year | AGE0 | AGE1 | AGE2 | AGE3 | AGE4 | AGE5 | AGE6 | AGE7 | AGE8 | AGE9 | AGE10 | AGE11 | AGE12 | AGE13 | AGE14 | AGE15 | AGE16 | AGE17 | AGE18 | AGE19 | AGE20 | effective sample size |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| HL+ | 1992 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2.3 | 3.0 | 3.0 | 1.8 | 0.0 | 1.0 | 0.3 | 0.2 | 0.3 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 12.0 |
| HL+ | 1993 | 0.0 | 0.0 | 0.0 | 0.1 | 3.3 | 3.4 | 8.7 | 8.0 | 6.7 | 0.8 | 1.1 | 0.1 | 0.1 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 32.8 |
| HL+ | 1994 | 0.0 | 0.0 | 0.0 | 0.4 | 6.6 | 9.5 | 5.0 | 2.3 | 1.8 | 0.9 | 0.8 | 0.4 | 0.4 | 0.4 | 0.3 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 29.2 |
| HL+ | 1995 | 0.0 | 0.0 | 0.0 | 0.0 | 0.5 | 4.8 | 9.1 | 5.7 | 2.7 | 1.3 | 0.8 | 0.3 | 0.4 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 26.1 |
| HL+ | 1996 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.2 | 4.5 | 3.5 | 1.3 | 1.3 | 0.3 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 12.3 |
| HL+ | 1997 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 2.4 | 2.5 | 2.1 | 1.7 | 0.1 | 0.1 | 0.0 | 0.0 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 9.5 |
| HL+ | 1998 | 0.0 | 0.0 | 0.0 | 0.5 | 1.5 | 3.8 | 2.0 | 2.9 | 1.3 | 1.4 | 1.8 | 0.0 | 0.1 | 0.0 | 0.0 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 15.6 |
| HL+ | 1999 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 2.8 | 2.1 | 3.1 | 7.6 | 3.5 | 2.4 | 0.4 | 0.4 | 0.7 | 0.1 | 0.0 | 0.0 | 0.4 | 0.0 | 0.0 | 0.0 | 23.8 |
| HL+ | 2000 | 0.0 | 0.0 | 0.0 | 0.0 | 9.5 | 4.4 | 8.6 | 5.5 | 4.9 | 3.7 | 2.9 | 2.1 | 0.9 | 1.1 | 0.8 | 0.6 | 0.0 | 0.0 | 0.0 | 0.3 | 0.3 | 45.8 |
| HL+ | 2001 | 0.0 | 0.0 | 0.0 | 0.0 | 4.6 | 45.2 | 13.9 | 19.3 | 11.9 | 4.0 | 8.7 | 5.3 | 2.9 | 3.8 | 1.7 | 1.2 | 2.3 | 1.2 | 0.3 | 0.0 | 0.1 | 126.8 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ■ | 1992 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 7.0 | 13.4 | 21.4 | 12.9 | 7.5 | 1.5 | 0.5 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.5 | 0.0 | 0.0 | 0.0 | 65.7 |
| ■ | 1993 | 0.0 | 0.0 | 0.0 | 0.0 | 5.5 | 12.9 | 28.5 | 25.9 | 9.7 | 6.5 | 3.2 | 1.9 | 0.6 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 95.4 |
| L | 1994 | 0.0 | 0.0 | 0.0 | 0.0 | 3.2 | 1.9 | 2.6 | 3.2 | 1.3 | 1.3 | 1.9 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 16.0 |
| L | 1995 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 3.5 | 19.3 | 11.3 | 18.0 | 10.7 | 7.3 | 1.7 | 0.6 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.1 | 74.0 |
| L | 1996 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 3.1 | 8.6 | 15.9 | 4.9 | 15.3 | 7.3 | 1.2 | 0.6 | 0.0 | 0.0 | 0.6 | 0.0 | 0.6 | 0.0 | 0.0 | 0.0 | 58.6 |
| ■ | 1997 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 0.3 | 0.3 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 2.0 |
| ■ | 1998 | 0.0 | 0.0 | 0.0 | 0.0 | 2.1 | 6.4 | 7.5 | 13.4 | 10.2 | 7.0 | 1.6 | 3.7 | 2.1 | 0.5 | 0.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 55.1 |
| L | 1999 | 0.0 | 0.0 | 0.0 | 0.0 | 1.8 | 17.0 | 37.0 | 61.1 | 97.7 | 80.4 | 37.5 | 23.9 | 12.8 | 5.8 | 3.0 | 1.8 | 0.6 | 0.0 | 0.6 | 0.0 | 0.6 | 381.6 |
| L | 2000 | 0.0 | 0.0 | 0.0 | 0.0 | 2.4 | 7.3 | 29.6 | 25.8 | 28.9 | 23.3 | 19.1 | 19.9 | 7.8 | 4.0 | 2.3 | 1.7 | 1.7 | 0.0 | 0.6 | 0.6 | 1.1 | 176.0 |
| L | 2001 | 0.0 | 0.0 | 0.0 | 0.0 | 1.2 | 58.9 | 30.5 | 56.9 | 30.5 | 17.7 | 24.6 | 19.6 | 15.4 | 4.2 | 6.1 | 5.3 | 2.0 | 1.7 | 1.1 | 1.2 | 3.5 | 280.5 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| rec | 1992 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 1.1 | 0.2 | 0.1 | 0.5 | 0.3 | 0.5 | 0.5 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 3.8 |
| rec | 1993 | 0.0 | 0.0 | 0.0 | 0.1 | 1.5 | 0.4 | 0.6 | 1.3 | 1.0 | 0.6 | 0.1 | 0.3 | 0.0 | 0.1 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 6.1 |
| rec | 1994 | 0.0 | 0.0 | 0.0 | 0.4 | 1.6 | 1.4 | 1.1 | 0.7 | 1.1 | 0.3 | 0.3 | 0.2 | 0.1 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 7.8 |
| rec | 1995 | 0.0 | 0.0 | 0.0 | 0.0 | 1.1 | 5.3 | 6.2 | 4.0 | 2.9 | 2.5 | 0.6 | 1.1 | 0.9 | 0.3 | 0.8 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.5 | 26.4 |
| rec | 1996 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 4.4 | 12.2 | 6.4 | 2.3 | 1.5 | 0.7 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 28.0 |
| rec | 1997 | 0.0 | 0.0 | 0.0 | 0.7 | 0.3 | 2.0 | 5.9 | 7.7 | 2.5 | 0.3 | 0.1 | 0.6 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 21.4 |
| rec | 1998 | 0.0 | 0.0 | 0.0 | 0.0 | 1.4 | 1.4 | 0.2 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 3.3 |
| rec | 1999 | 0.0 | 0.0 | 0.0 | 0.1 | 1.0 | 3.6 | 1.2 | 0.1 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 6.4 |
| rec | 2000 | 0.0 | 0.0 | 0.0 | 0.0 | 1.0 | 0.7 | 1.0 | 0.5 | 1.0 | 0.8 | 0.1 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 5.4 |
| rec | 2001 | 0.0 | 0.0 | 0.0 | 0.0 | 0.7 | 4.3 | 1.8 | 1.2 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.2 | 0.0 | 0.0 | 0.0 | 8.6 |

Table 5. Summary of reproductive parameters used in this assessment. Values for age 20 plus-group were computed by prorating the values for ages 20 to 30 assuming a total mortality rate of $0.4 \mathrm{yr}^{-1}$ (as done by Schirripa et al. 1999). The per capita fecundity is the product of the proportion female, proportion mature, and batch fecundity. The relative per capita fecundity is the per capita fecundity scaled by the average over all ages. The column labeled "Schirripa" is the vector used in the previous assessment (also scaled by its mean).

| Age | proportion <br> female | proportion <br> active | batch fecundity (millions <br> of ova) <br> by age <br> by length |  | relative per capita fecundity |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.940 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 | 0.929 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 | 0.916 | 0.505 | 0.109 | 0.110 | 0.223 | 0.245 | 0.043 |
| 4 | 0.901 | 0.537 | 0.128 | 0.144 | 0.274 | 0.334 | 0.122 |
| 5 | 0.884 | 0.565 | 0.151 | 0.183 | 0.332 | 0.438 | 0.252 |
| 6 | 0.864 | 0.590 | 0.177 | 0.226 | 0.399 | 0.554 | 0.426 |
| 7 | 0.842 | 0.612 | 0.208 | 0.274 | 0.474 | 0.678 | 0.626 |
| 8 | 0.816 | 0.632 | 0.244 | 0.325 | 0.557 | 0.806 | 0.835 |
| 9 | 0.788 | 0.650 | 0.287 | 0.380 | 0.650 | 0.933 | 1.032 |
| 10 | 0.756 | 0.667 | 0.338 | 0.436 | 0.752 | 1.053 | 1.206 |
| 11 | 0.721 | 0.682 | 0.397 | 0.493 | 0.862 | 1.162 | 1.345 |
| 12 | 0.683 | 0.695 | 0.466 | 0.550 | 0.979 | 1.255 | 1.445 |
| 13 | 0.643 | 0.708 | 0.548 | 0.608 | 1.103 | 1.327 | 1.504 |
| 14 | 0.601 | 0.719 | 0.644 | 0.664 | 1.230 | 1.377 | 1.523 |
| 15 | 0.557 | 0.730 | 0.757 | 0.719 | 1.360 | 1.402 | 1.505 |
| 16 | 0.512 | 0.740 | 0.890 | 0.772 | 1.490 | 1.403 | 1.454 |
| 17 | 0.467 | 0.749 | 1.046 | 0.823 | 1.617 | 1.381 | 1.378 |
| 18 | 0.423 | 0.757 | 1.230 | 0.871 | 1.739 | 1.338 | 1.282 |
| 19 | 0.379 | 0.765 | 1.446 | 0.917 | 1.854 | 1.277 | 1.172 |
| 20 | 0.338 | 0.773 | 1.699 | 0.960 | 2.105 | 1.035 | 0.850 |

Table 6. Annual commercial reported and calculated red grouper (mt) in the U.S. Gulf of Mexico catches and proportion of unclassified groupers considered red grouper in west Florida.

|  | proportion red <br> grouper in <br> unclassified | reported red <br> grouper (mt) | calculated red <br> grouper (mt) |
| ---: | ---: | ---: | ---: |
| 1986 | 0.680 | 2,869 | 2,937 |
| 1987 | 0.710 | 3,041 | 3,129 |
| 1988 | 0.660 | 2,123 | 2,220 |
| 1989 | 0.750 | 3,393 | 3,446 |
| 1990 | 0.629 | 2,166 | 2,197 |
| 1991 | 0.661 | 2,312 | 2,329 |
| 1992 | 0.607 | 1,919 | 1,930 |
| 1993 | 0.695 | 2,891 | 2,903 |
| 1994 | 0.641 | 2,233 | 2,241 |
| 1995 | 0.656 | 2,159 | 2,164 |
| 1996 | 0.659 | 2,028 | 2,030 |
| 1997 | 0.646 | 2,216 | 2,217 |
| 1998 | 0.540 | 1,801 | 1,804 |
| 1999 | 0.654 | 2,698 | 2,700 |
| 2000 | 0.625 | 2,635 | 2,636 |
| 2001 | 0.580 | 2,639 | 2,640 |

Table 7. Annual calculated red grouper from the commercial fishery by state. Calculated red grouper includes reported plus a portion of unclassified.

|  | AL | LA | MI | TX | wFL | total |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1986 | - | 1 | - | - | 2,936 | 2,937 |
| 1987 | - | 0 | - | - | 3,129 | 3,129 |
| 1988 | - | 0 | - | - | 2,220 | 2,220 |
| 1989 | 4 | 0 | - | - | 3,442 | 3,446 |
| 1990 | - | - | - | - | 2,197 | 2,197 |
| 1991 | - | 0 | - | - | 2,329 | 2,329 |
| 1992 | - | 0 | - | - | 1,930 | 1,930 |
| 1993 | - | 0 | - | - | 2,902 | 2,903 |
| 1994 | 0 | 1 | - | - | 2,240 | 2,241 |
| 1995 | - | - | - | - | 2,164 | 2,164 |
| 1996 | - | - | - | - | 2,030 | 2,030 |
| 1997 | - | 0 | - | - | 2,217 | 2,217 |
| 1998 | 3 | - | - | - | 1,801 | 1,804 |
| 1999 | - | 1 |  | 0 | - | 2,700 |
| 2000 | - | 0 | - | - | 2,700 |  |
| 2001 | - | 1 | - | - | 2,636 | 2,636 |
|  | - | - | - | 2,640 |  |  |

Table 8. U.S. Gulf of Mexico commercial landings of calculated red grouper (reported plus a portion of unclassified groupers) by gear (mt).

|  | longline | handline+ | total |
| ---: | ---: | ---: | ---: |
| 1986 | 1,202 | 1,735 | 2,937 |
| 1987 | 1,601 | 1,529 | 3,129 |
| 1988 | 956 | 1,264 | 2,220 |
| 1989 | 1,289 | 2,156 | 3,446 |
| 1990 | 1,225 | 972 | 2,197 |
| 1991 | 1,254 | 1,075 | 2,329 |
| 1992 | 663 | 1,267 | 1,930 |
| 1993 | 1,960 | 943 | 2,903 |
| 1994 | 1,237 | 1,004 | 2,241 |
| 1995 | 1,128 | 1,036 | 2,164 |
| 1996 | 1,364 | 665 | 2,030 |
| 1997 | 1,434 | 783 | 2,217 |
| 1998 | 1,302 | 502 | 1,804 |
| 1999 | 1,783 | 918 | 2,700 |
| 2000 | 1,346 | 1,290 | 2,636 |
| 2001 | 1,546 | 1,094 | 2,640 |

Table 9. Recreational harvest (number of fish) of red grouper in the Gulf of Mexico by state.

|  | TX | LA | MS | AL | wFL | total |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |
| 1986 | 5 | - | - | 1,210 | 821,209 | 822,424 |
| 1987 | 1 | - | - | 932 | 442,653 | 443,586 |
| 1988 | 10 | - | - | 648 | 738,163 | 738,821 |
| 1989 | 1 | 1 | - | 302 | 826,037 | 826,341 |
| 1990 | 1 | - | - | 272 | 219,595 | 219,868 |
| 1991 | - | 696 | - | 244 | 295,642 | 296,581 |
| 1992 | 2 | 1 | - | 172 | 469,424 | 469,599 |
| 1993 | 5 | 5 | - | 1,161 | 371,673 | 372,844 |
| 1994 | - | - | - | 723 | 320,697 | 321,420 |
| 1995 | 2 | - | - | 1,037 | 327,959 | 328,999 |
| 1996 | 2 | - | - | 2,177 | 148,109 | 150,288 |
| 1997 | - | - | - | 1,165 | 97,186 | 98,351 |
| 1998 | - | - | 9 | - | 841 | 117,742 |
| 1999 | - | 5 | - | 718,592 |  |  |
| 2000 | - | 3 | - | 1,063 | 257,869 | 258,935 |
| 2001 | - | 3 | - | 1,083 | 208,511 | 209,597 |

Table 10. Recreational harvest (number of fish)of red grouper in the Gulf of Mexico by mode.

|  | shore | headboat | charter | private | total |
| ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |
| 1986 | 5,863 | 34,512 | 111,071 | 670,984 | 822,430 |
| 1987 | 10,105 | 27,678 | 68,301 | 337,531 | 443,616 |
| 1988 | 7,601 | 28,333 | 71,071 | 631,814 | 738,819 |
| 1989 | - | 50,987 | 62,766 | 712,589 | 826,342 |
| 1990 | 13,506 | 18,653 | 70,962 | 116,750 | 219,871 |
| 1991 | 9,378 | 10,136 | 12,926 | 264,147 | 296,586 |
| 1992 | 24,264 | 9,654 | 53,094 | 382,585 | 469,598 |
| 1993 | 16,797 | 9,779 | 31,021 | 315,253 | 372,849 |
| 1994 | 3,770 | 10,303 | 38,186 | 269,162 | 321,421 |
| 1995 | 1,315 | 15,094 | 86,255 | 226,334 | 328,998 |
| 1996 | - | 16,216 | 28,042 | 106,029 | 150,286 |
| 1997 | 1,369 | 5,692 | 26,557 | 64,735 | 98,353 |
| 1998 | 901 | 6,531 | 29,546 | 81,619 | 118,597 |
| 1999 | - | 7,597 | 25,039 | 144,732 | 177,368 |
| 2000 | - | 10,226 | 30,858 | 217,853 | 258,937 |
| 2001 | - | 10,226 | 42,710 | 156,663 | 209,599 |

Table 11. Coefficients of variation about annual estimates of recreational harvest from MRFSS for red grouper from west Florida by mode.

|  | shore | charter | private |
| ---: | ---: | ---: | ---: |
| 1986 | 100 | 20 | 20 |
| 1987 | 95 | 24 | 13 |
| 1988 | 73 | 21 | 13 |
| 1989 |  | 24 | 12 |
| 1990 | 91 | 60 | 17 |
| 1991 | 37 | 37 | 13 |
| 1992 | 30 | 15 | 8 |
| 1993 | 45 | 18 | 13 |
| 1994 | 59 | 15 | 11 |
| 1995 | 71 | 38 | 14 |
| 1996 |  | 51 | 14 |
| 1997 | 100 | 47 | 17 |
| 1998 | 100 | 30 | 14 |
| 1999 |  | 15 | 13 |
| 2000 |  | 13 | 14 |
| 2001 |  | 12 | 14 |

Table 12. MRFSS estimates of the number of red grouper released alive by state in the Gulf of Mexico.

|  | LA | MS | AL | wFL | total |
| ---: | :---: | :---: | :---: | ---: | ---: |
| 1986 | - | - | - | 559,543 | 559,543 |
| 1987 | - | - | - | 489,823 | 489,823 |
| 1988 | - | - | - | 893,491 | 893,491 |
| 1989 | - | - | - | $2,034,825$ | $2,034,825$ |
| 1990 | - | - | 179 | $1,689,989$ | $1,690,168$ |
| 1991 | - | - | - | $3,012,905$ | $3,012,905$ |
| 1992 | - | - | - | $2,740,258$ | $2,740,258$ |
| 1993 | - | - | - | $1,708,137$ | $1,708,137$ |
| 1994 | - | - | - | $1,707,586$ | $1,707,586$ |
| 1995 | - | - | - | $1,712,438$ | $1,712,438$ |
| 1996 | - | - | - | $1,099,000$ | $1,099,000$ |
| 1997 | - | - | - | $1,129,070$ | $1,129,070$ |
| 1998 | - | - | - | $1,475,413$ | $1,475,413$ |
| 1999 | - | - | - | $1,994,645$ | $1,994,645$ |
| 2000 | - | - | - | $1,787,640$ | $1,787,640$ |
| 2001 | - | - | 126 | $1,557,902$ | $1,558,028$ |

Table 13. MRFSS estimates of the number of red grouper released alive by mode in the Gulf of Mexico.

|  | shore | charter | private | total |
| ---: | ---: | ---: | ---: | ---: |
| 1986 | 5,863 | 109,419 | 444,261 | 559,543 |
| 1987 | - | 86,356 | 40,467 | 489,823 |
| 1988 | 11,632 | 64,532 | 817,327 | 893,491 |
| 1989 | 1,794 | 155,246 | $1,877,785$ | $2,034,825$ |
| 1990 | 20,881 | 310,878 | $1,358,409$ | $1,690,168$ |
| 1991 | 33,429 | 56,522 | $2,922,955$ | $3,012,905$ |
| 1992 | 81,896 | 207,621 | $2,450,741$ | $2,740,258$ |
| 1993 | 7,567 | 79,105 | $1,621,466$ | $1,708,137$ |
| 1994 | 16,405 | 144,420 | $1,546,760$ | $1,707,586$ |
| 1995 | 5,099 | 226,189 | $1,481,149$ | $1,712,438$ |
| 1996 | 14,287 | 90,321 | 994,391 | $1,099,000$ |
| 1997 | 8,894 | 151,706 | 968,470 | $1,129,070$ |
| 1998 | 9,758 | 172,154 | $1,293,502$ | $1,475,413$ |
| 1999 | 6,049 | 231,609 | $1,756,987$ | $1,994,645$ |
| 2000 | 7,793 | 91,529 | $1,688,318$ | $1,787,640$ |
| 2001 | 3,234 | 122,511 | $1,432,283$ | $1,558,028$ |

Table 14. Proportions of the total estimate catch from MRFSS ( $\mathrm{A}+\mathrm{B} 1+\mathrm{B} 2$ ) which was released alive (B2).

| 1986 | 0.415 |
| ---: | ---: |
| 1987 | 0.541 |
| 1988 | 0.557 |
| 1989 | 0.724 |
| 1990 | 0.894 |
| 1991 | 0.913 |
| 1992 | 0.856 |
| 1993 | 0.825 |
| 1994 | 0.846 |
| 1995 | 0.845 |
| 1996 | 0.891 |
| 1997 | 0.924 |
| 1998 | 0.929 |
| 1999 | 0.922 |
| 2000 | 0.878 |
| 2001 | 0.887 |

Table 15. Alternative derived age composition for 1986-1997 assuming longline discard mortality of 0.33..

|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 69,996 | 53,005 | 88,446 | 195,692 | 64,583 | 104,927 | 95,944 | 59,366 | 59,624 | 62,270 | 40,525 | 39,312 |
| 2 | 452,470 | 363,692 | 549,670 | 678,459 | 86,783 | 163,360 | 115,829 | 112,367 | 175,566 | 166,524 | 168,730 | 168,682 |
| 3 | 539,337 | 414,122 | 546,550 | 762,851 | 125,286 | 205,332 | 164,645 | 180,262 | 181,724 | 199,596 | 162,886 | 163,988 |
| 4 | 406,850 | 305,758 | 360,896 | 507,807 | 152,701 | 244,790 | 243,894 | 257,899 | 233,433 | 245,109 | 183,134 | 191,404 |
| 5 | 268,678 | 218,195 | 225,998 | 310,060 | 150,482 | 222,250 | 234,708 | 256,802 | 220,872 | 226,052 | 163,369 | 169,430 |
| 6 | 170,920 | 155,552 | 140,650 | 186,796 | 128,081 | 167,424 | 179,387 | 205,742 | 169,947 | 172,837 | 122,490 | 126,164 |
| 7 | 108,950 | 109,262 | 88,455 | 113,107 | 99,372 | 116,133 | 124,832 | 148,200 | 118,989 | 120,304 | 84,134 | 87,028 |
| 8 | 70,188 | 75,342 | 56,136 | 69,380 | 72,586 | 77,753 | 83,346 | 101,111 | 79,560 | 79,934 | 55,357 | 57,886 |
| 9 | 45,671 | 51,200 | 35,825 | 43,165 | 51,010 | 51,235 | 54,599 | 67,087 | 52,007 | 51,959 | 35,714 | 37,823 |
| 10 | 29,926 | 34,455 | 22,958 | 27,196 | 35,008 | 33,530 | 35,487 | 43,921 | 33,667 | 33,482 | 22,880 | 24,524 |
| 11 | 19,696 | 23,047 | 14,771 | 17,318 | 23,694 | 21,898 | 23,021 | 28,607 | 21,739 | 21,542 | 14,654 | 15,872 |
| 12 | 13,002 | 15,366 | 9,545 | 11,127 | 15,912 | 14,310 | 14,952 | 18,621 | 14,057 | 13,888 | 9,417 | 10,287 |
| 13 | 8,600 | 10,229 | 6,195 | 7,203 | 10,643 | 9,370 | 9,737 | 12,141 | 9,117 | 8,989 | 6,077 | 6,686 |
| 14 | 5,700 | 6,806 | 4,038 | 4,691 | 7,105 | 6,150 | 6,363 | 7,939 | 5,937 | 5,842 | 3,942 | 4,363 |
| 15 | 3,783 | 4,530 | 2,643 | 3,071 | 4,740 | 4,048 | 4,173 | 5,206 | 3,880 | 3,814 | 2,569 | 2,856 |
| 16 | 2,515 | 3,017 | 1,737 | 2,019 | 3,162 | 2,672 | 2,745 | 3,425 | 2,548 | 2,499 | 1,682 | 1,878 |
| 17 | 1,673 | 2,011 | 1,145 | 1,332 | 2,111 | 1,766 | 1,811 | 2,260 | 1,677 | 1,645 | 1,106 | 1,237 |
| 18 | 1,116 | 1,341 | 757 | 882 | 1,409 | 1,171 | 1,198 | 1,494 | 1,108 | 1,086 | 729 | 817 |
| 19 | 744 | 895 | 501 | 585 | 943 | 779 | 794 | 992 | 734 | 718 | 483 | 543 |
| 20 | 1,483 | 1,784 | 985 | 1,156 | 1,880 | 1,534 | 1,563 | 1,952 | 1,439 | 1,406 | 942 | 1,064 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| total | 2,221,298 | 1,849,609 | 2,157,901 | 2,943,897 | 1,037,491 | 1,450,433 | 1,399,028 | 1,515,395 | 1,387,626 | 1,419,497 | 1,080,819 | 1,111,845 |

Table 16. Derived age composition for 1986-2001 assuming longline discard mortality of 0.33.

|  | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 70,061 | 53,057 | 88,612 | 195,788 | 64,630 | 105,219 | 95,916 | 59,656 | 59,689 | 60,818 | 41,527 | 40,435 | 52,501 | 70,972 | 62,991 | 55,642 |
| 2 | 452,845 | 363,862 | 556,278 | 679,458 | 88,962 | 166,424 | 117,179 | 112,843 | 176,159 | 163,151 | 190,996 | 179,627 | 134,740 | 172,371 | 164,271 | 141,881 |
| 3 | 539,656 | 414,268 | 551,324 | 763,931 | 127,713 | 208,692 | 165,410 | 177,377 | 180,507 | 196,676 | 180,113 | 173,773 | 134,296 | 190,344 | 200,688 | 192,575 |
| 4 | 407,096 | 305,900 | 361,437 | 508,644 | 153,858 | 247,239 | 242,048 | 254,897 | 232,098 | 242,615 | 202,139 | 203,363 | 157,310 | 235,189 | 242,387 | 253,875 |
| 5 | 268,766 | 218,296 | 225,574 | 310,691 | 150,673 | २23,926 | 232,925 | 255,473 | २20,२34 | 223,931 | 181,101 | 180,833 | 147,260 | 224,516 | 224,336 | 234,682 |
| 6 | 170,934 | 155,608 | 140,352 | 187,266 | 128,002 | 168,336 | 179,369 | 205,719 | 169,989 | 171,117 | 136,342 | 135,104 | 115,640 | 176,172 | 176,271 | 178,415 |
| 7 | 108,956 | 109,296 | 88,270 | 113,445 | 99,300 | 116,550 | 126,024 | 148,895 | 119,382 | 119,015 | 93,951 | 93,507 | 82,192 | 125,023 | 127,296 | 124,324 |
| 8 | 70,198 | 75,364 | 55,988 | 69,618 | 72,548 | 77,945 | 84,870 | 101,991 | 80,048 | 79,019 | 61,960 | 62,391 | 55,325 | 84,158 | 87,502 | 82,916 |
| 9 | 45,682 | 51,214 | 35,695 | 43,326 | 50,994 | 51,347 | 55,969 | 67,875 | 52,457 | 51,331 | 40,040 | 40,873 | 36,229 | 55,148 | 58,425 | 54,088 |
| 10 | 29,935 | 34,463 | 22,849 | 27,305 | 35,000 | 33,612 | 36,555 | 44,535 | 34,033 | 33,059 | 25,681 | 26,555 | 23,438 | 35,712 | 38,409 | 34,946 |
| 11 | 19,704 | 23,053 | 14,686 | 17,391 | 23,689 | 21,966 | 23,794 | 29,054 | 22,013 | 21,260 | 16,462 | 17,212 | 15,110 | 23,042 | 25,074 | 22,524 |
| 12 | 13,008 | 15,370 | 9,480 | 11,176 | 15,909 | 14,364 | 15,491 | 18,934 | 14,255 | 13,702 | 10,585 | 11,170 | 9,753 | 14,881 | 16,339 | 14,540 |
| 13 | 8,606 | 10,231 | 6,149 | 7,234 | 10,640 | 9,412 | 10,108 | 12,355 | 9,257 | 8,863 | 6,837 | 7,268 | 6,312 | 9,640 | 10,656 | 9,418 |
| 14 | 5,702 | 6,808 | 4,005 | 4,713 | 7,104 | 6,184 | 6,614 | 8,083 | 6,035 | 5,760 | 4,436 | 4,745 | 4,104 | 6,270 | 6,968 | 6,126 |
| 15 | 3,785 | 4,531 | 2,621 | 3,086 | 4,739 | 4,073 | 4,340 | 5,304 | 3,948 | 3,760 | 2,892 | 3,108 | 2,678 | 4,095 | 4,569 | 4,001 |
| 16 | 2,516 | 3,017 | 1,721 | 2,028 | 3,163 | 2,689 | 2,857 | 3,492 | 2,592 | 2,464 | 1,893 | 2,043 | 1,757 | 2,686 | 3,005 | 2,623 |
| 17 | 1,675 | 2,011 | 1,134 | 1,338 | 2,111 | 1,780 | 1,887 | 2,305 | 1,708 | 1,620 | 1,244 | 1,348 | 1,155 | 1,768 | 1,981 | 1,727 |
| 18 | 1,116 | 1,341 | 749 | 885 | 1,409 | 1,181 | 1,249 | 1,525 | 1,129 | 1,068 | 820 | 890 | 762 | 1,165 | 1,311 | 1,140 |
| 19 | 744 | 895 | 496 | 587 | 941 | 784 | 828 | 1,012 | 747 | 708 | 544 | 591 | 504 | 772 | 868 | 754 |
| 20 | 1,484 | 1,784 | 974 | 1,160 | 1,877 | 1,548 | 1,627 | 1,990 | 1,473 | 1,387 | 1,062 | 1,162 | 990 | 1,516 | 1,710 | 1,480 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| total | 2,2२2,469 | 1,850,369 | 2,168,394 | 2,949,070 | 1,043,264 | 1,463,270 | 1,405,060 | 1,513,317 | 1,387,754 | 1,401,324 | 1,200,625 | 1,185,999 | 982,056 | 1,435,439 | 1,455,058 | 1,417,677 |

Table 17. Indices of abundance and corresponding coefficients of variation (cv) used in the red grouper assessment.

| year | comm. handline (logbook) |  | comm. longline (logbook) |  | comm. trap <br> (logbook) |  | MRFSS private |  | trap survey |  |  | video survey |  | Mote tagging |  | operating units |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 |  |  |  |  |  |  | 0.4043 | 0.1555 |  |  |  |  |  |  |  | 1.5103 | 1.0000 |
| 1982 |  |  |  |  |  |  | 0.6877 | 0.0674 |  |  |  |  |  |  |  | 1.4992 | 1.0000 |
| 1983 |  |  |  |  |  |  | 0.7297 | 0.0950 |  |  |  |  |  |  |  | 1.2943 | 1.0000 |
| 1984 |  |  |  |  |  |  | 0.9563 | 0.0666 |  |  |  |  |  |  |  | 1.5662 | 1.0000 |
| 1985 |  |  |  |  |  |  | 0.6775 | 0.0830 |  |  |  |  |  |  |  | 1.8628 | 1.0000 |
| 1986 |  |  |  |  |  |  | 0.8396 | 0.0383 |  |  |  |  |  |  |  | 1.6760 | 1.0000 |
| 1987 |  |  |  |  |  |  | 0.6350 | 0.0486 |  |  |  |  |  |  |  | 1.3132 | 1.0000 |
| 1988 |  |  |  |  |  |  | 0.6703 | 0.0441 |  |  |  |  |  |  |  | 1.1015 | 1.0000 |
| 1989 |  |  |  |  |  |  | 0.9616 | 0.0392 |  |  |  |  |  |  |  | 1.8374 | 1.0000 |
| 1990 | 0.9733 | 0.3379 | 0.9658 | 0.2316 | 0.9125 | 0.2233 | 0.8866 | - 0.0446 |  |  |  |  |  |  |  | 0.7312 | 1.0000 |
| 1991 | 1.0297 | 0.0455 | 0.7363 | 0.1237 | 1.2077 | 0.1208 | 1.3364 | -0.0385 |  |  |  |  |  | 1.7901 | 0.1300 | 0.7996 | 1.0000 |
| 1992 | 1.0728 | 0.0427 | 0.6415 | 0.1316 | 0.8164 | 0.1200 | 1.2249 | 0.0282 | 0.8776 |  | 0.3480 | 0.2938 | 0.2730 | 1.4199 | 0.0800 | 0.9613 | 1.0000 |
| 1993 | 0.8378 | 0.0274 | 1.3691 | 0.1158 | 0.7712 | 0.1176 | 1.0972 | 20.0361 | 0.6883 |  | 0.3190 | 0.5911 | 0.1780 | 0.9547 | 0.1100 | 1.0210 | 1.0000 |
| 1994 | 0.9611 | 0.0255 | 0.8571 | 0.1149 | 0.7987 | 0.1192 | 1.1374 | 0.0333 | 3.0335 |  | 0.2430 | 0.6293 | 0.1730 | 1.2023 | 0.1900 | 0.8769 | 1.0000 |
| 1995 | 0.9771 | 0.0250 | 0.8526 | 0.1151 | 0.8893 | 0.1193 | 1.1088 | 0.0361 | 0.6452 |  | 0.3810 | 1.2454 | 0.1900 | 1.0659 | 0.2200 | 0.6804 | 1.0000 |
| 1996 | 0.7443 | 0.0254 | 0.9740 | 0.1152 | 0.6325 | 0.1204 | 0.8808 | 0.0385 | 0.6734 |  | 0.3720 | 1.2311 | 0.1650 | 0.7198 | 0.3000 | 0.4533 | 1.0000 |
| 1997 | 0.8389 | 0.0248 | 1.0230 | 0.1149 | 0.9336 | 0.1216 | 0.9294 | 0.0363 | 0.0819 |  | 0.8390 | 2.0093 | 0.1330 | 0.5098 | 0.1100 | 0.5482 | 1.0000 |
| 1998 | 0.7620 | 0.0244 | 1.1179 | 0.1150 | 0.7038 | 0.1273 | 1.0984 | 0.0299 |  |  |  |  |  | 0.6964 | 0.1000 |  |  |
| 1999 | 1.0286 | 0.0232 | 1.3124 | 0.1154 | 1.3769 | 0.1235 | 1.1310 | 0.0248 |  |  |  |  |  | 1.0283 | 0.1400 |  |  |
| 2000 | 1.2968 | 0.0226 | 0.9523 | 0.1155 | 1.6399 | 0.1229 | 1.0645 | 0.0285 |  |  |  |  |  | 0.8753 | 0.1300 |  |  |
| 2001 | 1.4776 | 0.0231 | 1.1980 | 0.1154 | 1.3175 | 0.1269 | 0.9982 | 0.0297 |  |  |  |  |  | 0.7375 | 0.1700 |  |  |

Table 18. Population estimates based on the original 1986-1997 catch/discard data (RFSAP 2000) contrasted with corresponding estimates obtained from the revised catch/discard data.

|  | Steepness $=0.7$ |  |  | Steepness $=0.8$ |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| revised | revised <br> revised <br> revised <br> $(93 \%)$ | original | $(33 \%)$ | $(90 \%)$ | original | Comments |  |
| MSY | 2.908 | 2.905 | 3.181 | 2.819 | 2.799 | 3.041 | $1000 '$ 's MT |
| $\mathrm{F}_{\text {current }}$ | 0.344 | 0.371 | 0.302 | 0.344 | 0.372 | 0.302 |  |
| $\mathrm{~F}_{\mathrm{MSY}}$ | 0.298 | 0.296 | 0.223 | 0.365 | 0.360 | 0.270 |  |
| $\mathrm{~F}_{\text {current }}$ |  |  |  |  |  | 1.117 |  |
| $/ \mathrm{F}_{\mathrm{MSY}}$ | 1.153 | 1.253 | 1.356 | 0.945 | 1.032 |  | $>1=$ overfishing |
| $\mathrm{SS}_{97}$ | 233.7 | 217.8 | 244.3 | 234.7 | 227.7 | 246.3 | MT gonad weight |
| $\mathrm{SS}_{\mathrm{MSY}}$ | 379.8 | 390.9 | 433.2 | 314.6 | 323.3 | 350.7 | MT gonad weight |
| $\mathrm{SS}_{97}$ |  |  |  |  |  | 0.702 |  |
| $/ \mathrm{SS}_{\mathrm{MSY}}$ | 0.615 | 0.557 | 0.564 | 0.746 | 0.704 |  | $<0.8=$ overfished |
| $\mathrm{T}_{\text {min }}$ | 2004 | 2004 | 2004 | 2002 | 2003 | 2002 | recovery at $\mathrm{F}=0$ |

Table 19. Population estimates from phase 2 (1986-2001data using methods of previous assessment) assuming 33\% or $90 \%$ mortality of longline releases. The quantitiy $S S$ is spawning stock in MT gonad weight. The $M S Y$, $O Y\left(F=0.75 F_{M S Y}\right)$, and $C_{\text {recover }}$ (catch expected to permit recovery to $S S_{M S Y}$ by 2012) are in MT.

|  | Release mortality $=33 \%$ <br> steepness $=0.7$ | Release mortality $=90 \%$ <br> steepness $=0.8$ |  |  |
| :--- | :---: | :---: | :---: | :---: |
|  | 3564 | 3283 | 3592 | 3292 |
| steepness $=0.7$ | steepness $=0.8$ |  |  |  |
| MSY | 3474 | 3211 | 3502 | 3220 |
| OY | 0.975 | 0.978 | 0.975 | 0.978 |
| $\mathrm{OY} / \mathrm{MSY}$ | 0.239 | 0.239 | 0.236 | 0.236 |
| $\mathrm{~F}_{0.1}$ | 0.482 | 0.482 | 0.472 | 0.473 |
| $\mathrm{~F}_{\text {max }}$ | 0.470 | 0.470 | 0.470 | 0.471 |
| $\mathrm{~F}_{30 \%}$ | 0.313 | 0.314 | 0.314 | 0.314 |
| $\mathrm{~F}_{40 \%}$ | 0.285 | 0.344 | 0.282 | 0.340 |
| $\mathrm{~F}_{\mathrm{MSY}}$ | 0.214 | 0.258 | 0.212 | 0.255 |
| $\mathrm{~F}_{\text {OY }}$ | 0.750 | 0.750 | 0.750 | 0.750 |
| $\mathrm{~F}_{\mathrm{OY}} / \mathrm{F}_{\mathrm{MSY}}$ | 0.360 | 0.364 | 0.372 | 0.376 |
| $\mathrm{~F}_{\text {current }}$ | 1.265 | 1.058 | 1.318 | 1.106 |
| $\mathrm{~F}_{\text {curren }} / \mathrm{F}_{\mathrm{MSY}}$ | 1.687 | 1.411 | 1.757 | 1.474 |
| $\mathrm{~F}_{\text {curren }} / \mathrm{F}_{\mathrm{OY}}$ | 484 | 382 | 501 | 395 |
| $\mathrm{SS}_{\mathrm{MSY}}$ | 600 | 473 | 620 | 488 |
| $\mathrm{SS}_{\mathrm{OY}}$ | 1.238 | 1.238 | 1.236 | 1.236 |
| $\mathrm{SS}_{\mathrm{OY}} / \mathrm{SS}_{\mathrm{MSY}}$ | 299 | 296 | 300 | 298 |
| $\mathrm{SS}_{2001}$ | 0.617 | 0.775 | 0.599 | 0.754 |
| $\mathrm{SS}_{2001} / \mathrm{SS}_{\mathrm{MSY}}$ | 0.498 | 0.626 | 0.485 | 0.610 |
| $\mathrm{SS}_{2001} / \mathrm{SS}_{\mathrm{OY}}$ | 2005 | 2004 | 2006 | 2004 |
| $\mathrm{~T}_{\text {min }}$ | 2690 | 3100 | 2630 | 3050 |
| $\mathrm{C}_{\text {recover }}$ |  |  |  |  |

Table 20. Population estimates from phase 3 (1986-2001data using methods of previous assessment plus sampled age composition) assuming $33 \%$ or $90 \%$ mortality of longline releases. The quantities $M S Y, O Y\left(F=0.75 F_{M S Y}\right)$, and $C_{\text {recover }}$ (catch expected to permit recovery to $S S_{M S Y}$ by 2012) are in MT gutted weight. $C_{\text {recover }}$ is replaced by $O Y$ when $S S / S S_{M S Y}>0.8$. The spawning stock $S S$ is in MT gonad weight in the case of the base model, but is unitless (relative) in the case of the new fecundity vector.

|  | base |  |  |  | use new fecundity |  |  |  | adjust cpue |  | downwt indices |  | split hist. cpue |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{RM}=33 \%$ |  | $\mathrm{RM}=90 \%$ |  | M=33\% |  | $\mathrm{M}=90 \%$ |  | $\begin{array}{rr} 33 \% & 90 \% \\ \mathrm{~h}=0.8 & \mathbf{h}=\mathbf{0 . 8} \end{array}$ |  | $\begin{array}{rr} 33 \% & 90 \% \\ \mathrm{~h}=0.8 & \mathbf{h}=\mathbf{0 . 8} \end{array}$ |  | $\begin{array}{rr} 33 \% & 90 \% \\ \mathrm{~h}=0.8 & \mathbf{h}=\mathbf{0 . 8} \end{array}$ |  |
|  | $\mathrm{h}=0.7$ | =0.8 | 0.7 | h=0.8 | $\mathrm{h}=0.7$ | $\mathrm{h}=0.8$ | 0.7 | $\mathrm{h}=0.8$ |  |  |  |  |  |  |
| MSY | 3571 | 344 | 596 | 353 | 374 | 281 | 388 | 284 | 2946 | 2957 | 219 | 239 | 3372 | 379 |
| OY | 3480 | 3270 | 3504 | 3279 | 3301 | 3217 | 3314 | 3220 | 2880 | 2891 | 3146 | 3166 | 3297 | 3304 |
| OY/MSY | 0.974 | 0.978 | 0.975 | 0.978 | 0.978 | 0.981 | 0.978 | 0.981 | 0.977 | 0.977 | 0.977 | 0.977 | 0.978 | 0.978 |
| $\mathrm{F}_{0}$ | 0.238 | 0.238 | 0.235 | 0.235 | 0.238 | 0.238 | 0.235 | 0.235 | 0.235 | 0.232 | 0.230 | 0.228 | 0.238 | 0.235 |
| $\mathrm{F}_{\text {n }}$ | 0.475 | 0.476 | 0.468 | 0.469 | 0.476 | 0.476 | 0.469 | 0.469 | 0.463 | 0.455 | 0.449 | 0.445 | 0.475 | 0.468 |
| $\mathrm{F}_{3}$ | 0.463 | 0.464 | 0.465 | 0.466 | 0.702 | 0.703 | 0.707 | 0.708 | 0.453 | 0.454 | 0.432 | 0.436 | 0.463 | 0.465 |
| $\mathrm{F}_{4}$ | 0.310 | 0.310 | 0.311 | 0.311 | 0.437 | 0.437 | 0.439 | 0.440 | 0.304 | 0.305 | 0.291 | 0.293 | 0.310 | 0.311 |
| $\mathrm{F}_{\mathrm{M}}$ | 0.281 | 0.340 | 0.280 | 0.33 | 0.329 | 0.382 | 0.326 | 0.37 | 0.332 | 0.329 | 0.321 | 0.320 | 0.340 | 0.337 |
| For | 0.211 | 0.255 | 0.210 | 0.253 | 0.247 | 0.286 | 0.245 | 0.283 | 0.249 | 0.247 | 0.241 | 0.240 | 0.255 | 0.253 |
| $\mathrm{F}_{\mathrm{OY}} / \mathrm{F}_{\mathrm{M}}$ | 0.750 | 0.750 | 0.750 | 0.75 | 0.750 | 0.750 | 0.750 | 0.75 | 0.750 | 0.750 | 0.750 | 0.750 | 0.750 | 0.750 |
| $\mathrm{F}_{\text {curi }}$ | 0.314 | 0.316 | 0.322 | 0.32 | 0.316 | 0.317 | 0.324 | 0.325 | 0.363 | 0.370 | 0.311 | 0.321 | 0.310 | 0.318 |
| $\mathrm{F}_{\text {current }} / \mathrm{F}$ | 1.117 | 0.930 | 1.152 | 0.962 | 0.961 | 0.829 | 0.993 | 0.860 | 1.092 | 1.126 | 0.969 | 1.005 | 0.913 | 0.946 |
| $\mathrm{F}_{\text {current }} / \mathrm{F}$ | 1.489 | 1.239 | 1.536 | 1.282 | 1.281 | 1.105 | 1.325 | 1.14 | 1.455 | 1.501 | 1.292 | 1.340 | 1.218 | 1.261 |
| $\mathrm{SS}_{\mathrm{N}}$ | 487 | 391 | 503 | 404 | 5746 | 5086 | 5925 | 5242 | 349 | 360 | 377 | 391 | 394 | 407 |
| $\mathrm{SS}_{\text {OY }}$ | 604 | 485 | 623 | 499 | 6816 | 5998 | 7019 | 6172 | 433 | 446 | 469 | 485 | 489 | 503 |
| $\mathrm{SS}_{\mathrm{OY}} / \mathrm{SS}$ | 1.239 | 1.239 | 1.238 | 1.237 | 1.186 | 1.179 | 1.185 | 1.177 | 1.241 | 1.239 | 1.244 | 1.242 | 1.239 | 1.237 |
| $\mathrm{SS}_{2001}$ | 338 | 337 | 342 | 341 | 5327 | 5321 | 5415 | 5409 | 279 | 282 | 344 | 346 | 343 | 347 |
| $\mathrm{SS}_{2001} / \mathrm{SS}_{\text {MSY }}$ | 0.694 | 0.862 | 0.680 | 0.845 | 0.927 | 1.046 | 0.914 | 1.032 | 0.799 | 0.784 | 0.911 | 0.886 | 0.870 | 0.852 |
| $\mathrm{SS}_{2001} / \mathrm{SS}_{\text {OY }}$ | 0.560 | 0.695 | 0.549 | 0.683 | 0.782 | 0.887 | 0.772 | 0.876 | 0.644 | 0.633 | 0.732 | 0.714 | 0.702 | 0.689 |
| $\mathrm{T}_{\mathrm{m}}$ | 2005 | 2004 | 2005 | 2004 | 2004 | 2001 | 2004 | 2001 | 2004 | 2005 | 2004 | 2004 | 2004 | 2004 |
| $\mathrm{C}_{\text {recover }}$ | 2980 | 3270 | 2950 | 3279 | 3301 | 3217 | 3300 | 3220 | 2610 | 2580 | 3146 | 3140 | 3297 | 3304 |



Figure 1. Weighted, sampled age composition from the longline fishery. The age composition for 1997 is not shown, because effective sample size was about 2 fish.



Figure 2. Estimated length (cm) at age for red grouper from Goodyear (1994) and the Richards curve from Lombardi-Carlson et al. (2002). In the upper panel each curve is shown for the approximate range of ages to which the curves were fit.


Figure 3. Relative fecundity of mature, active females. Circles are observations of batch fecundity (BFE) from Collins et al. 2002. The curve labeled "Shirripa" is the power function used by Schirripa et al 1999 to express gonad weight as function of length (converted to a function of age by use of the growth curve). The curve labeled "Collins (by length)" is the exponential function given by Collins et al. 2002 to express batch fecundity as function of length (converted to a function of age by use of the growth curve). The curve labeled "Collins (by age)" is the exponential function given by Collins et al. 2002 to express batch fecundity as function of age.
proportion of red grouper that are female

proportion of red grouper females that are active


Figure 4. Top panel) Descending logistic curve fit to proportion of sampled red grouper identified as female. Bottom panel) Asymptotic curve fit to proportion of females identified as active. Data (weighted by sample size) were obtained from Moe (1969), Koenig (1993), and Collins et al. (2002) as indicated by the legend.


Figure 5. Comparison of relative fecundity at age derived by Schirripa et al. (1999) and for this assessment.


Figure 6. Proportions of longline in the commercial catch calculated from the 1999 input to ASAP and from reef fish log books (2001 input to ASAP).


Figure 7. Proportions of MRFSS estimated catches which were released alive.


Figure 8. Relative derived age composition for even numbered years from 1986-2000.


Figure 9. Indices of abundance used in the red grouper assessment.


Figure 10. Relative indices of abundance derived from the Mote Marine Laboratory tagging data assuming no tag shedding and complete mixing (circles), high tag shedding rates of $0.8 \mathrm{yr}^{-1}$ with complete mixing (squares) and no tag shedding with little mixing such that the F on the tagged population is 10 times higher than the F on the untagged population (triangles). The major difference is in 1994, where the estimates of total loss rate were lowest (hence the effect of subtracting out the high tag shedding rate is proportionately the greatest).


Figure 11. Control rule diagram contrasting the stock status estimates provided to the 2000 RFSAP (assuming 33\% mortality of longline releases) with the corresponding estimates obtained using the revised catch/discard data (assuming either $33 \%$ or $90 \%$ mortality of longline releases). Lower solid line represents OY (where $F_{O Y}=0.75 F_{M S Y}$ ). The estimated fishing mortality rate for 1997 is used for $\mathrm{F}_{\text {current }}$.


Figure 12. Comparison of time-series trajectory of estimated $\mathrm{SS}_{\mathrm{S}} \mathrm{SS}_{\mathrm{MSY}}$ (Upper) and SS (lower) across ASAP model formulations using the same model structure as applied by the RFSAP for the catch and effort data through 1997 (phase 1) and the catch-effort data through 2001 (phase 2). The line indicated as Proj represents the projected SSB from the 1999 assessment based on catch information alone. The 2 panels on the left represent steepness 0.8 while those on the right represent steepness values of 0.7

## Recruits



Figure 13. Estimates of annual recruitment (age 1) from the models used by RFSAP (2000) contrasted with the estimates from the phase 2 models presented here.


Figure 14. Control rule diagram contrasting the stock status estimates from phase 2 (data from 1986-2001, methods as in previous assessment) assuming either $33 \%$ or $90 \%$ mortality of longline releases. Lower solid line represents OY (where $F_{O Y}=0.75 F_{M S Y}$ ). The estimated fishing mortality rate for 2001 is used for $\mathrm{F}_{\text {current }}$.


Figure 15. Spawning stock estimates from the four base runs of phase 3 (using directly aged samples of the catch) projected to the year 2020 at $\mathrm{F}_{\mathrm{OY}}=0.75 \mathrm{~F}_{\mathrm{MSY}}$ levels.


Figure 16. Estimates of abundance (log scale) and fishing mortality rate from phase 3 model with steepness $=0.8$ and longline release mortality $=33 \%$.


Figure 17. Comparison of time-series trajectory of estimated SS/Ssmsy (Upper) and SSB (lower) across ASAP model formulations using the same model structure as applied used by the RFSAP for the catch and effort data through 2001, with (phase 3) and without (phase 2) the direct age sample information in the model fit (the words 'age' in the legend indicate 'with direct age samples')..


Figure 18. Control rule diagram contrasting the stock status estimates from phase 3 (data from 19862001, methods as in previous assessment plus sampled age composition) assuming either $33 \%$ or $90 \%$ mortality of longline releases. Lower solid line represents OY (where $F_{O Y}=0.75 F_{M S Y}$ ). The estimated fishing mortality rate for 2001 is used for $\mathrm{F}_{\text {current }}$.

## ADDENDUM

The RFSAP (September 2002) reviewed the status of red grouper in the Gulf of Mexico using the September 6, 2002 version of this document and requested some additional information and some additional analyses. This addendum documents additional information and analyses provided to the RFSAP during and after the September 17-20 meeting in Miami.

Inputs
Estimated yield by sector is presented in Table A1.
The sampled age composition (from Lombardi et al. 2002) and the derived age composition (from catch, size composition and an ageing algorithm) are contrasted in Figure 1.

The weighting factors used in the ASAP runs are presented in Table A2. All but the weighting factors for the sampled age composition were the same as used for the runs for the RFSAP 2000 management advice unless otherwise noted in the text (some sensitivity runs). The sampled age composition was not available for the previous assessment, so those weighting factors were derived from as described in the main part of this report.

Assessment
The RFSAP (September, 2002) reviewed the results of the September 6, 2002 version of the above assessment and noted that they were most sensitive to the form of the fecundity at age vector. They agreed that it was reasonable to use the most recent information on age composition (Lombardi et al., 2002) and reproduction (Collins et al, 2002). However, concern was expressed regarding the most appropriate use of the new reproductive information. Accordingly, the RFSAP requested two additional sets of runs patterned after the base model described above but using either the batch fecundity at age relationship (FECvsAGE) discussed earlier or a new gonad weight at age relationship (GWTvsAGE) derived at the meeting (for details see Figures A2 and A3, Table A3). The release mortality for the longline fleet was assumed to be $33 \%$ and steepness was assumed to be either 0.7 or 0.8 (the RFSAP continues to believe there is little basis for deciding between the two steepness values).

The stock-status estimates obtained under the two alternative reproductive scenarios are compared to those of the original 'base' run in Figure A4 and Table A4. The base run, which used the FECvsTL vector, gave the most optimistic picture of stock status (not overfished and no overfishing regardless of steepness). The runs using the GWTvsAGE and FECvsAGE vectors were somewhat less optimistic, suggesting the stock was not overfished, but that overfishing was occurring (with steepness values of 0.7 or less). This result was somewhat unexpected inasmuch as the FECvsAGE vector ascribes relatively more productivity to age 15 and older animals than either of the competing fecundity vectors (see Figure 5 ) and it is the older animals which are estimated to have declined the most. However, even under virgin conditions only a small proportion of fish survive beyond age 15 where this divergence in relative fecundity estimates occurs, hence these older age groups contribute little to the overall fecundity of the stock (Figure A5). Instead, the assessment is most sensitive to the differences in the modeled fecundity of younger fish. A comparison of the four fecundity curves in Figure A3 reveals that the curves based on the new fecundity data (Collins et al., 2002) ascribe relatively greater productivity to ages 3-5 than the GWTvsTL curve derived by Schirripa et al. (1999). Of the new curves, the FECvsTL vector ascribes the most productivity to ages $3-5$, which are the age groups that have declined the least, therefore it leads to the most optimistic outcome.

The RFSAP believed that, of the four vectors examined, the GWTvsAGE fecundity vector was the most appropriate. It was preferred over the original GWTvsTL vector derived by Schirripa et al. (1999) because it incorporated additional data (Collins et al, 2002) and avoided the possible inaccuracies that may occur when converting length to age via a growth curve. It was favored over the corresponding batch fecundity curves because there were a great deal more observations of gonad weight than batch fecundity, making the relationship with age better determined, and also to maintain some continuity with the method used in the last assessment. Accordingly, the RFSAP requested projections of the constant harvest and constant effort scenarios that would lead to recovery above $\mathrm{SS}_{\text {MSY }}$ by 2012. These are shown in Figures A6 and A7, respectively.

Finally, there was some discussion regarding the danger of defining control rules in terms of benchmarks such as

MSY that depend upon the relationship between fecundity and age when that relationship is not well known. Concern was expressed that information collected in the future may lead to a different perception of the relative fecundity of younger red grouper, which in turn might lead to a different perception of stock status and different management advice. It was pointed out that alternative benchmarks based on yield per recruit theory, such as $\mathrm{F}_{0.1}$ and $\mathrm{SS}_{\mathrm{F} 0.1}$, were less variable across model formulations than the MSY related statistics (see Tables 20 and A4). In the present case, an $\mathrm{F}_{0.1}$-based control rule would suggest that the Gulf of Mexico red grouper stock is overfished and that overfishing is occurring regardless of the choice of fecundity vector (see Figure A8, Table A4). Nevertheless, the stock is estimated to recover to $\mathrm{SS}_{\mathrm{F} 0.1}$ levels by 2012 if F is reduced to slightly less than $\mathrm{F}_{0.1}$ (Figure A9).

Table A1. Red grouper landed (longline, handline+ and headboat) and harvested (charter and private+shore) yield. Handline+ indicates handline and other commercial gears exclusive of bottom and vertical longlines. Unless otherwise noted, data are in metric tons (mt). Headboat landings for 2000 and 2001 were estimated from recent years.

|  |  |  | private+ <br> shore | longline | handline+ | total |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  | mt | 1000 lb |
| 1986 | 57 | 142 | 866 | 1,202 | 1,735 | 4,003 | 8824 |
| 1987 | 45 | 83 | 444 | 1,601 | 1,529 | 3,701 | 8158 |
| 1988 | 48 | 108 | 975 | 956 | 1,264 | 3,351 | 7387 |
| 1989 | 65 | 86 | 980 | 1,289 | 2,156 | 4,577 | 10090 |
| 1990 | 54 | 206 | 373 | 1,225 | 972 | 2,830 | 6239 |
| 1991 | 29 | 34 | 727 | 1,254 | 1,075 | 3,120 | 6879 |
| 1992 | 26 | 151 | 1,153 | 663 | 1,267 | 3,260 | 7187 |
| 1993 | 37 | 82 | 875 | 1,960 | 943 | 3,897 | 8591 |
| 1994 | 27 | 108 | 753 | 1,237 | 1,004 | 3,129 | 6899 |
| 1995 | 45 | 260 | 684 | 1,128 | 1,036 | 3,152 | 6950 |
| 1996 | 40 | 81 | 307 | 1,364 | 665 | 2,458 | 5419 |
| 1997 | 16 | 78 | 196 | 1,434 | 783 | 2,507 | 5527 |
| 1998 | 19 | 85 | 233 | 1,302 | 502 | 2,141 | 4719 |
| 1999 | 25 | 73 | 422 | 1,783 | 918 | 3,220 | 7098 |
| 2000 | 29 | 91 | 646 | 1,346 | 1,290 | 3,402 | 7499 |
| 2001 | 29 | 123 | 450 | 1,546 | 1,094 | 3,242 | 7147 |

Table A2. Likelihood distributions and weighting factors $(\lambda)$ used in the present ASAP analyses. Values of $\lambda$ differ in some cases from those presented in Table 59 of Schirripa et al. (1999), however they are the same as the values used for the runs upon which the 2000 RFSAP based its advice. In the case of catch and discard at age proportions, the $\lambda$ 's are effective sample sizes. In the case of other data types, they are effectively the reciprocal of the variance

| Component | nobs | $\lambda$ | distribution |
| :---: | :---: | :---: | :---: |
| Total catch in weight |  |  |  |
| Commercial Longline | 16 | 100.5 | Lognormal |
| Commercial Other | 16 | 100.5 | Lognormal |
| Recreational | 16 | 100.5 | Lognormal |
| Total discards in weight |  |  |  |
| Commercial Longline | 16 | 25 | Lognormal |
| Commercial Other | 16 | 25 | Lognormal |
| Recreational | 16 | 25 | Lognormal |
| Derived age composition of catch |  |  |  |
| Commercial Longline | $16 \times 20$ | 100 | Multinomial |
| Commercial Other | $16 \times 20$ | 200 | Multinomial |
| Recreational | $16 \times 20$ | 150 | Multinomial |
| Derived age composition of discards |  |  |  |
| Commercial Longline | $16 \times 20$ | 25 | Multinomial |
| Commercial Other | $16 \times 20$ | 25 | Multinomial |
| Recreational | $16 \times 20$ | 100 | Multinomial |
| Sampled age composition of catch |  |  |  |
| Commercial Longline | 10x20 | annual values | Multinomial |
| Commercial Other | 10x20 | annual values | Multinomial |
| Recreational | 10x20 |  | Multinomial |
| Indices of abundance |  |  |  |
| Trap | 6 |  |  |
| Video | 6 lognormal distributed with |  |  |
| MRFSS | 16 |  |  |
| Logbook - Handline | $8 \lambda=1 / \log _{e}\left(1+C V^{2}\right)$ |  |  |
| Logbook - Longline | 8 㖪 |  |  |
| Logbook - Trap | 8 (annual values of $C V$ shown |  |  |
| Tagging | 7 in Table 17) |  |  |
| US Historical | 16 |  |  |
| Catchability deviations |  |  |  |
| Trap | 6 | 10000 | Lognormal |
| Video | 6 | 10000 | Lognormal |
| MRFSS | 16 | 10000 | Lognormal |
| Logbook - Handline | 8 | 10000 | Lognormal |
| Logbook - Longline | 8 | 10000 | Lognormal |
| Logbook - Trap | 8 | 10000 | Lognormal |
| Tagging | 7 | 10000 | Lognormal |
| US Historical | 16 | 10000 | Lognormal |
| Fmult deviations |  |  |  |
| Commercial Longline | 15 | 7 | Lognormal |
| Commercial Other | 15 | 11 | Lognormal |
| Recreational | 15 | 11 | Lognormal |
| Selectivity deviations |  |  |  |
| Commercial Longline | 14 | 1000 | Lognormal |
| Commercial Other | 14 | 1000 | Lognormal |
| Recreational | 14 | 1000 | Lognormal |
| N at age in year 1 | 19 | 4.48 | Lognormal |
| Stock-recruitment fit | 16 | 4.48 | Lognormal |
| Recruitment deviations | 16 | 8.96 | Lognormal |
| Selectivity curvature over age | 720 | 1000 | Lognormal |
| Selectivity curvature over time | 840 | 100.5 | Lognormal |

Table A3. Summary of reproductive parameters used to construct the GWTvsAGE vector derived during the September 2002 RFSAP meeting. Values for per capita fecundity of age 20 plus-group were computed by prorating the values for ages 20 to 30 assuming a total mortality rate of $0.4 \mathrm{yr}^{-1}$ (as done by Schirripa et al. 1999). The per capita fecundity is the product of the proportion female and gonad weight (the RFSAP elected not to incorporate the percent activity estimates, believing that the use of gonad weight observations implicitly included to some degree a measure of activity). The relative per capita fecundity is the per capita fecundity scaled by the average over all ages.

| Age | proportion <br> female | gonad weight <br> by age | relative per <br> capita fecundity |
| ---: | :---: | :---: | :---: |
| 1 | 0.940 | 0.000 | 0.000 |
| 2 | 0.929 | 0.000 | 0.000 |
| 3 | 0.916 | 27.130 | 0.148 |
| 4 | 0.901 | 44.574 | 0.239 |
| 5 | 0.884 | 65.513 | 0.344 |
| 6 | 0.864 | 89.740 | 0.461 |
| 7 | 0.842 | 117.092 | 0.585 |
| 8 | 0.816 | 147.439 | 0.715 |
| 9 | 0.788 | 180.674 | 0.845 |
| 10 | 0.756 | 216.704 | 0.973 |
| 11 | 0.721 | 255.449 | 1.094 |
| 12 | 0.683 | 296.840 | 1.205 |
| 13 | 0.643 | 340.813 | 1.302 |
| 14 | 0.601 | 387.314 | 1.382 |
| 15 | 0.557 | 436.290 | 1.443 |
| 16 | 0.512 | 487.696 | 1.483 |
| 17 | 0.467 | 541.488 | 1.501 |
| 18 | 0.423 | 597.628 | 1.499 |
| 19 | 0.379 | 656.079 | 1.477 |
| 20 | 0.338 | 716.807 | $* 1.308 *$ |

*plus group

Table A4. Comparison of stock status estimates from phase 3 model (1986-2001data using methods of previous assessment plus sampled age composition, $33 \%$ mortality of longline releases) using each of the three new fecundity vectors. The quantities $M S Y$, $O Y\left(F_{O Y}=0.75 F_{M S Y}\right)$, and $C_{\text {recover }}$ (catch expected to permit recovery to $S S_{M S Y}$ by 2012) are in MT gutted weight. Spawning stock $S S$ in the case of the GWTvsAGE model is in MT gonad weight, otherwise its is dimensionless (relative fecundity).

|  | FECvsTL |  | FECvsAGE |  | GWTvsAGE |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{h}=0.7$ | $\mathrm{h}=0.8$ | $\mathrm{h}=0.7$ | $\mathrm{h}=0.8$ | $\mathrm{h}=0.7$ | $\mathrm{h}=0.8$ |
| MSY | 3374 | 3281 | 3381 | 3276 | 3429 | 3295 |
| OY | 3301 | 3217 | 3308 | 3212 | 3350 | 3229 |
| OY/MSY | 0.978 | 0.981 | 0.978 | 0.981 | 0.977 | 0.980 |
| $\mathrm{F}_{0.1}$ | 0.238 | 0.238 | 0.238 | 0.238 | 0.238 | 0.238 |
| $\mathrm{F}_{\text {max }}$ | 0.476 | 0.476 | 0.476 | 0.476 | 0.476 | 0.476 |
| $\mathrm{F}_{30 \%}$ | 0.702 | 0.703 | 0.581 | 0.581 | 0.563 | 0.563 |
| $\mathrm{F}_{40 \%}$ | 0.437 | 0.437 | 0.343 | 0.344 | 0.354 | 0.354 |
| $\mathrm{F}_{\text {MSY }}$ | 0.329 | 0.382 | 0.312 | 0.371 | 0.306 | 0.364 |
| $\mathrm{F}_{\mathrm{OY}}$ | 0.247 | 0.286 | 0.234 | 0.278 | 0.229 | 0.273 |
| $\mathrm{F}_{\mathrm{OY}} / \mathrm{F}_{\mathrm{MSY}}$ | 0.750 | 0.750 | 0.750 | 0.750 | 0.750 | 0.750 |
| $\mathrm{F}_{2001}$ | 0.316 | 0.317 | 0.315 | 0.317 | 0.315 | 0.316 |
| $\mathrm{F}_{2001} / \mathrm{F}_{\mathrm{MSY}}$ | 0.961 | 0.829 | 1.012 | 0.853 | 1.031 | 0.869 |
| $\mathrm{F}_{2001} / \mathrm{F}_{\mathrm{OY}}$ | 1.281 | 1.105 | 1.350 | 1.137 | 1.374 | 1.159 |
| $\mathrm{F}_{2001} / \mathrm{F}_{0.1}$ | 1.329 | 1.332 | 1.328 | 1.332 | 1.327 | 1.331 |
| $\mathrm{SS}_{\text {MSY }}$ | 5746 | 5086 | 4585 | 3973 | 840 | 715 |
| $\mathrm{SS}_{\text {OY }}$ | 6816 | 5998 | 5482 | 4704 | 1015 | 859 |
| $\mathrm{SS}_{\mathrm{F} 0.1}$ | 6953 | 6595 | 5430 | 5121 | 993 | 930 |
| $\mathrm{SS}_{\mathrm{OY}} / \mathrm{SS}_{\mathrm{MSY}}$ | 1.186 | 1.179 | 1.196 | 1.184 | 1.208 | 1.201 |
| $\mathrm{SS}_{2001}$ | 5327 | 5321 | 4150 | 4146 | 705 | 704 |
| $\mathrm{SS}_{2001} / \mathrm{SS}_{\text {MSY }}$ | 0.927 | 1.046 | 0.905 | 1.043 | 0.840 | 0.985 |
| $\mathrm{SS}_{2001} / \mathrm{SS}_{\text {OY }}$ | 0.782 | 0.887 | 0.757 | 0.881 | 0.695 | 0.820 |
| $\mathrm{SS}_{2001} / \mathrm{SS}_{\mathrm{F} 0.1}$ | 0.766 | 0.807 | 0.764 | 0.810 | 0.710 | 0.758 |
| $\mathrm{T}_{\text {min }}$ | 2004 | 2001 | 2004 | 2001 | 2004 | 2002 |
| $\mathrm{C}_{\text {recover }}$ | 3301 | 3217 | 3230 | 3212 | 3190 | 3229 |



Figure A1. Comparison of sampled (left bar in each age) and derived age composition for red grouper caught by longline. The sampled age composition is weighted to account for geographic differences in the geographic distribution of the samples and longline catches.


Figure A2. Least-squares fits of the power function $\left(G W T=4.0739\right.$ age $\left.{ }^{1.7259}\right)$ and exponential function (46.130 $\left.e^{0.13425 a g e}\right)$ to gonad weight at age data collected by Collins et al (2002). The power function gave the best fit (lowest residual sum of squares). Only gonad weights of active females (vitellogenic or hydrated ova present, but excluding spent ovaries) sampled during the peak spawning months (March, April and May) were used. This is essentially an update of the data from the previous assessment (Schirripa and Legault 1999, Figure 8), but instead the curve fit is based only on those females whose ages were determined directly (Lombardi-Carlson et al. 2002) rather than estimated from length by use of a growth curve.


Figure A3. Comparison of the four measures of per capita fecundity on a relative scale. Note that the vectors based on gonad weight do not include estimates of the percent of females active in any given spawning year (the RFSAP felt that percent activity was, to some degree, implicitly accounted for in that gonad weight would be lower for inactive females).


Figure A4. Control rule diagram contrasting the stock status estimates from the runs requested by the 2002 RFSAP. Lower solid line represents control rule for management based on an OY target (point marked by the 'happy face'). $\mathrm{F}_{\text {current }}$ is the apical (fully-selected) fishing mortality rate in 2001. Vertical dashed line represents the minimum stock size threshold (0.8).


Figure A5. Relative fecundity of each age class under virgin condition assuming a natural mortality rate of $0.2 \mathrm{yr}^{-1}$ (this is the product of the per capita fecundity at age and the relative abundance of each age class, normalized to average a value of 1.0).

## Steepness $=0.7$





Steepness $=0.8$


Figure A6. Projections of landings and aggregate fecundity (spawning stock) under the constant level of landings that will permit recovery to $\mathrm{SS}_{\text {MSY }}$ by 2012 assuming fecundity at age follows the GWTvsAGE vector. In the case where steepness $=0.7, \mathrm{C}_{\text {recover }}=3190 \mathrm{mt}$ (gutted weight). However, if steepness $=0.8$ the stock is estimated to recover even if C > MSY, therefore a projection with $\mathrm{C}_{\text {recover }}$ set to OY ( $=3229$ mt ) is shown instead. Dashed lines represent equilibrium yield (landings) and spawning stock corresponding to $\mathrm{F}_{\mathrm{MSY}}$. All projections assumes the landings in 2002 are equal to the average from 2000 and 2001, future selectivity $=$ estimated selectivity for 2001, and future recruitment is a deterministic Beverton and Holt function of projected spawning stock.

## Steepness $=0.7$





Steepness $=0.8$


Figure A7. Projections of landings and aggregate fecundity (spawning stock) under the constant level of fishing mortality rate that permits recovery to $\mathrm{SS}_{\text {MSY }}$ by 2012 assuming fecundity at age follows the GWTvsAGE vector. In the case where steepness $=0.7, \mathrm{~F}_{\text {recover }}=0.298$. However, if steepness $=0.8$ the stock is estimated to recover even if $\mathrm{F}>\mathrm{F}_{\mathrm{MSY}}$, therefore a projection with $\mathrm{F}_{\text {recover }}$ set to $\mathrm{F}_{\mathrm{OY}}(=0.273)$ is shown instead. Dashed lines represent equilibrium yield (landings in gutted weight) and spawning stock corresponding to $\mathrm{F}_{\text {MSY }}$. All projections assume the landings in 2002 are equal to the average from 2000 and 2001, future selectivity $=$ estimated selectivity for 2001, and future recruitment is a deterministic Beverton and Holt function of projected spawning stock.


Figure A8. Control rule diagram expressed in terms of $\mathrm{F}_{0.1}$ benchmarks. Lower line would represent $\mathrm{F}_{\mathrm{OY}}$ $=0.75 \mathrm{~F}_{0.1} \cdot \mathrm{~F}_{\text {current }}$ is the apical (fully-selected) fishing mortality rate in 2001. Vertical dashed line is the minimum stock size threshold (0.8).

## Steepness $=0.7$

Steepness $=0.8$


Figure A9. Projections of landings and aggregate fecundity (spawning stock) under a constant level of fishing mortality rate equal to $\mathrm{F}_{0.1}\left(0.238 \mathrm{yr}^{-1}\right)$ assuming fecundity at age follows the GWTvsAGE vector. Dashed lines represent equilibrium yield (landings in gutted weight) and spawning stock corresponding to $\mathrm{F}_{0.1}$. All projections assume the landings in 2002 are equal to the average from 2000 and 2001, future selectivity = estimated selectivity for 2001, and future recruitment is a deterministic Beverton and Holt function of projected spawning stock

