

Research



Cite this article: Korczak-Abshire M, Hinke JT, Milinevsky G, Juárez MA, Watters GM. 2021 Coastal regions of the northern Antarctic Peninsula are key for gentoo populations. *Biol. Lett.* **17**: 20200708. <https://doi.org/10.1098/rsbl.2020.0708>

Received: 5 October 2020

Accepted: 4 January 2021

Subject Areas:

ecology, environmental science

Keywords:

Pygoscelis papua, climate change, range expansion, satellite telemetry

Author for correspondence:

Malgorzata Korczak-Abshire
e-mail: mka@ibb.waw.pl

Electronic supplementary material is available online at <https://doi.org/10.6084/m9.figshare.c.5271240>.

Coastal regions of the northern Antarctic Peninsula are key for gentoo populations

Malgorzata Korczak-Abshire¹, Jefferson T. Hinke², Gennadi Milinevsky^{3,4}, Mariana A. Juárez^{5,6} and George M. Watters²

¹Institute of Biochemistry and Biophysics, Polish Academy of Sciences, 02-106 Warsaw, Poland

²Antarctic Ecosystem Research Division, Southwest Fisheries Science Center, National Marine Fisheries Service, National Oceanic and Atmospheric Administration, La Jolla, CA 92037, USA

³Department of Atmospheric Physics and Geospace, National Antarctic Scientific Center of Ukraine, Kyiv 01601, Ukraine

⁴Physics Faculty, Taras Shevchenko National University of Kyiv, Kyiv 01033, Ukraine

⁵Departamento Biología de Predadores Tope, Instituto Antártico Argentino, San Martín, Buenos Aires B1650CSP, Argentina

⁶National Scientific and Technical Research Council (CONICET), Ciudad Autónoma de Buenos Aires, C1425FQB, Argentina

MK-A, 0000-0001-7695-0588; JTH, 0000-0002-3600-1414

Southern Ocean ecosystems are rapidly changing due to climate variability. An apparent beneficiary of such change in the western Antarctic Peninsula (WAP) is the gentoo penguin *Pygoscelis papua*, which has increased its population size and expanded its range southward in the last 20 years. To better understand how this species has responded to large-scale changes, we tracked individuals during the non-breeding winter period from five colonies across the latitudinal range of breeding sites in the WAP, including from a recently established colony. Results highlight latitudinal gradients in movement; strong associations with shallow, coastal habitats along the entire Antarctic Peninsula; and movements that are independent of, yet constrained by, sea ice. It is clear that coastal habitats essential to gentoo penguins during the breeding season are similarly critical during winter. Larger movements of birds from northern colonies in the WAP further suggest that leap-frog migration may influence colonization events by facilitating nest-area prospecting and use of new haul-out sites. Our results support efforts to develop a marine protected area around the WAP. Winter habitats used by gentoo penguins outline high priority areas for improving the management of the spatio-temporally concentrated krill (*Euphausia superba*) fishery that operates in this region during winter.

1. Introduction

Climate change fundamentally alters the structure and function of marine ecosystems by modifying ocean productivity, altering food-web dynamics and shifting species distributions [1]. Polar ecosystems are especially sensitive to climate perturbations because they are largely structured by the seasonal dynamics of sea ice [2], which have exhibited trends in extent and duration in both hemispheres [3]. An area of particular concern is the western Antarctic Peninsula (WAP), where increases in air and sea-surface temperatures over the last 40 years have reduced regional sea ice extent and duration [4]. Such physical perturbations are associated with changes in the distribution, abundance and survival of several species in the WAP ecosystem [5]. Further compounding risk to this ecosystem is the expanding fishery for Antarctic krill (*Euphausia superba*) [6,7], the largest, by mass, in the Southern Ocean [8].

Table 1. Tagging locations and mean deployment durations (range in parentheses). Tagging locations were at Lions Rump (LRP), Stranger Point (SPS), Cape Shirreff (CAS), Cierva Cove (CVA) and the Argentine Islands (AIS). All tags were released between 4 February and 29 March 2017.

age class	colony	longitude	latitude	<i>N</i>	duration (d)	maximum distance (km)
adult	LRP	−58.13	−62.14	15	105 (28–148)	109 (32–229)
	SPS	−58.62	−62.27	13	79 (45–157)	128 (29–186)
	CAS	−60.80	−62.46	9	100 (45–194)	151 (27–251)
	CVA	−60.98	−64.14	10	92 (16–221)	63 (28–141)
	AIS	−64.25	−65.24	14	126 (57–306)	62 (21–190)
juvenile	CAS	−60.80	−62.46	5	30 (12–76)	43 (23–113)
	CVA	−60.98	−64.14	5	28 (9–85)	63 (30–112)

Seabirds are important indicators of ecosystem status and are among the species impacted by climate change and fisheries [7,9–11]. Changes in population sizes and phenologies due to environmental variation in the Southern Hemisphere are evident [9,12–14]. For example, in the WAP, populations of ice-dependent Adélie penguins (*Pygoscelis adeliae*) and ice-tolerant chinstrap penguins (*P. antarcticus*) have declined [12,15], while the abundance and range of ice-avoiding gentoo penguins (*P. papua*) have increased [12]. Notably, range expansion and rapid population growth of gentoo penguins is occurring at the southern margin of their breeding range, where at least seven newly established colonies have been identified in last 20 years (electronic supplementary material, figure S1) [16]. Despite divergent population trends among the pygoscelid penguins, all three species have been affected by recent krill fishing during the non-breeding period [7], hereafter winter. Thus, although generally considered to be climate ‘winners’ [17], the risks to gentoo penguin populations should be further assessed to better inform conservation and management actions.

Across the WAP, gentoo penguins typically forage within 20 km of breeding sites during the austral summer [18–22]. During winter, gentoos are not constrained by the need to provision chicks and can undertake longer range movements. Prior tracking studies from the South Shetland Islands [18,21] suggested winter movements up to 10 times farther than during summer. Such dispersal to distant foraging areas is the primary mechanism by which range expansion could occur [23], but a lack of tracking data from colonies throughout the WAP limits understanding of how this seabird distributes during the winter and whether there is variation in movement among colonies. Given rapidly changing environmental and anthropogenic drivers in the WAP, identifying winter movements and patterns of habitat use by gentoo penguins are also useful for assessing population status and risks to the species. We therefore tracked the winter movements of gentoo penguins from five colonies across the latitudinal range of this species in the WAP, including from a recently established colony [24] near the southern limit of the species’ range.

2. Material and methods

We tracked 10 fledgling and 65 post-moult adult gentoo penguins from five colonies across the latitudinal range of breeding colonies in the WAP from February 2017 through January 2018 (table 1; electronic supplementary material, figure S1c). We used Sirtrack Kiwisat-202K2G-172A satellite

transmitters (60 × 27 × 17 mm, 34 g) to track fledglings and Wildlife Computers Spot-275 satellite transmitters (86 × 17 × 18 mm, 38 g) to track adults. All birds were captured on beaches and the transmitters were affixed to back feathers using glue and cable ties [25]. Transmitters were scheduled to transmit daily from 12:00 to 18:00 UTC, corresponding to daylight hours when birds should be foraging.

We processed raw location estimates by removing four adult deployments that were tracked less than 7 days, and all erroneous location estimates indicated by ‘Z’ quality codes or unspecified ellipse errors. Next, we applied a speed filter [26] assuming a conservative swim speed of 2.5 m/s. Remaining tracks were smoothed with a state-space model [27] using the R [28] package ‘crawl’ [29]. Model fits were used to generate 100 alternative tracks for each deployment, with locations estimated every 2 h. Alternative tracks were pooled and mapped to hexagonal polygons with centroids spaced 15 km apart (area ≈ 87 km²) to estimate habitat utilization distributions (HUDs) using the R package ‘crawl’ [30]. This spatial scale approximates daily foraging ranges by gentoo penguins during the breeding season [25].

We used several physical variables to compare the habitats used by birds from different colonies. We extracted bottom depths along each track from the ETOPO1 dataset [31]. We estimated distance to the nearest point of land using the ‘wrld_simpl’ database in the R [28] package ‘mapproj’ [32]. To examine near real-time experience of sea-surface temperatures (SST) and sea ice concentrations (SIC, expressed as per cent cover), we matched raw position estimates with daily SSTs from the multi-scale ultra-high-resolution SST data resolved on a 1 km grid [33] and daily SICs from the EUMETSAT Ocean and Sea Ice Satellite Application Facility, projected from a native 10 km grid to a 1 km grid [33].

We used the Tukey honest significant difference (HSD) to identify colony-level differences in movement and linear mixed-effects models using the R [28] package ‘lme4’ [34] to identify colony-level differences in physical habitat variables. We fitted separate models for each physical variable and included month as a fixed effect in all models to account for potential seasonal trends. Individuals were treated as random effects.

3. Results

Positions (*N* = 29 119) of fledgling and adult penguins were, respectively, reported for an average of 29 days (range: 9–86 days; table 1) and 100 days (range: 16–306 days). Maximum distances from tagging sites varied by colony (table 1). Gentoo penguins originating from the northern edge of their range in the WAP dispersed farther and with significantly greater shifts to the south (Tukey HSD $F_{4,66} = 12.8$,

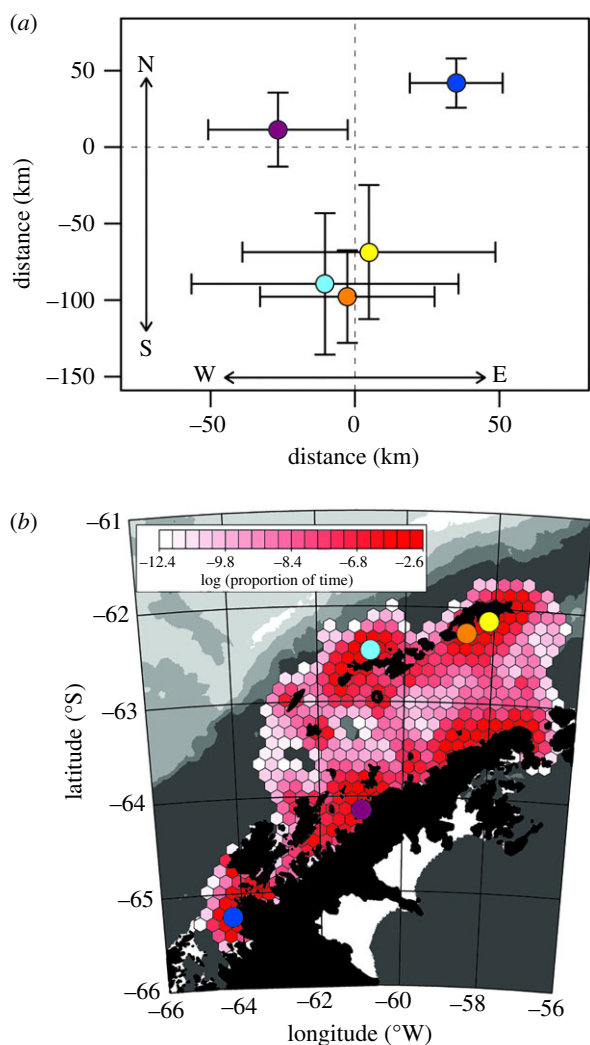


Figure 1. (a) Mean and 95% confidence intervals for positional shifts for birds from the Argentine Islands (dark blue), Cierva Cove (purple), Cape Shirreff (light blue), Stranger Point (orange) and Lions Rump (yellow). (b) HUDs for all tracked birds. Colony locations are indicated with circles coloured to match panel (a).

$p < 0.01$) than adults tagged at colonies further southwest (figure 1a). Longitudinal movements were not different across colonies (figure 1a, Tukey HSD $F_{4,66} = 1.9$, $p = 0.11$). Given high overlap of fledglings and adults (electronic supplementary material, figure S2), all tracks were pooled for further analyses.

The physical habitats encountered during winter were largely similar across colonies (electronic supplementary material, figure S3). The model-predicted distance from shore averaged 6.4 km with little variation among colonies (table 2). Near shore areas corresponded to shallow habitats that averaged 42 m deep across all colonies (table 2), noting that birds from Cape Shirreff and the Argentine Islands had the shallowest habitats. Birds from Cape Shirreff, with a more northern distribution for much of the winter, typically encountered warmer water than birds from other colonies (table 2). Responses to the distribution of SIC were not colony-specific when ice was encountered (table 2), and all birds usually occupied ice-free waters (figure 2).

The aggregate HUD for all tracks during winter (figure 1b) demonstrates an affinity for coastal areas. Winter HUDs were concentrated near tagging sites, around islands in the Bransfield Strait, and along the entire margin of the WAP from the

Argentine Islands to the tip of the Peninsula (figure 1b). Overlap of colony-level HUDs (figure 2) was common. This was particularly evident for birds from Stranger Point and Lions Rump, which overlapped extensively around King George/25 de Mayo Island and along the northern tip of the Peninsula from 59° W to 57° W. Similarly, birds from Cape Shirreff and Cierva Cove overlapped extensively along the central WAP from 63° W to 59° W. The HUDs for gentoo penguins from the Argentine Islands, the most southern colony, were largely isolated from other colonies.

Differences in colony-level movements from February to April occurred prior to the presence of sea ice near any colony (figure 2), demonstrating that sea ice was not the main driver of colony-specific, over-winter dispersal patterns. Nonetheless, the evolution of dense sea ice ($SIC > 50\%$) did affect the distribution of the birds during winter. For example, the HUDs of birds from Lions Rump contracted between June and July when high SICs blanketed the eastern Bransfield Strait. Likewise, birds from the Argentine Islands shifted northeast into the Gerlache Strait coincident with the expansion of dense sea ice in waters south of Anvers Island from August to October (figure 2). However, $SIC < 50\%$ did not preclude gentoo occupation of those areas (figure 2).

4. Discussion

We report clear latitudinal gradients in winter movements of gentoo penguins from five colonies of varying population sizes and trends along the WAP. Dispersal distances were larger for birds from northern colonies than from southern colonies. The HUDs of birds from different colonies overlapped in the relatively shallow, ice-free coastal margin of the WAP. Of particular importance were the coastal regions of the Antarctic Peninsula from the Argentine Islands to the tip of the Antarctic Peninsula. Our multi-colony tracking study demonstrates that, in the WAP, the coastal habitat essential to breeding gentoo penguins during summer is similarly critical during winter. The observed gradient in movement patterns suggests that birds from northern colonies are immigrant sources for the current range expansion of this species.

Marine top predators are often expected to change foraging behaviours, movement patterns and at-sea distributions [35] in response to climate-driven changes in prey distribution [9], rather than to direct changes in their physical environment. In the WAP, however, sea ice dynamics can fundamentally alter the availability of foraging habitats. As an 'ice-intolerant' species [36], we expected the movement patterns of gentoo penguins to be driven by avoidance of developing sea ice. However, all long-distance dispersal observed here was initiated before the presence of sea ice at study colonies. Indeed, winter movements of gentoo penguins in the Falkland/Malvinas Islands, where sea ice does not exist, were even farther than observed in the WAP [37]. Thus, sea ice is not the main driver of differences in colony-specific dispersal patterns among gentoo penguins. As sea ice extent and duration in the WAP are expected to decline under most climate-change scenarios [38], the observed latitudinal gradients in movement and the affiliation of gentoo penguins with coastal regions along the WAP may be expected in the future.

Dispersal of seabirds from breeding sites to winter foraging areas must be driven by reliable availability of prey.

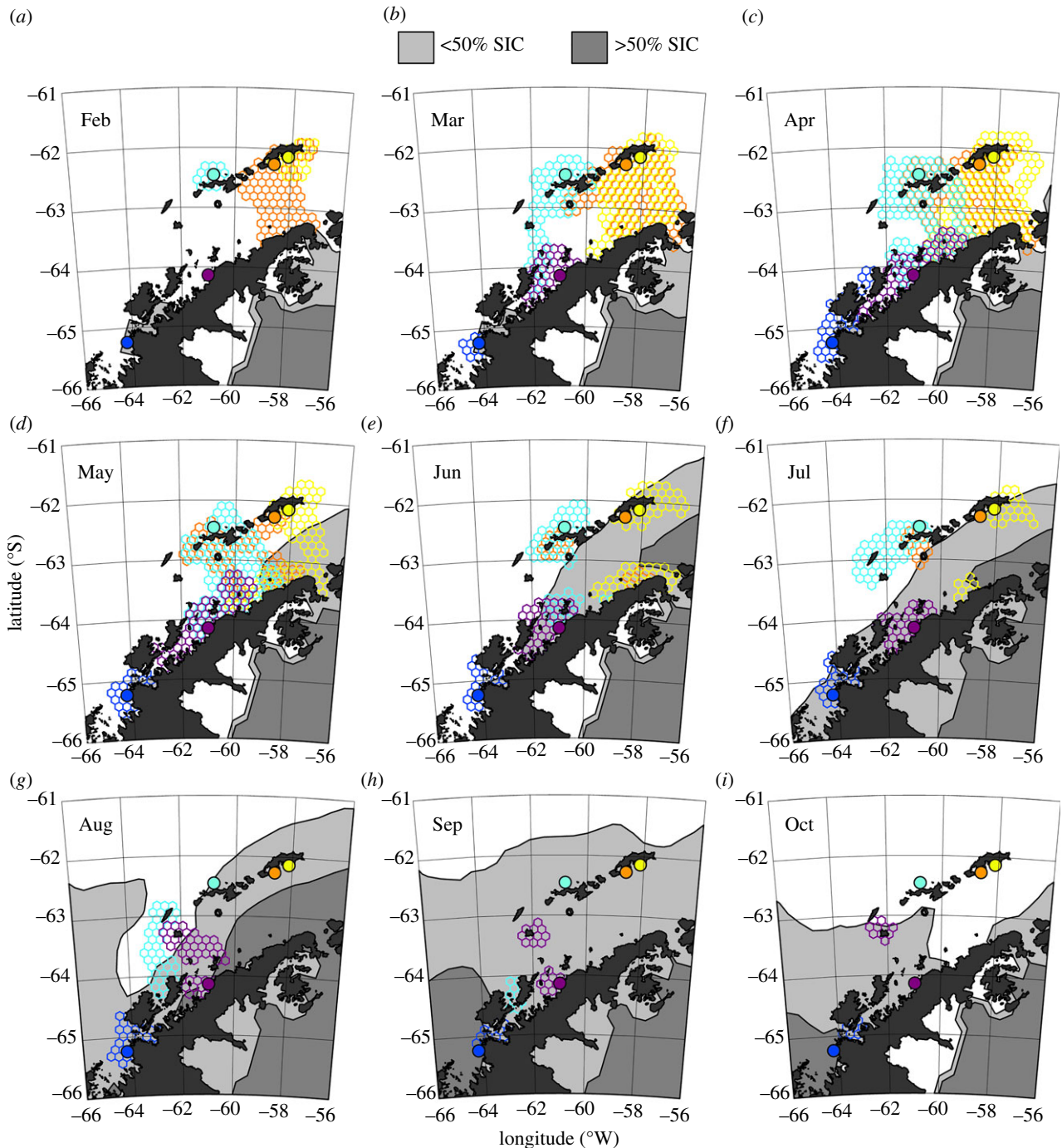


Figure 2. Monthly extents of colony-specific HUDs overlaid on monthly SIC. Colony locations are identified by filled circles. Colony locations and their HUDs are coloured as in figure 1.

Table 2. Mixed-effect model predictions for the fixed-effects (95% CI) of colony origin on distance from shore, depth, sea-surface temperatures (SST) and sea ice concentrations (SIC) for all tracked penguins.

colony	distance to shore (km)	depth (m)	SST (°C)	SIC (%)
LRP	7.6 (3.9–11.5)	52 (32–86)	1.03 (0.74–1.32)	39 (25–51)
SPS	6.0 (2.1–9.8)	67 (41–111)	1.0 (0.7–1.29)	38 (29–47)
CAS	6.3 (2.4–10.1)	24 (14–40)	1.62 (1.33–1.91)	34 (22–46)
CVA	4.3 (0.5–8.1)	53 (32–88)	0.83 (0.54–1.12)	37 (24–49)
AIS	7.9 (4.2–11.8)	18 (11–30)	1.06 (0.77–1.35)	45 (33–57)

Gentoo penguins typically exhibit benthic and pelagic foraging dives (less than 150 m depths) in coastal regions [20] with diets of crustaceans (mainly Antarctic krill), other

invertebrates and fishes that can vary by location [36]. Within the WAP, shorter dispersal ranges of gentoos from southern colonies relative to northern colonies suggest

greater food availability along the margins of the Antarctic Peninsula relative to the south Shetland Islands during winter. The continental shelves of the WAP, over which the HUDs were concentrated, are known to harbour high krill densities during winter (e.g. [39,40]), consistent with expectations of a southward contraction of krill distributions over larger scales [41]. These shallow, coastal areas also provide gentoo penguins with access to benthic and demersal resources, which are suspected to be important components of their winter diets [42].

The tracking data reported here suggest source populations for the ongoing range expansion of gentoo penguins in the WAP. Genetic analysis indicates that basin-scale dispersal and colonization events are rare for this species, and the Polar Front is an effective boundary between sub-Antarctic and Antarctic populations [23,43]. At smaller spatial scales, gentoo penguins are well-known colonizers of new breeding territory and quickly take advantage of ice-free breeding space [36]. Recent colonization events have been attributed to emigration from colonies at the southwestern edge of this species' range [12,44]. However, longer distance movements of birds from northern colonies suggest that rare dispersal events to the south could be instigated by birds from northern colonies. Such a leap-frog migration strategy has been reported for other seabirds [45]. Differential movement patterns of birds from different breeding colonies may be driven by variation in prey availability at breeding colonies but shared foraging habitat preferences that favour gradients in directed movement [45,46]. While we cannot test this hypothesis directly, the observed latitudinal gradient in the movement of gentoo penguins is consistent with a leap-frog strategy. In particular, larger scale movements during the winter provide an opportunity to prospect new haul-out sites and nesting areas that would support colonization events.

The effects of ongoing climate change in the WAP region are difficult to predict [47]. However, continued reduction in SIC during winter may be advantageous for coastal predators

and an increase in the availability of ice-free foraging habitats may facilitate southward expansion in the breeding range of gentoo penguins. Nonetheless, such expectations may be tempered by increases in the biomass of salps [48] with concomitant declines in krill biomass due to recruitment failures [41], local increases in the abundances of cetaceans (potential competitors with gentoo penguins for food) [49] and continued growth of the krill fishing industry [6,8]. Our study suggests that, in the WAP, a latitudinal gradient in the movement of gentoo penguins during winter might be a key to the dynamic of how gentoo populations cope with large-scale changes in the ecosystem.

Ethics. The research was permitted under U.S. Antarctic Conservation Act Permits (permit no. 2017-012), the Polish Permitting Authority IBB PAS (permit no. 01/2016; permit no. 05/2016) and the Argentine Dirección Nacional del Antártico Environmental Office (permit nos. 2017-10 and 2017-69). Field protocols were approved by the University of California San Diego Institutional Animal Care and Use Committee (S05480).

Data accessibility. Data for this article are available from the Dryad Digital Repository at: <https://doi.org/10.5061/dryad.5x69p8d20> [50].

Authors' contributions. M.K.A., J.T.H. and G.M.W. conceptualized the study. J.T.H., M.A.J., M.K.A. and G.M. curated the data. M.K.A. and J.T.H. conducted formal analysis. J.T.H., M.A.J., M.K.A., G.M. and G.M.W. acquired funding. J.T.H., M.A.J., M.K.A. and G.M. administered the project. M.K.A. and J.T.H. wrote the original draft. M.K.A., J.T.H., M.A.J., G.M. and G.M.W. reviewed and edited the final draft.

Competing interests. We declare we have no competing interests.

Funding. This study was funded by CCAMLR CEMP Special Fund, Commission for the Conservation of Antarctic Marine Living Resources, Henryk Arctowski Polish Antarctic Station and Instituto Antártico Argentino—Dirección Nacional del Antártico and NASA MEaSUREs programme.

Acknowledgements. Data used here were collected with support of Henryk Arctowski Polish Antarctic Station and the Akademik Vernadsky Ukrainian Antarctic Station. We thank the Instituto Antártico Argentino—Dirección Nacional del Antártico for support at Cierva Cove and Stranger Point. Reference to specific commercial products does not constitute recommendation by the United States Government.

References

- Hoegh-Guldberg O, Bruno JF. 2010 The impact of climate change on the world's marine ecosystems. *Science* **328**, 1523–1528. (doi:10.1126/science.1189930)
- Ducklow HW, Baker K, Martinson DG, Quetin LB, Ross RM, Smith RC, Stammerjohn SE, Vernet M, Fraser W. 2007 Marine pelagic ecosystems: the West Antarctic Peninsula. *Phil. Trans. R. Soc. B* **362**, 67–94. (doi:10.1098/rstb.2006.1955)
- Parkinson CL, DiGirolamo NE. 2016 New visualizations highlight new information on the contrasting Arctic and Antarctic sea-ice trends since the late 1970s. *Remote Sens. Environ.* **183**, 198–204. (doi:10.1016/j.rse.2016.05.020)
- Henley SF *et al.* 2019 Variability and change in the west Antarctic Peninsula marine system: research priorities and opportunities. *Prog. Oceanogr.* **173**, 208–237. (doi:10.1016/j.pocean.2019.03.003)
- Schofield O, Ducklow HW, Martinson DG, Meredith MP, Moline MA, Fraser WR. 2010 How do polar marine ecosystems respond to rapid climate change? *Science* **328**, 1520–1523. (doi:10.1126/science.1185779)
- Watters GM, Hill SL, Hinke JT, Matthews J, Reid K. 2013 Decision-making for ecosystem-based management: evaluating options for a krill fishery with an ecosystem dynamics model. *Ecol. Appl.* **23**, 710–725. (doi:10.1890/12-1371.1)
- Watters GM, Hinke JT, Reiss CS. 2020 Long-term observations from Antarctica demonstrate that mismatched scales of fisheries management and predator–prey interaction lead to erroneous conclusions about precaution. *Sci. Rep.* **10**, 1–9. (doi:10.1038/s41598-020-59223-9)
- Nicol S, Foster J. 2016 *The fishery for Antarctic krill: its current status and management regime*. Cham, Switzerland: Springer.
- Croxall JP, Trathan PN, Murphy EJ. 2002 Environmental change and Antarctic seabird populations. *Science* **297**, 1510–1514. (doi:10.1126/science.1071987)
- Forcada J, Trathan PN, Reid K, Murphy EJ, Croxall JP. 2006 Contrasting population changes in sympatric penguin species in association with climate warming. *Glob. Chang. Biol.* **12**, 411–423. (doi:10.1111/j.1365-2486.2006.01108.x)
- Descamps S, Tarrow A, Lorentsen SH, Love OP, Varpe Ø, Yoccoz NG. 2016 Large-scale oceanographic fluctuations drive Antarctic petrel survival and reproduction. *Ecography* **39**, 496–505. (doi:10.1111/ecog.01659)
- Lynch HJ, Naveen R, Trathan PN, Fagan WF. 2012 Spatially integrated assessment reveals widespread changes in penguin populations on the Antarctic Peninsula. *Ecology* **93**, 1367–1377. (doi:10.1890/11-1588.1)
- Lynch HJ, Fagan WF, Naveen R, Trivelpiece SG, Trivelpiece WZ. 2012 Differential advancement of breeding phenology in response to climate may alter staggered breeding among sympatric

- pygoscelid penguins. *Mar. Ecol. Prog. Ser.* **454**, 135–145. (doi:10.3354/meps09252)
14. Chambers LE *et al.* 2013 Phenological changes in the Southern Hemisphere. *PLoS ONE* **8**, e75514. (doi:10.1371/journal.pone.0075514)
 15. Trivelpiece WZ, Hinke JT, Miller AK, Reiss CS, Trivelpiece SG, Watters GM. 2011 Variability in krill biomass links harvesting and climate warming to penguin population changes in Antarctica. *Proc. Natl Acad. Sci. USA* **108**, 7625–7628. (doi:10.1073/pnas.1016560108)
 16. Humphries GRW, Naveen R, Schwaller M, Che-Castaldo C, McDowall P, Schrimpf M, Lynch HJ. 2017 Mapping application for penguin populations and projected dynamics (MAPPPD): data and tools for dynamic management and decision support. *Polar Rec. (Gr. Brit.)* **53**, 160–166. (doi:10.1017/S0032247417000055)
 17. Clucas GV, Dunn MJ, Dyke G, Emslie SD, Naveen R, Polito MJ, Pybus OG, Rogers AD, Hart T. 2014 A reversal of fortunes: climate change ‘winners’ and ‘losers’ in Antarctic Peninsula penguins. *Sci. Rep.* **4**, 1–7. (doi:10.1038/srep05024)
 18. Wilson RP, Alvarrez B, Latorre L, Adelung D, Culik B, Bannasch R. 1998 The movements of gentoo penguins *Pygoscelis papua* from Ardley Island, Antarctica. *Polar Biol.* **19**, 407–413. (doi:10.1007/s003000050266)
 19. Kokubun N, Takahashi A, Mori Y, Watanabe S, Shin HC. 2010 Comparison of diving behavior and foraging habitat use between chinstrap and gentoo penguins breeding in the South Shetland Islands, Antarctica. *Mar. Biol.* **157**, 811–825. (doi:10.1007/s00227-009-1364-1)
 20. Miller AK, Karnovsky NJ, Trivelpiece WZ. 2009 Flexible foraging strategies of gentoo penguins *Pygoscelis papua* over 5 years in the South Shetland Islands, Antarctica. *Mar. Biol.* **156**, 2527–2537. (doi:10.1007/s00227-009-1277-z)
 21. Hinke JT, Cossio AM, Goebel ME, Reiss CS, Trivelpiece WZ, Watters GM. 2017 Identifying risk: concurrent overlap of the Antarctic krill fishery with krill-dependent predators in the Scotia Sea. *PLoS ONE* **12**, 1–24. (doi:10.1371/journal.pone.0170132)
 22. Pickett EP, Fraser WR, Patterson-Fraser DL, Cimino MA, Torres LG, Friedlaender AS. 2018 Spatial niche partitioning may promote coexistence of *Pygoscelis* penguins as climate-induced sympatry occurs. *Ecol. Evol.* **8**, 9764–9778. (doi:10.1002/ece3.4445)
 23. Levy H, Clucas GV, Rogers AD, Leaché AD, Ciborowski KL, Polito MJ, Lynch HJ, Dunn MJ, Hart T. 2016 Population structure and phylogeography of the gentoo penguin (*Pygoscelis papua*) across the Scotia Arc. *Ecol. Evol.* **6**, 1834–1853. (doi:10.1002/ece3.1929)
 24. Dykyj IV *et al.* 2018 Features of chronology and breeding success of *Pygoscelis papua* and *P. adeliae* (Spheniscidae) penguins in the Wilhelm Archipelago (CCAMLR Subarea 48.1). *Ukr. Antarct. J.* **1**, 130–147. (doi:10.33275/1727-7485.1(17).2018.39)
 25. Hinke JT, Santos MM, Korczak-Abshire M, Milinevsky G, Watters GM. 2019 Individual variation in migratory movements of chinstrap penguins leads to widespread occupancy of ice-free winter habitats over the continental shelf and deep ocean basins of the Southern Ocean. *PLoS ONE* **14**, 1–19. (doi:10.1371/journal.pone.0226207)
 26. Freitas C, Lydersen C, Fedak MA, Kovacs KM. 2008 A simple new algorithm to filter marine mammal Argos locations. *Mar. Mamm. Sci.* **24**, 315–325. (doi:10.1111/j.1748-7692.2007.00180.x)
 27. Johnson DS, London JM, Lea MA, Durban JW. 2008 Continuous-time correlated random walk model for animal telemetry data. *Ecology* **89**, 1208–1215. (doi:10.1890/07-1032.1)
 28. R Core Team. 2019 *R: a language and environment for statistical computing*. Vienna, Austria: R Foundation for Statistical Computing. See <https://cran.r-project.org/>.
 29. Jonson DS, London JM. 2018 Crawl: an R package for fitting continuous-time correlated random walk models to animal movement data., Zenodo. (<https://doi.org/10.5281/zenodo.596464>)
 30. London JM. 2016 Crawl: utilities and visualization functions in support of crawl. R package version 0.1.0. See <https://github.com/jmlondon/crawl>.
 31. Amante C, Eakins BW. 2009 ETOPO1 1 Arc-Minute Global Relief Model: procedures, data sources and analysis. *NOAA Tech. Memo. NESDIS NGDC* **24**, 1–19. (doi:10.7289/V5C8276M)
 32. Bivand R *et al.* 2019 mapproj: Tools for handling spatial objects. R package version 0.9-8. *R Packag. version 0.9-5*. See <https://CRAN.R-project.org/package=mapproj>.
 33. Chin TM, Vazquez-Cuervo J, Armstrong EM. 2017 A multi-scale high-resolution analysis of global sea surface temperature. *Remote Sens. Environ.* **200**, 154–169. (doi:10.1016/j.rse.2017.07.029)
 34. Bates D, Mächler M, Bolker BM, Walker SC. 2015 Fitting linear mixed-effects models using lme4. *J. Stat. Softw.* **67**, 1–48. (doi:10.18637/jss.v067.i01)
 35. Trathan PN, Forcada J, Murphy EJ. 2007 Environmental forcing and Southern Ocean marine predator populations: effects of climate change and variability. *Phil. Trans. R. Soc. B* **362**, 2351–2365. (doi:10.1098/rstb.2006.1953)
 36. Bost C-A, Jouventin P. 1990 Evolutionary ecology of gentoo penguins (*Pygoscelis papua*). In *Penguin biology* (eds LS Davis, J Darby), pp. 85–112. New York, NY: Academic Press.
 37. Baylis AMM, Tierney M, Orben RA, González de la Peña D, Brickle P. 2021 Non-breeding movements of gentoo penguins at the Falkland Islands. *Ibis* (doi:10.1111/ibi.12882)
 38. Clarke A, Murphy EJ, Meredith MP, King JC, Peck LS, Barnes DKA, Smith RC. 2007 Climate change and the marine ecosystem of the western Antarctic Peninsula. *Phil. Trans. R. Soc. B* **362**, 149–166. (doi:10.1098/rstb.2006.1958)
 39. Reiss CS *et al.* 2017 Overwinter habitat selection by Antarctic krill under varying sea-ice conditions: implications for top predators and fishery management. *Mar. Ecol. Prog. Ser.* **568**, 1–16. (doi:10.3354/meps12099)
 40. Santa CF, Ernst B, Arata JA, Parada C. 2018 Spatial and temporal dynamics of the Antarctic krill fishery in fishing hotspots in the Bransfield Strait and South Shetland Islands. *Fish. Res.* **208**, 157–166. (doi:10.1016/j.fishres.2018.07.020)
 41. Atkinson A *et al.* 2019 Krill (*Euphausia superba*) distribution contracts southward during rapid regional warming. *Nat. Clim. Chang.* **9**, 142–147. (doi:10.1038/s41558-018-0370-z)
 42. Polito MJ, Lynch HJ, Naveen R, Emslie SD. 2011 Stable isotopes reveal regional heterogeneity in the pre-breeding distribution and diets of sympatrically breeding *Pygoscelis* spp. penguins. *Mar. Ecol. Prog. Ser.* **421**, 265–277. (doi:10.3354/meps08863)
 43. Tyler J, Bonfitto MT, Clucas GV, Reddy S, Younger JL. 2020 Morphometric and genetic evidence for four species of gentoo penguin. *Ecol. Evol.* **10**, 13 836–13 846. (doi:10.1002/ece3.6973)
 44. Lynch HJ, Naveen R, Fagan WF. 2008 Censuses of penguin, blue-eyed shag *Phalacrocorax atriceps* and southern giant petrel *Macronectes giganteus* populations on the Antarctic Peninsula, 2001–2007. *Mar. Ornithol.* **36**, 83–97.
 45. Ramos R, Sanz V, Militão T, Bried J, Neves VC, Biscoito M, Phillips RA, Zino F, González-Solis J. 2015 Leapfrog migration and habitat preferences of a small oceanic seabird, Bulwer’s petrel (*Bulweria bulwerii*). *J. Biogeogr.* **42**, 1651–1664. (doi:10.1111/jbi.12541)
 46. Fort J *et al.* 2012 Meta-population evidence of oriented chain migration in northern gannets (*Morus bassanus*). *Front. Ecol. Environ.* **10**, 237–242. (doi:10.1890/110194)
 47. Turner J *et al.* 2014 Antarctic climate change and the environment: an update. *Polar Rec.* **50**, 237–259. (doi:10.1017/S0032247413000296)
 48. Atkinson A, Siegel V, Pakhomov E, Rothery P. 2004 Long-term decline in krill stock and increase in salps within the Southern Ocean. *Nature* **432**, 100–103. (doi:10.1038/nature02996)
 49. Nowacek DP, Friedlaender AS, Halpin PN, Hazen EL, Johnston DW, Read AJ, Espinasse B, Zhou M, Zhu Y. 2011 Super-aggregations of krill and humpback whales in Wilhelmina bay, Antarctic Peninsula. *PLoS ONE* **6**, 2–6. (doi:10.1371/journal.pone.0019173)
 50. Korczak-Abshire M, Hinke JT, Milinevsky G, Juárez MA, Watters GM. 2021 Datasets supporting: Coastal regions of the northern Antarctic Peninsula are key for gentoo populations. Dryad Digital Repository. (<https://doi.org/10.5061/dryad.5x69p8d20>)