1 Estimating length composition of fish observed with stereo-video cameras: a

2 simulation study with application to red snapper (*Lutjanus campechanus*)

3

4 Erik H. Williams¹, Kyle W. Shertzer¹, and Nate Bacheler¹

5 1: Southeast Fisheries Science Center, National Marine Fisheries Service, National Oceanographic and Atmospheric
 6 Administration, Beaufort, NC, USA.

7

8 Abstract

Stereo-video cameras have become an important tool worldwide for enumerating abundance and 9 length compositions of marine fish. The two most common approaches for enumerating are 10 referred to as MaxN and MeanCount, where the former counts fish on the single video frame with 11 the most individuals observed and the latter uses the mean across multiple frames. Previous studies 12 have demonstrated that both approaches may work well for computing relative abundance. 13 However, basic fish population age-structure (e.g. more younger fish) and fish schooling 14 characteristics (e.g. fish of similar size/age swim together) suggest the potential for bias in length 15 compositions computed from the MaxN approach, as the single video frame with the most 16 individuals would tend to overrepresent smaller, more populous fish. To evaluate the two 17 approaches, we simulated a stationary video system placed on a sampling site inhabited by moving 18 19 fish, and we compared length compositions estimated from each approach to that of the true, 20 underlying population. Indeed, MaxN appears biased over a range of potential factors, in which the bias is toward estimates of mean size that are smaller than the true values. The factors leading 21 to biased age/length estimates from MaxN include 1) large fish counts in the single frame, 2) fish 22 schooling by age/length, especially at moderate school densities, and 3) small camera viewing area 23 relative to the site being sampled. In addition to the general evaluation, we applied the simulation 24 framework to video parameters derived from data on red snapper (Lutjanus campechanus) in U.S. 25 Atlantic waters. The results suggest that the MaxN length composition bias is small for red snapper 26 due primarily to the low counts of fish on site and low levels of schooling observed. 27 28 Keywords: sampling design, simulation, stereo-video methods, length composition data, red

- 29 snapper
- 30 Corresponding author: Erik H. Williams
- a. *Email:* erik.williams@noaa.gov
- b. *Phone:* 1-252-728-8603
- 33 c. *Fax:* 1-252-728-8789
- 34

35 1. Introduction

Sampling of fish using video has become more widespread as fishery scientists continue to seek 36 data improvements and more accurate stock assessments. Video has the advantages of being a non-37 38 lethal sampling technique for counting fish and is generally less selective than traditional sampling gears (Morrison and Carbines 2006, Bacheler et al. 2013), but also has the disadvantage of lacking 39 basic biological data that lethal sampling approaches have historically provided (e.g., length, age, 40 maturity, etc.). Knowing the length and age of sampled fish is vital to many modern stock 41 assessment and population demographic methods (Ono et al. 2015). Consequently, video 42 collection programs are increasingly using stereo-video, which allows the estimation of fish length 43 (Cappo et al. 2003, Cappo et al. 2004, Watson et al. 2005, Watson et al. 2010, Shortis et al. 2013, 44 Letessier et al. 2015, Langlois et al. 2015, Langlois et al. 2020, Schramm et al. 2020). This is an 45 important advance in video data collection, but may pose issues with video enumeration methods 46 that were established primarily to estimate abundance rather than length composition. 47

The enumeration of fish on video is typically based on a single frame when the most 48 individuals of the species of interest are observed ("MaxN"; Watson et al. 2010, Langlois et al. 49 2012; also referred to as MinCount, because the maximum observed on any frame is an estimator 50 of the minimum abundance on the site). The relationship between MaxN and true abundance may 51 52 be nonlinear, however, and methods that consider multiple frames have also been developed. For example, Schobernd et al. (2014) developed an alternative unbiased video metric called 53 "MeanCount," where fish abundance is estimated as the mean number of fish observed in a series 54 of 41 equally spaced frames over the viewing interval (see also Bacheler and Shertzer 2020). 55 Although there are different methods for enumerating fish on video, the measurement of fish 56

57 lengths from stereo-video has almost universally occurred at MaxN in order to avoid repeated measurements of the same fish (Langlois et al. 2012). 58

59

Reading stereo-video data for fish lengths versus enumerating fish counts may pose 60 different challenges with respect to statistical properties and sampling efficiency. Past studies have shown that using MaxN methods for computing relative abundance work well for tracking fish 61 populations through time under most conditions (Campbell et al. 2015), despite potential nonlinear 62 relationships with true abundance (Schobernd et al. 2014). Ideally, pilot studies and sampling 63 designs are conducted a priori to optimize efficiency in data collection (Mallet et al. 2021). 64 However, the statistical properties of length composition estimates from MaxN and MeanCount 65 video reading has not previously been studied. 66

Given that fish often school by size (Pitcher and Parrish 1993, Hoare et al. 2000) and 67 younger fish are typically more numerous than older fish for any particular species, the MaxN 68 approach may be biased when measuring fish on a single video frame with the most fish observed. 69 On the other hand, the MeanCount has less potential for bias given that it is not based on an extreme 70 value statistic, but has (to our knowledge) never been used for quantifying length distributions. 71

Determining the bias and precision of fish lengths from MaxN and MeanCount metrics is 72 important for accurately parameterizing stock assessment models. Length composition data 73 74 sampled from a fishery-independent video gear can provide important population dynamic information on the shape of the selectivity curves, the amount of total mortality, and cohort 75 strength. Video samples have the potential to provide important information on smaller and 76 younger fish, potentially serving as an independent estimate of recruitment. Unbiased video 77 samples could also be the basis for comparing length composition data from the fisheries, resulting 78 79 in improved selectivity estimates in stock assessments (Ono et al. 2015).

80 Here, we address potential error (e.g., bias) in stereo-video estimated length compositions by simulating a stationary video system placed on a sampling site inhabited by moving fish, and 81 then comparing length compositions estimated from MaxN and MeanCount approaches to those 82 83 of the true simulated population. We evaluate how length compositions based on the two methods are affected by local fish abundance, size/age dependent schooling behavior, and the viewing area 84 of the video system relative to the area of sampling site. Our goals are to identify conditions under 85 which estimated length compositions from either approach may be considered accurate and to 86 quantify bias where it occurs. We then apply the model using parameter estimates for red snapper 87 (Lutjanus campechanus) in U.S. Atlantic waters to demonstrate both how the model can be applied 88 to actual fish stocks to explore properties of estimated length compositions for this particular 89 species. 90

91

92 **2.** Methods

93 2.1 Simulated Population and Sampling Designs

We used a simple simulation study to investigate several potential factors that lead to bias in length 94 composition data collected from stereo-video capture systems using MaxN and MeanCount 95 methods. We simulated a stationary video sampling system placed on a single site inhabited by 96 97 moving fish. More specifically, a known age-structured population of fish (N) was simulated on the site with various levels of schooling properties and movement in and out of the camera field 98 of view. We configured the simulation to approximate the sampling system used by the Southeast 99 Reef Fish Survey (SERFS), which samples reef-associated species in Atlantic waters of the 100 101 southeast United States (Bacheler et al. 2013, Schobernd et al. 2014). All simulations were coded 102 using R (R Core Team, 2021).

103 The SERFS is a collaborative reef fish monitoring survey consisting of multiple vessels 104 sampling along the southeast United States Atlantic continental shelf from Cape Hatteras to Cape 105 Canaveral using standardized methods. Sampling occurs on hard-bottom sites between 15 and 115 106 m deep during daylight hours using baited chevron traps outfitted with video cameras (see 107 Bacheler and Shertzer 2020 for more details). Since 2019, stereo-video cameras have also been 108 attached to a subset of chevron traps to provide fish length measurements that could be used to 109 estimate selectivity patterns of video and trap gears (e.g., Langlois et al. 2015).

To simulate the video sampling process, we first generated a fish population with ages 110 ranging from 1 to 20 years (y) and an equilibrium (constant recruitment) age structure conditional 111 on an assumed annual natural mortality rate (M). For a given sampling site, a fish population (N)112 was randomly generated by drawing ages from a multinomial distribution with N fish and 113 probabilities determined by the equilibrium, exponentially decaying age structure. We implicitly 114 assumed selectivity or availability of fish to the video gear was 1.0 for all ages, although our 115 qualitative findings should be robust to this assumption. The length of the fish (L, with generic116 117 units) at each age (a) followed a von Bertalanffy growth curve with parameters L_{∞} , K, and t_0 (Table 1, Quinn and Deriso 1999). The height of fish (H) was fixed at 0.5 of L so that age specific two-118 dimensional fish area is L_aH_a . Variability in size at age for individual fish was modeled using a 119 120 normal distribution with an age-constant coefficient of variation (cv) (Table 1).

To simulate a video system for detecting fish, we made a few simplifying assumptions for ease of computation (Table 2). We assume there is no error in the measurement of fish lengths from videos. Fish were simulated in two dimensions, ignoring the distance of fish from the camera. Fish were simulated as if swimming on a 360-degree cylindrical plane around the camera. The distance of the fish from the camera is then a fixed quantity and was determined in our simulation 126 through a field of view parameter, specified as a width expressed in units of fish lengths of maximum size (L_{∞}) . In this case, we used a somewhat arbitrary field of view width (w) of eight 127 maximum sized fish, $8L_{\infty}$. We also fixed the proportion of the camera view (C_p) to 45%, or 162 128 129 degrees, based on the optical-physical properties of Go-Pro camera systems currently in use by the SERFS. The vertical dimension of the camera viewing system was assumed to be infinite or not 130 limiting (e.g. fish stayed in the vertical view at all times). Thus, the only way fish moved in and 131 out of camera view (detection) was through horizontal movement. Fish were assumed to be 132 detected by the camera when the mid-point of their body was within the fixed field of view. Using 133 the specified values of C_p and w, the simulated cylindrical plane is then defined in fish length units 134 of L_{∞} and is w/C_p wide and w high. The height was arbitrarily set for visual simplicity, but has no 135 impact on the simulation results. In natural systems, fish can leave the site or view of the camera 136 in ways we did not model, but this simply has the effect of reducing the effective camera view. In 137 that sense the effective viewing field of any video camera in practice is less than its optical 138 measurements alone. 139

Fish were distributed on a sample site surrounding the camera by first specifying the 140 random x-y positioning of a box to contain the school of fish. Fish schools were assumed to be 141 structured by age class, such that all fish of a particular age class would tend to school together. 142 143 The density of the schools (S_d) was specified by a value of the number of fish (n_a) per unit of fish area (L_aH_a) for each age (a). The value of S_d was assumed constant across ages. The value of S_d 144 constrains the positioning of the midpoint of each individual fish such that they stay within the 145 area defined by L_aH_a . A value of $S_d = 1$ implies the size of the school is confined to a box with 146 dimensions n_aL_a by n_aH_a fish length, such that 10 fish in a school would be contained within a box 147 of size equal to 10 fish lengths by 10 fish heights. A value of $S_d = 0$ implies no schooling or the 148

box for the school equals the entire cylindrical plane. A value of $S_d = 2$ implies 10 fish would be constrained to a box 5 fish lengths by 5 fish heights in size. Because we are only dealing with two dimensions in this simulation, $S_d \approx 9$ would be a theoretical maximum under a tight schooling assumption of fish maintaining a unit fish length distance between each other in three dimensions (i.e. sphere packing).

Movement of individual fish was governed by a two-stage randomization process. In the 154 155 first stage the box containing the school of fish was repositioned by adding a random movement step size that includes the entire cylindrical plane. In the second stage of the randomization process 156 the individual fish were repositioned with an additive random step size constrained by the box for 157 the fish school. Random uniform numbers were drawn in both the x and y dimensions of the entire 158 cylindrical plane to determine the movement step size for the fish school box positioning. A 159 wrapping function was used to force fish whose random movement went outside the right or left 160 side to follow a 360-degree cylinder. This randomization process was repeated at each time step, 161 which is arbitrary, but assumes fish position is independent with each time step. 162

Autocorrelation in movement of the fish school from one time step (frame) to the next time 163 step was controlled through a correlation parameter (ρ), specified as a value on the range [0-1]. 164 Mathematically this was accomplished by multiplying the random movement step size by $1-\rho$. A 165 value of $\rho = 1$ forces no school movement in each time step, while a value of $\rho = 0$ allows for full 166 movement anywhere within the cylindrical plane. In essence, the effect of this parameter is similar 167 to reduced sample sizes or shortened periods of video time. If fish did not move at all, then all 168 frames in the video would be identical and the same as a single snapshot. Thus, the movement 169 170 algorithm is not explicitly connected to the speed of swimming, but rather is controlled by the

171 correlation parameter. A visual example of a simulated video frame with fish on a site and viewing172 area of the camera is shown in Figure 1.

In summary, the video simulation generates *N* multinomial sampled random fish based on the equilibrium age structure from a known population. These fish are then randomly moved about the cylindrical plane of view for 41 video frames (time steps). All the frames together represent a *MeanCount* sample and the frame with the highest count represents the *MaxN* sample. Simulation settings are systematically changed to experimental values shown in Table 2 and compared to the 'true' population values.

179

180 2.2 Length Composition Data

Using the video simulation system described above, we implemented two types of fish counting methods to derive data sets, *MeanCount* and *MaxN*. These video sampling methods rely on counting the fish in multiple video frames during a soak interval of time. To match the SERFS methods we simulated 41 frames for counting fish, using all frames for *MeanCount* samples and the single frame with the maximum number of fish for *MaxN* samples.

To evaluate performance of the estimators, we created *MeanCount* and *MaxN* samples of 186 lengths and compared these estimates to the known lengths of the site's fish population. More 187 specifically, we compared mean length estimates and length distributions (compositions) between 188 the control (true) population and the simulated video samples. To compute compositions for the 189 MeanCount estimator, length samples from the 41 frames were pooled together; for the MaxN 190 estimator and the true composition, lengths were simply enumerated from the single frame or 191 population, respectively. Length measures were averaged to produce mean length estimates and 192 measures were binned into count vectors of 19 equal sized bins from fish sizes of 0 to L_{∞} (1000) 193

194 plus a bin for all fish over L_{∞} to produce proportional length compositions (see length data in 195 Figure 2).

Mean lengths were examined as the distribution of the differences between sampled and 196 197 simulated (true) mean lengths using the R package 'vioplot' (Adler and Kelly, 2021). Difference distributions with means not equal to 0 is an indicator of bias and the spread of the distributions is 198 an indicator of variance. We also report the proportion of runs (p-value) that were found to have 199 200 statistically significant similar length distributions (multinomial vectors) using the R package 'XNomial' (Engels 2015). This is considered an exact test based on a randomization process to 201 evaluate whether the multinomial vectors of true length composition and sampled length 202 composition are similar. In this study, we applied 100,000 randomization trials with the function 203 'xmonte.' The primary advantage is that it avoids potential issues of asymptotic approximations 204 (e.g. likelihood ratio or chi-square), but comes at the cost of extra computation time. 205

206

207 2.3 U.S. South Atlantic Red Snapper Case Study

208 The video simulation described above depends on multiple parameters that are likely to vary temporally during sampling, as well as by the size of the population and schooling nature of the 209 species being studied. As a case study, we used stereo-video data collected by SERFS on red 210 snapper (Lutianus campechanus). The data were collected from 2018 to 2019 with a sample size 211 of 72 videos that were read for lengths following the MeanCount protocols. The data included 212 measurements of individual fish sizes, as well as x-y-z positions of fish within a video frame. We 213 used this data set to estimate the range and statistical distribution of n fish per frame and schooling 214 density (S_d) based on the red snapper positions in the video frame. For simplicity we did not attempt 215 to break the red snapper data into size categories for determining S_d , rather S_d was computed for 216

all red snapper in the frame. Exponential, negative binomial, gamma, and lognormal statistical distributions were fit to the values of n and S_d measured for red snapper in each video frame. The best fit distributions were then used to draw random values for n and S_d in the simulation analysis described above, which was run for 20,000 trials.

221

222 **3. Results**

223 3.1 General results

Applying the fixed values of N = 40, M = 0.2, $S_d = 1$, $\rho = 0$, and $C_p = 0.45$ produced a clear bias in 224 the estimated age and length composition from the MaxN method (Figure 2). The multi-modal 225 pattern shown in the length composition was the result of modeling ages 1 and 2 with cv = 0.2, 226 resulting in distinct age classes showing in the length composition data. The pattern of positive 227 bias at the smaller size/ages and negative bias at the larger size/ages is a consistent result in our 228 study when bias in the MaxN method is present. This bias pattern results in smaller mean length 229 in the sample relative to the true value in every case. Therefore, a simpler representation of this 230 pattern is a measure of difference in mean length from the sample and true population. In contrast, 231 the *MeanCount* method indicated no bias in the composition estimates. 232

The effect of the number of fish on site was tested using a range of N = 5 to 200, with fixed parameters of M = 0.2, $S_d = 1$, $\rho = 0$, and $C_p = 0.45$. These simulations suggest the *MaxN* estimates are unbiased at the lowest number of fish on site (N = 5), but the negative bias in mean length increases with increasing number of fish on site. The increase in bias appears to asymptote at the highest levels of *N* fish on site. The *MeanCount* estimates appear to be unbiased for all values of *N* (Figure 3). The variance in mean length differences for *MaxN* and *MeanCount* decreases with increasing *N*, but is notably larger for the *MaxN* estimates overall (Figure 3). At the lowest value of N = 5, the proportion of length composition estimates that match the true value is much higher for *MaxN* compared to *MeanCount*, but for $N \ge 20$, the *MeanCount* length composition estimates match the true values much better than the *MaxN* estimates. At the highest value of N = 200, the *MaxN* length composition estimates become so biased that only 2% of estimates statistically match the true length compositions (Figure 3).

Next, we tested the effects of changes in M using a range of 0 to 0.5 (y^{-1}). In reality, M = 0245 246 is impossible, but in this scenario, we were essentially testing the magnitude of the differences between younger (smaller) and older (larger) fish as increasing M would tend to skew the 247 population toward younger individuals. MaxN estimates showed a pattern of increasing bias with 248 increasing M values up to M = 0.2 and then slightly decreasing bias out to values of M = 0.5 (Figure 249 4). Also, the proportion of length composition estimates from MaxN that matched the true values 250 increased steadily over the range of M values. MeanCount estimates remained unbiased, had a 251 lower variance in length differences, and had very good matches in the length composition 252 estimates over the full range of *M* values (Figure 4). 253

254 This simulation analysis was initially motivated by the suspected potential for bias of age/length composition estimates resulting from fish schooling (S_d). Both MaxN and MeanCount 255 estimates appear to be unbiased when there is no schooling effect in place (Figure 5). As the 256 257 schooling density increases, the simulations indicate that the bias in MaxN estimates increases quickly, reaching a saturation level at $S_d \ge 0.25$. For *MeanCount* the bias was near zero across the 258 range of schooling densities, but variance increased with higher schooling density. The length 259 composition estimates from the MaxN method showed an improved match percentage with 260 increasing values of S_d , while the *MeanCount* estimates seem to match almost perfectly across all 261 values of S_d (Figure 5). The case of $S_d = 0$ suggests that even when fish schooling is not present, 262

the *MaxN* method is unable to match the true length composition, with only 52% of the estimatesindicating statistically significant matches (Figure 5).

We next evaluated the effects of various levels of correlation (ρ) in movement on both *MaxN* and *MeanCount* estimates. Reduced movement, as indicated by higher ρ values, does not start to affect length estimates until it reaches extreme values of $\rho = 0.75$ and 1.0 (Figure 6). At the value of $\rho = 1.0$, which implies no fish movement, neither *MaxN* nor *MeanCount* are able to estimate the length composition reliably (Figure 6).

Last, we evaluated the effects of the camera view (C_p) on length composition estimates from the two reading methods. Not surprisingly, smaller C_p values increased variance in both *MeanCount* and *MaxN* samples, while increasing values approached complete sampling (a census) and thus perfect estimation (Figure 7). There is zero bias for the *MaxN* estimates at $C_p = 0.95$, as indicated by an absent violin plot and a matching statistic of 1.0 (Figure 7). It should be noted that even with a $C_p = 1.0$, suggesting a 360° camera system, the effective view is likely to be something less than 100% due to factors that affect detection, such as turbidity.

277

278 3.1 Red Snapper Results

Red snapper stereo-video data indicated considerable variability in the number of fish per frame (*n*) and S_d . Values for *n* ranged from 1 to 42 and values for S_d ranged from 0.2 to 4.5. The best fit for *n* came from an exponential distribution with a rate parameter equal to 0.242. To convert this to a value for *N* for use in our video simulation, we divided it by the proportion of viewing area covered by the video camera, $C_p = 0.45$. This distribution suggests that we should most frequently expect n<5 fish per frame, as was observed in the data. The best fit for the S_d measure came from a lognormal distribution with mean equal to -0.0086 and standard deviation equal to 0.5176. This distribution suggests a generally low schooling value, with a mean even lower than the value of S_d = 1 used in most of the simulation analyses above. The empirical estimates for red snapper were used in a stochastic simulation, drawing randomly from the fitted distributions described above using our video simulator for red snapper.

In the red snapper simulations, bias was observed in the *MaxN* estimates for n > 8, while the *MeanCount* method estimates remained unbiased for all values of n (Figure 8). However, more than 65% of the red snapper video samples had fewer than 10 fish per frame. Thus, when all simulated samples are pooled, the differences between age and length composition estimates from *MaxN* and *MeanCount* nearly disappeared (Figure 9).

295

296 4. Discussion

Our simulations showed that *MeanCount* produces unbiased estimates of length 297 composition data, and compared to MaxN, generally has a lower variance in mean length estimates. 298 MeanCount estimates also tended to statistically match the true length composition data more 299 300 frequently than MaxN estimates across a wide range of conditions. MaxN can produce biased length compositions, depending on conditions. The effect is most evident for species that show 301 strong age-dependent schooling behavior and, in our simulation, was always in the direction of 302 underestimating mean length. However, our case study for red snapper indicated very little 303 difference between length composition estimates from the MaxN and MeanCount methods, owing 304 to the low values for S_d and n. It should be noted that field observations typically report loosely 305 aggregated schooling behavior, which could explain a dampened bias effect for that species. 306

307 *MaxN* is the predominant method being used around the world to count and measure fish
308 from video data (Ellis and DeMartini 1995, Merritt et al. 2011, Cappo et al. 2004, Campbell et al.

2018). The primary reason for this lies in the efficiency of reading one video frame for lengths versus many frames for *MeanCount*. Our study suggests the optimal method for reading fish videos for length data may come down to a bias versus cost (manpower) trade-off. Our simulation shows that *MaxN* is unbiased when there is no fish schooling by age or size and when the numbers of fish per frame are low (n < 10). However, when fish school by age or length ($S_d \ge 0.25$) and the number of fish per frame increases (n > 10), then *MeanCount* will produce unbiased and statistically better length composition data than *MaxN*.

Our case study for red snapper indicates only minor differences in length composition 316 estimates from MaxN and MeanCount methodologies, because this species exhibits relatively low 317 S_d and *n* values. The differences are so small that it seems unlikely that this would affect stock 318 assessment or population estimates relying on length composition data. We chose red snapper for 319 convenience and data availability, but it is a large predator on U.S. South Atlantic reef sites and 320 therefore may have lower site abundance and exhibit less schooling behavior than other species in 321 the region, such as vermilion snapper (Rhomboplites aurorubens) or greater amberjack (Seriola 322 323 dumerili). More stereo-video analysis is needed for other fish species to indicate typical values of S_d and n. 324

In our simulation we focused on a video camera viewing proportion of 45%, based on current Go-Pro camera systems in use for SERFS. This value is likely an overestimate because of factors such as water clarity, vertical movements of fish out of range of camera (ignored in this simulation), and obscuring by overlapping fish, all of which will tend to reduce the effective viewing area of a camera. The simulation results here indicate lower view proportions will increase the bias induced by *MaxN* sampling when fish schooling is present. Our simulation also showed that the bias in length composition data from the *MaxN* method diminishes at high values of camera viewing ($C_p > 0.75$). This implies a nearly 360° camera viewing system may help overcome potential bias in length composition data from *MaxN*. Recent work has also shown that 360° cameras improve relative abundance estimation based on *MaxN* as well (Kilfoil et al. 2017, Campbell et al. 2018). This is another example of a potential bias versus cost trade-off in fish length composition data collection. The cost of a 360° stereo-video system could be quite high, but when balanced with the difference in cost between reading video data using *MaxN* versus *MeanCount*, the cost might be justified.

The biases in length composition estimates from MaxN appear over a range of potential 339 conditions. The bias is in one direction and is driven by the differential age or length composition 340 produced by a cohort-structured population, combined with fish schooling by age or length. 341 Natural fish age and length structures rarely match the equilibrium patterns shown in Figure 2 at 342 any particular time, but the average across years will generally tend toward an equilibrium. Thus, 343 the potential bias seen in this simulation will also follow the age/length structure fluctuations 344 created by varying recruitment and mortality processes, being high in some years and lower in 345 346 others. The pattern of bias illustrated in Figure 2 suggests that high recruitment will likely amplify the bias at younger ages, and then as that strong cohort decays, the bias will diminish. Fish 347 schooling or shoaling are highly dynamic processes that are likely habitat- and species-dependent. 348 In general, fish do tend to aggregate by similar length (Hoare et al. 2000, Pavlov and Kasumyan 349 2000). For some species, the tendency for schooling may dissipate with age. Our simulation 350 assumed the schooling density was constant across all ages and only the number fish declined with 351 age. However, schooling density is but one factor that affects sampling properties of video systems. 352 The number of fish on a particular site, the field of view of the camera, and the conditions of the 353 sampling site (e.g., currents, predators, bait) all have the potential to affect how fish schooling 354

behavior is expressed for a given video sample. Given the values for parameters from our 355 simulation that produced biased length composition estimates, the type of species most vulnerable 356 to bias would be fishes lower on the trophic scale, primarily forage fishes and other fishes known 357 358 to school together (e.g. sparids, scombrids, carangids). The species less likely to be affected would be top predatory fish that are more solitary. This suggests that understanding schooling behavior 359 of the target species is required to determine the optimal method for video data analysis. Without 360 361 that understanding, *MeanCount* estimates may be the best default approach, given that it was less biased over a broad range of conditions. 362

Our conclusions from this analysis are predicated on the simulation's accuracy at mimicking the natural system. Regardless of the many reasons why this simulation may not exactly reflect real fish swimming near a video camera, the range and number of factors we evaluated support the conclusion that there is potential for the *MaxN* sampling process to produce biased length composition data and that care should be taken when deciding what video reading method is applied to a given species.

369 Many video sampling systems are limited to clear water in topical and sub-tropical areas of the world. These systems often coincide with higher fish species diversity, thus making the 370 decision about the overall best data collection system for a multi-species system very difficult. As 371 372 video data collection systems shift toward stereo-video usage in fish sampling, the ability to measure the depth of viewing or z dimension opens up many possibilities, including possible 373 density or absolute abundance measures. Fully understanding the potential biases of MeanCount 374 and *MaxN* methods will be important if fish video sampling heads in that direction. In this paper, 375 376 we mimicked the 41 frame counts used by SERFS for the MeanCount method, however the simulation framework could be easily modified to test the cost-benefit of different numbers of 377

frames to use for *MeanCount* or *MaxN*. Until we improve computer automated video reading systems, the cost trade-offs that have to be considered for the best overall data collection systems are further complicated by video processing costs. With proper accounting of video processing costs, such an analysis could point to the optimal number of frames and most efficient video sampling method overall, be it *MeanCount* or *MaxN* (Bacheler and Shertzer 2015). This study and others that may follow will hopefully aid in that decision-making process.

In our red snapper case study, the conclusion was that MaxN estimates would be equally 384 valid as *MeanCount* estimates for use in demographic analysis or stock assessments. This may be 385 the case for many other species as well. However, should a species have relatively high abundance 386 and exhibit moderate to high levels of schooling, then MeanCount may provide more reliable 387 length composition estimates. In general, length composition data in stock assessments can be 388 important, and in particular video methods have the potential to fill important data gaps, including 389 information on the smallest individuals (Ono et al., 2015). If sampling systems shift away from 390 collection of physical specimens, and thus reduced age sampling, and toward video methods, the 391 392 importance of accurate length data in stock assessments will only increase.

393

394 Acknowledgements

395

We thank Matt Campbell, Kevin Craig and anonymous reviewers for critical comments on previous versions of this manuscript. Mention of trade names or commercial companies is for identification purposes only and does not imply endorsement by the National Marine Fisheries Service, NOAA. The scientific results and conclusions, as well as any views and opinions

- 400 expressed herein, are those of the authors and do not necessarily reflect those of any government
- 401 agency.

402 **References**

- 403 Adler, D., Kelly, S.T., 2021. vioplot: violin plot. R package version 0.3.7
- 404 https://github.com/TomKellyGenetics/vioplot
- 405 Bacheler, N.M., Schobernd, C.M., Schobernd, Z.H., Mitchell, W.A., Berrane, D.J., Kellison,
- 406 G.T., Reichert, M.J.M. 2013. Comparison of trap and underwater video gears for indexing
- 407 reef fish presence and abundance in the southeast United States. Fish Res 143: 81–88.
- 408 https://doi.org/10.1016/j.fishres.2013.01.013.
- 409 Bacheler N.M., Shertzer, K.W., 2015. Estimating relative abundance and species richness from
- video surveys of reef fishes. Fish Bull 113:15-26. https://doi.org/10.7755/FB.113.1.2
- 411 Bacheler N.M., Shertzer, K.W., 2020. Catchability of reef fish species in traps is strongly
- 412 affected by water temperature and substrate. Mar Ecol Prog Ser 642:179-190.
- 413 https://doi.org/10.3354/meps13337
- 414 Campbell, M.D., Pollack, A.G., Gledhill, C.T., Switzer, T.S., DeVries, D.A. 2015. Comparison
- 415 of relative abundance indices calculated from two methods of generating video count data.
- 416 Fish Res 170: 125–133. https://doi.org/10.1016/j.fishres.2015.05.011.
- Campbell, M.D., Salisbury, J., Caillouet, R., Driggers, W.B., Kilfoil, J. 2018. Camera field-ofvideo and fish abundance estimation: a comparison of individual-based model output and
- 419 empirical data. J Exp Mar Biol Ecol 501: 46–53.
- 420 Cappo M., Larsen, D.P., Hughes, R.M., 2004. Comparison of baited remote underwater video
- 421 stations (BRUVS) and prawn (shrimp) trawls for assessments of fish biodiversity in inter
- reefal areas of the Great Barrier Reef Marine Park. J Exp Mar Biol Ecol 302:123–152.
- 423 https://doi.org/10.1016/j.jembe.2003.10.006.

424	Cappo M., Harvey, E., Malcom, H., Speare, P., 2003. Potential of Video Techniques to Monitor
425	Diversity, Abundance and Size of Fish in Studies of Marine Protected Areas, North Beach,
426	Western Australia., pp. 455–464.

- 427 Ellis, D.M., DeMartini, E.E. 1995. Evaluation of a video camera technique for indexing
- 428 abundances of juvenile pink snapper, *Pristipomoides filamentosus*, and other Hawaiian
 429 insular shelf fishes. Fish Bull 93: 67–77.
- 430 Engels, B., 2015. XNomial: Exact Goodness-of-Fit Test for Multinomial Data with Fixed
- 431 Probabilities. R package version 1.0.4. https://CRAN.R-project.org/package=XNomial
- 432 Hoare, D.J., Krause, J., Peuhkuri, N., Godin, J.-G.J. 2000. Body size and shoaling in fish. J Fish
- 433 Biol 57: 1351–1366. https://doi.org/10.1111/j.1095-8649.2000.tb02217.x.
- 434 Kilfoil, J.P., Wirsing, A.J., Campbell, M.D., Kiszka, J.J., Gastrich, K.R., Heithaus, M.R., Zhang,
- 435 Y., Bond, M.E. 2017. Baited remote underwater video surveys undercount sharks at high
- densities: insights from full-spherical camera technologies. Mar Ecol Prog Ser 585: 113–121.
- 437 Langlois, T. J., Fitzpatrick, B. R., Fairclough, D. V., Wakefield, C. B., Hesp, S. A., McLean, D.
- 438 L., Harvey, E.S., Meeuwig, J. J., 2012. Similarities between line fishing and baited stereo-
- 439 video estimations of length-frequency: Novel application of kernel density estimates. PLoS
- 440 ONE 7(11):e45973. https://doi.org/10.1371/journal.pone.0045973
- 441 Langlois, T.J., Newman, S.J., Cappo, M., Harvey, E.S., Rome, B.M., Skepper, C.L., Wakefield,
- 442 C.B., 2015. Length selectivity of commercial fish traps assessed from in situ comparisons
- 443 with stereo video: Is there evidence of sampling bias? Fish Res 161:145–155.
- 444 https://doi.org/10.1016/j.fishres.2014.06.008

- Langlois, T., Goetze, J., Bond, T., et al., 2020. A field and video annotation guide for baited
- remote underwater stereo-video surveys of demersal fish assemblages. Methods Ecol Evol.
- 447 11: 1401–1409. https://doi.org/10.1111/2041-210X.13470.
- 448 Letessier, T.B., Juhel, J-B., Vigliola, L., Meeuwig, J.J., 2015. Low-cost small action cameras in
- stereo generates accurate underwater measurements of fish. J Exp Mar Biol Ecol. 466:120-
- 450 126.
- Mallet, D., Olivry, M., Ighiouer, S., Kulbicki, M., Wantiez, L., 2021. Nondestructive monitoring
 of soft bottom fish and habitats using a standardized, remote and unbaited 360° video
- 453 sampling method. Fishes 6(50):1-17. https://doi.org/10.3390/fishes6040050
- 454 Merritt, D., Donovan, M.K., Kelley, C., Waterhouse, L., Parke, M., Wong, K., Drazen, J.C.
- 455 2011. BotCam: a baited camera system for nonextractive monitoring of bottomfish species.
 456 Fish Bull 109: 56–67.
- 457 Morrison, M., Carbines, G. 2006. Estimating the abundance and size structure of an estuarine
- 458 population of the sparid *Pagrus auratus*, using a towed camera during nocturnal periods of
- inactivity, and comparisons with conventional sampling techniques. Fish Res 82: 150–161.
- 460 https://doi.org/10.1016/j.fishres.2006.06.024.
- 461 Ono, K., Licandeo, R., Muradian, M.L., et al., 2015. The importance of length and age
- 462 composition data in statistical age-structured models for marine species. ICES J Mar Sci
- 463 72:31-43. https://doi.org/10.1093/icesjms/fsu007
- Pavlov, D.S., Kasumyan, A.O., 2000. Patterns and mechanisms of schooling behavior in fish: a
 review. J. Icthy 40(2):S163-S231.
- 466 Pitcher, T.J., Parrish, J.K., 1993. Functions of shoaling behavior in teleosts. Pages 122-147 in
- 467 (Krebs, J.R., Davies, N.B., eds) Behavioral Ecology: An Evolutionary Approach. Oxford:

- 468 Blackwell Scientific Publications.
- 469 Quinn, T.J., Deriso, R.B., 1999. Quantitative fish dynamics. Oxford University Press, Oxford.
 470 560 p.
- 471 R Core Team., 2021. R: A language and environment for statistical computing. R Foundation for
- 472 Statistical Computing, Vienna, Austria. URL https://www.R-project.org/.
- Schobernd, Z. H., Bacheler, N. M., Conn, P. B., 2014. Examining the utility of alternative video
 monitoring metrics for indexing reef fish abundance. Can J Fish Aq Sci. 71:464–471.
- 475 https://doi.org/10.1139/cjfas-2013-0086
- 476 Schramm, K.D., Marnane, M.J., Elsdon, T.S., Jones, C., Saunders, B.J., Goetze, J.S., Driessen,
- 477 D., Fullwood, L.A.F., Harvey, E.S., 2020. A comparison of stereo-BRUVs and stereo-ROV
- 478 techniques for sampling shallow water fish communities on and off pipelines. Mar Env Res479 162:105198.
- 480 Shortis, M.R., Ravanbakhsh, M., Shafait, F., Harvey, E.S., Mian, A., Seager, J.W., Culverhouse,
- 481 P.F., Cline, D.E., Edgington, D.R., 2013. A review of techniques for the identification and
- 482 measurement of fish in underwater stereo-video image sequences. Videometrics, Range
- 483 Imaging, and Applications XII, SPIE. 8791. 87910G. 10.1117/12.2020941.
- 484 Watson, D.L., Harvey, E.S., Anderson, M.J., Kendrick, G.A., 2005. A comparison of temperate
- reef fish assemblages recorded by three underwater stereo-video techniques. Mar Biol.
- 486 148:415-425.
- 487 Watson, D.L., Harvey, E.S., Fitzpatrick, B.M., Langlois, T.J., Shedrawi, G., 2010. Assessing reef
- 488 fish assemblage structure: how different stereo-video techniques compare? Mar Biol.
- 489 157:1237-1250.

490 Table 1. Values of the biological parameters used in the population simulation.

Biological factor	Parameter	Values
Age range (<i>y</i>)	a	1 to 20
Natural mortality (y^{-1})	M	0.2
Asymptotic length (length units)	L_{∞}	1000
Growth rate (y^{-1})	Κ	0.25
Spawning adjustment (y)	t ₀	-0.5
Height of fish (length units)	Н	0.5L
Variability in length-at-age	cv	0.2

493 Table 2. Factors used in the simulation tests which were run 20,000

Video Factor	Parameter	Base Value	Experimental Range
Video width of view	W	$8L_{\infty}$	fixed
Number of fish on site	N	40	$\{5, 20, 40, 100, 200\}$
Natural mortality	M	0.2	$\{0, 0.1, 0.2, 0.35, 0.5\}$
Schooling density	Sd	1	$\{0, 0.25, 0.5, 1, 2\}$
Movement correlation	ρ	0	$\{0, 0.25, 0.5, 0.75, 1\}$
Proportion of camera view	C_p	0.45	{0.15, 0.3, 0.45, 0.65, 0.95}



Fig. 1. An example of a single video frame from a simulated fish population on a sample site. Fish are represented by ovals with the width and height representing the relative size at age, indicated by the number in the center of the oval. The video viewing frame is delineated between the vertical dashed lines with fish outside those lines avoiding detection. Simulation parameters are set to $S_d = 1$, N = 40, and growth and M values are set to the base values indicated in Tables 1 and 2.





Fig. 2. Average age and length composition estimates from the video sampling simulation with N = 40, M = 0.2, $S_d = 1$, $\rho = 0$, and $C_p = 0.45$ parameter settings and sampling repeated 20,000 times. *MeanCount* and *MaxN* age composition estimates are shown relative to the true population.



Fig. 3. Split violin plots showing the distribution of differences in mean length from the sample estimated and true on-site population for MaxN (light blue on left hand side) and MeanCount(dark blue on right hand side) methods as a function of number of fish on site for 20,000 random samples with N = 5 to 200, M = 0.2, $S_d = 1$, $\rho = 0$, and $C_p = 0.45$ parameter settings. Horizontal solid black lines represent medians and the vertical bold lines represent the interquartile range. Proportions along the top of each violin plot correspond to the proportion of multinomial samples with similarity P-values >= 0.95 when compared to the true multinomial values.



Fig. 4. Split violin plot showing the distribution of differences in mean length from the sample 517 estimated and true on-site population for MaxN (light blue on left hand side) and MeanCount 518 (dark blue on right hand side) methods as a function of natural mortality (M) values for 20,000 519 random samples with N = 40, M = 0.0 to 0.5, $S_d = 1$, $\rho = 0$, and $C_p = 0.45$ parameter settings. 520 Horizontal solid black lines represent medians and the vertical bold lines represent the 521 interquartile range. Proportions along the top of each violin plot correspond to the proportion of 522 multinomial samples with similarity P-values ≥ 0.95 when compared to the true multinomial 523 524 values.



Fig. 5. Split violin plot showing the distribution of differences in mean length from the sample estimated and true on-site population for MaxN (light blue on left hand side) and MeanCount(dark blue on right hand side) methods as a function of fish school density (S_d) values for 20,000 random samples with N = 40, M = 0.2, $S_d = 0$ to 2, $\rho = 0$, and $C_p = 0.45$ parameter settings. Horizontal solid black lines represent medians and the vertical bold lines represent the interquartile range. Proportions along the top of each violin plot correspond to the proportion of multinomial samples with similarity P-values >= 0.95 when compared to the true multinomial

533 values.



535 Fig. 6. Split violin plot showing the distribution of differences in mean length from the sample estimated and true on-site population for MaxN (light blue on left hand side) and MeanCount 536 (dark blue on right hand side) methods as a function of movement correlation (ρ) values for 537 20,000 random samples with N = 40, M = 0.2, $S_d = 1$, $\rho = 0.0$ to 1.0, and $C_p = 0.45$ parameter 538 settings. Horizontal solid black lines represent medians and the vertical bold lines represent the 539 interquartile range. Proportions along the top of each violin plot correspond to the proportion of 540 multinomial samples with similarity P-values ≥ 0.95 when compared to the true multinomial 541 542 values.





Fig. 7. Split violin plot showing the distribution of differences in mean length from the sample 544 estimated and true on-site population for MaxN (light blue on left hand side) and MeanCount 545 (dark blue on right hand side) methods as a function of video viewing proportion values for 546 20,000 random samples with N = 40, M = 0.2, $S_d = 1$, $\rho = 0$, and $C_p = 0.15$ to 0.95 parameter 547 settings. Horizontal solid black lines represent medians and the vertical bold lines represent the 548 interquartile range. Proportions along the top of each violin plot correspond to the proportion of 549 multinomial samples with similarity P-values ≥ 0.95 when compared to the true multinomial 550 551 values.



Fig. 8. Split violin plot showing the distribution of differences in mean length from the simulated Red Snapper population for MaxN (light blue on left hand side) and MeanCount (dark blue on right hand side) methods as a function of number of fish on-site for 20,000 random samples. Horizontal solid black lines represent medians and the vertical bold lines represent the interquartile range. Proportions along the top of each violin plot correspond to the proportion of multinomial samples with similarity P-values >= 0.95 when compared to the true multinomial values.





561 Fig. 9. Age and length composition estimates from the video sampling simulation of Red

Snapper with stochastic values matching those observed in video samples. Sampling repeated
 20,000 times. Average *MeanCount* and *MaxN* age and length composition estimates are shown

relative to the true population.