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Prepared by

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Prepared for

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### **PREFACE**

The Ohio State University's Center for Lake Erie Area Research has conducted an aquatic ecology monitoring program in Lake Erie in the vicinity of the Davis-Besse Nuclear Power Station for the Toledo Edison Company since April 1973. This effort has been supervised by Drs. Charles E. Herdendorf and Jeffrey M. Reutter. Dr. Herdendorf took responsibility for water quality analyses, and Dr. Reutter was responsible for biological analyses.

The following report provides an appraisal of the impacts of the operation of the Davis-Besse Nuclear Power Station, Unit 1, on the aquatic environment of Lake Erie in the vicinity of the Station. The primary responsibility for the preparation of the various components of the report are designated below:

## Charles E. Herdendorf

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- 3. Aquatic Environment
- 4. Impact Appraisal — Water Quality

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### EXECUTIVE SUMMARY

The Davis-Besse Nuclear Power Station is located in Ottawa County, Ohio, at Locust Point on the southwest shore of Lake Erie, about 21 miles east of Toledo. Unit 1 has a net electrical capacity of 906 MWe and a closed condenser cooling system which dissipates heat to the atmosphere by means of a natural-draft cooling tower, 493 feet high and 415 feet in diameter at its base. Make-up water for cooling purposes is drawn from Lake Erie from a submerged intake crib 3000 feet offshore through a buried eight-foot diameter conduit to a closed, but uncovered, intake canal. The canal is approximately 2950 feet long and terminates at the trash racks of the intake structure. Water is drawn through the intake crib and conduit by gravity. Design capacity for Unit 1 is 42,000 gpm with a resultant approach velocity through the crib ports of 0.25 ft/sec. Cooling tower blowdown is discharged at a point approximately 1200 feet offshore through a six-foot diameter buried conduit which terminates in a high velocity nozzle to promote rapid mixing. The maximum allowable  $\Delta T$  is 20<sup>0</sup>F.

Studies of the aquatic environment in Lake Erie in the vicinity of the intake and discharge of this station were initiated in 1973. From 1973 to 1979, with few exceptions, the following parameters were sampled, during ice-free times, at approximately monthly intervals: water quality, phytoplankton, zooplankton, benthic macroinvertebrates (60-day intervals in 1977, 1978, and 1979), fish, and ichthyoplankton (approximately 10-day intervals during the spring spawning season). Ichthyoplankton entrainment studies and fish impingement studies were initiated after the plant began operating in August 1977. As is to be expected when a new unit first goes "on line", Unit 1 was operated sporadically from August 1977 through December 1979. It is the purpose of this report to appraise the impact of unit operation on the aquatic environment by comparing results obtained prior to unit operation with those obtained from September 1977 through December 1979.

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Water Quality. Eighteen water quality parameters were monitored at approximately monthly intervals beginning in April 1974. In general the quality of Lake Erie water in the vicinity of the Station's discharge<br>structure has remained relatively constant over the past seven years. The structure has remained relatively constant over the past seven years. concentrations of dissolved and suspended substances were slightly higher during the operational period, particularly: chloride, magnesium, silica, sulfate, nitrate, turbidity, and suspended solids. Dissolved oxygen and phosphorus were slightly lower after operation. The magnitude of these differences was not great and appeared to be caused by the general condition of the nearshore waters of western Lake Erie rather than Unit operation.

Phytoplankton. Quantitative estimates of phytoplankton densities at Locust Point were obtained at approximately monthly intervals from 1974 through 1979. Operational phytoplankton densities were larger during the

spring and fall than pre-operational densities. This was a natural phenomenon occurring throughout the nearshore waters of western Lake Erie and not caused by unit operation.

Zooplankton. Quantitative estimates of zooplankton densities in Lake Erie at Locust Point were obtained at approximately monthly intervals from 1973 through 1979. With the exception of cladoceran densities, which were *very* similar during the pre-operational and operational studies, zooplankton operational densities, though generally similar to pre operational densities, were somewhat lower than the corresponding pre operational monthly density. However, these differences appeared to be due to natural phenomena occurring along the south shore of the Western Basin and not related to unit operation.

Benthic Macroinvertebrates. Benthic macroinvertebrate densities in Lake Erie at Locust Point were observed at approximately 30-day intervals from 1973-1976 and 60-day intervals from 1977-1979. Operational densities were within the ranges established during the pre-operational study for *every* month except September. Differences were attributable to natural variation.

Fish. Monthly gill net catches from Lake Erie near Locust Point from 1973-1979 were used to evaluate the impact of unit operation. Fish populations for each of the eight major species at Locust Point, alewife, channel catfish, freshwater drum, gizzard shad, spottail shiner, walleye, white bass, and yellow perch, and the density of all species combined showed little or no variation between pre-operational and operational results.

Ichthyoplankton. Ichthyoplankton densities from Lake Erie in the vicinity of the intake and discharge were monitored at approximately 10 day intervals from 1974 through 1979. Tremendous variability was observed from year to year. However, due to the similarity in densities observed at the intake and discharge and control stations, there is indication that the activities of the Power Station have not significantly altered these populations.

Entrainment. Ichthyoplankton entrainment estimates were not developed until the spring of 1978 as entrainment is an operational phenomenon, and there were few, if any, ichthyoplankters in Lake Erie to be entrained during the first fall and winter of the operational period (September 1977 - March 1978). During 1978 and 1979, the number of ichthyoplankters entrained was insignificant compared to lake populations. Furthermore, the off-shore intake, where larvae densities are lower, and the low intake water volume due to the cooling tower and closed condender cooling system, necessitate a very low-level impact on western Lake Erie fish populations.

Impingement. Fish impingement at the Davis-Besse Nuclear Power Station was estimated from measurements of approximately 24 hours taken approximately 3 times per week from January 1, 1978 to December 31, 1979. Goldfish was the species most commonly impinged, representing 49.9 percent (1978) and 78.6 percent (1979) of the total number of fish impinged. By number, the 6,607 fish impinged during 1978 were 0.04 percent of the Ohio 1978 sport fishing harvest, while the 4,385 fish impinged during 1979 were

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0.03 percent of the Ohio 1978 sport fishing harvest. By weight, impingement was less than 0.001 percent (both years) of the Ohio 1978 sport fishing harvest. These figures become even less significant when one realizes that the Ohio sport catch was only 83.4 percent of the Ohio 1978 commercial catch and only 15.9 percent of the 1978 commercial catch from all of Lake Erie.

Conclusion. Based upon the results obtained to date, there are indications that operation of the Davis-Besse Nuclear Power Station, Unit 1, has had no short-term deleterious effects on the Lake Erie ecosystem. Therefore, it is the conclusion of this appraisal that the Station has not significantly altered the aquatic environment at Locust Point and that long-term deleterious impacts are unlikely.

## INTRODUCTION

The Davis-Besse Nuclear Power Station, Unit 1, initiated commercial operation on August 29, 1977 (Table 1). The purpose of this report is to provide a preliminary appraisal of the impacts of station operation on the aquatic environment of Lake Erie. A pre-operational aquatic ecology monitoring program at the Station was begun in 1973-1974 and continued through the construction period. The program consisted of monitoring 18 water quality parameters and biological populations, including plankton, benthos and fish. Normally samples were taken monthly during the ice-free Once commercial operation was started, the monitoring program continued essentially unchanged, except for the addition of fish impingement/entrainment studies. attempt to compare natural water quality/biological variability, as measured during the pre-operational period, with values obtained during the operational period. The details of the monitoring program are found in Appendix B to License NPR-3 "Environmental Technical Specifications".

For the purposes of this report, the pre-operational period is considered to be from 1973 or 1974 (depending on when monitoring for a particular component began) to August 31, 1977. The operational period considered is from September 1, 1977 to December 31, 1979. The Station's operating history, including: 1) reactor power record, 2) electrical power record, 3) intake and discharge temperature records, 4) water pumping record, and 5) water discharge record are presented in Figures 1 to 10. It can be seen from these figures that during the period of operation being considered, average generation was approximately 33% of its potential capacity. This circumstance was largely due to several months of maintenance "outage" during the summer of 1978 and the "Three-Mile Island Incident" in 1979. Of the 28 operational months being considered in 1977, 1978, and 1979 water quality/biological sampling and mean unit output of greater than 453 MWe (50% capacity) coincided during six months.

#### STATION DESCRIPTION

#### Station Location

The Davis-Besse Nuclear Power Station, Unit 1 is located in Ottawa County, Ohio, on the southwest shore of Lake Erie, about 21 miles east of Toledo. The 954-acre site is located in Carroll Township adjacent to the mouth of the Toussaint River (coordinates:  $41^035^{\circ}57^{\circ}$  N and 83 $^0$ 05'28" W).The site has 7,250 feet of Lake Erie frontage (Figure 11). This section of shoreline is flat and marshy with a maximum elevation only a few feet above the lake level (U.S. Atomic Energy Commission, 1973).

# General Station Description

Unit 1 is a nuclear-powered electric generating facility with a net electrical capacity of 906 MWe. The facility utilizes a pressurized water reactor (PWR) manufactured by Babcock and Wilcox Company. Most of the heat from the turbine steam condenser is dissipated to the atmosphere by means of natural-draft cooling tower, 493 feet high and 415 feet in diameter at its base.

## Cooling Water Intake Design

The cooling water intake shown in Figure 12 is made up of three principle elements; the intake crib and conduit, intake canal, and intake structure. The Unit obtains its cooling water from Lake Erie through the intake crib. Water entering the intake crib flows by gravity through the eight-foot diameter intake conduit buried beneath the lake bottom to the intake canal. The water then flows through the intake canal to the intake structure located at the west end of the intake canal forebay. From the intake structure cooling water will be pumped to the various systems within the unit. These three principle components are described in detail in the following sections.

Intake Crib. The intake crib for the Davis-Besse Nuclear Power Station is located in the Western Basin of Lake Erie approximately 3000 feet offshore from the land area commonly known as Locust Point in approximately 11 feet of water at low water datum (568.6 ft. I.G.L.D.). The lake area off of Locust Point has been identified as an area of constant sand movement. The intake crib is a wooden cross shaped structure rising 3'-10" above the lake bottom with intake screens (ports) located in the ends of each of the four arms so that water enters the crib downward through the ports. At the design maximum flow of 42,000 gpm, the intake velocity has been calculated  $a\bar{t}$  0.25 ft/sec (U.S. Commission, 1975). Table 2 shows calculated intake velocities for various pumping rates. At the 42,000 gpm design flow rate, the velocity through the eight-foot diameter conduit would be approximately 1.8 ft/sec. This design is similar to the one used at the Oregon, Ohio, and Port Clinton, Ohio, municipal water intakes. Figure 13 shows the similarities of these intakes.

Normal practice in intake design has been to locate intake cribs in 20 to 50 feet of water to avoid ice formation and the possibility of blockage from ice jams. Inlet ports should be located four to eight feet off the bottom to minimize the uptake of sand, silt, and other sediment. However, adherance to these practices has not always been possible in the Western Basin of Lake Erie because of its shallowness. This is the case with the design chosen for the Davis-Besse intake crib. The Davis-Besse intake crib is located in relatively shallow water, 11 feet below low water datum, and five feet below the lowest water level experienced at the site, 562.9 IGLD computed from the Toledo gauging station records corrected to the site. Therefore, the intake design must be such that the crib will not be exposed by low water and the intake ports have to be high enough off the bottom that sand and sediment are not drawn into the crib. Locating the crib -in deeper water --was--investigated-but-found-not to be a viable alternative. Water depths of 20 feet are not reached in the vicinity of the site until approximately four to five miles from shore. The design finally chosen utilized a downward flow of water into the crib so that the intake ports could be located as far off the lake bottom as possible and still be under water during low lake level conditions.

During the design of the intake crib, consideration was given to using velocity caps to change the direction of the intake flow from vertical to horizontal. However, this did not turn out to be feasible, since under low lake level conditions the upper portion of the velocity caps would have been above water. Also, since the velocity caps would protrude above the top of the intake crib, they would be subjected to winter ice conditions. These ice conditions, floating ice, and wind blown ice masses, would most likely damage the velocity caps annually and in doing so could cause structural damage to the intake crib itself.

Intake Canal. The intake canal is an open channel with earthen embankments to convey water from the intake conduit (bringing water from the intake crib) to the intake structure located immediately east of Unit No. 1. The intake canal is approximately 2950 feet long including the forebay and is separated from the lake by a sand beach and beachfront dike constructed of large limestone rip-rap. The canal is approximately 40 to 45 feet wide at the bottom, with 3:1 side slopes and a water depth of 13 to 14 feet at normal lake levels except in the vicinity of the intake structure where it widens to form the forebay. At a flow rate of 42,000 gpm, the calculated velocity in the intake canal is approximately 0.11 ft/sec. The intake canal forebay is approximtely 800 feet long, 200 feet wide, at the bottom, with 3:1 side slopes and a water depth of 16 to 17 feet at normal lake levels.

Intake Structure. The intake structure is shown in Figure 14 and is located at the western end of the intake canal forebay. All of the water which is used by the unit is pumped via the pumps located in the intake structure. The following pumps are located in the intake structure.

Service Water Pumps - 2 operating, 1 standby Cooling Tower Makeup Pump - 2 used as required Dilution Pump - 1 used as required Water Treatment Feedpumps - 1 operating, 1 standby Screen Backwash Pumps - 2 used as required

These pumps are preceded by the trash racks and traveling screens. The trash racks are fixed screens, have 4-inch by 26-inch openings, and will be manually cleaned. The traveling screens have  $\frac{1}{4}$ -inch square openings and will be automatically cleaned either on a pre-set time<br>interval or differential pressure across the screens. The impinged interval or differential pressure across the screens. material washed from these screens is sluiced through a trough to a holding basin with an overflow weir discharge to allow monitoring of this material. Collections of impinged fish were made by placing a basket within the trough itself.

#### Water Use

The quantity of water used for cooling at the Davis-Besse Nuclear Power Station, Unit No. 1, has been minimized by using a closed condenser cooling water system and a natural draft cooling tower. The unit's water usage is also minimized by recycling the heated discharge from the service water system and using it as makeup to the closed condenser cooling water<br>system. This exceeds the requirement of 40 CFR 423.13. "Effluent system. This exceeds the requirement of 40 CFR  $423.13$ , limitation guidelines representing the degree of effluent reduction<br>attainable by the application of the best available technology application of the best available technology economically achievable" as well as 40 CFR 423.15, "New Source Performance Standards" which would permit the heated discharge from the service water system to be discharged, provided it meets chlorine limitations. Table 3 shows the unit's maximum, minimum, and average water usage for each month during 1978 at the intake crib.

### Discharge System

All station effluents (except storm water drainage and certain building drains which go to the Toussaint River) are mixed in the collection box prior to discharge into Lake Erie. Most of this mixture is cooling tower blowdown water and its associated dilution water which is added so that the concentration of dissolved solids in the discharge will be less than twice the concentration in the lake. The collection box has a small volume compared with the flow rates into it, and, therefore, the box merely serves to mix the various effluents. From the collection box, the station discharge flows through a six-foot diameter buried pipe to the slot-type jet discharge structure (4.5 feet wide x 1.5 feet high) 1200 feet offshore in Lake Erie (Figure 12). The elevation of the collection box provides the necessary head for discharge through the pipe to the lake under all predicted water level conditions. The slot-type discharge has an exit water velocity of about 6.5 ft/sec at the design maximum discharge flow of 20,000 gpm. The nominal calculated water velocity of 3.6 ft/sec, at the typical discharge rate of 11,000 gpm, promotes rapid entrainment and mixing with lake water. The lake bottom has been rip-rapped with rock for about 200 feet in front of the slot discharge to minimize scouring of the lake bottom and associated turbidity.

Chemical Discharge. All of the makeup water to the recirculating system (cooling tower) is partially neutralized with sulfuric acid, releasing carbon dioxide, and thereby reducing the amount of scale formed in the condenser. The only other chemical added to the circuits is<br>elemental chlorine for defouling. The recirculating cooling water elemental chlorine for defouling. blowdown contains the major fraction of all chemicals discharged to Lake<br>Erie. Due to the evaporation of water in the cooling tower, the Due to the evaporation of water in the cooling tower, the<br>ation of dissolved solids in the recirculating water is concentration of dissolved solids in the recirculating water approximately double that in the lake. Because of the addition of sulfuric acid and the loss of carbon dioxide, the sulfate ratio is slightly higher and the carbonate ratio is slightly lower in discharge water while ratios for various other chemicals are the same as in lake water.

Thermal Discharge. The discharge of cooling tower blowdown from the station's submerged discharge structure generates a thermal plume in Lake Erie. The plume is calculated to have a maximum surface area of 0.7 acres *f\** (U.S. Atomic Energy Commission, 1973). The temperature difference between cooling tower blowdown water and ambient lake water ranges as high as  $30^{\circ}$ F. Lake water is used to dilute the blowdown so that the effluent to the lake never exceeds 20<sup>0</sup>F above ambient lake water temperature.

### , AQUATIC ENVIRONMENT

### Habitat Description

Locust Point and Western Lake Erie. Locust Point is a gently curving headland on the south shore of western Lake Erie, approximately ten miles west of Port Clinton, Ohio (Figure 15). The Davis-Besse Nuclear Power Station is located on a 954-acre tract of land on this point. The terrain of the point is relatively flat and contains about 600 acres of marshland.<br>The Station has a 7.250-foot frontage on Lake Erie along the point. The The Station has a 7,250-foot frontage on Lake Erie along the point. point has a relatively stable barrier beach which separates Navarre marsh from the lake. The shore is not tending to straighten itself or advance over the wetland which is usual for barrier beaches with such a configuration. This may be in part due to the extensive rip-rap dike placed on the berm of the beach during the record-high water levels of the 1972 and 1973. The dike now protects the Station site, as well as the wetland, from the lake encroachment.

Hydrographic surveys show a *very* gentle slope of the lake bottom from the shore out for a distance of at least 4000 feet (Figure 15). Two sand bars typically lie in the nearshore zone, one at 120 feet offshore and the other at 280 feet from the beach. The deeper area between the beach and the first sand bar has a thin bottom layer of fluffy silt and shell fragments over the sand. The inshore slope of the first bar contains an abundant population of naiad clams. The sand bottom, generally medium- to fine-grained, extends to 800 feet offshore (5.0 feet water depth, IGLD, 1955). At this point the bottom deepens by 0.5 feet and is composed of At this point the bottom deepens by 0.5 feet and is composed of hard, glaciolacustrine clay which forms a 500 to 700-foot wide strip around the point. Lakeward the bottom again becomes sandy and the sand increases in thickness in a lakeward direction. The lake reaches a depth of ten feet at a distance of 200 feet offshore and 12 feet at 4000 feet offshore. The sand and gravel bottom, underlain by hard clay persists lakeward to the rocky reefs about three miles offshore (Figure 16).

The offshore reefs consist of bedrock and associated rock rubble and gravel. The topography of the reef tops ranges from rugged surfaces caused by bedrock pinnacles and large angular boulders, to smooth slabs of horizontally bedded rock. In places the exposed bedrock has the appearance of low stairs with steps dipping slightly to the east from the crest to the fringe of the submerged reef. All of the bedrock formations that form the reefs and shoals are carbonate rocks which contain abundant solution cavities, in many cases up to one or two cm in diameter. The bedrock itself is commonly masked by rubble composed of both autochthonous and glacial origin and ranging from small pebbles to boulders up to five feet in diameter. On the reefs, isolated patches of sand and gravel fill vertical joint cracks and small depressions in the bedrock; at the fringes<br>of reefs, sand and gravel beds or glacial till lap over the rock. During quiet periods the rocks are often covered by a thin layer of fluff, organic-rich silt, which can be several millimeters thick (Herdendorf, 1970).

Lakeward of the reefs the depths increase rapidly to 24 feet. Here the bottom is composed of mud (semi-fluid silt and clay-sized particles) and less than ten percent sand (Figure 16).

The lack of permanent siltation on the bedrock and gravel reefs make them the only suitable sites for "clean water" benthic organisms such as certain mayflies, caddisflies, isopods, and amphipods. These organisms are important in the food web of many of the commercial and game fish species of western Lake Erie. The absence of these invertebrate animals on or in the adjacent mud bottoms limits fish feeding to the reefs and inshore areas.

The reefs project above the bottom and they are generally areas of higher energy due to the force of waves and currents. These factors allow simulation of the environment found in the riffles of streams. Several species of fish, particularly walleye and white bass, appear to have enjoyed success in Lake Erie because of the availability of this type of habitat.

Because of the lack of shelter in the nearshore zone at Locust Point, except the intake and discharge structures, the area does not appear to support a large resident fish population. Monthly fish collections in this area (gill net, shore seine, and trawl) show great variability in species composition and relative abundance which strongly suggest a transient fish population. Results from 17 years of sampling at Locust Point indicate that 51 different species of fish have been captured (Table 4), but only ten species are of any real numerical or commercial significance. Alewife, carp, gizzard shad, white bass, emerald shiner, spottail shiner, yellow perch, channel catfish, freshwater drum, and walleye constitute over 97% of the total number that were captured (Reutter and Herdendorf, 1976).

The general flat or gently sloping lake bottom in the nearshore zone (within one mile of the shore) of Locust Point is broken only by the intake and discharge structures and uneven clay fill along the route of the buried pipelines. An ice barrier of rip-rap rock has been constructed on the lake side of the intake crib, and a scour prevention apron of similar material has been placed on the bottom lakeward of the discharge slot. In 1976, icthyoplankton sampling stations were established in the vicinity of the water intake discharge structure as well as control stations at similar distances offshore in an attempt to determine if these structures were inducing higher than normal fish spawning rates for their position offshore. The populations at these structures were within the normal range observed at the control station, indicating that the populations at the intake and discharge structures were not unusual for their position in the nearshore zone (Reutter and Herdendorf, 1976).

Intake Canal. In September 1974, the intake canal was poisoned to eliminate resident fish prior to the operation of the Station. During periods of 1972 and 1973 the intake canal was open to Lake Erie, and fish were free to enter the canal through an opening at the beachfront. In 1974 the canal was closed at the beach and the only water communication with the lake was via the 3000-foot-long, buried, intake pipe. Immediately prior to the poisoning, 22 trawls yielded 411 fish of 18 species. Trawls taken in the canal in October 1974, one month after poisoning, yielded only one fish, an adult carp, indicating that the kill was essentially complete. The benthic population was also destroyed in the process (Reutter and Herdendorf, 1975). Later trawls, in summer 1975, yielded 420 individuals of 13 species indicating some fish were entering the crib and traveling via the pipeline to the intake canal. The most common species found in the canal were white crappie, bullhead, black crappie, carp, yellow perch, and sunfish.

Trawls in the intake canal were not continued after 1975. However, there is evidence that white crappie, goldfish, and other species have developed resident populations in the intake canal, and these populations represent a sizeable percentage of the fish impinged on the traveling screens. The size, age classes and relative abundance of species impinged at the Station are markedly different than individuals captured with trawls and gill nets in the vicinity of the intake crib.

The intake canal is constructed of earthen walls and has a mud bottom over hard clay. The steep-sided walls of the canal preclude the development of extensive aquatic vegetation. The entire surface of the canal is unshaded. Velocities in the canal during 1978, are calculated to have had a maximum, minimum, and mean velocity of 0.16, 0.02, and 0.06 feet/sec, respectively.

#### Hydrology

Circulation Patterns. Western Lake Erie is dominated by the large in-flow of the Detroit River with a mean flow of approximately 210,000 cfs. The mid-channel flow of this river penetrates deep into the Western Basin, at times reaching the vicinity of Locust Point. The Maumee River, with an average flow of 4,700 cfs, is the second largest stream flowing into the lake and carries 37 percent of the sediment loading to the basin, but accounts for less than three percent of the total water drainage to Lake Erie. Maumee River water enters the lake through Maumee Bay where it divides into a northern flow along the Michigan shore and an eastern flow along the Ohio shore toward Locust Point. The Toussaint River, with an average flow of only 76 cfs, is a minor contributor to circulation patterns in the vicinity of Locust Point.

East of the dominating effect of the Detroit River, the prevailing southwest winds produce a clockwise surface flow around the Bass Islands to the northeast of Locust Point. However, this surface flow is often altered by changes in the direction, intensity, and duration of the wind. Strong winds from any direction can drive the surface currents over most of the basin toward the windward shore (Herdendorf, 1975). Current maps of western Lake Erie in the vicinity of Locust Point for various wind conditions are presented by Herdendorf (1970). Bottom currents have essentially the same pattern as surface flows in that part of the basin influenced by the Detroit River. However, in other parts of the basin bottom currents are commonly the reverse of and compensate for strong, wind-driven, surface currents.

Herdendorf and Braidech (1972) measured currents at 68 stations in the vicinity of Locust Point and the offshore reefs during a three-year study. The average recorded velocity for surface currents was 0.28 knots

(0.48 feet/sec) and 0.15 knots (0.26 feet/sec) for bottom currents. These velocities are not capable of eroding bottom material, but are able to transport fine sand, silt, clay, and fish eggs or larvae once they have Velocities in excess of 0.5 knots (0.84 feet/sec) were recorded on the reefs but not in the nearshore zone at Locust Point. The mean intake velocity for the Station is approximately half of the average bottom current velocity measured by Herdendorf and Braidech (1972).

Littoral Drift. Locust Point is at a position of diverging littoral (alongshore) drifts of sand which ordinarily would result in the beach being starved of sand because of movement east and west away from the headlands which form the point. However, the shore is apparently However, the shore is apparently maintained at near eguilibrium by replenishment from an extensive sand and gravel deposit which lies north of a narrow strip of compact glaciolacustrine clay that fronts the point beyond the sandy nearshore zone. Transportation of this material from offshore to the beach can be accomplished by at least three forces: 1) currents induced by wind action of Detroit River flow; 2) wave action; and 3) ice shove. Most of the sand probably migrates shoreward by wave action and currents generated by northeast and northwest storms. Evidence for the shoreward movement of sand can be found in the position of bars before and after major storms. For example, fathometer profiles of the lake bottom at Locust Point before (13 June 1972) and after (28 June 1972) tropical storm Agnes revealed that two offshore bars migrated 20 to 25 feet shoreward as a result of wave attack from the northwest storm (Herdendorf and Hair, 1972).

Thermal Conditions. Water temperatures in western Lake Erie range from  $32^{\circ}$  F in the winter to about 75<sup>0</sup> in late summer. The Western Basin frequently freezes from shore to shore in December and the ice cover breaks up in March and April. A shallow epilimnion develops early during the spring, but because the basin is so shallow, wind action causes efficient vertical mixing and by June the water becomes vertically isothermal. Diurnal microthermoclines are common in the summer, but prolonged periods of hot, calm weather can cause temporary thermal stratification, due to the heating of the surface water without the benefit of mixing. In 1953, such a situation resulted in severe oxygen depletion in the bottom water (Britt, 1955).

Water Quality. Nutrient overenrichment is the most significant water quality problem in western Lake Erie. Locust Point, being within the nearshore zone, is also characterized by low transparency, high concentrations of dissolved solids and warmer water temperature when compared with offshore water quality studies at Locust Point in July 1972 (Figure 17). Over the past 8 years most parameters have shown typical seasonal trends with only small variations from year to year. Trends for 8 water quality parameters from July 1972 through November 1979 are shown on<br>Figures 18, 19, and 20. Temperature and dissolved oxygen show normal Temperature and dissolved oxygen show normal seasonal trends for each year with only minor variations from one year to the next or over the entire period. DO appears to have undergone more depletion in 1976 and 1977 than in previous years or in 1978. Hydrogen-ion concentration (pH) and alkalinity remained fairly stable over the period. Transparency, turbidity, phosphorus, and conductivity have shown radical variations which are probably due to storms and dredging activities that

have disturbed the bottom sediments. Phosphorus levels were low in 1977, 1978, and 1979 compared to earlier years. In general however, no significant deviations from the normal quality of the water in this part of western Lake Erie have been observed during the past seven years.

### IMPACT APPRAISAL

#### Water Quality

### Procedures and Results

Water quality measurements during the period April 1974 to November 1979 were used for the purposes of this appraisal. The results of the water quality monitoring program are contained in semi-annual reports (1974-1976) and annual reports (1977 - 1979) of the Toledo Edison Company to the U.S. Nuclear Regulatory Commission. The data used included Station No. 13 (500 feet east of the discharge structure) and Station No. 8 (adjacent to the water intake crib). Station No. 13 serves as the station most likely to be impacted, while Station No. 8 serves as a control station (Figure 17). Each station was visited once a month during the ice-free period of the year (normally April-November). Surface and bottom water samples were taken at each station and were analyzed in accordance with the procedures listed in Table 5. Because the intake and discharge structures are located at or near the bottom, bottom samples were used for comparing pre-operational and operational conditions. Tables 6 to 23 summarize pre-operational and operational data for the 18 water quality parameters  $\ell^*$  at the intake and discharge stations. These data are displayed graphically for the discharge station on Figures 21 to 38. The following discussion summarizes the comparison for each of the parameters.

> Dissolved Oxygen. During both the pre-operational and operational period DO showed a typical trend of high values in the spring and fall with low concentrations in the summer. Operational concentrations were considerably lower than the pre-operational range in April and November, but not during the critical summer months (Figure 21).

> Hydrogen-ions (pH). Throughout the pre-operational and operational period pH values remained relatively stable, never exceeding 9.0 or falling below 7.5. The operational values showed more variability than the nearly straight-line mean concentration for the pre-operational period (Figure 22). However, both periods had a mean pH of 8.3.

> Transparency. Both the pre-operational and operational measurements showed the lowest water clarity in the spring, the best transparency in the summer, and intermediate clarity in the fall. In general, operational values were within the range of pre-operational values throughout the year (Figure 23).

> Turbidity. 'Being somewhat the reciprocal of transparency, the lowest readings occurred in the summer, the highest in spring and intermediate values in the fall for the pre-operational period. Operational values showed a general decreasing trend throughout the year, with only a slight rise in the fall. However, values for May, June, and

September well exceeded the pre-operational ranges for those months  $(Fiquare 24)$ .

Suspended Solids. This parameter, like turbidity showed a "U" shaped trend during the pre-operational period with summer concentrations being the lowest. Like transparency and turbidity, high particulate material in the water during the spring and fall months of the operational period yield readings in excess of the pre-operational ranges for these months (Figure 25).

Conductivity. This parameter is a measure of the ionized material in the water and it also shows high concentrations in the spring for both the pre-operational and operational periods. Only conductivity values in April for the operational period exceed the range for this month during the pre-operational period (Figure 26).

Dissolved Solids. The concentration of dissolved substances in the water during pre-operational and operational periods were relatively similar, with the operational data falling within or nearly within the pre-operational range for each month. Operational concentrations were somewhat lower than pre-operational conditions for April and October, while September was slightly higher (Figure 27).

Calcium. This element, one of the most common found in Lake Erie water, showed relatively consistent values during both the pre-operational and operational period. High concentrations typified the spring with considerably lower values in the summer and fall. Only in November did operational concentrations exceed the range of pre-operational data (Figure 28).

Chloride. Operational chloride concentrations were within the range of pre-operational concentrations during six of the eight months for which comparative data is available. The greatest discrepancy occurred in April<br>and November. Pre-operational data show a progressive decrease in Pre-operational data show a progressive decrease in concentration throughout the year, while operational data indicate a more "U" shaped trend (Figure 29).

Sulfate. Both pre-operational and operational sulfate data show relatively consistent concentrations throughout the year with somewhat higher values in the spring. Operational data were more erratic, with four months above the pre-operational range and one month below the range (Figure 30).

Sodium. A trend similar to that of sulfate was noted for sodium. Operational data again showed greater variability with two months above and one month below the range for pre-operational data. April and November yielded the highest concentrations for the operational period, both beyond the pre-operational range (Figure 31).

Magnesium. This parameter showed the least agreement between pre-operational and operational data of any of those tested. Operational concentrations exceeded the range of pre-operational data for all months except May. In April, the operational mean value was nearly double the pre-operational mean concentration (Figure 32).

Total Alkalinity. This parameter showed considerable variability in both the pre-operational and operational data, with the highest values occurring in the spring and fall during the pre-operational period and in the spring and summer during operation. April, July, August, and November were periods when operational values exceeded pre-operational ranges, while May and June were months of relatively low operational alkalinity (Figure 33).

Nitrate. Serving as a biological nutrient, this parameter<br>lates\_widely\_in\_response\_to\_plankton\_productivity. Concentrations fluctuates widely in response to plankton productivity. during both the pre-operational and operational periods were highest in the spring but decreased in the summer as this material was utilized by algae. Fall concentrations increased as algal productivity declined. Concentrations during both periods were relatively consistent with operational values being somewhat higher, particularly in June, August, and November (Figure 34).

Phosphorus. This parameter is also an important biological nutrient and, like nitrate, shows seasonal variations such as high spring and low summer concentrations. Pre-operational and operational data were relatively consistent throughout the year, except for May which showed a considerably higher mean concentration during the pre-operational period (Figure 35).

Silica. As a necessary material for diatom cells, silica also under goes seasonal changes in concentration. As the growing season progresses this material greatly declines in the water. Both pre-operational and operational data show the same seasonal trend. Operational concentrations exceeded the pre-operational ranges for May and November (Figure 36).

Biochemical Oxygen Demand. BOD levels were relatively consistent throughout the year for both the pre-operational and operational periods. Values were highest in the spring and lowest in the fall. All of the operational concentrations fall within the range of pre-operational data, except for June (Figure 37).

Temperature. Both pre-operational and operational data show typical seasonal temperature trends for Lake Erie; and both data sets are relatively consistent. Most of the operational values fall within the range of pre-operational data (Figure 38).

#### Appraisal

**/^\*1**

In general the quality of Lake Erie water in the vicinity of the Station's discharge structure has remained relatively constant over the past seven years (Figures 18, 19, and 20). In comparing the 18 water quality parameters during the ice-free months for the pre-operational versus the operational period (Figures 21 to 38), it can be seen that there is a 67% agreement (operational data within pre-operational range) between the two data sets. This is a relatively good agreement considering the short data base, particularly for the operational period (Figure 39).

Table 24 summarizes this comparison and provides an indication of the degree of difference between the two periods. In general the

concentrations of dissolved and suspended substances were higher during the operational period, particularly: magnesium, silica, nitrate,<br>turbidity, and suspended solids. Dissolved oxygen was lower after turbidity, and suspended solids. Dissolved oxygen was lower after operation. The magnitude of these differences was not great and seemed to be caused by the general condition of the nearshore waters of western Lake Erie rather than Station operation. For example, Table 17 shows that magnesium was not only high at the discharge (Sta. No. 13) but also high at the water intake (Sta. No. 8) which serves as a control station.

Table 25 indicates the percent change in water quality at the lake intake (Station 8) and discharge (Station 13) from the pre-operational period through the operational period. Dissolved oxygen and phosphorus showed the largest decreases in concentration (7 and 35 percent, respectively), while sulfate, magnesium, BOD, silica, chloride, turbidity, and suspended solids all had increases greater than 5%. In all cases where an increase in excess of 5% occurred at the discharge station, a similar increase was also observed at the control station. These observations further substantiate the conclusion that most of the changes are due to general lake conditions, and not localized changes resulting from Station<br>operation. The decrease in phosphorus concentration is consistent with The decrease in phosphorus concentration is consistent with other nearshore measurements in western Lake Erie which indicate a decline in this substance as a result of pollution abatement programs.

With the inherent limitations of the data base acknowledged, shortterm degradation of Lake Erie water quality can not be demonstrated as a result of Station operation. The stability of water quality in the *^* vicinity of Locust Point is well-documented; long-term deleterious impacts resulting from station operation are unlikely.

# Plankton Studies

#### Procedures

Plankton monitoring at the Davis-Besse Nuclear Power Station has been completed approximately monthly during ice-free periods since 1973<br>(Table 26), The stations at which samples were collected each vear are The stations at which samples were collected each year are listed in Table 27 and shown on Figure 17. In 1973 only quantitative zooplankton samples were collected, while both quantitative zooplankton and phytoplankton samples were collected in all other years. The preservation techniques have been modified occasionally as new techniques to make specimen identification easier appeared in the literature. However, no modifications which would have quantitatively affected the results were made, and formalin was always the final preservative. Two vertical tows, bottom to surface, were collected at each station for phytoplankton and zooplankton with a Wisconsin plankton net (12 cm mouth; 0.064 mm mesh in 1973 and 1974 and 0.080 mm mesh from 1975-1979). Each sample was concentrated to 50 ml and preserved. The volume of water sampled was computed by multiplying the depth of the tow by the area of the net mouth. Three 1-ml aliquots were withdrawn from each 50-ml sample and placed in counting cells.

Whole organism counts of the phytoplankton were made from 25 random Whipple Disk fields in each of the three 1-ml aliquots from each of the 2 samples. When filamentous forms number 100 or more in 10 Whipple fields, they were not counted in the remaining 15 fields. Identification was carried as far as practicable, usually to the genus or species level.

All zooplankters within each of the three 1-ml aliquots from each of the 2 samples were counted by scanning the entire counting cell with a microscope. Identification was carried as far as practicable, usually to the genus or species level.

# Phytoplankton Results

 $\sqrt{2}$ 

The results of the phytoplankton monitoring program were presented in the semi-annual reports (1974-1976) and annual reports (1977 - 1979) of the Toledo Edison Company to the U.S. Nuclear Regulatory Commission. This report summarizes the findings presented in these earlier reports through graphic presentations of monthly densities of the major phytoplankton components, Bacillariophyceae, Chlorophyceae, and Myxophyceae, encountered yearly from 1974-1979 (Figures 40 -45). Figure 46 presents the monthly estimates of the total phytoplankton density from 1974 through 1979.

Table 28 and Figures 47 - 50 summarize the above data in a different manner by combining all monthly density estimates from all years and all stations and comparing pre-operational means, minima, maxima, and standard deviations with operational results. Table 29 and Figures 51 - 53 use this same technique to compare the total phytoplankton densities observed at Station 8 (intake structure), Station 13 (plume area), and Station 3 (control station). A discussion of these comparisons follows.

Diatoms. Both pre-operational and operational densities were high during the spring and fall, and low during the summer (Figure 47). Spring densities were highest. This is typical for western Lake Erie and as one<br>would expect since diatoms are cold-water forms. Operational densities would expect since diatoms are cold-water forms. observed during the spring and fall were larger than the corresponding pre operational values. However, operational standard deviations (when able to be calculated) overlapped the pre-operational standard deviations.

Green Algae. Chlorophycean densities, in general, were much lower than diatom densities or blue-green algae densities during the pre operational and the operational studies. Furthermore, these green algae population densities are much less predictable seasonally than diatoms. Reutter (1976) has demonstrated that green algae densities parallel transparency closely and are opposite to turbidity and, therefore, are often controlled by factors such as the wind, which affects transparency by suspending bottom sediments through wave action. However, most of the suspending bottom sediments through wave action. monthly samples collected during the operational period fell within the range established during the pre-operational period, and for those which were outside the range (July, September, and November), the standard deviation of the operational period overlapped the standard deviation of the pre-operational period (Figure 48).

Blue-Green Algae. Myxophycean populations during both the pre operational and operational periods showed tendencies toward sudden,<br>large, mid-summer pulses (Figure 49). Operational densities were large, mid-summer pulses (Figure 49). generally larger than pre-operational densities. However, with the exception of October and November, the operational standard deviations always overlapped the pre-operational standard deviations.

Total Phytoplankton. The total phytoplankton density, i.e., the sum total of the 3 major component groups previously discussed and several other minor classes, was higher during most of the operational study than during the pre-operational study (Figure 50). However, with the exception of April and October, the standard deviations of the means observed.during the operational study overlapped the standard deviations from the pre operational study.

### Zooplankton Results

The results of the zooplankton monitoring program were presented in the semi-annual reports (1974-1976) and annual reports (1973, 1977, 1978 and 1979) of the Toledo Edison Company to the U.S. Nuclear Regulatory<br>Commission. This report summarizes the findings presented in these This report summarizes the findings presented in these earlier reports through graphic presentations of the monthly densities of the total zooplankton population and its major components, rotifers, copepods, and cladocerans encountered yearly from 1972 -1979 (Figures 54 -

Table 30 and Figures 58 - 61 summarize the data in a different manner by combining all monthly density estimates from all years and all stations<br>and comparing pre-operational means, minima, maxima, and standard and comparing pre-operational means, minima, maxima, and standard deviations with operational results. Table 31 and Figures 62 - 64 use this same technique to compare total zooplankton densities observed at Station 8 (intake structure), Station 13 (plume area), and Station 3 (control station). A discussion of these comparisons follows.

Total Zooplankton. The total zooplankton population density, i.e., a sum total of the major zooplankton groups (rotifers, copepods, and cladocerans) and any minor classes or orders, has usually exhibited two pulses, one in the late spring or early summer and a smaller pulse in the<br>fall. This is true of both pre-operational and operational results. This is true of both pre-operational and operational results, although operational densities were generally lower than pre-operational densities (Figure 58).

Rotifers. Rotifer densities at Locust Point during the operational period were lower for every month than the mean value from the preoperational period for the same month (Figure 59). However, the operational monthly mean was below the pre-operational monthly range only during June and November, and the operational monthly mean was always less than two standard deviations from the pre-operational mean.

Copepods. Copepod densities at Locust Point during the preoperational study generally exhibited spring pulses (Figure 60). This was also the case during the operational study, except the pulse was somewhat smaller than those observed during the pre-operational study. As observed with the rotifers, operational monthly densities were never more than two standard deviations from the pre-operational mean (Figure 60).

Cladocerans. Cladoceran densities at Locust Point during both the pre-operational and operational studies have exhibited spring (or early summer) and fall pulses (Figure 61). However, during the operational period the two pulses were less distinct. With the exception of August, none of the monthly operational densities were more than two standard deviations from the pre-operational mean.

### Appraisal

Prior to the appraisal of the effects of unit operation on the zooplankton and phytoplankton communities, some assistance in interpreting these results is warranted. First, one should bear in mind that when sampling the same population eight months each year for seven years, and plotting data with monthly minima and maxima, as in this report, eight minima and eight maxima will be generated. That is, there will be seven values for each of the eight months, or one value for each month from each of the seven years. Each of the eight months will have a minimum value and a maximum value, and, since there are eight months, there will be a total of eight minimum values and eight maximum values (one of each for each If there is nothing unusual about the environmental conditions which existed during any of the seven years, then each year would have an equal chance (probability) of producing several monthly minimum or maximum values. Assuming each year does have an equal probability of producing these minima and maxima, and since there are eight monthly minimum values and eight monthly maximum values, each year of the seven years would produce 1.14 of the monthly minimum values and 1.14 of the monthly maximum values. This is pointed out to demonstrate that it is natural for any year to produce a population extreme (monthly minimum or maximum value). Consequently, it should not be automatically viewed as a unit produced effect if any operational variable is above or below the pre-operational range.

Another point useful in the interpretation of these results involves the distance of the operational monthly mean from the pre-operational mean. A general "rule-of-thumb" is that when dealing with a normal mean. A general "rule-of-thumb" is that when dealing with a normal distribution, the area within one standard deviation on either side of the mean will contain approximately 66 percent of the values, two standard deviations would contain approximately 95 percent of the values, and three standard deviations would contain approximately 99 percent of the values.

As a final aid in interpreting these results, population densities are presented from a control station (unaffected) to allow comparison with the discharge where the impact should be greatest. This allows a distinction to be made between unusual values caused by unit operation and unusual results which are typical of the entire lake due to an unusual set of climatic or biological conditions -- natural variation.

Between September 1977 and the end of 1979, the operational period, plankton samples were collected on 18 occasions. On these dates, the station was operating at 90 percent capacity, 8 percent capacity, 100 percent capacity, 99 percent capacity, 48 percent capacity, and 0.0 percent capacity on the remaining 13 sampling dates.

Phytoplankton. Reutter and Fletcher (1980) summarized the results of phytoplankton sampling at Locust Point and concluded that "populations observed at Locust Point during 1979 are similar to those of previous years and appear typical for those occurring in the nearshore waters of the Western Basin of Lake Erie." This report has taken the results compiled by Reutter and Fletcher a step farther by computing means, ranges, and standard deviations for the pre-operational period and by adding the results from the last portion of 1977 to those of 1978 and 1979 to summarize the operational period.

Operational phytoplankton densities were somewhat larger than pre operational densities (Figure 50). This appears to be a general trend, as the operational values of the three major phytoplankton groups were never below the pre-operational range and often above it. Due to the unusually harsh winters of 1978 and 1979, it is likely that these differences were caused by natural weather conditions.

Figures 51 - 53 present phytoplankton densities at the station intake (Station 8), discharge (Station 13), and a control station (Station It would probably be safe to use the station intake as a control station, however, as an extra measure of caution Station 3, 3000 feet northwest of the discharge, was selected as a control. Using this comparative technique, any difference between pre-operational and operational data observed at the discharge which was also observed at the intake or Station 3 would obviously have been due simply to natural variation in population densities. The only large differences between operational and pre-operational data at the discharge were unusually high spring and fall population densities, and, since these were also observed at the intake and Station 3, they were obviously a natural phenomenon and not caused by unit operation.

In conclusion, to date, operation of the Davis-Besse Nuclear Power Station, Unit 1, has not had a significant effect on Lake Erie phytoplankton densities.

Zooplankton. Reutter and Fletcher (1980) summarized the results of zooplankton sampling at Locust Point through 1979 and concluded that "populations observed in 1979 should be considered typical for the south shore of the Western Basin of Lake Erie." This report has taken the results compiled by Reutter and Fletcher a step farther by computing means, ranges, and standard deviations for the pre-operational period and by adding the results from the last portion of 1977 to those of 1978 and 1979 to summarize the operational period.

Zooplankton operational densities, though generally similar to pre operational densities, were often somewhat lower than the corresponding pre-operational monthly density (Figures 58 - 61). However, as with the phytoplankton, these differences should not be interpreted as due to unit operation, for it appears that zooplankton densities even in unaffected areas were lower during the operational period (Figures 62 - 64).

Consequently, these differences were obviously attributable to natural variation and not unit operation.

The obvious conclusion is that to date, operation of the Davis-Besse Nuclear Power Station, Unit 1, has not had a significant effect on Lake Erie zooplankton densities.

### Benthic Studies

### Procedures

Benthic macroinvertebrate densities in the vicinity of the Davis-Besse Nuclear Power Station were monitored at approximately monthly intervals during ice-free periods (normally April through November) from 1973 through 1976, and at invervals of approximately 60 days during the ice-free periods of 1977, 1978, and 1979 (Table 32). The stations at which samples were collected each year are listed in Table 33 and shown on Figure 17. Population densities were sampled with a Ponar dredge (Area= $0.052 \text{ m}^2$ ). Three replicate grabs were collected at each station on each date from 1974 through 1979, whereas one sample was collected at each station on each date<br>during 1973, Samples were sieved on the boat through a U.S. #40 soil during 1973. Samples were sieved on the boat through a U.S. sieve, preserved in 10% formalin, and returned to the laboratory for identification and enumeration. Individuals were identified as far as practicable (usually to genus; to species when possible). Results were reported as the number of organisms per  $m^2$ .

#### Results

The results of the benthos monitoring program were presented in the semi-annual reports (1974 - 1976) and annual reports (1973, 1977, 1978, and 1979) of the Toledo Edison Company to the U.S. Nuclear Regulatory Commission. This report summarizes the findings presented in these earlier reports through a graphic presentation of the monthly benthic macroinvertebrate densities encountered yearly from 1972 - 1979 (Figure 65).

Table 34 and Figures 66 - 70 summarize the data in a different manner by combining all monthly density estimates for the major benthic groups from all years and all stations during the pre-operational study, and comparing these pre-operational monthly means, minima, maxima, and standard deviations to operational results. Table 35 and Figures 71 - 73 use this same technique to compare total benthic macroinvertebrate densities observed at Station 8 (intake structure), Station 13 (discharge area), and Station 3 (control station). A discussion of these comparisons follows.

Total Benthic Macroinvertebrates. The population densities of all benthic macroinvertebrates, i.e., the sum total of the major benthic groups (Coelenterata, Annelida, Arthropoda, and Mollusca), were generally the highest in the late summer and fall during the pre-operational study. During the operational study the highest densities occurred slightly earlier in the summer and fall (Figure 66). Operational densities were *very* close to the pre-operational mean during every month except September, when they were slightly lower than the pre-operational minimum.

Coelenterata. Pre-operational coelenterate population densities generally produced peaks in the spring and fall (Figure 67). During the operational study only a fall peak was observed. However, operational density estimates were always within one standard deviation of the pre-operational mean.

Annelida. Benthic annelid densities during both the pre-operational and operational studies showed peaks in late summer or early fall (Figure<br>68), However, all monthly operational results were within the 68). However, all monthly operational results were within the<br>pre-operational range or within one standard deviation of the pre-operational range or within one standard deviation of pre-operational mean, except May and September, when the operational densities were slightly lower.

Arthropoda. Both pre-operational and operational benthic arthrdpod densities peaked during the summer and fall (Figure 69). Operational densities were above the pre-operational maxima during May, June, and July, and below the minimum during October.

Mollusca. Benthic mollusc densities were extremely low (five was maximum during the seven-year study period) and variable, and, consequently, pre-operational/operational differences are difficult to detect (Figure 70). However, nothing unusual was observed during the operational period.

### Appraisal

Initially it should be pointed out, as discussed in the plankton appraisal (see page 18), that operational densities which fall outside the pre-operational range may be due to natural variation and not related to unit operation. To allow comparisons of ambient densities with densities at the unit discharge, population densities have been presented from Station 3, a control station located 3000 ft northwest of the unit discharge structure, the same distance from shore as the discharge and at approximately the same water depth. These comparisons allow one to more accurately assess the causes of observed differences - natural variation or unit operation.

During what is defined as the operational period, samples were collected on ten occasions. On these ten occasions, the unit was operating at 98 percent on one occasion, 100 percent on another, 99 percent on another, and not operating on the remaining seven dates. While this is *very* critical to water quality and plankton results, it is somewhat less important when observing benthic communities. Benthic communities are much less mobile than plankton or fish, and, therefore, are generally considered to be good pollution indicators, even of intermittent pollutants or environmental changes. The rationale is that even if the unit were not operating on the sampling date, a large portion of the community sampled would have been present when the unit was operating.
This is not true of plankters, and fish are capable of leaving when unfavorable conditions exist and then returning quickly when the conditions are improved.

Reutter (1980a) summarized the results of benthic macroinvertebrate sampling at Locust Point through 1979 and concluded that "populations found at Locust Point during 1979 must be considered typical for those of the nearshore waters of the Western Basin of Lake Erie ... no significant environmental changes due to unit operation were observed." This report has taken the results compiled by Reutter a step farther by computing means, ranges, and standard deviations for the pre-operational period and by adding the results from the last portion of 1977 to those from 1978 and 1979 to summarize the operational period.

Benthic macroinvertebrate densities observed during the operational study were within the limits established during the pre-operational study on all" but one occasion. A review of Figures 71 - 73 shows that variability in population densities was widespread and not related to unit operation. Operational densities observed at the discharge (Figure 72) more closely resembled pre-operational densities than did those observed at the intake (Figure 71) or Station 3 (Figure 73), which were designed to be the control stations. Results at Station 3, where no construction has ever occurred and well away from the intake and discharge, are graphic examples of the discussion at the beginning of this appraisal section, showing that natural variability can produce values far from the pre-operational densities. Furthermore, this type of variability is to be expected in the Locust Point vicinity, a shallow wave-swept zone with shifting substrate.

In conclusion, to date, operation of the Davis-Besse Nuclear Power Station, Unit 1, has not had a significant effect on Lake Erie benthic macroinvertebrate densities.

### Fisheries Population Studies

#### Procedures

Fish populations in Lake Erie at Locust Point in the vicinity of the Davis-Besse Nuclear Power Station were monitored at approximately monthly intervals during ice-free periods (normally April - November) from 1973<br>through 1979. Fish were collected by three sampling techniques, Fish were collected by three sampling techniques, experimental gill nets, shore seines, and trawls.

Experimental gill nets (125 feet long, consisting of five 25-ft contiguous panels of  $\frac{1}{2}$ , 3/4, 1, 1<sup>1</sup>/<sub>2</sub>, and 2-inch bar mesh) were set parallel to the intake pipeline at Station 8 (intake) and parallel to the discharge pipeline at Station 13 (discharge or plume area) from 1973 through 1979 (Table 36). During 1977, 1978, and 1979, nets were also placed at Stations 3 and 26 to serve as controls (Figure 17). Each net was fished at the lake bottom for approximately 24 hours. Results were reported as catch per unit effort (CPE), where one unit of effort was equal to one 24-hour set with one net.

Shore seining was conducted at Stations 23, 24, and 25 with a 100-ft bag seine  $(\frac{1}{4}$ -inch bar mesh). The seine was stretched perpendicular to the shoreline until the shore brail was at the water's edge. The far brail was then dragged through a 90<sup>0</sup> arc back to shore. Two hauls were made at each station in opposite directions.

Four 5-minute bottom tows with a 16-ft trawl (1/8-inch mesh bag) were conducted on a transect between Stations 8 (intake) and 13 (plume area) at a speed of 3 - 4 knots. Starting in 1977, tows were also made on a transect between Stations 3 and 26 for comparative purposes.

All fish captured by each technique were identified, enumerated, weighed, and measured (Trautman, 1957; Bailey, et al., 1970). All results were keypunched and stored on magnetic tape at The Ohio State University Computer Center.

#### Results

The results of the fishery population monitoring program are contained in the semi-annual reports (1974 - 1976) and the annual reports (1973, 1977, 1978, and 1979) of the Toledo Edison Company to the U.S. Nuclear Regulatory Commission. These reports have shown gill netting to be the superior sampling technique for measuring the impact of unit operation for several reasons:

- *afn\* 1. gill nets can be set right at the point of impact, are v. relatively unbiased sampling devices, and collect adequate sample sizes (quantities of fish);
	- 2. shore seines sample mainly young-of-the-year fish and, consequently, are subject to sudden pulses following spawning;
	- 3. shore seines sample at locations over 1000 feet from the point of discharge;
	- 4. trawls have been shown to collect too few fish.

Consequently, although the results of shore seining and trawling have greatly increased our ability to interpret yearly results, gill nets have proven to be the most effective assessment tool, and, therefore, these results and discussions will pertain mainly to this gear type.

Fifty-one fish species have been collected at Locust Point since However, the fish community at Locust Point has consistently been dominated by seven species: alewife, emerald shiner, freshwater drum, gizzard shad, spottail shiner, white bass, and yellow perch. These seven species generally constitute well over 90 percent of the annual catch by the sampling program. The monthly mean, minimum, maximum, and standard deviation of the number of each of these species, except emerald shiner, collected in the gill net set at the discharge have been presented in Table 37 and Figures 74 - 81. Emerald shiners are seldom collected in gill nets of these mesh sizes, so they were not included in<br>the tabulations. However, due to their economic importance, channel However, due to their economic importance, channel

catfish and walleye were added to the list. Table <sup>38</sup> and Figures <sup>82</sup> - <sup>85</sup> summarize the gill net results by presenting pre-operational means, minima, maxima, and standard deviations and comparing them to operational results at Stations 8 (intake), <sup>13</sup> (discharge or plume area), 3 and <sup>26</sup> (controls).

Alewife. Alewife densities in the vicinity of the unit discharge during both the operational and pre-operational periods were generally highest during the late summer and early fall (Figure 74). The maximum pre-operational catch was 322, while 136 was the maximum catch during the operational period (Table 37). Although operational catches were Although operational catches were generally lower than pre-operational catches, they were always within the pre-operational range.

Channel Catfish. Channel catfish catches during both the pre-<br>operational and operational studies were greatest during the summer (Figure 75). They were seldom a significant component of the catch, as <sup>18</sup> was the maximum pre-operational catch and 6 was the maximum operational catch (Table 37). The pre-operational and operational catches were quite similar, and all operational means were within the pre-operational range.

Freshwater Drum. During both the pre-operational and operational studies, freshwater drum were most abundant during the summer (Figure 76). The maximum catch during the pre-operational study was 50, while 75 was the maximum operational catch (Table 37). With the exception of June, which was higher, all operational catches were within the range established during the pre-operational study.

Gizzard Shad. Gizzard shad densities during both the preoperational and operational studies were always greatest during the late summer and fall (Figure 77). The maximum pre-operational catch was 184, while 291 was the maximum operational catch (Table 37). The monthly preoperational and operational mean catches were generally quite similar, and all but one of the operational means were within the pre-operational ranae  $(Figure 77)$ .

Spottail Shiner. Spottail shiners were always most abundant during the month of May (Figure 78). In fact, with the exception of April and June, the minimum catch in May was greater than the maximum catch of any of the other months during the pre-operational period. The operational catch was within the range established during the pre-operational period durinq all months but September.

Walleye. Walleye catches during both the pre-operational and operational studies were greatest during the summer (Figure 79). This species was never a significant portion of the catch, as 15 was the maximum prior to plant operation and 8was the maximum afterwards (Table 37). With the exception of August, when the operational catch was above the range of pre-operational catches, all catches after the unit began operation were within the range of catches prior to unit operation.

White Bass. White bass were generally most abundant during the summer (Figure 80 and Table 37). The magnitude of the pre-operational and operational catches were very similar, but the pre-operational peak occurred in August whereas the operational peak occurred in June. With the exception of June and July, when the operational catch was above the pre operational mean, all operational values were within the range established during the pre-operational study.

Yellow Perch. Yellow perch generally occurred in similar numbers from month to month during the pre-operational period with a slight increase in the early fall, followed by a decrease to low densities in November (Figure 81). Operational densities were of similar magnitude during all months but August when they were higher than the pre-operational mean but *very* close to the pre-operational maximum for September.

#### Appraisal

In the appraisals of the phytoplankton, zooplankton, and benthos sections, it was shown that extreme values, i.e., either maxima or minima, in addition to being potentially due to unit operation, will occur by chance alone, due to natural variation. Furthermore, the magnitude of the standard deviation gives one a good indication of the magnitude of natural variation to be expected.

The above statements are hardly necessary when evaluating the impact of unit operation on the fishery populations in the vicinity of the Davis-Besse Nuclear Power Station, for there was little or no variation out of the pre-operational range during the operational period for the eight major species (Figures 74 - 81). On the 17 sampling dates during the operational period, the unit was operating at above 90 percent capacity on four dates, 15.0 percent capacity on another, and not operating on the remaining twelve dates.

Another way to measure impact and an approach which allows us to include all species (not just the major eight) is to compare catches at the discharge (Station 13) and those at the intake (Station 8) with two control stations (Figures 82 - 85 and Table 38). This method shows that the only operational catches at the intake and discharge which were outside the pre-<br>operational range occurred during November (Figures 82 and 83). Both of operational range occurred during November (Figures 82 and 83). these catches were above pre-operational data which is an indication that it was either a lake-wide occurrance, or a case of fish being attracted to the rip-rap material which was placed around these structures to prevent bottom scouring and ice damage. However, since an identical November increase occurred at the control stations (Figures 84 and 85), natural variation, not unit operation, should be considered the cause.

In conclusion, to date, operation of the Davis-Besse Nuclear Power Station, Unit 1, has not had a significant effect on Lake Erie fish populations at Locust Point.

#### Ichthyoplankton

#### Procedures

Ichthyoplankton was sampled at Locust Point in the vicinity of the Davis-Besse Nuclear Power Station from 1974 through 1979 with a 0.75-meter

diameter oceanographic plankton net (No.00, 0.75 mm mesh). Each sample consisted of a 5-minute circular tow at 3 to 4 knots. Samples were collected at the surface and bottom of each station.

> Sampling was conducted at the following stations during the following years: 1974, Stations 8 and 12; 1975, Stations 8, 12, and Toussaint Reef (Figure 15); 1976, Stations 3, 8, 13, 26, 28, 29, and Toussaint Reef; 1977, 1978 and 1979, Stations 3, 8, 13, 29, and Toussaint Reef. Toussaint Reef was used for comparisons since the Ohio Division of Wildlife considers it a spawning location. Each sample was preserved in 5 percent formalin and returned to the laboratory for sorting and analysis. Samples were generally collected at approximately 10-day intervals from April through August. Sampling was terminated at the end of August to add a margin of safety to the USEPA (Grosse Ile Office) sampling program for the Western Basin of Lake Erie which terminated each year in July (Table 39).

> From 1974 to 1976, a single sample was collected at each depth of each station, and results were reported as the number of individuals per 5 minute tow. In 1977, 1978 and 1979, duplicate samples were collected at the surface and bottom of each station, and the net was equipped with a calibrated General Oceanics flowmeter to allow presentation of the results as the number of individuals per 100 m° of water. All specimens were identified and enumerated using the works of Fish (1932), Norden (1961a and b), and Nelson and Cole (1975).

#### Results

The results of the ichthyoplankton analyses have been thoroughly described in the semi-annual reports (1974 - 1976) and annual reports (1977, 1978, and 1979) of the Toledo Edison Company to the U.S. Nuclear Regulatory Commission. Since the reporting of results changed (catch per unit effort vs. no./100 m') during the course of the study, direct comparisons of results from 1977, 1978, and 1979 with those of the early pre-operational years, 1974 - 1976, are not possible. However, comparisons of the relative portions of the total density constituted by each species are possible.

Ichthyoplankton populations varied greatly from 1974 - 1979. Emerald shiners constituted 81 percent of the 1974 larvae, 1 percent of the 1975 larvae, 60 percent of the 1976 larvae, 3 percent of the 1977 larvae, 14 percent of the 1978 larvae, and 3 percent of the 1979 larvae. Yellow perch constituted 5 percent of the 1974 larvae, 70 percent of the 1975 larvae, 4 percent of the 1976 larvae, 26 percent of the 1977 larvae, 2 percent of the 1978 larvae, and 11 percent of the 1979 larvae. Gizzard shad appear to have increased significantly, reaching 34 percent of the 1976 larvae, 56 percent of the 1977 larvae, 69 percent of the 1978 larvae, and 82 percent of the 1979 larvae. It is felt that the above described variability is largely due to the fact that schooling populations are being sampled. Consequently, when the net is drawn through a school the density appears quite high. This is also quite dependent on the seasonal frequency of sampling. For example, if the weather allows more frequent spring sampling but prohibits summer sampling, then spring species such as perch and walleye appear relatively more abundant.

Nineteen seventy-eight was the second year that walleye constituted a significant portion of the catch. However, as noted in 1977, adult populations throughout the Western Basin are increasing greatly (Scholl, 1978). These walleye larvae contributed to the 53 percent<sub>s</sub> increase observed in larval densities from 1977 (mean density = 37.0/100 m<sup>3</sup>) to 1978 (mean density =  $56.6/100 \text{ m}^3$ ). However, gizzard shad were the major source of this increase as their mean densities increased from  $20.7/100$  m<sup>3</sup> in 1977 to 38.9/100 m3 in 1978. Yellow perch densities decreased significantly from 9.5/100 m $^{\circ}$  in 1977 to 1.2/100 m $^{\circ}$  in 1978. This decrease is similar to that observed by the Ohio Division of Wildlife for the adult population (Scholl, 1979).

The 1979 ichthyoplankton density (66.79/100 m°) was 18 percent greater than the 1978 density (56.6/100 m3) (Reutter, 1979). Although walleye densities decreased from 6.1/100 m $^{\circ}$  to 0.15/100 m $^{\circ}$ , the loss waş more than offset by yellow perch densities which increased from 1.2/100 m in 1978 to 7.46/100 m' in 1979 and gizzard shad densities which increased from 38.9/100 m $^{\circ}$  in 1978 to 54.64/100 m $^{\circ}$  in 1979. It appears that walleye and yellow perch densities will fluctuate yearly, however, a definite increasing trend is emerging for gizzard shad densities.

In 1976, control stations (3 and 29) were established on either side of the intake (Station 3)/discharge complex (Station 13) to determine if unusually large fish larvae populations were occurring due to possible spawning in the rip-rap material around these structures. This does not appear to be occurring to any significant degree as Station 13 (plume area) exhibited densities similar to Station 3 (control), and Station 8 (intake) exhibited the lowest densities. These lower densities observed at Station 8 are probably due to the fact that this station is the farthest from shore and in the deepest water.

#### Appraisal

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Ichthyoplankton at Locust Point in the vicinity of the Davis-Besse Nuclear Power Station, Unit 1, was sampled for two major reasons: 1) to determine if unit operation had a significant effect on densities in the area; and 2)to provide the ichthyoplankton densities to be used for the entrainment estimates. The first goal of the program is reasonable, and Reutter (1980b) stated, "due to the similarity between test and control stations, there is no indication that the activities of the plant have significantly altered these populations." To date, this assessment is true, although it should be tempered somewhat since on the 20 sampling dates during the operational study, the unit was operating at over 90 percent capacity on 3 dates, 40 percent capacity on another, 39 percent capacity on another one, 29 percent capacity on another, and not operating on the remaining 14.

The second reason for sampling ichthyoplankton is no longer valid as these results will not be used for entrainment estimates. Reutter and Cooper (1978) demonstrated that night samples at Locust Point produced density estimates 13.1 times greater than day estimates. Consequently, a

night ichthyoplankton sampling program was initiated, the results of which were to be used to estimate entrainment losses at the unit.

#### Fish Egg and Larvae Entrainment

#### Procedures

Fish egg and larvae (ichthyoplankton) entrainment at the Davis-Besse Nuclear Power Station was computed by multiplying the ichthyoplankton concentration observed at Station 8 (intake) by the intake volume. Ichthyoplankton densities were determined at approximately intervals from April - August of 1978 and 1979 from four 3-minute, oblique (bottom to surface) tows at 3 - 4 knots made at night on each date (Tables 40 and 41) with a 0.75 meter diameter heavy-duty oceanographic plankton net (No. 00, 0.75 mm mesh) equipped with a calibrated General Oceanics flowmeter. Oblique tows were selected as this is the technique required at intakes on Lake Erie by U.S. Environmental Protection Agency and U.S. Fish and Wildlife Service. Night sampling is also required by these agencies to minimize net avoidance by larvae and to more accurately assess populations of species which may cling to the bottom during daylight. Samples were preserved in 5% formalin and returned to the laboratory for sorting and analysis. All specimens were identified and enumerated using the works of Fish (1932), Norden (1961a and b), and Nelson and Cole<sub>3</sub>(1975). Densities were presented as number of ichthyoplankters per 100  $m<sup>3</sup>$  of water.

*gm* From the above estimates it was possible to determine an approximate *\* period of occurrence for each species and a mean density during that period. For example, during 1978 walleye were not found on April 30 or on June 7 or later (Table 40). They were present in samples from May 11 and May 21. Therefore, the period of occurrence was estimated to have been from May 6 (the midpoint between April 30 and May 11) to May 30 (the midpoint between May 21 and June 7) (Table 42). The mean ^density of walleye during this period was estimated  $\sharp$ o have been 41.6/100 m $^{\mathsf{S}}$ , computed from the concentration of 79.2/100 nr observed on May 11 and the concentration of  $4.0/100$  m<sup>3</sup> observed on May 21. It was this concentration, 41.6/100 m', which was multiplied by the volume of water drawn through the plant from May 6 to May 30. The same procedure was used in 1979 (Table 43). The daily intake volume was computed by multiplying the daily discharge volume by 1.3. The daily intake volumes were then added for all days within the period of occurrence of the species in question to determine the total intake volume during the period. All specimens were vouchered and all data were keypunched and stored at The Ohio State University's Center for Lake Erie Area Research, Columbus, Ohio.

#### Results

No pre-operational comparisons can be made since entrainment is associated with unit operation. Furthermore, since the operational period began in September 1977 (after the spawning season), no entrainment of fish and eggs occurred until 1978.

Ichthyoplankton densities observed at Station 8 (intake) during <sup>1978</sup> indicated that ichthyoplankters were entrained at the Davis-Besse Nuclear Power Station from May 6 to August 17 (Table 40). May 6 was selected as the first day since it is midway between April 30 and May 11. August 17 was selected as the last day because larvae were present in night samples on August <sup>11</sup> (Table 40) but were absent from day samples at Station 8 on August 23 and later.

During 1978 the mean larvae density from .all night samples at Station 8 (47.5/100 m<sup>3</sup>) was 49 percent greater than the mean density from all day samples collected at Station 8 (31.9/100 m<sup>3</sup>). Gizzard shad constituted 69 percent of the night ichthyoplankton population, followed by walleye at 22 percent, and emerald shiners at 5 percent (Table 40).

Based on the above results (Table 40), it is estimated that 6,311,371 larvae and 44,278 eggs were entrained at the Davis-Besse Nuclear Power Station during 1978 (Table 42). Of this total, gizzard shad constituted 76 percent, walleye 15 percent, and emerald shiners 5 percent.

Ichthyoplankton densities observed at Station 8 (intake) during 1979 indicated that ichthyoplankters were entrained at the Davis-Besse Nuclear Power Station from 26 April to 9 August (Table 41). April 26 was selected as the first day because several walleye were collected on the first sampling date (1 May) and 26 April is half of one sampling interval (10 days) ahead of this first collection. It should also be noted that in 1978 no ichthyoplankters were collected prior to 11 May. August 9 was selected as the last day since it is midway between 3 August, the last sampling date on which larvae were present, and 15 August, a sampling date on which no ichthyoplankters were collected.

During 1929 the mean larvae density from all night samples at Station 8 (142.97/100 m<sup>3</sup>) was 2.9 times greater than the mean density from all day samples collected at Station 8  $(36.7/100 \text{ m}^3)$ . Gizzard shad constituted 50 percent of the night ichthyoplankton population, followed by emerald shiners at 32 percent, yellow perch at 8 percent, freshwater drum at 5 percent, and smelt at 4 percent (Table 41).

Based on the results in Table 41, it is estimated that 20,620,799 larvae and 101,405 eggs were entrained at the Davis-Besse Nuclear Power Station during 1979 (Table 43). Of this total, gizzard shad constituted 49 percent, emerald shiners 33 percent, yellow perch 8 percent, freshwater drum 5 percent, and rainbow smelt 4 percent.

#### Appraisal

Ichthyoplankton entrainment at the Davis-Besse Nuclear Power Station during 1978 and 1979 was typical for an intake on the south shore of the Western Basin of Lake Erie -- it was strongly dominated by gizzard shad. As explained in the ichthyoplankton section of this report, gizzard shad are on the increase and, consequently, it would not be surprising if they *f* represented an even greater portion of the entrainment in future years.<br>Walleye and perch populations appear to be fluctuating. They will Walleye and perch populations appear to be fluctuating. obviously be entrained at this station. However, the number could vary greatly from year to year.

One way to put entrainment losses into perspective is to look at fecundity. Based on an average of 300,000 eggs/female gizzard shad (Hartley and Herdendorf, 1977), the 4,796,964 larvae entrained during 1978 could have been produced by 16 females; based on an average of 331,000 eggs/female walleye (Hartley and Herdendorf, 1977), the 916,738 larvae entrained during 1978 could have been produced by 3 females; and based on 44,000 eggs/female yellow perch (Hartley and Herdendorf, 1977) the 35,259 larvae entrained during 1978 could have been produced by 1 female. In actuality, the above estimates of the number of females required to produce the entrained larvae are quite low since they do not take mortality from eggs to larvae into account. If we assume 99 percent mortality from eggs to larvae to be safe (90 percent is probably more reasonable) then the entrained larvae could have been produced by 1,600 gizzard shad, 300 walleyes, and 100 perch. These values are less than 0.1 percent of the number of perch and walleye captured by Ohio sport fishermen in 1978 (Scholl, 1979). Furthermore, if one looks at the worst case, the value for the upper 95 percent confidence limit and assumes 99 percent mortality from eggs to larvae, the losses of perch and walleye larvae are still less than 0.25 percent of the number lost due to harvesting by Ohio sport fishermen.

Another way to determine the impact of entrainment losses is to estimate the number of adults the entrained larvae might have produced had they lived. This technique requires some knowledge of the mortality between larval stages and between year classes. Patterson (1976) has developed such estimates for yellow perch, and, since it is in the same family, the estimates will also be used here for walleye. Several assumptions are involved.

- I. All entrained larvae are killed.
- II. All larvae lost by entrainment are in their late larval stage. This provides a conservative or high estimate because it does not account for early larval mortality which may range from 83- 96 percent (Patterson, 1976).
- III. Yellow perch become vulnerable to commercial capture, and reach sexual maturity at age class III.
- IV. A one percent survival rate from late larvae to age III adults is assumed. Again, this is conservative since survival rates from:

late larvae to YOY = 4 to 17 percent; YOY to age class  $I = 12$  to 33 percent; age class I to age class II = 38 percent; age class II to age class III = 38 percent (Patterson, 1976, and Brazo,  $et$  al., 1975).

This trend translates to a survivorship ranging from 0.1 percent to one percent over the period from the late larval stage to age class III.

Based on the above assumptions, in 1978 the 916, 738 entrained walleye larvae might have produced 917-9,167 age class III adults and the 35,259 entrained yellow perch larvae might have produced 35-353 age class III adults. In 1979, the 41,648 entrained walleye larvae might have *\* produced 42 - 416 age class III adults and the 1,595,066 entrained yellow perch larvae might have produced 1,595 -15,951 age class III adults.

The author feels little weight should be placed on the above impact assessments since they are based on the number of entrained larvae which can vary greatly from year to year depending on the success of the hatch which in turn is dependent upon the size of the brood stock and weather conditions during spawning and incubation. In the case of Davis-Besse, the off-shore intake where larvae densities are lower and the low volume intake (1978 mean = 21,389 gpm) due to the cooling tower and closed condenser cooling system necessitate a very low-level impact on Western Basin fish populations.

## Fish Impingement

#### Procedures

As was the case with entrainment, impingement is an operational phenomenon and, consequently, pre-operational comparisons are impossible. Furthermore, since estimates are available for only a small portion of 1977 (Reutter, 1978), and since impingement should be viewed for an entire year to allow for seasonal interpretations, only the 1978 and 1979 results will be discussed.

Between January 1 and December 31, 1978 the traveling screens at the Davis-Besse Nuclear Power Station were operated 221 times, while between January 1 and December 31, 1979 the screens were operated 272 times. The date, time, and duration of each screen operation were recorded and keypunched, even when the impinged fish were not collected (Tables 44 and 45). Collections of impinged fish were made by Toledo Edison personnel during 144 of the 221 screen operations during 1978 and on 134 of the 272 screen operations in 1979 by placing a screen having the same mesh size as the traveling screens  $(\frac{1}{4} - \text{inch bar mesh})$  in the sluiceway through which the backwashed material passed. Fish collected in this manner were placed in plastic bags, labeled with the date and time of screen operation, and frozen. The samples were picked up by personnel of The Ohio State University's Center for Lake Erie Area Research (CLEAR) weekly. All specimens, or a representative number thereof, were also weighed and measured.

In addition to the information pertinent to traveling screen operation, the total number and total weight of each species and the length and weight of each individual fish were also keypunched. All these data were stored on magnetic tape at The Ohio State University for use with the<br>Statistical Analysis System: SAS (Barr et al., 1976) on an AMDAHL 370 SAS (Barr et al., 1976) on an AMDAHL 370 computer.

Since the time and duration of *every* screen operation was known, it was possible to determine the number of hours represented by each collection. From this a rate, fish impinged/hour, was developed and used to estimate impingement on days when samples were not collected.

#### Results

A total of 6,607 fish representing 20 species was impinged on the traveling screens at the Davis-Besse Nuclear Power Station from January 1 through December 31, 1978 (Table 46). Goldfish was the dominant species impinged representing 49.9 percent of the total. Only 6 other species represented more than 1 percent of the total: yellow perch, 23.9 percent; emerald shiner, 15.0 percent; gizzard shad, 5.9 percent; black crappie, 1.2 percent; freshwater drum, 1.2 percent; and rainbow smelt, 1.0 percent.

A total of 4,385 fish representing 19 species was impinged on the traveling screens at the Davis-Besse Nuclear Power Station from January 1 through December 31, 1979 (Table 47). Goldfish was the dominant species impinged representing 78.6 percent of the total. Only 4 other species represented more than 1 percent of the total: yellow perch, 6.5 percent; emerald shiner, 4.9 percent; gizzard shad, 3.7 percent; and freshwater drum, 2.6 percent.

Impingement was also computed on a monthly basis (Tables 48 and 49). Most of the impingement during 1978 occurred during April (43.5 percent) and December (35.3 percent). Of the 2,875 fish estimated to have been impinged during April, 834 (29.0 percent) were emerald shiners, 799 (27.8 percent) were goldfish, and 1,098 (38.2 percent) were yellow perch. Of the 2,330 fish estimated to have been impinged during December, 1,870 (80.3 percent) were goldfish and 360 (15.5 percent) were gizzard shad.

Most of the impingement during 1979 occurred during January (55.4 percent) and April (17.2 percent). Of the 2,429 fish estimated to have been impinged during January, 2,218 (91.3 percent) were goldfish, 103 (4.2 percent) were freshwater drum, and 80 (1.8 percent) were gizzard shad. Of the 753 fish estimated to have been impinged in April, 333 (44.2 percent) were goldfish, 200 (26.6 percent) were yellow perch, and 184 (24.4 percent) were emerald shiners.

#### Appraisal

With the exception of the blackside darter and the bluntnose minnow, all species impinged at the Davis-Besse Nuclear Power Station have been captured within the past 17 years at Locust Point (Table 4). However, both the blackside darter and bluntnose minnow have been reported from the island area of Lake Erie and most of the tributaries, including the Toussaint River and Turtle Creek near Locust Point (Trautman, 1957).

With the exception of goldfish, black and brown bullheads, and black and white crappies, the impinged fish occurred in relative numbers which were not unusual for populations in Lake Erie at Locust Point. These five species occurred in relative proportions well above that of the open lake. This indicates probable use of the intake canal as a permanent residence for these species. Furthermore, due to the small sizes of these fish (they were young-of-the-year) and results from previous trawling efforts

(Reutter and Herdendorf, 1975), it appears that these species are also spawning within the intake canal and, consequently, these losses should not be considered as a negative impact on the lake populations of these species.

Impingement losses at the Davis-Besse Nuclear Power Station during 1978 and 1979 were extremely low even when compared to other plants on the Western Basin with lower generating capacities (Reutter et al., 1978). Tables 50 - 52 present sport and commercial fish landings from the Ohio waters of Lake Erie and commercial landings from all of Lake Erie. Table 50 presents only 1978 results because 1979 sport fishing harvest estimates are not available for all species. However, they would probaly have been higher than 1978 because commercial fishing harvests increased by 13 percent from 1978 to 1979, and because the sport harvest of walleye increased from 1,652,000 in 1978 to 3,351,000 in 1979 (Ohio Department of Natural Resources, 1980). Although the fish impinged at Davis-Besse were primarily YOY (mean length, 74 mm and 71 mm in 1978 and 1979) and, consequently, much more abundant than the adults taken by commercial and sport fishermen, the total number impinged (including gizzard shad and goldfish which are not taken by sport fishermen) was only 0.04 percent (1978) and 0.03 percent (1979) of the number harvested by Ohio sport fishermen in 1978. This figure becomes even less significant when one realizes that the Ohio sport catch was only 83.4 percent of the Ohio 1978 commercial catch and only 15.9 percent of the 1978 commercial catch from all of Lake Erie (Tables 50 - 52).

The above comparisons make it obvious that impingement losses at the Davis-Besse Nuclear Power Station have an insignificant effect on Lake Erie fish stocks and further justification of this is unnecessary. However, it should be noted that although by number impingement losses were 0.04 percent (1978) and 0.03 percent (1979) of the Ohio 1978 sport fishing harvest, by weight impingement was less than 0.001 (1978 and 1979) percent<br>of the Ohio sport harvest. Furthermore, based on the estimates of of the Ohio sport harvest. Furthermore, based on the estimates of Patterson (1976) (see Entrainment Section) the impingement of 1,582 young of-the-year yellow perch (1978), aspecies which is *very* important to sport and commercial fishermen, might result in the loss of only <sup>28</sup> - <sup>75</sup> adults which is from 0.0002 to 0.0007 percent of the number captured by Ohio sport fishermen in 1978, while the impingement of 285 young-of-the-year perch in 1979 might result in the loss of 5-16 adults, which is from 0.00004 to 0.0001 percent of the total number of perch captured by Ohio sport fishermen in 1978. It should also be noted that no walleye were impinged.

The obvious conclusion is that impingement losses at the Davis-Besse Nuclear Power Station, Unit 1, have an insignificant effect on Western Basin fish stocks. Furthermore, although the plant did not operate at full capacity during much of these years, the circulating pumps were operated, and consequently, impingement estimates are based on the entire 2-year period and not just dates of generator operation.

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## TABLE 1

## MILESTONES FOR THE

## DAVIS-BESSE NUCLEAR POWER STATION, UNIT 1

 $\mathcal{A}^{\mathcal{A}}$  and  $\mathcal{A}^{\mathcal{A}}$  are  $\mathcal{A}^{\mathcal{A}}$ 



 $\mathcal{L}^{\text{max}}_{\text{max}}$ 

 $\mathcal{L}(\mathcal{L})$  .

 $\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L}}(\mathcal{L}^{\mathcal{L$ 

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## TABLE 2 CALCULATED INTAKE CRIB VELOCITIES FOR UNIT 1 FOR VARIOUS PUMPING RATES

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## TABLE 3

# MONTHLY PUMPING RATES AND<br>CALCULATED VELOCITIES AT THE DAVIS-BESSE<br>NUCLEAR POWER STATION WATER INTAKE CRIB<br>FOR 1978



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## TABLE 4

# SPECIES FOUND IN THE LOCUST POINT AREA 1963 - 1979 $^1$

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## TABLE 4 (CON'T)

# SPECIES FOUND IN THE LOCUST POINT AREA 1963 - 1979 $^{\mathbf{1}}$



1 Includes species collected in Federal Aid Project F-41-R at Locust Point

TABLE 5 PROCEDURES FOR WATER QUALITY DETERMINATION

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## TABLE 6 DISSOLVED OXYGEN DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES



Mean 10.0 2,1 9.1 1.3

 $\mathcal{L}^{\mathcal{L}}$  and the contribution of the contribution of  $\mathcal{L}^{\mathcal{L}}$ 

- 44 -

## TABLE 7 HYDROGEN-IONS (pH) DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES



 $\hat{\mathcal{A}}$ 

 $\mathbf{L}^{(1)}$  and  $\mathbf{L}^{(2)}$  and  $\mathbf{L}^{(3)}$  and  $\mathbf{L}^{(3)}$  and  $\mathbf{L}^{(3)}$  and

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## TABLE 8 TRANSPARENCY DATA FOR WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES

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 $\frac{1}{\sqrt{2}}\int_{0}^{\sqrt{2}}\left(\frac{1}{\sqrt{2}}\right)^{2}e^{-\frac{1}{2}(\sqrt{2}-\frac{1}{2})t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac{1}{2}t}e^{-\frac$ 

 $\mathcal{L}^{\pm}$ 

 $\bar{\beta}$ 



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## TABLE 9

TURBIDITY DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES

	INTAKE (STA. NO. 8)							
Month	Pre-Operational Data (F.T.U.)				Operational Data (F.T.U.)			
	Min	Max <sup>-</sup>	Mean	Std Dev	Min	Max	Mean	Std Dev
March	145.0	145.0	145.0	0.0				
April	$-12.0$	105.0	46.3	42.8	.67.0	67.0	67.0	0.0
May	5.5	21.0	14.9	6.7	46.0	55.0	50.5	6.4
June	10.0	53.0	26.3	18.6	40.0	57.0	48.5	12.0
July	3.0	53.0	16.9	24.2	14.0	53.0	33.0	26.9
August	2.0	23.0	10.5	9.0	13.0	18,0	15.5	3.5
September	5.0	10.0	9.3	4.0	10.0	27.0	18.3	8.5
October	7.0	18.0	11.7	5.7	13.0	32.0	20.7	10.0
November	13.0	36.0	21.7	12.5	8.0	58.0	26.0	27.8
December Mean	16.0	47.0	31.5 33.4	21.9 40.8			34.9	18.5
DISCHARGE (STA. NO. 13)								
March	148.0	148.0	148.0	0.0				
April	18.0	110.0	54.5	42.7	75.0	75.0	75.0	0.0
May	8,5	28.0	17.9	8.0	52.0	75.0	63.5	16.3
June	.7.0	$25.0^{\circ}$	17.5	8.2	49.0	54.0	51.5	3.5
July	4.5	45.0	19.4	18.6	15.0	34.0	24.5	13.4
August	2.0	24.0	12.3	9.5	16.0	17.0	$16.5^{\circ}$	0.7

 $6.0$ 

 $.7.2$ 

11.6

 $23.3$ 

41.9

47.0

42.0

64.0

 $\qquad \qquad \blacksquare$ 

11.0

 $7.0$ 

 $8.0$ 

 $\blacksquare$ 

28.7

 $23.3$ 

28.0

 $\blacksquare$ 

38.9

18.0

 $17.6$ 

 $31.2$ 

 $\blacksquare$ 

21.5

 $10.0$ 

 $13.7$ 

 $19.7$ 

 $37.5$ 

 $35.1$ 

16.0

 $22.0$ 

33.0

54.0

September

October

November December

Mean

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 $.4.0$ 

 $.9.0$ 

13.0

 $21.0$ 

 $\mathcal{A}$  is a set of the set of  $\mathcal{A}$  , and  $\mathcal{A}$ 

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## TABLE 10 SUSPENDED SOLIDS DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES



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 $\mathcal{L}(\mathcal{A})$  is a set of  $\mathcal{L}(\mathcal{A})$  . In the  $\mathcal{L}(\mathcal{A})$ 

## TABLE 11 CONDUCTIVITY DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES



 $\mathcal{L}=\overline{\mathcal{L}}$  . The set of  $\mathcal{L}=\{1,2,3,4,5\}$ 



 $\frac{1}{2}$ 



 $\bar{\mathcal{A}}$ 

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 $\mathcal{L}(\mathcal{L}(\mathcal{L}))$  and  $\mathcal{L}(\mathcal{L})$  and  $\mathcal{L}(\mathcal{L})$ 

 $\label{eq:2.1} \mathcal{L}_{\text{max}}(\mathcal{L}_{\text{max}}) = \mathcal{L}_{\text{max}}(\mathcal{L}_{\text{max}})$ 

 $\mathcal{L}^{\text{max}}_{\text{max}}$  and  $\mathcal{L}^{\text{max}}_{\text{max}}$ 





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 $\mathcal{L}(\mathcal{L}(\mathcal{L},\mathcal{$ 

 $\sim 100$  km s  $^{-1}$ 

 $\bigcap$ 

 $\mathbb{R}^2$ 

 $\label{eq:2} \frac{1}{\sqrt{2}}\sum_{i=1}^n\frac{1}{\sqrt{2}}\left(\frac{1}{\sqrt{2}}\sum_{i=1}^n\frac{1}{\sqrt{2}}\sum_{i=1}^n\frac{1}{\sqrt{2}}\right) \left(\frac{1}{\sqrt{2}}\sum_{i=1}^n\frac{1}{\sqrt{2}}\sum_{i=1}^n\frac{1}{\sqrt{2}}\right)$ 



 $\mathcal{L}_{\text{max}}$  and  $\mathcal{L}_{\text{max}}$  and  $\mathcal{L}_{\text{max}}$ 





 $\mathcal{L}^{\bullet}$ 

 $\mathbf{v}$ 

 $\hat{\mathcal{A}}$ 

 $\overline{\phantom{a}}$ 

 $\sim$ 

 $I<sub>0</sub>$  $\overline{\mathbf{a}}$  $-$ 

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 $\ddot{\phantom{a}}$ 

## TABLE 14

CHLORIDE DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES

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 $\ddot{\cdot}$ 

## TABLE 16 SODIUM DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES





- 54 -

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## TABLE 17 MAGNESIUM DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES

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## TABLE 18 TOTAL ALKALINITY DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES

 $\mathbb{E}^{\mathbb{Z}}$
# NITRATE DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES

 $\mathcal{L}_{\mathcal{A}}$  $\sim$ 





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## TABLE 20 PHOSPHORUS DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES







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 $\alpha$ 

# SILICA DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES



0.59

0.19

 $.0.16$ 

 $0.26$ 

 $\mathcal{L}^{\pm}$ 

0.06

 $[0, 03]$ 

 $0.16$ 

 $0.09$ 

 $0, 28$ 

 $0.13$ 

 $0.10$ 

 $0,21$ 

 $0.35$ 

 $0,28$ 

 $0.06$ 

 $0.07$ 

 $0.07$ 

 $0:28$ 

 $0.07$ 

 $0.07$ 

 $0.11$ 

 $0.36$ 

 $0.10$ 

 $0.64$ 

 $\blacksquare$ 

 $0.22$ 

 $0.09$ 

 $0.35$ 

 $0.47$ 

 $0.15$ 

 $0.02$ 

 $0.27$ 

 $\bullet$ 

 $0.40$ 

 $\overline{a}$  $\mathcal{L}^{\mathcal{L}}$  .  $\mathbb{R}^2$ 

September

October

November

December

Mean

 $\bullet$ 

# TABLE 22 BIOCHEMICAL OXYGEN DEMAND DATA FOR BOTTOM WATER IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES



November  $\begin{vmatrix} 2.00 \\ 2.00 \end{vmatrix}$  3.00  $\begin{vmatrix} 2.30 \\ 2.30 \end{vmatrix}$  0.60  $\begin{vmatrix} 1.0 \\ 1.0 \end{vmatrix}$  4.0  $\begin{vmatrix} 2.3 \\ 2.3 \end{vmatrix}$  1.5

Mean 2.57 0:56 .; 3.13 0.7

Decembe**r || 1.**00 **| 2.00 | 1.50 | 0.71 || -** || - || -

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 $\ddot{\phantom{a}}$ 

•"

### TABLE 23 TEMPERATURE DATA FOR BOTTOM WATER  $\sim$   $\sim$ IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES



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 $\bullet$ 

OPERATIONAL WATER QUALITY PARAMETERS FALLING OUTSIDE OF THE RANGE OF PRE-OPERATIONAL VALUES AT STATION 13



cn ro

### MEAN WATER QUALITY VALUES FOR PRE-OPERATIONAL AND OPERATIONAL PERIODS IN THE VICINITY OF LAKE INTAKE AND DISCHARGE STRUCTURES

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 $\mathcal{L}^{\text{max}}_{\text{max}}$ 



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 $\label{eq:2.1} \mathcal{L}(\mathcal{L}_{\mathbf{A}}) = \mathcal{L}(\mathcal{L}_{\mathbf{A}}) = \mathcal{L}(\mathcal{L}_{\mathbf{A}})$ 

 $\mathcal{L}^{\text{max}}_{\text{max}}$  , where  $\mathcal{L}^{\text{max}}_{\text{max}}$ 

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 $\epsilon_{\rm in}$ 

## TABLE 26



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PLANKTON AND WATER QUALITY SAMPLING DATES

 $<sup>1</sup>$  No phytoplankton collections</sup>

## PHYTOPLANKTON AND ZOOPLANKTON SAMPLING STRUCTURE, 1973-19791



All samples were collected by a vertical tow with a Wisconsin plankton net; 12cm mouth 0.064 mm mesh in 1973 and 1974 and 0.080 mm mesh from 1975-1979.

 $\boldsymbol{2}$ No phytoplankton sampling; Zooplankton only.  $\bullet$  .

# PRE-OPERATIONAL AND OPERATIONAL PHYTOPLANKTON DATA FROM LAKE ERIE IN<br>THE VICINITY OF THE DAVIS-BESSE NUCLEAR POWER STATION



<sup>1</sup>Results from samples collected from 1974 through August 1977. <sup>2</sup>Results from samples collected from September 1977 through 1979. 3April sample actually collected May 1.

### TABLE 28 (cont'd)

### PRE-OPERATIONAL AND OPERATIONAL PHYTOPLANKTON DATA FROM LAKE ERIE IN THE VICINITY OF THE DAVIS-BESSE NUCLEAR POWER STATION



Results from samples collected from 1974 through August 1977.

•Results from samples collected from September 1977 through 1979.

April sample actually collected May 1.

# PRE-OPERATIONAL AND OPERATIONAL PHYTOPLANKTON DATA<sup>1</sup> FROM THE VICINITY OF THE INTAKE AND DISCHARGE STRUCTURES AND A CONTROL STATION



<sup>1</sup>Data presented as number of whole organisms per liter.

<sup>2</sup>Data collected from 1974 through August 1977.

3<br>Data collected from September 1977 through 1979.

<sup>4</sup>April sample actually collected May 1, 1979.

### PRE-OPERATIONAL AND OPERATIONAL ZOOPLANKTON DATA FROM THE LOCUST POINT AREA



1Results from samples collected from 1973 through August 1977.

2<br>Results from samples collected from September 1977 through 1979.

3April sample actually collected May 1.

# TABLE 30 (cont'd)

# PRE-OPERATIONAL AND OPERATIONAL ZOOPLANKTON DATA FROM THE LOCUST POINT AREA



### **IUIAL**



<sup>1</sup>Results from samples collected from 1973 through August 1977.

<sup>2</sup>Results from samples collected from September 1977 through 1979.

3April sample actually collected May 1.

# PRE-OPERATIONAL AND OPERATIONAL ZOOPLANKTON DATA IN THE VICINITY OF THE INTAKE<br>AND DISCHARGE STRUCTURES AND A CONTROL STATION



<sup>1</sup>Data collected from 1973 through August 1977.

<sup>2</sup>Data collected from September 1977 through 1979.

3April sample actually collected May 1.

C



### BENTHIC MACROINVERTEBRATE , SAMPLING STRUCTURE, 1973-19791

**WARDS** 



 $^{\text{\text{1}}}$  Three replicate grab samples with a ponar dredge (A=0.052 m $^{\text{\text{2}}}$ ) were collected at the stations indicated each year except 1973 when only one grab was collected at each station.

Samples were collected only in April as water at this station was removed after this date to allow construction on the intake pumps.



# PRE-OPERATIONAL AND OPERATIONAL BENTHIC MACROINVERTEBRATE DENSITIES<sup>1</sup> FROM LAKE ERIE IN THE VICINITY OF THE DAVIS-BESSE NUCLEAR POWER STATION



Data presented as number of organisms per square meter.

2<br>Data collected from 1973 through August 1977.

3Data collected from September 1977 through 1979.

### TABLE 34 (cont'd)

# PRE-OPERATIONAL AND OPERATIONAL BENTHIC MACROINVERTEBRATE DENSITIES<sup>1</sup> FROM LAKE ERIE IN THE VICINITY OF THE DAVIS-BESSE NUCLEAR POWER STATION



1Data presented as number of organisms per square meter.

<sup>2</sup>Data collected from 1973 through August 1977.

3Data collected from September 1977 through 1979.

PRE-OPERATIONAL AND OPERATIONAL BENTHIC MACROINVERTEBRATE DATA<sup>1</sup> FROM<br>THE VICINITY OF THE INTAKE AND DISCHARGE STRUCTURES AND A CONTROL STATION



 $1$ Data presented as number of organisms per square meter.

<sup>2</sup>Data collected from 1973 through August 1977.

3Data collected from September 1977 through 1979.

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GILL NET SAMPLING DATES

 $\hat{\mathcal{L}}$ 

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 $\label{eq:2.1} \frac{1}{\sqrt{2}}\int_{0}^{\infty}\frac{1}{\sqrt{2\pi}}\left(\frac{1}{\sqrt{2}}\right)^{2}d\mu_{\rm{eff}}\,.$ 

 $\hat{\mathcal{A}}$ 

 $\frac{1}{\sqrt{2}}$ 

 $\label{eq:1} \frac{1}{\sqrt{2}}\left(\frac{1}{\sqrt{2}}\right)^{2} \left(\frac{1}{\sqrt{2}}\right)^{2}$ 

### - 78 - TABLE 37

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> PRE-OPERATIONAL AND OPERATIONAL GILL NET CATCHES<sup>-</sup> OF SELECTED SPECIES FROM LAKE ERIE IN THE VICINITY OF THE DAVIS-BESSE NUCLEAR POWER STATION DISCHARGE (STATION 13)



a 24-hour bottom set with an experimental gill net 125 ft long consisting of five 25-ft contiguous panels of  $\frac{1}{2}$ ,  $3/4$ ,  $1$ ,  $1\frac{1}{2}$ , and 2-inch bow mesh.

,<br>Results from samples collected from 1973 through August 1977.

### TABLE 37 (cont'd)

# PRE-OPERATIONAL AND OPERATIONAL GILL NET CATCHES<sup>1</sup> OF SELECTED SPECIES FROM LAKE ERIE IN THE VICINITY OF THE DAVIS-BESSE NUCLEAR POWER STATION DISCHARGE (STATION 13)



Results presented as the number of fish per unit effort, where one unit of effort equals<br>a 24-hour bottom set with an experimental gill net 125 ft long consisting of five 25-ft<br>contiguous panels of  $\frac{1}{2}$ , 3/4, 1, 1½,

<sup>2</sup>Results from samples collected from 1973 through August 1977.

### TABLE 37 (cont'd)

### PRE-OPERATIONAL AND OPERATIONAL GILL NET CATCHES' OF SELECTED SPECIES FROM LAKE ERIE IN THE VICINITY OF THE DAVIS-BESSE NUCLEAR POWER STATION DISCHARGE (STATION 13)



Results presented as the number of fish per unit effort, where one unit of effort equals a 24-hour bottom set with an experimental gill net 125 ft long consisting of five 25-ft contiguous panels of *h9* 3/4, 1, 1%, and 2-inch bow mesh.

,<br>Results from samples collected from 1973 through August 1977.

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### PRE-OPERATIONAL AND OPERATIONAL GILL NET DATA<sup>1</sup> FROM THE VICINITY OF THE DAVIS-BESSE NUCLEAR POWER STATION INTAKE, DISCHARGE, AND TWO CONTROL STATIONS



### STATION 8



<sup>1</sup>Results presented as number of fish per unit effort, where one unit of effort equals a 24-hour bottom set with an experimental gill net 125 ft long consisting of five 25-ft contiguous panels of *h\** 3/4, 1, l\*s, and 2-inch bar mesh.

 $2$ Results from samples collected from 1973 through August 1977.  $\cdot$ 

### TABLE 38 (cont'd.)

### PRE-OPERATIONAL AND OPERATIONAL GILL NET DATA<sup>-</sup> FROM THE VICINITY OF THE DAVIS-BESSE NUCLEAR POWER STATION INTAKE, DISCHARGE, AND TWO CONTROL STATIONS

Month	STATION 13							
	Pre-Operational Data $2^7$				Operational Data <sup>3</sup>			
	Min	Max	Mean	Std Dev	Min	Max	Mean	Std Dev
April	88	269	166	75			88	
May	120	1381	573	558	29	270	150	170
June	49	232	125	77	112	122	117	
July	94	254	163	82	85	138	112	37
August	136	$327 -$	237	84	186	387	287	142
September	73	382	270	141	122	366	206	138
October	104	691	337	312		433	178	225
November	6	208	76	94	85	1455	544	789
December		$- -$	14	---			---	
Mean	84	468	218	166	89	453	210	150

STATION 26



Results presented as number of fish per unit effort, where one unit of effort equals a 24-hour bottom set with an experimental gill net 125 ft long consisting of five 25-ft contiguous panels of  $\frac{1}{2}$ , 3/4, 1, 1 $\frac{1}{2}$ , and 2-inch bar mesh.

,<br>Results from samples collected from 1973 through August 1977.

**/0ms.**

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# TABLE 39

## ICHTHYOPLANKTON SAMPLING DATES



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- 84 -

### ICHTHYOPLANKTON DENSITIES IN THE VICINITY OF THE INTAKE OF THE DAVIS - BESSE NUCLEAR POWER STATION - 1978\*



Data presented as number of individuals per 100m and computed from 4 oblique tows (bottom to surface) collected at night.

 $- 85 -$ TABLE 41 ICHTHYOPLANKTON DENSITIES IN THE VICINITY OF THE INTAKE OF THE DAVIS-BESSE NUCLEAR POWER STATION - 1979\*



 $\mathbf i$ 

\*Data presented as number of individuals per 100m3 and computed from<br>4 oblique tows (bottom and surface) collected at night.

\*\*This is the subtotal of the larval stages. It is the mean of the surface<br>and bottom densities. Stage 1 = proto-larvae, no rays in fin/finfold.<br>Stage 2 = meso-larvae, first ray seen in median fins. Stage 3 = meta-<br>larvae,

### ICHTHYOPLANKTON ENTRAINMENT AT THE DAVIS-BESSE NUCLEAR POWER STATION - 1978



a Estimated from Table 1. See discussion on page 1.

b Estimated by multiplying daily discharge rate by 1.3 and adding all daily estimates for the specified period.

c Average concentration during their period of occurrence.

d Values which would have been less than zero were rounded back to zero.

### ICHTHYOPLANKTON ENTRAINMENT AT THE DAVIS-BESSE NUCLEAR POWER STATION - 1979



Estimated from Table 1. See discussion oh page 1.

Estimated by multiplying dally discharge rate by 1.3 and adding all dally estimates for the specified period.

cAverage concentration during their period of occurrence.

Values which would have been less than zero were rounded back to zero.

 $\ddot{\phantom{a}}$ 

### TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION FROM 1 JANUARY TO 31 DECEMBER 1978



- 88 -

## TABLE 44 (Con't.)

### TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION FROM 1 JANUARY TO 31 DECEMBER 1978

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- 89 -

### TABLE 44(Con't.)

### TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION FROM 1 JANUARY TO 31 DECEMBER 1978

 $\subset$ 



# TABLE 44 (Con't.)

### TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION FROM 1 JANUARY TO 31 DECEMBER 1978

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 $\sum_{i=1}^n$ 



# TABLE 44(Con't.)

# TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION<br>FROM 1 JANUARY TO 31 DECEMBER 1978



 $\subset$
### TABLE 45

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# TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION<br>FROM 1 JANUARY TO 31 DECEMBER 1979



# TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION<br>FROM 1 JANUARY TO 31 DECEMBER 1979

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## TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION FROM 1 JANUARY TO 31 DECEMBER 1979



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#### TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION FROM 1 JANUARY TO 31 DECEMBER 1979

#### TIME OF SCREEN OPERATION FISH HOURS SINCE DATE ON OFF COLLECTION **OPERATION** 20 August 20.30 21.30 N 27.85<br>21 August 21.00 17.00 18.00 Y 20.70 21 August 17.00 1 18.00 Y 20.70 22 August 17.50 18.50 N 24.50 23 August 17.45 18.45 " Y 23.95 24 August 20.55 22.00 N 27.55 25 August 17.00 18.00 Y 20.00 27 August 16.20 17.20 Y 47.20 28 August 18.50 19.50 N 26.30 29 August 16.45 17.45 Y 21.95 30 August 22.05 23.05 N 29.60 1 September 16.45 17.15 N 142.10<br>2 September 16.50 17.20 17.20 18 2 September 16.50 17.20 Y 24.05 September 16.45 17.15 N 23.95<br>September 16.50 17.20 Y 24.05 4 September 16.50 17.20 Y 24.05 5 September 16.50 17.20 N 24.00 6 September 16.45 17.15 Y 23.95 September 17.00 17.40 N 24.25<br>September 18.12 19.18 Y 25.78 8 September 18.12 19.18 19 Y 25.78<br>9 September 18.30 19.45 N 24.27 9 September 18.30 19.45 N 24.27 10 September 17.30 18.45 N 23.00<br>11 September 17.40 18.40 N 23.95 11 September 17.40 18.40 N 23.95 12 September 19.25 20.33 Y 25.93 13 September 16.40 | 18.15 | N 21.82<br>14 September 16.38 | 17.40 | Y 23.25 14 September 16.38 17.40 Y 23.25 15 September<br>16 September - 20.00 21.00 N 27.60<br>16 September - 16.31 17.02 N 20.02 16 September 16.31 17.02 N 20.02 17 September 16.35 17.05 N 24.03 18 September 19.02 19.35 Y 26.30 20 September 18.40 19.10 Y 47.75<br>21 September 16.25 16.55 N 21.45 21 September 16.25 | 16.55 | N | 21.45<br>22 September 16.35 | 17.05 | Y | 24.50 22 September 16.35 17.05 Y 24.50 23 September 16.15 16.50 N 23.45 24 September 16.54 17.27 Y 24.77 25 September 16.20 16.57 N<br>26 September 17.00 17.35 N 26 September 17.00 17.35 Y 24.78 28 September 16.40 17.10 N 23.75 29 September 16.11 16.44 Y 23.34 31 September 17.06 18.09 N 49.65<br>1 October 19.06 1.07 N 26.98 1 October 20.06 21.07 N 26.98 2 October 20.00 21.02 Y 23.95 4 October 17.14 18.25 Y 45.23 6 October 20.50 21.20 Y 50.95

#### TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION FROM 1 JANUARY TO 31 DECEMBER 1979

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#### TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION FROM 1 JANUARY TO 31 DECEMBER 1979

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## TRAVELING SCREEN OPERATION AT THE DAVIS-BESSE NUCLEAR POWER STATION<br>FROM 1 JANUARY TO 31 DECEMBER 1979



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TABLE 46

FISH SPECIES IMPINGED AT THE DAVIS-BESSE NUCLEAR POWER STATION: 1 January through 31 December 1978



\* Confidence Intervals could not be computed when no more than one representative of a given species occurred.

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FISH SPECIES IMPINGED AT THE DAVIS-BESSE NUCLEAR POWER STATION: 1 January through 31 December 1979



\* Confidence intervals could not be computed when no more than one

representative of a given species occurred.

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I *t—\** o I—» I TABLE 48

#### A SUMMARY OF MONTHLY FISH IMPINGEMENT AT THE DAVIS-BESSE NUCLEAR POWER STATIONS: 1 January through 31 December 1978

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TABLE 49

#### A SUMMARY OF MONTHLY FISH IMPINGEMENT AT THE DAVIS-BESSE NUCLEAR POWER STATIONS: 1 January through <sup>31</sup> December 1979

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TABLE 50 ESTIMATED 1978 SPORT AND COMMERCIAL FISH HARVEST FROM THE OHIO WATERS OF LAKE ERIE<sup>a</sup>



*a Scholl (1979).*

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*Estimated based on mean weight of sport fish.*

*c Data not available.*

*Thirty-eight percent carp.*

*e Excludes weight of "Others" caught by sport fishermen.*

*Closed to commercial fishing.*



COMMERCIAL FISH LANDINGS FROM THE OHIO WATERS OF LAKE ERIE: 1974-1979\*

\* Ohio Dept. of Natural Resources (1980). Data presented in kilograms.

\*\* Data not available.

## TABLE 52



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# COMMERCIAL FISH LANDINGS FROM<br>LAKE ERIE: 1975 - 1979ª

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## TABLE 52(Cont'd)

#### COMMERCIAL FISH LANDINGS FROM LAKE ERIE: 1975 - 1979a



 $a$  Muth (1980).

 $b$  Not taken commercially in Ohio and Michigan waters.

Included with "Others" during this year.

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*Figure 6. Water Temperature Record for Intake and Discharge for the Davis-Besse Nuclear Power Station, Unit 1 (1979).*

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FIGURE 11. STATION LOCATION MAP

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 $*18'-2''$  on the other side of the crib

### FIGURE 13. DETAILS OF WATER INTAKE CRIB





FIGURE 15. REEFS NEAR LOCUST POINT.



SEDIMENT DISTRIBUTION MAP OF WESTERN LAKE ERIE<br>IN THE VICINITY OF LOCUST POINT FIGURE 16.



FIGURE 17. BIOLOGICAL SAMPLING STATIONS AT THE DAVIS-BESSE NUCLEAR POWER STATION.

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### FIGURE 18. TRENDS IN MEAN MONTHLY TEMPERATURE, DISSOLVED OXYGEN, AND HYDROGEN ION<br>MEASUREMENTS FOR LAKE ERIE AT LOCUST POINT FOR THE PERIOD 1972-1979.  $\ddot{\phantom{a}}$





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## FIGURE 26. Comparison of Pre-operational and Operational Data for Conductivity of Bottom Water at Station Discharge (Station No. 13).





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*in Bottom Water at Station Discharge (Station No. 13).*

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FIGURE 35. Comparison of Pre-operational and Operational Data for Phosphorus in Bottom Water at Station Discharge (Station No. 13).



FIGURE 36. Comparison of Pre-operational and Operational Data for Silica

in Bottom Water at Station Discharge (Station No. 13).



*FIGURE 37. Comparison of Pre-operational and Operational Data of Biochemical Oxygen Demand of Bottom Water at Station Discharge (Station No. 13).*





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## MONTHLY MEAN BACILLARIOPHYCEAE, CHLOROPHYCEAE, AND<br>MYXOPHYCEAE POPULATIONS FOR LAKE ERIE AT LOCUST POINT, 1977.





MONTHLY MEAN BACILLARIOPHYCEAE, CHLOROPHYCEAE, AND<br>MYXOPHYCEAE POPULAȚIONS FOR LAKE ERIE AT LOCUST POINT, 1978.

## FIGURE 44



## FIGURE 45 MONTHLY MEAN BACILLARIOPHYCEAE, CHLOROPHYCEAE, AND MYXOPHYCEAE<br>POPULATIONS FOR LAKE ERIE AT LOCUST POINT, 1979.



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\*Dotted lines connect points (sampling dates) separated<br>by more than a full calendar month. Solid lines connect<br>points (dates) in consecutive months.



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Figure 48. Comparison of Pre-operational and Operational Data for Green Algae Densities in Lake Erie in the Vicinity of the Davis-Besse Nuclear *Power Station.*



Figure 49. Comparison of Pre-operational and Operational Data for Blue-green Algae Densities in Lake Erie in the Vicinity of the Davis-Besse Nuclear Power Station.



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Figure 53. Comparison of Pre-operational and Operational Data for Phytoplankton

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mean value for operational data, (September 1977 - December 1979).

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\*Dotted lines connect points (sampling dates) separated by more than a full calendar<br>month. Solid lines connect points (dates) in consecutive months.

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\*Dotted lines connect points (sampling dates) separated by more than a full calendar<br>month. Solid lines connect points (dates) in consecutive months.

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Figure 59. Comparison of Pre-operational and Operational Data for Zooplankton Rotifer Densities in Lake Erie in the Vicinity of the Davis-Besse Nuclear Power Station.





Figure 60.



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Figure 64. Comparison of Pre-operational and Operational Data for Zooplankton Densities at a Control Station (Sta. No. 3).



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*FigureJ57. Comparison of Pre-operational and Operational Data for Benthic Ceolenterate Densities in Lake Erie in the Vicinity of the Davis-Besse Nuclear Power Station.*







No. Organisms/meter<sup>2</sup>





*Figure 70. Comparison of Pre-operational and Operational Data for Benthic Mollusc Densities in Lake Erie in the Vicinity of the Davis-Besse Nuclear Power Station.*



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Figure 71. Comparison of Pre-operational and Operational Data for Benthic Macroinvertebrate Densities at the Station Intake (Sta. No. 8).



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*Figure 75. Comparison of Pre-operational and Operational Channel Catfish Catches in Gill Nets Set in the Vicinity of the Davis-Besse Nuclear Power Station Discharge (Station 13).*

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