

Satellite-measured water properties in high altitude Lake Tahoe

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Abstract: It has been difficult in satellite remote sensing to derive accurate water optical, biological, and biogeochemical products over high-altitude inland waters due to issues in satellite data processing (i.e., atmospheric correction). In this study, we demonstrate that accurate normalized water-leaving radiance spectra $nL_w(\lambda)$ can be derived for a high-altitude lake, Lake Tahoe, using improved Rayleigh radiance computations (Wang, M., *Opt. Express*, **24**, 12414–12429, 2016) which accurately account for water surface altitude effects in the Multi-Sensor Level-1 to Level-2 (MSL12) ocean color data processing system. Satellite observations from the Visible Infrared Imaging Radiometer Suite (VIIRS) onboard the Suomi National Polar-orbiting Partnership (SNPP) between 2012 and 2018 are used to evaluate and validate satellite-derived $nL_w(\lambda)$ spectra, and to quantitatively characterize water properties in the lake. Results show that VIIRS-derived $nL_w(\lambda)$ spectra are quite comparable with those from the in situ measurements. Subsequent retrievals of water biological and biogeochemical products show that chlorophyll-a (Chl-a) concentration and Secchi depth (SD) are reasonably well-estimated, and captured normal seasonal variations in the lake, e.g., the annual highest

27 Chl-a and SD normally occur in the winter while the lowest occur in the summer, which is
28 consistent with in situ measurements. Interannual variability of these water quality parameters
29 is also observed. In particular, Lake Tahoe experienced a significant environmental anomaly
30 associated with an extreme weather condition event in 2017, showing considerably decreased
31 $nL_w(\lambda)$ at the spectral bands of 410, 443, and 486 nm, and significantly reduced SD values in the
32 entire lake. The low SD measurements from VIIRS are consistent with in situ observations.
33 Following the event in the 2017–2018 winter, Lake Tahoe recovered and returned to its typical
34 conditions in spring 2018.

35 **Keywords:** High-altitude lakes; Lake Tahoe; VIIRS; satellite ocean color; normalized water-
36 leaving radiance $nL_w(\lambda)$; water optical properties.

37 1. Introduction

38 Located at an altitude of 1897 m between California and Nevada in the U.S., Lake Tahoe
39 (Fig. 1) is a large fresh water lake and is renowned for its water clarity. It is the second deepest
40 lake (after Crater Lake in southern Oregon) in the U.S. with a length of 35 km, width of 19 km,
41 average depth of 333 m, surface area of about 490 km², and total volume of approximately 156
42 km³. There are 63 tributaries that provide half of the water supply to Lake Tahoe, which has a
43 drainage area of about the same size as the lake. The evaporation in the region accounts for about
44 two thirds of the total water mass leaving the lake, and the other one third is through the lake's
45 only outlet, the Truckee River.

46 In the Lake Tahoe region, the climate is characterized by warm dry summers and chilly
47 winters with an average daily maximum temperature of ~25.5°C in July and an average daily
48 minimum temperature of ~4.6°C in December. The lake serves as an important resource for the
49 regional economy by attracting tourists from all over the world throughout the year with its
50 winter sports, summer recreations, and other activities and events. Thus, the conservation of lake
51 water quality and the surrounding environment has drawn major political, scientific, and public
52 interest in the region.

53 Lake Tahoe has a long history of water clarity monitoring using Secchi depth (SD)
54 measurements back to the 1960s (Goldman, 2000; Jassby et al., 1999). In the last five decades,
55 according to the 2018 Tahoe Environmental Research Center report (TERC, 2018), the water
56 clarity in terms of the SD declined consistently from ~30 m in the mid 1960s to ~20 m in recent
57 years (TERC, 2018). The SD decline can be attributed to increased phytoplankton concentration
58 and the amount of fine sediment particles in the lake (Jassby et al., 1999; Jassby et al., 2003).
59 Seasonally, one of the SD minimums occurs in June due to increased suspended sediment
60 discharges with the melting of snowpack, and another SD minimum occurs in December
61 primarily as a result of mixed-layer deepening (Jassby et al., 1999). The long-term SD model
62 also shows that the interannual variability of SD in the lake can be driven by weather changes
63 such as precipitation anomalies in the region (Jassby et al., 2003).

64 In addition, Lake Tahoe shows a long-term warming trend due to the upward trend of air
65 temperature and incoming longwave radiation in the region (Coats et al., 2006). Indeed, the
66 volume-weighted mean temperature increased about 0.015 °C/year between 1970 and 2002. The
67 thermal structure of the lake also shows changes, e.g., the decreasing depth of the thermocline
68 and the increase of resistance of the water in the lake to vertical layer mixing and stratification
69 (Coats et al., 2006; Sahoo et al., 2016). In Lake Tahoe, both upwelling and surface circulation
70 are observed and characterized with satellite remote sensing from the
71 Thermal Emission and Reflection Radiometer (ASTER) (Hook et al., 2007; Tonooka and
72 Palluconi, 2005; Tonooka et al., 2005), the Landsat Enhanced Thematic Mapper (ETM) (Barsi et
73 al., 2007; Hook et al., 2004), and the Moderate Resolution Imaging Spectroradiometer (MODIS)
74 (Hook et al., 2007). The upwelling of the intermediate-depth water was observed to occur
75 frequently in the spring and summer seasons. It brings the water from the depth of ~10 to ~30 m
76 to the surface, and leads to enhanced phytoplankton growth and decreased water clarity in Lake
77 Tahoe (Steissberg et al., 2005).

78 As one of the main drivers of the water clarity trend in Lake Tahoe, the rate of
79 phytoplankton growth has been steadily increasing over the past 50 years. Indeed, primary

80 productivity increased from $< \sim 50 \text{ g C m}^{-2} \text{ year}^{-1}$ in 1959 to $> \sim 200 \text{ g C m}^{-2} \text{ year}^{-1}$ in recent years
81 (TERC, 2018). In addition, the aerosol deposition provides most of the nutrients in the dissolved
82 inorganic particles and total nitrogen, as well as a significant amount of the total phosphorus
83 loading in Lake Tahoe (Jassby et al., 1994). In fact, the annual nitrogen (N), phosphorus (P),
84 and particular matter (PM) from the aerosol deposition in the lake are estimated to be about 185,
85 3, and 755 metric tons (Dolislager et al., 2012), respectively. Indeed, these nutrients from the
86 aerosol deposition contributed to the long-term increase of Chl-a concentration, water primary
87 productivity, and biomass in Lake Tahoe (Mackey et al., 2013). Specifically, it has been found
88 that the phytoplankton amount in Lake Tahoe is becoming phosphorus-limited from the nitrogen-
89 limited because the deposited aerosols in the lake contain nutrients with high N:P ratios (Chang
90 et al., 1992; Mackey et al., 2013). However, it is noted that, in addition to the phosphorus
91 limitation, phytoplankton growth in Lake Tahoe is also iron-limited (Chang et al., 1992).

92 In situ water optics measurements have long been conducted to characterize and quantify
93 optical and bio-optical properties in Lake Tahoe. The spectral irradiance and beam transmittance
94 were measured as a function of water depth in the lake (Smith et al., 1973). A blue color index
95 was also developed using the remote sensing reflectance $R_{rs}(\lambda)$ measurements to quantitatively
96 analyze the spatial and seasonal variations (Watanabe et al., 2016). It is noted that the remote
97 sensing reflectance $R_{rs}(\lambda)$ (as a function of the wavelength λ) is defined as $R_{rs}(\lambda) = nL_w(\lambda)/F_0(\lambda)$,
98 where $nL_w(\lambda)$ and $F_0(\lambda)$ are the normalized water-leaving radiance (Gordon, 2005; Morel and
99 Gentili, 1996; Wang, 2006) and the mean extraterrestrial solar irradiance (Thuillier et al., 2003),
100 respectively. As an oligotrophic subalpine lake, the role of ultraviolet (UV) radiation, and the
101 patterns of spatial and temporal variability of UV transparency were investigated (Rose et al.,
102 2009). It was found that UV transparency differs from the photosynthetically available radiation
103 (PAR). In fact, the combination of the UV and visible water transparency can provide a more
104 comprehensive understanding of the ecosystem changes and the biological and biogeochemical
105 processes in Lake Tahoe (Rose et al., 2009).

106 Satellite observations over global inland waters can provide an effective tool to monitor the
107 lake environmental changes such as algae bloom and water clarity. Using the shortwave infrared
108 (SWIR)-based atmospheric correction algorithm (Wang, 2007; Wang and Shi, 2005, 2007; Wang
109 et al., 2009b), it has been demonstrated that MODIS-derived water optical property data, e.g.,
110 $nL_w(\lambda)$ spectra, can be used to monitor and assess water property (quality) in turbid Lake Taihu
111 in China (Wang et al., 2011). The spatial and temporal water turbidity variations in Lake
112 Okeechobee are also characterized using MODIS measurements (Wang et al., 2012a).
113 Furthermore, the cyanobacteria blooms in Lake Taihu are quantitatively assessed and evaluated
114 from MODIS observations (Hu et al., 2010). In addition to water optical parameters such as
115 $nL_w(\lambda)$ spectra, lake surface temperature and many other water quality parameters, such as the
116 water diffuse attenuation coefficient at 490 nm $K_d(490)$ (Lee et al., 2005; Wang et al., 2009a),
117 Chl-a concentrations (Hu et al., 2012; O'Reilly et al., 1998; O'Reilly and Werdell, 2019; Wang
118 and Son, 2016), total suspended matter (TSM) (or suspended particulate matter (SPM))
119 concentrations (Knaeps et al., 2015; Nechad et al., 2010; Shi et al., 2018; Yu et al., 2019), SD
120 (Binding et al., 2015; Lee et al., 2016), float algae index (Hu, 2009), and inherent optical
121 properties (IOPs) (Lee et al., 2002; Shi et al., 2019; Werdell et al., 2013), can also be routinely
122 and reliably derived from satellite remote sensing observations. These satellite water quality
123 property data can be further used to study global lakes to characterize and quantify the long-term
124 water physical, optical, biological, and biogeochemical variability (Bolgrien and Brooks, 1992;
125 Shi and Wang, 2015; Shi et al., 2018; Son and Wang, 2019). In all of these studies, however,
126 satellite reflective solar radiance data (water color) are applied only to global sea-level inland
127 waters.

128 Lake Tahoe has long been used as one of the sites for validation of the absolute radiometric
129 calibration and surface geophysical products derived from various satellite sensors (mostly for
130 thermal bands and their corresponding applications) such as ASTER, MODIS, Landsat 5 and
131 ETM+ (Barsi et al., 2007; Hook et al., 2007; Steissberg et al., 2005; Tonooka and Palluconi,
132 2005; Tonooka et al., 2005), as well as for the sensor preflight and inflight calibration (Parada et

133 al., 1997; Thome et al., 1998). However, few satellite reflective solar radiance observations have
134 ever been used to characterize and quantify the lake ecosystem, and monitor water optical,
135 biological, and biogeochemical changes. This is mainly due to issues in the satellite data
136 processing, i.e., atmospheric correction (Gordon and Wang, 1994; IOCCG, 2010; Wang, 2007),
137 to accurately account for the effect of lake surface altitude (Wang, 2016). Specifically, the top-
138 of-atmosphere (TOA) Rayleigh scattering radiance computations were incorrect for high-altitude
139 lakes (Wang, 2016), leading to over-subtraction of the TOA Rayleigh-scattering radiance
140 contributions, and thereby deriving biased low $nL_w(\lambda)$ spectra (often negative values depending
141 on the lake surface altitude) (Gordon and Wang, 1994; Wang, 2016). Consequently, satellite-
142 derived optical, biological, and biogeochemical products (e.g., Chl-a, $K_d(490)$, SD, IOPs, or any
143 other products that use the inputs of satellite-measured $nL_w(\lambda)$ spectra) for global high-altitude
144 lakes were in error, and generally cannot be used. This includes satellite-derived water quality
145 products over global high-altitude lakes from the Sea-viewing Wide-field-view Sensor
146 (SeaWiFS), MODIS, and the Visible Infrared Imaging Radiometer Suite (VIIRS), etc. It is noted
147 that for global ocean color data processing the required ancillary data such as sea-level
148 atmospheric pressure, total column ozone amount, sea surface wind speed, and total column
149 water-vapor amount are routinely obtained from the National Center for Environmental
150 Prediction (NCEP) (Ramachandran and Wang, 2011).

151 Following a recent effort for improving the TOA Rayleigh radiance computations for
152 satellite ocean/water color remote sensing, in particular, for accounting for the water surface
153 altitude effect (Wang, 2016), the global VIIRS observations on both the Suomi National Polar-
154 orbiting Partnership (SNPP) and NOAA-20 satellites are processed using the improved Rayleigh
155 radiance computations in the Multi-Sensor Level-1 to Level-2 (MSL12) ocean color data
156 processing system (Wang et al., 2013). In this study, we use Lake Tahoe as an example to
157 demonstrate that high quality $nL_w(\lambda)$ spectra can now be derived from VIIRS measurements for
158 global high-altitude lakes to characterize and quantify the lake ecosystem as well as monitor
159 environmental changes. Specifically, VIIRS-SNPP-derived $nL_w(\lambda)$ spectra are compared with

160 those from the in situ measurements and shown to have good results. Furthermore, based on the
161 in situ SD measurements in the lake, a regional empirical SD algorithm has been developed
162 using the VIIRS-derived $nL_w(\lambda)$ at the wavelength of 551 nm. Therefore, the seasonal and
163 interannual variabilities of $nL_w(\lambda)$, Chl-a, and SD in Lake Tahoe are characterized and quantified.
164 In particular, we show that when using VIIRS-derived water property data the anomalous water
165 property in Lake Tahoe during the 2017–2018 abnormal event can be quantitatively investigated
166 and analyzed. Finally, the potential to use Lake Tahoe as an ocean color validation site for
167 evaluation and validation of satellite ocean color products, as well as routine sensor performance
168 monitoring (Wang et al., 2015), is discussed.

169 **2. Satellite-derived and in situ-measured water properties in Lake Tahoe**

170 *2.1. VIIRS-measured $nL_w(\lambda)$ spectra and other water color products*

171 Successfully launched in October 2011, VIIRS-SNPP provides continuous observations of
172 the Earth's atmosphere, land, cryosphere, and ocean properties with the 14 reflective solar bands
173 (RSBs) covering a spectral range of 410–2257 nm (Goldberg et al., 2013). For the satellite ocean
174 and inland water color remote sensing, the VIIRS-SNPP five visible bands (M1–M5) at the
175 nominal central wavelengths of 410, 443, 486, 551, and 671 nm, two imaging (I) bands (I1 and
176 I2) at 638 and 862 nm, two near-infrared (NIR) bands (M6 and M7) at 745 and 862 nm, and
177 three SWIR bands (M8, M10, and M11) at 1238, 1601, and 2257 nm are used to derive VIIRS
178 $nL_w(\lambda)$ spectra over the global ocean and coastal/inland waters (Wang and Jiang, 2018b; Wang et
179 al., 2013). It is noted that VIIRS measurements have spatial resolutions of 750 m and 375 m for
180 the M-bands and I-bands, respectively. Thus, VIIRS-derived $nL_w(\lambda)$ spectra at the spatial
181 resolution of 750 m are applicable for global open oceans (Wang et al., 2016a), while $nL_w(\lambda)$ at
182 the image bands (e.g., $nL_w(638)$ data) with the spatial resolution of 375 m are useful for various
183 coastal and inland water applications (Wang and Jiang, 2018b).

184 VIIRS global ocean and inland water color products have been routinely generated since
185 2012 using MSL12, which was originally developed for producing consistent global ocean color

186 data sets from multiple satellite sensors using a common data processing software (Wang, 1999;
187 Wang et al., 2002). It should be noted that for deriving accurate global ocean color products the
188 on-orbit vicarious calibration has been carried out for VIIRS using the in situ MOBY
189 measurements (Clark et al., 1997) in the waters off Hawaii (Wang et al., 2016b). During the
190 VIIRS period, MSL12 has been significantly modified and improved, including an atmospheric
191 correction algorithm (Jiang and Wang, 2014; Wang and Jiang, 2018a; Wang et al., 2012b), cloud
192 masking using the SWIR bands (Wang and Shi, 2006), straylight and cloud shadowing detection
193 (including the adjacency effect) (Bulgarelli et al., 2017; Hu et al., 2020; Jiang and Wang, 2013),
194 as well as improved satellite algorithms for various ocean/water color products, e.g., Chl-a
195 (Wang and Son, 2016), $K_d(490)$ (Wang et al., 2009a), IOPs (Shi and Wang, 2019), data quality
196 assurance (QA) (Wei et al., 2016), etc. In particular, improved TOA Rayleigh radiance
197 computations for accurately accounting for the effects of high-altitude inland waters have been
198 developed and implemented in MSL12 (Wang, 2016). It should be noted that in the MSL12
199 ocean/water color data processing the atmospheric diffuse transmittance (IOCCG, 2010; Yang
200 and Gordon, 1997) has also been modified to account for the high-altitude lake surface
201 atmospheric pressure changes (i.e., accounting for the Rayleigh optical thickness variations) for
202 both the solar and viewing directions. This is also an important fact impacted directly by the
203 Rayleigh optical thickness variation due to significant atmospheric pressure changes over high-
204 altitude lakes (IOCCG, 2010). However, it is assumed that there is negligible change in the
205 Rayleigh-aerosol interaction radiance term (Gordon and Wang, 1994; IOCCG, 2010; Wang,
206 2007) in the VIIRS data processing for a high-altitude lake. It should also be noted that VIIRS-
207 derived ocean color products have been well evaluated and validated over various ocean regions
208 and some sea-level inland waters through presentations in various conferences, meetings, and
209 workshops, as well as publications (Barnes et al., 2019; Hlaing et al., 2013; Mikelsons et al.,
210 2020; Wang et al., 2016a). However, VIIRS-derived inland water quality products have not been
211 evaluated over any high-altitude lakes such as Lake Tahoe.

212 One of the important issues for deriving accurate water property data over high-altitude
213 inland waters from remote sensing is to accurately account for the water surface altitude effect in
214 the TOA Rayleigh radiance computations (Wang, 2016), in addition to various other challenges
215 related to the data processing over global sea-level waters. This problem existed in previous
216 satellite ocean color data processing, which assumed the sea-level water surface property (i.e.,
217 atmospheric pressure measured and used at the sea-level) (Gordon et al., 1988a; Wang, 2002,
218 2005). For high-altitude lakes such as Lake Tahoe, this leads to an over-estimation of the TOA
219 spectral Rayleigh radiance contributions (therefore, over subtractions of Rayleigh radiance
220 values in atmospheric correction), thereby deriving biased low $nL_w(\lambda)$ spectra (or even negative
221 values) (Gordon and Wang, 1994; IOCCG, 2010; Wang, 2007). For example, the ratio of the
222 surface atmospheric pressure at 2 km altitude to that at sea-level is about 0.78, i.e., the TOA
223 Rayleigh-scattering radiance values over a 2 km altitude lake have an approximately similar
224 factor smaller than those over the sea-level lakes (Wang, 2016).

225 In fact, for high-altitude inland waters, the accurate calculation of the TOA spectral
226 Rayleigh scattering radiances is really important (particularly for the short visible bands) because
227 the TOA radiances contributed from aerosols are usually less important because aerosols are
228 mostly located close to sea-level (e.g., within ~2 km). With the improved TOA Rayleigh
229 radiance computation in MSL12 (Wang, 2016), the effects of high-altitude lakes (i.e.,
230 significantly low atmospheric pressure, and therefore, reduced spectral Rayleigh optical
231 thicknesses) are accurately accounted for. Therefore, water property data can be derived
232 accurately over global high-altitude waters including Lake Tahoe.

233 *2.2. In situ-measured $nL_w(\lambda)$ spectra*

234 As NASA's calibration site, four permanent moored buoys located at different locations in
235 Lake Tahoe routinely make in situ measurements such as longwave and shortwave radiations,
236 wind speed and direction, atmospheric pressure, aerosol optical depth (AOD), and water skin
237 temperature (<https://laketahoe.jpl.nasa.gov/measurements>). Specifically, the moored buoy at the
238 TB3 station as noted in Fig. 1 is located at [39°06'37"N, 120°04'31"W]. Hyperspectral radiance

239 measurements in the wavelength range of 360 nm to 875 nm were taken with the hyperspectral
240 radiometers installed at above water surface (0^+ m), 2 m, and 9 m depths on the TB3 buoy. Each
241 radiometer measured the downward spectral irradiance $E_d(\lambda, z)$ and upward spectral radiance
242 $L_u(\lambda, z)$ as a function of water depth z . Measurements were made hourly from local time 6:00 to
243 19:00. The water attenuation coefficient for $L_u(\lambda, z)$ was then calculated, and consequently, $L_u(\lambda,$
244 $0^+)$ (calculated from $L_u(\lambda, 0^-)$) was estimated. Finally, the remote sensing reflectance spectra
245 $R_{rs}(\lambda)$ were calculated with $L_u(\lambda, 0^+)$ and $E_d(\lambda, 0^+)$ (Watanabe et al., 2016).

246 During 2013, in situ in-water radiometric profile data were also taken at the TB3 station and
247 mid-lake stations with another vertical profiling hyperspectral radiometer system within two
248 hours of the local solar noon on March 1, April 28, July 25, September 9, October 25, and
249 December 16. Remote sensing reflectance spectra $R_{rs}(\lambda)$ were then calculated from the
250 measurements of the subsurface upward spectral radiance $L_u(\lambda, 0^-)$ and the downward spectral
251 irradiance $E_d(\lambda, 0^+)$ (Watanabe et al., 2016).

252 A total of 14 in situ $R_{rs}(\lambda)$ spectra measurements as described above at the station TB3 in the
253 lake were shown in Watanabe et al. (2016). Specifically, these in situ $R_{rs}(\lambda)$ spectra covered a
254 time period from December 16, 2013, to April 15, 2015. In this study, these 14 in situ $R_{rs}(\lambda)$
255 spectra acquired at the TB3 station in Lake Tahoe (Fig. 1) were converted to in situ $nL_w(\lambda)$
256 spectra to further compare, evaluate, and validate VIIRS-derived $nL_w(\lambda)$ spectra.

257 In addition to the in situ radiometric measurements at the TB3 station, continuous water
258 quality monitoring has been routinely conducted since 1968. The SD and Chl-a (at the depths of
259 0, 10, 50, 100 m) have been sampled routinely approximately every 34 days at the Mid-Lake
260 Tahoe Productivity (MLTP) station located at [39°8'30"N, 120°00'55.5"W] (noted in Fig. 1). The
261 in situ SD measurements and coincident VIIRS-derived $nL_w(\lambda)$ at the green band (551 nm) can
262 be used to develop an empirical algorithm to derive SD (Binding et al., 2008). The temporal
263 variations of VIIRS-derived Chl-a and SD data at the MLTP station are compared with those
264 from the in situ measurements to evaluate satellite algorithms performance for these two
265 important water property products.

266 3. Results

267 3.1. VIIRS-derived $nL_w(\lambda)$ spectra compared with those from in situ measurements

268 There are 10 (out of 14) radiance matchups for VIIRS-SNPP-derived and in situ-measured
269 $nL_w(\lambda)$ spectra at the TB3 station on the same date in Lake Tahoe. To compare the VIIRS-SNPP-
270 derived and in situ-measured $nL_w(\lambda)$, a box of 7×7 pixels (pixel at about 750 m spatial
271 resolution) centered at the TB3 station in the remapped $nL_w(\lambda)$ image was set, and the median of
272 $nL_w(\lambda)$ values in the box was calculated as the VIIRS-derived $nL_w(\lambda)$ value in order to compare
273 with the in situ $nL_w(\lambda)$ spectra. There were four cases with the cloud cover when VIIRS-SNPP
274 passed over the TB3 station, preventing the production of valid VIIRS $nL_w(\lambda)$ spectra. Fig. 2
275 provides examples of comparative results between VIIRS-derived and in situ-measured $nL_w(\lambda)$
276 spectra at the TB3 station on June 15, 2013 (Fig. 2a), July 15, 2013 (Fig. 2b), August 15, 2013
277 (Fig. 2c), and September 15, 2013 (Fig. 2d). The comparisons in Fig. 2 show that VIIRS-derived
278 $nL_w(\lambda)$ spectra match quite well with those from the in situ measurements, and the two data sets
279 are generally consistent (Fig. 2). However, there are some minor differences in the matchups in
280 $nL_w(\lambda)$ as VIIRS-derived $nL_w(410)$ and $nL_w(443)$ are slightly biased low on June 15, 2013,
281 relative to those from the in situ measurements. Overall, results in Fig. 2 demonstrate that
282 VIIRS-derived $nL_w(\lambda)$ spectra using the improved new Rayleigh radiance computation in MSL12
283 are generally accurate, showing significant improvements compared to those with incorrect
284 Rayleigh radiance computations for global high-altitude lakes (i.e., significantly biased low
285 $nL_w(\lambda)$ values, and even negative at the blue bands).

286 The overall accuracy of VIIRS-derived $nL_w(\lambda)$ spectra at the TB3 station is further evaluated
287 (Fig. 3). Since Lake Tahoe has high water clarity, values of $nL_w(638)$ and $nL_w(671)$ are generally
288 very small (close to zero). Differences in matchups for $nL_w(638)$ and $nL_w(671)$ are significantly
289 amplified in the logarithmic scale as shown in Fig. 3, although the differences are actually trivial.
290 For $nL_w(\lambda)$ at the blue and green bands (wavelengths at 410, 443, 486, and 551 nm), VIIRS-
291 derived $nL_w(\lambda)$ data agree quite well with the in situ measurements, with the mean $nL_w(\lambda)$ ratio
292 value between VIIRS-derived and in situ-measured $nL_w(\lambda)$ of 0.999 and the coefficient of

293 determination R^2 between the two data sets (in linear scale) of 0.857. Specifically, the mean
294 $nL_w(\lambda)$ ratio values between VIIRS-derived and in situ-measured $nL_w(\lambda)$ at the VIIRS
295 wavelengths of 410, 443, 486, 551, 638, and 671 nm are 1.019, 0.936, 1.076, 0.967, 1.297, and
296 1.286, respectively, while the mean $nL_w(\lambda)$ absolute difference values between the two for the
297 corresponding six VIIRS bands are 0.124, 0.109, 0.078, 0.037, 0.023, and 0.024 $\text{mW cm}^{-2} \mu\text{m}^{-1}$
298 sr^{-1} , respectively. It should be noted again that, for small values such as $nL_w(\lambda)$ at the VIIRS red
299 bands (638 and 671 nm), the mean absolute difference is a better and more meaningful measure
300 for describing the uncertainty. In fact, it shows the smallest mean absolute difference values for
301 $nL_w(\lambda)$ at the red bands although the $nL_w(\lambda)$ ratios at these bands are the largest due to very small
302 values. Results in Figs. 2 and 3 confirm that high-quality $nL_w(\lambda)$ spectra can be derived from
303 VIIRS-SNPP observations over Lake Tahoe. Therefore, the lake water optical, biological, and
304 biogeochemical property data (e.g., Chl-a and SD) can be derived from VIIRS-SNPP-measured
305 $nL_w(\lambda)$ spectra, and these data can be used to study the lake water property dynamics,
306 characterize and quantify the long-term ecosystem variability, monitor environmental change,
307 and detect hazardous events over the global high-altitude lakes. These possibilities and potentials
308 are evaluated and discussed below.

309 3.2. VIIRS-derived SD and Chl-a compared with the in situ data

310 In Lake Tahoe, the SD variations are correlated to the amount of fine sediment in the water
311 column (Jassby et al., 1999; Jassby et al., 2003). Loading and settling rates of such fine inorganic
312 particles ($< \sim 16 \mu\text{m}$ in diameter) were found to have the largest impact on the lake clarity (Sahoo
313 et al., 2010). Considering the fact that $nL_w(\lambda)$ at the red end of the spectrum (638 and 671 nm) are
314 usually close to zero for the lake, $nL_w(551)$ can be a sensitive indicator for the inorganic particle
315 concentrations and correlate with water clarity. Indeed, Binding et al. (2015) showed that the SD
316 in the Great Lakes can be well correlated to satellite-derived $nL_w(551)$ values. At the MLTP
317 station in Lake Tahoe (Fig. 1), there are total of 57 valid VIIRS $nL_w(\lambda)$ retrievals coincident with
318 the in situ SD measurements in the period between 2012 and 2018. Using these 57 sets of data,
319 an empirical SD model for the lake can then be developed and described as following:

320
$$SD = 10^{(a_0 - a_1 \times nL_w(551))}, \quad (1)$$

321 where coefficients $a_0 = 1.484$ and $a_1 = -0.551 \text{ mW}^{-1} \text{ cm}^2 \mu\text{m sr}$, which were derived from the
322 best fit to the in situ SD data to VIIRS-derived $nL_w(551)$. Fig. 4 shows a comparison between the
323 VIIRS-derived (using Eq. (1)) and in situ-measured SD at the MLTP station. The mean and
324 median ratios of the VIIRS-derived SD and in situ values are 1.0649 and 1.0365 with the
325 standard deviation (STD) of 0.1969. The comparison results (Fig. 4) show that VIIRS-derived
326 SD data are quite reasonable and can be used to estimate water clarity in the lake from satellite
327 observations.

328 Fig. 5 further provides the performance evaluation of VIIRS-derived Chl-a (Fig. 5a) and SD
329 (Fig. 5b) with the corresponding in situ measurements at the MLTP station during the period of
330 2012–2018. VIIRS-derived Chl-a data show the same seasonal trend as that from the in situ Chl-
331 a measurements, i.e., high Chl-a in the winter and low Chl-a in the summer. It should be noted
332 that in situ Chl-a data were derived as mean values from measured Chl-a data at the surface and
333 at 10 m water depth, considering SD values are normally between 20 and 30 m (therefore,
334 excluding in situ Chl-a data at 50 and 100 m). Results show that VIIRS-derived low Chl-a values
335 in the MLTP station are $\sim 0.2 \text{ mg m}^{-3}$ and consistent with those from in situ data. However, some
336 discrepancies between VIIRS-derived and in situ-measured Chl-a can be found for high Chl-a
337 values in the winter season. For example, VIIRS-derived Chl-a had the same peaks as the in situ
338 measurements in the winters of 2012–2013, 2013–2014, and 2017–2018. VIIRS-derived Chl-a
339 was particularly different from the in situ data in the winter of 2016–2017 (Fig. 5a).

340 VIIRS-derived SD data in the MLTP station show the same seasonal and interannual
341 variations as those from the in situ SD measurements (Fig. 5b). In the period between 2012 and
342 2016, VIIRS-derived SD data were consistent with the in situ measurements in terms of both the
343 magnitude and seasonal variation. However, relatively large SD discrepancies between them
344 were found with anomalously low in situ SD measurements in mid-2016 and late 2017, although
345 the SD seasonal variation from the two SD data sets was still the same.

346 Although there are some disagreements in the magnitude for VIIRS and in situ measured SD
347 (Figs. 4 and 5b) and Chl-a (Fig. 5a), results are generally consistent in terms of magnitude values
348 and particularly in their variations. It should be noted that the in situ SD and Chl-a measurements
349 also have their own uncertainties due to the instrument limitations and the operation differences,
350 e.g., human factors for the SD estimation. These uncertainties also contribute to the differences
351 between VIIRS and in situ measured Chl-a and SD data. Therefore, we can conclude that VIIRS-
352 derived Chl-a and SD data have reasonably good accuracy, and these data products along with
353 VIIRS-derived $nL_w(\lambda)$ spectra can be used to study water properties in Lake Tahoe.

354 3.3. VIIRS-measured climatology of $nL_w(\lambda)$, Chl-a, and SD

355 Noticeably, there are some significant differences in VIIRS-derived water property data in
356 2017 and early 2018, compared with those in the other years. Indeed, an anomalous event
357 happened in the period of 2017 to early 2018 in the region, and significantly impacted water
358 properties in Lake Tahoe (Staff-Report, 2018). Therefore, for the calculations of climatology
359 water properties in the lake, VIIRS measurements from 2012–2016 (excluding years 2017 and
360 2018) are used.

361 Fig. 6 provides the climatology images of $nL_w(410)$ (Fig. 6a), $nL_w(443)$ (Fig. 6b), $nL_w(486)$
362 (Fig. 6c), $nL_w(551)$ (Fig. 6d), $nL_w(638)$ (Fig. 6e), $nL_w(671)$ (Fig. 6f), Chl-a (Fig. 6g), and SD
363 (Fig. 6h), which were calculated as the median of all valid retrievals from VIIRS-SNPP
364 measurements between 2012 and 2016, providing the normal water optical, biological, and
365 biogeochemical conditions of the lake. Results from VIIRS observations show that Lake Tahoe
366 features include spatial uniformity in water properties and enhanced $nL_w(\lambda)$ at the blue bands,
367 except for small areas in the southern and northwestern parts of the lake where water depth is
368 less than ~20 m. Fig. 6 shows that spatial distributions of Chl-a, SD, and $nL_w(\lambda)$ in the lake are
369 quite uniform and there is little spatial difference with normal Chl-a ~0.25 mg m⁻³ and SD ~22 m
370 for the entire lake. The changes of $nL_w(\lambda)$ spectra are also small across the lake for $nL_w(\lambda)$ at the
371 all VIIRS bands. In fact, there is no specific spatial pattern for all climatology $nL_w(\lambda)$ spectra in
372 the lake. Spectrally, VIIRS-derived $nL_w(\lambda)$ spectra show the typical feature of clear blue waters

373 (Gordon et al., 1988b; Morel and Maritorena, 2001), showing $nL_w(443)$ (Fig. 6b) $\sim 0.8 \text{ mW cm}^{-2}$
374 $\mu\text{m}^{-1} \text{ sr}^{-1}$, while $nL_w(410)$ (Fig. 6a) and $nL_w(486)$ (Fig. 6c) are a little bit less than $nL_w(443)$. In
375 comparison, the climatology $nL_w(551)$ (Fig. 6d) in the lake is $\sim 0.3 \text{ mW cm}^{-2} \mu\text{m}^{-1} \text{ sr}^{-1}$. Both
376 $nL_w(638)$ (Fig. 6e) and $nL_w(671)$ (Fig. 6f) are less than $\sim 0.05 \text{ mW cm}^{-2} \mu\text{m}^{-1} \text{ sr}^{-1}$.

377 A transect line from the north to south across the lake is defined to further characterize the
378 uniformity of the lake spatial distributions in $nL_w(\lambda)$ spectra, Chl-a, and SD (Fig. 6a). Fig. 7
379 provides quantitative results of $nL_w(\lambda)$ spectra, Chl-a, and SD as a function of distance from the
380 north to south along the transect line noted in Fig. 6a. Fig. 7a shows that climatology $nL_w(443)$ is
381 generally stable at $\sim 0.8 \text{ mW cm}^{-2} \mu\text{m}^{-1} \text{ sr}^{-1}$ from the north of the transect line to 20 km in the
382 lake, and it trends a little lower to $\sim 0.6 \text{ mW cm}^{-2} \mu\text{m}^{-1} \text{ sr}^{-1}$ from 20 km to 30 km. However,
383 climatology $nL_w(551)$ value ($\sim 0.3 \text{ mW cm}^{-2} \mu\text{m}^{-1} \text{ sr}^{-1}$) is quite stable and does not show any
384 noticeable change along the transect line (Fig. 7a). In the southern end of the transect line, the
385 water type changes from a typical clear blue water to a typical shallow (bottom-affected) water
386 with significantly enhanced $nL_w(\lambda)$ in the visible bands (Fig. 7a).

387 The variations of climatology Chl-a and SD along the transect line are shown in Fig. 7b.
388 Chl-a shows little variation with a value of $\sim 0.25 \text{ mg m}^{-3}$ for the majority of the lake even though
389 it increases slightly to $\sim 0.3 \text{ mg m}^{-3}$ in the northern end region and spikes to over $\sim 1.0 \text{ mg m}^{-3}$ in
390 the southern end region (Figs. 6g and 7b). Similarly, SD is quite stable at $\sim 22 \text{ m}$ for the majority
391 of the transect line (typical oligotrophic waters) although decreased SD can be found in the
392 coastal region of southern Lake Tahoe (Figs. 6h and 7b).

393 It is noted that significantly enhanced Chl-a and $nL_w(\lambda)$ and reduced SD are observed in the
394 southern coast of the lake (Fig. 6), as well as in the southern end of the transect line (Fig. 7). In
395 fact, VIIRS-derived $nL_w(\lambda)$ spectra in the southern coastal region show effects like typical
396 bottom-affected water with significantly enhanced $nL_w(\lambda)$ spectra (Fig. 6). Although VIIRS-
397 derived $nL_w(\lambda)$ spectra in the region might still be valid with the NIR-SWIR combined
398 atmospheric algorithm (Wang and Shi, 2007), the lake bottom reflectance contributions to the
399 enhancements of the derived $nL_w(\lambda)$ spectra in the visible bands (especially in the blue bands)

400 can indeed lead to large errors in VIIRS-derived Chl-a and SD data in the southern end of the
401 region. Thus, VIIRS-derived Chl-a and SD in the bottom-impacted region may have significant
402 errors and should not be considered valid.

403 *3.4. Characterization of seasonal and interannual variability in water property*

404 Seasonal variations of $nL_w(\lambda)$ spectra, Chl-a, and SD in Lake Tahoe are shown in Fig. 8. For
405 $nL_w(443)$ (Fig. 8a, f, k, and p), it reaches its peak value in the summer season (Fig. 8f), and the
406 minimum occurs in the winter season (Fig. 8p). The seasonal change of $nL_w(551)$ (Fig. 8b, g, l,
407 and q) is relatively small (but quite noticeable) with highs in the summer and lows in the winter,
408 corresponding well to the seasonal variation in SD (Fig. 8e, j, o, and t) (as expected). Similarly,
409 the seasonal variation of $nL_w(671)$ in the lake is also very small (Fig. 8c, h, m, and r). Similar to
410 the climatology results, spatial variations of Chl-a in the lake for each season are generally small
411 (Fig. 8d, i, n, and s). The highest Chl-a value can be found in the winter (Fig. 8s), while the
412 lowest Chl-a occurs in the summer (Fig. 8i). The seasonal change of SD is also noticeable with
413 the lowest SD in the summer (Fig. 8j) and the highest SD in the winter (Fig. 8t).

414 In addition, Fig. 9 shows the interannual variations of $nL_w(\lambda)$ spectra, Chl-a, and SD from
415 2012 to 2018. VIIRS-derived $nL_w(443)$ in 2017 (Fig. 9z) decreased remarkably as compared to
416 the other years for the entire lake (Fig. 9a, f, k, p, u, and ee). Particularly, in the southern region
417 of the lake, $nL_w(443)$ dropped to below $\sim 0.5 \text{ mW cm}^{-2} \mu\text{m}^{-1} \text{ sr}^{-1}$. This indicates that there were
418 increased amounts of absorbing components in the water column. It is noted that the statistical
419 results in Fig. 9 should be quite reliable although there may be errors in the VIIRS-derived
420 $nL_w(\lambda)$ spectra. Although $nL_w(443)$ increased in 2018 (Fig. 9ee), the values were still lower than
421 those in the period of 2012–2016. Unlike $nL_w(443)$ (Fig. 9a, f, k, p, u, z, and ee), the interannual
422 variations of $nL_w(551)$ (Fig. 9b, g, i, q, v, aa, and ff) and $nL_w(671)$ (Fig. 9c, h, m, r, w, bb, and
423 gg) were not significant (not noticeable).

424 It is noted that VIIRS-derived Chl-a values in 2017 and 2018 may be over-estimated due to
425 the Chl-a algorithm issue in dealing with the increased amount of inorganic suspended particles
426 from high river runoff in that year. Thus, VIIRS-derived Chl-a values in 2017 and 2018 might be

427 biased, and therefore, are not shown in Fig. 9. Except for the years 2017 and 2018, the spatial
428 distributions of Chl-a in the other six years (Fig. 9d, i, n, s, and x) are similar. On the other hand,
429 different from the Chl-a interannual variation, SD showed notable decreases in 2016 and 2017
430 (Fig. 9y and dd) in comparison to normal (climatology) SD values (Fig. 6h) and to those in the
431 other years (Fig. 9e, j, o, t, and ii).

432 In addition, the seasonal and interannual variation in $nL_w(\lambda)$ spectra for the entire Lake
433 Tahoe are quantitatively evaluated (Fig. 10). Fig. 10a shows the clear seasonal change in $nL_w(\lambda)$
434 spectra. Specifically, $nL_w(410)$, $nL_w(443)$, and $nL_w(486)$ values are the highest in the summer and
435 the lowest in the winter. In the spring season, $nL_w(\lambda)$ spectrum is similar to the climatology (Fig.
436 10a). The seasonal differences in $nL_w(551)$ are smaller than those of the shorter wavebands, but
437 are still obvious with highs in the summer and lows in the winter, reflecting the SD seasonal
438 variation in the lake (Fig. 5b). It is noted again that $nL_w(638)$ and $nL_w(671)$ are both close to 0 for
439 all four seasons.

440 The interannual variability of $nL_w(\lambda)$ spectra for Lake Tahoe is also significant (Fig. 10b). In
441 particular, the $nL_w(\lambda)$ spectrum in 2017 was abnormal and outlying, showing significantly low
442 values for $nL_w(410)$, $nL_w(443)$, and $nL_w(486)$. For example, $nL_w(410)$ in 2017 was $\sim 0.5 \text{ mW cm}^{-2}$
443 $\mu\text{m}^{-1} \text{ sr}^{-1}$, compared to the normal value of $\sim 0.8 \text{ mW cm}^{-2} \mu\text{m}^{-1} \text{ sr}^{-1}$ in the other years. In 2018,
444 the depression of $nL_w(410)$ was alleviated, but still lower than those in the other normal years.
445 Fig. 10b shows that $nL_w(\lambda)$ spectra in 2012, 2014, 2015, and 2016 were all similar although some
446 slight interannual variations did exist. All of these seasonal and interannual variations in VIIRS-
447 derived $nL_w(\lambda)$ spectra are consistent with $nL_w(\lambda)$ spectral images as shown in Figs. 8 and 9.

448 3.5. The 2017–2018 anomaly event in Lake Tahoe

449 The interannual variability of water properties in Lake Tahoe in Figs. 9 and 10 shows that
450 anomalous $nL_w(\lambda)$ spectra and SD occurred in 2017, and this abnormal event extended to 2018.
451 Specifically, the median values of SD for the entire Lake Tahoe in each month between 2012 and
452 2018 were calculated to characterize and quantify the long-term temporal variations and the
453 environmental anomaly in 2017.

454 The SD variation in Lake Tahoe between 2012 and 2018 is shown in Fig. 11a. The seasonal
455 variation of SD was more pronounced in comparison to the interannual SD variation with SD
456 ranging between 18–19 m and 23–24 m. The lowest SD values were in the summers of 2013,
457 2016, and 2017. After removing the seasonal SD variation, Fig. 11b shows that SD anomaly
458 dropped below -2 m in the spring of 2017, and kept negative SD anomaly values for the entire
459 2017 year, consistent with the local media report (Staff-Report, 2018). In comparison, the
460 negative SD anomaly values in 2013 and 2016 were mild and lasted only a couple of months.
461 The temporal variation of the SD anomaly in Fig. 11b clearly shows that 2017 is the year with
462 the least water clarity in the studying period between 2012 and 2018.

463 Fig. 12 provides the quantitative spatial image details of the temporal evolvments of $nL_w(\lambda)$
464 spectra and SD in the period from early 2017 to spring 2018. In the period between December
465 2016 and February 2017, all lake water properties (Fig. 12a–d) were similar to those of
466 climatology for the winter season (Fig. 8p, q, r, and t). Starting in the spring of 2017 (March
467 2017–May 2017), the decreased $nL_w(443)$ (Fig. 12e) was found comparable to the typical
468 $nL_w(443)$ in the same season (Fig. 8a). In the summer of 2017 (June 2017–August 2017),
469 $nL_w(443)$ (Fig. 12i) was significantly lower than the same-season climatology $nL_w(443)$ (Fig. 8f).
470 In the autumn of 2017, $nL_w(443)$ reached the lowest (Fig. 12m). It is also noted that SD in this
471 season (Fig. 12p) was also lower than the SD climatology in the autumn (Fig. 8o).

472 Starting in the 2017–2018 winter season, $nL_w(\lambda)$ and SD went back to normal (Fig. 12q–t)
473 although anomalous $nL_w(443)$ and SD were still remarkable in comparison to their typical values
474 (Fig. 8p, q, r, and t). In the spring of 2018 (March 2018–May 2018), VIIRS-SNPP-measured
475 $nL_w(\lambda)$ and SD (Fig. 12u–x) in the lake were fully recovered from the abnormal environmental
476 event in 2017. Indeed, the lake’s $nL_w(\lambda)$ and SD images in this period were similar to their
477 typical ones (Fig. 8a, b, c, and e).

478 The water property anomaly in 2017 was observed also by the in situ water quality
479 monitoring system and reported elsewhere (TERC, 2018). The annual average of SD (18.2 m)
480 (TERC, 2018) was the lowest value ever recorded since Lake Tahoe long-term monitoring

481 started in 1968, while the average SD in 2014, 2015, and 2016 in the lake were 23.7, 22.3, and
482 21.1 m, respectively. This anomalous environmental variation was attributed to the extreme
483 weather and hydrologic events, all related to climatic change in the region. The total precipitation
484 in the region was 175 cm in 2017 in comparison to amounts below 50 cm in 2014 and 2015. This
485 high precipitation event led to high inflow carrying unusually high amount of sediments from
486 surrounding tributaries. As a result, the concentrations of inorganic particles in the lake were
487 elevated throughout the year, causing observed low water clarity, which is consistent with
488 VIIRS-SNPP measurements.

489 The in situ Chl-a measurements did not show anomalous values in 2017, although the in situ
490 Chl-a data exhibited relatively higher values in the winter. However, anomalous high Chl-a
491 values were derived from VIIRS-SNPP measurements in 2017, which were associated with low
492 $nL_w(\lambda)$ in the shorter wavebands (410, 443, and 486 nm) (Figs. 9z and 10b) and may be heavily
493 affected by absorption of high concentrations of inorganic sediments brought by high inflow as
494 discussed above. This discrepancy (likely due to the satellite Chl-a algorithm issue), despite the
495 reasonably good agreement of satellite-derived and in situ measurements in normal years,
496 suggested the importance of improving the understanding of detailed properties of optically
497 significant components in the water column and optical process of the lake for future
498 biogeochemical monitoring with higher accuracy.

499 **4. Discussions and Conclusion**

500 In the last three decades, satellite ocean color remote sensing (e.g., SeaWiFS, MODIS,
501 VIIRS, etc.) has been widely used in user communities for scientific researches and applications
502 (McClain, 2009) to understand global and regional water optical, biological, and biogeochemical
503 properties, and to evaluate their impact on climatic change, natural hazards, and various other
504 environmental variations. In fact, satellite ocean color data have been playing a critical role to
505 monitor and understand water quality over the global open ocean, coastal, and inland waters
506 (IOCCG, 2008, 2018).

507 The TOA Rayleigh scattering radiance computations, however, exhibited a significant error
508 in satellite-measured global high-altitude water property data due to an issue in atmospheric
509 correction to accurately account for the effects of the high-altitude water surface property (i.e.,
510 surface atmospheric pressure). The error resulted in considerably biased low $nL_w(\lambda)$ spectra
511 which are often unusable for water property monitoring. Indeed, there are few satellite water
512 color remote sensing studies over high-altitude lakes. It was shown that, with new improved
513 Rayleigh radiance computations (Wang, 2016), satellite remote sensing can now routinely
514 produce accurate and reliable water optical properties over high-altitude Lake Tahoe, similar to
515 the well-established remote sensing capability over global sea-level inland bodies of water. The
516 present results demonstrated that the new computation method can derive accurate $nL_w(\lambda)$ spectra
517 from high-altitude inland waters. It also showed that Chl-a and SD values estimated from $nL_w(\lambda)$
518 spectra were reasonable. Therefore, satellite-measured $nL_w(\lambda)$ spectra, Chl-a, SD, and potentially
519 other related products can be effectively used to characterize spatial and temporal variations of
520 the lake water properties. This is indeed an important development for future satellite-based
521 water quality monitoring of global inland waters.

522 Based on the in situ SD measurements, we developed a regional SD retrieval algorithm from
523 VIIRS observations for Lake Tahoe. Chl-a and SD from VIIRS observations were generally
524 consistent with the in situ measurements even though the algorithms are not perfect and some
525 discrepancies indeed exist. It is also noted that $nL_w(671)$ is higher in the center of the lake as in
526 Figs. 6, 8, and 9, while SD in the central lake is similar to SD in the coastal region. Although the
527 differences are minor, e.g., $nL_w(671)$ values of 0.031, 0.036, and 0.031 $\text{mW cm}^{-2} \mu\text{m}^{-1} \text{sr}^{-1}$ for the
528 transection line points of 5, 15, and 25 km in Fig. 7, respectively, it suggests that other water IOP
529 components such as colored dissolved organic matter (CDOM), particularly suspended inorganic
530 and organic particles may also contribute to the water property in the coastal region (reflected
531 particularly in the SD measurements).

532 VIIRS-SNPP observations over Lake Tahoe between 2012 and 2018 are used to quantify
533 water optical ($nL_w(\lambda)$ spectra), biological (Chl-a) and water quality (SD) properties in the lake.

534 Climatology $nL_w(\lambda)$, Chl-a, and SD show that the spatial variation of the lake is quite small with
535 Chl-a usually around $\sim 0.25 \text{ mg m}^{-3}$ and enhanced $nL_w(\lambda)$ at the blue bands. The present results
536 also showed that there was seasonal variability in the measured variables. The highest Chl-a and
537 SD normally occur in the winter season and the lowest values in the summer season. Significant
538 seasonal $nL_w(\lambda)$ variations can also be found at the VIIRS spectral bands of 410, 443, and 486
539 nm. In the summer, VIIRS-derived $nL_w(\lambda)$ at these three bands are generally above $\sim 0.8 \text{ mW cm}^{-2}$
540 $\mu\text{m}^{-1} \text{ sr}^{-1}$, while they are normally below $\sim 0.5 \text{ mW cm}^{-2} \mu\text{m}^{-1} \text{ sr}^{-1}$ in the winter. The seasonal
541 variation for VIIRS-derived $nL_w(551)$ is also obvious with highs in the summer and lows in the
542 winter.

543 In addition, there is significant interannual variability of water properties in Lake Tahoe
544 observed from VIIRS measurements from 2012–2018. In particular, 2017 was an abnormal year
545 with reduced SD and $nL_w(\lambda)$ at the blue bands. The anomaly started in the spring of 2017,
546 reached the minimum for SD and $nL_w(443)$ in the 2017 summer-autumn season. During the
547 2017–2018 winter, SD started recovering, and went back to the normal in the spring of 2018.
548 Anomalously low VIIRS-measured SD values were consistent with in situ measurements.
549 Following the anomalous climate condition in the region during the 2017–2018 period, the
550 increased suspended particles would probably be the cause of the low $nL_w(\lambda)$ in the blue end of
551 the spectrum (Fig. 10b), which ultimately led to the overestimation of Chl-a in the season. This
552 result suggests that a better understanding of optical processes of the lake with detailed
553 information about optically active components in the water column is desired to account for such
554 extreme conditions as well as future changes in environmental conditions. Still, the general trend
555 of water clarity and Chl-a in a normal year were well estimated by VIIRS-SNPP observations
556 with the current method.

557 The anomalous water properties observed from VIIRS-SNPP between 2017–2018 can be
558 attributed to the interannual climate variability in the region. As an element of the tributary
559 watershed drainage within the Truckee River Basin system, the Tahoe watershed covers more
560 than 1000 km^2 . Thus, the ecosystem can be significantly impacted by the precipitation variation

561 in the region. In early 2017, the western U.S. including California and Nevada experienced a
562 heavy precipitation period during the winter season following a multiyear drought period which
563 impacted the region significantly (<https://www.ncdc.noaa.gov/sotc/national/201713>). Indeed, the
564 total precipitation in the region was 175 cm in 2017 in comparison to below 50 cm in 2014 and
565 2015 (TERC, 2018). The anomalous precipitation amount and river runoff could transport more
566 inorganic and organic suspended matters, CDOM, and nutrients to the lake, leading to the
567 depressed SD in the 2017–2018 period.

568 The present research is the first study to derive, validate, and apply satellite-measured
569 $nL_w(\lambda)$ spectra from high-altitude lakes using new Rayleigh radiance computations, and results
570 shown here have important environmental and societal implications. The study demonstrates that
571 satellite ocean color remote sensing can provide long-term monitoring with high frequency
572 observations not only for the global ocean, coastal and sea-level inland waters, but also for
573 global high-altitude inland waters. For example, lakes in the Tibetan Plateau, which can be
574 considerably affected by the glacier melting due to global climate change, are rarely investigated,
575 because they are remotely scattered around in a vast area, and usually have an inclement
576 environment and weather conditions. Satellite remote sensing as demonstrated in this study can
577 provide an effective and efficient tool to evaluate the long-term environmental variability of
578 these systems in response to climate change.

579 Finally, it has been clearly shown that Lake Tahoe is generally a clear oligotrophic water
580 body with high spatial uniformity over almost the entire lake. These characteristics make it ideal
581 to be used as evaluation, validation, and a monitoring site for satellite ocean color remote
582 sensing. In fact, Lake Tahoe has been used for calibration and validation of thermal bands for
583 various satellite sensors (Barsi et al., 2007; Hook et al., 2007; Steissberg et al., 2005; Tonooka
584 and Palluconi, 2005; Tonooka et al., 2005). Advantages of Lake Tahoe as an additional
585 calibration and validation site for satellite ocean color sensors include (1) logistically easy to
586 access for maintaining the site (as the site has already been established and used for calibration
587 and validation for thermal bands), (2) located at high-altitude (~2 km) with less atmospheric

588 effects (i.e., less TOA radiance contributions from molecules and aerosols), and (3) weather
589 conditions are generally cooperative (e.g., no hurricanes) and instruments at the location are
590 more controllable. Therefore, Lake Tahoe can potentially serve as a good calibration and
591 validation site for satellite-derived ocean/water property products.

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597

598

Figure Captions

599 **Figure 1.** Map of Lake Tahoe located in California-Nevada of the U.S. The location for the in
600 situ $nL_w(\lambda)$ measurements at the TB3 station [39°06'37"N, 120°04'31"W] and in situ Chl-a and
601 SD at the MLTP station [39°8'30" N 120°00'55.5"W] are also marked.

602 **Figure 2.** (a) Comparisons of VIIRS-SNPP-derived and in situ-measured $nL_w(\lambda)$ spectra at the
603 TB3 station in Lake Tahoe for the measurement dates of (a) June 15, 2013, (b) July 15, 2013, (c)
604 August 15, 2013, and (d) September 29, 2013.

605 **Figure 3** Scatter plot of VIIRS-SNPP-derived $nL_w(\lambda)$ versus in situ-measured $nL_w(\lambda)$ at the
606 VIIRS spectral bands of 410, 443, 486, 551, 638, and 671 nm. Values of mean ratio (Mean Ratio,
607 unitless) and mean absolute difference (Mean Abs Diff, unit of $\text{mW cm}^{-2} \mu\text{m}^{-1} \text{sr}^{-1}$) between
608 VIIRS-derived and in situ-measured $nL_w(\lambda)$ at the six VIIRS spectral bands are also indicated in
609 the plot.

610 **Figure 4.** VIIRS-derived SD compared with those from the in situ measurements at the MLTP
611 station in Lake Tahoe.

612 **Figure 5.** VIIRS-derived water quality products compared with those from the in situ
613 measurements at the MLTP station between 2012 to 2018 for (a) Chl-a and (b) SD. Note that in
614 situ Chl-a data were derived with the mean from measured Chl-a values at the surface and 10 m
615 water depth.

616 **Figure 6.** VIIRS-derived climatology (2012–2016) water property data in Lake Tahoe for (a)
617 $nL_w(410)$, (b) $nL_w(443)$, (c) $nL_w(486)$, (d) $nL_w(551)$, (e) $nL_w(638)$, (f) $nL_w(671)$, (g) Chl-a, and (h)
618 SD. The transection line in panel (a) from the north to south at 120.02°W is marked for further
619 data analysis.

620 **Figure 7.** VIIRS-derived climatology (2012–2016) water property data along the transection line
621 (noted in Fig. 6a) in Lake Tahoe for products of (a) $nL_w(443)$, $nL_w(551)$, and $nL_w(671)$ and (b)
622 Chl-a and SD.

623 **Figure 8.** VIIRS-derived seasonal maps of water properties (2012–2016) in Lake Tahoe for
624 products of (along the row) $nL_w(443)$, $nL_w(551)$, $nL_w(671)$, Chl-a, and SD for the season of (a–e)
625 spring (March–May), (f–j) summer (June–August), (k–o) autumn (September–November), and
626 (p–t) winter (December–February).

627 **Figure 9.** VIIRS-derived interannual maps of water properties in Lake Tahoe for products of
628 (along the row) $nL_w(443)$, $nL_w(551)$, $nL_w(671)$, Chl-a, and SD for the year of (a–e) 2012, (f–j)
629 2013, (k–o) 2014, (p–t) 2015, (u–y) 2016, (z–dd) 2017, and (ee–ii) 2018. Note that Chl-a maps in
630 2017 and 2018 are not presented due to possible issue with the algorithm-caused overestimation.

631 **Figure 10.** VIIRS-derived $nL_w(\lambda)$ spectra in Lake Tahoe for (a) seasonal mean and climatology
632 (2012–2016) and (b) annual mean in 2012–2018.

633 **Figure 11.** VIIRS-derived temporal variation of the water property in Lake Tahoe between 2012
634 and 2018 for (a) SD and (b) SD anomaly. The red dotted line in plot (a) is the corresponding
635 climatology monthly mean values derived from 2012–2016.

636 **Figure 12.** VIIRS-derived Lake Tahoe water property of $nL_w(443)$, $nL_w(551)$, $nL_w(671)$, and SD
637 (along the row) and for the 2017–2018 abnormal event with specific time period of (a–d)
638 December 2016–February 2017, (e–h) March 2017–May 2017, (i–l) June 2017–August 2017,
639 (m–p) September 2017–November 2017, (q–t) December 2017–February 2018, and (u–x) March
640 2018–May 2018.

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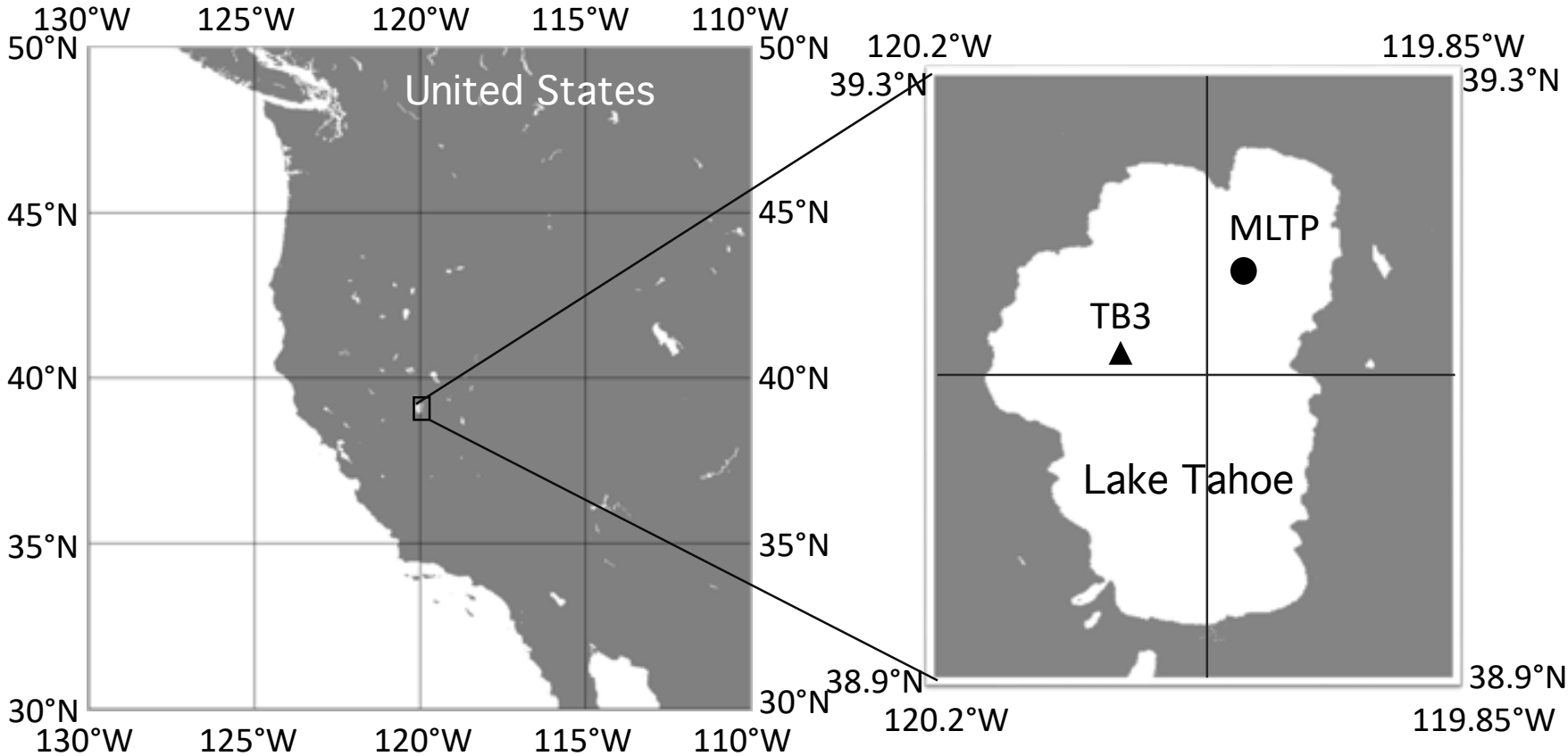
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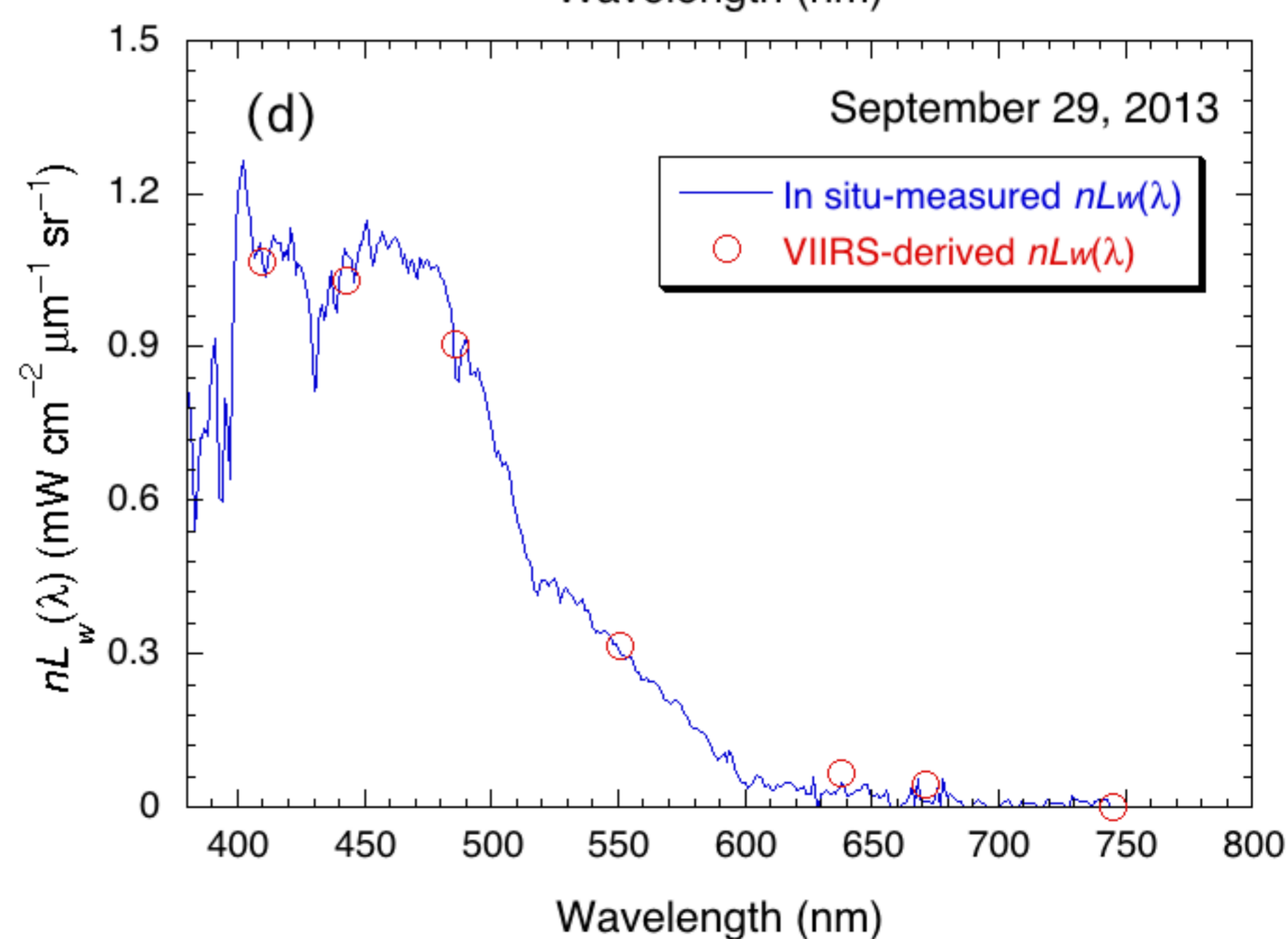
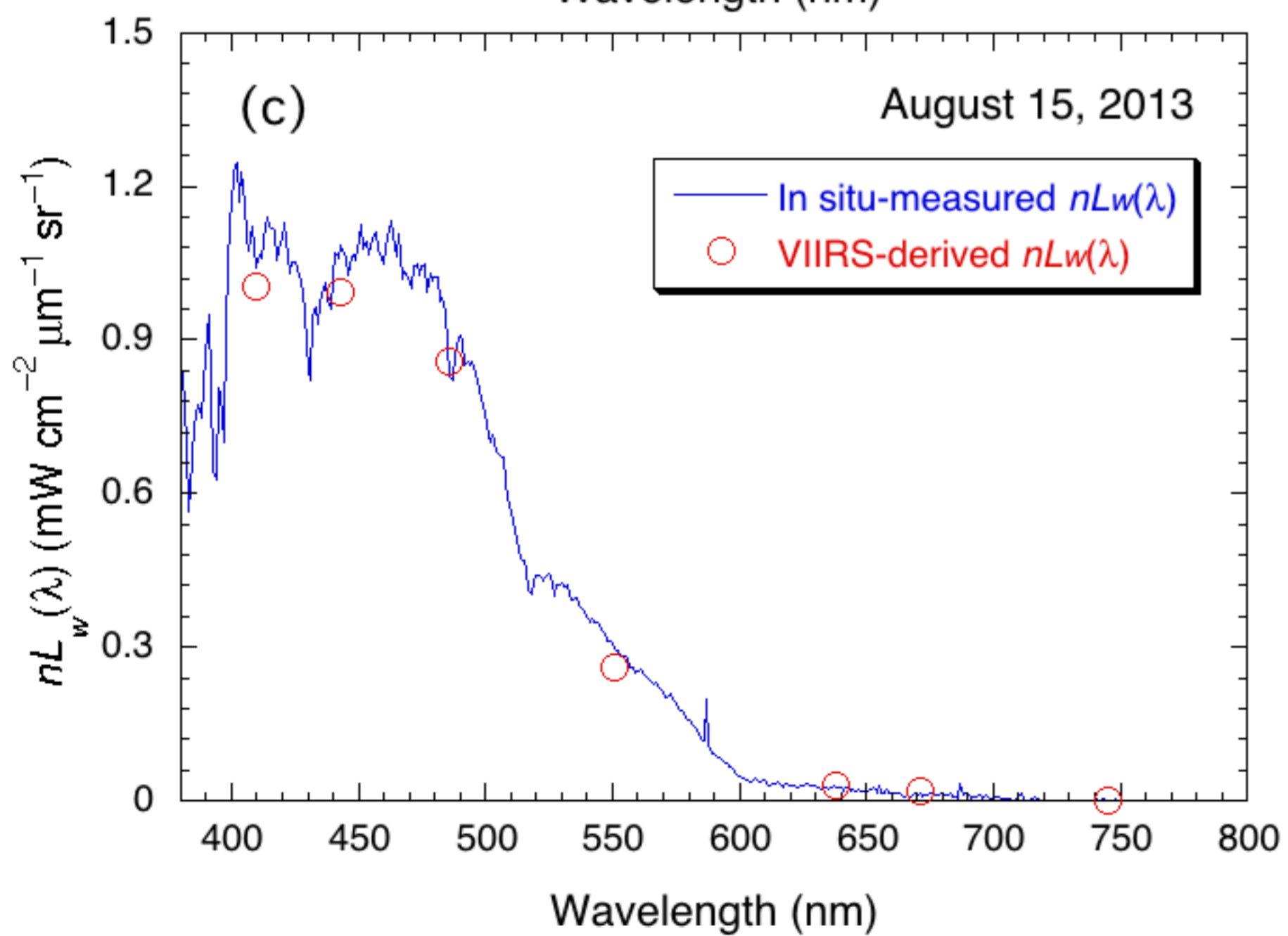
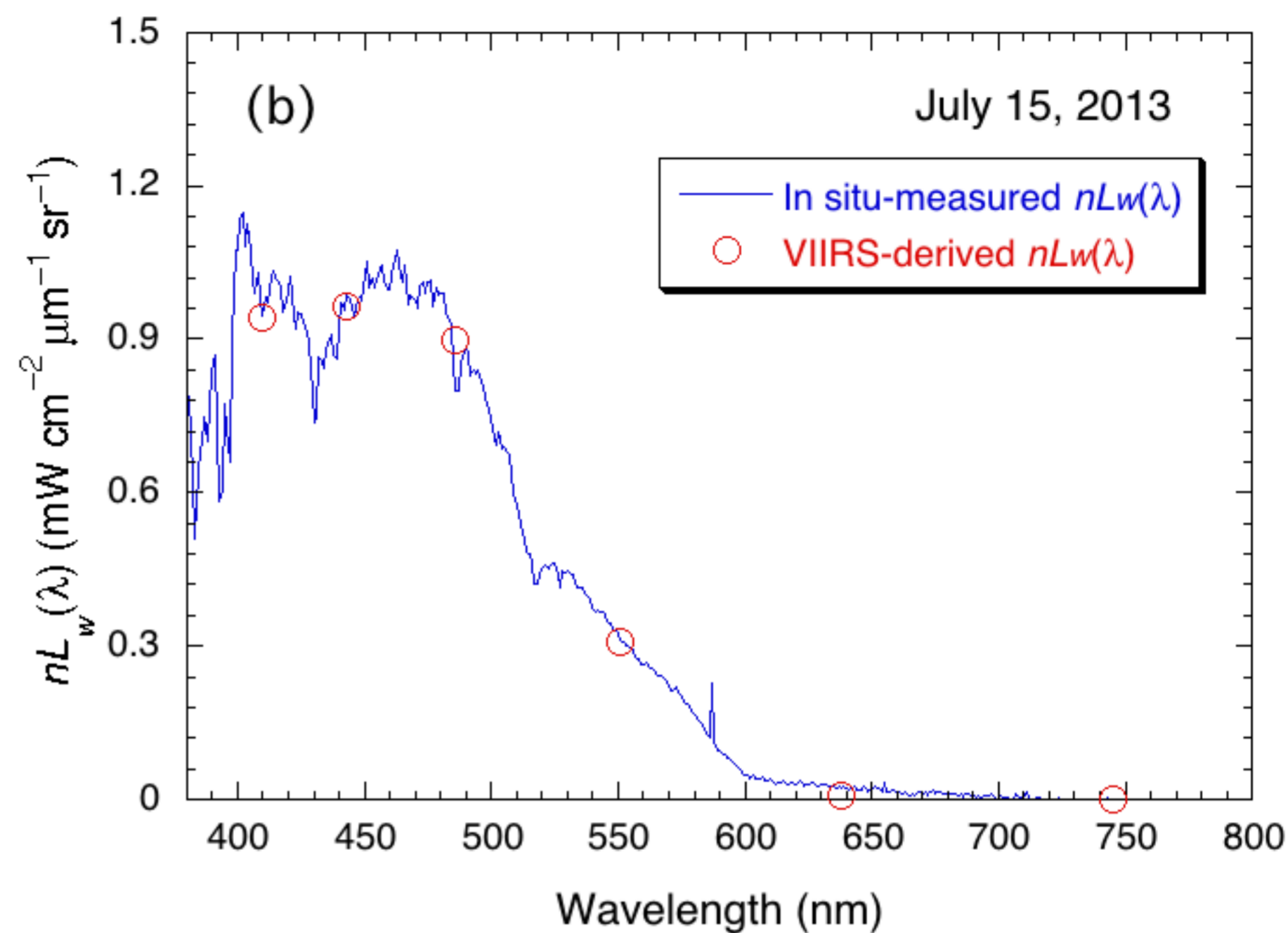
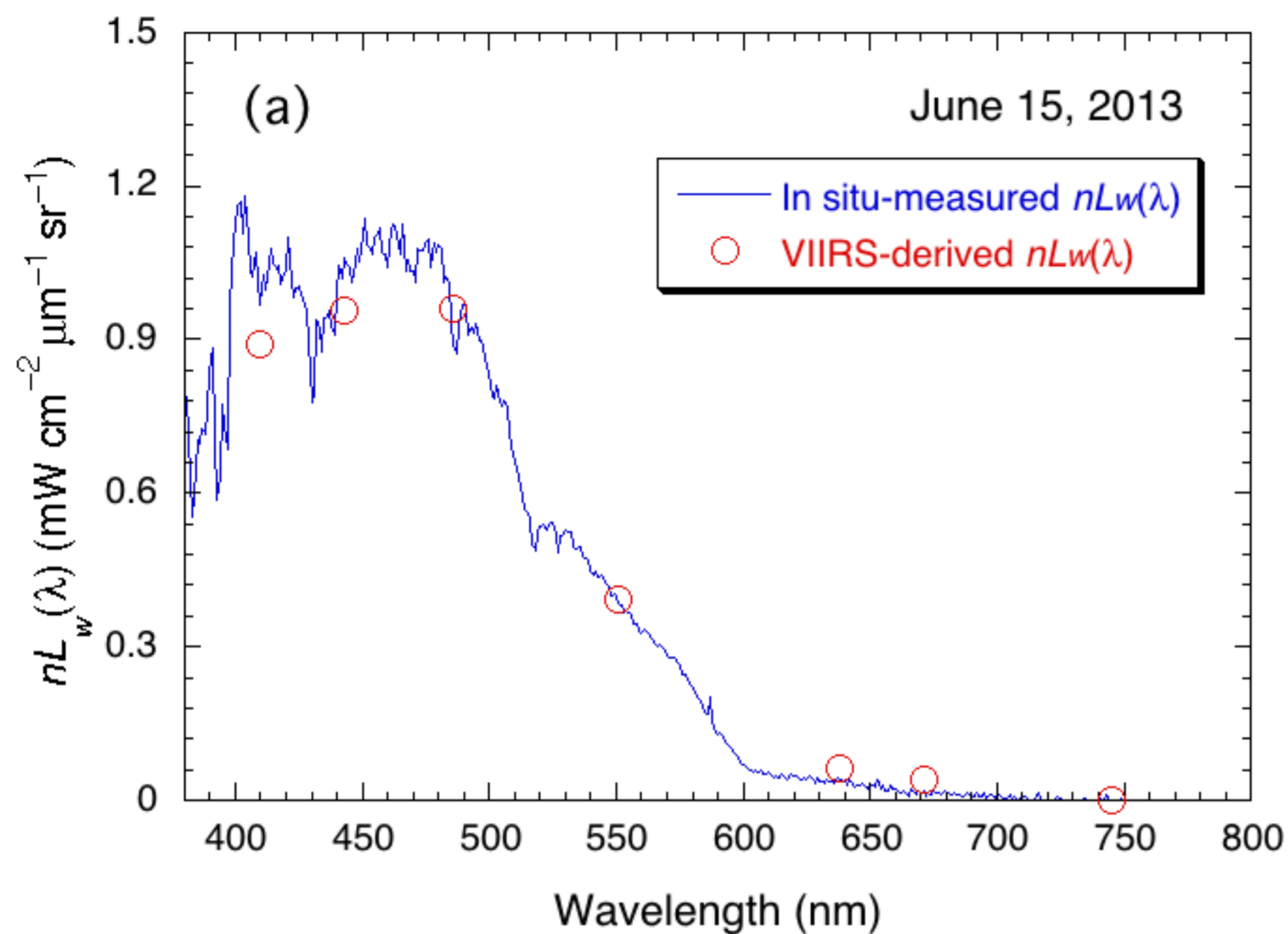
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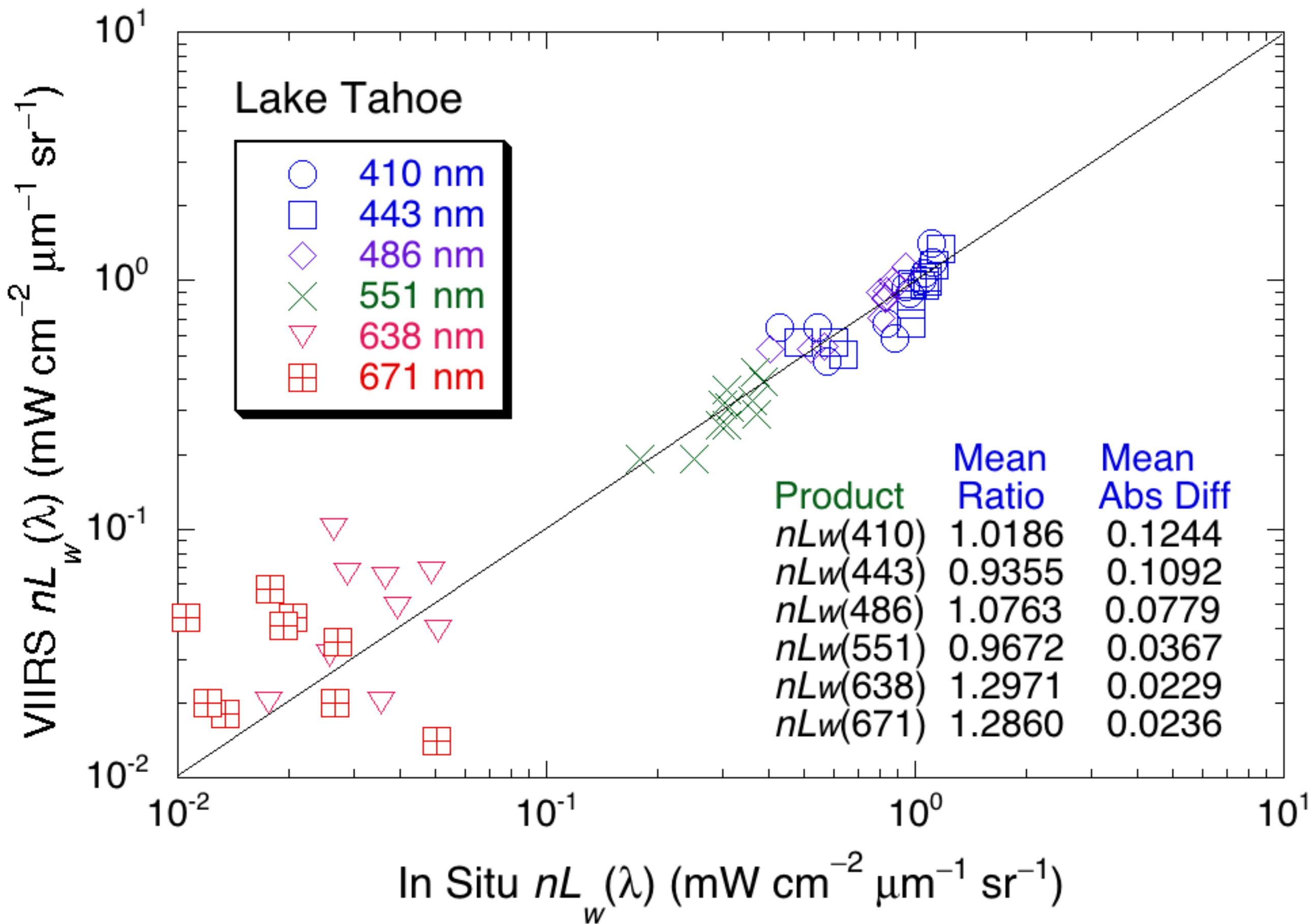
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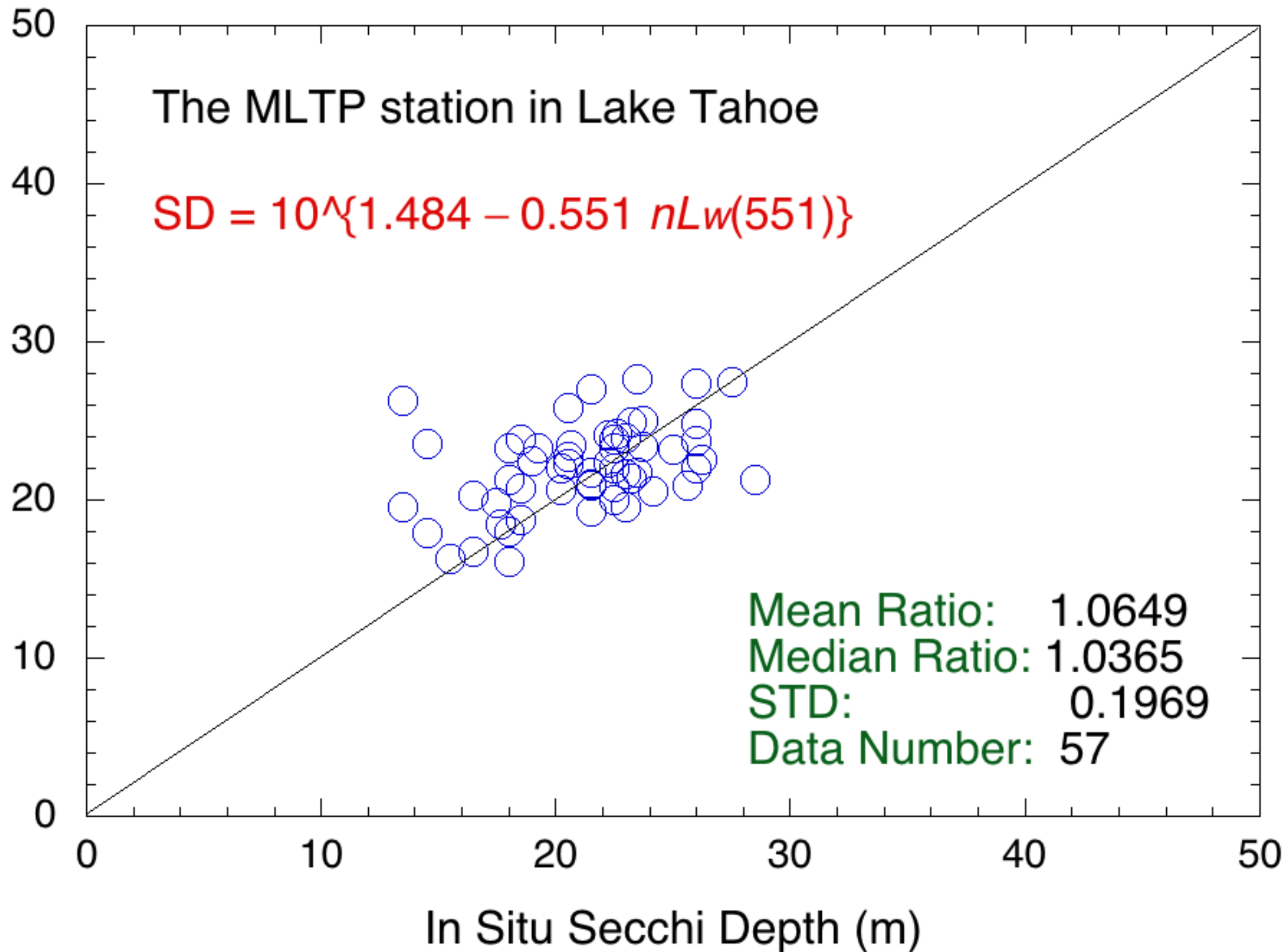


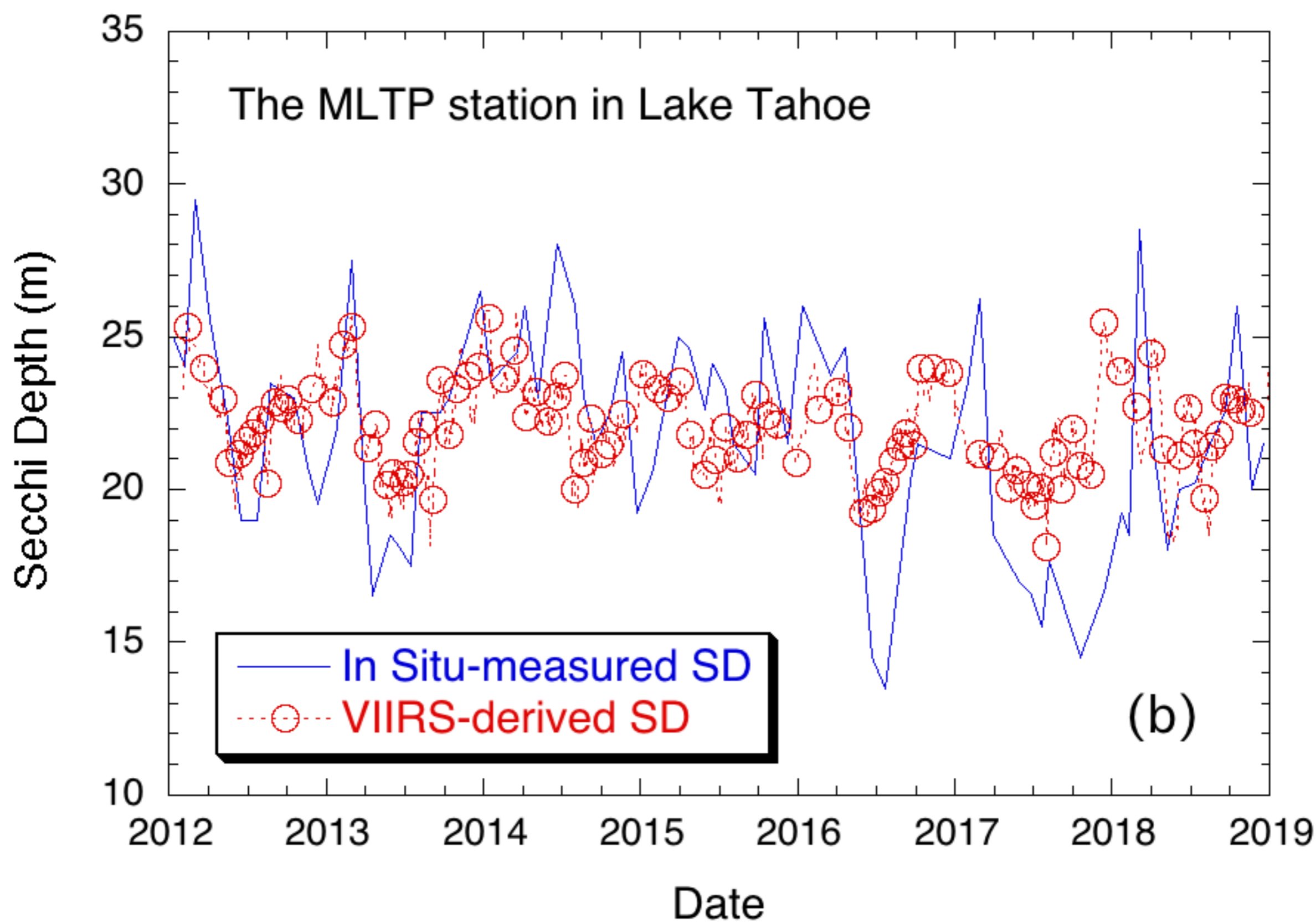
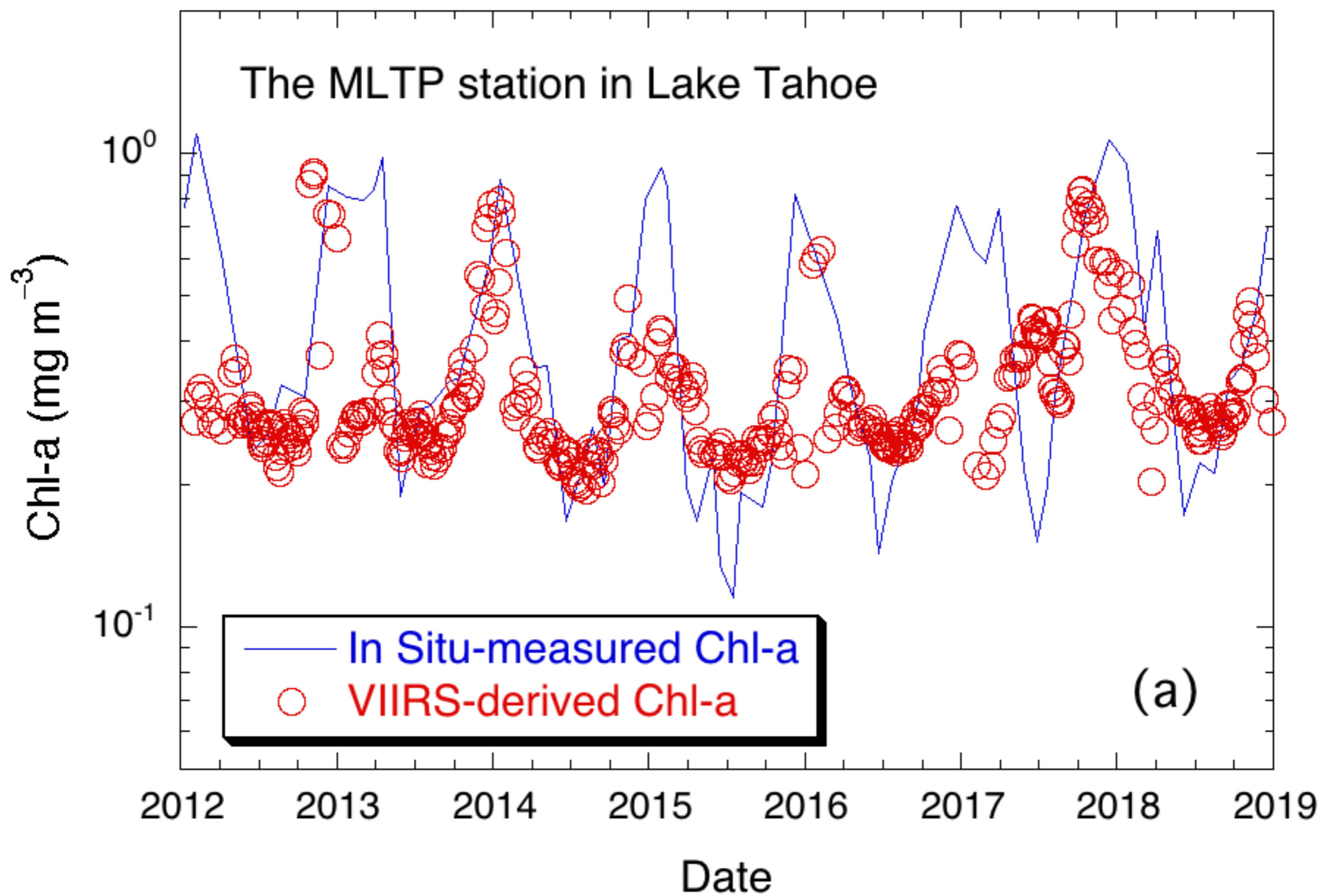


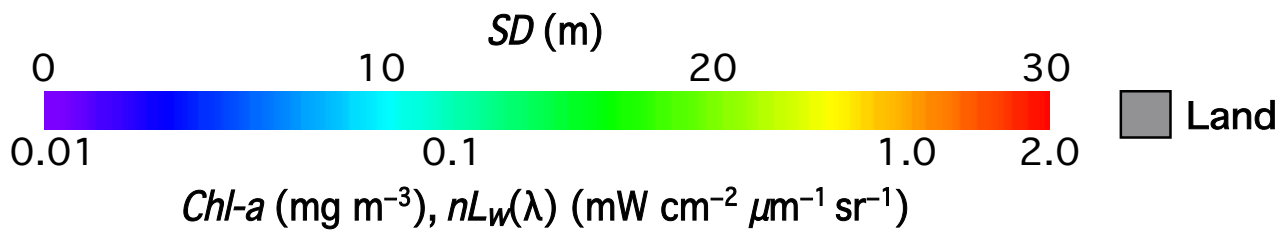
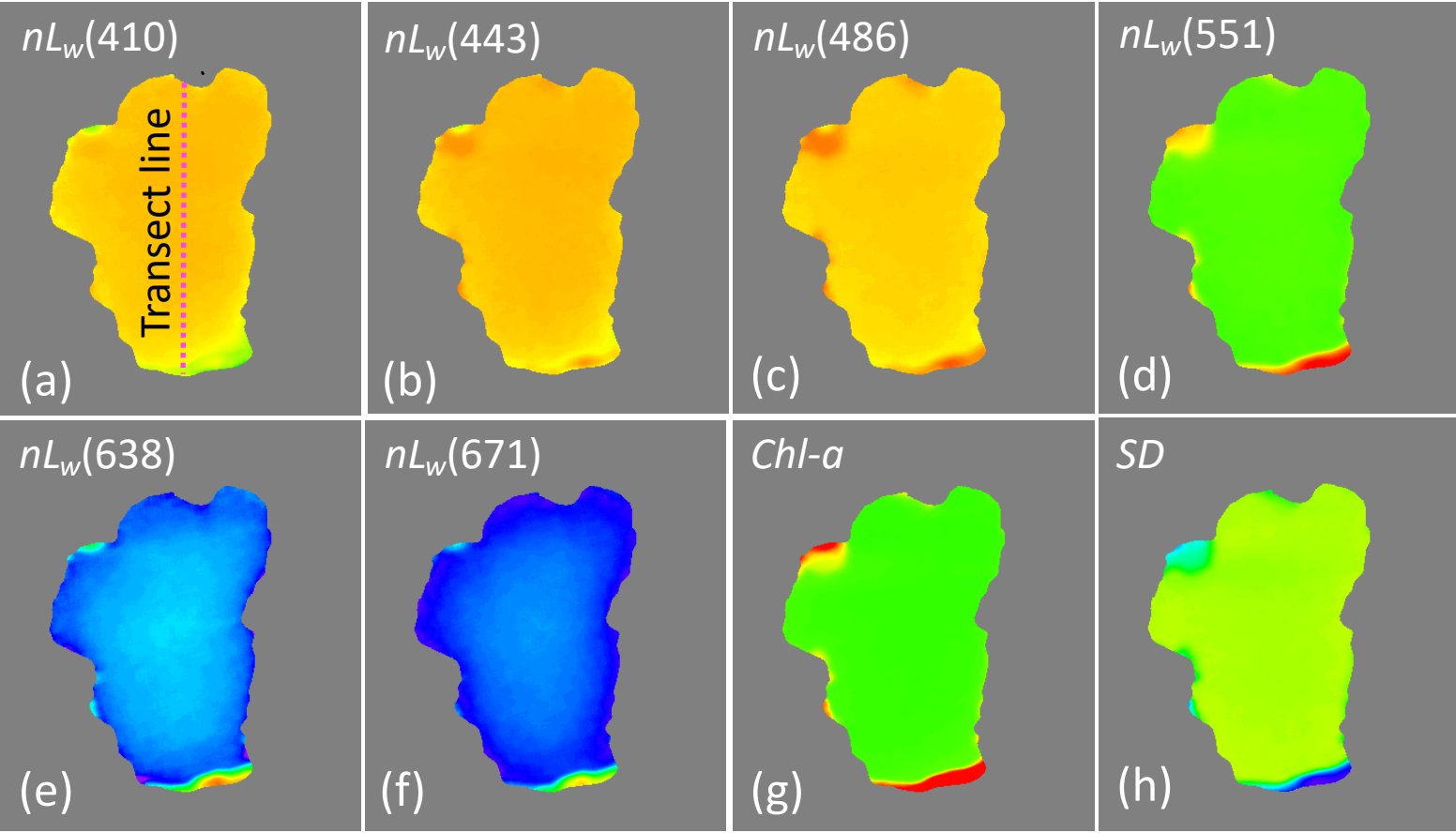
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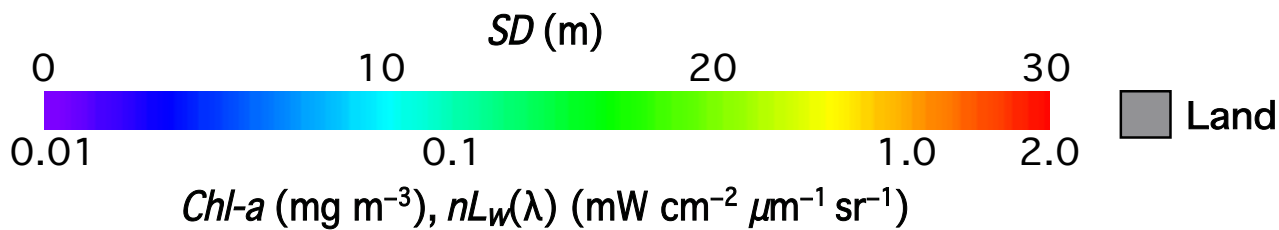
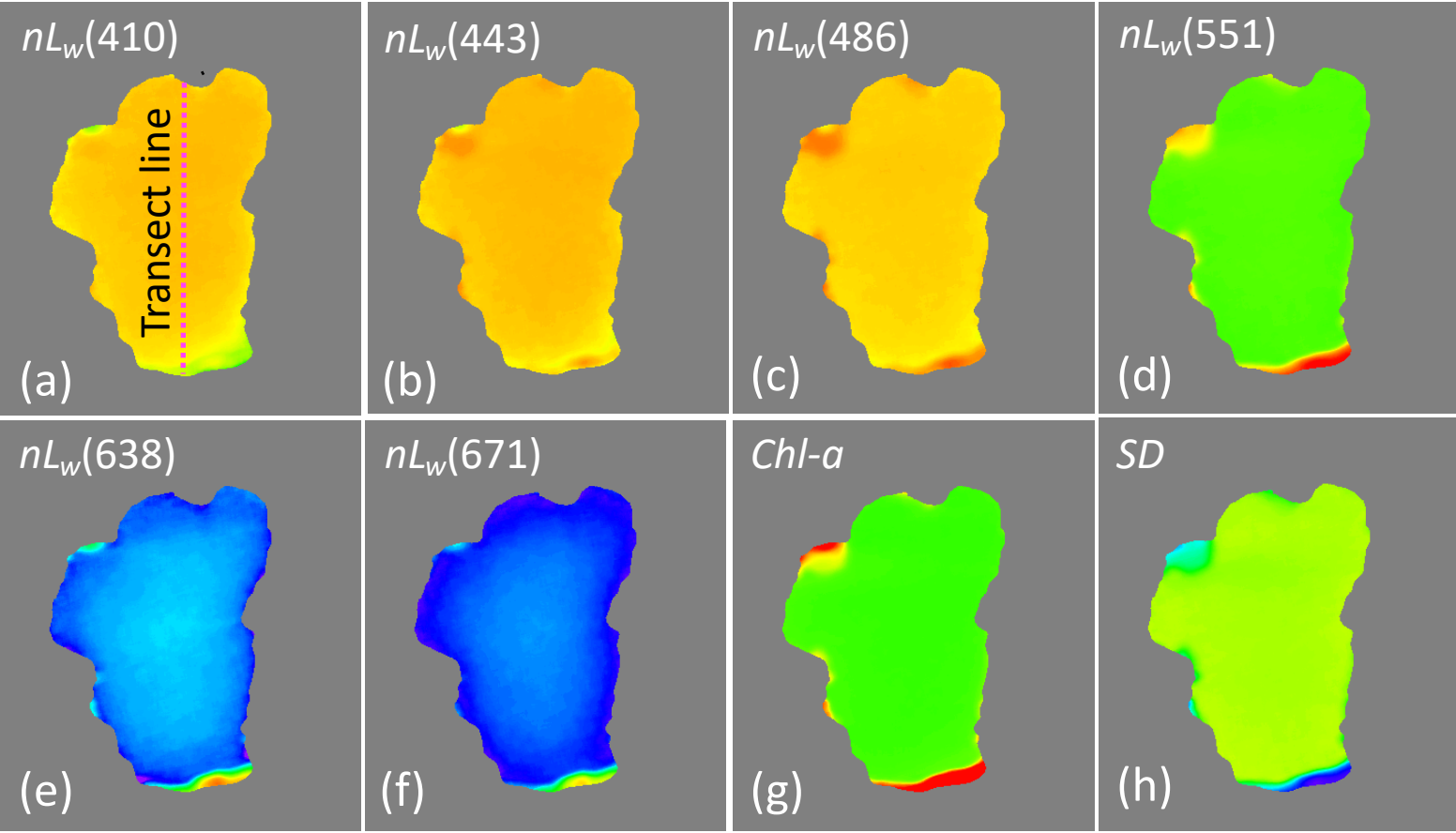
The MLTP station in Lake Tahoe

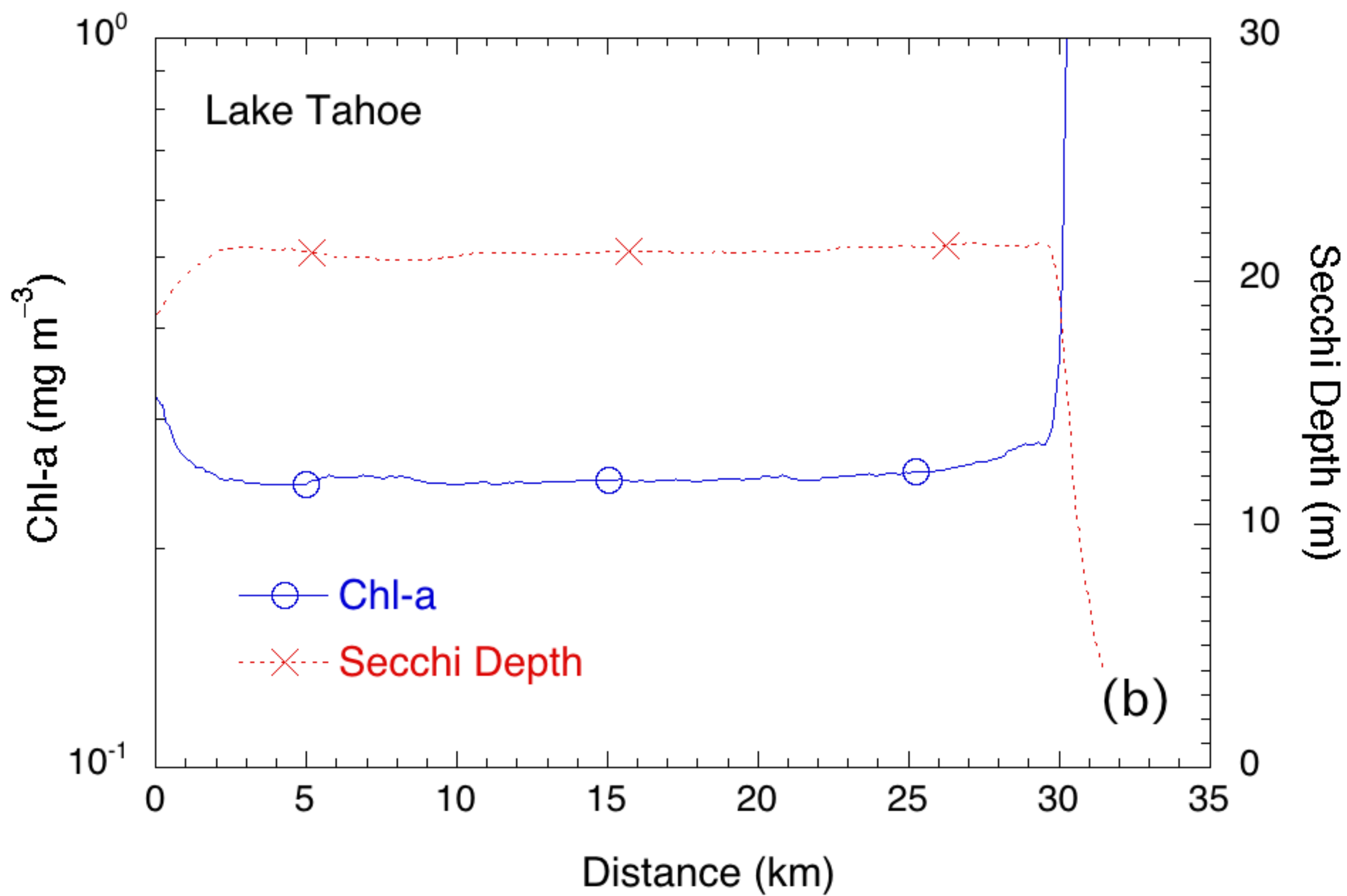
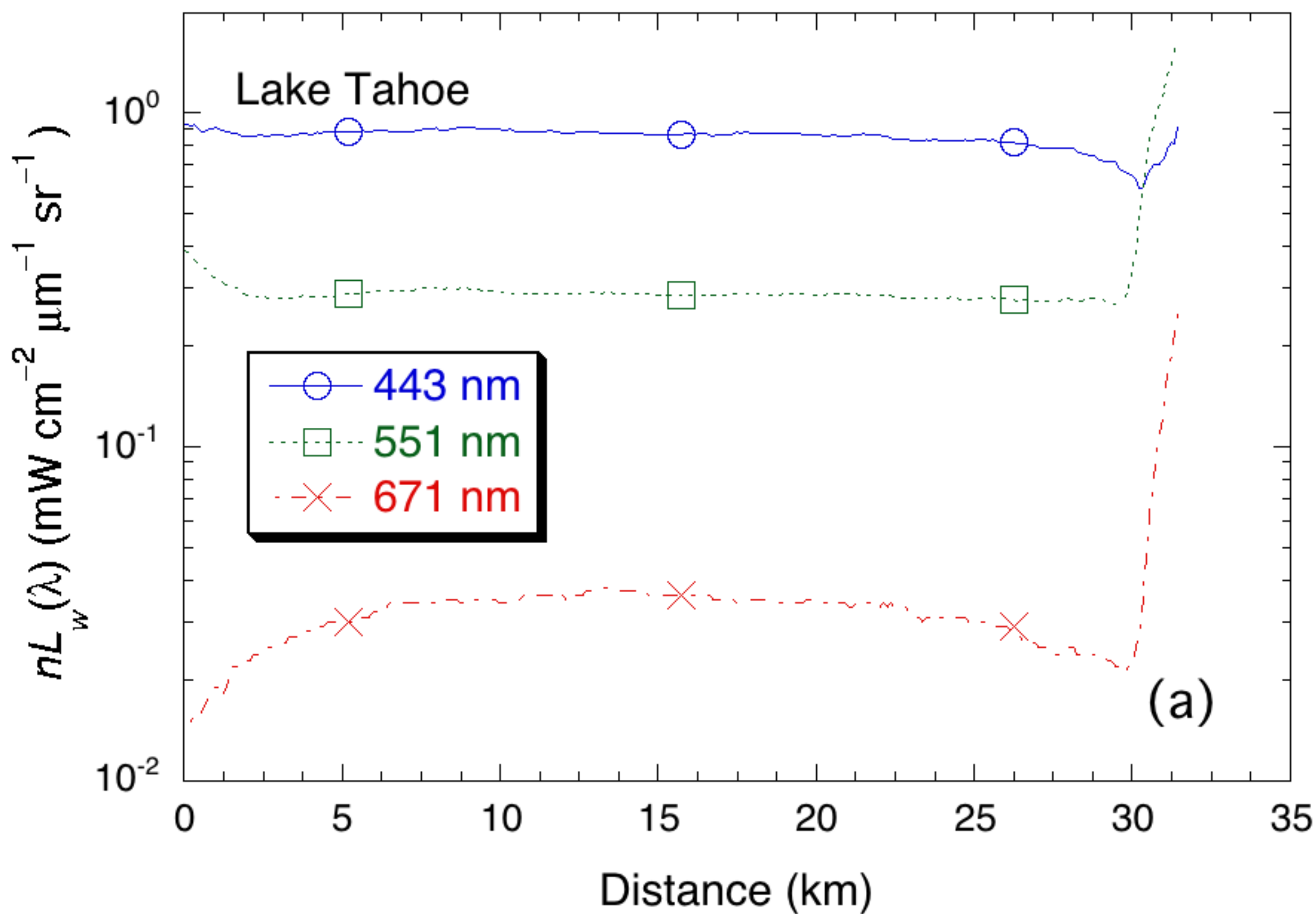
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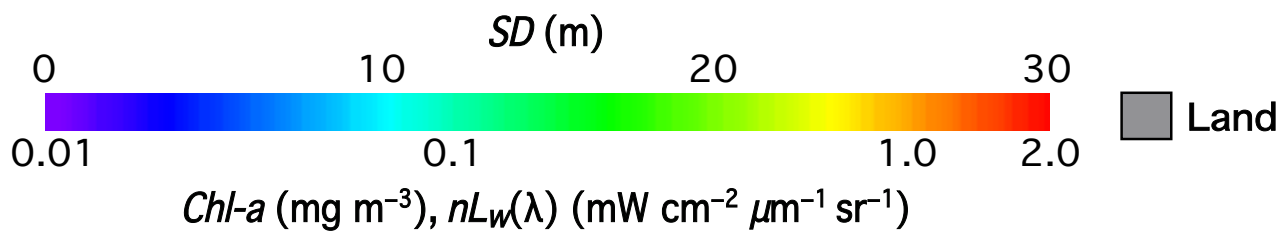
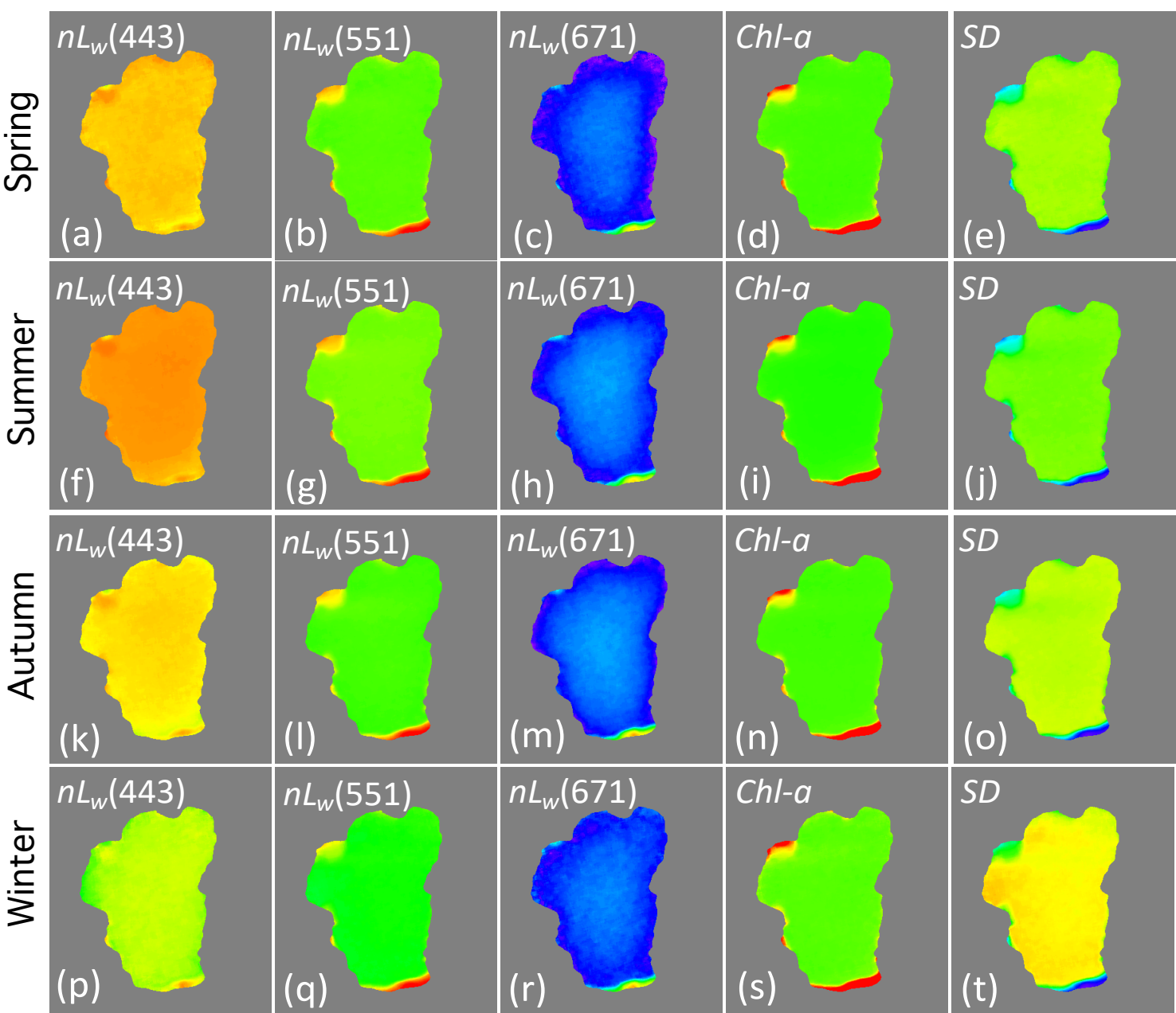


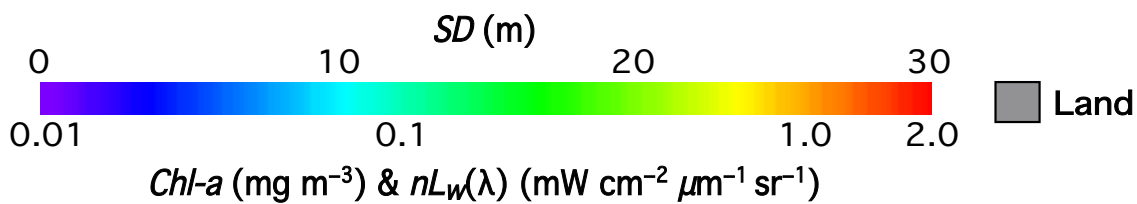
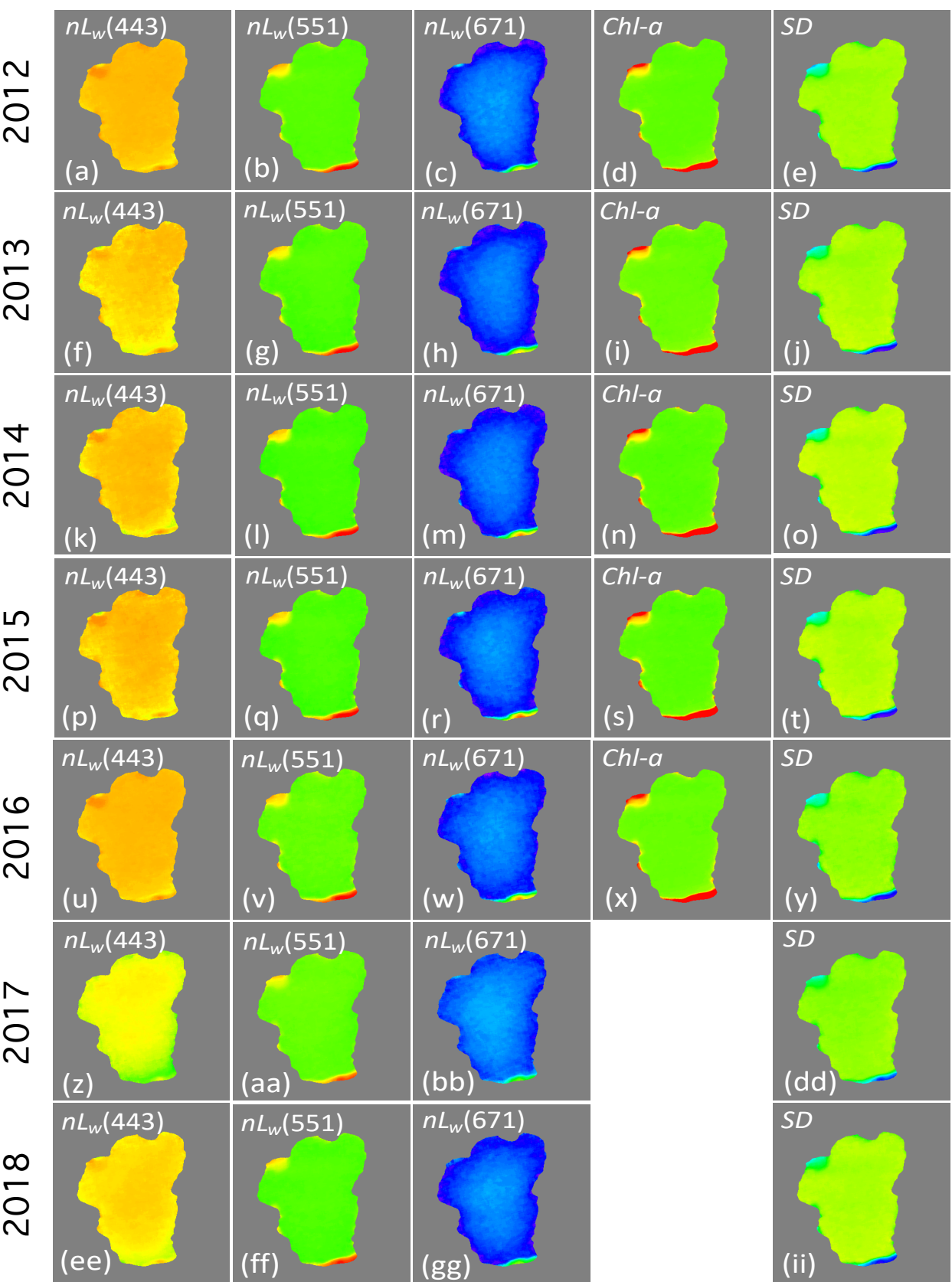


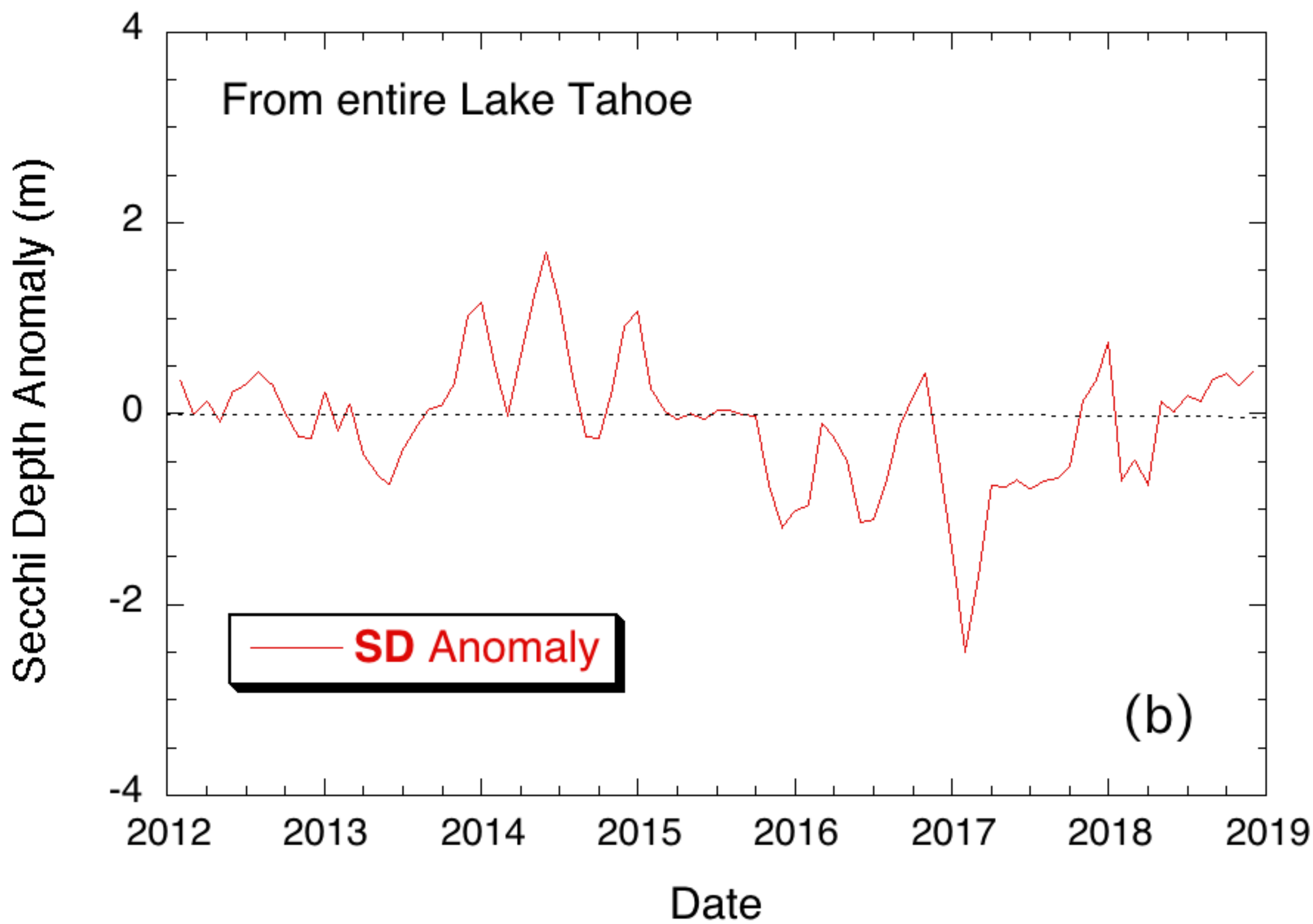
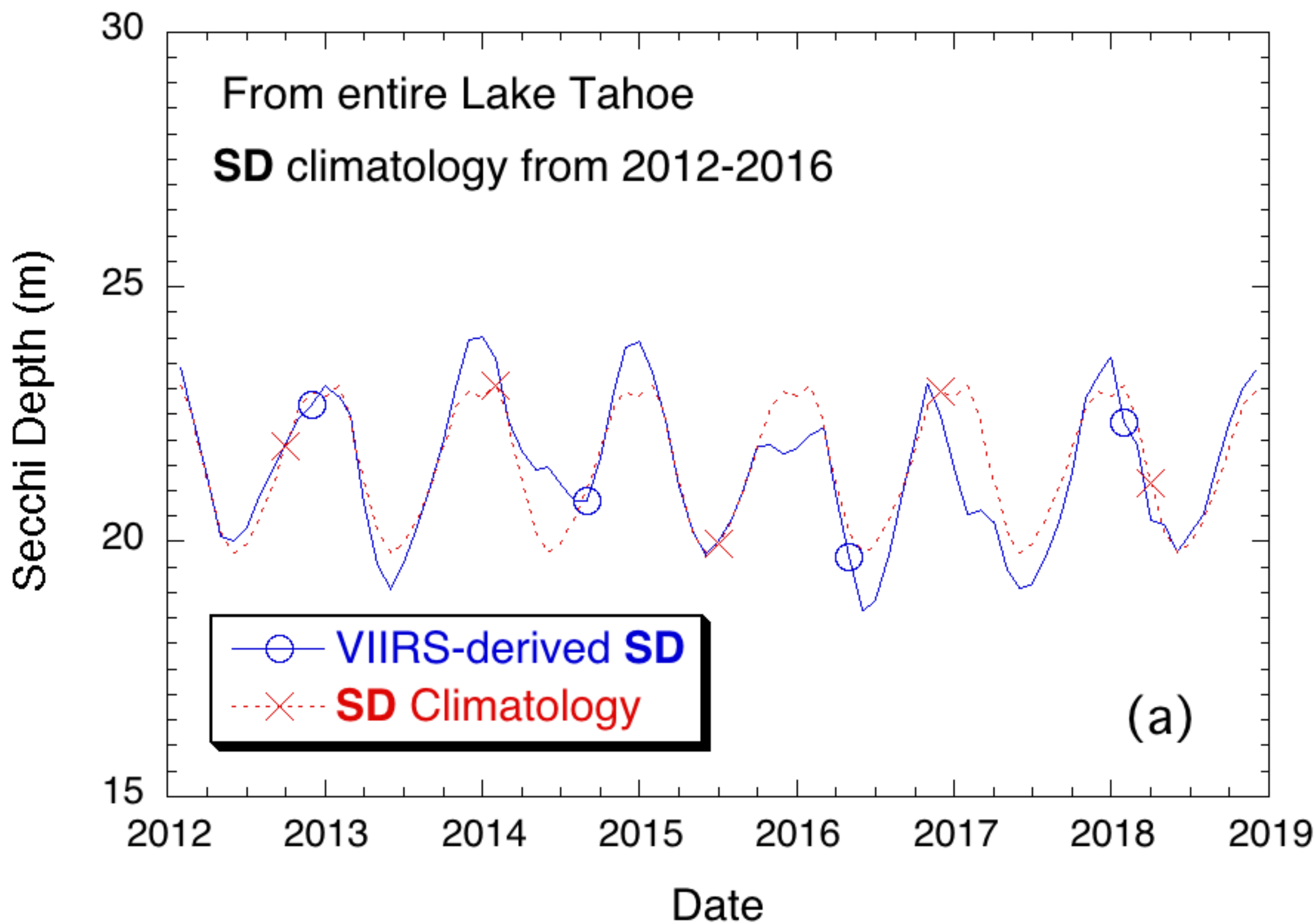


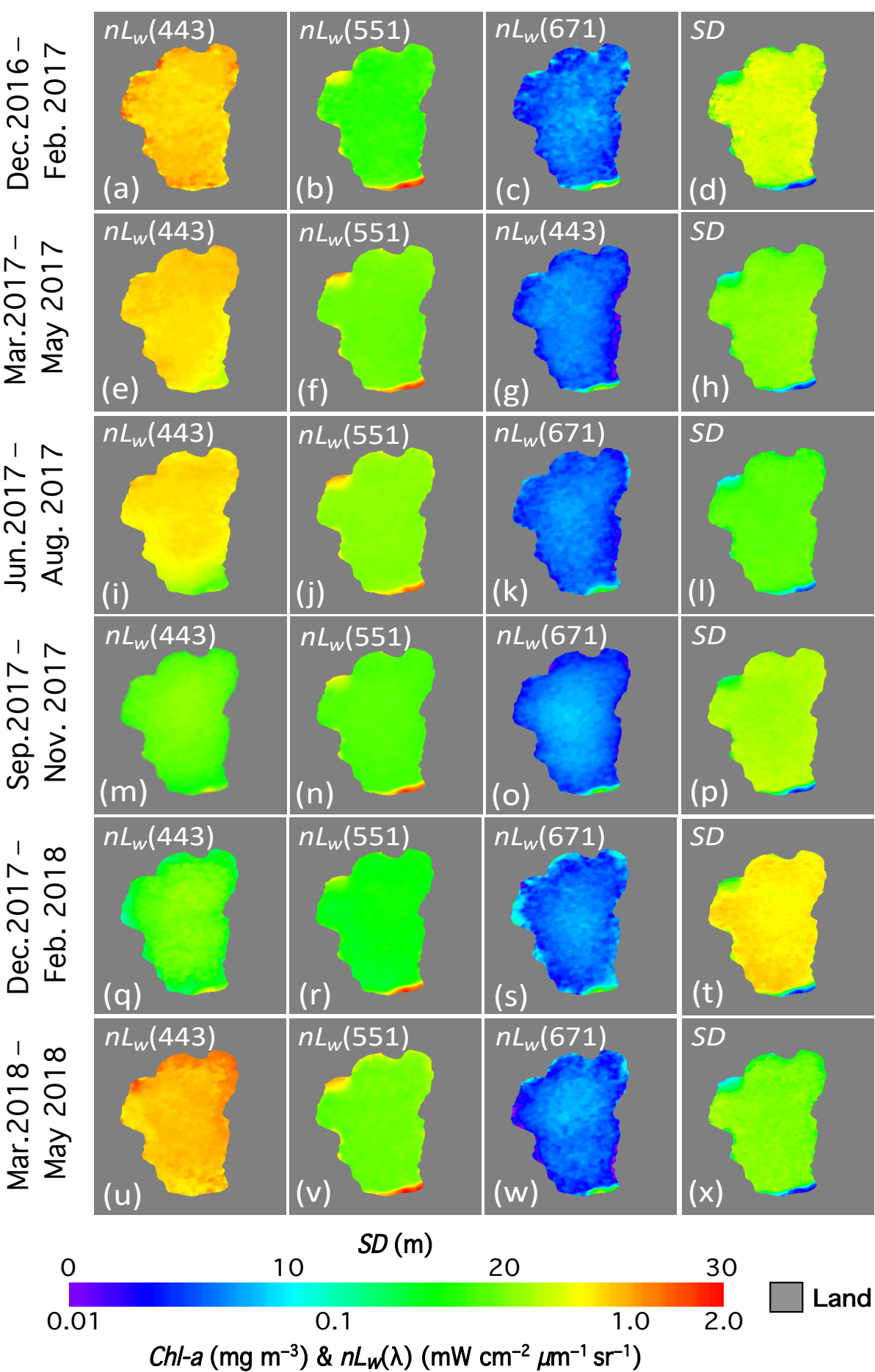












Graphical Abstract

The Visible Infrared Imaging Radiometer Suite (VIIRS)-measured climatology (2012–2016) water property data (the normalized water-leaving radiance spectra $nL_w(\lambda)$, chlorophyll-a (Chl-a) concentration, and water Secchi depth (SD)) in a high-altitude Lake Tahoe in US for (a) $nL_w(410)$, (b) $nL_w(443)$, (c) $nL_w(486)$, (d) $nL_w(551)$, (e) $nL_w(638)$, (f) $nL_w(671)$, (g) Chl-a, and (h) SD. The transection line in panel (a) from the north to south at 120.02°W in the lake is marked for further data analysis to show the spatial distribution of the lake water properties.