

1 **Estimating dredge catch efficiencies for the northern quahog (*Mercenaria***
2 ***mercenaria*) population of Narragansett Bay**

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26 **ABSTRACT**

27 The catch efficiency of a hydraulic dredge was tested on a population of the northern quahog
28 (*Mercenaria mercenaria*) in Narragansett Bay, RI, USA to understand gear limitations and
29 correct relative abundance time series data. In 2017 and 2018, 45 hydraulic dredge tows were
30 conducted following a longstanding fisheries-independent survey protocol, with the dredge
31 transects inspected on SCUBA to assess dredge catch efficiency. Bull raking and quadrat
32 samples taken on SCUBA were also conducted alongside the transects to compare sampling
33 methods. Average dredge catch efficiency across samples was 0.64 (± 0.29 standard deviation).
34 Bottom type was the most significant determinant of dredge catch efficiency, with higher catch
35 efficiency on hard bottom (0.73 efficiency) than on soft bottom (0.48 efficiency). The quadrat
36 and bull rake samples reflected higher catch rates than the dredge, but relationships between
37 relative abundance estimates from the alternate methods and the dredge were either weak or
38 insignificant. Bottom type, sediment classification, depth, and observed abundance were used to
39 model dredge catch efficiencies and predict fisheries-independent abundance indices to more
40 accurate estimates. Applying corrections using a generalized linear model scaled abundances
41 through time, with trends generally the same between time series both with and without the
42 corrections applied. This work provides an example of addressing gear efficiency concerns
43 through diverse collaborations and improving the science and management for a commercially
44 and recreationally significant marine resource.

45
46 **Short Title:** Dredge catch efficiency for northern quahog

47 **Key Words:** northern quahog, hydraulic dredge, catch efficiency, bottom type

48

49 **INTRODUCTION**

50 Fisheries-independent surveys are conducted to infer marine species population trends.
51 Such information is used as input data for stock assessment models and to inform fisheries
52 management regulations (Pennington and Stromme 1998, Rotherham et al. 2007). Within
53 statistical stock assessment models, fisheries-independent relative abundance indices are used to
54 estimate population sizes (Hilborn and Walters 1992). In its simplest form, the relationship
55 between a fisheries-independent survey abundance index (I) in a given year (t) and population
56 size (N) for a species is proportional, with a scaling factor incorporating time-invariant
57 catchability considerations (q) of the survey (Quinn and Deriso 1999, Hilborn 2001):

58
$$N_t = \frac{I_t}{q}$$

59 The catchability parameter is often estimated within stock assessment models (Miller et al. 2018)
60 yet comprises both the efficiency of the sampling gear and the availability of the population to
61 the survey (Cadrin et al. 2016), making it difficult to discern what the catchability parameter
62 alone represents. Catchability parameters can be further described as:

63
$$q = k \frac{a}{A}$$

64 where k represents fishing gear efficiency, and A the total resource area that is vulnerable to the
65 fishery or survey, a (Cadrin et al. 2016). To minimize the gear efficiency component of
66 catchability, testing gear configurations and sampling strategies to identify the most suitable
67 fishing gear to sample the target species is imperative (Rotherham et al. 2007).

68 The gear efficiency is often assessed using a paired or multi-gear experimental design
69 (Cadigan and Dowden 2010, Miller 2013). This approach utilizes multiple gears surveying
70 within nearly the same time and space, allowing other spatio-temporal confounding factors on
71 catch efficiency estimates to remain constant. In the fisheries-independent survey context, the

72 traditional sampling gear is typically one of the survey methods tested to understand the
73 uncertainty in survey trends used for science and management. Quantified catch efficiency data
74 can then be used to address gear efficiency concerns through correcting relative abundance
75 estimates *a priori* to stock assessment modeling.

76 Hydraulic dredges are a common fisheries-independent sampling gear used to monitor
77 soft bottom marine shellfish populations (Myer et al. 1981, Ganz et al. 1999, Thorarinsdottir et
78 al. 2010). Dredge configuration, water depth and sediment type have been found to influence
79 catch efficiency and can lead to inaccuracies in estimating local relative abundances
80 (Thorarinsdottir et al. 2010). Visual inspection of the dredge tracks with cameras or SCUBA are
81 optimal methods for estimating dredge catch efficiencies (Meyer et al. 1981) but can be
82 challenging to implement depending on the depth of sampling and turbidity caused by the
83 hydraulic dredge. Other gear types are also routinely used for assessing shellfish populations,
84 including diver or snorkel-based transects (Arnold et al. 1998), drop cameras (Bethoney et al.
85 2019), and hydraulic paten tongs (Southworth et al. 2010); similarly, these gears can also have
86 efficiency or selectivity concerns based on the species and gear.

87 The northern quahog (*Mercenaria mercenaria*) in Narragansett Bay, Rhode Island, USA
88 serves important roles in the ecosystem, as well as the economy and cultural heritage of Rhode
89 Island (Doering et al. 1986, Schuman 2015, McManus et al. 2019). In 1993, the Rhode Island
90 Department of Environmental Management Division of Marine Fisheries (RIDEM DMF) began
91 a hydraulic dredge survey to assess northern quahog population changes over time and inform
92 management measures. Proper estimates of gear efficiency for the dredge are paramount for the
93 fishery management decisions. A hydraulic dredge was chosen based of the sediment features of
94 Narragansett Bay and the sessile nature of the species; however, concerns have been raised,

95 particularly by the commercial quahog fishing industry, regarding the catch efficiency of the
96 dredge and the accuracy of derived relative abundance estimates used for fisheries management.

97 This study implemented multiple methods to quantify the relative abundance of quahogs
98 and to estimate the dredge catch efficiency. Methods used include hydraulic dredge sampling
99 with diver-based assessment of quahogs missed during dredge tows, diver-based quadrat
100 sampling, and commercial bull rake sampling. The objectives of this work were to (1) estimate
101 catch efficiency of the hydraulic dredge by market class of quahogs and environmental
102 covariates, (2) understand the relative catch efficiency between the various sampling methods,
103 and (3) compare traditional fisheries-independent relative abundance indices from the dredge to
104 those corrected based on the catch efficiency estimates. This work aimed to improve estimates of
105 quahog populations and fisheries management through collaboration with the quahog fishing
106 industry.

107 **METHODS**

108 *Hydraulic Quahog Dredge Survey*

109 The hydraulic dredge is equipped with 0.45-m (18-inch) width and 2.5-cm (1-inch)
110 spacings to target legal sized (\geq 1-inch hinge width) quahogs. The basal and rear ends of the
111 dredge have 1-inch bar spacing, with side and top panels containing 1-inch² wired mesh.
112 Sublegal quahogs are sporadically caught in the survey, but their sizes are not considered to be
113 selected by the gear. During sampling, the dredge is towed at a speed $<$ one knot over a target
114 transect length of 30.5-m (100-ft) for a swept area of approximately 13.9-m². All quahogs are
115 counted and measured (hinge width), and bycatch species identified and enumerated. In addition
116 to dredge catch and effort information, the day, year, position (latitude, longitude), bottom type
117 (hard/soft), depth, and finer sediment type classification are also recorded. The dredge blade

118 angle is adjustable and changing it could improve the efficiency depending on sediment;
119 however, it has been held constant at 15° across sites and years to maintain consistency in the
120 sampling.

121 The quahog dredge survey is conducted to monitor the quahog population of Narragansett
122 Bay using a fixed-rotational sampling design. The survey area is comprised of several strata,
123 each including individual stations. The strata were constructed to describe areas useful in
124 addressing several topics regarding quahogs, including their biology and life history, population
125 units and historical abundance gradients, spatial structure of the commercial quahog fishery, and
126 water quality considerations. The sampling frequency of each stratum has varied through time;
127 since 2010, a given stratum is targeted for sampling at minimum biannually, with all stations
128 within that stratum to be sampled.

129 *Multi-Gear Sampling*

130 Concurrently with hydraulic dredge tows, three additional sampling methods were used
131 to evaluate the sampling ability of the dredge. The first method estimated the dredge catch
132 efficiency by deploying divers on SCUBA to follow the towed dredge paths and collect quahogs
133 missed by the dredge laying in the dredge scars. Quahogs seen during the transect inspections
134 outside of the dredge scar were not considered to be missed by the dredge. Diver-based sampling
135 with quadrats and commercial bull raking were also used to compare quahog sampling methods
136 across varying gear types. Quadrat samples often covers a smaller swath of area than other
137 sampling methods but often rely on fewer gear efficiency issues. Up to three 1-m² quadrat
138 samples were collected at each dredge station via SCUBA, with their specific location adjacent
139 to the track chosen randomly. Divers manually sifted and dug through the upper six inches of
140 sediment within each quadrat and collected quahogs by hand. Bull rake pulls were taken

141 alongside the dredge tracks to compare a traditional scientific sampling method to that used by
142 the commercial fishery. The bull rake has been the preferred method of commercial quahog
143 digging after commercial harvest with dredges was prohibited in Narragansett Bay in 1956
144 (Mackenzie 1997). All bull rakes are comprised of a metal basket (or purse) with one open side
145 lined with protruding metal teeth (spaced at slightly less than 1 inch) used to dig into the
146 substrate. The basket is connected to a long, often telescopic pole (termed the stale) with handles
147 on the opposite end which the fishers hold and “rake” the benthos from the surface. Raking is
148 done onboard a fishing vessel which is typically freely drifting with the wind and currents.
149 Commercial fishers raked their bull rakes up to two times at a given station to provide insight on
150 catch variability within a site using this method. Bull rake pulls were conducted over typical rake
151 durations conducted by the fishery until the basket felt full, at which point the fishers halted
152 digging. Area swept for the bull rake tows were calculated using the width of the rake basket and
153 the distance towed, as measured using start and end points from handheld GPS or tablet systems.
154 The commercial diggers changed the specifications of their rake (e.g. basket width, weights, stale
155 length, and tooth length) at each site, based on their best judgement to optimize the efficiency of
156 the raking. Thus, with rake specification changes between sites, each bull rake tow was treated as
157 the most efficient sampling version of the bull rake. Quahogs sampled using each method were
158 enumerated by market size using their corresponding size classes: sublegal (< 25.4-mm width at
159 the hinge), littlenecks (25.4 to 34.9mm), topnecks (35 to 39.9 mm), cherrystones (40 to 43.9mm)
160 and chowders (\geq 44mm).

161 The multi-gear sampling was conducted at 45 stations throughout Narragansett Bay
162 over 2017 and 2018 (Figure 1). The stations were selected to cover a range of sediment types,
163 depths, local abundances of quahogs, and fisheries and water quality management areas. Each

164 sampling station had dredge tows conducted with corresponding dredge transect inspections and
165 quadrat sample collections using SCUBA. Bull rake sampling was done opportunistically, with
166 only some of the sampling dates and stations having pairwise bull rake data available (Table 1).

167 *Hydraulic Dredge Catch Efficiency*

168 Dredge catch efficiency was evaluated by comparing relative abundance estimates with
169 and without quahogs missed by the dredge included. Observed dredge relative abundance
170 estimates were corrected by summing the count of quahogs found lying in the dredge transect
171 during the inspection on SCUBA (N_T) to the number retained in the dredge (N_D):

172 Observed Dredge Relative Abundance = $\frac{N_D}{\text{Dredge Swept Area (m}^2\text{)}}$

173 Corrected Dredge Relative Abundance = $\frac{N_D + N_T}{\text{Dredge Swept Area (m}^2\text{)}}$

174 Corrected estimates were then compared to relative abundances obtained from the quadrat and
175 bull rake samples using linear regression and Wilcoxon Rank Sum Tests. Dredge catch
176 efficiency was estimated as the ratio of the observed dredge relative abundance to the corrected
177 estimate, ranging from 0 to 1:

178 Dredge Catch Efficiency = $\frac{\text{Observed Dredge Relative Abundance}}{\text{Corrected Dredge Relative Abundance}}$

179 Samples where quahogs were not found in the dredge nor on SCUBA inspection were excluded
180 from catch efficiency analyses. Catch efficiency was compared to covariates that were believed
181 to either influence local quahog abundance and/or influence the fishing ability of the dredge.

182 Significant differences in dredge catch efficiency within categorical variables were tested using
183 Kruskal-Wallis Rank Sum and post-hoc Nemenyi Tests, and continuous variables were tested
184 using linear regression.

185 The influence of these factors on dredge catch efficiency were further examined using
186 generalized linear models (Venables and Dichmont 2004). The models were constructed using R
187 package ‘glmmTMB’ (Brooks et al. 2017) with a beta error distribution. Given beta distributions
188 only apply to values within (and not inclusive of) 0 and 1, dredge efficiencies were transformed
189 to account for these observations using the sample size of the dataset following Smithson and
190 Verkuilen (2006):

$$191 \quad \text{Dredge Efficiency}_{\text{Transformed}} = \frac{\text{Dredge Efficiency}_{\text{Observed}} * (n - 1) + 0.5}{n}$$

192 Covariates assessed in modeling dredge efficiency included bottom type (hard vs. soft
193 bottom), sediment type, depth of sampling, and observed relative abundance. Bottom type,
194 sediment, and depth were included to standardize benthic characteristics that influence quahog
195 abundance (Pratt 1953, Pratt et al. 1992, Rice 1992) and may influence dredge efficiency. The
196 observed relative abundance was included in the modeling to better understand if the dredge
197 efficiency varies depending on the local density of quahogs in the dredge track (Ganz et al.
198 1994). Model variants using different combinations of these covariates were evaluated and
199 compared through minimization of the Akaike information criterion (AIC; Akaike 1973.) The
200 performance of the final model variant selected for use was evaluated by comparing the
201 predicted efficiencies for the samples of this study to those observed.

202 The selected model variant was then used to predict catch efficiencies for historical
203 RIDEM DMF hydraulic dredge survey tows. Samples with missing field data needed as
204 covariates in the model (e.g. bottom type, sediment, depth) could not produce efficiency
205 predictions and were excluded from the analysis. Each observed quahog relative abundance
206 estimate from the fisheries-independent survey through time was then divided by its
207 corresponding predicted efficiency to estimate corrected abundances. Annual average observed

208 and corrected relative abundance indices were produced to compare the differences in the trends.
209 The relative abundance indices were derived by averaging the observed and corrected relative
210 abundances by year and region. Time series were constructed for the spatial regions used in the
211 RIDEM DMF quahog stock assessment model: Greenwich Bay, Providence River, and
212 Narragansett Bay proper (Figure 1). The abundance time series generated represented relative
213 indices from the survey data; these indices did not represent the total population sizes nor swept
214 area abundances for spatial domains.

215 **RESULTS**

216 *Dredge Catch Efficiency and Comparison of Sampling Methods*

217 Of the 45 dredge tows with diver-based transect inspections, three tows had zero quahogs
218 caught in the dredge or found in the dredge scar on SCUBA. Of the 42 tows with positive
219 catches in either the dredge or dredge scar, the average quahog catch efficiency was 0.61 (0.28
220 standard deviation). Dredge efficiency did not vary statistically across market size classes
221 (Figure 2). The dredge exhibited a significantly greater catch efficiency for total quahog catch on
222 hard bottoms than soft bottoms (Figure 3; Kruskal-Wallis $\chi^2 = 7.95$, p-value=0.02), but not for
223 sediments where the bottom type was indiscernible or likely a mix. Dredge catch efficiency
224 varied greatly across and within sediment types, with mean efficiencies by sediment type ranging
225 from 0.44 to 1 (Figure 3, Table 2.) Generally, mud sediments had lower catch efficiency than
226 sand, but catch efficiencies by individual sediment type were not statistically different. Dredge
227 catch efficiency did not indicate a significant relationship with either depth or observed relative
228 abundance; catch efficiency was highly variable over depth and relative abundance, particularly
229 at the lower spectrums of these variables (Figure 3).

230 On average, chowders represented the most abundant size class caught in the dredge,
231 quadrat, and bull rake samples (Figure 4). The proportion of sublegal quahogs was often very
232 small ($\leq 5\%$), with those of littlenecks, cherrystones, and topnecks often comparable. Relative
233 abundances observed using the hydraulic dredge and those corrected based on quahogs found in
234 the transect were significantly correlated (Figure 5; $R^2=0.95$, $p\text{-value}<0.001$). Abundance
235 estimates from sampling with a quadrat were higher than corrected estimates from the dredge
236 (Figure 5; Wilcoxon Rank Sum Test $p<0.001$), with a mean abundance of $3.84 (\pm 6.44)$ quahogs
237 m^{-2} from quadrat samples compared to the corresponding dredge observed abundances of $1.91 (\pm$
238 $2.10)$ quahogs m^{-2} . The correlation between these paired samples was significant, but weak
239 ($R^2=0.12$, $p\text{-value}<0.001$). A similar comparison between corrected abundances from dredge and
240 bull rake sampling indicated a similar finding; bull rake abundances were on average greater than
241 those from the dredge, with a mean abundance of $5.60 (\pm 6.81)$ quahogs m^{-2} compared to the
242 corresponding corrected dredge abundances of $1.84 (\pm 2.07)$ quahogs m^{-2} (Figure 5; Wilcoxon
243 Rank Sum Test $p=0.006$). Linear regression indicated there was not a significant relationship
244 between the corrected dredge abundance estimates and bullrake sampling. Average abundance
245 estimates from quadrat sampling (3.26 ± 3.41 quahogs m^{-2}) and bull rake sampling (4.69 ± 7.72
246 quahogs m^{-2}) were not significantly different from each other (Wilcoxon Rank Sum Test > 0.05).

247 *Modeling Dredge Efficiency and Time Series Predictions*

248 When testing all prospective covariates within the models, only bottom type was a
249 significant covariate in predicting dredge catch efficiency, and the model variant using only
250 bottom type had the lowest AIC score (Table 3). For predictive purposes of incorporating
251 variability from both continuous and discrete variables on dredge catch efficiency, the model
252 variant with all covariates was used for dredge efficiency predictions. The correlation between

253 the final model predictions and observed efficiencies was significant (Figure 6; $R^2= 0.46$, p-value
254 < 0.001).

255 With corrections applied to the quahog abundance time series data, average annual
256 relative abundance indices increased for all years except for 2003 in Greenwich Bay (Figure 7.)
257 This instance is likely attributed to a low number of tows with required covariate data for
258 predictions, thus excluding many samples from the corrected annual 2003 abundance estimate.
259 Increases in relative abundances through time and region varied based on the sampling that
260 occurred in the given year but reached up to 1.98 times that of the original annual relative
261 abundance indices (Figure 7). Despite increased relative abundances from the model corrections,
262 the abundance trends for each region were similar when applying the dredge efficiency
263 corrections to samples. The Providence River remained the most abundant region, with
264 Greenwich Bay and Narragansett Bay proper indices still similar in magnitude (Figure 7).

265 **DISCUSSION**

266 Catch efficiency information for a hydraulic dredge is presented for the commercially and
267 recreationally significant northern quahog. The hydraulic dredge exhibited less than optimal
268 catch efficiency, with an average efficiency across the samples of 0.61. The measured dredge
269 efficiency in the present study is aligned with those reported for other fisheries incorporating a
270 hydraulic dredge to monitor infaunal bivalves (Medcof and Caddy 1971, Meyer et al. 1981,
271 Smolowitz and Nulk 1982, Michael et al. 1990, Thorarininsdottir et al. 2010). More specifically,
272 earlier work evaluating the quahog catch efficiency of hydraulic dredge sampling in Narragansett
273 Bay was 0.57 (Ganz et al. 1999).

274 Catch efficiencies were most strongly correlated with bottom type, with hard bottoms
275 having a higher efficiency on average than soft sediments (Table 2). These results indicated there

276 was a stronger difference in catch efficiency between hard and soft bottoms than previously
277 reported by Ganz et al. (1999); however, their work grouped bottom and sediment types together,
278 whereas this research analyzed bottom and sediment types separately. While sediment type did
279 not indicate there was a significant difference in catch efficiency between mud and sand, this
280 difference was likely manifested in the hard and soft bottom classification. Ganz et al. (1999)
281 noted that the hard bottoms in Narragansett Bay usually represent packed sand, and soft bottoms
282 are typically mud. Previous work suggested quahog catch efficiency varies with local quahog
283 density (Ganz et al. 1994), yet this study did not.

284 Dredge catch efficiency estimates by legal market classes of quahogs indicated that the
285 dredge sampled these classes equally well. In the absence of other gear data for comparison,
286 possible hypotheses for low relative abundance of sublegal quahogs could be gear selectivity or
287 small individuals being displaced from the sampling area due to the hydraulics of the dredge.
288 The similar market class catch compositions by gear type and lack of sublegal quahogs found in
289 the dredge scar during transect inspections suggest that the mesh gear selectivity hypothesis may
290 not hold. While the quadrat and bull rake samples had more sublegal quahogs than the dredge,
291 these gears also had fewer sublegal quahogs than older, legal market-sized, individuals. The lack
292 of capture of sublegal quahogs across the gears draws further questions regarding the settlement,
293 growth, mortality, and survey selectivity of the early life stages of quahogs, justifying future
294 research needs.

295 The paired diver sampling with quadrats and bull rake sampling caught more quahogs
296 than the dredge. It is possible that some quahogs were missed during the SCUBA inspections of
297 the dredge scar due to visibility or that quahogs observed outside the dredge scar were displaced
298 by the dredge and thus missed catch. These observer error hypotheses seem unlikely, as the

309 diligent SCUBA inspections likely led to minimal or negligible inspection error, and it is
300 improbable that the quahogs observed outside the dredge scar originated in the dredge path based
301 on their distance from the dredge scars. The disparity between the alternative methods and the
302 corrected dredge relative abundances may be attributed to the differing swept areas of the gear
303 types. Quahog distribution has often been described as super-dispersed or contagious, following
304 a negative binomial distribution (Saila and Gaucher 1996, Russell 1972.) The larger swept area
305 of the dredge (~14-m²) compared to that of the quadrats sampled (1-m² segments) may explain
306 this discrepancy, with the dredge representing a more integrated sample of the local standing
307 stock with a larger swept area. Bull rake swept areas were much more variable, as they were
308 conducted until the rake felt full to the fisher. This approach may lend itself to more efficient
309 sampling and ensure that either a rake head or rear cage of a hydraulic dredge does not overflow
310 before the end of a tow and cause quahog spillage (Meyer et al. 1981). The benefits of using an
311 industry fleet to sample with bull rakes include prospective improved efficiency, ability to
312 sample shallower and deeper depths based on boat outfitting the stale length changes, and
313 including industry members in the scientific process of data collection for management. Current
314 bull rake sampling drawbacks include the gear variability that quahoggers use between sites and
315 loss of standardization (e.g. changes in rake head width, rake tooth length, stale length, weights),
316 and the ability to ensure accurate distance or swept area calculations. Continued comparative
317 work with the hydraulic dredge, quadrats, and bull rakes would improve these inferences, as the
318 sample size for some of these comparisons was small and variability in quadrat and bull rake
319 replicates were large. An alternative solution may be to change the blade angle of the hydraulic
320 dredge by site to increase efficiency and not standardize the blade angle, which has been found
321 to improve dredge efficiency for sampling other shellfish (Meyer et al. 1981).

322 The adjustments in time series abundances from the dredge using this efficiency
323 information have more appropriately quantified northern quahog relative abundances in
324 Narragansett Bay. Improving these estimates aids in addressing longstanding concerns that the
325 hydraulic dredge survey does not accurately depict the local standing stock of legal sized
326 quahogs. While this work has improved relative abundance estimates, the overall trends in the
327 fisheries-independent abundance indices largely remained the same. This finding suggests that,
328 annually, the sample breakdown by environmental factors used to infer dredge efficiency have
329 largely been stable through time. The dredge efficiency corrections will improve estimates of the
330 Narragansett Bay quahog population size in stock assessment models, but may not affect
331 conclusions on population trajectories.

332 This work serves as an example of scientists, managers, and industry members
333 collaboratively addressing questions that improve fisheries science and management. The pair-
334 wise sampling provided dredge efficiency estimates and insight into the accuracy of a
335 longstanding hydraulic dredge survey. Further, other gear types that may be suitable for use in
336 the future were evaluated via supplementary fisheries-independent and industry-based sampling.
337 By incorporating commercial quahoggers into the sampling procedures and data collection, this
338 research improved both their understanding of how survey data are used by scientists and the
339 working relationship between industry and managers. The findings of this research have direct
340 applications, as they can be used to correct dredge gear efficiency issues prior to stock
341 assessment modeling.

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350

351 LITERATURE CITED

- 352 Akaike, H. 1973. Information theory and an extension of the maximum likelihood principle. In:
353 B. N. Petrov & F. Csaki. Proceedings of the 2nd international symposium on information
354 theory. Akadémiai Kiadó, Budapest. pp 267–281.
- 355 Arnold, W.S., D.C. Marelli, C.P. Bray, & M.M. Harrison. 1998. Recruitment of bay scallops
356 *Argopecten irradians* in Floridian Gulf of Mexico waters: scales of coherence. *Mar. Ecol.*
357 *Prog. Ser.* 170:143-157.
- 358 Bethoney, N.D., C. Cleaver, S.C. Samuel, S.R. Bayer, R.A. Wahle, & K.D.E. Stokesbury. 2019.
359 A comparison of drop camera and diver survey methods to monitor Atlantic sea scallops
360 (*Placopecten magellanicus*) in a small fishery closure. *J. Shell. Res.* 38(1):43-51.
- 361 Brooks, M. E., K. Kristensen, K. J. van Benthem, A. Magnusson, C. W. Berg, A. Nielsen, H. J.
362 Skaug, M. Machler, & B. M. Bolker. 2017. glmmTMB balances speed and flexibility
363 among packages for zero-inflated generalized linear mixed modeling. *R Journal* 9:378–
364 400.
- 365 Cadigan, N.G. & J.J. Dowden. 2010. Statistical inference about the relative efficiency of a new
366 survey protocol, based on paired-tow survey calibration data. *Fish. Bull.* 108:15–29.
- 367 Cadrin, S.X., G.R. Decelles, & D. Reid. 2016. Informing fishery assessment and management
368 with field observations of selectivity and efficiency. *Fish. Res.* 184:9-17.
- 369 Chestnut, A.F. 1952. Growth rates and movements of hard clams, *Venus mercenaria*. *Proc. Gulf*
370 *Caribb. Fish. Inst.* 4:49-59.
- 371 Doering, P.H., C.A. Oviatt, & J.R. Kelly. 1986. The effects of the filter-feeding clam *Mercenaria*
372 *mercenaria* on carbon cycling in experimental marine mesocosms. *J. Mar. Res.* 44: 839-
373 861
- 374 Ganz, A., N. Lazar, & A. Valliere. 1994. Quahaug Management Project, Phase I, Greenwich
375 Bay. A report to the Narragansett Bay Project. RI Division of Fish and Wildlife. 108 pgs.
- 376 Ganz, A., A. Valliere, M. Gibson, & N. Lazar. 1999. Narragansett Bay Quahaug Management
377 Plan. A report to the Narragansett Bay Project and the Rhode Island Marine Fisheries
378 Council. RI Division of Fish and Wildlife. 225 pp.

- 379 Hilborn, R. 2001. Calculation of biomass trend, exploitation rate, and surplus production from
380 survey and catch data. *Can. J. Fish. Aquat. Sci.* 58: 579-584.
- 381 Hilborn, R. & C.J. Walters, 1992. Quantitative Fisheries Stock Assessment: Choice, Dynamics,
382 & Uncertainty. Chapman and Hall, New York. 570 p.
- 383 Mackenzie, C.L. Jr. 1997. The U. S. molluscan fisheries from Massachusetts Bay through
384 Raritan Bay, N.Y. and N.J. NOAA Technical Report NMFS 0(127): 87-117
- 385 McManus, M.C., D.S. Ullman, S.D. Rutherford, & C. Kincaid. 2020. Northern quahog
386 (*Mercenaria mercenaria*) larval transport and settlement modeled for a temperate
387 estuary. *Limnol. Oceanog.* 65(2): 289-303
- 388 Medcof, J.C. & J.F. Caddy. 1971. Underwater observations on performance of clam dredges of
389 three types. I.C.E.S. CM. 1971/B:10.
- 390 Meyer, T.L., R.A. Cooper, & K.P. Pecci. 1981. The performance and environmental effects of a
391 hydraulic clam dredge. *Fish. Bull.* 43(9): 14-22.
- 392 Michael, K.P., G.P. Olsen, B.T. Hvid, & H.J. Cranfield. 1990. Design and performance of two
393 hydraulic subtidal clam dredges in New Zealand. New Zealand Fisheries Technical
394 Report No. 21.16 p.
- 395 Miller, T. J. 2013. A comparison of hierarchical models for relative catch efficiency based on
396 paired-gear data for US Northwest Atlantic fish stocks. *Can. J. Fish. Aquat. Sci.* 70:
397 1306–1316
- 398 Miller. T. J., D. R. Hart, K. Hopkins, N. H. Vine, R. Taylor, A. D. York, & S. M. Gallager. 2018.
399 Estimation of the capture efficiency and abundance of Atlantic sea scallops *Placopecten*
400 *magellanicus* from paired photographic-dredge tows using hierarchical models. *Can. J.*
401 *Fish. Aquat. Sci.* 76(6): 847-855.
- 402 Pratt, D.M. 1953. Abundance and growth of *Venus mercenaria* and *Callocardia morrhuana* in
403 relation to the character of bottom sediments. *J. Mar. Res.* 12: 60-74.
404
- 405 Pratt, S., A. Ganz, & M. Rice. 1992. A species profile of the quahog in Rhode Island. Rhode
406 Island Sea Grant Publication No. RIU-T-92-001 (P-1272). 117 p.
407
- 408 Quinn, T.J. & R.B. Deriso. 1999. Population Growth, Mortality, and the Fishing Process. *In*
409 *Quantitative Fish Dynamics*. Oxford Press University. New York, NY. 1-42.
410
- 411 Rice, M. 1992. The Northern quahog. Rhode Island Sea Grant Publication No. RIU-B-192-001
412 (P-1276). 60 p.
413
- 414 Rotherham, D., A.J. Underwood, M.G. Chapman, & C.A. Gray. 2007. A strategy for developing
415 scientific sampling tools for fishery independent surveys of estuarine fish in New South
416 Wales, Australia. *ICES J. Mar. Sci.* 64:1512–1516.
417

- 418 Russell, H.J. Jr. 1972. Use of a commercial dredge to estimate a hardshell clam population by
419 stratified random sampling. *J. Fish. Res. Bd. Can.* 29:1731-1735.
420
- 421 Saila, S.B. & T.A. Gaucher. 1966. Estimation of a sampling distribution and numerical
422 abundance of some molluscs in a Rhode Island salt pond. *Proc. Natl. Shellfish. Assoc.*
423 56:73-80.
- 424 Schuman, S. 2015. Rhode Island's shellfish heritage: an ecological history. Rhode Island Sea
425 Grant and University of Rhode Island's Coastal Resource Center. Narragansett, RI. 168
426 pp. <http://www.shellfishri.com/ecohistory/>.
- 427 Smithson, M. & J. Verkuilen. 2006. A better lemon squeezer? Maximum-likelihood regression
428 with beta-distributed dependent variables. *Psychol. Mod.* 11(1): 54-71.
- 429 Smolowitz, R.J. & V.E. Nulk. 1982. The design of an electrohydraulic dredge for clam surveys.
430 *Mar. Rev.* 44(4):1-18.
- 431 Southworth, M., J. Harding, J.A. Wesson, & R. Mann. 2010. Oyster (*Crassostrea virginica*,
432 Gmelin 1791) population dynamics on public reefs in the Great Wicomico River,
433 Virginia, USA. *J. Shell. Res.* 29(2):271-290.
- 434 Thorarinsdóttir, G. G., L. Jacobson, S.A. Ragnarsson, E.G. Garcia, & K. Gunnarsson. 2010.
435 Capture efficiency and size selectivity of hydraulic clam dredges used in fishing for
436 ocean quahogs (*Arctica islandica*): simultaneous estimation in the SELECT model. *ICES*
437 *J. Mar. Sci.* 67: 345–354.
- 438 Venables, W.N. & C.M. Ripley. 2004. GLMs, GAMs, and GLMMs: an overview of theory
439 for applications in fisheries research. *Fish. Res.* 70: 319-337.
- 440
- 441

442 Table 1. Dates and the number of stations sampled for each of the gear efficiency methods.

Date	Dredge Samples	Dredge Transect Inspection	Bull rake Samples	Quadrat Samples
10-Oct-2017	3	3	3	3
11-Oct-2017	5	5	-	5
18-Oct-2017	3	3	-	3
19-Oct-2017	4	4	3	4
27-Oct-2017	5	5	5	5
30-Jul-2018	5	5	-	5
1-Aug-2018	5	5	2	5
17-Sep-2018	5	5	-	5
5-Oct-2018	5	5	-	5
19-Oct-2018	5	5	2	5

443

444 Table 2. Descriptive statistics of mean dredge efficiency across bottom and sediment types.
 445 Sample sizes (n) and standard deviations around the means (sd) are presented by category. The
 446 categories with samples sizes of one do not have associated standard deviations. Significant
 447 differences between categories within a variable are denoted with lettered subscripts, as tested
 448 using Kruskal-Wallis Rank Sum and Nemenyi Tests.

Variable (n)	Dredge Efficiency (sd)
Bottom Type	
Hard (20) ^A	0.73 (0.23)
Soft (19) ^B	0.48 (0.28)
Unknown (3) ^{A,B}	0.69 (0.16)
Sediment	
Mud (14)	0.50 (0.28)
Mud, Shell (11)	0.67 (0.29)
Sand (1)	1.00
Sand, Mud (8)	0.73 (0.28)
Sand, Mud, Shell (3)	0.56 (0.20)
Sand, Rock, Shell (1)	0.44
Sand, Shell (4)	0.58 (0.22)

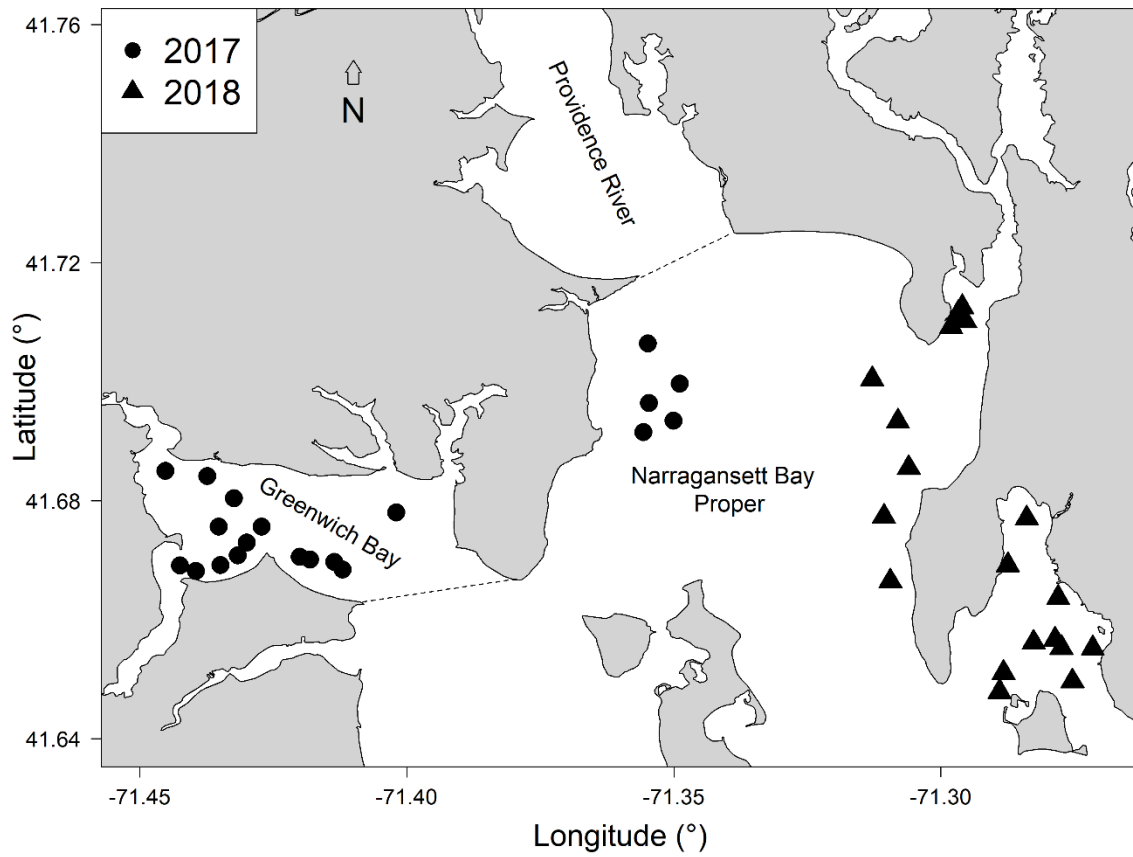
449

450 Table 3. Generalized linear model variants tested for predicting dredge catch efficiency, with ‘x’
 451 denoting variable inclusion in the model. Model variant in bold signifies the final model chosen
 452 for predictions.

Model	Covariate				ΔAIC	Degrees of Freedom
	Bottom Type	Sediment	Depth	Observed Abundance		
1	x				0	4
2	x	x			6.4	10
3	x	x	x		3.3	11
4	x	x	x	x	3.9	12

453

454



455

456 Figure 1. Station locations in Narragansett Bay for the multi-gear sampling in 2017 (circles) and
457 2018 (triangles). Regions modeled for the quahog stock assessment are labeled, with dashed
458 lines delineating the regions.

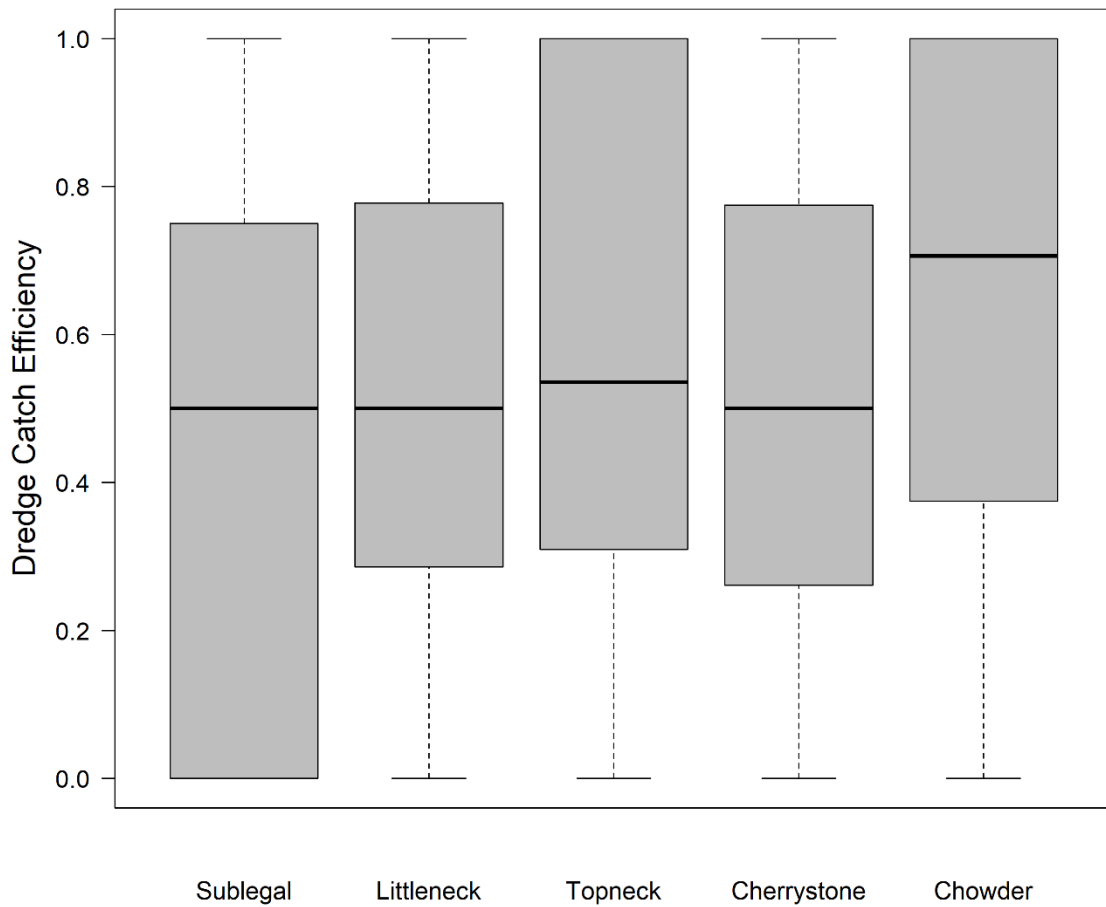


Figure 2. Dredge catch efficiencies for quahogs by market class. Dark lines, box heights, and tick marks indicate 50th, 25-75th, and 0-100th percentiles, respectively.

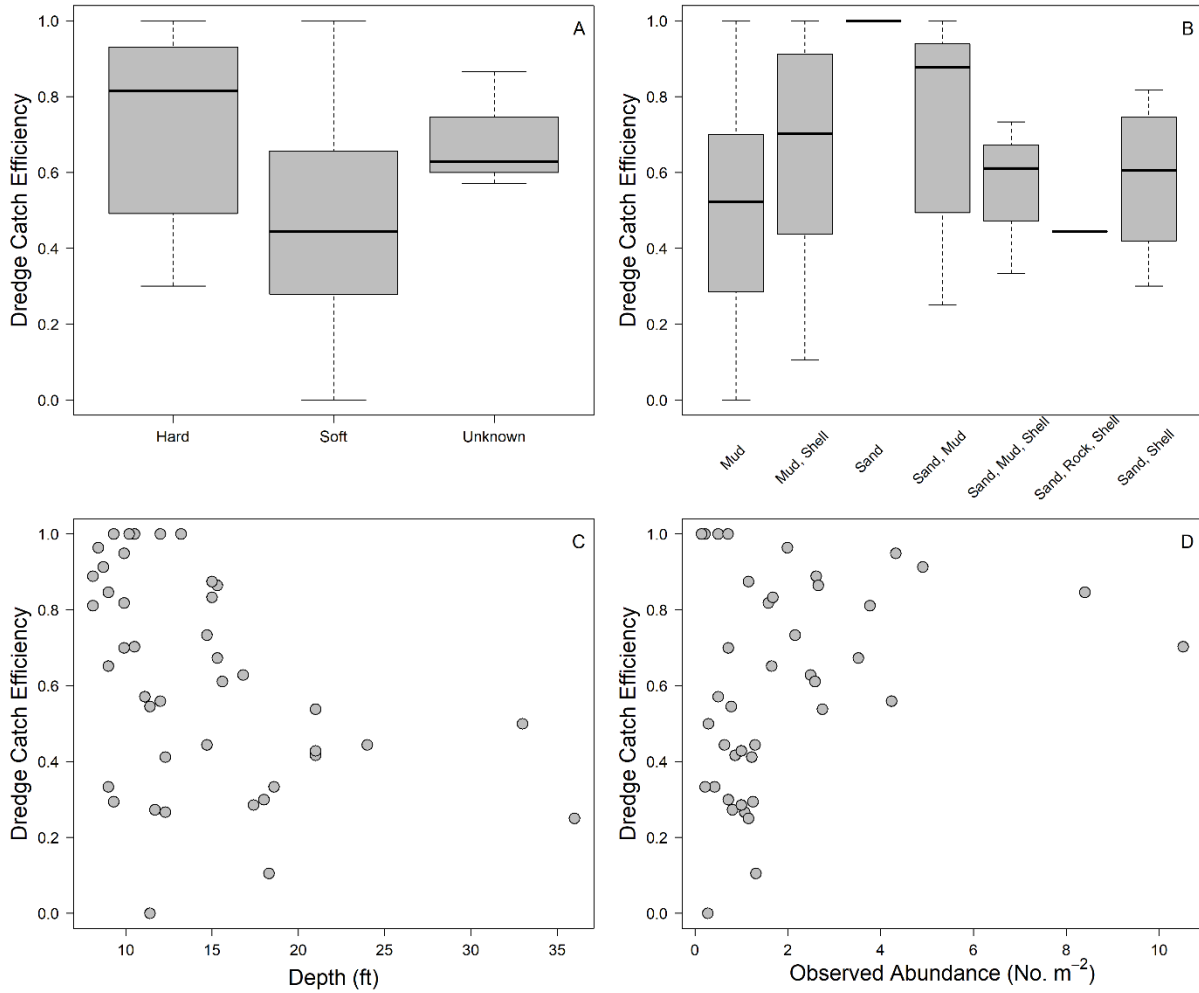


Figure 3. Dredge catch efficiencies for total quahog catch by (A) bottom type, (B) sediment type, (C) depth, and (D) total observed relative abundance. Dark lines, box heights, and tick marks for bottom and sediment types indicate 50th, 25-75th, and 0-100th percentiles, respectively.

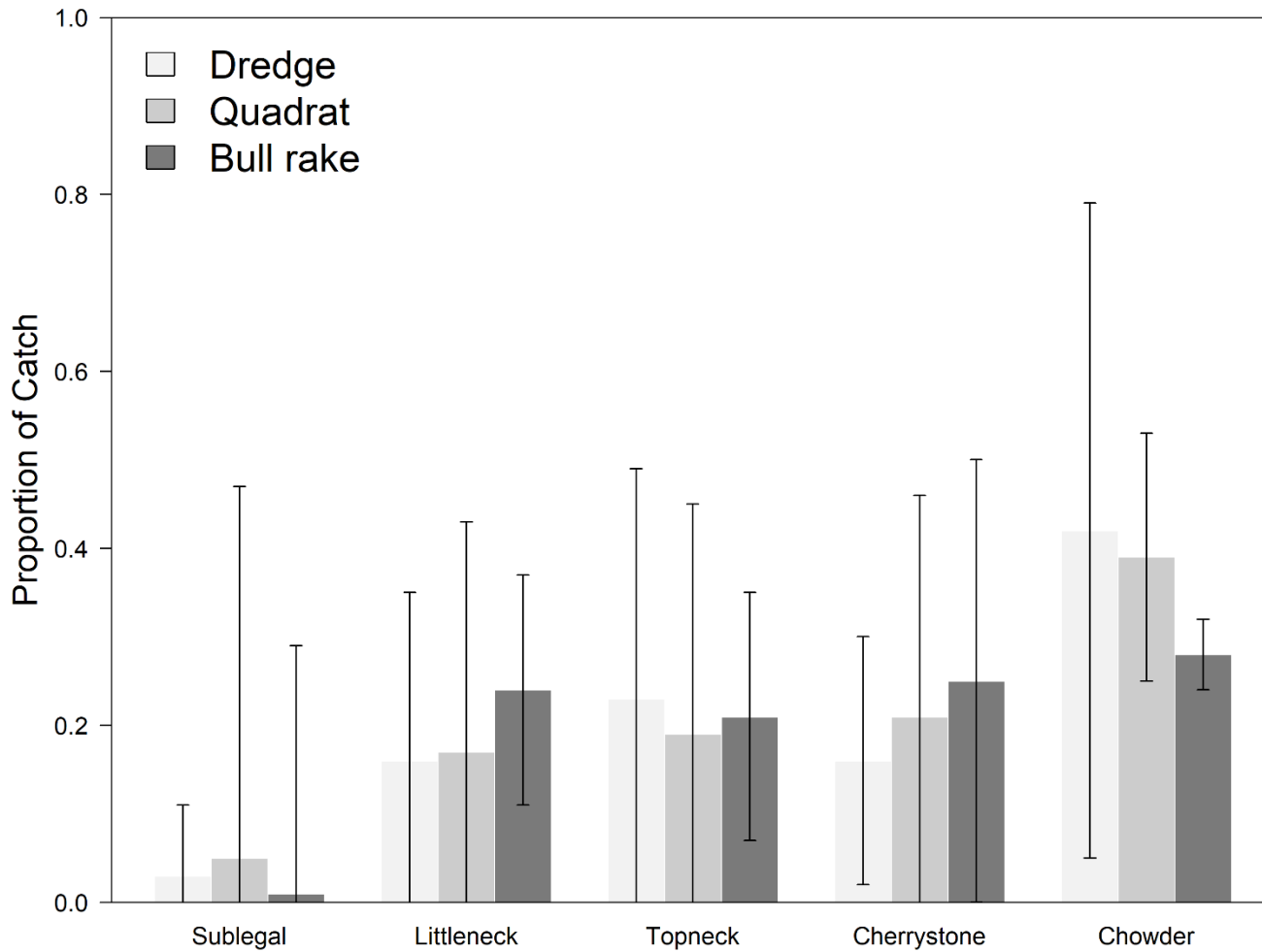


Figure 4. Average proportion of market classes within the samples of each sampling type: the hydraulic dredge, sampling with quadrats on SCUBA, and bull rake sampling. Proportions for quadrat and bull rake sampling include the replicates within a station. Bars are average proportions over all data within a sampling type, and bars represent the standard deviation range around the means.

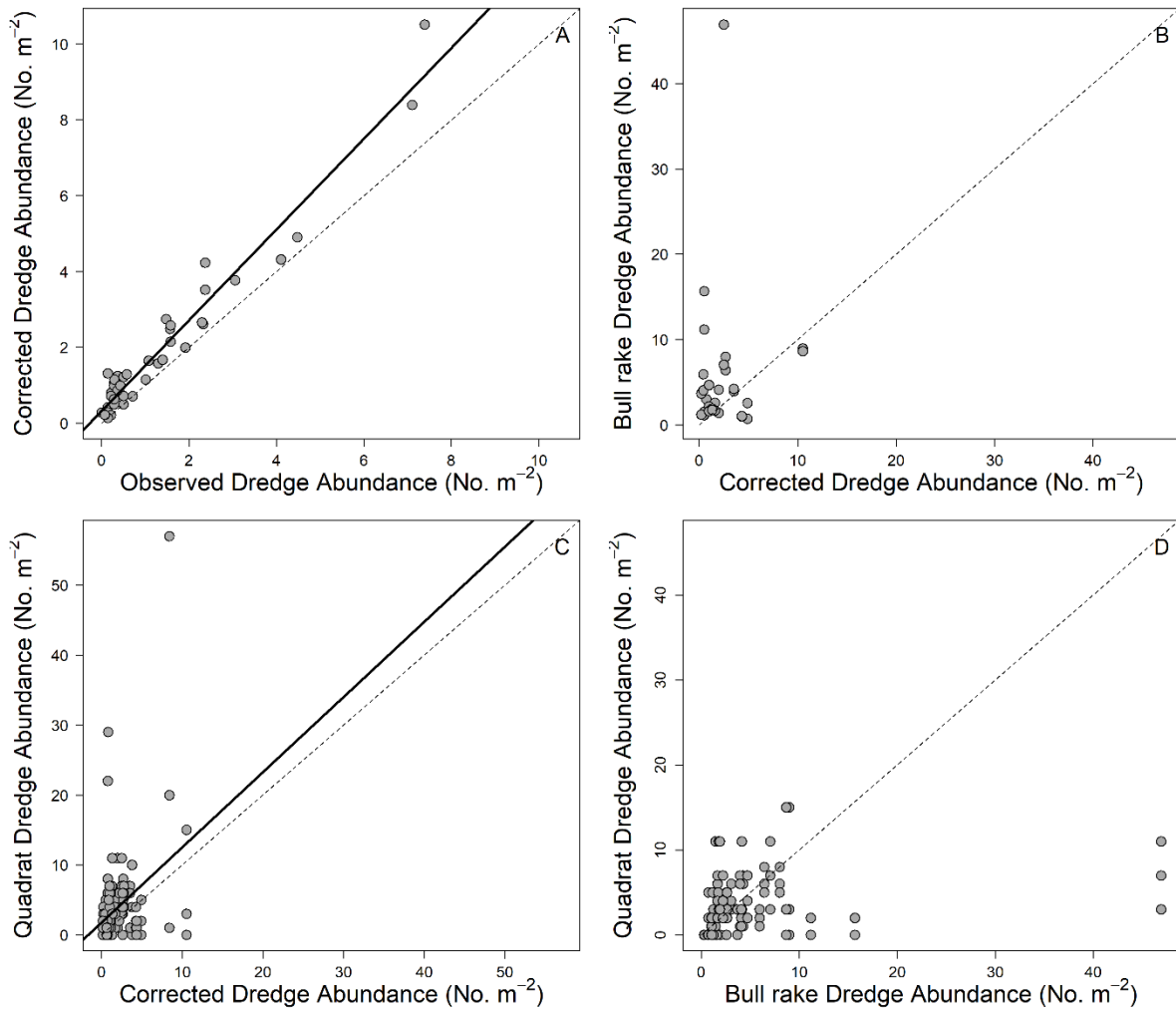


Figure 5. Relative abundance comparisons across gears and corrections. Observed quahog relative abundance from the hydraulic dredge compared to corrected abundances based on the transect inspections on SCUBA (A). Corrected quahog relative abundance from the hydraulic dredge compared to abundances collected using (B) bull rakes (n=32) and (C) quadrats on SCUBA (n=144), as well as (D) a comparison of bull rake and quadrat relative abundance estimates (n=114). Comparisons using bull rake and quadrat data were done at the replicate level. Solid lines represent linear fits for significant correlations. Dashed lines represent the 1:1 line, where abundances between the two estimates would be equal.

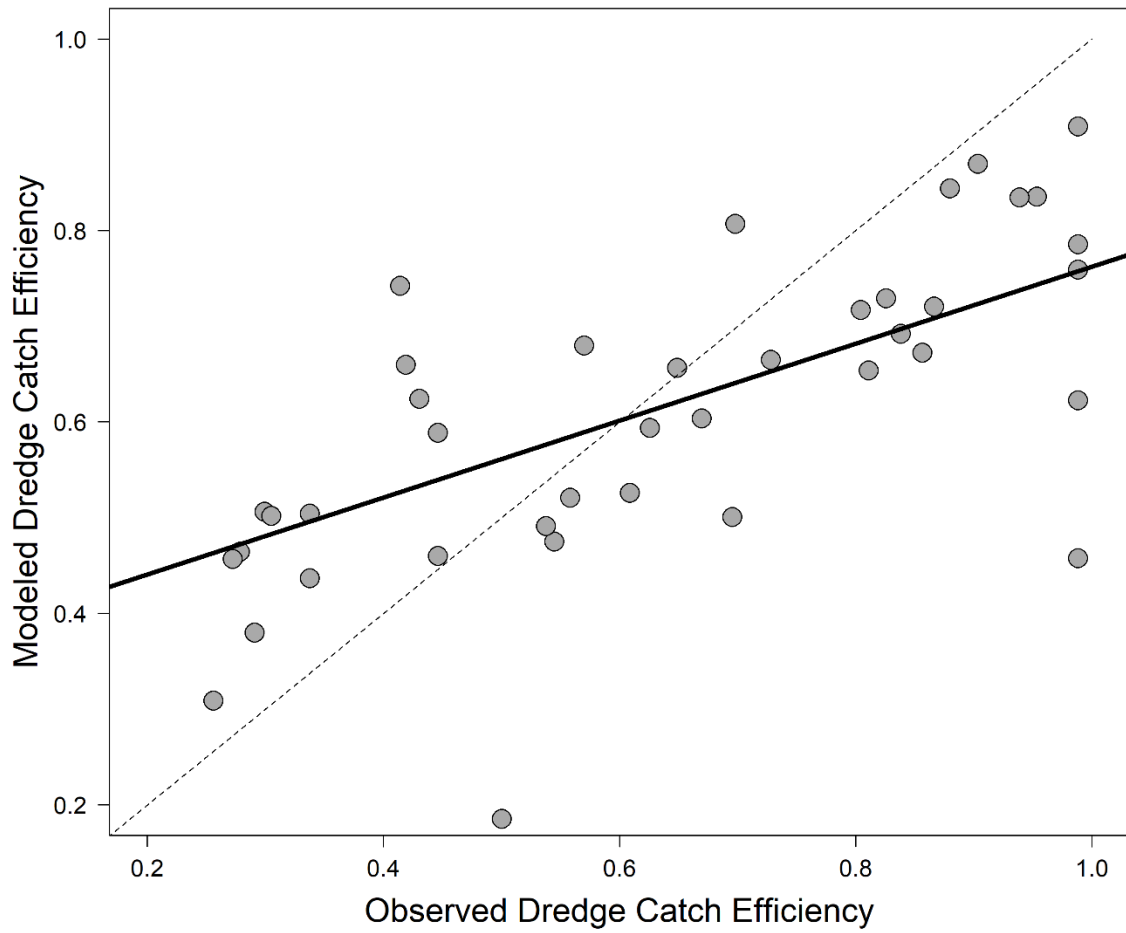


Figure 6. Observed dredge catch efficiency compared to that predicted from the model. The solid line represents the linear fit between the two datasets ($R^2=0.46$), and the dashed line represents a 1:1 line.

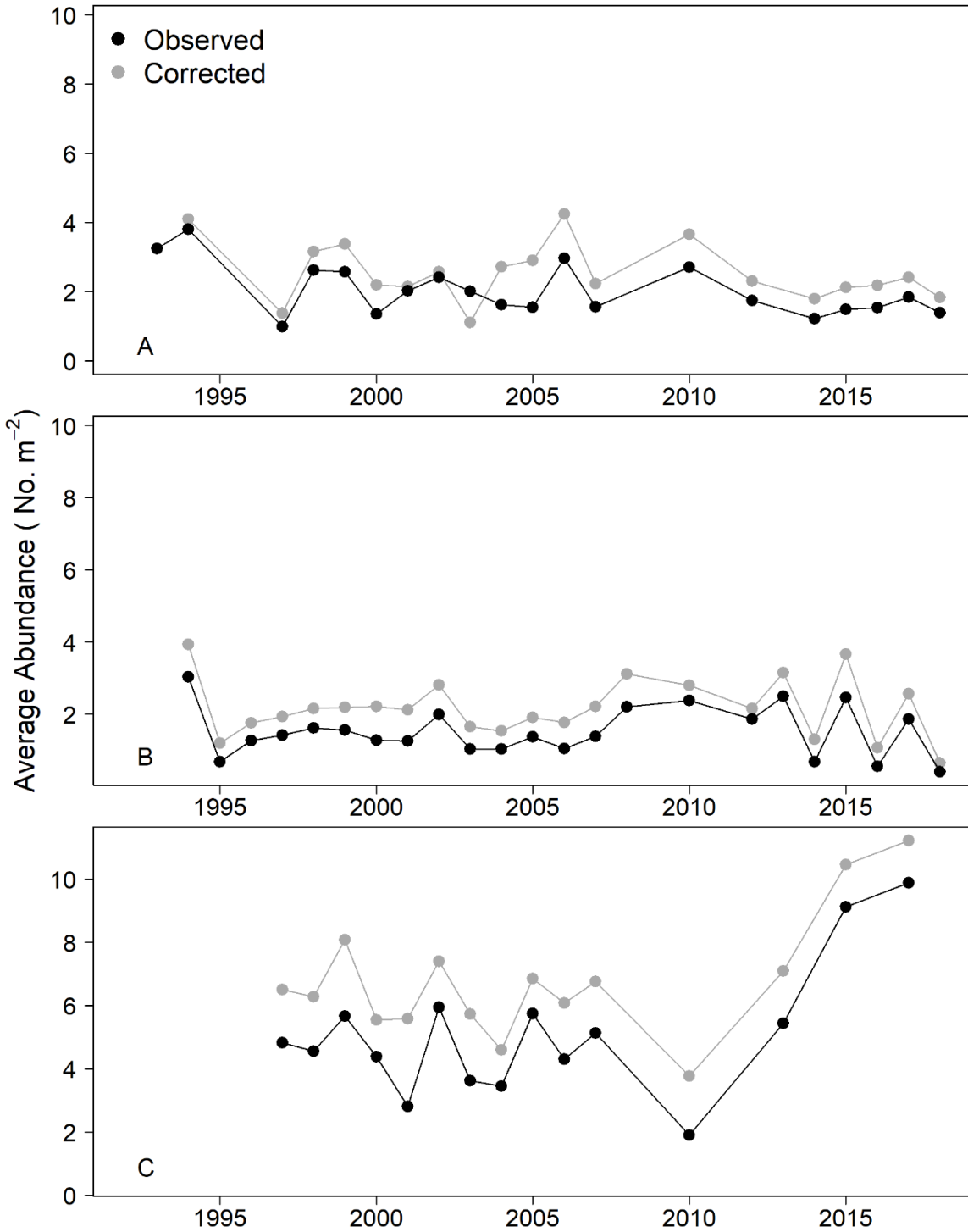


Figure 7. Quahog annual relative abundance indices from the RIDEM DMF quahog hydraulic dredge survey. Indices are presented for the three regions currently modeled for stock assessment: Greenwich Bay (A), Narragansett Bay proper (B), and the Providence River (C). Relative abundance time series are presented for those currently used ('Observed') and those corrected based on the dredge efficiency model ('Corrected').