

Title: A field test investigating the influence of brood stock origin and ploidy on the susceptibility of *Crassostrea virginica* to “triploid mortality” in the Chesapeake Bay

Authors: Joseph L. Matt*, Eric Guévelou, Jessica Moss Small, Standish K. Allen, Jr.

Authors' affiliation: Aquaculture Genetics and Breeding Technology Center, Virginia Institute of Marine Science, College of William & Mary, P.O. Box 1346, Gloucester Point, VA 23062. United States of America.

Corresponding author: Joseph L. Matt : Aquaculture Genetics and Breeding Technology Center, Virginia Institute of Marine Science, College of William & Mary, P.O. Box 1346, Gloucester Point, VA 23062. United States of America. E-mail address: jlmatt@vims.edu

Abstract

Mass mortalities of cultured triploid *Crassostrea virginica* in late spring, or “triploid mortality,” have been reported on farms in the Chesapeake Bay since 2012. Typical causes, such as disease or poor husbandry, were not responsible, and mortalities occurred without clear signs of biological or physical stressors. Previous comparisons of the effects of genetic origin on triploid mortality have been uncontrolled, initiating this investigation of the effect of brood stock source and ploidy on triploid mortality. Four triploid and four diploid crosses, produced by crossing different combinations of brood stock of Virginia, Louisiana, and Maine origin in early 2015, were tested at four commercial oyster farms in Virginia throughout 2016. From February to November, oysters from all crosses and sites were regularly sampled, and growth (shell height), condition (meat weight), and pathology were monitored, as were environmental conditions. Compared to diploids made from Virginia brood stock, diploids with

Maine genetic origin had high mortality and Virginia-Maine diploid hybrids exhibited mid-parent heterosis. A triploid mortality event occurred in late spring at only one site and only affected the triploid crosses. Evidence for substantial disease pressure from *Haplosporidium nelsoni* or *Perkinsus marinus* or of especially stressful environmental conditions based on temperature, salinity, pH, and dissolved oxygen was absent during the triploid mortality event. At 18 months, shell height was similar in the diploids and triploids with the most similar genetic origin. Triploids maintained meat weight through the summer, while meat weight in diploids dropped sharply. Triploids may be especially susceptible to late spring mortality events in the Chesapeake Bay, which justifies their classification as “triploid mortality,” and warrants further investigation on traits in triploids that may affect susceptibility.

Keywords: oyster; polyploid; 3n; summer mortality; gametogenesis

1. Introduction

Over the last two decades, the once prolific harvests for the eastern oyster, *Crassostrea virginica*, have been partially restored in the Chesapeake Bay from hatchery-based aquaculture. Pivotal to the rise in hatchery-based aquaculture has been selective breeding, which has made disease resistant and genetically improved oysters available (Dégremont et al., 2015; Frank-Lawale et al., 2014; Ragone Calvo et al., 2003). Production of *C. virginica* from aquaculture in the state of Virginia has grown rapidly in the last fifteen years. Seed production increased from 20 million to 264 million (13-fold) from 2005 to 2016, while the number of cultured oysters sold increased 40-fold (Hudson, 2017; Murray and Oesterling, 2006). Virginia now leads the East Coast of United States in aquaculture production of eastern oysters (Hudson, 2018).

An important genetic improvement in hatchery-based aquaculture in Virginia was the commercialization of triploid oysters. Over the last ten years, oyster aquaculture in Virginia has been

49 characterized by a near exclusive use of mated triploid oysters, which are triploids produced by crossing
50 tetraploid oysters to diploid oysters. Triploid oysters made up 80 to 95% of oyster crops on Virginia
51 farms between 2009-2017, and in 2015 and 2016, 94% of seed sold from Virginia hatcheries were
52 triploid (Hudson, 2018, 2017; Hudson and Murray, 2016, 2015, 2014; Murray and Hudson, 2013, 2012,
53 2011; Murray and Oesterling, 2010, 2009). The Virginia industry reports a preference for triploids
54 because they grow faster than diploids and maintain high meat quality during the spawning season
55 (Hudson, 2018). These advantages obtain from the reduced fecundity of triploid oysters, which alters
56 the energy allocation in favor of somatic growth (Purdom, 1972; Stanley et al., 1981). Lack of spawning
57 also improves market condition (Allen and Downing, 1986).

58 The apparent advantages of growing triploid oysters in the Chesapeake Bay may come with at
59 least one disadvantage, for example, so called “triploid mortality” (Guévelou et al., 2019). Since at least
60 2012, a number of oyster farms in the Chesapeake Bay have experienced mortality episodes, in which
61 substantial mortalities (> 20% of the crop) occur within a matter of weeks in the late spring, typically
62 between May and July (Guévelou et al., 2019). An acute incidence of triploid mortality in 2014 led to
63 losses as high as 50-80% on some farms (K. Hudson, pers. comm., Guévelou et al. 2019). Farmers
64 reported that oysters growing fastest and close to market size were susceptible. No pathogen was
65 identified in association with these mortalities, nor have any unusual environmental conditions been
66 consistently associated with the events, which made these mortalities unusual for oyster farmers in the
67 Chesapeake Bay (Guévelou et al., 2019). Reports of acute triploid mortality from several farms in
68 Virginia have also been received in 2019, ranging from 30% to 70% loss of crop (K. Hudson, pers.
69 comm.). Unusual mortality episodes involving triploids have also been reported from oyster farms in
70 Maryland, as well as Alabama and Louisiana since 2016 (Wadsworth et al., 2019).

71 Often, mass mortality events of oysters involve a pathogen, such as a protozoan parasite (e.g.
72 *Perkinsus* spp., *Haplosporidium* spp., *Bonamia* spp.), bacterium (e.g. *Vibrio* spp.), or virus (e.g. OsHV-1).

Of most recent importance in terms of global oyster aquaculture are mortalities caused from a certain genotype of Ostreid herpesvirus type I (OshV-1), OshV-1 μ Var (Segarra et al., 2010). The μ Var genotype was first identified following widespread mortalities on farms in France in the summer of 2008 and is now associated with mortalities during summer in other western European countries (Peeler et al., 2012), New Zealand (Keeling et al., 2014), and Australia (Paul-Pont, 2014). The severe economic impact of these mass mortalities has made resistance to OshV-1 μ Var a focus of applied research, including breeding research, which has demonstrated high potential for increased resistance through selection (Camara et al., 2017; Dégremont et al., 2015; Kube et al., 2018).

Mass mortalities of oysters have often occurred without an attributable pathogen or specific environmental stressor. Similar to *triploid* mortalities in Virginia, many cases of the long-reported *summer* mortalities of *Crassostrea gigas* involve only adult oysters and are not ascribed to a specific pathogen or environmental condition (Cotter et al., 2010; Glude, 1975; Koganezawa, 1975; Maurer and Comps, 1986; Samain and McCombie, 2008; Wendling and Wegner, 2013). Since the 1940s, possibly even 1912 (Takeuchi et al., 1960), episodes of mortalities of cultured *C. gigas* ranging from 10% up to 70% have been reported in summer in Japan (Koganezawa, 1975; Ventilla, 1984). Similar mortalities have been reported on the West Coast of the United States starting in the 1950s (Cheney, 2000; Glude, 1975; Perdue et al., 1981) and France in the 1970s (Goulletquer et al., 1998; Parache, 1989; Samain and McCombie, 2008). The conclusion in Koganezawa (1975) that summer mortalities were due to a physiological “disorder” induced by a suite of environmental conditions has been supported from follow up research on another continent (e.g. Samain and McCombie, 2008).

In Virginia, nearly all the farms affected by triploid mortality in 2014 were on the Chesapeake Bay-side of Eastern Shore of Virginia, and many farmers in the area were growing the “Northern cross” (Guévelou et al., 2019). The Northern cross was produced by mating tetraploids selectively bred from mid-Atlantic stocks with a proprietary line of diploids developed from New England stocks. Thus, a

hypothesis was developed that genes passed to the triploid by the “northern diploid” may contribute to late spring mortality in triploids. The initial investigation into triploid mortality in the Chesapeake Bay by Guévelou et al. (2019) included a version of this “Northern cross,” but the opportunistic nature of the field trial meant certain variables, like genetic provenance of the tested groups, could not be controlled.

This study is intended to be a more controlled complement to Guévelou et al. (2019) and shares many of the same objectives: testing whether genetic origin affects triploid mortality, measuring the survival of diploids during a triploid mortality event, examining the relationship between size and mortality, and investigating associations between triploid mortality and environmental conditions. Unique to this study are tetraploid brood stock of varying genetic origin. Considering tetraploids contribute two thirds of the chromosomes to the triploid genome, the genetics of the tetraploid parent may be especially influential of the triploid phenotype. Four genetically distinct diploid and triploid crosses were produced and deployed in a field test from February 2016 to November 2016. Triploid crosses were designed to produce a range of genotypes – origins in Virginia, Louisiana, and Maine – that would putatively affect susceptibility to triploid mortality. The diploid crosses comprised those of Virginia origin, which were produced for comparison with the triploids, as well as those of Maine origin, which were produced to further examine the performance of New England derived genotypes in the Chesapeake Bay. Special care was taken to mimic growing conditions in which previous triploid mortality events had been observed, specifically, early season spawns that yielded market sized adults at the cusp of the triploid mortality window 16-18 months after spawning (May – June).

2. Methods

2.1 Brood stock

Two lines of diploid and two lines of tetraploid *C. virginica* were used to produce diploid and triploid crosses (Figure 1). The diploid brood stock consisted of the Virginia Institute of Marine Science

(VIMS) DEBY line (Ragone Calvo et al., 2003) and a proprietary commercial line from Mook SeaFarms, Walpole, Maine. The tetraploid brood stock consisted of VIMS GEN and VIMS VBOY. The GEN line has been the principal source of tetraploids for the vast majority of commercial triploid production in the Chesapeake and has been propagated by the Aquaculture Genetics and Breeding Technology Center (ABC) since 2003. The VBOY line originated from the creation of a tetraploid line for the Oyster Research Lab of Louisiana State University (LSU). VBOY was made as a hybrid between a diploid line (OBOY, developed by Jerome LaPeyre, LSU) from Louisiana and the GEN line developed in Virginia. VBOY has also been held by ABC since 2013.

2.2 Crosses

All crosses were spawned simultaneously at the VIMS research hatchery in Gloucester Point, VA in February of 2015. Male tetraploids from the Virginia GEN line (chromosome set contribution: VV) and VBOY line with some Louisiana origin (chromosome set contribution: LL) were crossed to female diploids from the ABC DEBY line (chromosome set contribution: V) or the Mook line from Maine (chromosome set contribution: M) in a 2x2 matrix to produce triploid crosses (VVV, VVM, LLV, LLM). Eggs from the same V and M female diploids were also crossed to V and M male diploids in a 2x2 matrix to produce a diploid control group to gauge triploid mortality (VV), reciprocal Virginia-Maine hybrids (VM, MV), and Maine diploids (MM).

All crosses were conducted via strip spawning (Allen and Bushek, 1992). For each of the two diploid brood stocks (V & M), five to eight females were stripped and eggs were pooled. Each pool was then split into four aliquots. One aliquot of the egg pool from either V or M was allocated for each of the sperm sources (2N – V and M; 4N – VV and LL) (Figure 1). For each sperm source, five males were used, so each aliquot of diploid eggs was further subdivided five ways so that an individual male was fertilizing one fifth of the eggs from each egg source. After each subdivided batch of eggs had been allowed to complete fertilization, they were re-pooled into the eight major crosses depicted in Figure 1.

The early (February) spawn in 2015 was necessary to emulate the commercial experience of farms that had seen triploid mortality in near market size oysters in late spring.

2.3 Larvae and Seed

Larvae were reared at two hatcheries to ensure success in producing pediveliger larvae: the VIMS research hatchery (Gloucester Point, VA, USA) and a commercial hatchery, Oyster Seed Holdings (Gwynn's Island, VA, USA). Half of the embryos from each cross were transferred directly into tanks at the VIMS hatchery, and half were transported in a 50-milliliter centrifuge tube to the commercial hatchery, 50 kilometers away. Pediveliger larvae were successfully produced for all crosses at both hatcheries. Competent pediveligers were collected from each cross from both hatcheries were set on finely crushed oyster shell at the VIMS hatchery.

Several times throughout the spring, seed were graded to ensure fast growth to promote the earliest possible deployment to the field, which emulates the commercial process that gave rise to triploid mortality. By June of 2015, when enough seed were larger than 19 mm, 10-15 haphazardly selected individuals from each cross were measured for shell height (maximum dimension from the hinge to the bill), and a few thousand seed from each cross were deployed to four experimental field sites. At all sites, seed were reared as a bulk deployment in the intertidal zone in bags nested within single tier cages.

2.4 Sites and Experimental Deployment

All four experimental sites were commercial oyster farms. Three farms were on the bayside of the Eastern Shore of Virginia and one was on the western side of the Chesapeake Bay. For the three eastern sites – Nandua Creek (ND), Pungoteague (PG), Occohannock Creek (OC) – triploid mortality had been observed. For the western site – Rappahannock River (RR) – no such mortalities have been observed (Figure 2). Deployment of the experiment took place between February 29 and March 3 of 2016, in which oysters from all eight crosses were split into replicate bags at each site. From each cross

at each site, 450 oysters were haphazardly selected and equally split into 3 bags (150 x 3). At RR, ND, and OC, oysters were reared in single-tier cages in the subtidal zone, and each bag was randomly assigned a slot within the cages. Bags and cages were held in their assigned position throughout the experiment. For PG, oysters were reared in triple-tier cages in the subtidal zone. Bags and cages were rotated during the trial, but cages remained in the same area during the experiment.

2.5 Ploidy Verification

Oysters were sampled twice to verify ploidy via flow cytometry (FCM) (Allen and Bushek, 1992). All FCM measurements were made with a Sysmex-Partec Cyflow Space flow cytometer (Partec GmbH, Münster, Germany) using DAPI as a stain. The first sample occurred in April of 2015 when 25 two-month-old seed from each cross were haphazardly selected. The second ploidy verification was repeated at deployment (February/March 2016) by sampling 15 haphazardly selected individuals of each cross from each site.

2.6 Site Visits and Sampling

All sites were visited throughout the spring, summer, and fall of 2016. Sites were visited within a week of each other monthly except in May and June when they were visited twice a month. Live oysters were sampled, without replacement, at experimental deployment and during all site visits in the spring and summer of 2016 (April-August). For all sampling times after deployment, oysters were randomly sampled. Random sampling consisted of ordering oysters in piles of 10, selecting a pile with a random number generator, ordering the oysters from the pile, and then selecting the individual oysters with a random number generator. Five oysters from each bag were selected in April, July, and August. In May and June, five oysters were selected from each bag of diploids, and seven were selected from each bag of triploids. Shell height and meat weight were measured for all live samples. Meat weight was measured after shucking oysters, removing the somatic tissue from the shell, and letting the meat drain on a mesh net for a few minutes.

2.7 Mortality

Mortality within bags was assessed during each site visit. Live oysters and empty shells were counted from each bag and empty shells removed. Oysters that failed to seal their shell commissure were deemed moribund and were also counted and removed. For the first mortality assessment in April, the percent cumulative mortality per bag was calculated with the following equation:

$$\text{Cumulative mortality} = \frac{\text{Live}}{\text{Deployed}}$$

Where Live is the number of live oysters counted and Deployed is the sum of dead, live, and moribund oysters. For all subsequent sampling times, cumulative mortality per bag was calculated with the following equation:

$$\text{Cumulative mortality}_t = \frac{\text{Live}_t}{\text{Live}_{t-1} - \text{Samples Removed}_{t-1}} \times \text{Cumulative mortality}_{t-1}$$

where t indicates sampling time and Samples Removed are the number of live oysters sampled without replacement. The number deployed, assessed during counts in April, was uniform for most replicates, ranging from 146 to 155 oysters per bag. Two bags out of the 96 deployed had 171 and 198 oysters instead of 150. Mean cumulative mortality for each cross was calculated by averaging the cumulative mortalities among the three bags. Shell height was measured for empty shells and moribund oysters.

Mid-parent heterosis was the phenotypic difference between the mean of the parental crosses and the mean of the hybrids, defined as heterosis by Falconer (1981), and was measured for mortality between the Virginia and Maine diploids and the Virginia-Maine hybrids with the following equation:

$$H = \frac{(VV + MM)}{2} - \frac{(VM + MV)}{2}$$

where VV, MM, VM, and MV refer to the mortality of the corresponding crosses.

2.8 Histology and Pathology

Live samples were preserved for histology. For all live samples, the excised sample (slab) consisted of a 4 mm section of tissue cut perpendicular to the anterior –posterior axis, slightly ventral of the labial palps. Slabs were fixed in Davidson’s solution for 48 hours, then stored in 70% ethanol. Some fixed samples were selected *post-hoc* for pathology based on mortality results and were processed for histology by standard methods used at the VIMS Shellfish Pathology Laboratory (Carnegie and Bureson, 2011).

Infections were described by prevalence (percent of infected individuals) and weighted prevalence, which is based on the intensities of the infections. Intensities of *Haplosporidium nelsoni* infections were rated according to Carnegie and Bureson (2011), and intensities of *Perkinsus marinus* infections were rated according to Mann et al. (2014). Weighted prevalence (WP) was calculated with the following equation:

$$WP = \frac{5(\text{Heavy}) + 4(\text{Moderate to Heavy}) + 3(\text{Moderate}) + 2(\text{Light to Moderate}) + 1(\text{Rare or Light}) + 0(\text{Absent})}{\text{Number of samples}}$$

where Heavy, Moderate to Heavy, Moderate, Light to Moderate, and Rare or Light represent the number of oysters with these qualitative ratings of infection intensity. All pathology, via visual examination of histology slides, was completed by the VIMS Shellfish Pathology Laboratory.

2.9 Environmental Conditions

Temperature and salinity were monitored at each field site during the experiment by HOBO® conductivity logger (Onset Computer Corporation, Bourne, MA, USA) attached to a cage in the field trial at each site from January 1 to August 10, 2016. Some mechanical failure of the loggers produced gaps in the data. Calibration readings for conductivity were taken with a portable conductivity meter during each site visit, and conductivity data were converted to salinity in parts per thousand (ppt) using HOBOWare Conductivity Assistant (Onset Computer Corporation, Bourne, MA, USA). For RR, salinity

data was taken from daily recordings at a nearby VIMS research hatchery on Locklies Creek (VA, USA).

Average daily values were calculated for temperature and salinity at each site.

Conductivity, dissolved oxygen concentration, pH, chlorophyll *a* concentration, and turbidity

were monitored at ND and PG during the time triploid mortality was expected (May-July).

Measurements were taken in 15-minute intervals using the 6600 V2-4 Multi-Parameter Water Quality

Sonde (YSI Inc. / Xylem Inc., Yellow Springs, OH), between May 10 and July 12. In cases where data from

the conductivity logger and sonde were available at ND or PG, sonde data was reported.

2.10 Statistical Analysis

Mortality was analyzed using a mixed effects logistic regression. To compare “triploid mortality,” the following model was used for mortality from deployment to the end of spring (February 29-June 28):

$$\text{logit}(\pi_i) = \log\left(\frac{\pi_i}{1-\pi_i}\right) = \mu + \alpha_i + \beta_j + \alpha_i\beta_j + \gamma_{k(i,j)} + \varepsilon_{ijk}$$

where survival is a binary response variable (live or dead), μ is the mean probability of survival, α is the effect of cross (i is VV, VVV, VVM, LLV, and LLM), β is the effect of site (j is RR, ND, PG, and OC), $\alpha\beta$ is the interaction of cross and site, γ is the effect of bag nested within cross and site, and ε_{ijk} is the residual error. The mortality of the Virginia diploids, Virginia-Maine hybrids, and Maine diploids was analyzed with the same model except the data were from deployment to the end of the field trial (February 29-November 16) and i was VV, VM, MV, and MM. For both models, cross and site were fixed effects and bag was a random effect. Pair-wise differences were examined post-hoc using Tukey’s Honest Significant Difference (HSD) test. Mixed effects logistic regression was performed using the lme4 package in R (Bates et al., 2015).

Shell height in August of 2016 was analyzed using a two-way ANOVA:

$$Y_{ijk} = \mu + \alpha_i + \beta_j + \alpha_i\beta_j + \varepsilon_{ijk}$$

where Y_{ijk} is shell height of individual oysters in August 2016, μ is the overall mean, α is the effect of cross (i is VV, VM, MV, MM, VVV, VVM, LLV, and LLM), β is the effect of site (j is RR, ND, PG, and OC), $\alpha\beta$ is the interaction of cross and site, and ε_{ijk} is the residual error. Pair-wise differences were examined post-hoc using Tukey's HSD test. To determine if there was a difference in size among surviving triploids and triploids that died during the mortality event, shell heights of live triploids and dead triploids were analyzed using a nested ANOVA:

$$Y_{ijkl} = \mu + \alpha_i + \beta_j + \gamma_{k(j)} + \alpha_i\beta_j + \varepsilon_{ijkl}$$

where Y_{ijkl} is height of shells of live and dead triploids measured on June 6, 2016 at ND, μ is the overall mean, α is the effect of status (i is live or dead), β is the effect of cross (j is VVV, VVM, LLV, LLM), γ is the effect of bag nested within cross (k is replicate 1, 2, 3), $\alpha_i\beta_j$ is the interaction between status and cross, and ε_{ijkl} is the residual error.

Shell height data were visually assessed for normality and homogeneity of variance among factors, and residuals from models were visually examined for normality and homoscedasticity. For all tests, significance level = 0.05. All statistical analyses were performed in R (R Core Team, 2019).

3. Results

3.1 Ploidy

From flow cytometry, most samples from triploid crosses were verified triploid, and all samples from diploid crosses were verified diploid. In April of 2015, all oysters from triploid crosses were verified triploid (n=25) except VVM, where 24 of 25 were triploid and the other was diploid. In February/March 2016, all triploids were verified triploid (n=15) except for LLM at PG and LLM at RR, where 14 of 15 were triploid and the other was diploid.

3.2 Mortality

Mean cumulative mortality for crosses at all sites ranged from 6% to 76% by the end of the field trial (November 2016) (Figure 3). At all sites, MM had the highest and VV had the lowest cumulative

mortality. A large increase in cumulative mortality was observed in triploid crosses at ND between April 14 and June 7. During this time, cumulative mortality increased from 0% in all triploid crosses to 27% in VVV, 22% in VVM, and 17% in LLV (Figure 3). Increase in mortality for LLM was only 8%.

A significant interaction existed between site and cross during the window “triploid mortality” was expected (Table 1). At RR, PG, and OC, all crosses had a low probability of mortality (< 0.15) (Figure 4). At ND, the probability of mortality was 0.01 for VV, 0.29 for VVV, 0.27 for VVM, 0.20 for LLV, and 0.14 for LLM. The probability of mortality was significantly different between VV and the triploid crosses and was statistically indistinguishable among the triploid crosses (Figure 4).

There was a significant interaction between site and cross among the Virginia and Maine diploids (Table 1). Estimates of probability of mortality in Virginia and Maine diploids shared a similar pattern among sites (Figure 5). The VV cross always had the lowest probability of mortality (< 0.1) and was significantly different from other crosses within a site, with the only exception that VV and VM were statistically indistinguishable at OC. The hybrids (VM and MV) had a statistically indistinguishable probability of mortality within each site, and the probabilities for the hybrids ranged from 0.13 to 0.31. Maine diploids (MM) had the highest probability of mortality estimates, ranging from 0.46 to 0.76, which were significantly different from the other crosses within each site (Figure 5).

The Virginia-Maine hybrids exhibited mid-parent heterosis. In terms of cumulative mortality, heterosis accounted for 16% lower mortality at RR, 11% at ND, 5% at PG, and 20% at OC. Heterosis in probability of mortality accounted for 0.16 lower probability at RR, 0.11 at ND, 0.05 at PG, and 0.21 at OC.

3.3 Pathology

All triploids sampled on May 24 at ND were processed for histological analysis because of the spike in mortality in VVV, VVM, and LLV in late spring. Infection prevalence was low for *H. nelsoni* (\leq

10%) in all three crosses and infection intensity in all but one oyster was light, with one light-to-moderate from an LLM individual (Table 2). No infections of *P. marinus* or *Haplosporidium costale* were detected, and no evidence of bacterial infections was observed in the cross section. Prevalence of chlamydia ranged from 5% to 29% among crosses. Prevalence of hemocytosis was low ($\leq 10\%$) with light intensity (Table 2).

Samples of MM from ND and RR were processed for histological analysis because of high mortality. MM individuals were selected from August, a period immediately preceding the largest increase in cumulative mortality at these sites. No infections of *H. nelson* or *H. costale* were detected, however, infections of *P. marinus* were common, with incidences of 66% from RR and 60% from ND. For RR, intensities ranged from rare (90% of infections) to moderate-heavy (10% of infections). For ND, intensities ranged from rare to heavy, with moderate infections most common (44%).

3.4 Live vs. Dead Triploids at ND: Shell Height

Data for shell height of live and dead triploids collected on June 7, 2016 at ND were normally distributed and met the assumptions of homogeneity of variance among factors. Residuals from the ANOVA were normally distributed and homoscedastic. The number of dead oysters in a bag ranged from 3 to 26 (Table 3). The interaction of status (live or dead) and cross was significant ($p < 0.05$). For only the LLV cross, median shell length of dead oysters (91 mm) differed by more than 5 mm than that in live oysters (83 mm) (Figure 6).

3.5 Shell Height

Mean shell height of each cross ranged from 22 mm to 27 mm prior to field deployment in June 2015. In February/March at the start of the experiment, mean shell height of crosses at all sites ranged from 44 mm to 82 mm (Figure 7). VVV had the greatest mean shell height at RR with $82 \text{ mm} \pm 2$ (standard error) and ND with $79 \text{ mm} \pm 2$, LLV had the greatest at PG with $63 \text{ mm} \pm 3$, and LLM had the

greatest at OC with 74 mm \pm 4. At all sites in February/March 2016, MM had the lowest mean shell height (RR: 60 mm \pm 2), (ND: 57 mm \pm 2), (PG: 44 mm \pm 1), (OC: 58 mm \pm 2).

In August of 2016, mean shell height among crosses at each site ranged from 61 mm to 91 mm, with LLM at ND having the greatest mean shell height and MM at PG having the lowest mean shell height (Figure 7). Shell height data for August 2016 was normally distributed and met the assumption of homogeneity of variance. The ANOVA revealed cross ($F=16.9$, $p< 0.05$) and site ($F=41.3$, $p< 0.05$) were significant, but the interaction was not significant ($F=1.5$, $p=0.07$) (Table 4). Least square means for shell height of each cross and each site were determined from the ANOVA (Table 4). Overall, LLM had the greatest mean shell height (84 mm), and MM had the smallest mean shell height (68 mm). For sites, oysters at RR had the greatest mean shell height (84 mm) and oysters at PG had the smallest mean shell height (71 mm).

Results from Tukey's HSD test for comparisons for RR minus other sites were 2 mm, 13 mm, and 6 mm for ND, PG, and OC, respectively. The 95% confidence intervals that did not contain zero were for RR vs. PG and RR vs. OC (Table 4).

3.6 Meat Weight

Overall, mean meat weight ranged from 1.1 g (MM, PG, March) to 12.4 g (LLM, ND, June) throughout the experiment (Figure 8). Variation in the range of meat weight was about the same and highest at RR (9.1 g) and ND (9.5 g), followed by OC (7.2 g), and lowest at PG (4.9 g).

In general, the difference in mean meat weight between diploids and triploids was greater in summer than spring (Figure 8). In early May, mean meat weight of VVV was greater than VV by 27% at RR, 15% at ND, 20% at OC, and 6% at PG. Mean meat weight substantially decreased in most of the diploid crosses from early May to August. Among all diploid crosses, mean meat weight decreased by 29% at RR, 36% at ND, 37% at OC, and 33% at PG. During the same time, mean meat weight among all triploid crosses at RR, ND, OC, and PG changed by +8%, -12%, +1%, and +27%, respectively. By August,

the mean meat weight of VVV was greater than of VV by 51% at RR, 80% at ND, 53% at OC and 24% at PG.

3.7 Environmental Conditions

Average daily temperatures ranged from -1.7 °C during winter to 33.4 °C in the summer. No major differences were observed in average daily temperature among the experimental sites from the available data (Figure 9). Some comparisons among sites were not possible due to gaps in the data (Figure 9). During the period of the spike in mortality at ND (May 10-June 27), average daily temperatures were within 1°C at all sites (Table 5).

The mean of average daily salinities was within 1 ppt at ND, PG, and OC (Figure 10). Average daily salinities ranged from 16.7 ppt to 19.5 ppt at ND, 16.3 ppt to 19.7 ppt at PG, and 16.4 ppt to 21.3 ppt at OC. Average daily salinity was lower for RR, ranging from 11.4 ppt to 16.3 ppt. A large decrease in salinity from 16.3 ppt to 12.6 ppt was measured July 15 to July 16, caused by a rainfall event.

Dissolved oxygen concentration, pH, chlorophyll *a*, and turbidity monitored from May 10 to July 12 at ND and PG are reported in Appendix Figures A.1, A.2, A.3, and A.4. ND had the lowest average daily concentration of dissolved oxygen in May but was never lower than 6 mg/L. Average daily pH ranged from 7.7 to 8.3 at both sites and was lowest at ND on May 31 (7.7) and lowest at PG on July 4 (7.7). The range of average daily concentration (mg/L) of chlorophyll *a* was higher at ND (2.3 to 8.0) than PG (1.0 to 3.2). Turbidity, in Nephelometric Turbidity Units, was higher on average at ND (5.8) than PG (1.1) for the time both had measurements (May 10 to June 27), although the maximum turbidity was greater at PG (115.2) than ND (75.3).

4. Discussion

4.1 Triploid Mortality

The primary objective in this study was to investigate the cause of the recurring mortality events of triploid *C. virginica*, or triploid mortality, in the lower Chesapeake Bay. Triploid mortality observed in

380 this project was expected to match reports from farmers in previous years (K. Hudson, pers. comm.,
381 Guévelou et al., 2019). Most reports indicated mortalities would occur in near market sized oysters in
382 late spring (May-June), many (>20%) would die within a matter of weeks, and then relatively little
383 mortality would occur throughout the summer.

384 A mortality event fitting the description of triploid mortality was observed only at Nandua Creek
385 (ND). Between March and the end of June at ND, mortality ranged from 15 to 29% in the triploid
386 crosses. There was a small increase in cumulative mortality in these crosses for the subsequent four
387 months of the trial (< 10%). This mortality event was mild compared to previous reports from some
388 farmers in 2014 (Guévelou et al., 2019) and 2019 (K. Hudson, pers. comm.), however the timing and
389 brief period of the event suggest it is a signature example of triploid mortality.

390 The onset of the mortality event at ND occurred between May 10 to May 24, when the mean of
391 the average daily temperature was approximately 20°C, and ended between June 7 and June 27, when
392 the mean of the average daily temperature was approximately 27°C. Temperature and salinity,
393 however, were similar between ND and the other nearby Eastern Shore sites throughout the field trial.
394 DO concentration and pH were similar between ND and PG, but triploid mortality obtained only at ND.
395 As Guévelou et al. (2019) suggested, environmental conditions implicated in summer mortality of *C.*
396 *gigas*, such as quantity of food available and stressors from the sediment (Samain and McCombie,
397 2008), are likely worthy candidates of investigation for future study into the etiology of triploid
398 mortality.

399 The results from this trial demonstrate that triploid mortality in the Chesapeake Bay involves
400 more than a specific cross or brood stock source. The severity of triploid mortality was measured as the
401 probability of mortality between February/March to the end of June. Despite their heterogeneous
402 genetic lineage, the severity of triploid mortality among VVV, VVM, LLV, and LLM was statistically
403 indistinguishable at the only site with substantial mortality in the spring (>20%). Guévelou et al. (2019)

also measured mortality among different triploid crosses in late spring in the Chesapeake Bay, however the triploids differed only by the source of diploid brood stock. Guévelou et al. (2019) reported triploids made from a Maine diploid, or “Northern cross,” tended to have the lowest mortality across sites and triploids made from DEBY tended to have the highest. The test of genetic lineage in Guévelou et al. (2019), however, was not ideal because the crosses were not controlled, the seed were obtained from multiple commercial hatcheries, and the seed varied in age. Through simultaneous, controlled crosses, the present study found that neither a genetically distinct source of diploid brood stock (V or M) nor genetically distinct source of tetraploid brood stock (VV or LL) significantly affected the severity of triploid mortality.

Like Guévelou et al. (2019), the present study found no support for the hypothesis that genes passed by the “Northern diploid” are a cause of triploid mortality. Some of the other findings of Guévelou et al. (2019) are similar to this study: observed mortality on commercial sites on the bay-side of the Eastern Shore that matched previous reports of triploid mortality, found no relationship between mortalities and infections from *H. nelsoni* or *P. marinus*, and found, within crosses, the largest oysters were not especially susceptible to mortality in late spring.

The current study controlled for many of the factors that confounded the “rapid institutional response” reported by Guévelou et al. (2019). Still, the study presented here had its own design limitations. Each cross in the current study was produced by a single mass spawn. Replication of these spawns with multiple independent crosses would have provided more precise estimation of a brood stock effect (e.g. Dégremont et al. 2012). Regarding genetic effects, future research on triploid mortality should shift the focus from mass spawn studies to estimating variance in groups of siblings, or families, in order to possibly uncover a genetic basis for susceptibility to triploid mortality. High heritability for survival related to summer mortality has been observed in *C. gigas* (Dégremont et al., 2007).

Is “triploid mortality” truly *triploid* mortality? There has been scant data for mortality rates in diploids during a triploid mortality event. Results from Guévelou et al. (2019) were ambiguous: between the two sites with the signature triploid mortality pattern, Nassawadox Creek, diploids and triploids had similar mortality at Nassawadox Creek, while diploids had much lower mortality than triploids at Pungoteague Creek. In this study, where only one site had the signature triploid mortality, only the triploids were affected. The VV cross had less than 2% mortality during the triploid mortality event at ND. Clearly, the conditions causing the triploid mortality event at ND did not cause a similar mortality event in the diploid control group. From these results, we hypothesize that conditions inducing triploid mortality events in the Chesapeake Bay do not cause mass mortality in diploids and suggest classifying the events as “triploid mortality” is justified. Additional data from controlled field tests involving a greater variety of sites and diploids, and in which triploid mortality events occur, are required to further evaluate this hypothesis.

High mortality in triploid oysters on farms in the Southeastern United States have invited comparisons with triploid mortality in the Chesapeake Bay. In 2016, oyster farmers in Alabama and Louisiana observed unexpectedly high mortality in triploid stocks (Wadsworth et al., 2019). The mortality was investigated in a follow-up study by Wadsworth et al. (2019) in which the survival of diploids and triploids was tracked at commercial sites. Wadsworth et al. (2019) found high mortality at several of the sites and higher cumulative mortality in the triploids than diploids at most sites.

It is not clear if mortalities reported in Wadsworth et al. (2019) represent triploid mortality as it has been defined in the Chesapeake Bay. A major difference is that Wadsworth et al. (2019) recorded well-known stressors around the time of mortality events. Sudden drops in salinity, periods of very low salinity (< 5 ppt) and high temperature (> 30°C), and strong disease pressure from *P. marinus* were associated with mortality in Wadsworth et al. (2019a) , all of which have been absent in the recent investigations in triploid mortality in the lower Chesapeake Bay. Instead of classic “triploid mortality” as

defined in the Chesapeake Bay, differential mortality in Wadsworth et al. (2019) could have been due to the diploids having higher resistance to stress from low salinity and *P. marinus*.

Susceptibility to triploid mortality, like susceptibility to mortality owing to pathogens such as OsHV-1 μ var (Azéma et al., 2016; Petton et al., 2015), *Vibrio aestuarianus* (Azéma et al., 2016), and *Roseovarius crassostreae* (Bricelj et al., 1992; Ford and Borrero, 2001), likely varies based on ontogeny. Reports from farmers have suggested that while market or near-market oysters die during triploid mortality events, juvenile oysters or seed have low mortality. For this reason, only oysters older than one year were used during the field trials in Guévelou et al. (2019) and the current study. The factors involved in this ontogeny-based susceptibility are not clear, however one explanation is that susceptibility is positively related with reproductive effort. Reproductive effort, which has been linked to vulnerability to bacterial infections (De Decker et al., 2011; Wendling and Wegner, 2013) and “summer mortality” in *C. gigas* (e.g. Koganezawa, 1975; Samain and McCombie, 2008), is expected to increase with age in oysters (Bayne, 2017). Although triploid oysters are considered reproductively sterile, they may expend considerable reproductive effort during gametogenesis (Allen and Downing, 1990; Jouaux et al., 2010).

The closest analogy for triploid mortality in the Chesapeake Bay may be “summer mortality” observed in diploid *C. gigas*. Summer mortality episodes have regularly occurred in Japan (Koganezawa, 1975; Takeuchi et al., 1960; Ventilla, 1984), Western Europe (Cotter et al., 2010; Goulletquer et al., 1998; Parache, 1989; Samain and McCombie, 2008; Watermann et al., 2008), and the west coast of the US (Cheney, 2000; Glude, 1975; Perdue et al., 1981), often resulting in high crop losses (> 50%) of adult diploid *C. gigas* during the summer months. Like triploid mortality, many cases occur within a period of several weeks (e.g. Cheney 2000, Ropert et al. 2008, Cotter et al. 2010), and are not typically explained by extreme environmental conditions or intense disease pressure (Cheney, 2000; Glude, 1975; Koganezawa, 1975; Ropert et al., 2008).

The important overlap between summer mortality and triploid mortality may be in their generalized etiology. Summer mortality was originally suspected to be due to a “physiological disorder” that arose when warm temperatures and eutrophic conditions led to extensive gonad formation (Koganezawa, 1975), and additional research has repeatedly associated physiological processes of gametogenesis and spawning with summer mortality (Cotter et al., 2010; Perdue et al., 1981; Samain et al., 2007; Samain and McCombie, 2008). Physiological dysfunction related to gametogenesis and spawning may also be causing triploid mortality. Triploid mortality consistently occurs in late spring, when the physiology of oysters in the Chesapeake Bay is expected to be strongly influenced by gametogenesis and spawning. In oysters, it has been well documented that gametogenesis and spawning are also strongly influenced by environmental conditions (Delaporte et al., 2006; Dutertre et al., 2010, 2009; Liu et al., 2010) that can explain the strong site and inter-annual effects of summer mortality (Glude, 1975; Ropert et al., 2008) and triploid mortality.

A connection between gametogenesis and mortality in triploid *C. virginica* has been recently examined. Guévelou et al. (2019) raised the question of whether “the degree of triploid gametogenesis and reproduction” effects mortality, and Wadsworth et al. (2019) examined gonad development as a potential factor to explain mortality in triploids. Both found no evidence for an association between gonad development in triploids and mortality, however they both primarily classified gonad development using methods developed for diploid oysters. Triploid oysters have much different gonad development than diploids (Allen and Downing, 1990; Barber and Mann, 1991; Guévelou et al., 2019; Jouaux et al., 2010; Lee, 1988), and thus precision to detect differences may be lacking in a classification system developed for diploids. Further study in triploid mortality would benefit from a more specific classification system for gonad development in triploid *C. virginica*, which may reach the precision needed to detect a relationship between gametogenesis and susceptibility.

4.2 Mortality in Virginia and Maine Diploids

The control for triploid mortality in this study was the diploid DEBY (VV) cross. Diploids with a Maine genetic origin (VM, MV, and MM) were produced as an experiment. The VV cross exhibited low mortality while the Virginia-Maine hybrids and Maine diploids had significantly higher mortality. MM had especially high mortality and the most cumulative mortality at all sites by a wide margin (>20%).

In general, mortality in diploids with a Maine genetic origin was gradual and most of the mortality occurred in late summer. Gradual mortality over the summer is not unusual for oysters in the Chesapeake Bay, in part because disease pressure from *P. marinus* is expected to increase over summer and near a maximum by late summer (Burreson and Ragone Calvo, 1996). Diploids with Maine genetic origin were expected to be susceptible to *P. marinus* and have general unsuitability to warm estuarine conditions of the Chesapeake Bay. The Maine brood stock used in this study had been selected in a region with higher salinity, lower temperatures, and less *P. marinus* pressure than the sites in this study (e.g. Proestou et al., 2016). Many of the MM diploids (60 to 66%) were infected with *P. marinus* when examined in August, suggesting *P. marinus* may have been responsible for at least some of the mortality in the diploids with Maine genetic origin.

For mortality among the diploid constructs, Virginia-Maine reciprocal hybrids were statistically indistinguishable within sites. Differences between reciprocal hybrids have been observed in shellfish in the larval stage and may often be attributable to variance in egg quality (Cruz and Ibarra, 1997; Deng et al., 2005). Some studies, however, have also observed reciprocal effects in adults or juveniles, such as yield in *C. gigas* (Hedgecock and Davis, 2007), growth in the disk abalone *Haliotis discus hannai* (Deng et al., 2010), and growth and survival in the Atlantic deep-sea scallop *Placopecten magellanicus* (Wang and Côté, 2012).

The hybrids exhibited heterosis for mortality at all sites. Heterosis was defined by the deviation of the mean of the reciprocals from the mean of their parental populations (Falconer, 1981), often

referred to as mid-parent heterosis. Heterosis was for higher survival, in the direction of the selectively bred VV line, or DEBY line. DEBY has been selected for resistance to the major diseases in the Chesapeake Bay since about 1987, or about 14 generations (Ragone Calvo et al., 2003). Our hypothesis is that disease resistance from the DEBY contributes more than just an additive effect, perhaps by improving the overall health of the hybrid beyond fending off disease, per se. Callam et al. (2013) showed that triploids made from disease resistant tetraploids resulted in nearly completely resistant progeny no matter the source of the diploid parent, be it wild or selectively bred, also suggesting effects other than the additivity of disease resistance.

4.2 Growth & Condition

Shell height was a minor source of variation among crosses in this field trial and was similar in diploids and triploids. The variation can be attributed largely to the MM cross, the slowest growing cross by a wide margin. The most direct comparison between triploid and diploids was afforded by VVV versus VV crosses because of the similarity of pedigree between the tetraploid and diploid parents that comprised the triploid: mean shell height of VVV was only 3 mm greater than that of VV by August 2016 (18 months post-spawn).

Triploid *Crassostrea* spp. generally seem to grow substantially faster than their diploid counterparts based on shell height (Dégremont et al., 2012; Matthiessen and Davis, 1992; Nell and Perkins, 2005; Qin et al., 2019), however, several studies, including the current one, observed only a modest difference (Barber and Mann, 1991; Callam et al., 2016; Hand et al., 1998; Stone et al., 2013). Some authors have suggested the growth advantage in triploid oysters depends on the favorability of the growing conditions (Davis, 1994; Guo et al., 2009). Callam et al. (2016) suggested stress from low salinity inhibited triploids from growing faster than diploids at a low salinity site and drew a parallel with results from Davis (1994), which found *C. gigas* triploids only performed better than diploids at more

productive sites. This explanation, however, does not account for the small difference in shell height between triploids and diploids from selectively bred lines at the more mesohaline site in Callam et al. (2016), and contrasts with findings from Ibarra et al. (2017). Ibarra et al. (2017) found mated triploid *C. gigas* grew substantially faster than diploids at two tropical sites, characterized by low productivity and high temperatures, but not at a more productive temperate site. In this study, although PG seemed to be a less favorable growing site, there was only a small difference (> 10 mm) in mean shell height between VVV and VV at all sites by August.

Meat weight provided a clearer distinction between diploids and triploids in this field trial. Meat weight fluctuates during the year due to physiological changes in the oyster, such as, gametogenesis, spawning, and rebuilding of biochemical reserves (Loosanoff, 1942; Thompson et al., 1996). Mean meat weight decreased sharply in the diploids between May and August, which is likely attributable to spawning, which in *C. virginica* generally occurs between May and September in the Chesapeake Bay (Kennedy and Krantz, 1982). For triploids, mean meat weight was stable during this time, which demonstrates triploid *C. virginica* maintain high meat quality during the spawning season similar to previous findings in *C. gigas* (Allen and Downing 1986, Allen, 1988).

Despite the maintenance of body weight in triploids, spawning activity cannot be ruled out. Significant variation in fecundity of triploids has been observed (Allen and Downing, 1990; Jouaux et al., 2010), with some triploids producing significant amounts of gametes and spawning concurrently with diploids, while other triploids are virtually sterile. Allen and Downing (1990, 1986) found triploid *C. gigas* to have stable mean meat weight during the spawning interval for diploids, despite histological evidence that some triploids spawned. Normand et al. (2008) also inferred that triploid *C. gigas* spawned within the same time interval as diploids based on histological evidence. In their study, dry meat weight decreased substantially in triploids in parallel with diploids. Histological assessment is required to more accurately estimate the gonad development and spawning activity in the triploids.

Acknowledgements

Special thanks to Chris Rubino for his extensive efforts in the field and lab. Thank you to the following members of the Aquaculture Genetics and Technology Center for their assistance in the hatchery, field, and lab: Amanda Chesler, Nate Geyerhahn, Lauren Gregg, Kate Ritter, Karen Sisler, and Joana Sousa. Thanks to the following VIMS associates: Rita Crocket and Ryan Carnegie for histopathology, Cody Lapnow, Hunter Walker, and Mark Brush for assistance with the YSI sondes, and Sean Fate and Eddy Smith for assisting in the field work. Thank you to the following industry collaborators who volunteered their time and resources: Mike Congrove, Bob Boardman, Lee-Ann Fick, Matt May, Gerry Negley, and Wade Walker. This work was supported by Virginia Sea Grant and the National Oceanic and Atmospheric Administration (award number NA14OAR4170093) and the ABC 4.0 initiative supported by an anonymous donor through Fidelity Charitable Trust. The authors declare they have no conflict of interest. This is contribution No. xxxx of the Virginia Institute of Marine Science, College of William & Mary.

References

- Allen, Jr., S.K., 1988. Triploid Oysters Ensure Year-round Supply. *Oceanus* 31, 58–63.
- Allen, Jr., S.K., Bushek, D., 1992. Large-scale production of triploid oysters, *Crassostrea virginica* (Gmelin), using “stripped” gametes. *Aquaculture* 103, 241–251. [https://doi.org/10.1016/0044-8486\(92\)90170-P](https://doi.org/10.1016/0044-8486(92)90170-P)
- Allen, Jr., S.K., Downing, S.L., 1990. Performance of Triploid Pacific Oysters, *Crassostrea gigas*: Gametogenesis. *Can. J. Fish. Aquat. Sci.* 47, 1213–1222.
- Allen, Jr., S.K., Downing, S.L., 1986. Performance of triploid Pacific oysters, *Crassostrea gigas* (Thunberg). I. Survival, growth, glycogen content, and sexual maturation in yearlings. *J. Exp. Mar. Bio. Ecol.* 102, 197–208.
- Azéma, P., Travers, M.A., Benabdelmouna, A., Dégremont, L., 2016. Single or dual experimental infections with *Vibrio aestuarianus* and OsHV-1 in diploid and triploid *Crassostrea gigas* at the spat, juvenile and adult stages. *J. Invertebr. Pathol.* 139, 92–101. <https://doi.org/10.1016/j.jip.2016.08.002>
- Barber, B.J., Mann, R., 1991. Sterile triploid *Crassostrea virginica* (Gmelin, 1791) grow faster than diploids but are equally susceptible to *Perkinsus marinus*. *J. Shellfish Res.* 10, 445–450.
- Bates, D., Mächler, M., Bolker, B., Walker, S., 2015. Fitting linear mixed-effects models using {lme4}. *J. Stat. Softw.* 67, 1–48. <https://doi.org/10.18637/jss.v067.i01>

- Bayne, B.L., 2017. Biology of Oysters. Academic Press.
- Bricelj, V.M., Ford, S.E., Borrero, F.J., Perkins, F.O., Rivara, G., Hillman, R.E., Elston, R.A., Chang, J., Brook, S., York, N., Norris, P., Point, G., Street, W., Road, W.S.B., 1992. Unexplained mortalities of hatchery-reared, juvenile oysters, *Crassostrea virginica* (Gmelin). J. Shellfish Res. 11, 331–347.
- Burreson, E.M., Ragone Calvo, L., 1996. Epizootiology of *Perkinsus marinus* disease of oysters in Chesapeake Bay, with emphasis on data since 1985. J. Shellfish Res. 15, 17–34.
- Callam, B.R., 2013. Improvements in Triploid *Crassostrea virginica* Production: Characterizing the Diploid Parent. College of William and Mary.
- Callam, B.R., Allen, Jr., S.K., Frank-Lawale, A., 2016. Genetic and environmental influence on triploid *Crassostrea virginica* grown in Chesapeake Bay: Growth. Aquaculture 452, 97–106.
<https://doi.org/10.1016/j.aquaculture.2015.10.027>
- Camara, M.D., Yen, S., Kaspar, H.F., Kesarcodi-Watson, A., King, N., Jeffs, A.G., Tremblay, L.A., 2017. Assessment of heat shock and laboratory virus challenges to selectively breed for ostreid herpesvirus 1 (OsHV-1) resistance in the Pacific oyster, *Crassostrea gigas*. Aquaculture 469, 50–58.
<https://doi.org/10.1016/j.aquaculture.2016.11.031>
- Carnegie, R.B., Burreson, E.M., 2011. Declining impact of an introduced pathogen: Haplosporidium nelsoni in the oyster *Crassostrea virginica* in Chesapeake Bay. Mar. Ecol. Prog. Ser. 432, 1–15.
- Cheney, D.P., 2000. Summer mortality of Pacific oysters, *Crassostrea gigas* (Thunberg): initial finding on multiple environmental stressors in Puget Sound, Washington, 1998. J. Shellfish Res. 19, 353–359.
- Cotter, E., Malham, S.K., O’Keeffe, S., Lynch, S.A., Latchford, J.W., King, J.W., Beaumont, A.R., Culloty, S.C., 2010. Summer mortality of the Pacific oyster, *Crassostrea gigas*, in the Irish Sea: The influence of growth, biochemistry and gametogenesis. Aquaculture 303, 8–21.
<https://doi.org/10.1016/j.aquaculture.2010.02.030>
- Cruz, P., Ibarra, A.M., 1997. Larval growth and survival of two catarina scallop (*Argopecten circularis*, Sowerby, 1835) populations and their reciprocal crosses. J. Exp. Mar. Bio. Ecol. 212, 95–110.
[https://doi.org/10.1016/S0022-0981\(96\)02742-6](https://doi.org/10.1016/S0022-0981(96)02742-6)
- Davis, J.P., 1994. Studies on the influence of ambient temperature and food supply on growth rate, carbohydrate content and reproductive output in diploid and triploid Pacific oysters, *Crassostrea gigas* (Thunberg). <https://doi.org/10.16953/deusbed.74839>
- De Decker, S., Normand, J., Saulnier, D., Pernet, F., Castagnet, S., Boudry, P., 2011. Responses of diploid and triploid Pacific oysters *Crassostrea gigas* to *Vibrio* infection in relation to their reproductive status. J. Invertebr. Pathol. 106, 179–191. <https://doi.org/10.1016/j.jip.2010.09.003>
- Dégremont, L., Ernande, B., Bédier, E., Boudry, P., 2007. Summer mortality of hatchery-produced Pacific oyster spat (*Crassostrea gigas*). I. Estimation of genetic parameters for survival and growth. Aquaculture 262, 41–53. <https://doi.org/10.1016/j.aquaculture.2006.10.025>
- Dégremont, L., Nourry, M., Maurouard, E., 2015. Mass selection for survival and resistance to OsHV-1 infection in *Crassostrea gigas* spat in field conditions: Response to selection after four generations. Aquaculture 446, 111–121. <https://doi.org/10.1016/j.aquaculture.2015.04.029>
- Dégremont, L., Garcia, C., Frank-Lawale, A., Allen, Jr., S.K., 2012. Triploid oysters in the Chesapeake Bay: comparison of diploid and triploid *Crassostrea virginica*. J. Shellfish Res. 31, 21–31.
<https://doi.org/10.2983/035.031.0103>
- Dégremont, L., Garcia, C., Allen, Jr., S.K., 2015. Genetic improvement for disease resistance in oysters: A review. J. Invertebr. Pathol. 131, 226–241. <https://doi.org/10.1016/j.jip.2015.05.010>
- Delaporte, M., Soudant, P., Lambert, C., Moal, J., Pouvreau, S., Samain, J.F., 2006. Impact of food availability on energy storage and defense related hemocyte parameters of the Pacific oyster *Crassostrea gigas* during an experimental reproductive cycle. Aquaculture 254, 571–582.
<https://doi.org/10.1016/j.aquaculture.2005.10.006>
- Deng, Y., Liu, X., Zhang, G., Guo, X., 2005. Inbreeding depression and maternal effects on early

- performance of Pacific Abalone. N. Am. J. Aquac. 67, 231–236. <https://doi.org/10.1577/a04-021.1>
- Deng, Y., Liu, X., Zhang, G., Wu, F., 2010. Heterosis and combining ability: A diallel cross of three geographically isolated populations of Pacific abalone *Haliotis discus hannai* Ino. Chinese J. Oceanol. Limnol. 28, 1195–1199. <https://doi.org/10.1007/s00343-010-9903-7>
- Dutertre, M., Beninger, P.G., Barillé, L., Papin, M., Haure, J., 2010. Rising water temperatures, reproduction and recruitment of an invasive oyster, *Crassostrea gigas*, on the French Atlantic coast. Mar. Environ. Res. 69, 1–9. <https://doi.org/10.1016/j.marenvres.2009.07.002>
- Dutertre, M., Beninger, P.G., Barillé, L., Papin, M., Rosa, P., Barillé, A.L., Haure, J., 2009. Temperature and seston quantity and quality effects on field reproduction of farmed oysters, *Crassostrea gigas*, in Bourgneuf Bay, France. Aquat. Living Resour. 22, 319–329. <https://doi.org/10.1051/alr/2009042>
- Falconer, D.S., 1981. Introduction to quantitative genetics., Introduction to quantitative genetics. Longmans Green, London/New York.
- Ford, S.E., Borrero, F.J., 2001. Epizootiology and pathology of juvenile oyster disease in the Eastern Oyster, *Crassostrea virginica*. <https://doi.org/10.1006/jipa.2001.5052>
- Frank-Lawale, A., Allen, Jr., S.K., Dégremont, L., 2014. Breeding and Domestication of Eastern Oyster (*Crassostrea virginica*) Lines for Culture in the Mid-Atlantic, Usa: Line Development and Mass Selection for Disease Resistance. J. Shellfish Res. 33, 153–165. <https://doi.org/10.2983/035.033.0115>
- Glude, J.B., 1975. A summary report of Pacific coast oyster mortality investigations 1965–1972, in: Proceedings of the Third US–Japan Meeting on Aquaculture at Tokyo, Japan. pp. 1–28.
- Goulletquer, P., Soletchnik, P., Le Moine, O., Razet, D., Geairon, P., Faury, N., 1998. Summer mortality of the Pacific cupped oyster *Crassostrea gigas* in the Bay of Marennes-Oléron (France), in: CIEM Conseil International Pour l’Exploration de La Mer.
- Guévelou, E., Carnegie, R.B., Small, J.M., Hudson, K., Reece, K.S., Rybovich, M.M., Allen Jr., S.K., 2019. Tracking Triploid Mortalities of Eastern Oysters *Crassostrea virginica* in the Virginia Portion of the Chesapeake Bay. J. Shellfish Res. 38, 101–113.
- Guo, X., Wang, Y., Xu, Z., Yang, H., 2009. Chromosome set manipulation in shellfish, in: New Technologies in Aquaculture: Improving Production Efficiency, Quality and Environmental Management. Woodhead Publishing, pp. 165–194. <https://doi.org/10.1533/9781845696474.1.165>
- Hand, R.E., Nell, J.A., Maguire, G.B., 1998. Studies on triploid oysters in Australia . X . Growth and mortality of diploid and triploid Sydney rock oysters *Saccostrea commercialis*. J. Shellfish Res. 17, 1115–1127.
- Hudson, K., 2018. Virginia Shellfish Aquaculture Situation and Outlook Report: Results of the 2017 Virginia Shellfish Aquaculture Crop Reporting Survey. Reports. <https://doi.org/http://dx.doi.org/doi:10.21220/m2-qt5c-sk02>
- Hudson, K., 2017. Virginia Shellfish Aquaculture Situation and Outlook Report : Results of the 2016 Virginia Shellfish Aquaculture Crop Reporting Survey. <https://doi.org/https://doi.org/10.21220/V51K6T>
- Hudson, K., Murray, T., 2016. Virginia Shellfish Aquaculture Situation and Outlook Report : Results of the 2015 Virginia Shellfish Aquaculture Crop Reporting Survey. Reports. <https://doi.org/https://doi.org/10.21220/V5BD8N>
- Hudson, K., Murray, T., 2015. Virginia Shellfish Aquaculture Situation and Outlook Report : Results of the 2014 Virginia Shellfish Aquaculture Crop Reporting Survey. Reports. <https://doi.org/https://doi.org/10.21220/V56Q74>
- Hudson, K., Murray, T., 2014. Virginia Shellfish Aquaculture Situation and Outlook Report : Results of the 2013 Virginia Shellfish Aquaculture Crop Reporting Survey. Reports. <https://doi.org/https://doi.org/10.21220/V52X4T>
- Ibarra, A.M., Ascencio-Michel, R., Ramírez, J.L., Manzano-Sarabia, M., Rodríguez-Jaramillo, C., 2017.

- Performance of Diploid and Triploid *Crassostrea gigas* (Thunberg, 1793) Grown in Tropical Versus Temperate Natural Environmental Conditions . J. Shellfish Res. 36, 119–139.
<https://doi.org/10.2983/035.036.0113>
- Jouaux, A., Heude-Berthelin, C., Sourdain, P., Mathieu, M., Kellner, K., 2010. Gametogenic stages in triploid oysters *Crassostrea gigas*: Irregular locking of gonial proliferation and subsequent reproductive effort. J. Exp. Mar. Bio. Ecol. 395, 162–170.
<https://doi.org/10.1016/j.jembe.2010.08.030>
- Keeling, S.E., Brosnahan, C.L., Williams, R., Gias, E., Hannah, M., Bueno, R., McDonald, W.L., Johnston, C., 2014. New Zealand juvenile oyster mortality associated with ostreid herpesvirus 1-an opportunistic longitudinal study. Dis. Aquat. Organ. 109, 231–239. <https://doi.org/10.3354/dao02735>
- Kennedy, V., Krantz, L., 1982. Comparative gametogenic and spawning patterns of the oyster *Crassostrea virginica* (Gmelin) in central Chesapeake Bay. J. Shellfish Res. 2, 133–140.
- Koganezawa, A., 1975. Present status of studies on the mass mortality of cultured oyster in Japan and its prevention, in: Proceedings of the Third US-Japan Meeting on Aquaculture. pp. 29–34.
- Kube, P., Dove, M., Cunningham, M., Kirkland, P., Gu, X., Hick, P., O'Connor, W., Elliot, N., 2018. Genetic selection for resistance to Pacific Oyster Mortality Syndrome. CSIRO Agriculture and Food, NSW Department of Primary Industries, Australian Seafood Industries, Seafood CRC, and FRDC.
- Lee, M.M., 1988. Abnormal gametogenesis in triploid American oysters *Crassostrea virginica*. J. Shellfish Res. 7, 201–202.
- Liu, W., Li, Q., Gao, F., Kong, L., 2010. Effect of starvation on biochemical composition and gametogenesis in the Pacific oyster *Crassostrea gigas*. Fish. Sci. 76, 737–745.
<https://doi.org/10.1007/s12562-010-0274-y>
- Loosanoff, V.L., 1942. Seasonal gonadal changes in the adult oysters, *Ostrea virginica*, of Long Island Sound. Biol. Bull. 82, 195–206.
- Mann, R., Southworth, M., Carnegie, R.B., Crockett, R.K., 2014. Temporal Variation in Fecundity and Spawning in the Eastern Oyster, *Crassostrea virginica* , in the Piankatank River, Virginia . J. Shellfish Res. 33, 167–176. <https://doi.org/10.2983/035.033.0116>
- Matthiessen, G.C., Davis, J.P., 1992. Observations on growth rate and resistance to MSX (Haplosporidium nelsoni) among diploid and triploid eastern oysters (*Crassostrea virginica* (Gmelin, 1791) in New England. J. Shellfish Res. 11, 449.
- Maurer, D., Comps, M., 1986. Mortalités estivales de l'huître *Crassostrea gigas* dans le bassin d'Arcachon : facteurs du milieu, aspects biochimiques et histologiques. Pathol. Mar. Aquac. 29–41.
- Murray, T., Hudson, K., 2013. Virginia Shellfish Aquaculture Situation and Outlook Report : Results of the 2012 Virginia Shellfish Aquaculture Crop Reporting Survey. Reports.
<https://doi.org/https://doi.org/10.21220/V5Z71V>
- Murray, T., Hudson, K., 2012. Virginia Shellfish Aquaculture Situation and Outlook Report : Results of the 2011 Virginia Shellfish Aquaculture Crop Reporting Survey. Reports.
<https://doi.org/https://doi.org/10.21220/V5PQ7G>
- Murray, T., Hudson, K., 2011. Virginia Shellfish Aquaculture Situation and Outlook Report : Results of the 2010 Virginia Shellfish Aquaculture Crop Reporting Survey. Reports.
<https://doi.org/https://doi.org/10.21220/V5TD80>
- Murray, T., Oesterling, M., 2010. Virginia Shellfish Aquaculture Situation and Outlook Report : Results of the 2009 Virginia Shellfish Aquaculture Crop Reporting Survey. Reports.
<https://doi.org/https://doi.org/10.21220/V5JX45>
- Murray, T., Oesterling, M., 2009. Virginia Shellfish Aquaculture Situation and Outlook Report : Results of the 2008 Virginia Shellfish Aquaculture Crop Reporting Survey. Reports.
<https://doi.org/https://doi.org/10.21220/V5F716>
- Murray, T., Oesterling, M., 2006. Virginia Shellfish Aquaculture Situation and Outlook Report : Results of

- the 2004-2006 Virginia Shellfish Aquaculture Crop Reporting Survey. Reports.
<https://doi.org/https://doi.org/10.21220/V52123>
- Nell, J.A., Perkins, B., 2005. Studies on triploid oysters in Australia: farming potential of all-triploid Pacific oysters, *Crassostrea gigas* (Thunberg), in Port Stephens, New South Wales, Australia. Aquac. Res. 36, 530–536. <https://doi.org/10.1111/j.1365-2109.2005.01229.x>
- Normand, J., Le Pennec, M., Boudry, P., 2008. Comparative histological study of gametogenesis in diploid and triploid Pacific oysters (*Crassostrea gigas*) reared in an estuarine farming site in France during the 2003 heatwave. Aquaculture 282, 124–129.
<https://doi.org/10.1016/j.aquaculture.2008.06.026>
- Parache, A., 1989. Growth performance of oyster *Crassostrea angulata* and *Crassostrea gigas* reared in Arcachon Bay between 1950 and 1986: First results. Haliotis 19, 227–236.
- Paul-Pont, I., Evans, O., Dhand, N.K., Rubio, A., Coad, P., Whittington, R.J., 2014. Descriptive epidemiology of mass mortality due to Ostreid herpesvirus-1 (OsHV-1) in commercially farmed Pacific oysters (*Crassostrea gigas*) in the Hawkesbury River estuary, Australia. Aquaculture 422–423, 146–159. <https://doi.org/10.1016/j.aquaculture.2013.12.009>
- Peeler, E.J., Allan Reese, R., Cheslett, D.L., Geoghegan, F., Power, A., Thrush, M.A., 2012. Investigation of mortality in Pacific oysters associated with Ostreid herpesvirus-1 μ Var in the Republic of Ireland in 2009. Prev. Vet. Med. 105, 136–143. <https://doi.org/10.1016/j.prevetmed.2012.02.001>
- Perdue, J.A., Beattie, J.H., Chew, K.K., 1981. Some relationships between gametogenic cycle and summer mortality phenomenon in the Pacific oyster (*Crassostrea gigas*) in Washington State. J. Shellfish Res. 1, 1–10.
- Petton, B., Boudry, P., Alunno-Bruscia, M., Pernet, F., 2015. Factors influencing disease-induced mortality of Pacific oysters *Crassostrea gigas*. Aquac. Environ. Interact. 6, 205–222.
<https://doi.org/10.3354/aei00125>
- Proestou, D.A., Vinyard, B.T., Corbett, R.J., Piesz, J., Allen, Jr., S.K., Small, J.M., Li, C., Liu, M., DeBrosse, G., Guo, X., Rawson, P., Gómez-Chiarri, M., 2016. Performance of selectively-bred lines of eastern oyster, *Crassostrea virginica*, across eastern US estuaries. Aquaculture 464, 17–27.
<https://doi.org/10.1016/j.aquaculture.2016.06.012>
- Purdom, C.E., 1972. Induced polyploidy in plaice (*Pleuronectes platessa*) and its hybrid with the flounder (*Platichthys flesus*). Heredity (Edinb). 29, 11–24.
- Qin, Y., Zhang, Yuehuan, Mo, R., Zhang, Yang, Li, J., Zhou, Y., Ma, H., Xiao, S., Yu, Z., 2019. Influence of ploidy and environment on grow-out traits of diploid and triploid Hong Kong oysters *Crassostrea hongkongensis* in southern China. Aquaculture 507, 108–118.
<https://doi.org/10.1016/J.AQUACULTURE.2019.04.017>
- R Core Team, 2019. R: A Language and Environment for Statistical Computing.
- Ragone Calvo, L.M., Calvo, G.W., Burrenson, E.M., 2003. Dual disease resistance in a selectively bred eastern oyster, *Crassostrea virginica*, strain tested in Chesapeake Bay. Aquaculture 220, 69–87.
[https://doi.org/10.1016/S0044-8486\(02\)00399-X](https://doi.org/10.1016/S0044-8486(02)00399-X)
- Ropert, M., Mazurie, J., Bedier, E., Le Coz, F., Soletchnik, P., 2008. Evaluation of summer mortality risk factors in shellfish farming ecosystems, in: Samain, J.F., McCombie, H. (Eds.), Summer Mortality of the Pacific Oyster *Crassostrea Gigas* - The Morest Project. pp. 1–61.
- Samain, J.F., Dégremont, L., Soletchnik, P., Haure, J., Bédier, E., Ropert, M., Moal, J., Huvet, A., Bacca, H., Van Wormhoudt, A., Delaporte, M., Costil, K., Pouvreau, S., Lambert, C., Boulo, V., Soudant, P., Nicolas, J.L., Le Roux, F., Renault, T., Gagnaire, B., Geret, F., Boutet, I., Burgeot, T., Boudry, P., 2007. Genetically based resistance to summer mortality in the Pacific oyster (*Crassostrea gigas*) and its relationship with physiological, immunological characteristics and infection processes. Aquaculture 268, 227–243. <https://doi.org/10.1016/j.aquaculture.2007.04.044>
- Samain, J.F., McCombie, H., 2008. Summer mortality of Pacific oyster *Crassostrea gigas*: the Morest

- project. Eds Quae, Versailles, France.
- Segarra, A., Pépin, J.F., Arzul, I., Morga, B., Faury, N., Renault, T., 2010. Detection and description of a particular Ostreid herpesvirus 1 genotype associated with massive mortality outbreaks of Pacific oysters, *Crassostrea gigas*, in France in 2008. *Virus Res.* 153, 92–99. <https://doi.org/10.1016/j.virusres.2010.07.011>
- Stanley, J.G., Allen Jr, S.K., Hidu, H., 1981. Polyploidy induced in the American oyster, *Crassostrea virginica*, with cytochalasin B. *Aquaculture* 23, 1–10.
- Stone, B.W., Hadley, N.H., Kingsley-Smith, P.R., 2013. Evaluating the Potential Growth Advantage of Triploid Eastern Oysters (*Crassostrea virginica*) in South Carolina Relative to Commercially Cultured Diploid Native Stocks. *J. Shellfish Res.* 32, 647–655. <https://doi.org/10.2983/035.032.0304>
- Takeuchi, T., Takemoto, Y., Matsubara, T., 1960. Haematological study of bacterial affected oysters, Report of Hiroshima Prefectural Fisheries Experimental Station.
- Thompson, R.J., Newell, R.I.E., Kennedy, V.S., Mann, R., 1996. Reproductive processes and early development. *East. oyster Crassostrea virginica* 335–370.
- Ventilla, R.F., 1984. Recent developments in the Japanese oyster culture industry. *Adv. Mar. Biol.* 21, 1–57.
- Wadsworth, P., Casas, S., La Peyre, J., Walton, W., 2019. Elevated mortalities of triploid eastern oysters cultured off-bottom in northern Gulf of Mexico. *Aquaculture* 505, 363–373. <https://doi.org/10.1016/j.aquaculture.2019.02.068>
- Wang, C., Côté, J., 2012. Heterosis and combining abilities in growth and survival in sea scallops along the Atlantic Coast of Canada. *J. Shellfish Res.* 31, 1145–1149. <https://doi.org/10.2983/035.031.0425>
- Watermann, B.T., Herlyn, M., Daehne, B., Bergmann, S., Meemken, M., Kolodzey, H., 2008. Pathology and mass mortality of Pacific oysters, *Crassostrea gigas* (Thunberg), in 2005 at the East Frisian coast, Germany. *J. Fish Dis.* 31, 621–630. <https://doi.org/10.1111/j.1365-2761.2008.00953.x>
- Wendling, C.C., Wegner, K.M., 2013. Relative contribution of reproductive investment, thermal stress and *Vibrio* infection to summer mortality phenomena in Pacific oysters. *Aquaculture* 412–413, 88–96. <https://doi.org/10.1016/j.aquaculture.2013.07.009>

Table 1. Results from Type III Wald tests from the mixed effects logistic regression models to compare “triploid mortality” and mortality among diploids made from Virginia and Maine brood stock. df: degrees of freedom, Chisq: Chi-squared test statistic, p: p value at $\alpha=0.05$.

<u>Model</u>	<u>Factor</u>	<u>Chisq</u>	<u>df</u>	<u>p</u>
Triploid Mortality	Cross	15.5	4	< 0.05
	Site	9.1	3	< 0.05
	Cross x Site	33.9	12	< 0.05
Virginia and Maine Diploids	Cross	174.6	3	< 0.05
	Site	2.6	3	0.45
	Cross x Site	35.3	9	< 0.05

Table 2. Prevalence (P %) and weighted prevalence (WP) of infections of *Haplosporidium nelsoni*, *Perkinsus marinus*, *Haplosporidium costale*, nematopsis, chlamydia, and hemocytosis determined via histology for *Crassostrea virginica* sampled on May 24 and August 10 at Nandua Creek (ND) and August 8 at Rappahannock River (RR). Only prevalence was measured for nematopsis and chlamydia. n: number of observations. Abbreviations for crosses are found in Figure 1: L=Louisiana, M=Maine, V=Virginia.

<u>Date</u>	<u>Site</u>	<u>Cross</u>	<u>n</u>	<i>H. nelsoni</i>		<i>P. marinus</i>		<i>H. costale</i>	Nematopsis	Chlamydia	Hemocytosis	
				<u>P (%)</u>	<u>WP</u>	<u>P (%)</u>	<u>WP</u>	<u>P (%)</u>	<u>P (%)</u>	<u>P (%)</u>	<u>P (%)</u>	<u>WP</u>
May	ND	VVV	21	0	0	0	0	0	5	14	5	0.1
	ND	VVM	21	5	0.05	0	0	0	0	5	10	0.2
	ND	LLV	21	5	0.05	0	0	0	0	29	5	0.1
	ND	LLM	21	10	0.1	0	0	0	0	24	10	0.2
Aug	RR	MM	15	0	0	66	0.9	0	0	0	13	0.3
	ND	MM	15	0	0	60	1.5	0	7	13	20	0.4

Table 3. Number of dead oysters per bag from each cross and results from the ANOVA for shell height (mm) of live and dead triploid *Crassostrea virginica* sampled from Nandua Creek on June 7, 2016. Seven live oysters were sampled from each bag. df: degrees of freedom, SS: sums of squares, MS: mean sums of squares. Abbreviations for crosses are found in Figure 1: L=Louisiana, M=Maine, V=Virginia.

Dead Oysters					
<u>Cross</u>	<u>Bag 1</u>	<u>Bag 2</u>	<u>Bag 3</u>		
VVV	17	23	26		
VVM	24	14	9		
LLV	8	7	13		
LLM	8	7	3		
ANOVA					
<u>Factor</u>	<u>df</u>	<u>SS</u>	<u>MS</u>	<u>F</u>	<u>p</u>
Status	1	34	34	0.31	0.6
Cross	3	2625	875	7.85	< 0.05
Bag _{Cross}	2	166	83	0.75	0.5
Status x Cross	3	1120	373	3.35	< 0.05
Residuals	233	25977	111		

Table 4. Results from the ANOVA for shell height (mm) in August of 2016 for *Crassostrea virginica* crosses and sites, least square means from the ANOVA, and Tukey's Honest Significant Difference (HSD) test for VVV and VV among all other crosses and RR among all other sites. Results from Tukey's HSD test are reported with 95% family-wise confidence intervals in parentheses. df: degrees of freedom, SS: sums of squares, MS: mean sum of squares, \bar{x} : sample mean, ci: 95% confidence interval. Abbreviations for crosses are found in Figure 1: L=Louisiana, M=Maine, V=Virginia. Sites are RR=Rappahannock River, ND=Nandua Creek, PG=Pungoteague, and OC=Occahannock Creek.

Shell Height: August 2016								
ANOVA								
Factor	df	SS	MS	F	p			
Cross	7	11100	1586	16.9	< 0.05			
Site	3	11610	3870	41.3	< 0.05			
Cross x Site	21	2981	142	1.5	0.07			
Residuals	449	42069	94					
Least Square Means: (Cross)								
	\bar{x}	\underline{ci}						
LLM	84	82, 87						
VVV	82	80, 85						
LLV	82	79, 84						
VVM	80	78, 83						
VV	79	76, 81						
MV	77	74, 79						
VM	75	72, 77						
MM	68	66, 71						
Multiple Comparisons of Means: Tukey Contrasts (Cross)								
	VVV	VVM	LLV	LLM	VV	VM	MV	MM
VVV	--	2 (-3, 7)	0 (-5, 5)	-2 (-7, 3)	4 (-2, 9)	7 (2, 13)	5 (0, 10)	14 (9, 20)
VV	--	-2(-7, 3)	-3 (-9, 2)	-6 (-11, 0)	--	4 (-2, 9)	2 (-4, 7)	10 (5, 16)
Least Square Means: (Site)								
	\bar{x}	\underline{ci}						
RR	84	82, 85						
ND	82	80, 84						
OC	77	76, 79						
PG	71	69, 73						
Multiple Comparisons of Means: Tukey Contrasts (Site)								
	RR	ND	PG	OC				
RR	--	2 (-1, 5)	13 (10, 16)	6 (3, 10)				

Table 5. Mean of average daily temperature values (°C) at all sites during the mortality event at ND (May 10-June 27). Means are reported within sampling visits to ND: from May 10 to May 24, from May 24 to June 7, and from June 7 to June 27. *No data were available between May 24 and June 5 for RR. Sites are Rappahannock River (RR), Nandua Creek (ND), Pungoteague (PG), and Occohannock Creek (OC).

	May 10-May 24	May 24-June 7	June 7-June 27
RR	19.9*	--*	26.8
ND	19.6*	25.4*	26.5
PG	19.0*	25.4*	26.0
OC	19.3*	25.1*	26.3

Figure 1. The spawning design for *Crassostrea virginica* produced by crossing diploid oysters (2N) from Virginia (V) and Maine (M) in a 2x2 matrix, resulting in four diploid crosses (VV, VM, MV, and MM), as well as crossing the same female diploid oysters to male tetraploid oysters (4N) with Virginia (VV) and partial Louisiana (LL) origin to create four triploid crosses (VVV, VVM, LLV, and LLM).

♀ \ ♂		<u>2N</u>		<u>4N</u>	
		V	M	V	L
<u>2N</u>	V	VV	MV	VVV	LLV
	M	VM	MM	VVM	LLM

Figure 2. Map of commercial farms where the eight crosses of *Crassostrea virginica* were reared from June 2015 to November 2016. Map produced using ArcGIS® software by Esri. 1: Rappahannock River (RR); 2: Occohannock Creek (OC); 3: Nandua Creek (ND); 4: Pungoteague (PG).

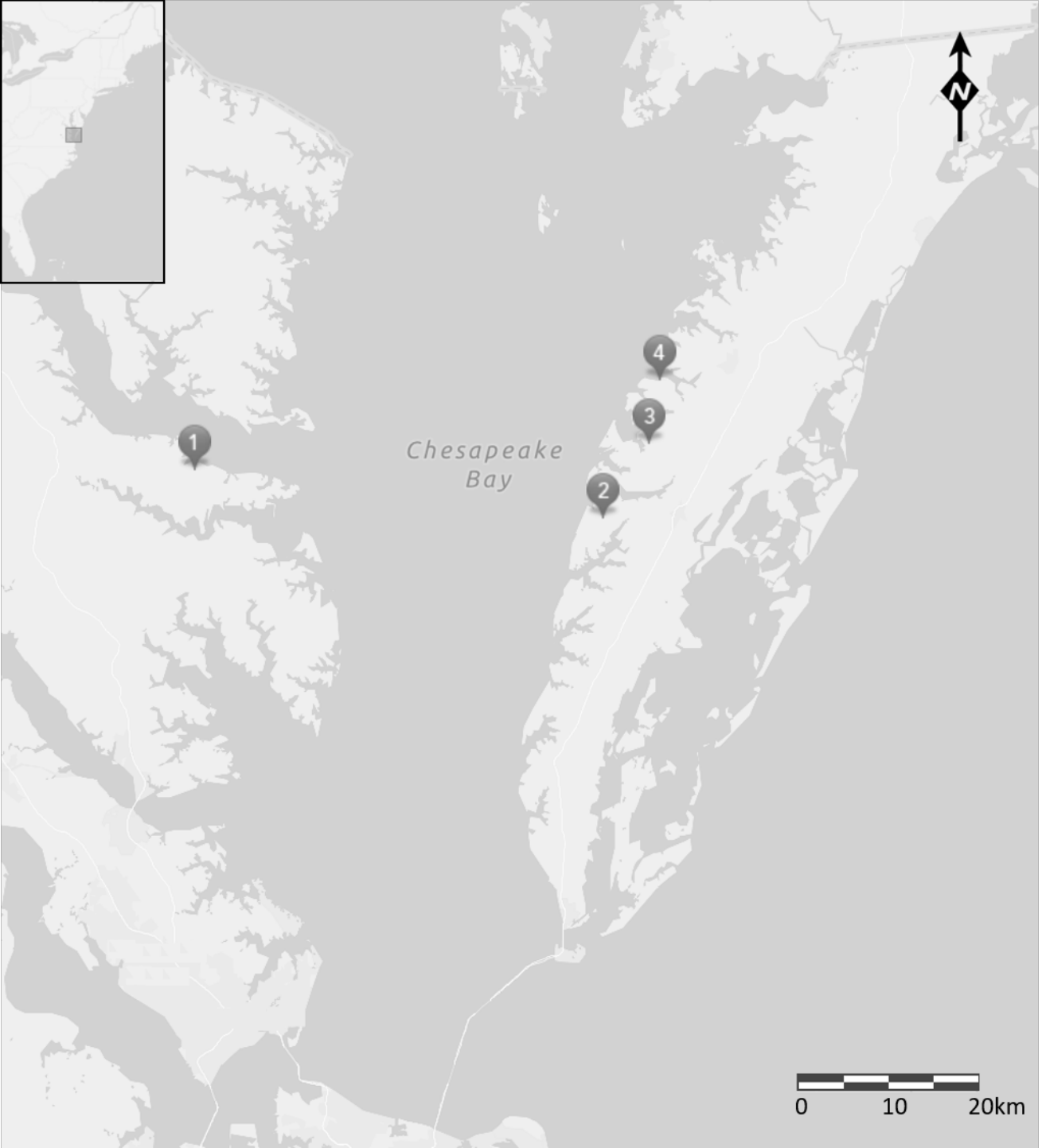
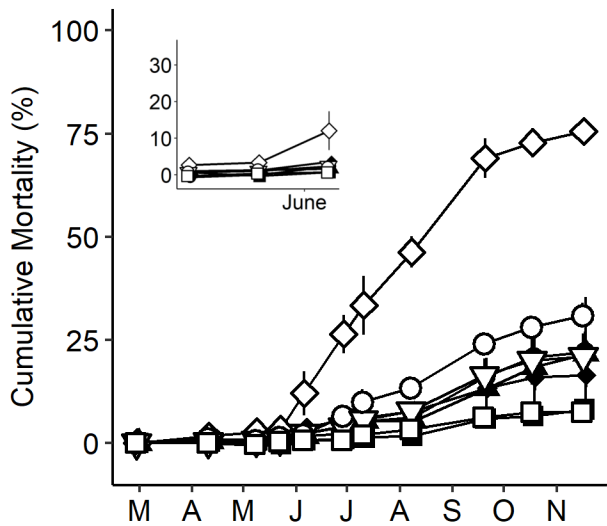
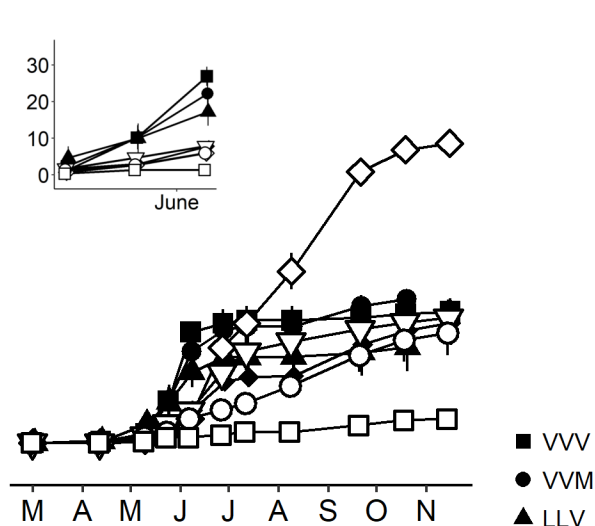


Figure 3. Mean cumulative mortality (%) in four triploid crosses (black) and four diploid crosses (white) of *Crassostrea virginica* at four sites in the Chesapeake Bay from April to November of 2016. Figure insets are closer views of cumulative mortality from early May to early June. Error bars represent \pm standard error. RR: Rappahannock River; ND: Nandua Creek; PG: Pungoteague; OC: Occohannock Creek. Abbreviations for crosses are found in Figure 1: L=Louisiana, M=Maine, V=Virginia.

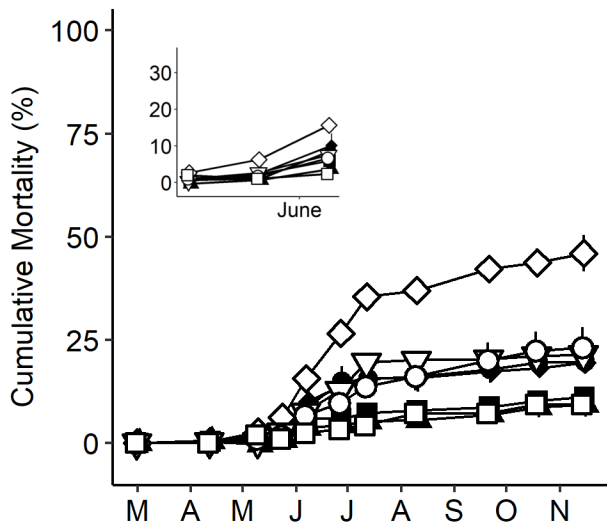
RR



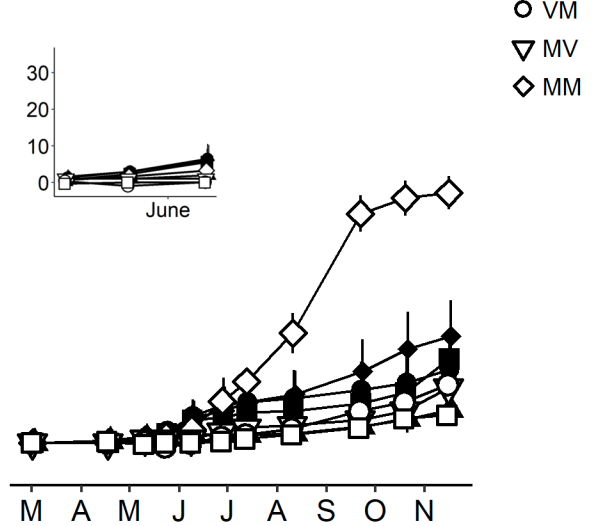
ND



PG



OC



- VVV
- VVM
- ▲ LLV
- ◆ LLM
- VV
- VM
- ▽ MV
- ◇ MM

Figure 4: Probability of mortality estimate, 95% confidence intervals, and pair-wise comparisons using Tukey's Honest Significant Differences test from the mixed effect logistic regression for mortality of a diploid cross (white) and triploid crosses (black) of *Crassostrea virginica* at four sites in the Chesapeake Bay from February through June of 2016. Sites are Rappahannock River (RR), Nandua Creek (ND), Pungoteague (PG), and Occohannock Creek (OC). Pair-wise comparisons were only made within sites. Crosses that share a letter are not significantly different at $\alpha = 0.05$. Abberviations for crosses are found in Figure 1: L=Louisiana, M=Maine, V=Virginia.

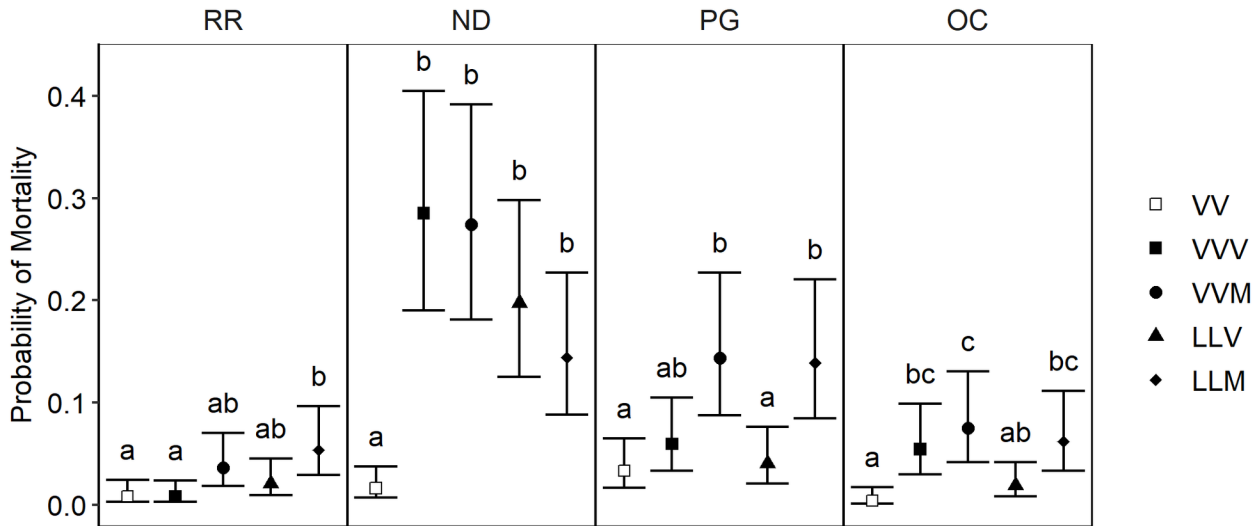


Figure 5: Probability of mortality estimates, 95% confidence intervals, and pair-wise comparisons using Tukey's Honest Significant Differences test from the mixed effect logistic regression for mortality of diploid crosses of *Crassostrea virginica* at four sites in the Chesapeake Bay from February to November of 2016. Sites are Rappahannock River (RR), Nandua Creek (ND), Pungoteague (PG), and Occohannock Creek (OC). Pair-wise comparisons were only made within sites. Crosses that share a letter are not significantly different at $\alpha = 0.05$. Abberviations for crosses are found in Figure 1: M=Maine, V=Virginia.

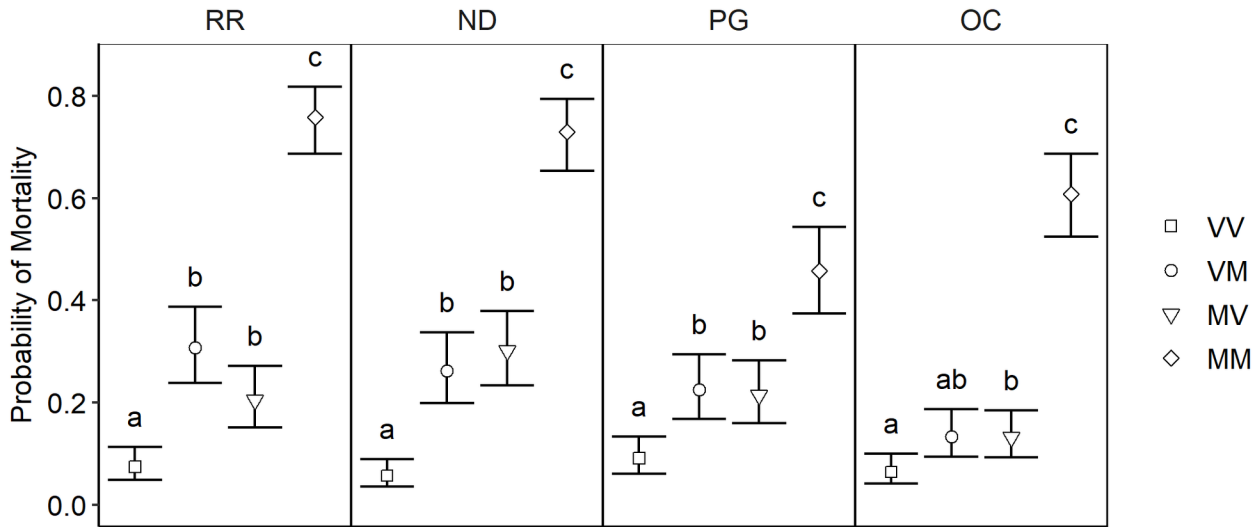


Figure 6. Shell length (mm) of live and dead *Crassostrea virginica* of four triploid crosses (VVV, VVM, LLV, LLM) sampled from Nandua Creek on June 7, 2016. Abbreviations for crosses are found in Figure 1: L=Louisiana, M=Maine, V=Virginia.

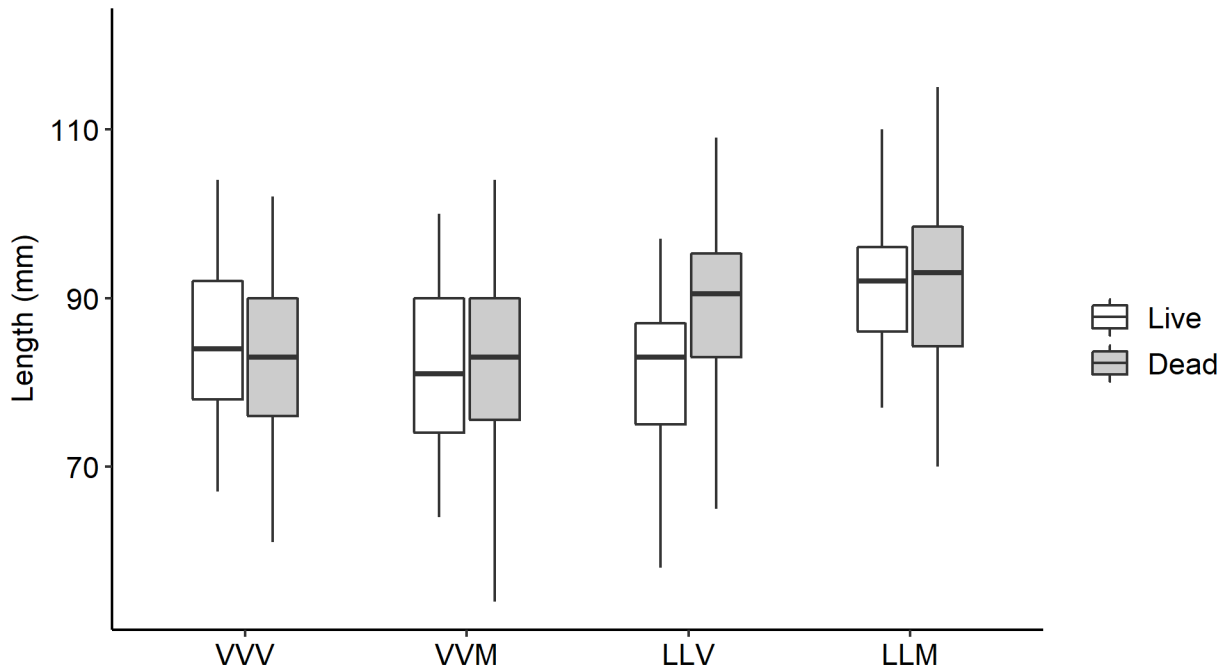


Figure 7. Mean shell height (mm) for four triploid crosses (black) and four diploid crosses (white) of *Crassostrea virginica* at each site at the start of the experiment in February/March 2016 and at the end of the experiment in August 2016. RR: Rappahannock River; ND: Nandua Creek; PG: Pungoteague; OC: Occohannock Creek. Abbreviations for crosses are found in Figure 1: L=Louisiana, M=Maine, V=Virginia.

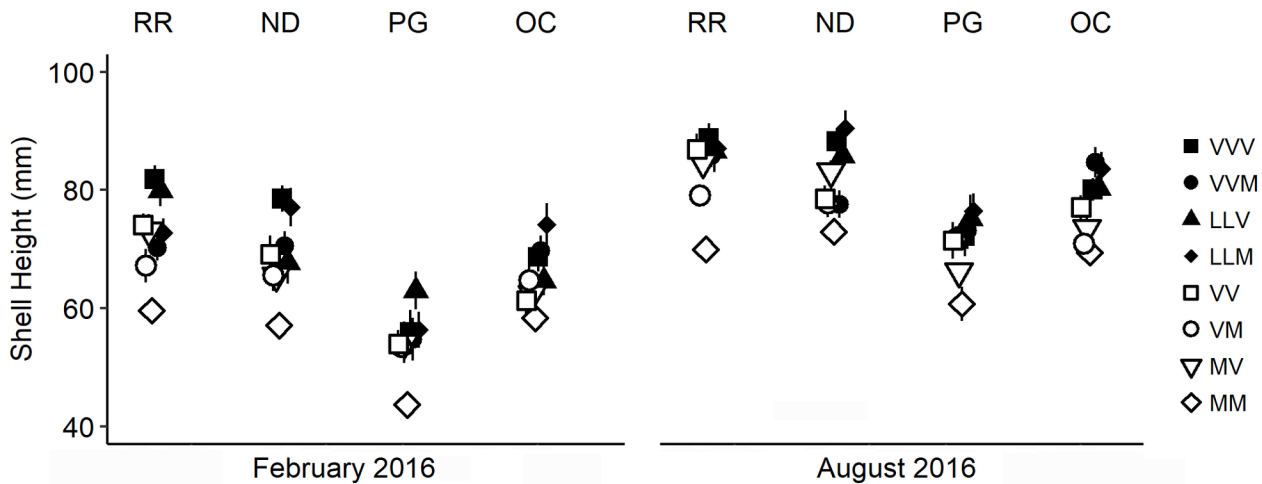


Figure 8. Mean meat weight (g) for four triploid (black) and four diploid (white) crosses of *Crassostrea virginica* reared at four sites from February-August of 2016. Error bars represent \pm standard error. RR: Rappahannock River; ND: Nandua Creek; PG: Pungoteague; OC: Occohannock Creek. Abbreviations for crosses are found in Figure 1: L=Louisiana, M=Maine, V=Virginia.

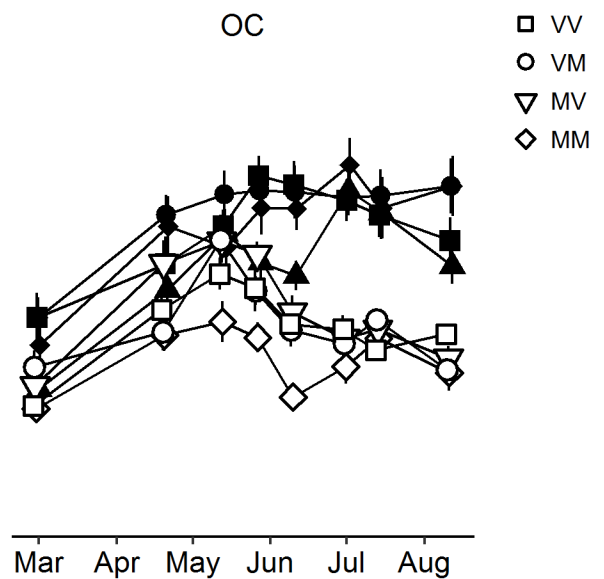
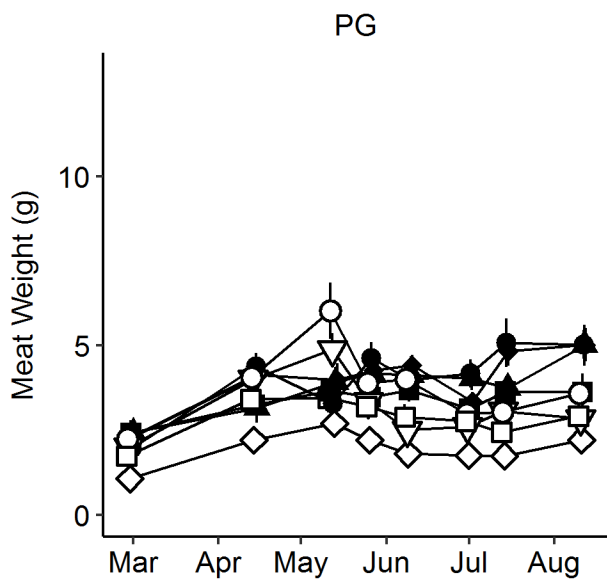
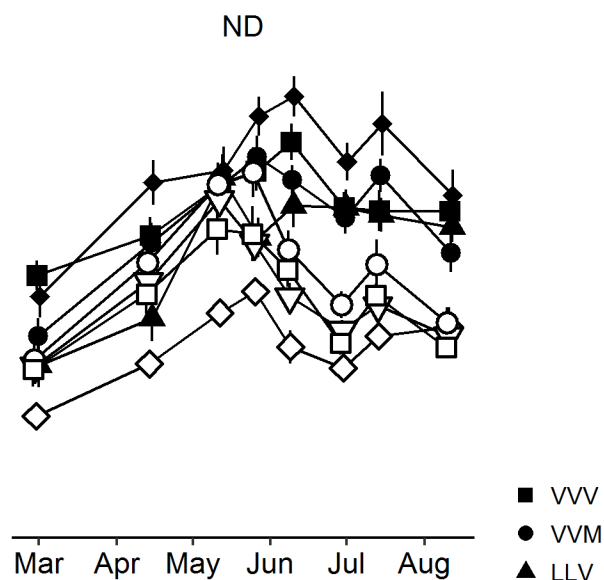
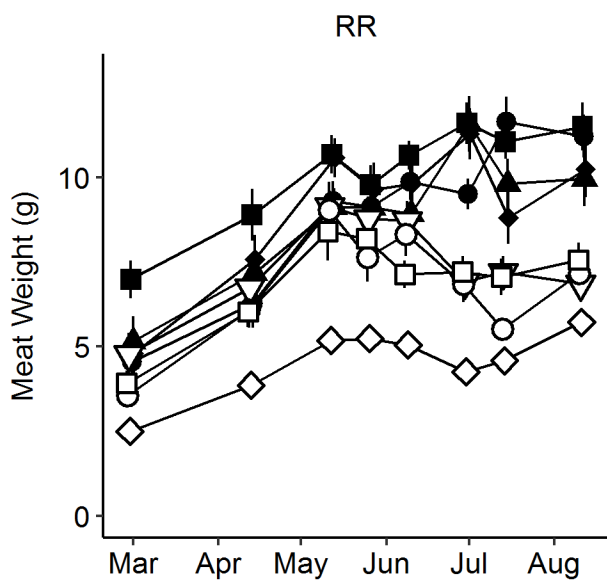


Figure 9. Average daily temperature (°C) from January 1 to August 10 of 2016 at each of the four sites where crosses of *Crassostrea virginica* were reared from June 2015 to August 2016. RR: Rappahannock River; ND: Nandua Creek; PG: Pungoteague; OC: Occohannock Creek.

Average Daily Temperature (°C)

30

○ RR

■ ND

▽ PG

△ OC

20

10

0

Jan

Feb

March

April

May

June

July

Aug

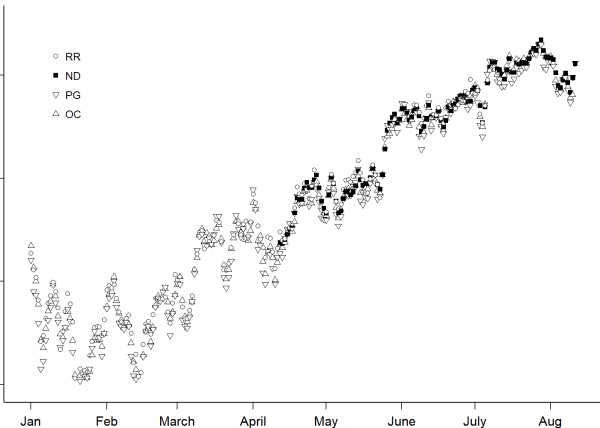


Figure 10. Average daily salinity (ppt) from May 10 to August 10, 2016 at each of the four sites where crosses of *Crassostrea virginica* were reared from June 2015-August 2016. RR: Rappahannock River; ND: Nandua Creek; PG: Pungoteague; OC: Occohannock Creek.

Average Daily Salinity (ppt)

