

1 **Modeling in an Integrated Ecosystem Research framework to** 2 **explore recruitment in Gulf of Alaska groundfish –** 3 **applications to management and lessons learned**

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23
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25 Research programmes

26 27 28 **Highlights**

- 29 • Model outputs should have utility for management decisions and regional stakeholders.
- 30 • Laboratory and field data should be appropriate to inform, improve, or validate models.
- 31 • Time must be allocated to incorporate new data and mechanistic understandings into
32 models.

33 34 **ABSTRACT**

35 The Gulf of Alaska Integrated Ecosystem Research Program (GOAIERP) supported multi-
36 disciplinary analyses integrating physical and biological oceanography and modeling to examine
37 how the environment influences survival and recruitment of early life stages of select commercially
38 and ecologically important groundfish species. Recruitment is an important component of
39 population variability, and understanding the processes influencing recruitment is central to
40 fishery management and ecosystem planning. Determining the relative impact of advection and
41 the environmental conditions experienced during transport between spawning and nursery areas
42 is an inherently interdisciplinary problem. It requires consideration of physical and lower trophic
43 level environments in concert with early life history dynamics. Here we discuss how Eulerian
44 ecosystem models and Lagrangian Individual-Based Models for groundfish were integrated within
45 the framework of an interdisciplinary observational program. Metrics (e.g. regionally-based
46 averaged water temperature, integrated primary production, probability of juvenile settlement)
47 were derived from model outputs as proxies for recruitment success. The recruitment indices
48 were then correlated to estimated recruitment from stock assessments. Using the GOAIERP as
49 a case study, we discuss the value that modeling can add to a field program and fisheries

50 management planning, the challenges faced, and steps that can be taken to maximize program
51 success. Coordination of model development, experimentation, and field sampling is necessary
52 but can be challenging. Consideration of the appropriate sequence during data analyses and
53 model development is critical. Careful consideration must be given to ensure that data collected
54 in the field will inform, improve, or validate models. Sufficient time must be allocated within the
55 program to incorporate field data collected during the program and mechanistic understandings
56 into the models. Model outputs should be designed to have utility to management decisions and
57 value to regional stakeholders. Collectively, the studies in this modeling program provide insight
58 as to how models might be used to better understand recruitment processes and lead to
59 recommendations to support the integration of ecosystem models into fisheries management.

61 **1. Introduction**

62
63 Marine fish population dynamics are responsive to environmental variation (e.g. Shelton and
64 Mangel, 2011; Vert-pre et al., 2013; Szuwalski et al., 2015) and climate regimes (Nye et al., 2014,
65 Barbeaux et al., 2020). However, the mechanisms influencing population variability are
66 complicated and remain poorly understood (Rothschild, 1986; Fogarty et al., 1991; Munch et al.,
67 2018). Determining how these mechanisms are affected by climate and fishing remains a primary
68 objective of research to support sustainable fisheries management (Fogarty, 2014) and a critical
69 component to effective ecosystem-based fisheries management (Link, 2002; Pikitch et al., 2004;
70 Essington and Punt, 2011). This is particularly true for the impact of the physical environment on
71 the biological and transport processes important to survival through the early life critical period
72 (Hjort, 1994), and the match-mismatch in the timing of larvae and their food sources (Cushing,
73 1990).

74 75 *1.1. Marine fish recruitment*

76
77 Recruitment, or the number of new individuals joining a population each year, is an integral
78 part of productivity and population fluctuations in marine fish. Understanding recruitment
79 processes can inform ecosystem approaches to management. Recruitment variability, however,
80 is poorly understood. Developing and integrating the datasets necessary to understand and
81 predict relevant environmental influences and ecological interactions, and determining how to
82 interpret the resulting dynamics in ways that inform stock assessments, remains a challenge.

83 The first weeks in life often determine survival in fish (Catalán et al., 2020). Many processes,
84 including environmental and multispecies interactions, determine the survival or mortality of fish
85 in this timeframe, the cumulative effects of which lead to variability in the success of annual larval
86 cohorts. Young larval fish are typically characterized by rapid dispersal due to advection and high
87 mortality rates due to starvation and predation. Variability in the transport of pelagic early life
88 stages of fish to suitable nursery and settlement habitats has been at the foundation of recruitment
89 analyses (Levin, 1994). Contributing factors to survival and mortality include fluctuations in the
90 spatial and temporal extent, magnitude, and availability of prey fields (Okamoto et al., 2012) and
91 predator densities (Leggett, 1986), oceanic and shelf transport (Norcross and Shaw, 1984; Fortier
92 and Leggett, 1985; Myers and Drinkwater, 1989), and availability of suitable habitat for settlement
93 (Wespestad et al., 2000; Johnson, 2007; Pirtle et al., 2019; Goldstein et al., 2020). In the context
94 of fisheries management, recruitment estimation using mechanistic modeling requires
95 determining linkages between environmental conditions and mortality estimates, understanding
96 transport and productivity processes, identifying critical habitat, and estimating settlement rates
97 (Houde, 1989; Chambers and Trippel, 2012; Stige et al., 2013). Beyond that, it requires
98 determining the relative contribution of early life stages of fish to overall stock dynamics.

99

100 *1.2. Fisheries management and modeling.*

101
102 Fishery management aims to produce sustainable biological, social, and economic benefits
103 from fisheries resources. This requires responding to variability and uncertainty, both in the state
104 of the resource, the assessment of that resource, and human activity (Walters, 1986; Hilborn and
105 Walters, 1992). Within the USA, all fish stocks of commercial importance are managed to a
106 greater or lesser extent, with rules being implemented and enforced by governing bodies to
107 prevent overfishing, allow for the recovery of overfished stocks, and achieve optimum yield.

108 Historically, fisheries management, including the annual total allowable catch, was determined
109 solely by the estimated adult population size, with little regard to how the prevailing environmental
110 conditions could impact survival and recruitment. Accounting for shifting distributions and
111 changing productivity is a critical need for the development of sound scientific advice (Karp et al.,
112 2019) for fishery management and such ecosystem considerations have been identified as a
113 priority for the management of Alaska groundfish (Witherell et al., 2000). The USA Magnuson-
114 Stevens Fishery Conservation and Management Act has identified the incorporation of ecosystem
115 considerations in fisheries management to be a longstanding priority and ongoing need. To date,
116 this has mostly been addressed in broadscale ecosystem analyses (Aydin et al., 2007; Gaichas
117 et al., 2011; Collie et al., 2016), through the development of indirect indices or indicators of
118 ecosystem state (Cury and Christensen, 2005; Link, 2005; Mueter et al., 2007), the application of
119 qualitative approaches to assess ecosystem status (Mace, 2000; Zador et al., 2017), or through
120 management strategy evaluations that include ecosystem drivers (A'mar et al., 2010; Fulton et
121 al., 2014; Punt et al., 2016a). Progress has been made in the design of operational management
122 strategies for achieving fishery ecosystem objectives in the face of ecosystem effects (Sainsbury
123 et al., 2000) and in the investigation of connections between ecosystem dynamics and
124 commercial and subsistence harvests (Haynie and Huntington, 2016). Recently, approaches
125 have been applied to develop spatial-temporal models in ecosystem assessments (Thorson et
126 al., 2019) and concerted efforts have been made to more explicitly link ecosystem processes
127 directly into single-species and multispecies assessment and forecasts (Holsman et al., 2019).

128 It is important to determine how models fit into the management approaches currently applied
129 when determining the utility of model products. Priorities in fishery management might be
130 distinguished as either tactical or strategic. Tactical priorities include the development of
131 immediate metrics, such as Overfishing Limits, Acceptable Biological Catches, and Annual Catch
132 Limits; these might also include determining whether overfishing is occurring or if the stock is
133 currently overfished or approaching an overfished state. Strategic issues reflect broadscale goals
134 for how an ecosystem or a fishery might operate in the future, and include determining broad
135 policy goals, and what control rules should be used to achieve those goals.

136
137 *1.3. Integrated Ecosystem Research Programs*

138
139 The North Pacific Research Board (NPRB) developed the Integrated Ecosystem Research
140 Program (IERP) to investigate mechanistic processes that structure ecosystems, drive
141 productivity, organize biological communities, shape species interactions and dynamics, and
142 influence processes important to human communities and industries (Baker and Smith, 2018).
143 Research promoted in this context is designed to advance hypothesis-driven multidisciplinary
144 research and to promote collaboration and integration across research disciplines (e.g. field
145 observations, laboratory investigation, and modeling) and ecosystem components (e.g. physics,
146 fishes, humans). IERPs were implemented in the Bering Sea (BS, 2007-2012; Wiese et al., 2012),
147 Pacific Arctic (2016-2021; Baker et al., 2020), and the Gulf of Alaska (GOA, 2010-2018; Dickson
148 and Baker, 2016; Ormseth et al., 2019; Lindeberg et al., 2022). Common objectives of these
149 integrated programs have been to understand the effects of climate variability and climate change
150 on the distribution, abundance, and production of marine organisms and to incorporate this

151 understanding into diagnostic and prognostic models that can then further inform our
152 understanding.

153

154 *1.4. Gulf of Alaska Integrated Ecosystem Research Program*

155

156 The GOA (Fig. 1) is a dynamic and productive region that supports several commercially
157 important fisheries. The international Global Oceans Ecosystem Dynamics (GLOBEC) program
158 (Fogarty and Powell, 2002) identified the coastal GOA as one of three regions of interest in the
159 US and established the GOA-GLOBEC program (Batchelder et al., 2005). The Gulf of Alaska
160 IERP (GOAIERP) built on the GLOBEC effort and engaged more than 50 scientists from 11
161 institutions (<https://www.nprb.org/gulf-of-alaska-project>). This multidisciplinary project examined
162 the oceanography, biology, and ecology of the GOA to better understand how the environment
163 influences the survival of larval and juvenile fish to the adult stage, and ultimately the success of
164 fisheries. Understanding recruitment variability in the GOA is made difficult by the complexity of
165 the physical system. The strongest currents found in the Northeast Pacific flow through the GOA
166 (Reed, 1984; Stabeno et al., 1995). This strong flow, combined with complicated topography
167 (Zimmermann and Prescott, 2015; Baker et al., 2019) and highly variable freshwater runoff
168 (Royer, 1982; Beamer et al., 2016; Danielson et al., 2020), contributes to a dynamic physical
169 system which in turn influences the entire ecosystem. To identify and quantify the physical and
170 biological factors that influence the productivity of groundfish species, five commercially and/or
171 ecologically important groundfish species that exhibit a broad range of life history strategies were
172 selected as the focus of the program: Walleye Pollock (WP, *Gadus chalcogrammus*), Pacific Cod
173 (PC, *Gadus macrocephalus*), Pacific Ocean Perch (POP; *Sebastes alutus*), Sablefish (SF,
174 *Anoplopoma fimbria*), and Arrowtooth Flounder (ATF, *Atheresthes stomias*). The suite of models
175 developed under the GOAIERP were employed to improve understanding of recruitment
176 fluctuations. These models covered a longer period (1996-2012) than the field sampling program
177 (2011, 2013), and were designed to address the program's central hypothesis that early life
178 survival is the primary factor determining the year-class strength of groundfish species in the
179 GOA. Here we discuss 'lessons learned' from modeling within a broader integrated ecosystem
180 research program such that our successes and shortcomings can be applied to future research
181 efforts and further the application and integration of ecosystem modeling in fisheries
182 management.

183

184

185 **2. Hypotheses and approach**

186

187 The GOAIERP was motivated by the foundational hypothesis that recruitment control occurs
188 in the early life history stages (e.g. "critical period hypothesis", Hjort, 1914) and is in part driven
189 by variation in the larval foraging environment (e.g. "match-mismatch hypothesis", Cushing,
190 1990). The GOAIERP proposed the following central hypothesis on survival and recruitment of
191 five focal groundfish species:

192

The Gauntlet –

The primary determinant of year-class strength for marine groundfishes in the Gulf of Alaska is early life survival. This is regulated in space and time by climate-driven variability in a biophysical gauntlet comprising offshore and nearshore habitat quality, larval and juvenile transport, and settlement into suitable demersal habitat.

193

194 Survival of an individual fish, from spawning to recruitment, is controlled by the complex and
195 variable biophysical environment encountered during egg and larval drift stages prior to reaching

196 habitat suitable for juvenile settlement. This does not preclude the effect of spawning biomass on
197 recruitment, but the working hypothesis in the GOA IERP was that post-spawning, the survival of
198 the earliest life stages of groundfish, during transport from offshore natal areas to nearshore
199 nursery habitats, is the principal influence affecting variability in recruitment given egg production.
200 As such, successful recruitment may depend on many interrelated factors affecting individual fish
201 along their transport pathways including those directly influencing survival (e.g. prey, predation),
202 as well as those influencing the physical environment and thus the pathways themselves. We
203 refer to the biophysical processes that occur along and influence transport pathways during the
204 first year of life, as “*the gauntlet*”.

205 An essential feature of the GOA IERP was its comprehensive spatial scope and
206 multidisciplinary and integrative structure. A collaborative research approach was applied,
207 involving four groups of scientists, three focusing on separate trophic levels (lower, middle, upper)
208 and one constructing models (Fig. 2). Integrated physical, chemical, and biological oceanographic
209 sampling was conducted along a comprehensive sampling grid in the GOA, extending from
210 Baranof Island in the east to Kodiak Island in the west (Fig. 1). Ecosystem surveys were
211 conducted at inshore and offshore sites and field surveys were complemented by laboratory
212 analysis of food habits and energetic condition, physiological experiments, and modeling.
213 Environmental conditions and processes influencing early ontogenetic stages of the focal fish
214 species were examined (Doyle and Mier, 2015), and results were used to develop species-
215 specific individual-based models (IBMs) to predict recruitment variability under various
216 environmental scenarios (Fig. 3; Gibson et al., 2019, Stockhausen et al., 2019a, 2019b, Hinckley
217 et al., 2019). Discussions before the start of the GOA IERP program conceptualized the likely
218 connections between model results and field data. Research design, data collection, analysis,
219 and interpretation were closely coordinated among the groups beginning early in the research
220 design phase.

221 To address the overarching gauntlet hypothesis and assess the impact of environmental
222 variability in driving transport and success of early life stages from spawning to settlement, the
223 GOA IERP modeling group integrated a series of modeling tools to address the following specific
224 hypotheses:

225
226 *H1: Recruitment variability of the five focal species is primarily influenced by variability in*
227 *the proportion of young fish transported from offshore spawning areas to nearshore*
228 *nursery areas (connectivity) due to interannual differences in the strengths of the physical*
229 *regimes that characterize the GOA environment.*

230
231 *H2: Recruitment variability is (secondarily) influenced by the survival of young fish*
232 *successfully transported to nursery areas, which varies due to differences in physical*
233 *factors (wind speed and direction, water temperature, runoff, mixing) and biological*
234 *processes (prey abundance, competition, predation) encountered along the transport*
235 *pathways.*

236
237

238 **3. Species-specific Individual-Based Models**

239
240 The suite of models to test H1 and H2 included a physical oceanographic model that simulated
241 the ocean environment, a lower trophic level model that simulated nutrients, phytoplankton,
242 zooplankton, production, and biomass fields, and early life history models for each of the five focal
243 species. Each of the GOA IERP models has been previously described in detail but is summarized
244 below.

245

246 3.1. Eulerian physical oceanographic and lower trophic level models

247

248 The well-established Regional Ocean Modeling System (ROMS, Haidvogel et al., 2008; Moore
249 et al., 2004; Shchepetkin and McWilliams, 2005) was used to simulate the time-varying, three-
250 dimensional hydrodynamics in the GOA. The implementation of ROMS for the GOA has a
251 horizontal resolution of approximately 3 km and can replicate common features in GOA circulation
252 that can influence transport—such as currents, eddies, meanders, and hydrographic fronts
253 (Cheng et al., 2012; Coyle et al., 2013; Dobbins et al., 2009; Hermann et al., 2016, 2009a;
254 Hinckley et al., 2009a). A lower trophic level Nutrient-Phytoplankton-Zooplankton (NPZ) model
255 was also developed and validated for the GOA (Hinckley et al., 2009b; Coyle et al., 2012, 2013)
256 and includes nitrate, ammonium, iron, large and small phytoplankton, microzooplankton, large
257 and small copepods, and euphausiid components (Fig. 4). The NPZ model was fully integrated
258 within the ROMS framework and the coupled model was run for 1996-2011 providing a 16-year
259 time series of model output for driving the IBMs.

260

261 3.2. Lagrangian larval fish models

262

263 Spatially explicit biophysical IBMs that use Lagrangian particle tracking algorithms have been
264 widely applied to study recruitment (e.g. Hinckley et al., 1996; Stockhausen and Lipcius, 2003)
265 and connectivity (e.g. Cowen et al., 2006, 2007). Each IBM used in the GOA IERP (Fig. 3) was
266 developed to reflect early life characteristics and behavior for the focal groundfish species. They
267 integrated biological processes affecting simulated individuals as they develop in time through
268 multiple early life stages, as well as due to advective and diffusive transport. The WP model was
269 developed under past research efforts (Parada et al., 2016, Hinckley et al., 2016) using the
270 Ichthyop 3.1 framework (Lett et al., 2008), while the remaining four IBMs were developed within
271 the GOA IERP (Gibson et al., 2019; Stockhausen et al., 2019a, 2019b, Hinckley et al., 2019) using
272 the Dispersal Model for Early Life Stages (DisMELS; Stockhausen, 2021) framework. Both IBM
273 platforms used stored oceanographic and lower trophic level output from the coupled ROMS-NPZ
274 model to simulate the environment experienced by early life stages. The IBMs were synthetic and
275 integrative, incorporating data obtained during and prior to the GOA IERP.

276 ATF, SF, and POP spawn in deep water at the edge of the continental shelf in the GOA. In
277 contrast, WP spawn primarily in Shelikof Strait while PC spawn on the GOA continental shelf. The
278 deep spawning species utilize shallow inshore habitats as juvenile nursery areas while juvenile
279 WP and PC can be found across the shelf. Each species-specific IBM was informed by a
280 conceptual model of its early life history (Fig. 3). The degree of complexity in each IBM reflected
281 the data available for each species. For example, pollock has historically been well studied in the
282 GOA so information on spatially explicit annual spawning biomass and predation pressure was
283 available for it, but not for the other focal species. A brief description of the early life history for
284 each species is outlined below and a summary of each life stage and the key model features are
285 shown in Table 1.

286 Tens of thousands of individual model fish were released in each simulation and each IBM was
287 run annually from spawning (first life stage is eggs) through to ‘settlement’, or the end of the year.
288 To support subsequent examination of individual histories, the entire history for each individual
289 ‘fish’ was retained on a daily time interval (i.e. location, depth, length, age, life stage), along with
290 information on their physical (i.e. temperature) and biological surroundings (i.e. biomass of
291 zooplankton prey fields) from the ROMS-NPZ model.

292

293 *Walleye Pollock (WP)*

294 WP are very fecund with highly variable mortality and growth rates in early life. Eggs are
295 released at 200-300 m depth in March and April and hatch two weeks later. Larvae begin diel
296 migration when they are ~7 mm in size, and gradually increase their swimming capacities until

297 their movement becomes independent of currents (Fig. 3a). Optimal WP prey depends on larval
298 size, temperature, light, turbulence, and turbidity (Porter et al., 2005). Predation on juvenile WP
299 may be important to recruitment, particularly as groundfish predator abundance has increased
300 since the 1980s. Environmental effects on larval survival were key to recruitment success prior to
301 the increase in predator biomass (Bailey, 2000). The prevailing hypothesis for WP is that Shelikof
302 Strait comprises the primary spawning area and the Shumagin Islands are the main nursery area
303 (Hinckley et al., 2001). Currents transport larvae southwest along the Alaska Peninsula (Yoklavich
304 and Bailey, 1990; Hinckley et al., 2001). The IBM for WP included four life stages (egg, yolk-sac
305 larvae, feeding larvae, and age-0 juveniles). Egg development was driven by age and temperature,
306 growth of yolk-sac larvae depended on degree days, the growth of feeding larvae and juveniles
307 depended on consumption estimated as a function of individual weight and temperature, and
308 predation on juveniles was based on groundfish predation data (Megrey and Hinckley, 2001;
309 Parada et al., 2016).

310

311 *Pacific Cod (PC)*

312 PC spawn between February and July in the GOA (Dunn and Matarese, 1987) over rocky
313 substrates at depths of 20-200 m (Hurst et al., 2009). Egg and larval dispersal may be limited
314 because eggs are demersal, and semi-adhesive (Alderdice and Forrester, 1971). Hatching of
315 pelagic yolk-sac larvae at ~3-4 mm standard length (SL) 21-26 days after fertilization and
316 individuals show strong surface orientation (Hurst et al., 2009). Juvenile nursery areas are
317 primarily shallow, coastal embayments (Abookire et al., 2007; Laurel et al., 2009). The IBM for
318 PC included six stages (egg, yolk-sac larvae, pre-flexion feeding larvae, post-flexion feeding
319 larvae, epipelagic juveniles, and settled juveniles), with stage-specific processes modeled for
320 growth, development, depth distribution, and diel migration (Hinckley et al., 2019; Fig 3b).

321

322 *Pacific Ocean Perch (POP)*

323 POP are members of the *Sebastes* genus, a primitive viviparous group (Love et al., 2002).
324 Reproduction occurs in April-May (Westerheim 1975) at depth (500-700 m). Larvae remain at
325 depth for a month or more, before moving to shallower depths (Love et al., 2002). They begin
326 feeding at 3-7 mm SL (Kendall and Lenarz 1987) and occupy the near-surface layers. The
327 duration of the larval stage is 1-2 months (Matarese et al., 2003) and this stage completes at 20-
328 30 mm SL. POP juveniles in the GOA remain in the water column for several months until fall, at
329 which time they use demersal subtidal habitats with complex topography and extensive cover
330 (Carlson and Haight, 1976). The IBM for POP consisted of five sequential early life stages:
331 preflexion larva, postflexion larva, pelagic juvenile, settlement-stage juvenile, and benthic juvenile
332 (Fig. 3c). The first four stages were defined in the IBM to facilitate ontogenetic changes in
333 "preferred" depth ranges, growth rates, and movement parameters. The final stage (benthic
334 juvenile) was simply a "marker" that indicated an individual had successfully settled in a benthic
335 nursery area. The IBM did not include bioenergetics or directed swimming. Similar to PC, the
336 GOA model domain was divided into 12 alongshore zones and several depth zones for analysis.

337

338 *Sablefish (SF)*

339 SF spawn pelagic eggs in winter near the edge of the continental shelf (Kendall and Matarese,
340 1987), with peak egg abundance in the western GOA in February (Doyle and Mier, 2015). Eggs
341 are found at depths >200 m and require 2-3 weeks to hatch (Mason, et al., 1983). Before hatching,
342 eggs sink to depths exceeding 400-500 m and maintain that position. The time from hatch to first
343 feeding is around two weeks (Boehlert and Yaklovich, 1985). Once the yolk sac is absorbed,
344 larvae swim to the surface and grow about 2 mm per day from about 10 to 80 mm SL (Kendall
345 and Matarese, 1987; Shenker and Olla, 1986). Following the transition to the juvenile stage,
346 individuals continue to inhabit the upper water column but undertake diel vertical migrations,
347 moving higher in the water column at night (Courtney and Rutecki, 2011; Sogard and Olla, 1998).

348 The IBM for sablefish included five sequential early life stages: egg, yolk sac larvae, feeding larva,
349 epipelagic juvenile, and settlement-stage juvenile (Fig. 3d). Each life stage was parametrized with
350 different growth rates, depth preferences, vertical swimming speeds, minimum and maximum
351 stage duration, and minimum size for stage transition.

352

353 *Arrowtooth Flounder (ATF)*

354 ATF spawn along the continental slope at depths of 100-500 m (Blood et al., 2007). Spawning
355 begins in December (Blood et al., 2007). Eggs are pelagic and the duration of the egg stage is
356 temperature-dependent. The mean size at hatching is 4.4 mm SL (Blood et al., 2007) and yolk
357 absorption is complete by 6.5-7 mm SL. Flexion occurs at 13.4 mm SL and transformation occurs
358 at 45 mm SL (Blood et al., 2007, Bouwens et al., 1999). Larvae ascend to shallower depths before
359 yolk sac absorption is complete. While most larvae are found along the outer shelf and slope,
360 many late-stage eggs and larger larvae have been found farther inshore and associated with
361 troughs and canyons, where downwelling relaxation and cross-shelf flow occur (Bailey and
362 Picquelle, 2002). Interannual variation in size is small compared to intra-annual variation,
363 suggesting that arrowtooth flounder hatch over an extended period (Bouwens et al., 1999).
364 Settlement begins in early August and finishes by the end of October. The IBM for ATF was
365 relatively simple, reflecting the limited knowledge of the early life stages for this species. Growth
366 rates were stage-dependent and movement was passive and undirected, except that individuals
367 moved vertically to remain within stage-specific “preferred” depth ranges. The model used eight
368 sequential early life stages: egg, small yolk sac larva, large yolk sac larva, small feeding preflexion
369 larvae, large feeding preflexion larvae, postflexion larvae, settlement-stage juveniles, and benthic
370 juveniles (Fig. 3e).

371

372

373 **4. Analytical approach and result highlights**

374

375 Several research questions (summarized in Table 2) were developed in discussions between
376 the four GOA IERP research groups early in the program. These were then clarified as specific
377 analyses that might be applied to address the primary hypothesis. Modeling hypothesis H1 was
378 addressed for all five focal species by performing connectivity analysis between spawning and
379 nursery areas. Hypothesis H2 was addressed through analysis of individual trajectories and
380 histories that provide details of the ‘gauntlet’ experienced during the first year of life as the young
381 fish were transported from offshore spawning areas to nearshore nursery areas. Because the
382 IBMs were of varying levels of complexity, depending on the information available to construct
383 them, we were able to apply the various types of trajectory analyses in more depth to some
384 species than others. Each type of analysis performed is outlined below and example results are
385 provided.

386

387 *4.1. Connectivity analysis*

388

389 Population connectivity, i.e. the relative strength in connectivity between each spawning and
390 nursery area (e.g. successful endpoint), is inherently a coupled bio-physical process and research
391 topic, involving physical processes (e.g. eddies, fronts, tides, geomorphology; Cowen and
392 Spoungle, 2009), as well as biological processes including behavior (e.g. vertical migration;
393 Cowen et al., 2002). The working theory for the connectivity analyses was that physical properties
394 are important in driving the size of the annual recruitment to the fishery (i.e. H1). The strength of
395 connectivity between areas is often expressed as the proportion of individuals released from a
396 spawning area that settled into a suitable nursery area, such that it is independent of spawning

397 stock size (Gibson et al., 2019; Stockhausen et al., 2019a). Connectivity can be impacted by the
398 simultaneous and combined effects of individual movement and physical forcing (e.g. currents).

399 The IBMs for PC, ATF, SF, and POP divided the GOA model domain into 12 alongshore
400 spawning and nursery zones (Fig. 5a), and several species-dependent depth zones. The GOA
401 model domain for WP was split into 45 bathymetric and topographically defined areas, allowing a
402 more detailed analysis. Connectivity values were calculated and analyzed at the end of each
403 simulation year for each spawning area-nursery area pair to develop a series of connectivity
404 matrices summarizing the interannual and median connectivity across the GOA for all species.
405 This allowed an examination of the interannual variability in “total connectivity”, the sum of all
406 probabilities in the connectivity matrix for each year, as well as the interannual variability in
407 connectivity between known/likely spawning and nursery areas. Connectivity indices were
408 subsequently correlated with stock assessment recruitment estimates (see section 5).

409 Mortality processes along individual trajectories were not estimated or included except for WP.
410 “Connectivity” therefore represents “maximum potential” connectivity between each spawning
411 and nursery area (i.e. Fig. 5) due to the interaction of physics and basic life history dynamics.
412 Understanding the connectivity between spawning and nursery areas in our study domain
413 enabled the identification of key spawning and recruitment sites for each species and
414 quantification of the degree to which physics alone can account for observed transport to nursery
415 areas. For example, results from the Sablefish IBM indicate that, in the absence of directed
416 horizontal movement, sablefish spawned throughout the GOA have the highest probability for
417 settlement in nursery areas in the central GOA (Fig. 5b). However, near-shore waters extending
418 from southeast Alaska to British Columbia are known to be some of the most important nursery
419 grounds for young sablefish (Sasaki, 1985), and juvenile sablefish are found consistently only
420 within St. John Baptist Bay (Fig. 1; Rutecki and Varosi, 1997). The probability of connectivity to
421 this region is not particularly high if spawning is assumed to occur evenly along the shelf break
422 throughout the GOA. This supports the hypothesis that sablefish spawning is likely more
423 concentrated in areas in the southeast GOA, or that settlement to this region depends on selective
424 behavioral traits of young sablefish not presently captured in the model. The spawning-nursery
425 area connectivity pattern for ATF (Fig. 5c) was similar to SF with a general east to west
426 connectivity and the highest probability for settlement in nursery areas in the central GOA.

427 The results of modelled connectivity patterns indicate that the early life history stages of PC
428 generally do not disperse far from their natal areas (Fig. 5d). Retention of modelled individuals in
429 areas where they were spawned was the strongest connectivity pattern seen.

430

431 *4.2. Trajectory analysis*

432

433 Contrasting the route taken by ‘successful’ individuals (i.e. those that reached a nursery area)
434 and ‘unsuccessful’ individuals (i.e. those that failed to) is useful in understanding what physical or
435 biological factors could make the difference in successful recruitment. Differences in the biotic
436 and abiotic environment experienced along individual trajectories throughout the simulation were
437 contrasted between successful recruits and non-settlers. Trajectory analysis undertaken included
438 (1) physical processes (e.g. eddies, currents) that directed fish towards or away from favored
439 nursery areas; (2) topographic features that influenced trajectories; (3) temperature histories of
440 successful settlers and non-settlers; (4) optimal duration of the pelagic stage; and (5) correlations
441 between successful settlement and movement over viable habitat.

442

443 *4.2.1. Visualizing individual paths*

444 Visually examining trajectories of individuals within the model grid can be considered the most
445 basic form of trajectory analysis, but large number of simulated individuals made it hard to discern
446 useful information. Displaying individual trajectories for only certain spawning/natal areas, or
447 individuals of a certain life stage was more informative (i.e. Stockhausen et al., 2019a; 2019b).

448 The difficulty in disentangling ‘spaghetti’ plots led to the development of a ‘mean trajectory’ from
449 each spawning area to reveal the path individuals took on average. Visual examination of
450 trajectories identified notable interannual differences in transport, for example, in the numbers of
451 young ATF carried offshore in large oceanic eddies (Stockhausen, 2019a); while some of these
452 fish returned to the shelf, many failed to reach a suitable nursery area and were considered
453 unsuccessful and lost to the system.

454 Visual inspection showed that while most of the individuals that exited the GOA model domain
455 did so along its western edge. Some individuals exited the grid to the southeast (Stockhausen et
456 al., 2019b; Gibson et al., 2019), which is somewhat counterintuitive as the dominant flow in the
457 GOA is counter-clockwise. These individuals were transported in this direction by ephemeral
458 mesoscale processes, such as wind-driven currents, at the edge of the model domain. Visual
459 inspection of WP trajectories allowed an exploration of the transport from known spawning areas
460 in the GOA to nursery grounds. From the Cook Inlet spawning area, individual particles were
461 transported to Shelikof and Semidi islands region where juvenile aggregations have been
462 observed (Fig. 6ai; Wilson et al., 1996). In contrast, the East Kodiak Island spawning region
463 produced trajectories that entrained individuals into mesoscale eddies off the shelf (Fig. 6aii) while
464 the Shelikof Strait spawning area produced individual trajectories that connected to the Shumagin
465 Islands region, a known nursery ground for WP (Fig. 6aiii).

466

467 4.2.2. Path analysis

468 “Path Analysis” was performed to search for common trajectories or areas of the GOA used
469 heavily by the focal species. For each year, individuals were grouped by spawning area and
470 recruitment status (successful/unsuccessful), and trajectories were examined to determine which
471 grid cell each particle was in at each time step. All instances of a particle being in a cell were
472 tallied (considering particles that were retained in the cell and those that entered or left the cell)
473 to derive a ‘particle day count’ for each grid cell that showed common routes or retention regions
474 for particles. The results were used to see if the transport of individuals was notably different in
475 low and high recruitment years. For example, the path analysis for sablefish spawned in the
476 southeast GOA indicates that in a high recruitment year (2000), individuals were initially retained
477 close to the shelf break in what appears to be a small eddy (Fig 6bi) and the common successful
478 path seems to be moving onto the shelf early, just north of Baranof Island. Conversely, in a lower
479 recruitment year (2011) most individuals appear to be retained, at least for some time, in a large
480 eddy that transported them off-shore (Fig. 6bii). The Path Analysis did not show individual
481 transport routes, but instead showed a pattern of use of the shelf. It also showed, as did the visual
482 inspection of trajectories, that some settlers could be transported quite far offshore and still return
483 and settle. This would appear to indicate that entrainment in the Alaska Stream or the large
484 mesoscale eddies is not necessarily fatal, assuming a sufficient food supply in these eddies.
485 Visual inspection and the path analysis of PC trajectories showed that Amatuli Trough could act
486 as a transport pathway off the shelf when fish were in the top 20 m of the water column. Often the
487 Amatuli Trough is thought of as a means for early life stages of some species located deeper in
488 the water column to transit from deep slope or oceanic areas up onto the continental shelf (Mordy
489 et al., 2019). Our finding underscores the importance of differences in the life histories of the focal
490 species. An overlap coefficient (OC; Hinckley et al., 2016) was used to quantify the overlap of
491 path analysis matrices in low, median, and high recruitment years (i.e. Hinckley et al., 2019).

492

493 4.2.3. Tortuosity Index

494 Computation of a “Tortuosity Index” for trajectories allowed examination of their twistiness, or
495 tortuosity, to determine how direct a route each individual took, as it was transported to (or away
496 from) nursery areas. We defined tortuosity as the arc-chord ratio, the ratio of the total length of
497 the trajectory (actual distance traveled), to the direct distance between the endpoints (i.e.
498 spawning and settlement location). This index was computed for each individual in each group

499 and then averaged over spawning groups and recruitment status (e.g. successful, unsuccessful).
500 The tortuosity analysis for WP showed the most convoluted trajectories occurred in 2002,
501 indicating that in this year the trajectory paths were much longer than the straight-line distance
502 between starting and ending points. A reduction in tortuosity was observed from 2005 to a
503 minimum in 2008. Examples of more tortuous and less tortuous trajectory indices are shown in
504 Fig. 7.

505

506 4.3. Summarizing the biophysical experience

507

508 Environmental conditions could facilitate or impede groundfish transport, development, and
509 successful recruitment. To address H2, i.e. that recruitment variability is influenced by the survival
510 of individuals as they are transported from spawning to nursery areas, we developed a series of
511 environmental indices. Indices representing large-scale environmental processes in the North
512 Pacific Ocean and GOA (i.e. the Arctic Oscillation and the Pacific Decadal Oscillation) were
513 considered, as were more regional scale (i.e. eastern/western onshore/offshore) indices
514 developed directly from the GOA ROMS-NPZ model. Other environmental indices were
515 developed from IBM trajectory analysis summarizing the environment directly experienced by
516 each individual. Environmental variables considered included regionally-averaged upper water-
517 column salinity and temperature, integrated primary production, cross-shelf flow, temperature and
518 salinity along the trajectory, zooplankton biomass encountered, life-stage pelagic duration, and
519 days spent over suitable settlement habitat. As was the case with the connectivity and trajectory
520 path analysis, each environment index was expressed on an annual time scale so that they could
521 be correlated to stock assessment measures of recruitment (see section 5). Environmental
522 conditions for individuals spawned in different areas were often highly correlated. For example,
523 2002 was colder and more saline for all SF individuals regardless of where they were spawned
524 (Fig. 8). Similarly, the Degree Days index calculated for PC trajectories, which summed
525 temperature at each location for each day over the entire trajectory, found the temperature
526 experienced by settlers and non-settlers to be similar, suggesting that experienced temperature
527 was not a direct determinant of survival to settlement.

528 For WP, the different conditions encountered by successful and unsuccessful individuals were
529 further explored using a multivariate Empirical Orthogonal Functions (EOF) analysis performed
530 on the normalized (z-score) values of environmental variables encountered along each trajectory
531 during the year 2008. While a typical physical EOF would look for the spatial structure of a *single*
532 variable on a grid, multivariate EOFs (individual/variable/time) were used to explore coupled
533 biophysical modes that varied through time across individual trajectories. The environmental
534 experience between those individuals that successfully recruited to the nursery areas (“winners”;
535 Fig. 9a) versus those who were unsuccessful, i.e. those that did not settle, or exited the grid
536 (“losers”; Fig. 9b) was then compared. The goal was to seek out characteristic life histories that
537 lead to a winning or losing result. The results of the formal EOF analysis and environmental
538 indices suggest that there are more ways to “win” than to “lose” (Fig. 10). The two dominant
539 factors (1st and 2nd mode of the EOF analysis) have time amplitudes that explain life history
540 changes during different parts of the year. The first mode has a modest negative amplitude
541 followed by an abrupt change near late July - early August, while the second mode has a gradual
542 negative to positive trend from spring into summer, followed by a decline to zero amplitude in the
543 fall. Overall, the second mode explained most of the collective variance across individuals and
544 their sampled environment in the summer, while the first mode explains most of that collective
545 variance in the fall. The sharp change in amplitudes in early fall is a natural consequence of the
546 dramatic change at that time in several of the included variables (e.g. neocalanus and life stage).
547 In both modes, the individual ‘winners’ exhibit a broad range of loadings on any life history variable
548 (not shown), and the near-zero average of loadings for each biophysical variable over all those
549 winning individuals (Fig. 10 c,d) indicates no characteristic pattern of winning life history. This

550 indicates that each individual has experienced a rather unique sequence of life history events
551 (e.g. food, temperature, depth) on their way to settlement. Conversely, the losers tend to have
552 similar loadings on any particular variable; averaging over losing individuals reveals common
553 features of their life histories (Fig. 10 e,f). This averaging reveals that losing individuals are
554 gradually advected into deeper areas over the spring and summer (note the strong positive
555 loadings on bathymetry and the positive trend through time) and experience lower values of
556 temperature, euphausiids, and copepods and higher values of salinity (consistent with offshore
557 conditions) in the fall.

558
559

560 **5. Using GOA IERP models to predict recruitment**

561

562 *5.1 Correlating model indices with recruitment*

563

564 We computed the Pearson's correlation coefficient (Table 3) between our indices (computed
565 for 1996-2011) and the stock assessment estimates of recruitment (A'mar and Palsson, 2014;
566 Dorn et al., 2016; Hanselman et al., 2016; Spies et al., 2016; Hulson et al., 2015) to determine
567 which of the annual indices from our connectivity, trajectory, and environmental analysis might be
568 important to understanding and predicting the recruitment of groundfish. For some species, the
569 analysis was extended beyond simple correlation to involve linear models that combined multiple
570 connectivity and environmental indices and had greater explanatory power. Correlations between
571 the time series (1996-2011) of percent of individuals settled in each nursery area derived from
572 the PC connectivity analyses showed that the fraction settled in the Shumagin Island was
573 positively correlated with observed PC recruitment, indicating that this may be an important
574 nursery area. Settlement in this region was positively correlated with the North Pacific Index (NPI)
575 ($r=0.53$, $p<0.05$) and negatively correlated ($r=-0.54$, $p<0.05$) with the Multivariate ENSO Index
576 (MEI). This indicates that PC success in settlement and recruitment may be increased when the
577 GOA gyre circulation is low, enhancing retention and short-distance transport, and minimizing
578 transport out of the GOA. For PC, the Degree Days index for the individuals that exited the GOA
579 was significantly negatively correlated with normalized PC recruitment (A'mar and Palsson, 2014).
580 Also for PC, the Tortuosity Index for settlers was significantly positively correlated with
581 recruitment. These indices also show that restricted transport of early stages is important.

582 The total annual connectivity between all spawning and nursery areas for SF was positively
583 correlated to recruitment ($r=0.45$, $p=0.08$). The strongest correlation ($r=0.56$, $p<0.05$) to SF
584 recruitment estimates was with the cross-shelf flow index. The tortuosity index generated using
585 the WP IBM output showed an inverse relationship with recruitment ($r=-0.42$, $p=0.11$). This implied
586 that more direct trajectories led to higher recruitment and that recruitment is positively affected by
587 efficient transport to nursery areas.

588 While the identification of indices to reliably predict recruitment in isolation has proven elusive,
589 these analyses have identified consistent mechanisms that may underlie successful recruitment.
590 Submesoscale eddies might play a role by concentrating or dispersing food availability in some
591 areas and affecting the direct arrival to potential nursery areas. These results also present new
592 hypotheses that should be tested in future research efforts.

593

594 *5.2. Models for mechanistic understanding vs. recruitment prediction tools*

595

596 Using ecosystem models and IBMs as recruitment prediction tools to inform stock assessment
597 presents several challenges but is a worthy goal. IBMs do not directly predict recruitment, rather
598 they predict indicators for potential recruitment (e.g. the probability that individuals will be
599 successfully transported to a suitable nursery ground). Delivery to suitable nursery habitats is

600 necessary, but not sufficient for recruitment. A more realistic strategy for IBM integration into stock
601 assessments is to incorporate the strength of correlations between model predictions and past
602 empirical observations of recruitment (Kough et al., 2013). De Oliveira and Butterworth (2005)
603 suggest that recruitment indices from an IBM can be considered useful for management if the
604 index, or combination of indices, can explain >50% of the variability in past recruitment. While we
605 were able to achieve this goal for some of our focal species, such as POP (Stockhausen et al.,
606 2019), sablefish (Gibson et al., 2019), and PC (Hinckley et al., 2019), we argue that the true value
607 of an IBM to both assessment scientists and fisheries managers lies in its ability to compare the
608 relative importance of potential mechanisms underlying recruitment variability (i.e. to suggest *why*
609 a correlation might be evident). This type of approach helps to narrow the range of potential
610 environmental predictors for each species (Table 3).

611 Insights from the GOA IERP as to how recruitment is affected by environmental conditions and
612 the spatial relationships between spawning and nursery grounds may improve future estimates
613 of stock structure and essential habitat. Our findings also might contribute to the development of
614 appropriate ecosystem-based management schemes when planning for future climate regimes.
615 For example, an IBM could provide the basis for environmentally-forced recruitment for a
616 Management Strategy Evaluation (MSE; Punt et al., 2016b), with candidate management
617 strategies evaluated against each other based on responses to a set of hypothesized
618 environmentally-linked recruitment regimes provided by the IBM.

619

620 *5.3 Model results to explore short-term impacts and long-term projections*

621

622 Managers are interested in not only the short-term recruitment potential of a stock, but the
623 stock response to climate variability, climate change, and regime shifts. As part of the
624 Intergovernmental Panel on Climate Change Fifth Assessment Report (AR5, IPCC 2013), global
625 forecasts have been developed and used to explore the coupled global atmospheric and oceanic
626 response to anticipated changes in atmospheric CO₂. It is relatively common to see these long-
627 range forecasts used to drive regional biophysical models to quantify expected changes in ocean
628 productivity. While this approach has utility in understanding the physical response of the ocean
629 and lower-trophic level organisms, its utility lessens as we consider organisms with more complex
630 lifecycles and behavioral responses. Furthermore, while the IPCC predictions generally agree
631 that there will be warming over the next 50 years, salinity, not temperature is the dominant driver
632 in the GOA. This region has a very narrow continental shelf bordered by steep glacial mountains
633 thus the variability in freshwater runoff is highly influential (impacting the buoyancy-driven Alaska
634 coastal current, water column stability and mixing, and shelf/ocean exchange (Royer 1982, 1998;
635 Hill et al., 2015). It has been postulated that the IPCC models generally do a relatively poor job of
636 mimicking the freshwater inputs (Stabeno, NOAA-AFSC, pers. comm.). We do not believe that
637 recruitment predictions using these long-range predictions of forcing will have strong predictive
638 power for GOA fisheries because freshwater is essential to understanding the physical, and
639 therefore the biological, dynamics of the GOA. The approach taken within the framework of the
640 GOA IERP (i.e. development of recruitment indices for focal species and attempting to understand
641 the recruitment response and underlying mechanisms of each species to any given physical
642 regime), is a more robust strategy and more likely to produce useful results that could inform the
643 adaptation of management schemes within the context of regime shifts.

644

645 *5.4. Linking spatial model outputs to systemwide recruitment*

646

647 A key result of the GOA IERP has been to describe the diversity of processes in the GOA,
648 especially the formal recognition of distinct dynamics in east and west (Zador and Yasumiishi,
649 2016) and connectivity between domains (Goldstein et al., 2019; Siddon et al., 2019). These
650 domains represent diverse user communities with different concerns and priorities. A “whole

651 region” model, such as a stock assessment or bulk biomass (non-spatial) food web model, may
652 be limited in its ability to deal with spatial structure, especially if links between east and west vary
653 over time.

654 An important part of management for some wide-ranging stocks in the GOA (e.g. SF) is area
655 apportionment of fishing quotas. To this end, the most management-useful recruitment indices
656 derivable from IBMs may not predict absolute recruitment, but rather spatially-relative recruitment,
657 how strongly east and west are coupled, or the connection into/out of the GOA, and whether these
658 change over time. IBMs are uniquely suited to produce such spatially derived indices, or for
659 producing spatial recruitment estimates for seeding a spatial whole-ecosystem model.

660

661 *5.5. The effects of the five focal species on the food web*

662

663 To bridge the gap between the ROMS-NPZ-IBM modeling efforts and management of fisheries
664 in the GOA, we initially proposed to incorporate our indices into an Ecosim food web model of the
665 western GOA (Gaichas et al., 2010, 2011). The NPZ components and five focal fish species of
666 GOAIERP are directly connected by the distance of single predator/prey links to almost all the
667 124 functional groups in the Ecosim model (Fig. 11), showing the central role of the focal species
668 in the ecosystem.

669 The food web model can be used to show the sensitivity of the food web to variability in the
670 biomass of the focal groundfish species, as well as explore uncertainties in our current knowledge,
671 and inform prediction, particularly in recruitment. Through a series of simulations with the food
672 web model, we lowered the recruitment production of each of the focal species by 10%, in turn,
673 and simulated the new ecosystem state resulting from the perturbation. Monte Carlo sampling
674 was used to generate a range of potential ecosystems within the error bounds of the input
675 parameters (biomass and diets). WP and ATF had substantial ecosystem impacts (Fig. 12), while
676 PC, POP, and SF led to fewer strong connections within the food web. One of the most uncertain
677 results is between juveniles and adults of the same species. This suggests that some
678 predator/prey relationships (e.g. competitive links between WP and POP) seem more certain than
679 the stock-recruitment relationship between juvenile and adult WP – the total overall production
680 (i.e. food available to all planktivores) may be more certain than the relative age structure of each
681 population. The implication for recruitment studies is that we should study a range of ecosystem
682 effects over time. Counter to conventional wisdom, recruitment predictions and stock/recruitment
683 relationships may be more uncertain than predictions based on predator/prey relationships. In
684 terms of focus for future research, it suggests that some species (e.g. POP, SF) are less
685 connected to the ecosystem, and the benefits of examining multispecies interactions for these
686 two species may be low compared to WP, PC, and ATF. This highlights the importance of
687 continuing to improve the skill of predictive dynamic models that assimilate the complexities of
688 the broader ecosystem and their impact on the success of early life stages.

689

690 **6. Application of model results to fisheries management**

691

692 *6.1 Use of recruitment estimates in stock assessments*

693

694 Fisheries management often aims to maintain a population at or above the biomass that
695 provides maximum sustainable yield (B_{MSY}). B_{MSY} is strongly influenced by the productivity of a
696 stock, and target biomasses for stocks in the North Pacific including the GOA, are set using either
697 estimated stock-recruitment relationships or a proxy for B_{MSY} (NPFMC, 2015). Confidence in
698 recruitment estimates (rather, estimates of cohort strength when a cohort is first monitored by
699 surveys) may be low depending on the age individuals recruit. Therefore, the management
700 reference points (e.g. target biomass and yield) for most federally-managed groundfish in Alaska

701 are not based on maximizing theoretical yield calculated from uncertain stock-recruitment
702 relationships, but rather on a more precautionary adult-based spawner/recruit biomass proxies
703 (NPFMC, 2015).

704 Recruitment estimates based on single-species stock assessments are model outputs,
705 reflecting several tunable parameters based on simplified assumptions, such as the value of age
706 and time-invariant natural mortality. The stock assessment models have been refined over time
707 given their basis for determining stock status (e.g. whether it is overfished) and sustainable
708 harvest levels. The uncertainty in stock assessment models has increased in recent years,
709 especially following unanticipated ecosystem and environmental impacts. As such, the Alaska
710 groundfish management system has adopted methods for reducing quotas using control rules
711 and assessment outcomes based on quantification of ecosystem risk (Dorn and Zador, 2020).
712 This “risk table” approach relies on indicators of ecosystem processes, both from direct
713 observations and as derived from ecosystem models. One goal of IERPs is to further understand
714 ecosystem processes to better predict population fluctuations that could be used to improve
715 recruitment predictions, quantify ecosystem risks not captured in single-species assessments,
716 inform fisheries management decisions, and forecast production. This is particularly relevant if
717 model results could improve estimates for context-dependent shifts in key input parameters to
718 stock assessment models (e.g. virgin recruitment, R_0 ; the natural mortality rate M ; and steepness
719 of the stock-recruitment relationship h ; Punt et al., 2020).

720 Predicting recruitment (at age-0), as our IBM’s currently do, may not be indicative of stock
721 projection. Still, there may be value to short-term forecasts for short-lived species, especially in
722 terms of evaluating potential risks to upcoming recruitment. For example, an unanticipated marine
723 heatwave occurred in the GOA between 2014-2016, leading to major declines in survival of
724 juvenile and adult Pacific cod that were not detected until those cohorts matured in later years
725 (Barbeaux et al., 2020). If models could provide recruitment estimates based on modeled
726 oceanographic conditions before recruitment is observed in the surveys for adults, managers
727 might develop control rules that use that information in advance. Additional information (e.g.
728 recruitment indicators from IBMs such as those developed in the GOA IERP) may be useful in
729 supplementing single-species stock assessments. Context and goals are important; a recruitment
730 indicator from a model such as an IBM should *not* be expected to provide tactically precise
731 estimation for maximizing yield. Rather, the predictions of an IBM (or most other ecosystem
732 models) might more effectively focus on improving adaptability in management, by either
733 anticipating or informing the response to events or by further accounting for environmental effects
734 and ecosystem interactions when evaluating and implementing management action.

735 Management strategy evaluation (MSE) is a simulation framework that quantifies the expected
736 performance of management processes, using summary statistics derived from operational
737 objectives (Sainsbury et al., 2000). It attempts to model the entire management system, including
738 the ‘true’ state of the resource, stock status, management advice, and decision processes, as well
739 as fleet dynamics. MSEs can be used to assess assumptions (Smith, 1994; Smith et al., 1999)
740 and consist of three components: an operating model, an estimation model, and a harvest control
741 rule (HCR). Operating models simulate populations and can be based on stock assessments
742 (Smith et al., 1999; Punt et al., 2016a), or indeed the ROMS physical system model, the NPZ
743 lower trophic level model, and IBMs of focal fish species developed in the GOA IERP. Estimation
744 models attempt to describe the dynamics of those populations based on generated observations.
745 Various operating models might be used to evaluate the impact of incorrect assumptions about
746 the population dynamics in the estimation model (e.g. Punt, 2003; Cur some non-essential text
747 A’mar et al., 2010). Ecosystem models and IBMs might provide diagnostics to identify data gaps
748 and/or conflicts. A critical component to this, and one that needs to be iteratively evaluated, is
749 model skill (see section 7.3).

750

751 *6.2 Use of dynamic ecosystem models*

752
753 Dynamic ecosystem models, such as the physical oceanography model and the lower trophic
754 level models used in the GOAIERP, can be useful to management in multiple ways. These
755 Eulerian models can be used to provide an alternate way to estimate environmental conditions
756 between surveys (in space and time) that is not based on interpolation of survey data. These
757 conditions provide a more complete picture of the ecosystem, thus allowing a broader context for
758 interpreting observations, for example through “risk tables”. In addition, relationships between
759 hindcast ocean conditions (validated against survey data) and fish production or stock status are
760 being used as part of the National Oceanic and Atmospheric Administration’s (NOAA) Rapid
761 Climate Assessment (Spencer et al., 2019) to prioritize future research based on which species
762 show the most sensitivity to measured variability as well as long-term forecast decisions. For
763 example, nowcasts of ocean conditions from a ROMS model of the Bering Sea (Kearney et al.,
764 2020) are currently included in the multispecies analysis of Bering Sea Walleye Pollock as part of
765 that region’s pollock stock assessment (Holsman et al., 2019). These modeled ocean conditions,
766 if related theoretically to fish production, may be used as indicators during the management
767 process to address uncertainty in measurements of the current ecosystem state. If current ocean
768 conditions do not favor recruitment that might add strength to a stock assessment estimate of
769 poor recruitment and support more precautionary measures until recruitment indices are later
770 validated through the sampled abundance of adult fish.

771 In addition to providing direct, tactical management advice based on the modeling of past and
772 current ocean conditions, the results of dynamic ecosystem models are important for strategic
773 planning, especially in light of ongoing and future climate change. For example, the direct coupling
774 of Bering Sea ROMS-NPZ outputs, projecting ocean conditions out to 2100, was used to drive a
775 food web model of the Bering Sea ecosystem to assess the impact of alternative management
776 strategies on the long-term sustainability of key fish stocks (Whitehouse et al., 2021). Such end-
777 to-end coupled modeling efforts can test the long-term resilience of current fisheries management
778 strategies (Holsman et al., 2020), and are an important component of planning management
779 strategies in the face of climate change, for example through formal fisheries ecosystem plans
780 (NPFMC, 2018).

781 Incorporation of ecosystem-level scientific advice into effective decision-making for marine
782 resource management is increasingly being part of a larger synthetic process including
783 stakeholders, scientists, decision-makers, and the public. Examples of such an approach are
784 Integrated Ecosystem Assessments (IEAs), which aim to be “a synthesis and integration of
785 information on relevant physical, chemical, ecological, and human processes in relation to
786 specified management objectives” (Levin et al., 2009). A suite of models for an ecosystem,
787 including ocean process models and a range of fisheries models, as described in Punt et al.
788 (2016b) and including single-species, multispecies, and IBMs, can contribute strongly to providing
789 strategic advice.

790 791 *6.3 Application of IBMs*

792
793 IBMs have been used for several purposes, including hypothesis generation and testing,
794 defining marine protected areas and spawning areas, stock structure studies, and connectivity
795 and recruitment studies. Among these, hypothesis generation and testing are probably the most
796 powerful. The IBMs developed under the GOAIERP provide a holistic way of looking at how the
797 early life stages of a groundfish species interact with the environment in a way that surveys and
798 process studies cannot. They encapsulate our best understanding of species' early life history
799 and, by providing a way to “scale-up” results from process studies and surveys to a much larger
800 area, IBMs can help develop a mechanistic understanding of the processes underlying fisheries
801 population trends.

802 Management of marine resources could also benefit from IBMs that quantify larval connectivity.
803 This knowledge might help guide policy because fished stocks are frequently not constrained
804 within the geopolitical boundaries in which they are managed. IBMs can reveal larval ‘corridors’
805 (i.e. spatial regions that regularly concentrate and nurture pelagic larvae during their ontogenetic
806 migration to nearshore environments). Stressed fisheries may benefit from the establishment of
807 a network of selectively-located marine protected areas corresponding to IBM-identified critical
808 regions e.g. spawning or nursery areas. Our findings, for example, indicate that there is likely at
809 least some connectivity between the pollock stocks in the GOA and the Bering Sea (Parada et
810 al., 2016) and between sablefish in Canadian waters south of the GOA, and the GOA population
811 (Gibson et al., 2019). These findings could be potentially transformational for the management of
812 these stocks and warrant further investigation.

813
814

815 **7. Model development within an Integrated Ecosystem Research** 816 **Program**

817

818 *7.1 Model development and design – approach and constraints to integration*

819

820 The development and application of numerical models to explore ecosystem dynamics in the
821 context of an IERP provides an opportunity to enhance overall program goals, focus research
822 questions, and provide a broader context for the interpretation of observational data and process
823 studies. The IERP framework is designed such that integration of the various research efforts is
824 a fundamental program goal. In this context, modeling could serve as a way to approach this
825 integration in a synthetic, comprehensive, and quantifiable fashion. Due to prohibitive costs, field
826 or laboratory studies directly address only specific, tightly-focused hypotheses regarding
827 environmental processes over generally limited temporal, spatial, and environmental extents. In
828 addition to directly addressing the project hypothesis, models within the GOA IERP aimed to
829 provide a broader spatial and temporal reference framework to aid in the interpretation of
830 observations and to identify areas of sampling. In this context, model results can guide sampling
831 design or follow-up field or lab studies by suggesting times of year or regions of interest, critical
832 to better resolving life history and population processes, potentially leading to reduced
833 uncertainty. Numerical models can also provide a means to synthesize field observations and lab
834 studies using an objective, coherent framework to integrate data across larger spatial and
835 temporal scales to predict regional-scale consequences.

836

837 *7.2 Timing*

838

839 Within an integrated research program, time needs to be allotted to allow incorporation of the
840 results of field and lab studies within the models (e.g. new parameterizations of model processes
841 or initial conditions). Fundamental questions related to timing and modeling in the context of an
842 integrated program are (1) sequencing, and (2) the purpose of the modeling effort. Most new data
843 from the GOA IERP project became available only at the end of the project. Within an IERP,
844 modeling, fieldwork, and lab studies should thus be approached as a set of interactive, iterative
845 components, and time and opportunities for interaction among these components needs to be
846 factored into program design to achieve true integration. Models might be used at the outset to
847 identify questions and processes for the field study provided time and effort had been previously
848 dedicated to model development and implementation – either under a prior program or under the
849 first phase of an IERP. Alternatively, sampling and field analyses might be conducted in advance
850 to aid in model conceptualization and inform model development and parameterization, or
851 validation. This issue of timing presents one of the greatest challenges in an IERP and requires

852 careful consideration. The appropriate timing of model development will depend on how the
853 modeling objectives are defined within the context of the overall study. If models are intended to
854 guide sampling, they need to be able to produce the required outputs early in the program, prior
855 to survey design. Within the GOA IERP, the coupled ROMS-NPZ model had been developed in
856 advance of the implementation of a field program and the modeling team was able to provide the
857 broader GOA IERP research community with output from multi-year, physical, and lower trophic
858 level simulations at the program onset. This early availability of model output was used to guide
859 the sample design of the observational program (Dickson and Baker, 2016). For example, the
860 initial sampling design proposed was focused on two grids on the eastern and western sides of
861 the GOA. Model outputs indicated that the large spatial gap in the central GOA was problematic
862 due to the along-shelf connectedness of currents in the GOA and the original sampling design
863 was modified accordingly.

864 Timing within the program duration is also an important consideration. The GOA IERP was
865 conducted over a relatively short, four-year time-frame. This and the inherent delay in the
866 availability of the observational data meant that there was only a limited amount of time for
867 conducting multiple iterations and improvements to any of the models. Only at the end of the
868 GOA IERP were we able to utilize the results from the observational components to improve the
869 representations of physical and ecological processes in the numerical models.

870

871 *7.3 Model validation*

872

873 *7.3.1 ROMS and NPZ model validation*

874 Validation of the lower trophic level and physical models used to drive the IBMs is possible to
875 at least some degree using bio-physical observations from shipboard sampling, moorings, and
876 satellite observations (Coyle et al., 2012, 2013, 2019; Hinckley et al., 2009b; Hermann et al.,
877 2009b). However, the mesoscale features of the GOA are influenced by fine-scale details of
878 atmospheric forcing (e.g. storms, episodic events) and chaotic (nonlinear) physical processes;
879 consequently predictions from the hydrographic and thus the lower trophic level models will never
880 precisely match field observations at any given location and time. Thus, point-by-point
881 comparisons of model and data can be misleading, especially when the data used for model
882 validation do not come with associated measures of error, and may lead to disregarding
883 informative results. Ideally, the models will capture summary statistics (e.g. the total kinetic
884 energy) of eddies and meanders in a given year, even if they cannot precisely capture the detailed
885 timing or locations (Coyle et al., 2012). Numerical models can only be as “good” as the field and
886 lab studies that support them and the utility of model-data comparisons is complicated by the
887 varying spatial and temporal resolution of the observations. Because of limits to data availability,
888 we were only able to assess the skill of the Eulerian hydrographic and lower trophic level models
889 by comparison of seasonal climatologies from aggregated data and model output (Hermann et
890 al., 2019; Coyle et al., 2019). Confidence in a model to replicate ecosystem dynamics on this
891 broad time scale should give confidence in its ability to suggest mechanisms and processes
892 important to ecosystem dynamics.

893

894 *7.3.1 IBM validation*

895 IBMs built on top of the lower trophic level and physical model add another layer of
896 assumptions and potential compounded errors. There are limited methods that can be used to
897 validate IBM predictions of larval dispersal and transport of individuals (North et al., 2009).
898 Comparing trajectories from model predictions to those from satellite-tracked drifters is of minimal
899 use due to the limited number of drifters that can be deployed, and the predisposition of drifters
900 to diverge from larval fish trajectories because of larval behavior. Chemical marking of individuals
901 might be a useful validation technique but is only applicable when populations are small, mortality
902 is low, and the likelihood of recapture is reasonable. One of the most straightforward ways to

903 corroborate dispersal and transport simulated by an IBM is by comparing modeled and empirical
904 spatial distributions of larvae and juveniles. Although it cannot be known if individuals from
905 spawning sites are the same individuals that are caught during later surveys, this comparative
906 approach can be useful when the sources of individuals caught in the field are relatively well
907 known. Differences in spatial distributions between IBM model output and survey data may be
908 attributed to several factors including differences in initial conditions (i.e. the location and timing
909 of modeled and actual spawning). Limitations in the hydrodynamic model, as well as limitations
910 in simulating larval and juvenile behaviors (i.e. the absence of directed swimming) could also
911 result in a miss-match between IBM output and survey data.

912 Even if IBMs might be evaluated based on their ability to match observations, there is an
913 enormous data deficit for model validation that cannot realistically be addressed, because of the
914 necessary spatial (i.e. high resolution, large extent) and temporal (i.e. high resolution, long term)
915 sampling required to resolve potentially-confounding effects. A better approach would be to
916 compare (appropriately averaged) model results under different scenarios of assumed critical
917 processes to generate testable hypotheses that can be subsequently addressed by targeted
918 fieldwork. In an IERP, part of the goal of the observational data collection should be to improve
919 the models' characterization of potentially-critical processes – by design, not just happenstance.
920 Such an approach would help ensure that the data collected are useful to the modeling effort and,
921 in turn, the model results are reflective of reality. Full consideration of the data required to support
922 model development and validation during the design phase of the field program would facilitate
923 better integration.

924 Model validation and model improvement should be treated as an iterative process where
925 mismatches between model and data can be used to improve the model and the surveys (if they
926 can be specifically designed to compare with model output). We recommend a multistep process
927 for the validation of spatial output of biophysical IBMs, starting with visual comparisons and simple
928 descriptive statistics, followed by the calculation of indices for features of interest, and finally,
929 using statistical and geostatistical approaches that can give measures of statistical significance
930 to the differences/similarity between spatial model output and data. For some purposes, i.e.
931 management applications, “success” in validation may need to be pre-defined by delineating
932 success thresholds. In addition to broad comparisons of GOAIERP IBM model output to
933 presence-absence data, the principal datasets available to compare with predictions were the
934 recruitment time series for each species, estimated as part of stock assessments conducted by
935 NOAA Fisheries (Spies et al., 2015).

936

937 *7.4 Interaction*

938

939 Determining the impact of advection and the environmental experiences of early life-stage
940 groundfishes as they are transported between spawning and nursery areas is an inherently
941 interdisciplinary problem that requires consideration of the physical and lower trophic level
942 environments in conjunction with early life history dynamics. Interaction and integration between
943 program components are essential. Here there are clear challenges. The GOAIERP involved over
944 50 researchers from a wide background of disciplines. Although the central hypothesis for
945 GOAIERP was established prior to solicitation of individual research projects, the specific
946 observational and modeling efforts in this program were proposed in isolation. As such, despite
947 the initial conceptualization of how the field data could guide modeling efforts, the models were
948 largely divorced from the field and lab studies conducted as part of the GOAIERP. We found that,
949 while many of the researchers were open to the idea that models could be a useful tool in survey
950 design and data interpretation, initially there was an underlying reluctance to fully embrace
951 modeling tools as useful additions to the program. This skepticism was partly due to mistrust in
952 the ability of models to adequately represent ecosystem and life history dynamics, and partly to
953 unfamiliarity with the proposed modeling approaches. More effort might be devoted to

954 demonstrating the utility of models, not only to explain mechanistic processes but also to suggest
955 further research topics and sampling design.

956 While the level of *interaction* between the modeling and field/lab components improved
957 substantially throughout the four-year GOIERP, there was a consensus that there was insufficient
958 time in the program allotted to promote effective *integration* between the modeling, field, and lab
959 components. In addition to the challenges of matching the timing of field programs and model
960 development, lack of ‘trust’ in the models may hinder full integration between modelers and
961 observationalists. As such, opportunities allowing (or forcing) team members from different IERP
962 components to come together, interact, and understand each other’s needs and contributions was
963 time well spent. Within the GOAIERP, directed program management and successive annual
964 meetings provided improved opportunities for such interactions: integrated presentations that
965 necessitated cross-disciplinary interaction prior to the meetings, required ‘speed dates’ between
966 small groups of researchers that may have never previously interacted, and time for general
967 discussion. In retrospect, we feel that the importance of this “face” time cannot be overestimated.
968
969

970 **8. Successes and lessons learned**

971

972 *8.1 Successes and shortfalls of the IBM modeling effort*

973

974 Here we discuss the general advances and limitations experienced in the GOAIERP modeling
975 effort related to model development and validation, developing a conceptual framework, and
976 synthesizing insights related to environmental indices and IBM outputs related to connectivity,
977 trajectory, and path analyses. Insights into each species and IBM framework and what we have
978 learned about potential recruitment mechanisms are summarized in Table 3.
979

979

980 *8.1.1. IBM methodological development*

981 One of the principal successes of the modeling effort was the methodological advancement in
982 IBMs. By providing a common objective and highly coordinated iterative process for exchange
983 between modelers approaching a common question, in a common system, with similar tools but
984 distinct targets (i.e. unique species and data resolution), the GOAIERP enabled a useful
985 exchange of ideas and approaches. By leveraging existing development code and analysis tools
986 GOAIERP was able to quickly produce five new or significantly updated, regionally-specific,
987 models.
988

988

989 *8.1.2. Insights from connectivity analyses*

990 Overall, model simulations showed agreement with our previous knowledge of prospective
991 viable spawning areas (Fig. 13). The exact spawning areas for PC, POP, SF, and ATF are not
992 known. IBM results for both POP, ATF, and SF indicate that spawning areas in the eastern GOA
993 would generally be more successful in terms of recruitment to inshore nursery areas than those
994 in the western GOA, due in large part to the counter-clockwise nature of the general GOA
995 circulation pattern and the offshore spawning areas of these species. Typical dispersion distances
996 on the order of several hundred km were found for these species. For PC, the IBM results suggest
997 settlement from all spawning areas is high, except for the West Shumagins area as individuals
998 spawned there were most likely to be swept offshore into the GOA basin. These results point to
999 the immediate need to refine our knowledge of actual spawning areas for PC, POP, SF, and ATF.
1000 More is known about WP spawning locations, and the prevailing hypothesis is that Shelikof Strait
1001 is the primary pollock spawning area in the GOA while the Shumagin Islands provide the main
1002 nursery area. The WP IBM suggests that the spawning areas in the GOA most likely to produce
1003 successful settlers are indeed in Shelikof Strait.

1004 Below we summarize what the GOA IERP IBMs tell us about likely connectivity between
1005 spawning and nursery areas for each of the five groundfish species. While connectivity appears
1006 to be able to explain a portion of recruitment variability, it is clear from our analysis that it is not
1007 the only factor.

- 1008 • *Walleye Pollock*: The WP IBM suggests a pattern in connection strength characterized
1009 by higher retention along the Inner Shelf domain (40-90%) and lower retention in the
1010 inner offshore region at the larval stage (30-40%). Connectivity between spawning
1011 grounds in the inner shelf domain (Kenai, North Kodiak, South Kodiak, Chirikof, East
1012 Shumagin, West Shumagin) and nursery grounds downstream (to the southeast) had a
1013 probability of ~0.2. A weaker connection (at the larval stage) with similar probability
1014 values was also demonstrated between spawning areas in the inner shelf domain and
1015 the inner offshore domain. The strongest connection (probability ~0.1 to 0.4) was
1016 between the inner offshore domain and the inner shelf domain and is evidence for cross-
1017 shelf transport to the inner shelf. For juveniles, a direct relationship was found between
1018 assessment-based recruitment anomalies and the PC1 from the EOF analysis ($r=0.41$,
1019 $p=0.11$), highlighting the importance of the environment through which the individuals
1020 were transported.
- 1021 • *Pacific Cod*: The PC IBM indicates that the young stages of PC generally do not disperse
1022 far from natal areas. Retention of individuals in areas where they are spawned was the
1023 strongest connectivity pattern followed by cross-shelf transport from the deep spawning
1024 areas to nearby shallow nursery areas. The restricted transport that we observe in our
1025 model results is, in part, because PC eggs are attached to the bottom for nearly a month
1026 of their early life and because the movement of simulated juveniles is restricted upon
1027 settlement. Settlement strength in Icy Bay and the West Shumagin area had the
1028 strongest correlations with recruitment estimates from stock assessments for PC. We
1029 found significant correlations between large-scale climate indices (i.e. the NPI and the
1030 MEI) and hypothesized that slower gyre circulation enhances retention and cross-shelf
1031 transport to nearshore nursery areas.
- 1032 • *Pacific Ocean Perch*: The POP IBM indicates that dispersal distances for individuals
1033 successfully settling were on the order of several hundred km. Connectivity is strongest
1034 (median~5%) between parturition areas in the southeast (Sitka and Cross Sound) and
1035 nursery areas in the central GOA, while somewhat weaker connections (median~3.8%)
1036 exist between Cross Sound and nursery areas in Icy Bay. The fraction of individuals
1037 originating from the Sitka and Prince William Sound spawning zones that successfully
1038 settled in any nursery area explained the greatest fraction of variance (66%) in
1039 assessment-based recruitment estimates (adjusted $r^2=0.62$; empirically-determined
1040 family-wise p -value <0.05).
- 1041 • *Sablefish*: The SF IBM suggests that the strongest connectivity (median~1.0%) in the
1042 GOA is between spawning areas over the continental shelf in the southeast (Sitka and
1043 Cross Sound) to shallow nursery areas in the central GOA. The direction of connection
1044 was generally from east to west, with very little retention of individuals within the same
1045 alongshore zones, or connectivity to regions to the east of a spawning zone. The total
1046 connectivity (i.e. the proportion of individuals settling to suitable nursery sites anywhere
1047 in the GOA, regardless of spawning area) correlated more strongly ($r=0.45$, $p=0.08$)
1048 with assessment-based recruitment estimates than the connectivity associated with any
1049 one spawning area or settlement site.
- 1050 • *Arrowtooth Flounder*: Similar to the results for both POP and SF, the ATF IBM suggests
1051 connectivity (median~3.5%) is the strongest between spawning in Sitka and Cross
1052 Sound and coastal nursery areas in the central GOA. However, in contrast to the results
1053 for those species, the second strongest connectivity (median~2.5%) was between

1054 spawning areas in Yakutat and Icy Bay and nursery areas in North Kodiak. The fraction
1055 of individuals successfully settling in nursery areas in North Kodiak, regardless of
1056 spawning area, was positively correlated ($\rho=0.64$, but the p-value, corrected for multiple
1057 comparisons, was not significant) with estimates of recruitment from stock assessments
1058 and accounted for 34% of the variance in recruitment.
1059

1060 *8.1.3 Insights from trajectory and path analyses*

1061 Environmental conditions experienced by the early life stages of the five focal groundfish
1062 species may be equally as important in determining recruitment success as connectivity. For
1063 example, in the case of SF, lower trophic level production during the first year of life, as individuals
1064 are transported from the deep offshore spawning areas to the shallow coastal nursery areas,
1065 correlated as strongly with recruitment as the connectivity index. Together these two variables
1066 were able to account for 50% of the recruitment variability as predicted by the stock assessments.
1067

1068 *8.2 Using ecosystem modeling approaches to inform fisheries management*

1069
1070 Depending on age at recruitment, confidence in the most recent year's predictions may not be
1071 strong. As a result, management strategies in Alaska currently use adult-based spawning
1072 biomass proxies (NPFMC, 2015). There have been increasing calls for EBFM (Link, 2002; Pikitch
1073 et al., 2004; Link, 2010; Fogarty, 2014; Fulton et al., 2014; NMFS, 2016) as it is clear that
1074 reductionist single-species stock assessment approaches that ignore broader environmental
1075 conditions overlook important implications and impacts on recruitment and survival (Hollowed et
1076 al., 2001; Duffy-Anderson et al., 2005). Model-informed estimates of the success of early stages
1077 and recruitment of juvenile fish into the adult population in the context of climate-driven variability
1078 could promote a broader ecosystem understanding and improved predictions of year-class
1079 strength for these GOA fish populations.
1080

1081 *8.2.1 Use and application of existing data and models to predict recruitment*

1082 GOA IERP models might be applied to inform recruitment estimates in the following ways:

- 1083 ROMS-NPZ outputs might be applied to understand patterns in annual recruitment using
1084 retrospective data and be used to develop new data streams to improve predictive
1085 capacity and determine environmental conditions that increase recruitment success or
1086 failure.
- 1087 IBMs for each of the focal groundfish species might be updated with emerging species-
1088 specific information and used to further identify data gaps and guide the development of
1089 process studies.
- 1090 IBM, ROMS, and NPZ models might be used to better inform differences in the distribution
1091 of recruitment across distinct spatial domains.
- 1092 Data and model outputs might be applied to characterize important ecosystem processes
1093 at distinct local and regional scales within the GOA.
1094

1095 *8.2.2. Development of a conceptual framework and species profiles*

1096 One success of the modeling effort has been to refine conceptual models and profiles for each
1097 of our focal groundfish species. This effort has allowed us to identify emergent properties of
1098 groundfish populations, distinguish significant phenomena determining the response of
1099 populations and communities to perturbation, and consider what aspects of biological or
1100 behavioral traits promote resilience species level (e.g. life history variation, phenotypic plasticity),
1101 community level (e.g. functional redundancy), ecosystem level (e.g. network connectivity). This
1102 information might help direct climate vulnerability analyses to assess the current status and future
1103 risk to populations, and also in determining which stocks would benefit from further integration of
1104 environmental data in their assessment.

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8.2.3. *Development of predictive environmental indices*

We had hoped the conclusion of our modeling work would enable us to develop a series of indices, or identify easily measurable variables that would be useful for predicting recruitment of the focal groundfish species. Our results suggest that recruitment variability is more complicated to predict. While the variability in successful transport of individuals from spawning areas to nursery grounds can explain some of the variability in recruitment, the variability in transport could not be tied to just one or two atmospheric or oceanic variables that could be easily measured. We had only a relatively short time series of model outputs to use in our multiple regression analyses so our statistical power was limited. With longer time-series it might be possible to identify more powerful indices or combinations of indices. Despite our inability to identify simple recruitment prediction indices, our work has helped shed light on the mechanisms that may underlie successful recruitment. For example, in the case of PC, slower gyre circulation may lead to increased retention and reduced transport, which may enhance recruitment. For species that spawn along the continental shelf break (POP, SF, ATF), conditions that promote on-shelf transport in the east and higher spring and summer production may favor increased recruitment.

8.2.4. *Ecosystem approaches to management*

The IEA process (Fig. 14) is a iterative approach to EBM or EBFM and involves several steps that integrate science and management. Ecosystem-based management goals should be established with strong engagement from managers, stakeholders, and the public. Ecosystem indicators must then be developed to track progress towards goals and ecosystem assessments conducted to determine ecosystem status, risk, and uncertainty. Ecosystem models, and species-specific models embedded in larger ecosystem models such as those presented here, can help to 1) identify policy tradeoffs inherent in goal setting (e.g. tradeoffs between predators and prey, or habitat and fishing); 2) identify or produce indicators (e.g. nowcast ocean conditions); 3) describe the theoretical or empirical relationship between indicators and goals (e.g. how productivity is linked to ocean conditions); 4) estimate ecosystem-based reference points during stock assessments (e.g. species vital rates, or multispecies sustainable yields); 5) quantify uncertainty and identify where established ecosystem relationships or management policies may no longer hold (e.g. ecosystem tipping points, thresholds or regime shifts); 6) fill in observation gaps, predict future conditions; and/or 7) serve as simulation testbeds for management strategies. Uses may be found for individual model outputs derived in the analyses described here to inform the effects of climate and oceanographic conditions on species outcomes and interactions.

8.2.5. *Ecosystem indicators report card*

As part of its annual stock assessment cycle, the National Oceanic and Atmospheric Administration (NOAA) Alaska Fishery Science Center (AFSC) produces Ecosystem Report Cards and Ecosystem Assessments for the Bering Sea, GOA, Aleutian Islands, and Alaskan Arctic (e.g. Zador et al., 2017b). These report cards contain a subset of relevant indicators chosen by ecosystem experts in consultation with stakeholders. Indicators are chosen to track ecosystem status from climate through living resources and ultimately humans. This information is presented to the North Pacific Fisheries Management Council immediately prior to quota-setting and has been used to support the adjustment of fishing quotas downward in cases where extra precaution was warranted (Dorn and Zador, 2020). Ideal indicators fill gaps in knowledge, can be characterized on annual scales, have long time-series, and are available “now” (Stephani Zador, NOAA-AFSC, personal communication). AFSC ecosystem assessment authors are working with PIs of the GOAIERP project to transition GOAIERP research into new indicators that might be produced on an ongoing basis.

9. Conclusions

The GOA is a dynamic and highly productive ecosystem that supports important fisheries and the communities dependent on them (Ormseth et al., 2019). The GOA IERP was one among many efforts made to develop large, coordinated, integrated research programs to improve understanding of ecosystem dynamics and fisheries population fluctuations in the North Pacific (Lindeberg et al., 2022). Several important findings from this effort have emerged.

9.1 *The gauntlet hypotheses*

The GOA IERP proposed the ‘gauntlet hypothesis’ – that the primary determinant of year-class strength for GOA groundfishes is early life survival. It was proposed that this is regulated in space and time by climate-driven variability, offshore and nearshore habitat quality, larval and juvenile transport, and settlement to suitable demersal habitat. We found that while the connectivity between spawning areas and nursery sites alone was able to explain significant amounts of variability in recruitment of some groundfish (i.e. SF ~20%, POP ~30%, ATF ~30%) it did not explain more than half of the recruitment variability for any species. While some transport dynamics may have been missed or underestimated, due to the resolution of the model and its skill in simulating fine-scale oceanographic and biological processes, our findings lead us to reject modeling hypothesis H1 and conclude that the proportion of young fish transported from offshore spawning areas to nearshore nursery areas cannot be the driving factor affecting recruitment. This points to the importance of environmental influences other than transport on young larvae (e.g. spatial predation) or to post-settlement processes in determining recruitment success. Because we could not generally incorporate mortality into the model, H2 remains largely unanswered.

Our modeling studies suggest that the eastern and western GOA are substantially different with respect to their contribution to important spawning and nursery area habitat. For the species that spawn on the shelf break (i.e. POP, SF, ATF), the eastern GOA appears to be much more likely to have spawning grounds that would produce successful recruits to populations in the GOA. Spawning areas for these species in the western GOA appear more likely to be providing recruits to populations (if any) downstream, perhaps to the southern side of the Aleutians or into the Bering Sea. Thus, it appears that the dominant east-west physical transport is key to driving these differences, and probably far more important than any biological differences in the food web in either region for these species. Conversely, the western GOA appears to be more important for WP and PC due to retention mechanisms. For PC, many fish that are spawned in the eastern GOA, especially those that spawn in deeper waters, were transported out of the GOA before being able to settle so were not recruited into the GOA population. For WP, historically the largest spawning concentrations occur in Shelikof Strait in the western GOA. However, secondary non-Shelikof aggregations in the western GOA have increased over time — potentially due to altered homing habits in response to changing environmental conditions (Ciannelli et al., 2007). Simulations have suggested the connectivity of GOA spawning regions to the Bering Sea through transport via Unimak Pass. However, the contribution of GOA spawning to Bering Sea recruitment is not yet understood.

9.2 *Integrated research*

The process of integration was challenging, but the GOA IERP program was intentionally designed to promote the integration of findings to better interpret results and enhance understanding of ecosystem dynamics. Following this experience of incorporating modeling into

1205 an IERP we present the following key factors that should be considered to maximize success
1206 when incorporating modeling into an IERP or other multidisciplinary research program:
1207

- 1208 Carefully consider the development status of the proposed models, including that
1209 additional development/tuning/validation time, will be needed within the IERP.
- 1210 Ensure the data needed for model development will be collected by the field program.
- 1211 Consider the time frame of the sampling program in relation to the model development to
1212 ensure the data are available in sufficient time for incorporation into the models.
- 1213 Schedule regular, guided meetings between model team members and field program
1214 participants to ensure maximum integration of these IERP components.
- 1215 Pre-determine how the 'success' of the models is to be judged.
1216

1217 Overall, the completion of the GOAIERP modeling should be seen in a management context
1218 as the start of a process (e.g. an IEA), rather than a single "delivery of results". Within
1219 management agencies such as NOAA, resources are being increasingly directed towards the
1220 operationalization of ecosystem-level models; working with these agencies, the GOAIERP
1221 modeling legacy may improve EBFM in the region. Following this experience, we highlight the
1222 following key factors for bringing models developed within the framework of an IERP into the
1223 management arena.
1224

- 1225 Close coordination with the local management agencies before and during the IERP.
1226 Determine who the ultimate end users will be and how model output and insights could be
1227 incorporated into the existing management strategy.
- 1228 Determination of key ecosystem drivers, and understanding of the models' abilities to
1229 capture variability in these drivers on short (annual) and longer (interannual->multi
1230 decadal) time scales.
- 1231 Successful incorporation of models into tactical and strategic decision-making requires
1232 support to keep the models maintained and updated.
1233

1234 More broadly, the development and application of numerical models to explore ecosystem
1235 dynamics in the context of an IERP can enhance overall program goals, focus research questions,
1236 and provide a broader context for the interpretation of observational data and process studies.
1237 Integrative numerical models can provide extrapolation to regional-scale implications for observed
1238 fine-scale processes. In turn, model results and sensitivity analyses can suggest biologically
1239 productive geographical areas or scientifically productive areas of inquiry, as well as sampling
1240 designs, for follow-up field or lab studies. Modeling, fieldwork, and lab studies thus need to be
1241 viewed as a set of interactive, iterative components within an IERP framework. Additionally, the
1242 IERP framework needs to provide the opportunity for interaction and the time for iteration among
1243 these components to achieve true integration.
1244

1245 *9.3 Informing management*

1246 An ongoing challenge to model development in the context of fishery management is to determine
1247 how to direct models and how to interpret results to better understand, hindcast, and predict stock
1248 status and trends. One challenge is determining the appropriate processes to consider and the
1249 assumptions involved. Another is distinguishing cause and effect versus correlation. Some of the
1250 most useful products from this effort may include the formal development of detailed species
1251 profiles, including information on early life history, known ecological trends, and hypothetical
1252 mechanisms related to transport and survival. These profiles should be maintained and updated
1253 with new information, as available. They might also inform future sampling efforts and survey
1254 designs (e.g. trawls, acoustics, larval sampling) as well as laboratory analyses (e.g. diet and
1255 energetics). IBMs have been successful in determining important habitats relevant to spawning

1256 and settlements. Results also identify key questions to explore further related to climatological
1257 effects, stock connectivity, and the potential for stock production and export beyond ecosystem
1258 boundaries (e.g. advection of GOA produced larvae to the Aleutian Islands and the eastern Bering
1259 Sea). Overall, the models provided insight to understand recruitment processes, important legacy
1260 products and will contribute to future GOA ecosystem studies.

1261 1262 *9.4 The art of modeling*

1263
1264 The GOAIERP modeling effort was useful, not only in informing understanding of recruitment
1265 in groundfish, and in integrating models into the program, but also as an exercise in model
1266 development. In any modeling approach, it is essential to determine not only what to include, but
1267 what not to include. Feasibility, simplicity, and parsimony are all important considerations. Several
1268 aspects will influence these decisions, including what data are available, what data are
1269 appropriate or accessible to the models, and the relative perceived importance of those data to
1270 model outputs. In this context, the elements critical to determining recruitment seem to include
1271 transport, prey fields, predation, and habitat. Fish are not passive particles and while the
1272 GOAIERP IBMs included stage-specific growth and vertical movement capabilities, future work
1273 might consider how the outputs presented would be influenced by additional behaviors, including
1274 more complex directed swimming (i.e. towards prey or geographical regions) and more complete
1275 bioenergetics for growth and natural mortality. Local-scale physical features (e.g. the extent to
1276 which canyons channel larvae and contrast with along-shelf flow) and episodic events (e.g. gap
1277 winds and associated increased cross-shelf flow facilitate cross-shelf transport) are also important
1278 to transport and deserve further examination. It would be worth considering how a physical model
1279 with finer spatial and temporal resolution (i.e. refined vertical water column complexity, temporally-
1280 resolved tidal currents and tidal stream transport, flow through canyons, and bottom currents)
1281 would influence IBM predictions. Finally, it is important to consider, on a theoretical level, how to
1282 best use snapshot-type data (e.g. oceanographic data at a single location and time) to inform,
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1748 **Tables**

1749

1750 Table 1. Summary of IBM details for each of the five focal species. Potential life stages include eggs (E),
 1751 yolk-sac-larvae (YSL), feeding larvae (FL), small and large preflexon feeding larvae (prF_{SM}, prF_{LG}),
 1752 postflexon feeding larvae (poF), Epipelagic juveniles (EPJ), juvenile settlement stage (J_{set}), juvenile benthic
 1753 stage (J_{ben}).
 1754

Species Trait	Pollock	Cod	ATF	POP	Sablefish
<i>Life Stages</i>	4:E, YSL, FL, J	6:E,YSL, prF, poF, EPJ, J _{ben}	8: E, YSL _{sm} ,YSL _{lg} , pre-FL _{sm} , pre-FL _{lg} post-FL, J _{set} , J _{ben}	5: FL _{pre} , FL _{post} , PJ, J _{set} , J _{ben}	5: E, YS, FL, EPJ, J _{set}
<i>Attaches to bottom</i>	No	Yes, E & J _{ben}	No	No	No
<i>Stage-specific growth</i>	Yes	Yes	Yes	Yes	YEs
<i>Temperature-dependent growth</i>	Yes-E, FL, stages	Yes-all	No	No	No
<i>Stage-specific depth preference</i>	Yes	Yes	Yes	Yes	Yes
<i>Terminal velocity/vertical position</i>	E, YSL	E, YSL	E, YSL	E, YSL	E, YSL
<i>Diel Migration</i>	FL,J	Yes- FL _{post} & EPJ	No	No	only J _{set}
<i>Vertical directed swimming</i>	Yes-J	Yes- YS, pre-FL _{pre} , FL _{post} , EPJ,	Yes	Yes	Yes
<i>Horizontal directed swimming</i>	Yes-J	No	No	No	No
<i>Consumption</i>	Yes-FL,J Based on historical data of prey density	No	No	No	No
<i>Predation</i>	Yes- on juveniles, based on ground fish predation data	No	No	No	No
<i>Competition</i>	No	No	No	No	No
<i>Unresolved Processes</i>	Juveniles directed swimming. variability in mortality – i.e. temporally-and spatially-explicit groundfish predation. Submesoscale processes	Potential juvenile movement. Consumption Mortality	Directional swimming towards preferred habitat; mortality consumption, mortality, temperature- dependent growth	Directional swimming with degree of natal homing in some life stage; mortality consumption, temperature- dependent growth, mortality	Directional swimming towards preferred habitat type consumption, mortality, temperature- dependent growth Transport mechanism to known and consistent nursery sites i.e. St. John Baptist Bay

1755

1756

1757 Table 2. Summary of research questions and the approach taken/proposed to answer the questions, and
 1758 which of the two hypotheses this helped address.
 1759

Research Question	Method	Hypothesis
Were the successful settlers concentrated into a narrow alongshore zone?	Connectivity analysis	H1
Where did the successful settlers come from?	Connectivity analysis	H1
Can connectivity between individual spawning and recruitment sites predict overall recruitment strength?	Connectivity analysis	H1
What were the paths followed by young fish that were spawned in different regions?	Trajectory analysis	H2
What were the differences in paths followed by the successful settlers by those that never settled, or by those that exited the GOA?	Trajectory analysis Path analysis	H2
What were the physical processes (e.g. eddies, currents) that took young fish towards or away from favored nursery areas?	Trajectory analysis	H2
What were the topographic features that steered them towards or away from these nursery sites?	Trajectory analysis	H2
What were the temperature histories of successful settlers vs. non-settlers?	Trajectory analysis	H2
Was there an optimum duration of the pelagic stage for successful settlers?	Trajectory analysis	H2
How direct or indirect were these trajectories?	Tortuosity analysis	H2
What was the prey field experienced by successful settlers vs. non-settlers?	Trajectory analysis	H2

1760
 1761

1762 Table 3. Summary of mechanisms important to groundfish recruitment, as identified by the IBMs. Notable
 1763 relationships (r or r^2) between model indices and estimates of recruitment from stock assessments are also
 1764 identified.
 1765

Connectivity Analyses		
IBM	Mechanisms Identified	Significant Correlations
WP	Connectivity showed that reaching the Shumagin West region as juveniles is more beneficial than having an early arrival as larvae to that nursery area, which is consistent with findings arising from examination of the tortuosity index	A positive correlation between recruitment anomalies and PC1 for larvae ($r=0.41$, $p=0.11$) An inverse relationship was found between recruitment anomalies and PC2 for larvae ($r=-0.58$, $p=0.02$)
PC	High larval retention in the Cook Inlet and along the Inner Shelf domain; the importance of Shumagin Islands as a nursery area	Settlement in Shumagin Islands vs MEI ($r=-0.54$, $p<0.05$), and vs. NPI ($r=0.53$, $p<0.05$)
POP	Connectivity was directed in a counterclockwise fashion, with typical dispersal distances ~100's km. More than 70% of individuals did not settle successfully. Individuals spawned in eastern GOA were much more likely to settle successfully. The probability of retention in the spawning zone was low.	Cross Sound to Prince William Sound (PWS) most strongly connected S-N pair ($r^2=0.62$, $p<0.05$)
SF	Sablefish settling in nursery areas in the GOA were most likely spawned in the eastern Gulf.	Settlement depth was the most influential parameter in determining Total Connectivity ($r^2=0.72$, $p<0.05$)
ATF	Similar to POP. More than 80% of individuals did not settle successfully.	
Trajectory Analyses		
IBM	Mechanisms Identified	Significant Correlations
WP	Efficient and direct trajectories bring a larger proportion of WP to nursery grounds Submesoscale eddies might play a role in concentrating or reducing dispersion offshore affecting the direct arrival to potential nursery areas. Climate indices correlated with stock assessment recruitment anomalies and WP biophysical indices (Principal components of connectivity)	The tortuosity index generated using the pollock IBM output showed an inverse relationship between the index and the stock assessment recruitment ($r=-0.42$, $p=0.11$). The number of anticyclonic and cyclonic submesoscale eddies are four times larger on the shelf than off-shelf in the western GOA, which is characterized by mesoscale eddies dispersing particles offshore MEI was inversely correlated with stock assessment recruitment anomalies ($r=-0.39$, $p=0.13$).

		PC1 larval connectivity positively correlates to PDO ($r=0.53$, $p<0.05$).and AO ($r=0.46$, $p=0.07$), while PC1 juvenile connectivity is directly correlates to PDO ($r=0.41$, $p=0.12$).
PC	Short mean trajectories for successful settlers. Settlers concentrated nearshore Unsuccessful settlers in PWS (too deep) Many PC exit the GOA to the southwest	
POP	Considerable transport off the shelf; Potential for export from the GOA to the AI and/or the EBS for individuals spawned in the western GOA.	Nursery areas most frequently reached by those successful individuals were in the central GOA
SF	Early on-shelf transport in the eastern GOA.	Southerly wind in eastern GOA (Jan-March) $r=0.5-0.7$, $p<0.05$
ATF	Considerable transport off the shelf; potential for export from GOA to AI and/or EBS for individuals spawned in western GOA.	Most effective nursery areas were in the central and western GOA
Predicting Recruitment		
IBM	Mechanisms Identified	Significant Correlations
WP	Pollock recruitment appears to be positively affected by physical conditions that result in efficient transport to the nursery areas.	Inverse relationship between the tortuosity index and the stock assessment recruitment ($r=-0.42$, $p=0.11$).
PC	Lower temperatures and curved paths improve recruitment.	Degree days ($r=-0.58$, $p<0.05$) Tortuosity index ($r=0.55$, $p<0.05$)
POP		Annual total fraction of successful individuals from natal zones 2 Sitka and 6 PWS ($r^2=0.62$, $p<0.05$)
SF	-Cross shelf transport -Primary production -Strength of S-SE wind. Total connectivity between all spawning sites and nursery areas had a stronger correlation with recruitment than the strength of connections to or from a specific region.	Total Connectivity $r^2=0.2$, $p=0.08$ Total Connectivity + Annual cross shelf flow + Primary Production $r^2=0.59$, $p=0.01$
ATF	No significant relationships identified between connectivity indices and recruitment estimates.	None

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1767

1768 **Figure Captions**

1769

1770 **Fig. 1.** Key features and location in the Gulf of Alaska study area. For orientation, the location of the detailed
1771 overview within the Gulf of Alaska is outlined (black box) in the top right inset. The full extent of the model
1772 domain (white box) is shown in the top left inset.

1773

1774 **Fig. 2.** The GOA IERP was a vertically-integrated hierarchical study of the physics, fisheries, and ecosystem
1775 of the GOA, partitioned among: the upper trophic level (UTL), the middle trophic level (MTL), the lower
1776 trophic level and physics (LTL), and modeling. Schematic illustrates the coupling of output from various
1777 models, coordination between process studies and modeling, choice and criteria for evaluating model
1778 outputs, interaction with human-induced impacts, and linkages between the scientific question and
1779 management needs. H1 and H2 refer to the two modeling hypotheses. The thick blue arrows indicate the
1780 flow of model output/input between the models used in the GOA IERP. The thin black arrows indicate the
1781 flow of information/data from the LTL, MTL, and UTL components into the models. The thicker black arrows
1782 indicate where model output informed H1 (dash) and H2 (solid).

1783

1784 **Fig. 3.** Conceptual models for IBMs of select Gulf of Alaska groundfish a) Walleye Pollock (WP), b) Pacific
1785 Cod (PC), c) Pacific Ocean Perch (POP), d) Sablefish (SF), and e) Arrowtooth Flounder (ATF). Life stages
1786 included in each IBM vary but generally include: egg, yolk sac larva, small feeding (SFPr) pre-flexion larva,
1787 large feeding pre-flexion (LFPr) larva, postflexion larva, pelagic juvenile, settlement-stage juvenile, and
1788 benthic juvenile (see: Gibson et al., 2019; Hinckley et al., 2019; Stockhausen et al., 2019a, 2019b. Larval
1789 drawings based on Matarese et al., 1989; illustrations by Beverly M. Vinter).

1790

1791 **Fig. 4.** Schematic illustrating the structure and direction of material flow in the GOA nutrient phytoplankton
1792 zooplankton model embedded in the ROMS physical oceanography model. Arrows indicate the direction of
1793 material flow.

1794

1795 **Fig. 5.** Connectivity matrices illustrate the annual median probability that individuals spawned as eggs in
1796 each alongshore zone (1-12) in the model domain (a) successfully settled in each alongshore nursery area
1797 for b) SF, c) ATF, and d) PC. Medians were computed across the 16-year simulation (1996-2011). For
1798 simplicity, the connectivity matrix shown for PC is for individuals spawned and settled within the < 70 m
1799 depth zone. PC individuals were also released and settled within additional depth zones, but the overall
1800 pattern of connectivity was similar. The connectivity to the GOA basin and areas outside of the model grid
1801 are not shown.

1802

1803 **Fig. 6.** Trajectories of Walleye Pollock spawned in Cook Inlet (ai), Kodiak Island (aii), and Shelikof Strait
1804 (aiii) between March to September. The trajectories are color-coded with simulated temperature values.
1805 Comparison of Sablefish path analysis for individuals released in spawning area 2 (Fig. 5a, eastern GOA)
1806 during a high recruitment year (2000) and a low recruitment year as identified by recruitment estimates from
1807 stock assessment. See text for description particle day count computation.

1808

1809 **Fig. 7.** Tortuosity index (Ti) for Walleye Pollock. Trajectory patterns show longer trajectories of individuals
1810 than the straight-line distance between starting to ending points for March 2002 (a), compared to more
1811 direct trajectories closer to 1 in 2011 (c). Time series of the tortuosity index including seasonal variability
1812 (b) and integrated the seasonal variability (d).

1813

1814 **Fig. 8.** Average temperature (a) and salinity (b) experienced by Sablefish individuals 'spawned' in area 2
1815 in the eastern Gulf (green), and in area 6 in the central Gulf (red). See Fig. 5a for the location of spawning
1816 regions.

1817

1818 **Fig. 9.** Sample trajectories of "winners" (a) and "losers" (b) for the EOF analyses in Pacific cod. 'Winners'
1819 are individuals that successfully settled, while 'losers' did not settle or were not retained within the central
1820 GOA model grid. The release location of individuals whose trajectories are shown is indicated by the white
1821 star. Individuals were released on Feb. 15th, model year 2008.

1822

1823 **Fig. 10.** Summary of biophysical modes detected by multivariate EOF analysis of normalized (z-scored) life
1824 history variables, comparing “winners” with “losers”. Top row: times series amplitude of modes 1 (a) and 2
1825 (b). Middle row: average over all “winning” individuals for each of the life history variables in modes 1 (c)
1826 and 2 (d). Bottom row: average over all “losing” individuals for each life history variable in modes 1 (e) and
1827 2 (f). Life history variables analyzed are as follows: larval stage (stage), larval diameter (diam), ambient
1828 euphausiids (eup), neocalanus (nca), copepods (cop), vertical mixing intensity (vmix), salinity (salt),
1829 temperature (temp), bathymetry (bath), depth (dep), latitude (lat) and longitude (lon).

1830
1831 **Fig. 11.** The Gulf of Alaska food web, as described by the Ecosim food web model (Gaichas et al., 2010;
1832 2012). GOAIERP focal species and NPZ components are shown in black; functional groups directly
1833 connected to a GOIERP group are shown in gray.

1834
1835 **Fig. 12.** Results (percent change) in species within the Ecosim food web model, resulting from a 10%
1836 (equilibrium) increase in the juvenile mortality of the five GOAIERP focal species. Bars display 50%, and
1837 lines represent 95% range of variation in Monte Carlo results based on uncertainty in data inputs.

1838
1839 **Fig. 13.** Important spawning (a) and nursery (b) sites for groundfish in the GOA, identified using IBMs.

1840
1841 **Fig. 14.** Simplified representation of the Integrated Ecosystem Assessment (IEA) process, as defined for
1842 use in NOAA ecosystem-based management efforts (Levin et al., 2009). Modified schematic of the NOAA
1843 IEA approach [<https://www.integratedecosystemassessment.noaa.gov/>]

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