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8 ABSTRACT

9 Geographic range size can span orders of magnitude for plant and animal species, with the study 10 of why range sizes vary having preoccupied biogeographers for decades. In contrast, there have 11 been few comparable studies of how range size varies across microbial taxa and what traits may 12 be associated with this variation. We determined the range sizes of 74,134 bacterial and archaeal 13 taxa found in settled dust collected from 1,065 locations across the United States. We found that 14 most microorganisms have small ranges and few have large ranges, a pattern similar to the range

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size distributions commonly observed for macrobes. However, contrary to expectations, those microbial taxa which were locally abundant did not necessarily have larger range sizes. The observed differences in microbial range sizes were generally predictable from taxonomic identity, phenotypic traits, genomic attributes, and habitat preferences, findings that provide

- 19 insight into the factors shaping patterns of microbial biogeography.
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Key Words: Geographic range size, biogeography, dispersal, microbiology, microbial dispersal,
dust-associated microbes

31 INTRODUCTION

32 Not all microbes are everywhere all the time. Due to both dispersal constraints and habitat

33 filtering, we know that many microbial taxa are restricted in their geographic and ecological

- 34 distributions (Martiny *et al.* 2006; Hanson *et al.* 2012). Microbial endemism has been
- demonstrated across a range of habitats including geothermal hot springs (Papke *et al.* 2003;

36 Whitaker *et al.* 2003), benthic ecosystems (Ruff *et al.* 2015), soil (Cho and Tiedje 2000; Vos and

37 Velicer 2008; Andam *et al.* 2016), and marine waters (Boucher *et al.* 2011; Ghiglione *et al.*

38 2012; Sul *et al.* 2013). Perhaps the best evidence for restricted microbial distributions comes

39 from decades of work on pathogens. Many pathogens of humans, domestic animals, and crops

40 are restricted to certain geographic areas and regions with specific environmental conditions

41 (Achtman 2008; Bebber *et al.* 2014; Just *et al.* 2014; Murray *et al.* 2015).

42

43 Like plants and animals, many microorganisms clearly have ranges – the geographic area where

44 a given taxon is found – and range sizes are likely to vary across bacterial and archaeal taxa. The

45 study of range size, and the factors that drive differences in range size and shape, have been

46 studied for more than a century by ecologists, biogeographers, and conservation biologists.

47 However, there is surprisingly little explicit documentation of microbial geographic range size,

48 taxonomic variation in range size, or the traits that might contribute to this variation.

49

50 Much of the variation in the geographic range size of plant and animal species is often predictable. For example, occupancy-abundance relationships are generally positive, and 51 52 organisms that are locally abundant also often have large geographic ranges (Gaston et al. 2000; 53 Holt et al. 2002; Roney et al. 2015), with causality likely flowing in both directions. Certain life 54 history strategies also vary predictably with range size. For example, species with greater 55 dispersal capabilities tend to have larger geographic ranges due to their ability to populate new 56 regions and to maintain gene flow among regions, as is the case for certain insects (McCauley et 57 al. 2014), birds (Laube et al. 2013), plants (Paul et al. 2009), and marine taxa (Macpherson 58 2003; Lester and Ruttenberg 2005; Lester et al. 2007). In addition, taxa able to live in many 59 habitat types, whether because they are generalists or have a high degree of phenotypic plasticity, 60 also tend to have larger geographic ranges (Pohlman et al. 2005; Pichancourt and van Klinken 61 2012; Morueta-Holme et al. 2013; Ofstad et al. 2016). Finally, closely related taxa often have similar geographic range sizes due to shared ecological attributes, as shown for species of birds 62 63 (Mouillot and Gaston 2009; Herrera-Alsina and Villegas-Patraca 2014).

64

65 With their small cell size, massive population numbers, and diverse physiologies, microbial taxa 66 have the potential for widespread dispersal and colonization, and consequently, large range sizes. 67 Evidence suggests there are unifying theories of biodiversity and biogeography across all domains of life (Green and Bohannan 2006; Locey and Lennon 2016). Thus, we predict that 68 69 many of the factors driving range size in plants and animals also influence microbial range size. For instance, we would expect that locally abundant microbial taxa would tend to have larger 70 71 geographic ranges than rare species in concordance with previous work which has demonstrated 72 a positive occupancy-abundance relationship for some microbial taxa living in specific 73 environments (Nemergut et al. 2011; Ruff et al. 2015). We would also expect that closely related 74 taxa should have more similar range size distributions due to a greater likelihood of sharing traits 75 that govern capacity for dispersal and colonization. As for macrobes, taxa that disperse well and 76 are able to tolerate a wide range of environmental conditions should have larger range sizes. We

77 predict that the relevant traits governing microbial dispersal may include dormancy or other 78 strategies related to stress (e.g. UV radiation, desiccation, extreme temperature) tolerance. For 79 example, endospore formation facilitates the dispersal of microbes through both time and hostile 80 conditions. This trait allows *Bacillus* species to travel across continents in the upper atmosphere 81 (Roberts and Cohan 1995) and thermophilic marine *Firmicutes* to persist in cold sediments 82 (Müller et al. 2014). Similarly, we predict that traits associated with the ability to colonize and grow in diverse environments may include genomic characteristics related to phenotypic 83 plasticity and habitat breadth. From previous work showing that genome size correlates with the 84 ability of soil bacteria to persist in a broad range of habitats (Barberán et al. 2014a; Cobo-Simón 85 86 and Tamames 2017), we would expect genome size to positively correlate with range size. 87 Hence, the suite of phenotypic traits and genomic attributes that influence the ecological 88 distribution of microbial taxa also likely influence range size.

89

90 To build a more comprehensive understanding of how and why microbial range sizes may vary, 91 we determined the range sizes and shapes of 74,134 bacterial and archaeal taxa found in settled 92 dust collected from outdoor building surfaces from 1,065 homes across the United States. We 93 focus on settled dust because it is found everywhere and easy to sample consistently. Likewise, 94 we know that those microbes found in settled dust were at one point airborne, allowing us to 95 identify organisms that can be dispersed through the atmosphere. Also, the settled dust found on 96 outdoor building surfaces is nutrient-limited and is unlikely to represent the original 97 environmental source of the taxa found therein. In other words, by examining the range sizes of 98 those microbes found in settled dust, we can more readily assess differences across taxa in their dispersal capabilities as opposed to differences related to colonization and establishment in a 99 more suitable environment for growth. 100

101

We calculated range size using both the area of occupancy (AOO) and the extent of occurrence (EOO) approximations, both of which are commonly used in macroecology (Gaston and Fuller 2009). Simplified, AOO is akin to a 'dot map' of observations across a grid overlay that are summed together, while EOO is comparable to 'connecting the dots' and calculating the area of the resulting shape. We determined if the distribution of range sizes for these microbial taxa is similar to plant and animal species, and to what extent the occupancy-abundance relationship

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explains variation in range size. Furthermore, we determined if taxonomy could predict differences in range size distributions. Finally, we mined the extensive wealth of information available in curated microbial databases to determine if phenotypic traits, genomic attributes, or habitat preferences can explain the measured variability in range size. This study represents one of the first comprehensive efforts to understand the variation in range size across a broad range of microbial taxa, whether microbes follow the same biogeographical patterns commonly observed for 'macrobes', and why some microbial taxa have larger range sizes than others.

115

116 METHODS

117 Sample collection and molecular analysis

118 Details of sample collection and molecular analysis have been described previously (Barberán et

al. 2015). Briefly, outdoor dust samples were collected from the upper trim on the outside

120 surface of an exterior door by participants of the Wild Life of Our Homes

121 (<u>http://robdunnlab.com/projects/wild-life-of-our-homes/</u>) citizen science project. Bacterial and

archaeal diversity was determined by sequencing the V4 hypervariable region of the 16S rRNA

123 gene with primers 515-F (GTGCCAGCMGCCGCGGTAA) and 806-R (GGA-

124 CTACHVGGGTWTCTAAT) (Fierer *et al.* 2012) using the direct PCR approach previously

described (Flores et al. 2012). Sequencing was done on the Illumina HiSeq or MiSeq platforms

126 with all reads trimmed to 100 bp. All reads were quality filtered (maximum e-value of 0.5),

dereplicated, and clustered into phylotypes at a 97% similarity threshold with the UPARSE

128 pipeline (Edgar 2013). Taxonomic identity was determined using the Ribosomal Database

129 Project classifier (Wang *et al.* 2007) trained on the Greengenes 13_8 16S rRNA database

130 (McDonald *et al.* 2012). All sequence data are accessible through the FigShare repository (

131 <u>https://doi.org/10.6084/m9.figshare.1270900.v8)</u>.

132

133 Eukaryotic sequences were removed, and those phylotypes present in >25% of negative control

134 samples (including phylotypes classified as Mycoplasma, Pseudomonas, Serratia, and

135 *Acinetobacter*) were also filtered prior to downstream analyses as they likely represent taxa

136 originating from reagent or amplification contamination (Salter *et al.* 2014). To minimize

amplicon sequencing biases between samples, low coverage samples (i.e. samples with <10,000

138 reads after quality filtering) were removed, and total sequence counts were normalized using a

139 cumulative-sum scaling approach (Paulson *et al.* 2013). We restricted our analyses to the

140 contiguous United States. and hence removed samples originating from Hawaii and Alaska.

141 Finally, we excluded rare phylotypes (i.e. phylotypes present in fewer than five samples, as at

142 least five observations are required to calculate range size using the minimum convex polygon

approach described below). In total, 74,134 phylotypes across 1,065 samples were included in all

- 144 downstream analyses.
- 145

146 Range size and shape calculations

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147 Latitude and longitude coordinates were inferred from sample locations (i.e. reported addresses), 148 and these coordinates were transformed into the Lambert conformal conic projection (LCC) for 149 all spatial analyses. Many plant species are dispersed by wind (Howe and Smallwood 1992; Clark et al. 2002), so given the potential similarities between plants and dust-associated 150 151 microbes in their dispersal dynamics, we used approaches to calculate range sizes commonly 152 employed by plant biogeographers. Range size was determined using both area of occupancy (AOO) (see Kolb et al. 2006; Kreft et al. 2006; Essl et al. 2009) and extent of occurrence (EOO) 153 (see Sérgio et al. 2007; Brummit et al. 2015). To determine AOO, we overlaid a 100 x 100 km² 154 grid that encompassed all sample locations and used the R package sp (Pebesma and Bivand 155 2005) to count the total number of grid cells in which each phylotype was observed. AOO range 156 157 size (km^2) was calculated by summing the area of total occupied grid cells. To determine EOO, 158 we used the R package adehabitatHR (Calenge 2006) to find the minimum convex polygon (MCP) after excluding 5% of the extreme points. EOO range size (km²) was calculated from the 159 160 area of the MCP circumscribing all observations for each phylotype. Range shape was 161 determined by calculating the maximum longitudinal and latitudinal dimensions of occurrence 162 for each phylotype. To control for biases introduced by uneven sampling intensity, we divided 163 the U.S. into six regions, sub-sampled 70 locations from each of these regions, and repeated the 164 range dimension analyses.

165

166 *Taxonomic signal and phenotypic, genomic, and habitat trait-based analyses*

167 Next, we assessed potential taxonomic determinants of range size. Phylotype range size was

168 ranked by Phylum, and phyla with fewer than 25 representative phylotypes were excluded. For

169 the most abundant phyla (i.e. Proteobacteria, Actinobacteria, Firmicutes, and Bacteroidetes),

phylotype range size was also ranked by Family, and families with fewer than 25 representativephylotypes were excluded.

172

173 Finally, we determined if differences in phenotypic traits, genomic attributes, or habitat 174 preferences could further explain variation in range size. We inferred putative traits of dust 175 phylotypes by matching their 16S rRNA gene sequences to those of reference strains from 176 curated, publicly available databases. Representative partial 16S rRNA gene sequences from each phylotype were matched against full length 16S rRNA gene sequences from the IJSEM 177 178 phenotypic database (Barberán et al. 2017) and from the Integrated Microbial Genomes (IMG) 179 database (Markowitz et al. 2014). Matches were determined using BLASTn (Altschul et al. 180 1990) at \geq 99% identity and \geq 95% coverage. We restricted these analyses to the top four most abundant phyla, which included Proteobacteria, Actinobacteria, Firmicutes, and Bacteroidetes. 181 182 We recognize that partial 16S rRNA gene sequences may not provide a level of resolution 183 sufficient for accurately identifying the phenotypic and genomic traits of all taxa. However, the 184 selected traits typically show a strong phylogenetic signal and are generally conserved across 185 broader taxa and lineages (Barberán et al. 2017).

186

187 We were able to match a total of 1,461 16S rRNA gene sequences of dust phylotypes (including 188 518 Proteobacteria, 428 Actinobacteria, 293 Firmicutes, and 222 Bacteroidetes) to 2,487 unique 189 full length 16S rRNA gene sequences in the IJSEM phenotypic database (Barberán et al. 2017). 190 We assessed how the AOO varied in relation to the following phenotypic traits: oxygen 191 tolerance, sporulation, pigmentation, Gram stain reaction, and source habitat. Here, source 192 habitat refers to the reported isolation source of a given strain from the IJSEM phenotypic 193 database (Barberán et al. 2017). We selected these traits because we expected that these traits 194 may influence dispersal and colonization capabilities. For discrete traits, we excluded phylotypes 195 with matches to multiple strains that had conflicting trait values.

196

197 We matched a total of 1,186 16S rRNA gene sequences of dust phylotypes (including 415

198 Proteobacteria, 276 Actinobacteria, 325 Firmicutes, and 170 Bacteroidetes) to 6,321 unique full

199 length 16S rRNA gene sequences in the IMG database (Markowitz *et al.* 2014). We assessed

200 how AOO varied with the following genomic attributes: G+C content, genome size, and 16S

rRNA operon copy number. We hypothesized that these attributes may influence phenotypic
plasticity and habitat preferences. G+C content and genome size are highly correlated in bacteria,
and large genomes are thought to confer broad niche breadth (Bently and Parkhill 2004).
Multiple copies of the 16S rRNA operon are common in microbial genomes and are reflective of
copiotrophic or oligotrophic life history strategies (Klappenbach *et al.* 2000), with those taxa
capable of higher maximum growth rates generally having a larger number of rRNA operons.

- For these genomic attributes, we determined mean values for phylotypes with matches tomultiple strains.
- 209

210 **RESULTS**

211 Microbial diversity and community composition

A total of 74,134 16S rRNA gene sequence phylotypes were observed across the 1,065 dust 212 213 samples (Fig. 1a), with each sample harboring 4,850 phylotypes on average. A total of 50 bacterial and archaeal phyla were recovered, and the dominant phyla were Proteobacteria, 214 215 Actinobacteria, Firmicutes, and Bacteroidetes with >75% of phylotypes assigned to these four 216 phyla (Appendix S1: Fig S1). On average a given phylotype was observed in 70 and 27 different 217 samples (mean and median, respectively). Nearly 24% of phylotypes were found in ≤ 10 samples, 218 and only 35 phylotypes were observed in $\geq 90\%$ of the samples (Fig. 1b). Community 219 composition was highly variable across the samples, with geographic distance as well as 220 environmental factors including soil pH, precipitation, primary productivity, and temperature 221 being the best predictors of overall differences in community composition (see Barberán et al. 2015 for details). 222

223

224 Range size and shape

The frequency distribution in range size as measured using the area of occupancy (AOO) was highly right-skewed and best described using a log-normal distribution; many taxa have small ranges and few have very large ranges (Fig. 1c). Across all phylotypes, the mean and median estimated AOO range sizes were 3,984 and 2,200 km², respectively. Alternatively, using the extent of occurrence (EOO), we found that the frequency distribution of EOO range size is best described as irregular and bimodal (Fig. 1d). Across all phylotypes, the estimated mean and median EOO range sizes were 4.2 and 4.3 million km², respectively, which is approximately half of the area of the contiguous U.S. We observed a strong positive correlation (Spearman's $\rho =$

233 0.89, P < 0.0001) between log AOO and EOO estimations of range size (Appendix S1: Fig S2).

Given the high potential for cell dispersal and the wide distribution of suitable habitats, we

expect that EOO is likely overestimating microbial range sizes here. Therefore, we used the more

- conservative AOO estimation for downstream analyses focused on determining what potential
- 237 factors might explain range size variation.

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239 We described range shape by calculating the maximum geographic spread in both longitudinal 240 and latitudinal dimensions. The frequency distribution of the longitudinal range is highly left-241 skewed; most phylotypes were found on both the eastern and western coasts and have a mean 242 and median east-west span of 3,869 km and 4,183 km, respectively (Appendix S1: Fig S3a). 243 While also left-skewed, there was a greater variation in the latitudinal range with a mean and 244 median north-south span of 1,855 km and 1,920 km, respectively (Appendix S1: Fig S3b). While 245 phylotypes with greater longitudinal spread also tend to have greater latitudinal spread 246 (Spearman's $\rho = 0.67$, P < 0.0001), range dimensions for most phylotypes (89.9%) are elongated 247 east-west as opposed to north-south (Fig. 2), and this pattern persists after normalization for the 248 irregular shape of the sampling region (i.e. the U.S. is larger east-west than north-south) and after 249 correcting for differences in sampling intensities across different regions (Appendix S1: Fig S4). 250 The bacterial and archaeal phylotypes are far more likely to have larger east-west distributions 251 than north-south distributions.

252

One of the most widely observed correlates of range size is local density or abundance, and species that are more abundant tend to have larger geographic ranges than rare species (Gaston 1996a; Holt *et al.* 2002). Interestingly, we find little support for this relationship for dustassociated microbes. Instead, we found only a weak correlation between the local relative abundance of a phylotype and its range size (Spearman's $\rho = 0.14$, P < 0.0001) (Appendix S1: Fig S5).

259

260 *Taxonomic differences in range size*

261 Given this broad distribution of range sizes across dust phylotypes, we next sought to determine

what additional factors could further explain this variation. To begin, we asked if range size

- 263 differed across taxonomic groups. We found that geographic range size has a strong taxonomic
- signal and varies significantly across phyla (one-way ANOVA; $F_{28, 70297} = 136.2$, P < 0.0001)
- 265 (Fig. 3). For example, phylotypes within the phylum Actinobacteria tend to have range sizes that
- are approximately 13% larger than the range sizes of phylotypes within the phylum
- 267 *Acidobacteria* (Tukey's test; P = 0.00022). Within the Archaea, range sizes of *Crenarchaeota* are
- approximately 69% larger than those of *Euryarachaeota* (Tukey's test; P < 0.0001).
- 269
- 270 At greater taxonomic resolution, the taxonomic signal for range size within the top four most
- abundant phyla was even more pronounced (Appendix S1: Fig S6). We observed significant
- differences across families of *Proteobacteria* (one-way ANOVA; $F_{45, 15833} = 84.9$, P < 0.0001)
- 273 (Appendix S1: Fig S6a), Actinobacteria (one-way ANOVA; $F_{40,8005} = 21.8$) (Appendix S1: Fig
- 274 S6b), *Firmicutes* (one-way ANOVA; $F_{21,7159} = 66.6$, P < 0.0001) (Appendix S1: Fig S6c) and
- 275 *Bacteroidetes* (one-way ANOVA; $F_{15, 5359} = 70.1$, P < 0.0001) (Appendix S1: Fig S6d). For
- 276 example, within the *Proteobacteria*, *Burkholderiaceae* share a similar range size with
- 277 *Rhizobiaceae* (Tukey's test; P = 0.98) and *Bradyrhizobiaceae* (Tukey's test; P = 0.052), and all
- three of these families have larger range sizes than *Neisseriaceae* or *Legionellaceae* (Tukey's
- 279 test; P < 0.0001) (Appendix S1: Fig S6a).
- 280

281 Phenotypic and genomic traits that vary with range size

282 Finally, we asked if certain phenotypic or genomic traits could predict variation in range size. We found that range size varies with oxygen tolerance (two-way ANOVA; $F_{3, 2324} = 67.5$, P < 283 284 0.0001); aerobes have geographic ranges approximately 63% larger than anaerobes (Tukey's test; P < 0.0001) (Fig. 4a). However, the strength of the relationship between oxygen tolerance 285 and range size differs between phyla (two-way ANOVA; $F_{9, 2324} = 2.6$, P = 0.0062) (Appendix 286 287 S1: Fig S7a). Unexpectedly, range sizes were approximately 19% smaller for those phylotypes 288 inferred to be capable of spore formation, even after restricting the analysis to obligate aerobes to 289 minimize potential biases incurred by many anaerobes being spore-formers (two-way ANOVA; $F_{1,1643} = 36.0, P < 0.0001$) (Fig. 4b), although the strength of this relationship differed between 290 phyla capable of spore formation (two-way ANOVA; $F_{1, 1643} = 5.0$, P = 0.025) (Appendix S1: 291 292 Fig S7b). Taxa that are pigmented tended to have ranges that are approximately 39% larger than 293 taxa that are not pigmented (two-way ANOVA; $F_{1,1802} = 55.4$, P < 0.0001) (Fig. 4c), and this

- pattern was independent of phylum identity (two-way ANOVA; $F_{3, 1802} = 1.5$, P = 0.22)
- 295 (Appendix S1: Fig S7c). Range size also varied with Gram stain; taxa with Gram stain positive
- cell walls have approximately 17% larger ranges than taxa with Gram stain negative cell walls
- 297 (two-way ANOVA; $F_{2,31207} = 32.4$, P < 0.0001) (Fig. 4d). Finally, range size varied with source
- habitat (two-way ANOVA; $F_{7,907} = 11.1$, P < 0.0001); taxa derived from soil and plants were
- 299 more likely to have larger ranges compared to taxa associated with aquatic environments such as
- seawater or marine sediments (Tukey's test; P < 0.005) (Fig. 4e).
- 301

302 With regards to genomic attributes, we found that range size was positively correlated with G+C

- 303 content (Pearson's r = 0.45, P < 0.0001) (Fig. 5a), but this relationship was largely driven by
- 304 *Proteobacteria* (Pearson's r = 0.39, P < 0.0001) and *Actinobacteria* (Pearson's r = 0.32, P <
- 305 0.0001) (Appendix S1: Fig S8a). Range size and genome size were also positively correlated
- 306 (Pearson's r = 0.22, P < 0.0001) (Fig. 5b), and this relationship was significant when we ran the
- analyses for each phylum individually (Appendix S1: Fig S8b). Finally, range size was
- negatively correlated with 16S rRNA operon copy number (Pearson's r = -0.28, P < 0.0001)
- 309 (Fig. 5c), but the direction and significance of this relationship varied when these analyses were
- 310 conducted within individual phyla (Appendix S1: Fig S8c).
- 311

312 **DISCUSSION**

- Geographic range size is a cornerstone of biogeography, and studies of how range sizes vary
 across taxa have contributed to the development of key paradigms in conservation biology,
 evolutionary biology, and ecology. Despite decades of studies investigating range size and range
 size determinants in plants and animals, comparable studies are rarely conducted with microbial
 taxa. We addressed this knowledge gap by investigating the range sizes and the potential factors
 associated with range size variation across a broad breadth of bacterial and archaeal taxa
 (Appendix S1: Fig S1) identified in dust samples collected from across the United States (Fig.
- **320** 1a).
- 321
- 322 The accurate evaluation of microbial range size distributions is challenging, and many of these
- 323 challenges also apply to the accurate estimation of plant or animal range sizes. First, most
- 324 microbial communities are highly diverse. Thus, adequate sampling depth is important, and it

325 remains challenging to determine with confidence whether a given taxon is truly absent in a 326 community or simply below the level of detection. High-throughput culture-independent 327 sequencing approaches, like the approach used here in which we identified microbial taxa in 328 samples by analyzing a mean of 59,831 16S rRNA gene sequences per sample, can help to 329 reduce the magnitude of this problem (Sogin et al. 2006; Lynch and Neufeld 2015). Even so, we 330 are undoubtedly underestimating the full extent of microbial diversity in individual samples. 331 Importantly, this problem of insufficient sampling depth, which limits our ability to confirm 332 which taxa are 'truly absent' in a given sample versus those taxa that were simply not detected, also plagues plant and animal surveys (MacKenzie et al. 2002; Cunningham and Lindenmayer 333 334 2005). Second, accurate estimations of range size are best achieved through extensive population 335 surveys across a broader geographic region of interest. While sampling efforts are inevitably 336 constrained by logistics, more is always better, and we were able to collect samples from 1,065 337 locations across the contiguous U.S. (Fig. 1a). Third, range sizes will undoubtedly vary as a 338 function of taxonomic resolution – the range sizes of sub-populations will likely be smaller than 339 range sizes of the broader species or genus. Most studies of plant and animal range size focus on 340 species or intra-species level resolutions. While the species definitions for plants and animals are 341 often arbitrary and somewhat inconsistent, microbiologists continue to intensely debate the 'microbial species concept' and even the mere existence of species (Roselló-Mora and Amann 342 343 2001; Gevers et al. 2005; Achtman and Wager 2008; Doolittle 2012). To remedy this, microbial 344 ecologists often define units of diversity, or phylotypes, based on similarity in marker gene 345 sequences. Such an approach was used here as we defined phylotypes as those taxa which shared 346 \geq 97% similarity in their 16S rRNA gene sequences, a threshold that roughly corresponds to a 347 bacterial 'species' (Stackebrandt and Goebel 1994; Kim et al. 2014). In short, the challenges 348 associated with estimating microbial range sizes are not unique to microbial ecology, and we 349 argue that robust investigation of microbial range size is possible with the sampling effort and 350 methodologies used here.

351

The AOO range size frequency distribution for dust phylotypes was highly right-skewed (Fig. 1c); many microbial taxa have small geographic ranges and fewer have large ones. This distribution of geographic range sizes, described as a 'hollow curve' that is approximately log-

normally distributed, is widely observed for many plant and animal species (Gaston 1996b;

356 Berry and Riina 2005; Orme et al. 2006; Agosta et al. 2013). In addition to range size, the shape 357 of a species' range is also commonly studied by plant and animal biogeographers (Brown et al. 358 1996). For instance, range shape can be used to identify the environmental variables that 359 determine patterns of range expansion (Pigot et al. 2010). Here, we described the range shapes of 360 these microbial taxa by measuring the maximum east-west and north-south spread of each 361 phylotype. We found that the north-south spread of taxa was more constrained that the east-west 362 spread (Fig. 2, Appendix S1: Fig S3). To put simply, many taxa are found on both eastern and 363 western coasts, but fewer are distributed across the southern and northern boundaries of the U.S. 364 This results in an east-west elongated range for a majority of dust phylotypes (Fig. 2, Appendix 365 S1: Fig S4), a pattern that is consistent with the east-west range elongation that is observed for 366 many North American plant and animal species (Brown et al. 1996; Rosenfield 2002; Schlachter 367 2010). This pattern may be a product of dispersal driven by the prevailing winds, which 368 predominately blow across North America from the west to the east. The migration of microbes 369 through the atmosphere has been previously linked to wind patterns and weather dynamics 370 (Yamagucki et al. 2012; Smith et al. 2013; Barberán et al. 2014b; Weil et al. 2017). This pattern 371 also suggests that there are latitudinal limits to dispersal, which could be the result of climatic 372 temperature constraints or historical biogeographical processes (Mittelbach et al. 2007). 373 Latitudinal constraints to dispersal are well documented across diverse plant and animal species 374 (Wiens et al. 2006; Svenning and Skov 2007; Salisbury et al. 2012), and more recently such 375 constraints have been documented in terrestrial soil bacteria (Andam et al. 2016; Choudoir et al. 376 2016). We think that future work integrating information on weather systems and other climate 377 variables to address mechanisms of microbial migration will be particularly insightful.

378

379 The frequency distribution in geographic range sizes and the spatial dimensions of range shape 380 for these dust-associated microbes are qualitatively similar to what is commonly observed for 381 plants and animals. In contrast, we find little support for the occupancy-abundance relationship 382 for dust-associated bacteria (Appendix S1: Fig S5). This finding goes against expectations as the 383 occupancy-abundance relationship has been widely observed for plants and animals (Gaston et 384 al. 2000). Although this relationship may somewhat be inflated by the challenges associated with 385 sampling rare taxa (Wenger and Freeman 2008; Sileshi et al. 2009), most bacterial phylotypes, 386 regardless of their local abundance, had small ranges, while phylotypes with high local

abundance were nearly as likely to have large ranges as rare taxa (Appendix S1: Fig S5). Thus,
abundance alone is not a useful predictor of microbial range sizes, and instead we expected that
much of the observed variation in microbial range size is likely due to evolutionary or ecological
traits affecting dispersal or habitat preferences.

391

392 Range size distributions varied across taxonomy, and mean range size differed significantly 393 between phyla (Fig. 3). Importantly, this relationship was not just driven by the most abundant 394 phyla. For example, range sizes for the *Crenarchaeota* and candidate phylum FBP are amongst 395 the largest in the dataset, yet these phyla are not ranked among the top ten most abundant phyla 396 (Appendix S1: Fig S1). We also see intra-group differences in range size distributions between 397 phyla. For example, range size approximations for *Proteobacteria* and *Bacteroidetes* encompass 398 values spanning the minimum and maximum of the entire dataset, while range size 399 approximations for candidate phyla WPS-2 or Chlorobi have a much narrower size distribution 400 (Fig. 3). Some of this variation in range size for *Proteobacteria* or *Bacteroidetes* is further 401 explained by clear differences in range size at the Family level of taxonomic resolution 402 (Appendix S1: Fig S6).

403

404 We identified a number of phenotypic traits, genomic attributes, and habitat preferences that 405 varied predictably as a function of geographic range size (Fig. 4, Fig. 5). Some of these traits are 406 consistent across phyla, while other traits explain more variation in range size within certain 407 phyla (Appendix S1: Fig S7, Appendix S1: Fig S8). For instance, we found that anaerobes were 408 more likely to have smaller range sizes than aerobes (Fig. 4a), potentially due to their inability to 409 survive dispersal through the oxygen-rich atmosphere. Contrary to expectations, we found non-410 spore forming aerobes had larger range sizes than spore-formers (Fig. 4b). This pattern was 411 consistent for Actinobacteria and Firmicutes, which are phyla with both spore-forming and non-412 spore forming members (Appendix S1: Fig S7b). Either there are other traits that are more 413 important than spore formation in determining dispersal capabilities, or we are limited in our 414 ability to accurately predict spore formation from the available in vitro data. Finally, we found 415 that pigmentation was associated with larger geographic ranges (Fig. 4c), potentially due to 416 pigment production offering UV protection to microbial cells during atmospheric dispersal. 417 Pigments have been shown to protect *Bacillus* endospores from radiation (Moeller *et al.* 2005),

and carotenoid pigments are also shown to protect proteobacterial phytopathogens from UV (To *et al.* 1994; Mohammadi *et al.* 2013).

420

421 Additionally, we found that both genome G+C content and genome size increased with range 422 size (Fig. 5a-b, Appendix S1: Fig S8a-b), although these genomic attributes are also positively 423 correlated with each other (Nishida 2012). Greater G+C content has been associated with 424 genome stability and thermal tolerance in some microbes (Nishio et al. 2003; Mann and Chen 425 2010). Larger genomes correspond to more genes and metabolic pathways that likely confer 426 greater physiological versatility and ability to survive diverse environmental conditions (Bently 427 and Parkhill 2004; Konstantinidis et al. 2006). Our findings are in line with recent studies 428 showing that larger genomes are linked to ubiquity and greater environmental and spatial 429 distributions (Barberán et al. 2014a; Cobo-Simón and Tamames 2017). Conversely, we observed 430 a negative correlation between 16S rRNA gene copy number and range size (Fig. 5c), suggesting 431 that oligotrophic life history strategies (see Klappenbach et al. 2000) are associated with greater 432 range sizes within some phyla. Finally, we found that the inferred habitat preferences of 433 microbes could explain some of the variation in range size. Soil and plant associated taxa had larger range size distributions than marine and aquatic habitat associated taxa (Fig. 4e). Not 434 435 surprisingly, these results suggest that those taxa that are likely found in widespread source 436 environments tend to have larger ranges. While we cannot explicitly determine the source origin 437 for each taxon, phyla that are dominant in soil, including Actinobacteria and Acidobacteria, have 438 some of the largest range size distributions (Fig. 3, Fig. 4e). Conversely, taxa from seawater and 439 other aquatic habitats tend to have smaller ranges (Fig. 4e), a pattern that may result from these 440 source habitats not being as widespread across the sampled region, limited aerosolization of 441 microbial cells from these source environments, or a reduced capacity for these aquatic taxa to survive desiccation. 442

443

Together our results illustrate a wide variation in range size of diverse bacterial and archaeal taxa
found in settled outdoor dust. The shape of the range size frequency distribution of these
microbes is similar to many plants and animals, suggesting similar processes can drive observed
biogeographical patterns. However, the canonical occupancy-abundance relationship explains
little of the variation observed here. Instead, we found range size to vary between major phyla

- and identified phenotypic traits and genomic attributes that also vary across taxonomy. These
- 450 traits likely influence dispersal capabilities or the ability to colonize and establish in an
- 451 environment following a dispersal event. Many dust-associated taxa are of ecological,
- 452 agricultural, and medical importance, and integrating range size calculations and range size
- 453 determinants into microbial ecology will advance our understanding of the spatial distributions
- 454 of taxa of interest. Together, this work highlights the importance of both dispersal dynamics and
 455 habitat distribution in generating patterns in microbial biogeography.
- 456

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- 464 Environment Program.
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466 **REFERENCES**

- 467 Achtman, M. 2008. Evolution, population structure, and phylogeography of genetically
- 468 monomorphic bacterial pathogens. *Annual Review of Microbiology* **62**:53–70.
- Achtman, M., and M. Wagner. 2008. Microbial diversity and the genetic nature of microbial
 species. *Nature Reviews Microbiology* 6: 431–440.
- 471 Agosta, S. J., J. Bernardo, G. Ceballos, and M. A. Steele. 2013. A macrophysiological analysis
 472 of energetic constraints on geographic range size in mammals. *PLoS ONE* 8:e72731.
- Altschul, S. F., W. Gish, W. Miller, E. W. Myers, and D. J. Lipman. 1990. Basic local alignment
 search tool. *Journal of Molecular Biology* 215:403–410.
- 475 Andam, C. P., J. R. Doroghazi, A. N. Campbell, P. J. Kelly, M. J. Choudoir, and D. H. Buckley.
- 476 2016. A latitudinal diversity gradient in terrestrial bacteria of the genus *Streptomyces. mBio*477 7:e02200–15.

- 478 Barberán, A., J. Ladau, J. W. Leff, K. S. Pollard, H. L. Menninger, R. R. Dunn, and N. Fierer.
- 479 2015. Continental-scale distributions of dust-associated bacteria and fungi. *Proceedings of*480 *the National Academy of Sciences* 112:5756–5761.
- 481 Barberán, A., K. S. Ramirez, J. W. Leff, M. A. Bradford, D. H. Wall, and N. Fierer. 2014a. Why
- 482 are some microbes more ubiquitous than others? Predicting the habitat breadth of soil
 483 bacteria. *Ecology Letters* 17:794–802.
- 484 Barberán, A., J. Henley, N. Fierer, and E. O. Casamayor. 2014b. Structure, inter-annual
- recurrence, and global-scale connectivity of airborne microbial communities. *The Science of the Total Environment* **487**:187–195.
- Barberán, A., Velazques, H. C., Jones, S., and Fierer, N. 2017. Hiding in plain sight: mining
 bacterial species records for phenotypic trait information. *mSphere* 2:e00237–17.
- Bebber, D. P., T. Holmes, and S. J. Gurr. 2014. The global spread of crop pests and pathogens. *Global Ecology and Biogeography* 23:1398–1407.
- Bentley, S. D., and J. Parkhill. 2004. Comparative genomic structure of Prokaryotes. *Annual Review of Genetics* 38:771–791.
- Berry, P. E., and R. Riina. 2005. Insights into the diversity of the Pantepui flora and the
 biogeographic complexity of the Guayana Shield. *Biologiske Skrifter* 55:145–167.
- 495 Boucher, Y., O. X. Cordero, A. Takemura, D. E. Hunt, K. Schliep, E. Bapteste, P. Lopez, C. L.
- Tarr, and M. F. Polz. 2011. Local mobile gene pools rapidly cross species boundaries to
 create endemicity within global *Vibrio cholerae* populations. *mBio* 2:e00335–10.
- Brown, J. H., G. C. Stevens, and D. M. Kaufman. 1996. The geographic range: size, shape,
 boundaries. and internal structure. *Annual Review of Ecology and Systematics* 27:597–623.
- 500 Brummitt, N., S. P. Bachman, E. Aletrari, H. Chadburn, J. Griffiths-Lee, M. Lutz, J. Moat, M. C.
- 501 Rivers, M. M. Syfert, and E. M. Nic Lughadha. 2015. The Sampled Red List Index for
- Plants, phase II: ground-truthing specimen-based conservation assessments. *Philosophical Transactions of the Royal Society of London B, Biological sciences* 370:20140015.
- Calenge, C. 2006. The package "adehabitat" for the R software: a tool for the analysis of space
 and habitat use by animals. *Ecological Modelling* 197:516–519.
- 506 Cho, J.-C., and J. M. Tiedje. 2000. Biogeography and degree of endemicity of fluorescent
- 507 *Pseudomonas* strains in soil. *Applied and Environmental Microbiology* **66**:5448–5456.

- 508 Choudoir, M. J., J. R. Doroghazi, and D. H. Buckley. 2016. Latitude delineates patterns of
 509 biogeography in terrestrial *Streptomyces*. *Environmental Microbiology* 18:4931–4945.
- 510 Clark, J. S., Beckage, B., HilleRisLambers, J., Ibanez, I., LaDeau, S., McLachlan, J., Mohan, J.,
- and M. Rocca. 2002. Plant dispersal and migration. Pages 81–93 in R. E. Munn, editor.
- 512 *Encyclopedia of Global Change*. John Wiley & Sons, Ltd., Chichester, U.K.
- Cobo-Simón, M., and J. Tamames. 2017. Relating genomic characteristics to environmental
 preferences and ubiquity in different microbial taxa. *BMC genomics* 18: 499.
- 515 Cunningham, R. B., and D. B. Lindenmayer. 2005. Modeling count data of rare species: some
 516 statistical issues. *Ecology* 86:1135–1142.
- 517 Doolittle, F. 2012. Population genomics: how bacterial species form and why they don't exist.
 518 *Current Biology* 22:R449–R451.
- 519 Edgar, R. C. 2013. UPARSE: highly accurate OTU sequences from microbial amplicon reads.
 520 *Nature Methods* 10:996–998.
- Essl, F., M. Staudinger, O. Stöhr, L. Schratt-Ehrendorfer, W. Rabitsch, and H. Niklfeld. 2009.
 Distribution patterns, range size and niche breadth of Austrian endemic plants. *Biological Conservation* 142:2547–2558.
- 524 Fierer, N., J. W. Leff, B. J. Adams, U. N. Nielsen, S. T. Bates, C. L. Lauber, S. Owens, J. A.
- 525 Gilbert, D. H. Wall, and J. G. Caporaso. 2012. Cross-biome metagenomic analyses of soil
- microbial communities and their functional attributes. *Proceedings of the National Academy*of *Sciences* 109:21390–21395.
- Gaston, K. J., and R. A. Fuller. 2009. The sizes of species' geographic ranges. *The Journal of Applied Ecology* 46:1–9.
- 530 Gaston, K. J. 1996a. The multiple forms of the interspecific abundance-distribution relationship.
 531 *Oikos* 76:211.
- 532 Gaston, K. J., T. M. Blackburn, J. J. D. Greenwood, R. D. Gregory, R. M. Quinn, and J. H.
- Lawton. 2000. Abundance–occupancy relationships. *The Journal of Applied Ecology* 37:39–
 59.
- Gaston, K. J. 1996b. Species-range-size distributions: patterns, mechanisms and implications.
 Trees 11:197–201.

- 537 Gevers, D., Cohan, F. M., Lawrence, J. G., Spratt, B. G., Coenye, T., Feil, E. J., Stackebrandt, E.,
- Van de Peer, Y., Vandamme, P., Thompson, F. L., and J. Swings. 2005. Re-evaluating
 prokaryotic species. *Nature Reviews Microbiology* 3:733–739.
- 540 Ghiglione, J.-F., P. E. Galand, T. Pommier, C. Pedros-Alio, E. W. Maas, K. Bakker, S. Bertilson,
- 541 D. L. Kirchman, C. Lovejoy, P. L. Yager, and A. E. Murray. 2012. Pole-to-pole
- 542 biogeography of surface and deep marine bacterial communities. *Proceedings of the*
- 543 *National Academy of Sciences* **109**:17633–17638.
- Green, J., and B. J. M. Bohannan. 2006. Spatial scaling of microbial biodiversity. *Trends in Ecology and Evolution* 21: 501–507.
- Hanson, C. A., J. A. Fuhrman, M. C. Horner-Devine, and J. B. H. Martiny. 2012. Beyond
 biogeographic patterns: processes shaping the microbial landscape. *Nature Reviews Microbiology*. 10:497–506.
- Herrera-Alsina, L., and R. Villegas-Patraca. 2014. Biologic interactions determining geographic
 range size: a one species response to phylogenetic community structure. *Ecology and Evolution* 4:968–976.
- Holt, A. R., K. J. Gaston, and F. He. 2002. Occupancy-abundance relationships and spatial
 distribution: a review. *Basic and Applied Ecology* 3:1–13.
- Howe, H. F., and J. Smallwood. 982. Ecology of seed dispersal. *Annual Review of Ecology and Systematics* 13:201–228.
- Just, M. G., J. F. Norton, A. L. Traud, T. Antonelli, A. S. Poteate, G. A. Backus, A. Snyder-
- Beattie, R. W. Sanders, and R. R. Dunn. 2014. Global biogeographic regions in a humandominated world: the case of human diseases. *Ecosphere* 5:art143.
- Kim, M., H.-S. Oh, S.-C. Park, and J. Chun. 2014. Towards a taxonomic coherence between
 average nucleotide identity and 16S rRNA gene sequence similarity for species demarcation
- of prokaryotes. *International Journal of Systematic and Evolutionary Microbiology* 64:346–
 351.
- Klappenbach, J. A., J. M. Dunbar, and T. M. Schmidt. 2000. rRNA operon copy number reflects
 ecological strategies of bacteria. *Applied and Environmental Microbiology* 66:1328–1333.
- Kolb, A., F. Barsch, and M. Diekmann. 2006. Determinants of local abundance and range size in
 forest vascular plants. *Global Ecology and Biogeography* 15:237–247.

- 567 Konstantinidis, K. T., A. Ramette, and J. M. Tiedje. 2006. The bacterial species definition in the
- genomic era. *Philosophical transactions of the Royal Society of London B, Biological Sciences* 361:1929–1940.
- 570 Kreft, H., J. H. Sommer, and W. Barthlott. 2006. The significance of geographic range size for
 571 spatial diversity patterns in Neotropical palms. *Ecography* 29:21–30.
- Laube, I., H. Korntheuer, M. Schwager, S. Trautmann, C. Rahbek, and K. Böhning-Gaese. 2013.
 Towards a more mechanistic understanding of traits and range sizes: avian traits and range
 size. *Global Ecology and Biogeography* 22:233–241.
- Lester, S. E., and B. I. Ruttenberg. 2005. The relationship between pelagic larval duration and
 range size in tropical reef fishes: a synthetic analysis. *Proceedings of the Royal Society B*, *Biological Sciences* 272:585–591.
- Lester, S. E., B. I. Ruttenberg, S. D. Gaines, and B. P. Kinlan. 2007. The relationship between
 dispersal ability and geographic range size. *Ecology Letters* 10:745–758.
- Locey, K. J., and J. T. Lennon. 2016. Scaling laws predict global microbial diversity. *Proceedings of the National Academy of Sciences* 113: 5970–5975.
- Lynch, M. D. J., and J. D. Neufeld. 2015. Ecology and exploration of the rare biosphere. *Nature Reviews Microbiology* 13:217–229.
- 584 MacKenzie, D. I., J. D. Nichols, G. B. Lachman, S. Droege, J. Andrew Royle, and C. A.
- Langtimm. 2002. Estimating site occupancy rates when detection probabilities are less than
 one. *Ecology* 83:2248–2255.
- 587 Macpherson, E. 2003. Species range size distributions for some marine taxa in the Atlantic
 588 Ocean. Effect of latitude and depth. *Biological Journal of the Linnean Society* 80:437–455.
- Mann, S., and Y.-P. P. Chen. 2010. Bacterial genomic G+C composition-eliciting environmental
 adaptation. *Genomics* 95:7–15.
- 591 Markowitz, V. M., I.-M. A. Chen, K. Palaniappan, K. Chu, E. Szeto, M. Pillay, A. Ratner, J.
- 592 Huang, T. Woyke, M. Huntemann, I. Anderson, K. Billis, N. Varghese, K. Mavromatis, A.
- Pati, N. N. Ivanova, and N. C. Kyrpides. 2014. IMG 4 version of the integrated microbial
 genomes comparative analysis system. *Nucleic Acids Research* 42:D560–D567.
- 595 Martiny, J. B. H., B. J. M. Bohannan, J. H. Brown, R. K. Colwell, J. A. Fuhrman, J. L. Green, M.
- 596 C. Horner-Devine, M. Kane, J. A. Krumins, C. R. Kuske, P. J. Morin, S. Naeem, L. Øvreås,

- A.-L. Reysenbach, V. H. Smith, and J. T. Staley. 2006. Microbial biogeography: putting
 microorganisms on the map. *Nature Reviews Microbiology* 4:102–112.
- McCauley, S. J., C. J. Davis, E. E. Werner, and M. S. Robeson. 2014. Dispersal, niche breadth
 and population extinction: colonization ratios predict range size in North American
 dragonflies. *The Journal of Animal Ecology* 83:858–865.
- 602 McDonald, D., M. N. Price, J. Goodrich, E. P. Nawrocki, T. Z. DeSantis, A. Probst, G. L.
- 603 Andersen, R. Knight, and P. Hugenholtz. 2012. An improved Greengenes taxonomy with 604 explicit ranks for ecological and evolutionary analyses of bacteria and archaea. *The ISME*
- 604 explicit ranks for ecological and evolutionary analyses of bacteria and archaea. *The ISME*605 *Journal* 6:610–618.
- 606 Mittelbach, G. G., D. W. Schemske, H. V. Cornell, A. P. Allen, J. M. Brown, M. B. Bush, S. P.
- Harrison, A. H. Hurlbert, N. Knowlton, H. A. Lessios, C. M. McCain, A. R. McCune, L. A.
- 608 McDade, M. A. McPeek, T. J. Near, T. D. Price, R. E. Ricklefs, K. Roy, D. F. Sax, D.
- Schluter, J. M. Sobel, and M. Turelli. 2007. Evolution and the latitudinal diversity gradient:
 speciation, extinction and biogeography. *Ecology Letters* 10:315–331.
- 611 Moeller, R., G. Horneck, R. Facius, and E. Stackebrandt. 2005. Role of pigmentation in
- 612 protecting *Bacillus* sp. endospores against environmental UV radiation. *FEMS Microbiology* 613 *Ecology* 51:231–236.
- Mohammadi, M., L. Burbank, and M. C. Roper. 2012. Biological role of pigment production for
 the bacterial phytopathogen *Pantoea stewartii* subsp. *stewartii*. *Applied and Environmental Microbiology* 78:6859–6865.
- 617 Morueta-Holme, N., B. J. Enquist, B. J. McGill, B. Boyle, P. M. Jørgensen, J. E. Ott, R. K. Peet,
- 618 I. Šímová, L. L. Sloat, B. Thiers, C. Violle, S. K. Wiser, S. Dolins, J. C. Donoghue, N. J. B.
- 619 Kraft, J. Regetz, M. Schildhauer, N. Spencer, and J.-C. Svenning. 2013. Habitat area and
- 620 climate stability determine geographical variation in plant species range sizes. *Ecology*
- 621 *Letters* **16**:1446–1454.
- Mouillot, D., and K. Gaston. 2009. Spatial overlap enhances geographic range size conservatism.
 Ecography 32:671–675.
- 624 Murray, K. A., N. Preston, T. Allen, C. Zambrana-Torrelio, P. R. Hosseini, and P. Daszak. 2015.
- 625 Global biogeography of human infectious diseases. *Proceedings of the National Academy of*
- 626 *Sciences* 112:12746–12751.

627	Müller, A	Albert L.,	de Rezende.	Júlia Rosa.	Hubert,	Casey	RJ. K	lieldsen.	Kasper	Urup.
-	7						- 2	,		

628 Lagkouvardos, Ilias, Berry, David, Jørgensen, Bo Barker, and Loy, Alexander. 2014.

Endospores of thermophilic bacteria as tracers of microbial dispersal by ocean currents. *The ISME Journal* 8:1153–1165.

- 631 Nemergut, D. R., E. K. Costello, M. Hamady, C. Lozupone, L. Jiang, S. K. Schmidt, N. Fierer,
- A. R. Townsend, C. C. Cleveland, L. Stanish, and R. Knight. 2011. Global patterns in the
 biogeography of bacterial taxa. *Environmental Microbiology* 13:135–144.
- Nishida, H. 2012. Evolution of genome base composition and genome size in bacteria. *Frontiers in Microbiology* 3:420.
- 636 Nishio, Yousuke, Nakamura, Yoji, Kawarabayasi, Yutaka, Usuda, Yoshihiro, Kimura, Eiichiro,

637 Sugimoto, S, Matsui, K, Yamagishi, A, Kikuchi, H, Ikeo, K, and Gojobori, T. 2003.

638 Comparative complete genome sequence analysis of the amino acid replacements

- responsible for the thermostability of *Corynebacterium efficiens*. *Genome Research* **13**:1572–1579.
- Ofstad, E. G., I. Herfindal, E. J. Solberg, and B.-E. Sæther. 2016. Home ranges, habitat and body
 mass: simple correlates of home range size in ungulates. *Proceedings of the Royal Society Bm Biological Sciences* 283:20161234.
- Orme, C. D. L., R. G. Davies, V. A. Olson, G. H. Thomas, T.-S. Ding, P. C. Rasmussen, R. S.
 Ridgely, A. J. Stattersfield, P. M. Bennett, and I. P. F. Owens. 2006. Global patterns of
 geographic range size in birds. *PLoS Biology* 4:e208.
- Papke, R. T., N. B. Ramsing, M. M. Bateson, and D. M. Ward. 2003. Geographical isolation in
 hot spring cyanobacteria. *Environmental Microbiology* 5:650–659.
- Paul, J. R., C. Morton, C. M. Taylor, and S. J. Tonsor. 2009. Evolutionary time for dispersal
 limits the extent but not the occupancy of species' potential ranges in the tropical plant
- 651 genus *Psychotria* (Rubiaceae). *The American Naturalist* **173**:188–199.
- Paulson, J. N., O. C. Stine, H. C. Bravo, and M. Pop. 2013. Differential abundance analysis for
 microbial marker-gene surveys. *Nature Methods* 10:1200–1202.
- Pebesma, E., and R. S. Bivand. 2005. Classes and methods for spatial data in R. *R News* 5.
- 655 Pichancourt, J.-B., and R. D. van Klinken. 2012. Phenotypic plasticity influences the size, shape
- and dynamics of the geographic distribution of an invasive plant. *PloS ONE* **7**:e32323.

- Pigot, A. L., I. P. F. Owens, and C. D. L. Orme. 2010. The environmental limits to geographic
 range expansion in birds. *Ecology Letters* 13:705–715.
- Pohlman, C. L., A. B. Nicotra, and B. R. Murray. 2005. Geographic range size, seedling
 ecophysiology and phenotypic plasticity in Australian Acacia species. *Journal of*
- 661 *Biogeography* **32**:341–351.
- Roberts, M. S., and F. M. Cohan. 1995. Recombination and migration rates in natural
 populations of *Bacillus subtilis* and *Bacillus mojavensis*. *Evolution* 49:1081–1094.
- Roney, N. E., A. Kuparinen, and J. A. Hutchings. 2015. Comparative analysis of abundance–
 occupancy relationships for species at risk at both broad taxonomic and spatial scales. *Canadian Journal of Zoology* 93:515–519.
- Roselló-Mora, R., and R. Amann. 2001. The species concept for prokaryotes. *FEMS Microbiology Reviews* 25:39–67.
- Rosenfield, J. A. 2002. Pattern and process in the geographical ranges of freshwater fishes. *Global Ecology and Biogeography* 11:323–332.
- Ruff, S. E., J. F. Biddle, A. P. Teske, K. Knittel, A. Boetius, and A. Ramette. 2015. Global
 dispersion and local diversification of the methane seep microbiome. *Proceedings of the National Academy of Sciences* 112:4015–4020.
- Salisbury, C. L., N. Seddon, C. R. Cooney, and J. A. Tobias. 2012. The latitudinal gradient in
 dispersal constraints: ecological specialisation drives diversification in tropical birds. *Ecology Letters* 15:847–855.
- 677 Salter, S. J., M. J. Cox, E. M. Turek, S. T. Calus, W. O. Cookson, M. F. Moffatt, P. Turner, J.
- Parkhill, N. J. Loman, and A. W. Walker. 2014. Reagent and laboratory contamination can
 critically impact sequence-based microbiome analyses. *BMC biology* 12:87.
- Schlachter, K. J. 2010. Range shape and range elongation of North American trees. *Physical Geography* 31:40–57.
- Sergio, C., R. Figueira, D. Draper, R. Menezes, and A. Sousa. 2007. Modelling bryophyte
 distribution based on ecological information for extent of occurrence assessment. *Biological Conservation* 135:341–351.
- 685 Sileshi, G., G. Hailu, and G. I. Nyadzi. 2009. Traditional occupancy–abundance models are
- inadequate for zero-inflated ecological count data. *Ecological Modelling* **220**:1764–1775.

- 687 Smith, D. J., H. J. Timonen, D. A. Jaffe, D. W. Griffin, M. N. Birmele, K. D. Perry, P. D. Ward,
- and M. S. Roberts. 2013. Intercontinental dispersal of bacteria and archaea by transpacific
 winds. *Applied and Environmental Microbiology* **79**:1134–1139.
- 690 Sogin, M. L., H. G. Morrison, J. A. Huber, D. M. Welch, S. M. Huse, P. R. Neal, J. M. Arrieta,
- and G. J. Herndl. 2006. Microbial diversity in the deep sea and the underexplored "rare
 biosphere." *Proceedings of the National Academy of Sciences* 103:12115–12120.
- 693 Stackebrandt, E., and B. M. Goebel. 1994. Taxonomic note: a place for DNA-DNA reassociation
 694 and 16S rRNA sequence analysis in the present species definition in bacteriology.
- 695 *International Journal of Systematic and Evolutionary Microbiology* **44**:846–849.
- Sul, W. J., T. A. Oliver, H. W. Ducklow, L. A. Amaral-Zettler, and M. L. Sogin. 2013. Marine
 bacteria exhibit a bipolar distribution. *Proceedings of the National Academy of Sciences*110:2342–2347.
- Svenning, J.-C., and F. Skov. 2007. Could the tree diversity pattern in Europe be generated by
 postglacial dispersal limitation? *Ecology Letters* 10:453–460.
- To, K.-Y., E.-M. Lai, L.-Y. Lee, T.-P. Lin, C.-H. Hung, C.-L. Chen, Y.-S. Chang, and S.-T. Liu.
 1994. Analysis of the gene cluster encoding carotenoid biosynthesis in *Erwinia herbicola*Eho13. *Microbiology* 140:331–339.
- Vos, M., and G. J. Velicer. 2008. Isolation by distance in the spore-forming soil bacterium
 Myxococcus xanthus. Current Biology 18:386–391.
- Wang, Q., G. M. Garrity, J. M. Tiedje, and J. R. Cole. 2007. Naive Bayesian classifier for rapid
- assignment of rRNA sequences into the new bacterial taxonomy. *Applied and Environmental Microbiology* 73:5261–5267.
- Weil, T., C. De Filippo, D. Albanese, C. Donati, M. Pindo, L. Pavarini, F. Carotenuto, M.
- 710 Pasqui, L. Poto, J. Gabrieli, C. Barbante, B. Sattler, D. Cavalieri, and F. Miglietta. 12/2017.
- 711 Legal immigrants: invasion of alien microbial communities during winter occurring desert
 712 dust storms. *Microbiome* 5:23.
- Wenger, S. J., and M. C. Freeman. 2008. Estimating species occurrence, abundance, and
 detection probability using zero-inflated distributions. *Ecology* 89:2953–2959.
- 715 Whitaker, R. J. 2003. Geographic barriers isolate endemic populations of hyperthermophilic
- 716 Archaea. *Science* **301**:976–978.

Wiens, J. J., C. H. Graham, D. S. Moen, S. A. Smith, and T. W. Reeder. 2006. Evolutionary and 717 718 ecological causes of the latitudinal diversity gradient in hylid frogs: treefrog trees unearth 719 the roots of high tropical diversity. The American Naturalist 168:579–596. 720 Yamaguchi, N., T. Ichijo, A. Sakotani, T. Baba, and M. Nasu. 2012. Global dispersion of 721 bacterial cells on Asian dust. Scientific Reports 2:535. 722 723 DATA AVAILABILITY 724 725 Data associated with this study are available on figshare: https://doi.org/10.6084/m9.figshare.1270900.v8 726 727 728 729 **FIGURE LEGENDS** 730 731 Figure 1. 732 Map of the contiguous United States with the locations of the 1,065 outdoor dust samples shown 733 with blue points (1a). Geographic range size was calculated for dust taxa using two approaches, the area of occupancy (AOO) and the extent of occurrence (EOO) approximations (see 734 735 Methods). Kernel density distributions for occupancy (i.e. total observations across sample sites) 736 (1b), area of occupancy (AOO) range estimations (1c), and extent of occurrence (EOO) range 737 estimations (1d) for dust phylotypes. 738 Figure 2. 739 740 Points show the maximum longitudinal and corresponding latitudinal range for each phylotype. 741 Phylotypes with greater east-west spread also tend to have greater north-south spread (Spearman's $\rho = 0.67$, P < 0.0001). Since the United States has greater east-west than north-742 south dimensions, the blue dashed line normalizes for this difference and depicts the ratio of 743 744 possible maximum spread. Points above this line (10.1%) indicate ranges elongated north-south, 745 and points below this line (89.9%) indicate ranges elongated east-west (see inset). See Appendix 746 S1: Fig S3 for the density distributions of longitudinal and latitudinal ranges. 747

- 748 Figure 3.
- Boxplots illustrating range size distributions for dust taxa ranked by Phylum. Log_{10} AOO range size estimations vary significantly between phyla (one-way ANOVA; $F_{28, 70297} = 136.2$, P <
- 751 0.0001).
- 752 753 Figure 4.
- 754 Phenotypic traits and source habitats of dust bacteria were inferred by matching representative
- partial 16S rRNA phylotype sequences to full length 16S rRNA sequences in the IJSEM
- 756 phenotype database (see Methods). Boxplots illustrate the relationship between the AOO range
- size estimation and oxygen tolerance (4a), spore formation in obligate aerobes (4b), pigmentation
- (4c), Gram stain (4d), and habitat (4e) for the most abundant phyla including *Proteobacteria*,
- 759 Actinobacteria, Firmicutes, and Bacteroidetes. Range size varies significantly with oxygen
- 760 tolerance (ANOVA; $F_{3, 2324} = 67.5$, P < 0.0001) (4a), spore formation (ANOVA; $F_{1, 1643} = 5.0$, P
- 761 = 0.025) (4b), pigmentation (ANOVA; $F_{1, 1802}$ = 55.4, P < 0.0001) (5c), Gram stain (ANOVA;
- 762 $F_{2,31207} = 32.4, P < 0.0001$) (4d), and habitat (ANOVA; $F_{7,907} = 11.1, P < 0.0001$). See Appendix
- 763 S1: Fig S7 for phenotypic traits by phyla.
- 764

765 **Figure 5.**

- 766 Genomic attributes of dust taxa were inferred by matching representative partial 16S rRNA
- requences to full length 16S rRNA sequences in the IMG database (see Methods). Panels depict
- the relationship between AOO range size estimation and mean G+C content (5a), genome size
- (5b), and log₁₀ 16S rRNA copy number (5c) for the most abundant phyla including
- 770 Proteobacteria, Actinobacteria, Firmicutes, and Bacteroidetes. Points depict the AOO range size
- estimations and the mean values of genomic traits. Blue lines show the linear regression with
- gray shading indicating 95% confidence intervals. Pearson's product-moment correlation r is
- reported. See Appendix S1: Fig S8 for genomic traits by phyla.
- 774

Figure 1.







anusc Figure 4. utl

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Figure 5.

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