





## ORIGINAL RESEARCH

# Does perspective matter? A case study comparing Eulerian and Lagrangian estimates of common murre (*Uria aalge*) distributions

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## Abstract

Studies estimating species' distributions require information about animal locations in space and time. Location data can be collected using surveys within a predetermined frame of reference (i.e., Eulerian sampling) or from animal-borne tracking devices (i.e., Lagrangian sampling). Integration of observations obtained from Eulerian and Lagrangian perspectives can provide insights into animal movement and habitat use. However, contemporaneous data from both perspectives are rarely available, making examination of biases associated with each sampling approach difficult. We compared distributions of a mobile seabird observed concurrently from ship, aerial, and satellite tag surveys during May, June, and July 2012 in the northern California Current. We calculated utilization distributions to quantify and compare variability in common murre (*Uria aalge*) space use and examine how sampling perspective and platform influence observed patterns. Spatial distributions of murre were similar in May, regardless of sampling perspective. Greatest densities occurred in coastal waters off southern Washington and northern Oregon, near large murre colonies and the mouth of the Columbia River. Density distributions of murre estimated from ship and aerial surveys in June and July were similar to those observed in May, whereas distributions of satellite-tagged murre in June and July indicated northward movement into British Columbia, Canada, resulting in different patterns observed from Eulerian and Lagrangian perspectives. These results suggest that the population of murre observed in the northern California Current during spring and summer includes relatively stationary individuals attending breeding colonies and nonstationary, vagile adults and subadults. Given the expected growth of telemetry studies and advances in survey technology (e.g., unmanned aerial systems), these results highlight the importance of considering methodological approaches, spatial extent, and synopticity of distribution data sets prior to integrating data from different sampling perspectives.

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## KEYWORDS

aerial survey, common murre, satellite telemetry, ship-based survey, species distribution, survey design

## 1 | INTRODUCTION

Distribution and abundance data of mobile species are useful for identifying important foraging, migration, and breeding habitats (Elith & Leathwick, 2009; MacArthur, 1972). Data can be obtained from observations collected during surveys within a predetermined frame of reference (i.e., Eulerian sampling) or by sampling discrete locations estimated using animal-borne tracking devices (i.e., Lagrangian sampling; Rutz & Hays, 2009; Tremblay et al., 2009).

Eulerian survey designs sample at  $x$ - $y$  coordinates at predetermined stations or along contiguous transects, often replicated through time. The primary objective of Eulerian sampling approaches is to obtain information about animal distribution and abundance in a predefined area and time period. In the ocean, vessel-based Eulerian surveys regularly use direct sightings to quantify the distributions of marine mammals (Ainley, Dugger, Toniolo, & Gaffney, 2007; Ballance & Pitman, 1998; Keiper, Ainley, Allen, & Harvey, 2005) and seabirds (Ainley et al., 2005; Ballance, Pitman, & Reilly, 1997). Ships can survey coastal and offshore ecosystems for relatively long (i.e., weeks to months) periods across hundreds to thousands of kilometers, and simultaneously sample in situ abiotic and biotic factors including seawater temperature, chlorophyll concentration, and prey species abundance and composition, which allows quantification of animal-habitat relationships (Ainley, Ribic, & Woehler, 2012; Fiedler et al., 1998). However, ships are slow relative to the movement of mobile species including seabirds, and the flux of birds into or out of a survey area, as well as vessel avoidance or attraction by some species, may bias distribution and abundance estimates by convoluting spatial patterns with the passage of time (van Franeker, 1994; Wahl & Heinemann, 1979). Aerial surveys (e.g., airplanes and drones) are another Eulerian sampling approach that sample along transects in a relatively short (i.e., hours to days) period and, because the movement of seabirds is slow relative to an aircraft, provide a synoptic estimate of species distribution and abundance (Briggs, Tyler, & Lewis, 1985a; Buckland et al., 2001; Certain & Bretagnolle, 2008). Aircraft can survey areas often inaccessible to ships (e.g., nearshore shallow habitats and ice fields), but may not be able to transit as far offshore to survey pelagic habitats beyond the continental shelf (Henkel, Ford, Tyler, & Davis, 2007; Hodgson, Baylis, Mott, Herrod, & Clarke, 2016). Accordingly, ship-based and aerial survey data are limited by the spatial and temporal extent and sampling resolution of the survey (Watanuki et al., 2016). Species detectability can also be an issue, as smaller, rare, or cryptic species may not be accurately represented in a data set (Barbraud & Thiebot, 2009; Monk, 2014). Further, for many species, breeding status, sex, and age of individual seabirds cannot be discerned from sighting data, constraining most analyses to the population level. Despite these limitations, transect

surveys from ships transiting the world's oceans were an early and significant contributor to studies of pelagic seabird distributions (Brown, 1980; Murphy, 1936; Wynne-Edwards, 1935), and ship and aircraft surveys continue to be an important component of seabird research (Ainley et al., 2009; Certain & Bretagnolle, 2008; Hunt et al., 2018).

In contrast to Eulerian approaches, Lagrangian survey designs track seabirds through space and time using data logging or tracking devices attached to individuals (Burger & Shaffer, 2008; Hart & Hyrenbach, 2009; Hooker, Biuw, McConnell, Miller, & Sparling, 2007). Satellite-linked tags that provide near-real-time, continuous, and independent sampling are a common tool for Lagrangian sampling (Adams, MacLeod, Suryan, Hyrenbach, & Harvey, 2012; Hatch, Meyers, Mulcahy, & Douglas, 2000). Depending on mobility of the species, a Lagrangian sampling approach may increase the spatial extent and resolution of the survey area compared with an Eulerian perspective (Block, Costa, Boehlert, & Kochevar, 2002). Fine-scale (i.e., 1–10 km) movements of individuals can be measured with satellite tags and then matched as closely as possible to remotely sensed, modeled, or in situ environmental data to gain insights on correlations between movement and habitat use (Adams & Flora, 2010; Phillips, Horne, Adams, & Zamon, 2018). Further, many tracking devices now carry additional sensors that provide insight on physiology and foraging behavior (Burger & Shaffer, 2008; Ropert-Coudert & Wilson, 2005; Wilson et al., 2002). While telemetry provides high-resolution data at an individual level, transmitter cost and logistical challenges can limit the number of tags deployed (i.e., sample size; Lindberg & Walker, 2007). Individual heterogeneity, often attributed to sex and age differences (Gutowsky, Leonard, Conners, Shaffer, & Jonsen, 2015; Hedd, Montevecchi, Phillips, & Fifield, 2014), complicates population-level inferences (Krietsch et al., 2017), and presence-only data often require additional steps to develop habitat models (Lobo, Jiménez-Valverde, & Hortal, 2010; Phillips et al., 2009). Despite sampling constraints, significant advances in tracking technology during the last two decades have resulted in important insights into seabird movement and distribution (Shaffer et al., 2006; Votier, Bicknell, Cox, Scales, & Patrick, 2013; Weimerskirch, Bishop, Jeanniard-du-Dot, Prudor, & Sachs, 2016) and tracking tags are now used on many wide-ranging avian species (Hart & Hyrenbach, 2009; Tremblay et al., 2009).

As the number of Eulerian and Lagrangian studies of marine mammals and seabirds increases (Block et al., 2016; Drew, Piatt, & Renner, 2015), efforts to combine data from these two perspectives have increased. This is due in part to the potential to expand spatial and temporal sampling scales, which could enhance studies of species' distributions and inform conservation efforts (Fujioka et

al., 2014; Watanuki et al., 2016). Data from Eulerian and Lagrangian perspectives or platforms may be complementary, but integration can be complicated by biases inherent in data collected from different sampling approaches, including a mismatch in spatiotemporal sampling coverage. Concurrent and spatially overlapping data from both Eulerian and Lagrangian perspectives are rare, consequently differences in species distribution patterns attributable to sampling perspective are difficult to evaluate.

We used contemporaneous data from Eulerian and Lagrangian surveys to examine whether sampling perspective or platform influences estimates of a seabird's distribution. We quantified and compared common murre (*Uria aalge*) density distributions observed during May, June, and July 2012 from ship, aerial, and satellite telemetry surveys in the northern California Current. Murres are one of the most numerous seabird species along the west coast of North America (Briggs, Tyler, Lewis, & Carlson, 1987; Carter et al., 2001; Thomas & Lyons, 2017), with ~532,000 individuals attending colonies and breeding along the Oregon and Washington coasts during spring and summer (April–August; Naughton, Pitkin, Lowe, So, & Strong, 2007; Speich & Wahl, 1989). Nesting adult murres are central place foragers that search for prey within ~100 km of their colony (Davoren, Montevecchi, & Anderson, 2003; Decker & Hunt, 1996; Hatch et al., 2000). Thus, the expected movement constraints of murres and the availability of concurrent ship, plane, and telemetry data sets allowed us to compare spatial patterns of murres observed during the breeding season using different sampling perspectives and platforms.

## 2 | METHODS

All sampling was conducted in continental shelf waters along the northern Oregon and Washington coasts, with a focus near the mouth of the Columbia River and colonies adjacent to this geographic feature.

### 2.1 | Eulerian sampling

#### 2.1.1 | Ship-based surveys

We used ship-based data from an ongoing ecosystem research program examining the ocean ecology of salmon off the Washington and Oregon coasts (Brodeur, Myers, & Helle, 2003). Using standard strip transect survey methods (Tasker, Jones, Dixon, & Blake, 1984) during daylight hours in May and June 2012, we collected direct sightings of flying or floating murres (Figure 1) within 300 m of a chartered, commercial fishing vessel (for full details, see Phillips, Horne, & Zamon, 2017). Each sighting was spatially and temporally indexed with Global Positioning System (GPS) coordinates using SeeBird software (v 4.1.5.0; NOAA Fisheries Southwest Fisheries Science Center, La Jolla, California, USA). Each east-west transect was ~40 km in length, with survey efforts beginning offshore and the ship traveling shoreward for 2 hr at

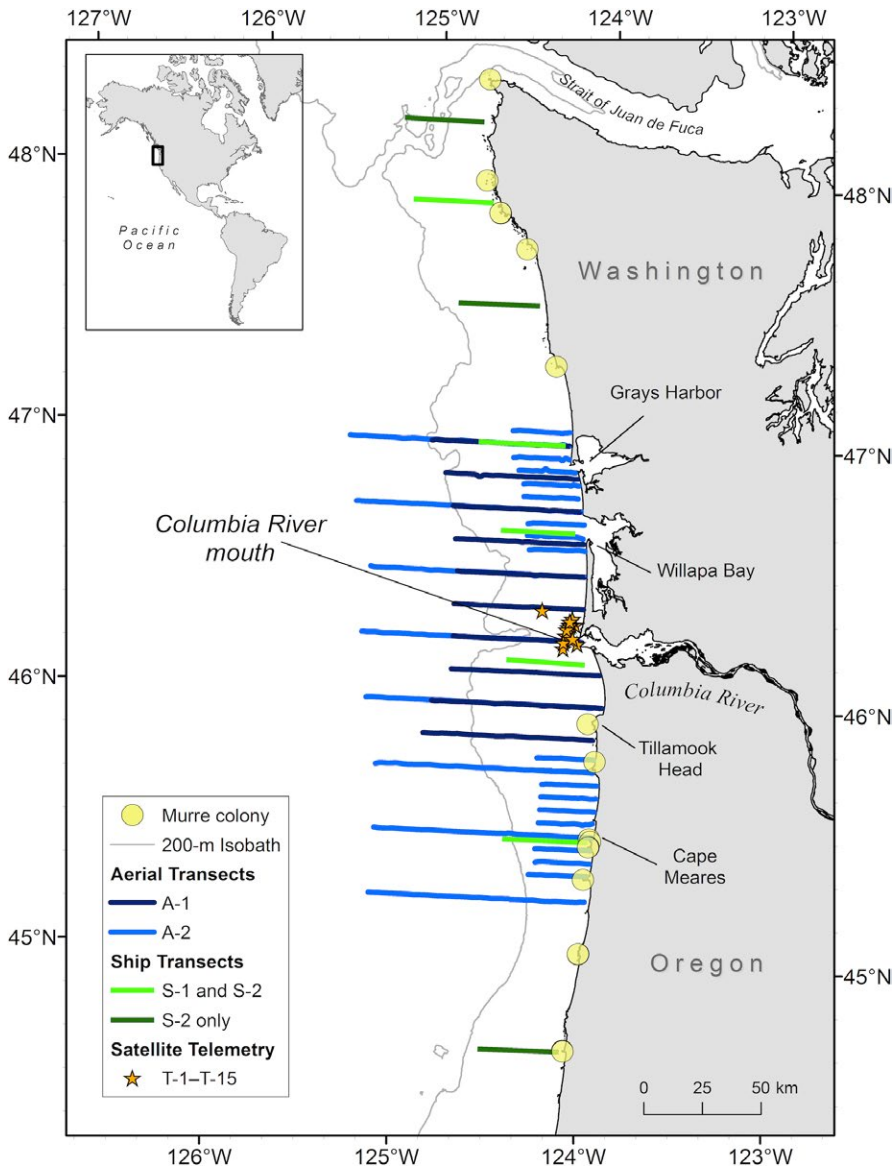


**FIGURE 1** Common murres (*Uria aalge*) observed floating on the surface of the water from a ship survey. Photograph credit: J.E. Zamon/NOAA Fisheries

~5 m/s to within ~3–5 km of shore (Figure 2). To sample a large latitudinal range (44.7–48.2°N) of the northern California Current, the north–south distance between transects ranged from 35 to 90 km. Data were collected along five transects during a survey in late May to early June 2012 (S-1) and on eight transects during late June 2012 (S-2; Table 1).

#### 2.1.2 | Aerial surveys

We used data from aerial surveys of the northern California Current conducted by the U.S. Geological Survey (Adams, Felis, Mason, & Takekawa, 2014). Sightings of murres were recorded from twin-engine, high-wing aircraft (Partenavia P-68, Aspen Helicopters, Oxnard, CA, or Commander AC-500, GoldAero, Arlington, WA) along predetermined, systematic, east-west-oriented transects flown at 160 km/h from the 2000-m isobath to shore (~90 km; Figure 2). Using aerial survey methods, modified slightly from Mason et al. (2007), two observers counted all birds observed in 150-m strip transects (75 m per side) from 60 m above sea level. The low-elevation survey methods were reviewed by NOAA's National Marine Fisheries Service, who granted a Letter of Concurrence to the U.S. Geological Survey. The number and location of individual murres were linked using observation time with GPS data that allowed simultaneous collection of coordinates. Sampling occurred in a latitudinal range from 45.2 to 47.0°N. Data were collected on 10 transects on 19 May 2012 (A-1) and on 24 transects on 1 and 4 July 2012 (A-2; Table 1). Transects flown during A-1 were spaced 13.9 km apart and extended 72.4-km offshore, whereas A-2 included a mix of broad survey transects (27.8-km spacing, up to 93.6-km offshore) and two focal-area surveys (each with ten, 25-km-long parallel transect lines spaced 6 km apart) nested within the broad survey transects (Adams et al., 2014; Figure 2). For this study, we treated counts of murres obtained during the two July surveys as one survey for analyses (i.e., all transects were analyzed together) unless otherwise noted.



**FIGURE 2** Study area off the Washington and Oregon coast, with geographical points of interest labeled. All surveys were conducted during spring–summer 2012. Ship transects surveyed during 30 May–3 June (S-1) and 21–28 June (S-2) are shown in light green; transects surveyed only during S-2 are shown in dark green. Aerial transects flown on 19 May (A-1) are shown in dark blue; broad survey transects and focal-area surveys flown on 1 and 4 July (A-2) are shown in light blue. Locations of tag deployments for satellite-tracked common murre (*Uria aalge*; T-1–T-15) released near the mouth of the Columbia River on 4–5 May are shown as orange stars. Major murre colonies (>2,000 birds) are shown as yellow points, and those identified in the text are labeled

## 2.2 | Lagrangian sampling

### 2.2.1 | Satellite telemetry

We used locations collected from satellite tags (Telonics TAV-2617 platform terminal transmitters [PTTs]) deployed on 12 murre captured and released at night at sea near the mouth of the Columbia River on 4 and 5 May 2012 (Figure 2). Authority for satellite telemetry was provided by USGS Bird Banding Laboratory Auxiliary Marking Authority no. 22911 (J.A.) and no. 23682 (J.E.Z.), and State of Washington Scientific Collection Permit no. 05-500 (J.E.Z.). Capture and tagging methods were approved under the USGS Animal Care and Use Committee #WERC-2007-03. PTTs were programmed to transmit every 60 s for 4 hr in the morning (08:00–12:00 hours) and 4 hr in the evening (14:00–18:00 hours), which coincided with Eulerian surveys that were conducted during daylight hours. Locations of individual birds were determined using the ARGOS system ([www.argos-system.org](http://www.argos-system.org); CLS, 2013) and

archived via the Satellite Tracking and Analysis Tool (STAT; Coyne & Godley, 2005). To resolve tag attachment or instrument failure, we removed data from tags that did not transmit for more than 2 weeks, had intermittent transmissions (e.g., 5-day gap in transmissions), or showed evidence of halted movement (i.e., when median daily movements fell below the 95% confidence interval of average movement of birds for the sampling year; S. Loredó pers. comm.). To maximize location accuracy, all ARGOS location class data (LC-3 through LC-B, excluding LC-Z) were filtered using speed, distance, and angle, resulting in a nominal spatial accuracy of 3 km (*mfilter* function in R package *argosfilter*, Freitas, Lydersen, Fedak, & Kovacs, 2008; for full details see Phillips et al., 2018). We also plotted all tag locations in ArcMap 10.3 (ESRI, Redlands, CA) over a high-resolution land layer to determine whether any tagged murre utilized colonies during the study period. To determine sex of tagged birds, we collected blood from each murre during tag deployments by aseptic puncture of the medial metatarsal vein and placed one drop of blood on a buffered molecular sexing card for analysis by Zoogen, Inc. (Davis, CA).

**TABLE 1** Description of ship, plane, and satellite telemetry-based data collections for common murres (*Uria aalge*) in 2012 including sampling perspective and platform type, survey identity, date range, duration, track length, and total sightings or tag locations used for analyses

Perspective	Platform	ID	Date range	Duration (days)	Track length (km)	Total sightings/locations
Eulerian	Ship	S-1	5/30–6/3	5	145.8	428
		S-2	6/21–6/28	8	262.1	749
	Plane	A-1	5/19	1	600.6	618
		A-2	7/1, 7/4	2	1,160.2	880
Lagrangian	PTT	T-1	5/5–6/11	37	1541.6	182
		T-2	5/5–5/23	19	859.3	112
		T-3	5/6–7/15	70	4,649.3	317
		T-4	5/5–6/3	29	1,478.3	156
		T-5	5/5–7/10	66	2,677.6	281
		T-6	5/5–7/17	72	3,323.5	348
		T-7	5/5–7/9	65	2,501.0	301
		T-8	5/5–7/13	68	3,290.1	337
		T-9	5/6–7/4	60	2,162.9	295
		T-12	5/5–5/23	18	670.4	100
		T-13	5/6–7/23	79	4,409.6	341
		T-15	5/5–7/11	67	2,890.4	317

### 2.3 | Data analysis

For the two Eulerian data sets, we first compared overall density of murres observed during ship and aerial surveys. We calculated densities of murres observed during ship-based surveys by dividing the total number of murres counted in 3-km bins (~10-min increments) by the strip area searched (0.9 km<sup>2</sup>) to obtain murres/km<sup>2</sup>. Similarly, we calculated densities of murres observed during aerial surveys by dividing the total number of murres counted in 2.4-km bins (~1-min increments) by the strip area (either 0.18 km<sup>2</sup> [one observer] or 0.36 km<sup>2</sup> [two observers]) to obtain murres/km<sup>2</sup>. To determine whether mean densities differed within data sets, we compared densities observed during S-1 and S-2, and A-1 and A-2 using *t* tests (Zar, 1999). To determine whether offshore distribution patterns varied by survey method, we evaluated histograms of the frequency of murres observed as a function of distance from shore. We removed the focal-area survey data from S-2 histogram plots as these transects did not extend beyond 25 km of shore.

Because absolute densities cannot be estimated from locations of satellite-tagged murres, we calculated Brownian bridge utilization distributions (Horne, Garton, Krone, & Lewis, 2007) to estimate each tagged murre's probability of occurrence using the *kernelbb* function in R package *adehabitat* (Calenge, 2006). A utilization distribution (UD) is a probability distribution that gives the probability density that an animal is found at a given point in space. It is estimated by sampling the location of individuals in space through time. The Brownian bridge UD approach provides an estimate of space use from animal trajectories with serial autocorrelation of relocations (Horne et al., 2007). We created an overall 99% UD for all 12 murres by first calculating 99% UD for each individual bird (i.e., 99%

cumulative probability that an individual murre would be present in all 3-km<sup>2</sup> cells) and then proportionately weighting the individual UD by its tracking duration (i.e., tracking days per individual divided by total tracking days for all individuals) and summing with the rest of the individually weighted UD. The overall UD represents a summed probability density surface of tagged murre space use during the full duration of tag transmissions, with a spatial resolution of 3 km<sup>2</sup>. Because UD values are calculated from a population of individuals and have a spatial context, they are similar to mapped densities and can be compared. To estimate concurrent tagged murre distributions during each ship or aerial survey, we calculated separate UD of tagged murres during each survey time period, using the full spatial extent of tag locations. For the ship surveys, this included a UD during 30 May–3 June (S-1; *n* = 10 tagged birds, *n* = 233 locations) and 21 June–28 June (S-2; *n* = 8 tagged birds, *n* = 298 locations). To compare with the aerial surveys, we calculated a UD on 19 May (A-1; *n* = 12 tagged birds, *n* = 60 locations) and on 1 and 4 July (A-2; *n* = 8 tagged birds, *n* = 157 locations). Because telemetry data were available for the periods before, between, and after each ship or aerial survey, we calculated separate UD during these periods to determine whether tagged murre distributions were different earlier or later in the season when Eulerian survey data were unavailable. Finally, the distance from shore of satellite-tagged murre locations was tabulated and plotted to compare with offshore distributions of murres observed during ship and aerial surveys.

To compare distributions of murres observed from ship and aerial surveys with the satellite telemetry-derived UD, we created interpolated, continuous-surface density distributions using the *kernel interpolation with barriers* tool in ArcMap 10.3. Kernel density estimation (KDE) is a simple nonparametric statistical technique



that estimates a real-valued function as the weighted average of neighboring observed data (Worton, 1989). The weight is defined by the kernel, such that closer points are given greater weights, and smoothness is set by the kernel bandwidth (Worton, 1989). We used a first-order polynomial kernel function and kernel bandwidths set to the minimum north–south distance between transects to create a smooth prediction surface from ship and aerial transect observations. This is similar to approaches used in other studies of seabird distributions derived from transect survey data (O'Brien, Webb, Brewer, & Reid, 2012; Perrow, Harwood, Skeate, Praca, & Eglington, 2015). To quantify similarities in murre spatial distributions among ship, plane, and telemetry data sets, we calculated percent overlap of concurrent kernel density (KD) and UD surfaces using the *tabulate intersection* tool in ArcMap 10.3. This approach calculates the spatial overlap based on the surface area of each predicted distribution. We limited overlap analyses to the area of each KD surface, thereby excluding UD surfaces from tagged birds that extended beyond the area surveyed during each ship or aerial survey. We compared overlap of full (99%) and 50% (i.e., core use areas) UD distributions that occurred within the full and 50% KD during each ship or aerial survey. We also calculated and plotted the geographic mean center, or center of gravity (CG), of satellite-tagged murre locations and murre locations observed during each ship and aerial survey (Bez & Rivoirard, 2001; Woillez, Poulard, Rivoirard, Petitgas, & Bez, 2007; Woillez, Rivoirard, & Petitgas, 2009), and measured the Euclidean distance between CGs for each survey comparison.

## 3 | RESULTS

### 3.1 | Eulerian sampling

#### 3.1.1 | Ship-based surveys

We counted a total of 428 murre during 43.4 km<sup>2</sup> of survey effort during S-1 and 749 murre during 78.8 km<sup>2</sup> of survey effort during S-2 (Table 1). Murre were found across most of the extent of ship surveys (4.7–44.8 km from shore), with greatest numbers of individuals occurring between 10 and 20 km of shore (Figure 3). Mean densities of murre were not significantly different between S-1 (9.9 murre/km<sup>2</sup>) and S-2 (9.5 murre/km<sup>2</sup>;  $t_{72.3} = -0.076$ ,  $p = 0.940$ ). Murre densities were consistently greatest adjacent to a large murre colony on the Cape Meares (CM) transect and on the Columbia River (CR) transect (Figure 4). During S-1, mean densities of murre on the CM transect (26.0 murre/km<sup>2</sup>) and on the CR transect (13.9 murre/km<sup>2</sup>) were approximately five to nine times greater than the mean density observed on the other three transects (2.8 murre/km<sup>2</sup>). During S-2, mean densities on the CM (30.0 murre/km<sup>2</sup>) and CR (23.3 murre/km<sup>2</sup>) transects were also greater than the other transects. We calculated intermediate murre densities along the central Washington coast near Grays Harbor and Willapa Bay during S-1 and S-2 (2.7–4.2 murre/km<sup>2</sup>). Similar intermediate densities of murre were observed off the northern Washington coast on S-2 (3.2–3.7 murre/km<sup>2</sup>), with the

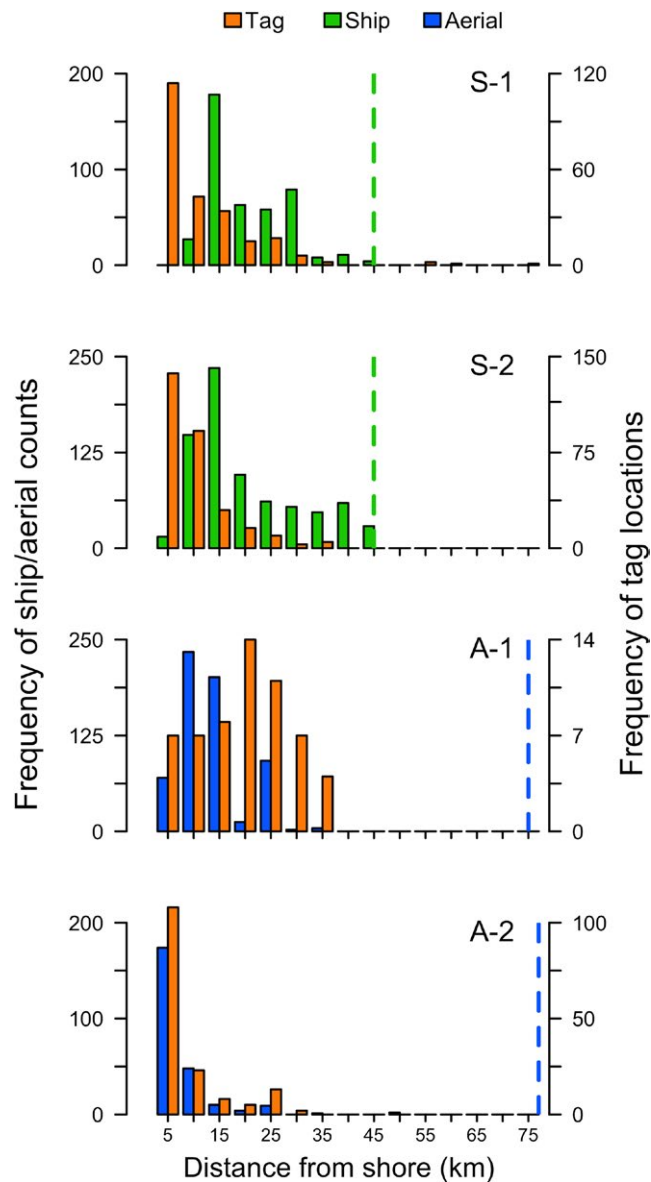
exception of greater mean densities near La Push (10.2 murre/km<sup>2</sup>; Figure 4). Low densities (mean: <3.0 murre/km<sup>2</sup>) occurred in the southernmost portion of the survey area off the central Oregon coast on S-2 (Figure 4).

#### 3.1.2 | Aerial surveys

We counted a total of 618 murre during 45.1 km<sup>2</sup> of survey effort during A-1 and 880 murre during 162.5 km<sup>2</sup> of survey effort during A-2 (Table 1). During aerial surveys, we observed murre between 0.3 and 50 km from shore (Figure 3). The offshore distribution of murre during A-1 was primarily between 5 and 25 km from shore, with greatest numbers of individual murre located 10–15 km from shore. During A-2, most murre occurred within 5 km of shore. Mean densities did not differ between A-1 (13.7 murre/km<sup>2</sup>) and A-2 (5.4 murre/km<sup>2</sup>;  $t_{370.7} = 1.54$ ,  $p = 0.125$ ). During A-1, densities of murre were greatest on the three transects along the northern Oregon coast (mean: 35.1 murre/km<sup>2</sup>), including the transect adjacent to the Tillamook Head murre colony and near the mouth of the Columbia River (Figure 4). Although we observed lesser densities of murre off the southern Washington coast, relatively greater densities (4.4 murre/km<sup>2</sup>) were observed near Grays Harbor during A-1. During A-2, greatest densities (mean: 9.6 murre/km<sup>2</sup>) were also observed on transects off the northern Oregon coast near murre colonies at Tillamook Head and Cape Meares, and least densities (mean: 0.48 murre/km<sup>2</sup>) were observed off the southern Washington coast near Willapa Bay and Grays Harbor (Figure 4).

#### 3.1.3 | Lagrangian sampling

We tracked satellite-tagged murre for an average of 54.2 ± 21.9 days (mean ± SD) between early May and early July. Tracking duration ranged from 18 to 73 days, with 7 of 12 (58%) tags transmitting for ≥63 days (Table 1). Fifty-eight percent ( $n = 7$ ) of tagged murre were female, 33% ( $n = 4$ ) were male, and the sex of one murre could not be determined. Most murre locations occurred within 5–10 km from shore (range: 3–76 km), and were closer to shore than murre observed during ship or aerial surveys, except during A-1 when the number of tag locations was smaller (Figure 3). Tracked murre occupied a vast at-sea area (114,900,000 km<sup>2</sup>; i.e., >2 times the area of the State of California). Overall, the 99% utilization distribution (UD) indicated a broad latitudinal use of nearshore coastal waters between British Columbia and central California (Figure 5). Highest use areas were located off the northern Oregon and southern Washington coasts, and the west coast of southern Vancouver Island, British Columbia, Canada (Figure 5a). We observed some use of waters near a small colony along the central Washington coast (Grenville Arch), as well as Tillamook Head and Cape Meares in northern Oregon, but obvious central place foraging behavior, such as repeated trips to land, was not observed in the tracking data. One male murre traveled ~1,500 km to southern-central California and spent most of its time during the study between Monterey Bay and the Santa Barbara Channel; two female murre moved ~950 km north to the west coast



**FIGURE 3** Distribution of common murre (*Uria aalge*) distances from shore during ship (S-1, S-2) and broad aerial surveys (A-1, A-2), and satellite telemetry tag locations during each ship or aerial survey. Dashed vertical lines indicate the offshore extent of each ship or aerial survey. During A-2, the plane surveyed 94-km offshore, but the x-axis was truncated because no murre were observed more than 75 km from shore

of Canada near the southeast Alaska–British Columbia border (i.e., Celestial Reef in Dixon Entrance). Two additional females and one male murre moved into waters along the west coast of Vancouver Island, while the rest of the tagged murre (three females, two males, and one undetermined sex) remained in Washington and Oregon waters for the duration of tag transmissions.

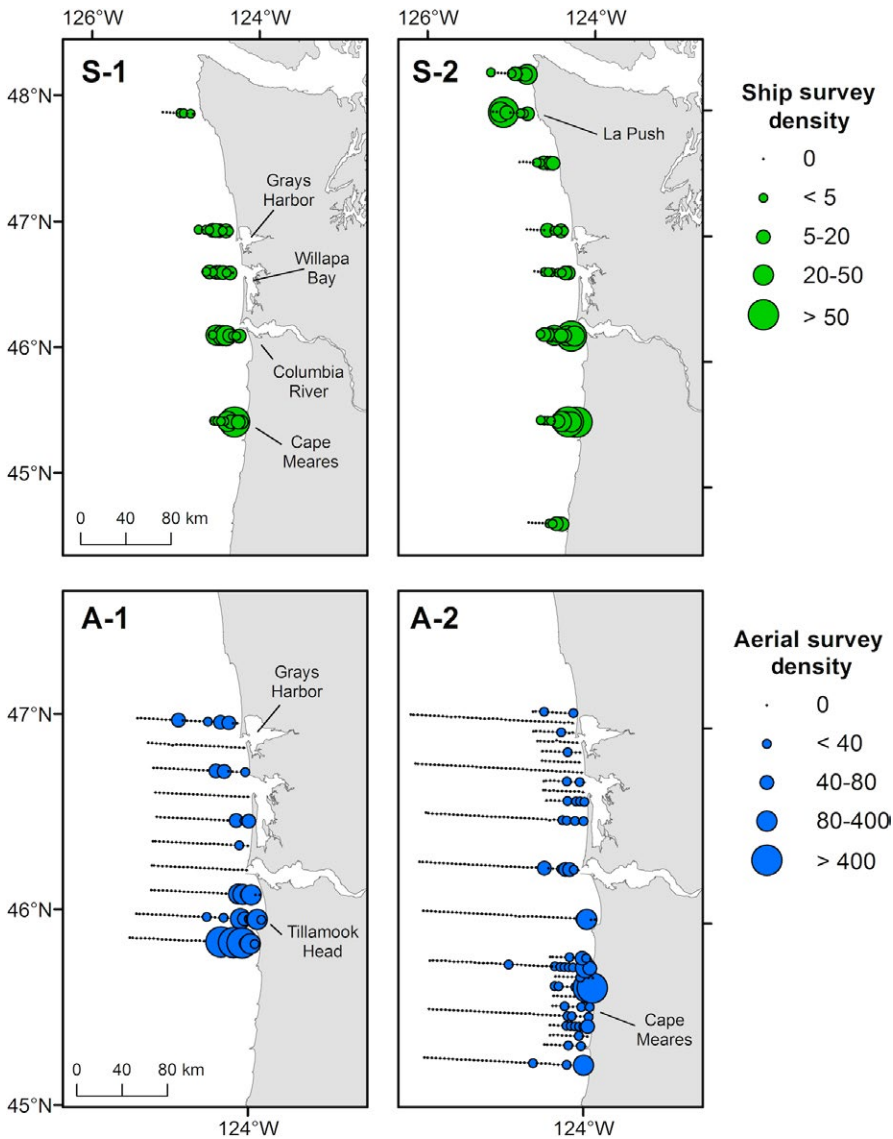
The UD of tagged birds calculated for the period before the first Eulerian survey occurred indicated that satellite-tracked murre exhibited high spatial use of waters along the southern Washington and northern Oregon coast, with a high-use area near the mouth of the Columbia River (Figure 5b). During this time, one male murre flew

south into southern Oregon and northern California waters. With the exception of the murre that moved into California waters, the UD calculated during S-1 indicated that most tagged murre remained aggregated off Grays Harbor and Willapa Bay, and near the mouth of the Columbia River, similar to observations made during the ship survey (Figure 6a, c). Spatial overlap between the 99% UD and the full kernel density surface (KD) during S-1 was 35%, and 25% of core use areas (50% UD and KD) overlapped. The geographic mean centers of gravity (CGs) were 37 km apart. A similar spatial distribution of tagged murre was observed during the 16 days between ship surveys S-1 and S-2, although the UD revealed that some tagged murre shifted north during this period into Canadian waters along the west coast of Vancouver Island (Figure 5c). During S-2, tagged murre were more broadly distributed throughout Washington coastal waters, with greatest spatial use near Grays Harbor (Figure 6b, d). Overlap between the 99% UD and the full KD during S-2 was 30%, and 27% of core use areas overlapped. The CGs were separated by 22 km. Spatial distributions of murre observed during both ship surveys were similar, with most murre observed in northern Oregon waters near Cape Meares; CGs between the two ship surveys were separated by 44 km. After the ship surveys were completed, some tagged murre continued moving north into Canadian waters, with two birds moving as far north as Dixon Entrance (Celestial Reef) near the Alaska–British Columbia, Canada border (Figure 5d). Some murre also remained within Washington coastal waters, although the UD indicated minimal spatial use of this area (Figure 5d).

The UD during A-1 indicated that the majority of tagged murre used waters near Grays Harbor, Willapa Bay, and off the mouth of the Columbia River (Figure 7a). During A-1, 56% of the 99% UD overlapped with the full KD, and 39% of core use areas overlapped. The CGs of murre were separated by 21 km (Figure 7c). Spatial distributions of tagged murre were similar to A-1 during the 41 days between aerial surveys. However, during A-2 tagged murre showed high use of waters along the west coast of Vancouver Island and low use of waters near Grays Harbor (Figure 7b). In comparison, murre densities observed from the plane were greatest farther south on the Oregon coast, and overlap between the UD and KD during A-2 was only 12%, and only 4% of core use areas overlapped (Figure 7d). The distance between the CGs during A-2 was 302 km, reflecting the northward movement of tagged murre and southerly distribution of murre observed during the aerial survey. In comparison, the CGs of murre observed during A-1 and A-2 were only 33 km apart, and located off northern Oregon, similar to the murre CG locations observed during ship surveys. After aerial surveys were completed, locations of tagged murre were widespread in coastal Washington and Canadian waters, primarily along the west coast of southern Vancouver Island and farther north near Dixon Entrance (Celestial Reef; Figure 5d).

## 4 | DISCUSSION

We used concurrent data from ships, planes, and satellite telemetry to illustrate that seabird distributions inferred from independent,



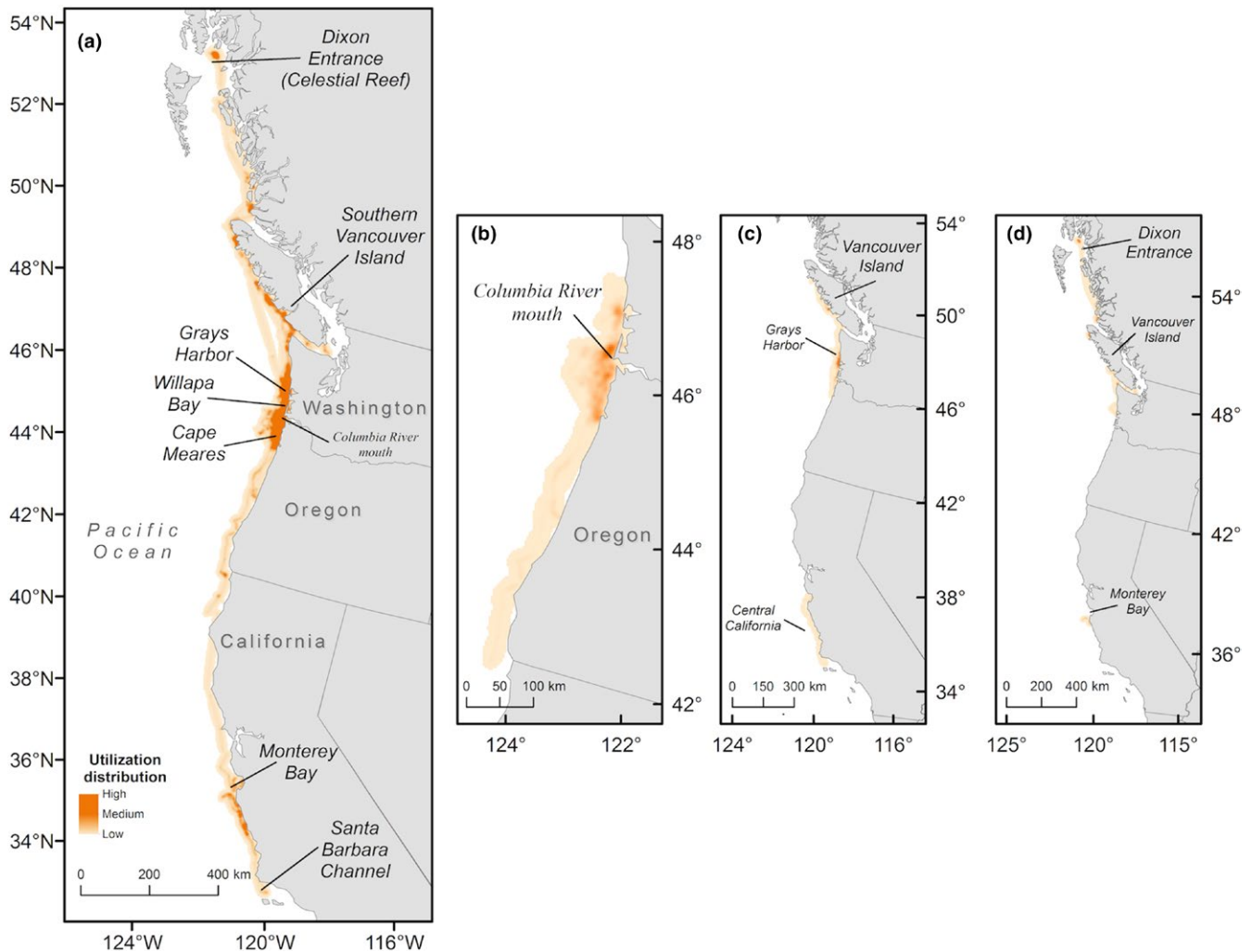
**FIGURE 4** Density distributions of common murre (*Uria aalge*) observed in 2012 during ship surveys (green points; 3-km bins) on 30 May–3 June (S-1) and 21–28 June (S-2), and aerial surveys (blue points; 2.4-km bins) on 19 May (A-1) and 1 and 4 July (A-2) in the northern California Current. Geographic locations identified in the text are labeled

contemporaneous data sets can indicate similar high-use areas, but differences in survey perspective and spatiotemporal extent can influence observed patterns. At the spatiotemporal scale of the northern California Current during May 2012, distributions of murre observed in all three data sets were similar and indicated high use of nearshore waters along the Washington coast, near the mouth of the Columbia River, and in northern Oregon near some of the largest murre colonies along the coast, including Cape Meares and Tillamook Head (Carter et al., 2001; Naughton et al., 2007). This is not surprising given that the study period coincided with the breeding season for murre (April–August), a time when both breeding and nonbreeding murre aggregate on the water near colonies before and after foraging bouts (Ainley, Nettleship, Carter, & Storey, 2002; Zador & Piatt, 1999). Regardless of latitude, all murre occurred primarily within 0–25 km of the coast, with tagged murre locations generally occurring closer (3–5 km) to shore than murre observed during ship surveys, which did not survey in shallow water within ~5 km of shore due to hull draft. Aerial surveys revealed nearshore distributions of murre more

similar to the telemetry data, particularly during A-2. Consistently similar densities of murre observed from ship and aerial surveys during May, June, and July demonstrate that large numbers of murre occupy the northern California Current during spring and summer, and that both Eulerian methods can effectively survey the regional distribution of this relatively large-bodied, coastal seabird (Briggs, Tyler, & Lewis, 1985b; Henkel et al., 2007). Satellite telemetry results during the early part of the study indicated similar spatial distributions of murre across independent data sets, but we documented a broader latitudinal distribution of tagged murre later in the study period as individual birds moved beyond the boundaries of the Eulerian survey transects.

The relatively stable density distributions of murre observed in the ship and aerial survey data contrast with the dynamic distributions observed in the telemetry data and illuminate how different survey perspectives can reveal differing patterns of species' distributions. Estimates of murre distributions observed from all three platforms during May indicated high use of waters in southern Washington and aggregation near colonies in northern Oregon.



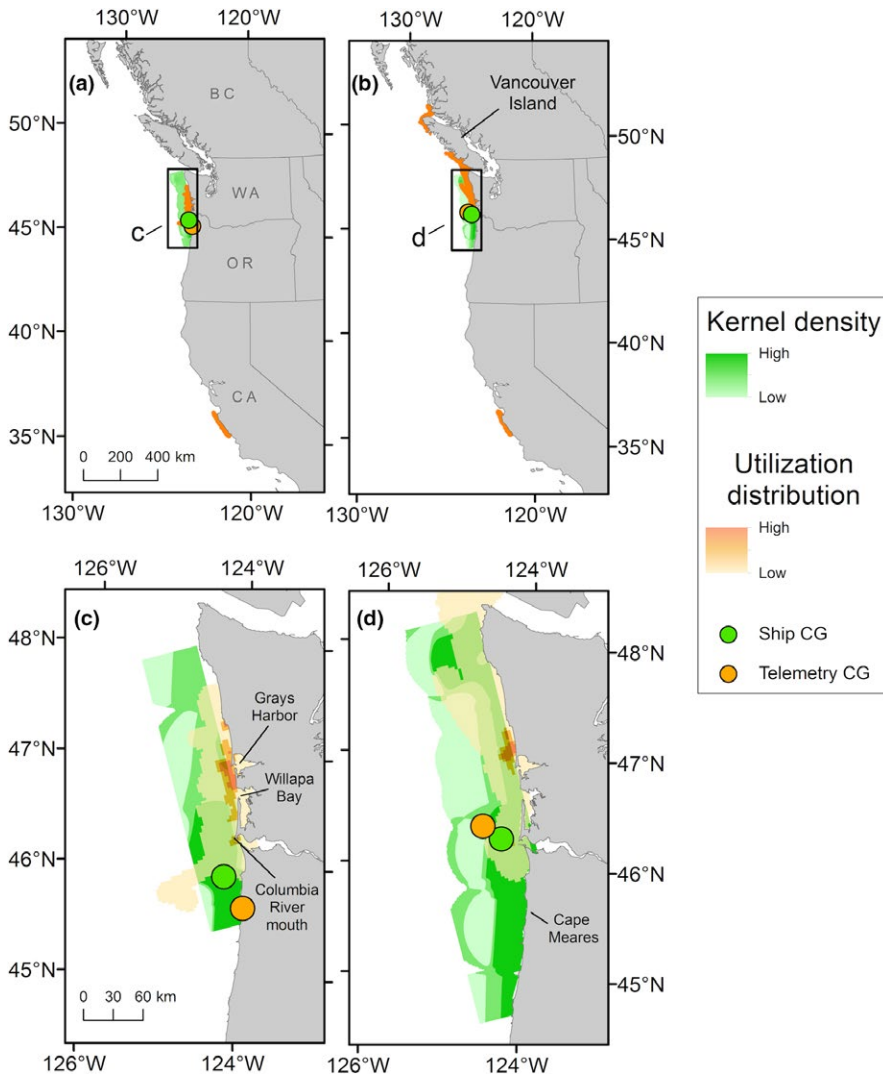


**FIGURE 5** Utilization distribution (99%) of 12 satellite-tagged common murre (*Uria aalge*) observed in 2012 during (a) the full study period (4 May–23 July), (b) before ship or aerial surveys began (4 May–18 May), (c) between ship surveys (4 June–20 June) and (d) after the second aerial survey (A-2) was completed (5 July–23 July). Locations of high-use areas identified in the text are labeled

Relatively high overlap of core utilization distributions and close association of geographic centers of gravity (<40 km apart) suggest that most of the murre observed in each survey data set were collocated in a relatively small region of the northern California Current. One of the highest use areas occurred near the mouth of the Columbia River, which is a productive area that supports a variety of prey fish for seabirds and attracts murre (Litz, Emmett, Bentley, Claiborne, & Barceló, 2013; Phillips et al., 2017). The consistent occurrence of murre aggregations near the mouth of the Columbia River, and the relative ease of capturing murre from the water, is the primary reason that all of the at-sea captures and tag deployments during the study occurred in this area. There are no active murre colonies along the coast between the mouth of the Columbia River and Grays Harbor, so our results suggest that murre observed in this area were breeding birds that commuted at least 60 km north from large colonies in northern Oregon or moved a minimum of 50–100 km south from colonies along the Washington coast (e.g., Bodelteh Islands, Grenville Arch Rock; Thomas & Lyons, 2017). Alternatively, as the

telemetry data suggest, murre observed in this area may not be associated with a colony (i.e., nonbreeders) and therefore able to continually occupy productive waters near the river mouth without returning to coastal colonies.

Although the data from May suggest that common murre in the northern California Current are locally resident, examination of the telemetry data from June and July demonstrates unexpected high mobility among tagged murre, a shift in high-use areas with time, and greater use of distant coastal waters in California and British Columbia later in the study period. Thus, at least a portion of the murre population occupying the Washington and Oregon coasts are transient, with a predominantly northward flux of individuals occurring between May and July. The ship and aerial surveys may have observed a portion of the murre population that are locally resident from May through July, or new transient individuals that moved into the study area as other murre moved out of the area. While the proportion of murre that are resident or transient is unknown, we conclude that the population of murre occupying the northern

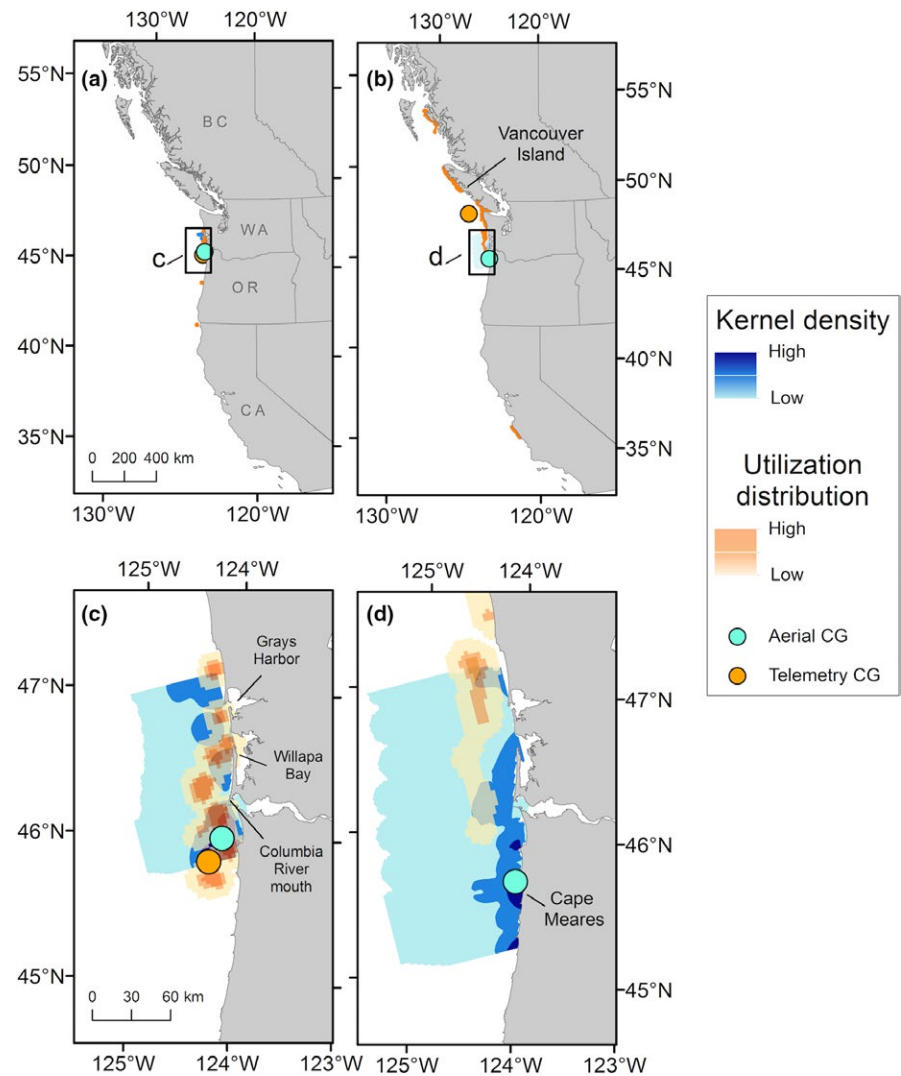


**FIGURE 6** Common murre (*Uria aalge*) density distributions observed during ship surveys (green surface) and satellite telemetry (orange surface) during the same time period in 2012. Distributions of tagged murre observed during S-1 (30 May–3 June) are shown (a) at the broad scale and (c) within the area surveyed by the ship. Distributions of tagged murre observed during S-2 (21 June–28 June) are shown (b) at the broad scale and (d) within the area surveyed by the ship. The corresponding geographic centers of gravity (CG) are shown as green or orange points. Locations identified in the text are labeled

California Current likely consists of a mixed group of central place foraging adults and vagile, nonbreeding adults and subadults that differ in their occupancy and use of the California Current. The murre tracked in this study may have been young birds, nonbreeding adults, or failed breeders because these groups exhibit greater dispersals away from colonies than breeding birds (Hatch et al., 2000). Alternatively, the unexpected mobility could indicate that tagging caused individuals to change their movement and/or breeding behavior (see Phillips et al., 2018 for a discussion).

While regional densities of murre observed from ship and aerial surveys were similar during the study period, and the surveys were relatively synchronous, the differences in survey timing and spatial resolution may explain fine-scale disparities in spatial patterns (van Franeker, 1994; Ronconi & Burger, 2009; Ryan & Cooper, 1989). Ship surveys were designed to sample the entire coast from central Oregon to northern Washington, and transects were separated by 35–90 km, which allowed for observations of murre densities across a wider range of the northern California Current but also may have obscured higher-resolution variability. In comparison, the aerial surveys were more limited in their overall latitudinal extent

but the greater number of more closely spaced transects, especially the focal-area surveys which were only 6 km apart, may have captured higher-resolution variability in hourly and daily murre distributions than in ship surveys. Murre are known to aggregate near convergent fronts formed along the boundary between fresh and saltwater near the mouth of the Columbia River (Phillips et al., 2018), where prey fish distributions are also concentrated (Litz et al., 2013; Phillips et al., 2017). Variation in Columbia River plume circulation and the formation of convergent fronts occur at temporal periods of hours to days (Jay, Pan, Orton, & Horner-Devine, 2009; Jay, Zaron, & Pan, 2010), which is often not detectable at the sampling resolution of the ship surveys. Aerial surveys may have occurred during a period when Columbia River plume circulation or prey distributions caused lower densities of murre near Willapa Bay and Grays Harbor compared with oceanographic conditions when the ship surveys occurred. Based on our observations, the area to the north and south of the Columbia River mouth is a high-use area for all murre in this study, although fine-scale variation in distributions and changes throughout the study period suggest that different groups of birds may use this habitat differently during the spring and summer.



**FIGURE 7** Common murre (*Uria aalge*) density distributions observed during aerial surveys (blue surface) and satellite telemetry (orange surface) during the same time period in 2012. Distributions of tagged murrees observed during A-1 (19 May) are shown (a) at the broad scale, and (c) within the area surveyed by the plane. Distributions of tagged murrees observed during A-2 (1 and 4 July) are shown (b) at the broad scale and (d) within the area surveyed by the plane. The corresponding geographic centers of gravity (CG) are shown as blue or orange points. The CG of tagged murrees observed during A-2 was 302 km north of the CG of murrees observed from the plane and only shown on the broad scale map. Geographic locations identified in the text are labeled

The use of satellite telemetry in this study offered the opportunity to expand the spatial extent and resolution by recording near-continuous information about each individual murre's location, thus eliminating sampling constraints imposed by predetermined ship or aerial transects. This enabled us to demonstrate that the spatial extent of individual murrees during the breeding season can encompass nearshore waters of California, Oregon, Washington, and British Columbia, essentially the full range of the California Current. Tags transmitted continuously for an average of two months between May and July, so we attained more continuous sampling of murre distributions compared with discrete ship and aerial surveys. Because we captured birds at sea, rather than at a colony, breeding status prior to tagging is unknown. There were no major differences in the sex ratio of tagged murrees and their movement patterns, suggesting a somewhat random sample, but whether murrees segregate at sea in relation to age or breeding status, or colony of origin, is unknown. Future research on this topic would provide important insight on murre conservation and management in the northern California Current (Thomas & Lyons, 2017). Tracking a small number of individuals can lead to large variability in observed habitat use (Fossette

et al., 2014; Hays et al., 2016; Lindberg & Walker, 2007), and sample size may have also influenced the observed results. Of the 12 murrees tagged, one flew to California, and five flew to Canada. To better understand the spatial and temporal extent of tagged animal distributions, Lindberg and Walker (2007) used simulations to estimate that at least 20–30 tagged individuals were necessary to reveal population patterns. Increasing the number of individual tag deployments may also provide better insight into comparability of different perspectives and platforms used to evaluate animal distributions.

The choice of survey perspective, platform, and spatiotemporal extent can be influenced by study objectives, accessibility of the area, sampling logistics, and available resources (Ainley et al., 2012). This research demonstrates that for surveys where objectives include obtaining accurate population abundance estimates and spatial use of coastal areas by a large-bodied seabird, an Eulerian perspective using either ship or aerial survey methods produces similar results, although the spatial extent of survey transects can limit inferences on a population's full spatial extent (see also Briggs et al., 1985b; Henkel et al., 2007). Ships are ideal platforms to sample concurrent abiotic and/or biotic parameters such as sea surface

temperature and prey density (Ainley et al., 2012), and therefore offer potentially more opportunities for ecological studies. Aerial surveys, however, can accomplish a survey in a much smaller amount of time, are not as limited by sea surface conditions and ocean depth, and may capture higher-resolution variation in density distributions. In comparison, a Lagrangian perspective using satellite telemetry enables a much larger spatiotemporal sampling range compared to Eulerian surveys, allowing for a more extensive analysis of habitat use throughout a seabird's potential range. However, these results demonstrate that data from satellite telemetry of birds captured and tagged at sea may not be representative of the full population of interest (Priddel et al., 2014), and space use may not be necessarily related to actual density at sea (Ainley et al., 2012). By collecting and comparing concurrent data from three independent platforms, we obtained a more comprehensive understanding of the distribution of the murre population during the breeding season in the northern California Current, including connectivity to populations in British Columbia and California.

While ship-based, aerial, and telemetry surveys can provide complementary information on species distributions, the results of this study indicate that a thorough assessment of the spatial extent and synopticity of relevant data is an important first step before integrating methodological perspectives. Depending on a study's objective, the spatiotemporal mismatch between independent data sets may bias observed species' distributions and relationships to habitat features. For example, a comparison between a ship-based (Santora, Ralston, & Sydeman, 2011) and a telemetry-based (Adams et al., 2012) survey of sooty shearwater (*Ardenna grisea*) distributions off the central California coast demonstrated that shearwaters observed from either perspective use the same general habitat, but that tagged birds were concentrated nearshore where larger vessels could not survey (c.f., Watanuki et al., 2016). Whether observed differences in spatial distributions of shearwaters were related solely to a spatiotemporal mismatch in sampling coverage, or possibly to differential habitat use or prey availability, remains unknown.

Efforts to combine Eulerian and Lagrangian perspectives using seabird counts within quantitative models have been conducted (Hyrenbach, Keiper, Allen, Ainley, & Anderson, 2006; Louzao et al., 2009; Yamamoto et al., 2015), and methods continue to be refined (Watanuki et al., 2016). Development of separate habitat models using data from each sampling perspective, and then comparing and integrating results across models, presents a powerful tool to quantify factors influencing marine mammal and seabird distributions and habitat use (Watanuki et al., 2016). This integrative approach has facilitated ongoing efforts to identify and delineate marine protected areas for multiple mobile marine predators (Ballard, Jongsomjit, Veloz, & Ainley, 2012; Camphuysen, Shamoun-Baranes, Bouten, & Garthe, 2012; Perrow et al., 2015), as well as dynamic ocean management approaches (Hazen et al., 2016, 2018; Maxwell et al., 2015). This type of habitat modeling could be a useful next step for the data presented here, especially in combination with Eulerian survey data from areas used by

tagged murres in California and British Columbia to provide a comprehensive analysis of common murre spatial distributions along the west coast. Given the expected growth of telemetry studies (Hart & Hyrenbach, 2009) and efforts to integrate independent data sets (Watanuki et al., 2016), our results serve as a case study on how sampling perspective and choice of platform can influence spatiotemporal observations of species distributions.

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## CONFLICT OF INTEREST

None declared.

## AUTHOR CONTRIBUTIONS

This study was conceived and designed by EMP, JA, and JEZ. Data were collected by EMP, JA, JEZ, and JJF. EMP conducted data analyses with assistance from JA, JEZ, JJF, and JKH. JA, JEZ, JJF, and JKH contributed to writing the manuscript with EMP. EMP edited the final manuscript, and all authors approve of its submission.

## DATA ACCESSIBILITY

Data supporting this manuscript have been deposited in Dryad <https://doi.org/10.5061/dryad.1hg2n56>.

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## REFERENCES

- Adams, J., Felis, J., Mason, J. W., & Takekawa, J. Y. (2014). *Pacific Continental Shelf Environmental Assessment (PaCSEA): aerial seabird and marine mammal surveys off northern California, Oregon, and Washington, 2011–2012* (p. 266). Camarillo, CA: U.S. Dept. of the Interior, Bureau of Ocean Energy Management, Pacific OCS Region.
- Adams, J., & Flora, S. (2010). Correlating seabird movements with ocean winds: Linking satellite telemetry with ocean scatterometry. *Marine Biology*, *157*, 915–929. <https://doi.org/10.1007/s00227-009-1367-y>
- Adams, J., MacLeod, C., Suryan, R. M., Hyrenbach, K. D., & Harvey, J. T. (2012). Summer-time use of west coast US National Marine Sanctuaries by migrating sooty shearwaters (*Puffinus griseus*). *Biological Conservation*, *156*, 105–116. <https://doi.org/10.1016/j.biocon.2011.12.032>
- Ainley, D. G., Dugger, K. D., Ford, R. G., Pierce, S. D., Reese, D. C., Brodeur, R. D., ... Barth, J. A. (2009). Association of predators and prey at frontal features in the California Current: Competition, facilitation, and co-occurrence. *Marine Ecology Progress Series*, *389*, 271–294. <https://doi.org/10.3354/meps08153>
- Ainley, D. G., Dugger, K. M., Toniolo, V., & Gaffney, I. (2007). Cetacean occurrence patterns in the Amundsen and Southern Bellingshausen Sea Sector, Southern Ocean. *Marine Mammal Science*, *23*, 287–305. <https://doi.org/10.1111/j.1748-7692.2007.00109.x>
- Ainley, D. G., Nettleship, D. N., Carter, H. R., & Storey, A. E. (2002). Common Murre (*Uria aalge*). The Birds of North America Online.
- Ainley, D. G., Ribic, C. A., & Woehler, E. J. (2012). Adding the ocean to the study of seabirds: A brief history of at-sea seabird research. *Marine Ecology Progress Series*, *451*, 231–243. <https://doi.org/10.3354/meps09524>
- Ainley, D. G., Spear, L. B., Tynan, C. T., Barth, J. A., Pierce, S. D., Glenn Ford, R., & Cowles, T. J. (2005). Physical and biological variables affecting seabird distributions during the upwelling season of the northern California Current. *Deep Sea Research Part II: Topical Studies in Oceanography*, *52*, 123–143. <https://doi.org/10.1016/j.dsr2.2004.08.016>
- Ballance, L. T., & Pitman, R. L. (1998). Cetaceans of the western tropical Indian Ocean: Distribution, relative abundance, and comparisons with cetacean communities of two other tropical ecosystems. *Marine Mammal Science*, *14*, 429–459. <https://doi.org/10.1111/j.1748-7692.1998.tb00736.x>
- Ballance, L. T., Pitman, R. L., & Reilly, S. B. (1997). Seabird community structure along a productivity gradient: Importance of competition and energetic constraint. *Ecology*, *78*, 1502–1518. [https://doi.org/10.1890/0012-9658\(1997\)078\[1502:SCSAAP\]2.0.CO;2](https://doi.org/10.1890/0012-9658(1997)078[1502:SCSAAP]2.0.CO;2)
- Ballard, G., Jongsomjit, D., Veloz, S. D., & Ainley, D. G. (2012). Coexistence of mesopredators in an intact polar ocean ecosystem: The basis for defining a Ross Sea marine protected area. *Biological Conservation*, *156*, 72–82. <https://doi.org/10.1016/j.biocon.2011.11.017>
- Barbraud, C., & Thiebot, J.-B. (2009). On the importance of estimating detection probabilities from at-sea surveys of flying seabirds. *Journal of Avian Biology*, *40*, 584–590. <https://doi.org/10.1111/j.1600-048X.2009.04653.x>
- Bez, N., & Rivoirard, J. (2001). Transitive geostatistics to characterise spatial aggregations with diffuse limits: An application on mackerel ichthyoplankton. *Fisheries Research*, *50*, 41–58. [https://doi.org/10.1016/S0165-7836\(00\)00241-1](https://doi.org/10.1016/S0165-7836(00)00241-1)
- Block, B. A., Costa, D. P., Boehlert, G. W., & Kochevar, R. E. (2002). Revealing pelagic habitat use: The tagging of Pacific pelagics program. *Oceanologica Acta*, *25*, 255–266. [https://doi.org/10.1016/S0399-1784\(02\)01212-4](https://doi.org/10.1016/S0399-1784(02)01212-4)
- Block, B. A., Holbrook, C. M., Simmons, S. E., Holland, K. N., Ault, J. S., Costa, D. P., ... Kochevar, R. E. (2016). Toward a national animal telemetry network for aquatic observations in the United States. *Animal Biotelemetry*, *4*, 6. <https://doi.org/10.1186/s40317-015-0092-1>
- Briggs, K. T., Tyler, W. B., & Lewis, D. B. (1985a). Aerial surveys for seabirds: Methodological experiments. *The Journal of Wildlife Management*, *49*, 412–417.
- Briggs, K. T., Tyler, W. B., & Lewis, D. B. (1985b). Comparison of ship and aerial surveys of birds at sea. *The Journal of Wildlife Management*, *49*, 405–411.
- Briggs, K. T., Tyler, W. B., Lewis, D. B., & Carlson, D. R. (1987). Bird communities at sea off California, 1975–1983. *Studies in Avian Biology*, *11*, 1–74.
- Brodeur, R. D., Myers, K. W., & Helle, J. H. (2003). Research conducted by the United States on the early ocean life history of Pacific salmon. *North Pacific Anadromous Fish Commission Bulletin*, *3*, 89–131.
- Brown, R. G. B. (1980). Seabirds as marine animals. In J. Burger, B. L. Olla, & H. E. Winn (Eds.), *Behavior of marine animals* (pp. 1–39). Berlin, Germany: Springer US.
- Buckland, S. T., Anderson, D. R., Burnham, K. P., Laake, J. L., Borchers, D. L., & Thomas, L. (2001). *Introduction to distance sampling: estimating abundance of biological populations*, 1st ed. New York, NY: Oxford University Press, Oxford.
- Burger, A. E., & Shaffer, S. A. (2008). Application of tracking and data-logging technology in research and conservation of seabirds. *The Auk*, *125*, 253–264. <https://doi.org/10.1525/auk.2008.1408>
- Calenge, C. (2006). The package “adehabitat” for the R software: A tool for the analysis of space and habitat use by animals. *Ecological Modelling*, *197*, 516–519. <https://doi.org/10.1016/j.ecolmodel.2006.03.017>
- Camphuysen, K. (C. J.), Shamoun-Baranes, J., Bouten, W., & Garthe, S. (2012). Identifying ecologically important marine areas for seabirds using behavioural information in combination with distribution patterns. *Biological Conservation*, *156*, 22–29. <https://doi.org/10.1016/j.biocon.2011.12.024>
- Carter, H. R., Wilson, U. W., Lowe, R. W., Rodway, M. S., Manuwal, D. A., Takekawa, J. E., & Yee, J. L. (2001). Population trends of the common murre (*Uria aalge californica*). *Biology and Conservation of the Common Murre in California, Oregon, Washington, and British Columbia* (pp. 33–132). Washington, DC: US Department of the Interior, US Geological Survey, Biological Resources Division USGS/BRD/ITR-2000-0012.
- Certain, G., & Bretagnolle, V. (2008). Monitoring seabirds population in marine ecosystem: The use of strip-transect aerial surveys. *Remote Sensing of Environment*, *112*, 3314–3322. <https://doi.org/10.1016/j.rse.2008.01.019>
- CLS (Collecte Localisation Satellites). (2013). ARGOS. Retrieved from <http://www.argos-system.org/>. <http://www.argos-system.org/>
- Coyne, M. S., & Godley, B. J. (2005). Satellite tracking and analysis tool (STAT): An integrated system for archiving, analyzing and mapping animal tracking data. *Marine Ecology Progress Series*, *301*, 1–7. <https://doi.org/10.3354/meps301001>
- Davoren, G. K., Montevecchi, W. A., & Anderson, J. T. (2003). Distributional patterns of a marine bird and its prey: Habitat selection based on prey and conspecific behaviour. *Marine Ecology Progress Series*, *256*, 229–242. <https://doi.org/10.3354/meps256229>
- Decker, M. B., & Hunt, G. L. (1996). Foraging by murre (*Uria* spp.) at tidal fronts surrounding the Pribilof Islands, Alaska, USA. *Marine Ecology Progress Series*, *139*, 1–10. <https://doi.org/10.3354/meps139001>
- Drew, G. S., Piatt, J. F., & Renner, M. (2015). User's guide to the North Pacific Pelagic Seabird Database 2.0: U.S. Geological Survey Open-File Report 2015–1123, p. 52. <http://dx.doi.org/10.3133/ofr20151123>.
- Elith, J., & Leathwick, J. R. (2009). Species distribution models: Ecological explanation and prediction across space and time. *Annual Review of Ecology, Evolution, and Systematics*, *40*, 677–697. <https://doi.org/10.1146/annurev.ecolsys.110308.120159>
- Fiedler, P. C., Reilly, S. B., Hewitt, R. P., Demer, D., Philbrick, V. A., Smith, S., ... Mate, B. R. (1998). Blue whale habitat and prey in the California Channel Islands. *Deep Sea Research Part II: Topical*



- Studies in Oceanography*, 45, 1781–1801. [https://doi.org/10.1016/S0967-0645\(98\)80017-9](https://doi.org/10.1016/S0967-0645(98)80017-9)
- Fossette, S., Witt, M. J., Miller, P., Nalovic, M. A., Albareda, D., Almeida, A. P., ... Godley, B. J. (2014). Pan-Atlantic analysis of the overlap of a highly migratory species, the leatherback turtle, with pelagic long-line fisheries. *Proceedings of the Royal Society of London B: Biological Sciences*, 281, 20133065. <https://doi.org/10.1098/rspb.2013.3065>
- Freitas, C., Lydersen, C., Fedak, M. A., & Kovacs, K. M. (2008). A simple new algorithm to filter marine mammal Argos locations. *Marine Mammal Science*, 24, 315–325. <https://doi.org/10.1111/j.1748-7692.2007.00180.x>
- Fujioka, E., Kot, C. Y., Wallace, B. P., Best, B. D., Moxley, J., Cleary, J., ... Halpin, P. N. (2014). Data integration for conservation: Leveraging multiple data types to advance ecological assessments and habitat modeling for marine megavertebrates using OBIS-SEAMAP. *Ecological Informatics*, 20, 13–26. <https://doi.org/10.1016/j.ecoinf.2014.01.003>
- Gutow, S. E., Leonard, M. L., Conners, M. G., Shaffer, S. A., & Jonsen, I. D. (2015). Individual-level variation and higher-level interpretations of space use in wide-ranging species: An albatross case study of sampling effects. *Frontiers in Marine Science*, 2, <https://doi.org/10.3389/fmars.2015.00093>
- Hart, K. M., & Hyrenbach, K. D. (2009). Satellite telemetry of marine megavertebrates: The coming of age of an experimental science. *Endangered Species Research*, 10, 9–20. <https://doi.org/10.3354/esr00238>
- Hatch, S. A., Meyers, P. M., Mulcahy, D. M., & Douglas, D. C. (2000). Seasonal movements and pelagic habitat use of murrelets and puffins determined by satellite telemetry. *The Condor*, 102, 145–154. [https://doi.org/10.1650/0010-5422\(2000\)102\[0145:SMAPHU\]2.0.CO;2](https://doi.org/10.1650/0010-5422(2000)102[0145:SMAPHU]2.0.CO;2)
- Hays, G. C., Ferreira, L. C., Sequeira, A. M. M., Meekan, M. G., Duarte, C. M., Bailey, H., ... Thums, M. (2016). Key questions in marine megafauna movement ecology. *Trends in Ecology & Evolution*, 31, 463–475. <https://doi.org/10.1016/j.tree.2016.02.015>
- Hazen, E. L., Palacios, D. M., Forney, K. A., Howell, E. A., Becker, E., Hoover, A. L., ... Bailey, H. (2016). WhaleWatch: A dynamic management tool for predicting blue whale density in the California Current. *Journal of Applied Ecology*, <https://doi.org/10.1111/1365-2664.12820>
- Hazen, E. L., Scales, K. L., Maxwell, S. M., Briscoe, D. K., Welch, H., Bograd, S. J., ... Lewison, R. L. (2018). A dynamic ocean management tool to reduce bycatch and support sustainable fisheries. *Science Advances*, 4, eaar3001. <https://doi.org/10.1126/sciadv.aar3001>
- Hedd, A., Montevecchi, W. A., Phillips, R. A., & Fifield, D. A. (2014). Seasonal sexual segregation by monomorphic sooty shearwaters *Puffinus griseus* reflects different reproductive roles during the pre-laying period. *PLOS ONE*, 9, e85572. <https://doi.org/10.1371/journal.pone.0085572>
- Henkel, L. A., Ford, R. G., Tyler, W. B., & Davis, J. N. (2007). Comparison of aerial and boat-based survey methods for marbled murrelets *Brachyramphus marmoratus* and other marine birds. *Marine Ornithology*, 35, 145–151.
- Hodgson, J. C., Baylis, S. M., Mott, R., Herrod, A., & Clarke, R. H. (2016). Precision wildlife monitoring using unmanned aerial vehicles. *Scientific Reports*, 6, 22574. <https://doi.org/10.1038/srep22574>
- Hooker, S. K., Biuw, M., McConnell, B. J., Miller, P. J. O., & Sparling, C. E. (2007). Bio-logging science: Logging and relaying physical and biological data using animal-attached tags. *Deep Sea Research Part II: Topical Studies in Oceanography*, 54, 177–182. <https://doi.org/10.1016/j.dsr2.2007.01.001>
- Horne, J. S., Garton, E. O., Krone, S. M., & Lewis, J. S. (2007). Analyzing animal movements using Brownian bridges. *Ecology*, 88, 2354–2363. <https://doi.org/10.1890/06-0957.1>
- Hunt, G. L., Renner, M., Kuletz, K. J., Salo, S., Eisner, L., Ressler, P. H., ... Santora, J. A. (2018). Timing of sea-ice retreat affects the distribution of seabirds and their prey in the southeastern Bering Sea. *Marine Ecology Progress Series*, 593, 209–230. <https://doi.org/10.3354/meps12383>
- Hyrenbach, K. D., Keiper, C., Allen, S. G., Ainley, D. G., & Anderson, D. J. (2006). Use of marine sanctuaries by far-ranging predators: Commuting flights to the California Current System by breeding Hawaiian albatrosses. *Fisheries Oceanography*, 15, 95–103. <https://doi.org/10.1111/j.1365-2419.2005.00350.x>
- Jay, D. A., Pan, J., Orton, P. M., & Horner-Devine, A. R. (2009). Asymmetry of Columbia River tidal plume fronts. *Journal of Marine Systems*, 78, 442–459. <https://doi.org/10.1016/j.jmarsys.2008.11.015>
- Jay, D. A., Zaron, E. D., & Pan, J. Y. (2010). Initial expansion of the Columbia River tidal plume: Theory and remote sensing observations. *Journal of Geophysical Research-Oceans*, 115, C00B15. <https://doi.org/10.1029/2008jc004996>
- Keiper, C. A., Ainley, D. G., Allen, S. G., & Harvey, J. T. (2005). Marine mammal occurrence and ocean climate off central California, 1986 to 1994 and 1997 to 1999. *Marine Ecology Progress Series*, 289, 285–306. <https://doi.org/10.3354/meps289285>
- Krietsch, J., Hahn, S., Kopp, M., Phillips, R., Peter, H., & Lisovski, S. (2017). Consistent variation in individual migration strategies of brown skuas. *Marine Ecology Progress Series*, 578, 213–225. <https://doi.org/10.3354/meps11932>
- Lindberg, M. S., & Walker, J. (2007). Satellite telemetry in avian research and management: Sample size considerations. *The Journal of Wildlife Management*, 71, 1002–1009. <https://doi.org/10.2193/2005-696>
- Litz, M. N. C., Emmett, R. L., Bentley, P. J., Claiborne, A. M., & Barceló, C. (2013). Biotic and abiotic factors influencing forage fish and pelagic nekton community in the Columbia River plume (USA) throughout the upwelling season 1999–2009. *ICES Journal of Marine Science*, 71, 5–18. <https://doi.org/10.1093/icesjms/fst082>
- Lobo, J. M., Jiménez-Valverde, A., & Hortal, J. (2010). The uncertain nature of absences and their importance in species distribution modelling. *Ecography*, 33, 103–114. <https://doi.org/10.1111/j.1600-0587.2009.06039.x>
- Louzao, M., Bécarea, J., Rodríguez, B., Hyrenbach, K. D., Ruiz, A., & Arcos, J. M. (2009). Combining vessel-based surveys and tracking data to identify key marine areas for seabirds. *Marine Ecology Progress Series*, 391, 183–197. <https://doi.org/10.3354/meps08124>
- MacArthur, R. H. (1972). *Geographical ecology: Patterns in the distribution of species*, 1st ed. New York, NY: Harper & Row.
- Mason, J. W., McChesney, G. J., McIver, W. R., Carter, H. R., Takekawa, J. Y., Gollightly, R. T., ... McCrary, M. D. (2007). At-sea distribution and abundance of seabirds off southern California: A 20-year comparison. *Studies in Avian Biology*, 33, 1–112.
- Maxwell, S. M., Hazen, E. L., Lewison, R. L., Dunn, D. C., Bailey, H., Bograd, S. J., ... Crowder, L. B. (2015). Dynamic ocean management: Defining and conceptualizing real-time management of the ocean. *Marine Policy*, 58, 42–50. <https://doi.org/10.1016/j.marpol.2015.03.014>
- Monk, J. (2014). How long should we ignore imperfect detection of species in the marine environment when modelling their distribution? *Fish and Fisheries*, 15, 352–358. <https://doi.org/10.1111/faf.12039>
- Murphy, R. C. (1936). *Oceanic birds of South America*. New York, NY: American Museum of Natural History.
- Naughton, M. B., Pitkin, D. J., Lowe, R. W., So, K. J., & Strong, C. S. (2007). Catalog of Oregon seabird colonies. USFWS Biological Technical Publication BTP-R1009-2007.
- O'Brien, S. H., Webb, A., Brewer, M. J., & Reid, J. B. (2012). Use of kernel density estimation and maximum curvature to set Marine Protected Area boundaries: Identifying a Special Protection Area for wintering red-throated divers in the UK. *Biological Conservation*, 156, 15–21. <https://doi.org/10.1016/j.biocon.2011.12.033>
- Perrow, M. R., Harwood, A. J. P., Skeate, E. R., Praca, E., & Eglington, S. M. (2015). Use of multiple data sources and analytical approaches to derive a marine protected area for a breeding seabird.

- Biological Conservation*, 191, 729–738. <https://doi.org/10.1016/j.biocon.2015.08.031>
- Phillips, E. M., Horne, J. K., Adams, J., & Zamon, J. E. (2018). Selective occupancy of a persistent yet variable coastal river plume by two seabird species. *Marine Ecology Progress Series*, 594, 245–261. <https://doi.org/10.3354/meps12534>
- Phillips, E. M., Horne, J. K., & Zamon, J. E. (2017). Predator–prey interactions influenced by a dynamic river plume. *Canadian Journal of Fisheries and Aquatic Sciences*, 74, 1375–1390. <https://doi.org/10.1139/cjfas-2016-0302>
- Phillips, S. J., Dudík, M., Elith, J., Graham, C. H., Lehmann, A., Leathwick, J., & Ferrier, S. (2009). Sample selection bias and presence-only distribution models: Implications for background and pseudo-absence data. *Ecological Applications*, 19, 181–197. <https://doi.org/10.1890/07-2153.1>
- Priddel, D., Carlile, N., Portelli, D., Kim, Y., O'Neill, L., Bretagnolle, V., ... Rayner, M. J. (2014). Pelagic distribution of Gould's Petrel (*Pterodroma leucoptera*): Linking shipboard and onshore observations with remote-tracking data. *Emu*, 114, 360–370. <http://dx.doi.org/10.1071/MU14021>
- Ronconi, R. A., & Burger, A. E. (2009). Estimating seabird densities from vessel transects: Distance sampling and implications for strip transects. *Aquatic Biology*, 4, 297–309. <https://doi.org/10.3354/ab00112>
- Ropert-Coudert, Y., & Wilson, R. P. (2005). Trends and perspectives in animal-attached remote sensing. *Frontiers in Ecology and the Environment*, 3, 437–444. [https://doi.org/10.1890/1540-9295\(2005\)003\[0437:TAPIAR\]2.0.CO;2](https://doi.org/10.1890/1540-9295(2005)003[0437:TAPIAR]2.0.CO;2)
- Rutz, C., & Hays, G. C. (2009). New frontiers in biologging science. *Biology Letters*, <https://doi.org/10.1098/rsbl.2009.0089>
- Ryan, P. G., & Cooper, J. (1989). Observer precision and bird conspicuousness during counts of birds at sea. *South African Journal of Marine Science*, 8, 271–276. <https://doi.org/10.2989/02577618909504567>
- Santora, J. A., Ralston, S., & Sydeman, W. J. (2011). Spatial organization of krill and seabirds in the central California Current. *ICES Journal of Marine Science*, 68, 1391–1402. <https://doi.org/10.1093/icesjms/fsr046>
- Shaffer, S. A., Tremblay, Y., Weimerskirch, H., Scott, D., Thompson, D. R., Sagar, P. M., ... Costa, D. P. (2006). Migratory shearwaters integrate oceanic resources across the Pacific Ocean in an endless summer. *Proceedings of the National Academy of Sciences*, 103, 12799–12802. <https://doi.org/10.1073/pnas.0603715103>
- Speich, S. M., & Wahl, T. R. (1989). Catalog of Washington seabird colonies. *U.S. Fish and Wildlife Services Biological Report*, 88(6), 510.
- Tasker, M. L., Jones, P. H., Dixon, T., & Blake, B. F. (1984). Counting seabirds at sea from ships: A review of methods employed and a suggestion for a standardized approach. *The Auk*, 101, 567–577.
- Thomas, S. M., & Lyons, J. E. (2017). Population trends and distribution of common murre *Uria aalge* colonies in Washington, 1996–2015. *Marine Ornithology*, 45, 95–102.
- Tremblay, Y., Bertrand, S., Henry, R. W., Kappes, M. A., Costa, D. P., & Shaffer, S. A. (2009). Analytical approaches to investigating seabird–environment interactions: A review. *Marine Ecology Progress Series*, 391, 153–163. <https://doi.org/10.3354/meps08146>
- van Franeker, J. A. (1994). A comparison of methods for counting seabirds at sea in the Southern Ocean. *Journal of Field Ornithology*, 65, 96–108.
- Votier, S. C., Bicknell, A., Cox, S. L., Scales, K. L., & Patrick, S. C. (2013). A bird's eye view of discard reforms: Bird-borne cameras reveal seabird/fishery interactions. *PLoS ONE*, 8, e57376. <https://doi.org/10.1371/journal.pone.0057376>
- Wahl, T. R., & Heinemann, D. (1979). Seabirds and fishing vessels – co-occurrence and attraction. *Condor*, 81, 390–396. <https://doi.org/10.2307/1366965>
- Watanuki Y., R. M. Suryan, H. Sasaki, T. Yamamoto, E. L. Hazen, M. Renner, ...W. J. Sydeman (Eds.) (2016). Spatial ecology of marine top predators in the North Pacific: Tools for integrating across datasets and identifying high use areas. PICES Sci. Rep. No. 50:55 pp.
- Weimerskirch, H., Bishop, C., Jeanniard-du-Dot, T., Prudor, A., & Sachs, G. (2016). Frigate birds track atmospheric conditions over months-long transoceanic flights. *Science*, 353, 74–78. <https://doi.org/10.1126/science.aaf4374>
- Wilson, R. P., Grmillet, D., Syder, J., Kierspel, M. A. M., Garthe, S., Weimerskirch, H., ... Nel, D. (2002). Remote-sensing systems and seabirds: Their use, abuse and potential for measuring marine environmental variables. *Marine Ecology Progress Series*, 228, 241–261. <https://doi.org/10.3354/meps228241>
- Wuillez, M., Poulard, J.-C., Rivoirard, J., Petitgas, P., & Bez, N. (2007). Indices for capturing spatial patterns and their evolution in time, with application to European hake (*Merluccius merluccius*) in the Bay of Biscay. *ICES Journal of Marine Science*, 64, 537–550. <https://doi.org/10.1093/icesjms/fsm025>
- Wuillez, M., Rivoirard, J., & Petitgas, P. (2009). Notes on survey-based spatial indicators for monitoring fish populations. *Aquatic Living Resources*, 22, 155–164. <https://doi.org/10.1051/alr/2009017>
- Worton, B. J. (1989). Kernel methods for estimating the utilization distribution in home-range studies. *Ecology*, 70, 164–168. <https://doi.org/10.2307/1938423>
- Wynne-Edwards, V. C. (1935). *On the habits and distribution of birds on the North Atlantic*. Boston, MA: Boston Society of Natural History.
- Yamamoto, T., Watanuki, Y., Hazen, E. L., Nishizawa, B., Sasaki, H., & Takahashi, A. (2015). Statistical integration of tracking and vessel survey data to incorporate life history differences in habitat models. *Ecological Applications*, 25, 2394–2406. <https://doi.org/10.1890/15-0142.1>
- Zador, S. G., & Piatt, J. F. (1999). Time-budgets of common murres at a declining and increasing colony in Alaska. *The Condor*, 101, 149–152. <https://doi.org/10.2307/1370455>
- Zar, J. H. (1999). *Biostatistical analysis*, 4th ed. Upper Saddle River, NJ: Prentice Hall.

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