Ecosystem-based management and the wealth of ecosystems

Seong Do Yun[a], Barbara Hutniczak[b], Joshua K. Abbott[c], and Eli P. Fenichel[a,1]

[a]School of Forestry and Environmental Studies, Yale University, New Haven, CT 06511; [b]National Oceanic and Atmospheric Administration National Marine Fisheries Service, Silver Spring, MD 20910; and [c]School of Sustainability, Arizona State University, Tempe, AZ 85287

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We merge inclusive wealth theory with ecosystem-based management (EBM) to address two challenges in the science of sustainable management of ecosystems. First, we generalize natural capital theory to approximate realized shadow prices for multiple interacting natural capital stocks (species) making up an ecosystem. These prices enable ecosystem components to be better included in wealth-based sustainability measures. We show that ecosystems are best envisioned as portfolios of assets, where the portfolio's performance depends on the performance of the underlying assets influenced by their interactions. Second, changes in ecosystem wealth provide an attractive headline index for EBM, regardless of whether ecosystem wealth is ultimately included in a broader wealth index. We apply our approach to the Baltic Sea ecosystem, focusing on the interacting community of three commercially important fish species: cod, herring, and sprat. Our results incorporate supporting services embodied in the shadow price of a species through its trophic interactions. Prey fish have greater shadow prices than expected based on market value, and predatory fish have lower shadow prices than expected based on market value. These results are because correctly measured shadow prices reflect interdependence and limits to substitution. We project that ecosystem wealth in the Baltic Sea fishery ecosystem generally increases conditional on the EBM-inspired multispecies maximum sustainable yield management beginning in 2017, whereas continuing the current single-species management generally results in declining wealth.

natural capital | inclusive wealth | fisheries | Baltic Sea

The influential management consultant Peter Drucker is credited with saying "what gets measured, gets managed."* Ecosystem-based management (EBM) is frequently upheld as a holistic management alternative to traditional, decoupled, single-species management approaches (2, 3). However, implementation of EBM lags, in part, because of the lack of a conceptually sound framework and measurable headline indicator for assessing management tradeoffs among multiple species, their habitat, and human wants and needs from the ecosystem. Instead, managers are often faced with a proliferation of potentially contradictory measures, making it difficult to measure the performance of managed ecosystems. Furthermore, the proliferation of indicators undercuts the holistic aims of EBM and often leads to "ecological" vs. "economic" arguments appealing to distinct indicators. A headline "bioeconomic" indicator that measures the value to society created through a more inclusive management approach is critical to help make good on the promise of EBM and provide policymakers, tasked with implementation of EBM, with a concrete measure of progress. Such an indicator can be developed by linking EBM with bioeconomic modeling and concepts and measures from sustainability assessment.

Writing concerning societal wellbeing, Stiglitz et al. (4) note that "[w]hat we measure affects what we do, and if our measurements are flawed, our decisions may be distorted." Arrow et al. (5), Dasgupta (6), and Barbier (7) argue that sustainability requires maintaining the capacity for future human wellbeing (including but not limited to market consumption), which requires constant or increasing "wealth"—the price-weighted sum of all societal capital stocks valued at appropriate asset prices. Wealth accounting (8–10) (e.g., inclusive, comprehensive, or genuine wealth) has a long history in economics (11) and provides a rigorous economic paradigm for measuring sustainability (5, 12–15). There is broad acknowledgment that ecosystems are important stores of wealth. For example, the 2014 Inclusive Wealth report suggests that 28% of global wealth is contained in ecosystems, which may be an underestimate (9). Resolving EBM's need for a headline indicator of a system's ecological–economic sustainability and national accountant's need for accurate prices at which to value ecosystems is important for environmental stewardship and sustainability (16). At the core of this problem is the accurate measurement of natural capital asset prices, also known as "shadow prices" (17).

We contribute to the EBM, wealth accounting, and natural capital literatures by extending (18, 19) to accurately measure natural capital asset prices in systems with multiple interacting stocks of natural capital. The extension enables a wealth metric that can guide resource management at local scales by generating a headline EBM index—the change in natural capital wealth in an ecosystem—while providing appropriate data for national accounts. Although the theory of wealth accounting was developed in the context of national accounts, it may be exceedingly valuable at the ecosystem management scale (20–22). Moreover, we show that properly derived and measured natural capital asset prices (shadow prices) reflect the localized limits of substitution among ecosystem components. Shadow prices capture ecological

Significance

Ecosystems store vast quantities of wealth, but difficulties measuring wealth held in ecosystems prevent its inclusion in accounting systems. Ecosystem-based management endeavors to manage ecosystems holistically. However, ecosystem-based management lacks headline indicators to evaluate performance. We unify the inclusive wealth and ecosystem-based management paradigms, allowing apples-to-apples comparisons between the wealth of the ecosystem and other forms of wealth, while providing a headline performance index for evaluating the performance of ecosystem-based management. We project that the Baltic Sea fishery ecosystem yields increasing stores of wealth over the next 50 y under the ecosystem-based management-inspired multispecies maximum sustainable yield management beginning in 2017, whereas the previous single-species management generally results in declining wealth.

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[1]To whom correspondence should be addressed. Email: Eli.Fenichel@yale.edu.

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*Attribution is actually uncertain, but Sir William Thomson, Lord Kelvin, seems to have expressed a similar idea well before Drucker (1).

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and economic interactions between capital stocks, implying that a wealth index need not be merely a weak sustainability index (23, 24) but is capable of capturing the limits and opportunities of substitution, thereby advancing Barbier's (7) "capital sustainability" concept. Our empirical analysis provides a roadmap for the application of natural capital asset measurement and wealth accounting in more complex EBM contexts.

Valuing an Ecosystem Full of Natural Capital

Operationalizing wealth accounting requires natural capital asset prices or shadow prices (6, 8, 9, 14, 17). The Systems of Environmental–Economic Accounting framework (7, 13, 25) suggests that entire ecosystems can be valued as assets. However, ecosystems are better viewed as a fund or portfolio of distinct but interacting natural assets (26, 27). The performance of the ecosystem fund is a function of the performance of the underlying assets. Unlike a financial fund, the capital stocks held in an ecosystem fund directly interact and affect each other's performance. As with financial assets, the way that the assets in an ecosystem are jointly managed is critical to their valuation—thereby linking EBM with ecosystem wealth accounting. Management interventions typically operate differentially on components of an ecosystem and may have complex feedbacks through the ecosystem. It is important to disaggregate the bundled ecosystem portfolio into the most important or material assets to track the wealth held in an ecosystem and assess management's influence on changes in ecosystem wealth through time.

Fenichel and Abbott (FA) (19) and Fenichel et al. (18) adapted Jorgenson's (28) capital asset pricing approach to a single stock of natural capital to provide "apples-to-apples" comparisons between traditional capital asset prices and natural capital asset prices. FA enables valuation of natural capital assets under a given management approach without requiring the analyst to assume an optimized policy. FA prices are scarcity measures and account for modeled feedbacks from human behavior and institutions. These features allow resource managers to use wealth accounting approaches to examine how management changes affect capital asset values apart from any change in the immediate state of the ecosystem. The FA approach can be extended to measure shadow prices for an ecosystem with N interacting stocks of natural capital with each stock indexed by i (derivation is in *SI Text*):

$$p^i(\boldsymbol{s}) = \frac{MD^i(\boldsymbol{s},\boldsymbol{x}(\boldsymbol{s})) + \left(\frac{\partial p^i}{\partial s^i}\dot{s}^i + \sum_{j\neq i}^{N}\frac{\partial p^j}{\partial s^i}\dot{s}^j\right) + \sum_{j\neq i}^{N} p^j\frac{\partial \dot{s}^j}{\partial s^i}}{\delta - [MG^i(\boldsymbol{s}) - MHI^i(\boldsymbol{s},\boldsymbol{x}(\boldsymbol{s}))]}. \quad [1]$$

Eq. **1** takes the same general form as the FA single-stock natural capital asset pricing equation (18). The left-hand side, $p^i(\boldsymbol{s})$, is the shadow price for natural capital stock $s^i \in \boldsymbol{s}$, where $\boldsymbol{s}$ is an N-length vector of capital stocks. The shadow price measures the extra current and discounted future benefits arising from conserving an additional unit of a natural capital stock, its marginal social value, given the current institutions and assumed management plan. As in FA, the shadow price is a function of the marginal dividends, MD^i, an index of net benefit flows to society from an additional unit of natural capital stock i; a series of price and stock effects $((\partial p^i/\partial s^i)\dot{s}^i + \sum_{j\neq i}^{N}(\partial p^j/\partial s^i)\dot{s}^j) + \sum_{j\neq i}^{N} p^j \partial \dot{s}^j/\partial s^i$, which we discuss momentarily; a "baseline" discount rate δ ; a marginal growth rate of natural resource i, MG^i; and a rate of marginal human impact, MHI^i. As in FA, MD, MG, and MHI are functions of the stocks of natural capital and humans' institutionally conditioned responses to these stocks—the economic program $\boldsymbol{x}(\boldsymbol{s})$.

Despite Eq. **1** taking the same general form as FA's single-stock natural capital asset price equation, there are three important differences. First, the shadow price of natural capital stock i is a function of the entire N-length vector of stocks, $\boldsymbol{s}$. The shadow price of a stock of natural capital depends, in general, on the state of the entire ecosystem. Second, although FA described the economic program as a scalar function of a single capital stock, here we generalize it as a mapping from an N-length vector of stocks to an L-length vector of human actions: $\boldsymbol{x}(\boldsymbol{s})$. This vector-valued function captures all human feedbacks that are relevant to benefits flows (MD) or human impacts to natural capital stocks in the ecosystem (MHI) and allows for multiple inputs and joint production, which are common in ecosystems (29).

The third change compared to FA has the greatest ramifications for linking wealth accounting and EBM. The terms following MD in the numerator of Eq. **1** are complicated relative to those in FA. These terms capture expectations of the influence of conserving an increment of natural capital i on the future productivity of the ecosystem fund in delivering valued services (i.e., capital gains) given the biophysical dynamics and economic program. The second numerator term $((\partial p^i/\partial s^i)\dot{s}^i + \sum_{j\neq i}^{N}(\partial p^j/\partial s^i)\dot{s}^j)$ reflects the effects of investment in s^i on capital gains through its effects on the shadow prices of all assets in the ecosystem fund (i.e., "price effects"). In FA's single-stock case, this term collapses to $dp/dt = \dot{p} = (\partial p/\partial s)\dot{s}$. The third numerator term in Eq. **1**, $\sum_{j\neq i}^{N} p^j(\partial \dot{s}^j/\partial s^i)$, captures an effect that is only present in multiasset systems: the effects of an investment in s^i on the physical growth of other natural capital stocks in the fund (i.e., "cross-stock effects"). Together, price effects and cross-stock effects enable shadow prices to capture nonlinear substitution possibilities and complementarities among stocks as their quantities change. These substitution possibilities can arise from consumption, production, or ecological interactions, and it is these substitution and complementarity relationships that are at the core of the sustainability problem (7, 30).

Headline Indicator for EBM

The potential present and future wellbeing stored in a set of capital stocks at time t, $\boldsymbol{s}(t)$, can be expressed as the net present value (NPV) of dividend flows, D, from natural and other capital stocks: $V(\boldsymbol{s}(t)) = \int_t^{\infty} e^{-\delta(\tau-t)} D(\boldsymbol{s}(\tau),\boldsymbol{x}(\boldsymbol{s}(\tau)))d\tau$. The quantity V incorporates the trajectory of human behavior and stocks through time according to the ecological dynamics and economic program. Although changes in V are the ultimate theoretical basis for assessing welfare (23, 31), it can be convenient to work with a linear accounting index known as inclusive wealth (6, 8, 9): $IW(t) = \sum_{i=1}^{N} p^i(t)s^i(t)$. For a discrete time interval Δt, let $\Delta IW = \sum_{i=1}^{N} \bar{p}^i \Delta s^i$, where $\bar{p}^i$ is a weighted average of the asset price for stock i at the beginning and end of the accounting period (21). Dasgupta (23), Dasgupta and Maler (32), and Arrow et al. (5) show that $\Delta V/\Delta t \approx \Delta IW/\Delta t$ for a sufficiently short time period. A necessary condition for sustainability over Δt is that $\Delta V \approx \Delta IW \geq 0$, because this condition ensures maintenance of the potential to provide valued service flows today and in the future. Our approach to calculating natural capital shadow prices leads to direct computation of ΔIW and ΔV. Furthermore, when the shadow prices come from Eq. **1**, ΔIW or ΔV provides a headline ecological–economic index of the sustainable management of an ecosystem under EBM and a quantity for national accountants to use to value changes in ecosystems.

A general concern with green accounting and especially inclusive wealth is that species and abiotic components of ecosystems are not perfect substitutes. On the surface, the linear-in-stocks form of IW suggests that declines in asset s^i can be indefinitely substituted at a constant rate for investments in another capital stock s^j, such that IW is unchanged. This view (24) may originate from and be sustained by efforts to calculate inclusive wealth using adjusted market prices to illustrate the IW concept (10). However, recognizing from Eq. **1** that $IW = \sum_{i=1}^{N} p^i(\boldsymbol{s})s^i$, the inclusive wealth index is actually nonlinear in stocks. Therefore, stocks are not necessarily perfect substitutes—assuming that shadow prices are updated between evaluation points to reflect changes in substitutability. As stocks change, their shadow prices change

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along with the shadow prices of other stocks. If $\partial p^i/\partial s^j > 0$, then s^i and s^j are capital complements. When stocks are capital complements, increasing one stock raises the value of the other. For example, increases in a prey stock make a predator stock more productive and hence, more valuable, whereas an increase in predator stocks creates a need for more prey and hence, makes prey more valuable. If $\partial p^i/\partial s^j < 0$, then s^i and s^j are capital substitutes. Stocks of natural capital are substitutes when increases in one stock reduce the value of the other. For example, two species that fill similar roles in an ecological–economic system (e.g., two competing prey species) may be capital substitutes. However, the substitutability or complementarity of capital stocks depends on more than their ecological interactions; the interactions of capital stocks in the provision of dividends, *MD*, and human responses to changes in the ecosystem portfolio $\boldsymbol{x}(\boldsymbol{s})$ also matter. Overall, these capital substitutes and complement relationships reflect joint ecological–economic interactions and relative scarcity within the ecosystem. The ability to capture these interactions makes ΔIW and ΔV strong candidates as headline indicators for EBM.

Case Study

To illustrate the usefulness of ΔIW and ΔV as indicators of sustainable management for EBM and respond to recent calls for assessing marine natural capital (33), we compute shadow prices for the three most ecologically and economically important natural capital stocks in the Baltic Sea commercial fishery: cod (*Gadus morhua*), herring (*Clupea herengus*), and sprat (*Sprattus sprattus*). Predatory cod shapes this ecosystem (34). By contrast, prey availability, including herring and sprat, influences cod's consumption rates and growth, thus creating a feedback loop that strongly ties these three species together. For simplicity, we assume that environmental factors are invariant, and we do not consider the sensitivity of cod to the salt water inflows from the North Sea. Tahvonen et al. (35) show that stochasticity is not critical to maximum sustained yield in this system.

Cod, sprat, and herring generate ∼80% of the Polish Baltic Sea fleet's revenue, which is representative of other Baltic Sea nations (36). For the case study, we focus on net revenue, a monetary index for an important benefit flow in this system, but the methodology could include a broader set of benefit flows, beyond food provision, in other applications. It is important to consider how changes over time in net revenue and stocks are affected by and affect ecosystem structure, fishers' responses to regulations, and feedbacks among the fish species (37). In 2014, Poland began managing all three species with binding individual vessel-level allocations, which are shares of the total allowable catch (TAC) for each stock. The shares are set as the fraction of the assessed fish stock in each year. Joint production (29) best describes the fishery, because the three species are jointly harvested (38). The allocation system, ex-vessel prices, fishing costs, and the market economy that drive fishers to attempt to maximize net revenue together constitute the economic program, which is almost certainly inefficient. In this multispecies fishery, we assume limited ability to target particular species, so that the economic program—the supply of effort in the fishery—can be written as a scalar function $x(\boldsymbol{s})$, where $x(\boldsymbol{s})$ is a function of all three stocks. If effort could be perfectly targeted (39), then $\boldsymbol{x}(\boldsymbol{s})$ could be a three-element vector of the effort levels targeting each stock. Fig. 1 shows the historical stock performance and projected biomass starting from 2013, assuming single-species management and the business as usual (BAU) TAC allocation regime based on the single-species stock assessments vs. an alternative, recently accepted, multispecies maximum sustainable yield (MMSY) TAC allocation regime (EC 2016/1139) that follows EBM principles (40, 41).

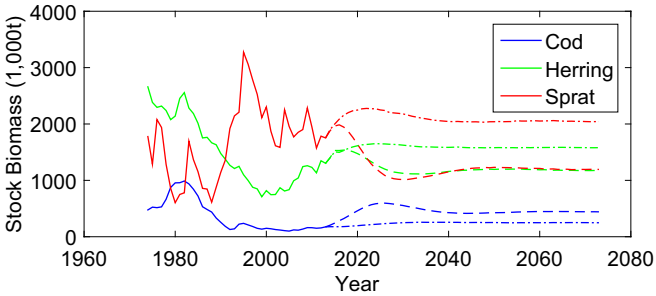

Fig. 1. Historical stock biomass (solid lines), predicted stock biomass under single-species BAU (dashed lines), and forecasted stock biomass under the EBM MMSY (dot-dash lines) from the multispecies interaction model.

Natural Capital Prices in an Ecosystem "Fund." We estimate three interconnected shadow price functions, one for each fish stock. To illustrate the three-to-one shadow price function, we hold two stocks constant (Fig. 2). As expected, the price curves for each species slope downward in own stock; prices increase as quantity declines, reflecting the relative scarcity of each stock. The shape of these price curves is affected by cross-stock effects seen by comparing the position of the price curves at alternative stock levels for the other two species.

The price curves show that a species in an ecosystem can be an asset, having a positive shadow price, or a liability, having a negative shadow price, depending on its abundance (42). In such cases, the label of "asset" or "liability" applies to the marginal value; a species may provide overall benefits, although its marginal value in conservation is negative. The shadow price of cod is positive at 2013 stock levels but negative at the system's BAU steady state, whereas the shadow prices for prey fish are positive at the BAU steady state but negative at 2013 levels. At 2013 levels, cod are scarce, so scarce that their implicit value in conservation for reproduction (the shadow price) exceeds the ex-vessel price of harvesting the cod (illustrated by the black dots in Fig. 2). However, in the BAU equilibrium, the shadow price for cod is negative, suggesting that the BAU strategy fails to adequately account for the predation effects on the prey stock, resulting in an overinvestment in cod. In contrast to BAU management, the MMSY management program maintains all three stocks as assets, with positive shadow prices, in equilibrium (Fig. 2).

The price curves for herring and sprat reflect their direct value for their respective fisheries as well as their indirect value as "inputs" to or "supporting services" for the growth and sustenance of cod stocks. In 2013, herring and sprat stocks are above their steady-state levels. This abundance paired with low cod stocks that cannot benefit from additional prey saturation drive the 2013 shadow values of prey stocks below zero. Prey stocks decline in equilibrium, which in combination with growth in the cod population, ensures that they become assets in the long run.

Whether stocks are capital substitutes or complements ultimately depends on the complex interplay of trophic interactions, how changes in species abundance affect harvester behavior, and the way that stocks enter the *MD* function. To examine these patterns in the Baltic, we map shadow price contours for herring and cod (Fig. 3) under the BAU scenario. Upward (downward) sloping contours indicate complementary (substitute) relationships between the stocks on the axes (*SI Text*). The complementary and substitute relationships reflected in the shadow prices are strongly influenced by ecological relationships as well as economic and behavioral factors. It is these kinds of relationships that EBM is charged with balancing (43).

In the Baltic Sea case, the substitutability or complementarity between species is tightly linked to trophic interactions. The prey species, herring and sprat (Fig. 3*A*), have the downward sloping

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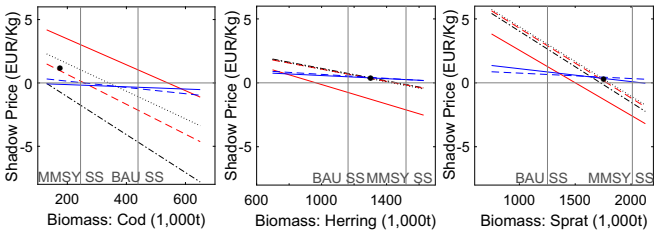

Fig. 2. Shadow prices when the two other species are fixed at the 2013 (solid lines) and equilibrium (dashed lines) levels for the single-species BAU (red) and EBM MMSY (blue) scenarios. Black curves show sensitivity for the BAU case when other stocks are held at 80% (dot-dash lines) and 110% (dotted lines) of equilibrium levels. Black dots show 2013 ex-vessel prices (*y* axis) at 2013 stock levels (*x* axis). SS, steady states.

contours of capital substitutes. As the stock of herring declines, increasing its shadow price, a compensating increase in sprat drives the shadow price of herring back to its previous level. Sprat and herring are substitutes to fishers in the linear profit function entering D and substitutes as prey for cod. The slopes of the contours are not constant, implying that the extent of substitutability between herring and sprat varies depending on relative abundances. The shadow price reflects the imperfect ecological and economic niche overlap. This nonconstant shadow price relationship means that declines in sprat cannot be made up with a fixed ratio increase in herring and vice versa.

Cod and either prey species have upward sloping contour lines, implying complementarity, an extreme lack of substitutability (Fig. 3 *B–D*). Increasing abundance of prey species enhances the asset value of cod and vice versa. Although cod and prey are substitutes in terms of their immediate contribution to fishing profits, the predator–prey ecological relationship drives their capital complementarity through supporting ecosystem services, which come through the MG, cross-stock, and cross-price terms. Specifically, the wealth gains from investing in cod are enhanced with an accompanying increase in a prey stock. The marginal social value of cod stock increases faster with increases in sprat than with increases in herring, which is illustrated by the flatter contours in Fig. 3*C* relative to Fig. 3*D*.

Changes in Baltic Sea Wealth. Consider how management coupled with ecological dynamics stores wealth in the Baltic Sea fish populations. As noted above, nondeclining intertemporal wellbeing, $\Delta V \geq 0$, is a necessary condition for sustainability (23), where the change in wealth, ΔIW, serves as an approximation for ΔV when ΔV is not measurable (44). The theoretical equivalence of ΔV and ΔIW holds for relatively short intervals where the shadow prices of assets are effectively constant. When shadow prices change significantly over evaluation intervals, a weighted average of before and after prices is more appropriate; however, finding the optimal weighting to minimize the divergence between ΔV and ΔIW may quickly become intractable for highly nonlinear systems (21). We measure the potential divergence between ΔV and ΔIW using a first-order midpoint approximation for the shadow price in ΔIW [i.e., $\Delta IW = 0.5(p_{t-1} + p_t)(s_t - s_{t-1})$].

We calculate V and IW (the latter indexed to V for the base year 2013) under BAU (black lines in Fig. 4) and EBM-based MMSY (blue lines in Fig. 4) scenarios. We also plot the paths of benefits flows (D) (dot-dash lines in Fig. 4). V, as the NPV of future D, acts as a "leading indicator" for fluctuations in benefits flows from the ecosystem, anticipating future declines in wellbeing in terms of reduced D (31). The IW and V indices qualitatively track over the range of the data, and the signs of the slopes (changes) generally align. For the historical (pre-2013) period, the sign of ΔIW does not match that of ΔV on a year to year basis in two of the years. Nevertheless, the ΔV and ΔIW over the historical period generally tell a consistent story of increasing wealth and intertemporal wellbeing, suggesting that past management generally passes a sustainability test. The post-2013 forecasts show that the sign of ΔIW matches that of ΔV 100% of the time for

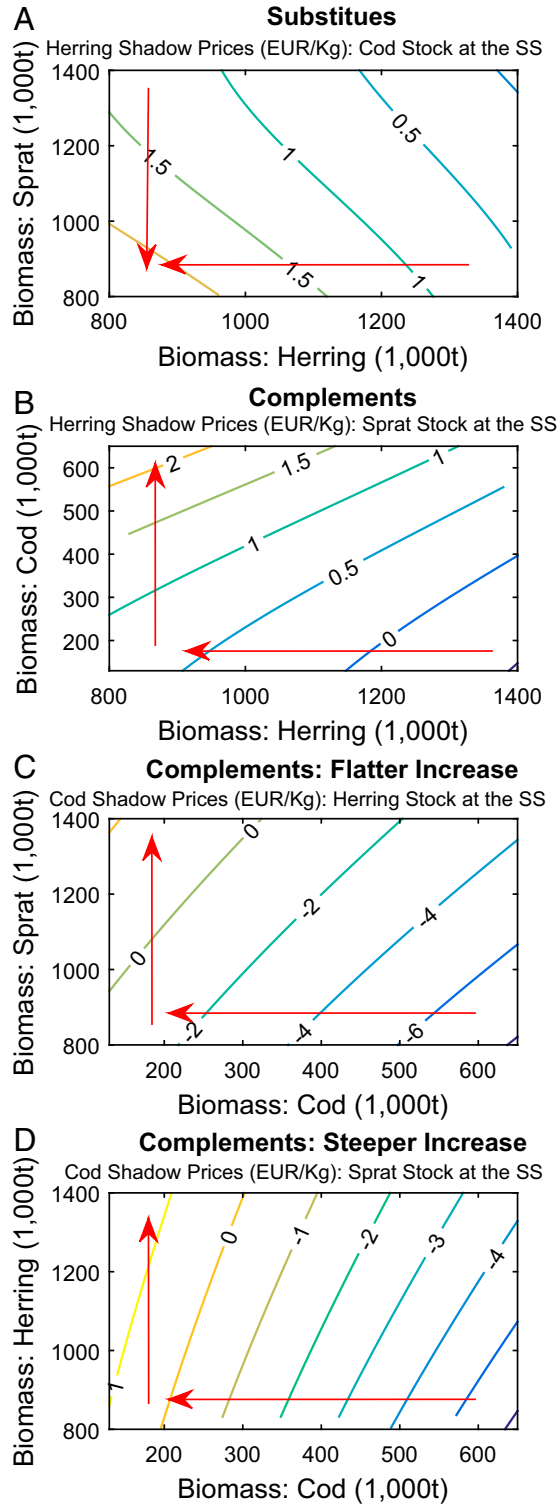

Fig. 3. Contours of shadow prices (single-species BAU scenario). Shadow prices are calculated by fixing the cod stock at the steady state in *A*, the sprat stock at the steady state (SS) in *B* and *D*, and the herring stock at the steady state in *C*. The arrows are the increasing direction of shadow prices.

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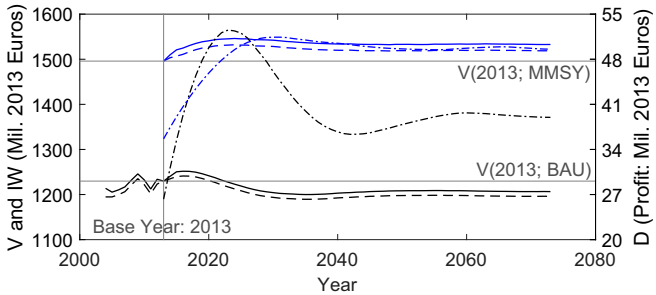

Fig. 4. $V(t)$ (solid lines), $IW(t) = \Delta IW + IW(t-1)$ indexed to 2013 so that $IW(2013) = V(2013)$ (dashed lines), and profit $D(t)$ (dot-dash lines) as a function of time for single-species BAU (black) and EBM-based MMSY scenarios (blue). Historical data (pre-2013) are also shown in black.

the BAU and the MMSY scenarios, suggesting that ΔIW is a strong proxy for ΔV in the Baltic Sea ecosystem over the range of historical and simulated data.

Viewed over a 60-y horizon (2013–2073), which leads to a system steady state, the analysis of ΔV and ΔIW for the single-species BAU scenario suggests that this management may be suspect on sustainability grounds. By contrast, if MMSY was implemented in 2013, we project that ΔV and ΔIW would be positive relative to 2013. Furthermore, moving from BAU to MMSY management in 2013 instantaneously increases V, suggesting that MMSY management is preferred on efficiency and sustainability grounds. The MMSY scenario is in line with recently proposed and accepted changes to the Baltic Sea fisheries regulations set to begin in 2017 (EC 2016/1139) that take into account interactions between species and shift toward EBM. Single-species BAU management ultimately leads to a greater cod stock. The MMSY scenario leads to fewer cod but greater prey fish abundance. Changes in the inclusive wealth and intertemporal well-being indices capture the dual roles of herring and sprat in their roles in directed fisheries and as ecological inputs in the production of cod. Capturing these direct and indirect roles of species in decision-making is an important tenant of EBM.

Although the Baltic Sea fishery ecosystem is more likely to be sustainable under the EBM MMSY regime than the single-species BAU regime, neither is strictly sustainable across all intervals of the 60-y forecast. Because of the nonmonotonicity of V and IW over time, there are intervals where wealth declines, which is inconsistent with wealth or welfare-based notions of sustainability. Furthermore, theoretically, the equivalence between ΔV and ΔIW can deteriorate over longer time intervals (greater Δt) when stocks and shadow prices are in flux. In this case, although BAU and MMSY IW trajectories track V remarkably well on a year to year basis, small quantitative deviations between ΔV and ΔIW emerge immediately after 2013, opening a wedge between V and IW. Such divergences can lead to deviations in the sign of ΔV and ΔIW over multiyear periods when the path of V and hence, IW is nonmonotonic—as evidenced here by different signs for ΔV and ΔIW (relative to 2013) over the period 2020–2023 in the BAU case. The gap emerges between the indices, because the linear price averaging embodied in IW is not sufficiently precise when there is rapid change in the system and price curves are not sufficiently linear over the region of system change.

Discussion

Measurement is a necessary condition for effective management, but it is not sufficient. One must also synthesize and weigh the tradeoffs reflected in the measurements themselves. Failure to do so risks leaving decision-makers adrift in a sea of indicators. Assessing progress toward sustainable management requires informing managers about the consequences to present and future stakeholders of investment and divestment decisions across a complex fund of interacting species and their habitats—demanding a fundamentally integrative decision approach. Assessing the change in the wealth stored in an ecosystem facilitates the aims of EBM by breaking down artificial barriers between economic and ecological indicators. Our case study reveals that wealth generally rises (falls) under the EBM (single-species management). Under EBM, cod, which provides the product of greatest commercial value, is maintained at a lower level than under the single-species management scenario, but prey fish, which provide a lower-valued commercial product, are maintained at a greater level—in part for their role in sustaining cod stocks. This insight comes from balancing complex interactions and nonlinear production processes—precisely the mission of EBM.

Our example responds to calls for a headline EBM indicator in fisheries and marine systems (33). However, this framework and pricing approach are applicable across many domains (18, 45). Furthermore, grounding EBM in the theory of sustainability assessment could raise support for EBM in new domains other than those where it has already gained significant acceptance.

There is broad interest in valuing ecosystems as unitary stocks. However, changes in the wealth of ecosystems cannot be valued without considering changes in the value of the constituent stocks that make up the ecosystem. It is reasonable and useful to include the value of the "ecosystem fund" into regional or national wealth accounts, filling an important gap in the sustainability accounting literature (46, 47). However, the wealth stored in ecosystems is only a portion of the wealth of humanity, which also includes human and built capital. Linking natural capital accounting to broader wealth accounting efforts is ultimately essential to fully assess sustainability and guide decision-making.

Limited substitutability and complementarities are at the heart of the sustainability question (7, 30, 48). Correctly measured shadow prices capture the limits of substitutability and important complementarities. The linearity of wealth-based indicators combined with the use of fixed shadow prices occasionally have led to misconceptions that wealth-based indicators impose an assumption that capital stocks are perfect substitutes for one another—that it is a weak sustainability index (12, 16, 24). However, the shadow prices resulting from Eq. **1** are functions of the underlying capital stocks, and the interactions among stocks are reflected through cross-price and cross-stock effects and MG and MHI terms. Wealth metrics are not measures of weak sustainability. Shadow prices and associated changes in wealth reflect the limits of substitution if derived from a coupled systems modeling approach that adequately incorporates ecological and economic knowledge. Our case study shows that ecological interactions can dictate whether stocks are capital substitutes or capital complements; therefore, ecological relationships take on an even more vital role in natural capital asset prices in systems with interacting stocks than they did in FA's single-asset approach (18, 19). Although our case study focuses on the substitutability between stocks of natural capital and provisioning services, the framework can be extended to the analysis of substitutability between stocks of natural and human or produced capital and can incorporate a broader set of services. Indeed, there is nothing in theory or practice that requires first nesting stocks of interest within a system into the bins of natural, human, and produced capital. An advantage of our approach is that the measures of substitutability between capital stocks emerge endogenously from the structure of the bioeconomic model rather than being imposed arbitrarily. EBM essentially charges decision-makers with the responsibility of managing a complex fund of natural assets in a way that yields sustainable benefits to stakeholders. This Olympian task can be facilitated by providing managers with informed measurements of substitution possibilities within their purview and an approach to assess tradeoffs. Shadow prices for natural capital, derived from integrative bioeconomic

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models and embedded within a wealth accounting framework, provide the essential bridge from measurement to management.

Methods

The vector of shadow prices, p^i, associated with $\boldsymbol{s}$ can be recovered using similar approximation techniques as in ref. 18. The derivation of natural capital asset prices for multiple interacting stocks and the numerical method to approximate the suggested shadow prices are explained in *SI Text*. We implemented an empirical analysis of the Baltic Sea fishery ecosystem based on ref. 40 using a curve-fitting method detailed in *SI Text*. The recovery of natural capital asset prices is performed using the R package {*capn*} (environment.yale.edu/profile/eli-fenichel/software).

We capture predator–prey dynamics and regulation-constrained vessel responses in the Baltic Sea fishery with a modified version of the ecological–economic vessel-level fishing choice model in ref. 40. We assume that Polish management is representative of the Baltic Sea and scale results based on Poland's fleet to the system. In the model (details are in *SI Text*, and parameterization in Tables S1–S7), the fishing fleet consists of 411 vessels that optimize individual behavior subject to regulations, feasibility constraints (maximum days that the vessel can spend at sea), owned capital, production structure, and individual technical efficiency. The economic component allows estimation of the profit-seeking response to the quota allocations based on the established harvest control rule. The biological submodel describes the feedback within the ecosystem based on interacting components derived from established food web links. We first build a detailed age-structured model with eight age classes and predator–prey relationships for each stock. Then, we use approximation techniques to compress the dynamics to focus on the three natural capital stocks (Fig. S1).

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1. Thomson W (1891) *Popular Lectures and Addresses* (Macmillan and Co., London).
2. Christensen NL, et al. (1996) The report of the Ecological Society of America committee on the scientific basis for ecosystem management. *Ecol Appl* 6:665–691.
3. Pikitch EK, et al. (2004) Ecology. Ecosystem-based fishery management. *Science* 305: 346–347.
4. Stiglitz JE, Sen A, Fitoussi J-P (2010) *Mis-Measuring Our Lives: Why GDP Doesn't Add Up, the Report by the Commission on the Measurement of Economic Performance and Social Progress* (New Press, New York).
5. Arrow K, et al. (2004) Are we consuming too much? *J Econ Perspect* 18:147–172.
6. Dasgupta P (2014) Measuring the wealth of nations. *Annu Rev Resour Economics* 6: 17–31.
7. Barbier EB (2011) *Capitalizing on Nature* (Cambridge Univ Press, New York).
8. World Bank (2011) *The Changing Wealth of Nations* (World Bank, Washington, DC).
9. UNU-IHDP and UNEP (2014) *Inclusive Wealth Report 2014, Measuring Progress Toward Sustainability* (Cambridge Univ Press, Cambridge, UK).
10. Arrow KJ, Dasgupta P, Goulder LH, Mumford KJ, Oleson K (2012) Sustainability and the measurement of wealth. *Environ Dev Econ* 17:317–353.
11. Fisher I (1906) *The Nature of Capital and Income* (Norwood Press, Norwood, MA).
12. Hanley N, Dupuy L, McLaughlin E (2015) Genuine savings and sustainability. *J Econ Surv* 29:779–806.
13. Barbier EB (2013) Wealth accounting, ecological capital and ecosystem services. *Environ Dev Econ* 18:133–161.
14. Hamilton K, Hartwick JM (2014) Wealth and sustainability. *Oxf Rev Econ Policy* 30: 170–187.
15. Obst C, Hein L, Edens B (2016) National accounting and the valuation of ecosystem assets and their services. *Environ Resour Econ* 64:1–23.
16. Polasky S, et al. (2015) Inclusive wealth as a metric of sustainable development. *Annu Rev Environ Resour* 40:6.1–6.22.
17. Smulders S (2012) An arrow in the Achilles' heel of sustainability and wealth accounting. *Environ Dev Econ* 17:368–372.
18. Fenichel EP, et al. (2016) Measuring the value of groundwater and other forms of natural capital. *Proc Natl Acad Sci USA* 113:2382–2387.
19. Fenichel EP, Abbott JK (2014) Natural capital from metaphor to measurement. *J Assoc Environ Res Econ* 1:1–27.
20. Pearson LJ, Biggs R, Harris M, Walker B (2013) Measuring sustainable development: The promise and difficulties of implementing Inclusive Wealth in the Goulburn-Broken Catchment, Australia. *Sustainability* 9:16–27.
21. Fenichel EP, et al. (2016) Wealth reallocation and sustainability under climate change. *Nat Clim Chang* 6:237–244.
22. Dovern J, Quaas MF, Rickels W (2014) A comprehensive wealth index for cities in Germany. *Ecol Indic* 41:79–86.
23. Dasgupta P (2007) *Human Well-Being and the Natural Environment* (Oxford Univ Press, New York).
24. Pearce D, Atkinson G (1993) Capital theory and the measurement of sustainable development: An indicator of "weak" sustainability. *Ecol Econ* 1993:103–108.
25. United Nations, et al. (2014) *System of Environmental–Economic Accounting 2012 Central Framework* (United Nations, New York).
26. Ando AW, Mallory ML (2012) Optimal portfolio design to reduce climate-related conservation uncertainty in the Prairie Pothole Region. *Proc Natl Acad Sci USA* 109: 6484–6489.
27. Schindler DE, et al. (2010) Population diversity and the portfolio effect in an exploited species. *Nature* 465:609–612.
28. Jorgenson DW (1963) Capital theory and investment behavior. *Am Econ Rev* 53: 247–259.
29. Nalle DJ, Montgomery CA, Arthur JL, Polasky S, Schumaker NH (2004) Modeling joint production of wildlife and timber. *J Environ Econ Manage* 48:997–1017.
30. Quaas MF, van Soest D, Baumgartner S (2013) Complementarity, impatience, and the resilience of natural-resource-dependent economies. *J Environ Econ Manage* 66: 15–32.
31. Fleurbaey M, Blanchet D (2013) *Beyond GDP Measuring Welfare and Assessing Sustainability* (Oxford Univ Press, New York).
32. Dasgupta P, Maler K-G (2000) Net national product, wealth, and social well-being. *Environ Dev Econ* 5:69–93.
33. Lu Y (2016) Sustainable development: Rate oceans' capital to help achieve SDGs. *Nature* 537:34.
34. Casini M, et al. (2008) Multi-level trophic cascades in a heavily exploited open marine ecosystem. *Proc Biol Sci* 275:1793–1801.
35. Tahvonen O, Quaas MF, Voss R (2016) *What Difference Does It Make? Age Structure, Gear Selectivity, Stochastic Recruitment, and Economic Vs. MSY Objectives in the Baltic Cod Fishery* (Department of Economics, Department of Forest Sciences, University of Helsinki, Helsinki, Finland).
36. STECF, ed (2015) *The 2015 Annual Economic Report on the EU Fishing Fleet* (STECF, Copenhagen).
37. Lade SJ, et al. (2015) An empirical model of the Baltic Sea reveals the importance of social dynamics for ecological regime shifts. *Proc Natl Acad Sci USA* 112:11120–11125.
38. Hutniczak B (2014) Increasing pressure on unregulated species due to changes in individual vessel quotas: An empirical application to trawler fishing in the Baltic Sea. *Mar Resour Econ* 29:201–217.
39. Squires D (1987) Fishing effort: Its testing, specification, and internal structure in fisheries economics and management. *J Environ Econ Manage* 14:268–282.
40. Hutniczak B (2015) Modeling heterogeneous fleet in an ecosystem based management context. *Ecol Econ* 120:203–214.
41. Collie JS, et al. (2016) Ecosystem models for fisheries management: Finding the sweet spot. *Fish (Oxf)* 17:101–125.
42. Horan RD, Bulte EH (2004) Optimal and open access harvesting of multi-use species in a second-best world. *Environ Resour Econ* 28:251–272.
43. Sanchirico JN, Smith MD, Lipton DW (2008) An empirical approach to ecosystem-based fishery management. *Ecol Econ* 64:586–596.
44. Arrow KJ, Dasgupta P, Maler K-G (2003) Evaluating projects and assessing sustainable development in imperfect economies. *Environ Resour Econ* 26:647–685.
45. Bond CA (2017) Valuing coastal natural capital in a bioeconomic framework. *Water Econs Policy* 3:2.
46. Obst C, Vardon M (2014) Recording environmental assets in the national accounts. *Oxf Rev Econ Policy* 30:126–144.
47. Barbier EB (2014) Economics: Account for depreciation of natural capital. *Nature* 515: 32–33.
48. Seppelt R, Manceur AM, Liu J, Fenichel EP, Klotz S (2015) Synchronized peak-rate years of global resources use. *Ecol Soc* 19:50.
49. Dasgupta P, Maler K-G, Barrett S (1999) Intergenerational equity, social discount rates and global warming. *Discounting and Intergenerational Equity*, eds Portney P, Weyant J (Resources for the Future, Washington, DC), pp 51–78.
50. OMB (2003) *Circular A-4: Regulatory Analysis* (Executive Office of the President, Washington, DC), Budget OoMa.
51. Shapiro C, Stiglitz JE (1984) Equilibrium unemployment as a worker discipline device. *Am Econ Rev* 74:433–444.
52. Judd KL (1998) *Numerical Methods in Economics* (MIT Press, Cambridge, MA).
53. Miranda MJ, Fackler PL (2002) *Applied Computational Economics and Finance* (MIT Press, Cambridge, MA).
54. Press WH, Teukolsky SA, Vetterling WT, Flannery BP (2007) *Numerical Recipes 3rd Edition: The Art of Scientific Computing* (Cambridge Univ Press, New York).
55. Vlassenbroeck J, Van Dooren R (1988) A Chebyshev technique for solving nonlinear optimal control problems. *IEEE Trans Automat Contr* 33:333–340.
56. ICES (2014) *Report of the Baltic Fisheries Assessment Working Group (WGBFAS)* (ICES, Copenhagen).
57. ICES (2013) *Report of the Benchmark Workshop on Baltic Multispecies Assessment (WKBALT)* (ICES, Copenhagen).
58. Churchhill R, Owen D (2010) *The EC Common Fisheries Policy* (Oxford Univ Press, Oxford).
59. Gelb A, Platte RB, Rosenthal WS (2008) The discrete orthogonal polynomial least squares method for approximation and solving partial differential equations. *Commun Comput Phys* 3:734–758.
60. Hewitt E, Hewitt RE (1979) The Gibbs-Wilbraham phenomenon: An episode in Fourier analysis. *Arch Hist Exact Sci* 21:129–160.

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